

Adult longevity and its relationship with conservation status in European butterflies

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Abstract Many European butterfly species are currently experiencing serious declines, and may be threatened with extinction. Nevertheless, due to limited knowledge on the species biology and ecology, detailed assessments of endangerment level are not possible, and instead identifying species of conservation concern has to rely on proxies. Earlier studies suggested several characteristics, including host plant specificity, overwintering stage, patch size requirements or mobility, as potentially useful indicators of butterfly species vulnerability, but the usefulness of adult longevity in this respect has not been considered so far. Based on the information gathered through an extensive literature search we investigated the relationship between adult life span, flight period length or the temporal fragmentation index calculated as the ratio of the two parameters, and conservation status of European butterflies. We found that the species classified in one of the IUCN conservation concern categories (i.e. Endangered, Near Threatened, or Vulnerable) lived shorter as adults and were characterised by higher values of the temporal fragmentation index, while there was no particular pattern concerning flight period length. We believe that the apparent effects detected reflect the fact that shorter adult life span, and thus increased temporal fragmentation, in combination with protandry, i.e. earlier emergence of males, decrease individual chances of finding mating partners. Such a situation leads

to lower effective population size and reduced viability, especially in the case of small populations. All concerned, the investigated parameters reflecting adult longevity may serve as ‘early warning’ indicators, helping to flag-up butterfly species possibly at risk.

Keywords Extinction risk · Flight period · Life span · Species vulnerability · Temporal fragmentation · Threat level

Introduction

Butterfly populations in Europe have declined drastically in recent decades (Thomas et al. 2004; EEA 2011). These negative trends have prompted the launching of numerous programmes for butterfly conservation (Warren and Bourn 2011). However, for effective conservation, it is necessary to properly identify species threatened with extinction, and Red Lists are compiled for this purpose. Assessments of species positions on such lists should ideally be based on thorough knowledge of their biology and ecology and how these affect the species vulnerability to threats (Margules and Pressey 2000; Mattila et al. 2006). Despite the fact that butterflies comprise one of the most studied invertebrate groups, such knowledge is nevertheless available for only a very limited number of butterfly species (van Swaay 2002; Wenzel et al. 2006; Müller et al. 2010). Therefore, for their successful conservation, it is important to identify traits that predispose butterfly species to extinction risk (Mattila et al. 2006). Life history traits and/or ecological characteristics could be used as indicators of potential vulnerability to threats as many of these characteristics are common among species of conservation concern (Statzner et al. 2001; Mattila et al. 2006; Nylin and Bergström 2009).

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Studies conducted to date have identified a wide range of butterfly characteristics that can potentially act as proxies for extinction risk. One of the most commonly addressed aspects in this respect is the division into generalists and specialists (Nylin and Bergström 2009; Ali and Agrawal 2012; Bartoňová et al. 2014). The latter group includes a disproportionately high number of threatened species due to their stricter habitat and host plant requirements (Hodgson 1993; Purvis et al. 2000; Fontaine et al. 2007). Another well-established pattern relates to life history and voltinism, where univoltine species and/or those overwintering in the egg or larval stage are more susceptible to climate change, which has recently become one of the most serious drivers of butterfly declines (Hodgson 1993; Conrad et al. 2004; Mattila et al. 2006; Nylin and Bergström 2009). In addition, low mobility was typically reported for threatened butterflies (Kotiaho et al. 2005; Mattila et al. 2006; Nylin and Bergström 2009; Habel et al. 2015). This is not surprising, because less mobile species with low colonization success rates are more vulnerable to the effects of habitat fragmentation, which is nowadays a crucial threat for butterflies (Thomas 1995; Novacek and Cleland 2001; Baguette and Schtickzelle 2006; Franzen and Johanneson 2007). Also as a consequence of habitat fragmentation, butterflies with greater patch size requirements are highly represented among species of conservation concern (Cowley et al. 1999; Kotiaho et al. 2005; Baguette and Stevens 2013).

It has recently been pointed out that the extinction risk of butterfly populations is likely to depend not only upon classic spatial fragmentation of their habitats, but also upon their fragmentation over time (Nowicki et al. 2005b). The latter derives from the fact that the individual life span of the adult butterflies is usually much shorter than is the length of adult occurrence season dubbed as the flight period. Consequently, groups of individuals from different parts of a season do not have the chance to mate with one another. This problem is further exacerbated by protandry, i.e. earlier emergence of males in the season as compared with females, which is typical for butterflies (Wiklund and Fagerström 1977). The extent of and a temporal fragmentation (*sensu* Nowicki et al. 2005b) depends upon adult life span and flight period length and so it is highly variable among butterflies. Surprisingly, the effects of temporal fragmentation on butterfly species extinction risk have not been investigated so far.

