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1	Leaf traits and photosynthetic responses of Betula pendula saplings to a range of
2	ground-level ozone concentrations at a range of nitrogen loads
3	
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15	
16	Summary
17	Ground-level ozone (O ₃) concentrations and atmospheric nitrogen (N) deposition rates have
18	increased strongly since the 1950s. Rising ground-level O3 concentrations and atmospheric N
19	deposition both affect plant physiology and growth, however, impacts have often been
20	studied in isolation rather than in combination. In addition, studies are often limited to a
21	control treatment and one or two elevated levels of ozone and/or nitrogen supply. In the

- 22 current study, three-year old *Betula pendula* saplings were exposed to seven different O_3
- profiles (24 hour mean O_3 concentration of 36-68 ppb in 2013, with peaks up to an average
- of 105 ppb) in precision-controlled hemispherical glasshouses (solardomes) and four different
- N loads (10, 30, 50 or 70 kg N ha⁻¹ y⁻¹) in 2012 and 2013. Here we report on the effects of

26	enhanced O ₃ concentrations and N load on leaf traits and gas exchange in leaves of varying
27	age and developmental stage in 2013. The response of leaf traits to O_3 (but not N) vary with
28	leaf developmental stage. For example, elevated O ₃ did not affect the chlorophyll content of
29	the youngest fully expanded leaf, but it reduced the chlorophyll content and photosynthetic
30	parameters in aging leaves, relatively more so later than earlier in the growing season.
31	Elevated O ₃ enhanced the N content of senesced leaves prior to leaf fall, potentially affecting
32	subsequent N cycling in the soil. Enhanced N generally stimulated the chlorophyll content
33	and photosynthetic capacity. Whilst elevated O ₃ reduced the light-saturated rate of
34	photosynthesis (A_{sat}) in aging leaves, it did not affect stomatal conductance (g_s) . This
35	suggests that photosynthesis and g_s are not closely coupled at elevated O_3 under-light
36	saturating conditions. We did not observe any interactions between O_3 and N regarding
37	photosynthetic parameters ($V_{c,max}$, J_{max} , A_{sat}), chlorophyll content, g_s , N content in senesced
38	leaves and leaf number. Hence, the sensitivity of these leaf traits to O_3 in young silver birch
39	trees is neither reduced nor enhanced by N load.
40	
41	Keywords: Air pollution; chlorophyll content; leaf age; nitrogen content; photosynthetic
42	capacity; stomatal conductance.
43	

44 Abbreviations

45 $A_{sat} = light-saturated rate of photosynthesis at ambient CO₂ (390 ppm)$

46 A/Ci curve = plot of net photosynthesis (A) against leaf internal CO_2 concentration (Ci)

47 AICc = Akaike Information Criterion, corrected for small sample size

48 $g_s = stomatal conductance$

49 ID = identity

50 $J_{max} = maximum$ rate of electron transport

51 LMM = linear mixed model

52 ppb = parts per billion

53 $V_{c,max}$ = maximum rate of carboxylation

54

55 Introduction

Tropospheric ozone (O₃) concentrations have been increasing from a background of ca. 10-15 56 ppb in the early 20^{th} century, due to increased emissions of O₃ precursors from anthropogenic 57 sources (Cooper et al., 2014; Royal Society, 2008; Solberg et al. 2005; Volz and Kley, 1988). 58 In Europe, background concentrations have roughly doubled between 1950 – 2000, followed 59 60 by a decade with no further rise or even a reduction at some sites (Cooper et al., 2014; Parrish et al., 2012). Present day annual average background concentrations are levelling off at ca. 30 61 - 40 ppb in Europe and North America, (Cooper et al., 2014; Royal Society, 2008), but are 62 still rising significantly in East Asia (Cooper et al., 2014). In addition, atmospheric 63 concentration and deposition of reactive nitrogen (N) has increased strongly since the 1950s 64 with the application of the Haber-Bosch process to produce N fertilizers on an industrial scale 65 (Sutton et al., 2011). Despite the rise in both ground-level O₃ concentrations and atmospheric 66 N deposition since the 1950s, impacts on vegetation have often been studied in isolation 67 rather than in a combination of both factors. The two pollutants are closely interlinked as a 68 major pathway for the formation of tropospheric O₃ is the photochemical reaction with N 69 oxides (Royal Society, 2008). 70

71 Generally, O₃ tends to have a negative impact on sensitive plant species, resulting in a decline 72 in photosynthesis, biomass production or crop yield, and accelerated senescence (Ainsworth et al., 2012). A meta-analysis by Wittig et al. (2007) showed that the increase in O₃ since the 73 74 industrial revolution has been responsible for a reduction in photosynthesis of approximately 11% in trees, which reduced tree productivity by approximately 7% (Wittig et al. 2009). 75 Further reductions were observed with a rise in O₃ concentration above the current ambient 76 level (Wittig et al., 2007, 2009). Partly due to the successful implementation of air pollution 77 abatement policies in Europe, reductions in O₃ precursor emissions have resulted in a decline 78 79 in peak O₃ concentrations in the summer in recent years at some sites across Europe (Simpson et al., 2014; Torseth et al., 2012). However, mean annual O₃ concentrations have 80 hardly changed in recent decades (EMEP, 2016) due to a rise in the frequency of low-range 81 82 O₃ concentrations (Simpson et al., 2014). Background O₃ concentrations are currently at levels where they chronically affect sensitive plant species (Mills et al., 2011a). 83 In contrast, N tends to have a fertilising effect on plant growth and photosynthesis in areas 84 85 where N is a limiting nutrient. However, above the critical load or level, the impact of N on sensitive species will become negative in most species (Bobbink and Hettelingh, 2011; Cape 86 et al., 2009). For example, forest observations in Switzerland suggested an increase in 87 phosphorus limitation with increasing N deposition, resulting in a decline in growth in Fagus 88 89 sylvatica (Braun et al., 2010). N pollution is now a major threat to terrestrial biodiversity 90 (Bobbink et al., 2010; Dise et al., 2011). Relatively few studies have investigated the combined impacts of both O₃ and N on 91 92 vegetation at realistic ozone concentrations (Mills et al., 2016). Evidence suggests that O₃ and 93 N can have both synergistic and antagonistic effects on tree species. Pääkkönen and Holopainen (1995) and Landolt et al. (1997) showed that high N supply ameliorates the 94 negative impact of O₃ in *Betula pendula*. Similar responses were reported for *Larix kaempferi* 95

96 (Watanabe et al., 2006), Populus tremula x Populus tremuloides (Häikiö et al., 2007) and *Ouercus kelloggii* (Handley and Grulke, 2008). However, enhanced N supply increased 97 growth sensitivity to O₃ of *Pinus sylvestris* (Utriainen and Holopainen, 2001b) and *Fagus* 98 99 crenata (Wanatabe et al., 2012; Yamaguchi et al., 2007) seedlings. The effect of N supply on the sensitivity to O₃ was not significant for Picea abies (Thomas et al., 2005; Utiainen and 100 Holopainen, 2001a) or Quercus robur seedlings (Marzuoli et al., 2016). Complex interactions 101 102 were observed between O₃ and N impacts on growth of young trees of *Fagus sylvatica* (Thomas et al., 2006). In a recent review, Mills et al. (2016) showed that the beneficial effect 103 104 of N on root development was lost at high O₃ exposure whilst the effects of increasing O₃ on root biomass became more pronounced as N supply increased. 105 106 Many previous studies investigating the interactive impacts of O₃ and N on trees have been 107 limited to two to three O₃ and N exposure levels. In the current study, three year old silver 108 birch (Betula pendula) seedlings were exposed to seven O₃ regimes (24 h mean of 36 – 68 ppb) and four levels of N load $(10 - 70 \text{ kg N ha}^{-1} \text{ y}^{-1})$ for two years (2012 and 2013). Silver 109 birch is known to be sensitive to O₃ (LRTAP Convention, 2015; Mills et al., 2011b) and has a 110 widespread distribution in Europe, from northern Finland to Italy and from Ireland to 111 Kazakhstan. We investigated whether O₃ sensitivity of leaf traits was affected by N supply. 112 We report in detail on the impacts of O₃ and N on chlorophyll content, stomatal conductance 113 (g_s) , leaf photosynthetic traits, N concentration in senesced leaves and leaf number during the 114 115 second year of exposure. Based on the cumulative nature of O_3 impacts on plants (LRTAP) Convention, 2015), we hypothesised that negative effects of O_3 on leaf traits will increase 116 with time of exposure, with older leaves being more affected than younger leaves. Based on 117 118 previously reported results (Landolt et al., 1997; Pääkkönen and Holopainen, 1995) for silver birch, we hypothesised that enhanced N will reduce negative effects of O₃. 119

