1	Enhanced Phosphorus Locking by Novel Lanthanum/Aluminum-Hydroxide				
2	Composite: Implication for Eutrophication Control				
3	Rui Xu ^{\dagger, \ddagger} , Meiyi Zhang ^{\dagger} , Robert J.G. Mortimer [§] , and Gang Pan ^{*, \dagger, §}				
4	[†] Research Center for Eco-Environmental Sciences, Chinese Academy of Sciences,				
5	Beijing 100085, China				
6	[‡] University of Chinese Academy of Sciences, Beijing 100049, China				
7	[§] School of Animal, Rural and Environmental Sciences, Nottingham Trent University,				
8	Brackenhurst Campus, NG25 0QF, UK				
9	* Corresponding author: gpan@rcees.ac.cn (GP)				
10					
11					
12					
13					
14					
15					
16					
17					
18					
19					
20					
21					
22					
23					
24					
25					

26 Abstract: Lanthanum (La) bearing materials have been widely used to remove phosphorus (P) in water treatment. However, it remains a challenge to enhance 27 phosphate (PO₄) adsorption capacity and La usage efficiency. In this study, La was co-28 precipitated with aluminum (Al) to obtain a La/Al-hydroxide composite (LAH) for P 29 adsorption. The maximum PO₄ adsorption capacities of LAH (5.3% La) were 76.3 and 30 45.3 mg P g⁻¹ at pH 4.0 and 8.5, which were 8.5 and 5.3 times higher than those of 31 commercially available La-modified bentonite (Phoslock[®], 5.6% La), respectively. P 32 K-edge X-ray absorption near edge structure analysis showed that PO₄ was 33 34 preferentially bonded with Al under weakly acid conditions (pH 4.0), while tended to associate with La under alkaline conditions (pH 8.5). La L_{III}-edge extended X-ray 35 absorption fine structure analysis indicated that PO₄ was bonded on La sites by forming 36 37 inner sphere bidentate-binuclear complexes and oxygen defects exhibited on LAH surfaces, which could be active adsorption sites for PO₄. The electrostatic interaction, 38 ligand exchange and oxygen defects on LAH surfaces jointly facilitated PO₄ adsorption 39 40 but with varied contribution under different pH conditions. The combined contribution of two-component of La and Al may be an important direction for the next generation 41 42 of commercial products for eutrophication mitigation.

43 **TOC**

Oxygen defects Oxygen defects occupied 🔵 Oxygen 🔵 Phosphorus 🔘 Hydroger

- 44 45
- 46
- 47

48 Introduction

Excessive phosphorus (P) discharge can trigger severe eutrophication in aquatic 49 environments,¹⁻³ which may further lead to harmful algal blooms, deterioration of water 50 quality, and eventually, ecosystem collapse.^{4, 5} Over recent decades, various materials 51 have been used for P removal in water bodies.² In particular, La-modified materials or 52 La-bearing compounds are often considered as effective geo-engineering tools to lock 53 P from aquatic ecosystems.^{6, 7} For instance, the commercial product Phoslock[®], a La-54 modified bentonite clay, has been used for internal Ploads control in 200 water bodies.^{2,} 55 8 56

Previous studies suggest that the binding ability between La and PO₄ anions 57 dominates P locking by La-bearing materials.⁹⁻¹¹ Ligand exchange is generally 58 considered as the main mechanism of PO₄ affinity to La-bearing materials.^{9, 10} It is well 59 known that the ability of ligand exchange depends on the surface charge of the 60 adsorbent.¹² However, existing La-modified materials (including Phoslock[®]) often have 61 a negative surface charge in natural waters,^{10, 13} which is not favorable for attracting 62 $H_2PO_4^-$ or HPO_4^{2-} from the viewpoint of electrostatic interaction. If the surface charge 63 of La-modified material could be made positive, it would attract more PO₄ anions.¹² 64 We hypothesize that synthesizing a La-bearing adsorbent with positive surface charge 65 would enable a two component mechanism involving both electrostatic interaction and 66 67 ligand exchange to improve PO_4 adsorption capacity and La use efficiency. This requires a new host material for La in order to develop a more efficient PO₄ adsorbent. 68 Aluminum (Al) compounds are also widely used for P removal in water treatment.¹⁴ 69 Al (hydro)oxide is effective at adsorbing P selectively in the presence of competing 70 anions such as Cl⁻, NO₃⁻, HCO₃⁻ and SO₄²⁻.¹⁵ It is also abundant in nature, low cost and 71 has a high point of zero charge (pH_{PZC}).¹⁴ Al compounds could potentially be a host 72

73 material for La in synthesizing a new adsorbent with positive surface charge. On the other hand, it has been reported that the effectiveness of Al treatment depends on the 74 pH in waters,¹⁶ because Al compounds are prone to desorption of PO₄ under alkaline 75 conditions.^{17, 18} Moreover, the ecological effect of aluminum has become a concern in 76 aquatic research.^{17, 19} Owing to the high affinity between La and PO₄ anions, if an Al 77 compound is incorporated in the fabrication with La, the PO₄ adsorbed by La/Al 78 composite may be more stable over a wide pH range, which is important for practical 79 remediation engineering under field conditions. 80

As an element-specific and in situ method to detect the molecular structures of 81 reacting species, X-ray absorption fine structure (XAFS) spectroscopy, including X-ray 82 absorption near edge structure (XANES) and extended X-ray absorption fine structure 83 84 (EXAFS) spectroscopy, has been developed for determining the properties of PO₄ bonding on solid surfaces.^{11, 20, 21} P K-edge XANES has been used to study PO₄ 85 adsorption in single or binary systems of Fe/Al-hydroxide materials, which can 86 distinguish PO₄ contents associated with Fe- or Al- hydroxide.^{20, 22, 23} EXAFS could 87 provide additional insights for PO₄ bonding mode by detecting the local structure of 88 metal elements reacted with PO₄.¹¹ By using La L_{III}-edge EXAFS spectroscopy, the PO₄ 89 bonded on Phoslock[®] was proved to be an insoluble rhabdophane (LaPO₄·nH₂O, n \leq 90 3).¹¹ However, to the best of our knowledge, little information is available about the 91 effect of La local structure on PO₄ adsorption capacity and its transformation after 92 93 reaction with PO₄.

