

Contributions of a global network of tree diversity experiments to sustainable forest plantations

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First Author Biography

Kris Verheyen is an Associate Professor at the Department of Forest and Water Management, Ghent University. His research interests include studies on (1) the link between biodiversity and ecosystem functioning and (2) the impact of global changes on biodiversity and ecosystem functioning. Using these insights, he tries to develop (3) guidelines for ecological restoration and (4) management strategies for the optimal delivery of multiple ecosystem services.

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1 Contributions of a global network of tree diversity experiments to sustainable forest plantations

2

3 Abstract

4 The area of forest plantations is increasing worldwide helping to meet timber demand and protect natural
5 forests. However, with global **change monospecific** plantations are increasingly vulnerable to abiotic
6 and biotic disturbances. As an adaption measure we need to move to plantations that are more diverse
7 in genotypes, species and structure, with a design underpinned by science. TreeDivNet
8 (www.treedivnet.ugent.be), a global network of tree diversity experiments, responds to this need by
9 assessing the advantages and disadvantages of mixed species plantations. The network currently consists
10 of 18 experiments, distributed over 36 sites and five ecoregions. With plantations 1 to 15 years old,
11 TreeDivNet can already provide relevant data for forest policy and management. In this paper, we
12 highlight some early results on the carbon sequestration and pest resistance potential of more diverse
13 plantations. Finally, suggestions are made for new, innovative experiments in understudied regions to
14 complement the existing network.

15

16 **Keywords:** biodiversity experiments, functional biodiversity research, plantation forest, sustainable
17 forest management, ecological restoration.

18

19 1. A global call for sustainable forest plantations

20 Although the global forest area declined by c. 13 million ha per year between 2000 and 2010, the forest
21 plantation area actually increased annually by c. 5 million ha in the same time period, representing c. 7
22 %, i.e. 264 million ha, of the global forest area in 2010 (FAO 2010). Afforestation rates may increase
23 further due to incentives for carbon sequestration and the global pledge to protect the remaining natural
24 forests of the world against **degradation, e.g. as part of REDD+**. Forest plantations already provide up
25 to 33% of the total industrial roundwood volume harvested annually in the world, and are projected to
26 make up as much as 50% of the global **industrial** roundwood production by 2040 (Kanninen 2010).
27 Beyond wood production, plantations also provide a range of other ecosystem services, including carbon
28 sequestration and water retention (Pawson et al. 2013). Moreover, when incorporated into integrated
29 landscape management, plantations can play a large role in achieving biodiversity conservation
30 objectives by offsetting the need to extract resources from natural forests (Paquette and Messier 2010).

31 Currently, plantation forests are almost exclusively planted as monocultures (Nichols *et al.* 2006, Panel
32 1). Yet, several reviews published recently provide evidence, from both natural forests and plantations,
33 that biomass production and the delivery of other ecosystem services can improve with tree diversity
34 (Nadrowski et al. 2010; Scherer-Lorenzen 2014). Furthermore, global change may increase disturbance
35 frequencies and intensities in both natural forest (Woods et al. 2005) and plantations (Pawson et al.
36 2013), significantly affecting wood supply chains with severe economic consequences (Hanewinkel et
37 al. 2012). Forest plantations that are diverse in genotypes, species, structure and function, should be
38 better able to adapt to changing environmental conditions than monocultures (**van Hensbergen 2006**;
39 Bauhus et al. 2010). This calls for the development of novel, more diversified forest plantations that can
40 improve plantations' stability, productivity and delivery of ecosystem services. Since plantations are
41 often established near human settlements, they are the primary window through which society looks at
42 forest management. Changing the way we manage plantations and set objectives for them can therefore
43 have profound and rapid impacts on the social acceptance of forestry (Paquette and Messier 2013). It
44 has been noted, however, that foresters currently resist establishing mixed plantations, in large parts

45 because of the perception that mixing genotypes and species reduces yield and complicates forest
46 management operations (Carnol *et al.* 2014).

47 TreeDivNet, a new global network of tree diversity experiments, responds to the need for a solid,
48 science-based framework for documenting and understanding the benefits and drawbacks of mixed
49 plantations. In this paper, we explain the need for new afforestation trials and present the TreeDivNet
50 network of experimental plantations. We show some early results from the network and formulate
51 suggestions for additional experimental plantations that may cover existing research gaps.

52

53 **2. The need for a 21st century generation of forest plantation trials**

54 In the 18th and 19th century, foresters such as von Carlowitz, Hartig and Cotta developed the concepts
55 of sustainable forest management as a response to the increasing overexploitation of European forests
56 (Morgenstern 2007). To base these concepts upon science, the first long-term silvicultural trials were
57 established to identify the most productive species and provenances to plant in novel forests. The trials
58 were definitely a success for the development of production-oriented management; large-scale forest
59 plantations were established with fast-growing tree species. The trials were often designed as common
60 garden experiments comparing the growth and performance of different species and provenances at one
61 site, i.e., under similar environmental conditions. Despite the lively debate about the advantages and
62 disadvantages of pure versus mixed forests (even in that early era), most of the trials consisted of
63 monocultures or, less frequently, two-species mixtures (Scherer-Lorenzen 2014). Presently, 300 years
64 after von Carlowitz's proposition of sustainability and given recent advances in biodiversity science
65 (e.g. Cardinale *et al.* 2012), we need to know which mixtures provide **higher levels of biomass**
66 **production and of other ecosystem services** and how environmental conditions affect the relationship
67 between tree diversity and forest functioning, both in space and time.

