

31 **Abstract**

32 A diagenetic model is used to simulate the diagenesis and burial of particulate organic carbon (C_{org})
33 and phosphorus (P) in marine sediments underlying anoxic versus oxic bottom waters. The latter are
34 physically mixed by animals moving through the surface sediment (bioturbation) and ventilated by
35 burrowing, tube-dwelling organisms (bioirrigation). The model is constrained using an empirical
36 database including burial ratios of C_{org} with respect to organic P ($C_{org}:P_{org}$) and total reactive P
37 ($C_{org}:P_{react}$), burial efficiencies of C_{org} and P_{org} , and inorganic carbon-to-phosphorus regeneration ratios.
38 If P_{org} is preferentially mineralized relative to C_{org} during aerobic respiration, as many previous studies
39 suggest, then the simulated P_{org} pool is found to be completely depleted. A modified model that
40 incorporates the redox-dependent microbial synthesis of polyphosphates and P_{org} (termed the
41 microbial P pump) allows preferential mineralization of the bulk P_{org} pool relative to C_{org} during both
42 aerobic and anaerobic respiration and is consistent with the database. Results with this model show
43 that P burial is strongly enhanced in sediments hosting fauna. Animals mix highly labile P_{org} away
44 from the aerobic sediment layers where mineralization rates are highest, thereby mitigating diffusive
45 PO_4^{3-} fluxes to the bottom water. They also expand the redox niche where microbial P uptake occurs.
46 The model was applied to a hypothetical shelf setting in the early Paleozoic; a time of the first
47 radiation of benthic fauna. Results show that even shallow bioturbation at that time may have had a
48 significant impact on P burial; an effect that increases with subsequent deeper bioturbation. Our
49 model provides support for a recent study that proposed that faunal radiation in ocean sediments led to
50 enhanced P burial and, possibly, a stabilization of atmospheric O_2 levels. The results also help to
51 explain $C_{org}:P_{org}$ ratios in the geological record and the persistence of P_{org} in ancient marine sediments.

52 1. Introduction

53 Permanent burial of phosphorus (P) in continental margin sediments is a major control on the marine
54 P inventory, primary productivity and possibly interglacial CO₂ concentrations (e.g. Broecker, 1982;
55 Wallmann, 2014). On Myr time scales, P burial exerts a strong, if not dominating, influence on
56 atmospheric O₂ levels (Van Cappellen and Ingall, 1996; Lenton and Watson, 2000; Boyle et al.,
57 2014). Diagenetic transformations of P and the factors controlling P burial are thus intensively studied
58 (reviewed by Benitez-Nelson, 2000; Paytan and McLaughlin, 2007; Slomp, 2011; Ruttenberg, 2014).

59 Particulate organic P (P_{org}) comprises more than 90 % of P that rains to the seafloor (Delaney, 1998).
60 Phosphate (PO₄³⁻) that is solubilized from P_{org} in the sediments can be sequestered into authigenic
61 mineral phases, mainly as carbonate fluorapatite (CFA) and P bound or adsorbed to iron
62 oxyhydroxides (FeP) (Berner et al., 1993). These transformations, more generally described as ‘sink
63 switching’, increase the overall benthic retention efficiency of P (Ruttenberg and Berner, 1993;
64 Anderson et al., 2001). Observations from the North Atlantic suggest that sink-switching is enhanced
65 in sediments that are bioturbated, that is, reworked by infaunal deposit and detritus feeders (Slomp et
66 al., 1996). The sedimentary retention capacity of P further appears to be sensitive to the ambient redox
67 conditions (Algeo and Ingall, 2007). For example, FeP content tends to be lower under anoxic bottom
68 waters due to reduced ferrous iron oxidation and co-sequestration of P (Sundby et al., 1992; Jensen et
69 al., 1995; McManus et al., 1997). CFA content may be higher under these conditions, especially in
70 modern oxygen minimum zones where phosphorites may form (Papineau, 2010).

71 The impact of oxygen levels on P_{org} burial is less well understood. Data on P_{org} burial efficiencies
72 (PBE) are scarce, yet point toward lower values under anoxic versus oxic bottom waters, that is,
73 preferential mineralization of P_{org} in anoxic settings. This trend is opposite to the organic carbon
74 burial efficiency (CBE), which is apparently high under anoxic bottom waters and low under oxic
75 waters (see Table 1 and further discussion by Burdige, 2007). It thus follows that molar organic
76 carbon to phosphorus ratios (C_{org}:P_{org}) in laminated anoxic facies commonly exceed the Redfield ratio
77 (106:1), whereas bioturbated sediments have ratios that are around Redfield or lower (Table 1). This
78 difference is maintained, although less pronounced, for the ratio of C_{org} to reactive phosphorus (P_{reac} =
79 P_{org} + FeP + CFA). Consistent with these findings, the inorganic carbon-to-phosphorus regeneration
80 ratio derived from in situ flux measurements, (C:P)_{REG}, is often above Redfield in oxic settings and
81 below it in sediments underlying intermittently or permanently oxygen-deficient waters (Table 1).
82 Consequently, C:P burial ratios are used as a proxy for bottom water redox conditions and biological
83 productivity in the paleo-ocean (Ingall and Jahnke, 1994; Slomp and Van Cappellen, 2007).

84 The factors that control the C_{org}:P_{org} ratio are unclear and controversial (Van Cappellen and Ingall,
85 1996; 1997; Colman et al., 1997). Basic questions remain, such as the extent of preferential

86 mineralization of P_{org} relative to C_{org} under oxic versus anoxic conditions (Colman and Holland,
87 2000), not least because the kinetics of C_{org} and P_{org} mineralization are very poorly understood. Some
88 workers have proposed that microbial sequestration of P can help to explain redox-dependent $C_{org}:P_{org}$
89 burial ratios (Ingall and Van Cappellen, 1990; Ingall et al., 1993; Ingall and Jahnke, 1997; Sannigrahi
90 and Ingall, 2005; Diaz et al., 2008; Goldhammer et al., 2010; Diaz et al., 2012). Many genera of
91 bacteria are known to accumulate and store P as polyphosphate under aerobic and nitrogenous
92 conditions (Shapiro, 1967; Gächter and Meyer, 1993; Davelaar, 1993; Kulaev and Kulakovskaya,
93 2000). Polyphosphates are long chains of orthophosphate units linked by high energy
94 phosphoanhydride bonds (Gächter and Meyer, 1993). Experimental observations indeed confirm that
95 biologically associated P is higher in oxidized sediments (Gächter et al., 1988; Aller, 1994). It has
96 been suggested that organic P biomolecules synthesized under oxic conditions, possibly via
97 polyphosphate intermediates, are converted into nearly non-metabolizable organic P such as some
98 phosphate esters and phosphonates that are subsequently permanently buried (Ingall et al., 1990;
99 Berner et al., 1993; Ingall et al., 1993; Ingall and Jahnke, 1994; Van Cappellen and Ingall, 1994;
100 Ingall and Jahnke, 1997). This would constitute a microbially-mediated P sink that is more efficient
101 under oxic bottom waters, favoring an increase in P_{org} burial relative to C_{org} (i.e. low $C_{org}:P_{org}$ burial
102 ratios, Table 1). Refractive microbial P compounds could also help to explain the puzzling persistence
103 of P_{org} in ancient marine sediments (Ruttenberg, 2014).

104 In this study, we aim to unify these different perspectives of P cycling using a diagenetic model for
105 continental margin sediments, specifically, a shallow marine shelf with a water depth of ~100 m. The
106 observational database in Table 1 is used to (i) ascertain the extent of preferential mineralization of
107 P_{org} relative to C_{org} in oxic and anoxic sediments, and (ii) quantify the impact of sediment mixing by
108 animals on P burial. Both of these are open questions with an important bearing on the interpretation
109 of the sedimentary record.

110 Our interest in the role of fauna on P burial stems from the radiation of deposit feeders and burrowers
111 in the terminal Ediacaran to early Paleozoic (ca. 542 - 420 Ma). This may have occurred against a
112 backdrop of variability in ocean redox conditions, with conceivable impacts on, and geochemical
113 feedbacks with, the spread of early animals (Canfield et al., 2007; Shen et al., 2008; Butterfield, 2009;
114 Johnston et al., 2012; Lyons et al., 2012; Och and Shields-Zhou, 2012). One possible feedback loop
115 relates to the argument that the onset of bioturbation at this time led to enhanced burial of P and
116 stabilization of atmospheric O_2 levels (Boyle et al., 2014). The major control on sedimentary P burial
117 in global models of ancient oceans is traditionally assumed to be the O_2 availability rather than
118 bioturbation per se, with any mechanistic linkage with bioturbation being merely implicit (Van
119 Cappellen and Ingall, 1994; 1996; Lenton and Watson, 2000). The validity of this assumption is tested

120 here with the diagenetic model. We find that indwelling fauna, in combination with microbial P
 121 synthesis, would have strongly enhanced P burial in the early Palaeozoic Era.

