# **Movement of feeder-using songbirds: the influence**

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# of urban features

3 Daniel T. C. Cox<sup>1</sup>\*, Richard Inger<sup>1</sup>, Steven Hancock<sup>2</sup>, Karen Anderson<sup>1</sup> & Kevin J. Gaston<sup>1</sup>

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<sup>1</sup>Environment & Sustainability Institute, University of Exeter, Penryn, Cornwall TR10 9EZ,

6 U.K.

- <sup>2</sup>Global Ecology Lab, University of Maryland, Maryland, MD20742, U.S.
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- 9 \*Corresponding author. E-mail: <u>dan.t.cox@googlemail.com</u>

10 Private gardens provide vital opportunities for people to interact with nature. The most 11 popular form of interaction is through garden bird feeding. Understanding how landscape 12 features and seasons determine patterns of movement of feeder-using songbirds is key to 13 maximising the well-being benefits they provide. To determine these patterns we established 14 three networks of automated data loggers along a gradient of greenspace fragmentation. Over 15 a 12-month period we tracked 452 tagged blue tits *Cyantistes caeruleus* and great tits *Parus* 16 *major* moving between feeder pairs 9,848 times, to address two questions: (i) Do urban 17 features within different forms, and season, influence structural (presence-absence of 18 connections between feeders by birds) and functional (frequency of these connections) 19 connectivity? (ii) Are there general patterns of structural and functional connectivity across 20 forms? Vegetation cover increased connectivity in all three networks, whereas the presence 21 of road gaps negatively affected functional but not structural connectivity. Across networks 22 structural connectivity was lowest in the summer when birds maintain breeding territories, 23 however patterns of functional connectivity appeared to vary with habitat fragmentation. 24 Using empirical data this study shows how key urban features and season influence 25 movement of feeder-using songbirds, and we provide evidence that this is related to 26 greenspace fragmentation.

### 28 Introduction

29 As urbanization increases globally, greenspaces in cities and towns are becoming of greater importance for the provision of ecosystem services<sup>1,2</sup>. Domestic gardens are a major 30 component of these green spaces<sup>3-5</sup>. They constitute easily accessible and immediate 31 32 locations where people can interact with nature, enabling access to the broad range of health and well-being benefits that nature provides<sup>6, 7</sup>. Birds are a key component of garden wildlife<sup>8</sup> 33 34 and for many people their interactions with wild birds may form the main daily wildlife 35 interaction<sup>9</sup>. Watching birds and listening to their song have been shown positively to influence human psychological well-being<sup>10-15</sup>. Given these benefits, it is perhaps 36 37 unsurprising that the provision of supplementary food is the most popular form of wildlife gardening (reviewed by <sup>11</sup>). 38 39 40 Domestic green spaces are often characterised by numerous small and densely packed gardens that are utilised and managed by an equivalent number of households<sup>4,5</sup>. This results 41 42 in individual birds typically moving between multiple gardens to forage and visit feeders, 43 where they will be seen by, and so provide benefits to, multiple people. The ability of birds to 44 move between gardens thus increases the potential benefits that any individual bird can 45 provide, with actual movement being determined by the structures of the gardens themselves,

46 including their geographical location in relation to one another, the habitat for birds within

47 the gardens, and the surrounding urban features.

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Previous studies in urban areas have estimated connectivity for birds within and between
public green spaces<sup>16-20</sup>. These studies suggest that vegetation between green spaces
preserves connectivity, while multiple barriers, such as roads and rivers, cumulatively
decrease landscape permeability. However, despite the clear importance of domestic gardens

53 in generating connectivity in their own right and for facilitating connectivity between larger green spaces<sup>1,4</sup>, fragmented land ownership and management mean that their role in shaping 54 connectivity is largely unexplored empirically $^{21,22}$ . Indeed, how structural patterns of key 55 56 features that distinguish different urban forms affect the flow of birds around the landscape is 57 currently unknown. In the wider landscape there is seasonal variation in connectivity that is related to habitat quality and quantity $^{23,24}$ , therefore we might expect this also to occur across 58 59 domestic gardens and to vary across different urban forms. For example, birds defend smaller 60 home ranges when breeding in summer than when foraging more widely in winter.

