1	
2	
3	In situ determination of trace elements in <i>Fucus</i> spp. by
4	field-portable-XRF
5	
6	
7	Andrew Turner*, Hiu Poon, Alex Taylor & Murray T. Brown <sup>1</sup>
8	
9	School of Geography, Earth and Environmental Sciences and <sup>1</sup> School of Biological and
10	Marine Sciences, Plymouth University, Drake Circus, Plymouth PL4 8AA, UK
11	
12	
13	
14	*Corresponding author. Tel: +44 1752 584570; Fax: +44 1752 584710; e-mail:
15	aturner@plymouth.ac.uk
16	
17	
18	Accepted 9 March 2017
19	http://dx.doi.org/10.1016/j.scitotenv.2017.03.091
20	Embargo period 1/9/2019
21	
22 23	

#### 24 Abstract

Fresh and freeze-dried sample sections of the coastal macroalgae, *Fucus serratus* and *F*. 25 26 vesiculosus, and the brackish water macroalga, F. ceranoides, have been analysed for trace 27 elements by field-portable-x-ray fluorescence (FP-XRF) spectrometry using a Niton XL3t in 28 a low density mode with thickness correction. When analysed fresh in a laboratory accessory 29 stand for a period of 200 seconds, As, Br, Fe and Zn were registered in the apex, mid-frond and lower stipe of all species, with detection limits of a few  $\mu g g^{-1}$  (As) or a few tens of  $\mu g g^{-1}$ 30 <sup>1</sup> (Br. Fe, Zn); when analysed dry under the same conditions, concentrations returned were 31 32 systematically higher and Cu and Pb were detected in a number of F. ceranoides sections. 33 Concentrations arising from both approaches on a dry weight basis were highly correlated, 34 with deviations from unit slope attributed to the absorption of fluorescent x-rays by internal 35 and surficial water when analysed fresh. With algorithms correcting for the effects of water 36 on mass and x-ray absorption, sections of F. vesiculosus and F. ceranoides were analysed in 37 situ with the XRF connected to a mobile stand and laptop. Dry weight concentrations 38 returned for As and Zn were significantly correlated with respective concentrations 39 subsequently determined by ICP-MS following acid digestion and with a slope close to 40 unity; lower concentrations of Fe returned by ICP were attributed to the incomplete acid 41 digestion of silt particles that evaded an initial cleaning step, while Br concentrations could 42 not be verified independently because of loss of volatile forms during digestion. The in situ 43 determination of trace elements in fucoids by FP-XRF provides a rapid and non-destructive 44 means of monitoring environmental quality and identifying hot-spots of contamination, and enables a research strategy to be developed iteratively that is informed by immediate results. 45 46

47 **Keywords:** macroalgae; fucoids; FP-XRF; monitoring; trace elements; ICP-MS

#### 48 **1. Introduction**

Marine macroalgae represent a large and diverse group of primary-producing organisms in 49 50 the coastal zone that create habitat structure, provide food, promote biodiversity and serve as 51 a carbon sink (Duarte et al., 2013; Matias et al., 2015). Macroalgae also have an economic 52 value, acting as a source of food, nutrients and medicines for human consumption and 53 offering a means of bioremediation and a potential bioenergy resource (Bruhn et al., 2011; 54 Tabarsa et al., 2012). Being sessile, macroalgae are influenced directly by ambient 55 environmental conditions, and in this respect the distribution and occurrence of certain 56 species may often reflect local water quality (Guinda et al., 2008). Moreover, because of 57 their thick cell walls and high polysaccharide content, macroalgae are also able to 58 accumulate many aqueous contaminants, and in particular trace metals and metalloids, to 59 concentrations several thousand times higher than the ambient water column (Zbikowski et 60 al., 2006). Consequently, many species act as vehicles for the transfer of contaminants up the 61 food chain (Chan et al., 2003; Mulholland and Turner, 2011) and serve as useful biomonitors 62 that provide a direct and integrated assessment of bioavailable contaminants over a period of 63 time (Cairrão et al., 2007; Boubonari et al., 2008). The littoral brown fucoids are particularly 64 useful in the latter respect because of their extensive distribution, ease of identification and 65 sampling, tolerance of wide variations of temperature and salinity, abundance all year round 66 and limited ability to regulate contaminant concentrations (Martin et al., 1997; Rainbow, 67 2006; Sondergaard et al., 2014). Accordingly, Fucus spp. have been selected for inclusion in 68 the Environmental Specimen Banks (ECBs) of several European countries in order to 69 monitor long-term changes in anthropogenic contamination (Viana et al., 2010; Rüdel et al., 70 2010).

71

As commonly employed biomonitor organisms, there is a requirement for the routine
determination of trace elements in macroalgae. Conventionally, analysis is performed on

74 dried samples that have been digested in hot, concentrated acid by, for example, atomic 75 absorption spectrometry (Reis et al., 2014) or inductively coupled plasma (ICP) 76 spectrometry (Brito et al., 2012). This approach can, however, be time-consuming, labour-77 intensive and costly, and the destruction of samples has implications for the long-term 78 viability of archived specimen banks. Recently, we investigated the feasibility of field 79 portable-x-ray fluorescence (FP-XRF) spectrometry as a rapid, non-destructive means of 80 determining trace elements in various species of macroalgae (Bull et al., 2017). Specifically, 81 we employed a Niton XL3t spectrometer configured in a low density, 'plastics' mode and 82 with a corrective algorithm for sample thickness to measure dried samples housed in a 83 laboratory accessory stand. For the elements that were detected, there was a significant 84 correlation between concentrations measured directly and those returned independently by 85 ICP-mass spectrometry following HNO<sub>3</sub> digestion, with relationships satisfying the EPA definitive level criterion for As and quantitative screening level for Cu and Zn 86 87 (Environmental Protection Agency, 2007).