68 To address these issues, several scientific approaches are available. Given the long lifespan and size of
69 trees, simulation models that predict ecosystem service output along a range of tree diversities and

70 environmental conditions are an obvious approach. However, such models need parameterization, which
71 is an enormous challenge given how poorly we understand biotic interactions among species. Parameters
72 can be estimated based on experiments or observational studies, but both the types and ranges of tree
73 diversities we seek to study are not always present. Still, highly interesting and relevant work has been
74 accomplished with simulation tools (e.g. Morin et al. 2011). Observational studies are invaluable for
75 providing real-world reference data (Baeten et al. 2013), but also have many drawbacks because tree
76 species composition strongly depends on environmental factors or management. Experiments avoid
77 these issues, but there are still relatively few experiments with replicated stands of mixed species
78 (Scherer-Lorenzen 2014), and many of these use only a small number of **(nevertheless commercially**
79 **important)** tree species.

80

81 **3. TreeDivNet and examples of its potential to contribute to sustainable forest plantations**

82 In response to the need for in-depth knowledge of the functioning of mixed plantations and the services
83 they provide, tree diversity experiments have been planted worldwide over the past 15 years. These
84 experiments have now been integrated within the global network TreeDivNet
85 (www.treedivnet.ugent.be). The unifying characteristic of TreeDivNet experiments is that tree species
86 are grown in both monoculture and mixtures, and that tree diversity levels are replicated in a randomized
87 design, allowing for the effects of diversity to be tested. Tree diversity experiments can yield reliable
88 estimates of ecosystem functioning as the experimental design controls the levels and range of tree
89 diversity and allows accounting for potentially confounding factors due to site conditions and local
90 environmental gradients. In addition, long-term monitoring of the performance of individual trees and
91 multiple ecosystem processes in experiments will provide a rich record of the development of the forest
92 ecosystem and its overall functioning (see for example Potvin and Gotelli 2008). This will lead to a
93 deeper understanding of the influence of the diversity, composition and structure of a forest on its
94 functioning and a more complete picture of the relationships between productivity and other ecosystem
95 functions and services. Long-term monitoring will also allow us to better understand how forest

96 diversity, structure and composition influence forest stability. We will then be able to plant and manage
97 forests in a way that increases their resistance and resilience to, e.g., predicted changes in climate.
98 Different aspects of tree diversity, i.e., species richness, genetic diversity, structural and functional
99 diversity, will be used as tools to face the key challenges of modern sustainable afforestation.

100 At present, TreeDivNet consists of 18 experiments, located at 36 sites and in five ecoregions (Table 1;
101 Figure 1). More than 1 000 000 trees have been planted in the experiments on a total surface area of c.
102 800 ha, which makes TreeDivNet one of the largest research infrastructures in ecology worldwide. The
103 oldest experiment (Satakunta, Finland) was planted in 1999. The experiments included in TreeDivNet
104 manipulate woody plant diversity – in terms of species richness (taxonomic diversity), evenness,
105 composition, genetic and functional diversity – over wide diversity gradients and are designed to allow
106 separation of diversity and identity effects (see Figure 2 for an example, and Bruelheide et al. 2014).
107 The tree species in the TreeDivNet experiments are both widely planted commercial species, but also
108 many less-frequently used species. One important additional component is the inclusion of tree
109 provenances from different regions (e.g., BiodiversiTREE, US; FORBIO, Belgium; and Climate Match,
110 UK), providing a valuable opportunity to test whether assisted migration enhances the services provided
111 by diverse plantations in the face of climate change (Pedlar et al. 2012).

112 TreeDivNet functions according to the guidelines for globally distributed experiments (cf. Borer et al.
113 2014). At present, the network has no central funding. Participation is entirely voluntary, but has clear
114 benefits for the participants. TreeDivNet offers unique opportunities for multidisciplinary and
115 multifunctional research on the relationship between tree diversity and ecosystem functioning in major
116 forest types around the world and enables synthesis studies across the globe. Thus, TreeDivNet
117 contributes to the lively field of functional biodiversity research, which has delivered a wealth of
118 knowledge about the biotic control of ecosystem functioning over the last two decades. However, most
119 of this knowledge was gained in smaller-stature, shorter-lived vegetation such as grasslands; forests
120 came into the focus of this research field only recently. Despite the young age of most experiments,
121 TreeDivNet can already provide results relevant for policy and management, as illustrated in the
122 following two examples.

123

124 **3.1 Species identity, plot diversity, and mixture composition as determinants of aboveground**
125 **carbon sequestration**

126 The possibility of using afforestation to create carbon sinks while taking biodiversity concerns into
127 account provides a good example of the potential contributions of experimental tree plantations within
128 TreeDivNet. Sequestering both above and belowground carbon has been recognized in the context of
129 the Clean Development Mechanism of the Kyoto protocol (Thomas et al. 2010), and has gained
130 momentum with the development of an international mechanism for reducing emissions from
131 deforestation and forest degradation known as REDD+ (Cerbu et al. 2011). However, the choice of
132 provenance/genotype and species, **each with different carbon sequestration time profiles**, and the
133 positive or negative effects of mixtures for maximizing carbon sequestration rates in forest plantations
134 at different sites across the globe are still open to debate.