122 Material and methods

123 Plant material, experimental site and treatments

On 13th April 2012, three-year old trees of silver birch (*Betula pendula*, UK provenance) 124 were planted in 6.5 litre pots (diameter 21 cm, height 25 cm) filled with John Innes No. 1 125 compost (J. Arthur Bowers). Mean height of the trees when planted was 68 cm and they had 126 not branched yet. All plants were inoculated with mycorrhiza (obtained from Buckingham 127 Nurseries, UK). The pots were randomly distributed between seven hemispherical 128 129 glasshouses (solardomes; 3 m diameter, 2.1 m height) at Abergwyngregyn nr. Bangor, North Wales, United Kingdom. After an acclimation period in the solardomes, treatments were 130 started on 30th April 2012. Plants were exposed to O₃ based on an episodic profile recorded at 131 132 a rural O₃ monitoring site during July 2006 (Aston Hill, Wales, UK, 52°50'N, 3°03'W). The Aston Hill profile was applied as the highest treatment, and for the other treatments the O₃ 133 concentration was reduced by 10 ppb during the peaks and 3 ppb during periods with 134 background O₃ (Figure 1; Hewitt et al., 2016). The profiles cover a range of concentrations 135 found in Europe and this is the only facility that has sufficiently sophisticated control of O₃ to 136 be able to deliver O₃ treatments that increase in small increments. The 24 hour mean 137 concentration was between 36 and 68 ppb in 2013, with peaks up to on average 105 ppb 138 (Figure 1). Plants were exposed to O₃ during the growing season, starting on 30th April and 139 finishing on 12th October in 2012, and starting on 17th May and finishing on 16th September 140 in 2013, when the trees were harvested. At harvest, the roots didn't show any signs of being 141 pot-bound. Between 17th October 2012 and 14th May 2013, the trees were kept outside under 142 ambient conditions. The solardomes were ventilated at a rate of two air changes per minute 143 and charcoal-filtered air was injected with controlled amounts of O₃. O₃ was provided by a 144 G11 O₃ generator (Ozone Industries, UK) equipped with oxygen concentrator (Dryden Aqua, 145

146 UK). Concentrations were determined by a computer-controlled O₃ injection system (Lab VIEW version 8.6, National Instruments, Texas, US). O3 was distributed to each solardome 147 via polytetrafluoroethylene (PTFE) tubing, with the concentration inside each solardome 148 measured for 5 min every 30 minutes using two O₃ analyzers (400a, Enviro Technology 149 Services, Stroud, UK) of matched calibration. Four N treatments were applied weekly in each 150 solardome as ammonium nitrate, in 200 ml water containing other macronutrients (P, K, S, 151 Ca, Mg) and various micronutrients (Fe, Mn, Mo, Cu, Zn, B). N addition rates were 10, 30, 152 50 and 70 kg N ha⁻¹ yr⁻¹. N and nutrient additions were started at the same time as the O_3 153 154 exposure (see above) and continued for the duration of the experiment, including in the winter. Plants were watered twice a week or as required to maintain soil moisture content 155 near field capacity. Each dome contained 24 pots, i.e. six pots per N supply. The O₃ profiles 156 157 were assigned randomly to the solardomes. The weekly O₃ profiles for 2012 were described by Hewitt et al. (2014), the weekly O₃ profiles for 2013 were very similar and are shown in 158 Figure 1. A summary of O₃ treatments and climate conditions for 2013 is provided in Table 1. 159 In one solardome, ambient air temperature, photosynthetically active radiation (PAR), 160 temperature and relative humidity were continuously monitored by an automatic weather 161 station (Skye Instruments Ltd, Llandridod Wells, UK) and soil moisture content was 162 continuously monitored using Theta Probes (Delta-T Devices Ltd, Cambridge, UK). 163 164 Although O_3 treatments were not replicated, numerous previous studies (Hayes et al., 2012; 165 Hewitt et al., 2014, 2016; Mills et al., 2009) have established the statistical validity of conducting un-replicated experiments using the solardome facility, and no solardome effects 166 were detectable in independent measurements of air or leaf temperature (Hewitt et al., 2016). 167 168

169 Chlorophyll content and stomatal conductance (g_s) measurements

170 Chlorophyll content (determined non-destructively using CCM200, ADC, UK, and expressed as an index) and g_s (AP4 Porometer, Delta T, UK) measurements were conducted at about 171 four-weekly intervals on the third leaf (youngest fully expanded leaf) from the top of the 172 canopy throughout the growing season in five O_3 treatments. The maximum g_s for each 173 treatment was determined as the 95-percentile. Silver birch is heterophyllous (Clausen and 174 Kozlowski, 1965), i.e. has early and late leaves which started to flush in April and July 175 respectively in our study. In addition, the chlorophyll content was determined on leaves used 176 for leaf gas exchange measurements (including g_s), i.e. a designated leaf throughout the 177 178 growing season (see below), and determined near the end of the growing season on the eight leaves on the highest branch of trees (six replicates per treatment). 179

180

181 *Leaf gas exchange measurements*

182 Youngest fully expanded leaves were identified on 21 May 2013. Gas exchange

measurements were conducted on these designated leaves exposed to the lowest (24 hr mean 183 = 36 ppb) and highest O_3 concentration (24 hr mean = 68 ppb) and supplied with the lowest 184 (10 kg N ha⁻¹ yr⁻¹) and highest N rate (70 kg N ha⁻¹ yr⁻¹). Measurements on the designated 185 leaves were conducted at 6-weekly intervals during the following periods: 10 - 14 June, 22 -186 25 July and 2 - 5 September 2013. To determine photosynthetic capacity of birch leaves, the 187 response curve of the net photosynthetic rate (A) to the intercellular CO₂ concentration (Ci), 188 189 i.e. the A/Ci curve, was constructed using a portable gas analyser with PLC6 universal leaf cuvette (Ciras-2, PPsystems). The measurements were conducted at light saturation 190 (photosynthetic photon flux density = $1500 \mu mol m^{-2} s^{-1}$; LED light source), a leaf 191 temperature of 22 °C (determined by built-in infrared temperature sensor), and a leaf-to-air 192 vapour pressure deficit of 0.6 - 1.1 kPa. The A/C_i curve was constructed by measuring A at 193 11 air CO₂ concentrations between ca. 50 and 1900 µmol mol⁻¹. A_{sat} and associated g_s were 194

determined at ambient CO₂ (390 µmol mol⁻¹). Leaves were inserted into a broad leaf chamber 195 $(area = 2.5 \text{ cm}^2)$ and allowed to stabilise for 15 min at growth CO₂. At each CO₂ 196 concentration, gas exchange parameters were recorded after readings had stabilised (ca. 2-3197 198 min). The maximum rate of carboxylation (V_{c.max}) and the maximum rate of electron transport (J_{max}) were calculated from the A/Ci curve according to the methodology described 199 by Sharkey et al. (2007). V_{c,max} and J_{max} were calculated for a leaf temperature of 25 °C, using 200 the measurement leaf temperature and atmospheric pressure as input parameters. The ratio 201 between J_{max} and V_{c.max} was also determined. In addition to the A_{sat} measurements from the 202 203 A/Ci curve, A_{sat} was also measured in late August/early September at all six O₃ concentrations at low and high N supply (4 - 5 replicates per treatment). 204 205 206 Leaf number and N content of leaves 207 Leaf number was counted on four days during the 2013 growing season: 14 May, 31 July, 2 and 24 September (day before harvest). Throughout September, leaves that were ready to 208 209 drop off were collected for N analysis. The N content of mature, green leaves was also determined in the middle of July. The leaves were dried at 70 °C, ground with a (Cyclotec) 210 and 10 mg was weighed in tin capsules for analysis. Leaves were analysed for N content at 211 the Chemical Analysis Laboratory, Forest Research, UK. The total N content was determined 212 by dry combustion (ISO 13878, using Carbo Erba CN analyser, Flash1112 series). 213

214

215 *Statistical analyses*

The precise O_3 control system used in the solardomes allowed relatively small changes in O_3 profile to be simulated, facilitating dose-response analyses rather than just comparison of the impacts at low and high ozone exposure. The large number of O_3 treatments allowed for effects of N on the slope of dose-response functions to be determined. We note that the lack 220 of treatment replication may raise concerns, however, we believe that the benefit of using more treatments outweights this limitation, as published previously (Hayes et al., 2012; 221 Hewitt et al., 2014, 2016; Mills et al., 2009). Air flow rates were matched between 222 223 solardomes, and recorded climatic conditions did not vary significantly from solardome to solardome (Hewitt et al., 2016). For all linear mixed models (LMMs; normal error) applied 224 below, a model set was created using the package lme4, v1.1-7 (Bates et al., 2015) in R 225 226 version 3.2.3 (R core team, 2014), sequentially removing interactions and predictor variables from the global model, and model selection was carried out by examining the change in 227 228 Akaike Information Criterion, corrected for small sample size (AICc). The model with the lowest AICc value is optimal, with models differing in 2 - 7 AICc units from the best model 229 having little empirical support (Burnham and Anderson, 2002). Bayesian Information 230 231 Criterion (BIC) values, which penalise the number of model parameters more strictly, were also examined. For the optimal model selected, p-values were obtained for each term in the 232 model using the R package lmerTest, v2.0-20 (Kuznetsova et al., 2014). For all models, 233 statistical assumptions (normality and even spread of residuals) were checked using residual 234 plots. Response variables were transformed where necessary (log or square root). 235

