

1 **The role of nutrients, productivity, and climate in determining tree fruit**
2 **production in European forests**

3 **Authors:** Marcos Fernández-Martínez ^{*ab}, Sara Vicca ^c, Ivan A. Janssens ^c, Josep Maria
4 Espelta ^b, Josep Peñuelas ^{ab}.

5

6 **Addresses:**

7 ^a CSIC, Global Ecology Unit, CREAM-CSIC-UAB, Cerdanyola del Vallés 08193, Catalonia,
8 Spain

9 ^b CREAM, Cerdanyola del Vallès 08193, Catalonia, Spain.

10 ^c Centre of Excellence PLECO (Plant and Vegetation Ecology), Department of Biology,
11 University of Antwerp, 2610 Wilrijk, Belgium.

12

13 ***Corresponding author:**

14 Marcos Fernández-Martínez

15 Tel: +34 935 814 851

16 E-mail address: m.fernandez@creaf.uab.cat

17

18

Post-print of Fernández-Martínez, Marcos et al. «The role of nutrients, productivity, and climate in determining tree fruit production in European forests» in The New Phytologist (Wiley), Vol. 213, Issue 2 (January 2017), p. 669-679. The final version is available at DOI [10.1111/nph.14193](https://doi.org/10.1111/nph.14193)

19 Summary

20• Fruit production (NPP_f), the amount of photosynthates allocated to reproduction ($\%GPP_f$) and
21their controls for spatial and species-specific variability (e.g., nutrient availability, climate) are
22poorly studied in forest ecosystems. We characterised fruit production and its temporal
23behaviour for several tree species and resolved the effects of gross primary production (GPP),
24climate, and foliar nutrient concentrations.

25• We used data for litterfall and foliar nutrient concentration from 126 European forests and
26related it to climatic data. GPP was estimated for each forest using a regression model.

27• Mean NPP_f ranged from approximately 10 to 40 g C m⁻² y⁻¹ and accounted for 0.5-3% of the
28GPP. Forests with higher GPPs produced larger fruit crops. Foliar zinc (Zn) and phosphorus
29(P) concentrations were associated positively with NPP_f , while foliar Zn and K were negatively
30related to its temporal variability. Maximum NPP_f and interannual variability of NPP_f was higher
31in *Fagaceae* than in *Pinaceae* species.

32• NPP_f and $\%GPP_f$ were similar amongst the studied species despite the different reproductive
33temporal behaviour of *Fagaceae* and *Pinaceae* species. We report, for the first time, that foliar
34concentrations of P and Zn play an important role in determining $\%GPP_f$, NPP_f and its temporal
35behaviour.

36**Keywords:** crop size, fertility, fruiting, reproductive effort, seeds, spatial variability

371. Introduction

38Reproductive behaviour in plant species is a crucial part of ecosystem functioning, and,
39therefore, patterns of production of fruit has received much attention from a wide range of
40ecologists (Koenig & Knops, 2005). *Masting* is a reproductive behaviour, presented mostly by
41anemophilous and long-lived plant species, consisting on the synchronised alternation of years
42with extremely large fruit crops and years with little or no fruit production (Kelly & Sork, 2002).
43Because of the consequences that masting can entail on the ecosystems (e.g., important
44cascading effects throughout the food web (Ostfeld & Keesing, 2000)), a large array of studies
45have tried to understand how and why this extreme interannual variability in fruit production
46happens (Fernández-Martínez *et al.*, 2016b). Nonetheless, it remains unclear why some
47species or populations produce larger fruit crops or have more temporally variable fruit
48production than others. Most theories have focused on the ultimate causes (e.g., avoiding
49seed predation) for these particular life-history traits (Silvertown, 1980; Kelly, 1994), but the
50proximate causes (e.g., the mechanistic effect of temperature on pollination) of the spatial
51variability in fruit production and its temporal behaviour have rarely been explored. Some
52authors have suggested that the size of the fruit crop from trees subjected to poor growing
53conditions (e.g. extreme or highly variable climates or poor soil properties such as low nutrient
54availability) should exhibit higher interannual variability (Kelly & Sork, 2002) because of the
55benefits of the *economy of scale* (Norton & Kelly, 1988; Smith *et al.*, 1990). However, little
56empirical evidence has yet been found to confirm that high interannual variability in fruit
57production is associated with poor growing conditions, such as drought (Fernández-Martínez
58*et al.*, 2012) or nutrient limitations.

59Most of the studies focusing on fruit or seed production have quantified the number of fruits
60produced per tree (Sork *et al.*, 1993; Espelta *et al.*, 2008; Crone *et al.*, 2011; Fernández-
61Martínez *et al.*, 2015), and only few have provided data in units of mass produced per area
62(Campioli *et al.*, 2010; Herbst *et al.*, 2015; Pérez-Ramos *et al.*, 2015). This is essential for
63elucidating the contribution of reproduction to ecosystemic net primary production (NPP),
64which is the sum of aboveground (wood, leaves, fruits) and belowground (roots) biomass
65production, root exudates and the emission or organic volatile compounds. The lack of a
66quantitative assessment of C allocation to fruit production has been an important drawback for
67correctly estimating fruit NPP (NPP_f) and total biomass production in, for instance, synthesis

68studies on plant carbon allocation (e.g. Vicca *et al.*, 2012; Campioli *et al.*, 2015a). Given that
69the mean residence time of carbon is very dependent on the compartment where carbon is
70allocated to (Fernández-Martínez *et al.*, 2016a), measures of NPP_f and the percentage of
71gross primary production (GPP, which equals net ecosystem photosynthesis) allocated to
72reproduction would facilitate carbon balance closure in forest ecosystems and would improve
73in-depth understanding of plant strategies under different environmental conditions and
74responses to changes in e.g., climate and nutrient availability.

75Endogenous factors such as stand age, basal area, or taxonomical family and exogenous
76factors such as climate, nutrient availability, and anthropogenic impacts have been
77demonstrated to largely control GPP, NPP and carbon allocation to different NPP fractions at
78the global scale (Vicca *et al.*, 2012; Campioli *et al.*, 2015), and different components of NPP
79(e.g. foliage, stems, or roots) may be controlled by different factors (Fernández-Martínez *et al.*,
802014a). These endogenous and exogenous factors may thus also have an effect on NPP_f .
81Nutrient availability, in particular, may play an important role in determining NPP_f and its
82temporal behaviour, because fruits and seeds are nutrient-richer (mostly in nitrogen [N],
83phosphorus [P], and potassium [K]) than vegetative tissues (Reekie & Bazzaz, 1987). Fruit
84production may thus have stronger relationships with nutrient availability than vegetative
85tissues (Sala *et al.*, 2012). Nitrogen, for example, is the primary limiting nutrient for vegetative
86growth (Elser *et al.*, 2007; LeBauer & Treseder, 2008) and has also been positively correlated
87with investment in reproduction (Han *et al.*, 2013; Miyazaki *et al.*, 2014). Nitrogen deposition
88has been suggested to increase forest productivity (Luyssaert *et al.*, 2010; de Vries *et al.*,
892014), but when combined with sulphur (S) deposition, the acidification of soils can have a
90negative impact on productivity and soil processes (Oulehle *et al.*, 2011; Büntgen *et al.*, 2013).
91Plants growing in P-deficient environments may also have a lower investment in sexual
92reproduction than those living in P-rich environments (Fujita *et al.*, 2014). Other macro- and
93micronutrients, often rarely studied in wild plants, may also play an important role in
94determining fruit production and its temporal behaviour (Swietlik, 2002).

95While elemental concentrations are important, the ratios between the various elements may be
96even more relevant. The foliar N:P ratio was suggested as an important indicator of plant
97nutritional status and vegetative (Güsewell, 2004; McGroddy *et al.*, 2004; Sardans *et al.*, 2012,
982016b) and reproductive (Fujita *et al.*, 2014) production. For example, species with low foliar

99N:P ratios tend to be fast-growing and more competitive for soil resources than species with
100higher N:P ratios (Willby *et al.*, 2001; Elser *et al.*, 2003; Peñuelas *et al.*, 2013). Other studies
101have suggested that, within the same species, plants or shoots producing flowers may have
102lower N:P ratios than those that do not produce flowers (Eckstein & Karlsson, 1997; Güsewell,
1032004). Other elemental stoichiometries, such as C:N or C:P ratios, may also indicate
104limitations of soil nutrients (Fernández-Martínez *et al.*, 2014b) and may therefore also be
105associated with reproductive behaviour in plants.

106The general aim of this study was to characterise fruit production and its temporal behaviour
107for several of the most abundant European tree species and to distinguish species-specific
108variability from the effects of the taxonomical family, productivity, foliar nutrient concentrations,
109climate, and atmospheric deposition of N and S. We specifically aimed i) to estimate average
110NPP_f and the percentage of GPP (hereafter %GPP_f) allocated to fruit production of various tree
111species distributed across Europe, ii) to parameterize masting by assessing the temporal
112variation of NPP_f of these species by calculating the coefficient of variation (CV), the
113consecutive disparity index (D), and the temporal autocorrelation at a one-year time lag (AR1:
114the correlation between fruit production in year y and year $y-1$), and iii) to identify the possible
115determinants of the spatial and species-specific variability of fruit production and its temporal
116behaviour in relation to GPP, foliar nutrient (C, N, P, K, S, Fe, Mg, Ca, Zn, Mn, and Cu)
117concentrations and C:N:P stoichiometries (C:N, N:P, and C:P ratios), climate (mean annual
118temperature and precipitation and their interannual variabilities), and atmospheric deposition of
119N and S. We also explored the potential differences in fruiting behaviour between species of
120the *Pinaceae* and *Fagaceae* families.

1212. Materials and methods

1222.1. Data collection

1232.1.1. Data for litterfall and foliar nutrient concentrations

124We downloaded data for litterfall (Pitman *et al.*, 2010) and foliar nutrient concentrations (Rautio
125*et al.*, 2010) from the ICP Forests database (International Co-operative Programme on
126Assessment and Monitoring of Air Pollution Effects on Forest, operated under the UNECE
127Convention on Long-range Transboundary Air Pollution, <http://icp-forests.net/>), containing data
128for several forest species in Europe. Fruit and foliar litterfall for each site were summed over

129entire years, and foliar nutrient concentrations (C, N, P, K, S, Fe, Mg, Ca, Zn, Mn, and Cu)
130were averaged per site. Average foliar C:N:P stoichiometries per site were calculated on a
131mass basis. We used foliar nutrient concentrations instead of soil nutrient availability as the
132measure of nutrient availability for plants because foliar nutrient concentrations often better
133reflect nutrient availability for plants than nutrient concentration of soils because sometimes
134nutrients in soils are not available for plants because of multiple factors such as
135bioimmobilization (Aber & Melillo, 1982) or complex formation (related to too high or too low
136pH; Truog, 1947; Comerford, 2005).

137The original data for litterfall were available in units of g of dry weight $\text{m}^{-2} \text{y}^{-1}$ and we used
138carbon concentration data (provided by the same database) to convert these data to units of g
139C $\text{m}^{-2} \text{y}^{-1}$. We also used site information such as mean site diameter to calculate mean basal
140area, as an estimation of the mean size of the trees at each site. To avoid eventual artefacts of
141anomalous years dominating too short data sets, of the 210 forests originally available, we
142used only 126 forests with time series with four or more years of data (**Figure S1**). We also
143excluded *Ceratonia siliqua* and *Larix decidua* because of the scarcity of replicates (two and
144one forest, respectively), as well as 12 multi-species forests. We provided summary values for
145*Pinus nigra*, but did not include this species in the statistical models, because only three
146replicates were available. Because of missing data for some of the predictor variables, only 97
147forests were finally used to fit the statistical models.

1482.1.2. Climatic data

149We extracted mean annual temperature and precipitation data (MAT and MAP, respectively) for
150our forests from the WorldClim database (Hijmans *et al.*, 2005). This database provides
151climatic data with a high spatial resolution (30 arc seconds, ca. 1 km at the equator) and
152contains robust mean monthly climatic data derived from lengthy time series (1950-2000).

1532.1.3. GPP data and the calculation of the NPP_f -to-GPP percentage

154We aimed to understand the relative reproductive investment of trees (relative to GPP), but the
155ICP forest network did not measure GPP. We therefore developed and tested a model that
156predicts average GPP from independent drivers. We used a global forest database (Luyssaert
157*et al.*, 2007; Fernández-Martínez *et al.*, 2014b) containing carbon fluxes and productivity data
158from forests. We extracted data from sites that reported both mean GPP (mainly derived from

159 eddy covariance CO₂ fluxes) and foliar NPP, as well as MAT, MAP, and leaf type. We chose
160 this set of predictors because these variables were also available for each of the ICP-forest
161 sites used in this study, allowing the estimation of GPP for each site. The model, containing
162 data for 84 forests, accounted for 74% of the variance in GPP. We cross-validated the model
163 using 75% of the data as a sampling set and the remaining 25% as the testing set and
164 repeated this procedure 1000 times randomising both subsets. The cross-validation indicated
165 that the model performed acceptably well: average errors of the model predictions were lower
166 than 12% for the testing set. The model and cross-validation are summarised in the
167 Supplementary Material (**Figure S2** and **Section 2: Estimating GPP**). We calculated the
168 percentage of GPP allocated to NPP_f (hereafter %GPP_f) as: $\%GPP_f = 100 \times NPP_f \times GPP^{-1}$.

