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5

6 **THE RESPONSE OF SUB-MEDITERRANEAN GRASSLANDS TO RAINFALL**
7 **VARIATION IS INFLUENCED BY EARLY SEASON PRECIPITATION**

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23 **Running head:** Effects of precipitation variation on productivity

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31 **Abstract**

32 **Question:** Climate change will likely modify patterns of precipitation, with an expected increase of
33 intra-annual variability and increased frequency and magnitude of extreme events. The
34 Mediterranean area is expected to be very sensitive to such events as water availability is already
35 limited. However, the effect of precipitation variability on ecosystem services, such as plant
36 productivity, is widely unknown.

37 What is the short-term effect of an experimental precipitation gradient on the above ground net
38 primary productivity (ANPP) of two contrasting sub-Mediterranean grassland ecosystems? **Are**
39 **there effects of different intra-annual rainfall patterns, i.e. dry or wet early season (spring), on the**
40 **ANPP?** Do the functional groups of grasses and forbs differ in their reaction?

41 **Location:** Torricchio Nature Reserve, Central Apennines, Italy.

42 **Methods:** We selected two grasslands characterized by contrasting conditions in geophysical and
43 soil chemical parameters (north- and south-facing slopes). In both sites, during two climatically
44 different years, mid-season (summer) precipitation was manipulated in order to obtain a gradient of
45 rainfall availability, comprising additional rainfall, ambient rainfall conditions and rainfall
46 reduction. The above-ground biomass, subdivided according to the functional groups of forbs and
47 grasses, was collected at the end of each treatment period.

48 **Results:** A significant increase of the ANPP due to experimental increase in summer rainfall
49 appeared in the year with the wet spring, but only in the mesic north-facing slope. This response
50 was driven by the increased productivity of perennial forbs while grasses showed a stable
51 aboveground production. On the contrary, we found no positive effect of experimental increase in
52 summer precipitation on ANPP in the year with the dry spring. The variability of the ANPP
53 increased significantly in the xeric south-facing slope in the year with the wet spring, most likely
54 reflecting indirect effects of small-scale heterogeneity such as variations in soil depth.

55 **Conclusions:** Intra-annual precipitation variation can have noticeable implications for sub-
56 Mediterranean montane grassland agriculture: **livestock pressure should be limited** in years with an
57 irregular spring drought, regardless of summer precipitation, especially in mesic grasslands.

58

59 **Keywords:** ANPP; Climate change; Experiment; Functional group; Precipitation variation; Intra-
60 annual rainfall; Drought.

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62

63 **INTRODUCTION**

64 Water availability is the main constraint to plant productivity in many terrestrial biomes (Heisler-
65 White 2008; Hsu et al. 2012) and it is an ecosystem driver that will be strongly affected by climate
66 change (Houghton et al. 2001). In addition to changes in mean annual precipitation, general
67 circulation models predict increases in the intra-annual (seasonal) variability of precipitation (Hsu et
68 al. 2012), with potential effects on plant productivity according to the timing and size of
69 precipitation inputs (Swemmer et al. 2007; Heisler-White et al. 2009). In particular, the impact of
70 precipitation variability on grassland productivity represents a topic of current concern due to its
71 relevance for agricultural activities (Grime et al. 2000).

72 Endeavours to understand the effect of precipitation variability on productivity of grasslands
73 include different methodologies, i.e. studying temporal and spatial natural gradients (Nippert et al.
74 2006; Golodets et al. 2013) as well as manipulative experiments (Fay et al. 2003; Holub et al. 2013;
75 De Boeck et al. 2015). Experimental studies on the effects of precipitation variation have mainly
76 been performed in European temperate grasslands and the North American temperate tallgrass
77 prairie. From the European systems it has been concluded that aboveground productivity is hardly
78 affected by rainfall variability (Grime et al. 2008; Jentsch et al. 2011; Dengler et al. 2014). Effects
79 were seen in the North American tallgrass prairie where changes in rainfall, i.e. less frequent rainfall
80 and/or a reduction of the amount of rain per rainfall event, had a negative influence on plant