122

123 2. Model

124 2.1. Architecture and application

125 A vertically-resolved 1-D model is used to investigate the coupled biogeochemical cycling of C, N, P,
 126 Mn, Fe and S in the uppermost meter of fine-grained continental shelf sediments (ca. 100 m water
 127 depth). Full details of the model can be found in the Supplementary Material and in Dale et al.
 128 (2015a). For the present study, we limit the biogeochemical model description to the previously
 129 unpublished P cycle. Key model boundary conditions and parameters are listed in Table 2.

130 A total of 13 solutes and 17 solid species are considered in the model. Particulate matter is transported
 131 dynamically through the sediment column by accumulation (burial) and bioturbation. Solutes are
 132 transported by molecular diffusion and bioirrigation. The coupling of biogeochemistry and transport is
 133 summarized by the following generic mass-conservation equations (Berner, 1980; Boudreau, 1997):

134

$$135 \quad \varphi \frac{\partial C_a}{\partial t} = \frac{\partial}{\partial z} \left(\varphi D \frac{\partial C_a(z,t)}{\partial z} \right) - \frac{\partial \varphi(L) \omega_{acc} C_a}{\partial z} + \alpha(0) \cdot \gamma \cdot \exp \left(-\frac{z}{z_{bio}} \right) \varphi (C_a(0) - C_a) + \Sigma \varphi R \quad (1a)$$

136

$$137 \quad (1 - \varphi) \frac{\partial C_s}{\partial t} = \frac{\partial}{\partial z} \left((1 - \varphi) D_b(0) \cdot \exp \left(-\frac{z^2}{2 \cdot z_{bt}^2} \right) \frac{\partial C_s}{\partial z} \right) - \frac{\partial (1 - \varphi(L)) \omega_{acc} C_s}{\partial z} + \Sigma (1 - \varphi) R \quad (1b)$$

138

139 where C_a and C_s are the time- and depth-dependent concentrations of solutes in mmol cm^{-3} (of pore
 140 water) and particulate species in weight percent of dry sediment (%), respectively, z (cm) denotes
 141 depth in the sediment, φ is porosity, ω_{acc} (cm yr^{-1}) is the sedimentation rate, D ($\text{cm}^2 \text{yr}^{-1}$) is the
 142 molecular diffusion coefficient, and ΣR is the sum of the rate of change of concentration due to
 143 biogeochemical reactions. Constitutive equations describing the depth dependency of transport
 144 parameters are provided in Table S1, and the complete list of biogeochemical reactions, rate
 145 expressions and parameters are given in Table S2 to S5. In this study, model results represent steady
 146 state simulations, that is, where concentrations are invariable with time ($\partial C / \partial t = 0$).

147 The third term on the right-hand-side of Eq. (1a) describes the exponential decrease of bioirrigation
 148 with sediment depth. Bioirrigation is the non-local exchange of seawater with burrow water by the
 149 pumping activity of tube-dwelling animals (Aller and Aller, 1992). Mathematically, the process is
 150 described using the parameters $\alpha(0)$ (yr^{-1}) and $C_a(0)$, which are the maximum bioirrigation coefficient

151 (under oxic bottom waters) and solute concentration at the sediment-water interface, respectively. z_{bio}
152 (cm) is the attenuation coefficient that parameterizes the depth to which burrows extend below the
153 sediment surface. The dimensionless parameter γ is a scaling parameter for ferrous iron to reflect
154 rapid oxidation on burrow walls (see Dale et al., 2015a).

155 Bioturbation, the first term on the right-hand-side of Eq. (1b), is treated as biodiffusion, analogous to
156 molecular diffusion, using a Gaussian function to describe the decrease in sediment mixing with
157 sediment depth (Boudreau, 1996). The maximum rate of particle reworking in the surface mixed layer
158 is defined by the bioturbation enhanced diffusion coefficient, $D_b(0)$ ($\text{cm}^2 \text{yr}^{-1}$). The parameter z_{bt} (cm)
159 defines the depth at which bioturbation intensity is half of $D_b(0)$, according to the Gaussian decrease
160 in bioturbation intensity. A relation between $D_b(0)$ and z_{bt} has, to our knowledge, not been
161 demonstrated empirically, and these parameters are independent of one another.

162 Our treatment of biodiffusion can be viewed as reverse conveyor-belt feeding, i.e. downward
163 transport of particles (reviewed by Burdige, 2006; Meysman et al., 2003). It is also a local process,
164 because it mixes sediment between adjacent layers. However, under natural conditions, most modes
165 of sediment transport adhere to non-local exchange formalisms, that is, transport of material between
166 non-adjacent sediment layers (Meysman et al., 2003). Non-local sediment mixing can occur both
167 upwards by conveyor-belt (head-down) feeding and downwards (e.g. ingestion at surface, egestion at
168 depth). Non-local conveyor-belt feeding tends to retain radiotracers such as ^{210}Pb in the surface mixed
169 layer, and steady-state tracer profiles resulting from this type of transport tend to resemble those
170 resulting from the biodiffusion model (Boudreau, 1986). Non-local reverse conveyor-belt feeding, on
171 the other hand, gives rise to subsurface maxima in tracer profiles (Smith et al., 1986). Although the
172 exact pattern of tracer distribution will depend on the specific mixing mechanism, each type of
173 feeding behavior will lead to a mixing of aged and fresh organic matter within the bioturbated. In this
174 paper, we adopt the biodiffusion model because it has been proven to be a robust empirical model for
175 sediment mixing and is straightforward to solve numerically. More complex treatments have been
176 developed elsewhere (Robbins, 1986; Boudreau, 1986).

177 The value of z_{bt} is set to 3 cm, to reflect a mean mixed layer thickness in the modern ocean of 5 to 10
178 cm (Boudreau, 1997; Teal et al., 2008). The parameters z_{bt} and z_{bio} are deemed to be independent of
179 one another since the ecology of sediment ‘bulldozers’ and irrigators is different. Here, we simply set
180 z_{bio} to 2 cm so that the irrigation depth is similar to the bioturbation depth (Archer et al., 2002). It is
181 important to note that these parameters correspond to the sediment mixed layer where reworking or
182 bulldozing by animals is sufficiently intense to homogenize the sediment (Tarhan et al., 2015). In
183 contrast, burrows may extend many decimeters below the mixed layer and impact biogeochemical
184 cycling there. The ghost shrimp *N. californiensis*, for example, can enhance denitrification by
185 pumping seawater nitrate into deep sediment layers (Bertics et al., 2010). In this paper, we do not

186 consider deep irrigation because it is in all likelihood much weaker than surface irrigation (Fossing et
187 al., 2000). Our focus on the surface sediment will facilitate comparisons with the early Paleozoic
188 scenarios where the maximum burrow depth is limited to a few cm (Tarhan et al., 2015). The
189 sensitivity of the model to bioturbation and bioirrigation is tested later.

190 The model is applied across a range of bottom water dissolved O₂ concentrations from anoxic (0.1
191 μM) to oxic (150 μM). The former threshold is not strictly anoxic, but instead represents the micro-
192 aerobic conditions that exist within some ‘anoxic’ oxygen minimum zones such as in the Eastern
193 South Pacific Ocean (Thamdrup et al., 2012). Under oxic conditions, the parameters $D_b(0)$ and $\alpha(0)$
194 take values of 28 cm² yr⁻¹ and 465 y⁻¹, respectively (Supplement). They are allowed to decrease
195 gradually toward zero if O₂ drops below a 20 μM threshold. At this concentration, a decrease in
196 species richness and bioirrigation has been observed (Levin and Gage, 1998; Dale et al., 2013) with
197 surface deposit feeders dominating over burrowing macrofauna (Middelburg and Levin, 2009).

198 We begin by focusing on modern sediments, for which data are relatively abundant, in order to
199 parameterize the P cycle. To avoid the risk of being too site-specific, P turnover is mainly constrained
200 by the data in Table 1 that originate from a wide range of marine settings. In that sense, the model is
201 more emblematic of an average shallow continental margin setting. Parameterization of transport and
202 biogeochemical processes draws from a large body of previous empirical studies from shelf
203 environments (Supplementary Material). In particular, this includes realistic organic matter and iron
204 fluxes to the seafloor, organic matter reactivity, reaction rate constants, sedimentation rates, and
205 bioturbation and bioirrigation intensities. Most of these are taken directly from the shelf scenario in
206 Dale et al. (2015a) that is confirmed against a global database of dissolved iron fluxes. Only a
207 minimal amount of additional parameter tuning has been used for organic matter degradation as
208 described below. The model is then applied to the early Paleozoic shelf environment (ca. 542 – 420
209 Ma) by altering the boundary conditions accordingly. We define sediments underlying anoxic waters
210 as ‘anoxic’ whereas bioturbated sediments under oxic waters are termed ‘oxic’, whilst noting that the
211 latter quickly turn anoxic and anaerobic below the thin surface oxidized layer (< ca. 1 cm, Glud et al.,
212 2008 and Results).

