

1 **Seasonality of cladoceran and bryozoan resting stage $\delta^{13}\text{C}$ values**
2 **and implications for their use as palaeolimnological indicators of**
3 **lacustrine carbon cycle dynamics**

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19 **Abstract**

20

21 The stable carbon isotope composition, expressed as $\delta^{13}\text{C}$ values, of chitinous resting
22 stages of planktivorous invertebrates can provide information on past changes in carbon
23 cycling in lakes. For example, the $\delta^{13}\text{C}$ values of cladoceran ephippia and bryozoan
24 statoblasts have been used to estimate the past contribution of methane-derived carbon to
25 lake food webs and variations in the $\delta^{13}\text{C}$ value of planktonic algae. Limited information,
26 however, is available concerning seasonal variations in $\delta^{13}\text{C}$ values of these organisms and
27 their resting stages. We measured the seasonal variation in $\delta^{13}\text{C}$ values of *Daphnia*
28 (Branchiopoda: Cladocera: Daphniidae) and their floating ephippia over a 2-year period in
29 small, dimictic Lake Gerzensee, Switzerland. Floating ephippia of *Ceriodaphnia*
30 (Branchiopoda: Cladocera: Daphniidae) and statoblasts of *Plumatella* (Phylactolaemata:
31 Plumatellida: Plumatellidae) were analysed during parts of this period. Furthermore, $\delta^{13}\text{C}$
32 values of remains from all three organism groups were analysed in a 62-cm-long sediment
33 core. Throughout the year, *Daphnia* $\delta^{13}\text{C}$ values tracked the $\delta^{13}\text{C}$ values of particulate
34 organic matter (POM), but were more negative than POM, indicating that *Daphnia* also
35 utilize a relatively ^{13}C -depleted carbon source. *Daphnia* ephippia $\delta^{13}\text{C}$ values did not show
36 any pronounced seasonal variation, suggesting that they are produced batch-wise in autumn
37 and/or spring and float for several months. In contrast, $\delta^{13}\text{C}$ values of *Ceriodaphnia* ephippia
38 and *Plumatella* statoblasts followed variations in $\delta^{13}\text{C}_{\text{POM}}$ values, *Ceriodaphnia* values being
39 the most negative of the resting stages. Average cladoceran ephippia $\delta^{13}\text{C}$ values in the
40 flotsam agreed well with ephippia values from Gerzensee surface sediments. In contrast,
41 average *Plumatella* statoblast $\delta^{13}\text{C}$ values from the flotsam were 4 ‰ more negative than in
42 the surface sediments. In the sediment core, $\delta^{13}\text{C}$ values of the two cladocerans remained
43 low (mean -39.0 ‰ and -41.9 ‰) throughout the record. In contrast, *Plumatella* had distinctly
44 less negative $\delta^{13}\text{C}$ values (mean -32.0 ‰). Our results indicate that in Gerzensee, *Daphnia*
45 and *Ceriodaphnia* strongly relied on a ^{13}C -depleted food source throughout the past 150

46 years, most likely methane-oxidising bacteria, whereas this food source was not a major
47 contribution to the diet of bryozoans.

48 **Introduction**

49

50 Chitinous remains of aquatic invertebrates are readily preserved in lake sediments and their
51 stable carbon isotope compositions, expressed as $\delta^{13}\text{C}$ values, have been used as a proxy
52 to study past changes in the availability and importance of different carbon pathways in
53 lakes (Frossard et al. 2014; Rinta et al. 2016; van Hardenbroek et al. 2010; Wooller et al.
54 2012). Two such pathways are the fixation of dissolved CO_2 by algae and the recycling of
55 carbon from methane (CH_4) by methane-oxidising bacteria (MOB). Up to now,
56 palaeolimnological studies that analysed the $\delta^{13}\text{C}$ values of chitinous invertebrate remains
57 have focused mainly on benthic invertebrates, most prominently chironomid larvae (Insecta:
58 Diptera: Chironomidae), whose chitinous head capsules are regularly found in sediments
59 (Belle et al. 2014; Heiri et al. 2012; van Hardenbroek et al. 2010). Additionally, exoskeleton
60 fragments of planktonic invertebrates such as *Bosmina* (Branchiopoda: Cladocera:
61 Bosminidae) have been analysed for their $\delta^{13}\text{C}$ values (Perga 2009, 2011). It has recently
62 been suggested that similar information may be gained from $\delta^{13}\text{C}$ analysis of resting stages
63 of planktivorous invertebrates (Schilder et al. 2015a, 2015b; van Hardenbroek et al. 2013,
64 2014; Wooller et al. 2012). Examples of these resting stages are ephippia, produced by
65 *Daphnia* (Branchiopoda: Cladocera: Daphniidae) and other planktonic cladocerans, and
66 statoblasts, produced by bryozoans that form sessile colonies on hard substrates in the
67 shallow parts of lakes (Wood and Okamura 2005). Ephippia and statoblasts are abundant in
68 lake sediments (Francis 2001; Korhola and Rautio 2001), providing a potential archive for
69 palaeoecological reconstructions.

70 In modern lake food web studies, the analysis of the $\delta^{13}\text{C}$ value of different food
71 components is a widely used technique to differentiate between carbon sources (Bunn and
72 Boon 1993). The main organic carbon sources for filter-feeding zooplankton are algae,
73 detritus, and heterotrophic bacteria (Edmondson 1957; Lampert 2011). Using characteristic
74 isotopic signatures of these different sources, carbon flow can be traced through a lake's

75 food web (Fry 2006; Taipale et al. 2007). Fundamental for such studies is that the carbon
76 isotopic composition of an organism closely reflects the isotopic signature of its diet (DeNiro
77 and Epstein 1978; Peterson and Fry 1987). For *Daphnia* and their ehippia, Perga (2011)
78 and Schilder et al. (2015b) showed that ehippia $\delta^{13}\text{C}$ values closely reflect the $\delta^{13}\text{C}$ values
79 of the living *Daphnia* at the time of ehippia production. Van Hardenbroek et al. (2016)
80 recently demonstrated that the $\delta^{13}\text{C}$ values of bryozoan statoblasts are related to the $\delta^{13}\text{C}$
81 values of the colonies that produced them. More work is required, however, to show that
82 $\delta^{13}\text{C}$ values of *Daphnia* ehippia and bryozoan statoblasts reflect the diet of the parent
83 organisms during or shortly before resting stage formation.

84 Understanding the extent to which lakes change their capacity to sequester or
85 release carbon in response to global warming and eutrophication is a key issue for
86 palaeoecologists today (Seddon et al. 2014). New insights into lake carbon cycling can be
87 gained from analysing $\delta^{13}\text{C}$ values of aquatic invertebrate remains. Recently, it was
88 suggested that carbon from CH_4 may be an important additional transfer pathway of carbon
89 in lake food webs, indicated by remarkably low $\delta^{13}\text{C}$ values of aquatic invertebrates when
90 they incorporate CH_4 -derived, ^{13}C -depleted carbon (Bastviken et al. 2003; Bunn and Boon
91 1993; Grey et al. 2016). The $\delta^{13}\text{C}$ values of biogenic CH_4 in small European lakes range
92 between -86 and -61 ‰ (Rinta et al. 2015), which is markedly depleted in ^{13}C relative to
93 algae with $\delta^{13}\text{C}$ values that typically fall in the range from -35 to -25 ‰ (Jones et al. 1999;
94 Peterson and Fry 1987). The large difference in $\delta^{13}\text{C}$ values thus allows for differentiation
95 between photosynthetically produced organic matter and MOB as potential carbon sources
96 for primary consumers in lakes. Significant relationships have been observed between $\delta^{13}\text{C}$
97 of *Daphnia* ehippia in surface sediments and diffusive CH_4 flux (Van Hardenbroek et al.
98 2013) and within-lake CH_4 concentrations (Schilder et al. 2015a). Cladoceran remains may
99 thus record changes in past diffusive CH_4 flux in lakes, though they do not provide direct
100 information on other forms of CH_4 fluxes such as ebullition or plant-mediated transport.

101 Palaeoecological studies using $\delta^{13}\text{C}$ values of chitinous remains have largely
102 overlooked the use of bryozoan statoblasts, which have been suggested as a resource for

103 reconstructing $\delta^{13}\text{C}$ values of primary production (Turney 1999; van Hardenbroek et al. 2014,
104 2016). As filter feeders, Bryozoa rely on algae, particulate organic matter (POM) and
105 associated microorganisms (e.g. bacteria) as their main food source (Kaminsky 1984).
106 Bryozoa are mainly bound to their substrate and do not have access to food sources from
107 deeper water layers and are therefore not, or only to a small degree, influenced by CH_4 -
108 derived carbon (van Hardenbroek et al. 2016). By combining $\delta^{13}\text{C}$ measurements of mobile
109 filter feeders like *Daphnia* with sessile filter feeders like *Plumatella* (Phylactolaemata:
110 Plumatellida: Plumatellidae), it is possible to separate changes in the importance of CH_4 -
111 derived carbon from changes in algal $\delta^{13}\text{C}$ values, or more generally, changes in lake
112 productivity (van Hardenbroek et al. 2014; Rinta et al. 2016). At present, however, no
113 information is available about seasonal changes in $\delta^{13}\text{C}$ values of cladoceran and bryozoan
114 resting stages or about the extent to which this seasonality influences the $\delta^{13}\text{C}$ values of
115 statoblast and ephippia remains in lake sediments. Furthermore, it is unknown whether the
116 $\delta^{13}\text{C}$ values of deposited organism remains reflect integrated $\delta^{13}\text{C}$ values for statoblasts and
117 ephippia that float on the lake surface, or whether the sedimentary assemblages are
118 characterized by systematically higher or lower values. Such differences could appear as a
119 consequence of degradation, transport processes, or production of resting stages in cryptic
120 habitats that do not contribute to floating resting stages at the lake surface (e.g. from deeper
121 littoral, lower epilimnetic or profundal habitats). These uncertainties currently hinder the
122 development of $\delta^{13}\text{C}$ analysis of invertebrate resting stages as a proxy for reconstructing
123 past variations in carbon cycling in lakes.