169 2.1.4. Atmospheric deposition data

170 We obtained data for atmospheric deposition from the EMEP gridded datasets and ICP
171 Forests. Modelled N and S deposition data from the EMEP gridded maps were correlated with
172 the ICP data for each forest ($R=0.32$ and 0.67 , $P < 0.001$, for N and S deposition,
173 respectively). We used the EMEP data throughout in our statistical models because the ICP
174 data were incomplete for our database.

175 2.2. Data analyses

176 We calculated all the average metrics of fruit production per site from the annual values per
177 site: average NPP_f, average %GPP_f, maximum NPP_f recorded in each forest, NPP_f CV, NPP_f D
178 (Martín-Vide, 1986), and NPP_f AR1. D was calculated in addition to CV because it was recently
179 suggested to improve the assessment of temporal variability relative to the CV, especially in
180 negatively autocorrelated time series. Further information on the D index can be found in the
181 Supplementary Material, **Section 1: The consecutive disparity index (D)**. AR1 indicates the
182 temporal correlation between fruit crop sizes for years $y-1$ and y . Negative AR1 values in fruit
183 production time series may be a sign of depletion of resources after large fruit crops (Sork *et*
184 *al.*, 1993; Fernández-Martínez *et al.*, 2015). These values of fruit CV, D and AR1 were further
185 averaged across sites per species to characterise the temporal behaviour of NPP_f of that
186 species. Using average site values as replicates, we identified differences amongst species
187 and families using ANOVAs and Tukey's HSD tests for multiple comparisons.

188 We fitted linear mixed models (LMMs) to characterise the relationships of the exogenous (MAT,
189 MAP, and N and S atmospheric deposition) and endogenous (foliar nutrient concentrations,
190 GPP, basal area, and family – *Pinaceae* or *Fagaceae*) variables with the temporal behaviour of
191 fruit production (mean NPP_f , $\%GPP_f$, maximum NPP_f , and the CV, D, and AR1 of NPP_f). Using
192 the species as a random effect (to take into account, for example, the species effect on
193 stoichiometry), we tested whether the fruit production metrics were correlated with productivity
194 and with forest characteristics, such as GPP, basal area, family of the species, climate, foliar
195 nutrient concentrations, and atmospheric deposition of N and S. The saturated model used
196 GPP, basal area, N and S deposition, MAT, MAP, the CVs of MAT and MAP, foliar
197 concentrations of C, N, P, K, S, Fe, Ca, Cu, Mg, and Zn, and foliar C:N, N:P, and C:P ratios as
198 fixed effects. We used D of MAT and MAP instead of the CV for the models predicting D of
199 NPP_f . Variables were selected using the backwards-forwards stepwise method, starting from
200 the saturated model. Once the most important variables were selected, we calculated the first-
201 order interactions of the variables within the model and removed the non-significant terms.
202 Models were fitted with the *lme* function of the R package “nlme” (Pinheiro *et al.*, 2013) using
203 restricted maximum likelihood and a Gaussian distribution. To provide a measure of goodness
204 of fit of our models we assessed the marginal (fixed factors) and conditional (fixed + random
205 factors) variance explained by the model (pseudo R^2) using the method proposed by
206 Nakagawa & Schielzeth (2013) implemented in the MuMIn R package (Barton, 2015). We also
207 used generalised additive models (GAMs) to explore the non-linearities between the response
208 variables and the predictors using spline functions. NPP_f and D of NPP_f were log-transformed
209 to meet the assumption of normality and heteroscedasticity in the model residuals. All analyses
210 were performed using R statistical software (R Core Team, 2015).

2113. Results

2123.1. Species-specific variability in fruit production and its temporal behaviour

213 Mean fruit production amongst the studied species ranged from (mean \pm standard error) $6.1 \pm$
214 $1.7 \text{ g C m}^{-2} \text{ y}^{-1}$ in *Pseudotsuga menziesii* to $40.6 \pm 9.9 \text{ g C m}^{-2} \text{ y}^{-1}$ in *Pinus nigra* (**Figure 1**,
215 **Table S1**). In fact, mean NPP_f per year varied little amongst species and differed significantly
216 only between the two above species and between *Pinus sylvestris* and *Quercus petraea*
217 (**Figure 1**). Nonetheless, when grouped *Pinaceae* species had a higher average NPP_f than
218 *Fagaceae* species (23.2 ± 2.1 and $15.0 \pm 1.6 \text{ g C m}^{-2} \text{ y}^{-1}$, respectively; ANOVA, $P = 0.003$).

219 Similarly, the percentage of GPP invested in NPP_f (%GPP_f) varied little amongst species
220 (**Figure 1**) and, on average, was <3% in all species (**Table S1**). The only differences amongst
221 species were found between *P. sylvestris*, *P. menziesii*, and *Q. petraea*, with %GPP_f higher in
222 *P. sylvestris* than in the other two species. On the other hand, %GPP_f was different between
223 families, being higher for *Pinaceae* than *Fagaceae* forests (1.7 ± 0.2 and $1.1 \pm 0.1\%$,
224 respectively; ANOVA, $P = 0.004$). Maximum NPP_f was similar across species and averaged at
225 $46.9 \pm 3.1 \text{ g C m}^{-2} \text{ y}^{-1}$ (**Figure 2, Table S1**).

226 Temporal variability in fruit production differed substantially across species and between
227 families (**Figure 3, Table S1**). The *Fagaceae* species *Fagus sylvatica*, *Q. petraea*, and *Q.*
228 *robur* had the highest temporal variabilities, with an average CV >1.3 and D >1.5 (**Table S1**).
229 Both these indices of temporal variability indicated higher variability in fruit production in
230 *Fagaceae* than in *Pinaceae* forests (1.38 ± 0.05 and 0.73 ± 0.05 for CV and 1.81 ± 0.09 and
231 0.85 ± 0.05 for D, respectively; $P < 0.001$). Also, the ANOVA indicated that the first
232 autocorrelation coefficient (AR1) differed amongst species (**Figure 3, Table S1**). Tukey's HSD
233 test, however, found significant differences in AR1 only between *P. menziesii* and *F. sylvatica*.
234 The former species had almost no autocorrelation, and the latter had a strong negative
235 autocorrelation. Autocorrelation in fruit production generally tended to be negative, suggesting
236 that years with higher (lower) fruit production tended to be followed by a year with lower
237 (higher) fruit production.

238 3.2. Determinants of fruit production and of its temporal behaviour

239 Our statistical linear mixed models (LMMs) indicated that GPP and foliar Zn and P
240 concentrations (**Figure 4**) were positively associated with NPP_f within and across species
241 (**Table 1**). We also found a statistically significant positive interaction between foliar Zn and P
242 concentrations ($P = 0.01$), indicating that the positive effect of Zn increased with increasing
243 foliar P concentration (and vice versa) (**Figure 5**). Our analysis thus identified a synergistic
244 effect of foliar Zn and P concentrations, leading to higher fruit production when both nutrients
245 occurred at high concentrations in the leaves. The LMM accounted for 56% of the variance in
246 NPP_f , with 40% explained by the fixed effects (GPP, Zn and P) and 16% by differences
247 amongst species.

248 Similar to NPP_f , $\%GPP_f$ was higher in forests with high foliar Zn and P concentrations (**Table 1**).
249 The interaction between foliar Zn and P concentrations was also statistically significantly
250 positively associated with $\%GPP_f$ ($P = 0.025$). The model for $\%GPP_f$ explained 44% of the
251 variance in $\%GPP_f$, with 31% accounted for by the fixed effects and the remaining 13% by
252 interspecific variability. Maximum recorded NPP_f was associated positively with GPP, basal
253 area and foliar Zn concentrations, and negatively with MAP and foliar N and C:P
254 concentrations (**Table 1**). Also, *Fagaceae* species presented larger maximum NPP_f than
255 *Pinaceae* species ($P = 0.004$). The model accounted for 41% of the variance in maximum
256 NPP_f , with 31% explained by the fixed effects and 10% by interspecific variability.

257 Temporal variability in fruit production differed depending on the index used to evaluate
258 interannual variation (CV and D), but temporal variability in NPP_f was clearly larger for
259 *Fagaceae* than for *Pinaceae* species ($P < 0.01$ for both indices; **Table 1, Figure 4**). The model
260 predicting the temporal variability of fruit production using the CV index indicated that more
261 productive forests with higher foliar Zn concentrations exhibited lower interannual variability in
262 fruit production. In contrast, the model predicting D of NPP_f indicated that D was higher in
263 forests with larger basal areas and lower foliar K concentrations (**Table 1**). Overall, the CV and
264 D models explained 62% and 56%, respectively, of the variance in the CV and D of fruit
265 production, with 58% and 55% explained by the fixed effects and the remaining 4% and 1% by
266 interspecific variability.

267 Finally, the model that best predicted temporal AR1 indicated a tendency for more negative
268 values of temporal autocorrelation in forests with high loads of N deposition than in forests less
269 exposed to N deposition ($P = 0.055$). However, N deposition explained only 3% of the variance
270 in the temporal autocorrelation of fruit production, and species-specific variability explained
271 20% of the variability in AR1.

2724. Discussion

273 Estimates of NPP_f and their role in the forest C balance

274 Our results provide the first estimates of carbon allocation to fruit production for some of the
275 most abundant tree species in European forests. Despite the high variability in the estimates,
276 due mainly to the differences in GPP and foliar nutrient concentrations (**Table 1**), average NPP_f
277 in our sample of European forests ranged from 10 to 40 g C m⁻² y⁻¹ and accounted for 0.5-3%

278of the GPP, both being higher in *Pinaceae* than in *Fagaceae* species. These estimates of NPP_f
279and GPP allocation are lower than those in previous studies. Herbst et al. (2015) reported
280mean NPP_f estimates of 95 ± 37 and 73 ± 25 g C m⁻² y⁻¹ for managed and unmanaged stands
281of *F. sylvatica*, respectively. GPP allocation to NPP_f averaged 6.7 ± 3.4 and $4.6 \pm 2.0\%$ in the
282managed and unmanaged stands respectively. In addition, %GPP_f reached 23% in the
283managed stand during an exceptional mast year. Campioli et al., (2010) reported an average
284NPP_f for *P. sylvestris* of 76 ± 8 g C m⁻² y⁻¹ and an average %GPP_f of $6.0 \pm 0.6\%$.

285Various reasons might be behind the differences between our results and previous reports of
286NPP_f and allocation to fruit. Differences in the endogenous characteristics of the stands, such
287as GPP or foliar nutrient concentrations, are of major importance for fruiting NPP (**Figure 4,**
288**Table 1**). Moreover, exogenous factors such as the management of the stand can also
289influence carbon allocation to fruiting (Herbst et al., 2015). The assessment of average values,
290however, is very dependent on the period of measurement due to the high irregularity of fruit
291production, especially when analysing relatively short time series of masting species such as
292those of the *Fagaceae* family. Different sampling methodologies could also be responsible for
293these differences in average NPP_f.

294NPP_f is usually a relatively small component of the carbon balance of forest ecosystems, but
295its high interannual variability in masting species, such as *F. sylvatica* or *Quercus* sp., identifies
296NPP_f as an important component to consider for a better assessment of the ecosystem carbon
297balance (Herbst et al., 2015) or the cascading effects that fruit NPP can entail on the entire
298food web of an ecosystem (Ostfeld & Keesing, 2000). Mean maximum NPP_f was 2- to 3-fold
299higher than mean NPP_f for all species (**Table S1**). %GPP_f can thus increase substantially in
300years with large fruit crops. Further analyses with longer time series are clearly needed to
301obtain more robust estimates of the role of fruit production in the forest carbon cycle.
302Nonetheless, our results do provide new insights on the fate of photosynthesised C in forest
303ecosystems. However, because of the way in which we estimated GPP for our sites and the
304error propagation, our estimates of %GPP_f may be subjected to larger error estimates than
305those we calculated by mean \pm SE. Therefore, any interpretation of the values we report
306should take that uncertainty into consideration.

