

1 Dynamic Reservoir-Condition Microtomography of Reactive Transport in
2 Complex Carbonates: Effect of Initial Pore Structure and Initial Brine pH

3

4 *Menke, H.P.¹; Bijeljic, B.¹; Blunt, M.J.¹.

5

6 ¹Qatar Carbonates and Carbon Storage Research Centre (QCCSRC), Department
7 of Earth Science and Engineering, Imperial College London, SW7 2AZ, London,
8 United Kingdom

9 *Corresponding Author. email: h.menke12@imperial.ac.uk, phone:+44
10 07543547195

11

12 ABSTRACT

13

14 We study the impact of brine acidity and initial pore structure on the
15 dynamics of fluid/solid reaction at high Péclet numbers and low Damköhler
16 numbers. A laboratory μ -CT scanner was used to image the dissolution of Ketton,
17 Estailades, and Portland limestones in the presence of CO₂-acidified brine at
18 reservoir conditions (10 MPa and 50°C) at two injected acid strengths for a
19 period of 4 hours. Each sample was scanned between 6 and 10 times at $\sim 4 \mu\text{m}$
20 resolution and multiple effluent samples were extracted. The images were used
21 as inputs into flow simulations, and analysed for dynamic changes in porosity,
22 permeability, and reaction rate. Additionally, the effluent samples were used to
23 verify the image-measured porosity changes.

24 We find that initial brine acidity and pore structure determine the type of
25 dissolution. Dissolution is either uniform where the porosity increases evenly
26 both spatially and temporally, or occurs as channelling where the porosity
27 increase is concentrated in preferential flow paths. Ketton, which has a relatively
28 homogeneous pore structure, dissolved uniformly at pH = 3.6 but showed more
29 channelized flow at pH = 3.1. In Estailades and Portland, increasingly complex
30 carbonates, channelized flow was observed at both acidities with the channel
31 forming faster at lower pH. It was found that the effluent pH, which is higher
32 than that injected, is a reasonably good indicator of effective reaction rate during
33 uniform dissolution, but a poor indicator during channelling. The overall
34 effective reaction rate was up to 18 times lower than the batch reaction rate
35 measured on a flat surface at the effluent pH, with the lowest reaction rates in

36 the samples with the most channelized flow, confirming that transport
37 limitations are the dominant mechanism in determining reaction dynamics at the
38 fluid/solid boundary.

39

40 1. INTRODUCTION

41

42 Carbon Capture and Storage (CCS) has been proposed as the primary
43 solution for reducing anthropogenic carbon dioxide (CO₂) emissions. Accurate
44 predictive modelling is imperative for assessing CO₂ storage security (Herzog et
45 al., 2003; Metz et al., 2005). When supercritical CO₂ mixes with carbonate aquifer
46 brine at great depth it forms carbonic acid that can partially dissolve the host
47 rock and alter the storage capacity and permeability of the formation (Langmuir
48 et al., 1997; Morse and Mackenzie, 1990; Peng et al., 2013). The amount and type
49 of dissolution will determine both storage capacity (Li et al., 2006; Rötting et al.,
50 2015) and how the CO₂ and CO₂-acidified brine moves in the subsurface
51 (Johnson et al., 2002; Molins et al., 2012; Steefel et al., 2013). A complete
52 understanding of the brine-rock system is needed to understand and predict the
53 rate of fluid movement and distribution in the aquifer (Daccord et al., 1993a;
54 Daccord et al., 1993b; Maheshwari et al., 2013).

55 Carbonate dissolution by CO₂ acidified brine is controlled by many factors
56 including pore space heterogeneity, initial pH, and flow conditions (El-Maghraby
57 et al., 2012; Fredd and Fogler, 1998; Luquot and Gouze, 2009). At reservoir
58 conditions the acidity of the brine is strongly dependent on temperature and
59 pressure (Peng et al., 2013). Additionally, the acid pH changes depending on the
60 amount of calcium dissolved in solution (Leal et al., 2013; Peng et al., 2015).

61 Imaging of dissolution in limestone rock has been performed at the core
62 (~cm) scale using Computed Tomography (CT). Ott et al. (2012) investigated
63 dissolution where, for several rock types and flow regimes, it was found that
64 reaction of the dissolved CO₂ with the solid matrix increases physical
65 heterogeneity. Ellis et al. (2011) used core-scale imaging to show that exposure
66 to CO₂ progressively reduced the sealing capacity of fractured claystone
67 caprocks. Smith et al. (2012) imaged dissolution of a small dolomitic core before
68 and after reaction with CO₂ acidified brine. Lamy-Chappuis et al. (2014) used CT

69 scanning to measure calcite dissolution in calcareous gritstone and found that
70 the permeability increase with porosity did not agree with classical models and
71 that the dissolution rate was largely due to the acid supply. Tutolo et al. (2014)
72 examined dolomite dissolution in the presence of CO₂ saturated brine and found
73 evidence of dissolution at low temperatures followed by exsolution and
74 precipitation as the temperature increased. Ott and Oedai (2015) examined
75 carbonate dissolution in the presence of two-phase flow and found large
76 differences in dissolution regimes depending on the CO₂ saturation.

77 Core-scale imaging is valuable because it is three-dimensional and can be
78 temporally resolved. However, it has a maximum resolution of ~50 μm and thus
79 can only detect the fluid/solid boundary of large features such as fractures and
80 wormholes. The greater portion of the pore space is only resolved as an average
81 greyscale value that blurs or eliminates the solid/liquid interface thus making
82 segmentation and analysis of local changes in connectivity and reaction rate
83 difficult. There are relatively few experiments that investigate the impact of
84 coupled dissolution and flow phenomena at the micron scale – observations that
85 are necessary for the digitisation of the pore network for direct numerical
86 simulations of both flow and transport.

87 X-ray microtomography (μ-CT) is an accepted method of studying pore-
88 scale processes (Blunt et al., 2013; Gouze and Luquot, 2011). μ-CT creates three-
89 dimensional images with a high spatial resolution (~1 μm) non-invasively. Two
90 types of light sources are used for μ-CT imaging: synchrotron and laboratory.
91 Synchrotron light sources are only sporadically accessible, but have powerful x-
92 ray beams that can take images very quickly. Laboratory scanners are less
93 expensive and more widely available, but have a weaker x-ray source that must
94 be carefully fine-tuned to study dynamic processes. Carbonate rocks are
95 extremely diverse and complex (Archie, 1952) and thus it requires many
96 experiments to determine the principal controls on dissolution for a range of
97 carbonate rock types. Therefore, the development of a laboratory-based
98 technique that allows for reservoir condition imaging of complex pore-structures
99 with a temporal resolution sufficient to study dynamic processes is imperative
100 for acquiring the breadth of data needed for complete understanding of pore-
101 scale processes in carbonate rocks. In an imaging experiment, the reaction rate

102 can be calculated from the measured change in porosity. However, it is not
103 possible to assess the accuracy of this approach without effluent analysis. Thus
104 it is also important to complement image-based measurements with more
105 conventional geochemical sampling. In this paper we present a benchmark
106 comparison between imaging and geochemical sampling using a new laboratory
107 method described below.

108 There have been several studies that have successfully used μ -CT to study
109 dynamic pore-scale processes. Berg et al. (2013) and Andrew et al. (2015) both
110 used a synchrotron light source to look at dynamic pore-scale displacement
111 events between multiple fluid phases with around ~ 1 min temporal resolution.
112 Other work has looked at fluid/solid reaction in complex carbonates, but were
113 either not performed at reservoir conditions, or were limited to the study of pre
114 and post reaction images (Hao et al., 2013; Luquot et al., 2014; Noiriél et al.,
115 2013; Smith et al., 2013). Menke et al. (2015) imaged the dissolution dynamics of
116 a limestone with a relatively simple pore structure at reservoir conditions and
117 found that the dissolution was uniform throughout the sample along the axis of
118 flow. Menke et al. (2016) examined the dissolution of two heterogeneous
119 carbonates at two flow rates and found a new dissolution regime, channelling, in
120 which conditions for formation were dependent on initial pore topology rather
121 than flow conditions. However, all used an initial brine pH of 3.1 and neither set
122 of experiments included effluent analysis, while the samples were either limited
123 to homogenous rocks that were relatively easy to image, or were performed
124 using a synchrotron light source with specialized equipment.

125 This paper describes reactive dissolution processes at reservoir
126 conditions at the pore scale in both homogenous and heterogeneous carbonates
127 using a laboratory scanner and investigates the impact of initial brine pH on
128 dissolution regimes combined with effluent analysis. We will focus on measuring
129 pore-scale effective reaction rates between two CO₂-acidified brines of pH = 3.1
130 and 3.6 and limestone rock. We observe two distinct dissolution regimes –
131 uniform and channelling – whose emergence and rate of development are
132 dependent on initial structural heterogeneity and initial brine pH.

133 To investigate the effect of pore-scale heterogeneity we have selected
134 three limestones: Ketton, Estailades and Portland. Ketton has been shown to

135 have homogeneous flow characteristics, in that the distribution of velocities
136 computed on an image before reaction is narrower than observed for Estailades
137 and Portland (Bijeljic et al., 2013a; Bijeljic et al., 2013b). We present a dynamic *in*
138 *situ* study of carbonate dissolution at reservoir conditions with effluent sampling
139 to track the calcium concentration. Image sequences are binarized and then
140 analysed for porosity and surface area changes. Furthermore, the images are
141 used as input into flow models to elucidate the evolution of permeability and
142 pore-scale velocity fields. Additionally, effective reaction rates are calculated for
143 both effluent samples and CT images and compared, demonstrating both the
144 accuracy of μ -CT imaging of fluid-rock reaction and the formation process of
145 different dissolution structures.

146

147 2. METHODS

148

149 2.1 Sample selection

150

151 To investigate the effect of pore-scale heterogeneity three limestone
152 carbonates are examined: Ketton, Estailades, and Portland Basebed. These
153 samples were chosen because they were readily available, relatively chemically
154 pure, and diverse in pore structure. Core plugs used for bulk scale analysis were
155 drilled from the same 1 m³-block of rock as experimental samples to minimize
156 local heterogeneity. The pore throat radii calculated from the Mercury Intrusion
157 Capillary Pressure (MICP) curves are show in Figure 1 and the bulk rock
158 characteristics are listed in Table 1. Portland has a unimodal porosity while both
159 Ketton and Estailades have bimodal pore structures. However, only the larger
160 porosity modes in Ketton and Estailades are accessible to μ -CT scanning at 4- μ m
161 resolution. All pores below the resolution of 4- μ m are classified as microporosity
162 in this study.

163

164 2.2 Experimental apparatus

165

166 The Zeiss Versa XRM-500 μ -CT scanner was used to image reaction
167 between calcite and CO₂ saturated brine at reservoir conditions. Using our *in situ*

168 apparatus [Figure 2], ~1 cm long 4 mm diameter carbonate cores of Ketton,
169 Estailades, and Portland Basebed limestone were reacted by injecting both pH
170 3.6 and pH 3.1 supercritical (sc) CO₂ saturated brine.

171 Supercritical CO₂ and 1% *wt* KCl 5% *wt* NaCl brine were pre-equilibrated
172 at experimental conditions (10 MPa & 50°C) in a heated Hastelloy reactor with
173 an entrainment stirrer. Powdered particles of the host rock were used to raise
174 the pH to 3.6 for three of the experiments. An effluent sample was taken from the
175 reactor at the beginning of each experiment to verify initial pH (V5 & V3 in
176 Figure 2); however, as we show later, once reaction was established the pH in
177 the rock, as determined from the effluent, was around 1 unit higher than the
178 injected value.

179 Once fully saturated, the brine was pulled through the core at a constant
180 flow rate of 0.1 mL.min⁻¹. Tomographic images were taken successively for 4
181 hours at the centre of the core with a 4 mm field of view. Only the centre of the
182 core was imaged in order to take as many dynamic images as possible during the
183 experimental time.

184 A thermocouple inserted into the radial port of the Hassler cell coupled
185 with Kapton® heating tape and a proportional integral derivative controller
186 regulated cell temperature to within 1°C. The high-pressure ISCO
187 (www.isco.com) syringe pumps regulated flow accurately to 0.001 mL.min⁻¹. The
188 core was wet initially with a highly-absorbing 25% *wt* KI unreactive brine that
189 made arrival of the low absorbing reactive brine from the reactor easy to
190 observe and made dead volume calculations unnecessary.

191 Sample heterogeneity determines the speed at which accurate scans can
192 be taken, as more complex and smaller structures require larger numbers of
193 projections to image with sufficient edge sharpness for segmentation. Each rock
194 was first scanned without any fluid at 2400 projections to create an extremely
195 sharp map of the pore space before reaction. Then reactive flow was started and
196 between 6 and 10 scans were taken successively as fast as the scanner allowed
197 with between 400 and 800 projections to resolve the large changes in structure
198 over the course of 4 hours.