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62 Birds that utilize feeders provide an ideal group for exploring the relationships among urban 63 form, connectivity and cultural service delivery. Radio Frequency Identification (RFID) 64 technology provides a means of doing so. This can produce a continuous record of the time 65 and date of when an individual bird carrying a Passive Integrated Transponder (PIT) tag 66 visits a resource patch. Networks of RFID readers can thus be used to record individual 67 movement in time and space as birds visit different feeders. This allows the influence of 68 different urban features on structural and functional connectivity to be determined, structural 69 connectivity here being the presence-absence of connections made by birds moving between 70 feeders and functional connectivity the frequency of those connections. We set up three 71 networks of custom-designed low-powered RFID equipped bird feeders within domestic 72 gardens in the Cranfield triangle in Southern England, UK, with each network within a 73 different urban form that is common in Europe; these had, respectively, low, medium or high 74 green space fragmentation. We used hyperspectral and LiDAR data to characterise the 75 landscape structure through which tagged birds were likely to move between feeder pairs 76 within each network. There were 17 feeders per network, and these were operated 77 continuously over a 12-month period to explore two general questions:

- (i) How do different features within each urban form, and season, influence general
  patterns of structural and functional connectivity for birds?
- 80 (ii) Are there general patterns of structural and functional connectivity across these81 forms?
- 82

## 83 **Results**

84 In total we tagged 452 individuals of two common species of feeder-using birds (blue tit 85 Cyanistes caeruleus and great tit Parus major) between June 2013 and August 2014 (see 86 Supplementary Table S1 and Fig. S1). We divided the year into four equal seasons: summer, 87 autumn, winter, spring. We then considered that a tagged bird visiting first one and then a 88 different feeder within each network and within each season made a connection, with data 89 collection starting on the 1<sup>st</sup> September 2013. Across the three networks, 51% (±2 SD) of 90 tagged individuals made one or more connections between feeders (n = 9,848; Fig. 1). 91 Eighty-eight percent of connections occurred within two days (n = 8,652; See Supplementary 92 Fig. S2). We discarded from the analysis connections that took longer to make because we 93 considered there was a high probability that birds travelled to the second feeder via a non-94 direct route. Using hyperspectral and LiDAR data we categorised the habitat in an ellipsoid 95 between feeder pairs to establish variation in urban form across the three networks of RFID 96 readers (Table 1). For each feeder pair we calculated the distance between feeders, the 97 shortest distance between feeders and a bird catching site, and finally within each ellipsoid 98 we calculated vegetation cover and the number of road gaps (Table 1; Fig. 1). 99 100 URBAN FEATURES WITHIN FORMS AND SEASON

101 The first stage of our analysis tested for the effect of different urban features and season on

102 structural connectivity (the presence or absence of a connection between feeder pairs in any

season) and functional connectivity (the frequency of these movements between feeder pairsin any season) within each of the networks.

106	For structural connectivity, in each network the likelihood of a connection being present
107	between feeders (i.e. connectivity) increased with the percentage vegetation cover (Table 2a;
108	Fig. 2a), while the presence of one or more road gaps did not affect movement (Table 2a; Fig.
109	2b). In the networks of low and medium fragmentation, connectivity decreased with distance
110	between feeders. In the network of low fragmentation, connectivity was lowest in summer
111	and highest in the autumn and winter (Table 2a; Fig. 2c). In the network of medium
112	fragmentation, connectivity was higher across the year relative to summer (Table 2a; Fig. 2c).
113	In the network of high fragmentation, connectivity was highest in spring relative to the other
114	seasons (Table 2; Fig. 2c). Great tits moved between fewer feeder pairs than blue tits in the
115	medium and high fragmentation networks, while there was decreased movement with
116	increasing distance from the ringing site in the network of medium fragmentation (Table 2).
117	
118	For functional connectivity, vegetation cover increased the frequency of movement across all
119	networks (Table 2b). In the network of low and medium fragmentation the frequency of
120	connections decreased with increasing distance between feeders, while decreasing in all
121	networks in the presence of road gaps (Table 2b; Fig. 2d). There was seasonal variation in
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	connectivity relative to summer that varied across networks; in the network of low
123	connectivity relative to summer that varied across networks; in the network of low fragmentation connectivity was higher across the year relative to summer, while in the
123 124	connectivity relative to summer that varied across networks; in the network of low fragmentation connectivity was higher across the year relative to summer, while in the network of medium fragmentation connectivity was lowest in autumn and winter (Table 2b;
123 124 125	connectivity relative to summer that varied across networks; in the network of low fragmentation connectivity was higher across the year relative to summer, while in the network of medium fragmentation connectivity was lowest in autumn and winter (Table 2b; Fig. 2e). In the network of low fragmentation connectivity was lowest in autumn and highest
123 124 125 126	connectivity relative to summer that varied across networks; in the network of low fragmentation connectivity was higher across the year relative to summer, while in the network of medium fragmentation connectivity was lowest in autumn and winter (Table 2b; Fig. 2e). In the network of low fragmentation connectivity was lowest in autumn and highest in spring (Table 2b; Fig. 2e). Movement decreased with distance to ringing site in the

