1	Synthesis
2	
3	
4	What we use is not what we know: environmental predictors in plant distribution models
5	
6	
7	Heidi K. Mod, Daniel Scherrer, Miska Luoto & Antoine Guisan
8	
9	Mod, H.K. (corresponding author, heidi.mod@helsinki.fi) ^{a, b, †}
10	Scherrer, D. (daniel.scherrer@unil.ch) ^{b, †}
11	Luoto, M. (miska.luoto@helsinki.fi) ^{a, ‡}
12	Guisan, A. (antoine.guisan@unil.ch) ^{b, c, ‡}
13	
14	^a Department of Geosciences and Geography, University of Helsinki, P.O. Box 64 (Gustaf
15	Hällstöminkatu 2a), FI-00014 University of Helsinki, Finland
16	^b Department of Ecology and Evolution, University of Lausanne, Biophore, CH-1015 Lausanne,
17	Switzerland
18	^c Institute of Earth Sciences, University of Lausanne, Géopolis, CH-1015 Lausanne, Switzerland
19	
20	[†] Shared first authorship.
21	[‡] Shared last authorship.
22	
23	Printed journal page estimate: 9837 words (12.3 pages), 1 table 0.5 pages, 5 figures 2 pages, total
24	15 pages

25 ABSTRACT

Questions: The choice of environmental predictor variables in correlative models of plant species 26 distributions (hereafter SDMs) is crucial to ensure predictive accuracy and model realism, as 27 highlighted in multiple earlier studies. Because variable selection is directly related to a model's 28 capacity to capture important species' environmental requirements, one would expect an explicit 29 prior consideration of all ecophysiologically meaningful variables. For plants, these include 30 temperature, water, soil nutrients, light, and in some cases, disturbances and biotic interactions. 31 However, the set of predictors used in published correlative plant SDM studies varies considerably. 32 No comprehensive review exists of what environmental predictors are meaningful, available (or 33 missing), and used in practice to predict plant distributions. Contributing to answer these questions 34 is the aim of this review. 35

Methods: We carried out an extensive, systematic review of recently published plant SDM studies (years 2010-2015; n = 200) to determine the predictors used (and not used) in the models. We additionally conducted an in-depth review of SDM studies in selected journals to identify temporal trends in the use of predictors (years 2000-2015; n = 40).

40 Results: A large majority of plant SDM studies neglected several ecophysiologically meaningful
41 environmental variables, and the number of relevant predictors used in models has stagnated or
42 even declined over the last 15 years.

Conclusions: Neglecting ecophysiologically meaningful predictors can result in incomplete niche quantification and can thus limit the predictive power of plant SDMs. Some of these missing predictors are already available spatially or may soon become available (e.g., soil moisture). However, others are not yet easily obtainable across whole study extents (e.g., soil pH and nutrients), and their development should receive increased attention. We conclude that more effort should be made to build ecologically more sound plant SDMs. This requires a more thorough rationale for the choice of environmental predictors needed to meet the study goal, and the

50	development of missing ones. The latter calls for increased collaborative effort between ecological
51	and geo-environmental sciences.
52	
53	Keywords: covariate; environment; habitat suitability; independent variable; model; niche; plant;
54	predictor; species distribution;
55	
56	Abbreviations: DEM = digital elevation model, GIS = geographic information system, SDM =
57	correlative species distribution modelling, WoS = ISI Web of Science
58	
59	Running title: Variable selection and species distribution models
60	
61	
62	INTRODUCTION
63	
64	Correlative species distribution modelling (SDM; also called ecological niche, habitat suitability,
65	and (bio)climatic envelope modelling as well as various other names, hereafter all included under
66	the acronym 'SDM'; see Guisan et al. 2013) is a topical approach in ecology and biogeography
67	
	(Franklin 2009, Peterson et al. 2011, Moquet et al. 2015). Over the last decades (Booth et al. 2014),
68	(Franklin 2009, Peterson et al. 2011, Moquet et al. 2015). Over the last decades (Booth et al. 2014), the number of correlative SDM studies has steadily increased, and SDM is currently one of the most
68 69	(Franklin 2009, Peterson et al. 2011, Moquet et al. 2015). Over the last decades (Booth et al. 2014), the number of correlative SDM studies has steadily increased, and SDM is currently one of the most popular methods used to study the impact of various threats to biodiversity and to support related
68 69 70	(Franklin 2009, Peterson et al. 2011, Moquet et al. 2015). Over the last decades (Booth et al. 2014), the number of correlative SDM studies has steadily increased, and SDM is currently one of the most popular methods used to study the impact of various threats to biodiversity and to support related conservation decisions (Guisan et al. 2013). In addition to a large number of case studies on species
68 69 70 71	(Franklin 2009, Peterson et al. 2011, Moquet et al. 2015). Over the last decades (Booth et al. 2014), the number of correlative SDM studies has steadily increased, and SDM is currently one of the most popular methods used to study the impact of various threats to biodiversity and to support related conservation decisions (Guisan et al. 2013). In addition to a large number of case studies on species distributions for conservation and risk assessment (Broennimann & Guisan 2008; Araújo et al.

72 2011; Jiménez-Valverde et al. 2011; Alagador et al. 2014), there is on-going discussion on

theoretical and technical issues, including modelling techniques, selection and evaluation of models,

handling of spatial autocorrelation and, most importantly, variable selection (Franklin 1995; Austin

2002, 2007; Guisan & Thuiller 2005; Araujo & Guisan 2006; Guisan et al. 2006, Dormann 2007;
Elith & Leathwick 2009; Zimmermann et al. 2010; Austin & Van Niel 2011a; Thibaud et al. 2014).
As SDMs statistically relate environmental variables to the presence/absence (or presence-only) of a
species to predict species distributions (Guisan & Zimmermann 2000), the selection of the most
appropriate set of environmental variables as predictors is essential (Dormann 2007).

80

Many of the SDM (sensu lato) reviews published within the last 20 years have called for the use of 81 more ecologically meaningful predictors (Franklin 1995, 2009; Guisan & Zimmermann 2000; 82 83 Guisan & Thuiller 2005; Guisan et al. 2006, Elith & Leathwick 2009; Austin & Van Niel 2011a, Peterson et al. 2011). For plants, seven environmental factors are generally considered essential for 84 growth and survival: temperature, water, nutrients, light, disturbances, biotic interactions and CO₂ 85 86 (Körner 2014, see also Guisan & Zimmermann 2000; Austin & van Niel 2011a and Appendix S1). However, although CO₂ is crucial for plant survival and productivity, it is not a limiting resource 87 under natural growth conditions at current and future atmospheric concentrations (e.g., Körner 88 2006; Norby & Zak 2011; Inauen et al. 2012; Bader et al. 2013). Under such conditions, the nutrient 89 90 cycle and climatic constraints control carbon capture, and therefore CO₂ is usually omitted in 91 correlative analyses of species distributions, such as SDMs, and will not be considered further in 92 this review. All of the other factors can be resources (i.e., can be consumed by the species; e.g. nutrients, water, light) or regulators (i.e., can affect metabolic processes; e.g. temperature; see 93 94 Huston 2002) and can have direct (proximal) and indirect (distal) effects on plants (Austin 2002). Thus, in standard SDMs, where species occurrence (and absence) is modelled principally as a 95 function of environmental conditions, the most realistic and accurate predictions should only be 96 achieved when all factors defining a species' niche and thus driving its distribution are accounted 97 for at the geographic scale considered (Pearson & Dawson 2003; McGill 2010). When considering 98 99 the environmental factors shaping species distribution from a niche modelling perspective, it is also

100 important to distinguish between bionomic (dynamically altered by the species through being consumed or modified) and scenopoetic (constant, not affected by the species) variables (see 101 Hutchinson 1978; Peterson et al. 2011). In this review, by considering the environmental niche 102 (Grinnell 1917; Hutchinson 1957) of plants (Austin 1980; Austin & Smith 1989) in a wide sense, 103 we include both regulator and resource predictors, but because precise data on the dynamics of 104 environmental variables are scarce, we consider resources to remain constant (i.e. we do not 105 consider what could be consumed by the species itself) over the location and time period of the 106 study. 107

108

In addition to the importance of ecological justification for the use of ecophysiologically relevant 109 variables in SDMs, Austin (2002) and later Araujo & Guisan (2006) highlighted the importance of 110 111 acknowledging the biological significance of the selected variables, despite the diverse automated and mathematically optimized variable selection methods developed for SDMs. Additionally, 112 Petitpierre et al. (in review) showed that selecting variables based on expert knowledge rather than 113 an automated selection from huge numbers of predictors can lead to better predictive performances 114 and be more reflective of biological and ecological understanding, especially for fine-scale studies 115 116 (see also Pearson & Dawson 2003 for the hypothesized higher importance of non-climatic variables 117 at finer scales; but see Harwood et al. 2014).

118

Although ecophysiological theory (Lambers et al. 2008; Körner 2014), community assembly
experiments (Fukami et al. 2005; Scherber et al. 2010) and biogeographical models (e.g. Franklin
1995; Bertrand et al. 2012; Dubuis et al. 2013; Wisz et al. 2013) stress the importance of various
groups of ecophysiologically essential predictors (Fig. 1), it seems that a large majority of SDMs
are built without consideration of the ecophysiological relevance and comprehensiveness of the set
of predictors (Pearson & Dawson 2003; Guisan & Thuiller 2005; Austin & Van Niel 2011a). The

most prominent explanation for this incomplete choice of predictors is the unavailability of some 125 data. It seems that largely available variables are frequently used in models (e.g., WorldClim; 126 Hijmans et al. 2005), while the use of less easily available or lacking environmental data is 127 understandably less frequent or absent in SDMs, respectively. This is however a working 128 hypothesis. Making further progress in SDM science therefore requires understanding the primary 129 causes of incomplete use of environmental information. Species distribution models are potentially 130 powerful tools to analyse and predict plant species and community distributions, but their strength, 131 validity and accuracy depend largely on the input data used. Yet, despite a long-standing knowledge 132 of which predictors should theoretically be used, no study has comprehensively reviewed which 133 ecophysiologically meaningful variables are currently used and not used or missing, so that 134 recommendations can be made on where further development is required to obtain all important 135 136 predictors in a spatially explicit form.