199

200 2.3 Effluent extraction and calculation of pH

201

202 Effluent extraction was carried out during the entire 240 minutes imaging
203 period. 2 mL of effluent was extracted from the effluent stream using a VICI
204 (www.vici.com) electronically actuated 2-position 6-port high pressure high
205 temperature HPLC valve (Figure 2, V12) outside the μ -CT structure
206 approximately every 40 minutes. As the sampling loop contained 2 mL of fluid
207 and the flow rate in the experiments was $0.1 \text{ mL}\cdot\text{min}^{-1}$, each sample represents
208 the average fluid over a 20-minute period. Samples were immediately acidified
209 using a 2% *wt.* HNO_3 solution and then analysed for $[\text{Ca}^{2+}]$ using an ICP-MS. The
210 pH was then computed using the geochemical equations found in Leal et al.
211 (2013) using the Ca^{2+} concentration in the effluent as input and assuming that
212 the brine remains fully saturated with CO_2 .

213

214 2.4 Image processing

215

216 The raw projections were reconstructed and artefacts and beam
217 hardening removed with the Zeiss (www.zeiss.com) proprietary imaging
218 software. The 2000^3 images were binned on the camera to increase signal to
219 noise resulting in a 1000^3 image of $\sim 4 \mu\text{m}$ voxel size. Further processing was
220 performed by the image processing modules in Avizo 9.1 program
221 (www.vsg.com).

222 The images [Figure 3] were first filtered using the non-local means filter
223 (Buades et al., 2005; Buades et al., 2008). The first scan in the series is the dry,
224 unreacted scan of the core, has 2400 projections and constitutes time = 0 min.
225 This high projection rate means that the image is very sharp, small pores are
226 very clearly resolved and thus can easily be segmented. High projection rates are
227 only possible for experiments where scan times are long and thus it is not
228 possible to do *in situ* imaging at such high rates. However, taking a high-quality
229 scan before starting the experiment has the advantage of preserving this spatial
230 information for later integration.

231 All images were registered to the first scan in each series (hereafter called
232 the reference scan) with normalized mutual information. The scans were then
233 resampled to the reference scan grid using the robust Lanczos method (Lanczos,

234 1950) so that dynamic changes could be computed on a voxel by voxel basis. The
235 greyscale reference scan was then segmented using the watershed segmentation
236 method (Schlüter et al., 2014) resulting in a binary image of the unreacted core.

237 Dynamic scans of highly heterogeneous rocks during dissolution are
238 difficult to segment because the rock-fluid boundaries are not sharp both due to
239 the boundaries changing during the scan and the low projection rate which
240 means that the small pores and small surface features are not clearly resolved.
241 Thus, we use the difference image approach to segmentation (Menke et al.,
242 2016). A difference image was computed between the greyscale reference and
243 reacted scans that was then filtered and segmented by watershed created a
244 segmented version of the changing pore space. The segmented difference image
245 was then subtracted from the segmented dry scan to achieve a segmented image
246 for each time step.

247 The pore space was then separated into areas with and without reaction
248 for analysis of the effect of local flow on reaction rate [Figure 4]. The segmented
249 difference between two consecutive images [Figure 4a,b] was obtained [Figure
250 4c]. The entire pore space of the first image [Figure 4d] was then individually
251 labelled for each unconnected channel on each slice [Figure 4e]. These labels
252 were then separated into labels that intersect with the reacted voxels in the
253 segmented difference image and labels that do not intersect [Figure 4f]. In this
254 way it is possible to isolate the channels with reaction from channels without
255 reaction.

256

257 2.5 Flow Modelling

258

259 Flow through the pore space of the segmented images was computed by
260 solving both Navier-Stokes and the volume conservation equations using the
261 finite volume method in OpenFoam as described in Bijeljic et al. (2013b); Raeini
262 et al. (2012). We apply a constant pressure boundary condition at the inlet and
263 the outlet faces of the images, whereas no-slip boundary conditions are
264 implemented on the other faces. Each voxel in the segmented images is a grid
265 block in the flow simulator that computes velocities at the grid block centres.
266 Later in the Results section we visualise the ratios of the magnitude of u at the

267 voxel centres divided by the average pore velocity $U_{av}=q/\phi$, where q is the Darcy
268 velocity in $\text{m}\cdot\text{s}^{-1}$ and ϕ is the porosity, to study dynamic changes in the velocity
269 fields both in the full pore space, and only in the channels which later react.

270

271

272 3. RESULTS

273

274 We study the impact of reaction rate and initial pore structure on
275 dissolution for three limestones: Ketton, Estailades, and Portland Basebed
276 [Table 2]. First, in experiments 1 and 2, brines of $\text{pH} = 3.1$ and 3.6 are injected
277 into Ketton limestone to investigate the impact of reaction rate on dissolution.
278 Then in experiments 3-6 the more heterogeneous limestones, Estailades and
279 Portland, are reacted using the same conditions to study the impact of physical
280 heterogeneity.

281 The images are first binarized and analysed for changes in porosity and
282 surface area. Then a flow solver is run on the segmented images to examine the
283 dynamic changes in permeability and pore velocity. Additionally, we calculate
284 the time-resolved changes in the dimensionless numbers that are traditionally
285 used to describe flow and reactive conditions. Moreover, we examine the
286 predicted overall reaction rate from effluent analysis versus the reaction rate
287 observed by dynamic x-ray microtomography.

288

289 3.1 Porosity, Permeability, and Surface Area

290

291 Each segmented image series was analysed for changes in porosity by
292 counting the number of voxels of pore and rock. Figure 5 depicts the slice-
293 averaged porosity along the axis of flow for each image through time. Porosity
294 increases with time during dissolution and in all six experiments the porosity
295 increased relatively uniformly along the axis of flow with the exception of
296 Estailades, which showed slightly more dissolution near the inlet. In experiment
297 1 the porosity also increases uniformly through time. Conversely, in experiments
298 2-6 the majority of the porosity increase occurs in the first 120 minutes. As
299 expected, the porosity increases faster with a lower pH when experiments 3, 4, 5,

300 and 6 are compared. However, the total increase in porosity is highest for
301 experiment 1 (24%) and lowest for experiment 6 (6%) and thus the total amount
302 of dissolution does not appear to be strongly correlated with the pH of the
303 reactive fluid.

304 Permeability was calculated directly on the segmented images using
305 Darcy's law: we use the fluid viscosity, the average pressure gradient across the
306 sample, and the average flow speed at the outlet computed using the Navier-
307 Stokes solver.

308 Figure 6 depicts the evolution of porosity and permeability for Ketton,
309 Estailades, and Portland carbonates at an injected brine pH of 3.1 and 3.6. The
310 change in porosity with time is relatively steady for Experiment 1 [Figure 6a].
311 However, there are dramatic decreases in the rate of change at later times for
312 experiments 2-6 with this being more marked for those rocks with the greatest
313 initial pore-space heterogeneity.

314 Figure 6d depicts the evolution of the specific surface area, S during
315 dissolution. S was calculated by summing the number of void voxel faces
316 adjacent to grains on the segmented images. S [m^{-1}] is the surface area [m^2]
317 divided by the bulk volume [m^3]. S initially increases sharply for all experiments,
318 and then remains relatively steady at later times for experiments 2-6, but
319 continues on an upward trend for experiment 1. We note that the initial specific
320 surface area is almost two times higher for Estailades, which may partially
321 explain the increased dissolution at the sample inlet as there are more reactive
322 surface sites to encourage dissolution along major flow paths.

323

324 3.2 Pore-scale Velocity Fields

325

326 The pore-scale velocity fields are visualised for Ketton, Estailades, and
327 Portland carbonates [Figure 7] by calculating the ratios of the magnitude of u at
328 the voxel centres divided by the average pore velocity, U_{av} [$\text{m}\cdot\text{s}^{-1}$], which is the
329 Darcy velocity q divided by the scan porosity ϕ_{CT} taken from the segmented
330 image. As the flow rate is held constant by the pumps throughout the
331 experiment, the Darcy velocity is also constant. In experiment 1, there are
332 several flow paths through the rock that widen through time. Conversely in

333 experiments 2, 3, 4 there are many flow paths that consolidate to a single
334 preferential pathway. In experiments 5 & 6, there are initially a small number of
335 pathways that widen at later times.

336 Figure 8 depicts the probability density functions (PDFs) of velocity. The
337 velocity histogram is computed on the pore-space images and sampled uniformly
338 in 256 bins of $\log(|\mathbf{u}|/U_{av})$, which are calculated as the ratios of the magnitude of
339 u at the voxel centres divided by the average pore velocity U_{av} . Ketton has the
340 narrowest distribution, but this still spans approximately 5 orders of magnitude.
341 With increasing heterogeneity, we see a wider distribution for Estailades, while
342 for Portland we see voxel velocities over 10 orders of magnitude. For experiment
343 1, in Ketton, there is very little change in the velocity distribution, which
344 indicates that the dissolution is relatively uniform. However, by the end of
345 experiment 2 there is evidence of a sharp peak at high velocity and a rounded
346 peak at lower velocities that suggests a preferential flow path and regions of
347 slower flow. In experiments 3-6 [Figure 8c-f], the PDFs initially show two
348 distinct regions of stagnant and high velocity flow. However, as dissolution
349 progresses the greater proportion of velocities gradually shift to the high
350 velocity region and the stagnant region diminishes. At lower pH this shift toward
351 higher velocity happens more quickly.

352

353 3.3 Dimensionless numbers and Reaction Rates

354

355 We define the Péclet number Pe , Damköhler number Da . Pe [-] is the pore-
356 scale ratio of diffusive to advective times:

357

$$358 \quad Pe = \frac{U_{av}L}{D} \quad (3)$$

359

360 where D is the molecular diffusion coefficient of Ca^{2+} at 25°C ($7.5 \times 10^{-10} \text{ m}^2.\text{s}^{-1}$),
361 U_{av} [$\text{m}.\text{s}^{-1}$] is the average pore velocity, and L [m] is the characteristic length. $L =$
362 π/S where S [m^{-1}] is the specific surface area (Mostaghimi et al., 2012).

363

364 We use the flow rate of $0.1 \text{ mL}.\text{min}^{-1}$ and the specific surface area
measured from the segmented images to calculate a characteristic length L of

365 between 251 μm and 458 μm and a Pe of between 394 and 822 for the six
 366 experiments [Table 3]. L and Pe are lowest for Estailades which has a high
 367 specific surface area and highest for Portland which has the lowest porosity (and
 368 thus the highest U_{av}). At these high Péclet numbers fresh reactant is readily
 369 available throughout the core since the spatially averaged advection is much
 370 faster than diffusion.

371 The Damköhler number [-] is defined as the ratio of advective to reactive
 372 times:

$$374 \quad Da = \frac{\pi r}{U_{av} n} \quad (4)$$

375
 376 where r is the reaction rate constant of pure calcite, n is calculated using $n = \rho_{\text{calcite}} [1 - \phi_{\text{total}}] / M_{\text{calcite}}$, ρ_{calcite} is the density of pure calcite ($2.71 \times 10^3 \text{ kg.m}^{-3}$), and
 377 M_{calcite} is the molecular mass of calcite (0.1 kg.mol^{-1}). ϕ_{total} , the total porosity, is
 378 measured using helium pycnometry for Ketton ($\phi_{\text{total}} = 0.2337$) and Estailades
 379 ($\phi_{\text{total}} = 0.295$) (Andrew et al., 2014) and measured using nitrogen pycnometry
 380 for Portland ($\phi_{\text{total}} = 0.213$). Here we describe both the porosity detectable at the
 381 resolution of the images, ϕ_{CT} , and micro-porosity. We assume that the intra-
 382 granular micro-porosity is unaffected by reaction, which is consistent with less
 383 than 0.5% observable difference in raw grey-scale within the grains. Thus we
 384 define $\phi_{\text{total}} = \phi_{\text{grain}} (1 - \phi_{\text{CT}}) + \phi_{\text{CT}}$ to calculate ϕ_{grain} at the beginning of the
 385 experiment: we assume that the value of ϕ_{grain} remains constant throughout the
 386 experiment and recalculate ϕ_{total} as ϕ_{CT} increases during dissolution.

387
 388 It has been proposed by Compton et al. (1989) and experimentally
 389 confirmed by Peng et al. (2015) that in an entirely surface controlled reaction
 390 regime of $\text{pH} < 4$, calcite dissolution kinetics can be modelled as a first order
 391 heterogeneous reaction with the equation:

$$393 \quad r_{\text{exp}} = k \alpha_{\text{H}^+} \quad (5)$$

394
 395 where r_{exp} is the batch reaction rate, k is the reaction rate constant and α_{H^+} is the
 396 activity of H^+ which is equal to $10^{-\text{pH}}$. In this model the system is assumed to be at

397 equilibrium at the surface and thus the reaction quotient is 1. Using the r_{exp} at
 398 our experimental conditions (10 MPa and 50°C) measured by Peng et al. (2015)
 399 at our experimental conditions and a pH 3.1 of $8.1 \times 10^{-4} \text{ mol.m}^{-2}.\text{s}^{-1}$ we can
 400 calculate r_{exp} at pH 3.6 to be $2.56 \times 10^{-4} \text{ mol.m}^{-2}.\text{s}^{-1}$: we use these values to
 401 estimate Da for the experiments in Table 3.