Here, a La/Al-hydroxide composite (LAH) with high PO₄ adsorption efficiency was
synthesized by a co-precipitation method. Three LAH composites with different La:Al
molar ratios of 1:30, 1:20 and 1:10 were prepared in order to investigate the role of La
in LAH for PO₄ adsorption. Batch experiments were conducted to determine the

98 adsorption and desorption properties of PO₄ on LAH. P K-edge XANES spectroscopy was used to investigate the proportions of PO_4 associated with Al relative to La. La L_{III}-99 edge EXAFS spectroscopy was used to detect the local atomic structure changes around 100 101 the La (III) ion before and after PO₄ adsorption to elucidate the adsorption mechanism of PO₄ on LAH surfaces. The objective of this study is to synthesize a highly efficient 102 PO₄ adsorption material and to investigate its mechanism at the molecular level, so that 103 the methodology could be used to guide new development of geo-engineering materials 104 for eutrophication management. 105

106 Materials and methods

LAH synthesis and characterization. Predetermined amounts of analytical grade 107 LaCl₃·7H₂O and AlCl₃ were dissolved into 200 mL deionized water to obtain La/Al 108 109 molar ratios of 1:30, 1:20 and 1:10 (LAH-1/30, LAH-1/20 and LAH-1/10). Under stirring, 2 mol L⁻¹ NaOH aqueous solution was added dropwise to the mixed solutions 110 at 333 K until pH 9.0. Then the mixture was stirred for 2 h and aged for 24 h at 333 K. 111 112 Precipitates were washed three times with deionized water and freeze-dried for 24 h. Powders of Al(OH)₃ and La(OH)₃ were synthesized by the same process to act as 113 controls (Details are provided in the Supporting Information). 114

To obtain the exact content of La and Al in the LAH, samples were digested in 6 mol 115 L^{-1} HCl and the chemistry of the resultant solutions was determined using an 116 117 inductively coupled plasma optical emission spectrometer (ICP-OES, Optima 8300, PerkinElmer, USA). Brunauer-Emmett-Teller (BET) surface areas and pore size 118 distributions were examined by a Micromeritics ASAP 2020 static volumetric analyzer. 119 120 LAH morphologies were observed using field emission scanning electron microscopy (FESEM, SU 8020, Hitachi, Japan). X-ray powder diffraction (XRD) data of the LAH 121 were recorded on an X'Pert PRO MPD X-ray diffractometer (PANalytical, The 122

Netherlands) with a Cu-K radiation ($\lambda = 1.5408$) at 40 kV and 40 mA in the 2 θ range of 5° to 90°. Electron paramagnetic resonance (EPR) spectrum of LAH-1/10 was recorded on a JEOL JES-FA200 EPR spectrometer (Japan) at room temperature. Zeta (ζ) potential analysis of LAH was measured by a Zetasizer Nano ZS potential analyzer (Malvern Instrument Ltd., United Kingdom). Phoslock[®] (Phoslock[®] Water Solutions Ltd, Australia) was also characterized by the same procedures described above.

Adsorption and desorption experiments. In order to identify the influence of pH on the equilibrium adsorption capacity, adsorption experiments of PO₄ on LAH were conducted in polypropylene tubes at pH 4.0 ~ 10.0. LAH (1 g L⁻¹) were added to 50 mL polypropylene tubes with initial P concentration 80 mg P L⁻¹ in 0.01 mol L⁻¹ NaCl background and shaken at 170 rpm at 298 K for 48 h. The solution pH was maintained constant with 0.01 mol L⁻¹ HCl and NaOH.

Adsorption isotherm experiments were conducted in 50 mL polypropylene tubes by 135 mixing various concentrations of KH_2PO_4 solutions with 1 g L⁻¹ adsorbents in 0.01 mol 136 L⁻¹ NaCl background. The mixture was stirred at 298 K for 48 h allowing the adsorption 137 to reach equilibrium. pH 4.0 (pH for maximum PO₄ adsorption in this study, Figure S3) 138 and pH 8.5 (up limit pH for lakes)²⁴ were chosen to determine the adsorption and 139 desorption behaviors of PO₄. At the end of adsorption isotherm experiments, the 140 exhausted solutions of LAH samples which had initial P concentration 80 mg P L⁻¹ were 141 centrifuged to preserve 10 mL suspension at the bottom, and then 0.01 mol L⁻¹ NaCl 142 solution was added to keep the same ionic strength. The experimental conditions of 143 desorption experiments were conducted following adsorption studies. 144

The PO₄ concentration was measured by ascorbic acid method to monitor the
absorbance at 880 nm with spectrophotometer (UV-756 PC, Shanghai Sunny Hengping
Scientific Instrument CO. LTD, China). The concentrations of released Al³⁺ and La³⁺

148 were analyzed by ICP-OES (Optima 8300, PerkinElmer, USA).

XAFS data collection and analysis. Phosphorus K-edge XANES data were 149 collected on Beamline 4B7A at the Beijing Synchrotron Radiation Facility (BSRF), 150 China. The storage ring was operating at 2.5 GeV with current from 250 to 150 mA. A 151 double crystal Si (111) monochromator offered an energy resolution of about 0.6 eV 152 with a beam size of 1.0×3.0 mm². The incident X-ray energy was calibrated by using 153 the white line of K₂SO₄ (S K-edge) at 2482.4 eV before spectra collection. Spectra were 154 collected in the fluorescence yield (FLY) mode between -30 to +90 eV relative to P K-155 edge energy at 2152 eV with a minimum step size of 0.3 eV between 2140 and 2180 156 eV. The P1s peak position of KH₂PO₄ was used to calibrate the P K-edge XANES 157 spectra. La L_{III}-edge EXAFS data were collected on beamline 1W2B at the BSRF, 158 159 China. The monochromator energy calibration was monitored to 5483 eV using a vanadium foil as an internal standard with a transmission chamber detector. The moist 160 samples were fixed in a Teflon cell and sealed with Mylar film during EXAFS 161 measurement.²⁵ An average of three scans was performed to achieve suitable 162 single/noise, and no obvious change in spectral data was observed during the three 163 scans. The spectral data were processed following the standard procedures of 164 background absorption removal, normalization, k-space conversion and Fourier 165 transformation. Details for data collection and analysis were given in the Supporting 166 167 Information.