135

136 According to FAO's Global Planted Forest Assessment database (FAO 2006), the total number of
137 species used in plantations ranges from four in Finland to twenty in China, France, India, and Ukraine.
138 Yet, studies in TreeDivNet experimental plantations suggest that the carbon sequestration rates of tree
139 species that are rarely planted in forestry may be higher than for species that are traditionally planted
140 for wood production. In Sardinilla, Panama, for instance, only one of the four species with the highest
141 carbon stocks after 10 years of growth, *Dalbergia retusa*, is currently used as a timber-producing species
142 (Figure 3a). In BEF-China, *Choerospondias axillaris*, *Nyssa sinensis*, *Triadica cochinchinensis*, *Melia*
143 *azedarach* and *Schima superba*, which are not currently used for commercial timber, were found to
144 sequester more carbon two years after planting than the commercially planted timber species
145 *Cunninghamia lanceolata* or *Pinus massioniana*. Early observations thus support the presence of species
146 identity effects, which highlights the importance of increasing the number of species used in plantation
147 projects. **Nevertheless, widespread application of these new species is probably contingent on their**
148 **potential use as timber species.**

149 TreeDivNet experiments also allow comparing the provisioning of ecosystem services from mixed as
150 opposed to monoculture plantations. A recent meta-analysis, using data from a TreeDivNet experiment
151 and elsewhere, indicates that woody mixtures sequester at least as much aboveground carbon as the most
152 productive monocultures in any given location (Hulvey et al. 2013). This suggests that plantations could
153 use mixtures of multiple species selected outside of traditional forestry practice to maximize above-
154 ground carbon storage. **if the latter would be the primary interest**. Furthermore, early TreeDivNet results
155 indicate that the performance of high carbon sequestering species might be contingent upon the diversity
156 level of the plot in which they are growing. In BangorDIVERSE, UK, *Alnus glutinosa* and *Betula*
157 *pendula* were more efficient at storing carbon after nine years than **some** traditional timber-producing
158 species, with *A. glutinosa* performing better in mixture than in monoculture (Figure 4). In Sardinilla,
159 mixtures established with three and six species overyielded compared with monocultures and this effect
160 of diversity increased with time over 10 years (Sapijanskas et al. 2013). However, variability among
161 plots with the same species richness level also suggests that certain combinations of species are
162 apparently able to sequester more carbon than others.

163

164 We propose that, in order to more easily identify species and mixtures that sequester high levels of
165 carbon, relationships between carbon sequestration rates and common life history traits could be useful.
166 Early data collected at TreeDivNet experiments suggest that these relationships may be site-specific, as
167 has been found in natural forests (Stegen et al. 2009).

168

169 **3.2 Which mixtures optimize insect pest control in young tree plantations?**

170 Although often less spectacular than abiotic disturbances such as storms or fires, biotic damage can
171 dramatically alter the functioning of forest ecosystems and reduce their productivity. For instance, every
172 year, on average 15 - 20% of the trees in European forests are affected by pest and pathogen damage,
173 resulting in increased tree mortality or reduced tree growth. Climate change with increasing
174 temperatures and more frequent drought events is expected to aggravate forest pest damage through
175 increased pest proliferation or reduced plant defense (Jactel et al. 2012). It is therefore critical to better

176 understand the significance of forest diversity for the forest's resistance to pest insects and its resilience
177 to their outbreaks.

178

179 Meta-analyses have shown that, overall, mixed forests are less prone to pest insect damage than
180 monocultures (Jactel and Brockerhoff 2007), supporting the associational resistance hypothesis. This
181 hypothesis states that focal trees surrounded by heterospecific neighbours are less likely to be found and
182 affected by insect herbivores. However, these reviews have several limitations: (1) they focused on the
183 effects of single pest species, whereas the entire community of insect herbivores interacts with the trees;
184 (2) the long-term effects of insect herbivory have not been studied; and (3) the ecological mechanisms
185 underlying associational resistance could not be investigated in detail.

186

187 By contrast, the design of the TreeDivNet experiments makes it possible to address these issues. Indeed,
188 early results on diversity - herbivore resistance relationships from BIOTREE (Germany), FORBIO
189 (Belgium), Satakunta (Finland), and ORPHEE (France) indicate that the identity of the focal (Figure 5)
190 and associated tree species appeared to be more important than plot species richness *per se* in explaining
191 the effects of tree diversity on insect herbivory damage. Interestingly, there were more cases found for
192 associational susceptibility, which might be due to the young age of the experiments and/or the
193 assessment of all insect damage rather than a focus on few pests, as done in other studies. Insect damage
194 is now a staple protocol in most TreeDivNet experiments and so more results over a greater span of
195 conditions will be available soon.

196

197 A recent meta-analysis, which included data from several TreeDivNet experiments, has shown that both
198 phylogenetic relatedness of tree species in mixtures and insect herbivore feeding specialization are
199 important predictors of forest diversity effects on insect pests (Castagneyrol et al. 2014). The degree of
200 dilution of a focal tree species among non-host trees was also important in associational resistance
201 (Castagneyrol et al. 2013). Moreover, reduced host-tree apparency recently emerged as a main driver of
202 resistance in mixed stands as neighbouring heterospecific trees can disrupt host-finding behavior in
203 insect herbivores (Castagneyrol et al. 2013). Finally, mixed forests can provide natural enemies with

204 more feeding resources or microhabitats and thus enhance the biological control of pest insects
205 (Riihimaki et al. 2005).

206

207 These preliminary findings provide a basis for several recommendations for the design of mixed species
208 plantations that can be more resistant to insect pests: (1) mixing more functionally and phylogenetically
209 dissimilar tree species, such as conifers and broadleaves, can result in a more effective reduction in
210 herbivore damage (Castagneyrol et al. 2014), but (2) a significant reduction in the proportion of host
211 trees in mixtures is required to reduce damage by specialist herbivores (Jactel and Brockerhoff 2007).

212

213 **4 Ideas for additional experimental tree diversity plantations**

214 We are now entering the second decade of experimental manipulations of tree diversity. The TreeDivNet
215 experiments have been designed to understand mechanisms and to quantify a large suite of ecosystem
216 functions and services relevant to 21st century forest plantations. Gaps remain, however, in both the
217 scale and scope of the existing experiments. We outline some important aspects here to guide future tree
218 diversity experiments (see also Bruelheide et al. 2014).