124 The main aim of this study was to assess the influence of seasonality on $\delta^{13}\text{C}$ values
125 of zooplankton remains and evaluate their potential use as palaeoecological indicators for
126 lacustrine carbon cycle dynamics. This study provides a first assessment of seasonal
127 variations in $\delta^{13}\text{C}$ values of cladoceran and bryozoan resting stages in a small temperate
128 European lake, with respect to physical, chemical and biological variables. In particular, the
129 focus was on assessing seasonal variations in transfer pathways of carbon in the lake, and
130 their influence on the $\delta^{13}\text{C}$ values of organisms in the water column, as well as their

131 chitinous resting stages (i.e. planktonic Cladocera and their ephippia, bryozoan statoblasts).
132 In a field campaign, $\delta^{13}\text{C}$ values of *Daphnia* and their ephippia were analysed over a two-
133 year period, and $\delta^{13}\text{C}$ values of floating resting stages of *Ceriodaphnia* (Branchiopoda:
134 Cladocera: Daphniidae) and *Plumatella* were analysed when present during this period. In a
135 second step, we investigated the implications of seasonal changes for interpretation of $\delta^{13}\text{C}$
136 values in fossil invertebrate remains. To that end, ephippia of the cladocerans *Daphnia* and
137 *Ceriodaphnia*, and statoblasts of the bryozoan *Plumatella* were analysed in the lake surface
138 sediment and in a short sediment core covering roughly the past 150 years. Our study was
139 conducted on Lake Gerzensee (7°33'E, 46°50'N, 606 m a.s.l.), a small temperate lake in the
140 foreland of the Swiss Alps, about 20 km south of Bern (Fig. 1). The lake is characterised by
141 exceptionally high lake water CH_4 concentrations (Rinta et al. 2015; Schilder et al. 2016).
142 Lake Gerzensee has a surface area of 0.27 km², a total water volume of 0.16 km³, and a
143 maximum depth of 10 m. The catchment area (2.6 km²) consists of 80 % agricultural land,
144 5 % wooded land, and 15 % urban areas (Lotter et al. 2000). The mean annual temperature
145 is 8.8 °C (Bern Zollikofen 1981-2010, Bundesamt für Meteorologie und Klimatologie
146 MeteoSchweiz 2014). Today, Lake Gerzensee is eutrophic, with summer anoxia in the
147 hypolimnion (Zeh et al. 2004).

148

149 **Materials and methods**

150

151 Lake Gerzensee was sampled on 15 one-to-three-day visits between October 2012 and July
152 2014, throughout all seasons (Electronic Supplementary Material [ESM] Table S1). In the
153 first period, between October 2012 and September 2013, measurements were taken every
154 other month (6 visits). Based on the first year of data, the fieldwork protocol was adjusted to
155 gain more detailed information about deeper-water POM and chlorophyll *a* concentrations
156 (ESM Table S1). In the second year, intervals between field visits were reduced to two
157 weeks from September 2013 to December 2013, to cover the destratification period (4 visits),
158 and sampling continued with 2-4 week intervals after the lake ice thawed in February 2014,

159 until July 2014 (5 visits). Data for March and July 2014 were collected along with other
160 fieldwork activities, using a shortened protocol (ESM Table S1).

161

162 Sample collection

163

164 Vertical profiles of temperature and dissolved oxygen concentration ($[O_2]_{aq}$) were measured
165 for every metre in the water column at the lake centre (location C1, Fig. 1), using a multi-
166 sensor probe (WTW Cellox[®] 325 oxi1970i, Germany). *Daphnia* individuals were collected
167 from the oxic part of the water column at the lake centre (location C1) in multiple vertical
168 hauls, with a 40- μ m mesh plankton net. Flotsam was collected from the lake surface with a
169 hand net (mesh size \sim 120 μ m). Water samples were collected at C1 in 0.7 m and 8 m water
170 depth (top epilimnion and bottom hypolimnion, respectively), using a 5-L water sampler
171 (UWITEC, Austria). For these water samples, pH was measured (Waterproof pHTestr 20,
172 Oakton, USA), and 60 ml of water was injected with a syringe into a 118-ml glass vial
173 through a 10-mm-thick butyl rubber stopper (Apodan, Denmark) to determine the abundance
174 and $\delta^{13}C$ value of the dissolved inorganic carbon (DIC). The vials were prepared beforehand
175 with 200 μ l of H_3PO_4 (85 %), closed, and repeatedly vacuumed and flushed with N_2 to
176 ensure that no CO_2 remained in the vials (Rinta et al. 2015; Schilder et al. 2015a). Samples
177 for the $\delta^{13}C$ analysis of POM were collected from the lake water at C1 (0.7 m water depth)
178 by passing the water through a 250- μ m sieve before manually pushing water through a glass
179 fibre filter (Whatman GF/C 25 mm, pore size 1.2 μ m) with a syringe (water volume recorded
180 in the field). A second sample of POM was obtained in a similar fashion, immediately put into
181 90 % undenatured ethanol, and covered with aluminium foil for chlorophyll *a* analysis.
182 Immediately after returning from fieldwork, POM filters designated for chlorophyll *a* analysis
183 were heated to 70 °C in 90 % undenatured ethanol for 10 min, put in an ultrasound bath for
184 5 min, and stored in a refrigerator for at least 48 hours. The samples were then filtered
185 through a membrane filter to remove suspended particles. Chlorophyll *a* measurements
186 were based on absorbance at characteristic wavelengths (Schwoerbel 1994). Samples were

187 injected into glass cuvettes (1 cm light path) and absorbance at wavelengths (λ) 665 nm and
188 750 nm was measured with a spectrophotometer (Jenway, UK). Chlorophyll a content in the
189 lake water was calculated following EAWAG (1995).

190

191 Isotope analysis

192

193 *$\delta^{13}\text{C}$ analysis of DIC and POM*

194

195 The concentration of CH_4 and CO_2 in the headspace of the DIC samples was measured by
196 gas chromatography with a flame ionisation detector and methanizer (GC-FID; Shimadzu
197 GC8, PoropackN column, see Rinta et al. 2015 for details). Lake water DIC concentrations
198 and dissolved CH_4 concentrations ($[\text{CH}_4]_{\text{aq}}$) were back-calculated from measured headspace
199 CO_2 and CH_4 concentrations. Dissolved CO_2 concentrations ($[\text{CO}_2]_{\text{aq}}$) (as the sum of
200 dissolved CO_2 and H_2CO_3) were calculated following Stumm and Morgan (1996), accounting
201 for lake water pH, temperature, and DIC concentrations.

202 Filters containing lake water POM were freeze-dried and a maximum amount of filter
203 material was separated from the seston. The seston was then transferred into ultra-clean tin
204 cups. For $\delta^{13}\text{C}$ analysis of DIC, 15-20 ml of gas from the headspace of the vial was allowed
205 to escape into a 60-ml syringe. The gas was then injected into a pre-vacuumed 12-ml glass
206 vial such that slight overpressure was applied. The procedure was repeated for $\delta^{13}\text{CH}_4$
207 measurements. Isotope samples of CO_2 , CH_4 , and POM were analysed at the Stable
208 Carbon Isotope Facility of the University of California-Davis on an Elementar Vario EL Cube
209 or Micro Cube elemental analyser interfaced to a PDZ Europa 2020 isotope ratio mass
210 spectrometer (IRMS). Analytical uncertainties for gas sample $\delta^{13}\text{C}$ values were $< 0.1 \text{‰}$ (one
211 standard deviation) for CO_2 for two to three replicate measurements of three laboratory
212 standards ($\delta^{13}\text{C} = -40.73 \text{‰}$, -10.39‰ , and -3.59‰) and $< 0.1 \text{‰}$ (one standard deviation)
213 for CH_4 for replicate measurements of a laboratory standard ($\delta^{13}\text{C} = -36.7 \text{‰}$, $n = 6$). Results
214 are reported in conventional δ -notation relative to the international standard Vienna PeeDee

215 Belemnite (V-PDB).

216

217 $\delta^{13}\text{C}$ analysis of zooplankton

218

219 Living *Daphnia* were kept in approximately 1 L of unfiltered lake water for 1-2 days. They
220 were then separated from other organisms under a dissecting microscope (magnification
221 20x–50x), and were frozen in demineralised water. Samples were freeze-dried and weighed
222 into ultra clean tin cups (Lüdi Swiss AG, Switzerland). Because *Daphnia* ephippia cannot be
223 identified to species level under the microscope, the species of living *Daphnia* was not
224 determined for this analysis. Flotsam was sieved at 100 μm and examined under a
225 dissecting microscope. Ephippia of the planktonic cladocerans *Daphnia* and *Ceriodaphnia*,
226 and statoblasts of the bryozoan *Plumatella* were the only invertebrate resting stages found
227 regularly and abundantly enough for $\delta^{13}\text{C}$ analysis. These resting stages were identified
228 according to Vandekerkhove et al. (2004) (Cladocera) and Wood and Okamura (2005)
229 (Bryozoa), exposed to 10 % potassium hydroxide (KOH) for 2 hours (van Hardenbroek et al.
230 2010), rinsed 5-10 times with demineralised water, and picked into ultra-clean tin cups.
231 Isotope samples of *Daphnia* as well as ephippia and statoblasts were also analysed at the
232 Stable Carbon Isotope Facility of the University of California-Davis on an Elementar Vario EL
233 Cube or Micro Cube elemental analyser interfaced to a PDZ Europa 2020 IRMS. Sample
234 sizes were in the range of 120-250 individuals for *Daphnia* (150-500 μg), 150-200 for
235 *Daphnia* ephippia (150-250 μg), 200-300 for *Plumatella* statoblasts (150-300 μg), and 450-
236 600 (150-200 μg) for *Ceriodaphnia* ephippia. Analytical uncertainties for invertebrate $\delta^{13}\text{C}$
237 measurements were ≤ 0.6 ‰ (one standard deviation) for replicate measurements ($n = 3$ -36)
238 of five laboratory standards (Bovine Liver ($\delta^{13}\text{C} = -21.7$ ‰), USGS-41 Glutamic Acid ($\delta^{13}\text{C} =$
239 37.6 ‰), Nylon 5 ($\delta^{13}\text{C} = -27.7$ ‰), Peach Leaves ($\delta^{13}\text{C} = -26.1$ ‰), Glutamic Acid ($\delta^{13}\text{C} = -$
240 28.9 ‰)).

241

242 Sediment analysis

243

244 *Sediment coring and chronology*

245

246 In October 2012, a 62-cm-long sediment core (GER12) was recovered from the centre of
247 Lake Gerzensee (9.5 m water depth) using a gravity corer (UWITEC, Austria). Upon arrival
248 in the laboratory, core GER12 was sampled at 1-cm intervals and freeze-dried until further
249 analysis. For ^{210}Pb and ^{137}Cs dating, a total of 15 freeze-dried samples from the upper 48 cm
250 of the core were analysed using gamma spectrometry at the Department of Chemistry and
251 Biochemistry at the University of Bern, Switzerland. ^{210}Pb (46.5 keV), ^{241}Am (59.5 keV),
252 ^{226}Ra progenies (351.9 and 609.3 keV), and ^{137}Cs (661.7 keV) were measured using a
253 Broad Energy Germanium (BEGe) Canberra detector with low background and high
254 absolute full-energy peak efficiencies for close on-top geometries of > 20 % and ~ 5 % for
255 ^{210}Pb and ^{137}Cs , respectively.

