1 Application of the Red List Index as an indicator of habitat change

- 2 Aino Jusléna*, Juha Pykäläb, Saija Kuuselab, Lauri Kailaa, Jaakko Kullberga, Jaakko Mattilaa, Jyrki Muonaa,
- 3 Sanna Saaria, Pedro Cardosoa.
- ^a Finnish Museum of Natural History, University of Helsinki, P.O. Box 17, FI-00014 University of Helsinki
- 5 ^b Finnish Environment Institute, P.O. Box 140, FI-00251 Helsinki
- 6 **Corresponding author:**
- 7 Aino Juslén: aino.juslen@helsinki.fi; tel. +358 50 3109703, Finnish Museum of Natural History, University of Helsinki,
- 8 P.O. Box 17, FI-00014 University of Helsinki

Abstract

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For the first time ever, the International Union for Conservation of Nature Red List Index for habitat types was calculated for an entire country, Finland. The RLIs were based on species threat assessments from 2000 and 2010 and included habitat definitions for all 10 131 species of 12 organism groups. The RLIs were bootstrapped to track statistically significant changes. The RLI changes of species grouped by habitats were negative for all habitat types except for forests and rural biotopes which showed a stable trend. Trends of beetles and true bugs were positive in rural and forest habitats. Other 16 observed trends of species group and habitat combinations were negative. Several trends observed were in accordance with studies focusing on particular taxa and habitats, and drivers for their change. This study demonstrates the usefulness of the RLI as a tool for observing habitat change based on species threat assessment data.

Keywords: biodiversity indicator, biodiversity loss, habitat, Finland, RLI, threatened species

1. Introduction

The 2010 conference of the parties to the Convention on Biological Diversity (CBD) in Aichi, Japan declared the 2010-2020 decade as a Decade on Biodiversity. Twenty biodiversity targets were set to be met by the year 2020 (Tittensor et al. 2014). Among these, target 12 says "By 2020 the extinction of known threatened species has been prevented and their conservation status, particularly of those most in decline, has been improved and sustained". The IUCN Red List of Threatened Species is the most widely used information source on the extinction risk of species (Rodrigues et al. 2006; Mace et al. 2008; but see Cardoso et al. 2011, 2012). The IUCN Red List Index (RLI) (Butchart et al. 2004, 2007), which reflects overall changes in IUCN Red List status over time of a group of taxa, was agreed by the parties to the CBD to be used as an overall index of change, to quantify to what extent target 12 is being met. The RLI uses weight scores based on the Red List status of each of the assessed species. These scores range from 0 (Least Concern) to Extinct/Extinct in the Wild (5). Summing these scores across all species and relating them to the worst-case scenario - all species extinct - gives us an indication of how biodiversity is doing. Importantly, the RLI is based on true improvements or deteriorations in the status of species, i.e. "genuine changes". It excludes category changes resulting from, e.g., new knowledge (Butchart et al. 2007). The RLI approach helps to develop a better understanding of which taxa, regions or ecosystems are declining or improving. The aim is to provide policy makers, stakeholders, conservation practitioners and the general public with sound knowledge of biodiversity status and change, and tools with which to make informed decisions. At a global level, the IUCN Red List Index has been calculated for birds (Butchart et al. 2004; Hoffman et al. 2010), mammals (Hoffmann et al. 2010, 2011), amphibians (Hoffman et al. 2010), corals (Butchart et al. 2010), and cycads (The Millenium Development Goals Report 2015). An ongoing project is heading to present a sampled Red List Index (SRLI, Baillie et al. 2008) of plants (Brummitt et al. 2015) and efforts towards a SRLI of butterflies (Lewis and Senior 2011) and Odonata are made (Clausnitzer et al. 2009). At a regional and national level, RLIs or SRLIs have been presented for certain groups (Lopez et al. 2011; Szabo et al. 2012; Moreno Saiz et al. 2015; Woinarski et al. 2015) or multiple species groups (Gärdenfors 2010; Juslén et al. 2013; Rondinini et al. 2014). A parallel set of criteria was proposed to be applied to ecosystems in lieu of species, with much the same

objectives, the IUCN Red List of Ecosystems (RLE, Rodríguez et al. 2011). This has not been widely

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adopted as of yet, either at global or regional scales. National assessments of threatened habitat types have been carried out, for example in Finland (Raunio et al. 2008; Kontula and Raunio 2009). Kontula and Raunio (2009) even presented a procedure for assigning IUCN Red List categories for habitat types. However, this assessment has been carried out only once in Finland, and temporal trends cannot be presented as of yet. Until repeated assessments of risk of collapse of particular ecosystem types are available using the Red List of Ecosystems approach, it will not be possible to produce a Red List Index for different ecosystems using the RLE approach. However, as a proxy for ecosystem or habitat change, it is possible to calculate RLIs for sets of species characteristic of particular ecosystem or habitat types.