219

220 First, while biodiversity research has made considerable advances on theoretical grounds, there is still a
221 lack of linkages to applied sciences and industrial practices, even though it has been shown that different
222 management types and intensities affect diversity-function relationships (e.g. Weigelt et al. 2009). In
223 addition, the provision of wood is always listed among the ecosystem services a forest, planted or not,
224 can provide. The outreach of next-generation experiments would be tremendously increased if practical
225 issues were added already during the design phase, for example treatments testing and costing different
226 planting patterns, maintenance methods, and harvesting techniques in a multi-species context, both in
227 plantations and in naturally-regenerated forests (see also Nichols et al. 2006). There is hence an
228 important need for mixed species demonstration experiments, set-up in collaboration with forest
229 managers and industries, and established at operational scales using available equipment and techniques.

230 This could apply to both forestry and agroforestry systems, including short-rotation coppices and all
231 variations of selection and multi-cohort stands. Moreover, to be practically relevant, future experiments
232 may need to focus more strongly on testing or finding well-functioning genotypic and species
233 compositions.

234

235 A second big issue in the design of tree diversity experiments is the scale, both temporal and spatial.
236 Because of the high costs of large plots and the long-term time commitments, most plots in TreeDivNet
237 experiments are, with a few exceptions, ¼ hectare or smaller (Table 1). Many processes affecting forest
238 dynamics, e.g., competition and mortality, are scale-dependent, and many of the forest ecosystem
239 services, including the provision of timber, biodiversity, water purification, carbon storage, and
240 recreational opportunities, are supplied at different spatial and temporal scales. Hence, there is an urgent
241 need for tree diversity experiments that capture these larger scale processes, similar to seminal
242 watershed-level studies such as Hubbard Brook (www.hubbardbrook.org). Studies spanning multiple
243 scales could provide pivotal information regarding the spatial and temporal scales at which forest
244 biodiversity influences ecosystem functions and services. Comparing watersheds with different
245 manipulated tree diversities would be a truly important step forward. Such large-scale experiments could
246 be inspired by a land-sharing vs. land-sparing approach, such as the functional zoning in forestry (e.g.
247 Messier et al. 2009). Furthermore, as effects of biodiversity on ecosystem functioning appear to be time-
248 dependent and to grow larger with time (Reich et al. 2012), longer-term studies are also required. While
249 some of our experiments are planned with such long-term temporal perspective, others focus on early
250 phases of establishment. Still missing are experiments where species are planted at different points in
251 time, with pioneer and mid- to late-successional species, which without doubt would enhance our
252 predictive capabilities of diversity effects along successional trajectories.

253

254 Third, theory and empirical evidence suggest that biodiversity is particularly important to buffer
255 ecosystems against stressors and to increase their stability (Loreau and de Mazancourt 2013), but to date

256 few TreeDivNet experiments explicitly incorporate stress as an experimental factor. The ORPHEE
257 (France) and IDENT (Canada, Italy) experiments have incorporated a water availability treatment, and
258 the IDENT site in Germany and Ridgefield (Australia) incorporate nutrient addition treatments, but the
259 inclusion of other stressors would clearly broaden the inferences of TreeDivNet experiments. For
260 example, results from smaller-scale experiments have shown that including factors such as mammalian
261 herbivory (Cook-Patton et al. 2014) and fire (Adair et al. 2009) can influence the direction and
262 magnitude of diversity effects.

263 **Fourth and finally**, although TreeDivNet includes experiments in tropical, temperate, and boreal
264 systems, the distribution of experiments is skewed as relatively few are located in other important
265 biomes/climate regions. For example, only two experiments lie in Central/South America and one in
266 Africa, but these are not located in the largest forested areas and biodiversity hotspots on either continent
267 (i.e. in the Amazon or Congo Basin). In addition, despite covering large areas on the globe, shrublands
268 are also underrepresented.

269 The foresters of the 19th century demonstrated an impressive long-term perspective when they
270 established the first forestry trials to find answers to the pressing questions of that time. Globally
271 distributed experiments, such as TreeDivNet, could become new important research pillars to face the
272 great challenges that global changes will put on forest ecosystems and to deliver highly relevant
273 guidelines for forest policy and management worldwide. This is particularly important since plantations
274 are likely to increase tremendously in area worldwide in the next decades.

275

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405 **Panel 1**

406 Multi-species tree plantations are still relatively rare worldwide, but is this topic important within the
407 forest research communities and is there an increasing interest in the last 10 years? We investigated
408 these questions using the software WORDSTAT 6.0 (Péladeau 2003) by comparing the percentage of
409 abstracts containing the word “plantation” that also contained the words “species mixture, mixed system,
410 mixed plantation, mixed-species plantation or multi-species plantation” between the proceedings of the
411 IUFRO World Congresses* of 2005 and 2014. In the proceedings of 2014, we found 2426 abstracts of
412 which 267 used the term “plantation”. Of these 267 abstracts, 20 (or 7.5%) also used at least one of the
413 terms referring to mixed plantation mentioned above. In the proceedings of 2005, we found 1454
414 abstracts of which 238 used the term “plantation”. Of these 238 abstracts, only 1 (or 0.4%) also used at
415 least one of the terms referring to mixed plantation. This clearly shows that the interest in multi-species
416 tree plantations is increasing, which bodes well for the future of such plantations worldwide.