### 129 PATTERNS OF MOVEMENT ACROSS URBAN FORMS

130 The second stage of our analysis explored general patterns of structural and functional 131 connectivity across the three networks. We found that structural connectivity decreased 132 significantly across the three networks with increasing green space fragmentation (low 133 fragmentation, 77%; medium fragmentation, 68%; high fragmentation, 55% of feeder pairs had connections; ANOVA of quasi-binomial model, network  $X^2 = 20.4$ , df = 2, P = <0.0001; 134  $R^2 = 0.15$ ; Table 3a; Fig. 3a). Functional connectivity was greatest in the network of low 135 136 fragmentation, whilst there was no difference between the networks of medium and high fragmentation (ANOVA of quasi-Poisson model, network  $X^2 = 1708.0$ , df = 2, P = <0.0001; 137  $R^2 = 0.39$ ; Table 3b; Fig. 3b). In both models connectivity decreased with distance between 138 139 feeder pairs (Table 3).

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## 141 **Discussion**

142 Understanding how the spatial and temporal heterogeneity of urban green spaces determines 143 patterns of connectivity is critical for manipulating the flow of ecosystem services around 144 where people live and work. This is the first study that uses empirical data to model both 145 structural (a measure of the ability of birds to move through the landscape) and functional (a 146 measure of the frequency of movement of individuals) connectivity. Due to the labour and 147 time intensive nature of a study of this kind it is not possible within realistic budgets to test 148 variation in movement across large numbers of networks (e.g. replicating different urban 149 forms or across a gradient of forms). However, we show that key urban features and season 150 influenced movement within three distinct urban forms, and (recognising the limitations of a 151 three-site comparison) there is evidence that overall levels of connectivity varied across 152 forms with increased movement being associated with reduced green space fragmentation.

154 In the network of low fragmentation movement focused around a large central cluster with 155 birds using green corridors to move to, and between, feeders away from ringing sites, and 156 movement then decreased as green spaces became more fragmented (Fig. 1a). This was 157 supported in our models with both forms of connectivity increasing with vegetation cover, 158 and decreasing with distance between feeder pairs. Functional but not structural connectivity 159 was negatively correlated with the presence of road gaps, suggesting that birds do travel 160 between green fragments but roads cause resistance to frequent movement. Connectivity 161 varied by season, being lowest in the summer probably as a consequence of breeding season 162 territoriality. Structural connectivity was then greatest in autumn, at a time when birds are 163 engaged in natal and post-breeding dispersal, before decreasing in winter as garden songbirds become more settled in their wintering territories<sup>25</sup>. Relative to the summer, functional 164 165 connectivity increased during the year, peaking in winter when food supplies were 166 constrained and birds moved frequently between known feeders. The increased availability of 167 vegetation and green corridors in this network may allow ecological processes most closely to mimic those we expect to see in more natural habitats<sup>25,26</sup>. 168 169 170 In the network of medium fragmentation movement mostly occurred within the central 171 woodland area and along vegetation corridors that largely originated from the wooded area, 172 while there was little movement between feeders within suburban gardens (Fig. 1b). 173 Connectivity increased with vegetation cover, but decreased with distance from ringing site 174 indicating that suburban gardens reduced landscape permeability. Functional connectivity but 175 not structural connectivity decreased in the presence of road gaps supporting this conclusion. 176 Structural connectivity decreased in summer when birds remain in their breeding territories, 177 but was higher across the rest of the year possibly because birds were more likely to explore

178 into gardens. Functional connectivity decreased in autumn and winter relative to spring and 179 summer, suggesting that birds visited feeders opportunistically as they passed through a garden as opposed to having established wintering territories<sup>27</sup>. Great tits showed lower 180 181 structural connectivity relative to blue tits in the networks of medium and high fragmentation, 182 possibly because they were less abundant around these networks (see Supplementary Table 183 S1). Great tits tend to forage in mature trees and so may have been more reluctant to move to 184 feeders away from vegetated corridors possibly as a consequence of relatively high blue tit 185 populations<sup>28</sup>.