402 We can then use dynamic imaging to assess the accuracy of this predicted
 403 reaction rate by observing the evolution of the porosity throughout the
 404 experiment. As both Da and $PeDa \ll 1$, the reaction rate is slow compared to
 405 both the advection and diffusion rates. We then use the change in overall
 406 porosity to estimate the average effective reaction rate [$\text{mol.m}^{-2}.\text{s}^{-1}$] for each
 407 image time-step:

$$409 \quad r_{eff} = \frac{\rho_{grain}(1-\phi_{grain})}{M_{calcite}S} \frac{\Delta \phi_{CT}}{\Delta t} \quad (6)$$

410
 411 where $\Delta \phi_{CT}$ is the image-measured change in porosity within the time between
 412 the start of each scan, Δt . S [m^{-1}] is the specific surface area at the beginning of
 413 that time.

414 Figure 9 depicts the evolution of Pe , Da , $PeDa$, and r_{eff} for each experiment.
 415 Pe [Figure 9a] decreases with time for each experiment with the largest Pe
 416 decrease in the samples with the most dissolution and the fastest decline of Pe in
 417 the samples with the most heterogeneity. Conversely Da [Figure 9b] has a
 418 general increase with all experiments with the greatest increase in experiment 4
 419 and the smallest in experiments 3 & 5. Thus, $PeDa$ [Figure 9c] has a small dip at
 420 the beginning of reaction for experiments 2-6 but remains relatively constant at
 421 later times for all experiments.

422 The image-measured effective reaction rate declines with time for all
 423 experiments, but remains the highest for experiment 1. Experiments 2, 4, & 6
 424 conducted at pH 3.1 have slightly higher initial reaction rates than the
 425 experiments conducted at pH 3.6. With the exception of experiment 1 where the
 426 reaction rate remains high, the reaction rates decline sharply at later times.

427 Local Péclet statistics are calculated on each voxel from the velocity
 428 distributions of the direct simulations. Each void voxels velocity is normalized to

429 the experimental Darcy velocity and then used with equation 3 to calculate a
430 local Péclet number. Figure 10a shows the maximum Pe in the experiments: it is
431 evident that extremely high values can be reached in the centres of the fast-
432 flowing channels. Therefore, there are regions of the pore space where transport
433 is clearly advection-dominated. However, Figure 10b depicts the fraction of void
434 voxels that have a local Péclet number of <1 and have diffusion-dominated
435 transport. The fraction of voxels with diffusion-dominated transport is greatest
436 in the experiments with more heterogeneous pore structure: indeed, they
437 constitute the majority of the pore space in experiments 2-6. These slower-
438 flowing regions are situated at or close to the solid surface: where dissolution
439 occurs, the fluid is relatively stagnant and there are significant diffusive
440 limitations on transport and reaction.

441

442

443 3.5 Effluent Analysis

444

445 Figure 11 depicts the calcium concentration and pH calculated from
446 effluent samples. The concentration of Ca^{2+} [Figure 11a] is measured using an
447 ICP-MS from effluent samples taken during the experiment. We observe the
448 highest Ca^{2+} concentration initially for experiment 1, which then declines slightly
449 through time. In experiments 2-6, the Ca^{2+} concentration starts high (and is
450 higher with a lower injected pH), declines sharply, and then remains relatively
451 steady as the experiment continues.

452 As expected, the pH follows the same trends as the Ca^{2+} concentration
453 with the highest steady values for experiment 1 and lower values for
454 experiments 2-6. Notably, however, the pH is higher than that of the injected
455 brine. It is highest for Ketton, then Estailades and then Portland, indicating
456 different amounts of transport and reaction in the experiments: this explains
457 why injection pH is not a good indicator of the behaviour for different rocks. The
458 effluent pH is consistent with the ordering of effective reaction rates, Figure
459 9(d): where the reaction is faster, the pH in the effluent is higher.

460

461 4. DISCUSSION

462

463 4.1 Dynamic Porosity-Permeability Relationships

464

465 We fit the data in Figure 6a to an empirical power-law relationship
466 between permeability and porosity. The best fit to the Ketton experiments 1 and
467 2 were $K = 8.03 \times 10^{-8} \phi^{5.45}$ and $K = 4.13 \times 10^{-6} \phi^{6.93}$ respectively. For the
468 Estailades experiments 3 and 4 the best fits were $K = 2.36 \times 10^{-2} \phi^{13.4}$ and $K =$
469 $5.01 \times 10^{-5} \phi^{9.99}$ respectively. For the Portland experiments 5 and 6 we found K
470 $= 2.17 \times 10^2 \phi^{16.2}$ and $K = 6.32 \times 10^{-9} \phi^{23.8}$, respectively. The exponents are higher
471 than the exponent of 3 used in the Kozeny-Carmen relationship and the
472 exponents reported in Menke et al. (2015), but are similar to those reported in
473 Noiriél et al. (2004) for limestone dissolution and Smith et al. (2013) for
474 dissolution of low permeability carbonates. The porosity-permeability
475 relationship follows a trend of increasing exponents with increasing initial
476 heterogeneity with a marked sensitivity to porosity. The pH of the injected brine
477 is a poor indicator of the permeability change. However, the exponents rise with
478 decreasing initial permeability, which suggests a strong correlation between
479 pore-space heterogeneity and the total permeability change.

480

481 4.2 Dynamic Velocity Correlations

482

483 To test our hypothesis that initially present fast flow paths dissolve more
484 quickly than stagnant regions, we use the correlation coefficient presented in
485 Menke et al. (2016). This is defined as:

486

$$487 \quad C(t) = \frac{\sum_{i=1}^n v_i(0)v_i(t)}{n} - \frac{\sum_{i=1}^n v_i(0)}{n} \frac{\sum_{i=1}^n v_i(t)}{n} \quad (1)$$

488

489 with a normalized function:

490

$$491 \quad \bar{C}(t) = \frac{C(t)}{C(0)} \quad (2)$$

492

493 where $v_i(t)$ is the velocity in voxel i at time t . The computation is only made on
 494 voxels that are void in the reference (initial) image at time $t=0$. $C(0)$ is the
 495 variance of the initial velocity.

496 Figure 12 depicts the normalised correlation for all experiments: in all
 497 cases this increases over time, indicating that dissolution preferentially occurs in
 498 those regions of the pore space that initially had high flow speeds. If instead we
 499 had truly uniform dissolution, with reaction occurring at the same rate
 500 everywhere, we would expect the correlation function to decline over time.
 501 Experiment 1 has the slowest rise in the correlation function and thus is closest
 502 to uniform dissolution. Experiments 2-6 show a large rise in C , which provides a
 503 quantitative signature for channel formation in regions of the pores space with
 504 the highest velocities initially (Menke et al., 2016). Additionally, the correlations
 505 are in good agreement with the experiments performed at the same conditions
 506 in Menke et al. (2016). We also observe that injected pH is a poor indicator of the
 507 correlation in all cases except perhaps in experiments 1 and 2 where the pH
 508 changes the dissolution regime and thus the frequency of channelized flow paths.

509

510 4.3 Methods of Measuring Reaction Rates

511

512 We can compare the mass of dissolved solid predicted from the effluent
 513 sample to the mass of dissolved solid observed by μ -CT. The normalised
 514 difference, Figure 13, is found from:

515

$$516 \Delta CaCO_3(s)_{\#normalised} = \frac{CaCO_3_{effluent\#} - CaCO_3(s)_{CT\#}}{CaCO_3_{effluent\#}} \quad (7)$$

517

518 where the mass of solid calcium carbonate remaining in the sample is predicted
 519 by effluent analysis by subtracting the cumulative mass of calcium measured in
 520 the effluent samples from the mass of calcium carbonate in the initial CT image
 521 (time zero) at each time step:

522

$$523 CaCO_3_{effluent\#} = CaCO_3(s)_{CT_{time=0}} - \frac{M_{calcite}}{M_{Ca^{2+}}} \sum_{i=0}^{i=\#} ([Ca^{2+}]_{effluent} -$$

$$524 [Ca^{2+}]_{reactor}) F \Delta t_i \quad (8)$$

525

526 where t is time [s], F is the flow rate [$\text{m}^3 \cdot \text{s}^{-1}$], $M_{\text{Ca}^{2+}}$ is the molecular mass of Ca^{2+} ,
527 M_{calcite} is the molecular mass of calcite, $\#$ is the effluent sample number, and
528 $\text{CaCO}_{3(s)CT}$ is the total mass of solid calcium carbonate in the core calculated
529 from the initial CT scan. Here the concentrations are measured in $\text{moles} \cdot \text{m}^{-3}$.

530 The total mass of calcium carbonate is calculated at each time from the CT
531 images with:

532

$$533 \quad \text{CaCO}_{3(s)CT_t} = (1 - \phi_{total_t}) \rho_{\text{calcite}} \pi d^2 h \quad (9)$$

534

535 where d is the core radius, h is the total core length and ρ_{calcite} is the density of
536 calcite. The ϕ_{total} in the imaged 4 mm portion of the core is used as a proxy for
537 the entire 10 mm length. We believe this is reasonable as the porosity change is
538 relatively constant along the axis of flow in the imaged portion of the core
539 [Figure 5].

540 We observe an agreement between the CT observed dissolution and that
541 predicted by effluent analysis to within 6%, Figure 13. Most of the time the
542 difference is positive, indicating that effluent analysis may be missing some of
543 the dissolution. This could be due to small particles that are left behind when the
544 effluent sample is transferred from the sampling syringe into test tubes for ICP-
545 MS analysis or slightly differing amounts of dissolution in the unimaged portions
546 of the core.

547 Moreover, we can use the pH calculated from the effluent samples in the
548 experiment and equation 5 to compare the image-measured reaction rate to
549 assess whether batch- measured effective reaction rates can be used to predict
550 pore-scale effective reaction rates from the effluent-calculated pH [Table 4]. We
551 use the values of pH calculated from the effluent Ca^{2+} concentration and image-
552 measured r_{eff} values from ~ 180 minutes, which is when we observe that the rate
553 of reaction has reached a quasi-equilibrium state for all experiments. The surface
554 area used in our r_{eff} calculations is the image-measured surface area. Peng
555 (2015) found that the geometric surface area of a cleaved piece of calcite mineral
556 was accurate to within 0.2% for calculating batch reaction rates. We find that the

557 values of pH calculated from the effluent Ca^{2+} concentration are a reasonably
558 good indicator of image-observed reaction rate to within a factor of two for
559 experiments 1 and 3. However, as reactive flow complexity increases and the
560 dissolution regime changes from uniform to channelling the prediction becomes
561 less accurate. Experiments 2 and 4 are predicted to have an effective reaction
562 rate that is 6 times higher than those observed, and experiment 5 and 6 are
563 predicted to be 18 times higher. Had we used the injected pH the discrepancies
564 would have been even larger, with an over-estimate of likely rate by up to a
565 factor of 100.

566 We believe that the discrepancy between predicted and measured
567 reaction rates is due to flow limitations in removing reaction products from the
568 rock surface. Figure 14 depicts the interplay between flow and transport
569 processes during uniform dissolution and channel formation. During uniform
570 dissolution [Figure 14a] the flow velocity distribution is relatively narrow and
571 the majority of the pore space experiences relatively high velocities. Thus, the
572 reaction products are efficiently moved away from the surface of the grains and
573 are swept away down the advective flow streamlines such that fresh reactive
574 fluid is continually exposed to the entire rock surface. This causes the reaction to
575 be relatively uniform, the observed effective reaction rate to be high, and the
576 calculated effluent pH to be high, as the Ca^{2+} concentration in the effluent is
577 representative of the entire pore fluid. This also implies that for these samples
578 the resolution of μ -CT scanning adequately captures the surface area available
579 for reaction in the surface-reaction-controlled regime. However during
580 channelling [Figure 14b], the velocity distribution spans many orders of
581 magnitude and there are many regions of the pore space that do not experience
582 fast flow: see Figure 10b. The majority of the injected reactive fluid moves
583 through the centre of the fast flow paths without encountering or reacting with
584 the rock grains except on the channel walls. Thus, during uniform dissolution the
585 effluent pH calculated is indicative of the average pore-space reaction, while for
586 channelling the effluent pH calculated is only an indicator of the reaction within
587 the fast flow paths, which is considerably less due to the smaller surface area
588 available for reaction. This is confirmed by the recalculation of the image
589 observed reaction rate using only the surface area of the reactive channels [Table

590 4] which shows much better agreement between the predicted and observed
591 effective reaction rates.

592

593 5. CONCLUSIONS

594

595 We have found that the initial brine pH and initial pore structure have a
596 clear influence on the reaction-induced development of porosity and
597 permeability at the pore scale. We observe uniform dissolution in a homogenous
598 carbonate at an injection pH of 3.6 and channelling at pH = 3.1 and high
599 heterogeneity. Additionally, in heterogeneous samples we see faster channelling
600 with increasing heterogeneity and decreasing pH of the injected brine. We find
601 that effluent pH is a reasonably good indicator of effective reaction rate during
602 uniform dissolution. However, pH is a poor indicator of the effective reaction
603 rate during channelling, but can be used to help identify dissolution regime with
604 a lower effluent pH being more likely to be caused by channelized flow.
605 Moreover, we observe that the overall reaction rates are up to 18 times lower
606 than batch reaction rates at the effluent pH due to transport and flow limitations
607 at the fluid/solid interface: the majority of the pore space is in a diffusion-limited
608 regime even when the average flow rate indicates advection-dominated
609 conditions.