Butchart et al. (2004) has already used such an approach for birds. In practice, any index based on species trends that includes additional information such as habitat types can be used to perceive trends on species groups other than taxonomic. Besides the RLI, we can mention the Living Planet Index (LPI), which is based on population trends of vertebrates from around the world and that has been used in multiple ways, including for quantifying habitat trends (Loh et al. 2005; Collen et al. 2009). The LPI does however require much more information than the RLI, hence its focus on vertebrates.

Here we propose and develop the first national RLI applied to ecosystem level, using Finnish species and their habitats as an example. The approach is intended to complement both the taxon-based RLI and the ecosystem-based RLE, bridging the gap between the two.

2. Material and methods

2.1 Species data

There are approximately 45 000 known species in Finland, and about 21 400 of these had adequate data for threat assessments both in 2000 and 2010 (Rassi et al. 2001, 2010). The present study is based on 10 131 taxa assessed in both years, (Table 1), as we restricted the analyses to species groups well covered in both assessments: beetles (3 384 species), butterflies & moths (below denoted as butterflies) (2 247),

lichens (1 392), vascular plants (1 197), bryophytes (873), true bugs (463), birds (237), polypores (220), mammals (57), dragonflies and damselflies (51) and herptiles (10).

As a part of the method, back casting was used to identify the species with genuine threat category changes. The 2000 Red List categories were adjusted retrospectively based on current information and taxonomy when needed. The RLI calculations include only category changes due to genuine changes in species statuses (Butchart et al. 2007). Back casting was performed already for species groups other than Lepidoptera by Juslén et al. (2013). The reasons for any category change are listed in Rassi et al. 2010 for the species in threatened categories regionally extinct (RE), critically endangered (CR), endangered (EN), vulnerable (VU), near-threatened (NT) and data deficient (DD). The working documentation lists reasons for the Least Concern (LC) species. Any challenging back casting cases were separately discussed with experts of the group in question. Regarding Lepidoptera, LK and JK have made the back casting purposely for the study now presented. Altogether 529 genuine changes were found in the 12 groups studied (Table 1).

2.2. Habitat data

The habitats for species listed in the Finnish Red Data Book (i.e. for those categorized as RE, CR, EN, VU, and NT) were published by Rassi et al. (2010). For LC species we followed the unpublished habitat classification listed at the threat assessment documentation or other working documentation produced by expert groups during two years (except beetles and butterflies, for which no classification was produced previously).

The habitat classification categories were: forests, mires, aquatic habitats, shores, rock outcrops (including erratic boulders), alpine heaths and meadows above tree-level, and rural biotopes and cultural habitats.

Definitions of the habitats are given in table 2, and more detailed subcategorizations are published in Rassi et al. 2010. These differ from the standard classifications by IUCN (http://www.iucnredlist.org/technical-documents/classification-schemes/habitats-classification-scheme-ver3) in two ways (see also Tables 2 and 5). First, mires were separated from other aquatic habitats due to their exceptional extension in Finland and importance for many Finnish species. Second, marine intertidal and coastal areas were merged due to the difficulty in separating them given the characteristics of Finnish geology and marine hydrology.

The habitat classification for Least Concern beetles and butterflies was conducted in this study. Habitats of the Least Concern species of Coleoptera were based on published sources (Koch 1989a, 1989b, 1992) and checked by Jaakko Mattila and Jyrki Muona. Besides own expertise, we used a database consisting of 670 000 observations of beetles in Finland. This database is not public, but the Finnish Coleoptera Atlas based on the database has been published (The Finnish Expert Group on Coleoptera 2010). The habitats of the least concern species of Lepidoptera were defined by experts Lauri Kaila and Jaakko Kullberg, who also had a database of Lepidoptera of 1.600 000 observations supporting their work (Hyönteistietokanta).

Additionally, a few missing habitats for the other ten groups of organisms were obtained with the help of the Finnish expert groups of species. The whole habitat classification data per species is given in Appendix 1.

Often species occur and establish sustaining populations on several habitat types. Yet, one habitat could always be pointed out by experts as the primary habitat type. This might be the original habitat of the species, for instance, *Thymus serpyllum* is classified to forests, as its original main habitat in Finland is esker forests (Hämet-Ahti et al. 1998), although it nowadays also occurs on sandy riverbanks and sometimes on sandy road banks. Or it might be the habitat where the species occurs in higher abundance. For high-mobility animals, that may occur in different habitats seasonally or during their life cycles, e.g. birds, the primary habitat was the preferential nesting habitat. Habitats of holomethabolic insects were defined according to the larvae preference, as most of their life-cycle is spent on this stage.

2.3. The Red List Index for habitats

Based on the red-list status of species occupying each habitat, we calculated the RLI for habitats.

The RLI value was calculated by multiplying the number of taxa in each red-list category by the category weight (0 for LC, 1 for NT, 2 for VU, 3 for EN, 4 for CR, 5 for RE/EX). These products were summed and then divided by the number of taxa multiplied by the maximum weight 5 ("maximum possible denominator"). To obtain the RLI value, this sum is subtracted from 1. The index value varies between 0 and 1 (Butchart et al. 2007). The lower the value, the closer the set of taxa is heading towards extinction. If the value is 0 all the taxa are (regionally) extinct. If the value is 1 all the taxa are assessed as Least Concern. The

instructions for national and regional use by Bubb et al. (2009) exclude the species that have been assessed as Extinct (EX) in the earlier assessment. We calculated the RLIs including the taxa assessed as Regionally Extinct (RE) in 2000, as some of these taxa were rediscovered in Finland during the observed period (see also Juslén et al. 2013).