213

214 2.2. Phosphorus biogeochemistry

215

216 The P cycle initially centers on three particulate species (Fig. 1): organic P (P_{org}), iron-associated P
217 (FeP) and carbonate fluorapatite (CFA). Dissolved P takes the form of inorganic orthophosphate
218 (PO₄³⁻). The P model is similar to previous applications (Van Cappellen and Berner, 1988; Reed et al.,
219 2011; Kraal et al., 2012) and, like these other models, highly simplifies the representation of P cycling
220 in sediments, especially with regard to authigenic P mineral formation. However, the main diagenetic

221 pathways are included. Our model structure differs from the other studies by the inclusion of P uptake
222 by microorganisms (Section 3.3), plus a more realistic description of organic matter degradation
223 kinetics. P concentrations are reported in $\mu\text{mol g}^{-1}$ of P (0.01 wt. % = $3.23 \mu\text{mol g}^{-1}$).

224 Mineralization of deposited organic matter, chemically defined as $\text{CN}_{r_{NC}}\text{P}_{r_{PC}}$, drives all diagenetic
225 reactions that take place in the model. Parameters r_{NC} and r_{PC} are the (Redfield) atomic ratios of
226 particulate organic nitrogen and phosphorus to carbon (Table 2). Dissolution of organic nitrogen and
227 phosphorus is coupled to the rate of C_{org} mineralization. C_{org} is oxidized by several pathways; aerobic
228 respiration, nitrate and nitrite reduction (denitrification), manganese oxide reduction, dissimilatory
229 iron reduction, sulfate reduction and methanogenesis in that order (R_1 to R_7 , Table S4). The kinetics
230 of C_{org} mineralization is described using continuum kinetics, where the rate constant for C_{org}
231 degradation is time dependent (Middelburg, 1989; Boudreau and Ruddick, 1991):

$$232 \quad k(t) = v (a + t)^{-1} \quad (2)$$

233 In this expression, a (yr) is the average lifetime of the reactive components and v (dimensionless) is
234 the distribution of C_{org} reactivity. Low v values indicate that C_{org} is dominated by refractory
235 components, whereas higher values correspond to a more even distribution of reactive types.
236 Similarly, C_{org} characterized by low a will be rapidly degraded whereas high a implies less reactive
237 material that is more likely to be buried to deeper sediments. Reactive continuum-type models predict
238 a more realistic profile of C_{org} reactivity with sediment depth than those based on ‘multi-G’ kinetics.
239 The latter require a somewhat arbitrary number of discrete C_{org} fractions, or ‘G’, to be assigned
240 individual rate constants that are poorly defined at the global scale (Boudreau and Ruddick, 1991).
241 The choice of degradation kinetics is important because it determines the sediment redox structure
242 and the depth at which P fractions are dissolved or precipitated and, hence, phosphate fluxes across
243 the sediment-water interface. However, continuum models are difficult to apply to bioturbated
244 sediments since the age and reactivity of organic matter within the bioturbated zone tend to be poorly
245 constrained (Boudreau and Ruddick, 1991; Middelburg, 1989). Ages of radioactive tracers within the
246 bioturbated zone depend not only on the burial velocity and bioturbation rate but also on the decay
247 rate of the tracer itself (Meile and Van Cappellen, 2005).

248 We used a previous approach (Dale et al., 2015a) that calculates discrete C_{org} fractions from the
249 continuum parameters a and v that, when summed together, give the same C_{org} reactivity profile as the
250 continuum model. More accurate results are produced as the number of defined C_{org} fractions
251 increases. Fourteen fractions were found to give an excellent agreement with the continuum model
252 (Dale et al., 2015a), using an initial parameterization of a (3×10^{-4} yr) and v (0.125) constrained from
253 oxic degradation experiments with fresh phytoplankton (Boudreau et al., 2008). The rate of P_{org}
254 mineralization for each pathway is initially determined as:

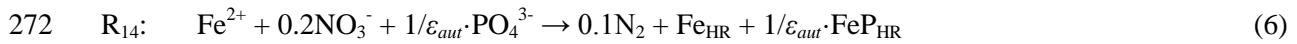
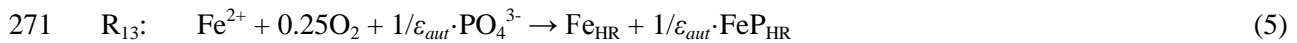
255 $R_{P_{org}} = r_{CP} \cdot R_{POC} \cdot \psi$ (3)

256 where RPOC is the total rate of POC degradation and ψ represents the sum of oxidant limitation and
 257 inhibition terms (see Supplement). Later, the rate of P_{org} mineralization will be adjusted depending on
 258 whether organic matter is respired aerobically or anaerobically. Organic phosphorus is mineralized to
 259 dissolved PO_4^{3-} .

260 Four reactive fractions of particulate iron oxides were defined according to the classification scheme
 261 based on wet chemical extractions (Canfield et al., 1992; Poulton et al., 2004). These are defined as
 262 highly reactive (Fe_{HR} , e.g. nano-goethite), moderately reactive (Fe_{MR} , e.g. goethite and hematite), and
 263 poorly reactive (Fe_{PR} , e.g. iron silicates). Fe_{HR} , Fe_{MR} and Fe_{PR} are all dissolvable by sulfide, but at
 264 different rates (see Supplementary Material). The model also includes detrital (unreactive) iron.

265 Reactions that couple the Fe and P cycle are the dissolution of Fe_{HR} by dissimilatory iron reduction
 266 (R_5), authigenic precipitation of Fe_{HR} via aerobic and anaerobic oxidation of ferrous iron (R_{13} , R_{14} ,
 267 R_{24}), reductive dissolution of Fe_{HR} and Fe_{MR} by sulfide (R_{26}), and the ageing and crystallization of
 268 Fe_{HR} into more stable Fe_{MR} phases (R_{28}). These are given as (see Supplement for more details):

269



276

277 Detrital P bound within highly crystalline iron fractions and other minerals is not considered in the
 278 model. The iron module also includes precipitation of iron sulfide minerals, although without
 279 associated P turnover (Krom and Berner, 1980).

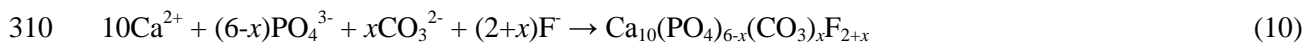
280 The change in FeP content is calculated using the particulate fraction of P to Fe (θ , mol P / mol Fe)
 281 and/or Fe:P enrichment ratios (ϵ , mol Fe / mol P) (Dale et al., 2013). The capacity of poorly
 282 crystalline iron oxides to sequester large amounts of ambient PO_4^{3-} is well recognized (Torrent et al.,
 283 1992; Slomp et al., 1996; Anschutz et al., 1998; Feely et al., 1998). The Fe:P ratio in amorphous iron
 284 oxides in surface sediments, principally nano-particulate goethite, seems to be constant at around 10

285 (Sundby et al., 1992; Jensen and Thamdrup, 1993; Slomp et al., 1996; Anschutz et al., 1998; van der
 286 Zee et al., 2003). Hence, we take this value for ϵ_{aut} in R_{13} , R_{14} and R_{24} . The rate expressions for these
 287 reactions include a rate-limiting term that slows down the rate of PO_4^{3-} uptake into Fe_{HR} if PO_4^{3-}
 288 concentrations are diminished to low levels (Table S5). Authigenic Fe_{HR} can therefore have a Fe:P
 289 ratio that is higher than 10 if the ambient PO_4^{3-} concentration is very low at the site where Fe^{2+}
 290 crystallizes from solution.

291 The FeP pool can be preserved by ageing processes (Lijklema, 1980; Borggaard, 1991). However,
 292 crystallization decreases the mineral surface area and the sorptive capacity of PO_4^{3-} , releasing PO_4^{3-} to
 293 solution (Houben, 2003). Based on observations by Poulton and Canfield (2006), the Fe:P ratio of
 294 recrystallized iron, ϵ_{age} , is set to be a factor of 4 higher than for the highly reactive fractions (40).