256

257 *$\delta^{13}\text{C}$ analysis of sedimentary invertebrate remains*

258

259 For invertebrate $\delta^{13}\text{C}$ analysis, core GER12 was sampled every fourth centimetre. When
260 sample mass of the invertebrate remains was not sufficient for $\delta^{13}\text{C}$ analysis, the lower
261 adjacent centimetre was added to the sample. Nonetheless, some samples in the lower half
262 of the core had to be pooled with the next regular-interval sample to obtain a sufficient
263 number of remains for analysis. Hence, invertebrate $\delta^{13}\text{C}$ values represent remains from up
264 to 6 cm of sediment. For each sample, 50 % by weight of the freeze-dried material was
265 deflocculated in 10 % KOH for 2 hours and sieved at 100 μm (van Hardenbroek et al. 2010).
266 *Daphnia* ephippia, *Ceriodaphnia* ephippia, and *Plumatella* statoblasts were identified
267 according to Vandekerkhove et al. (2004) for Cladocera and Wood and Okamura (2005) for
268 Bryozoa and separated from the sediment. Remains were treated with 2M NH_4Cl solution
269 buffered with 0.35 NaOH for 20 hours to remove carbonates (Verbruggen et al. 2010), and
270 picked into pre-weighed silver cups (6 x 4 mm; Säntis, Switzerland). All fossil invertebrates

271 were analysed at the Alaska Stable Isotope Facility of the University of Alaska, Fairbanks,
272 on a Costech ESC 4010 elemental analyzer interfaced via a ThermoConflo III to a Thermo
273 Delta V IRMS. Analytical uncertainties were < 0.1 ‰ for replicate measurements of a
274 laboratory standard (peptone, $\delta^{13}\text{C} = -15.8 \text{ ‰}$, $n = 24$). Sample sizes were in the range of
275 20-55 remains for *Daphnia* ephippia and *Plumatella* statoblasts, and 100-200 for
276 *Ceriodaphnia* ephippia. In addition to the sediment core, material from a sediment trap
277 located close to the coring site at the centre of Gerzensee during 2012/13 was available
278 from another study (C. Adolf, unpublished). The trap was placed 3 m above the lake floor
279 and remained in the lake for 12 months. The material was processed in the same way as the
280 sediment core, and one sample for *Ceriodaphnia* ephippia was analysed for its $\delta^{13}\text{C}$ value.

281

282 Statistical analysis

283

284 Statistical analyses on material collected alive and as floatsam were performed in R (R
285 Development Core Team 2008). For Pearson correlation tests, normality of the variables
286 was tested prior to analysis using the Shapiro-Wilk test. When a normal distribution of the
287 data was rejected, Spearman's rank correlation was used to test for correlations. The
288 average value reported for fieldwork parameters was calculated as a weighted mean of all
289 samples collected during the campaign and each sample was weighted by the time interval
290 for which it is most representative.

291

292 **Results**

293

294 Physical and chemical lake conditions

295

296 All variables showed characteristic seasonal variations throughout the two years of
297 monitoring. Measured surface water temperatures ranged between 4.6 °C in December

298 2013 and 25.8 °C in July 2013, with a mean of 12.4 °C (Fig. 2a). The bottom water (8 m
299 depth) was anoxic ($[O_2]_{aq} < 1 \text{ mg L}^{-1}$) from June to November (Fig. 2b). Chlorophyll *a* values
300 ranged from 15.2 $\mu\text{g chl } a \text{ L}^{-1}$ in autumn 2013 to 7.2 $\mu\text{g chl } a \text{ L}^{-1}$ in February 2014, but only
301 one measurement was taken during summer (July 2014). Deep-water chlorophyll *a*
302 measured at 7 m depth in May and July 2014 was comparable (July, 14.7 $\mu\text{g chl } a \text{ L}^{-1}$) or
303 higher (May, 14.2 $\mu\text{g chl } a \text{ L}^{-1}$) than surface water measurements (12.6 and 8.6 $\mu\text{g chl } a \text{ L}^{-1}$,
304 respectively). $[CO_2]_{aq}$ in the epilimnion ranged from peak values just after autumn mixing and
305 in spring (96.7-192.8 μM) to 9.6 μM in September (ESM Fig. S1). $[CH_4]_{aq}$ in the water
306 column reached high values (1082 μM) in the hypolimnion during summer stratification.
307 Surface water $[CH_4]_{aq}$ was highest in November during autumn turnover (49.8 μM , ESM Fig.
308 S1).

309

310 Stable carbon isotopes

311

312 With the exception of *Daphnia* ephippia, all measured variables showed seasonal variation
313 in $\delta^{13}\text{C}$ values, with the lowest values right after autumn lake mixing and throughout winter,
314 and the highest values in summer (Fig. 3a).

315 The isotopic composition of $[CO_2]_{aq}$ ($\delta^{13}\text{C}_{CO_2}$) for location C1 at the lake surface was
316 most ^{13}C -depleted in February and March ($\delta^{13}\text{C}_{CO_2} \sim -21 \text{ ‰}$), whereas the highest $\delta^{13}\text{C}_{CO_2}$
317 value of -12.1 ‰ was recorded in July 2013 (Fig. 3a). The seasonal cycle in $\delta^{13}\text{C}_{CO_2}$ values
318 was less pronounced in the hypolimnion (data not shown). During summer, a vertical
319 gradient in $\delta^{13}\text{C}_{CO_2}$ values developed between the epilimnion and the hypolimnion, which
320 was eliminated at lake mixing in autumn. At the lake centre, the $\delta^{13}\text{C}$ value of surface water
321 POM ($\delta^{13}\text{C}_{POM}$) was highest during summer (-28.5 ‰), and reached its minimum in March
322 2013 (-41.6 ‰, mean = -34.7 ‰; Fig. 3a). Bottom water $\delta^{13}\text{C}_{POM}$, measured at 7 m depth,
323 was in the range of surface water $\delta^{13}\text{C}_{POM}$ for measurements in April and May 2014 (ESM
324 Fig. S2d). However, during summer stratification in July 2014, bottom water $\delta^{13}\text{C}_{POM}$ values
325 were distinctly more negative than surface water $\delta^{13}\text{C}_{POM}$ (-38.0 ‰ and -30.0 ‰,

326 respectively). Surface water $\delta^{13}\text{C}_{\text{POM}}$ was on average 17.0 ‰ (range 13.2 to 21.0 ‰) more
327 negative than surface water $\delta^{13}\text{C}_{\text{CO}_2}$, with the largest differences measured in March 2013
328 and May 2014 (Fig. 3a).

329 *Daphnia* $\delta^{13}\text{C}$ values ($\delta^{13}\text{C}_{\text{Daph}}$) in the lake centre were on average 3.4 ‰ more
330 negative than $\delta^{13}\text{C}_{\text{POM}}$ values (Fig. 3a). The average $\delta^{13}\text{C}_{\text{Daph}}$ value was -39.4 ‰. Values
331 ranged from -44.2 ‰ in early spring to -29.8 ‰ in summer. $\delta^{13}\text{C}_{\text{Daph}}$ values were positively
332 correlated with chlorophyll *a* concentrations in surface water POM (Pearson correlation $r =$
333 0.86, $p < 0.05$, $n = 6$; ESM Fig. S2a), and $[\text{CH}_4]_{\text{aq}}$ in bottom waters (log-transformed,
334 Pearson correlation $r = 0.86$, $p < 0.01$, $n = 13$; ESM Fig. S2b). Moreover, bottom water
335 $[\text{CH}_4]_{\text{aq}}$ showed a negative correlation with the offset between $\delta^{13}\text{C}_{\text{POM}}$ and $\delta^{13}\text{C}_{\text{Daph}}$ ($\Delta^{13}\text{C}_{\text{POM-}}$
336 Daph) (Spearman's rank correlation $r = 0.67$, $p < 0.05$, $n = 13$; ESM Fig. S2c).

337 In contrast to $\delta^{13}\text{C}_{\text{Daph}}$ values, the $\delta^{13}\text{C}$ values of floating *Daphnia* ephippia ($\delta^{13}\text{C}_{\text{DaphFlot}}$)
338 did not show seasonal variations, but instead remained relatively constant around -39.5 ‰
339 (range -41.7 to -38.8 ‰, Fig. 3a and b). Interestingly, floating *Ceriodaphnia* ephippia $\delta^{13}\text{C}$
340 values ($\delta^{13}\text{C}_{\text{CerioFlot}}$) showed changes over time, with the lowest $\delta^{13}\text{C}$ values down to -50.1 ‰
341 in winter (Fig. 3b). On average, $\delta^{13}\text{C}_{\text{CerioFlot}}$ was more ^{13}C -depleted (mean = -42.9 ‰, range -
342 50.1 to -34.3 ‰) than $\delta^{13}\text{C}_{\text{DaphFlot}}$, whereas *Plumatella* statoblast $\delta^{13}\text{C}$ ($\delta^{13}\text{C}_{\text{PluFlot}}$) was least
343 ^{13}C -depleted (mean = -36.3 ‰, range -40.0 to -34.6 ‰). In-lake abundance of the floating
344 remains could not be assessed by the applied sampling technique, and no structural
345 degradation of the remains was visible under the microscope on any of the fieldwork dates.
346 *Ceriodaphnia* ephippia from the sediment trap had a $\delta^{13}\text{C}$ value of -42.4 ‰.

347

348 Sediment core and chronology

349

350 For core GER12, activity of unsupported ^{210}Pb was transformed into an age-depth model for
351 the upper 48 cm using the constant rate of supply (CRS) model (Appleby 2001; Appleby and
352 Oldfield 1978) (Fig. 4). Activity measurements of ^{137}Cs showed a distinct peak at 28.5 cm
353 and a second smaller peak at 20.5 cm. Many European lakes show two maxima in ^{137}Cs

354 activity, which are associated with aboveground nuclear bomb tests in AD 1963 and the
355 Chernobyl reactor accident in AD 1986 (Appleby 2001). For Swiss lakes, the latter peak is
356 usually greater than the first (Albrecht et al. 1998; Lotter et al. 1997). This was not the case
357 in Gerzensee sediments where only one measurement point defined the anticipated
358 Chernobyl peak. The Chernobyl peak may thus not be fully revealed in the record because
359 of the relatively low sampling resolution, i.e. every 4 cm. Therefore the ^{210}Pb model was not
360 modified to fit the ^{137}Cs profile. The lower peak in ^{137}Cs activity, expected to coincide with the
361 nuclear bomb peak in AD 1963, was confirmed by ^{241}Am and agrees well with the ^{210}Pb
362 chronology (Fig. 4).