236

237 <u>Chlorophyll content and stomatal conductance (g_s) </u>

238 To investigate factors influencing chlorophyll content of the youngest fully expanded leaf, the

LMM included added N, 24 hour mean O_3 concentration (2013, n = 5 levels) and

240 measurement period as fixed, continuous variables and the random effects of dome and plant

identity (ID). As there was a suggestion of a quadratic relationship between the chlorophyll

content and time, a quadratic time term was included. For eight leaves on the top branch (near

the end of the growing season), a further LMM with chlorophyll content as the response

variable, the continuous fixed effects of 24 hour mean O_3 (2013, n = 4) and leaf number, N

(factor: high and low), and a random effect of plant ID was also run. A quadratic O_3 term was included to allow for the suggested non-linear relationship between chlorophyll content and O_3 . The LMM for g_s included added N, 24 hour mean O_3 concentration (2013, n = 5 levels) and measurement period as fixed, continuous variables. A quadratic term for measurement period and random effects of dome and plant ID were also included. Extreme outliers in the g_s data (thought to be due to measurements from damp leaves) were removed using the median absolute deviation (mad) method (using a rejection criteria of median ± 5.2 mad).

253 <u>Leaf gas exchange measurements</u>

For each parameter (e.g. V_{c,max}, J_{max}) of the designated leaf, a LMM including the fixed 254 effects of time, N and O₃ (as factors, low and high), and the random effect of plant ID, was 255 256 applied. The relationship between Asat and chlorophyll content and Asat and gs was tested using LMMs, including the fixed effects of O₃ and N and the random effect of plant ID. The 257 A_{sat} measurements taken in late August were analysed using a LMM with a continuous fixed 258 effect for 24 hour mean O_3 (2013, n = 6 levels), N as a factor (high and low), and a random 259 effect of dome. Two further Asat model sets were also run, with chlorophyll content and gs 260 included as continuous fixed predictors. 261

262

263 <u>N content of leaves and leaf number</u>

The effect of O₃ and added N on the N content of mature (mid July) and senesced leaves
(September) was analysed using a general linear model (glm) containing an interaction
between the continuous variables of added N and 24 hour mean O₃ concentration (2013, n = 7
levels). As the variance explained by the random effect of dome was zero, a mixed effects
model was not used. To investigate changes in leaf number, a generalised linear mixed effect
model (GLMM), with leaf count as the response variable and a Poisson error distribution,

was run using the glmmPQL function in package MASS (Venables and Ripley, 2002) to deal
with over-dispersion in the data. Fixed effects included 24 hour mean O₃ concentration
(2012-2013, n = 7 levels) and added N as continuous variables and month as a factor, while
dome and plant ID were included as random effects. As the glmmPQL function does not
provide AIC values, p-values and estimated standard errors were used to determine which
fixed effects were needed in the model.

276

277

278 **Results**

279 Chlorophyll content and gs of the youngest fully expanded leaf

The chlorophyll content of the third leaf (youngest fully expanded leaf) declined in the first 280 281 part of the growing season but then increased again between 23 July and 23 August due to new growth of late leaves (p < 0.001 for 'Time'; Figure 2, Table 2). The third leaf developed 282 from new growth (late leaf) had a similar chlorophyll content as the third leaf near the start of 283 284 the season (early leaf). Remobilisation of N within the trees is likely to have started as older leaves began to drop off from the middle of July onwards as new leaves developed. 285 Throughout the season, the chlorophyll content was higher at 70 than at 10 kg N ha⁻¹ y⁻¹ (p < 286 0.001; Figure 2, Table 2), with the difference being lowest at the end of July and highest at 287 288 the end of August. O₃ concentration did not affect the chlorophyll content of the youngest 289 fully expanded leaf. While visual examination of the raw data suggested that stimulation of the chlorophyll content at high N supply was diminished at the highest O₃ exposure at the end 290 of May, two or three-way interactions were not observed (Table 2). The g_s changed with time 291 292 and was highest at the end of August (late leaves) compared to earlier in the year (early leaves); gs was not affected by O3 or N treatment and there were no significant two or three-293 way interactions (Figure 3; Table 2). The maximum gs reached its highest value at 50 kg N 294

295 ha⁻¹ y⁻¹, and was 256, 307, 339 and 337 mmol H₂O m⁻² s⁻¹ at 10, 30, 50 and 70 kg N ha⁻¹ y⁻¹ 296 respectively.

297

298 Chlorophyll content of leaves along the first top branch near the end of the growing season To assess the response to treatment of the chlorophyll content of leaves of different 299 developmental stages, the chlorophyll content along a branch was measured late in the season 300 (4th September) when a wide range of chlorophyll content was present. The chlorophyll 301 content was higher at 70 kg N ha⁻¹ y⁻¹ than at 10 kg N ha⁻¹ y⁻¹ (p < 0.001) and decreased with 302 303 increasing leaf age (i.e. leaf number; p < 0.001; Figure 4, Tables 3 and S1). The effect of O₃ was dependent on leaf age, with young leaves showing stimulation in the chlorophyll content 304 (although the highest chlorophyll content was already reached in the youngest leaf at 39 ppb 305 306 O₃) and medium-aged leaves showing no clear O₃ effect. In the oldest leaves (leaf 8), the chlorophyll content was highest at 36 ppb O₃ and 70 kg N ha⁻¹ y⁻¹ and lowest at 68 ppb O₃ 307 and 10 kg N ha⁻¹ y⁻¹. At 68 ppb O₃ the chlorophyll content declined faster with increasing leaf 308 309 age than at 36 ppb O_3 .

310

311 Photosynthetic capacity and other leaf traits of a leaf aging throughout the season

The photosynthetic capacity (V_{c,max} and J_{max}) of an aging leaf was reduced by elevated O₃ and 312 stimulated by high N supply, with no significant interaction being observed between O₃ and 313 N (Figure 5, Table 4 and S2). Whilst elevated O₃ increased the J_{max}/V_{c,max} ratio (ca. 7% on 314 average), enhanced N supply reduced the $J_{max}/V_{c,max}$ ratio (ca. 8% on average; Table S2). 315 Although the photosynthetic capacity declined with time, this was primarily due to a 316 317 significant decline between June and July, with no further decline being observed between July and September. It should be noted that some of the designated leaves had already 318 dropped off the trees in September, hence their photosynthetic capacity could not be assessed 319

320 in September. In addition, variation for all measured parameters was highest near the end of the growing season. The negative effect of O₃ on photosynthetic capacity increased with time 321 (Table S2), resulting in an O₃ x time interaction for J_{max} (for $V_{c,max} p = 0.08$ for the interaction 322 323 term in the model). The response of light-saturated photosynthesis (A_{sat}) to treatment was similar to that of J_{max} , except that the response to N was not significant (p= 0.12). The 324 response of the chlorophyll content to treatment was also similar to J_{max} (a measure of the 325 maximum rate of electron transport in the photosystems). The gs of leaves under ambient CO₂ 326 and light saturation was not affected by time, O_3 exposure or N supply. There was a linear 327 328 relationship between A_{sat} and chlorophyll content (p < 0.001; Figure 6a), indicating that nondestructive measurement of the chlorophyll content provides a good estimate of A_{sat}. This 329 relationship was not affected by O3 and N treatment. There was also a linear relationship 330 331 between photosynthetic capacity ($V_{c,max}$ and J_{max}) and chlorophyll content (p < 0.001; Figure S1). There was a weak relationship (p = 0.038) between A_{sat} and g_s, with a lot of scatter in the 332 data, and the relationship was not affected by O₃ or N treatment (Figures 6b, S2). At the end 333 of August, A_{sat} was also measured at additional O₃ treatments at an N supply of 10 and 70 kg 334 ha⁻¹ y⁻¹. These measurements showed a linear decline in A_{sat} with O_3 exposure later in the 335 season (p < 0.001; Figure 7). Enhanced N supply stimulated A_{sat} (p = 0.012): the average A_{sat} 336 (+/- one SE) was 4.8 (+/- 0.3) and 6.0 (+/- 0.4) μ mol m⁻² s⁻¹ at 10 and 70 kg N ha⁻¹ y⁻¹, 337 respectively. There was no interaction between O₃ exposure and N supply. Again, there was a 338 339 strong linear relationship between A_{sat} (p < 0.001) and chlorophyll content and a weak linear relationship (p = 0.033) between A_{sat} and g_s (Figure S3). 340

341

342 N content of leaves

343 The N content of green, mature leaves in the middle of July was not affected by treatment;

the mean N content of these leaves was 1.2% on average. Near the end of the growing season,

the N content of leaves that were ready to drop off had declined to 0.5 - 0.7%. An increase in O₃ exposure resulted in a significant (p < 0.001) increase in N content of these leaves (Figure 8). N supply did not significantly affect the N content of senesced leaves and there was no significant O₃ x N interaction.