137

Here, we evaluate whether the predictors used in correlative plant SDM studies correspond to the 138 known ecophysiological needs of plant species and whether additional constraints, such as biotic 139 factors and disturbances, are included. Simultaneously, we aim to identify which of the 140 141 ecophysiologically relevant variables are missing and whether their omission is due to the unavailability of data in a mapped format or to other causes. We do not either intend to review 142 exhaustively the literature to exemplify good from bad modelling practices, nor to provide examples 143 from our own analyses. We concentrate on niche-based species distribution models of plants 144 (vascular plants and bryophytes) and mainly consider direct abiotic variables - both regulator and 145 resource (sensu Austin 1980) - as well as biotic and disturbance variables. Plants form the basis of 146 primary production and the food chain and, as such, are important for other species, biodiversity 147 and environmental conservation in general. Focusing solely on plants also allows for a more in-148 149 depth review. We acknowledge the importance of other, non-niche processes influencing plant

- distributions, such as dispersal and (evolutionary) history (Soberón & Peterson 2005), but we do not
- examine these processes explicitly here, as we consider them to be outside the scope of this review,
- 152 which centres on environmental niche predictors. Further, although efforts towards incorporating
- the environmental predictors discussed here are also in progress in the field of mechanistic
- modelling (see, e.g., D'Amen et al. in press), this review only considers correlative SDMs.

155 MATERIALS AND METHODS

156

We performed two web searches to extract original articles (excluding reviews, opinions and 157 perspectives) dealing with SDMs of vascular plants and bryophytes. The target of the first search 158 was to record recently published (2010-2015) articles in high-quality ecological journals (see 159 Appendix S2 for the journals used), while the target of the second search was to examine the 160 temporal changes in the variables used in the SDMs. The first search was performed using the query 161 ("species distribution model*" OR "habitat model*" OR "ecological niche model" OR "niche 162 model*" OR "habitat distribution model*" OR "habitat suitability model*" OR "niche-based 163 model*" OR "bioclimatic envelope model*") AND (vegetation OR plant* OR vascular OR 164 bryophyte*) following Guisan et al. (2013) in the ISI Web of Science (WoS), restricting the time 165 range and journals to meet the filters specified above. This search resulted in 745 papers (hereafter 166 called the 'recent search'). The second WoS search used the same search words, but the results were 167 limited to two journals, Journal of Vegetation Science and Journal of Biogeography, after 168 preliminary queries showed the high number of plant SDM studies published in these journals, 169 accounting for the years 2000-2015. The second search was also repeated in other search engines to 170 171 increase the number of articles and to complement missing years, resulting in a total of 171 articles 172 (hereafter called the 'temporal search').

173

For all of the selected articles, we recorded the environmental predictors that were used in the SDMs. To standardize the results, we divided the predictors into eight variable categories, partially following Austin and Van Niel (2011a, see also Appendix S1): temperature, water, substrate (including nutrients but not moisture), radiation, biotic interactions, disturbance (including anthropogenic factors), topography and land use (Table 1, see detailed list of different variables in Appendix S3). The temperature and water categories were further divided into mean, extreme and 180 seasonality variables, and the water category had two additional sub-classes: water balance and soil moisture. The substrate-related category was divided into two classes: bedrock/pH and nutrients. 181 The category of biotic variables accounted for all variables expressing the influence of other 182 biological agents (e.g., cover of vegetation or certain plant species, species richness, and presence 183 or abundance of animal species). The disturbance category accounted for processes that primarily 184 destroy vegetation, such as fire, geomorphological disturbance and human activities, although these 185 processes can also have a positive impact on certain species (e.g., ruderals; Grime 1977). 186 Topographic and land-use related variables do not represent direct or resource variables for plants, 187 but because these are regularly included in SDMs (Franklin 1995) and have an indirect impact on 188 plant distribution through altering the distribution of temperature, moisture, nutrients and light, they 189 were also recorded here (Moeslund et al. 2013). All generally ecophysiologically meaningful 190 191 predictor variables could be assigned to 16 classes (Table 1). Predictors that were meaningful for the target of the original study but not for our review (such as fragmentation and distance to 192 waterbodies) were not recorded but are included in the total number of predictors. 193

194

195 From each selected SDM study, we further recorded the taxonomic group of species of interest and 196 the resolution of the input/environmental data. Only studies that used species distribution data (presence-absence or presence-only) were included in further analyses, i.e. studies on species 197 richness or abundance were not considered. To avoid bias in our analyses due to the tendency to 198 199 highlight the use of climate variables only, we restricted our searches to studies conducted up to a resolution of 1 km² (~30 arc seconds). Studies at coarser resolution (and often larger scale) 200 effectively tend to include only climatic variables due to data availability and the scale-dependence 201 of different predictors (Pearson & Dawson 2003, Thuiller et al. 2004; but see Harwood et al. 2014). 202 From the 745 'recent' articles found in the WoS, 182 met our requirements (that is, they involved 203 actual SDMs concerning plants and had a maximum 1 km² resolution). Hereafter, however, our 204

analyses include 200 studies due to some articles using distinct sets of predictors for different
species or different spatial resolutions. Each of these studies were divided into separate studies. Of
the 'temporal' articles, forty pertained to plants and were conducted at a maximum resolution of 1
km². The resulted dataset was used to examine the number and type of predictors included in the
models. Especially, this was done in order to distinguish which predictors are frequently used in the
SDMs, and on the other hand, which predictors are not used and might require further developing.

211

To account for environmental and spatial coverage, we recorded the continent and biome of origin of the data. The articles included study areas from all continents. Most studies were from Europe (n= 84) and North America (n = 53), with fewer studies from Australia (n = 25), Africa (n = 20), Latin America (n = 15) and Asia (n = 12). All biomes were covered with an expected bias towards European and North American biomes (temperate, boreal, Mediterranean, alpine, arctic) where more studies have been conducted overall.

218

219 **RESULTS**

220

221 In the 'recent' articles, the average number of predictors included in the models was eleven (Fig. 2). The number of predictors considered in the models varied from one to 75. The different classes of 222 variables covered in the models varied from one to thirteen (out of the 16 defined in this study), 223 224 with only two studies covering all eight of our categories (Fig. 2). Several variables under one class and/or category were often simultaneously included as predictors. Variables from the five most 225 essential categories (temperature, water, substrate, radiation, biotic interactions) were included in 226 seven studies, with all of these also including disturbance, topography and/or land-use related 227 variables. Overall, the reviewed studies represent considerable variability in the different variables 228

used. In particular, the 'water balance' and 'biotic' classes included various sets of different typesof factors (see Appendix S3).

231

Most of the 'recent' studies included temperature- and water-related variables (both were included 232 in 88.5 % of studies). Each of the temperature sub-classes appeared in more than half of the SDMs. 233 The most frequently included water-related variables were monthly or annual mean precipitation 234 (68.5 %), with extreme and seasonal precipitation and water balance appearing in approximately 235 one third of the studies (Fig. 3). Approximately one third of the studies included only climatic 236 237 variables (derived from temperature and/or precipitation). Measurements or approximations of actual or potential soil water or soil moisture were incorporated in 15 studies. 238 239 Substrate-related variables were used in ~ 40 % of the studies, and variables directly representing 240 bedrock/pH or nutrients were included in approximately one quarter of the studies. Only 60 studies 241 involved variables representing light. One fifth of the studies included some biotic component as a 242 predictor variable. Variables representing natural disturbances were included in 17 studies. 243 Variables related to human activity were included in 19 studies. 244 245

After climatic variables, topographic factors were most commonly included in the SDMs screened
in this study (44.5 %). Land use was included in 32 studies, with one study using land use as a mask
to exclude certain areas.

249

250 There were no significant differences in the number of variable classes used among the continents

251 (Fig. 4). Only Latin America (LAm) had a significantly lower number of variable categories

compared with the other continents.

The 'temporal search' showed no increase in the number of categories accounted for in the SDMs through time (2000-2015). On the contrary, the number of variables from different categories showed a decreasing trend (Spearman's rank correlation -0.40*; Fig. 5). Exceptions were the SDM studies from 2011 (by Austin and Van Niel (2011b), Meier et al., Mellert et al. and Ohmann et al.), which increased the number of categories included; all studies discussed the importance of selecting variables on an ecological basis or the impacts of omitting meaningful predictors in the models and thus included variables from multiple categories.

- 261
- 262

263 **DISCUSSION**

264

265 Ecological theory, supported by experimental and correlative studies, stresses that multiple environmental factors drive the distribution of species (e.g., Larcher 1975, Fitter & Hay 2002, 266 Schulze et al. 2005, see also e.g., Guisan & Zimmermann 2000; Elith & Leathwick 2009; Franklin 267 2009; Austin & Van Niel 2011a; Bertrand et al. 2012; Dubuis et al. 2013; le Roux et al. 2013a, b), 268 particularly temperature, water, nutrients, light, biotic interactions and disturbances (see Appendix 269 270 S1). In recently published SDM studies, many of these factors were omitted or replaced with rough surrogates (e.g., precipitation for plant available water). Indeed, more than half (53 %) of the plant 271 SDM studies reviewed here based their predictions solely on the categories of temperature and 272 273 water or on those two categories plus one additional variable, thus potentially neglecting several other ecophysiologically relevant aspects (e.g., substrate, radiation and/or biotic interactions. 274 Although it is important to highlight that not all of these categories might be meaningful for all 275 SDMs; see the next paragraph). While data availability is likely a potential reason for the omission 276 of ecophysiologically meaningful predictors, the wide range of variables used in some exemplar 277 278 studies (see next sections and Appendix S3) indicates that some influential and available predictors may tend to be neglected. Furthermore, there was no difference in the number of predictor classes
used in studies from the "data rich" continents (Europe, North America) and the "data poor"
continents (Fig. 4), suggesting that data availability may not be a sufficient explanation for the
absence of important predictors in the models.