81 productivity (Fay et al. 2003). However, up to our knowledge and according to the review of
82 Miranda et al. (2011), there are no studies on grasslands in the Mediterranean basin which
83 represents a climatic transition zone between the temperate mid-latitude and the tropical dry
84 climate. This represents a crucial research gap, given the assumed high sensitivity of Mediterranean
85 systems to climatic alterations (Jongen et al. 2011). In the Mediterranean environment, water to
86 support primary productivity is a limiting factor, making it highly likely that grassland productivity
87 is affected by precipitation variation (Suttle et al. 2007; Jongen et al. 2011). Considering the
88 relevance of plant functional aspects in this context (Wellstein et al. 2011), different plant functional
89 groups, such as forbs and grasses, can respond differently to precipitation variation, which may in
90 turn influence productivity (Zavaleta et al. 2003). Also, xeric versus mesic site conditions can make
91 a difference for the response of productivity to precipitation variation (Holub et al. 2013; Knapp et
92 al. 2008; Swemmer et al. 2007). Mediterranean montane regions are characterized by large
93 topographic complexity as well as marked water availability gradients (Lavorel et al. 1998) calling
94 for the consideration of local site conditions when studying precipitation effects on grassland
95 productivity.

96 In this contribution, we evaluate the effects of experimentally induced gradients of water
97 availability on aboveground net primary productivity (ANPP) of a representative sub-Mediterranean
98 montane grassland landscape of the central Apennine Mountains (Torricchio Nature Reserve, Italy).
99 We compare the effects of experimental precipitation variation in two contrasting calcareous
100 perennial grasslands with mesic (north-facing slope with deeper soil) and xeric (south-facing slope
101 with extremely shallow soil) site conditions (Wellstein et al. 2013). Contrasting spring weather
102 conditions during the two-year study period (2011, 2012) allowed us to examine the influence of
103 spring precipitation on the response of ANPP to summer rainfall variation.

104

105 We expect (H1) that ANPP should increase with increasing summer precipitation, although (H2)
106 varying spring precipitation and (H3) dry (S-facing) vs. mesic (N-facing) grassland systems may

107 influence the response. In addition, (H4) we expect that different functional groups should respond
108 differently, with forbs being more responsive. Lastly, (H5) we assume that the variability of ANPP
109 should be highest on the dry and heterogeneous S-facing slope.

110

111 **MATERIALS AND METHODS**

112 **Study area**

113 The Torricchio Nature Reserve (Central Apennines, Italy; Appendix S1) provides montane
114 calcareous grasslands on Jurassic-Cretaceous limestone and is under protection since 1970;
115 previously, the grasslands were grazed. Mean annual precipitation reaches 1,250 mm and mean
116 annual temperature is around 11 °C (Halassy et al. 2005).

117 We selected two study sites with an area of about one hectare each, representing the contrasting
118 environmental conditions of the mesic north- and the xeric south-facing slope (Wellstein et al. 2013,
119 Appendix S1, S2, S3). The south-facing site is characterized by a shallow/skeletal soil with low
120 water holding capacity and higher soil heterogeneity compared to the north-facing slope (Wellstein
121 et al. 2013). For details on the vegetation see Wellstein et al. (2014).

122

123 **Experimental design**

124 In both sites, the precipitation manipulations consisted of a gradient of rainfall availability,
125 comprising additional rainfall (A, additional), ambient rainfall (C, control) and rainfall reduction
126 (D, decreased rainfall). At each site, five plots were established with a distance of at least twenty
127 meters between each other. Each plot was composed of three 1 m x 1 m sub-plots: one sub-plot in
128 the centre of a 4 m² roof to simulate rainfall reduction (D); one sub-plot downstream the rainout
129 shelter to simulate additional rainfall by receiving the additional rain which has fallen on the slope-
130 parallel inclined roof and dropped in the centre of the sub-plot (A); one sub-plot under ambient
131 rainfall conditions (C). Roofs were constructed with a steel frame and covered with transparent 3
132 mm plastic foil that permitted over 93% penetration of photosynthetically active radiation (PAR).

133 The duration of precipitation manipulations has been estimated applying the method of extreme
134 value distributions (1000-year event; Jentsch et al. 2007) using climate data series covering 50
135 years. For the Torricchio Nature Reserve, the 1000-year drought event resulted in 58.5 days without
136 precipitation. The experiment covered two consecutive years (2011 and 2012). According to the
137 determined length of the extreme drought period, sub-plots with shelters were roofed from May 31st
138 to July 27th 2011, and from May 22nd to July 20th 2012.