168 **Results**

Characterization. The physiochemical properties of the three synthesized LAH
materials, commercial product Phoslock[®] and Al(OH)₃ were summarized in Table 1.
The La mass proportion of LAH-1/30 was 5.3%, which was close to that of Phoslock[®]
(5.6%). The La mass proportions of LAH-1/20 and LAH-1/10 were 7.8% and 13.1%,

respectively. The BET surface areas of LAH-1/30, LAH-1/20 and LAH-1/10 were 174 107.1, 121.0 and 99.3 m² g⁻¹, respectively, which were much higher than those of 175 Phoslock[®] (38.2 m² g⁻¹) and Al(OH)₃ (51.2 m² g⁻¹).

A da a vla a vrá	La (wt %)	Al (wt %)	surface area	pore volume
Adsorbent			$(m^2 g^{-1})$	$(cm^3 g^{-1})$
LAH-1/30	5.3	31.7	107.1	0.34
LAH-1/20	7.8	31.4	121.0	0.39
LAH-1/10	13.1	26.5	99.3	0.31
Phoslock®	5.6	7.8	38.2	0.14
Al(OH) ₃	-	34.6	51.2	0.06

176 Table 1. General characteristics of LAH, Phoslock[®] and Al(OH)₃

The XRD analysis suggested that LAH were mainly composed of Al(OH)₃ (PDF NO. 74-1119) and La(OH)₃ (PDF NO. 36-1481) (Figure 1a). The characterization peaks of La(OH)₃ were weaker with increasing La contents, indicating that more amorphous La(OH)₃ was formed. The EPR spectrum of LAH-1/10 showed a clear signal at g =2.0509 (Figure 1b) corresponding to oxygen defects on the surfaces.^{26, 27} FESEM images showed that the average particle size of LAH composite was about 8-16 µm and thin flakes appeared on LAH surfaces with increasing La contents (Figure S1).



Figure 1. (a) XRD patterns of LAH, in comparison with model compounds of Al(OH)₃
and La(OH)₃. Asterisks (*), Al(OH)₃; Plus (+), La(OH)₃. (b) EPR spectrum of LAH1/10 at room temperature.

The ζ potential of Phoslock[®] and LAH samples before and after PO₄ adsorption were 188 shown in Figure 2. The LAH had positive surface charge from pH 4.0 to 9.7, while 189 Phoslock[®] had negative potential from -5.4 to -43.1 mV in the pH range of 4.0 to 10.0. 190 The ζ potential of all LAH materials decreased after PO₄ adsorption at initial P 191 concentration 50 mg P L⁻¹. For LAH-1/30, the ζ potential decreased significantly from 192 about +50 mV (pH 4.0) and +10 mV (pH 8.5) to +11 mV (Figure 2, dot a) and -21 mV 193 (Figure 2, dot b) (p < 0.001, paired t-test, SPSS 19.0) after PO₄ adsorption, respectively. 194 The ζ potential of La(OH)₃ and Al(OH)₃ were also determined, and the pH_{pzc} value of 195 La(OH)₃ and Al(OH)₃ was approximately 9.4 (Figure S2). 196



197

Figure 2. ζ potential of LAH and Phoslock[®] as a function of pH (solid line), and that of LAH with adsorbed PO₄ at pH 4.0 (dot a) and 8.5 (dot b).

200 **Phosphate adsorption and desorption.** The effect of pH on PO₄ adsorption to LAH 201 was investigated in the pH range of $4.0 \sim 10.0$ (Figure S3). LAH-1/10 had the highest 202 PO₄ adsorption capacity of 71.6 mg P g⁻¹ at pH 4.0 and decreased to 26.1 mg P g⁻¹ at 203 pH 10.0. The data showed LAH exhibited a maximum PO₄ adsorption capacity at pH 204 4.0 and dropped with increasing pH values. Consequently, pH 4.0 (pH for maximum

PO₄ adsorption in this study) and pH 8.5 (up limit pH for lakes)²⁴ were chosen in
following experiments to investigate the macro phenomena and micro mechanisms of
PO₄ adsorption on LAH.

Phosphate adsorption isotherms for LAH were L-curves (Figure 3) and were fitted 208 with the Langmuir model (Table S1). LAH-1/10 had the maximum PO₄ adsorption 209 capacities (O_m) of 128.2 and 70.4 mg P g⁻¹ at pH 4.0 and pH 8.5 (Table S1). The O_m of 210 LAH-1/30 (5.3% La, Table 1) were 8.5- and 5.3-fold higher than those of Phoslock[®] 211 (5.6%, Table 1) at pH 4.0 and pH 8.5, respectively. The PO₄ specific adsorptions on 212 Phoslock[®] and LAH were calculated on the basis of Q_m (Table S1) and BET surface 213 area (Table 1). LAH-1/10 exhibited higher PO₄ specific adsorptions of 1.29 and 0.71 214 mg m⁻² than those of LAH-1/20 (1.00 and 0.53 mg m⁻²) and LAH-1/30 (0.71 and 0.42 215 mg m⁻²) at pH 4.0 and 8.5 (Table S1), respectively. The PO₄ desorption rates of LAH-216 1/30, LAH-1/20 and LAH-1/10 were 2.54%, 2.01% and 4.05% at pH 4.0, which were 217 reduced to 0.28%, 0.22% and 0.29% at pH 8.5 (Figure S4), respectively. 218



219

Figure 3. Langmuir isotherm fitting plots of LAH and Phoslock[®] at pH 4.0 (a) and pH 8.5 (b). The experimental condition: temperature = 298 K, adsorbent dosage = 1 g L⁻¹, reaction time = 48 h.

In order to assess the ecological risk of LAH to natural water bodies, concentrations of released Al^{3+} and La^{3+} were measured in natural water pH ranges from 7.0 to 9.0

(Figure 4). The maximum concentration of released Al³⁺ was 0.078 mg L⁻¹ by LAH1/30 at pH 9.0 (Figure 4a). LAH-1/10 had the highest released La³⁺ concentration of
0.077 mg L⁻¹ at pH 7.0 (Figure 4b). The dissolutions of Al and La from LAH at pH 4.0
to 10.0 were also determined (Figure S5). The maximum Al and La dissolved
proportions were 0.58% by LAH-1/30 and 0.55% by LAH-1/10 at pH 4.0, respectively.