283

The intentional use of an ecophysiologically incomplete set of predictors in correlative modelling is 284 acceptable, for instance, if the study deliberately focuses on the climatic niche or climatic range 285 only, provided that this is clearly acknowledged. Therefore, it is important to distinguish here 286 287 between two classes of studies according to their ultimate goal: studies which aim would require including all potentially important variables (e.g. fine-scale predictions for conservation, or 288 addressing aspects of species' ecology in general), and studies which aim does not necessarily 289 require more than one type of predictors (e.g. climate-change studies only interested in fitting 290 species' climatic niches and climatic ranges). Also, in some other cases, a comprehensive set of 291 meaningful predictors may not be essential in SDMs (e.g., when illustrating the development of 292 new methodologies, or if models representing a specific aspect of the niche are explicitly desired; 293 294 Thuiller et al. 2005). Nevertheless, in all type of SDMs, it is important to justify the choice of 295 predictors, and interpret the results in accordance with used predictors. Indeed, only few of the 296 studies reviewed here acknowledged the ecophysiologically incomplete set of environmental drivers used as predictors (e.g., Bertrand et al. 2012; Aguirre-Gutiérrez et al. 2013; Ikeda et al. 2014; 297 298 Riordan & Rundel 2014, Petitpierre et al. in review), and many studies provided no ecological rationale for the choice of predictors. In the next sections, focusing our discussion on SDMs aiming 299 to comprehensively capture species ecological niche, we aim to provide such rationale, discuss 300 ways to account for the needed predictors in SDMs, and identify missing predictors for which 301 302 development and mapping are still needed at a fine scale. However, we do not provide any 303 estimates of an adequate number of predictors, which depends on the number and distribution of

species occurrences and the algorithm or approach used (see e.g., Wisz et al. (2008) and Franklin(2009)).

306

307 **Temperature**

308

Temperature and water-related variables were the most commonly used predictors among the 309 reviewed studies (Fig. 3). While temperature is frequently accounted for in the models and plays an 310 indisputable role in regulating plant species growth and thus, distribution (see Appendix S1), two 311 312 noteworthy issues concerning temperature were identified from our literature analyses. First, there is a large variety of temperature data products available, with the class of temperature variable used 313 having an impact on model performance (Barbet-Massin & Jetz 2014; Slavich et al. 2014). For 314 example, the impact of mean temperature on plants differs from that of extremes or seasonality in 315 both ecological meaning and modelling performance (Zimmermann et al. 2009). In seasonally 316 variable environments especially, annual mean temperature does not represent the growing season 317 or over-wintering conditions, which potentially play a more central role in governing the 318 distribution of plants (Aerts et al. 2006; Paulsen & Körner 2014). One solution to choose between 319 320 different temperature-related variables might be to include multiple variables in a model, as exemplified by many studies using climatic data provided by WorldClim (Hijmans et al. 2005). 321 However, this raises problems of multicollinearity (Graham 2003; Dormann et al. 2013) and 322 conflicts with the objective of parsimony (Mac Nally 2000). Ultimately, the environmental 323 conditions of the study area and the requirements of the species should determine the most suitable 324 temperature-related variable(s) – a viewpoint only rarely considered or tested in the modelling 325 studies. 326

328 Second, while there is a multitude of temperature data readily available for modelling, their resolution and accuracy can be coarse compared with the species data (Dingman et al. 2013; 329 Franklin et al. 2013; Potter et al. 2013; Pradervand et al. 2014). Temperature measurements are 330 typically obtained by interpolating sparse measurements and neglecting the impact of local 331 topography, land cover or water bodies on local temperatures experienced by plants (Scherrer & 332 Körner 2011; Franklin et al. 2013; Aalto et al. 2014; Slavich et al. 2014). Alternatively, improved 333 temperature maps could be obtained by a combination of increased field measurements (e.g., 334 thermal loggers), predictive methods, high-resolution digital elevation models (DEMs) and thermal 335 336 remote sensing rather than spatial interpolations (Scherrer and Körner 2010, Dingman et al. 2013; Pradervand et al. 2014). Thus, while the availability of temperature data is not a primary problem, 337 their usability and ecological significance in SDMs could be improved by increasing their 338 339 resolution and accuracy.

340

341 Water

342

Predictors representing water availability for plants are often derived from precipitation, a class of 343 344 climatic predictors inheriting similar challenges to those discussed for temperature. In addition, precipitation is a poor surrogate for plant available water, especially in high-resolution studies that 345 cover small areas, due to the effects of local topography and soil substrate on the amount and 346 347 distribution of soil moisture (le Roux et al. 2013c; Piedallu et al. 2013). Therefore, while water as a category of predictors is almost always acknowledged in the models, the ecophysiological 348 significance of the water predictors being used might be poor in many cases. Some studies have 349 used water balance (precipitation minus evapotranspiration), which represents a more accurate 350 measure of plant available water compared with precipitation. Some soil moisture indices derived 351 352 from climate data and geographic information systems (GIS) modelling are available (e.g.,

Trabucco & Zomer 2010), but these proxies also neglect the impact of terrain on plant available moisture. Using high-resolution topographic information in combination with climate and soil measurements could provide a more promising basis for modelling high-resolution soil moisture data (Aalto et al. 2013; Pradervand et al. 2014).

357

Ideally, soil moisture measurements taken in the field should most accurately represent the water 358 available to plants. Studies that incorporate field-quantified soil moisture values in their models 359 have improved predictive power, especially at high spatial resolutions (le Roux et al., 2013c). 360 However, collecting these high-resolution and accurate soil moisture data over large areas is rarely 361 feasible. Remote sensing combined with GIS provides ready-to-use (coarse-scale) indices of 362 moisture or wetness (e.g., the surface saturation degree of ASCAT soil wetness indices, see Brocca 363 364 et al. 2010; Lakshmi 2013; Wagner et al. 2013), and other recent developments such as Synthetic Aperture Radars (Elbialy et al. 2014), hyperspectral aerial images (Pottier et al. 2014) and spatial 365 modelling (Aalto et al. 2013) show promise in estimating actual soil moisture at higher resolutions. 366 To conclude, although often accounted for in SDMs with distal predictors, water-related variables 367 could be improved through combined approaches mixing refined field measures, GIS modelling and 368 369 remote sensing.

370

371 Nutrients

372

The role of soil and its nutrients on plant performance is acknowledged by most ecologists (Epstein & Bloom 2005; see also Appendix S1) as well as their role on model performance by many modellers (almost half in our study; see also Coudun et al. 2006; Coudun & Gégout 2007; Bertrand et al. 2012; Dubuis et al. 2013). It seems hardly feasible to obtain high-resolution field

377 measurements of nutrient content and geo-chemical properties of soils across a whole study area.

378 Thus, most studies that included substrate variables used either geological or geomorphological surrogates such as bedrock, pH or landforms, or factors related to soil structure, such as texture or 379 soil depth (Bertrand et al. 2012; Dubuis et al. 2013). This highlights the need for more sophisticated 380 indices of soil nutrient content, analogous to those being developed for soil moisture. The use of 381 soil ecological indicator values (e.g., Ellenberg) also highlights such a need (Coudun et al. 2006). 382 Improved spatial predictors of soil characteristics are thus still required, such as those derived from 383 remote sensing (Parviainen et al. 2013) or potentially from statistical modelling (Lagacherie 1992), 384 to further improve plant SDMs (Dubuis et al. 2013). 385

- 386
- 387 Light

388

389 The importance of light for plants and its use as a predictor in SDMs were previously discussed by Austin and Van Niel (2011a). Solar radiation can be calculated using DEM and, if available, canopy 390 cover in efficient GIS tools (McCune & Keon 2002). However, light-related variables were only 391 included in less than one third of the studies we reviewed, meaning that more than two thirds of the 392 393 reviewed studies neglected an important factor controlling plant distributions, especially at local 394 scales. In the studies accounting for light, it was mostly represented by the sum of (potential) solar radiation over various seasons. In these cases, the radiation variable actually expresses heat rather 395 than photosynthetically active radiation (PAR) and therefore acts similarly to temperature. To 396 397 obtain a real measure of PAR, light must be measured specifically, and the effects of cloud cover and canopy interception must be taken into account (Aguiar et al. 2012; Wang et al. 2014). 398 Nevertheless, inclusion of a solar radiation variable often improves model prediction by adding 399 information on fine-scale energy input, especially in topographically heterogeneous areas (Austin & 400 Van Niel 2011a). At a given elevation, slopes with different aspects can have very different soil and 401 402 vegetation temperatures (Scherrer & Körner 2010; Gunton et al. 2015). In contrast to average

temperatures based mostly on adiabatic lapse rates, solar radiation can include information
regarding aspect, relief shading and daylight period (Kumar et al. 1997; Austin & Van Niel 2011a).
However, as mentioned before, the use of solar radiation as a predictor can lead to misleading
interpretations, as its impact on plants might strongly depend on season, canopy structure and cloud
cover. Thus, the radiation variables should firstly be incorporated into SDMs, seasonal variations
should be accounted for, and the effects of canopy and cloud cover should be included when
studying understory vegetation (Nieto-Lugilde et al. 2015).

410

411 **Biotic interactions**

412

Biotic interactions play a role in altering the potential environmental niche, for example, through 413 competition, facilitation and herbivory (Brooker & Callaghan 1998; Callaway et al. 2002; Araújo & 414 Luoto 2007; Pellissier et al. 2010; Mod et al. 2014). As the importance of biotic interactions and 415 how to measure their importance (Godsoe & Harmon 2012) and account for them in SDMs are still 416 under discussion (Kissling et al. 2012; Wisz et al. 2013), many SDMs do not include biotic factors. 417 Implicitly, these SDMs assume that the important biotic interactions (in a given area or habitat) are 418 419 already indirectly accounted for at the sampling stage (when gathering observations) because biotic interactions influence the realized distribution of the species (McGill et al. 2006) and are thus 420 captured in the realized environmental niche (Araùjo & Guisan 2006). Nonetheless, biotic 421 components were used in approximately one-fifth of the studies, indicating their increasing 422 importance in SDMs. However, explicit information on biological interactions remains difficult to 423 obtain in a spatially explicit form, as the biotic factors governing the assemblage of individual 424 species into communities are still largely unknown (Kissling et al. 2012, Wisz et al. 2013), and 425 associated assembly rules remain to be developped (Guisan & Rahbek 2011). However, surrogates 426 427 such as dominant species cover have been shown to provide some measure of biotic interactions (le

Roux et al. 2014), and incorporating these surrogates has improved both the explanatory and
predictive power of SDMs (Meier et al. 2010; Pellissier et al. 2010). Various methods to account for
biotic interactions in SDMs are presented in Kissling et al. (2012), Wisz et al. (2013) and Pollock et
al. (2014).