610 Bench-top x-ray microtomography coupled with a reservoir condition
611 flow apparatus and effluent sampling provides an accessible experimental tool
612 that is adaptable to many applications. Our method integrates experiment and
613 modelling to explore the dynamics of reservoir condition reactive flow at the
614 micron scale.

615

616 ACKNOWLEDGEMENTS

617

618 We gratefully acknowledge funding from the Qatar Carbonates and Carbon
619 Storage Research Centre (QCCSRC), provided jointly by Qatar Petroleum, Shell,
620 and Qatar Science & Technology Park. We also thank Ali Raeini and Joao Pereira
621 Nunes for help with the flow simulations.

622

623 REFERENCES

624

- 625 Andrew, M., Bijeljic, B., Blunt, M.J., 2014. Pore-scale imaging of trapped
626 supercritical carbon dioxide in sandstones and carbonates. *International*
627 *Journal of Greenhouse Gas Control*, 22: 1-14.
- 628 Andrew, M., Menke, H., Blunt, M.J., Bijeljic, B., 2015. The Imaging of Dynamic
629 Multiphase Fluid Flow Using Synchrotron-Based X-ray Microtomography
630 at Reservoir Conditions. *Transport in Porous Media*: 1-24.
- 631 Archie, G.E., 1952. Classification of carbonate reservoir rocks and petrophysical
632 considerations. *AAPG Bulletin*, 36(2): 278-298.
- 633 Berg, S. et al., 2013. Real-time 3D imaging of Haines jumps in porous media flow.
634 *Proceedings of the National Academy of Sciences of the United States of*
635 *America*, 110(10): 3755-3759.
- 636 Bijeljic, B., Mostaghimi, P., Blunt, M.J., 2013a. Insights into non - Fickian solute
637 transport in carbonates. *Water Resources Research*, 49(5): 2714-2728.
- 638 Bijeljic, B., Raeini, A., Mostaghimi, P., Blunt, M.J., 2013b. Predictions of non-
639 Fickian solute transport in different classes of porous media using direct
640 simulation on pore-scale images. *Physical Review E*, 87(1): 013011.
- 641 Blunt, M.J. et al., 2013. Pore-scale imaging and modelling. *Advances in Water*
642 *Resources*, 51: 197-216.
- 643 Buades, A., Coll, B., Morel, J.-M., 2005. A non-local algorithm for image denoising.
644 *Computer Vision and Pattern Recognition, 2005. CVPR 2005. IEEE*
645 *Computer Society Conference on*, 2: 60-65.
- 646 Buades, A., Coll, B., Morel, J.-M., 2008. Nonlocal image and movie denoising.
647 *International Journal of Computer Vision*, 76(2): 123-139.
- 648 Compton, R.G., Pritchard, K.L., Unwin, P.R., 1989. The dissolution of calcite in acid
649 waters: mass transport versus surface control. *Freshwater Biology*, 22(2):
650 285-288.
- 651 Daccord, G., Lenormand, R., Lietard, O., 1993a. Chemical Dissolution of a Porous-
652 Medium by a Reactive Fluid .1. Model for the Wormholing Phenomenon.
653 *Chemical Engineering Science*, 48(1): 169-178.
- 654 Daccord, G., Lietard, O., Lenormand, R., 1993b. Chemical Dissolution of a Porous-
655 medium by a Reactive Fluid .2. Convection vs Reaction, Behavior Diagram.
656 *Chemical Engineering Science*, 48(1): 179-186.
- 657 Dong, H., Blunt, M.J., 2009. Pore-network extraction from micro-computerized-
658 tomography images. *Physical review E*, 80(3): 036307.
- 659 El-Maghraby, R., Pentland, C., Iglauer, S., Blunt, M., 2012. A fast method to
660 equilibrate carbon dioxide with brine at high pressure and elevated
661 temperature including solubility measurements. *The Journal of*
662 *Supercritical Fluids*, 62: 55-59.
- 663 Ellis, B. et al., 2011. Deterioration of a fractured carbonate caprock exposed to
664 CO₂ - acidified brine flow. *Greenhouse Gases: Science and Technology*,
665 1(3): 248-260.
- 666 Fredd, C., Fogler, S., 1998. Influence of Transport and Reaction on Wormhole
667 Formations in Porous Media. *AIChE*, 44(9).
- 668 Gouze, P., Luquot, L., 2011. X-ray microtomography characterization of porosity,
669 permeability and reactive surface changes during dissolution. *Journal of*
670 *Contaminant Hydrology*, 120-21: 45-55.

671 Hao, Y., Smith, M., Sholokhova, Y., Carroll, S., 2013. CO₂-induced dissolution of
672 low permeability carbonates. Part II: Numerical modeling of experiments.
673 *Advances in Water Resources*, 62: 388-408.

674 Herzog, H., Caldeira, K., Reilly, J., 2003. An issue of permanence: Assessing the
675 effectiveness of temporary carbon storage. *Climatic Change*, 59(3): 293-
676 310.

677 Johnson, J.W., Nitao, J.J., Steefel, C.I., Knuass, K.G., 2002. Reactive transport
678 modeling of geologic CO₂ sequestration in saline aquifers: the influence of
679 intra-aquifer shales and the relative effectiveness of structural, solubility,
680 and mineral trapping during prograde and retrograde sequestration,
681 Lawrence Livermore National Laboratory, Livermore, CA.

682 Lamy - Chappuis, B., Angus, D., Fisher, Q., Grattoni, C., Yardley, B.W., 2014. Rapid
683 porosity and permeability changes of calcareous sandstone due to CO₂ -
684 enriched brine injection. *Geophysical Research Letters*, 41(2): 399-406.

685 Lanczos, C., 1950. An iteration method for the solution of the eigenvalue problem
686 of linear differential and integral operators. United States Governm. Press
687 Office.

688 Langmuir, D., Hall, P., Drever, J., 1997. *Aqueous Environmental Geochemistry*.

689 Leal, A.M., Blunt, M.J., LaForce, T.C., 2013. A robust and efficient numerical
690 method for multiphase equilibrium calculations: Application to CO₂-
691 brine-rock systems at high temperatures, pressures and salinities.
692 *Advances in Water Resources*, 62: 409-430.

693 Leal, A.M., Blunt, M.J., LaForce, T.C., 2014. Efficient chemical equilibrium
694 calculations for geochemical speciation and reactive transport modelling.
695 *Geochimica et Cosmochimica Acta*, 131: 301-322.

696 Li, Z.W., Dong, M.Z., Li, S.L., Huang, S., 2006. CO₂ sequestration in depleted oil and
697 gas reservoirs - caprock characterization and storage capacity. *Energy
698 Conversion and Management*, 47(11-12): 1372-1382.

699 Luquot, L., Gouze, P., 2009. Experimental determination of porosity and
700 permeability changes induced by injection of CO₂ into carbonate rocks.
701 *Chemical Geology*, 265(1-2): 148-159.

702 Luquot, L., Rodriguez, O., Gouze, P., 2014. Experimental Characterization of
703 Porosity Structure and Transport Property Changes in Limestone
704 Undergoing Different Dissolution Regimes. *Transport in Porous Media*,
705 101(3): 507-532.

706 Maheshwari, P., Ratnakar, R., Kalia, N., Balakotaiah, V., 2013. 3-D simulation and
707 analysis of reactive dissolution and wormhole formation in carbonate
708 rocks. *Chemical Engineering Science*, 90: 258-274.

709 Menke, H.P., Bijeljic, B., Andrew, M.G., Blunt, M.J., 2015. Dynamic Three-
710 Dimensional Pore-Scale Imaging of Reaction in a Carbonate at Reservoir
711 Conditions. *Environmental science & technology*, 49(7): 4407-4414.

712 Menke, H.P., Bijeljic, B., Blunt, M.J., 2016. Reservoir Condition Imaging of Reactive
713 Transport in Heterogeneous Carbonates Using Fast Synchrotron
714 Tomography – Effect of Initial Pore structure and Flow Conditions.
715 *Chemical Geology*, 428(1): 15-26.

716 Metz, B., Davidson, O., De Coninck, H., Loos, M., Meyer, L., 2005. IPCC, 2005: IPCC
717 special report on carbon dioxide capture and storage. Prepared by
718 Working Group III of the Intergovernmental Panel on Climate Change.
719 Cambridge, United Kingdom and New York, NY, USA, 442 pp.

720 Molins, S., Trebotich, D., Steefel, C.I., Shen, C., 2012. An investigation of the effect
721 of pore scale flow on average geochemical reaction rates using direct
722 numerical simulation. *Water Resources Research*, 48(3).

723 Morse, J.W., Mackenzie, F.T., 1990. *Geochemistry of sedimentary carbonates*.

724 Mostaghimi, P., Bijeljic, B., Blunt, M.J., 2012. Simulation of flow and dispersion on
725 pore-space images. *SPE J*, 17(4): 1131-1141.

726 Noiriel, C., Gouze, P., Bernard, D., 2004. Investigation of porosity and
727 permeability effects from microstructure changes during limestone
728 dissolution. *Geophysical research letters*, 31(24).

729 Noiriel, C., Gouze, P., Made, B., 2013. 3D analysis of geometry and flow changes in
730 a limestone fracture during dissolution. *Journal of Hydrology*, 486: 211-
731 223.

732 Ott, H. et al., 2012. Core-flood experiment for transport of reactive fluids in rocks.
733 *Review of Scientific Instruments*, 83(8): 084501.

734 Ott, H., Oedai, S., 2015. Wormhole formation and compact dissolution in single -
735 and two - phase CO₂ - brine injections. *Geophysical Research Letters*,
736 42(7): 2270-2276.

737 Peng, C., 2015. *Chemical Interactions between Carbon Dioxide Acidified Aqueous*
738 *Fluids and Carbonate Minerals*, Imperial College London, 275 pp.

739 Peng, C., Crawshaw, J.P., Maitland, G.C., Trusler, J.M., 2015. Kinetics of calcite
740 dissolution in CO₂-saturated water at temperatures between (323 and
741 373) K and pressures up to 13.8 MPa. *Chemical Geology*, 403: 74-85.

742 Peng, C., Crawshaw, J.P., Maitland, G.C., Trusler, J.M., Vega-Maza, D., 2013. The pH
743 of CO₂-saturated water at temperatures between 308K and 423K at
744 pressures up to 15MPa. *The Journal of Supercritical Fluids*, 82: 129-137.

745 Raeini, A.Q., Blunt, M.J., Bijeljic, B., 2012. Modelling two-phase flow in porous
746 media at the pore scale using the volume-of-fluid method. *Journal of*
747 *Computational Physics*, 231(17): 5653-5668.

748 Rötting, T.S., Luquot, L., Carrera, J., Casalinuovo, D.J., 2015. Changes in porosity,
749 permeability, water retention curve and reactive surface area during
750 carbonate rock dissolution. *Chemical Geology*, 403: 86-98.

751 Schlüter, S., Sheppard, A., Brown, K., Wildenschild, D., 2014. Image processing of
752 multiphase images obtained via X - ray microtomography: A review.
753 *Water Resources Research*, 50(4): 3615-3639.

754 Smith, M.M., Sholokhova, Y., Hao, Y., Carroll, S.A., 2012. Evaporite caprock
755 integrity: An experimental study of reactive mineralogy and pore-scale
756 heterogeneity during brine-CO₂ exposure. *Environmental science &*
757 *technology*, 47(1): 262-268.

758 Smith, M.M., Sholokhova, Y., Hao, Y., Carroll, S.A., 2013. CO₂-induced dissolution
759 of low permeability carbonates. Part I: Characterization and experiments.
760 *Advances in Water Resources*, 62: 370-387.