139 Following Ashcroft and Gollan (2013) we measured the microclimatic conditions of the soil. Soil
140 temperature and relative humidity (RH) of the soil air were measured during the second year every
141 two hours at high resolution (DS1923 iButton Hygrochron Temperature/Humidity logger, Maxim
142 Integrated, San Jose, CA, USA; precision of 0.0625 °C/0.04% RH) at both slopes directly below the
143 soil surface (-1 cm) with a replication of two per treatment. General meteorological data was
144 provided by the local meteorological station of the Torricchio Nature Reserve.

145 The total above-ground biomass was collected in a sampling area of 400 cm² (10 cm × 40 cm) in
146 each sub-plot once a year at the end of the treatment period (peak biomass harvest, a common
147 measure for ANPP). The size of the sampling area was used in previous similar studies in
148 Mediterranean grasslands (Golodets et al. 2013). Biomass samples were subdivided into the
149 functional groups of forbs and grasses, oven dried (48 hours at 80 °C) and weighed.

150

151 **Seasonal patterns of ambient rainfall**

152 A detailed report on the timing and amount of ambient precipitation in each study year is given
153 in Appendices S4 and S5. From these data it emerges that the rainfall pattern was completely
154 different between study years, with the first year (2011) having a higher shortage of water
155 underneath the roofs and a higher amount of water in the additional rainfall sub-plots during
156 treatment. With respect to the local average precipitation of the season (Venanzoni 2003), the spring
157 of 2011 was very dry while that of 2012 showed a higher level of precipitation (see Appendix S5 for
158 details).

159

160 **Data analysis**

161 The hypotheses H1, H2 and H3 were tested using for each study year (2011, 2012) and each
162 system (mesic, xeric) a trend-test for a monotonic trend, namely the Jonckheere-Terpstra Test (R
163 package *DescTools*) (Jonckheere 1954; Terpstra 1952). To test H4, statistical tests were repeated for
164 the functional groups of grasses and forbs separately. H5 was tested using a trend-test for monotonic
165 trend in variance suggested by Neuhauser and Hothorn (2000) (R package *lawstat*) for the total as
166 well as for the functional group ANPP .

167 For each trend test, we applied Bonferroni correction as we tested two grassland systems and two
168 years, resulting in a significance threshold of $\alpha = 0.0125$ (= 0.05 divided by 4 tests). Furthermore, to
169 test the effect of ambient rainfall on ANPP, we compared the productivity in control sub-plots
170 between the two years in each slope.

171 For the relative humidity of the soil air, daily mean values were calculated for each site (slope),
172 and treatment. For the soil temperature, differences between roofed and non-roofed sub-plots were
173 tested using the Mann-Whitney U-test. Prior to analysis, both temperature and humidity
174 observations of the iButtons were corrected using internal, sensor specific, factory supplied
175 calibration data (iButton – DS1923). As the iButtons can saturate under humid conditions producing
176 values higher than 100%, the humidity observations were corrected applying a scaling and
177 correction procedure (Ashcroft and Gollan 2013).

178 All tests were performed using the software R 3.0.3 (R Development Core Team 2014).

179

180 **RESULTS**

181 **Effects of treatments on micro-climate**

182 The relative humidity of the soil air generally decreased during the season from spring to
183 summer in all treatments during the experiment (Fig. 1). Differences between treatments were
184 visible at both slopes with lower relative humidity values occurring in rainfall reduction treatment

185 and higher values in the additional rainfall treatment. Shelters had a significant effect on soil
186 temperature. The mean difference between roofed and non-roofed plots in the north-facing slope (-1
187 cm) was 0.66°C ($p = 0.009$) and in the south-facing slope 1.02°C ($p = 0.014$).

188

189 **Effects of treatments on productivity**

190 A comparison of the ambient rainfall plots (C sub-plots) was made to show the differences
191 between the two study years under ambient conditions: the year with the wet spring (2012) showed
192 significantly higher total ANPP under ambient rainfall than the year with the dry spring (2011).
193 These significant differences were seen in both grassland systems but were higher in the north-
194 facing slope (ANPP increase by 75%; $p = 0.019$) than in the south-facing slope (ANPP increase by
195 16%; $p = 0.049$).