417

418 *: IUFRO is the International Union of Forest Research Organizations and organizes its world congress every 4 or 5 years
419 (www.iufro.org)

420

Table 1 The 18 experiments of TreeDivNet are established around the globe (see Figure 1) to investigate the relations between different aspects of forest ecosystem functioning and tree diversity: species richness (SR), functional diversity (FD), genetic diversity (GD), phylogenetic diversity (PD), and evenness (EV). See www.treedivnet.ugent.be for more information on the experiments.

| ID | ecoregion | name | plant year | no sites | no plots | species pool | plot size (m ²) | tree diversity ^a | SR gradient | FD variables | GD gradient |
|-----|-----------|---------------------|------------|----------|----------|--------------|-----------------------------|-----------------------------|---------------|---------------------|---|
| bo1 | boreal | Satakunta | 1999 | 4 | 163 | 5 | 400 | SR, GD, PD | 1, 2, 3, 5 | - | 1, 2, 4, 8 clones (<i>Betula</i>) |
| te1 | temperate | BiodiversiTREE | 2013 | 1 | 75 | 16 | 1225 | SR, FD, GD | 1, 4, 12 | AM, EM fungi | 1,2 provenances |
| te2 | temperate | BangorDIVERSE | 2004 | 1 | 92 | 7 | 45-196 | SR, FD | 1, 2, 3 | shade tolerance | - |
| te3 | temperate | Climate Match | 2011 | 2 | 177 | 4 | 144, 1152 | SR, GD | 1, 4 | - | 1, 2, 3, 4 provenances |
| te4 | temperate | FORBIO ^b | 2010, 2012 | 3 | 127 | 10 | 1296, 1575, 1764 | SR, GD | 1, 2, 3, 4 | - | 1, 3 provenances (<i>Quercus, Fagus</i>) |
| te5 | temperate | ORPHEE | 2008 | 1 | 256 | 5 | 400 | SR, FD | 1, 2, 3, 4, 5 | deciduous/evergreen | - |

| | | | | | | | | | | |
|------|---------------|-------------------------|------------------------|---|------|------|------------------------|-------------------|--|------------------------------------|
| te6 | temperate | Communitree | 2009 | 1 | 90 | 1 | 0.24 GD | - | - | 1, 2, 3, 4 half-sib families |
| te7 | temperate | ECOLINK-Salix | 2014 | 3 | 99 | 1 | 92 GD | - | - | 1, 2, 3, 4 clones (<i>Salix</i>) |
| te8 | temperate | Kreinitz | 2005 | 1 | 98 | 6 | 25 SR, FD | 0, 1, 2, 3, 5, 6 | litter decomposition rate | - |
| te9 | temperate | B-Tree | 2013 | 1 | 44 | 4 | 170 - 300 SR, FD | 1, 2, 4 | AM, EM fungi | - |
| te10 | temperate | BIOTREE ^b | 2003, 2004 | 4 | 117 | 19 | 300 - 12000 SR, FD, EV | 1, 2, 3, 4, 6, 10 | 9 traits | - |
| te11 | temperate | IDENT ^b | 2009, 2010, 2012, 2013 | 5 | 1192 | 1919 | 8-16 SR, FD, PD | 1, 2, 4, 6, 12 | native/exotic c. 20 traits | - |
| me1 | Mediterranean | IDENT ^b | 2014 | 1 | 308 | 12 | 10 SR, FD, PD | 1, 2, 4, 6 | evergreen/deciduous drought resistance | - |
| me2 | Mediterranean | Ridgefield ^b | 2010 | 1 | 124 | 8 | 447 SR, FD | 0, 1, 2, 4, 8 | nutrient acquisition | - |

| | | | | | | | | | | growth form | | |
|-----|-------------|------------------------|-----------|---|-----|----|----------|------------|-------------------|------------------------|-----------------------|--------------------------|
| | | | | | | | | | | 0, 1, 2, 4, 8, 16, | random extinction | 3 - 38 half-sib families |
| st1 | subtropical | BEF-China ^b | 2009/2010 | 2 | 566 | 60 | 667 | SR, GD | 24 tree sp. | scenarios and directed | (for 13 tree species) | |
| | | | | | | | | | crossed with 0, | scenarios based on | 1 or 4 seed families | |
| | | | | | | | | | 2, 4, 8 shrub sp. | SLA and rarity | per species | |
| tr1 | tropical | Agua Salud | 2008 | 1 | 267 | 10 | 1755 | SR | 1, 2, 5, 6 | - | - | |
| tr2 | tropical | Sardinilla | 2001/2003 | 2 | 32 | 26 | 675-2025 | SR, FD | 1, 3, 6, 9, 18 | shade tolerance | - | |
| tr3 | tropical | Gazi Bay | 2004 | 1 | 32 | 3 | 36 | SR | 1, 2, 3 | - | - | |
| tr4 | tropical | Sabah ^b | 2010 | 1 | 124 | 16 | 40000 | SR, FD, GD | 1, 4, 16 | tree height | 2, 4 genera | |

^a extra treatments investigated: water availability (ORPHEE, IDENT), fertilization with N, P, N+P (IDENT), N deposition and non-native weed cover

(Ridgefield), liana removal (Sabah)

^b extensive info on the design of these experiments can also be found in Bruelheide et al. (2014; BEF-China), Hector et al. (2011; Sabah), Perring et al. (2012; Ridgefield), Scherer-Lorenzen et al. (2007; BIOTREE), Tobner et al. (2014; IDENT), and Verheyen et al. (2013; FORBIO).