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187 In the network of high fragmentation, green space corridors between streets with terraced 188 housing appeared to funnel movement (Fig. 1c). This could in part be a consequence of the 189 ringing locations being located at the periphery of the network, however, there was very little 190 movement into the relatively vegetation-impoverished central green islands located between 191 the two ringing locations suggesting that crossing terraced streets causes resistance to 192 movement relative to moving along green spaces. In this highly fragmented urban form 193 structural and functional connectivity were positively correlated with vegetation but 194 unaffected by distance between feeder pairs (at the scale of the study). Again, functional but 195 not structural connectivity was negatively correlated with the presence of road gaps. The 196 presence and frequency of connections were greatest in spring, possibly because birds were 197 searching for breeding territories<sup>29</sup>. At a time when birds are undergoing post breeding 198 dispersal connectivity was least in autumn, suggesting that tagged birds did not linger but 199 instead were passing through.

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There was variation in structural connectivity across networks, with 22% more feeder pairsforming connections in the network of low compared to that of high fragmentation (Fig. 3a).

203 Despite the large numbers of possible drivers behind the movement of feeder-using birds 204 (such as non-experimental bird feeding, density of cat populations or intra-specific 205 territoriality) this study shows a gradient in the ability of birds to move between different 206 feeders with green space fragmentation. Of particular concern is that over the whole 12-207 month period tagged birds failed to visit feeders within the relatively impoverished green 208 fragments in the highly fragmented network, suggesting that residents in these areas will 209 effectively be cut off from this form of nature connection. High quality green space maintains 210 the flow of birds across a broad range of gardens, providing opportunities to a greater number 211 of households to interact with birds.

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213 Functional connectivity was also greatest in the network of low fragmentation. However, we 214 found no difference in this regard between the networks of medium and high fragmentation, 215 and their relatively low functional connectivity suggests that the associated gardens contained 216 relatively poor quality habitat so that birds either passed through the networks, or the 217 networks were too fragmented to establish territories. The networks of medium and low 218 fragmentation both contained green corridors surrounding residential houses, mainly 219 differing in the relative absence of trees in the medium fragmented network. Large trees are known to be keystone structures in urban areas for increasing connectivity<sup>30</sup>. Their loss may 220 221 be a key factor in contributing towards the reduced movement into gardens in this network. 222 This is of concern because this urban form is representative of new developments (within the 223 last 25 years) that are common in European cities.

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The phenomenon of garden bird feeding is growing in many developed regions, with bird

feeders now being a common component of urban areas<sup>11</sup>. The provision of large quantities

of supplementary food drives movement patterns and abundances of feeder-using birds<sup>31,32</sup>.

228 Understanding how individual birds move between gardens with bird feeders, thus provides 229 real world insights into actual movement in the urban landscape. Watching birds at garden 230 feeders provides people with a sense of increased psychological well-being, feelings of relaxation<sup>13</sup> and of being connected to the natural world<sup>12,13</sup>. Listening to bird song and 231 232 watching birds in the garden have been shown to contribute towards attention restoration and recovery from stress<sup>10,15</sup>. Greater connectivity will increase the probability that birds will be 233 234 seen by and so provide pleasure to multiple households, thus multiplying the benefits 235 provided by any individual bird even though it is usually impossible for households to 236 distinguish between these individuals.

237

238 Understanding and quantifying the relationship between the movement of wildlife and urban 239 features is key for ecologically sensitive planning to aid the flow of cultural services within 240 existing cities. Given that many existing urban areas have relatively inflexible urban forms, 241 improved movement could be achieved through targeted greening at focal points of 242 connectivity (e.g. in specific parks and gardens). The applied use of high quality remote 243 sensing techniques, landscape ecology principles and theory (e.g. patch and matrix 244 frameworks) and systematic conservation planning approaches to identify and exploit these 245 focal points has the potential disproportionately to increase the movement of birds into 246 impoverished areas. Targeted greening could be achieved through a combination of 247 management by local authorities, while also raising public awareness of the importance of 248 best practise habitat management in their own gardens. Future research needs to focus on 249 producing real world tools for ease of use by public and private stakeholders for mapping 250 connectivity and identifying and exploiting these focal points. How we improve existing 251 forms and design new residential areas will have a large impact on the daily nature exposure

- of the people that live there, and thus has the potential to mitigate many of the negative
- 253 impacts of urbanisation.