349

350 *Leaf number*

Independent of time of the season, enhanced N supply stimulated the number of leaves on silver birch (p < 0.001; Figure 9). On the other hand, while there was no effect of O₃ on leaf number at the start of the season, leaf number decreased with higher O₃ at the end of the season (p < 0.001). There was no evidence of an interaction between O₃ and nitrogen.

355

356 Discussion

In the current study, conducted at a range of O₃ and N exposure levels, no significant 357 interactions between the impacts of O₃ and N on photosynthesis and other leaf traits were 358 observed, either for young leaves or aging leaves. This suggests that the sensitivity of these 359 traits to O_3 in young silver birch trees is neither ameliorated (as shown for some trees species, 360 e.g. Häikiö et al., 2007; Handley and Grulke, 2008; Watanabe et al., 2006) nor enhanced by 361 N fertilization (as shown for some other tree species, e.g. Utriainen and Holopainen, 2001b; 362 Wanatabe et al., 2012; Yamaguchi et al., 2007). Although Pääkkönen and Holopainen (1995) 363 364 suggested that sufficient N supply might reduce O₃-sensitivity in silver birch, interactions between O₃ and N supply were not consistent between experiments and only observed 365 consistently for certain leaf parameters (e.g. leaf area and leaf dry mass per area). They had 366 conducted experiments at only two O₃ exposure levels (ambient and ca. 1.6 x ambient) and 367 up to three N loads (37, 74 and 150 kg N ha⁻¹ y⁻¹). Our results, however, confirmed that the 368 response of leaf traits to O_3 is dependent on the developmental stage and age of the leaf 369

370 studied, in agreement with various other studies (Bagard et al., 2008; Hoshika et al., 2013; Mäenpää et al., 2011; Oksanen et al., 2003). As the magnitude of the impact of O₃ is related 371 to the accumulated O₃ flux (Phytotoxic Ozone Dose) entering the stomata (Mills et al., 372 373 2011b), relatively little impact was expected early in the growing season, with impacts increasing with leaf age and longer-term exposure. Indeed, O₃ x time interactions were 374 observed for most photosynthetic parameters and the chlorophyll content of aging leaves. In 375 contrast, the gs of aging leaves was not affected by O3 and no O3 x time interactions were 376 found. 377

378

379 Traits of a young leaf throughout the growing season

Although the leaf primordia for early leaves were developed during the previous year, there 380 381 did not seem to be a carry-over effect of ozone on the number of leaves developed in 2013. Enhanced N, however, seems to have stimulated the number of leaves present at the start of 382 2013, before O_3 and N treatments were started again (Figure 8). The late leaves started to 383 384 develop in July and the first older leaves started to drop off near the end of July. The heterophyllous character of silver birch explains why the chlorophyll content index of the 385 third leaf (youngest fully expanded leaf) declined first between the end of May and 23rd July, 386 and subsequently rose again by the 23rd of August to a level slightly higher than at the end of 387 May (Figure 2). It is likely that remobilisation of N occurred during this period from the older 388 389 to the young, newly grown leaves. As to be expected, the chlorophyll content of the young leaves increased significantly with increasing N supply, and this increase was independent of 390 the Julian day. O₃, however, did not affect the chlorophyll content of the young leaves, which 391 might be due to the relatively short time of exposure to O_3 (as discussed below, elevated O_3 392 reduced the chlorophyll content of aging leaves). Linear relationships between chlorophyll 393 content and V_{c.max}, J_{max} and A_{sat} (see Figures 6, S1 and S2) suggest that a similar response to 394

395	O_3 and N can be expected for these photosynthetic traits for the third leaf when developing
396	throughout the season. Indeed, Riikonen et al. (2009) did not find an effect of O_3 on $V_{c,max}$,
397	J_{max} and A_{sat} in <i>Betula pendula</i> saplings on the youngest fully expanded leaf, however, the
398	level of O_3 exposure was very low in their study (mean O_3 concentration of $32 - 34$ ppb). On
399	the other hand, g_s was not affected by O_3 exposure or N supply in our study. Although many
400	studies have reported a decline in g_s in trees with increasing O_3 (Wittig et al., 2007, and
401	references therein), including for Betula pendula (Riikonen et al., 2009), other studies have
402	reported no change or even an increase in g_s in response to O_3 exposure (Mills et al., 2016,
403	and references therein). The varying responses might well be explained by inherent
404	differences in responses of species/genotypes to O ₃ , variation in exposure facilities,
405	magnitude and duration of O ₃ exposure. For example, Wittig et al. (2007) observed no
406	change at low to medium O_3 concentrations (30 – 59 ppb), a decline at medium to high O_3
407	concentration ($60 - 119$ ppb) and no change at very high O ₃ concentrations (>120 ppb).
408	

409 Traits of an aging leaf throughout the growing season

Leaf photosynthetic capacity ($V_{c,max}$, J_{max}) and A_{sat} declined between $10 - 14^{th}$ June and $22 - 14^{th}$ 410- 25^{th} July (Figure 5), but did not decline any further between $22 - 25^{\text{th}}$ July and $2 - 5^{\text{th}}$ 411 September, apart from at 10 kg N ha⁻¹ y⁻¹ and 68 ppb O_3 , where the decline continued. The 412 halt in decline later in the season may be influenced by the fact that older, senesced leaves 413 started to drop off with some new growth and remobilisation of N occurring (see above). 414 High O₃ exposure reduced the photosynthetic capacity and A_{sat} of leaves, and more so later 415 than early in season. A significant interaction between O₃ and time (although not significant 416 for $V_{c,max}$) confirms that effects of O_3 are dependent on the accumulated phytotoxic O_3 dose 417 (LRTAP Convention, 2015; Mills et al., 2011b). Noormets et al. (2010) also found that 418 elevated O₃ decreased photosynthetic parameters and leaf chlorophyll content, particularly in 419

420	the lower canopy, and that the negative impact of O_3 increased with time. A similar response
421	to O ₃ was found for A _{sat} in <i>Fagus crenata</i> seedlings, with a significant interaction between O ₃
422	and N only observed in July (Yamaguchi et al., 2007). However, it is not clear for which
423	leaves Asat was determined in Fagus crenata. Whilst Marzuoli et al. (2016) reported a
424	significant decline in $V_{c,max}$ due to elevated O_3 in <i>Quercus robur</i> seedlings in the first year, no
425	significant effect of elevated O_3 on $V_{c,max}$ was observed in the second year of O_3 exposure. In
426	contrast to Yamaguchi et al. (2007), we did not find any evidence that N impacts dominate
427	early in the season and that O_3 impacts dominate later in the season. Although elevated N
428	supply stimulated photosynthetic capacity ($V_{c,max}$ and J_{max}) as to be expected, N supply did
429	not significantly affect Asat. Whilst high N supply seems to have alleviated to some extent the
430	impact of O_3 on photosynthetic capacity observed at low N supply later in the growing
431	season, there was no significant interaction between O3 and N treatment, not even when the
432	data for September were analysed separately. The latter might be due to the low number of
433	replication (four instead of six as some leaves had already dropped off the tree) and the
434	higher variability later in the season. The higher decline in Asat compared to photosynthetic
435	capacity in September at high O3 exposure and high N supply can be explained by a decline
436	in g_s (Figure 5). Nevertheless, O_3 exposure, N supply and time did not affect g_s in the aging
437	leaf in our study when measured at light saturation. In contrast, Yamaguchi et al. (2007)
438	reported a significant increase in g _s in September due to high O ₃ exposure in <i>Fagus crenata</i> ,
439	with no O_3 effect observed earlier in the year, however, they did not define clearly which leaf
440	was measured. We observed a significant and strong linear relationship between A_{sat} and the
441	chlorophyll content of leaves. Although significant, the linear relationship between A_{sat} and g_s
442	was weak ($p = 0.038$), with the data showing a lot of scatter. This suggests that impacts of
443	treatment on A _{sat} were mainly driven by impacts on photosynthetic capacity rather than
444	changes in g_s . Indeed, previous studies have shown that O_3 -induced reductions in A_{sat} are