363 The abundance of invertebrate remains in the sediments of Gerzensee allowed for the
364 analysis of eight samples of *Daphnia ephippia* covering the upper 27 cm, nine samples of
365 *Plumatella statoblasts* (0-43 cm), and 15 samples of *Ceriodaphnia ephippia* (0-62 cm). The
366 $\delta^{13}\text{C}$ values of all three invertebrate genera showed only small variations throughout the
367 record (*Daphnia ephippia* -39.8 to -37.9 ‰, *Ceriodaphnia ephippia* -43.3 to -39.2 ‰, and
368 *Plumatella statoblasts* -33.3 to -29.7 ‰; Fig. 5). $\delta^{13}\text{C}$ values of *Plumatella statoblasts* were
369 on average distinctly less negative (-32.0 ‰) than the remains of the two mobile filter
370 feeders *Daphnia* and *Ceriodaphnia* (-39.0 ‰ and -41.9 ‰, respectively). The most negative
371 $\delta^{13}\text{C}$ values were measured for *Ceriodaphnia ephippia* with the minimum of -43.3 ‰ at 41-
372 42 cm depth (AD ~1920). Remains of *Daphnia* and *Ceriodaphnia* showed a very similar
373 pattern, with one distinct peak of less negative $\delta^{13}\text{C}$ values at about 18 cm depth (AD ~1980).
374 In contrast, $\delta^{13}\text{C}$ values of *Plumatella statoblasts* remained constant during this period, but
375 show a 3 ‰ increase around 30 cm (AD ~1955). Besides differences in $\delta^{13}\text{C}$ values,
376 *Plumatella statoblasts* had a lower C:N ratio (mean = 4.3) compared to values for *Daphnia*
377 and *Ceriodaphnia ephippia* (6.3 and 6.5, respectively).

378

379 Discussion

380

381 Seasonal variation in *Daphnia* $\delta^{13}\text{C}$ values

382

383 *Daphnia* $\delta^{13}\text{C}$ values were consistently below -38.5‰ and reached minimum values, less
384 than -44‰ , in the period from autumn to spring. This is well below commonly reported
385 values for algae, which usually lie above -35‰ (France 1995; Peterson and Fry 1987;
386 Vuorio et al. 2006). Nonetheless, $\delta^{13}\text{C}_{\text{Daph}}$ values closely track the seasonal cycle of surface
387 water $\delta^{13}\text{C}_{\text{POM}}$ values (Fig. 3a), but with an average difference of 3.4‰ between $\delta^{13}\text{C}_{\text{POM}}$
388 and $\delta^{13}\text{C}_{\text{Daph}}$. Hence, $\Delta^{13}\text{C}_{\text{POM-Daph}}$ was slightly higher than values reported by del Giorgio and
389 France (1996) for zooplankton in Canadian lakes and other published values discussed by
390 these authors (mean difference 2.6‰). A positive correlation between $\delta^{13}\text{C}_{\text{Daph}}$ values and
391 surface water chlorophyll *a* concentrations, which can serve as a proxy for algal
392 concentrations (Sartory and Grobbelaar 1984), was observed between autumn and spring
393 (ESM Fig. S2a). Hence, low $\delta^{13}\text{C}_{\text{Daph}}$ values coincide with relatively low concentrations of
394 algae in surface water POM. During times when algae are less abundant, other food sources
395 may contribute to the carbon uptake of *Daphnia* (Taipale et al. 2008). Several studies have
396 shown that *Daphnia* and other invertebrates can incorporate CH_4 -derived carbon to
397 supplement their diet (Bastviken et al. 2003; Devlin et al. 2015; Kankaala et al. 2006). In a
398 study of a small Finnish lake, Taipale et al. (2008) found that MOB contributed to *Daphnia*'s
399 diet throughout the year, comprising up to 50 % of the diet in autumn. Considering the high
400 CH_4 concentrations in Gerzensee, uptake of ^{13}C -depleted carbon by feeding on MOB seems
401 the likely explanation for the exceptionally low $\delta^{13}\text{C}_{\text{Daph}}$ values measured in Gerzensee.

402

403 The relationship between *Daphnia* $\delta^{13}\text{C}$ values and $[\text{CH}_4]_{\text{aq}}$

404

405 Field studies that compared the carbon isotopic composition of invertebrate remains in
406 surface sediment samples with in-lake CH_4 abundance suggest that a relationship exists
407 between the $\delta^{13}\text{C}$ values of some invertebrate groups and lake water CH_4 concentrations
408 (Schilder et al. 2015a; van Hardenbroek et al. 2013). Within our two-year measurement

409 period at Gerzensee, $\delta^{13}\text{C}_{\text{Daph}}$ was positively correlated with $[\text{CH}_4]_{\text{aq}}$ in the bottom waters, i.e.
410 we recorded $\delta^{13}\text{C}_{\text{Daph}}$ values more similar to $\delta^{13}\text{C}_{\text{POM}}$ during times of high bottom water
411 $[\text{CH}_4]_{\text{aq}}$ (ESM Fig. S2b). This is also indicated by the negative correlation between $[\text{CH}_4]_{\text{aq}}$ in
412 the bottom waters and $\Delta^{13}\text{C}_{\text{POM-Daph}}$ (ESM Fig. S2c). Hence, during lake stratification, when
413 $[\text{CH}_4]_{\text{aq}}$ is increasing in the hypolimnion, *Daphnia* utilise carbon whose isotopic signature is
414 more similar to $\delta^{13}\text{C}_{\text{POM}}$ than at times when the water column is mixed. This suggests that
415 during lake stratification, algae are the main food source of *Daphnia*, whereas *Daphnia* may
416 rely more heavily on other food sources when the lake is mixed. In a multi-lake study,
417 Schilder et al. (2015a) found a negative correlation between late summer $[\text{CH}_4]_{\text{aq}}$ in both
418 bottom and surface waters, and $\delta^{13}\text{C}$ values of *Daphnia* ehippia isolated from surface
419 sediment samples, indicating that *Daphnia* are more likely to incorporate CH_4 -derived carbon
420 in lakes with high CH_4 abundance. With regard to this relationship, our results imply that in
421 lakes with high CH_4 accumulation during summer stratification, more ^{13}C -depleted carbon
422 becomes accessible to the food web upon mixing and oxygenation of CH_4 in the water
423 column, and this signal is seen in the floating ehippia. Other processes, e.g. increased
424 transport of allochthonous organic carbon and higher associated CH_4 production may also
425 contribute to the importance of CH_4 -derived carbon for *Daphnia* in the autumn months.

426

427 Floating cladoceran and bryozoan remains

428

429 In a field study, Perga (2011) showed that there is no significant carbon isotope fractionation
430 between *Daphnia* and their ehippia. This was confirmed in a laboratory experiment by
431 Schilder et al. (2015b), which showed that *Daphnia* ehippia $\delta^{13}\text{C}$ values closely resemble
432 those of *Daphnia* during ehippia production under different environmental conditions (-0.2
433 ± 0.4 ‰ for 12°C , 1.3 ± 0.3 ‰ for 20°C). In Gerzensee, $\delta^{13}\text{C}_{\text{DaphFlot}}$ remained very similar
434 across the annual cycle (Figs. 3a, b). If no fractionation is assumed during ehippia
435 production, this suggests that the ehippia were produced batch-wise and then floated on
436 the lake for several months. In the data set presented here, periods when $\delta^{13}\text{C}_{\text{DaphFlot}}$

437 coincided with $\delta^{13}\text{C}_{\text{Daph}}$ include December 2012, June 2013, November 2013, and April 2014.
438 This is in line with the observation that ephippia production commonly occurs in early
439 summer and late autumn (Cáceres 1998), when environmental conditions deteriorate, e.g.
440 there is temperature decline, crowding, oxygen depletion, or limited food availability (Korhola
441 and Rautio 2001). Hence, in Gerzensee, changes in $\delta^{13}\text{C}_{\text{DaphFlot}}$ reflect $\delta^{13}\text{C}_{\text{Daph}}$ during
442 ephippia production, i.e. in late autumn and/or early spring, but $\delta^{13}\text{C}_{\text{DaphFlot}}$ does not reflect
443 seasonal changes in the diet of *Daphnia*. This observation differs from the findings of
444 Schilder (2014) at Dutch Lake De Waay, where *Daphnia* ephippia followed the seasonal
445 variation of *Daphnia* $\delta^{13}\text{C}$, indicating several production periods per year. The reason for the
446 different timing of ephippia production is unclear, although it could be explained by
447 differences in species composition between the two lakes. Regardless of the cause, our
448 results indicate that different production intervals, and possibly switching between different
449 production intervals over time, potentially influences *Daphnia* ephippia $\delta^{13}\text{C}$ values in
450 sediment records.

451 Floating *Ceriodaphnia* ephippia show larger variability in $\delta^{13}\text{C}$ values than *Daphnia*
452 ephippia (Fig. 3b). To our knowledge, no prior study has assessed the fractionation between
453 maternal *Ceriodaphnia* and the chitinous structure of their resting eggs. Thus, following the
454 simplest assumption of no (or constant) fractionation, as is observed for *Daphnia*, our results
455 suggest that *Ceriodaphnia* ephippia are produced at several times or continuously
456 throughout the year. No living *Ceriodaphnia* were analysed in this study, but the broad range
457 of $\delta^{13}\text{C}_{\text{CeriodaphFlot}}$ values (-34.3 to -50.1 ‰), which is comparable in span to the seasonal cycle of
458 living *Daphnia* (-29.8 to -44.2 ‰), may be an indication that the changes in $\delta^{13}\text{C}_{\text{CeriodaphFlot}}$
459 represent (a part of) the seasonal cycle of *Ceriodaphnia* $\delta^{13}\text{C}$.

460 Floating *Plumatella* statoblast $\delta^{13}\text{C}$ values also show more pronounced seasonal
461 changes than $\delta^{13}\text{C}_{\text{DaphFlot}}$, but the variability is not as large as for $\delta^{13}\text{C}_{\text{CeriodaphFlot}}$ (Fig. 3b). Since
462 no living *Plumatella* zooids were analysed, no pattern of statoblast production can be
463 determined. However, a recent study by van Hardenbroek et al. (2016) suggests that the
464 $\delta^{13}\text{C}$ values of *Plumatella* statoblasts collected from colonies are significantly correlated with

465 *Plumatella* zooid $\delta^{13}\text{C}$. Okamura and Hatton-Ellis (1995) found that production of statoblasts
466 may start in mid-summer, but highest production occurs in late summer and early autumn.
467 Overwintering statoblasts are released when colonies collapse in late autumn. Following
468 these findings, $\delta^{13}\text{C}$ values of floating *Plumatella* statoblasts may reflect zooid $\delta^{13}\text{C}$ values
469 during statoblast production from late summer to late autumn.