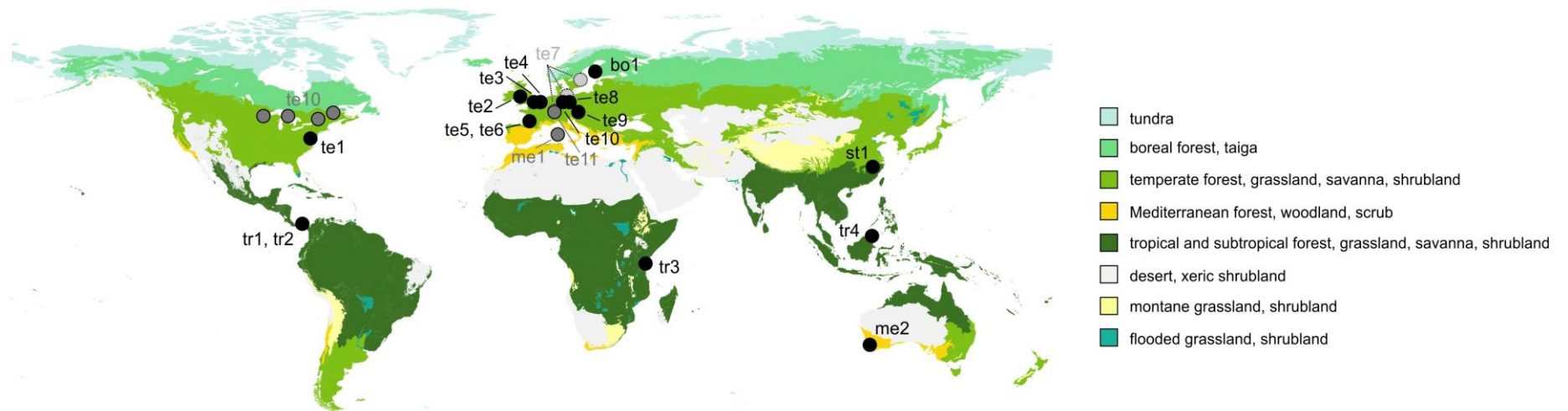


Figure 1. The 18 experiments of TreeDivNet in the boreal (bo), temperate (te), Mediterranean (me), subtropical (st) and tropical (tr) regions of the world. The dark grey dots represent the IDENT experiment; the light grey dotted ones are the ECOLINK-Salix experiment; the other experiments are in black. See Table 1 for the characteristics of the experiments. Map based on Olson et al. (2001), data from <http://www.worldwildlife.org/publications/terrestrial-ecoregions-of-the-world>.

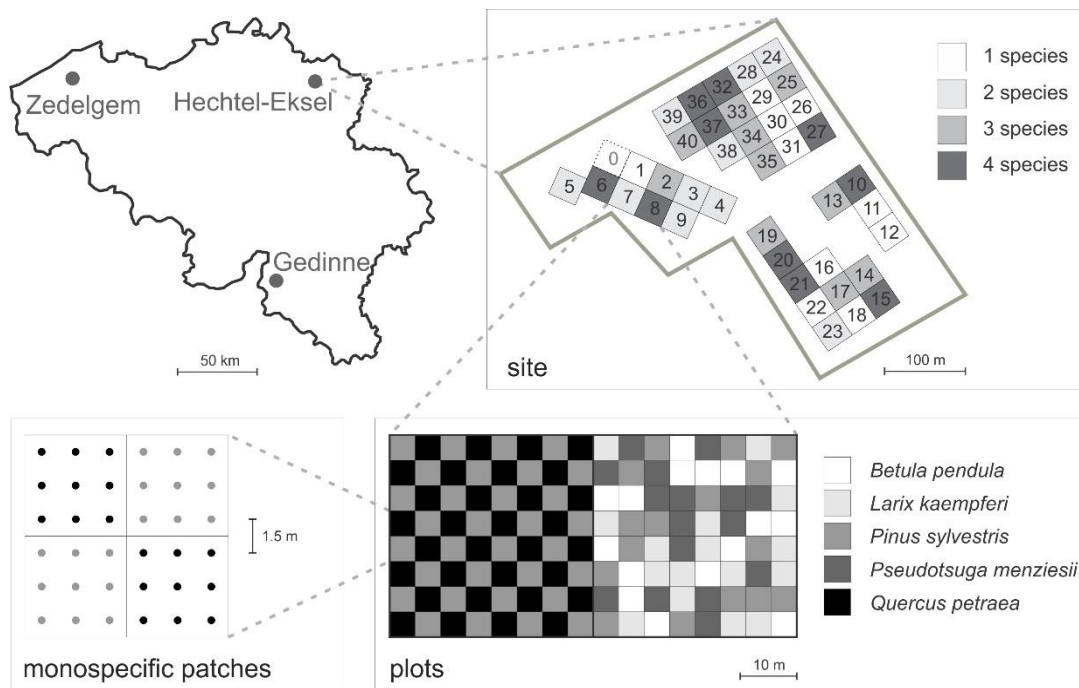


Figure 2. Example of the design of one of the TreeDivNet experiments. The FORBIO experiment was established at three sites in Belgium. The tree species diversity per plot ranges from one to four species. The within-plot design is shown for a two-species and a four-species plot. The trees were planted on a 1.5 m x 1.5 m grid, in small monospecific patches of 3 x 3 trees. These patches are arranged in a checkerboard pattern in the two-species mixtures and randomly attributed to the species in the three- and four-species mixtures (see Verheyen et al. 2013 for more details).