196 When testing the hypotheses on the effect of the rainfall gradient on ANPP, the total ANPP
197 showed no significant response to the gradient of rainfall availability in the year with the dry spring
198 in either study system (Table 1, Fig. 2a,b). A significant increase of ANPP with increasing rainfall
199 availability was seen, however, in the year with the wet spring (2012) and this was only apparent in
200 the grassland system of the north-facing slope (Table 1, Fig. 2a). There the total ANPP increased by
201 more than one half (52%) with increasing rainfall availability while no significant changes of ANPP
202 were seen at the south-facing slope (Table 1, Fig. 2b). When testing the functional groups
203 separately, a significant increase was found only for the forbs in the north-facing slope while the
204 grasses showed no changes in aboveground production (Table 1, Fig. 3).

205 The total ANPP showed a significant increase in variability with increasing rainfall availability in
206 the second year of precipitation manipulation only on the south facing slope (Table 2). When
207 looking at the functional groups of forbs and grasses on this slope, a significant increase in
208 variability with increasing rainfall availability was seen only for the grasses (Table 2). The forbs
209 showed a significant increase in the variability of ANPP only on the north-facing slope in the
210 second year (Table 2).

211

212 **DISCUSSION**

213 *Microclimate*

214 During the two seasons of weather manipulation, fixed shelters proved to be effective tools for
215 altering the amount of rainfall. Judging from the limited sample size of relative humidity of the soil
216 air measurements, the treatment led to continuous differences in the levels of relative humidity of
217 the soil air directly below the soil surface (Fig. 1). **While values of relative humidity of the soil air**
218 **close to 100% have no effect on plants**, strong drops indicate strong changes in the water
219 availability for plants as the permanent wilting point of plant species is highly sensitive to changes
220 in relative humidity of the soil air (Lal and Shukla 2004).

221 The increase in relative soil air humidity during rainfall events in the rainfall reduction treatment
222 (Fig. 1) depends probably on water runoff during major precipitation events. However, we did not
223 aim to test for total exclusion of water availability but rather aimed to establish a gradient of water
224 availability.

225 Furthermore, temperature alterations observed under the rainout shelters were relatively low and
226 generally comparable to those seen in other fixed shelter designs (Fay et al. 2000 and references
227 therein).

228

229 *Early season precipitation conditioning ANPP*

230 The positive response of ANPP towards experimental precipitation increase was seen only in the
231 year with the wet spring (2012), rejecting H1 and confirming H2, and only on the mesic north-
232 facing slope, confirming H3. This effect was seen despite the lower summer water availability
233 during the experiment in 2012 compared to 2011 (Appendix S5). This confirms the findings of
234 Swemmer et al. (2007) and Heisler-White et al. (2009) reporting that the distribution and size of
235 precipitation events can affect ANPP independently of the precipitation amount. The fact that the
236 response was significant only in 2012 is more likely related to the climatic differences of the spring

237 prior to the experiment than to the recurrence of the treatments. This is supported by the
238 significantly higher ANPP under ambient conditions in 2012 than in 2011. The response of ANPP
239 towards experimental precipitation increase in the year with the wet spring is in line with the
240 findings of both experimental and modelling studies. Epstein et al. (1999) found that the
241 precipitation seasonality is the most important factor accounting for variation in ANPP in a model
242 simulation and Suttle et al. (2007) report dramatic changes in Mediterranean grassland productivity
243 after spring water addition.

244 Looking into the mechanisms behind the relevance of the water availability in spring for plant
245 growth, we hypothesize that in the studied sub-Mediterranean context the physiological and
246 morphological plant adaptation at the beginning of the growing season leads to a higher growth
247 potential. This capacity can then be used to raise productivity under increased precipitation in
248 summer. The review of Zeppel et al. (2014) emphasizes that the seasonal distribution of
249 precipitation influences processes triggering plant growth such as tiller production, root-shoot
250 biomass, root depth, canopy leaf area, stomatal conductance and photosynthesis.

251 Evidence of adaptation capacity of functional growth traits to fine-scale environmental variation
252 (Wellstein et al. 2013) makes adaptations to environmental variations over time likely, such as
253 changes in plant water availability due to altered precipitation patterns.