88

89 In order to further the XRF approach for measurements of trace elements in macroalgae in 90 situ, the effects of water, as a contributor to both sample mass and x-ray absorption, need to 91 be accounted for and factored in to any calibration. To this end, the present study compares 92 trace element concentrations returned by FP-XRF for different species of macroalgae analysed both in the fresh and freeze-dried states. Specifically, fucoids were selected because 93 94 of their importance in coastal biomonitoring and their relatively high thickness (for x-ray 95 absorption) compared with other species of macroalgae (Bull et al., 2017). With the effects 96 of water empirically quantified, the practicalities and challenges of deploying the XRF in the 97 field for the direct measurement of trace elements in macroalgae are discussed.

98

99

#### 101 **2. Materials and methods**

# 102 2.1. Sampling and sample preparation

103 Whole samples of fucoid were handpicked at low tide in July 2016 from two sites within 25 104 minutes' driving distance of the laboratory at Plymouth University (Figure 1). At Firestone 105 Bay, a small, pebble-sand beach in the coastal embayment of Plymouth Sound, five 106 specimens of two coastal fucoids, F. serratus and F. vesiculosus, were collected from the 107 rocky substrates of the littoral zone and stored in a cool box in a series of zip-lock 108 polyethylene bags. From the intertidal mudflats of the upper Tavy Estuary, a tidal tributary 109 of the Tamar Estuary and an environment impacted by historical mining activities, ten 110 specimens of F. ceranoides, a brackish water fucoid, were collected and stored likewise. In 111 the laboratory, samples of F. serratus and F. vesiculosus were divided and subjected to two 112 different methods of clearing sediment and epiphytes from the surface; thus, one half was 113 cleaned in Millipore Milli-Q water (MQW) with the aid of a Nylon brush and subsequently 114 scraped with a plastic spatula after applying a 10% solution of ethanol to the tissue surface, 115 while the other half was cleaned with MQW only. After blotting dry with 3-ply hygiene roll, 116 all plants were dissected on a plastic tray using a stainless steel blade, with ~ 5 cm sections 117 of the apex, mid-frond and lower stipe (just above the holdfast) from each plant retained and 118 stored in individual specimen bags. Because of the smaller size of F. ceranoides and results 119 arising from the cleaning methods of the two coastal macroalgae, samples of the brackish 120 water fucoid were cleaned in MQW only before being dissected likewise.

- 121
- 122
- 123
- 124
- 125
- 126



128 Figure 1: The sampling locations for the fucoid macroalgae.

129

# 130 2.2. FP-XRF analysis

Sample sections processed in the laboratory (n = 90) were analysed for trace elements (As, 131 132 Br, Cd, Cr, Cu, Fe, Hg, Ni, Pb and Zn) directly and without drying by energy dispersive FP-133 XRF using a battery-powered, 1.3 kg Niton analyser (model XL3t 950 He GOLDD+) housed 134 in a ThermoScientific accessory stand of steel construction and tungsten-plastic shielding (PN 420-017; weight ~ 10 kg, chamber volume =  $4000 \text{ cm}^3$ ). Analysis was performed in a 135 136 low density mode that uses a fundamental parameters-based alpha coefficient correction 137 model (Turner and Solman, 2016). Because the intensity of fluorescence generated by low 138 density and weakly absorbing samples is dependent on the thickness of material, a corrective 139 algorithm (down to 50 µm) was also applied after section thickness had been measured in 140 mm and to two decimal places using digital callipers. With plastic tweezers, samples were

141 placed onto a SpectraCertified Mylar polyester 3.6 µm film, which was then positioned 142 carefully such that the smoothest and flattest part of the macroalgal section lay directly and centrally above the 8 mm XRF detector window. After closing the accessory stand lid, the 143 144 XRF was activated remotely and via USB using a Fujitsu laptop computer. Analysis was 145 tested for a variety of conditions of which a collimation of 8 mm and a counting period of 146 200 seconds, comprising 150 seconds at 50 kV and 40 µA and 50 seconds at 20 kV and 100 147  $\mu$ A, appeared to be optimal in terms of detection, error and sample throughput. To check the 148 performance of the XRF and as an analytical quality control, Niton polyethylene reference discs impregnated with known concentrations of various trace elements (PN 180-619, 149 150 LOT#T-18 and PN 180-554, batch SN PE-071-N) were analysed throughout each 151 measurement session. On completion of measurements, spectra and elemental concentrations (in  $\mu g g^{-1}$  and with a counting error of  $2\sigma$ ) were downloaded to the laptop using Niton Data 152 153 Transfer PC software.