432

433 Disturbance

434

The type and necessity of including disturbance variables in models are highly environment-435 specific. Frost-related disturbances can strongly impact vegetation in arctic and alpine areas by 436 destroying some species and subsequently, creating space for other species (le Roux et al. 2013a; le 437 Roux & Luoto 2014). In dryer areas, fire may play such a role (Tucker et al. 2012, but see 438 Crimmins et al. 2013). Disturbance has been incorporated in some models, for example, as the 439 proportion of the area that is disturbed (le Roux et al. 2013a), as an index of geomorphic 440 disturbances (Randin et al. 2009a), or as time elapsed since the last fire (Moretti et al. 2006). The 441 use of predictors related to natural disturbances in SDMs may be particularly important when 442 analysing the potential impacts of changing climate because changes in the intensity of these 443 444 processes associated with climatic shifts may represent key mechanisms by which changes in temperature and rainfall patterns affect vegetation assemblages (le Roux & Luoto 2014, although 445 see Crimmins et al. 2013). Similar to other disturbances, the use of anthropogenic predictors is 446 447 situational, depending on the study environment, species and study target. For semi-natural or urban landscapes and/or species highly associated with humans, the use of anthropogenic predictors might 448 be crucial to obtain reasonable predictions (Kouba et al. 2011; Senan et al. 2012). 449

450

451 **Topography and land use**

453 Variables representing topography are often included in plant distribution models (see also Franklin 1995). Including these variables has been demonstrated to improve plant SDMs (e.g., Sormunen et 454 al. 2011), but interpreting the actual drivers of plant distributions related to these variables can be 455 difficult. Because the effects of topographic variables on plant distributions are distal (i.e., they do 456 not directly impact plants, but they do alter light, moisture, temperature and nutrient conditions; 457 Moeslund et al. 2013), it is not possible to interpret the causal relationships between these variables 458 and the target species (Austin 2007). Correlation between indirect gradients and species distribution 459 results only from location dependence (Austin 2002). Despite the demonstrated ability of 460 topographic variables to improve local models, the use of these indirect variables hampers 461 understanding of proximal species-environment relationships and reduces transferability (Randin et 462 al. 2006). Field quantification of environmental variables or the use of purely proximal variables 463 (sensu Austin 2002) would assist in identifying the actual environmental factors that species 464 respond to and would thus provide more detailed understanding of species distributions and 465 ultimately, yield more realistic SDMs. Therefore, using in-situ measured direct and resource 466 variables instead of indirect gradients (such as elevation, aspect and topographic position) would be 467 advisable (Austin 2002; Pradervand et al. 2014), especially when SDMs are also used to explain 468 469 species distributions. Land use was occasionally included in the models we reviewed. Its inclusion usually improves the explanatory and predictive power of SDMs (Von Holle & Motzkin 2007) but 470 only for predicting species abundances in some cases (Randin et al. 2009b). However, interpreting 471 472 the proximal impact of land-use predictors on plant distributions suffers the same problems discussed for topographic variables (i.e., being often not proximal). 473

474

475 Implications for future studies

477 As hypothesized, limited data availability could be one justification for omitting potentially influential ecophysiologically predictors in SDMs despite their demonstrated advantages for the 478 explanatory and predictive power (e.g., Austin & Van Niel 2011b, Bertrand et al. 2012, le Roux et 479 al. 2014). The other hypothesized explanation was the intended omission, e.g., in studies of climatic 480 niches and ranges (e.g., Thuiller 2005, Petitpierre et al. 2012). However, data unavailability and 481 intended omission can hardly explain all instances (especially in data-rich areas of Europe, North-482 America and Australia, Fig. 4) where important non-climatic factors were excluded (see similar 483 statement made 20 years previously by Franklin 1995). Indeed, many of the studies provided no 484 justification for the choice of predictors or only provided a reference to another study relying on a 485 similar set of predictors without considering the influence of the study area or the ecophysiological 486 requirements of the studied species to determine a meaningful set of predictors. Furthermore, 487 488 despite increasing recognition of the importance of a variety of environmental variables for predicting plant distributions (e.g. Austin & Van Niel 2011a, Dubuis et al. 2013) and the increasing 489 availability of numeric data (including from remote sensing), the number of ecophysiologically 490 significant variable categories considered in SDMs seems rather to have decreased during the 21st 491 century. Therefore, we argue that in the future, an ecologically sound reasoning for the choice of 492 493 predictors in the SDMs should become common practice, and the models and predictions should always be interpreted in perspective of the set of predictors used. 494

495

In addition, our literature review highlighted that some variable classes are poorly represented in terms of data quantity (e.g. global coverage) and quality (e.g. resolution). More attention should be paid to ensure that all relevant environmental predictors are made available for modelling at the scale investigated. Although measuring or deriving proximal predictors over large areas can be difficult for single researchers, large international efforts are increasingly developed to use remote sensing products for such purpose (Zimmermann et al. 2007, Estes et al. 2010). More research 502 should also be dedicated to produce finer-scale and more proximal data to improve our understanding of the factors driving species distributions (Gunton et al. 2015) and therefore, the 503 production of more realistic predictions. Here too, remote sensing and GIS can produce promising 504 data products (Bradley et al. 2012, Pottier et al. 2014, He et al. 2015), and ecologists and ecological 505 modellers should give more attention to collaborative research within the geo-environmental 506 sciences. 507 508 509 **CONCLUSIONS** 510 511 Our study reveals that the rationale, selection and use of environmental predictors in many plant 512 species distribution models do not systematically match established ecophysiological theory, 513 perspectives on ecologically meaningful variable selection or demonstrated improvements in 514 SDMs, and therefore calls for the need to add several meaningful variables in SDMs. Except for the 515 pure climatic niche studies and methodological experiments, many plant SDMs so far have omitted 516 important environmental variables, and the number of predictors representing the essential 517 518 ecophysiological aspects pertaining to plants has not increased during the 21st century, despite increased numerical data availability. In particular, nutrients, actual light, disturbance and biotic 519 interactions should be incorporated more systematically into SDMs, together with the most 520 commonly used temperature and water variables. Furthermore, the type of temperature and water 521 variables to be used should also be given more careful attention. The development of new 522 environmental variables will require improved collaborative research between ecological and geo-523 environmental sciences as well as access to advanced technology, such as remote sensing and GIS 524 modelling approaches. Developing new sets of ecophysiologically more meaningful predictors 525 526 provides the basis for a paradigm change in SDM research.

E	2	7
Э	2	1

529

530

531

549

ACKNOWLEDGEMENTS

23

532	Kordelin foundation. M.L. was partly funded by the Academy of the Finland project nr 286950.
533	A.G. received financial support from the Swiss National Science Foundation (Grant 31003A-
534	152866/1).
535	
536	
537	AUTHOR CONTRIBUTIONS
538	
539	A.G. and M.L. conceived the idea and outline for this manuscript; H.M. and D.S. performed the
540	literature review; H.M. and D.S. led the writing, with A.G. and M.L.
541	
542	
543	
544	
545	
546	REFERENCES
547	
548	Aalto, J., le Roux, P.C. & Luoto, M. 2013. Vegetation mediates soil temperature and moisture in

H.M. received funding from the Research Foundation of the University of Helsinki and the

Aalto, J., le Roux, P.C. & Luoto, M. 2014. The meso-scale drivers of temperature extremes in high-550

arctic-alpine environments. Arctic, Antarctic, and Alpine Research 45: 429-439.

latitude fennoscandia. Climate Dynamics 42: 237-252. 551

552	Aerts, R., Cornelissen, J.H.C. & Dorrepaal, E. 2006. Plant performance in a warmer world: General
553	responses of plants from cold, northern biomes and the importance of winter and spring
554	events. Plant Ecology 182: 65-77.
555	Aguiar, L.J.G., Fischer, G.R., Ladle, R.J., Malhado, A.C.M., Justino, F.B., Aguiar, R.G. & Costa,
556	J.M.N. 2012. Modeling the photosynthetically active radiation in South West Amazonia
557	under all sky conditions. Theoretical and Applied Climatology 108: 631-640.
558	Aguirre-Gutiérrez, J., Carvalheiro, L.G., Polce, C., van Loon, E.E., Raes, N., Reemer, M. &
559	Biesmeijer, J.C. 2013. Fit-for-purpose: Species distribution model performance depends on
560	evaluation criteria – Dutch hoverflies as a case study. PLoS One 8: e63708.
561	Alagador, D., Cerdeira, J.O. & Araújo, M.B. 2014. Shifting protected areas: Scheduling spatial
562	priorities under climate change. Journal of Applied Ecology 51: 703-713.
563	Araujo, M.B. & Guisan, A. 2006. Five (or so) challenges for species distribution modelling. Journal
564	of Biogeography 33: 1677-1688.
565	Araújo, M.B. & Luoto, M. 2007. The importance of biotic interactions for modelling species
566	distributions under climate change. Global Ecology and Biogeography 16: 743-753.
567	Araújo, M.B., Alagador, D., Cabeza, M., Nogués-Bravo, D. & Thuiller, W. 2011. Climate change
568	threatens european conservation areas. Ecology Letters 14: 484-492.
569	Austin, M.P. 1980. Searching for a model for use in vegetation analysis. Vegetatio 42: 11-21.
570	Austin, M.P. 2002. Spatial prediction of species distribution: An interface between ecological
571	theory and statistical modelling. Ecological Modelling 157: 101-118.
572	Austin, M.P. 2007. Species distribution models and ecological theory: A critical assessment and
573	some possible new approaches. <i>Ecological Modelling</i> 200: 1-19.
574	Austin, M.P. & Smith, T.M. 1989. A new model for the continuum concept. Vegetatio 83: 35-47.
575	Austin, M.P. & Van Niel, K.P. 2011a. Improving species distribution models for climate change
576	studies: Variable selection and scale. Journal of Biogeography 38: 1-8.