In the present study we evaluated the relationships between adult life span, flight period length, and a temporal fragmentation index (defined as the ratio of flight period length to adult life span) on one hand, and the species conservation status, as reported in the European Red List (van Swaay et al. 2010), on the other hand. Butterflies that have short life spans and long flight season length will

have fewer opportunities for males and females representing different daily cohorts to meet together (Nowicki et al. 2005b). Consequently, we hypothesised that a higher level of threat should be associated with: (a) shorter adult life span; (b) longer flight period length; and (c) higher values of the temporal fragmentation index. We tested the above hypotheses using the data gathered through an extensive literature review.

Methods

Literature search

In order to gather information on adult life span and flight period length in butterflies, we searched for mark-recapture studies (which typically assess both these parameters) within the following databases: ISI Web of Science (<http://apps.webofknowledge.com/>), Scopus (<http://www.scopus.com/home.url>), and Google Scholar (<http://scholar.google.com/>). We used “(re)capture” and “butterfly/butterflies” or “Lepidoptera” as the searched keyword combinations. We restricted our search results to only European species, because relatively little information was available for all others, with the slight exception of a small number of North American species. We also utilised relevant grey literature on the subject, in particular academic theses or reports known to us.

We considered only adult life span data originating from field studies, and thus we excluded literature on adult life span measured in controlled conditions, i.e. physiological longevity. This is because when butterflies are raised in enclosures or laboratories, their recorded longevity is known to overestimate the life span actually achieved in nature (Karlsson and Wiklund 2005; Nowicki et al. 2005b). We likewise excluded mark-recapture estimates of butterfly residence time assessed in clearly open populations, which are subject to substantial emigration and thus considerably underestimate butterfly longevity (Nowicki et al. 2005a). An exception was made for the cases in which the emigration rate could be estimated and accounted for, e.g. using the Virtual Migration model (Hanski et al. 2000). Consequently, we believe that the life span estimates we used in our study represent true adult longevities and not just adult residence times within the surveyed populations.

Quite often the literature sources did not explicitly report adult life span but instead they gave adult survival rate estimates. In such cases we converted survival rate (ϕ) into adult life span (e) using the following formula: $e = (1 - \phi)^{-1} - 0.5$ (Nowicki et al. 2005b). Whenever the studies reported adult life span or survival estimates for males and females separately, we used the mean of the two values because they typically differed rather little, i.e. by

less than 10%. It should be noted that adult life span values provided by mark-recapture studies are restricted to a single season, and thus they do not account for the fact that certain individuals flying during the season might also be on the wing in the previous or in the following year in species overwintering as adults. Nevertheless, due to the high mortality of overwintering adults such individuals are relatively few. Furthermore, their existence does not undermine the usefulness of single-season data for our study, in which adult life span is understood as the time during which an adult flies and breeds within a single generation, and not as its total life expectancy, including the period of inactive overwintering.

Regarding flight period length, we calculated it as the total number of days of adult occurrence inclusive of the first and last day. Thus, for example, a flight period from 1 to 30 July corresponds to a length of 30 days (and not 29 days). We excluded the studies that did not cover the entire flight period, which was either explicitly mentioned by the authors or was clear from the reported results, with relatively high daily numbers of individuals recorded at the beginning or the end of the study period. In a few cases, we also used information on flight period length from studies other than mark-recapture surveys, e.g. behavioural observations conducted from the very start to the very end of the season. On the other hand, we decided not to use the information on flight period provided by several general books on butterflies (e.g. Settele et al. 1999; Beneš et al. 2002). Even though such information was easily available for most European species, it turned out to be too superficial to be applicable in our analyses. Specifically, the information was given in the form of general statements, mentioning that e.g. species X flies from early July to mid-August, thus allowing only very coarse assessments of flight period length with margins of error of as much as 10–15 days.

Data handling and analysis

The extent of temporal fragmentation was calculated as the ratio between flight period length and average adult life span, both measured in days. In the case of studies spanning several years (e.g. Schtickzelle et al. 2002; Nowicki et al. 2009), we treated the data from each year separately. In making the calculations, we endeavoured to use data on life span and flight period length from the same population and year. This usually was possible because most mark-recapture studies reported both. Otherwise, we paired together data on adult life span and flight period from the closest-lying locations, but in rare cases these locations were quite distant from one another and represented different biogeographic regions. For multivoltine species, we relied on information provided by studies conducted on the first generation whenever possible. This is because the

population size of the first generation is generally much smaller than those of later generations and it is thus critical for species persistence, and furthermore the occurrence of later generations is sometimes facultative (Franzen and Johannesson 2007; Fric et al. 2010; Nabielec and Nowicki 2015).