154

Immediately after sample measurement, individual macroalgal sections were weighed using a five-figure Sartorius analytical balance before being returned to their original specimen bags and freeze-dried for 48 h using an Edwards Super Modulyo. Dried sections were then re-analysed by XRF under the operating conditions described above and after appropriate (dry) thickness correction, before being re-weighed, returned to their specimen bags and stored under desiccation pending acid digestion (see below).

161

# 162 2.3. Macroalgae digestion and analysis by ICP-MS

163 As an independent measure of trace elements in the macroalgae, all freeze-dried sample

164 sections were subsequently acid-digested and analysed by inductively coupled plasma-mass

spectrometry (ICP-MS). Thus, samples of about 0.1 g were accurately weighed into

166 individual Teflon tubes to which 2.5 ml aliquots of HNO<sub>3</sub> (Fisher Chemical TraceMetal<sup>TM</sup>

Grade) were added. The contents were digested in a CEM MARS 5 XPRESS microwave at
1600 W for 45 min before being allowed to cool to room temperature. Digests were then
washed into individual 10 ml volumetric flasks and diluted to mark with ultra-pure Millipore
Milli-Q water. For an assessment of digestion efficacy and analytical accuracy, a fucoid
reference material (*Fucus vesiculosus*, ERM-CD200; certified for As, Br, Cd, Cu, Fe, Hg,
Pb, Se and Zn) was digested in triplicate likewise.

173

174 Digests were analysed for elements that had been detected by XRF using a collision cell-175 ICP-MS (Thermo X-series II, Thermoelemental, Winsford, UK) with a concentric glass 176 nebuliser and conical spray chamber. RF power was set at 1400 W and coolant, auxiliary, nebuliser and collision cell gas flows rates were 13 L Ar min<sup>-1</sup>, 0.70 L Ar min<sup>-1</sup>, 0.72 L Ar 177 min<sup>-1</sup> and 3.5 mL 7% H<sub>2</sub> in He min<sup>-1</sup>, respectively. The instrument was calibrated externally 178 179 using four mixed standards prepared by dilutions of a QC 26 multi-element solution (CPI International, Amsterdam) in 0.1 M HNO<sub>3</sub>, and internally by the addition of 100  $\mu$ g L<sup>-1</sup> of In 180 181 and Ir to all samples and standards. Data were acquired over a dwell period of 10 ms, with 182 50 sweeps per reading and three replicates.

183

Aqueous concentrations derived from ICP-MS were converted to dry weight concentrations (in  $\mu$ g g<sup>-1</sup>) from the volume of diluted digest and mass of macroalga dissolved in acid. Limits of detection on this basis were < 2.5  $\mu$ g g<sup>-1</sup> for all trace elements analysed, and measured concentrations in the reference macroalga were within 15% of published values with the exception of Br and Fe (recoveries of about 50% and 70%, respectively).

189

190 2.4. Statistical analysis

191 Correlation analysis was performed on paired data series using the Data Analysis Toolpak in

192 Excel 2016, with the strength of association reported as Pearson's moment correlation

193 coefficient (*r*) and the significance of the relationship as the probability of *r* not being

194 different from zero (*p*, and where  $\alpha = 0.05$ ). One-way ANOVA and Tukey's post-hoc test

195 were used in Minitab 17 to identify significant differences ( $\alpha = 0.05$ ) in mean elemental

196 concentrations and water contnets among macroalgae and parts thereof, and in mean

197 elemental concentrations arising from the three analytical methods.

198

## **3. Results and Discussion**

#### 200 3.1. Macroalgal water content and thickness

201 Quantification of the water content of the macroalgal sections is critical for converting elemental concentrations from a fresh weight basis to a dry weight basis and for evaluating 202 203 the impact of the fluid on x-ray behaviour and intensity (mainly through photoelectric 204 absorption, but also via Compton scattering and internal reflections; Parsons et al., 2013). 205 Mean percentage water, calculated from the fresh and dry weights of each section and shown 206 in Table 1, ranged from about 50% to nearly 90%, and for all species the order of descending 207 water content was: apex > mid-frond > lower stipe. There was no statistical difference in 208 water content between common sections of F. vesiculosus and F. serratus, but the water 209 content of sections of F. ceranoides were significantly greater than corresponding sections of 210 the former two species. The method of tissue cleaning made a difference to mean water 211 content that was significant only for the lower-stipe of *F. vesiculosus* from Firestone Bay. 212 Thus, here, cleaning in MQW resulted in a higher percentage compared with sections having 213 undergone additional cleaning with ethanol 214 215 Fucoid section thickness, measured for XRF data correction, did not display a clear

216 dependency on species, location with respect to the frond or means of tissue cleaning. On

217 average, however, sections were thicker while wet  $(1.02 \pm 0.15 \text{ mm})$  than when freeze-dried

218  $(0.85 \pm 0.19 \text{ mm}).$ 

- 219
- 220 Table 1: Percentage water content of the fucoid macroalgal sections undergoing cleaning in
- 221 Milli-Q water (MQW) and ethanol, and/or MQW only. The mean and standard deviation of

	<i>F.</i>	serratus (n=5)	F. ve	siculosus (n=5)	F. ceranoides (n=10)
	MQW	MQW+ethanol	MQW	MQW+ethanol	MQW
арех	81.2 <u>+</u> 1.8	76.3 <u>+</u> 2.4	77.7 <u>+</u> 2.2	73.1 <u>+</u> 1.7	86.7 <u>+</u> 1.9
mid-frond	67.7 <u>+</u> 4.0	63.7 <u>+</u> 5.1	62.6 <u>+</u> 3.7	60.3 <u>+</u> 2.8	76.9 <u>+</u> 4.8
lower stipe	61.1 <u>+</u> 1.7	54.2 <u>+</u> 5.3	61.8 <u>+</u> 1.3	51.8 <u>+</u> 2.1	70.5 <u>+</u> 3.6

222 *n* measurements is given in each case.