230

Figure 4. Concentrations of released Al^{3+} and La^{3+} from pH 7.0 to 9.0. The experimental condition: temperature = 298 K, adsorbent dosage = 1 g L⁻¹, reaction time = 48 h.

P K-edge XANES. Normalized P K-edge XANES spectra for LAH-1/30 samples 234 after PO₄ adsorption with initial P concentrations (C_0) 50 and 80 mg P L⁻¹ at pH 4.0 and 235 8.5 were shown in Figure 5. LCF results showed that the proportions of PO₄ bonded 236 with La-hydroxide (La-P) for LAH-1/30 with adsorbed PO₄ at pH 4.0 were 32.6% and 237 45.0%, which were less than those with Al-hydroxide (Al-P) of 67.4% and 55.0% 238 (Figure 5a, b). However, for LAH-1/30 with adsorbed PO₄ at pH 8.5, the proportions 239 of PO₄ boned with La-hydroxide (La-P, 52.6% and 54.1%) were higher than those with 240 Al-hydroxide (Al-P, 47.4% and 45.9%) (Figure 5c, d). 241



Figure 5. Linear combination fitting for LAH-1/30 samples after PO₄ adsorption with weighted components of La-P and Al-P. The experimental condition: (a) pH 4.0 with $C_0 50 \text{ mg P L}^{-1}$, (b) pH 4.0 with $C_0 80 \text{ mg P L}^{-1}$, (c) pH 8.5 with $C_0 50 \text{ mg P L}^{-1}$ and (d) pH 8.5 with $C_0 80 \text{ mg P L}^{-1}$.

242

La L_{III}-edge EXAFS. In order to investigate the La local coordination environment, 247 the La L_{III}-edge EXAFS spectra analysis was conducted (Figure 6). Spectra of 248 commercial La(OH)₃ and Phoslock[®] were also collected (Figure S6, S7). The first shell 249 La-O of LAH was at a distance of 2.55 ~ 2.60 Å (Table S2), which was very close to 250 that of the La-O shell of La(OH)₃ (2.5 ~ 2.6 Å).²⁸ The La-O CN of LAH-1/30, LAH-251 1/20 and LAH-1/10 were 6.8 ± 0.7 , 6.5 ± 0.4 and 6.1 ± 0.3 (Table S2), respectively. Al 252 atoms were detected at 3.10 ~ 3.11 Å with La-Al CN of 2.6 ± 0.9 , 2.2 ± 0.5 and $1.4 \pm$ 253 0.5 to LAH-1/30, LAH-1/20 and LAH-1/10 (Table S2), respectively. The La-O CN of 254 LAH samples after PO₄ adsorption were 7.3 ~ 7.6 (Table S3). A new La-P shell in LAH 255 samples with adsorbed PO₄ was detected at a distance of $3.20 \sim 3.31$ Å and the average 256

257 CN of La-P shell were about 2.5 (Table S3). Additionally, the original CN of La-O shell 258 in Phoslock[®] was 8.9 ± 0.8 (Table S2), which was about 8.8 after PO₄ adsorption (Table 259 S3). With increased La loading, the intensity of La-O peak became weaker in LAH 260 composite (Figure S8b). Besides, the attenuation rate of $\chi(k)$ of LAH was faster than 261 that of La(OH)₃ (Figure S8a). In accordance with XRD spectra (Figure 1a), La L_{III}-edge 262 EXAFS spectra also proved that amorphous La(OH)₃ was formed in LAH during the 263 La/Al co-precipitation process.



264

Figure 6. Normalized k²-weighted La L_{III}-edge EXAFS spectra (a1, b1, c1), the corresponding Fourier transformed magnitude (a2, b2, c2) and real (q) part of Fourier transform (c1, c2, c3) of pristine LAH, LAH with adsorbed PO₄ at pH 4.0 (LAH-PO₄, pH 4.0) and LAH with adsorbed PO₄ at pH 8.5 (LAH-PO₄, pH 8.5). Samples of LAH-1/30 (black), LAH-1/20 (red) and LAH-1/10 (blue). The dotted lines were observed spectra and the solid lines were fitting curves.

Electrostatic attraction and ligand exchange. The ζ potential data showed that the 272 pH_{PZC} value of LAH-1/30, LAH-1/20 and LAH-1/10 was approximately 9.7 (Figure 2). 273 274 When solution pH was lower than 9.7, LAH were protonated and positively charged. Compared to negatively charged Phoslock[®], the positively charged surfaces of LAH 275 could attract more negatively charged PO₄ anions by electrostatic interaction over a 276 wide pH range. The results showed that the ζ potential of LAH decreased with 277 increasing pH values (Figure 2). Lower pH is beneficial for the protonation of LAH 278 279 surfaces, which could enhance the electrostatic attraction between LAH and PO₄ anions. Owing to the higher surface charge and more protonated active sites, LAH-1/30 showed 280 a higher PO₄ adsorption capacity (76.3 mg P g⁻¹, Table S1) at pH 4.0 than that (45.3 mg 281 P g⁻¹, Table S1) at pH 8.5. Although La(OH)₃ is also positively charged under natural 282 lake pH conditions (Figure S2),²⁹ the utility of La (hydro)oxide alone in eutrophication 283 restoration is not recognized as an economic way in view of its high price and relative 284 285 scarcity. The incorporation of Al hydroxide with La in LAH materials not only maintains high pH_{PZC} values, but also reduces costs compared to using La (hydro)oxide 286 alone. 287

Changes of solution pH during PO₄ adsorption on LAH-1/30 at initial pH 4.0 and 8.5 288 were monitored (Figure S9). The solution pH increased from initial pH 4.0 to nearly 289 290 6.5 after 24 h and it increased little when initial pH was 8.5 (Figure S9). The increase in pH value was due to the release of OH⁻ after ligand exchange interaction, indicating 291 that ligand exchange played a primary role in PO₄ adsorption.^{9, 10} The surface charges 292 of pristine LAH-1/30 decreased significantly after PO₄ adsorption (Figure 2), 293 suggesting that PO₄ was bonded on LAH surfaces by forming inner-sphere complexes 294 via ligand exchange.³⁰ In order to demonstrate ligand exchange interaction at the 295