445 mainly driven by reductions in photosynthetic capacity in the chloroplasts (Farage et al., 1991; Hoshika et al., 2013; Niu et al., 2014; Yamaguchi et al., 2007). In agreement with 446 Noormets et al. (2010) and Bagard (2008), the biochemical component of photosynthesis 447 $(V_{c,max})$ was damaged more by O₃ than electron transport (J_{max}), as the ratio of J_{max} to V_{c,max} 448 was enhanced at elevated O₃ exposure. Decreases of both Rubisco content and activation 449 state have been reported for trees and crops (Ainsworth et al., 2012), often associated with 450 down regulation of photosynthetic genes and upregulation of genes involved in programmed 451 cell death and/or tissue senescence (Bohler et al., 2007; Ernst, 2013; Heath, 2008; Pell et al., 452 453 1997). In general, terrestrial plants allocate as much as 50% of their leaf N content to Rubisco because of inefficient catalytic activity of the enzyme (Spreitzer and Salvucci, 2002). If Asat 454 is mainly determined by the amount of Rubisco in birch leaves, then the O₃-induced 455 456 reduction in Asat might well be due to a decrease in allocation of leaf N to Rubisco 457 (Yamaguchi et al., 2007). Because photosynthetic measurements are often made on the most recently fully expanded upper canopy leaves, it is often difficult to demonstrate differences in 458 459 net photosynthesis due to O₃ until near or after the time when production of new upper canopy leaves has ceased. The growth of new leaves observed in July might have enabled 460 birch to compensate for some of the O₃ damage to the photosynthetic apparatus occurring in 461 older leaves later in the growing season. 462

463

464 Developmental stage of the leaves affects the response to treatments

The results discussed above indicate that the response of leaf traits to treatments is affected
by the developmental stage of leaves, with different responses being observed in young
leaves compared to aging leaves (Bagard et al., 2008; Hoshika et al., 2013; Mäenpää et al.,
2011; Oksanen et al., 2003). This was further confirmed when the chlorophyll content of
different leaves along a branch was determined. Whereas the chlorophyll content declined

470 with O₃ concentration in older leaves, in younger leaves the highest chlorophyll content was found at intermediate O₃ concentrations at high N supply (Table 2, Figure 4). For medium-471 aged leaves, there was no effect of O_3 on the chlorophyll content. Hoshika et al. (2013) 472 473 reported differences in stress resistance to O₃ between early and late leaves in white birch (Betula platyphylla var. japonica), with early leaves not showing an O₃-induced reduction in 474 photosynthesis, and an O₃-induced decline in photosynthesis and early defoliation occurring 475 476 in late leaves. Hence, late leaves were more sensitive to O_3 than early leaves in white birch. However, our results indicate that differences in response to O_3 are primarily determined by 477 478 the age/developmental stage of leaves and exposure time to O₃ rather than the type of leaf. Responses to treatments are further confounded when nearer the end of the growing season N 479 480 is remobilised from old leaves to young leaves, maintaining the photosynthetic capacity of 481 the young leaves. The various responses to O₃, and the interaction between O₃ and N, 482 reported in the literature might well be due to measurements on leaves at different developmental stages, which confounds comparisons between experiments. In agreement 483 484 with Uddling et al. (2005), we observed that enhanced O_3 impairs the resorption of N in leaves of silver birch near the end of the growing season. Similar results were also reported 485 for Fagus crenata (Yamaguchi et al., 2007). An increase in leaf N, partly due to the release of 486 NH4⁺ from both protein degradation and enhanced activity of the shikimate pathway, is 487 488 known to occur at high ozone dose (early senescence) and a part of this N could be 489 transiently re-assimilated via an increased activity of the cytosolic glutamine synthetase (Bohler et al., 2007; Dizengremel et al., 2012; Galant et al., 2012) or stored in gamma-490 aminobutyric acid (Dizengremel et al., 2012). The enhanced N content in litter at elevated O₃ 491 492 is likely to affect the subsequent cycling of N in the soil.

493

494 Conclusions

495 At the leaf level, we did not observe any interactions between O₃ and N regarding photosynthetic parameters, chlorophyll content, N content in senesced leaves and leaf 496 number. This suggests that the sensitivity of these leaf traits to O₃ in young silver birch trees 497 498 was neither reduced nor enhanced by N fertilization under our experimental conditions. Our results suggest that the response of leaf traits to O₃ is dependent on the developmental 499 stage/age of the leaf studied. Whilst O₃ reduced the net rate of photosynthesis in aging birch 500 leaves later in the season, O₃ did not affect the g_s of those leaves. This suggests that there was 501 only a weak coupling between photosynthesis and g_s in aging silver birch leaves under light-502 503 saturating conditions. When studying the impacts of O₃ on leaf traits, it is important to define the developmental stage of the leaves, as this will affect the measured response to O_3 . 504

505

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513 **References**

514

Ainsworth, E.A., Yendrek, C.R., Sitch, S., Collins, W.J., Emberson, L.D., 2012. The effects
of tropospheric ozone on net primary production and implications for climate change. Annual
Review of Plant Biology 63, 637-661.

- 519 Bagard, M., Le Thiec, D., Delacote, E., Hasenfratz-Sauder, M., Banvoy, J., Gérard, J.,
- 520 Dizengremel, P., Jolivet, Y., 2008. Ozone-induced changes in photosynthesis and
- 521 photorespiration of hybrid poplar in relation to the developmental stage of the leaves.
- 522 Physiologia Plantarum 134, 559–574.
- 523
- Bates, D., Maechler, M., Bolker, B., Walker, S., 2015. Fitting linear mixed-effects models
 using lme4. Journal of Statistical Software 67, 1-48.
- 526
- 527 Bobbink, R., Hettelingh, J.-P. (Eds.), 2011. Review and revision of empirical critical loads
- 528 and dose response relationships. Proceedings of an international expert workshop,
- 529 Noordwijkerhout, 23-25 Juni 2010, RIVM-report 680359002. Coordination Centre for
- 530 Effects, RIVM, Bilthoven, The Netherlands.
- 531
- 532 Bobbink, R., Hicks, K., Galloway, J.N., Spranger, T., Alkemade, R., Ashmore, M.,
- 533 Bustamante, M., Cinderby, S., Davidson, E., Dentener, F., Emmett, B., Erisman, J.W., Fenn,
- 534 M., Gilliam, F., Nordin, A., Pardo, L., De Vries, W., 2010. Global assessment of nitrogen
- deposition effects on terrestrial plant diversity: a synthesis. Ecological Applications 20, 30-536 59.
- 537
- Bohler, S., Bagard, M., Oufir, M., Planchon, S., Hoffmann, L., Jolivet, Y., Hausman, J.-F.,
- 539 Dizengremel, P., Renaut, J., 2007. A DIGE analysis of developing poplar leaves subjected to
- ozone reveals major changes in carbon metabolism. Proteomics 7, 1584-1599.
- 541

542	Braun, S., Thomas, V.F.D, Quiring, R., Flückiger, W., 2010. Does nitrogen deposition
543	increase forest production? The role of phosphorus. Environmental Pollution 158, 2043-
544	2052.

546 Burnham, K.P., Anderson, D.R., 2002. Model selection and inference: a practical547 information-theoretic approach. Springer, New York.

548

Cape, J.N., Van der Eerden, L.J., Sheppard, L.J., Leith, I.D., Sutton, M.A., 2009. Evidence
for changing the critical level for ammonia. Environmental Pollution 157, 1033-1037.

551

Clausen, J.J., Kozlowski, T.T., 1965. Heterophyllous shoots in *Betula papyrifera*. Nature
205: 1030-1031.

554

- 555 Cooper, O.R., Parrish, D.D., Ziemke, J., Balashov, N.V., Cupeiro, M., Galbally, I.E., Gilge,
- 556 S., Horowitz, L., Jensen, N.R., Lamarque, J.-F., Naik, V., Oltmans, S.J., Schwab, J., Shindell,
- 557 D.T., Thompson, A.M., Thouret, V., Wang, Y., Zbinden, R.M., 2014. Global distribution and
- trends of tropospheric ozone: An observation-based review. Elementa Science of the

559 Anthropocene, 2, 000029, doi:10.12952/journal.elementa.000029.

- 561 Dise, N.B., Ashmore, M., Belyazid, S., Bleeker, A., Bobbink, R., De Vries, W., Erisman,
- J.W., Spranger, T., Stevens, C.J., Van den Berg, L., 2011. Nitrogen as a threat to European
- terrestrial biodiversity. In: Sutton, M. A., Howard, C. M., Erisman, J.W., Billen, G., Bleeker,
- A., Grennfelt, P., Van Grinsven, H., Grizzetti, B. (Eds). The European nitrogen assessment:
- sources, effects and policy perspectives, Cambridge University Press, Cambridge, UK, pp.
- 566 463-494.