295 The only allochthonous source of FeP is associated with the flux of Fe_{HR} , which requires knowledge
 296 of the Fe:P atomic ratio in iron oxides deposited on the shelf (ϵ_{all}). Most reactive iron oxide deposited
 297 there originates as colloidal coagulates formed in the low salinity region in estuaries (Boyle et al.,
 298 1977). The fraction of P that is adsorbed to these colloids decreases with increasing salinity, pH and
 299 decreasing concentrations of suspended particulate matter (Boyle et al., 1977; Bale and Morris, 1981;
 300 Lebo, 1991; Spiteri et al., 2008). Jordan et al. (2008) reported an increase in the Fe:P ratio in the
 301 Patuxent River estuary from ca. 9 in the freshwater end member to ca. 13 at a salinity of 7.
 302 Measurements in the St. Lawrence Estuary show a Fe:P ratio of around 18 to 23 close to the sea water
 303 end member (Lucotte and D'Anglejan, 1983). Further P depletion in iron oxides can occur if the
 304 suspended material has first been subjected to reductive dissolution in the estuarine sediment (Berner
 305 and Rao, 1994). These studies demonstrate a wide range of the Fe:P ratio in deposited iron oxides. For
 306 our model we assume that $\epsilon_{all} = 25$, that is, depleted in P relative to authigenic Fe_{HR} .

307 Finally, the model includes the irreversible crystallization of dispersed CFA directly from porewater
 308 PO_4^{3-} without a precursor phase (Froelich et al., 1988; Van Cappellen and Berner, 1988). We assume
 309 the following idealized stoichiometry (Froelich et al., 1988):



311 where a value of $x = 0.26$ is assigned, such that the F:P ratio in the crystallized apatite is 0.39
 312 (Froelich et al., 1988; Van Cappellen and Berner, 1988). Changes in the extent of substitution of
 313 carbonate and fluoride due to factors such as the pore fluid composition are ignored (Froelich et al.,
 314 1988). The rate of apatite precipitation is described as a kinetically-controlled process. The rate
 315 limiting factor is the concentration of PO_4^{3-} in excess of the concentration in thermodynamic
 316 equilibrium with CFA, C_{eq} (Van Cappellen and Berner, 1988). Apatite formation is inhibited if PO_4^{3-}
 317 $< C_{eq}$; taken as $10 \mu M$ (Van Cappellen and Berner, 1988). This assumes that the PO_4^{3-} concentration is
 318 the limiting factor for CFA precipitation rather than Ca^{2+} or F^- , which is reasonable since apatite is

319 more enriched in P than F, and seawater Ca^{2+} concentrations are orders-of-magnitude higher (Froelich
320 et al., 1983; Schuffert et al., 1994). A limiting term for F^- is nonetheless included in the rate
321 expression to ensure that CFA cannot crystallize if F^- is completely consumed (Table S5). The rate
322 constant for CFA precipitation is assigned to be 1 yr^{-1} based on studies in Arabian Sea sediments
323 (Kraal et al., 2012). A more rigorous expression for CFA precipitation that considers the effect of
324 porewater pH and carbonate content should be explored in future work (Jahnke et al., 1983).

325 The depositional flux of apatite to the seafloor is set to zero (Table 2). We have chosen to omit detrital
326 apatite from the model since it is believed to be unreactive (Berner et al., 1993). Vivianite
327 precipitation is also not considered since this mineral is likely to be undersaturated in margin
328 sediments (see Supplement). Estimates of the contribution by biogenic P (typically in the form of
329 CFA fish bones and scales) to porewater dissolved P in sediments are scarce. Stoichiometric
330 porewater models suggest that biogenic P contributes insignificantly to PO_4^{3-} turnover in sediments
331 outside of upwelling areas (Suess, 1981). Froelich et al. (1982) also calculated that burial of biogenic
332 fish debris is a minor sink for P at the global scale. The fact that global fish production (ca. 2.5 g m^{-2}
333 yr^{-1} , Jennings et al., 2008) is equivalent to less than 2 % of primary production (ca. $150 \text{ g m}^{-2} \text{ yr}^{-1}$,
334 Sarmiento and Gruber, 2006) seems to support this idea. For these reasons, biogenic P is considered to
335 be inconsequential and is not included in the model (Froelich et al., 1982). Nonetheless, more work is
336 needed to properly determine the importance of biogenically-derived P to benthic P fluxes.

337

338 **3. Parameterization of the phosphorus model**

339 *3.1. C_{org} and P_{org} mineralization*

340 A mechanistic understanding of P burial in sediments first requires a careful analysis of the poorly
341 understood kinetics of P_{org} mineralization. Since P_{org} burial is often discussed in terms of $C_{\text{org}}:P_{\text{org}}$
342 ratios, the first step in the modelling procedure was to constrain the mineralization of C_{org} using the
343 available database of organic carbon burial efficiencies (CBE) (Table 1).

344 The CBE in sediments on the continental margin under oxic bottom waters is $25 \pm 15 \%$ (Table 1)
345 versus $>40 \%$ for oxygen-deficient bottom waters ($\text{O}_2 < 20 \mu\text{M}$). This difference has been argued to be
346 driven by preferential preservation of C_{org} in the absence of oxygen (Demaison and Moore, 1980;
347 Hartnett et al., 1998; Hedges et al., 1999). Running the anoxic model with C_{org} mineralization defined
348 using the degradation kinetics for fresh phytoplankton (see 2. *Methods*) predicts a low CBE of 15 %.
349 These kinetics thus over-predict C_{org} mineralization in natural anoxic sediments, probably because the
350 experiments made no allowance for anaerobic mineralization (Westrich and Berner, 1984). The model
351 was therefore adjusted by decreasing the value of the parameter describing the C_{org} reactivity

352 distribution, ν , from 0.125 to 0.05, implying a greater fraction of less degradable components within
353 the bulk organic mixture. Re-running the model gives a CBE of 47% for both anoxic and oxic
354 sediments. Whilst this agrees with observations in anoxic sediments (Table 1), the predicted CBE in
355 oxic (bioturbated and bioirrigated) sediments is too high. Enhanced C_{org} degradation under oxic
356 bottom waters was thus achieved by multiplying the rate of aerobic C_{org} degradation by an
357 acceleration factor, $foxC$ (see Table S5). With $foxC = 2$, for example, the model predicts a more
358 realistic CBE of 16% for oxic sediments. This adjustment fits with the long-standing paradigm of
359 preferential mineralization under oxic conditions (Hartnett et al., 1998; Hedges et al., 1999).

360 PBE is lowest in sediments underlying oxygen deficient bottom waters, with values in the range of 2
361 to 11 % (Table 1). To achieve this with the model, it was necessary to enhance the dissolution of P_{org}
362 during anaerobic respiration *relative* to C_{org} using an acceleration factor, $fanoxP$ (Table S5). A value
363 of 1.7 provides a PBE of 12 % for the anoxic scenario that is in good agreement with previous studies
364 with bottom water $O_2 < 20 \mu M$. The corresponding regeneration ratio, $(C:P)_{REG}$, is 64, that is, P-
365 enriched relative to the Redfield ratio of the organic matter raining to the seafloor. This also agrees
366 with measured $(C:P)_{REG}$ of around 10-60 for anoxic margins (Table 1). Furthermore, the $C_{org}:P_{org}$
367 burial ratio is 423, again well within the range of the observations (300 - 600).

368 If an equivalent acceleration factor of 1.7 is used for P_{org} mineralization relative to C_{org} during aerobic
369 respiration ($foxP$), then P_{org} degradation is too high and the sediment loses too much PO_4^{3-} by
370 diffusion to the water column. The result is a PBE of only 5% and a $C_{org}:P_{org}$ burial ratio of 335. PBE
371 estimates for sediments under oxic bottom waters are scarce, and probably in the region of ca. 20 to
372 40 %, whilst burial ratios range from 30 to 115 (Table 1). Therefore, $foxP$ must be less than $fanoxP$
373 for the current model set-up. With $foxP = 0.5$, that is, slower than the rate of aerobic C_{org}
374 mineralization, the model predicts a PBE of 33% and a $C_{org}:P_{org}$ burial ratio of 53 for oxic sediments,
375 respectively, both of which are supported by measured data. Furthermore, $(C:P)_{REG}$ is 140, not
376 dissimilar to observations under oxic bottom waters of 118 ± 24 (Wallmann, 2010).

377

378 3.2. Preferential mineralization of P_{org} versus C_{org}

379 To summarize, the model can reproduce PBEs, $(C:P)_{REG}$ and $C_{org}:P_{org}$ burial ratios that are within the
380 ranges of the field data if P_{org} mineralization proceeds at roughly half the rate of C_{org} mineralization
381 during aerobic respiration and twice as fast during anaerobic respiration. In that case, we should
382 expect to see $C_{org}:P_{org}$ ratios in oxic environments that are P enriched relative to Redfield ($\ll 106:1$).
383 However, field data rather indicate that P_{org} is mineralized preferentially to C_{org} under oxic conditions.
384 For instance, the molecular composition of particulate P sinking through oxic water columns down to
385 ca. 4000 m in diverse oceanic settings has been analyzed using ^{31}P NMR (nuclear magnetic

386 resonance) spectroscopy and sequential extractions (Paytan et al., 2003; Faul et al., 2005). These
387 studies showed that phosphate diesters are preferentially solubilized, with $C_{org}:P_{org}$ ratios ranging from
388 276 to 1138 and a mean value of 318. Selective P remineralization also occurs in the dissolved
389 organic fraction. Clark et al. (1998) measured an increase in the C:P ratio of dissolved organic matter
390 with depth in the Pacific Ocean from 247:1 at the surface to 539:1 at 4000 m. Further examples of
391 preferential mineralization of P_{org} in oxic waters have been discussed by Ruttenberg (2014). The
392 current model formulation of aerobic mineralization is thus inconsistent with the observations in the
393 water column.