2.4. Statistical analysis

We conducted independent analyses with different species groupings by taxon, by habitat and a combination of these. For each group of species in the three groupings we calculated three values: RLI 2000, RLI 2010 and the change between the years (i.e. RLI 2010 – RLI 2000). A simple arithmetic analysis would not show whether the group indices were statistically different or the change between the years was significantly different from a null hypothesis of no change. We therefore resampled all the values with non-parametric bootstrapping. For each group, species were randomly sampled with replacement until the original number of species was attained. For each of the 10.000 resampling events the RLI 2000, RLI 2010 and the respective differences were calculated. The confidence limits (α = 0.05) of the RLI values per group and year were the 2.5 and 97.5 percentiles of the respective 10.000 randomizations. The change between the years was considered statistically significant if more than 95% of the randomization values had the same sign (either increase or decrease) as the true values. Statistics were performed using the R 3.1.2 statistical environment (R Core Team 2014).

3. Results

The number of taxa in different primary habitats was 4 031 in forests, 513 in mires, 633 in aquatic habitats, 1 257 in shores, 969 in rock outcrops, 411 in alpine heaths and meadows, and 2 317 in rural biotopes (Table 3).

The RLI value for all Finnish species combined was 0.882 in 2000 and 0.879 in 2010. The minor changes observed against Juslén et al. (2013) were due to the inclusion of Lepidoptera in the dataset. The new bootstrap analyses showed that dragonflies, true bugs and beetles were statistically less threatened than the other groups, whose confidence limits mostly overlap (Fig. 1). The RLI changes between the years

were significantly negative for bryophytes, lichens, vascular plants, butterflies and birds and positive for beetles and true bugs (Table 4). Dragonflies, herptiles, mammals and polypores show no significant trend.

Alpine habitats followed by rock outcrops present the most threatened species on average, with aquatic habitats, forests and mires hosting the least threatened (Fig. 2). The RLI changes between the years were significantly negative for all habitat types except forests and rural biotopes, which show no significant trends (Table 5).

Significant RLI trends between 2000 and 2010 were found for 20 combinations of groups of organisms with primary habitats (Table 6; Appendix 2). The trends of beetles and true bugs were positive in rural and forest habitats, otherwise observed trends were all negative. Trends of bryophytes were negative in six habitats and of vascular plants in five. Negative trends were also recovered in two habitats for both birds and lichens, and in one habitat for butterflies. In dragonflies and damselflies, polypores, herptiles and mammals no positive or negative trends were observed (Appendix 2).

4. Discussion

This study demonstrates that it is useful to calculate the RLI for species grouped by habitat, in addition to the usual taxonomic grouping. Several trends were revealed in accordance with published studies focusing on particular taxa and habitats. In general, more negative trends were found, with positive trends being possibly due to the effects of climate warming on several insect species that are expanding northwards. Few scientific papers analyzing reasons for population changes among the Finnish threatened species other than birds exist. Only in one habitat type (forests) several papers focused on recent trends in threatened species were available, such as the simulation study by Fedrowitz et al. (2012) showing continuous decrease of threatened epiphytic lichens. Our main findings, grouped by habitats, are elaborated in the table 5 with likely drivers and references with supporting notes.

We suggest that the habitat-based RLI may show a different, complementary view to the ecosystem-based RLE. Even though some habitats may not be improving, their constituent species may show positive trends

due to other factors such as the climate change. The habitat-based RLI clearly bridges the gap between the taxon-based RLI and the RLE.

The RLI has been used in multiple ways, usually to evaluate the impact of contrasting policies on the threat status of different taxonomic groups. Hoffmann et al. (2011) used it to attempt to quantify the impact of conservation efforts on the extinction risk of two groups of mammals. Young et al. (2014) quantified the impact of a conservation organisation's programmes on extinction risk of a set of species. Visconti et al. (2015) used the RLI for projecting the likely impact of different policy decisions. Moreno Saiz et al. (2015) tested it as a tool to assess the success of national conservation policies.

The latter authors recommended using various indicators as basis for planning regional conservation measures and evaluating their success. However, they also listed several challenges in using and interpreting the RLI. Above all, they recognize it is a summary statistic, which may mask the individual patterns under a global trend. For example, if 10 species increase and 10 decrease in their status the index will reveal the exact same value as if no species change at all, although these are quite different situations. Researchers and stakeholders should therefore always search for individual species that may be at odds with the general trend of the group and try to understand why this might happen. Although this is also verified in the present study, our results show the RLI to be useful for evaluating species trends in different habitat types.