223

## 224 3.2. XRF detection limits for trace elements in macroalgae

225 XRF detection limits for trace elements in the fucoids, defined as three counting errors for a 226 200-second counting time, are presented in Table 2. Here, limits for all species, sectional 227 locations and cleaning methods have been pooled and are shown for samples analysed in 228 both the fresh state and after freeze-drying; with regard to the former, limits are shown on a 229 fresh weight basis and, after correction for water content, a dry weight basis. Note that for 230 some elements (Cd, Cr, Cu, Hg, Ni, Pb) limits have been averaged from at least fifteen 231 measurements in which the element was not detected by the instrument but a value of  $3\sigma$ was returned directly; where less than fifteen sample sections were undetectable (As, Br, Fe, 232 233 Zn), limits were based on the values of  $2\sigma$  returned on detection and after multiplication by 234 1.5.

235

Mean detection limits are generally lower when samples are analysed fresh than when freeze-dried, presumably because the greater flexibility of wet macroalgal sections allows them to be placed closer to the detector window of the instrument. However, when wet weight concentrations are converted to a dry weight basis, detection limits are higher than samples analysed dry. Here, we surmise that the effects of water on elemental dilution and x-

241	ray absorption and scattering outweigh the benefits of increased proximity to the detector.
242	Overall, mean detection limits are lowest and average less than 10 $\mu g~g^{\text{-1}}$ (on both a dry
243	weight and wet weight basis) for As and Pb and are less than 25 $\mu$ g g <sup>-1</sup> for Br, Cu, Hg, Ni
244	and Zn, and are similar to corresponding limits reported for dried sections of F. serratus
245	reported by Bull et al. (2017). Within these constraints, As and Fe were detected in all fucoid
246	section analyses performed in the present study ( $n = 180$ ), while Br and Zn were detected in
247	178 and 172 cases, respectively, with non-detection always associated with the analysis of
248	fresh samples. Note that although Cu and Pb were detected in some samples of $F$ .
249	<i>ceranoides</i> , the number of cases ( $n = 7$ and $n = 5$ , respectively) was too few for establishing
250	relationships between the different analytical approaches and differences among the three
251	sectional components of the macroalga.

Table 2: A summary (as mean  $\pm$  one standard deviation; n > 15) of the Niton XRF detection limits for trace elements in fucoid macroalgae analysed fresh and dry and for a 200-second counting time (dw = dry weight; fw = fresh weight).

	As	Br	Cd	Cr	Cu	Fe	Hg	Ni	Pb	Zn
dry, μg g <sup>-1</sup> dw	4.7 <u>+</u> 1.0	17.9 <u>+</u> 4.6	26.3 <u>+</u> 2.3	6.1 <u>+</u> 2.2	16.2 <u>+</u> 4.2	24.5 <u>+</u> 6.7	12.3 <u>+</u> 2.8	14.4 <u>+</u> 3.8	5.9 <u>+</u> 1.0	11.7 <u>+</u> 2.7
fresh, µg g⁻¹ fw	1.6 <u>+</u> 0.2	4.7 <u>+</u> 1.2	14.4 <u>+</u> 2.4	7.9 <u>+</u> 2.0	5.3 <u>+</u> 1.0	6.8 <u>+</u> 1.3	5.0 <u>+</u> 1.2	5.0 <u>+</u> 1.0	2.3 <u>+</u> 0.5	3.5 <u>+</u> 1.1
fresh, μg g⁻¹ dw	6.6 <u>+</u> 2.0	20.1 <u>+</u> 9.5	61.0 <u>+</u> 23.4	32.3 <u>+</u> 11.9	22.2 <u>+</u> 8.5	28.2 <u>+</u> 10.1	21.2 <u>+</u> 9.0	21.0 <u>+</u> 8.9	9.7 <u>+</u> 3.8	14.3 <u>+</u> 5.5