470

471 Linking $\delta^{13}\text{C}$ values of floating resting stages to those of fossil invertebrate remains

472

473 The average $\delta^{13}\text{C}$ value of floating *Daphnia* ephippia ($-39.5\text{‰} \pm 0.9$ [one standard
474 deviation]) is in excellent agreement with the $\delta^{13}\text{C}$ value of *Daphnia* ephippia in the
475 uppermost 1-3 cm of the sediment at the centre of the lake (-39.8‰ ; Fig. 5), and with the
476 average $\delta^{13}\text{C}$ value of living *Daphnia* collected during the two-year fieldwork period (-39.4‰
477 ± 4.6). This confirms that $\delta^{13}\text{C}$ values of *Daphnia* ephippia in the sediment of Gerzensee
478 reflect the $\delta^{13}\text{C}$ signal of floating *Daphnia* ephippia during the year. Seasonal changes in
479 $\delta^{13}\text{C}_{\text{Daph}}$, however, are not recorded by floating or deposited *Daphnia* ephippia $\delta^{13}\text{C}$ values in
480 the lake (see previous sections). This has implications for the interpretation of the sediment
481 core record of *Daphnia* ephippia $\delta^{13}\text{C}$ values. In Gerzensee, nearly constant $\delta^{13}\text{C}_{\text{DaphFlot}}$
482 values throughout the year indicate batch-wise production of ephippia in autumn and/or early
483 spring. Changes in the $\delta^{13}\text{C}$ value of sedimentary *Daphnia* ephippia may thus record
484 changes in the importance of MOB in the diet of *Daphnia* (only) during spring and/or autumn.
485 The latter period is indirectly coupled with the accumulation of $[\text{CH}_4]_{\text{aq}}$ during summer
486 stratification, which is the reason for high CH_4 abundance in the water column during autumn
487 lake mixing. Hence $\delta^{13}\text{C}$ values of sedimentary *Daphnia* ephippia may indicate the
488 accumulation of $[\text{CH}_4]_{\text{aq}}$ during summer stratification in small European lakes like Lake
489 Gerzensee, as suggested by Schilder et al. (2015a).

490 The *Ceriodaphnia* ephippia $\delta^{13}\text{C}$ value recorded in the sediment trap during 2012/13 (-
491 42.5‰) is in excellent agreement with the surface sediment measurement presented here (-
492 42.9‰ , 1-3 cm; Fig. 5). The average flotsam $\delta^{13}\text{C}$ value ($-42.9\text{‰} \pm 4.8$) also agrees very

493 well with the surface sediment measurement, although no full annual cycle is covered by the
494 flotsam measurements (October to May only, Fig. 3b). Our study is the first to assess and
495 confirm that the $\delta^{13}\text{C}$ values of *Ceriodaphnia* ehippia in the sediment can provide
496 information about the $\delta^{13}\text{C}$ value of floating ehippia at the time of deposition.

497 The *Plumatella* statoblast $\delta^{13}\text{C}$ value in the surface sediment (-32.7 ‰, pooled 1-3 and
498 5-7 cm, this study) was distinctly less negative than the average $\delta^{13}\text{C}$ value of floating
499 statoblasts (-36.3 ‰ \pm 1.7; Fig. 5). It should, however, be noted here that summer
500 measurements of $\delta^{13}\text{C}_{\text{PluFlot}}$ are underrepresented in our dataset, with only one measurement
501 in July 2013 (Fig. 3b). It therefore remains uncertain whether a better representation of the
502 summer period may resolve this mismatch between average flotsam values and the surface
503 sediment measurement.

504

505 Interpreting fossil invertebrate $\delta^{13}\text{C}$ values

506

507 *Daphnia* ehippia $\delta^{13}\text{C}$ values in our sediment record (-39.8 to -37.9 ‰) were continuously
508 below -35 ‰, whereas algae are commonly reported to have $\delta^{13}\text{C}$ values above -35 ‰
509 (France 1995; Peterson and Fry 1987; Vuorio et al. 2006). Our values are comparable to
510 sedimentary *Daphnia* ehippia $\delta^{13}\text{C}$ values found in Lake De Waay, the Netherlands
511 (Schilder 2014), but are distinctly more negative than cladoceran $\delta^{13}\text{C}$ values reported from
512 sediment records of Lake Strandsjön, Sweden (van Hardenbroek et al. 2014) and Lake
513 Annecy, France (Frossard et al. 2014). In sediment records covering longer time periods,
514 van Hardenbroek et al. (2013) and Wooller et al. (2012) found $\delta^{13}\text{C}$ values over the range of
515 all of the studies mentioned above. In a recent study in Lake Mekkojärvi, Finland, Rinta et al.
516 (2016) showed that *Daphnia* ehippia $\delta^{13}\text{C}$ values changed abruptly from values below -
517 45 ‰ to values greater than -40 ‰, shifts that these authors interpreted as representing
518 changes in the availability of CH_4 in this lake. As discussed earlier, $\delta^{13}\text{C}$ values in *Daphnia*
519 ehippia have been shown to correlate negatively with CH_4 abundance in the hypolimnia of
520 small European lakes (Schilder et al. 2015a). Therefore, the observed *Daphnia* ehippia

521 $\delta^{13}\text{C}$ values below -35 ‰ suggest that $[\text{CH}_4]_{\text{aq}}$ remained high at Gerzensee throughout the
522 record. Only the least negative $\delta^{13}\text{C}$ values, around 20 cm (AD ~1970-1980), could indicate
523 a short period with a reduced influence of ^{13}C -depleted carbon in the diet of *Daphnia*, but
524 $\delta^{13}\text{C}$ values remained clearly more negative than commonly reported algal $\delta^{13}\text{C}$ values.
525 Lower *Daphnia* $\delta^{13}\text{C}$ values observed for the sediments of eutrophic and stratified lakes
526 Gerzensee and De Waay can potentially be explained by a greater importance of CH_4 -
527 derived carbon compared to that in less nutrient-rich Lake Annecy and non-stratified Lake
528 Strandsjön, where algae are the main carbon source of *Daphnia*. Mekkojärvi is a very small,
529 stratified, humic lake with low oxygen concentrations within 1 m of the lake surface (Rinta et
530 al. 2015), which may explain the high relevance of CH_4 -derived carbon in its planktonic food
531 web.

532 *Ceriodaphnia ephippia* were not analysed in any of the above-mentioned studies.
533 *Ceriodaphnia ephippia*, however, had even more negative $\delta^{13}\text{C}$ values than *Daphnia* in the
534 Gerzensee sediment (range -43.3 to -39.4 ‰; Fig. 5). This indicates a similar, but potentially
535 more dominant source of ^{13}C -depleted carbon for *Ceriodaphnia ephippia*. The systematic
536 ^{13}C -depletion of *Ceriodaphnia ephippia* relative to *Daphnia ephippia* may be caused by a
537 difference in the average particle size that these taxa filter from the water. *Ceriodaphnia* feed
538 on smaller particles than *Daphnia* (Geller and Müller 1981), potentially containing a higher
539 proportion of MOB and other bacteria. The relative ^{13}C -depletion of *Ceriodaphnia* could also
540 be caused by differences in body composition, e.g. lipid content, between taxa, as has been
541 suggested by Matthews and Mazumder (2005) for other zooplankton species. However, C:N
542 ratios, which may be an indicator of lipid content (Matthews and Mazumder 2005), do not
543 show large differences between *Daphnia* and *Ceriodaphnia ephippia* (mean = 6.3 and 6.4 for
544 the uppermost 27 cm of the sediment record, respectively). Therefore, low $\delta^{13}\text{C}$ values
545 throughout the record suggest that, similar to *Daphnia*, *Ceriodaphnia* rather consistently
546 incorporated a ^{13}C -depleted carbon source over the past 150+ years, at least during the
547 season(s) of ephippia production.

548 *Plumatella statoblast* $\delta^{13}\text{C}$ values in our sediment record were distinctly less negative

549 compared to *Daphnia* and *Ceriodaphnia* ehippia (-33.3 to -29.7 ‰; Fig. 5). This suggests
550 that *Plumatella* zooids accessed carbon sources that were less ^{13}C -depleted compared to
551 those on which the two cladoceran taxa relied. Similar $\delta^{13}\text{C}$ values were reported by van
552 Hardenbroek et al. (2014), for *Plumatella* statoblasts in shallow Lake Strandsjön in Sweden
553 (-33.1 to -28.0 ‰). Our $\delta^{13}\text{C}$ values for *Plumatella* statoblasts are well in the range of
554 commonly reported $\delta^{13}\text{C}$ values for algae. Hence, it appears that $\delta^{13}\text{C}$ values of sedimentary
555 *Plumatella* statoblasts can potentially provide information about the temporal evolution of
556 algal $\delta^{13}\text{C}$, and thus help track past changes in the lake's carbon cycle. However, a single
557 statoblast flotsam sample from our study, collected during autumn mixing in November 2013
558 (Fig. 3b), yielded a $\delta^{13}\text{C}$ value of -40.0 ‰, well below the values expected for algae in Lake
559 Gerzensee. Similarly, individual bryozoan samples with very negative $\delta^{13}\text{C}$ values have
560 been reported in previous studies (e.g. Rinta et al. 2016; van Hardenbroek et al. 2016). This
561 suggests that these organisms may also ingest ^{13}C -depleted carbon sources such as MOB
562 under circumstances when they are abundant in their habitats.

563 In the sediment core from Gerzensee, *Daphnia* and *Ceriodaphnia* ehippia show very
564 similar variations in their $\delta^{13}\text{C}$ values in the part of the core where both groups are present
565 (0-27 cm; Fig. 5). This suggests that *Daphnia* and *Ceriodaphnia* were subject to similar
566 changes in their diet over time. In contrast, *Plumatella* statoblasts show a peak in $\delta^{13}\text{C}$
567 values that does not coincide with the peak in the ehippia of *Daphnia* and *Ceriodaphnia*.
568 The two groups also differ in the range of their $\delta^{13}\text{C}$ values, which points towards different
569 food sources for *Plumatella* and the cladoceran taxa. Access to different carbon sources
570 may occur because of differences in both habitat and particle-size preference. *Plumatella*
571 are sessile organisms in shallow areas of lakes, whereas free-living invertebrates like
572 *Daphnia* and *Ceriodaphnia* can reach deep-water food sources. Zooplankton in deeper
573 water layers may feed directly on MOB, or on other organisms that feed on MOB (e.g.
574 ciliates), which can lead to distinctly negative zooplankton $\delta^{13}\text{C}$ values (Jones and Grey
575 2011). In addition, Kaminski (1984) estimated that *Plumatella* can ingest particles between 5
576 and 17 μm in diameter, whereas *Daphnia* are able to filter particles between 0.5 and 30 μm

577 in diameter (Geller and Müller 1981). Hence, bacteria (< 2 µm) that are not attached to
578 larger particles may be too small to be caught in the lophophores of *Plumatella*, but may be
579 accessible as a food source for *Daphnia*. An alternative explanation for the observed
580 differences between cladoceran and bryozoan remains may be that both groups rely on the
581 same carbon sources (e.g. algae, detritus, heterotrophic bacteria, and possibly MOB), but
582 their relative importance in the diet is different. Whether one, both or more reasons are
583 causing the observed differences in $\delta^{13}\text{C}$ values cannot be conclusively answered by this
584 study.