307*The role of nutrients in NPP_f and its temporal behaviour*

308Our forests were in the lower range of mean concentrations of foliar Zn (**Table S2**), indicating
309that some of the forests might have been Zn deficient (Swietlik, 2002). Zn deficiency has been
310well studied in agricultural crops, but to the best of our knowledge, we are the first to report a
311potential role in forest ecosystems. Zn deficiency usually occurs in plants growing on alkaline
312soils because of the reduced bioavailability of Zn (Ma & Lindsay, 1990), but can also occur in acidified
313weathered soils (Alloway, 2009). Zn limitation can have a negative impact on plant vegetative
314growth and especially fruit yield (Swietlik, 2002), because of the roles Zn plays in several key
315metabolic processes such as protein synthesis (as a component of ribosomes, Prask and
316Plocke, 1971) or in metalloenzymes such as carbonic anhydrases (Dell & Wilson, 1985)
317involved in the conversion of carbon dioxide and water to bicarbonate and protons
318(photosynthesis and biomass production). Cakmak and Marschner (1988) found that Zn
319deficiency in *Gossypium hirsutum*, *Triticum aestivum*, *Lycopersicon esculentum*, and *Malus*
320*domestica* increased the exudation of K, amino acids, sugars, and phenolics from roots
321because of increased membrane permeability. By exuding more carbon-rich compounds in the
322soil, Zn-deficient plants may thus have less photosynthates available for aboveground
323compartments (Vicca *et al.*, 2012; Fernández-Martínez *et al.*, 2014b, 2016a) and fruit
324production (**Figures 5 and 6**). These processes may also be responsible for increased
325temporal variability in seed production, because carbon reserves in Zn-limited trees might be
326insufficient for regular production of large fruit crops (Isagi *et al.*, 1997).

327However, despite foliar Zn concentrations have an effect within species, the effect of foliar Zn
328is also associated to the different reproductive behaviour amongst taxonomical families, since
329*Pinaceae* species present higher foliar Zn concentrations than *Fagaceae* species (Table S3)
330and both families exhibit different patterns of fruit production (i.e., *Pinaceae* species invest
331more on reproduction [Figure 1] while *Fagaceae* species present higher interannual variability
332in NPP_f [Figure 3]). In that sense, high foliar Zn concentrations may be related to fast growing
333species (opportunistic, the r-selected species), producing fruits every year, while low foliar Zn
334concentrations may be linked to slow growing species (good competitors, the k-selected
335species) investing more heavily in fewer reproducing events. Thus, masting behavior could be
336related to an evolutionary strategy of the k-selected species that would confer an evolutionary
337advantage over their competitors (Kelly & Sork, 2002).

338 Forests with higher foliar P concentrations had higher mean fruit production and allocation to
339 fruit production in our study, supporting the finding by Fujita *et al.* (2014). P had the strongest
340 effect in forests with high foliar Zn concentrations (**Table 1**). This synergistic effect of foliar P
341 and Zn concentrations could be a consequence of an induced limitation of one nutrient when
342 the other is increased, (Elser *et al.*, 2007) in agreement with von Liebig's law of the minimum
343 (Von Liebig, 1840). The close association of P (Elser *et al.*, 2003) and Zn (Prask & Plocke,
344 1971) with ribosomes, and therefore with metabolism and biomass production, further supports
345 a positive interaction between these elements. Our results also point out that both families,
346 *Pinaceae* and *Fagaceae*, present a similar relationship between foliar P concentrations and
347 NPP_f despite their morphological and functional differences. These results also agree with
348 previous reports suggesting that higher foliar P concentrations enable larger fruit crops
349 compared to those with lower concentrations (e.g. in olive trees (Erel, 2008) and amongst
350 multiple plant species (Fujita *et al.*, 2014)) or, more generally, that P limitation constrains NPP_f
351 in forests (Perkins, 2004; Plassard & Dell, 2010) and in freshwater, marine, and other
352 terrestrial ecosystems (Elser *et al.*, 2007; Peñuelas *et al.*, 2013). Additionally, higher foliar C:P
353 ratios in our study were associated with lower maximum NPP_f (**Table 1**). This finding is also
354 consistent with P limitation for fruit production: P-deficient forests may not be able to produce
355 as large fruit crops as those with sufficient P availability.

356 According to our statistical models, trees with higher foliar N concentrations not only did not
357 have higher NPP_f but had lower maximum NPP_f than those with lower N concentrations (**Table**
358 **1**). This is intriguing given the fact that foliar N concentrations are higher in *Fagaceae* species
359 than in *Pinaceae* species (**Table S2**), despite the latter presents lower maximum NPP_f (**Table**
360 **1**). This may suggest that the negative relationship between foliar N concentration and
361 maximum NPP_f happens only within species and not amongst species. N has been identified
362 as the primary limiting nutrient for vegetative growth (Elser *et al.*, 2007; LeBauer & Treseder,
363 2008) and reproduction (Han *et al.*, 2013; Miyazaki *et al.*, 2014), but the lack of association
364 between foliar N concentration and NPP_f might be a consequence of other nutrients limiting
365 production, such as Zn or P. Most of Europe is exposed to high rates of N deposition, which
366 may increase N availability for plants, as indicated by the relatively high average values of
367 foliar N concentrations in European forests (**Table S2**).

368The negative relationship between atmospheric N deposition and AR1 of NPP_f may indicate
369that forests with higher N reserves are able to produce fruits crops with a periodic recurrence
370(Isagi *et al.*, 1997) when other factors are not limiting growth during a particular year (such as
371weather, other nutrients such as Zn or P, or a combination of both (Fernández-Martínez *et al.*,
3722015). On the other hand, our analyses revealed that foliar K concentration were negatively
373related to interannual variability in fruit production (**Table 1**). This relationship may not remain
374amongst species because foliar K concentrations were lower in the *Pinaceae* family, which
375presents significantly lower interannual variability (**Table 1**). Foliar K concentrations are related
376to plant water regulation, thus being an important nutrient especially in arid environments,
377where water availability is amongst the most important factors driving fruit and secondary
378production (Ogaya & Peñuelas, 2007; Pérez-Ramos *et al.*, 2010; Garbulsky *et al.*, 2013). K
379bioavailability increases with increasing annual precipitation, but, through conservation
380mechanisms such as resorption, plants can increase their K concentrations (Sardans &
381Peñuelas, 2015). Hence, we hypothesise that trees with higher foliar K concentrations produce
382fruits more regularly because they can cope better with eventual reduced water availability and
383are thus less sensitive to interannual changes in water availability (Fernández-Martínez *et al.*,
3842015).

385Our results thus indicate that nutrient-rich forests produce more fruits, allocate a larger fraction
386of the GPP to fruit production, and produce fruit more regularly than nutrient-limited forests.
387This supports previous findings that suggested higher C allocation to aboveground NPP in
388nutrient-rich forests (Vicca *et al.*, 2012; Fernández-Martínez *et al.*, 2014b, 2016a), as well as
389theory and observations suggesting that high interannual variability in fruit production is more
390pronounced in environments where resources are scarce (Kelly & Sork, 2002; Fernández-
391Martínez *et al.*, 2012). Nutrient availability may thus limit NPP_f even more than it limits
392vegetative NPP, because fruits are more nutrient-demanding than vegetative tissues (Reekie &
393Bazzaz, 1987). Amongst all investigated nutrients, foliar Zn and P concentrations were most
394strongly associated with higher fruit production. However, trees growing on soils with low
395nutrient availability might be forced to be more conservative in the use of nutrients and thus
396present higher concentrations in their tissues, acting like nutrient pools. Hence, further
397research is needed to disentangle the effects of soil nutrient availability and foliar nutrient
398concentrations on ecosystem functioning.

399 Other determinants of NPP_f and its temporal behaviour

400 Nutrient availability and stoichiometry played a very important role in determining NPP_f and its
401 temporal behaviour, but site productivity and the taxonomical family were also key factors.
402 Higher photosynthesis (GPP) was associated with higher NPP_f , higher maximum NPP_f , and
403 lower temporal variability (CV). This result is logical, because forests with larger GPP fluxes
404 also typically have larger NPP fluxes (Fernández-Martínez *et al.*, 2014a). Whether all NPP
405 components increase equally with increasing GPP, however, remains unresolved. In addition,
406 higher GPP might reduce the CV of fruit production because more photosynthates can be
407 allocated to storage to later be used for reproduction and allowing more regular production of
408 fruit crops. However, due to the positive relationship between NPP_f and GPP and the
409 dependence of the CV on the mean, this relationship might be spurious. This second
410 hypothesis is supported by the lack of a relationship between D (which is less sensitive to the
411 mean than the CV (**Supplementary Information 1: The consecutive disparity index**) of fruit
412 production and GPP. A larger basal area, however, was also associated with higher maximum
413 NPP_f , perhaps due to the competitive advantage of large trees to acquire various resources
414 (e.g. nutrients, water, and sunlight).

415 In addition to the strong influence of foliar nutrient concentrations on NPP_f and its interannual
416 variability, we found a significant taxonomic effect on NPP_f . In particular, *Fagaceae* and
417 *Pinaceae* species presented a markedly different behaviour in fruit production patterns, with
418 the latter exhibiting lower maximum NPP_f (probably because they are generally established
419 over nutrient poor soils) and lower interannual variability in NPP_f . Based on our results, we can
420 consider that the *Fagaceae* species studied here present a clear masting behaviour (i.e., high
421 interannual variability, negative autocorrelation in fruit production (Fernández-Martínez *et al.*,
422 2016b)) while it is not that clear for some of the *Pinaceae* species, especially *P. sylvestris*, *P.*
423 *nigra* and *P. menziesii*, which present relatively low interannual variability and no negative
424 autocorrelation in NPP_f (**Table S1**). These differences may be related to different life-history
425 traits and evolutionary strategies related to avoidance of seed predation or to different patterns
426 of resource allocation. The different nature of fruits produced by the two families (cones in
427 *Pinaceae* and nuts in *Fagaceae*) might be responsible for the different allocation of resources
428 to fruit production in terms of biomass production. *Pinaceae* species spend many resources
429 when producing cones (bearing the seeds), while in *Fagaceae* species each seed (nut) is

430 usually much cheaper to produce in terms of C. This difference in the amount of resources
431 needed to produce fruits between *Pinaceae* and *Fagaceae* families might be behind the
432 different patterns of fruit production in our studied species. Also, the different availability of
433 nutrients between families, such as N (higher in *Fagaceae*, **Table S2**) or Zn (higher in
434 *Pinaceae*, **Figure 4, Table S2**), might also condition the different patterns of fruit production
435 because of different nutrient-use efficiencies amongst species from different families (Sardans
436 *et al.*, 2016a).

437 Finally, our results also point out that masting behaviour can be well parameterised using the D
438 index, because it takes into account not only the variability of the time series but also its
439 temporal autocorrelation (**Table S1**). Since masting behaviour is related to high interannual
440 variability and negative autocorrelation in fruit production, which is exactly what the D index
441 accounts for (**Supplementary Information 1: The consecutive disparity index**), D seems to
442 be a good candidate to characterise masting behaviour using a single index.

4435. Conclusions

444 On average, NPP_f ranged from approximately 10 to 40 g C m⁻² y⁻¹ and accounted for 0.5-3% of
445 the GPP, with little differences amongst species or between families (*Fagaceae* – *Pinaceae*).
446 However, mean maximum NPP_f and interannual variability in NPP_f differed specially between
447 families, being higher for *Fagaceae* than for *Pinaceae*. These differences are likely to be linked
448 to different life-history traits and evolutionary strategies related to avoidance of seed predation
449 or to different patterns of resource allocation, given the different nature of their fruits.

450 More productive (higher GPPs) and nutrient rich forests produced larger and more regular fruit
451 crops and allocated a larger percentage of photosynthates into fruit production, which
452 highlights the paramount role of available reserves of nutrients and carbohydrates to allocate
453 into reproduction. Especially foliar zinc (Zn) and phosphorus (P) concentrations were
454 associated positively with fruit crop size, while foliar Zn and K were negatively related to its
455 temporal variability. To the best of our knowledge we report, for the first time, Zn deficiency in
456 forests.

457 Acknowledgements

458 We thank the ICP Forest organisation and collaborators for providing the litterfall data used in
459 this study. This research was supported by the European Research Council Synergy grant
460 ERC-2013-SyG 610028-IMBALANCE-P, the Spanish Government grant CGL2013-48074-P,
461 and the Catalan Government projects SGR 2014-274 and FI-2013. SV is a postdoctoral fellow
462 of the Research Foundation – Flanders (FWO).

463 **Author contribution**

464 M.F-M, S.V., I.J. and J.P planned and designed the research. M.F-M wrote the manuscript and
465 analysed the data. All co-authors contributed substantially in writing the paper.