394 Although temporal and spatial variability of water column particle fluxes creates uncertainty in the
395 data analysis (e.g. Paytan et al., 2003), it seems reasonable to assume that preferential mineralization
396 of P_{org} relative to C_{org} continues in oxidized surface sediments. Identification of preferential P_{org}
397 mineralization in sediments is more problematic because the thickness of the diffusive oxic layer at
398 the sediment-water interface is typically only a few mm thick (Glud et al., 2008). Furthermore, P_{org}
399 dynamics are overprinted by inorganic P cycling and sediment reworking and mixing by bioturbation.
400 Diagenetic models can help to eliminate such interferences. Using such an approach, Krom and
401 Berner (1981) inferred preferential loss of P_{org} in surface sediments in Long-Island Sound by
402 comparing the calculated $C_{org}:P_{org}$ stoichiometry of mineralized organic matter versus the bulk
403 $C_{org}:P_{org}$ composition of the phytoplankton community. They further showed that P_{org} mineralization
404 in the oxic layer was more rapid than in the underlying anoxic sediments. The empirical model
405 developed by Ingall and Van Cappellen (1990) also required preferential regeneration of P_{org} to
406 simulate $C_{org}:P_{org}$ burial ratios in a wide range of marine environments. In contrast, Reed et al. (2011)
407 reported opposite trends for P turnover in a modern (seasonally) hypoxic basin. They tuned their
408 model to data assuming selective P_{org} mineralization relative to C_{org} under anoxic conditions only.

409 Despite the ambiguity that exists concerning the extent of preferential mineralization of P_{org} relative to
410 C_{org} under oxic conditions (Ruttenberg, 2014), the consensus seems to be that P_{org} is mineralized more
411 rapidly than C_{org} during the initial stages of mineralization, and certainly not more slowly.
412 Consequently, if P_{org} is indeed preferentially degraded in oxic sediments, then the PO_4^{3-} that is
413 released close to the sediment water interface must be sequestered and entrained in the sediment
414 before it can escape to the water column. Constraints provided by $C_{org}:P_{org}$ burial ratios and PBE
415 strongly indicate that this sink is organic in nature.

416

417 *3.3. Expanding the model: P sequestration by microorganisms*

418 The model is extended to include a simple microbial P cycle (Fig. 1) to investigate whether the
419 hypothesized authigenic synthesis of refractive biomolecules (see Introduction) can explain

420 preferential mineralization of P_{org} relative to C_{org} and the observed C:P burial ratios in different marine
 421 settings. Two additional species are considered; polyphosphate (P_{poly}) and unreactive microbial
 422 organic P (P_{org-U}). The microorganisms are not modeled explicitly, and P_{poly} and P_{org-U} are assumed to
 423 be particulate species transported passively through the sediments by bioturbation and burial. A
 424 deposition flux of zero is prescribed at the sediment surface for P_{poly} and P_{org-U} (Robin boundary), with
 425 a zero gradient at the bottom (Neumann boundary).

426 Microorganisms assimilate porewater PO_4^{3-} in aerobic and nitrogenous sediments leading to an
 427 increase in the P_{poly} pool (Davelaar, 1993). A fraction (f_{poly}) of P_{poly} is converted to P_{org-U} to simulate
 428 the synthesis of refractive P biomolecules. This fraction is buried without further reaction and
 429 constitutes a microbial P pump that removes bioavailable P from the ocean. The remainder ($1 - f_{poly}$) is
 430 recycled back to the bulk P_{org} pool on the understanding that labile dead bacterial biomass is
 431 remineralized. P_{poly} that is transported into anaerobic sediment layers is rapidly hydrolyzed back to
 432 PO_4^{3-} . This simulates the breakdown of P_{poly} by bacteria to gain energy to survive anaerobiosis as well
 433 as the lack of P_{poly} in anoxic sediments (Gächter and Meyer, 1993; Davelaar, 1993; Sannigrahi and
 434 Ingall, 2005; Schulz and Schulz, 2005; Brock and Schulz-Vogt, 2010). Phosphate that is released to
 435 the porewater can be trapped in authigenic phases (sink switching) or diffuse upwards and re-
 436 assimilated by bacteria as P_{poly} . Transformation of P_{poly} into CFA is not considered (Diaz et al., 2008),
 437 although we recognize that this may take place in oxygen minimum zones (Goldhammer et al., 2010).

438 The proposed cycle is by necessity highly simplified since the distribution of different classes of
 439 organically-bound P molecules in sediments is not well understood and polyphosphates are typically
 440 extracted in operationally defined P_{org} pools (Sannigrahi and Ingall, 2005; Diaz et al., 2012). X-ray
 441 spectromicroscopy has nonetheless revealed that polyphosphates are an important P phase at oxic-
 442 anoxic interfaces in marine sediments, and that they may constitute a larger fraction of the
 443 operationally-defined detrital P than previously assumed (Diaz et al., 2008; Kraal et al., 2015). With
 444 this in mind, the following equation describing the rate of change of P_{poly} in the sediment is proposed:

$$445 \quad \frac{dP_{poly}}{dt} = + \text{synthesis} - \text{hydrolysis} - P_{org-U} \text{ synthesis} - P_{org} \text{ synthesis} \quad (11)$$

446 The kinetics and parameterization of these pathways are only vaguely constrainable by field and
 447 experimental observations. We make the assumption that the rate of P_{poly} synthesis (R_{30}) is dependent
 448 on the concentration of PO_4^{3-} in the porewater as well as the concentration of NO_X ($NO_3^- + NO_2^-$):

$$449 \quad R_{30} = k_{30} \cdot PO_4^{3-} \cdot NO_X \quad (12)$$

450 where k_{30} is the rate constant. No dependency on O_2 is included because it is exhausted before NO_X ,
 451 such that the sediment penetration depth of NO_X is higher than for O_2 (discussed below). The impact
 452 of C limitation on P_{poly} synthesis is not considered (Steenbergh et al., 2012).

453 Hydrolysis of P_{poly} is defined using a dimensionless oxygen threshold, δ :

$$454 R_{31} = k_{31} \cdot P_{poly} \cdot \delta \quad (13)$$

455 where $\delta = 0$ for $O_2 \geq 1 \mu M$ and $\delta = 1$ for $O_2 < 1 \mu M$. Transformation of P_{poly} to P_{org} (R_{32}) and P_{org-U}
456 (R_{33}) are described as first-order processes:

$$457 R_{32} = k_{32} \cdot (1 - f_{poly}) \cdot P_{poly} \quad (14)$$

$$458 R_{33} = k_{33} \cdot f_{poly} \cdot P_{poly} \quad (15)$$

459 The rate coefficients k_{31} to k_{33} are set to high values of 50 yr^{-1} . Rapid turnover of P_{poly} has been
460 inferred from the transient development of large subsurface PO_4^{3-} peaks attributed to the breakdown
461 of bacterial P_{poly} (Schulz and Schulz, 2005; Dale et al., 2013). The rate limiting step in the microbial P
462 cycle is P_{poly} synthesis, such that the turnover of microbial P pump can be regulated by adjusting k_{30}
463 whereas the fraction channeled into refractive P_{org-U} is determined by f_{poly} . Fast transformation of PO_4^{3-}
464 into P_{poly} by sulfur oxidizing bacteria has been measured using ^{33}P tracer (Goldhammer et al., 2010).
465 We acknowledge the possibility, and likelihood, that a different set of parameters or microbial P
466 cycling pathways could lead to equivalent results. Our main interest here is whether P_{org} synthesis by
467 bacteria is a concept that is supported by the data.

468 On the understanding that aerobic P_{org} mineralization is more rapid than C_{org} mineralization, we
469 initially set $foxP$, the enhancement of aerobic P_{org} mineralization relative to C_{org} , equal to 2. We then
470 adjusted the rate of P_{poly} synthesis to bring the model back into the range of the observations.