585

586 **Conclusions**

587

588 In Gerzensee, *Daphnia* $\delta^{13}\text{C}$ values closely follow the seasonal cycle of POM $\delta^{13}\text{C}$ values,
589 with the most negative values of *Daphnia*, down to -44.2 ‰, measured in early spring.
590 Interestingly, *Daphnia* ephippia $\delta^{13}\text{C}$ values did not show any pronounced seasonal
591 variations. Seasonal changes in *Daphnia* $\delta^{13}\text{C}$ values, and hence potential seasonal
592 changes in the diet of *Daphnia*, are therefore not recorded in the $\delta^{13}\text{C}$ values of *Daphnia*
593 ephippia in the sediments of Gerzensee. Very low $\delta^{13}\text{C}$ values of *Ceriodaphnia* ephippia in
594 the flotsam, especially in winter (-50.1 ‰), confirm that there is a contribution of CH₄-derived
595 carbon to the diet of the crustacean zooplankton in Gerzensee. *Plumatella* statoblasts are
596 less depleted in ¹³C than the two cladoceran taxa, indicating little or no influence of CH₄-
597 derived carbon in the diet of the bryozoan, at least during the time(s) when statoblasts are
598 produced.

599 The average *Ceriodaphnia* and *Daphnia* ephippia $\delta^{13}\text{C}$ values in the flotsam agree
600 very well with the values in surface sediments, and for *Ceriodaphnia*, with measurements
601 from a sediment trap. This shows that in Gerzensee the $\delta^{13}\text{C}$ values of ephippia in the
602 surface sediment, and potentially in general at a certain sediment depth, reflect an
603 integrated value for ephippia floating on the lake during a particular time span, possibly over

604 one to two years. In contrast, average flotsam *Plumatella* statoblast $\delta^{13}\text{C}$ values were about
605 4 ‰ more negative than the surface sediment measurement. The reason for this mismatch
606 may be an under-representation of the summer period in our data set. Nonetheless, this
607 issue deserves further investigation in Gerzensee and other lakes.

608 Low $\delta^{13}\text{C}$ values of the two cladocerans throughout the 62-cm-long sediment core,
609 comparable to those measured in the flotsam, indicate that in Gerzensee, *Daphnia* and
610 *Ceriodaphnia* relied on a ^{13}C -depleted carbon source to supplement their diet throughout the
611 past 150 years. *Daphnia* and *Ceriodaphnia* ephippia $\delta^{13}\text{C}$ values showed similar variations,
612 with the highest values for the two taxa around the same depth. This suggests that *Daphnia*
613 and *Ceriodaphnia* have been subject to similar changes in their diet over time. In contrast,
614 *Plumatella* statoblast $\delta^{13}\text{C}$ values in the sediment record were again distinctly less negative,
615 and showed a 3 ‰ increase at a different depth than the $\delta^{13}\text{C}$ values of the two cladocerans.
616 This difference might be explained, in part, by the feeding behaviour and different habitats of
617 the organism groups, which influences the extent to which they can incorporate CH_4 -derived
618 carbon. Our study confirms the findings of earlier studies, which indicated that taxon-specific
619 $\delta^{13}\text{C}$ values measured on aquatic invertebrate remains can provide insights into long-term
620 changes in the relative importance of different carbon sources. This technique can be
621 applied to better understand the impacts of, for example, land-use change, eutrophication,
622 and climate change on the carbon cycling in lakes. Additional studies similar to the one
623 presented here will be needed to further constrain the effects of seasonality on the $\delta^{13}\text{C}$
624 values of fossil statoblast and ephippia samples, and to assess the extent to which seasonal
625 patterns observed in Gerzensee are representative of those in other lake ecosystems.

626

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628

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635

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782 **Figure Captions**

783

784 **Fig. 1** Map of Switzerland showing the location of Lake Gerzensee in the Swiss foreland of
785 the Alps (triangle) and the bathymetry of the lake. Black circles indicate the fieldwork
786 measuring station (C1) and coring site

787

788 **Fig. 2** (a) Temperature and (b) dissolved oxygen concentrations ($[O_2]_{aq}$) measured at 1-m
789 depth intervals in the water column at the lake centre (location C1) over the fieldwork period
790 in 2012-2014. Colour figure online.

791

792 **Fig. 3** (a) $\delta^{13}C$ values of CO_2 , particulate organic matter (POM), *Daphnia* for location C1
793 (pelagic zone) and *Daphnia ephippia* (whole lake) over the fieldwork period. (b) $\delta^{13}C$ values
794 of *Daphnia ephippia*, *Ceriodaphnia ephippia*, and *Plumatella* statoblasts (whole lake) over
795 the fieldwork period. *Daphnia* and flotsam $\delta^{13}C$ values represent single measurements
796 except for 25 February 2014 and 27 March 2014 when enough material was available for
797 three measurements of *Daphnia* $\delta^{13}C$. In those cases average values are presented

798

799 **Fig. 4** Age-depth model (black line, triangles) for core GER12 based on ^{210}Pb activity. The
800 grey line shows the accumulation of ^{137}Cs activity in the record (based on the ^{210}Pb -inferred
801 accumulation rate), with the diamond showing the location of the lower activity peak of ^{137}Cs
802 in the ^{210}Pb -based age model. This peak is correlated with atmospheric nuclear bomb testing
803 in AD 1963 and coincides with the only measurable ^{241}Am activity in the core (1.9 ± 0.2 Bq/kg).

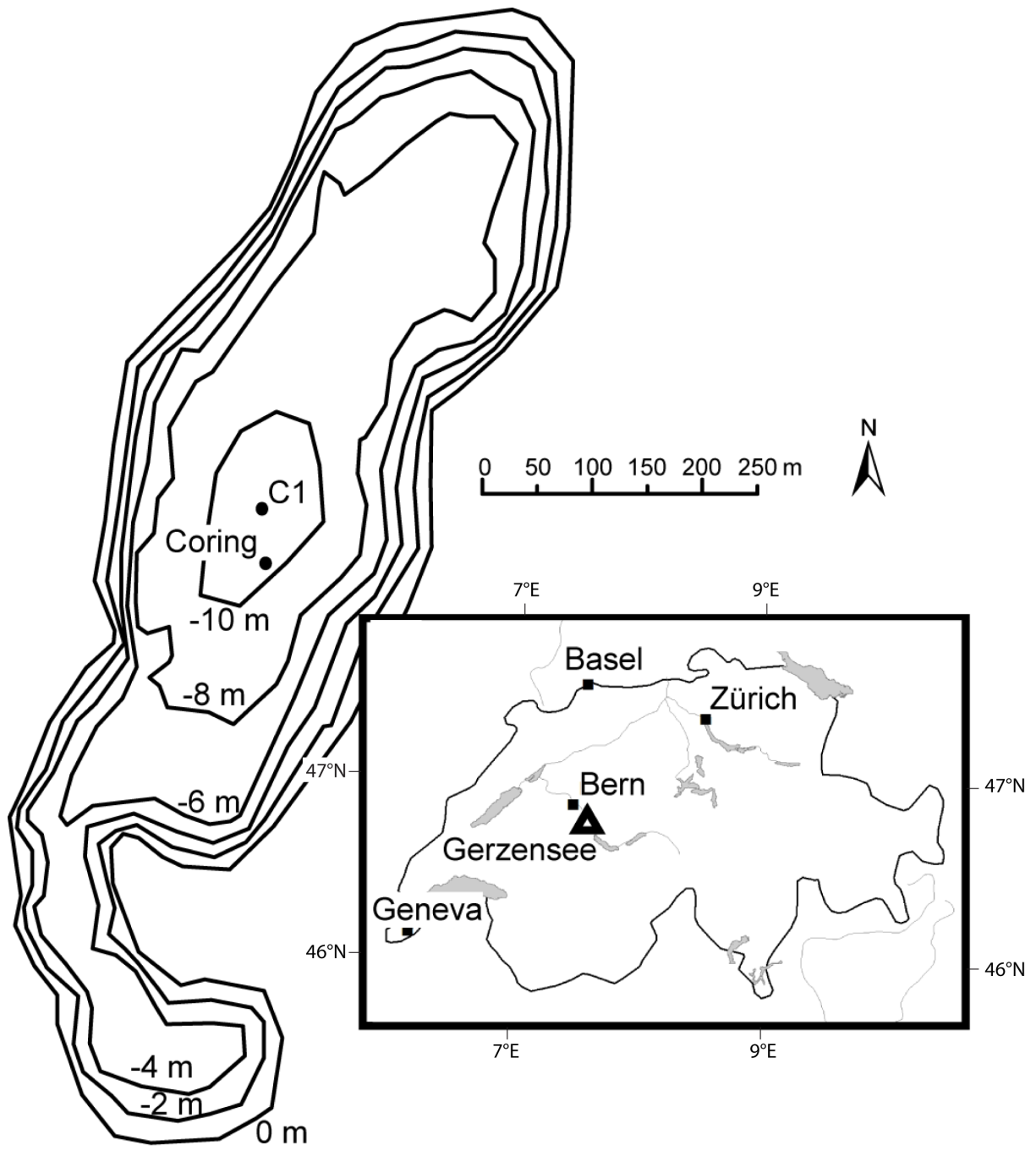
804

805 **Fig. 5** $\delta^{13}C$ values of fossil invertebrate remains of *Daphnia*, *Ceriodaphnia* and *Plumatella* in
806 sediment core GER12 and from the flotsam of the lake (flotsam values plotted above 0 cm
807 sediment core depth). Symbols in the downcore record represent the average depth of the
808 measurements in cases for which samples were pooled to obtain sufficient weight for $\delta^{13}C$

809 analysis. Measurements may represent remains from up to 6 cm of sediment (see text for
810 details.