568	Dizengremel, P., Vaultier, MN., Le Thiec, D., Cabané, M., Bagard, M., Gérant, D., Gérard,
569	J., Dghim, A.A., Richet, N., Afif, D., Pireaux, JC., Hasenfratz-Sauder, MP., Jolivet, Y.,
570	2012. Phosphoenolpyruvate is at the crossroads of leaf metabolic responses to ozone stress.
571	New Phytologist 195, 512-517.
572	
573	EMEP (2016). Air pollution trends in the EMEP region between 1990 and 2012.
574	EMEP/CCC-Report 1/2016. ISBN 978-82-425-2833-9.
575	
576	Ernst, D., 2013. Integrated studies on abiotic stress defence in trees: the case of ozone. In:
577	Matyssek, R. et al. (Eds). Climate change, air pollution and global challenges: Understanding
578	and perspectives from forest research. Developments in Environmental Science 13. Elsevier
579	Ltd., Amsterdam, pp. 289-307.
580	
500	
581	Farage, P.K., Long, S.P., Lechner, E.G., Baker, N.R., 1991. The sequence of change within
581 582	Farage, P.K., Long, S.P., Lechner, E.G., Baker, N.R., 1991. The sequence of change within the photosynthetic apparatus of wheat following short-term exposure to ozone. Plant
581 582 583	Farage, P.K., Long, S.P., Lechner, E.G., Baker, N.R., 1991. The sequence of change within the photosynthetic apparatus of wheat following short-term exposure to ozone. Plant Physiology 95, 529-535.
581 582 583 584	Farage, P.K., Long, S.P., Lechner, E.G., Baker, N.R., 1991. The sequence of change within the photosynthetic apparatus of wheat following short-term exposure to ozone. Plant Physiology 95, 529-535.
581 582 583 584 585	 Farage, P.K., Long, S.P., Lechner, E.G., Baker, N.R., 1991. The sequence of change within the photosynthetic apparatus of wheat following short-term exposure to ozone. Plant Physiology 95, 529-535. Galant, A., Koester, R.P., Ainsworth, E.A., Hicks, L.M., Jez, J.M., 2012. From climate
581 582 583 584 585 586	 Farage, P.K., Long, S.P., Lechner, E.G., Baker, N.R., 1991. The sequence of change within the photosynthetic apparatus of wheat following short-term exposure to ozone. Plant Physiology 95, 529-535. Galant, A., Koester, R.P., Ainsworth, E.A., Hicks, L.M., Jez, J.M., 2012. From climate change to molecular response: redox proteomics of ozone-induced responses in soybean.
581 582 583 584 585 586 586	 Farage, P.K., Long, S.P., Lechner, E.G., Baker, N.R., 1991. The sequence of change within the photosynthetic apparatus of wheat following short-term exposure to ozone. Plant Physiology 95, 529-535. Galant, A., Koester, R.P., Ainsworth, E.A., Hicks, L.M., Jez, J.M., 2012. From climate change to molecular response: redox proteomics of ozone-induced responses in soybean. New Phytologist 194, 220-229.
581 582 583 584 585 586 586 587 588	 Farage, P.K., Long, S.P., Lechner, E.G., Baker, N.R., 1991. The sequence of change within the photosynthetic apparatus of wheat following short-term exposure to ozone. Plant Physiology 95, 529-535. Galant, A., Koester, R.P., Ainsworth, E.A., Hicks, L.M., Jez, J.M., 2012. From climate change to molecular response: redox proteomics of ozone-induced responses in soybean. New Phytologist 194, 220-229.

590 of elevated ozone and nitrogen on growth and photosynthesis of European aspen (*Populus*

- *tremula*) and hybrid aspen (*P. tremula* x *Populus tremuloides*) clones. Canadian Journal of
 Forest Research 37, 2326–2336.
- 593

Handley, T., Grulke, N.E., 2008. Interactive effects of O₃ exposure on California black oak
(*Quercus kelloggii* Newb.) seedlings with and without N amendment. Environmental
Pollution 156, 53-60.

597

Hayes, F., Wagg, S., Mills, G., Wilkinson, S., Davies, W., 2012. Ozone effects in a drier
climate: implications for stomatal fluxes of reduced stomatal sensitivity to soil drying in a
typical grassland species. Global Change Biology 18, 948-959.

601

Heath, R.L., 2008. Modification of the biochemical pathways of plants induced by ozone:
What are the varied routes to change? Environmental Pollution 155, 453-463.

604

Hewitt, D.K.L, Mills, G., Hayes, F., Wilkinson, S., Davies, W., 2014. Highlighting the threat
from current and near-future ozone pollution to clover in pasture. Environmental Pollution
189, 111-117.

608

- Hewitt, D.K.L, Mills, G., Hayes, F., Davies, W., 2016. The climate benefits of high-sugar
- grassland may be compromised by ozone pollution. Science of the Total Environment 565:95-104.

- Hoshika, Y., Watanabe, M., Inada, N., Mao, Q., Koike, T., 2013. Photosynthetic response of
- 614 early and late leaves of white birch (*Betula platyphylla* var. *japonica*) grown under free-air
- ozone exposure. Environmental Pollution 182, 242-247.

617	Kuznetsova, A., Brockhoff, P.B., Christensen, R.H.B., 2014. ImerTest: Tests in Linear Mixed					
618	Effects Models. R package version 2.0-20. http://CRAN.R-project.org/package=lmerTest					
619						
620	Landolt, W., Günthardt-Goerg, M.S., Pfenninger, I., Einig, W., Hampp, R., Maurer, S.,					
621	Matyssek, R., 1997. Effect of fertilization on ozone-induced changes in the metabolism of					
622	birch (Betula pendula) leaves. New Phytologist 137, 389-397.					
623						
624	LRTAP Convention, 2015. Manual on methodologies and criteria for modelling and mapping					
625	Critical Loads and Levels and air pollution effects, risks and trends. Chapter 3: Mapping					
626	critical levels for vegetation. http://icpvegetation.ceh.ac.uk					
627						
628	Mäenpää, M, Riikonen, J., Kontunen-Soppela, S., Rous, M., Oksanen, E., 2011. Vertical					
629	profiles reveal impact of ozone and temperature on carbon assimilation of Betula pendula and					
630	Populus tremula. Tree Physiology 31, 808-818.					
631						
632	Marzuoli, R., Monga, R., Finco, A., Gerosa, G. (2016). Biomass and physiological responses					
633	of Quercus robur (L.) young trees during 2 years of treatments with different levels of ozone					
634	and nitrogen wet deposition. Trees 30, 1995-2000.					
635						
636						
637	Mills, G., Harmens, H., Wagg, S., Sharps, K., Hayes, F., Fowler, D., Sutton, M., Davies, W.,					
638	2016. Ozone impacts on vegetation in a nitrogen enriched and changing climate.					
639	Environmental Pollution 208, 898-908.					
640						

- Mills, G., Hayes, F., Simpson, D., Emberson, L., Norris, D., Harmens, H., Büker, P., 2011a.
 Evidence of widespread effects of ozone on crops and (semi-)natural vegetation in Europe
 (1990 2006) in relation to AOT40 and flux-based risk maps. Global Change Biology 17:
 592-613.
- 645
- Mills, G., Hayes, F., Wilkinson, S., Davies, W.J., 2009. Chronic exposure to increasing
 background ozone impairs stomatal functioning in grassland species. Global Change Biology
 15, 1522-1533.
- 649
- 650 Mills, G., Pleijel, H., Braun S., Büker, P., Bermejo, V., Calvo, E., Danielsson, H., Emberson,
- L., González Fernández, I., Grünhage, L., Harmens, H., Hayes, F., Karlsson, P.-E., Simpson,
- D., 2011b. New stomatal flux-based critical levels for ozone effects on vegetation.
- Atmospheric Environment 45, 5064-5068.
- 654
- Noormets, A., Kull, O, Sôber, A., Kubiske, M.E., Karnosky, D.F., 2010. Elevated CO₂
- response of photosynthesis depends on ozone concentration in aspen. Environmental
- 657 Pollution 158, 992–999.
- 658
- 659 Oksanen, E., 2003. Responses of selected birch (*Betula pendula*) clones to ozone change
- over time. Plant, Cell and Environment 26, 875–886.
- 661
- 662 Parrish, D.D., Law, K.S., Staehelin, J., Derwent, R., Cooper, O.R., Tanimoto, H., Volz-
- Thomas, A., Gilge, S., Scheel, H.-E., Steinbacher, M., Chan, E., 2012. Long-term changes in
- lower tropospheric baseline ozone concentrations at northern mid-latitudes. Atmospheric
- 665 Chemistry and Physics 12, 11485-11504.