As mentioned, besides the RLI other indices can be disaggregated into different groups so that different aspects of biodiversity change can be studied. These might be taxonomic groups (the subject of most RLI studies), habitat types (the subject of this study), or many other. Dividing species into functional groups may be a particularly useful way of using the RLI, as function is related with ecosystem services and thus trends in particular groups may reveal or even precede changes in services, many of them critical for human well-being.

References 218 219 Albrecht A, Mattila K, Rinne V, Rintala T, Söderman G (2010) Hemiptera. In: Rassi P, Hyvärinen E, Juslén 220 A, Mannerkoski I. (eds) The 2010 red list of Finnish species. Ympäristöministeriö and Suomen 221 ympäristökeskus, Helsinki, pp 397-416 222 Baillie JEM, Collen B, Amin R, Akçakaya HR, Butchart SHM, Brummitt NA, Meagher TR, Ram M, Hilton-223 Taylor C, Mace GM (2008) Toward monitoring global biodiversity. Conservation Letters 1: 18–26 224 225 Brommer J, Lehikoinen A, Valkama J (2012) The Breeding Ranges of Central European and Arctic Bird 226 Species Move Poleward. PloS One (Article: e43648). DOI: 10.1371/journal.pone.0043648 227 228 Brummitt N, Bachman SP, Aletrari E, Chadburn H, Griffiths-Lee J, Lutz M, Moat J, Rivers MC, Syfert MM, 229 Nic Lughadha EM (2015) The sampled red list index for plants, phase II: ground-truthing specimen-based 230 conservation assessments. Philos T Roy Soc B 370 (Article: 20140015). DOI: 10.1098/rstb.2014.0015 231 232 Bubb PJ, Butchart SHM, Collen B, Dublin H, Kapos V, Pollock C, Stuart SN, Vié JC (2009) IUCN Red List 233 index—guidance for national and regional use. IUCN, Gland 234 235 Butchart SHM, Stattersfield AJ, Bennun LA, Shutes SM, Akçakaya HR, Baillie JEM, Stuart SN, Hilton-236 Taylor C, Mace GM (2004) easuring global trends in the status of biodiversity: Red List Indices for birds. 237 PLOS Biology (Article: e383). DOI:10.1371/journal.pbio.0020383 238 239 Butchart SHM, Akçakaya HR, Chanson J, Baillie JEM, Collen B, Quader S, Turner WR, Amin R, Stuart SN, 240 Hilton-Taylor C (2007) Improvements to the Red List index. PLOS One (Article: e140). DOI: 241 10.1371/journal.pone.0000140. 242 243

Butchart SHM et al (2010) Global biodiversity: indicators of recent declines. Science 328: 1164-1168

Cardoso P, Borges PAV, Triantis KA, Ferrández MA, Martín JL (2011) Adapting the IUCN Red List criteria
for invertebrates. Biol Conserv 144: 2432–2440
Cardoso P, Borges PAV, Triantis KA, Ferrández MA, Martín JL (2012) The underrepresentation and
misrepresentation of invertebrates in the IUCN Red List. Biol Conserv 149: 147–148
Observation and According to the Control of the Con
Clausnitzer V et al. (2009) Odonata enter the biodiversity crisis debate: The first global assessment of an
insect group. Biol Conserv 142: 1864–1869
Oallan D. Lah. L. Ma-Daa L. Militaraa O. Aasia D. Daillia (FNA (0000) Manitaria na hanna in cantabanta
Collen B, Loh J, McRae L, Whitmee S, Amin R, Baillie JEM (2009) Monitoring change in vertebrate
abundance: the Living Planet Index. Conservation Biology 23: 317-327
Fedrowitz K, Kuusinen M, Snäll T (2012) Metapopulation dynamics and future persistence of epiphytic
cyanolichens in a European boreal forest ecosystem. J Appl Ecol 49: 493-502
The Finnish Expert Group on Coleoptera (2010) The Finnish Expert Group on Coleoptera 2010. Atlas of the
Beetles of Finland. http://koivu.luomus.fi/elaintiede/kovakuoriaiset/catlas1.htm.
Gärdenfors U (ed.) (2010) Rödlistade arter i Sverige 2010 - the 2010 red list of Swedish species.
ArtDatabanken, Uppsala
Hanski I (2000) Extinction debt and species credit in boreal forests: modelling the consequences of
different approaches to biodiversity conservation. Ann Zool Fenn 37: 271–280
Hanski I (2005) The shrinking world: Ecological consequences of habitat loss. International Ecology
Institute, Oldendorf
Hautala H, Laaka-Lindberg S, Vanha-Majamaa I (2011) Effects of retention felling on epixylic species in
boreal spruce forests in southern Finland. Restor Ecol 19: 418–429