296 molecular level, we measured P K-edge XANES spectra of LAH-1/30 samples with adsorbed PO₄ (Figure 5). According to LCF analysis, the proportions of PO₄ bonded 297 with Al-hydroxide were 2.07 and 1.22 times higher than those with La-hydroxide at pH 298 4.0 (Figure 5a, b). On the contrary, the proportions of PO₄ bonded with Al-hydroxide 299 (Al-P, 47.4% and 45.9%) were less than those with La-hydroxide (La-P, 52.6% and 300 54.1%) at pH 8.5 (Figure 5c, d). The fitting data showed PO₄ was preferentially bonded 301 with Al in weakly acidic conditions (pH 4.0), while most PO₄ tended to associate with 302 La under alkaline conditions (pH 8.5). The results showed that both La-hydroxide and 303 304 Al-hydroxide in LAH composite could bind with PO₄ anions by forming inner-sphere complexes.^{31, 32} In this study, the XANES measurement provided strong evidence that 305 a ligand exchange mechanism played a role both at pH 4.0 and 8.5. 306

307 It was reported that higher alkaline condition (pH 9.5) might result in the dissolution of Al(OH)₃ and the release of PO₄ concomitantly.¹⁷ However, the Al³⁺ and PO₄ released 308 proportions in LAH were both very low at alkaline conditions (Figure S4, S5). 309 According to La L_{III}-edge EXAFS analysis, a La-Al shell was formed in LAH samples 310 (Table S2). As the addition of La to alumina could generate new-type Lewis acid site 311 on the surface,³³ the increased La loading in LAH might affect the properties of Al-312 hydroxide in LAH composite. The Al³⁺ and La³⁺ ions all showed a higher dissolution 313 proportions at lower alkalinity (Figure S5).³⁴ It has been hypothesized that natural 314 organic matter can impact PO₄ adsorption via the formation of OM-Fe(III)-PO₄ or OM-315 Al(III)-PO₄ ternary complexes.^{35, 36} Hence, we induced that the ternary LAH-Al³⁺/La³⁺-316 PO₄ complex might also be formed when pH changed to the low value of 4.0. 317

The oxygen defects on La compound. EXAFS analysis showed that La-O CN of LAH-1/30, LAH-1/20 and LAH-1/10 were 6.8 ± 0.7 , 6.5 ± 0.4 and 6.1 ± 0.3 (Table S2),

- respectively. The average La-O CN of LAH composite were approximately 6.5, which

were all less than the saturated La-O CN of 9.0 in La(OH)₃.²⁸ It is generally recognized 321 that lower CN of surface ions may depend on oxygen defects on the surfaces of metal 322 compound.³⁷⁻³⁹ Among LAH materials, LAH-1/10 with highest La content (13.1%, 323 Table 1) might had the most amounts of oxygen defects on the surfaces. Additionally, 324 EPR spectrum demonstrated the presence of oxygen defects on the surfaces (Figure 1b). 325 It was reported that the surface defect sites were more reactive than the perfect 326 surfaces.^{40, 41} LAH-1/10 with more La loading exhibited higher PO₄ specific 327 adsorptions of 1.29 (pH 4.0) and 0.71 mg m⁻² (pH 8.5) than those of LAH-1/20 and 328 329 LAH-1/30 (Table S1), respectively, which might be also attributed to more amounts of oxygen defects on the surfaces. Compared to the mixture of pure La(OH)₃ and pure 330 Al(OH)₃ to a La/Al molar ratio 1:30, the PO₄ specific adsorption on LAH-1/30 at pH 331 8.5 (0.42 mg m⁻², Table S1) was 1.6 times higher than the mixture (0.26 mg m⁻², Figure 332 S10), further supporting that the oxygen defects on LAH surfaces provided special 333 adsorption towards PO₄ anions. The La-O CN of Phoslock[®] with PO₄ bonded was 8.8 334 335 (Table S3), which was similar to its original CN of 8.9 (Table S2). While, the average La-O CN of LAH after PO₄ adsorption increased to 7.5 (Table S3) compared to its 336 original average value of 6.5 (Table S2). The change of La-O in LAH before and after 337 PO₄ adsorption proved the role of oxygen defects on LAH to facilitate PO₄ adsorption. 338 It was reported that the oxygen defects on Al compound were not observed until the 339 material was calcined at above 473 K.⁴² Therefore, the role of oxygen defects on La 340 compound for PO₄ adsorption, not Al compound, was the major consideration in 341 present study since the prepared temperature of LAH was 333 K. 342 After PO₄ adsorption, a new La-P shell was detected in LAH samples (Table S3).

The distance of La-P shell was at 3.20 ~ 3.31 Å (Table S3), which was shorter than that 344

343

of LaPO₄•1.4H₂O (3.52 Å) and LaPO₄ (4.16 Å).¹¹ With the bond distance and the 345

average CN (~ 2.5, Table S3) of La-P shell, the PO₄ surface configuration was in
agreement with previous EXAFS studies of PO₄ adsorption on goethite/water
interface.⁴³ Hence, the most likely configuration of PO₄ on La site was a bidentatebinuclear complex.⁴³ The La-PO₄ bonding was a direct evidence to demonstrate the
formation of inner-sphere complex of P on La site.

Toxicity analysis. The release of metal ions is a major concern when using lake 351 restoration materials, which may induce potentially environmental risk and limit their 352 usage as P removers in some environments.² It has been demonstrated that excessive 353 concentrations of Al^{3+} and La^{3+} may be toxic to the sensitive species in waters.^{2, 17} 354 Daphnia magna has been used to indicate eco-toxicity potential when using geo-355 engineering materials.⁴⁴ According to previous studies, the median effect 356 concentrations (EC₅₀) of Al^{3+} and La^{3+} to *Daphnia magna* were above 3.9 and 23 mg 357 L^{-1} .^{45, 46} In this study, the maximum concentrations of released Al³⁺ and La³⁺ after 358 adsorption procedure were 0.078 mg L^{-1} by LAH-1/30 and 0.077 mg L^{-1} by LAH-1/10 359 under natural water pH conditions (Figure 4), which were much lower than the reported 360 thresholds of 3.9 and 23 mg L⁻¹, respectively. Moreover, it has been reported that La³⁺ 361 and Al^{3+} ions could also bind with CO_3^{2-} and humics in aquatic environments thus 362 possibly reducing their toxicity to some extent.⁴⁷⁻⁴⁹ Although further investigation is 363 needed under a wide range of environmental conditions, the initial findings suggest that 364 365 it is possible to use LAH materials at their optimal dosage for control of internal P loads without causing substantial adverse effects to aquatic ecosystem. 366