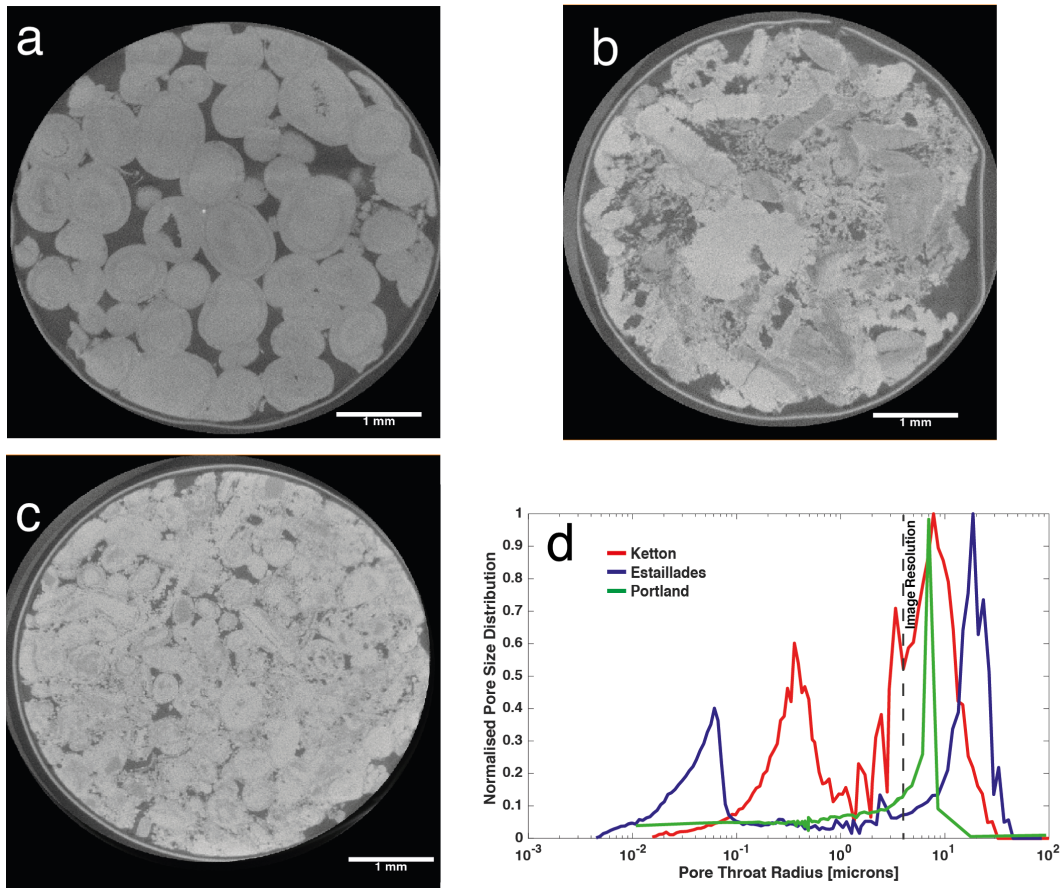
761 Steefel, C.I., Molins, S., Trebotich, D., 2013. Pore scale processes associated with
762 subsurface CO₂ injection and sequestration. *Reviews in Mineralogy and*
763 *Geochemistry*, 77(1): 259-303.

764 Tutolo, B.M., Luhmann, A.J., Kong, X.-Z., Saar, M.O., Seyfried Jr, W.E., 2014.
765 Experimental observation of permeability changes in dolomite at CO₂
766 sequestration conditions. *Environmental science & technology*, 48(4):
767 2445-2452.

768

769

770



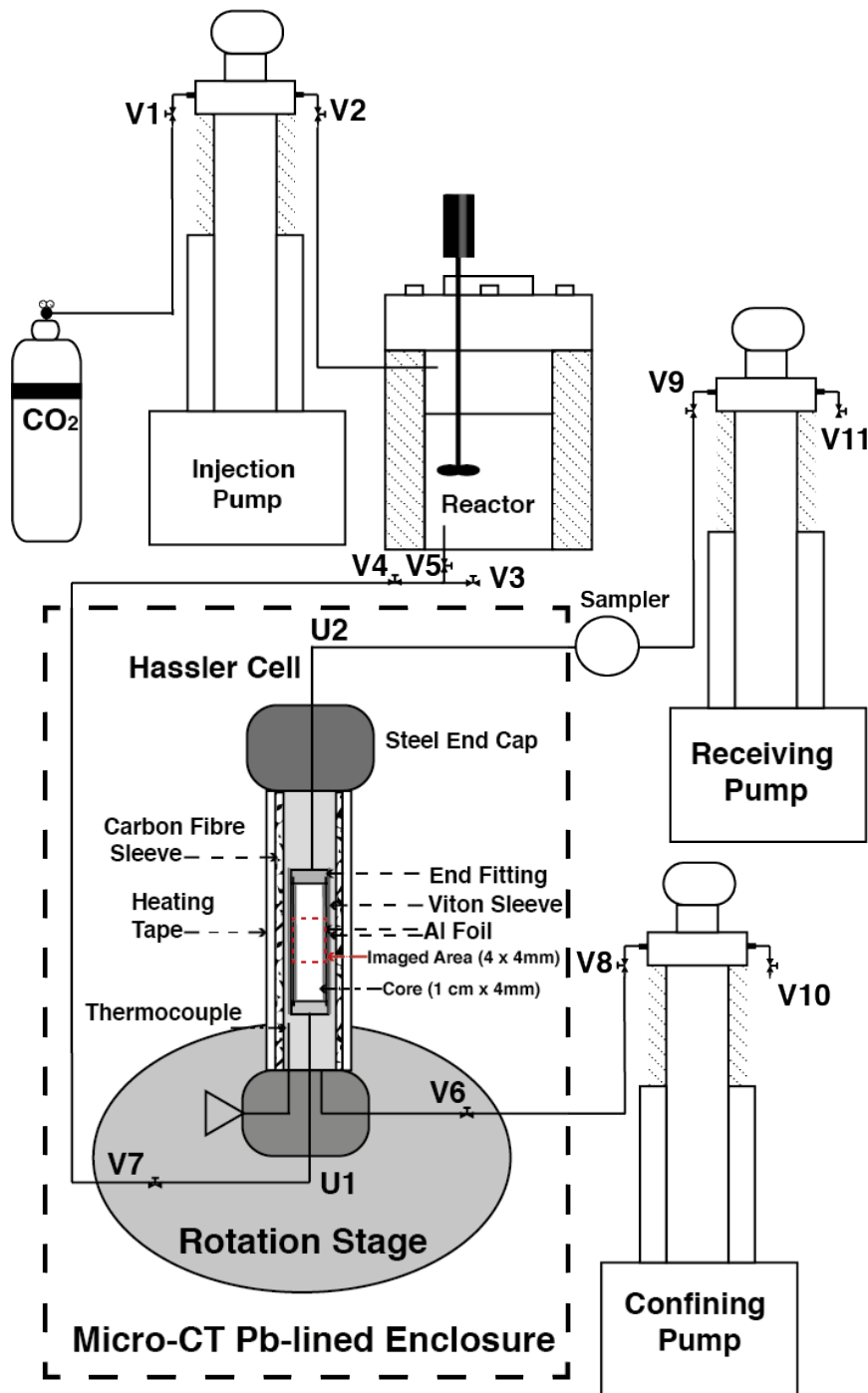
771

772 **Figure 1** 2D μ -CT slices of Ketton (a), Estailades (b), and Portland Basebed (c)

773 with their normalised pore throat radii distributions (d). The μ -CT image

774 resolution threshold is the black dotted line.

775



776

777 **Figure 2** The *in situ* experimental apparatus. CO₂ is pressurized by the injection
 778 pump and used to equilibrate brine in the reactor. Reactive brine is pulled
 779 through core assembly by the receiving pump. The cell is confined by deionized
 780 water in the confining pump and heated using heating tape controlled by a
 781 thermocouple in the confining fluid. The experimental system is connected
 782 together using tubing and fluid flow is directed using Valves (V) and Unions (U).
 783 Effluent from the flow cell is sampled using Valve 12. Modified from Menke et al.
 784 (2015).

785 **Table 1** Summary of physical rock properties. Chemical composition was found
 786 using X-ray diffraction.

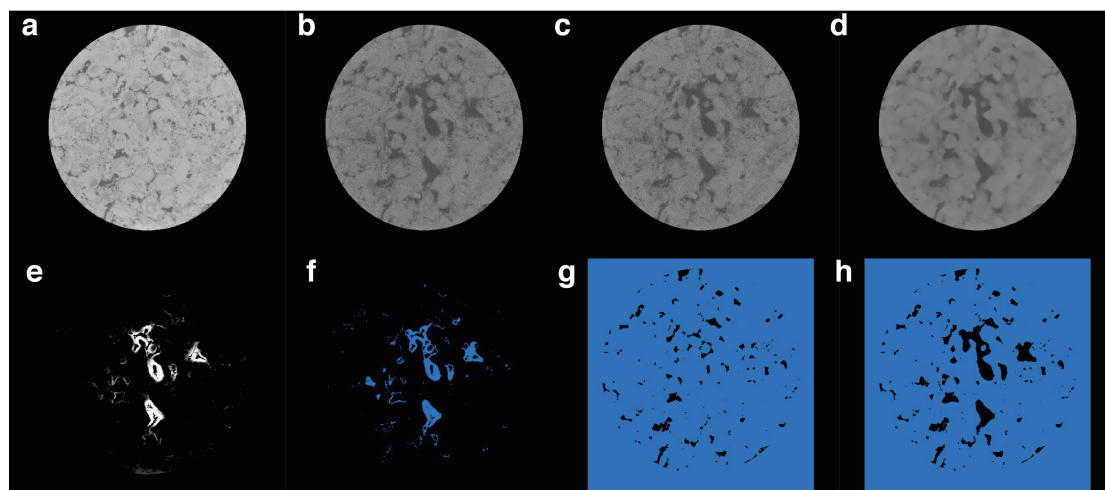
	Ketton	Estailades	Portland Basedbed
Rock type	Limestone	Limestone	Limestone
Origin	Ketton, Rutland, UK	Oppede, France	Portland, UK
Age	169-176 million Years	22 million years	150-190 million years
Bulk-measured total porosity ϕ_T []	0.2337	0.295	0.213
Bulk-measured absolute permeability [m²]	2.81×10^{-12}	1.490×10^{-12}	5.24×10^{-15}
% Calcite	99.1	97.2	97.9
% Quartz	0.9	2.8	2.1
% Clay	0	0	0
% Feldspar	0	0	0

787
788

789 **Table 2** Experimental Parameters

Experiment Number	1	2	3	4	5	6
Rock Name	Ketton Limestone	Ketton Limestone	Estailades Limestone	Estailades Limestone	Portland Basebed Limestone	Portland Basebed Limestone
Flow rate [mL/min]	0.1	0.1	0.1	0.1	0.1	0.1
Injected Fluid pH	3.6	3.1	3.6	3.1	3.6	3.1
Temperature [°C]	50	50	50	50	50	50
Pressure [MPa]	10	10	10	10	10	10
Number of Scans	6	10	6	6	6	6
Voxel Size [μ m]	4.15	4.07	4.15	4.07	4.15	4.59
Projections	800	400	800	800	800	800
Scan Time [min]	40	20	40	40	40	40
Exposure Time [sec]	1.4	1.4	1.3	1.3	1.4	1.4