466

467References

- 468Aber JD, Melillo JM. 1982. Nitrogen Immobilization in Decaying Hardwood Leaf Litter As a Function of
469Initial Nitrogen and Lignin Content. *Canadian Journal of Botany-Revue Canadienne De Botanique* **60**:
4702263–2269.
- 471Alloway BJ. 2009. Soil factors associated with zinc deficiency in crops and humans. *Environ Geochem*
472*Health* **31**: 537–548.
- 473Barton K. 2015. MuMIn: Multi-model inference. R package version 1.17.1. [http://CRAN.R-](http://CRAN.R-project.org/package=MuMIn)
474project.org/package=MuMIn.
- 475Büntgen U, Tegel W, Kaplan JO, Schaub M, Hagedorn F, Bürgi M, Brázdil R, Helle G, Carrer M,
476Heussner K-U, et al. 2013. Placing unprecedented recent fir growth in a European-wide and
477Holocene-long context. *Frontiers in Ecology and the Environment*: 131218051117007.
- 478Cakmak I, Marschner H. 1988. Increase in Membrane Permeability and Exudation in Roots of Zinc
479Deficient Plants. *Journal of Plant Physiology* **132**: 356–361.
- 480Capioli M, Gielen B, Granier a., Verstraeten a., Neiryneck J, Janssens I a. 2010. Carbon
481allocation to biomass production of leaves, fruits and woody organs at seasonal and annual scale in a
482deciduous- and evergreen temperate forest. *Biogeosciences Discussions* **7**: 7575–7606.
- 483Capioli M, Vicca S, Luysaert S, Bilcke J, Ceschia E, Chapin III FS, Ciais P, Fernández-Martínez
484M, Malhi Y, Obersteiner M, et al. 2015. Biomass production efficiency controlled by management in
485temperate and boreal ecosystems. *Nature Geoscience* **8**: 843–846.
- 486Comerford NB. 2005. Soil factors affecting nutrient bioavailability. *Nutrient acquisition by plants: an*
487*ecological perspective* **CDXXXI**: 1–14.
- 488Crone EE, McIntire EJB, Brodie J. 2011. What defines mast seeding? Spatio-temporal patterns of
489cone production by whitebark pine. *Journal of Ecology*: no–no.
- 490Dell B, Wilson SA. 1985. Effect of zinc supply on growth of three species of Eucalyptus seedlings and
491wheat. *Plant and Soil* **88**: 377–384.
- 492Eckstein RL, Karlsson PS. 1997. Above-ground growth and nutrient use by plants in a subarctic
493environment: effects of habitat, life-form and species. *Oikos*: 311–324.
- 494Elser JJ, Acharya K, Kyle M, Cotner J, Makino W, Markow T, Watts T, Hobbie S, Fagan W, Schade
495J, et al. 2003. Growth rate-stoichiometry couplings in diverse biota. *Ecology Letters* **6**: 936–943.
- 496Elser JJ, Bracken MES, Cleland EE, Gruner DS, Harpole WS, Hillebrand H, Ngai JT, Seabloom
497EW, Shurin JB, Smith JE. 2007. Global analysis of nitrogen and phosphorus limitation of primary
498producers in freshwater, marine and terrestrial ecosystems. *Ecology letters* **10**: 1135–42.
- 499Erel R. 2008. Flowering and fruit set of olive trees. *American Society* **133**: 639–647.
- 500Espelta JM, Cortés P, Molowny-horas R, Sánchez-humanes B, Retana J. 2008. Masting Mediated
501by Summer Drought Reduces Acorn Predation in Mediterranean Oak Forests. *Ecology* **89**: 805–817.
- 502Fernández-Martínez M, Belmonte J, María Espelta J, Espelta JM. 2012. Masting in oaks:
503Disentangling the effect of flowering phenology, airborne pollen load and drought. *Acta Oecologica* **43**:
50451–59.
- 505Fernández-Martínez M, Garbulsky M, Peñuelas J, Peguero G, Espelta JM. 2015. Temporal trends in
506the enhanced vegetation index and spring weather predict seed production in Mediterranean oaks.
507*Plant Ecology* **216**: 1061–1072.
- 508Fernández-Martínez M, Vicca S, Janssens IA, Capioli M, Peñuelas J. 2016a. Nutrient availability

- 509and climate as the main determinants of the ratio of biomass to NPP in woody and non-woody forest
510compartments. *Trees, structure and function* **30**: 775–783.
- 511Fernández-Martínez M, Vicca S, Janssens IA, Espelta JM, Peñuelas J. 2016b. The North Atlantic
512Oscillation synchronises fruit production in western European forests. *Ecography* **39**.
- 513Fernández-Martínez M, Vicca S, Janssens IA, Luysaert S, Campioli M, Sardans J, Estiarte M,
514Peñuelas J. 2014a. Spatial variability and controls over biomass stocks, carbon fluxes and resource-
515use efficiencies in forest ecosystems. *Trees, structure and function* **28**: 597–611.
- 516Fernández-Martínez M, Vicca S, Janssens IA, Sardans J, Luysaert S, Campioli M, Chapin III FS,
517Ciais P, Malhi Y, Obersteiner M, et al. 2014b. Nutrient availability as the key regulator of global forest
518carbon balance. *Nature Climate Change* **4**: 471–476.
- 519Fujita Y, Venterink HO, van Bodegom PM, Douma JC, Heil GW, Hölzel N, Jabłońska E, Kotowski
520W, Okruszko T, Pawlikowski P, et al. 2014. Low investment in sexual reproduction threatens plants
521adapted to phosphorus limitation. *Nature* **505**: 82–6.
- 522Garbulsky MF, Peñuelas J, Ogaya R, Filella I. 2013. Leaf and stand-level carbon uptake of a
523Mediterranean forest estimated using the satellite-derived reflectance indices EVI and PRI.
524*International Journal of Remote Sensing* **34**: 1282–1296.
- 525Güsewell S. 2004. N:P ratios in terrestrial plants: Variation and functional significance. *New Phytologist*
526**164**: 243–266.
- 527Han Q, Kabeya D, Iio A, Inagaki Y, Kakubari Y. 2013. Nitrogen storage dynamics are affected by
528masting events in *Fagus crenata*. *Oecologia* **174**: 679–687.
- 529Herbst M, Mund M, Tamrakar R, Knohl A. 2015. Differences in carbon uptake and water use between
530a managed and an unmanaged beech forest in central Germany. *Forest Ecology and Management*
531**355**: 101–108.
- 532Hijmans RJ, Cameron SE, Parra JL, Jones PG, Jarvis A. 2005. Very high resolution interpolated
533climate surfaces for global land areas. *International Journal of Climatology* **25**: 1965–1978.
- 534Isagi Y, Sugimura K, Sumida a., Ito H. 1997. How Does Masting Happen and Synchronize? *Journal*
535*of Theoretical Biology* **187**: 231–239.
- 536Kelly D. 1994. The evolutionary ecology of mast seeding. *Trends in Ecology & Evolution* **9**: 465–470.
- 537Kelly D, Sork VL. 2002. Mast seeding in perennial plants: why, how, where? *Annual Review of Ecology*
538*and Systematics* **33**: 427–447.
- 539Koenig W, Knops J. 2005. The mystery of masting in trees. *American Scientist* **93**: 340–347.
- 540LeBauer DS, Treseder KK. 2008. Nitrogen limitation of net primary productivity in terrestrial
541ecosystems is globally distributed. *Ecology* **89**: 371–379.
- 542Von Liebig J. 1840. Die organische Chemie in ihrer Anwendung auf Agricultur und Physiologie:
543Friedrich Vieweg Sohn.
- 544Luysaert S, Ciais P, Piao SL, Schulze E-D, Jung M, Zaehle S, Schelhaas MJ, Reichstein M,
545Churkina G, Papale D, et al. 2010. The European carbon balance. Part 3: forests. *Global Change*
546*Biology* **16**: 1429–1450.
- 547Luysaert S, Inglima I, Jung M, Richardson a. D, Reichstein M, Papale D, Piao SL, Schulze E-D,
548Wingate L, Matteucci G, et al. 2007. CO₂ balance of boreal, temperate, and tropical forests derived
549from a global database. *Global Change Biology* **13**: 2509–2537.
- 550Ma Q, Lindsay WL. 1990. Divalent Zinc Activity in Arid-Zone Soils Obtained by Chelation. *Soil Science*
551*Society of America Journal* **54**: 719.

- 552 **Martín-Vide J. 1986.** Notes per a la definició d'un índex de «desordre» en pluviometria. *Societat*
553 *Catalana de Geografia*: 89–96.
- 554 **McGroddy ME, Daufresne T, Hedin OL. 2004.** Scaling of C : N : P Stoichiometry in Forests
555 Worldwide : Implications of Terrestrial Redfield- Type Ratios. *Ecology* **85**: 2390–2401.
- 556 **Miyazaki Y, Maruyama Y, Chiba Y, Kobayashi MJ, Joseph B, Shimizu KK, Mochida K, Hiura T,**
557 **Kon H, Satake A. 2014.** Nitrogen as a key regulator of flowering in *Fagus crenata*: understanding the
558 physiological mechanism of masting by gene expression analysis. *Ecology letters* **17**: 1299–1309.
- 559 **Nakagawa S, Schielzeth H. 2013.** A general and simple method for obtaining R² from generalized
560 linear mixed-effects models (RB O'Hara, Ed.). *Methods in Ecology and Evolution* **4**: 133–142.
- 561 **Norton DA, Kelly D. 1988.** Mast Seeding Over 33 Years by *Dacrydium cupressinum* Lamb. (rimu)
562 (Podocarpaceae) in New Zealand: The Importance of Economies of Scale. *Functional Ecology* **2**: 399–
563 408.
- 564 **Ogaya R, Peñuelas J. 2007.** Species-specific drought effects on flower and fruit production in a
565 Mediterranean holm oak forest. *Forestry* **80**: 351–357.
- 566 **Ostfeld R, Keesing F. 2000.** Pulsed resources and community dynamics of consumers in terrestrial
567 ecosystems. *Trends in Ecology & Evolution* **15**: 232–237.
- 568 **Oulehle F, Evans CD, Hofmeister J, Krejci R, Tahovska K, Persson T, Cudlin P, Hruska J. 2011.**
569 Major changes in forest carbon and nitrogen cycling caused by declining sulphur deposition. *Global*
570 *Change Biology* **17**: 3115–3129.
- 571 **Peñuelas J, Poulter B, Sardans J, Ciais P, van der Velde M, Bopp L, Boucher O, Godderis Y,**
572 **Hinsinger P, Llusia J, et al. 2013.** Human-induced nitrogen-phosphorus imbalances alter natural and
573 managed ecosystems across the globe. *Nature communications* **4**: 2934.
- 574 **Pérez-Ramos IM, Ourcival JM, Limousin JM, Rambal S. 2010.** Mast seeding under increasing
575 drought: results from a long-term data set and from a rainfall exclusion experiment. *Ecology* **91**: 3057–
576 68.
- 577 **Pérez-Ramos IM, Padilla-Díaz CM, Koenig WD, Marañón T. 2015.** Environmental drivers of mast-
578 seeding in Mediterranean oak species : does leaf habit matter ? *Journal of Ecology*: 1–10.
- 579 **Perkins JL. 2004.** *Pinus albicaulis* seedling regeneration after fire.
- 580 **Pinheiro J, Bates D, DebRoy S, Sarkar D, Team} {R Core. 2013.** nlme: Linear and Nonlinear Mixed
581 Effects Models.
- 582 **Pitman R, Bastrup-Birk A, Breda N, Rautio P. 2010.** Sampling and Analysis of Litterfall. *Manual on*
583 *methods and criteria for harmonized sampling, assessment, monitoring and analysis of the effects of*
584 *air pollution on forests*: 16.
- 585 **Plassard C, Dell B. 2010.** Phosphorus nutrition of mycorrhizal trees. *Tree physiology* **30**: 1129–39.
- 586 **Prask J a, Plocke DJ. 1971.** A Role for Zinc in the Structural Integrity of the Cytoplasmic Ribosomes of
587 *Euglena gracilis*. *Plant physiology* **48**: 150–155.
- 588 **R Core Team. 2015.** R: A Language and Environment for Statistical Computing.
- 589 **Rautio P, Fürst A, Stefan K, Raitio H, Bartels U. 2010.** Sampling and Analysis of Needles and
590 Leaves. *ICP Forests Manual XII*: 19.
- 591 **Reekie EG, Bazzaz FA. 1987.** Reproductive effort in plants. II Does carbon reflect the allocation of
592 other resources? *American Naturalist* **129**: 897–906.
- 593 **Sala A, Hopping K, McIntire EJB, Delzon S, Crone EE. 2012.** Masting in whitebark pine (*Pinus*