For species with more than one record available, we calculated median values of adult life span, flight period length, and temporal fragmentation index and used them in the subsequent analyses, since the distributions of records were often non-normal (right-skewed). In turn, it is worth noting that the distribution of median values among species was normal for all of the three parameters considered.

For the purpose of the analyses, we adopted the Red List status of European butterflies in accordance with van Swaay et al. (2010). Despite our original expectations, the sample sizes of species for which we were able to gather data were small in the three conservation concern categories, i.e. Endangered (EN), Vulnerable (VU), and Near Threatened (NT). We therefore decided to pool these into a single ‘conservation concern’ category (CC) and analyse it against the species categorized as Least Concern (LC).

We analysed the relationship between adult life span, flight period length or temporal fragmentation index, and conservation status (CC vs. LC, dichotomous dependent variable) using logistic regression analysis. For each predictor, we conducted two separate analyses, using the full data set gathered as well as what we term a ‘core data set’. The core data set excluded species for which data quality was problematic for various reasons. These included cases of (a) species for which data on adult life span and flight period length came from distant populations, representing different biogeographic regions; (b) species for which only the data for the second generation was available; and (c) *Maculinea alcon*, the conservation status of which is questionable, apparently due to its uncertain systematic status, with two distinct forms existing, namely *M. alcon* ‘alcon’ and *M. alcon* ‘rebeli’ (Als et al. 2004; Steiner et al. 2005; Pecsénye et al. 2007; Sielezniew et al. 2012). It is classified as LC by van Swaay et al. (2010), however many authors regard both forms to be under threat in Europe (Wallis-DeVries 2004; Tartally et al. 2008; Czekes et al. 2014).

Obviously, a common problem with cross-species analyses is that records for related species may not be fully independent from one another. The standard solution in such cases is controlling for phylogenetic autocorrelations (Martins and Hansen 1996). However, this was not possible in our study, because a full phylogenetic tree with scalable inter-specific distances is not yet available for European butterflies (cf. Cowley et al. 2001; Bartoňova et al. 2014). Therefore, in order to verify the risk of phylogenetic autocorrelation biases, we instead applied the intraclass correlation coefficients (Stanish and Taylor 1983; Lessells and

Boag 1987) to test for potential repeatability of our records within species as well as higher taxa, namely tribes, subfamilies, and families (but not at the genus level, because we rarely had data for more than one species per genus). The testing yielded significant results for species (life span: $r_I=0.7379$, $P<0.0001$; flight period length: $r_I=0.5513$, $P<0.0001$), but not for any higher taxa (life span: $r_I=0.0539$, $P=0.6691$ for tribes; $r_I=0.1166$, $P=0.4962$ for subfamilies; $r_I=0.0408$, $P=0.6386$ for families; flight period length: $r_I=0.1922$, $P=0.2835$ for tribes; $r_I=0.1401$, $P=0.4468$ for subfamilies; $r_I=0.1312$, $P=0.4213$ for families). Such an outcome indicates that our data records were highly repeatable within species, but fairly independent among them. This, in combination with the fact that despite a relatively small sample size we managed to gather data for a wide range of European butterfly species, makes us believe that the results of our analyses are not biased by potential phylogenetic autocorrelations.

Results

We successfully gathered relevant information for 50 species of European butterflies, including 4 classified as EN, 5 as VU, and 6 as NT, as well as 35 species classified as LC (Table 1). The average adult life span of these species ranged from ca. 2.5 to 15 days. The flight period length was between 20 and 50 days in most cases, although some clear outliers could also be noticed. The shortest adult occurrence season was reported for *Pseudophilotes bavius* (median value of 16.5 days), whereas in satyrid butterflies it sometimes approached or exceeded 70 days (in *Maniola jurtina* and *Coenonympha pamphilus* respectively). There was no apparent correlation between flight period length and adult life span (Pearson's correlation: $r=0.1794$, $P=0.2125$), and consequently the ratio of the two parameters, which we defined as the temporal fragmentation index, varied greatly from ca. 2 to more than 12 (Table 1).

As indicated by Red List categories, extinction risk generally increased with increasing adult life span (Fig. 1a). Similarly, butterflies in the three categories collectively characterized as CC had higher temporal fragmentation index values (Fig. 1c). On the other hand, there was no clear pattern concerning flight period length, which turned out to be slightly elevated among NT species. The latter result was partly due to strong variation within this particular group (Fig. 1b).