275	
276	Hedenås H, Hedström P (2007) Conservation of epiphytic lichens: Significance of remnant aspen (<i>Populus</i>
277	tremula) trees in clear-cuts. Biol Conserv 135: 388-395
278	
279	Heino J, Virtanen R, Vuori KM, Saastamoinen J, Ohtonen A, Muotka T (2005) Spring bryophytes in
280	forested landscapes: land use effects on bryophyte species richness, community structure and persistence.
281	Biol Conserv 124: 539–545
282	
283	Hoffman M et al. (2010) The impact of conservation on the status of the world's vertebrates. Science 330:
284	1503–1509
285	
286	Hoffman M, Belant JL, Chanson JS, Cox NA, Lamoreux J, Rodrigues ASL, Schipper J, Stuart SN (2011)
287	The changing fates of the world's mammals. Philos T Roy Soc B 366: 2598–2610
288	
289	Hyvärinen E, Kouki J, Martikainen P (2006) Fire and green-tree retention in conservation of red-listed and
290	rare deadwood-dependent beetles in Finnish boreal forests. Conserv Biol 20: 1710–1719
291	Hyönteistietokanta (http://hyonteiset.luomus.fi/insects/main/EntDatabase.html) [Entomological database].
292	Finnish Museum of Natural History, Helsinki, 27.10.2015
232	Tillingi Mascall of Nataral History, Ficiginal, 27. To.2010
293	Hämet-Ahti L, Suominen J, Ulvinen T, Uotila P (eds.) (1998) Retkeilykasvio, 4. uudistettu painos,
294	Luonnontieteellisen keskusmuseon kasvimuseo, Helsinki.
295	
296	Ikonen I (2011) Integrated coastal planning and management in southwest Finland. In: Reihmanis J (ed.)
297	Nordic-Baltic-Belarus solutions in farming for biodiversity. Latvijas Dabas Fonds, Riga, pp 34–41
298	
299	Ilmonen J, Mykrä H, Virtanen R, Paasivirta L, Muotka T (2012) Responses of spring macroinvertebrate and
300	bryophyte communities to habitat modification: community composition, species richness, and red-listed

species. Freshwater Science 31: 657-667

Johansson P (2008) Consequences of disturbance on epiphytic lichens in boreal and near boreal forests.
Biol Conserv 141: 1933–1944
Juslén A, Hyvärinen E, Virtanen LK (2013) Application of the Red-List Index at a national level for multiple
species groups. Conserv Biol 27: 398–406
Jutila H (2001) How does grazing by cattle modify the vegetation of coastal grasslands along the Baltic
Sea? Ann Bot Fenn 38: 181–200
Koch K (1989a) Die Käfer Mitteleuropas. Ökologie. Band 1. Krefeld, Goecke & Evers
Karla K (4000h) Dia Kitan Mittalawan an Öhalaria Dand O Krafald Oasala 9 Essas
Koch K (1989b) Die Käfer Mitteleuropas. Ökologie. Band 2. Krefeld, Goecke & Evers
Koch K (1992) Die Käfer Mitteleuropas. Ökologie. Band 3. Krefeld, Goecke & Evers
ricon richologica di anticio di cara di ricona, decente di Evene
Kontula T, Raunio A (2009) New method and criteria for national assessments of threatened habitat types
Biodivers Conserv 18: 3861–3876
Kouki J, Arnold K, Martikainen P (2004) Long-term persistence of aspen – a key host for many threatened
species – is endangered in old-growth conservation areas in Finland. J Nat Conserv 12: 41–52
Laaksonen T, Lehikoinen A (2013) Population trends in boreal birds: Continuing declines in agricultural,
northern, and long-distance migrant species. Biol Conserv 168: 99–107
Lewis OT, Senior MJM (2011) Assessing conservation status and trends for the world's butterflies: the
Sampled Red List Index approach. J Insect Conserv 15: 121–128

Loh J, Green RE, Ricketts T, Lamoreux JF, Jenkins M, et al. (2005) The Living Planet Index: using species population time series to track trends in biodiversity. Philosophical Transactions of the Royal Society of London B, 360: 289-295 Lopez L (2011) Estado de las Aves del Paraguay. Guyra Paraguay, BirdLife International. Asunción, Paraguay. Mace GM, Collar NJ, Gaston KJ, Hilton-Taylor C, Akçakaya HR, Leader-Williams N, Milner-Gulland EJ, Stuart SN (2008) Quantification of extinction risk: IUCN's system for classifying threatened species. Conserv Biol 22: 1424-1442 Martikainen P (2001) Conservation of threatened saproxylic beetles: significance of retained aspen Populus tremula on clearcut areas. Ecol Bull 49: 205-218 The Millenium Development Goals Report 2015. United Nations, New York. Moreno Saiz JC, Domínguez Lozano F, Marrero Gómez M, Bañares Baudet Á (2015) Application of the Red List Index for conservation assessment of Spanish vascular plants. Conserv Biol 29: 910-919 Olofsson J, Oksanen L (2005) Effects of reindeer density on vascular plant diversity on North Scandinavian mountains. Rangifer 25: 5-18 Penttilä R, Siitonen J, Kuusinen M (2004) Polypore diversity in a managed and old-growth boreal Picea abies forests in southern Finland. Biol Conserv 117: 271-273 Pykälä J (2000) Mitigating human effects on European biodiversity through traditional animal husbandry. Conserv Biol 14: 705-712