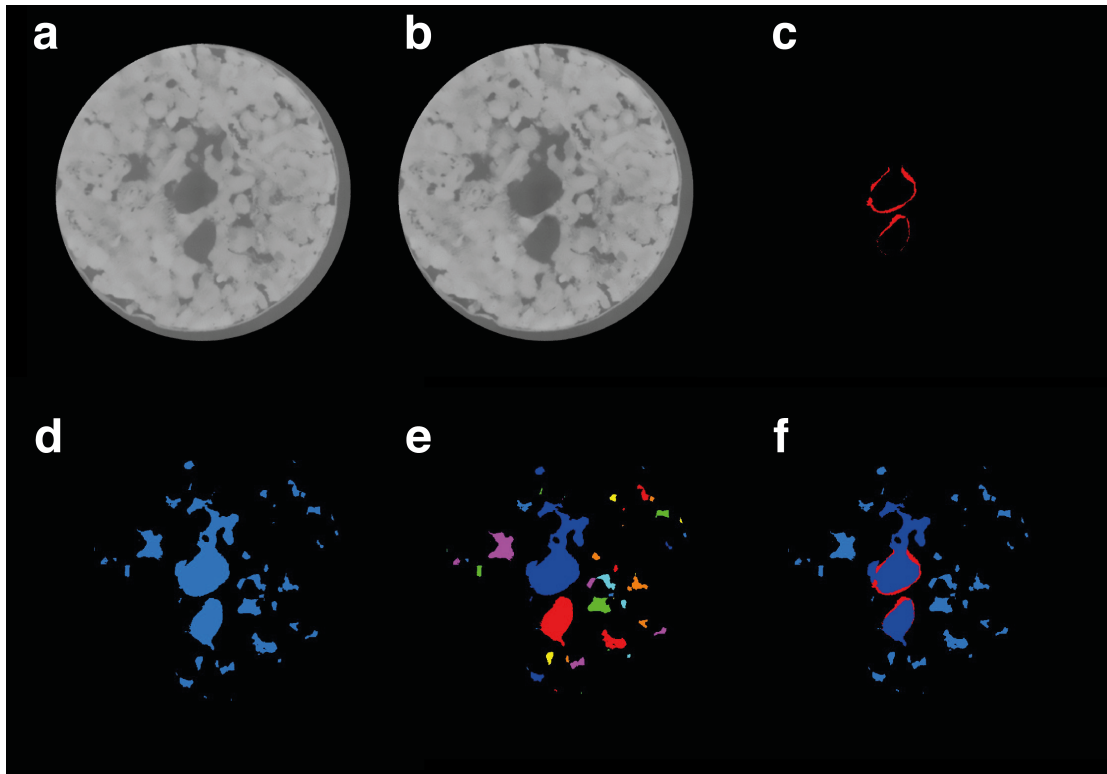
790



791

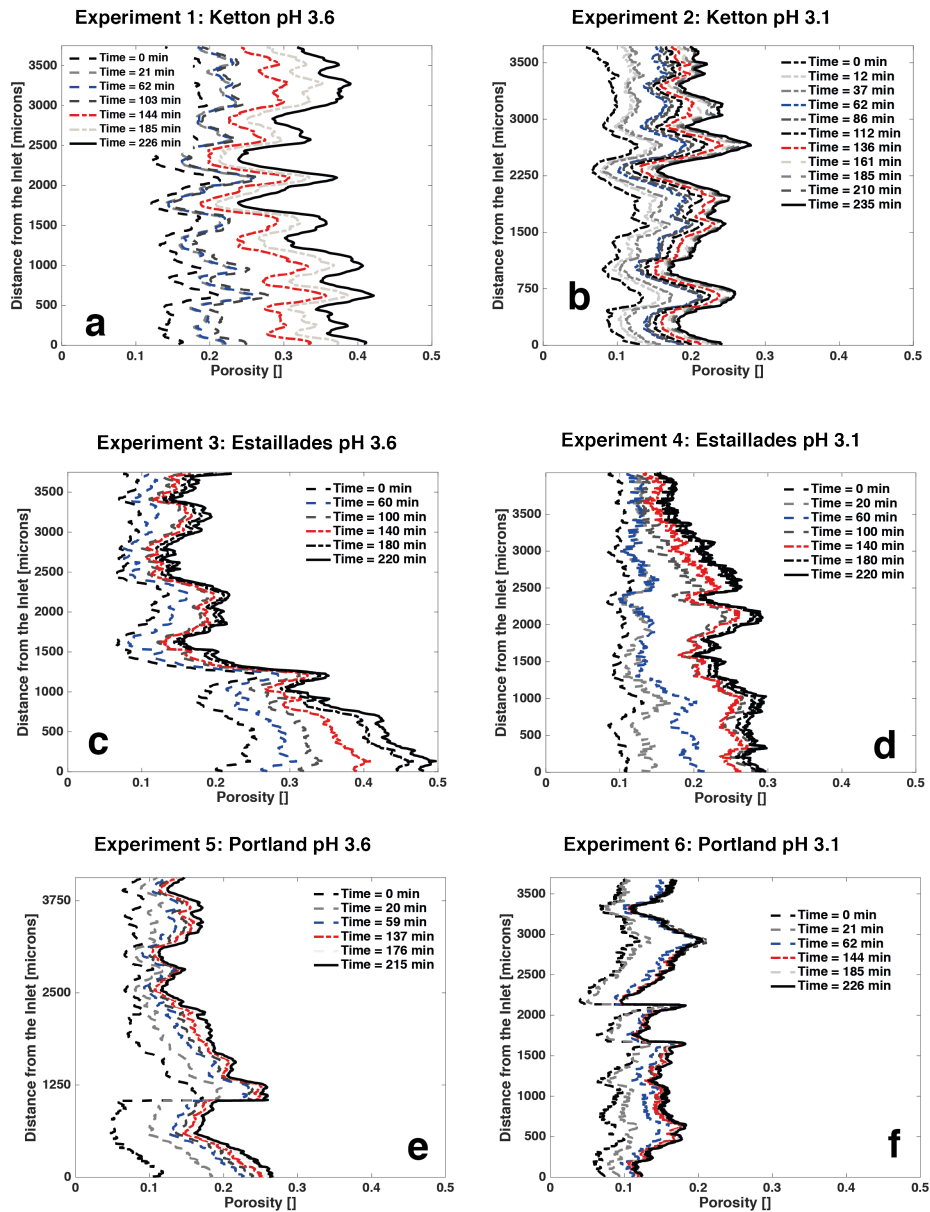
792 **Figure 3** The image analysis workflow. Raw dry scans (a) and wet scans (b) are
 793 registered and resampled (c) and then filtered (d). The wet scan is subtracted
 794 from the dry scan to create a difference image (e). The difference image is the

795 segmented (f) along with the dry scan (g). (f) is then subtracted from (g) to
796 obtain the segmented wet scan (h).
797



798
799 **Figure 4** The channel analysis workflow. Raw scans pre (a) and post (b)
800 reaction, the segmented difference (c), the entire segmented pore space (d), the
801 segmented pore space labelled by channel (e), and the pore space separated into
802 channels with reaction (dark blue) and channels without reaction (light blue)
803 with the reacted voxels overlain red (f).

804
805



806

807

Figure 5 Sliced-averaged porosity versus distance from the sample inlet for

808

Ketton at pH 3.6 (a) and pH 3.1 (b), Estailades at pH 3.6 (c) and pH 3.1 (d) and

809

Portland at pH 3.6 (e) and pH 3.1 (f). The blue and blue lines indicate ~1 and ~2

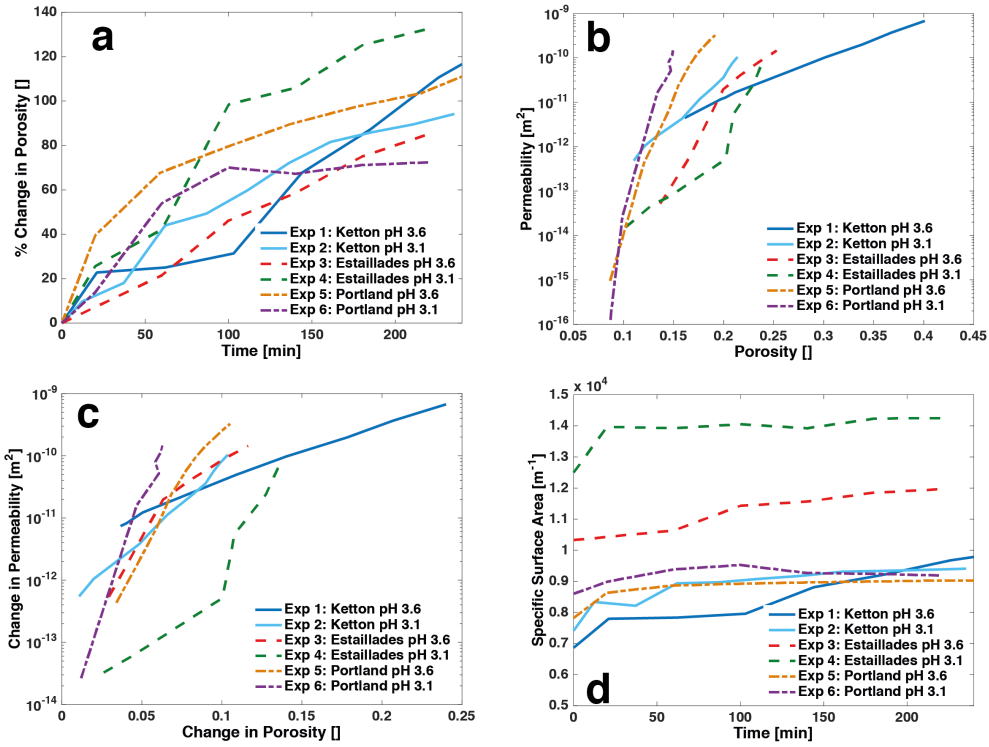
810

hours into the experiment. The inlet is defined as the beginning of the imaged

811

portion of the sample closest to the inlet face of the core.

812



813

814 **Figure 6** The percent change in average porosity across the whole sample versus

815 time (a), permeability versus porosity (b), the change in permeability as a

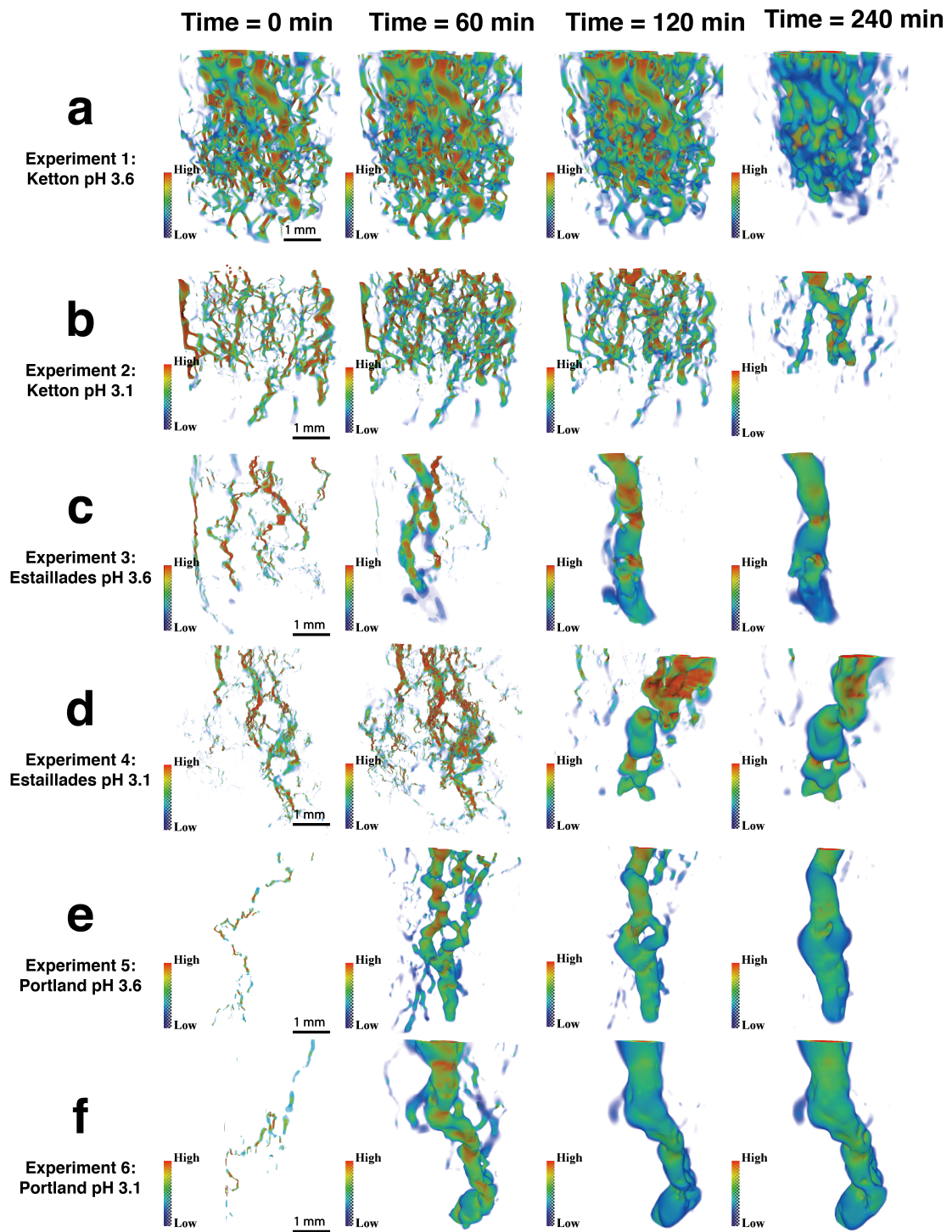
816 function of the change in porosity (c), and the specific surface area versus time

817 (d) of Ketton, Estailades, and Portland carbonates during dissolution. The

818 porosity and surface area are calculated directly from the segmented images.