The logistic regression analyses confirmed the above patterns, revealing significant relationships with species conservation status (CC vs. LC) in the case of adult life span and temporal fragmentation index, but no effect whatsoever for flight period length (Table 2). It is noteworthy that the effect of adult longevity and temporal

fragmentation increased (despite considerably smaller sample sizes) when the analyses were restricted to the core data set, thus excluding species for which the quality of the data gathered was problematic. The threshold value for adult life span at which a species had 50% probability of being listed in one of the CC categories was 3.53 days for the full data set and 4.31 days for the core data set. In the case of temporal fragmentation index, the respective thresholds were 9.19 and 8.01.

Discussion

The selection of species on which we based our investigation may not be fully representative for the whole spectrum of European butterflies. In particular, an underrepresented group are the species from Mediterranean region as well as from northern Europe, the conservation status of which is potentially less related to adult longevity and more to the life history parameters beyond the scope of the present study, such as wintering stage and voltinism (Mattila et al. 2006; Nylin and Bergström 2009). Another limitation of our database is the fact that it lacked very common species, mostly of Nymphalidae and Pieridae families, which are neglected in mark-recapture studies, apparently because of the lack of scientific and conservation interest in them. Nevertheless, based on anecdotal information such species are reported to have long-living adults (Settele et al. 1999; Beneš et al. 2002), and thus we believe that their inclusion would have actually strengthened the outcome of our analyses.

Our results indicated the existence of clear relationships between adult life span and temporal fragmentation index on one hand and conservation status of European butterflies on the other. Moreover, the relationships proved to be significant, regardless of whether the full data set or the core data set was used, which increases our confidence in these findings. In contrast, butterfly conservation status was not linked in any way to flight period length as defined in our study.

Butterfly adult life span is a part of an adaptive life history which involves mating and egg laying strategy (Carey 2001; Beck and Fiedler 2009). The average life span reported in our study typically reached only a few days. Such a short life span implies that the mating must take place shortly after eclosion to minimise delay and allow most of the females to oviposit their eggs before they die (Scott 1973; Beck and Fiedler 2009). Hence, in species with short life spans adult butterflies have a very narrow time window to copulate. Adult life span also affects realised fecundity, i.e. the number of eggs laid, which in turn can have a critical impact on population viability (Fischer et al. 2006; Pijpe 2007; Haeler et al. 2014). Low quantities

Table 1 Summary information on adult life span, flight period length, and the temporal fragmentation index (i.e. ratio of flight period length to adult life span) gathered for European butterflies