- 594albicaulis) depletes stored nutrients. *The New phytologist* **196**: 189–99.
- 595Sardans J, Alonso R, Carnicer J, Fernández-Martínez M, Vivanco MG, Peñuelas J. 2016a. Factors
596influencing the foliar elemental composition and stoichiometry in forest trees in Spain. *Perspectives in*
597*Plant Ecology, Evolution and Systematics* **18**: 52–69.
- 598Sardans J, Alonso R, Janssens I, Carnicer J, Vereseoglou S, Rillig MC, Fernández-Martínez M,
599Sanders TGM, Peñuelas J. 2016b. Foliar and soil concentrations and stoichiometry of nitrogen and
600phosphorous across European *Pinus sylvestris* forests: relationships with climate, N deposition and tree
601growth. *Functional Ecology* **30**: 676–689.
- 602Sardans J, Peñuelas J. 2015. Potassium: a neglected nutrient in global change. *Global Ecology and*
603*Biogeography* **24**: 261–275.
- 604Sardans J, Rivas-Ubach A, Peñuelas J. 2012. The C:N:P stoichiometry of organisms and ecosystems
605in a changing world: A review and perspectives. *Perspectives in Plant Ecology, Evolution and*
606*Systematics* **14**: 33–47.
- 607Silvertown JW. 1980. The evolutionary ecology of mast seeding in trees. *Biological Journal of the*
608*Linnean Society* **14**: 235–250.
- 609Smith CC, Hamrick JL, Kramer CL. 1990. The Advantage of Mast Years for Wind Pollination. *The*
610*American Naturalist* **136**: 154.
- 611Sork VL, Bramble J, Sexton O. 1993. Ecology of mast-fruited in three species of North American
612deciduous oaks. *Ecology* **74**: 528–541.
- 613Swietlik D. 2002. Zinc Nutrition of Fruit Crops. *Hort technology* **12**: 45–50.
- 614Truog E. 1947. Soil Reaction Influence on Availability of Plant Nutrients. *Soil Science Society of*
615*America Journal* **11**: 305.
- 616Vicca S, Luysaert S, Peñuelas J, Campioli M, Chapin FS, Ciais P, Heinemeyer A, Högberg P,
617Kutsch WL, Law BE, et al. 2012. Fertile forests produce biomass more efficiently. *Ecology letters* **15**:
618520–6.
- 619de Vries W, Du E, Butterbach-Bahl K. 2014. Short and long-term impacts of nitrogen deposition on
620carbon sequestration by forest ecosystems. *Current Opinion in Environmental Sustainability* **9-10**: 90–
621104.
- 622Willby NJ, Pulford ID, Flowers TH. 2001. Tissue nutrient signatures predict herbaceous-wetland
623community responses to nutrient availability. *New Phytologist* **152**: 463–481.

624

625 **Figure captions**

626 **Figure 1:** Box-and-whisker plots showing NPP_f (upper panel) and NPP_f -to-GPP percentage
627 (lower panel) for nine European tree species. Dark grey boxes indicate coniferous species and
628 light grey boxes indicate broadleaved species. The P values of the ANOVAs for differences
629 amongst species are shown inside each panel. Different letters indicate significant differences
630 ($P < 0.05$) amongst species (Tukey's HSD test for multiple comparisons). Numbers above the
631 letters indicate the number of replicates. Average values are presented in Table S1.

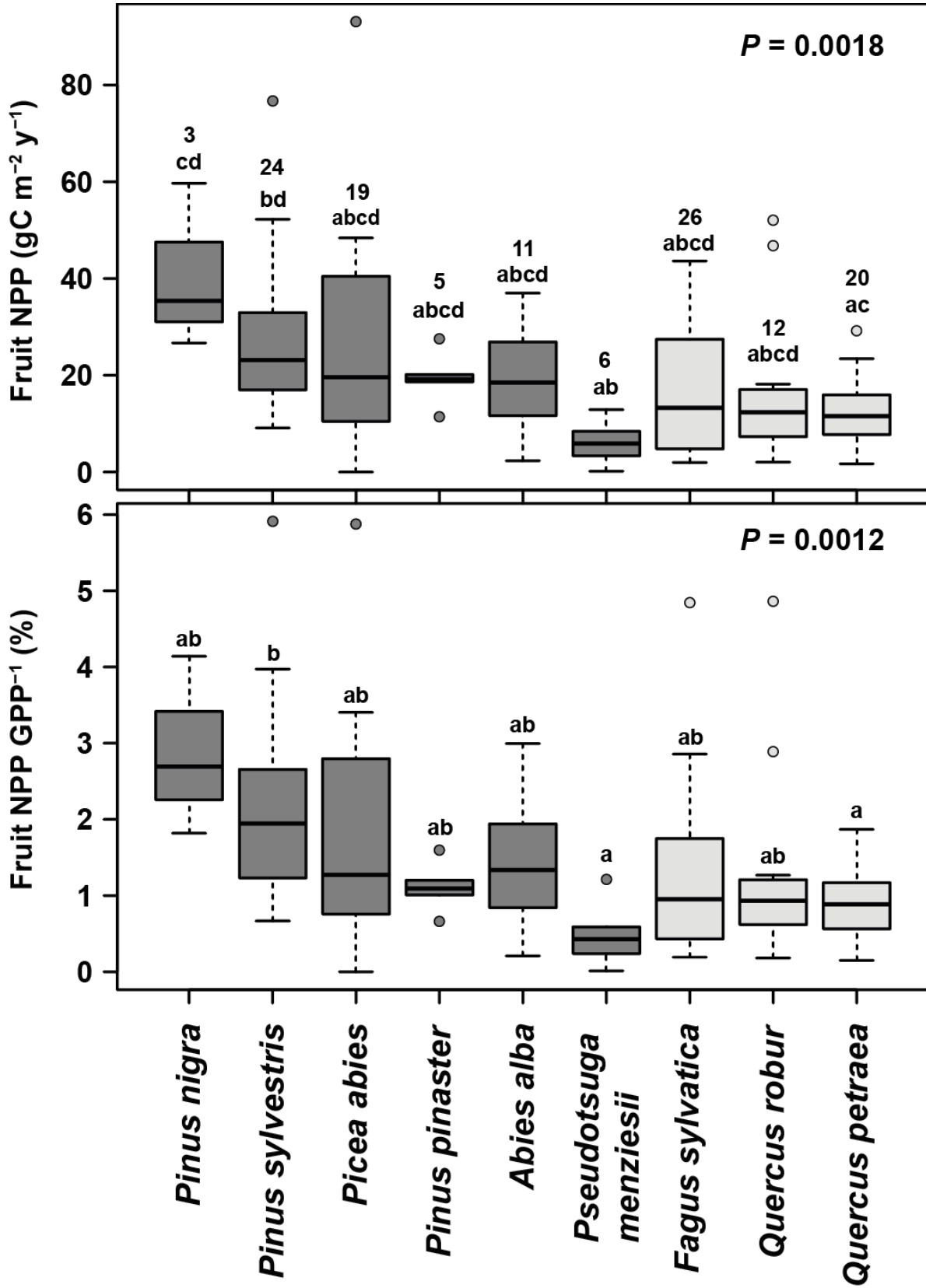
632 **Figure 2:** Box-and-whisker plot showing maximum fruit net primary production (NPP) (upper
633 panel) for nine European tree species. Dark grey boxes indicate coniferous species and light
634 grey boxes indicate broadleaved species. The P value of the ANOVA for differences amongst
635 species is shown inside the plot. Average values are presented in Table S1.

636 **Figure 3:** Box-and-whisker plots showing the coefficient of variation (CV, upper panel),
637 disparity (D, middle panel), and autocorrelation coefficient for the first lag (AR1, lower panel) of
638 fruit net primary production (NPP) for nine European tree species. Dark grey boxes indicate
639 coniferous species and light grey boxes indicate broadleaved species. The P values of the
640 ANOVAs for differences amongst species are shown inside each panel. Different letters
641 indicate significant differences ($P < 0.05$) amongst species (Tukey's HSD test for multiple
642 comparisons). Average values are presented in Table S1.

643 **Figure 4:** Partial residuals plots for the models of fruit net primary production (NPP_f) and the
644 CV of NPP_f without interactions. Upper panels show the relationships of gross primary
645 production (GPP) (panel a), foliar Zn (panel b) and P (panel c) concentrations with NPP_f .
646 Lower panels show the relationships of gross primary production (GPP) (panel d), foliar Zn
647 concentration (panel e) and family (panel f) with CV of NPP_f . Beta weights ($\beta \pm SE$) and their
648 significance (P) within the model without interactions are shown in each panel. Light blue
649 shading indicates 95% confidence bands.

650 **Figure 5:** 3D plot showing the significant interaction between foliar P and Zn concentrations for
651 predicting fruit net primary production (NPP). The response surface was calculated using a
652 generalised additive model explaining 28% of the variance in NPP_f . Blue and green colours of
653 the response surface indicate, respectively, low and high NPP_f .