257

256

### 258 3.3. Comparison of elemental concentrations when analysed wet and dry

Figure 2 compares the dry weight concentrations of readily detectable elements (As, Br, Fe, Zn) in the fucoids that were returned by the Niton XRF when analysed dry, [XRF-*dry*], and when analysed fresh and individually corrected for water content, [XRF-*fresh*]. Note that here, data for each element are not discriminated by species, location on the frond or means of tissue cleaning. Also shown are the best-fit equations (forced through the origin) that define each element, along with corresponding Pearson's moment correlation coefficients 265 and the line signifying unit slope. In all cases, and despite changes in thickness and morphology incurred by freeze-drying, elemental concentrations arising from both 266 267 approaches were highly correlated (p < 0.01) with gradients exceeding unit value; that is, concentrations returned when analysed dry were, on average, higher than concentrations 268 269 returned when analysed fresh but dry-weight corrected. This suggests that the presence of 270 internal and surficial water suppresses the strength of fluorescent x-rays reaching the 271 detector window of the FP-XRF through absorption and scattering. Consistent with this 272 assertion, deviation from unit slope is greatest for Fe, whose characteristic x-rays are of low 273 energy ( $K_{\alpha} = 6.405 \text{ keV}$ ;  $L_{\alpha} = 0.705 \text{ keV}$ ) and relatively easily absorbed by water, and least 274 for Br, whose characteristic x-rays are of higher energy ( $K_{\alpha} = 11.924 \text{ keV}$ ;  $L_{\alpha} = 1.481 \text{ keV}$ ) 275 and, therefore, less easily absorbed.



Figure 2: Dry weight elemental concentrations in the coastal and brackish water fucoid
macroalgae returned by the Niton XRF when analysed dry and fresh. Shown inset for each
element are equations of best fit when forced through the origin.

3.4. Inter- and intra-species variations in elemental concentrations and comparison with
 ICP-MS

283 Figures 3 to 6 show the dry-weight concentrations of As, Br, Fe and Zn in the different parts 284 of each species of fucoid and as determined by the two XRF approaches (that is, analysis of 285 fresh sections versus analysis of freeze-dried sections) and by ICP-MS following acid digestion. Note that all data presented are for tissues cleaned in MQW only and that results 286 287 arising from samples subjected to additional cleaning with ethanol were very similar. 288 Regarding As, mean concentrations were not statistically different among the different 289 methods of determination with the exception of the lower stipe in *F. serratus*, where 290 concentrations were lower when analysed by XRF in the fresh state than by ICP, and the 291 apex in F. ceranoides, where concentrations were higher when analysed dry by XRF. Among the different parts of the frond, mean concentrations were generally higher in the 292 293 apex than the mid-frond and lower stipe, an effect that was evident from each analytical 294 approach in at least one species of fucoid. Overall, absolute concentrations of As were 295 greatest in the apex of F. ceranoides, and concentrations were significantly greater in the mid-frond and lower stipe of *F. ceranoides* than in corresponding parts of both *F. serratus* 296 297 and F. vesiculosus according to at least one analytical approach.

298

With respect to Br, results arising from ICP-MS analysis have been neglected due to loss of
volatile forms (e.g. HBr and Br<sub>2</sub>) during acid-oxidizing digestion, an effect that is often
significant when opening the digestion vessel at the end of the mineralisation process (Di
Narda et al., 2001). Mean concentrations of the halogen were never statistically different
between the two XRF approaches and concentrations were not different among the three
sectional components of *F. serratus*. Concentrations were, however, significantly lower in

305 the stipe of *F*. *vesiculosus* than in its apex, and significantly higher in the stipe of *F*.

*ceranoides* than in the apex where the lowest overall mean concentrations were observed.

308	Among the elements readily detected, concentrations of Fe were most variable among
309	replicates. Consequently, there were no statistical differences observed between the two
310	XRF approaches, despite mean concentrations returned being double when analysed dry in
311	some cases. Determination by ICP-MS returned significantly lower concentrations than one
312	or both XRF approaches (and by factors up to an order of magnitude) for the mid-fronds of
313	both F. serratus and F. ceranoides and the apex and lower stipe of the latter.
314	
315	Statistical differences in the mean concentrations of Zn were observed among the three
316	analytical approaches only for the lower stipe of F. vesiculosus (lower by XRF after section
317	drying), and the apex of F. serratus and apex and mid-frond of F. ceranoides (higher when
318	analysed by XRF after drying than by both other approaches). With the exception of the apex
319	analysed by XRF when fresh, mean concentrations of Zn were always statistically higher in
320	F. ceranoides than corresponding concentrations in F. vesiculosus. In fewer cases, mean
321	concentrations were higher in <i>F. ceranoides</i> than in <i>F. serratus</i> and in <i>F. serratus</i> than in <i>F.</i>
322	vesiculosus.
323	
324	
325	
326	
327	
328	
329	
330	

Figure 3: Dry weight concentrations of As in the different parts of the fucoid species and as
returned by FP-XRF analysis of fresh sections (grey bars) and dry sections (open bars) and
by ICP-MS analysis following acid digestion (hatched bars). Errors represent the standard







337 Figure 4: Dry weight concentrations of Br in the different parts of the fucoid species and as

returned by FP-XRF analysis of fresh sections (grey bars) and dry sections (open bars).

339 Errors represent the standard deviation about the mean of *n* measurements.

Figure 5: Dry weight concentrations of Fe in the different parts of the fucoid species and as returned by FP-XRF analysis of fresh sections (grey bars) and dry sections (open bars) and by ICP-MS analysis following acid digestion (hatched bars). Errors represent the standard deviation about the mean of *n* measurements.



366 Figure 6: Dry weight concentrations of Zn in the different parts of the fucoid species and as

367 returned by FP-XRF analysis of fresh sections (grey bars) and dry sections (open bars) and

368 by ICP-MS analysis following acid digestion (hatched bars). Errors represent the standard

369 deviation about the mean of *n* measurements.