819 Permeability is found from applying Darcy's Law from the computed flow field.

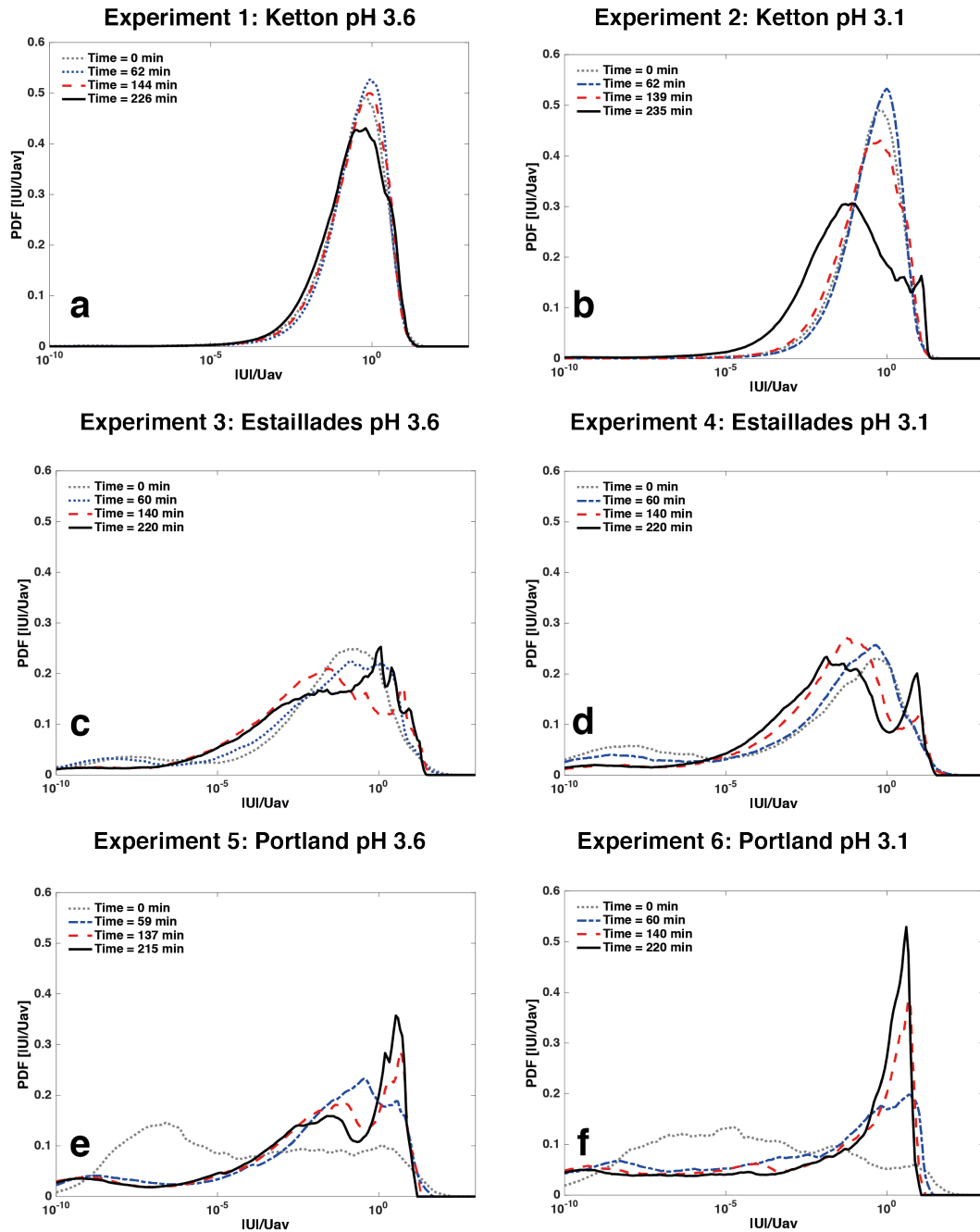
820



821

822 **Figure 7** The ratios of the magnitude of u at the voxel centres divided by the
 823 average pore velocity U_{av} are coloured using a linear scale where low velocity is
 824 blue, medium velocity is yellow, and high velocity is red.

825



826

827 **Figure 8** Probability density functions (PDFs) of pore voxel velocity during (a)
 828 experiment 1, (b) experiment 2, (c) experiment 3, (d) experiment 4, (e)
 829 experiment 5, (f) experiment 6. The velocity histogram is computed on the pore-
 830 space images and sampled uniformly in 256 bins of $\log(|\mathbf{u}|/U_{av})$, which are
 831 calculated as the ratios of the magnitude of \mathbf{u} at the voxel centres divided by the
 832 average pore velocity U_{av} . The distributions are shown at times 0 (dotted line),
 833 ~ 60 (blue), ~ 120 (red), ~ 240 (solid black line) minutes.

834

835 **Table 3** Computed Flow and Reaction Parameters for times 0, ~60, ~120 and
 836 ~240 min.

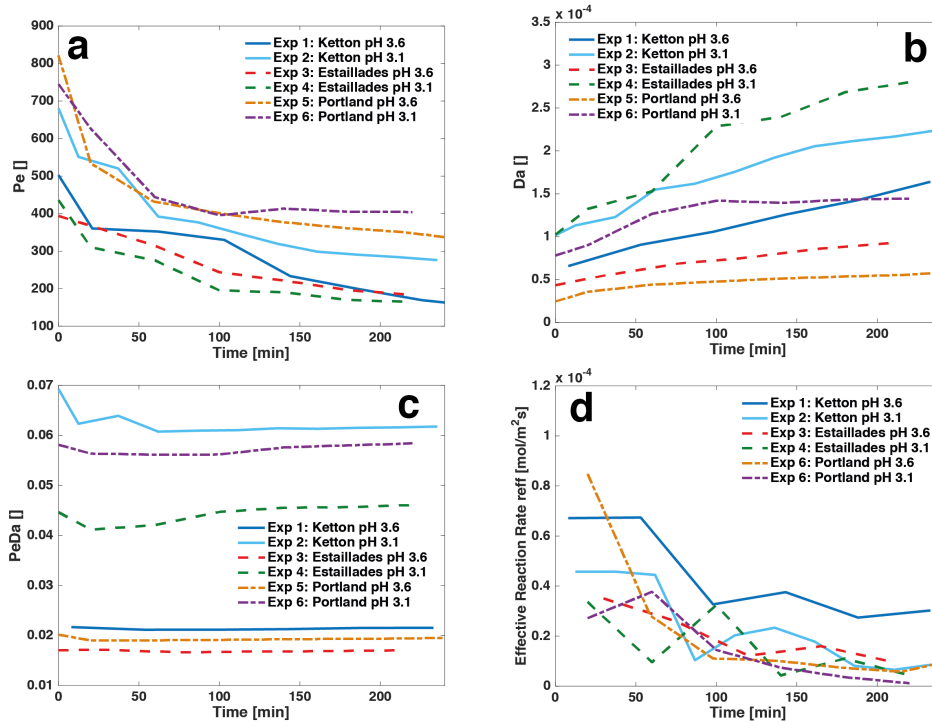
Time [min]	Property	Experiment 1: Ketton pH 3.6	Experiment 2: Ketton pH 3.1	Experiment 3: Estailades pH 3.6	Experiment 4: Estailades pH 3.1	Experiment 5: Portland pH 3.6	Experiment 6: Portland pH 3.1
0							
	absolute permeability (direct simulation) [m ²]	4.42×10^{-12}	4.81×10^{-13}	5.17×10^{-14}	1.53×10^{-14}	9.52×10^{-16}	1.25×10^{-16}
	image measured porosity ϕ_{CT} []	0.161	0.110	0.137	0.102	0.086	0.087
	image measured specific surface area S [m ⁻¹]	6.86×10^3	7.41×10^3	1.03×10^4	1.25×10^4	7.82×10^3	8.60×10^3
	calculated L [m]	4.58×10^{-4}	4.24×10^{-4}	3.04×10^{-4}	2.51×10^{-4}	4.01×10^{-4}	3.65×10^{-4}
	calculated Pe [] [eq 3]	5.03×10^2	6.81×10^2	3.94×10^2	4.36×10^2	8.22×10^2	7.45×10^2
	calculated Da [] using batch reaction rate [eq 4]	4.70×10^{-5}	1.02×10^{-4}	4.33×10^{-5}	1.02×10^{-4}	2.46×10^{-5}	7.80×10^{-5}
	calculated $PeDa$ []	2.36×10^{-2}	6.93×10^{-2}	1.71×10^{-2}	4.46×10^{-2}	2.02×10^{-2}	5.81×10^{-2}
	calculated reaction rate r_{eff} [mol.m ⁻² .s ⁻¹] [eq 6]	6.71×10^{-5}	4.57×10^{-5}	3.5×10^{-5}	1.85×10^{-5}	8.48×10^{-5}	2.71×10^{-5}
60							
	absolute permeability (direct simulation) [m ²]	1.26×10^{-11}	4.28×10^{-12}	5.88×10^{-13}	7.33×10^{-14}	6.31×10^{-12}	1.59×10^{-11}
	image measured porosity ϕ_{CT}	0.201	0.159	0.166	0.145	0.145	0.133

	\square						
	image measured specific surface area S [m^{-1}]	7.83×10^3	8.94×10^3	1.06×10^4	1.39×10^4	8.86×10^3	9.38×10^3
	calculated L [m]	4.01×10^{-4}	3.52×10^{-4}	2.95×10^{-4}	2.26×10^{-4}	3.54×10^{-4}	3.35×10^{-4}
	calculated Pe [] [eq 3]	3.52×10^2	3.92×10^2	3.15×10^2	2.75×10^2	4.33×10^2	4.44×10^2
	calculated Da [] using batch reaction rate [eq 4]	6.17×10^{-5}	1.55×10^{-4}	5.45×10^{-5}	1.53×10^{-4}	4.40×10^{-5}	1.27×10^{-4}
	calculated $PeDa$ []	2.18×10^{-2}	6.07×10^{-2}	1.72×10^{-2}	4.21×10^{-2}	1.93×10^{-2}	5.62×10^{-2}
	calculated reaction rate r_{eff} [$\text{mol} \cdot \text{m}^{-2} \cdot \text{s}^{-1}$] [eq 6]	6.74×10^{-5}	1.71×10^{-5}	2.59×10^{-5}	1.07×10^{-5}	2.89×10^{-5}	1.44×10^{-5}
120							
	absolute permeability (direct simulation) [m^2]	5.39×10^{-11}	2.17×10^{-11}	4.03×10^{-11}	5.50×10^{-12}	5.66×10^{-11}	7.98×10^{-11}
	image measured porosity ϕ_{CT} []	0.270	0.190	0.217	0.210	0.164	0.145
	image measured specific surface area S [m^{-1}]	8.81×10^3	9.17×10^3	1.16×10^4	1.39×10^4	8.96×10^3	9.27×10^3
	calculated L [m]	3.57×10^{-4}	3.42×10^{-4}	2.72×10^{-4}	2.62×10^{-4}	3.50×10^{-4}	3.39×10^{-4}
	calculated Pe [] [eq 3]	2.34×10^2	3.20×10^2	2.22×10^2	1.90×10^2	3.78×10^2	4.13×10^2
	calculated Da [] using batch reaction rate [eq 4]	9.06×10^{-5}	1.92×10^{-4}	7.57×10^{-5}	2.39×10^{-4}	5.09×10^{-5}	1.39×10^{-4}
	calculated	2.12×10^{-2}	6.14×10^{-2}	1.68×10^{-2}	4.55×10^{-2}	1.93×10^{-2}	5.76×10^{-2}

	$PeDa$ []						
	calculated reaction rate r_{eff} [mol.m ⁻² .s ⁻¹] [eq 6]	3.27×10^{-5}	1.78×10^{-5}	7.01×10^{-6}	4.26×10^{-6}	5.77×10^{-6}	3.57×10^{-6}
240							
	absolute permeability (direct simulation) [m ²]	2.01×10^{-10}	1.04×10^{-10}	1.45×10^{-10}	6.44×10^{-11}	2.52×10^{-10}	1.47×10^{-10}
	image measured porosity ϕ_{CT} []	0.339	0.214	0.253	0.237	0.186	0.150
	image measured specific surface area S [m ⁻¹]	9.67×10^3	9.41×10^3	1.20×10^4	1.43×10^4	9.19×10^3	1.43×10^4
	calculated L [m]	3.25×10^{-4}	3.34×10^{-4}	2.63×10^{-4}	2.20×10^{-4}	3.38×10^{-4}	3.42×10^{-4}
	calculated Pe [] [eq 3]	1.69×10^2	2.76×10^2	1.84×10^2	1.64×10^2	3.30×10^2	4.04×10^2
	calculated Da [] using batch reaction rate [eq 4]	1.26×10^{-4}	2.24×10^{-4}	9.28×10^{-5}	2.80×10^{-4}	5.95×10^{-5}	1.44×10^{-4}
	calculated $PeDa$ []	2.13×10^{-2}	6.18×10^{-2}	1.70×10^{-2}	4.61×10^{-2}	1.96×10^{-2}	5.84×10^{-2}
	calculated reaction rate r_{eff} [mol.m ⁻² .s ⁻¹] [eq 6]	3.76×10^{-5}	8.71×10^{-6}	9.61×10^{-6}	4.26×10^{-6}	5.74×10^{-6}	1.16×10^{-6}

837

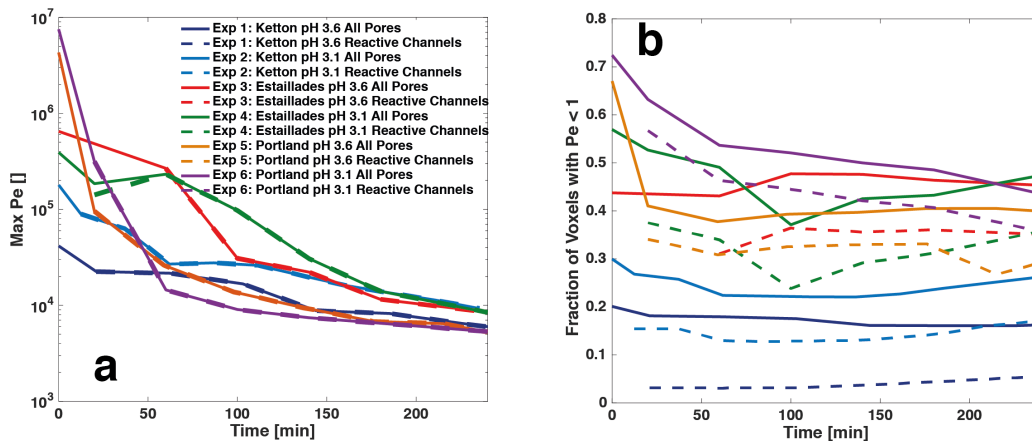
838



839

840 **Figure 9** (a) Péclet number (Pe), (b) Damköhler number (Da), (c) Péclet-
 841 Damköhler number ($PeDa$), and (d) effective reaction rates (r_{eff}) for each
 842 experiment through time.