254

255 ***Response of ANPP to increased rainfall in mesic vs. xeric grassland***

256 While in the year with the wet spring, the mesic north-facing slope showed a strong increase (>
257 50%) of the ANPP with experimentally increased precipitation, we did not find such positive effects
258 on the xeric south-facing slope, most likely due to the extremely shallow soil with low water
259 holding capacity (Wellstein et al. 2013). This is in line with the results of Buckland et al. (1997)
260 who claimed that plants growing on shallower soils benefit less of water addition. However,
261 Zavaleta et al. (2003) found that the timing of rainfall events and midseason droughts can influence
262 species productivity in a relatively xeric Mediterranean grassland. **In conditions with much higher**

263 amount of precipitation (climatic gradient from 90 to 780 mm mean annual precipitation) than in
264 our study, Golodets et al. (2013) demonstrate that arid pastures profited more from increased
265 precipitation during the growing season than mesic ones. Similarly, Holub et al. (2013) found a
266 significant effect of increased water availability at the driest site, while the moistest site did not
267 respond. These contrasting findings reinforce the hypothesis that it might generally depend on
268 ecosystem properties such as soil characteristics as well as on the amount of precipitation variation
269 whether ANPP reacts to rainfall variation or not.

270

271 *Forbs vs. grasses response in ANPP*

272 Looking into the functional groups, we found that forbs profit more from increased water
273 availability than grasses as a significant ANPP increase was found only for forbs (Tab. 2, Fig. 3)
274 confirming H4 for the mesic north-facing slope in the year with the wet spring. This finding is in
275 line with results of another study showing that increased precipitation enhanced forb production but
276 affected grasses little (Zavaleta et al. 2003). The study of Suttle et al. (2007) also points towards the
277 hypothesis of the responsiveness of forbs to precipitation demonstrating that the strongest initial
278 productivity response after an extended spring rainfall is by forbs followed by a more complex
279 community response in the subsequent years. Our result of a stable productivity of grasses was also
280 seen in other systems (Fay et al. 2003). Our findings emphasize the importance to include plant
281 functional aspects when studying the effect of climate change on ecosystem services (Wellstein et
282 al. 2011).

283

284 *ANPP variability*

285 Only on the xeric south-facing slope the variability of ANPP increased with experimentally
286 increased precipitation in the year with the wet spring (Tab. 2) confirming hypothesis H5 of a
287 higher variability in the system with higher soil heterogeneity (Wellstein et al. 2013). Depending on
288 the fine-scale variability of the water retention capacity of the soil (Wellstein et al. 2013), the plants

289 could or could not profit from rainfall events making a high variability in the ANPP response likely.
290 Looking into the differences of functional groups, we found an increase of ANPP variability with
291 increasing precipitation for the forbs only on the north-facing slope. This result was unexpected as
292 the N-facing slope does not have as high soil heterogeneity as the S-facing slope. . We suggest that
293 plant-plant interactions such as the competition with the dense grass carpet in the mesic system
294 (Wellstein et al. 2014) might be responsible for the significant ANPP variation of forbs meaning that
295 the growth of some, but not all, forb species was affected by competition. However, on the S-facing
296 slope grasses ANPP variability increased, most likely triggered by soil heterogeneity.

297

298 **CONCLUSIONS**

299 Climate models predict an increase of intra-annual precipitation variability including both the
300 quantity and the timing for the European Mediterranean basin (Bolle 2012). The results of our study
301 show that intra-annual precipitation variation can have important implications for sub-
302 Mediterranean montane grasslands. The relatively low yield of the mesic grasslands could be
303 strongly increased in wet springs. On the contrary, dry springs could inhibit this positive effect by
304 affecting the growth capacity of plants. In particular, functional groups should be considered as they
305 can play a key role in driving these responses.

306 This might have implications for adapting montane agriculture to climatic change: livestock
307 pressure should be limited in years with an irregular spring drought, regardless of summer
308 precipitation, especially in mesic grasslands. This is in line with the recommendation of Catorci et
309 al. (2012) to strongly reduce grazing in dry periods in sub-Mediterranean grasslands.