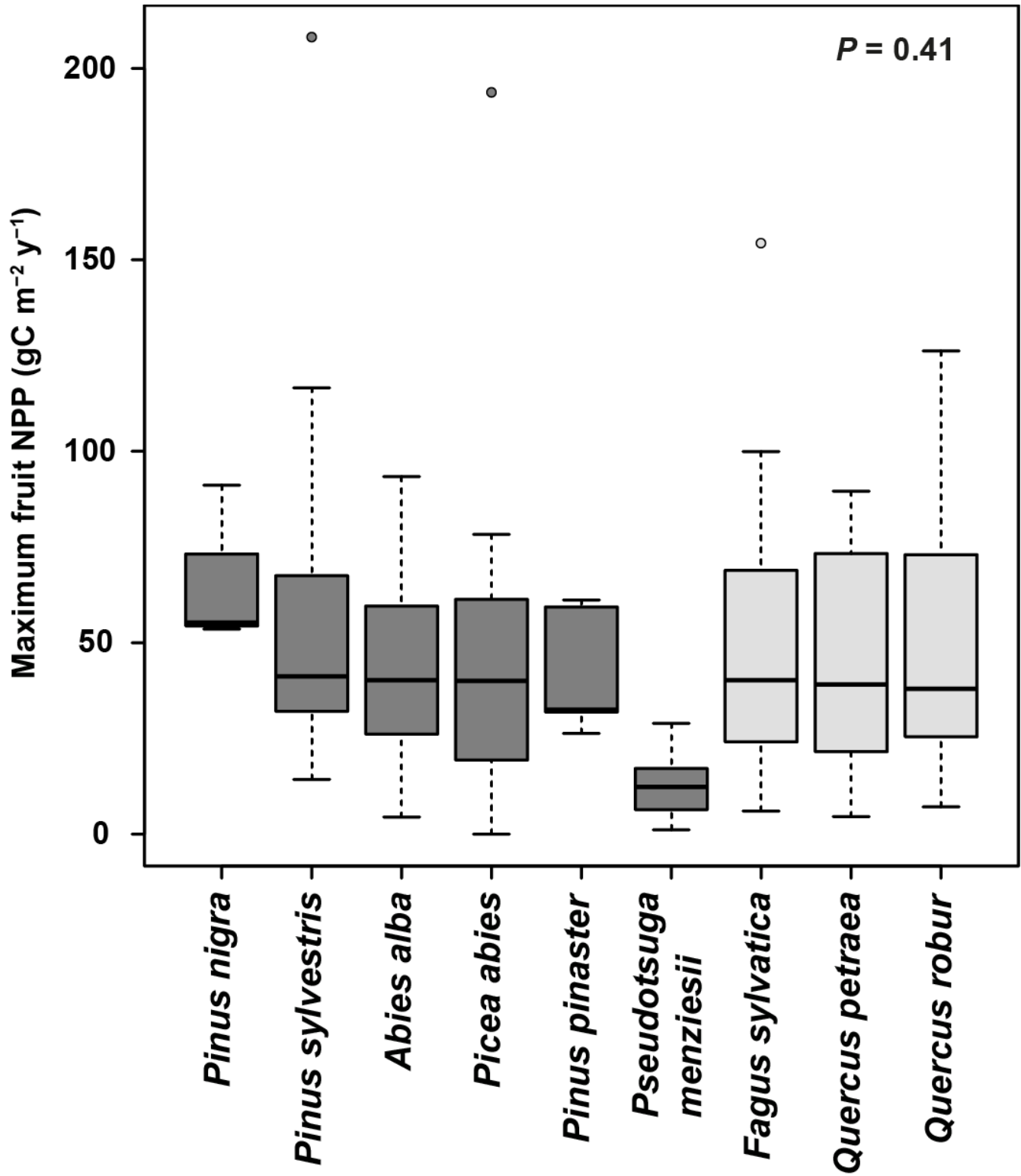
655 Figure 1



656

45
46

657 Figure 2

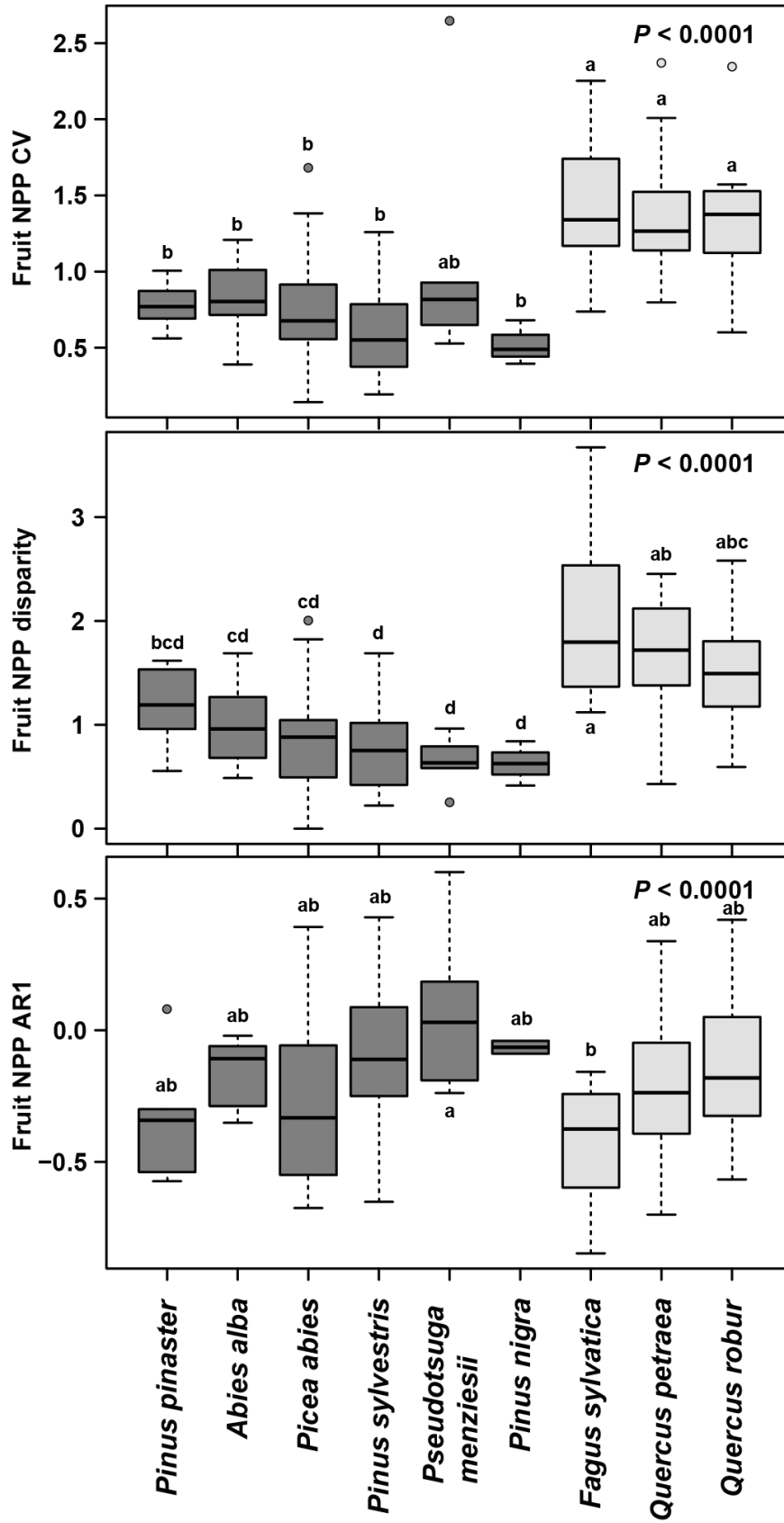


658

659

47
48

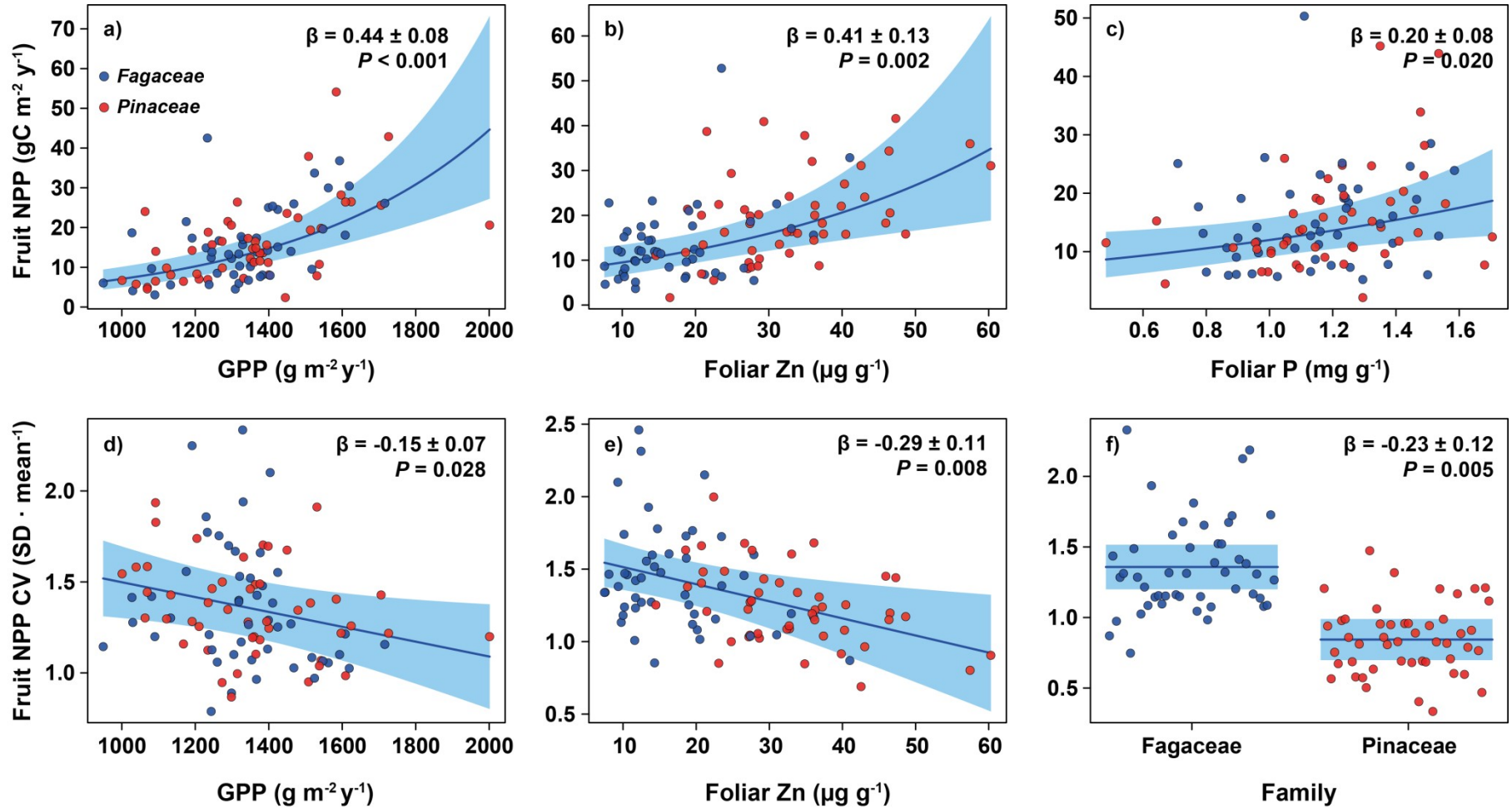
660Figure 3



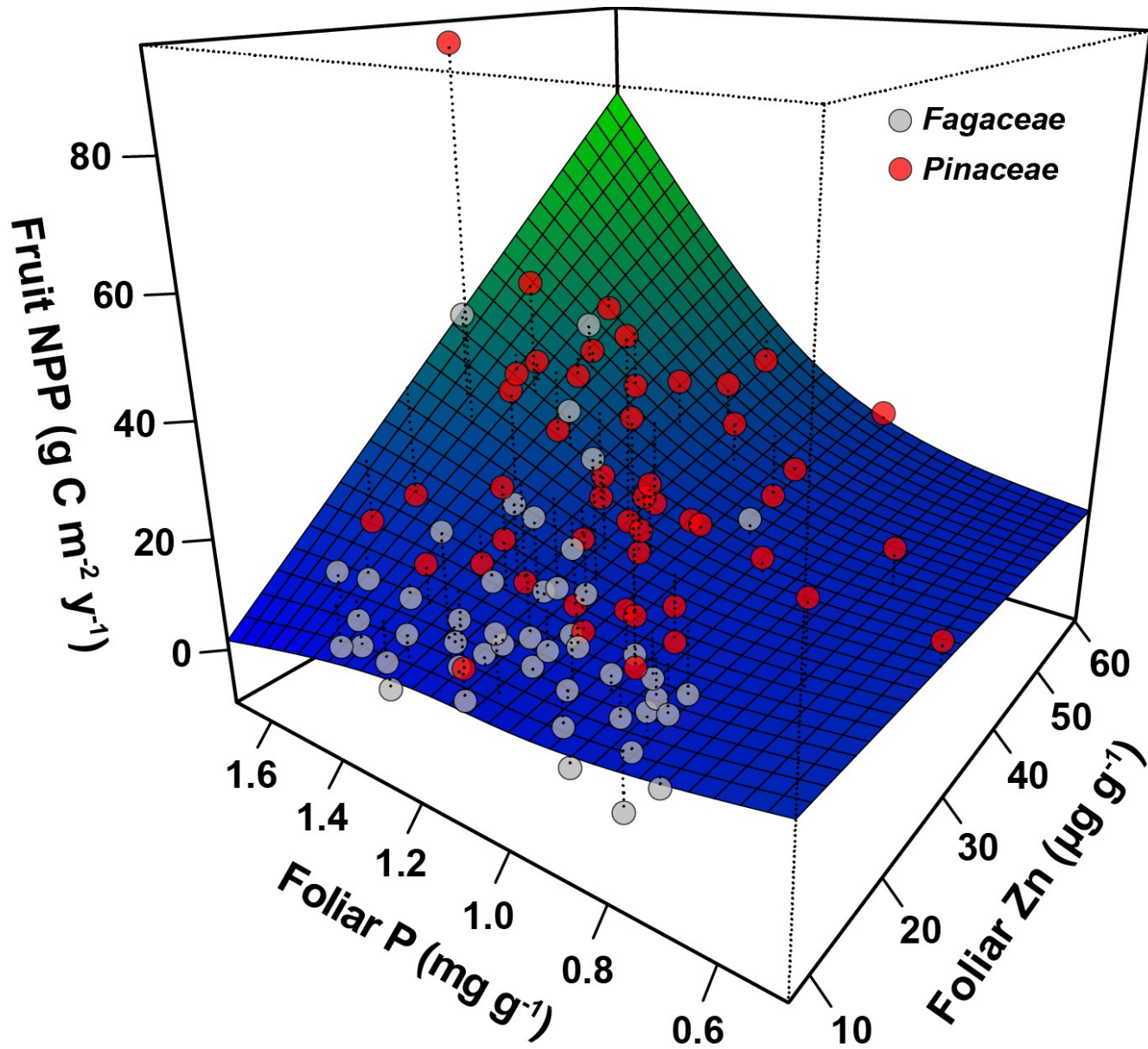
661

49
50

662 Figure 4



663
664 Figure 5



665

53
54

27

666**Table 1:** Summary of the mixed models predicting fruit net primary production (NPP), the percentage of gross primary
667production allocated to NPP_f (%GPP_f), maximum NPP_f, and the CV, D, and AR1 of NPP_f. Values are beta weights ± SE. For
668family, the coefficient shown indicates the change from *Fagaceae* to *Pinaceae*. Variance explained for fixed factors corresponds
669to marginal R^2 , and interspecific variance is the variance explained by the random effects. Ln indicates the response variable
670was transformed to the natural logarithm. Abbreviations: mean annual precipitation (MAP), coefficient of variation (CV), disparity
671(D), and autocorrelation coefficient at lag 1 (AR1). † $P < 0.1$, * $P < 0.05$, ** $P < 0.01$, *** $P < 0.001$. All models included species as
672a random variable.