Pykälä J (2004) Effects of new forestry practices on rare epiphytic macrolichens. Conserv Biol 18: 831–838

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349

350

351

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353

354

355

356

357

359	
360	Pöyry J, Luoto M, Heikkinen RK, Kuussaari M, Saarinen K (2009) Species traits explain recent range shifts
361	of Finnish butterflies. Glob Change Biol 15: 732–743
362	
363	Pöysä H, Rintala J, Lehikoinen A, Väisänen, R.A., 2012. The importance of hunting pressure, habitat
364	preference and life history for populations trends of breeding waterbirds in Finland. Eur J Wildlife Res 59:
365	245–256
366	
367	Raatikainen KM, Heikkinen RK, Pykälä J (2007) Impacts of local and regional factors on vegetation of
368	boreal semi-natural grasslands. Plant Ecol 189: 155–173
369	
370	Rassi P, Alanen A, Kanerva T, Mannerkoski I (eds) (2001) Suomen lajien uhanalaisuus 2000.
371	Ympäristöministeriö & Suomen ympäristökeskus, Helsinki
372	
373	Rassi P, Hyvärinen E, Juslén A, Mannerkoski I (eds) (2010) Suomen lajien uhanalaisuus—Punainen kirja
374	2010. The 2010 red list of Finnish species. Ympäristöministeriö ja Suomen ympäristökeskus, Helsinki
375	
376	Raunio A, Schulman A, Kontula T, (eds) (2008) Assessment of threatened habitat types in Finland – part 1
377	Results and basis for assessment (in Finnish). Suomen ympäristö 8: 1–264
378	
379	R Core Team (2014) R: A language and environment for statistical computing. R Foundation for Statistical
380	Computing, Vienna, Austria. Available at http://www.R-project.org/
381	
382	Reinikainen A, Mäkipää R, Vanha-Majamaa I, Hotanen JP (eds) (2000) Kasvit muuttuvassa
383	metsäluonnossa. Metsäntutkimuslaitos & Kustannusosakeyhtiö Tammi, Helsinki
384	
385	Rodrigues ASL, Pilgrim JD, Lamoreux JF, Hoffmann M, Brooks TM (2006) The value of the IUCN Red List
386	for conservation. Trends Ecol Evol 21: 71–76

387	Rodríguez JP et al (2011) Establishing IUCN Red List criteria for threatened ecosystems. Conserv Biol 25:
388	21–29
200	Pandinini C. Pattistani A. Taefili C. 2014 La atata della Diadivaraità in Italia: l'applicazione dell'appraesio
389	Rondinini, C., Battistoni, A., Teofili, C. 2014. Lo stato della Biodiversità in Italia: l'applicazione dell'approccio
390	Sampled Red List e Red List Index
391	
392	Szabo JK, Butchart SHM, Possingham HP, Garnett ST (2012) Adapting global biodiversity indicators to the
393	national scale: a red list index for Australian birds. Biol Conserv 148: 61–68
394	
395	Tikkanen OP, Martikainen P, Hyvärinen E, Junninen K, Kouki J (2006) Red-listed boreal forest species of
396	Finland: associations with forest structure, tree species, and decaying wood. Ann Zool Fenn 43: 373–383
397	
398	Tittensor DP et al. (2014) A mid-term analysis of progress toward international biodiversity targets. Science
399	346: 241–244
400	
401	Toivanen T, Kotiaho JS (2007) Mimicking natural disturbances of boreal forests: the effects of controlled
402	burning and creating dead wood on beetle diversity. Biodivers Conserv 16: 3193–3211
403	
404	Virkkala R, Korhonen KT, Haapanen R, Aapala K (2000) Protected forests and mires in forest and mire
405	vegetation zones in Finland based on the 8th National Forest Inventory (In Finnish with an English
406	summary). The Finnish Environment 395: 1–49
407	
408	Virkkala R, Heikkinen RK, Leikola N, Luoto M (2008) Projected large-scale range reductions of northern-
409	boreal land bird species due to climate change. Biol Cons 141: 1343–1353
410	
411	Virkkala R, Rajasärkkä A (2011) Climate change affects populations of northern birds in boreal protected
412	areas. Biological Letters 7: 395–398
413	
414	Virkkala R, Rajasärkkä A (2012) Preserving species populations in the boreal zone in a changing climate:

contrasting trends of bird species groups in a protected area network. Nature Conserv 3: 1-20

416	
417	Visconti P et al (2015) Projecting global biodiversity indicators under future development scenarios.
418	Conserv Lett DOI: 10.1111/conl.12159
419	
420	Woinarski JCZ, Burbidgec AA, Harrison PL (2015) Ongoing unraveling of a continental fauna: Decline and
421	extinction of Australian mammals since European settlement. PNAS 112: 4531-4540
422	
423	Young RP, Hudson MA, Terry AMR, Jones CG, Lewis RE, Tatayah V, Zuël N, Butchart SHM (2014)
424	Accounting for conservation: Using the IUCN Red List Index to evaluate the impact of a conservation
425	organization. Biol Conserv 180: 84–96
426	
427	Figure captions
428	
429 430 431	Figure 1. The RLI trends between 2000 and 2010 showing the confidence limits for RLI values of each group of organisms.
432 433	Figure 2. The RLI trends between 2000 and 2010 showing the confidence limits for RLI values of each primary habitat.