- Austin, M.P. & Van Niel, K.P. 2011b. Impact of landscape predictors on climate change modelling
 of species distributions: A case study with *Eucalyptus fastigata* in southern New South
 Wales, Australia. *Journal of Biogeography* 38: 9-19.
- 580 Bader, M.K.F., Leuzinger, S., Keel, S.G., Siegwolf, R.T.W., Hagedorn, F., Schleppi, P. & Körner,
- 581 C. 2013. Central European hardwood trees in a high-Co₂ future: Synthesis of an 8-year
 582 forest canopy Co₂ enrichment project. *Journal of Ecology* 101: 1509-1519.
- Barbet-Massin, M. & Jetz, W. 2014. A 40-year, continent-wide, multispecies assessment of relevant
 climate predictors for species distribution modelling. *Diversity and Distributions* 20: 12851295.
- Bertrand, R., Perez, V. & Gegout, J.C. 2012. Disregarding the edaphic dimension in species
 distribution models leads to the omission of crucial spatial information under climate
 change: The case of *Quercus pubescens* in France. *Global Change Biology* 18: 2648-2660.
- Booth, T. H., Nix, H. A., Busby, J. R. & Hutchinson, M. F. 2014. BIOCLIM: the first species
- distribution modelling package, its early applications and relevance to most current
 MAXENT studies. *Diversity and Distributions* 20: 1-9.
- 592 Bradley, B.A., Olsson, A.D., Wang, O., Dickson, B.G., Pelech, L., Sesnie, S.E. & Zachmann, L.J.
- 593 2012. Species detection vs. Habitat suitability: Are we biasing habitat suitability models
 594 with remotely sensed data? *Ecological Modelling* 244: 57-64.
- Brocca, L., Melone, F., Moramarco, T., Wagner, W. & Hasenauer, S. 2010. Ascat soil wetness
 index validation through in situ and modeled soil moisture data in central italy. *Remote Sensing of Environment* 114: 2745-2755.
- Broennimann, O. & Guisan, A. 2008. Predicting current and future biological invasions: Both
 native and invaded ranges matter. *Biology Letters* 4: 585-589.
- Brooker, R.W. & Callaghan, T.V. 1998. The balance between positive and negative plant
- 601 interactions and its relationship to environmental gradients: A model. *Oikos* 81: 196-207.

602	Callaway, R.M., Brooker, R.W., Choler, P., Kikvidze, Z., Lortie, C.J., Michalet, R., Paolini, L.,
603	Pugnaire, F.I., Newingham, B., Aschehoug, E.T., Armas, C., Kikodze, D. & Cook, B.J.
604	2002. Positive interactions among alpine plants increase with stress. Nature 417: 844-848.
605	Coudun, C. & Gégout, JC. 2007. Quantitative prediction of the distribution and abundance of
606	Vaccinium myrtillus with climatic and edaphic factors. Journal of Vegetation Science, 18:
607	517-524.
608	Coudun, C., Gégout, JC., Piedallu, C. & Rameau, JC. 2006. Soil nutritional factors improve
609	models of plant species distribution: An illustration with Acer campestre (l.) in France.
610	Journal of Biogeography 33: 1750-1763.
611	Crimmins, S.M., Dobrowski, S.Z., Mynsberge, A.R. & Safford, H.D. 2013. Can fire atlas data
612	improve species distribution model projections? Ecological applications 24: 1057-1069.

- D'Amen, M., Rahbek, C., Zimmermann, N.E. & Guisan A. in press. Spatial predictions at the
 community level: from current approaches to future frameworks. *Biological Reviews*10.1111/brv.12222.
- Dormann, C. 2007. Promising the future? Global change projections of species distributions. *Basic and Applied Ecology* 8: 387-397.
- Dormann, C.F., Elith, J., Bacher, S., Buchmann, C., Carl, G., Carré, G., Marquéz, J.R.G., Gruber,

B., Lafourcade, B., Leitão, P.J., Münkemüller, T., McClean, C., Osborne, P.E., Reineking,

B., Schröder, B., Skidmore, A.K., Zurell, D. & Lautenbach, S. 2013. Collinearity: A review

- of methods to deal with it and a simulation study evaluating their performance. *Ecography*36: 27-46.
- Dubuis, A., Giovanettina, S., Pellissier, L., Pottier, J., Vittoz, P. & Guisan, A. 2013. Improving the
 prediction of plant species distribution and community composition by adding edaphic to
 topo-climatic variables. *Journal of Vegetation Science* 24: 593-606.

626	Elbialy, S., Mahmoud, A., Pradhan, B. & Buchroithner, M. 2014. Application of spaceborne
627	synthetic aperture radar data for extraction of soil moisture and its use in hydrological
628	modelling at gottleuba catchment, Saxony, Germany. Journal of Flood Risk Management 7:
629	159-175.

- Elith, J. & Leathwick, J.R. 2009. Species distribution models: Ecological explanation and prediction
 across space and time. *Annual Review of Ecology, Evolution, and Systematics* 40: 677-697.
- Epstein, E. & J., B.A. (2005) *Mineral nutrition of plants: principles and perspectives*, 2 edn.
 Sinauer Associates, Sunderland.
- Estes, L. D., P. R. Reillo, Mwangi, A. G., Okin, G. S. & Shugart, H. H. 2010. Remote sensing of
 structural complexity indices for habitat and species distribution modeling. *Remote Sensing of Environment* 114: 792-804.
- 637 Fitter, A.H. & Hay, R.K.M. (2002) Environmental physiology of plants, 3 ed. Academic Press,
- London & San Diego.Franklin, J. 1995. Predictive vegetation mapping: Geographic
 modelling of biospatial patterns in relation to environmental gradients. *Progress in Physical*
- 640 *Geography* 19: 474-499.
- Franklin, J. 2009. *Mapping species distributions: Spatial inference and prediction*. Cambridge
 University Press Cambridge.
- Fukami, T., Martijn Bezemer, T., Mortimer, S.R. & van der Putten, W.H. 2005. Species divergence
 and trait convergence in experimental plant community assembly. *Ecology Letters* 8: 12831290.
- Godsoe, W. & Harmon, L.J. 2012. How do species interactions affect species distribution models? *Ecography* 35: 811–820.
- Graham, M.H. 2003. Confronting multicollinearity in ecological multiple regression. *Ecology* 84:
 2809-2815.

650	Grime, J.P. 1977. Evidence for the existence of three primary strategies in plants and its relevance
651	to ecological and evolutionary theory. The American naturalist 111: 1169-1194.
652	Grinnell, J. 1917. The niche-relationships of the california thrasher. The Auk: 427-433.
653	Guisan, A. & Zimmermann, N.E. 2000. Predictive habitat distribution models in ecology.
654	Ecological Modelling 135: 147-186.
655	Guisan, A. & Thuiller, W. 2005. Predicting species distribution: Offering more than simple habitat
656	models. Ecology Letters 8: 993-1009.
657	Guisan, A. & Rahbek, C. 2011. SESAM - a new framework integrating macroecological and

- species distribution models for predicting spatio-temporal patterns of species assemblages. *Journal of Biogeography* 38: 1433-1444.
- Guisan, A., Lehmann, A., Ferrier, S., Austin, M., Overton, J.M.C., Aspinall, R. & Hastie, T. 2006.
 Making better biogeographical predictions of species' distributions. *Journal of Applied Ecology* 43: 386-392.
- Guisan, A., Tingley, R., Baumgartner, J.B., Naujokaitis-Lewis, I., Sutcliffe, P.R., Tulloch, A.I.T.,
- 664 Regan, T.J., Brotons, L., McDonald-Madden, E., Mantyka-Pringle, C., Martin, T.G.,
- 665 Rhodes, J.R., Maggini, R., Setterfield, S.A., Elith, J., Schwartz, M.W., Wintle, B.A.,
- Broennimann, O., Austin, M., Ferrier, S., Kearney, M.R., Possingham, H.P. & Buckley,
- Y.M. 2013. Predicting species distributions for conservation decisions. *Ecology Letters* 16:
 1424-1435.
- Gunton, R.M., Polce, C. & Kunin, W.E. 2015. Predicting ground temperatures across european
 landscapes. *Methods in Ecology and Evolution* 6: 532–542.
- Harwood, T.D., Mokany, K. & Paini, D.R. 2014. Microclimate is integral to the modeling of plant
 responses to macroclimate. *Proceedings of the National Academy of Sciences of the United States of America*, 111: E1164-E1165.

674	He, K.S., Bradley, B.A., Cord, A.F., Rocchini, D., Tuanmu, M-N., Schmidtlein, S., Turner, W.,
675	Wegmann, M. & Pettorelli, N. 2015. Will remote sensing shape the next generation of
676	species distribution models? Remote Sensing in Ecology and Conservation 1: 4-18.

- Hijmans, R.J., Cameron, S.E., Parra, J.L., Jones, P.G. & Jarvis, A. 2005. Very high resolution
 interpolated climate surfaces for global land areas. *International journal of climatology* 25:
 1965-1978.
- Huston, M.A. (2002) Introductory essay: Critical issues for improving predictions. *Predicting species occurrences: Issues of accuracy and scale* (eds. by J.M. Scott, P.J. Heglund and
- 682 M.L. Morrison), pp. 7-21.
- Hutchinson, G.E. 1957. Concluding remarks. *Cold Spring Harbor Symposia on Quantitative Biology* 22: 415-427.
- Hutchinson, G.E. 1978. *An introduction to population ecology*. Yale University Press, New Haven
 and London.
- Ikeda, D.H., Grady, K.C., Shuster, S.M. & Whitham, T.G. 2014. Incorporating climate change and
 exotic species into forecasts of riparian forest distribution. *Plos One* 9: e107037.
- Inauen, N., Körner, C. & Hiltbrunner, E. 2012. No growth stimulation by CO₂ enrichment in alpine
 glacier forefield plants. *Global Change Biology* 18: 985-999.