667	Pääkkönen, E., Holopainen, T., 1995. Influence of nitrogen supply on the response of clones
668	of birch (Betula pendula Roth.) to ozone. New Phytologist 129, 595-603.
669	
670	Pell, E.J., Schlagnhaufer, C., Arteca, R.N. 1997. Ozone-induced oxidative stress:
671	Mechanisms of action and reaction. Physiologia Plantarum 100: 264-273.
672	
673	R Core Team, 2014. R: A language and environment for statistical computing. R Foundation
674	for Statistical Computing, Vienna, Austria. URL: <u>http://www.R-project.org/</u> .
675	
676	Riikonen, J., Mäenpää, M, Alavillamo, M., Silfver, T., Oksanen, E., 2009. Interactive effect
677	of elevated temperature and O ₃ on antioxidant capacity and gas exchange in <i>Betula pendula</i>
678	saplings. Planta 230, 419–427.
679	
680	Royal Society, 2008. Ground-level ozone in the 21st century: future trends, impacts and
681	policy implications. Science Policy Report 15/08. The Royal Society, London, UK.
682	
683	Sharkey, T.D., Bernacchi, C.J., Farquhar, G.D., Singsaas, E.L., 2007. Fitting photosynthetic
684	carbon dioxide response curves for C_3 leaves. Plant, Cell and Environment 30, 1035–1040.
685	
686	Simpson. D., Arneth, A., Mills, G., Solberg, S., Uddling, J., 2014. Ozone — the persistent
687	menace: interactions with the N cycle and climate change. Current Opinion in Environmental
688	Sustainability 9–10, 9–19.
689	

- 690 Solberg, S., Bergström, R., Langner, J., Laurila, T., Lindskog, A., 2005. Changes in Nordic
- 691 surface ozone episodes due to European emission reductions in the 1990s. Atmospheric
- 692 Environment 39, 179-192.
- 693 Spreitzer, R.J., Salvucci, M.E., 2002. Rubisco: structure, regulatory interactions, and
- 694 possibilities for a better enzyme. Annual Review of Plant Biology 53: 449 -475.
- 695
- 696 Sutton, M. A., Howard, C. M., Erisman, J.W., Billen, G., Bleeker, A., Grennfelt, P., Van
- 697 Grinsven, H., Grizzetti, B., 2011. The European nitrogen assessment: sources, effects and
- 698 policy perspectives, Cambridge University Press, Cambridge, UK.
- 699
- Thomas, V.F.D., Braun, S., Flückiger, W., 2005. Effects of simultaneous ozone exposure and
 nitrogen loads on carbohydrate concentrations, biomass, and growth of young spruce trees
- 702 (*Picea abies*). Environmental Pollution 137, 507–516.
- 703
- 704 Thomas, V.F.D., Braun, S., Flückiger, W., 2006. Effects of simultaneous ozone exposure and
- nitrogen loads on carbohydrate concentrations, biomass, growth and nutrient concentrations
- of young beech trees (*Fagus sylvatica*). Environmental Pollution 143, 341-354.
- 707
- 708 Tørseth, K., Aas, W., Breivik, K., Fjæraa, A.M., Fiebig, M., Hjellbrekke, A.G., Lund Myhre,
- C., Solberg, S., Yttri, K.E., 2012. Introduction to the European Monitoring and Evaluation
- 710 Programme (EMEP) and observed atmospheric composition change during 1972–2009.
- 711 Atmospheric Chemistry and Physics 12, 5447–5481.
- 712
- 713 Uddling, J., Karlsson, P.E., Glorvigen, A., Sellden, G., 2005. Ozone impairs autumnal
- resorption of nitrogen from birch (*Betula pendula*) leaves, causing an increase in whole-tree
- nitrogen loss through litter fall. Tree Physiology 26,113–120.

~	1	C
		n
	-	~

717	Utriainen, J., Holopainen, T., 2001a. Influence of nitrogen and phosphorus availability and
718	ozone stress on Norway spruce seedlings. Tree Physiology 21, 447-456.
719	
720	Utriainen, J., Holopainen, T., 2001b. Nitrogen availability modifies the ozone responses of
721	Scots pine seedlings exposed in an open-field system. Tree Physiology 21, 1205–1213.
722	
723	Venables, W. N., Ripley, B. D., 2002. Modern Applied Statistics with S. Fourth Edition.
724	Springer, New York.
725	
726	Volz, A. and Kley, D., 1988. Evaluation of the Montsouris series of ozone measurements
727	made in the nineteenth century. Nature 332, 240-242.
728	
729	Watanabe, M., Yamaguchi, M., Iwasaki, M., Matsuo, N., Naba, J., Tabe, C., Matsumura, H.,
730	Kohno, Y., Izuta, T., 2006. Effects of ozone and/or nitrogen load on the growth of Larix
731	kaempferi, Pinus densifloram and Cryptomeria japonica seedlings. Journal of Japan Society
732	for Atmospheric Environment 41, 320–334.
733	
734	Watanabe, M., Yamaguchi, M., Matsumura, H., Kohno, Y., Izuta, T., 2012. Risk assessment
735	of ozone impact on Fagus crenata in Japan: consideration of atmospheric nitrogen
736	deposition. European Journal of Forest Research 131, 475-484.
737	
738	Wittig, V.E., Ainsworth, E.A., Long, S.P., 2007. To what extent do current and projected
739	increases in surface ozone affect photosynthesis and stomatal conductance of trees? A meta-

analytic review of the last three decades of experiments. Plant, Cell and Environment 30,1150-1162.

742

743	Wittig, V.E.	, Ainsworth, E.A.	, Naidu, S.L.,	Karnoski, D.F., Lor	ng, S.P., 200	9. Quantifying the
	0,	, , ,	, , , ,	, , ,	0, ,	

impact of current and future tropospheric ozone on tree biomass, growth, physiology and

biochemistry: a quantitative meta-analysis. Global Change Biology 15, 396-424.

746

747 Yamaguchi, M., Watanabe, M., Iwasaki, M., Tabe, C., Matsumura, H., Kohno, Y., Izuta, T.,

748 2007. Growth and photosynthetic responses of *Fagus crenata* seedlings to O₃ under different

749 nitrogen loads. Trees 21, 707–718.

Ozone treatment	1	2	3	4	5	6	7
24 hr mean (ppb)	35.7	39.3	42.3	47.8	53.1	59.1	67.9
Daylight mean (ppb)	36.1	40.0	43.6	50.0	54.7	63.9	71.5
AOT40 (ppm h⁻¹)	2.4	4.2	7.1	12.1	16.1	25.1	32.5
Season max. (ppb)	62.4	71.3	81.3	100.4	111.3	113.5	126.3
Air temperature (°C) VPD (kPa)				PAR (umol m ⁻² s ⁻¹)			
24 h	Daylight	Daily max.	24 h	Daylight	Daily max.	Daylight	Daily max.
mean	mean	mean	mean	mean	mean	mean	mean
19.8	24.0	25.6	0.67	1.24	1.55	515	808

Table 1. Summary of ozone treatments and climate conditions in 2013.

753

Table 2. Summary statistics of effects of treatments and time¹ on the chlorophyll content and

755 stomatal conductance (G_s) of the third leaf (youngest fully expanded) measured throughout

the growing season.

Variable	Chl. content	Gs		
O ₃	n.e.	n.e.		
Ν	p<0.001	n.e.		
Time	p<0.001	p<0.01		
Time ²	p<0.001	p<0.01		
¹ No two or three-way interactions w				

¹ No two or three-way interactions were found.

759 n.e. = no significant effect at p = 0.05.

Time² = time squared.

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Table 3. Summary statistics of effects of treatments and leaf number on the chlorophyll
content of silver birch leaves of different age along the first top branch near the end of the
growing season (on 4 September 2013).

765

Variable	Chl. content			
0 ₃	n.e.			
Ν	p<0.001			
Leaf number	p<0.001			
O ₃ * N	n.e.			
O ₃ * Leaf number	p<0.001			
N * Leaf number	n.e.			
$O_3 * N * Leaf number$	n.e.			
n.e. = no significant effect at p = 0.05.				

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Table 4. Summary statistics of effects of treatments and time¹ on photosynthetic parameters,
 stomatal conductance (G_s) and chlorophyll (Chl.) content (index) of the same birch leaves
 measured throughout the season, i.e. aging with time.