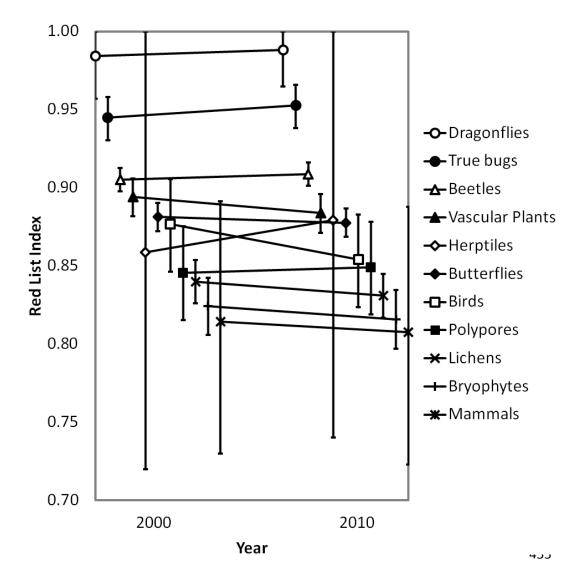


Figure 1.

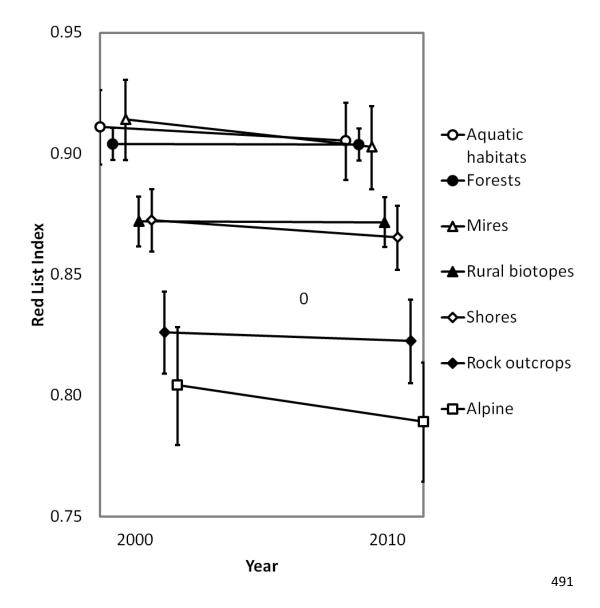


Figure 2.

Table 1. Number of species known in Finland (Total) by organism groups included in our study, number of taxa included in the red-list assessment of 2010, number of RE, CR, EN, VU, and DD taxa together in 2010, number of taxa excluded from the study because they were Data Deficient or not assessed in 2000 as not having an established population, number of taxa included in the present study and those that genuinely changed red-list category between 2000 and 2010.

Organism group	Total	Assessed (% total)	RE, CR, EN, VU, NT, or DD (% assessed)	Excluded as Data Deficient or other reasons (% assessed)	Included (% assessed)	Genuinely changed (% included)
Beetles (Coleoptera)	3 697	3 416 (92.4)	737 (21.6)	32 (0.9)	3 384 (99.1)	138 (4.1)
Birds (Aves) Bryophytes (Bryophyta,	249	241 (96.8)	89 (36.9)	4 (1.7)	237 (98.3)	66 (27.8)
Marchantiophyta and Anthocerophyta)	906	896 (98.9)	364 (40.6)	23 (2.6)	873 (97.4)	35 (4.0)
Butterflies (Lepidoptera)	2 576	2 313 (89.8)	707 (30.6)	66 (2.9)	2 247 (97.1)	130 (5.8)
Dragonflies and damselflies (Odonata)	55	52 (94.5)	1 (1.9)	1 (1.9)	51 (98.1)	1 (2.0)
Herptiles (Reptilia and Amphibia)	12	10 (83.3)	3 (30.0)	0	10 (100)	1 (10.0)
Lichens (Lichenes)	1 832	1 545 (84.3)	686 (44.4)	153 (9.9)	1 392 (90.1)	59 (4.2)
Mammals (Mammalia) Polypores	72	59 (81.9)	22 (37.3)	2 (3.4)	57 (96.6)	4 (7.0)
(Aphyllophorales and Heterobasidiomycetes)	237	225 (94.9)	95 (42.2)	5 (2.2)	220 (97.8)	9 (4.1)
True bugs (Heteroptera)	506	469 (92.7)	64 (13.6)	6 (1.3)	463 (98.7)	19 (4.1)
Vascular plants (Tracheophyta)	ca. 3 550	1 206 (40.0)	334 (27.7)	9 (0.7)	1 197 (99.3)	67 (5.6)
All species	ca. 13 692	10 432 (76.2)	3 102 (29.7)	304 (2.9)	10 131 (97.1)	529 (5.2)