Environmental implications. In recent decades, cost-efficient and ecologically benign P locking materials have been in urgent demand in geo-engineering for eutrophication control.^{6, 7, 50} This study may promote the development of novel Lamodified materials for highly efficient P removal from water bodies. To remove one Kg

P in water bodies, the cost of LAH is 18.5 - 25.6 \$/Kg P, which is less than that of 371 Phoslock[®] (27.9 \$/Kg P) (Table S4). The LAH composite has the potential to save up 372 to ~ 34% cost compared to that of Phoslock[®]. Moreover, LAH have much higher P 373 adsorption capacity and La use efficiency than other La-modified materials reported 374 before (Table S5). After LAH are applied in natural waters (pH 8.5), the composition 375 of La-hydroxide in LAH may play a primary role in locking P than Al-hydroxide 376 (Figure 5). The combined contribution of two-component of La and Al may be an 377 important direction to improve the P removal efficiency for the next generation of 378 379 commercial products. LAH exhibited great practical potential due to its low PO₄ desorption rates and low metal ions released concentrations under natural lake pH 380 conditions. The positively charged LAH may also have the potential in removing other 381 382 negatively charged pollutants in waters, such as algal cells, arsenate and chromate. With the combination of LAH and modified local soil,⁵¹ the toxic algae and P in waters may 383 be removed, and the floc resuspension and P release from lake sediments can also be 384 capped and blocked as well.⁵⁰ Considering the challenge of understanding long-term 385 effects, further studies are needed to evaluate its impacts on aquatic ecological 386 responses. With additional research, LAH may be a compelling candidate as a geo-387 engineering tool for P control. 388

389 ASSOCIATED CONTENT

Supporting Information

Details on the synthesis of La(OH)₃ and Al(OH)₃, equations used in this study and XAFS data collection and analysis. Figures showing the FESEM images of LAH, ζ potential of La(OH)₃ and Al(OH)₃, effect of pH on phosphate adsorption efficiency, phosphate desorption rates of LAH, Al/La dissolutions from LAH at varying pH 4.0-10.0, EXAFS experimental and fitted spectra, changes of solution pH and Langmuir adsorption isotherm on mixture of La(OH)₃ and Al(OH)₃. Tables showing Langmuir
isotherm parameters and phosphate specific adsorptions, EXAFS data of La(OH)₃,
LAH and Phoslock[®], cost of LAH and Phoslock[®] and comparison of La-modified
materials. This information is available free of charge via the Internet at
http://pubs.acs.org.

401 AUTHOR INFORMATION

402 **Corresponding Author**

- 403 *Corresponding author: Tel.: +86 10 62849686; Fax: +86 10 62849686; E-mail address:
- 404 gpan@rcees.ac.cn
- 405 Notes
- 406 The authors declare no competing financial interest.

407 ACKNOWLEDGEMENTS

The research was supported by the Strategic Priority Research Program of CAS (XDA09030203) and National Natural Science Foundation of China (21377003). We thank the Beijing Synchrotron Radiation Facility (BSRF, China) for providing the beam time of 4B7A and 1W2B. Dr. Zheng Lei, Ma Chenyan, An Pengfei, Hu Yongfeng and Liu Wei are acknowledged for their valuable assistance with the XANES and EXAFS experiments and data analysis.

414 **References**

- 415 (1) Conley, D. J.; Paerl, H. W.; Howarth, R. W.; Boesch, D. F.; Seitzinger, S. P.; Havens,
- 416 K. E.; Lancelot, C.; Likens, G. E. Controlling eutrophication: nitrogen and phosphorus.
- 417 *Science* **2009**, *323* (20), 1014-1015.
- 418 (2) Douglas, G. B.; Hamilton, D. P.; Robb, M. S.; Pan, G.; Spears, B. M.; Lürling, M.
- 419 Guiding principles for the development and application of solid-phase phosphorus
- 420 adsorbents for freshwater ecosystems. *Aquat. Ecol.* **2016**, *50* (3), 385-405.

- 421 (3) Wang, L. J.; Pan, G.; Shi, W. Q.; Wang, Z. B.; Zhang, H. G. Manipulating nutrient
- 422 limitation using modified local soils: a case study at Lake Taihu (China). *Water Res.*423 **2016**, *101*, 25-35.
- (4) Carpenter, S. R. Eutrophication of aquatic ecosystems: bistability and soil
 phosphorus. *Proc. Natl. Acad. Sci. USA* 2005, *102* (29), 10002-10005.
- 426 (5) Schindler, D. W.; Hecky, R. E.; Findlay, D. L.; Stainton, M. P.; Parker, B. R.;
- 427 Paterson, M. J.; Beaty, K. G.; Lyng, M.; Kasian, S. E. Eutrophication of lakes cannot
- 428 be controlled by reducing nitrogen input: results of a 37-year whole-ecosystem
- 429 experiment. *Proc. Natl. Acad. Sci. USA* **2008**, *105* (32), 11254-11258.
- 430 (6) Spears, B. M.; Dudley, B.; Reitzel, K.; Rydin, E. Geo-engineering in lakes--a call
- 431 for consensus. *Environ. Sci. Technol.* **2013**, *47* (9), 3953-3954.
- 432 (7) Spears, B. M.; Maberly, S. C.; Pan, G.; Mackay, E.; Bruere, A.; Corker, N.; Douglas,
- 433 G.; Egemose, S.; Hamilton, D.; Hatton-Ellis, T.; Huser, B.; Li, W.; Meis, S.; Moss, B.;
- 434 Lürling, M.; Phillips, G.; Yasseri, S.; Reitzel, K. Geo-engineering in lakes: a crisis of
- 435 confidence? *Environ. Sci. Technol.* **2014**, *48* (17), 9977-9979.
- 436 (8) Copetti, D.; Finsterle, K.; Marziali, L.; Stefani, F.; Tartari, G.; Douglas, G.; Reitzel,
- 437 K.; Spears, B. M.; Winfield, I. J.; Crosa, G.; D'Haese, P.; Yasseri, S.; Lürling, M.
- 438 Eutrophication management in surface waters using lanthanum modified bentonite: A
- 439 review. *Water Res.* **2016**, *97* (15), 162-174.
- 440 (9) Huang, W. Y.; Li, D.; Liu, Z. Q.; Tao, Q.; Zhu, Y.; Yang, J.; Zhang, Y. M. Kinetics,
- 441 isotherm, thermodynamic, and adsorption mechanism studies of La(OH)₃-modified
- 442 exfoliated vermiculites as highly efficient phosphate adsorbents. *Chem. Eng. J.* 2014,
- 443 *236* (2), 191-201.
- 444 (10) Huang, W. Y.; Zhu, Y.; Tang, J. P.; Yu, X.; Wang, X. L.; Li, D.; Zhang, Y. M.
- 445 Lanthanum-doped ordered mesoporous hollow silica spheres as novel adsorbents for