Variable	V _{c,max}	J _{max}	J _{max} /V _{c,max}	A _{sat}	G_s (at A_{sat}	Chl. content
O ₃	p<0.01	p<0.01	p=0.05	p<0.01	n.e. ²	(p=0.055)
Ν	p<0.01	p=0.02	p=0.01	n.e.	n.e.	p=0.03
Time	p<0.001	p<0.001	(p=0.09)	p<0.001	n.e.	p<0.001
O ₃ * Time	e (p=0.08)	p<0.01	n.e.	p=0.001	n.e.	p=0.04

¹ No $O_3 \times N$, N x Time or $O_3 \times N \times T$ ime interactions were found.

776 ² n.e. = no significant effect at p = 0.05.



Figure 1. Weekly ozone profile in seven solardomes at Abergwyngregyn, nr. Bangor, NorthWales in 2013.





Figure 2. Effect of ozone exposure and nitrogen supply on the chlorophyll content (index) ofsilver birch leaves (third leaf on main stem, fully expanded) from late May to late August2013. Values are means \pm one SE (n = 5 - 6). Solid line: 10 kg N ha⁻¹ y⁻¹, broken line: 70 kgN ha⁻¹ y⁻¹.





Figure 3. Effect of ozone exposure and nitrogen supply on the stomatal conductance (G_s) of silver birch leaves (third leaf on main stem, fully expanded) from late May to late August 2013. Values are means \pm one SE (n = 5 – 6). Solid line: 10 kg N ha⁻¹ y⁻¹, broken line: 70 kg N ha⁻¹ y⁻¹.





Figure 4. Effect of ozone exposure and nitrogen supply on the chlorophyll content (index) of 801 silver birch leaves of different age along the first top branch near the end of the growing 802 season (4 September 2013). a) Data are means (n = 6); the first number in the legend refers to 803 the 24 hr mean ozone concentration (ppb) and the second number to the nitrogen supply (kg 804 N ha⁻¹ y⁻¹); trend lines are shown for the lowest (black lines) and highest (grey lines) ozone 805 exposure for low (solid lines) and high (broken lines) nitrogen supply. See Table S1 for 806 values and standard errors. b) Data are means \pm one SE (n = 24), with ozone exposure data 807 808 averaged for low (filled circles, solid line) and high (open circles, broken line) nitrogen supply. Leaf number one is the youngest leaf near the top of the branch, whereas leaf number 809 eight is the oldest measured leaf situated near the origin of the branch. 810



Figure 5. Photosynthetic capacity ($V_{c,max}$ and J_{max}), light-saturated rate of photosynthesis (A_{sat}), chlorophyll content (index), and stomatal conductance (G_s ; at ambient CO₂) of aging silver birch leaves, i.e. the same leaves were measured throughout the growing season. Trees were exposed to a 24 hr mean ozone concentration of 36 or 68 ppb and supplied with 10 or 70 kg N ha⁻¹ yr⁻¹ in 2013. Data are means ± one SE (n = 4 - 6).





Figure 6. Relationship between light-saturated rate of photosynthesis (A_{sat}) and a)
chlorophyll content (index) and b) stomatal conductance (G_s) of aging silver birch leaves (see

Figure 5). Trees were exposed to a 24 hr mean ozone concentration of 36 or 68 ppb and supplied with 10 or 70 kg N ha⁻¹ yr⁻¹ in 2013.





Figure 7. Light-saturated rate of photosynthesis (A_{sat}) of aged silver birch leaves near the end of the growing season (late August, early September). Trees were exposed to a 24 hr mean ozone concentration of 36 - 68 ppb and supplied with 10 or 70 kg N ha⁻¹ yr⁻¹ in 2013. In a) mean values (± one SE; n = 4 - 5) are shown for different nitrogen supply rates, whereas in b) mean values (± one SE; n = 8 - 10) for different nitrogen supply rates were pooled.



Figure 8. Impacts of ozone exposure on the leaf nitrogen concentration of silver birch leaves
just before they fell of the tree. Values are the mean of four different nitrogen supply rates (±
one SE).

843



844 845

Figure 9. Impact of ozone and nitrogen on leaf number near the start (5th May) and end of the 2013 growing season (24th September). Data are means \pm one SE (n = 5 - 6). Any effect early in the season reflect a carry-over effect from the previous growing season. Trend lines are

shown for the lowest (solid line) and highest (broken line) nitrogen supply.

Supplement

Table S1. Effect of ozone exposure and nitrogen supply on the chlorophyll content of silver birch leaves of different age along the first top branch near the end of the growing season (4 September 2013). Leaf number one is the youngest leaf near the top of the branch.

			Chlorop	hyll content
Ozone (ppb)	N (kg ha ⁻¹ y ⁻¹)	Leaf no.	Mean	Standard error
35.7	10	1	11.4	2.2
		2	11.0	2.1
		3	8.6	1.2
		4	6.4	0.8
		6	5.9	0.9
		7	5.5	0.8
		8	5.0	0.8
	70	1	14.1	2.9
		2	14.1	2.7
		3	12.7	1.9
		- 5	10.9	1.0
		6	10.1	1.7
		7	9.3	1.8
		8	7.7	1.0
39.3	10	1	14.4	1.1
		2	12.2	1.3
		3	9.0	1.0
		5	8.0	1.8
		6	7.1	1.7
		7	6.4	1.1
		8	6.1	1.0
	70	1	20.9	2.0
		2	17.3	1.2
		3	14.5	1.0
		5	10.6	1.7
		6	7.6	0.8
		7	6.3	0.8
		8	5.4	0.9
47.8	10	1	13.8	2.3
		2	9.0	2.9
		4	7.7	1.6
		5	6.3	1.4
		6	5.6	1.2
		7	5.0	0.2
	70	8	4.7	0.4
	70	1	19.0	2.4
		2	17.6	2.5
		4	15.1	2.3
		5	13.0	2.5
		6	9.9	1.9
		7	8.6	1.8
67.0	10	8	6.0	0.7
07.9	10	2	13.7	2.0
		3	10.5	2.0
		4	8.1	1.7
		5	6.6	1.5
		6	4.9	0.9
		7	4.3	0.5
	70	0 1	3.8 19.5	0.4 2 3
	70	2	15.8	2.1
		3	12.4	2.6
		4	10.5	2.3
		5	8.8	2.1
		6	7.5	1.9
		7	5.9 1 G	1.1
		0	4.0	1.1

Table S2. Relative difference (%) in photosynthetic parameters, chlorophyll (Chl.) content

and stomatal conductance (g_s) at high (24 hr mean of 68 ppb) compared to low ozone (24 hr
 mean of 36 ppb) at different times during the growing season. Replace Chlorophyll index

11 with Chlorophyll content in table.

Month	N (kg ha ⁻¹ y ⁻¹)	A _{sat}	V _{c,max}	J_{max}	J _{max} /V _{c,max}	Chl. index	gs
June	10	9.4	10.2	8.5	-2.0	-11.2	-9.8
	70	-0.1	7.1	1.8	-7.0	8.1	0.3
July	10	27.4	32.4	30.0	-5.5	13.4	-2.5
	70	3.6	16.5	4.0	-14.8	11.0	13.6
September	10	45.0	59.0	50.7	-14.9	26.4	18.8
	70	39.9	25.7	34.0	7.7	29.5	30.4
June	Combined ¹	5.0	10.4	6.1	-5.3	3.2	-4.8
July	Combined ¹	15.3	24.6	16.8	-11.2	13.0	7.6
September	Combined ¹	42.2	40.8	41.5	-3.6	28.1	24.8

¹ Data for low and high N were combined as there was no significant $O_3 \times N$ interaction.

Negative values indicate a higher value of the variable at higher compared to low O₃ exposure.





Figure S1. Relationship between a) V_{c,max} and b) J_{max} and chlorophyll content of aging silver
 birch leaves (see Figure 5). Trees were exposed to a 24 hr mean ozone concentration of 36 or
 68 ppb and supplied with 10 or 70 kg N ha⁻¹ yr⁻¹ in 2013. See also Figure 5.





Figure S2. Relationship between light-saturated rate of photosynthesis (A_{sat}) stomatal
conductance of aging silver birch leaves (see Figure 6b). Trees were exposed to a 24 hr mean
ozone concentration of 36 (black lines) or 68 ppb (grey lines) and supplied with 10 (solid
lines) or 70 kg N ha⁻¹ yr⁻¹ (broken lines) in 2013.





Figure S3. Relationship between light-saturated rate of photosynthesis (A_{sat}) and and a)
 chlorophyll content and b) stomatal conductance of aged silver birch leaves near the end of

41 the growing season (late August, early September). Trees were exposed to a 24 hr mean

42 ozone concentration of 36 - 68 ppb and supplied with 10 or 70 kg N ha⁻¹ yr⁻¹ in 2013. See

43 also Figure 7.