471

472 **4. Results**

473 The model with the microbial P pump is able to simulate the entire suite of field observations in oxic
474 and anoxic sediments with $k_{30} = 3 \times 10^7 \text{ M}^{-1} \text{ yr}^{-1}$ and a slightly higher $fanoxP$ of 2 (Table 2 and 3). The
475 latter adjustment from the previous value of 1.7 compensates for the limited amount of microbial P_{org}
476 that is synthesized in the nitrogenous layer in the anoxic sediment. Furthermore, f_{poly} is equal to 0.25,
477 such that 25 % of the P_{poly} pool that is not hydrolyzed is converted to P_{org-U} and 75 % to P_{org} . The
478 relative fractionation of P_{poly} into labile and unreactive P_{org} depends somewhat on the poorly
479 constrained values of $foxP$. Whilst there is obviously some latitude in these numbers, we note that the
480 model would not be able to simulate the database if P_{poly} were entirely converted to P_{org} or P_{org-U} alone.
481 An important conclusion from this model set-up is that P_{org} is preferentially mineralized relative to
482 C_{org} during aerobic *and* anaerobic respiration.

483 The mass fluxes in Fig. 2 show that P turnover is more intense in oxic sediments. P_{org} mineralization
484 rates are $136 \mu\text{mol m}^{-2} \text{d}^{-1}$, versus $89 \mu\text{mol m}^{-2} \text{d}^{-1}$ in anoxic sediments. This is partly compensated by
485 higher rates of PO_4^{3-} sequestration into P_{poly} (82 versus $19 \mu\text{mol m}^{-2} \text{d}^{-1}$). By way of comparison,
486 measured P sequestration rates by sulfur oxidizing bacteria are of a similar magnitude (210 to 780
487 $\mu\text{mol m}^{-2} \text{d}^{-1}$; Goldhammer et al., 2010). As a result, PO_4^{3-} concentrations are lower in oxic sediments
488 and P_{org} and P_{reac} concentrations are higher (green and black curves, Fig. 3). Identical qualitative
489 trends have been observed at neighboring oxic and anoxic sites in Effingham Inlet (Ingall et al., 2005;
490 Sannigrahi and Ingall, 2005). Removing the microbial P pump leads to complete dissolution of P_{org} in
491 oxic sediments and low accumulation of other authigenic P phases (red curves, Fig. 3).

492

493 5. Discussion

494 5.1. P cycling with the microbial P pump

495 The simulated P dynamics considering microbial P sequestration and burial are consistent with the
496 observed burial efficiencies, benthic fluxes and burial ratios in diverse marine settings, and suggest an
497 important microbial control of P cycling in sediments. Permanent microbial P sequestration can be
498 viewed as a sink-switching mechanism that permits preferential burial of P relative to C_{org} (Ruttenberg
499 and Berner, 1993; Filippelli and Delaney, 1996; Anderson et al., 2001). Microorganisms may,
500 therefore, act as a barrier to PO_4^{3-} fluxes in the same way as P sequestration by iron oxides and apatite
501 crystallization (Sundby et al., 1992; Gächter and Meyer, 1993; Anschutz et al., 1998; Slomp et al.,
502 1996; Sannigrahi and Ingall, 2005). Microbial sink-switching leads to $C_{org}:P_{reac}$ burial ratios of 44 and
503 196 in oxic and anoxic sediments, respectively (Table 3). These agree well with those measured
504 above and within Mediterranean sapropels (21 versus $54 - 161$; Slomp et al., 2004) and with values of
505 140 ± 50 and 180 ± 90 from (anoxic) black shale sequences from the Cayuga Basin and Yorkshire
506 coast (Ingall et al., 1993; Anderson et al., 2001).

507 In the current oxic configuration, PO_4^{3-} trapping into iron oxides (ca. $9 \mu\text{mol m}^{-2} \text{d}^{-1}$; Fig. 2) removes a
508 relatively small fraction of PO_4^{3-} from the porewater compared to microbial P. This result arises from
509 tuning the model to the observational database in Table 1, and involves no tweaking of other
510 parameters to lessen the importance of Fe-P cycling. The iron cycle has been previously constrained
511 from a global empirical database (Dale et al., 2015a). A lesser contribution (quantitatively speaking)
512 of iron-associated P cycling on P burial relative to P_{org} is consistent with previous findings from some
513 settings (Ruttenberg and Berner, 1997; Ingall and Jahnke, 1997; Sannigrahi and Ingall, 2005) but not
514 others (Jensen et al., 1995; Slomp et al., 1996). This can be attributed potentially to numerous factors,
515 such as regional variability in external iron inputs, bottom water O_2 levels, and bioirrigation and
516 bioturbation intensities. Model sensitivity analysis (not shown) reveals that the fraction of P that is

517 buried in association with iron oxides is particularly sensitive to the Fe:P molar ratio in the iron
518 fraction of Fe_{MR} that is crystallized from Fe_{HR} by ageing (ϵ_{age}). This ratio is set to 40, compared to a
519 molar ratio of 10 for freshly precipitated Fe_{HR} . The higher value represents an increase in the
520 crystallinity of Fe_{MR} relative to Fe_{HR} and subsequent loss of P binding sites (see 2. *Model*). The rate of
521 PO_4^{3-} that is sequestered into Fe_{MR} by ageing is proportional to ϵ_{age} , such that a halving of ϵ_{age} doubles
522 the burial flux of P associated with Fe_{MR} . Thus, precipitation and ageing of FeP could conceivably
523 sequester more PO_4^{3-} than currently assumed by the model. It cannot, however, explain the entire sink
524 for PO_4^{3-} because this would be inconsistent with the $\text{C}_{\text{org}}:\text{P}_{\text{org}}$ burial ratios and P_{org} burial efficiencies.

525 The fact that total P_{org} in oxic settings is roughly twice that in anoxic sediments demonstrates that
526 recalcitrant microbial P_{org} constitutes around half of the P_{org} buried below the bioturbated layer (Fig.
527 3). A higher turnover of microbial biomass under oxic versus anoxic conditions has been noted before
528 (Sun et al., 2002), and conceivably could be coupled to a more intense microbial P pump. In this
529 study, we did not explicitly examine the $\text{C}_{\text{org}}:\text{P}_{\text{org}}$ ratio of the microorganisms or simulate microbial
530 C_{org} . The $\text{C}_{\text{org}}:\text{P}_{\text{org}}$ ratio of bacteria has been often cited to be between 29 and 63 (Berner et al., 1993),
531 which is close to the simulated ratio in oxic sediments of 73 (Table 3). Yet, X-ray microanalysis on
532 intact cells from surface sediment samples in the Baltic Sea revealed high bacterial $\text{C}_{\text{org}}:\text{P}_{\text{org}}$ ratios up
533 to 400, although this may be related to substrate (C_{org}) limitation (Steenbergh et al., 2012). This type
534 of quantitative information could serve to further validate the model if the degradation kinetics of
535 cellular P_{org} were known, for both oxic and anoxic sediments. Compared to FeP and CFA, microbial P
536 cycling in marine sediments has hardly been studied, resulting in a severe lack of information on the
537 genesis and reactivity of sedimentary P_{org} . Consequently, the significance of the benthic microbial P
538 pump to the global P cycle can only be inferred currently using observations of the bulk P pool.

539

540 5.2. The impact of animals on P burial

541 Our results show that sediments on the modern marine shelf that are mixed and ventilated by
542 bioturbation and bioirrigation are around 50 % more efficient at retaining P than sediments under
543 anoxic bottom waters, with net P_{reac} accumulation rates of 35 and 23 $\mu\text{mol m}^{-2} \text{d}^{-1}$, respectively (Fig.
544 2). The P_{reac} burial flux scales positively with O_2 availability (solid black curve, Fig. 4a; the grey
545 shaded area shows the range of P burial expected for all values of f_{poly}), and the effect of O_2 is stronger
546 for sediments that are more intensively reworked by bioturbation and bioirrigation (dashed black
547 curves). The direct contribution of bioturbation and bioirrigation is shown by observing the lower P_{reac}
548 accumulation rates in simulations without benthic infauna (dashed blue curves in Fig. 3 and Fig. 4a).
549 It is interesting to note that here the burial flux initially decreases for $\text{O}_2 < 20 \mu\text{M}$ due to a weak
550 microbial P pump and preferential dissolution of P_{org} by aerobic respiration (Fig. 4a). A reversal in
551 P_{reac} burial when $\text{O}_2 > 20 \mu\text{M}$ marks the point where synthesis and burial of $\text{P}_{\text{org-U}}$ begins to mitigate

552 the PO_4^{3-} flux to the bottom water. P_{reac} burial in model runs with mixing but without microbial P
553 uptake are even lower (dashed red curves in Fig. 3 and Fig. 4).