- 446 efficient phosphate removal. J. Mater. Chem. A **2014**, 2 (23), 8839-8848.
- 447 (11) Dithmer, L.; Lipton, A. S.; Reitzel, K.; Warner, T. E.; Lundberg, D.; Nielsen, U. G.
- 448 Characterization of phosphate sequestration by a lanthanum modified bentonite clay: a
- solid-state NMR, EXAFS, and PXRD study. *Environ. Sci. Technol.* 2015, *49* (7), 45594566.
- 451 (12) Zhang, L.; Zhou, Q.; Liu, J. Y.; Chang, N.; Wan, L. H.; Chen, J. H. Phosphate
- adsorption on lanthanum hydroxide-doped activated carbon fiber. *Chem. Eng. J.* 2012, *185-186*, 160-167.
- (13) Ross, G.; Haghseresht, F.; Cloete, T. E. The effect of pH and anoxia on the
 performance of Phoslock[®], a phosphorus binding clay. *Harmful Algae* 2008, 7 (4), 545550.
- 457 (14) Kumar, E.; Bhatnagar, A.; Hogland, W.; Marques, M.; Sillanpää, M. Interaction of
 458 anionic pollutants with Al-based adsorbents in aqueous media A review. *Chem. Eng.*
- 459 *J.* **2014,** *241* (4), 443-456.
- 460 (15) Tanada, S.; Kabayama, M.; Kawasaki, N.; Sakiyama, T.; Nakamura, T.; Araki, M.;
- Tamura, T. Removal of phosphate by aluminum oxide hydroxide. *J. Colloid Interf. Sci.*2003, 257 (1), 135-140.
- (16) Lijklema, L. Interaction of orthophosphate with iron(III) and aluminum hydroxides. *Environ. Sci. Technol.* **1980**, *14* (5), 537-541.
- (17) Reitzel, K.; Jensen, H. S.; Egemose, S. pH dependent dissolution of sediment
 aluminum in six Danish lakes treated with aluminum. *Water Res.* 2013, 47 (3), 14091420.
- 468 (18) Berkowitz, J.; Anderson, M. A.; Amrhein, C. Influence of aging on phosphorus
 469 sorption to alum floc in lake water. *Water Res.* 2006, *40* (5), 911-916.
- 470 (19) Gensemer, R. W.; Playle, R. C. The bioavailability and toxicity of aluminum in

- 471 aquatic environments. *Crit. Rev. Env. Sci. Tec.* **1999**, *29* (4), 315-450.
- (20) Liu, Y. T.; Hesterberg, D. Phosphate bonding on noncrystalline Al/Fe-hydroxide
- 473 coprecipitates. *Environ. Sci. Technol.* **2011**, *45* (15), 6283-6289.
- 474 (21) Khare, N.; Martin, J. D.; Hesterberg, D. Phosphate bonding configuration on
- 475 ferrihydrite based on molecular orbital calculations and XANES fingerprinting.
- 476 *Geochim. Cosmochim. Ac.* **2007,** *71* (18), 4405-4415.
- 477 (22) Khare, N.; Hesterberg, D.; Martin, J. D. XANES investigation of phosphate
- 478 sorption in single and binary systems of iron and aluminum oxide minerals. *Environ*.
- 479 *Sci. Technol.* **2005**, *39* (7), 2152-2160.
- 480 (23) Khare, N.; Hesterberg, D.; Beauchemin, S.; Wang, S. L. XANES determination of
- adsorbed phosphate distribution between ferrihydrite and boehmite in mixtures. Soil
- 482 Sci. Soc. Am. J. 2004, 68 (2), 460-469.
- 483 (24) Xiong, W. H.; Peng, J. Development and characterization of ferrihydrite-modified
- 484 diatomite as a phosphorus adsorbent. *Water Res.* **2008**, *42* (19), 4869-4877.
- 485 (25) Zhang, M. Y.; He, G. Z.; Pan, G. Binding mechanism of arsenate on rutile (110)
- and (001) planes studied using grazing-incidence EXAFS measurement and DFT
- 487 calculation. *Chemosphere* **2015**, *122*, 199-205.
- 488 (26) Lunsford, J. H.; Jayne, J. P. Electron paramagnetic resonance of oxygen on ZnO
- and ultraviolet-irradiated MgO. J. Chem. Phys. **1966**, 44 (4), 1487-1492.
- 490 (27) Liu, Z. E.; Cheng, J. J. EPR study on $YBa_2Cu_{3-x}M_xO_{7-\delta}$ (M=Zn²⁺, Co³⁺) oxide
- 491 ceramics. *Bull. Chin. Ceramic Soc.* **1991**, *1*, 6-9.
- 492 (28) Ali, F.; Chadwicka, A. V.; Smit, M. E. EXAFS analysis of the structural evolution
- 493 of gel-formed La₂O₃. J. Mater. Chem. **1997**, 7 (2), 285-291.
- 494 (29) Parks, G. A. The isoelectric points of solid oxides, solid hydroxides, and aqueous
- 495 hydroxo complex systems. *Chem. Rev.* **1965**, *65* (2), 177-195.