Species	Status	Life span (days)	Flight period (days)	Temporal fragmentation	Sources
<i>Colias myrmidone</i> ^a	EN	3.54 (3.32–3.75)	26 (21–31)	7.47 (5.60–9.34)	Szentirmai et al. (2014)
<i>Lycaena helle</i>	EN	7.19 (5.45–7.83)	39 (29–64)	5.87 (3.87–8.90)	Fischer et al. (1999), Bauerfeind et al. (2009), Reymond (2014), Turlure et al. (2014), Nabielec and Nowicki (2015)
<i>Phengaris (=Maculinea) arion</i>	EN	3.53 (3.07–4.26)	37 (32–39)	10.28 (9.15–10.92)	Bonelli et al. (2013)
<i>Coenonympha oedippus</i>	EN	3.00 (2.50–4.20)	27.5 (18–28)	7.98 (6.67–10.00)	Örvössy et al. (2013)
<i>Phengaris (=Maculinea) teleius</i>	VU	3.01 (1.61–4.16)	38 (28–56)	12.01 (9.12–40.00)	Nowicki et al. (2005a, b, 2009, 2014), Vodá et al. (2010)
<i>Euphydryas maturna</i>	VU	6.53 (4.75–8.30)	42.5 (36–49)	7.33 (4.34–10.32)	Wahlberg et al. (2002), Konvička et al. (2005)
<i>Lopinga achine</i>	VU	6.30 (5.95–6.64)	28 (25–31)	4.43 (4.20–4.67)	Bergman and Landin (2002), Streitberger et al. (2012)
<i>Coenonympha tullia</i>	VU	3.05 (2.80–3.30)	24.5 (21–28)	7.99 (7.50–8.48)	Turner (1963), Warren (1992), Komonen et al. (2004)
<i>Erebia sudetica</i>	VU	4.00	29	7.25	Nowicki et al. (2005a)
<i>Thymelicus acteon</i> ^b	NT	7.00	41	5.86	Thomas (1983), Buszko and Mastowski (2008)
<i>Parnassius mnemosyne</i>	NT	9.06 (5.05–11.15)	51 (31–51)	5.66 (4.57–10.10)	Schmidt (1989), Seufert (1990), Konvička and Kuras (1999)
<i>Parnassius apollo</i>	NT	3.73 (3.20–4.26)	29 (25–33)	8.08 (5.87–10.30)	Brommer and Fred (1999), Komonen et al. (2004), Fred et al. (2006)
<i>Iolana iolas</i>	NT	5.88 (3.63–8.12)	49.5 (49–50)	9.83 (6.16–13.50)	Rabasa et al. (2005, 2007), Heer et al. (2013)
<i>Phengaris (=Maculinea) nausithous</i>	NT	2.84 (2.02–5.74)	40 (23–56)	12.64 (4.36–23.76)	Pfeifer et al. (2000, 2007), Nowicki et al. (2005a, b, 2014), Vodá et al. (2010)
<i>Euphydryas desfontainii</i>	NT	5.55	36	6.49	Pennekamp et al. (2014)
<i>Pyrgus sidae</i>	LC	9.20	25	2.72	Hernández-Roldán et al. (2009)
<i>Hesperia comma</i>	LC	4.40 (3.00–10.30)	35 (23–50)	7.95 (2.23–15.63)	Thomas (1983), Komonen et al. (2004), Soulsby and Thomas (2012)
<i>Zerynthia polyxena</i>	LC	5.28 (4.40–6.17)	29 (20–38)	5.94 (3.24–8.64)	Örvössy et al. (2005), Batáry et al. (2008), Celik (2012)
<i>Leptidea sinapis</i> ^a	LC	8.35 (6.50–10.20)	38.5 (33–44)	5.00 (3.24–6.77)	Warren et al. (1986), Komonen et al. (2004), Friberg et al. (2008)
<i>Leptidea reali</i> ^a	LC	7.60	44	5.79	Friberg et al. (2008)
<i>Anthocharis cardamines</i>	LC	6.95 (5.60–8.30)	28.5 (21–34)	4.30 (2.53–6.07)	Courtney and Duggan (1983), Dempster (1997)
<i>Lycaena virgaureae</i>	LC	6.43 (6.20–6.65)	31 (29–33)	4.84 (4.36–5.32)	Fjellstad (1998), Komonen et al. (2004), Haaland (2015)
<i>Lycaena hippothoe</i>	LC	9.60 (7.00–10.00)	28 (28–32)	2.92 (2.80–4.57)	Fischer (1998), Fischer and Fiedler (2001), Komonen et al. (2004)
<i>Satyrrium w-album</i> ^b	LC	6.90	28	4.06	Warren (1992), Komonen et al. (2004)
<i>Cupido minimus</i> ^b	LC	15.00	31	2.07	Morton (1985), Komonen et al. (2004)
<i>Pseudophilotes bavius</i>	LC	2.80 (2.40–5.40)	16.5 (12–28)	5.47 (2.50–11.67)	Crişan et al. (2014)
<i>Phengaris (=Maculinea)alcon</i> ^c	LC	2.44 (1.62–5.98)	29 (18–60)	11.88 (5.59–29.27)	Seufert (1993), Nowicki et al. (2005a, 2009), Timuş et al. (2013)
<i>Plebejus argus</i>	LC	3.35 (3.20–3.50)	30 (20–40)	9.11 (5.71–12.50)	Warren (1992), Lewis et al. (1997), Cormont et al. (2011)

Table 1 (continued)