## 371 *3.5. Summary and implications of findings*

In a previous article, we demonstrated the potential of FP-XRF for determining trace element 372 373 concentrations in different species of dried coastal macroalgae in a laboratory accessory 374 stand (Bull et al., 2017). The technique has distinct advantages over conventional methods 375 involving sample digestion that include reduced time and costs, non-destruction of material 376 (of particular significance to archived specimen banks), increased sample throughput, 377 minimal operator training, capability of exploring tissue spatial variability and avoidance of 378 hazardous wastes. Because monitoring in situ requires direct analysis without drying, 379 however, the present study evaluated the effects of the presence of internal and surficial 380 water on elemental concentrations returned for various fucoids. Thus, a comparison of 381 results arising from the analysis of fresh sections that had been dry-weight normalised and 382 the analysis of sections that had been subsequently freeze-dried revealed a greater sensitivity 383 of the latter approach but results that were highly correlated for all elements considered. 384 Lower dry-weight concentrations returned when analysed fresh are attributed to the absorption of characteristic x-rays by water contained within or at the surface of the 385 386 macroalga.

387

388 Since variations in the percentage water in a given section of fucoid were small, with 389 relative standard deviations of less than 5% in most cases, instantaneous, quantitative 390 correction for macroalgal water content may be readily accomplished through species- and 391 section-specific algorithms; alternatively, it is possible that the fluorescence of Cl ( $K_{\alpha} = 2.62$ ) 392 keV,  $K_{\beta} = 2.82$  keV) could be used as a direct proxy for water content if local salinity is 393 known (Tjallingii et al., 2007). Additional, element-specific corrections for x-ray absorption 394 by water based on the gradients of the relationships between samples analysed fresh and dry 395 (Figure 2) would also be required for complete quantification of concentrations on a dry 396 weight basis. In practice, corrections for the effects of water may be stored in the Niton XRF

397 software as alternative calibrations in the low density mode by adding appropriate slopes398 and, if necessary, intercepts.

399

In most cases, dry-weight concentrations of As and Zn obtained by the analysis of fresh and 400 401 dried sections of fucoids by FP-XRF were not statistically different to corresponding 402 concentations derived independently by ICP-MS following acid digestion. For Fe in the 403 estuarine macroalga, F. ceranoides, however, we attribute significantly lower results arising 404 from ICP analysis to the incomplete release of Fe from the macroalga and to the presence of 405 silt particles on the tissue surface that evaded cleaning and that were detected by the XRF 406 but not completely digested by HNO<sub>3</sub>. Among the elements analysed, the latter effect would 407 be most significant for Fe given its high concentration in fine sediment from the Tavy Estuary (about 60,000  $\mu$ g g<sup>-1</sup> determined on dried, intertidal silt by FP-XRF in a higher 408 density, 'mining' mode, and compared with As and Zn concentrations of 90 and 250  $\mu$ g g<sup>-1</sup>, 409 respectively). The heterogeneous dispersion of silt on the tissue surface would also account 410 411 for the relatively high variability of Fe concentrations measured by XRF among replicates of 412 the same sample section.

413

# 414 *3.4. Deployment of the XRF in situ*

415 With the effects of macroalgal water evaluated and quantified, the feasibility of employing 416 the Niton FP-XRF spectrometer in situ was tested. Thus, the Tavy Estuary was revisited and 417 sections from F. ceranoides and F. vesiculosus analysed under the operating conditions described above (instrument mode, counting time, energy ranges) after cleaning in MQW, 418 419 dissection, blotting dry and thickness measurement with callipers. Initial attempts using the 420 XRF handheld against sections placed on a solid but smooth surface (e.g. a plastic tray on a 421 flat rock) and activated manually via the tilting touchscreen proved unsuccessful for a 422 number of reasons. For example, positioning the XRF window such that it covered the

423 macroalgal section completely was difficult, despite the aid of live video-footage generated 424 by a colour charge-coupled device camera and sampling imaging system adjacent to the 425 detector; moreover, once positioning had been accomplished, holding the instrument still for 426 a suitable length of time against the slimy, fucoid surface was not possible. A moving x-ray 427 source over a low density, irregular sample also poses a safety hazard to the operator though 428 radiation scattering; although this hazard could be minimised by using a backscatter collar-429 shield around the nose of the instrument (Figure 7a), the additional size and weight of 430 equipment further inhibited accurate and steady positioning of the detector window.

431

432 Successful application of the XRF in the field was, however, accomplished when coupled to a lightweight (~ 2.5 kg) and small-volume (300 cm<sup>3</sup>) mobile test-stand (ThermoScientific, 433 434 PN 430-032) and laptop (Figure 7b). Here, the test-stand was placed on a level, stable 435 surface and the instrument subsequently securely fixed to the steel baseplate with the nose 436 pointing upwards. Individual sample sections were placed on polyester film and positioned 437 centrally over the detector window with the aid of plastic tweezers and, if necessary, held 438 flat and in place with weights (e.g. small stones) at each end (Figure 7c). Once the shielded 439 (tungsten-plastic) stand lid was closed, measurements were activated remotely using the 440 laptop and via USB.