#### 254 Materials and methods

#### 255 STUDY NETWORKS

256 The focal geographical area for this study is what has been termed the 'Cranfield triangle'

257 (52°07'N, 0°61'W). Located ~60km to the north of London, UK, the main urban areas

consisted of Milton Keynes, Luton and Bedford, having a combined population of c. 524,000.

Each study network occupied approximately 0.5 km<sup>2</sup>, with its precise location determined by

the presence of suitable areas in which to mist-net feeder-using birds. Relative to each other,

the networks consisted of one characterized by physically well connected green spaces with

262 high levels of vegetation cover and detached houses (in Milton Keynes; Network of low

263 fragmentation; Fig. 1a), a network with intermediate levels of connected green spaces and

264 vegetation cover with semi-detached houses (in Luton; Network of medium fragmentation;

Fig. 1b), and a network with terraced houses consisting of fragmented green spaces with low

levels of vegetation cover (in Bedford; Network of high fragmentation; Fig. 1c).

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## 268 EXPERIMENTAL DESIGN

269 Bird feeders with integrated RFID reader and antenna were constructed utilising custom-

270 made Arduino components (Relchron LTD, Kirkcaldy, Scotland; <sup>33</sup>). Rectangular antennas

271 (40 mm x 32 mm) recording at 125 kHz were fixed using cable ties and epoxy plastic to the

underside of a single perch on a standard medium sized seed feeder (360 mm, The Royal

273 Society for the Protection of Birds, Sandy, UK). Other perches on the feeder were removed

and associated feeding ports sealed closed to ensure all visitations were recorded. Data-

- loggers were programmed to alternate 400 ms of recording with 400 ms of pause. When a
- 276 PIT tag was within range of the antenna the data logger recorded the time and date along with
- the tag's unique identification number onto a 4gb memory card (SanDisk, Milpitas, USA).

278 The readers were powered 24 hours a day by a 12v battery (Xplorer 88 amp deep cycling,

279 Alpha-batteries, Rochdale, UK), allowing continuous monitoring of feeder usage.

280

281	Each network considered of 17 RFID bird-feeding stations. Each station was set up in a
282	private garden averaging 81 m ( $\pm 20$ m SD) from its nearest neighbour station. Within the
283	constraint of there being suitable locations within gardens, feeders were placed $\sim 0.5$ m from
284	cover for birds, although the actual position was inevitably influenced by property residents
285	who volunteered use of their gardens (Fig. 1). Feeders were installed in gardens up to six
286	weeks before the experiment began to allow birds to familiarise with them. Every feeding
287	station consisted of a bird feeder with an RFID reader, antenna and power source, a bird
288	feeder stand (Kingfisher, Paddington, UK) and a universal squirrel baffle (Gardman,
289	Peterborough, UK). The bottom of the feeder was suspended $\sim 1m$ above the ground. We used
290	the same seed mix in all feeders throughout the experiment (Summer Seed Mix, Haithes, Bird
291	Food specialists, UK). The feeder was maintained by the property residents to ensure that
292	birds could access seed at all times, and a researcher visited all stations every 30 days to
293	replace the battery and to download the data. All 51 feeding stations were fully operational
294	between 1 <sup>st</sup> September 2013 and 31 <sup>st</sup> August 2014.
295	

# 296 TAGGING IN THE FIELD

Birds were caught and tagged in private green spaces at two locations in each network,

298 chosen to maximize catching rates and with the nearest RFID feeding station approximately

299 15 m from the closest net (Fig. 1). Mist-nets and tape lures were used to catch two garden

- 300 species that commonly visit bird feeders: blue tit *Cyanistes caeruleus* and great tit *Parus*
- 301 *major*. Mist netting was carried out intensively during the experimental set-up period and
- 302 then monthly in each network, with birds being fitted with British Trust for Ornithology

303 (BTO) metal rings and with a PIT tag, which was fully moulded into an 8mm plastic ring (IB

304 Technology, Aylesbury, U.K.). All bird ringing was carried out under BTO license A/5780

305 with a special endorsement to attach PIT tags to target species. This research was

306 conducted with approval from, and in accordance with, the University of Exeter