Species	Status	Life span (days)	Flight period (days)	Temporal fragmentation	Sources
<i>Aricia eumedon</i> ^b	LC	3.59	19	5.30	Seufert (1993), Komonen et al. (2004)
<i>Polyommatus icarus</i> ^a	LC	4.40 (3.40–5.40)	26 (18–34)	5.80 (5.29–6.30)	Dowdeswell et al. (1940), Scott (1973), Komonen et al. (2004)
<i>Polyommatus bellargus</i>	LC	9.10 (8.10–10.10)	26.5 (24–29)	2.98 (2.38–3.58)	Davis et al. (1958)
<i>Polyommatus coridon</i>	LC	5.52 (3.70–8.70)	56.5 (24–63)	10.79 (4.21–14.86)	Davis et al. (1958), Nowicki et al. (2005a), Schmitt et al. (2006)
<i>Argynnis paphia</i> ^b	LC	11.50	38	3.30	Magnus (1954), Komonen et al. (2004)
<i>Argynnis aglaja</i>	LC	8.20	58	7.07	Zimmermann et al. (2009)
<i>Brenthis ino</i>	LC	9.74 (5.70–13.79)	40.5 (35–46)	5.30 (2.54–8.07)	Zimmermann et al. (2005), Fric et al. (2010)
<i>Boloria eunomia</i>	LC	8.23 (3.00–11.55)	35.5 (26–45)	4.32 (2.68–11.00)	Schtickzelle et al. (2002), Turlure et al. (2010)
<i>Boloria euphrosyne</i>	LC	9.00 (5.97–11.10)	29 (24–32)	3.39 (2.61–4.02)	Baguette and Neve (1994), Komonen et al. (2004), Al Dhaheer (2009)
<i>Boloria aquilonaris</i>	LC	4.26	21	4.93	Turlure et al. (2010)
<i>Euphydryas aurinia</i>	LC	6.40 (2.24–15.37)	31 (15–42)	3.16 (2.60–12.95)	Munguira et al. (1997), Wahlberg et al. (2002), Anthes et al. (2003), Komonen et al. (2004), Schtickzelle et al. (2005), Fric et al. (2010), Zimmermann et al. (2011), Casacci et al. (2015)
<i>Melitaea cinxia</i>	LC	5.80	33	5.69	Wahlberg et al. (2002)
<i>Melitaea Didyma</i>	LC	7.00 (5.50–8.00)	46 (30–51)	5.75 (5.45–7.29)	Vogel and Johannesen (1996)
<i>Melitaea diamina</i>	LC	8.59 (6.35–10.61)	29 (29–51)	4.57 (3.38–4.81)	Hanski et al. (2000), Wahlberg et al. (2002), Fric et al. (2010)
<i>Melitaea athalia</i>	LC	10.00 (5.45–11.26)	35 (30–55)	4.88 (3.50–5.50)	Warren (1987), Wahlberg et al. (2002), Fric et al. (2010), Cormont et al. (2011)
<i>Pararge aegeria</i> ^a	LC	9.50	23	2.42	Warren (1992), Komonen et al. (2004)
<i>Lasiommata megera</i> ^a	LC	4.20	40	9.52	Parr et al. (1968), Harker and Shreeve (2008)
<i>Coenonympha pamphilus</i> ^a	LC	7.30	76	10.41	Wickman (1985)
<i>Aphantopus hyperantus</i>	LC	3.95 (3.60–4.29)	36.5 (35–38)	9.29 (8.86–9.72)	Sutcliffe et al. (1997), Soulsby and Thomas (2012)
<i>Maniola jurtina</i>	LC	6.55 (6.51–10.83)	67 (51–78)	7.84 (6.19–11.91)	Tudor and Parkin (1979), Brakefield (1982), Lörtscher et al. (1997), Cormont et al. (2011)
<i>Erebia aethiops</i>	LC	6.50	33	5.08	Slámová et al. (2013)
<i>Minois dryas</i>	LC	4.47 (3.50–6.74)	25.5 (19–32)	5.17 (4.75–6.75)	Pellet and Gander (2009), Bilnicki (2015)

Whenever more than one record was available for a species, we present the median with the range (min–max) in parentheses. Conservation status follows the European Red List of Butterflies (van Swaay et al. 2010): *EN* endangered, *VU* vulnerable, *NT* near threatened, *LC* least concern. Superscripts indicate species which were excluded from the core data set used in the analyses for various reasons