441

Essentially, this is the same approach as that employed in the laboratory using the accessorystand. Additional benefits of performing measurements in situ, however, include the development of a strategy or focus that is iterative or directly informed by immediate results, identification of contamination hot-spots, elements of concern or the effects of a pollution incident, determination of which samples to return to the laboratory for further characterisation, and little or no degradation of macroalgae should transport to the laboratory be otherwise time-consuming. With three people in the field and working concurrently on

separate tasks (sampling, sample processing and analysis), algal section throughput for a 200 second counting time was about 15 per hour, and with a single, fully-charged battery, the XRF could be deployed for a period of up to six hours. Given the weight of equipment involved (about 15 kg for the XRF, stand, laptop and cases), set up and measurement are also possible with a single operator, although throughput would be significantly reduced because of the requirement for an individual to conduct multiple tasks successively or concurrently.

456

457 For different sections analysed in situ, concentrations measured directly were converted to 458 dry weight concentrations using the average (generic) percentage water for a given type of 459 section of a particular species and subsequently corrected for x-ray water absorption by 460 applying the element-specific gradients defining the relationship between samples analysed 461 fresh and dry (Figure 2). In Figure 8, results for As and Zn derived accordingly, [XRF-in situ], are shown for samples in which concentrations were subsequently determined by ICP-462 463 MS following drying and acid digestion. For both elements, correlations were significant (p 464 < 0.05) with r values exceeding the US EPA quantitative screening criterion of 0.7 465 (Environmental Protection Agency, 2007). For Fe, concentrations derived in situ were 466 significantly correlated with but considerably higher than those derived independently by 467 ICP-MS for reasons outlined above. With respect to F. vesiculosus in the Tavy, mean concentrations of As and Zn derived in situ (67 and 200  $\mu$ g g<sup>-1</sup>) are also similar to mean 468 values reported in the literature for the upper estuary (86 and 382  $\mu$ g g<sup>-1</sup> respectively; 469 470 Rainbow et al., 2011).

471

472

473

474



- 492 Figure 7: (a) The Niton XL3t plus backscatter shield; (b) configuration of the instrument in
- 493 situ and coupled to the portable stand and laptop; (c) a fucoid section placed above the
- 494 detector window and within the stand.

499 Figure 8: Relationship between As and Zn concentrations in *F. ceranoides* and *F.* 

500 *vesiculosus* determined by ICP-MS following acid digestion and by FP-XRF deployed in





## 515 **4. Conclusions**

516 Although FP-XRF does not have the capability of sub-part per million analyses to replace 517 atomic or mass spectrometry, this study has shown that the Niton XL3t provides a rapid, 518 cost-effective and non-destructive means of measuring various trace elements in both fresh 519 and dry fucoid species of macroalgae, provided that a low density mode with thickness 520 correction is employed. The analytical conditions described (mode of application, 521 collimation, counting time, energy ranges) allow the ready quantification of As to dry weight concentrations down to a few  $\mu g g^{-1}$  and Br, Fe and Zn to concentrations of a few tens of  $\mu g$ 522 g<sup>-1</sup>; measurement of Cu and Pb in fucoids is also possible in moderately to highly 523 contaminated sites. Coupled to a mobile test-stand and laptop, the instrument can be 524

525	deployed in situ for rapid diagnostic and strategic purposes and to evaluate intra- and inter-
526	specific concentration variations, with full quantification possible after empirical adjustment
527	of data for the effects of water on sample weight and x-ray absorption.
528	
529	Acknowledgements
530	We are grateful to Dr Andrew Fisher for technical support during the ICP-MS analysis. This
531	study was funded partly by a UoP HEIF V Marine Institute grant.
532	
533	References
534	Brito, G.B., de Souza, T.L., Bressy, F.C., Moura, C.W.N., Korn, M.G.A., 2012. Levels and
535	spatial distribution of trace elements in macroalgae species from the Todos os Santos Bay,
536	Bahia, Brazil. Marine Pollution Bulletin 64, 2238-2244.
537	
538	Boubonari, T., Malea, P., Kevrekidis, T., 2008. The green seaweed Ulva rigida as a
539	bioindicator of metals (Zn, Cu, Pb and Cd) in a low-salinity coastal environment.
540	Botanica Marina 51, 472-484.
541	
542	Bruhn, A., Dahl, J., Nielsen, H.B., Nikolaisen, L., Rasmussen, M.B., Markager, S., Olesen,
543	B., Arias, C., Jensen, P.D., 2011. Bioenergy potential of Ulva lactuca: Biomass yield,
544	methane production and combustion. Bioresource Technology 102, 2595-2604.
545	
546	Bull, A., Brown, M.T., Turner, A., 2017. Novel use of field-portable-XRF for the direct
547	analysis of trace elements in marine macroalgae. Environmental Pollution 220, 228-233.