307 Biosciences ethical review committee, project number 2013/72.

308

# 309 LANDSCAPE CHARACTERIZATION

310 We characterized the spatial pattern of vegetation and non-vegetation, and canopy height in 311 the landscape for each network using airborne remote sensing, specifically with hyperspectral 312 and LiDAR data (see Supplementary Appendix S1). We described the habitat that birds are 313 likely to use when moving between any pair of feeders within a network by applying a buffer 314 around and between each pair of feeders equal to 0.25 times the distance between them. 315 Based on previous studies, such an ellipsoid, with a constant length to width ratio of three, 316 represents a reasonable area for a bird seeking to move between feeders (for example, see Supplementary Fig. S3; see also <sup>19,34</sup>). Within each buffer we calculated the percentage of 317 318 pixels containing tall vegetation (> 0.7m; vegetation cover according to the pixel values 319 exceeding a basic threshold for a vegetation index; Table 1; see Supplementary Appendix 320 S1). Finally, we counted the number of road gaps between each feeder pair, where a road gap 321 was considered to be present when a road dissected the buffer (for example; see 322 Supplementary Fig. S3). At the spatial extent of study, three or more road gaps were rare and 323 so these were pooled with two road gaps. This was then treated as a three-level factor of 0-2 324 road gaps (Table 1). We calculated the distance in meters of the closest feeder of each pair 325 with the closest ringing site (termed here as ringing site distance; mean = 42m, SD = 46m). We then divided the year into four equal seasons: autumn (1<sup>st</sup> Sep-30<sup>th</sup> Nov), winter (1<sup>st</sup> Dec-326 28<sup>th</sup> Feb), spring (1<sup>st</sup> Mar–31<sup>st</sup> May) and summer (1<sup>st</sup> Jun–30<sup>th</sup> Aug 2014). To assess the 327

probability of whether a bird passed within the ellipsoid between feeders, we calculated thetime taken to make each connection.

330

## 331 STATISTICAL ANALYSIS

## 332 URBAN FEATURES WITHIN FORMS AND SEASON

All statistical analyses were performed in R version  $3.1.2^{35}$ , with mixed effects models built

334 using the lme4 package<sup>36</sup>. To explore how different features within the three networks

determined seasonal movement of feeder-using birds we used a hurdle modelling

framework<sup>37</sup>. The hurdle model consisted of a binomial model (presence-absence of at least

one connection in any season by species as the response variable, with one replication per

338 species, per feeder pair, per season) and a count effect model (frequency of these connections

in any season as the response variable, with one replication per individual tagged bird, per

feeder pair where connections  $\geq 1$ ) based on a Poisson distribution truncated at 0 (i.e. no

341 stochastic absence of connection). A hurdle modelling approach differentiated the effects of

342 network on structural and on functional connectivity, and better accommodated marked

343 overdispersion in our response variable than using a Poisson model<sup>37</sup>

344



346 binomial and Poisson mixed effect models for each network separately. We standardized the

347 continuous variables vegetation cover, distance between feeder pair and ringing site distance

348 (i.e. each was rescaled to have a mean of zero and a standard deviation of 1). To explore

349 structural connectivity we built a generalized linear mixed model (GLMM) with a binomial

as on the effect of vegetation cover, distance between feeder pair,

351 number of road gaps and season on the presence-absence of connections between feeder pairs

352 (see Table 1 for a summary of covariates by network). We also included species and the

353 distance from ringing sites as covariates. Finally, to control for replication of feeder pairs

across the seasons, for each feeder station we included a unique identification number

355 (FeederID) within network as two random effects. To explore functional connectivity we

356 built a GLMM with a Poisson error distribution to test the effect of the same predictor

357 variables on the frequency of connections between feeders at the individual level (i.e. where

358 connections  $\geq$  1). We included three random effects; FeederID for each feeder station to

359 control for replication of feeder pairs across the season and a unique tag number to control

360 for multiple individuals moving between the same feeder pair in any season.

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## 362 PATTERNS OF MOVEMENT ACROSS URBAN FORMS

363 To explore general patterns of movement across the networks, we pooled connections by

364 season before again using a hurdle modelling framework. The hurdle model consisted of a

365 quasi-binomial model (presence/absence of at least one connection between feeder pairs, with

366 one replication per feeder pair) and a count effect model (frequency of connections, with one

367 replication per feeder pair where connections  $\geq 1$ ). It was assumed that general patterns of

368 movement and associated cultural service provision by feeder-using songbirds were

369 dependent on individual birds and not species, so we pooled connections by species. In each

370 case we then tested for the effect of network type (included as a three level factor: low

371 fragmentation; medium fragmentation; high fragmentation) on the response variable. We

included distance between feeder pairs within feeder groups (in meters) as a covariate to

373 control for slight variation between networks (Table 1).