- 496 (30) Wan, B.; Yan, Y. P.; Liu, F.; Tan, W. F.; He, J. J.; Feng, X. H. Surface speciation of
- 497 *myo*-inositol hexakisphosphate adsorbed on TiO₂ nanoparticles and its impact on their

498 colloidal stability in aqueous suspension: A comparative study with orthophosphate.

499 Sci. Total. Environ. 2016, 544, 134-142.

- 500 (31) Liu, J. Y.; Zhou, Q.; Chen, J. H.; Zhang, L.; Chang, N. Phosphate adsorption on
- hydroxyl-iron-lanthanum doped activated carbon fiber. *Chem. Eng. J.* 2013, 215-216,
 859-867.
- 503 (32) Rajan, S. S. S. Changes in net surface charge of hydrous alumina with phosphate
- souther states the second states and sources and sourc
- 505 (33) Takashi, Y.; Takaki, H.; Takahiro, M.; Tsunehiro, T.; Takuzo, F. Structures and
- acid-base properties of La/Al_2O_3 role of La addition to enhance thermal stability of r-
- 507 Al₂O₃. *Chem. Mater.* **2003**, *15* (25), 4830-4840.
- 508 (34) Spears, B. M.; Lürling, M.; Yasseri, S.; Castro-Castellon, A. T.; Gibbs, M.; Meis,
- 509 S.; McDonald, C.; McIntosh, J.; Sleep, D.; Van Oosterhout, F. Lake responses following
- 510 lanthanum-modified bentonite clay (Phoslock[®]) application: an analysis of water
- column lanthanum data from 16 case study lakes. *Water Res.* **2013**, *47* (15), 5930-5942.
- 512 (35) Gerke, J. Phosphate adsorption by humic Fe-oxide mixtures aged at pH 4 and 7
- and by poorly ordered Fe-oxide. *Geoderma* **1993**, *59* (1-4), 279-288.
- (36) Kizewski, F. R.; Boyle, P.; Hesterberg, D.; Martin, J. D. Mixed anion
 (phosphate/oxalate) bonding to Iron(III) materials. *J. Am. Chem. Soc.* 2010, *132* (7),
 2301-2308.
- 517 (37) Tokarz-Sobieraj, R.; Witko, M.; Gryboś, R. Reduction and re-oxidation of
- molybdena and vanadia: DFT cluster model studies. *Catal. Today* 2005, *99* (1-2), 241253.
- 520 (38) Wu, Q. P.; Zheng, Q.; van de Krol, R. Creating oxygen vacancies as a novel

- strategy to form tetrahedrally coordinated Ti^{4+} in Fe/TiO₂ nanoparticles. J. Phys. Chem.
- 522 *C.* **2012,** *116* (12), 7219-7226.
- 523 (39) Perera, D. N.; Harrowell, P. Stability and structure of a supercooled liquid mixture
- 524 in two dimensions. *Phys. Rev. E* **1999**, *59* (5 Pt B), 5721-5743.
- 525 (40) Brückman, K.; Grabowski, R.; Haber, J.; Mazurkiewicz, A.; Słoczyński, J.;
- 526 Wiltowski, T. The role of different MoO₃ crystal faces in elementary steps of propene
- 527 oxidation. J. Catal. **1987**, 104 (1), 71-79.
- 528 (41) Menetrey, M.; Markovits, A.; Minot, C. Reactivity of a reduced metal oxide
- surface hydrogen, water and carbon monoxide adsorption on oxygen defective rutile
- 530 TiO₂ (110). Surf. Sci. 2003, 524 (s 1-3), 49-62.
- 531 (42) Wang, J. A.; Bokhimi, X.; Morales, A.; Novaro, O.; López, T.; Gómez, R.
- Aluminum local environment and defects in the crystalline structure of sol-gel alumina
 catalyst. *J. Phys. Chem. B* 1999, *103* (2), 299-303.
- 534 (43) Abdala, D. B.; Northrup, P. A.; Arai, Y.; Sparks, D. L. Surface loading effects on
- orthophosphate surface complexation at the goethite/water interface as examined by
- extended x-ray absorption fine structure (EXAFS) spectroscopy. J. Colloid Interf. Sci.
- 537 **2015**, *437*, 297-303.
- (44) Wang, Z. B.; Zhang, H. G.; Pan, G. Ecotoxicological assessment of flocculant
 modified soil for lake restoration using an integrated biotic toxicity index. *Water Res.*2016, 97, 133-141.
- (45) Sneller, F. E. C.; Kalf, D. F.; Weltje, L.; Van Wezel, A. P. Maximum permissible
 concentrations and negligible concentrations for rare earth elements (REEs). RIVM
 report 601501011. 2000.
- 544 (46) Biesinge, K. E.; Christen, G. M. Effects of various metals on survival, growth,
- reproduction, and metabolism of Daphnia-magna. J. Fish. Res. Board Can. 1972, 29

- 546 (12), 1691-1700.
- 547 (47) Lee, J. H.; Byrne, R. H. Examination of comparative rare earth element
 548 complexation behavior using linear free-energy relationships. *Geochim. Cosmochim.*549 Ac. 1992, 56 (3), 1127-1137.
- 550 (48) Dithmer, L.; Nielsen, U. G.; Lundberg, D.; Reitzel, K. Influence of dissolved
- organic carbon on the efficiency of P sequestration by a lanthanum modified clay. *Water*
- 552 *Res.* **2016**, *97*, 39-46.
- 553 (49) Borggaard, O. K.; Raben-Lange, B.; Gimsing, A. L.; Strobel, B. W. Influence of
- humic substances on phosphate adsorption by aluminium and iron oxides. *Geoderma*
- **555 2005**, *127* (3-4), 270-279.
- 556 (50) Pan, G.; Dai, L. C.; Li, L.; He, L. C.; Li, H.; Bi, L.; Gulati, R. D. Reducing the
- 557 recruitment of sedimented algae and nutrient release into the overlying water using
- 558 modified soil/sand flocculation-capping in eutrophic lakes. *Environ. Sci. Technol.* 2012,
- *46* (9), 5077-5084.
- 560 (51) Pan, G.; Yang, B.; Wang, D.; Chen, H.; Tian, B. H.; Zhang, M. L.; Yuan, X. Z.;
- 561 Chen, J. In-lake algal bloom removal and submerged vegetation restoration using
- 562 modified local soils. *Ecol. Eng.* **2011,** *37* (2), 302-308.