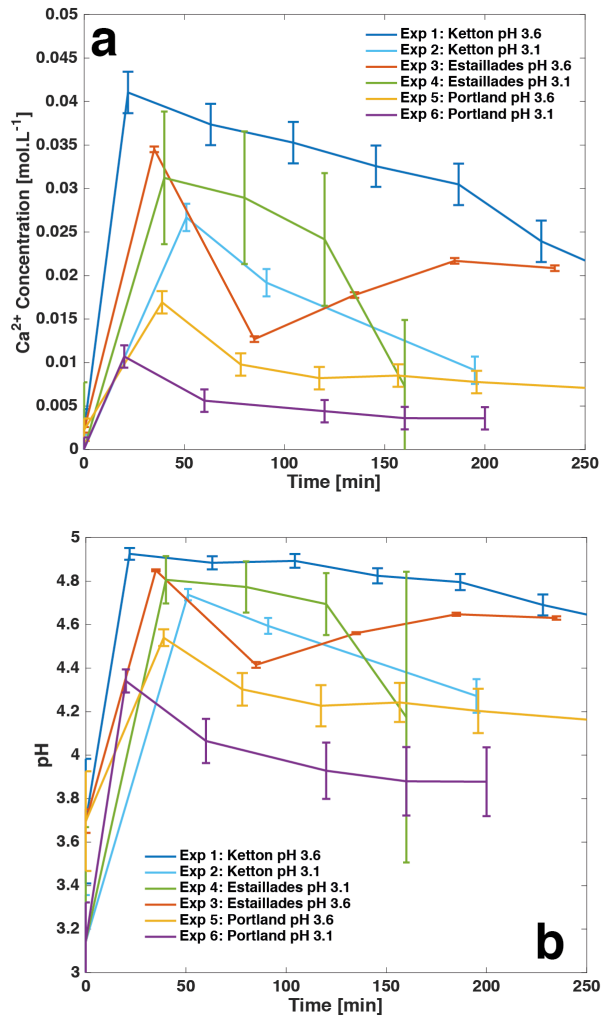
843



844

845 **Figure 10** (a) Maximum Péclet number (Pe) and (b) fraction of void voxels with
 846 Péclet number (Pe) ≤ 1 , measured by direct simulation for each experiment
 847 through time both in the entire connected pore space (solid lines) and only in the
 848 channel which have reacted during the previous tomographic scan period
 849 (dashed lines).

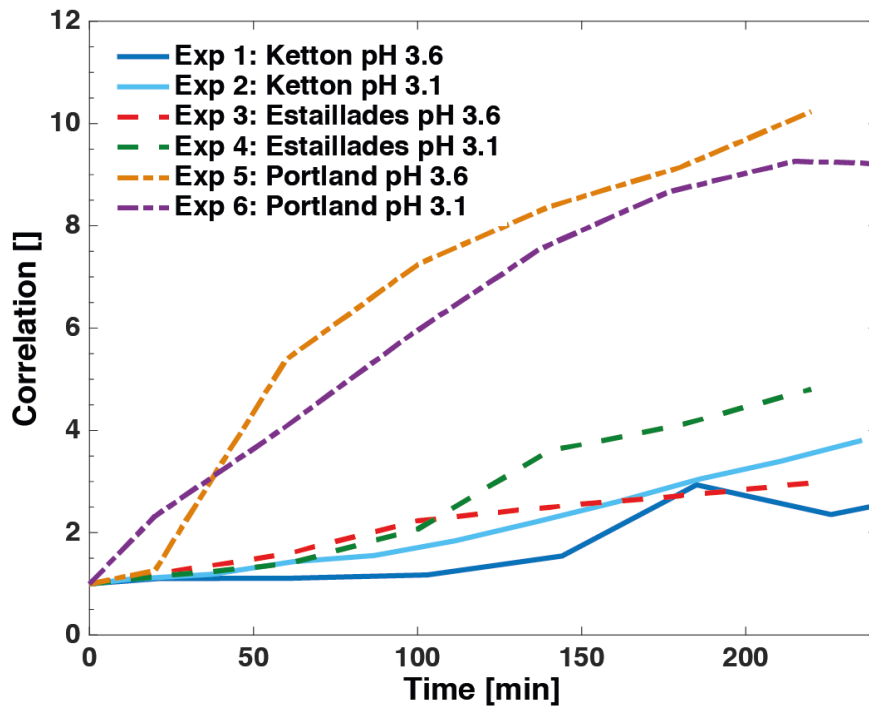
850



851

852 **Figure 11** (a) Ca²⁺ concentration [mol.L⁻¹] measured in effluent and (b) pH as
 853 calculated from Ca²⁺ concentration by the method presented in Leal et al. (2014)
 854 versus time [min]. The error bars represent the standard error where $\sigma_x = \frac{\sigma}{\sqrt{n}}$,
 855 of the concentration and pH values. Each sample was taken over a period of 20
 856 minutes. The points on the graph represent the beginning time of each 20-
 857 minute sampling period and the average values of concentration and pH during
 858 that period.

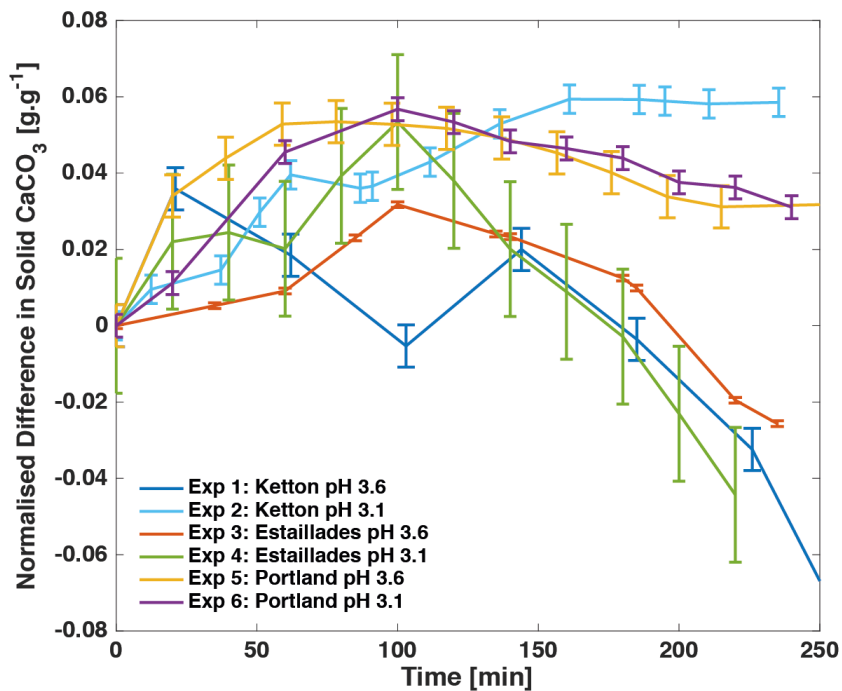
859



860

861 **Figure 12** Normalized correlation, Eqs. (1) and (2), between the initial voxel
 862 velocities (time=0) and each those from successive image through time for each
 863 experiment.

864



866

867 **Figure 13** A comparison of effluent measured and CT observed fluid-solid
 868 reaction plotted as the difference in dissolved solid CaCO_3 versus time. The error
 869 bars represent the standard uncertainty of the effluent calcium content.

870

871 **Table 4** The effective reaction rates calculated using the observed surface area
 872 of the pore space, the observed surface area of the reactive channels and the pH
 873 predicted from the effluent at time ~ 180 minutes. The error represents the
 874 standard error of the effluent calcium content.

Experiment #	Image observed	Image observed	Effluent Calculated pH	Effluent Calculated $r_{exp} [\text{mol.m}^{-2}\text{s}^{-1}]$ [eq 6]
	All Pores $r_{eff} [\text{mol.m}^{-2}\text{s}^{-1}]$ [eq 5]	Reactive Channels $r_{eff} [\text{mol.m}^{-2}\text{s}^{-1}]$ [eq 5]		
Experiment 1: Ketton pH 3.6	3.27×10^{-5}	5.52×10^{-5}	4.80 ± 0.04	$1.6 \times 10^{-5} \pm 1.4 \times 10^{-6}$
Experiment 2: Ketton pH 3.1	8.09×10^{-6}	1.31×10^{-5}	4.27 ± 0.08	$5.5 \times 10^{-5} \pm 9.2 \times 10^{-6}$
Experiment 3: Estailades pH 3.6	1.59×10^{-5}	3.10×10^{-5}	4.65 ± 0.01	$2.3 \times 10^{-5} \pm 5.3 \times 10^{-7}$
Experiment 4: Estailades pH 3.1	1.10×10^{-5}	1.41×10^{-5}	4.17 ± 0.67	$6.8 \times 10^{-5} \pm 5.4 \times 10^{-5}$
Experiment 5: Portland pH 3.6	3.57×10^{-6}	3.33×10^{-5}	4.20 ± 0.10	$6.4 \times 10^{-5} \pm 1.3 \times 10^{-5}$

Portland pH 3.6

Experiment 6:

Portland pH 3.1

7.46×10^{-6}

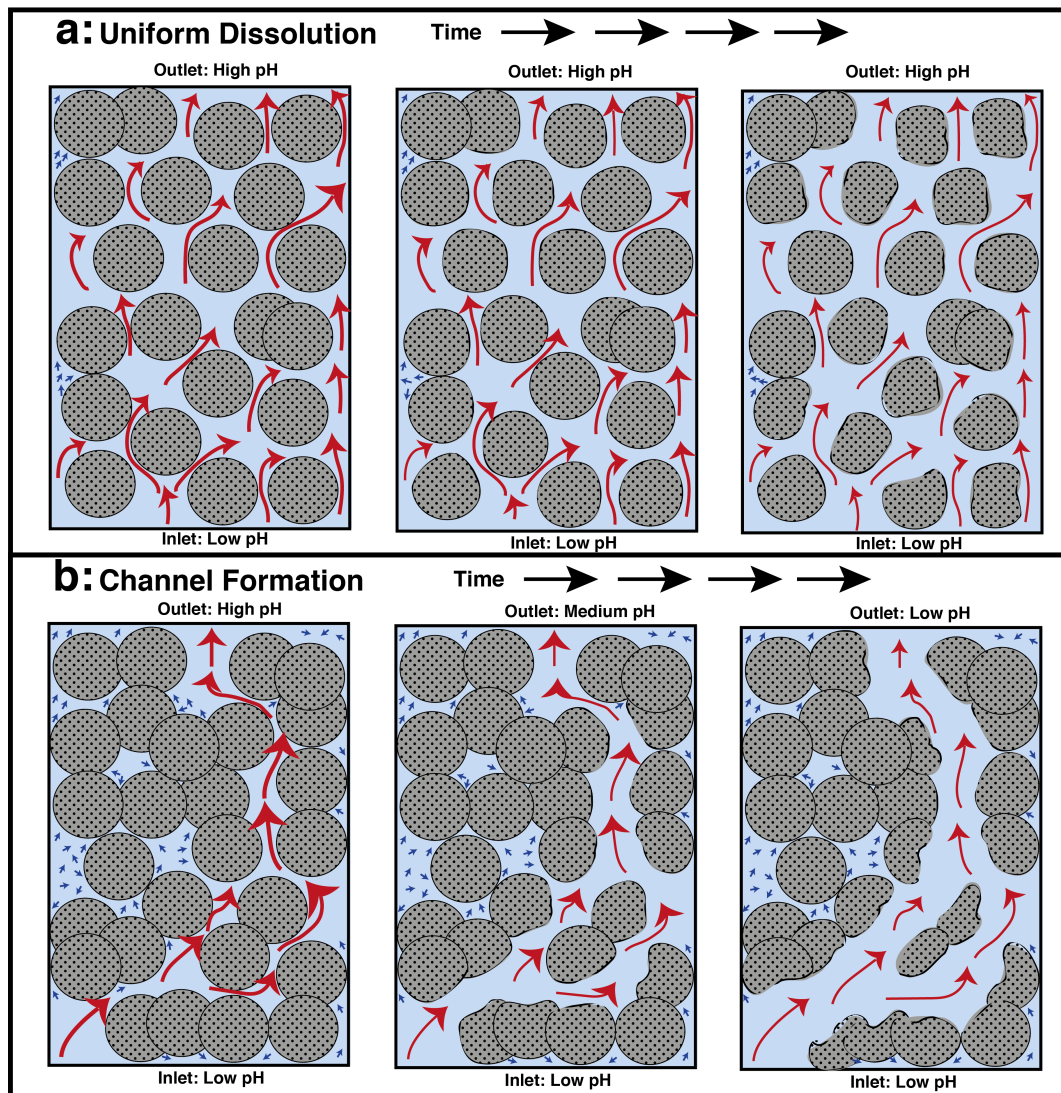
4.65×10^{-5}

3.88 ± 0.16

$1.4 \times 10^{-4} \pm 4.2 \times 10^{-5}$

875

876



877

878 **Figure 14** A schematic of the flow and transport processes through time during

879 (a) uniform dissolution and (b) channel formation. Rock grains (grey circles) are

880 dissolved in advectively-dominated regions of fast flow (red arrows). Diffusive

881 regions (blue arrows) show little to no reaction.

882