^aData available only for the second or third generation

^bData on adult life span and flight period length came from different regions

^cQuestionable conservation status—see [Methods](#) for the rationale

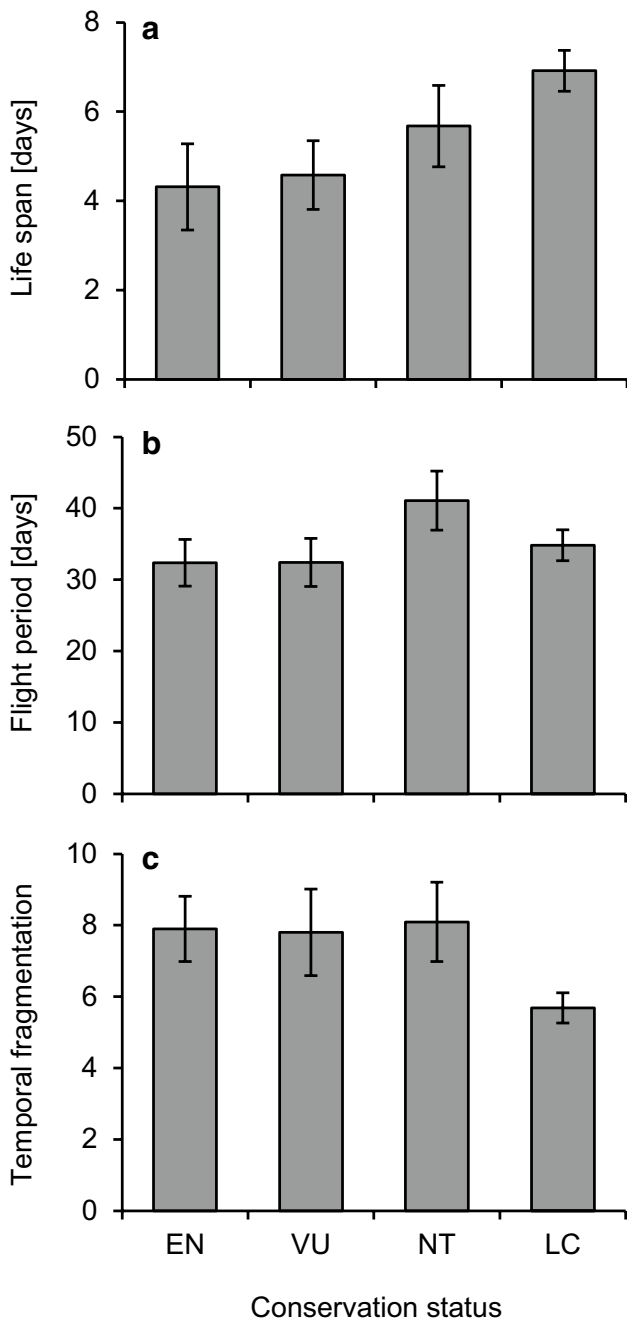


Fig. 1 Adult life span (a), flight period length (b), and temporal fragmentation index (c) in relation to the conservation status of European butterflies (EN endangered, VU vulnerable, NT near threatened, LC least concern). The values shown represent means (with their SEs) across all the investigated species in each category. Different conservation concern categories (EN, VU, NT) are treated separately only for the graphic presentation, but they were pooled together in the analyses (see Table 2)

of oviposited eggs over the long term are bound to result in population decline. It is thus not surprising that the results of our study confirmed that short life span corresponds to higher threat level in European butterfly species.

Alternative hypotheses explaining the relationship between adult life span or temporal fragmentation index and conservation status involve mobility and predation. Since both emigration probability and movement distance are typically time-dependent (Hanski et al. 2000), short-lived butterflies may be expected to emigrate in lower numbers and move shorter distance throughout their adult lifetime. Species with such characteristics are more likely to experience the negative effects of habitat fragmentation, and hence be more prone to being threatened (Kotiaho et al. 2005; Franzen and Johannesson 2007; Habel et al. 2015), especially in the highly fragmented landscapes of Europe. However, the plausibility of the above explanation is undermined by empirical studies indicating that dispersal capabilities and butterfly life span may be negatively correlated, since investing in mobility and longevity is subject to a developmental trade-off (Hanski et al. 2006; Niitepõld and Hanski 2013).

Additionally, the better conservation status of species with long-living adults might possibly be attributed to their lower mortality due to predation. It has been reported that butterflies with anti-predator defence features (such as aposematism, eye-spots, etc.) have longer life spans, but it must be stressed that the effect of these anti-predator defences on longevity was rather weak and mostly restricted to tropical butterflies (Beck and Fiedler 2009). Furthermore, although predation on adult butterflies may sometimes be considerable, it is nonetheless of relatively little importance for population dynamics as compared with predation experienced at the egg, larval or pupal stages (Dempster 1984; Warren 1992).

An increased level of temporal fragmentation was also found to correspond to higher species vulnerability in our study. A possible explanation is that if butterflies live for only a limited part of the flight period then they may have lower chances of finding mates. These low mating opportunities are further decreased by the fact that males emerge earlier in the season than females, usually by several days, due to protandry (Pfeifer et al. 2000; Petit et al. 2001; Nowicki et al. 2005b). This phenomenon is common in insects, particularly in species for which the flight season is very long and the majority of the population occurs at the beginning of the flight season, such as is the case of the mayfly (Gibbs and Siebenmann 1996; Takemon 2000). Protandry prevents inbreeding and ensures that only strong males survive long enough to mate. Moreover, it also supports immediate female fertilisation, which minimises the risk of females dying before mating (Fagerström and Wiklund 1982; Zonneveld and Metz 1991; Zonneveld 1992; Morbey and Ydenberg 2001).