Jiménez-Valverde, A., Peterson, A.T., Soberón, J., Overton, J.M., Aragón, P. & Lobo, J.M. 2011.
 Use of niche models in invasive species risk assessments. *Biological Invasions* 13: 2785 2797.

- 694 Kissling, W.D., Dormann, C.F., Groeneveld, J., Hickler, T., Kühn, I., McInerny, G.J., Montoya,
- 595 J.M., Römermann, C., Schiffers, K., Schurr, F.M., Singer, A., Svenning, J.-C.,
- 696Zimmermann, N.E. & O'Hara, R.B. 2012. Towards novel approaches to modelling biotic
- 697 interactions in multispecies assemblages at large spatial extents. *Journal of Biogeography*
- **698 39**: 2163-2178.

699	Kouba, Y., Alados, C.L. & Bueno, C.G. 2011. Effects of abiotic and anthropogenic factors on the
700	spatial distribution of <i>Quercus faginea</i> in the spanish central pyrenees. <i>Plant Ecology</i> 212:
701	999-1007.

Kumar, L., Skidmore, A.K. & Knowles, E. 1997. Modelling topographic variation in solar radiation
 in a gis environment. *International Journal of Geographical Information Science* 11: 475 497.

- Körner, C. 2006. Plant CO₂ responses: An issue of definition, time and resource supply. *New Phytologist* 172: 393-411.
- 707 Körner, C. (2014) Pflanzen im lebensraum. Strasburger lehrbuch der pflanzenwissenschaften (eds.
- by J.W. Kadereit, C. Körner, B. Kost and U. Sonnewald), pp. 759-810. Springer Spektrum,
 Berlin.
- 710 Lagacherie, P. 1992. Formalisation des lois de distribution des sols pour automatiser la
- 711 *cartographie pédologique à partir d'un secteur pris comme référence*. PhD Thesis,
- 712 Université de Montpellier, Montpellier.
- Lakshmi, V. 2013. Remote sensing of soil moisture. *ISRN Soil Science* 33, 1-33.
- Lambers, H., Chapin III, F.S. & Pons, T.L. 2008. *Plant water relations*. Springer, New York, US.
- Larcher, W. 1975. *Physiological plant ecology*. 2nd ed. Springer, London, UK.

le Roux, P.C. & Luoto, M. 2014. Earth surface processes drive the richness, composition and
occurrence of plant species in an arctic–alpine environment. *Journal of Vegetation Science*25: 45-54.

- le Roux, P.C., Virtanen, R. & Luoto, M. 2013a. Geomorphological disturbance is necessary for
 predicting fine-scale species distributions. *Ecography* 36: 800-808.
- le Roux, P.C., Aalto, J. & Luoto, M. 2013c. Soil moisture's underestimated role in climate change
 impact modelling in low-energy systems. *Global Change Biology* 19: 2965-2975.

723	le Roux, P.C., Pellissier, L., Wisz, M.S. & Luoto, M. 2014. Incorporating dominant species as
724	proxies for biotic interactions strengthens plant community models. Journal of Ecology 102:
725	767-775.

- le Roux, P.C., Lenoir, J., Pellissier, L., Wisz, M.S. & Luoto, M. 2013b. Horizontal, but not vertical,
 biotic interactions affect fine-scale plant distribution patterns in a low energy system.
 Ecology 94: 671-682.
- Mac Nally, R. 2000. Regression and model-building in conservation biology, biogeography and
 ecology: The distinction between and reconciliation of 'predictive' and 'explanatory'
 models. *Biodiversity & Conservation* 9: 655-671.
- McCune, B. & Keon, D. 2002. Equations for potential annual direct incident radiation and heat
 load. *Journal of Vegetation Science* 13: 603-606.
- 734 McGill, B.J. 2010. Matters of scale. *Science* 328: 575-576.
- McGill, B.J., Enquist, B.J., Weiher, E. & Westoby, M. 2006. Rebuilding community ecology from
 functional traits. *Trends in Ecology & Evolution* 21: 178-185.
- 737 Meier, E.S., Edwards Jr, T.C., Kienast, F., Dobbertin, M. & Zimmermann, N.E. 2011. Co-
- occurrence patterns of trees along macro-climatic gradients and their potential influence on
- the present and future distribution of *fagus sylvatica* 1. *Journal of Biogeography* 38: 371-
- 740 382.
- Meier, E.S., Kienast, F., Pearman, P.B., Svenning, J.-C., Thuiller, W., Araújo, M.B., Guisan, A. &
 Zimmermann, N.E. 2010. Biotic and abiotic variables show little redundancy in explaining
 tree species distributions. *Ecography* 33: 1038-1048.
- 744 Mellert, K.H., Fensterer, V., Kuechenhoff, H., Reger, B., Koelling, C., Klemmt, H.J. & Ewald, J.
- 745 2011. Hypothesis-driven species distribution models for tree species in the Bavarian Alps.

Journal of Vegetation Science 22: 635-646.

747	Mod, H.K., le Roux, P.C. & Luoto, M. 2014. Outcomes of biotic interactions are dependent on			
748	multiple environmental variables. Journal of Vegetation Science 25: 1024-1032.			
749	Moeslund, J.E., Arge, L., Bøcher, P.K., Dalgaard, T. & Svenning, JC. 2013. Topography as a			
750	driver of local terrestrial vascular plant diversity patterns. Nordic Journal of Botany 31: 129-			
751	144.			
752	Moretti, M., Conedera, M., Moresi, R. & Guisan, A. 2006. Modelling the influence of change in fire			
753	regime on the local distribution of a mediterranean pyrophytic plant species (Cistus			
754	salviifolius) at its northern range limit. Journal of Biogeography 33: 1492-1502.			
755	Mouquet, N., Lagadeuc, Y., Devictor, V., Doyen, L., Duputié, A., Eveillard, D., Faure, D., Garnier,			
756	E., Gimenez, O., Huneman, P., Jabot, F., Jarne, P., Joly, D., Julliard, R., Kéfi, S., Kergoat,			
757	G.J., Lavorel, S., Le Gall, L., Meslin, L., Morand, S., Morin, X., Morlon, H., Pinay, G.,			
758	Pradel, R., Schurr, F.M., Thuiller, W. & Loreau, M. 2015. Predictive ecology in a changing			
759	world. Journal of Applied Ecology 52: 1293-1310.			
760	Nieto-Lugilde D., Lenoir J., Abdulhak S., Aeschimann D., Dullinger S., Gégout JC., Guisan A.,			
761	Pauli H., Renaud J., Theurillat JP. Thuiller, W., Van Es, J., Vittoz, P., Willner, W.,			
762	Wohlgemuth, T., Zimmermann, N.E. & Svenning, JC. 2015. Tree cover at fine and coarse			
763	spatial grains interacts with shade tolerance to shape plant species distributions across the			
764	Alps. <i>Ecography</i> 37: 1-12.			
765	Norby, R.J. & Zak, D.R. 2011. Ecological lessons from free-air Co ₂ enrichment (face) experiments.			
766	Annual Review of Ecology, Evolution, and Systematics 42: 181.			
767	Ohmann, J.L., Gregory, M.J., Henderson, E.B. & Roberts, H.M. 2011. Mapping gradients of			
768	community composition with nearest-neighbour imputation: extending plot data for			
769	landscape analysis. Journal of Vegetation Science 22: 660-676.			

770	Parviainen, M., Zimmermann, N., Heikkinen, R. & Luoto, M. 2013. Using unclassified continuous
771	remote sensing data to improve distribution models of red-listed plant species. Biodiversity
772	and Conservation 22: 1731-1754.
773	Paulsen, J. & Körner, C. 2014. A climate-based model to predict potential treeline position around
774	the globe. Alpine botany 124: 1-12.
775	Pearson, R.G. & Dawson, T.P. 2003. Predicting the impacts of climate change on the distribution of
776	species: Are bioclimate envelope models useful? Global Ecology and Biogeography 12:
777	361-371.
778	Pellissier, L., Bråthen, K.A., Pottier, J., Randin, C.F., Vittoz, P., Dubuis, A., Yoccoz, N.G., Alm, T.,
779	Zimmermann, N.E. & Guisan, A. 2010. Species distribution models reveal apparent
780	competitive and facilitative effects of a dominant species on the distribution of tundra plants.
781	<i>Ecography</i> 33: 1004-1014.
782	Peterson, A.T., Soberón, J., Pearson, R.G., Anderson, R.P., Martínez-Meyer, E., Nakamura, M. &

- Araújo, M.B. 2011. *Ecological niches and geographic distributions*. Princeton University
 Press, Princeton, USA.
- Petitpierre, B., C. Kueffer, Broennimann, O., Randin, C., Daehler, C. & Guisan, A. 2012. Climatic
 niche shifts are rare among terrestrial plant invaders. *Science* 335: 1344-1348.
- Petitpierre, B., Broenniman, O., Kueffer, C., Daehler, C. and Guisan A. in review. Selecting the best
 set of predictors to build transferable species distribution models. *Global Ecology and*
- 789 <u>Biogeography</u>.
- Piedallu, C., Gegout, J.-C., Perez, V. & Lebourgeois, F. 2013. Soil water balance performs better
 than climatic water variables in tree species distribution modelling. *Global Ecology and Biogeography* 22: 470-482.
- 793 Pollock, L.J., Tingley, R., Morris, W.K., Golding, N., O'Hara, R.B., Parris, K.M., Vesk, P.A. &
- McCarthy, M.A. 2014. Understanding co-occurrence by modelling species simultaneously