554 These results demonstrate that enhanced P burial may not only be driven by O_2 availability, as
555 assumed in some modeling studies (Van Cappellen and Ingall, 1994; Slomp and Van Cappellen,
556 1994), but requires the intervention of P-storing microorganisms and animals to churn and ventilate
557 the sediment. Enhanced P burial in oxic sediments with faunal communities is explained
558 mechanistically as follows. First, bioturbation limits diffusive PO_4^{3-} fluxes to the bottom water by
559 shunting highly labile P_{org} away from the aerobic sediment layers where mineralization rates are
560 elevated. This can be appreciated by noting that P_{org} mineralization rates in the top millimeter of
561 bioturbated sediments are several-fold lower than for non-bioturbated sediments, whereas the opposite
562 is true below this depth (Fig. 5). Sub-surface mineralization of P_{org} helps to trap P within the sediment
563 by sink-switching. Without bioturbation, P_{org} can only be transported downwards by the relatively
564 slow process of burial, allowing more P_{org} to be solubilized to PO_4^{3-} at the sediment-water interface
565 and released back to the ocean. Slomp et al. (1996) similarly proposed that P trapping in North
566 Atlantic slope sediments is favored by the downward mixing of iron-bound P by infauna followed by
567 sink-switching to CFA. Our result is partly dependent on the reverse conveyor belt feeding
568 mechanism of bioturbation (see 2. *Model*). It is likely, however, that other bioturbation formalisms
569 result in a similar redistribution of labile P_{org} , although we have not tested these explicitly.

570 Second, seawater pumping into the sediment by bioirrigation enhances nitrification and creates a more
571 favorable niche for P-storing bacteria by deepening the nitrogenous sediment zone relative to
572 oxidized, non-mixed sediments (compare green and dashed blue curve, Fig. 3b). This result is partly
573 dependent on the kinetics of nitrate (NO_x) consumption and the irrigation coefficient, $\alpha(0)$, that are
574 somewhat poorly constrained at the global scale. Deeply-burrowing polychaetes can result in large
575 nitrate penetration depths (Dale et al., 2011; Renz and Forster, 2014), although not all irrigated
576 sediments show the same features (Devol and Christensen, 1993; Dale et al., 2014). In reality, patterns
577 of solute transport in and around burrows are much more complex than portrayed with the simple
578 irrigation model used here. The thickness of the oxic and nitrogenous zones will depend on many
579 physical and biological factors such as the sediment permeability, bottom water currents and the
580 ecology of each animal species. Nonetheless, the deeper nitrogenous zone with versus without
581 animals supports the proposed rate formulation for P_{poly} synthesis (Eq. (12)). Again, it is important to
582 stress that Eq. (12) is purely conceptual, although it does conform to the known redox window over
583 which microbial P sequestration occurs (Davelaar, 1993). It should also be noted that O_2
584 concentrations with animals are also on average 50 % higher than in the case without them (Fig. 3a).
585 Inclusion of O_2 into Eq. (12), for example by making P_{poly} synthesis dependent on the sum of $\text{NO}_x +$
586 O_2 , would require recalibration of the rate constant, k_{30} , but would not fundamentally alter the main
587 results.

588

589 5.3. Implications for *P* burial in the early Paleozoic

590 We modified the model to assess the dynamics of *P* burial in the early Paleozoic to test the hypothesis
591 advanced by Boyle et al. (2014) that colonization of the continental shelves by bioturbating and
592 burrowing animals at some time between the Ediacaran and early Paleozoic resulted in enhanced *P*
593 burial. A more nuanced view is now emerging in the palaeontological literature, in which the full
594 scale of the “Cambrian substrate revolution” may have been delayed, perhaps until as late as the
595 Ordovician-Silurian boundary or late Silurian (Tarhan et al., 2015). In other words, there may have
596 been a delay between the appearance of the first trace fossils and quantitatively significant mixing of
597 continental shelf sediments (Tarhan and Droser, 2014).

598 Although the timing and trajectory of the spread of faunal invasion of sediments is still being worked
599 out, it seems certain that at some point between ca. 550 Ma and 420 Ma the bioturbation intensity and
600 burrow depth increased. There is high uncertainty in assigning real numbers to these parameters, and
601 the consensus on this issue is evolving. Mángano and Buatois (2014) reported an increase in
602 maximum burrow depth from around 1 to 6 cm across the Ediacaran/ Paleozoic boundary based on
603 trace fossil specimen analysis. In contrast, sedimentological data suggest that burrow depths in the
604 lower Cambrian never exceeded 3 cm, and bed thicknesses point toward millimetre-scale mixing
605 depths in the lower to middle Cambrian (ca. 542 – 507 Ma) increasing to 1 – 2 cm in the Ordovician –
606 Silurian (ca. 450 – 420 Ma) (Tarhan and Droser, 2014). Clearly, sediment mixing and burrowing
607 depths during the early Cambrian were significantly lower than they are on ocean margins today.

608 In accordance with these latest data, the mixed depth was reduced to 0.5 cm ($z_{bt} = 0.25$ cm).
609 Bioirrigation was also reduced so that the burrow flushing intensity at 2 cm depth was ~10% of the
610 surface value ($z_{bio} = 1$ cm). *P* burial is insensitive to lower z_{bio} values since solute transport in the
611 uppermost sediment layer is dominated by diffusive exchange with the bottom water. To account for
612 the lower bioturbation intensity in sediments this time, the sediment mixing intensity ($D_b(0)$) was
613 decreased arbitrarily from 27 to 5 cm² yr⁻¹ and bioirrigation ($\alpha(0)$) from 465 y⁻¹ to 50 y⁻¹. Due to the
614 near-impossibility of constraining these latter values accurately, the model focuses on a qualitative
615 comparison between sediments with and without indwelling fauna in the early Paleozoic.

616 The paleo simulations also account for the different boundary conditions at the sediment-water
617 interface. Geochemical evidence that is interconnected with ancient ocean redox status allows some
618 essential features to be reasonably well described (Och and Shields-Zhou, 2012). The deep ocean in
619 the early Paleozoic was ferruginous, with euxinic conditions restricted to the biologically productive
620 margins (Canfield et al., 2008; Li et al., 2010; Lyons et al., 2014). The shallow seas, in contrast, may

621 have been similar to productive regions in the contemporary ocean and experienced oxic bottom
622 waters (Kendall et al., 2012; Lyons et al., 2014).

623 Constraints on global oxygen levels are qualitative at best for this period of Earth's history. Late
624 Ediacaran / early Paleozoic O₂ concentrations and may have been several tens of μM (Canfield et al.,
625 2007; Bjerrum and Canfield, 2011), with anoxic episodes persisting well into the early Paleozoic
626 (Saltzman et al., 2015). As a reflection of this uncertainty, we assessed P burial over the same range
627 of bottom water O₂ concentrations as previously (0 to 150 μM).

628 Sulfate (SO₄²⁻) concentrations in the early Paleozoic ocean were also much lower than today.
629 Concentrations following the Great Oxygenation Event (2.4 Ga) may have remained at only 1 – 3 mM
630 until the Cambrian or even later, when bioturbation itself may have contributed to a rise in SO₄²⁻ via
631 increased oxidation of sedimentary sulfide (Canfield and Farquhar, 2009; Li et al., 2010; Tarhan et al.,
632 2015). We used a sulfate concentration of 3 mM, and kept all other model parameters at their modern
633 values, including the sedimentation rate. This assumption relates to fact that continental weathering
634 rates at the end of the Ediacaran appear to have been similar to those of today (Maloof et al., 2010;
635 Peters and Gaines, 2012). Sensitivity analysis (not shown) indicates that P burial scales proportionally
636 with sedimentation rate.

637 Knowledge of the flux and lability of organic matter deposited on the sea floor in the early Paleozoic
638 is completely lacking, and we prescribe the contemporary values in the absence of better information.
639 Given that photosynthetic eukaryotes evolved more than 1500 Ma, the biochemical composition and
640 cell size of photosynthesizing biota may have been comparable to today, even though species
641 diversity was lower (Falkowski et al., 2004; Butterfield, 2007; Och and Shields-Zhou, 2012). In that
642 case, particle fluxes would be still be eukaryote-dominated, enhanced by aggregation and ballasting
643 resembling the modern ocean (Lenton et al., 2014). Running the model using the same organic matter
644 flux and reactivity as for the modern scenario will elucidate the impact of ocean chemistry and
645 sediment mixing on P burial. These assumptions can be relaxed if more data becomes available.

646 Model results of the early Paleozoic scenario with bottom water O₂ concentrations of 25 μM show
647 that, in general, the redox structure is not dramatically different from the modern scenario (Fig. S1
648 and S2). Notably, though, the aerobic and nitrogenous layers are less well developed due to the lower
649 bottom water O₂ concentrations, sediment mixing and irrigation rates. Over the range of O₂ levels
650 tested, this causes a decrease in P burial relative to the modern setting mainly as result of lower
651 microbial P_{org} accumulation (black curve in Fig. 4b and Supplement). The results nonetheless show
652 that, even with a much thinner and weakly mixed surface layer, the onset of bioturbation at the low O₂
653 concentrations of some 10s μM believed to characterize this time period (Canfield et al., 2007;
654 Bjerrum and Canfield, 2011) would have increased P burial, shown schematically by the blue arrow in

655 Fig. 4b. Further increases in the depth of the mixed layer in the late Cambrian lead to even greater
656 burial of P (black arrow).