Figure 3. Aboveground biomass (Mg C ha^{-1}) after 10 years of growth in the Sardinilla experiment (Panama). The common timber species are indicated in green in the figure and underlined here. Species abbreviations are the first letter of the genus and species name: *Albizia adinocephala*, *Anacardium excelsum*, *Astronium graveolens*, *Cordia alliodora*, *Calycophyllum candidissimum*, *Colubrina glandulosa*, *Cedrela odorata*, *Dalbergia retusa*, *Diphysa robinoides* (DRO), *Enterolobium cyclocarpum*, *Erythrina fusca*, *Gliricidia sepium*, *Guazuma ulmifolia*, *Hura crepitans*, *Inga punctate*, *Luehea seemannii*, *Ormosia macrocalyx*, *Pachira quinata*, *Pseudosamanea guachapele*, *Spondias mombin*, *Tabebuia rosea*. The biomass was calculated using the equation of Chave (2005) equation for tropical moist forest, and mean tree biomass per species was scaled up to one hectare assuming 1000 trees per plot. Estimations were done for the species represented in the Sardinilla planted forest by at least five individuals.

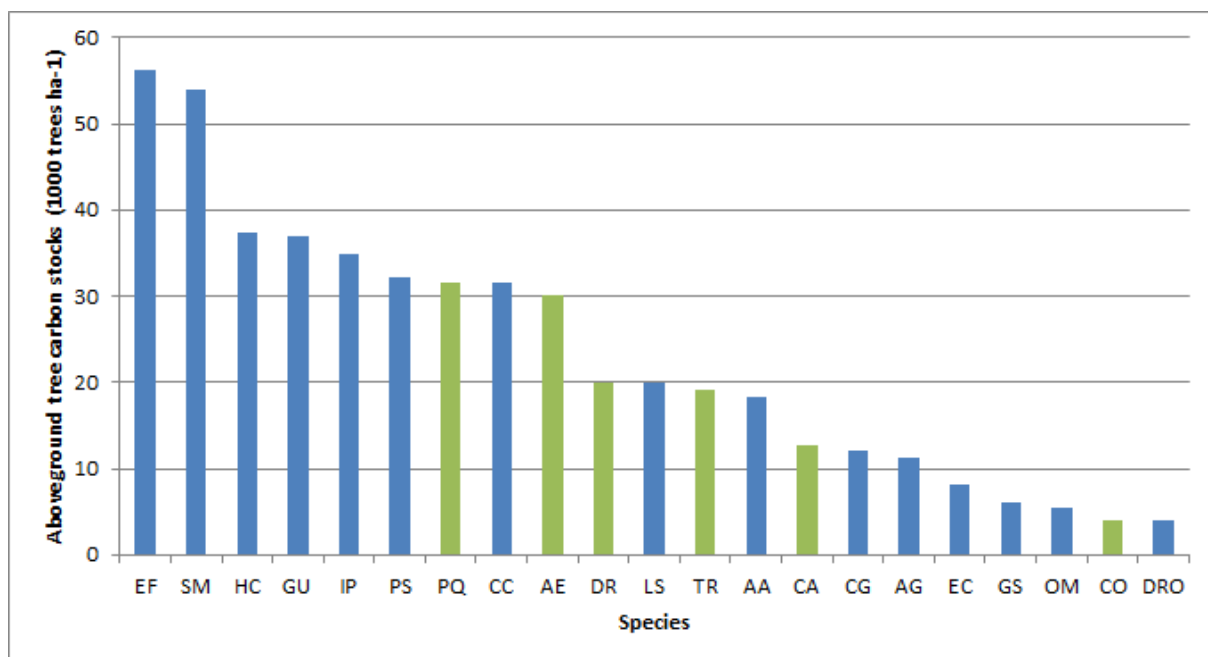
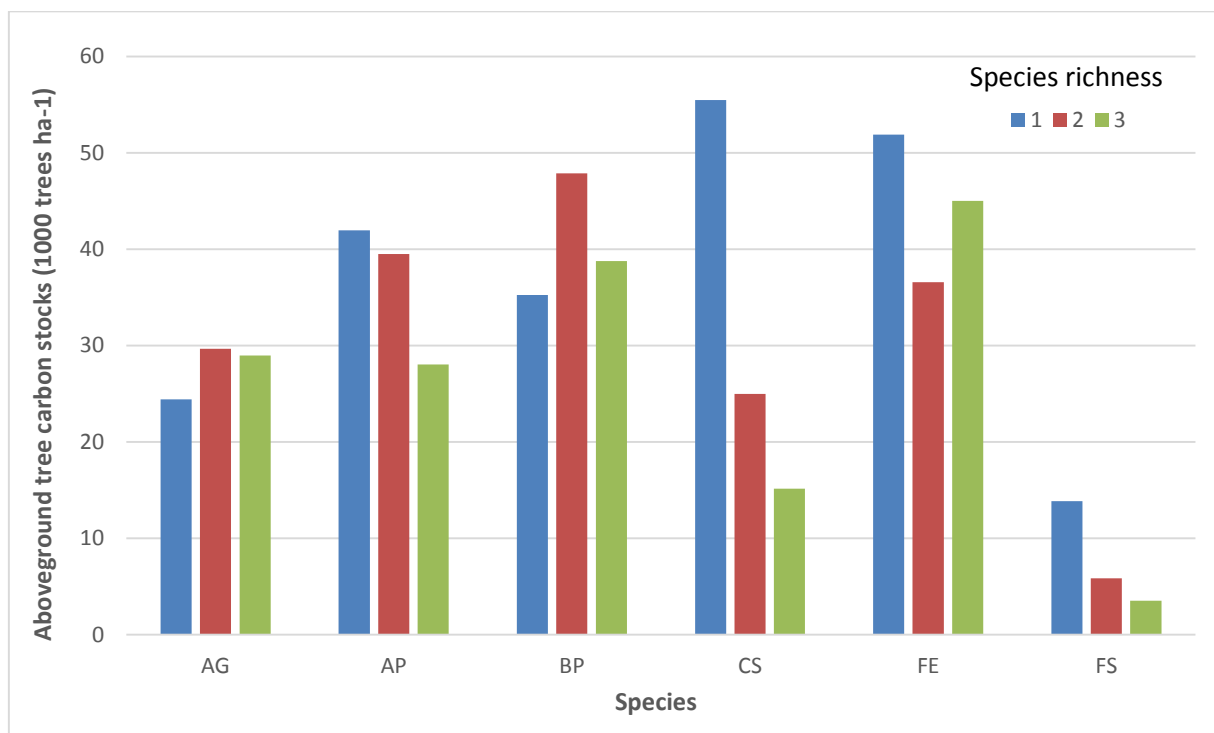