811 **Figures**

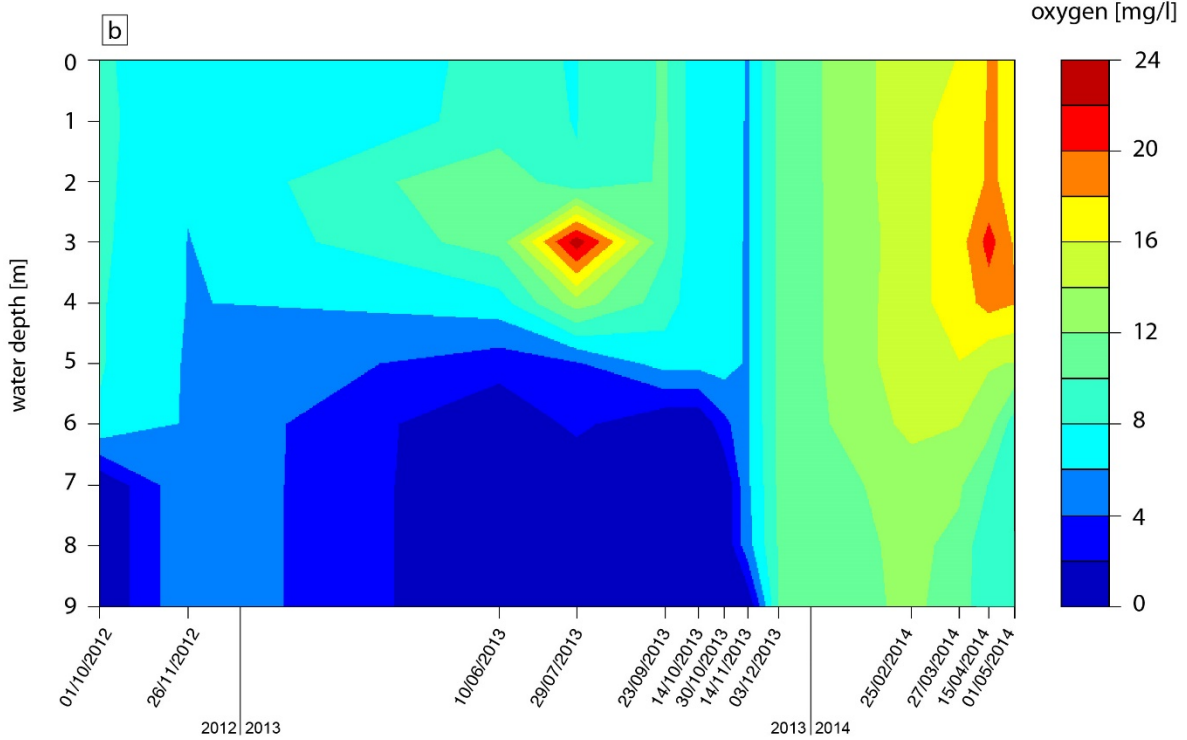
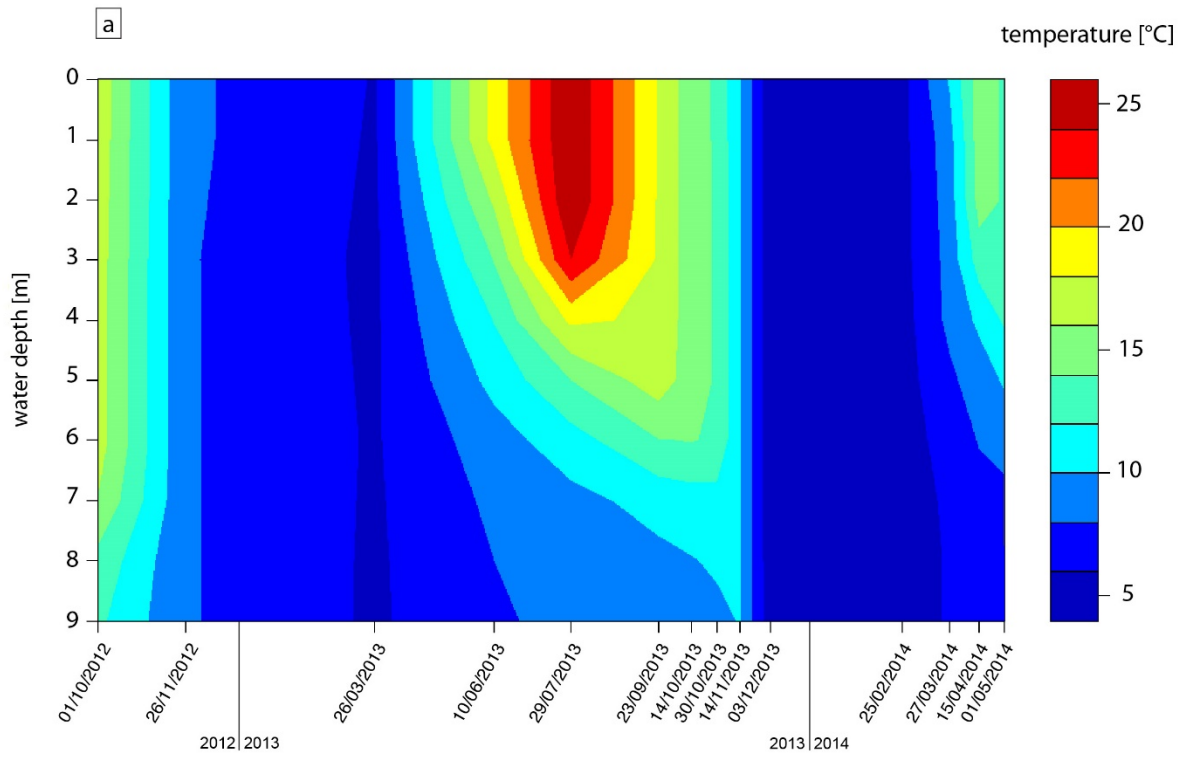
812 Figure 1



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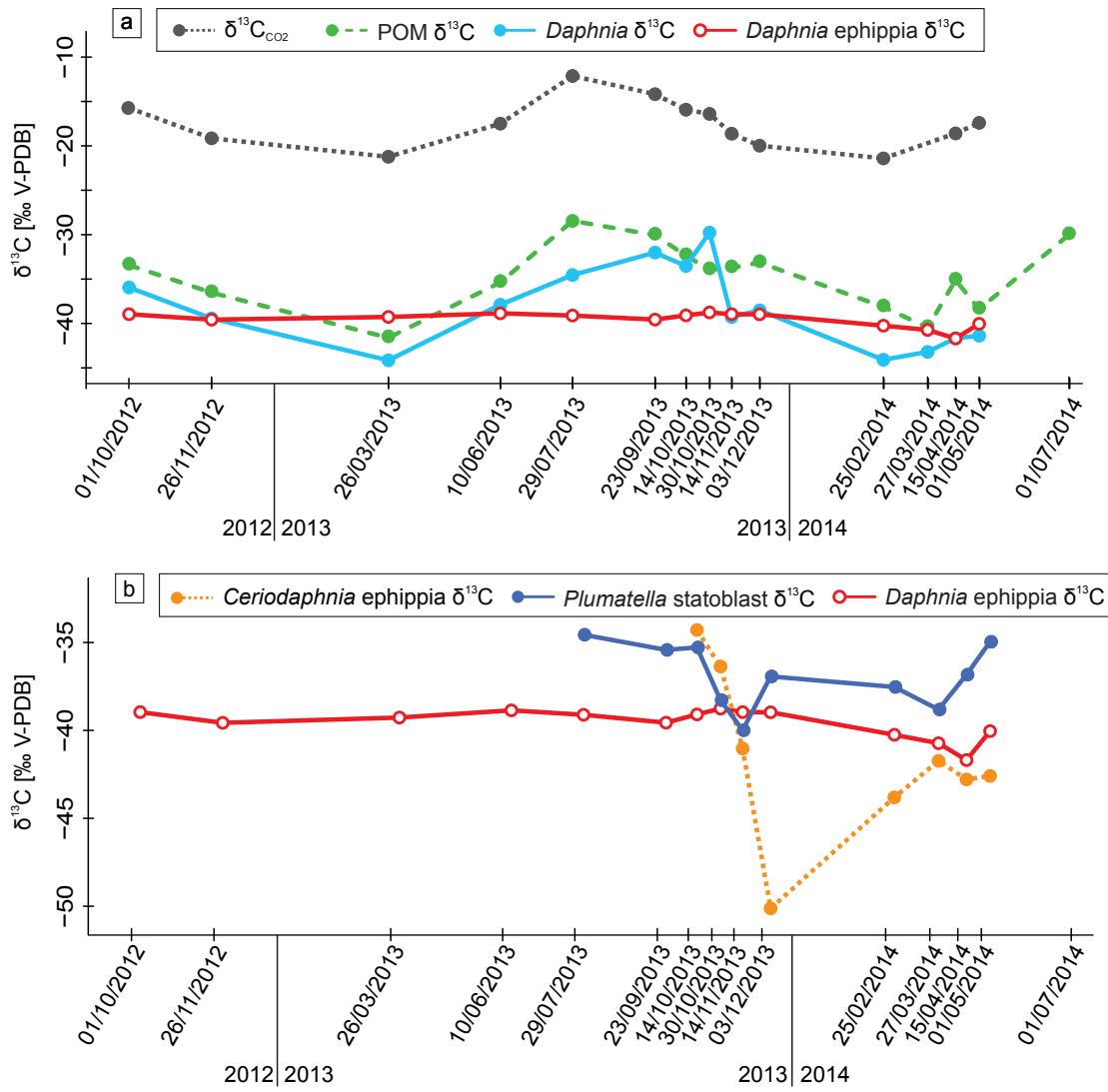
815 Figure 2



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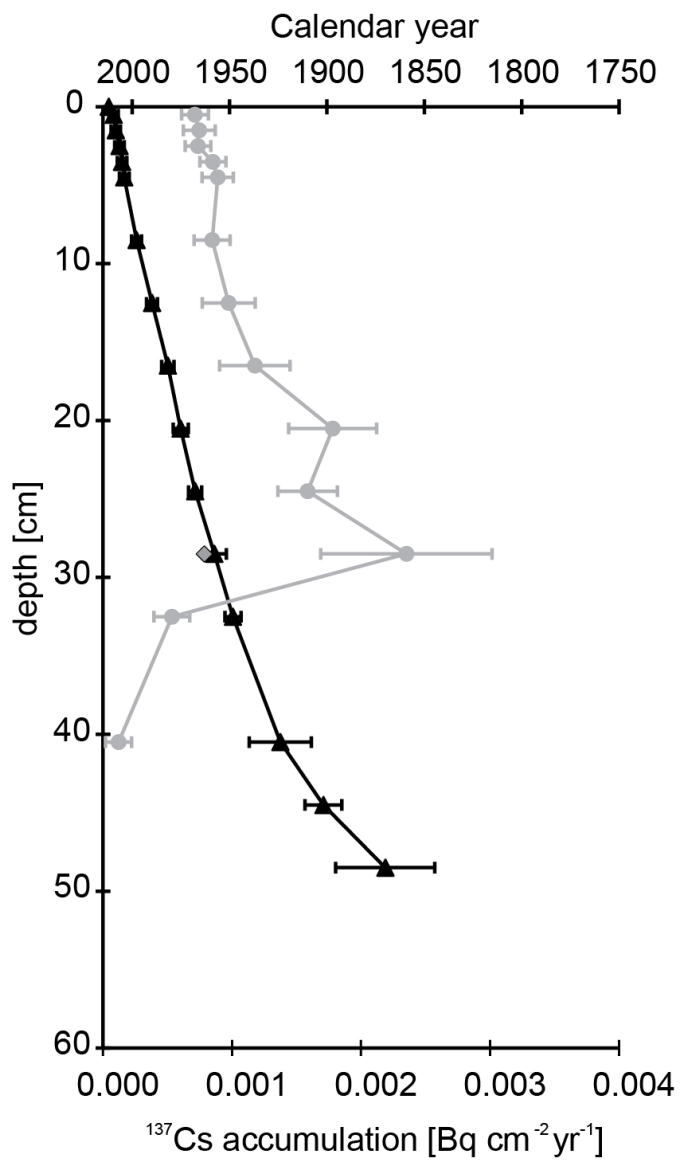
818 Figure 3