	Ln NPP _f	Ln %GPP _f	Max NPP _f	CV	D	AR1
Endogenous						
Family			- ± 0.3 **	- ± 0.1 **	- ± 0.0 ***	
GPP	0.4 ± 0.0 ***		0.2 ± 0.1 *	- ± 0.0 *		
Basal area			0.3 ± 0.1 **		0.1 ± 0.0 *	
Foliar N			- ± 0.2 *			
Foliar P	- ± 0.1	- ± 0.2 *				
Foliar Zn	- ± 0.4	- ± 0.4 *	0.3 ± 0.1 *	- ± 0.1 **		
Foliar C:P			- ± 0.1 **			
Foliar K					- ± 0.0 ***	
Exogenous						
MAP			- ± 0.1 *			
N deposition						- ± 0.1 †
Interactions						
Foliar P:Foliar Zn	1.2 ± 0.4 **	1.1 ± 0.5 **				
Variance						
Fixed factors	0.40	0.31	0.31	0.58	0.55	0.03
Interspecific	0.16	0.13	0.10	0.04	0.01	0.20
Total	0.56	0.44	0.41	0.62	0.56	0.23

673Supplementary material

674Figure captions:

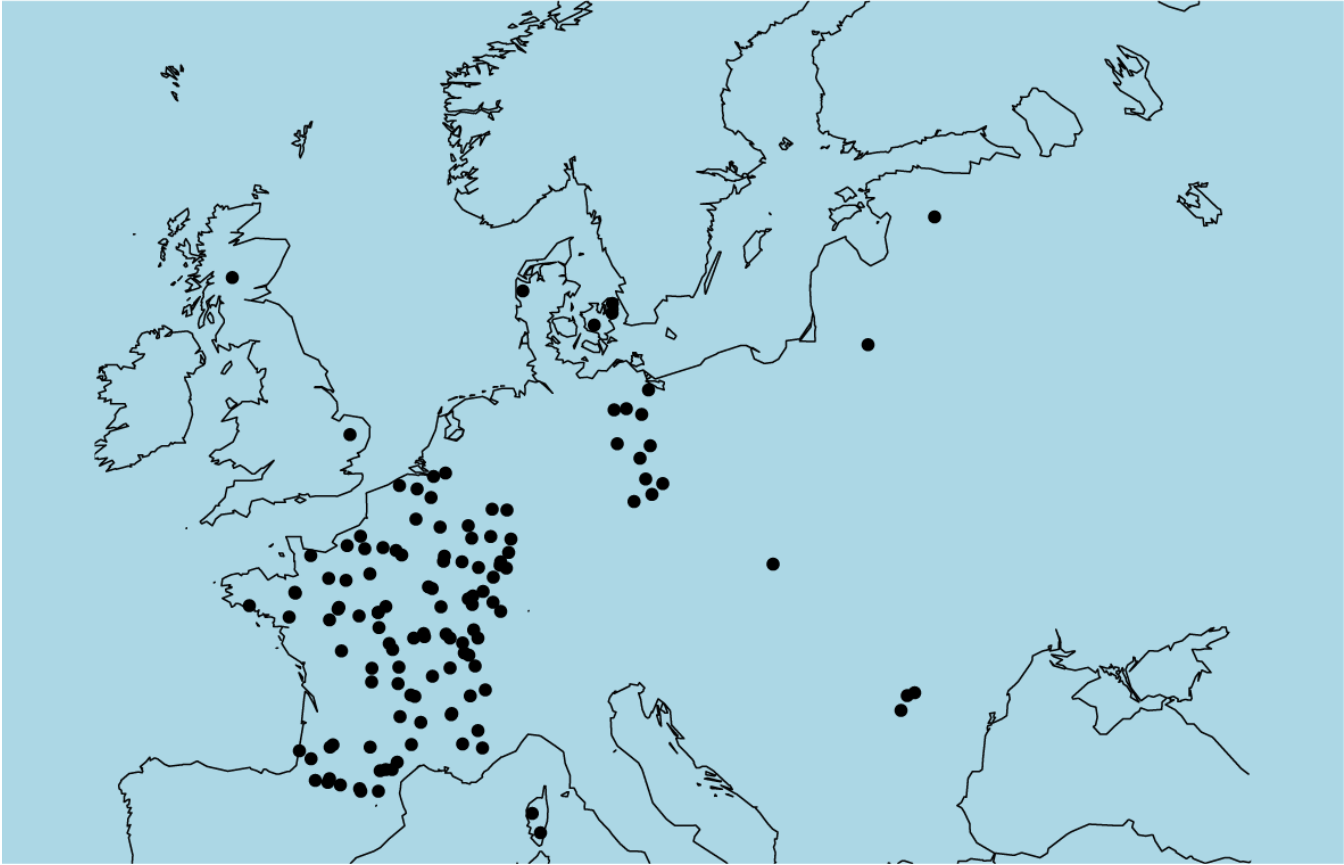
675**Figure S1:** Map showing the location of the 126 forests monitored in this study.

676**Figure S2:** Observed versus predicted GPP values from the model used to estimate GPP for the study sites. See materials and
677methods and Section 1: Estimating GPP in the supplementary material for further information about model adjustment.

678**Figure S3:** Comparison of the CV and D indices using two time series with equal means and standard deviations but different
679autocorrelation structures. AR1, autocorrelation coefficient for lag 1; CV, coefficient of variation (standard deviation • mean-1);
680and D, disparity index (see **Eqs. 1 and 2** in **Supplementary Material Section 1: The consecutive disparity index (D)**).

681

682Figure S1:

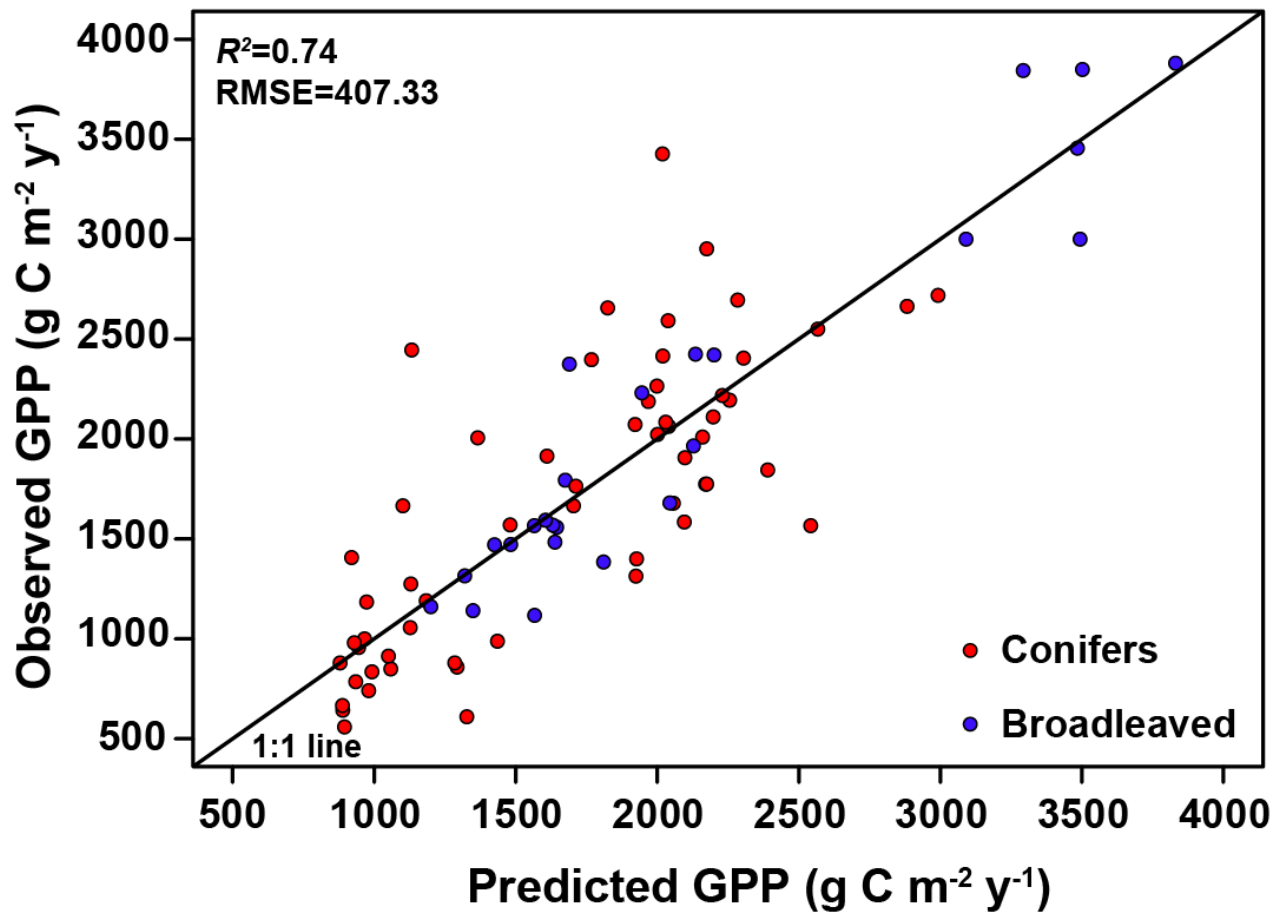


683

684

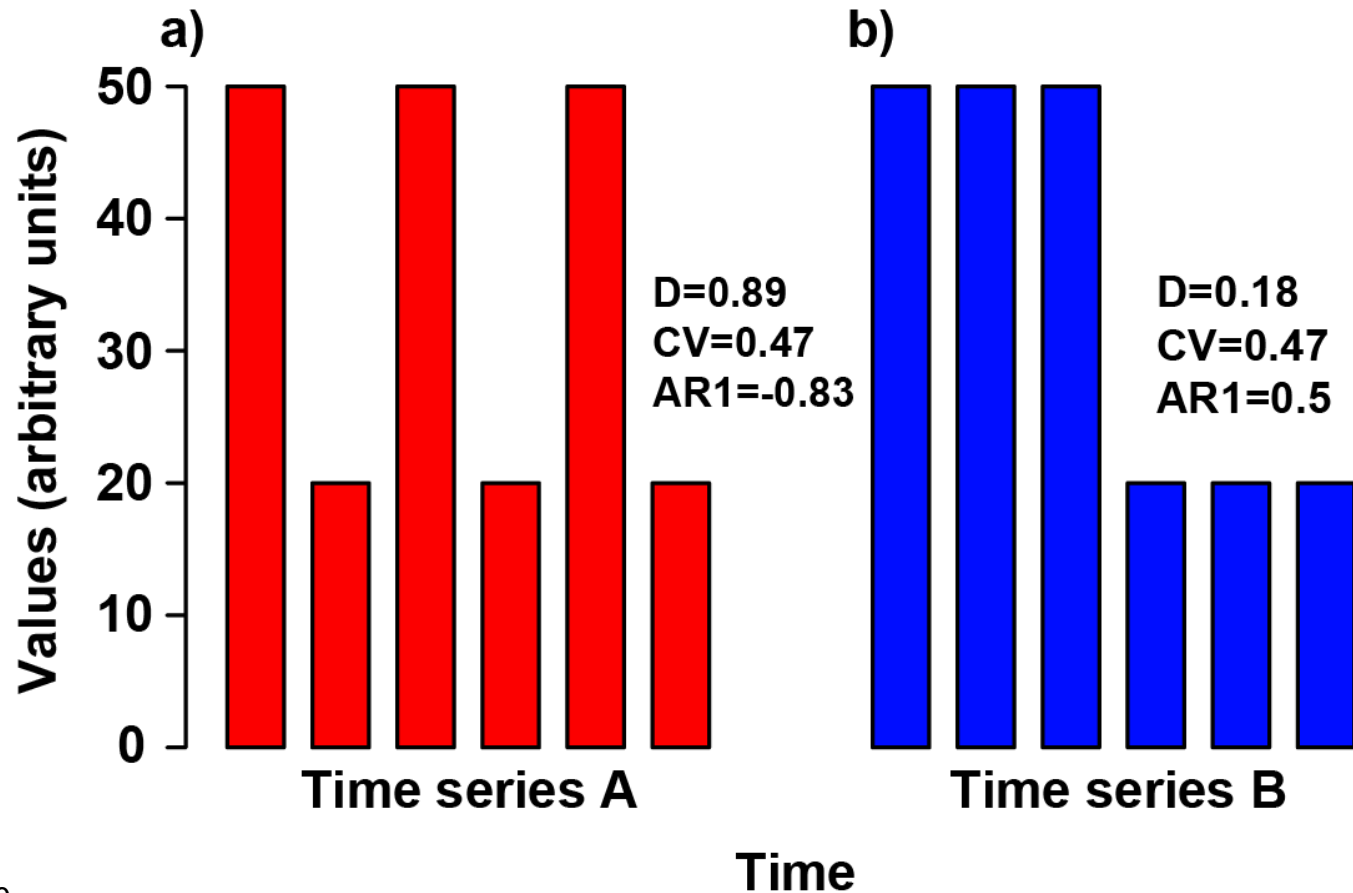
685

686Figure S2



687

61
62



691fruit net primary production (NPP) ($\text{g C m}^{-2} \text{y}^{-1}$), allocation to NPP_f ($NPP_f \cdot GPP^{-1} = \%GPP_f$), mean maximum NPP_f (Max NPP), and
 692coefficient of variation (CV), consecutive disparity (D), and first autocorrelation coefficient (AR1) of NPP_f . AR1 P shows the significance of
 693the AR1 coefficients from a *t*-test, and *n* indicates the number of sites per species. Different letters indicate significant differences ($P < 0.05$)
 694amongst species (Tukey's HSD test for multiple comparisons).