549	Cairrão, E., Pereira, M.J., Pastorinho, M.R., Morgado, F., Soares, A.M.V.M., Guilhermino,
550	L., 2007. Fucus spp. as a mercury contamination bioindicator in costal areas (Northwestern
551	Portugal). Bulletin of Environmental Contamination and Toxicology 79, 388-395.
552	
553	Chan, S.M., Wang, W.X., Ni, I.H., 2003. The uptake of Cd, Cr, and Zn by the macroalga
554	Enteromorpha crinita and subsequent transfer to the marine herbivorous rabbitfish, Siganus
555	canaliculatus. Archives for Environmental Contamination and Toxicology 44, 298-306.
556	
557	Di Narda, F., Toniolo, R., Bontempelli, G., 2001. Improved microwave digestion procedure
558	for inductively coupled plasma mass spectrometric determinations of inorganic bromide
559	residues in foodstuffs fumigated with methyl bromide. Analytica Chimica Acta 436, 245-
560	252.
561	
562	Duarte, C.M., Losada, I.J., Hendricks, I.E., Mazarrasa, I., Marba, N., 2013. The role of
563	coastal plant communities for climate change mitigation and adaptation. Nature Climate
564	Change 3, 961-968.
565	
566	Environmental Protection Agency, 2007. Method 6200 - Field portable x-ray fluorescence
567	spectrometry for the determination of elemental concentrations in soil and sediment.
568	http://www3.epa.gov/epawaste/hazard/testmethods/sw846/pdfs/6200.pdf. Accessed 7/16.
569	
570	Guinda, X., Juanes, J.A., Puente, A., Revilla, J.A., 2008. Comparison of two methods for
571	quality assessment of macroalgae assemblages, under different pollution types. Ecological
572	Indicators 8, 743-753.
573	

574	Martin, M.H.,	Nickless, G.	, Stenner,	R.D.,	1997.	Concentrations of	cadmium,	copper,	lead
-----	---------------	--------------	------------	-------	-------	-------------------	----------	---------	------

575 nickel and zinc in the alga *Fucus serratus* in the Severn estuary from 1971 to 1995.

576 Chemosphere 43, 25–334.

577

- 578 Matias, M.G., Arenas, F., Rubal, M., Pinto, I.S., 2015. Macroalgal composition determines
- 579 the structure of benthic assemblages colonizing fragmented habitats. PLOS ONE 10, article

580 e0142289, DOI: 10.1371/journal.pone.0142289.

581

Mulholland, R., Turner, A., 2011. Accumulation of platinum group elements by the marine
gastropod *Littorina littorea*. Environmental Pollution 159, 977-982.

584

585 Parsons, C., Grabulosa, E.M., Pili, E., Floor, G.H., Roman-Ross, G., Charlet, L., 2013.

586 Quantification of trace arsenic in soils by field-portable x-ray fluorescence spectrometry:

587 Considerations for sample preparation and measurement conditions. Journal of Hazardous

588 Materials 262, 1213-1222.

589

Rainbow, P.S., 2006. Biomonitoring of trace metals in estuarine and marine environments.
Australasian Journal of Ecotoxicology 12, 107-122.

592

Rainbow, P.S., Kriefman, S., Smith, B.D., Luoma, S.N., 2011. Have the bioavailabilities of
trace metals to a suite of biomonitors changed over three decades in SW England estuaries
historically affected by mining? Science of the Total Environment 409, 1589-1602.

- 597 Reis, P.A., Cassiano, J., Veiga, P., Rubal, M., Sousa-Pinto, I., 2014. Fucus spiralis as
- 598 monitoring tool of metal contamination in the northwest coast of Portugal under the

599	European Water Framework Directives. Environmental Monitoring and Assessment 186,
600	5447-5460.

602	Rüdel, H., Fliedner, A., Kosters, J., Schroter-Kermani, C., 2010. Twenty years of elemental
603	analysis of marine biota within the German Environmental Specimen Bank-a thorough look
604	at the data. Environmental Science and Pollution Research 17, 1025-1034.
605	
606	Sondergaard, J., Bach, L., Gustavson, K., 2014. Measuring bioavailable metals using
607	diffusive gradients in thin films (DGT) and transplanted seaweed (Fucus vesiculosus), blue
608	mussels (Mytilus edulis) and sea snails (Littorina saxatilis) suspended from monitoring
609	buoys near a former lead-zinc mine in West Greenland. Marine Pollution Bulletin 78, 102-
610	109.
611	
612	Tabarsa, M., Rezaei, M., Ramezanpour, Z., Waaland, J.R., 2012. Chemical compositions of
613	the marine algae Gracilaria salicornia (Rhodophyta) and Ulva lactuca (Chlorophyta) as a
614	potential food source. Journal of the Science of Food and Agriculture 92, 2500-2506.
615	
616	Tjallingii, R., Röhl, U., Kölling, M., Bickert, T., 2007. Influence of the water content on x-
617	ray fluorescence core-scanning measurements in soft marine sediments. Geochemistry,
618	Geophysics, Geosystems 8, Q02004, doi:10.1029/2006GC001393.
619	
620	Turner, A., Solman, K.R., 2016. Analysis of the elemental composition of marine litter
621	by field-portable-XRF. Talanta 159, 262-271.

- 623 Viana, I.G., Aboal, J.R., Fernàndez, J.A., Real, C., Villares, R., Carballeira, A., 2010. Use of
- 624 macroalgae stored in an Environmental Specimen Bank for application of some European
- 625 Framework Directives. Water Research 44, 1713–1724.
- 626
- 627 Zbikowski, R., Szefer, P., Latala, A., 2006. Distribution and relationships between selected
- 628 chemical elements in green alga *Enteromorpha* sp from the southern Baltic. Environmental
- 629 Pollution 143, 435-448.