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466		

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474

# 475 Author contributions

- 476 DTCC, RI, KJG conceived and designed the study. DTCC carried out the fieldwork, analysed
- 477 the data and wrote the paper. SH and KA processed and provided the remote sensing data.
- 478 All authors edited the paper.
- 479

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**Table 1.** Summary of urban features per feeder pair in each of the three networks: mean

485 distance between pairs of feeders, mean vegetation cover within the buffer and the total

486 number of road gaps crossing buffers (as a measure of overall green space fragmentation).

487 Associated standard errors are shown in brackets.

488

489	Network	Distance between feeders (metres)	% Vegetation cover	Total number of road gaps
	Low fragmentation	203 (±92)	45.8 (±8.4)	121
490	Medium fragmentation	218 (±98)	28.1 (±10.4)	182
	High fragmentation	213 (±98)	19.3 (±7.6)	302

492	<b>Table 2.</b> The relationships between a) structural (binomial) and b) functional (Poisson)
493	connectivity and the presence of key urban features and season by network, for two feeder-
494	using songbirds. We show parameter estimates with standard errors and confidence intervals
495	(CI) for factor levels relative to a comparative base factor level (0 road gaps, summer and
496	blue tits, respectively). Significant variables and factor levels are shown as: * $P < 0.05$ ; ** $P$
497	<0.01; *** $P$ <0.001. We show the marginal $R^{2}_{GLMM(m)}$ and conditional $R^{2}_{GLMM(c)}$ .

	Low fragm	nentation	Medium fragmentation		High fragmentation	
	Estimate (±se)	CI (2.5%; 97.5%)	Estimate (±se)	CI (2.5%; 97.5%)	Estimate (±se)	CI (2.5%; 97.5%)
a) Structural connec	ctivity					
Intercept	-1.7 (±0.4)***	-2.7; -0.7	-1.8 (±0.4)***	-2.6; -0.8	-2.4 (±0.6)***	-3.7; -1.0
Vegetation cover	0.3 (±0.1)**	0.1; 0.6	0.8 (±0.2)***	0.4; 1.1	0.5 (±0.2)**	0.1; 0.9
Distance	-1.3 (±0.1)***	-1.6;-1.0	-1.2 (±0.2)***	-1.6; -0.9	-0.2 (±0.3)	-0.7; 0.4
1 road gap	-0.1 (±0.3)	-0.6; 0.4	-0.3 (±0.3)	-0.8; 0.3	0.1 (±0.4)	-0.8; 0.9
2 road gaps	-0.2 (±0.4)	-1.0; 0.7	-0.0 (±0.4)	-0.7; 0.7	-0.7 (±0.6)	-1.8; 0.4
Autumn	1.5 (±0.2)***	1.0; 2.0	0.8 (±0.3)**	0.3; 1.4	0.3 (±0.3)	-0.4; 1.0
Winter	0.8 (±0.2)**	0.3; 1.3	0.6 (±0.3)*	0.0; 1.1	0.6 (±0.3)	-0.1; 1.2
Spring	0.1 (±0.2)	-0.4; 0.6	0.8 (±0.3)**	0.3; 1.3	0.9 (±0.3)**	0.3; 1.6
Species	0.0 (±0.2)	-0.3; 0.4	-1.1 (±0.2)***	-1.4; -0.7	-1.6 (±0.2)***	-2.1; -1.1
Ring site distance	-0.2 (±0.3)	-0.7; 0.2	-0.5 (±0.2)**	-0.8; -0.1	-0.3 (±0.3)	-0.8; 0.4
$R^{2}_{GLMM(m)}$		0.29		0.4		0.28
$R^{2}_{GLMM(c)}$		0.57		0.55		0.62
b) Functional conne	ectivity					
Intercept	1.2 (±0.3)***	0.6; 1.7	1.3 (±0.2)***	1.0; 1.6	1.2 (±0.3)***	0.5; 1.7
Vegetation cover	0.1 (±0.0)*	0.0; 0.2	0.2 (±0.1)**	0.1; 0.3	-0.2 (±0.1)*	-0.5; -0.1
Distance	-0.5 (±0.0)***	-0.6; -0.5	-0.3 (±0.1)***	-0.4; -0.1	-0.2 (±0.1)	-0.5; 0.0
1 road gap	-0.4 (±0.1)***	-0.6; -0.3	-0.2 (±0.1)**	-0.4; -0.1	-0.1 (±0.2)***	-1.5; -0.9
2 road gaps	0.3 (±0.2)	-0.1; 0.7	-0.3 (±0.2)*	-0.6; -0.0	-2.4 (±0.4)***	-2.9; -1.4
Autumn	0.3 (±0.1)***	0.1; 0.4	-0.5 (±0.1)***	-0.7; -0.4	-0.3 (±0.1)**	-0.5; -0.1
Winter	0.4 (±0.1)***	0.3; 0.6	-0.5 (±0.1)***	-0.7; -0.4	-0.0 (±0.1)	-0.2; 0.2
Spring	0.2 (±0.1)**	0.1; 0.4	0.1 (±0.1)	-0.1; 0.3	0.6 (±0.1)***	0.4; 0.8
Species	0.2 (±0.2)	-0.1; 0.5	-0.1 (±0.1)	-0.3; 0.1	-0.0 (±0.2)	-0.5; 0.4
Ring site distance	-0.1 (±0.1)	-0.2; 0.0	-0.2 (±0.1)**	-0.3;-0.0	-0.1 (±0.2)	-0.4; 0.2
$R^{2}_{GLMM(m)}$		0.13		0.15		0.17
$R^{2}_{GLMM(c)}$		0.22		0.17		0.23