Although all the aforementioned effects are considered positive, some negative consequences of protandry are also known. Since a fraction of males may not survive

Table 2 Results of multiple logistic regression analyses of factors affecting conservation status (conservation concern vs. least concern) of European butterflies

Predictor	Data set (with sample size)	Parameter value (\pm SE)		Model fit		
		Intercept	Estimate	χ^2	<i>P</i>	<i>R</i> ²
Life span	Full (<i>n</i> = 50)	1.29 \pm 0.91	-0.37 \pm 0.16	6.83	0.0089	0.27
	Core (<i>n</i> = 37)	1.82 \pm 1.10	-0.42 \pm 0.19	6.05	0.0138	0.33
Flight period length	Full (<i>n</i> = 50)	-1.12 \pm 0.98	0.01 \pm 0.03	0.09	0.7685	0.01
	Core (<i>n</i> = 37)	-1.18 \pm 1.20	0.02 \pm 0.03	0.25	0.6192	0.01
Temporal fragmentation	Full (<i>n</i> = 50)	-3.18 \pm 0.99	0.35 \pm 0.13	7.85	0.0051	0.29
	Core (<i>n</i> = 37)	-3.85 \pm 1.31	0.48 \pm 0.18	9.44	0.0021	0.41

long enough to mate with the later-emerging females, some females will consequently go unmated, and will thus be unable to lay eggs. This lost reproductive potential can significantly decrease population size and could also lead to population extinctions. The protandry effect might therefore be evolutionarily advantageous at higher population densities but, if the population density decreases, it can be harmful (Calabrese and Fagan 2004). Additionally, weather conditions could affect the proper timing of male and female emergence by up to several days (Schtickzelle et al. 2002; Robinet and Roques 2010). For instance, rainy days at the beginning of the flight period in 1983 were found to influence individual development in *Euphydryas editha bayensis*, resulting in the first females emerging 14 days after males, which had a clear negative effect on the population dynamics of the species (Dobkin et al. 1987; Baughman 1991). Furthermore, a high level of temporal fragmentation, namely short individual life span in relation to long flight period, in combination with protandry seriously reduced the effective population size (*sensu* Lande and Barrowclough 1987; Hill 1972), thus accelerating the loss of genetic variability in small populations.

In contrast to several earlier studies, which suggested that extinction risk in butterflies decreases with lengthening flight period (Komonen et al. 2004; Kotiaho et al. 2005; Franzen and Johannesson 2007), our analyses did not reveal any link between flight period length alone and species conservation status. The most straightforward explanation for such a result could be that flight period length per se is unimportant for species extinction risk, and it only matters in combination with adult longevity. Nevertheless, we hypothesise that the situation is more complex, namely a longer time of adult occurrence has both positive and negative consequences for species viability. On the one hand, as discussed above, it increases the temporal fragmentation of butterfly populations and may reduce mating opportunities for both sexes. On the other hand, a longer flight period allows for the compensation of the negative effects of stochastic changes during the flight period such as unfavourable weather conditions, and inappropriate management interventions, e.g. inappropriate timing of meadow mowing (Cormont et al. 2011). An extended adult occurrence

season improves the resilience of butterfly populations to catastrophic events such as floods or fires (cf. Konvička et al. 2002; Kajzer-Bonk et al. 2013; Nowicki et al. 2015), because only a small fraction of individuals is affected if a short-term disturbance happens during the flight period.

Our findings demonstrate that both adult life span and temporal fragmentation index may serve as useful 'early warning' indicators, helping to flag-up butterfly species possibly at risk from among those for which detailed information essential for evaluating threat level is lacking. Regrettably, as our literature search implies, the estimates of adult life span and flight period length (needed for calculating temporal fragmentation index) are not readily available for most butterflies either. However, they are relatively easy to get through mark-recapture studies. These are simple to plan and conduct, and they may be carried out with the help of amateur naturalists. This gives the longevity-related parameters analysed in the present study a substantial advantage over the proxies for species vulnerability previously suggested by other authors (Kotiaho et al. 2005; Mattila et al. 2006; Nylin and Bergström 2009), such as host plant specificity, overwintering stage, patch size requirements or mobility (see Introduction for their rationale), because assessing the latter characteristics typically requires specialist expertise.

Obviously, the main drawback of mark-recapture studies is their labour-intensity. Therefore, it would be highly desirable if the information on adult longevity and flight period length could be extracted from the well-established butterfly monitoring schemes based on transect counts. Evaluating flight period length with transect counts requires increased frequency of the transect surveys, because biweekly counts, as currently adopted in most monitoring schemes (van Swaay et al. 2008), are not enough for this purpose. In turn, assessing adult longevity with transect counts appears more difficult to accomplish. Zonneveld (1991) developed a theoretical model for the estimation of life span from transect count data, but its applicability has so far been hampered by rigorous assumptions, which are difficult to meet in real world situations (Nowicki et al. 2008). Nevertheless, more recent developments based on this model, such as the Insect Count Analyzer (INCA)

software (Longcore et al. 2003), are promising and give some hope that transect counts can be reliably used to derive butterfly life span estimates in the near future.

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