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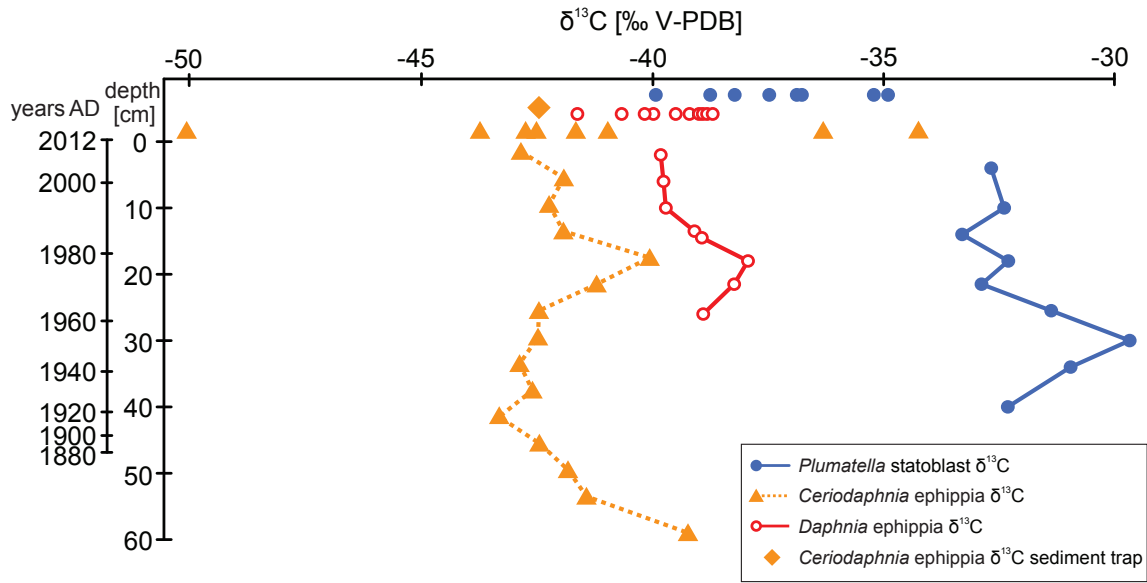
821 Figure 4



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824 Figure 5



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826

827 **Electronic Supplementary Material**

828

829 **Figure Captions**

830

831 **Fig. S1**

832 Dissolved CO₂ ([CO₂]_{aq}) and CH₄ ([CH₄]_{aq}) concentrations for surface (0.7 m water depth)
833 and bottom waters (8 m water depth) at the lake centre (location C1) over the fieldwork
834 period in 2012-2014

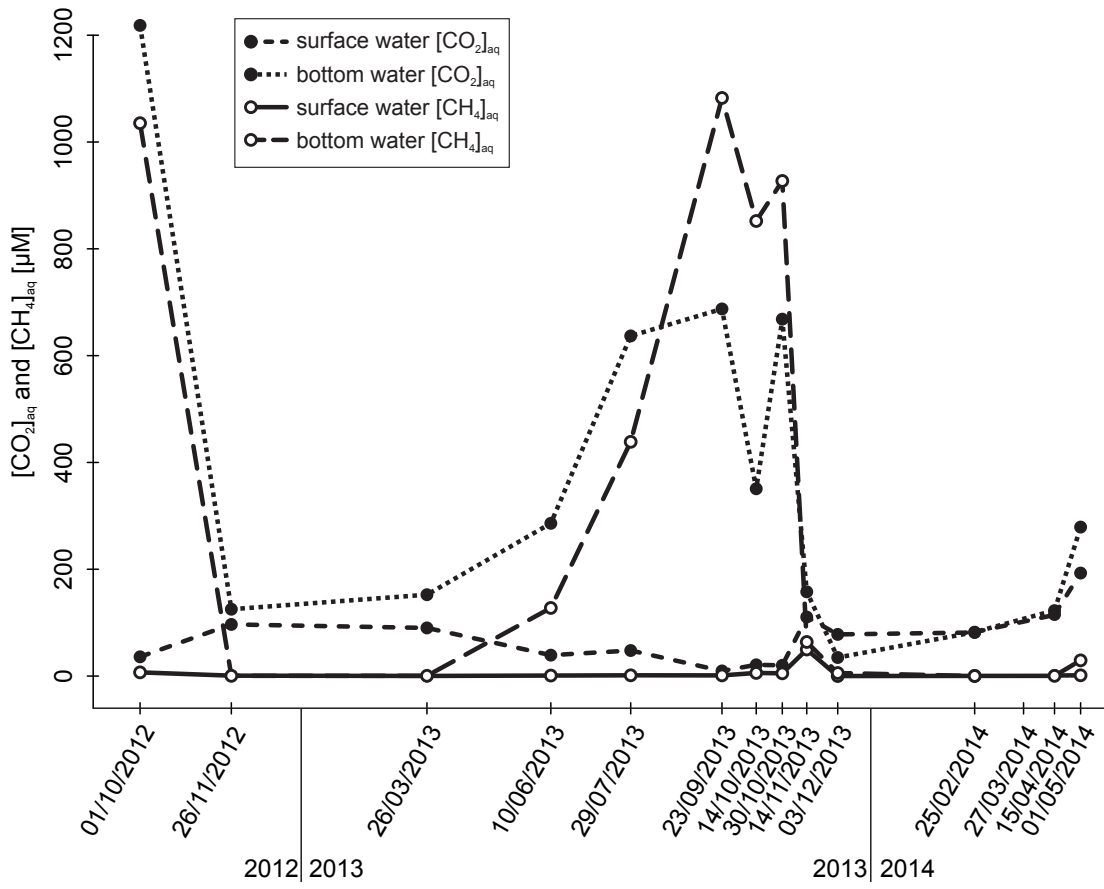
835

836 **Fig. S2**

837 (a) Relationship between surface water chlorophyll *a* concentrations (0.7 m water depth) and
838 *Daphnia* δ¹³C at the lake centre (location C1); (b) relationship between log-transformed
839 bottom water dissolved CH₄ concentrations ([CH₄]_{aq}) (8 m water depth) and *Daphnia* δ¹³C at
840 the lake centre (location C1); (c) relationship of the difference between particulate organic
841 matter (POM) δ¹³C and *Daphnia* δ¹³C in the surface water (Δ¹³C_{POM-Daph}) and bottom water (8
842 m water depth) [CH₄]_{aq}; (d) surface (0.7 m water depth) and bottom water (8 m water depth)
843 POM δ¹³C at the lake centre (location C1)

844

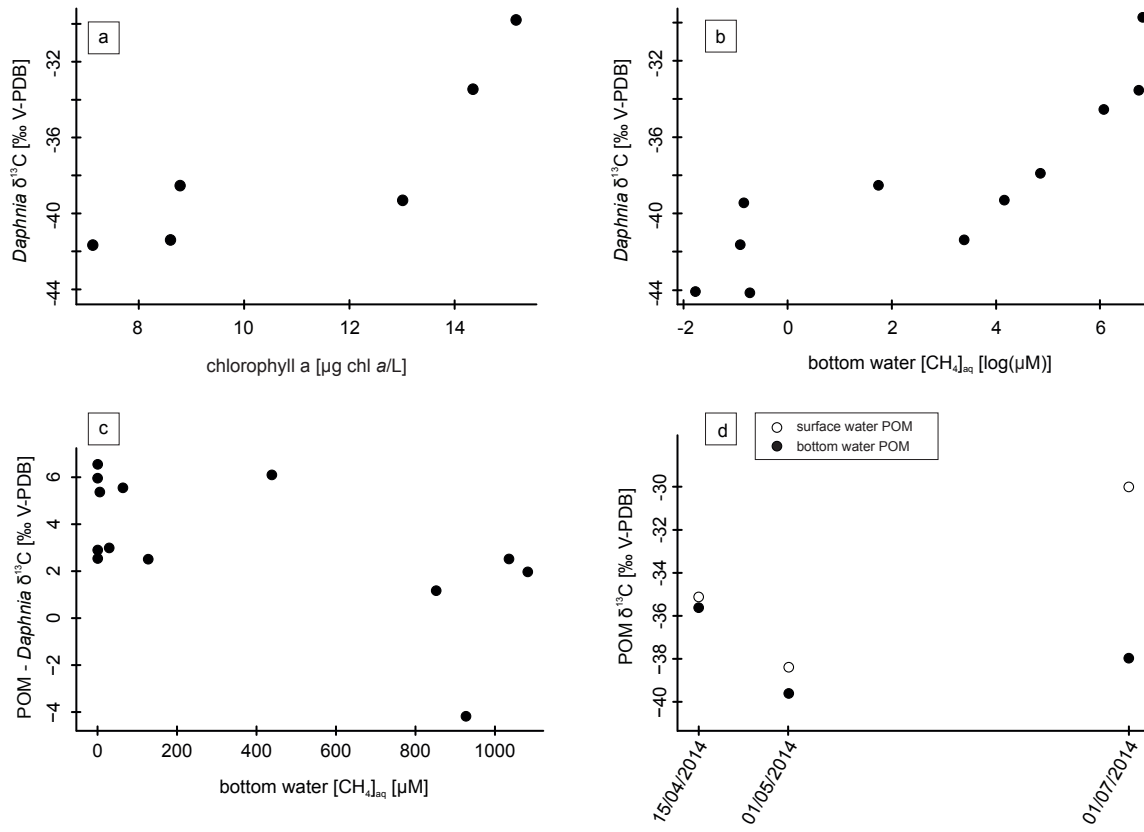
845 Figure S1



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847

848 Figure S2



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850

851 **Table S1** Samples collected and analysed in the course of repeated fieldwork on Gerzensee between October 2012 and July 2014. “x” denotes
 852 that a sample was collected, “-“ shows where no sample was collected

853

Fieldwork dates [d.m.yr]	Temp	[O ₂] _{aq}	<i>Daphnia</i> δ ¹³ C	<i>Daphnia</i> ephippia δ ¹³ C	<i>Ceriodaphnia</i> ephippia δ ¹³ C	<i>Plumatella</i> statoblast δ ¹³ C	DIC	δ ¹³ C _{POM} (0.7 m)	δ ¹³ C _{POM} (7 m)	Chl a [µg/l]	[CH ₄] & [CO ₂]
01.-03.10.2012	x	x	x	x	-	-	x	x	-	-	x
26.-28.11.2012	x	x	x	x	-	-	x	x	-	-	x
26.-28.03.2013	x	x	x	x	-	-	x	x	-	-	x
10.-13.06.2013	x	x	x	x	-	-	x	x	-	-	x
29.-31.07.2013	x	x	x	x	-	x	x	x	-	-	x
23.-25.09.2013	x	x	x	x	x	x	x	x	-	-	x
14.10.2013	x	x	x	x	x	x	x	x	-	x	x
30.10.2013	x	x	x	x	x	x	x	x	-	x	x
14.11.2013	x	x	x	x	x	x	x	x	-	x	x
03.12.2013	x	x	x	x	x	x	x	x	-	x	x
25.02.2014	x	x	x	x	x	x	x	x	-	-	x
26.03.2014	x	-	-	x	x	x	-	-	-	-	x
15.04.2014	x	x	x	x	x	x	x	x	x	x	x
01.05.2014	x	x	x	x	x	x	x	x	x	x	x
01.07.2014	-	-	-	-	-	-	-	x	x	x	x

854