- with a Joint Species Distribution Model (JSDM). *Methods in Ecology and Evolution* 5: 397406.
- Potter, K.A., Woods, H.A. & Pincebourde, S. 2013. Microclimatic challenges in global change
 biology. *Global Change Biology* 19: 2932-2939.
- Pottier, J., Malenovský, Z., Psomas, A., Homolová, L., Schaepman, M.E., Choler, P., Thuiller, W.,
 Guisan, A. & Zimmermann, N.E. 2014. Modelling plant species distribution in alpine
 grasslands using airborne imaging spectroscopy. *Biology Letters* 10
- Pradervand, J.-N., Dubuis, A., Pellissier, L., Guisan, A. & Randin, C. 2014. Very high resolution
 environmental predictors in species distribution models: Moving beyond topography?
 Progress in Physical Geography 38: 79-96.
- Randin, C.F., Vuissoz, G., Liston, G.E., Vittoz, P. & Guisan, A. 2009a. Introduction of snow and
 geomorphic disturbance variables into predictive models of alpine plant distribution in the
 western swiss alps. *Arctic Antarctic and Alpine Research* 41: 347-361.
- Randin, C.F., Jaccard, H., Vittoz, P., Yoccoz, N.G. & Guisan, A. 2009b. Land use improves spatial
 predictions of mountain plant abundance but not presence-absence. *Journal of Vegetation Science* 20: 996-1008.
- Randin, C.F., Dirnböck, T., Dullinger, S., Zimmermann, N.E., Zappa, M. & Guisan, A. 2006. Are
 niche-based species distribution models transferable in space? *Journal of Biogeography* 33:
 1689-1703.
- Riordan, E.C. & Rundel, P.W. 2014. Land use Compounds habitat losses under projected climate
 change in a threatened california ecosystem. *Plos One* 9: e86487.
- 816 Scherber, C., Eisenhauer, N., Weisser, W.W., Schmid, B., Voigt, W., Fischer, M., Schulze, E.-D.,
- 817 Roscher, C., Weigelt, A., Allan, E., Beszler, H., Bonkowski, M., Buchmann, N., Buscot, F.,
- 818 Clement, L.W., Ebeling, A., Engels, C., Halle, S., Kertscher, I., Klein, A.-M., Koller, R.,
- 819 Konig, S., Kowalski, E., Kummer, V., Kuu, A., Lange, M., Lauterbach, D., Middelhoff, C.,

820	Migunova, V.D., Milcu, A., Muller, R., Partsch, S., Petermann, J.S., Renker, C., Rottstock,
821	T., Sabais, A., Scheu, S., Schumacher, J., Temperton, V.M. & Tscharntke, T. 2010. Bottom-
822	up effects of plant diversity on multitrophic interactions in a biodiversity experiment. Nature
823	468: 553-556.
824	Scherrer, D. & Körner, C. 2010. Infra-red thermometry of alpine landscapes challenges climatic
825	warming projections. Global Change Biology 16: 2602-2613.
826	Scherrer, D. & Körner, C. (2011). Topographically controlled thermal-habitat differentiation buffers
827	alpine plant diversity against climate warming. Journal of Biogeography 38: 406-416.
828	Schulze, E., Beck, E. & Muller-Hohenstein, K. 2005. Plant Ecology, 1st ed. Springer-Verlag,
829	Berlin.Senan, A.S., Tomasetto, F., Farcomeni, A., Somashekar, R.K. & Attorre, F. 2012.
830	Determinants of plant species invasions in an arid island: Evidence from Socotra island
831	(Yemen). Plant Ecology 213: 1381-1392.
832	Slavich, E., Warton, D.I., Ashcroft, M.B., Gollan, J.R. & Ramp, D. 2014. Topoclimate versus
833	macroclimate: how does climate mapping methodology affect species distribution models
834	and climate change projections? Diversity and Distributions, 20: 952-963.
835	Soberón, J. & Peterson, A.T. 2005. Interpretation of models of fundamental ecological niches and
836	species' distributional areas. Biodiversity Informatics 2: 1-10.
837	Sormunen, H., Virtanen, R. & Luoto, M. 2011. Inclusion of local environmental conditions alters
838	high-latitude vegetation change predictions based on bioclimatic models. <i>Polar Biology</i> 34:
839	883-897.
840	Thibaud, E., Petitpierre, B., Broennimann, O., Davison, A.C. & Guisan, A. 2014. Measuring the
841	relative effect of factors affecting species distribution model predictions. Methods in
842	Ecology and Evolution 5: 947-955.
843	Thuiller, W., Araújo, M.B. & Lavorel, S. 2004. Do we need land-cover data to model species
844	distributions in Europe? Journal of Biogeography 31: 353-361.

845	Thuiller, W., Richardson, D.M., Pysek, P., Midgley, G.F., Hughes, G.O. & Rouget, M. 2005.
846	Niche-based modelling as a tool for predicting the risk of alien plant invasions at a global
847	scale. Global Change Biology 11: 2234-2250.

- 848 Trabucco, A. & Zomer, R.J. 2010. Global Soil Water Balance Geospatial Database. CGIAR
- 849 Consortium for Spatial Information. Published online, available from the CGIAR-CSI
 850 GeoPortal at: http://www.cgiar-csi.org
- Tucker, C.M., Rebelo, A.G. & Manne, L.L. 2012. Contribution of disturbance to distribution and
 abundance in a fire-adapted system. *Ecography* 35: 348-355.
- 853 Wagner, W., Hahn, S., Kidd, R., Melzer, T., Bartalis, Z., Hasenauer, S., Figa-Saldaña, J., de
- 854 Rosnay, P., Jann, A., Schneider, S., Komma, J., Kubu, G., Brugger, K., Aubrecht, C., Züger,
- J., Gangkofner, U., Kienberger, S., Brocca, L., Wang, Y., Blöschl, G., Eitzinger, J.,
- Steinnocher, K., Zeil, P. & Rubel, F. 2013. The ascat soil moisture product: A review of its
 specifications, validation results, and emerging applications. *Meteorologische Zeitschrift* 22:
 5-33.
- Wang, L., Gong, W., Ma, Y., Hu, B. & Zhang, M. 2014. Photosynthetically active radiation and its
 relationship with global solar radiation in Central China. *International Journal of Biometeorology* 58: 1265-77.
- Wisz, M.S., Hijmans, R. Li, J., Peterson, A.T., Graham, C.H. & Guisan, A. 2008. Effects of sample
 size on the performance of species distribution models. *Diversity and Distributions* 14: 763773.
- Wisz, M.S., Pottier, J., Kissling, W.D., Pellissier, L., Lenoir, J., Damgaard, C.F., Dormann, C.F.,
- 866 Forchhammer, M.C., Grytnes, J.A., Guisan, A., Heikkinen, R.K., Hoye, T.T., Kuhn, I.,
- 867 Luoto, M., Maiorano, L., Nilsson, M.C., Normand, S., Ockinger, E., Schmidt, N.M.,
- Termansen, M., Timmermann, A., Wardle, D.A., Aastrup, P. & Svenning, J.C. 2013. The

869	role of biotic interactions in shaping distributions and realised assemblages of species:					
870	Implications for species distribution modelling. Biological Reviews 88: 15-30.					
871	Von Holle, B. & Motzkin, G. 2007. Historical land use and environmental determinants of					
872	nonnative plant distribution in coastal southern new england. Biological Conservation 136:					
873	33-43.					
874	Zimmermann, N. E., Edwards, T. C. Moisen, G. G., Frescino, T. S. &Blackard, J. A. 2007. Remote					
875	sensing-based predictors improve distribution models of rare, early successional and					
876	broadleaf tree species in Utah. The Journal of Applied Ecology 44: 1057-					
877	1067.Zimmermann, N.E., Edwards, T.C., Graham, C.H., Pearman, P.B. & Svenning, J.C.					
878	2010. New trends in species distribution modelling. <i>Ecography</i> 33: 985-989.					
879	Zimmermann, N.E., Yoccoz, N.G., Edwards, T.C., Meier, E.S., Thuiller, W., Guisan, A., Schmatz,					
880	D.R. & Pearman, P.B. 2009. Climatic extremes improve predictions of spatial patterns of					
881	tree species. Proceedings of the National Academy of Sciences of the United States of					
882	America 106 Suppl 2: 19723-19728.					
883						
884						
885						
886	Supporting Information					
887	Online Supporting Information may be found in the online version of this article:					
888	Appendix S1 Ecophysiological meaning of different categories of variables for plant species					
889	Appendix S2 Journals and numbers of studies included in the paper.					
890	Appendix S3 Variables included in the different classes and categories					

TABLES

Table 1. Classification of predictors into eight categories and 16 classes (see Appendix 3 for details

of the variables). The five first columns represent the most important categories, which we refer to

as 'the five most essential categories' in the text.

			540001400	Raulation	inter- actions	Distui bance	graphy	use
	mean (annual, seasonal, monthly) temperature	mean / summed (annual, seasonal, monthly) precipitation	pH, bedrock	radiation, clouds	variables related to other organisms	geomorpho- logical processes, fire	slope, aspect, elevation,	land-use classes
Classes	extreme temperatures	extreme precipitation	nutrients			anthropo- genic variables		
	seasonality	seasonality						
		water balance						
		soil moisture						





904

Fig. 1. Example of a conceptual framework of relationships between resources, direct and indirectenvironmental gradients and their influence on the growth, performance, and geographical

907 distribution of vascular plants and vegetation. ET = Evapotranspiration, P.A.R = Photosynthetically

active radiation. Adapted from Guisan & Zimmermann 2000.



Fig. 3. Proportion of studies in which each predictor class was used: 1a mean temperature; 1b
extreme temperature; 1c seasonality of temperature; 2a mean precipitation; 2b extreme
precipitation; 2c seasonality of precipitation; 2d water balance; 2e soil moisture; 3a pH/bedrock; 3b
nutrients; 4 radiation; 5 biotic interactions; 6a natural disturbances; 6b human disturbances; 7 land
use; 8 topography.



Fig. 2. Frequency of the number of variables, classes (16) and categories (see Table 1) accounted
for in the plant species distribution modelling studies. One outlier value (75) was removed from the
histogram representing the number of variables in the SDMs.