310 Lastly, our results call for caution when interpreting the outcome of experimental climate
311 manipulations as the climatic conditions of the period prior to the experiment can influence the
312 responses of plant species and vegetation. For this reason we encourage studies dealing with
313 precipitation manipulation at the beginning of the growing season.

314

315

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416 **List of electronic appendices**

417 **Appendix S1.** Land use types, topography and position of the plots in in the north-facing and the
418 south-facing slope of the study area.

419 **Appendix S2.** Geo-physical characterization of the sampling sites.

420 **Appendix S3.** Mean number of species per treatment per year for each sampling unit of each slope

421 **Appendix S4.** Estimation of the rainfall, for each treatment and year, from the beginning of the
422 growing season to the end of the experiment.

423 **Appendix S5.** General overview of the precipitation patterns in the two experimental years.

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442 **TABLES**

443 **Table 1**

444 Results of the **trend test for monotonic decline in aboveground net primary productivity (ANPP)**
 445 **with reduced rainfall availability** (Jonckheere-Terpstra test) in the mesic north-facing (site N) and
 446 the xeric south-facing (site S) grassland system in the two years of precipitation manipulation.

447 Results are shown for the total ANPP and for the functional groups of grasses and forbs.

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site	year	functional group	Aboveground biomass mean (g/m ²)			JT test-statistic	p-value
			reduced rainfall	ambient rainfall	additional rainfall		
N	2011	all	182	189	215	49	0.118
		grasses	135	155	160	33	0.703
		forbs	47	34	55	42	0.348
N	2012	all	226	332	475	66	0.001
		grasses	192	269	273	46	0.199
		forbs	34	63	202	72	0.000
S	2011	all	150	187	241	43	0.306
		grasses	119	58	169	47	0.180
		forbs	31	128	72	52	0.066
S	2012	all	139	216	299	54	0.044
		grasses	103	58	157	36.5	0.554
		forbs	36	158	142	59	0.013

Significant p-values are given in bold (p < 0.0125 after Bonferroni correction)

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454 **Table 2**

455 Results of the Neuhauser-Hothorn test for monotonic trend in variances of aboveground net primary
 456 productivity (ANPP) with increasing rainfall availability in the mesic north-facing (site N) and the
 457 xeric south-facing (Site S) grassland system in the two years of precipitation manipulation.
 458 Significant p-values suggest increasing variance with increasing water availability. Results are
 459 shown for the total ANPP and for the functional groups of grasses and forbs.

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site	year	functional group	test-statistic	p-value
		all	0.683	0.340
N	2011	grasses	2.572	0.023
		forbs	0.600	0.370
		all	1.915	0.066
N	2012	grasses	1.346	0.151
		forbs	3.277	0.008
		all	1.800	0.078
S	2011	grasses	1.171	0.191
		forbs	2.168	0.044
		all	3.715	0.004
S	2012	grasses	2.999	0.012
		forbs	1.763	0.083

Significant p-values are given in bold ($p < 0.0125$ after Bonferroni correction)