**Table 3.** Hurdle model testing for the relationships between networks on overall levels of
movement of feeder-using garden bird: a) Structural connectivity; b) Functional connectivity.

501 We show parameter estimates and associated standard errors, t values and confidence

- 502 intervals (CI) for medium and low fragmentation networks relative to the base factor level of
- 503 the high fragmentation network. Significant factor levels are shown as: \*P < 0.05; \*\*P < 0.01;
- 504 \*\*\*P < 0.001. The pseudo  $R^2$  is McFaddens.

	Estimate (±se)	t value	CI (2.5%; 97.5%)
a) Structural connectivity			$pR^2 = 0.15$
Intercept	1.7 (±0.3)	5.6***	1.2; 2.4
Medium fragmentation	1.0 (±0.3)	3.7***	0.5; 1.6
Low fragmentation	1.1 (±0.3)	4.1***	0.6; 1.7
Distance	-0.008 (±0.001)	-6.9***	-0.01; -0.006
b) Functional connectivity			$pR^2 = 0.39$
Intercept	5.0 (±0.2)	20.2***	4.9; 6.0
Medium fragmentation	0.02 (±0.3)	0.1	-0.6; 0.5
Low fragmentation	0.6 (±0.2)	2.4*	0.2; 1.1
Distance	-0.1 (±0.001)	-7.5***	-0.02; -0.01

**Figure 1.** The frequency of connections (i.e. functional connectivity) of two species of garden bird moving between bird feeders, within a) the network of low fragmentation, b) the network of medium fragmentation, c) the network of high fragmentation. Connections occurred over a 12-month period. The upper panel rasters were generated using hyperspectral and LiDAR data (Appendix S1), we show the location of rfid bird feeders in red. Habitat classification: white; vegetation free surfaces at ground level, light grey, buildings; medium grey, grass & low lying vegetation, dark grey; vegetation (at 2m resolution). The lower panels show the frequency of each connections (black line, >100; >50-100, dark grey line; >10-50, medium grey line; 1-10 light grey line) and the total number of connections made by each feeder (divided into 4 categories denoted by increasing size and brightness of the red circle: 0; 10; 50; 100; >200).  $\blacklozenge$  Bird catching locations. Images were created in R version  $3.1.2^{34}$ .

512

\* To increase the clarity of the image only those connections that occurred between feeder pairs that were less than the mean distance between all

feeder pairs are shown (<213 m); this only loses 9% of the total connections made, and does not exclude any feeder pairs with =>10 connection.

516 **Figure 2.** The effect of urban features, and season, on structural (a-c) and functional (d-e)

- 517 connectivity across networks, for feeder-using birds. Structural connectivity (presence-absence of
- 518 connections): a) the percentage vegetation cover of feeder pairs with connections present and absent
- 519 by network; b) the number of feeder pairs against the number of road gaps and c) the number of
- 520 feeder pairs that formed connections in each season. Functional connectivity (frequency of
- 521 connections where  $\geq 1$  connection was made (log 10 on y-axis): d) frequency against the number of
- 522 road gaps by network, and e) frequency against season.

- 523 Figure 3. Comparison of the movements of tagged birds between feeder pairs across networks: a)
- 524 Structural connectivity (the numbers of feeders that each feeder is connected to), and b) Functional
- 525 connectivity (the total number of connections made to each feeder). Pseudo R<sup>2</sup> from quasi-models
- 526 shown.