Table S1: Species (mean \pm SE)

	NPP_f		%GP P_f		Max NPP_f		CV		D		AR1		AR1 P		n
<i>Abies alba</i>	18.7 ±	abc	1.4 ±	a	43.1 ±	a	0.83 ±	b	1.00 ±	cd	-0.15 ±	0.071	a	1	
	3.3	d	0.3	b	7.9		0.08		0.11		0.04	7	b	1	
<i>Fagus sylvatica</i>	16.1 ±	abc	1.2 ±	a	49.0 ±	a	1.42 ±	a	2.02 ±	a	-0.42 ±	<0.00	b	2	
	2.5	d	0.2	b	6.9		0.08		0.15		0.04	01	b	6	
<i>Picea abies</i>	25.0 ±	abc	1.8 ±	a	46.2 ±	a	0.76 ±	b	0.89 ±	cd	-0.27 ±	0.001	a	1	
	5.2	d	0.4	b	9.8		0.09		0.12		0.08	1	b	9	
<i>Pinus nigra</i>	40.6 ±	cd	2.9 ±	a	66.6 ±	a	0.52 ±	b	0.63 ±	d	-0.06 ±	0.732	a	3	
	9.9		0.7	b	12.2		0.08		0.12		0.02	2	b	3	
<i>Pinus pinaster</i>	19.4 ±	abc	1.1 ±	a	42.2 ±	a	0.78 ±	b	1.17 ±	bc	-0.33 ±	0.006	a	5	
	2.6	d	0.2	b	7.4		0.08		0.19	d	0.12	0	b	5	
<i>Pinus sylvestris</i>	26.8 ±	bd	2.1 ±	b	54.6 ±	a	0.60 ±	b	0.77 ±	d	-0.09 ±	0.153	a	2	
	3.2		0.3		8.3		0.05		0.09		0.06	8	b	4	
<i>Pseudotsuga menziesii</i>	6.1 ±	ab	0.5 ±	a	13.1 ±	a	1.06 ±	a	0.64 ±	d	0.07 ±	0.524	a	6	
	1.7		0.2		4.1		0.32	b	0.10		0.12	9	a	6	
<i>Quercus petraea</i>	12.3 ±		0.9 ±	a	44.5 ±	a	1.36 ±	a	1.71 ±	ab	-0.20 ±	0.001	a	2	
	1.5	ac	0.1		6.2		0.08		0.11		0.06	2	b	0	
<i>Quercus robur</i>	16.9 ±	abc	1.3 ±	a	49.6 ±	a	1.35 ±	a	1.51 ±	ab	-0.10 ±	0.243	a	1	
	4.6	d	0.4	b	10.8		0.12		0.16	c	0.09	1	b	2	

695

696**Table S2:** Species (mean \pm SE) foliar C, N, P, and K concentrations and stoichiometries (C:N, C:P, and N:P). Concentrations have units of
697mg g⁻¹ except for C, which is per cent of dry weight. C:N, C:P, and N:P ratios are calculated on a mass basis. Different letters indicate
698significant differences ($P < 0.05$) amongst species (Tukey's HSD test for multiple comparisons), and n indicates the number of sites per
699species.

	C		N		P		K		C:N		C:P		N:P		n
<i>Abies alba</i>	52.42 \pm	a	12.94 \pm	b	1.16 \pm	b	5.61 \pm		40.02 \pm		452.6 \pm		11.32 \pm		1
	0.11	b	0.31	c	0.04	c	0.25	bc	0.75	b	15.8	b	0.44	c	1
<i>Fagus sylvatica</i>	53.07 \pm		24.23 \pm		1.17 \pm	b	7.04 \pm	ab	22.11 \pm		479.9 \pm		21.73 \pm		2
	0.43	a	0.53	a	0.05	c	0.32	c	0.53	c	18.5	b	1.02	a	6
<i>Picea abies</i>	51.35 \pm	a	13.61 \pm	b	1.34 \pm		5.60 \pm		38.10 \pm		397.0 \pm		10.48 \pm		1
	0.37	b	0.30	c	0.07	b	0.25	bc	0.74	b	17.1	b	0.41	c	9
<i>Pinus nigra</i>	53.50 \pm		14.90 \pm	b	1.17 \pm	b	6.37 \pm	ab	41.47 \pm		428.0 \pm		12.47 \pm		
	NA	a	3.62	c	0.11	c	0.49	c	NA	b	NA	b	2.17	c	3
<i>Pinus pinaster</i>	52.41 \pm	a	9.06 \pm		0.70 \pm		4.60 \pm		60.34 \pm		808.1 \pm		13.59 \pm	b	
	0.26	b	0.55	c	0.08	c	0.81	c	5.00	a	97.3	a	1.88	c	5
<i>Pinus sylvestris</i>	52.51 \pm	a	15.87 \pm		1.28 \pm		5.43 \pm		34.41 \pm		429.0 \pm		12.67 \pm		2
	0.15	b	0.58	b	0.04	b	0.13	bc	1.23	b	15.7	b	0.70	c	4
<i>Pseudotsuga menziesii</i>	53.02 \pm	a	16.30 \pm		1.21 \pm		7.44 \pm		32.63 \pm		443.4 \pm		13.65 \pm	b	
	0.22	b	0.43	b	0.06	b	0.44	ab	0.77	b	23.3	b	0.83	c	6
<i>Quercus petraea</i>	52.38 \pm	a	23.66 \pm		1.10 \pm	b	7.16 \pm	ab	22.60 \pm		502.1 \pm		22.11 \pm		2
	0.18	b	0.53	a	0.05	c	0.27	c	0.47	c	24.4	b	0.86	a	0
<i>Quercus robur</i>	52.56 \pm	a	24.92 \pm		1.33 \pm		7.80 \pm		22.35 \pm		420.5 \pm		19.99 \pm	a	1
	0.71	b	1.47	a	0.07	b	0.76	a	1.99	c	34.8	b	1.85	b	2

700

701

702**Table S3:** Species (mean \pm SE) foliar S, Fe, Ca, Mg, Mn, Zn, and Cu concentrations. S, Ca, and Mg concentrations have units of mg g⁻¹,
703and Fe, Mn, Zn, and Cu have units of μ g g⁻¹. Different letters indicate significant differences ($P < 0.05$) amongst species (Tukey's HSD test
704for multiple comparisons), and n indicates the number of sites per species.

	S		Fe		Ca		Mg		Mn		Zn		Cu	
<i>Abies alba</i>	0.95 \pm	d	47.98 \pm	b	8.33 \pm	a	1.35 \pm	ab	917.14 \pm	b	29.42 \pm	bc	4.01 \pm	c
<i>Fagus sylvatica</i>	1.50 \pm	ab	95.30 \pm	a	7.24 \pm	ab	1.12 \pm	ab	1390.59 \pm	a	23.91 \pm	bc	7.16 \pm	a
<i>Picea abies</i>	0.91 \pm	d	54.76 \pm	b	5.14 \pm	bc	1.00 \pm	ab	823.44 \pm	b	22.54 \pm	cd	2.98 \pm	c
<i>Pinus nigra</i>	0.96 \pm	d	89.07 \pm	a	2.79 \pm	d	0.95 \pm	ab	512.33 \pm	b	37.15 \pm	ab	4.50 \pm	b
<i>Pinus pinaster</i>	0.84 \pm	d	53.20 \pm	b	3.25 \pm	d	1.48 \pm	ab	182.33 \pm	b	25.53 \pm	bc	3.01 \pm	c
<i>Pinus sylvestris</i>	1.01 \pm	d	59.98 \pm	b	3.30 \pm	d	0.83 \pm	ac	621.43 \pm	b	42.64 \pm	a	4.18 \pm	c
<i>Pseudotsuga menziesii</i>	1.11 \pm	cd	66.14 \pm	b	3.47 \pm	cd	1.41 \pm	c	904.38 \pm	b	22.83 \pm	d	4.35 \pm	b
<i>Quercus petraea</i>	1.36 \pm	bc	89.69 \pm	a	6.40 \pm	ab	1.60 \pm	a	1920.72 \pm	a	11.27 \pm	e	6.84 \pm	a
<i>Quercus robur</i>	1.64 \pm	a	105.82 \pm	a	6.84 \pm	ab	1.76 \pm	a	1129.10 \pm	a	14.70 \pm	de	7.61 \pm	a

705

706 **1. The consecutive disparity index (*D*)**

707

708 *D* assesses the consecutive variations in a time series and so is sensitive to real time-step to time-step
709 variations, conversely to the CV index, which is insensitive to temporal autocorrelation. *D* has been
710 used in climate research to better assess interannual variability in the highly irregular precipitation time
711 series of the Iberian Peninsula (Martín-Vide, 1986) and is calculated as:

$$712 \quad D = \frac{1}{n-1} \cdot \sum_{i=1}^{n-1} \left| \ln \frac{p_{i+1}}{p_i} \right| \quad \text{Equation 1}$$

713 where p_i is the series value and n is the series length. To avoid numerical indetermination (division by 0)
714 when a time series contains zeros, we can sum a constant (k , usually a unit) to the entire time series
715 as:

$$716 \quad D = \frac{1}{n-1} \cdot \sum_{i=1}^{n-1} \left| \ln \frac{p_{i+1} + k}{p_i + k} \right| \quad \text{Equation 2}$$

717 The core of *D* lies in the assessment of the variability by taking into account the consecutive changes in
718 a time series (see Eqs. 1 and 2). Additionally, on the contrary to CV (standard deviation · mean⁻¹), the
719 calculation of *D* does not take the mean of the time series into account, which makes it less dependent
720 on the mean. An easy example of the differences between the CV and the *D* indices can be seen in
721 **Figure S3**. Both time series have the same CV but completely opposite temporal behaviours. The first
722 time series (A) fluctuates every year, but the second time series (B) is stable during the first half and
723 then shifts to a second state. The higher consecutive interannual variability makes the *D* index to be
724 higher in time series A.

725 2. Estimating GPP

726 To estimate GPP for each forest, we constructed a linear model using data from a global forest
 727 database (Luysaert et al., 2007; Fernández-Martínez et al., 2014b) containing carbon fluxes and
 728 productivity data from forests. We extracted data from sites that reported annual GPP (mainly derived
 729 from eddy covariance CO₂ fluxes) and foliar NPP, as well as MAT, MAP, and leaf type. Abbreviations:
 730 mean annual precipitation (MAP), mean annual temperature (MAT), standard errors (SE), standardized
 731 coefficients (β), root mean squared error (RMSE), degrees of freedom (df).

732 Model summary

	Estimate	SE	β	SE	<i>t</i>	Pr(> <i>t</i>)	
(Intercept)	985.63	284.82	0.000	0.000	3.46	0.0008	***
	7	8				83	
	-						
Leaf type - conifers	239.18	236.92	-0.138	0.136	-1.01	0.3158	
	6	5				77	
						0.1238	
MAP	-0.422	0.271	-0.484	0.311	1.556	0.0624	
						63	
						0.6242	
MAT	-9.244	18.794	-0.068	0.137	0.492	0.6242	
						26	
						1.42E-	
Foliar NPP	4.627	0.798	0.517	0.089	5.795	0.0007	***
Leaf type - conifers:map	0.378	0.186	0.453	0.223	2.032	0.0456	*
						15	
						0.0077	
MAP:MAT	0.035	0.013	0.829	0.303	2.734	0.0077	**
						62	
	R²						
Leaf type	0.029		R²	0.7374			
MAP	0.219		R²_{adj}	0.7169			
MAT	0.222		RMSE	407.33			
			Error				
Foliar NPP	0.241		%	9.15%			
Leaf type:MAP	0.009		df	77			
				<0.00			
MAP:MAT	0.016		P	01			

Model crossvalidation (75% data as training test - 25% validation data)

	2.50%	50%	97.50%	Mean	SE
R²	0.636	0.707	0.767	0.706	0.033
R²_{adj}	0.616	0.691	0.754	0.690	0.034
RMSE on crossvalidation	268.76	396.57	526.76	396.70	67.72
	5	7	4	0	6
% error on crossvalidation		11.52	17.08	11.40	2.63
	7.38%	%	%	%	%