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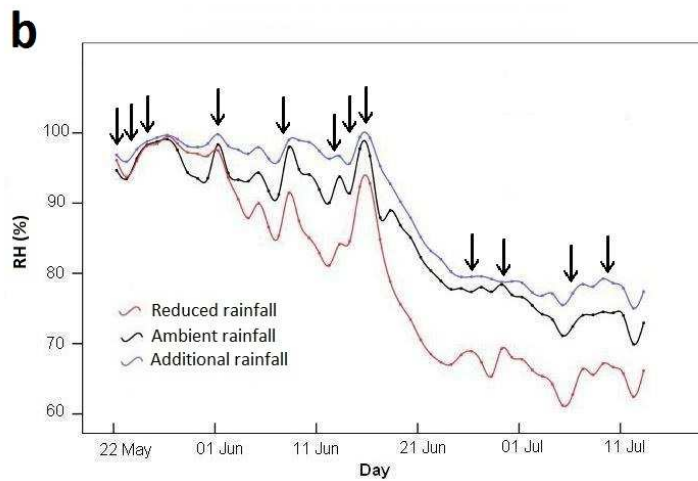
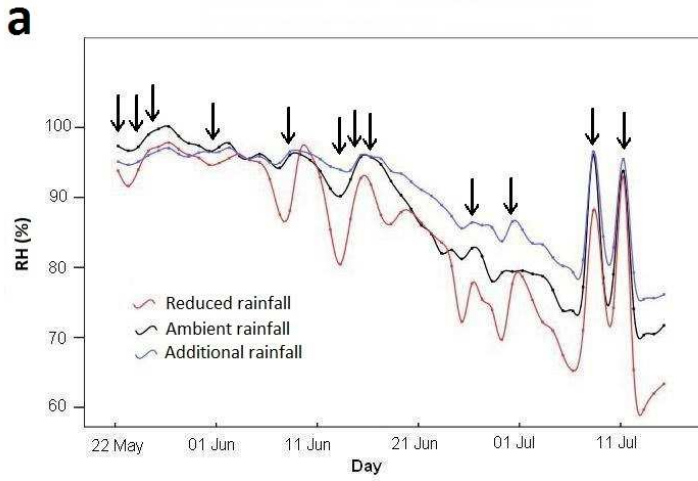
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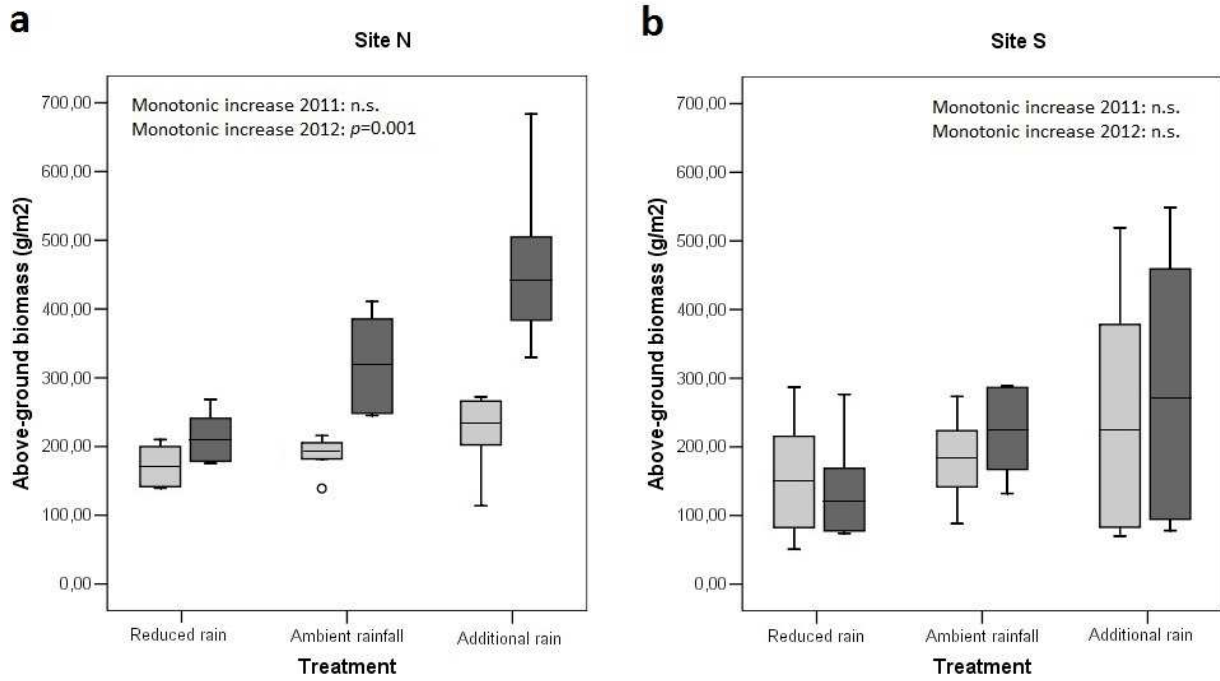
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466 **FIGURES**

467 **Fig. 1.** Daily relative humidity of soil air (relative humidity of the soil air %) in the year 2012,
468 measured directly below the soil surface in each treatment at (a) the north-facing slope and (b) the
469 south-facing slope. Arrows represent rainfall events.



473 **Fig. 2.** Aboveground net primary productivity (ANPP) in the mesic north-facing (a, site N) and the
474 xeric south-facing (b, site S) slope for three treatments of precipitation manipulation in each year
475 (2011 in light grey, 2012 in dark grey).