Figure 4. Aboveground carbon (Mg C ha^{-1}) after 9 years of growth at the BangorDIVERSE experiment (UK). Species abbreviations are the first letter of the genus and species name: *Alnus glutinosa*, *Acer pseudoplatanus*, *Betula pendula*, *Castanea sativa*, *Fraxinus excelsior*, *Fagus sylvatica*. The biomass was calculated using general European temperate forest equations from Zianis et al. (2005) and site-specific equations from Smith et al. (2013). Mean tree biomass per species was scaled up to one hectare assuming 1000 trees per plot. Biomass estimations were based on the average species diameter of each replicate plot ($n=3$).



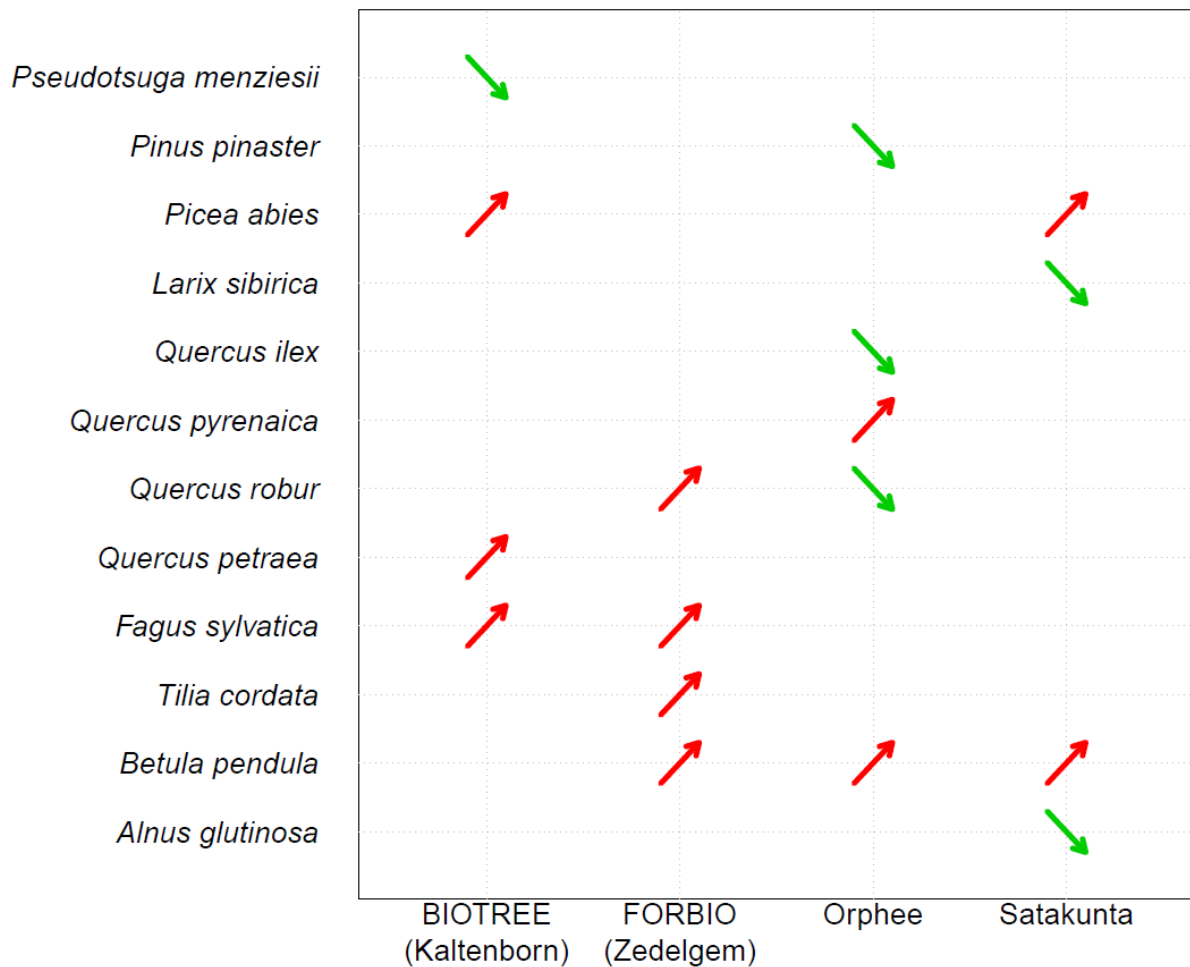


Figure 5. Species-specific responses of defoliation (chewing + skeletonizing damage) to tree diversity in four TreeDivNet experiments. Green and red arrows indicate reduced and increased herbivory in mixed plots as compared to monocultures, i.e. associational resistance and associational susceptibility, respectively. It was estimated based on the site-specific difference in mean damage on a given species grown in mixtures and mean damage on corresponding monocultures. Data was taken from Setiawan et al. (2014) for the FORBIO experiment and from Haase et al. (2015) for the BIOTREE, ORPHEE and Satakunta experiment.