Fig. 4. The number of categories and classes accounted for in the plant species distribution models
(SDMs) using data from different continents. The boxes represent the median and the 25/75
percentile, and the whiskers are 2 SD. The mean is indicated by a black square, and significant
differences are marked with an asterisk.



Fig. 5. Number of variable categories (as presented in Table 1) used in the SDM studies published
in two journals from 2000-2015. Spearman's rank correlation between the years and categories
included is -0.40*. Black squares indicate the mean values of all studies published within a year,
and the grey dots indicate individual studies.

Supporting information to the paper

Heidi K. Mod, Daniel Scherrer, Miska Luoto & Antoine Guisan. What we use is not what we know: environmental predictors in plant distribution models. *Journal of Vegetation Science*.

Appendix S1. Ecophysiologically relevant variables for plant distribution

Seven environmental factors are generally considered as essential for plant growth and survival: light, water, temperature, nutrients, biotic interactions, disturbance and CO₂ (Guisan & Zimmermann, 2000, Kadereit *et al.*, 2014). All these factors can have direct and indirect effects on plants and in combination with dispersal and historical factors, they define the abundance and distribution of plant species (Soberon & Peterson, 2005).

Temperature is the most common regulatory factor considered in SDM's. Temperature directly effects the speed of growth and in case of strong seasonality defines the growing season length. Additionally, minimum and maximum temperatures can reflect physiological thresholds for plants by frost or heat resistance.

Water has several essential functions in plants including photosynthesis, cooling by transpiration and maintaining turgor. In SDMs "water" is usually reflected by either precipitation alone or in combination with evapotranspiration (e.g. water balance). These environmental variables are considered a proxy for plant available water. However, this might not be the case if soils and topography are heterogeneous, as plant available water is strongly influence by both soil type and topographic position. The seasonality of available water/precipitation might lead to temporal flooding, drought or snow cover and thus requires special adaptations by the present plant species.

Nutrients are taken up with water by roots (often with the help of mycorrhiza). Many micronutrients are essential for plant survival including potassium, calcium, magnesium, sulphur, boron, chlorine, manganese, molybdenum and zinc but most significant for productivity are usually the contents of nitrogen and phosphorus. Nutrients in a wider sense can also influence the pH of the soils, whereas bedrock together with living organism are the primary regulators of available nutrients in soils. Therefore, while deriving nutrient content of the soils might not be effective, bedrock, soil pH and soil texture are often used as surrogates in the SDMs.

Light is often expressed as global radiation and therefore energy (W/m2) driving temperature (air, leaf, and soil) and evapotranspiration. However, for plants light reflects also photo active radiation (PAR) and is thereby directly related to photosynthesis. While radiation can be easily modelled and is relatively independent of the vegetation, PAR is strongly affected by the canopy structure of the vegetation. Therefore, the available light for photosynthesis might be very different in a forest compared to open grassland at otherwise similar global radiation (energy). Additionally, light might contain important signals for plant development (e.g. germination and photoperiodism).

Biotic interactions act among and between species, and have both positive and negative impact by prohibiting or ameliorating growth. Impact of other species can be direct (e.g. competition, herbivory) or indirect (e.g. ameliorating harsh microclimatic conditions, shading, nutrient addition by manure). Biotic interactions have been included to the SDMs as e.g. presence or cover of dominant species, remote sensed vegetation index or interaction matrices for multispecies co-occurrence datasets.

Disturbance's impact is mainly negative for species as soil, water, air or snow movement, fire or anthropogenic activities destroy vegetation. However, some ruderal species benefit from disturbances indirectly as they decrease competition and create space by destroying dominant species, and some specialist species require disturbances, as fire and water-logging for germination. Disturbances have also secondary impact on vegetation, by indirectly impacting soil properties: e.g. cryoturbation bring nutrients closer to soil surface.

CO2 the carbon source for plants and therefore essential for their survival and productivity. However, the levels of CO2 among sites don't vary enough to be limiting or having a significant influence on species composition and therefore are ignored in correlative models such as SDM's.

Topography and land use do not have a direct impact on plants, but they affect the distribution of ecophysiolosically meaningful factors (e.g. temperature, light). Topography and land use related variables are easily available and incorporating them often improve SDMs.

References:

Guisan, A. & Zimmermann, N.E. 2000. Predictive habitat distribution models in ecology. *Ecological Modelling* 135: 147-186.

Kadereit, J.W., Körner C., Kost B., Sonnewald U. 2014. Strasburger - Lehrbuch der Pflanzenwissenschaften, Berlin, Springer

Spektrum.

Soberón, J. & Peterson, A.T. 2005. Interpretation of models of fundamental ecological niches and species' distributional areas.

Biodiversity Informatics 2: 1-10.

Appendix S2. Journals and number of studies included by the query (and subsequent analyses).

Recent search:

Ambio 3 (1)	Global Change Biology 58 (16)
American Naturalist 4 (1)	Global Ecology and Biogeography 45 (11)
Annals of Botany 9 (2)	Journal of Applied Ecology 13 (1)
Applied Vegetation Science 9 (5)	Journal of Biogeography 59 (17)
Biodiversity and Conservation 22 (5)	Journal of Ecology 14 (2)
Biological Conservation 49 (10)	Journal of Vegetation Science 24 (8)
Biology Letters 2 (2)	Landscape Ecology 12 (1)
Climatic Change 7 (3)	Methods in Ecology and Evolution 15 (1)
Conservation Biology 9 (1)	Nature Communications 1 (0)
Conservation Letters 1 (0)	New Phytologist 5 (2)
Diversity and Distribution 62 (19)	Oecologia 1 (0)
Ecography 52 (21)	Oikos 5 (1)
Ecological Applications 20 (5)	Perspectives in Plant Ecology 7 (0)
Ecological Modelling 50 (20)	Plant Ecology 8 (7)
Ecological Monographs 3 (3)	Plos One 113 (29)
Ecology 9 (2)	Proceedings of National Academy of Sciences 10 (1)
Ecology Letters 11 (1)	Proceedings of Royal Society B 14 (2)
Ecosystems 1 (0)	Science 2 (0)
Functional Ecology 3 (0)	Trends in Ecology and Evolution 1 (0)

Temporal search

Journal of Vegetation Science 39 (12) Journal of Biogeography 122 (28) Appendix S3. Variables included in different classes and categories.

TEMPERATURE

mean temperature

- (annual / monthly) mean temperature (of coldest / warmest / driest / wettest quarter / summer / winter)
- soil temperature
- warmth index (the annual sum of positive differences between monthly mean temperatures and e.g. 5 degrees, i.e. a measure of the effective warmth for plants)

extreme temperature

- (annual) min / max temperature (of coldest / warmest driest / wettest quarter / month / season)
- mean temperature of coldest / warmest / driest / wettest month
- mean daily max / min temperature (for DJF / MAM / JJA / SON)

temperature seasonality

- seasonality, annual / diurnal range
- growing degree days (all thresholds) / freezing degree days (FDD) (soil / air) / non-FDD / chilling degree days
- isothermality
- heat units (annual sum of daily temperatures exceeding X degrees)
- frost duration
- winter / summer cold / heat wave duration

WATER

mean precipitation

- (annual / monthly) mean / summed precipitation (of coldest / warmest / driest / wettest quarter / season)
- days with rain > 1 mm
- rainfall intensity

extreme precipitation

- mean / summed / min / max precipitation of coldest / warmest / driest / wettest month
- highest 5-day precipitation

precipitation seasonality

- seasonality, annual range
- snow (cover duration, annual snowfall)

- dry / wet season /day length / intensity / frequency
- % of annual precipitation falling during the growing season
- average flood duration
- the standard deviation of hydrographs

water-balance

- (annual / seasonal / monthly) water balance
- (annual / seasonal) evapo-transipiration, vapour pressure
- (mean / annual / seasonal / soil) water / moisture deficit / surplus / availability /stress
- (annual / seasonal / plant available) water/ wetness / moisture / aridity index
- water content
- flow accumulation
- average water level
- soil moisture (days; days when soil moisture air temperature ratio is favourable for plant growth)
- waterlogging index

soil water capacity

- soil water capacity, measured soil moisture
- soil drainage class
- hydraulic soil presence class

SUBSTRATE

bedrock / ph

- bedrock, lithology, rock type
- pH
- surface geology, geological substrate

nutrients

- nutrients, fertility, Cation-exchange capacity, calcareous
- soil material / depth / order / quality / texture / type
- organic matter, loaminess, alluvial, clay / silt / sand content, salt, gypsum
- soil grain size, bulk density
- FAO soil group
- remote sensed Normalized difference soil index, soil production index

• water regime (ordered classes from dry to waterlogged)

LIGHT

- solar radiation (daily, annual, seasonal)
- most / least radiated quarter
- mean hours of sunshine
- clouds

BIOTIC

- NDVI, Landsat bands, Enhanced Vegetation Indices, remote sensed vegetation (indices / classes)
- vegetation height / density / volume/ cover
- canopy / forest / tree cover
- productivity, Net Primary Production
- ecological classification, succession time
- pollinators
- litter
- distance to moorland, moorland presence / absence
- stand basal area
- % of sparsely / dense vegetated brownfield
- % of brownfield with low / high vegetation

DISTURBANCE

natural

- fire, volcanic ash
- geomorphological disturbance
- trampling, grazing
- % area of disturbed terrain

anthropogenic

- population / settlement / building density
- distance to urban areas / roads / harbour / roads
- agriculture, afforestation, soil drainage, roads, human perturbation, forest / etc. management

- human footprint, anthropization degree
- brick rubble
- ownership status (measure of land management)
- predominance of exotic species

TOPOGRAPHY

- altitude (range), terrain curvature, topographic position, slope, flatness, meso-topography, % of steep topography, slope type
- aspect, eastness, northness
- rockiness, ruggedness, topographic wetness index,
- topographic diversity

LAND-USE

- Corine, land-use classes (if only "biotic" land-use -> 'biotic' class)
- distance to potential forest, age of forest