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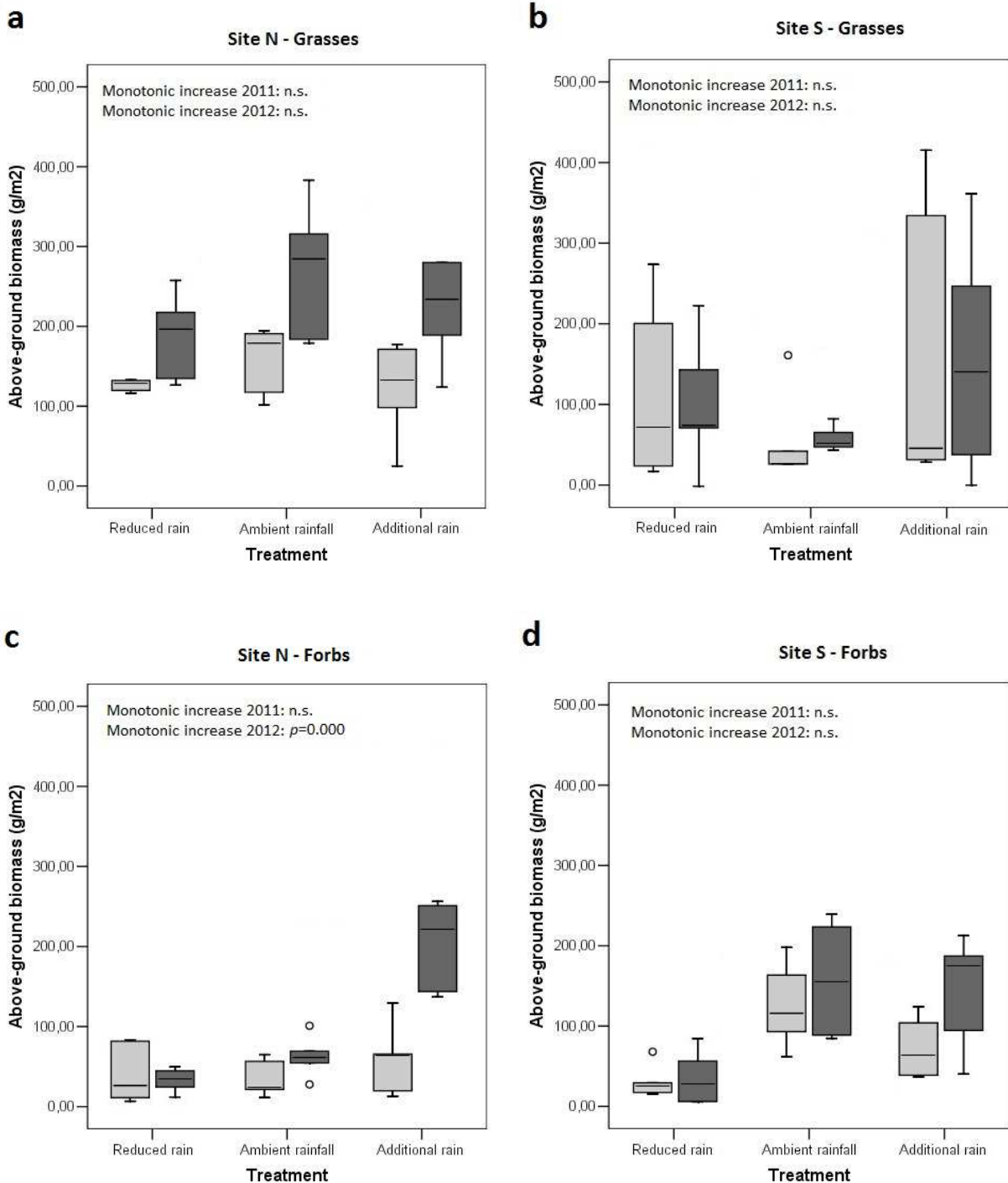
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490 **Fig. 3.** Aboveground net primary productivity (ANPP) of the functional group of grasses in the
 491 north- (a) and south-facing slope (b) and the functional group of forbs in the north- (c) and south-
 492 facing slope (d). The ANPP is displayed for the three treatments of precipitation manipulation in
 493 each year (2011 in light grey, 2012 in dark grey).



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