Table 2. Habitat classification used in Finnish Red Data Books 2001 and 2010 (Rassi et al. 2001: 2010)

Habitat	Additional explanation	Corresponding IUCN habitat
Alpine	Alpine heaths and meadows above tree-level	Native grassland
Aquatic habitats	Baltic Sea, lakes and ponds, small ponds, rivers, brooks and streams, rapids, spring complexes	Wetlands
Aquatio Habitato	Rich fens, fens, pine mires, spruce	Wetlands (subcategory: bogs,
Mires	mires	marshes, swamps, fens, peatlands
Forests	Heath forests, herb-rich forests, mountain birch forests	Forests
Rock outcrops	Rock outcrops, including erratic boulders	Inland rocky areas
Rural biotopes and cultural habitats	Seminatural grasslands, wooded pastures and pollard meadows, ditches, arable land, parks, yeards, gardens, roadsides, railway embankments, buildings	Artificial
	Shores of the Baltic Sea, lake shores	Marine/Intertidal and Marine
Shores	and river banks	Coastal/Supratidal

Table 3. The number of taxa in different primary habitats used in the study.

Organism group	Alpine heaths and meadows	Aquatic habitats	Forests	Mires	Rock outcrops	Rural biotopes and cultural habitats	Shores	All habitats
Beetles (Coleoptera)	889	285	1559	72	1	33	545	3 384
Birds (Aves)	36	56	78	20	2	20	25	237
Bryophytes (Bryophyta,								
Marchantiophyta and Anthocerophyta)	81	83	138	123	269	108	71	873
Butterflies (Lepidoptera)	688	0	1 143	137	27	54	198	2 247
Dragonflies and damselflies (Odonata)	0	46	0	5	0	0	0	51
Herptiles (Reptilia and Amphibia)	1	2	5	0	0	0	2	10
Lichens (Lichenes)	57	3	537	17	600	79	99	1 392
Mammals (Mammalia)	12	7	32	1	0	2	3	57
Polypores (Aphyllophorales and Heterobasidiomycetes)	15	0	198	0	0	0	7	220
True bugs (Heteroptera)	191	44	138	9	1	2	78	463
Vascular plants (Tracheophyta)	346	107	203	129	69	114	229	1 197
All species	2 316	633	4 031	513	969	412	1257	10 131

Table 4. The RLI in 2000 and 2010 and respective change in different groups of organisms and statistical significance of this change.

Group	RLI 2000	RLI 2010	Change	p-value
Beetles	0.905	0.909	0.003	<0.001
Birds	0.877	0.854	-0.023	0.012
Bryophytes	0.824	0.816	-0.008	< 0.001
Butterflies	0.881	0.878	-0.004	0.005
Dragonflies	0.984	0.988	0.004	0.372
Herptiles	0.859	0.879	0.020	0.342
Lichens	0.840	0.831	-0.009	< 0.001
Mammals	0.814	0.807	-0.007	0.224
Polypores	0.846	0.849	0.004	0.144
True bugs	0.945	0.953	0.008	0.001
Vascular Plants	0.894	0.884	-0.010	<0.001

Table 5. The RLI changes between 2000 and 2010 in different primary habitats (Finnish Red Data Book classification and IUCN Habitat classification) and the statistical significance, statistically significant changes in different organism groups and habitat combinations; and the likely drivers behind the RLI trends shown in the study with supporting notes and references.

Table 6. The changes of RLI for 11 groups of organisms in different primary habitats between 2000 and 2010. Statistically significant combinations are marked with asterisks (* = p < 0.05, ** = p < 0.01, *** = p < 0.001).

Group	Alpine	Aquatic	Forests	Mires	Rock	Rural	Shores
Beetles	0	-0.003	0.006***	-0.003	0	0.004*	0
Birds	-0.04	-0.028	-0.002	-0.08*	0	0	-0.048*
Bryophytes	-0.02***	-0.012**	-0.007**	-0.008**	-0.003*	0	-0.02***
Butterflies	-0.015	0	-0.001	-0.007	0.007	-0.003	-0.016**
Dragonflies	0	0.004	0	0	0	0	0
Herptiles	0	0	0	0	0	0	0.101
Lichens	-0.003	0	-0.017***	-0.012	-0.004***	-0.007	0
Mammals	0	0	-0.006	0	0	-0.017	0
Polypores	0	0	0.003	0	0	0.013	0
True bugs	0	0.005	0.013***	0	0	0.008*	0
Vascular Plants	-0.019***	-0.006*	-0.006	-0.014***	-0.006	-0.009***	-0.014***

Appendix 1.

The species included in the study, their main habitats and the IUCN threat classification in 2000 (backcasted) and 2010.

Appendix 2.

RLI values for all combinations of taxonomic groups and habitat types (Appendix) are available online. The authors are solely responsible for the content and functionality of these materials. Queries (other than absence of the material) should be directed to the corresponding author.

^{*}p < 0.05 ** p < 0.01

^{***} p < 0.001