657 These results are of qualitative value only due to the assumptions involved in constraining the
658 boundary conditions in ancient sediments. We also assumed for simplicity that the physical aspects of
659 bioturbation, such as particle disaggregation and modification of the sediment fabric as well as
660 feeding mode (local versus non-local), resemble modern sediments. We tested the effect of changes in
661 sediment water content arising from colonization by indwelling fauna. Deckere et al. (2001) observed
662 a lower water content of defaunated versus faunated sediment experiments, and other natural
663 observations have shown that seasonal anoxia and lack of bioirrigation reduces sediment porosity by
664 around 10% (Dale et al., 2013). Anoxic baseline simulations with a 10% lower porosity show a
665 modest decrease in P_{reac} burial from 23 to 18 $\mu\text{mol m}^{-2} \text{d}^{-1}$, driven by an increase in solid fraction
666 mineralization close to the sediment water interface and subsequent loss of PO_4^{3-} . Early sediment
667 colonization may thus have had a greater effect on P_{reac} burial than the baseline model currently
668 predicts.

669 The oceanic P residence time (ca. 20 kyr, Wallmann, 2010) is relatively short compared to the $>10^5$
670 years needed for the stabilizing feedbacks on the oceanic P inventory via redox-dependent P burial
671 (Van Cappellen and Ingall, 1994; Lenton and Watson., 2000). Thus, the radiation of indwelling
672 sediment fauna would have helped to lower the oceanic bioavailable P pool over these long time
673 scales, leading to a negative and stabilizing feedback on atmospheric O_2 levels via primary production
674 (Boyle et al., 2014). This could have contributed to a wider stability in the Earth system that may have
675 been necessary for the proliferation of recently developed animal forms that characterize the early
676 Cambrian (Butterfield, 2007; 2009). The results presented here show mechanistically how animals
677 may have ‘engineered’ the physical, geochemical and microbiological structure of their habitats,
678 leading to stabilizing negative feedbacks with the O_2 levels upon which they depend, with wider
679 implications for the coevolution of the earth system and animal forms (Mermillod-Blondin and
680 Rosenberg, 2006; Butterfield, 2007; 2009; Boyle et al., 2014). In combination with microbial catalysis
681 of bioavailable P into organic compounds, animals very likely exert an important control on the
682 cycling of burial of P in the modern and ancient ocean.

683 **5. Conclusions**

684 Observed C and P burial and recycling rates from the literature interpreted with a diagenetic model
685 strongly suggest that uptake and recycling of P by microorganisms enhances P burial in bioturbated
686 and bioirrigated marine sediments alongside iron-associated phosphorus and authigenic apatite
687 formation. Whilst microbial cycling of P has been recognized for many years by limnologists, and
688 inferred from $C_{\text{org}}:P_{\text{org}}$ burial ratios and benthic fluxes by marine geochemists, redox-dependent
689 storage and burial of refractory microbial P (termed the microbial P pump) has thus far not been

690 included in diagenetic models of P cycling in marine sediments. Microbial P sequestration in
691 conjunction with bioturbation intensity and bottom water O₂ levels (summarized in Fig. 4) should be
692 considered when parameterizing redox-dependent benthic P release in global biogeochemical models.
693 O₂ concentration by itself may be insufficient to explain benthic phosphate fluxes in periods in Earth's
694 history when bioturbating and burrowing fauna colonized or recolonized the seafloor. Our results
695 support the notion that the presence of P_{org} in ancient marine sediments could, in part, be due to the
696 synthesis of refractive microbial compounds in surface sediments.

697 More accurate information regarding the controls on microbial sequestration of porewater PO₄³⁻ by,
698 for example, organic carbon content and reactivity, sedimentation rate and bottom water redox
699 conditions and well as physiological controls on P uptake in natural sediments awaits further study.
700 New approaches to quantify microbial P biomarkers or polyphosphates such as X-ray spectroscopy
701 (XANES) will be invaluable to expand the database on the microbial P content of surface sediments
702 in settings displaying a range of bottom water redox conditions (e.g. Kraal et al., 2015).

703

704 **Acknowledgements**

705 The authors thank Lidya Tarhan and an anonymous referee for comprehensive and thoughtful reviews
706 and Tim Lyons for the editorial handling. This work is a contribution of the Sonderforschungsbereich
707 754 "Climate – Biogeochemistry Interactions in the Tropical Ocean" (www.sfb754.de) which is
708 supported by the Deutsche Forschungsgemeinschaft.

709

710 **References**

- 711 Algeo, T. J. and Ingall, E. (2007) Sedimentary Corg:P ratios, paleocean ventilation, and Phanerozoic
712 atmospheric pO₂. *Palaeogeogr. Palaeoclimatol. Palaeoecol.* **256**, 130-155.
- 713 Aller, R. C. and Aller, J. Y. (1992) Meiofauna and solute transport in marine muds. *Limnol. Oceanogr.* **37**,
714 1018-1033.
- 715 Anderson, L. D., Delaney, M. L. and Faul, K. L. (2001) Carbon to phosphorus ratios in sediments: Implications
716 for nutrient cycling. *Glob. Biogeochem. Cy.* **15**, 64-79.
- 717 Anschutz, P., Zhong, S. and Sundby, B. (1998) Burial efficiency of phosphorus and the geochemistry of iron in
718 continental margin sediments. *Limnol. Oceanogr.* **43**, 53-64.
- 719 Archer, D. E., Morford, J. L. and Emerson, S. R. (2002) A model of suboxic sedimentary diagenesis suitable for
720 automatic tuning and gridded global domains. *Glob. Biogeochem. Cy.* **16**, 1017,
721 doi:10.1029/2000GB001288.
- 722 Bale, A. J. Morris, A. W. (1981) Laboratory simulation of chemical processes induced by estuarine mixing: The
723 behaviour of iron and phosphate in estuaries. *Estuar. Coast. Shelf Sci.* **13**, 1-10.
- 724 Benitez-Nelson, C. R. (2000) The biogeochemical cycling of phosphorus in marine systems. *Earth-Sci. Rev.* **51**,
725 109-135.
- 726 Berner, R. A. (1980) *Early Diagenesis: A Theoretical Approach*. Princeton University Press, Princeton, 241 pp.
- 727 Berner, R. A. and Rao, J. L. (1994) Phosphorus in sediments of the Amazon River and estuary: Implications for
728 the global flux of phosphorus to the sea. *Geochim. Cosmochim. Acta* **58**, 2333-2339.

729 Berner, R. A., Ruttenger, K. C., Ingall, E. D. and Rao, J. L. (1993) The nature of phosphorus burial in modern
730 marine sediments. In *Interactions of C, N, P and S Biogeochemical Cycles and Global Change*, NATO ASI
731 Ser., Ser. I, vol. 4, (eds. R. Wollast, F. T. Mackenzie, and L. Chou), Springer, Berlin, pp. 365– 378.

732 Bertics, V. J., Sohm, J. A., Treude, T., Chow, C. E. T., Capone, D. G. Fuhrman, J. A., and Ziebis, W. (2010)
733 Burrowing deeper into benthic nitrogen cycling: The impact of bioturbation on nitrogen fixation coupled to
734 sulfate reduction, *Mar. Ecol. Prog. Ser.*, **409**, 1–15.

735 Bjerrum, C. J. and Canfield, D. E. (2011) Towards a quantitative understanding of the late Neoproterozoic
736 carbon cycle. *Proc. Natl. Acad. Sci. U.S.A.* **108**, 5542-5547.

737 Borggaard, O. K. (1991) Effects of phosphate on iron oxide dissolution in ethylenediamine-N,N,N'-tetraacetic
738 acid and oxalate. *Clays Clay Miner.* **39**, 324-328.

739 Boudreau, B. P. (1986) Mathematics of tracer mixing in sediments; I. Nonlocal mixing and biological conveyor-
740 belt phenomena. *Am. J. Sci.* **286**, 161-198.

741 Boudreau, B. P. (1997) *Diagenetic Models and Their Implementation: Modelling Transport and Reactions in*
742 *Aquatic Sediments*, Springer-Verlag, Berlin, 414 pp.

743 Boudreau, B. P., Arnosti, C., Jorgensen, B. B. and Canfield, D. E. (2008) Comment on "Physical Model for the
744 Decay and Preservation of Marine Organic Carbon". *Science* **319**, 1616-1617.

745 Boudreau, B. P. and Ruddick, B. R. (1991) On a reactive continuum representation of organic matter diagenesis.
746 *Am. J. Sci.* **291**, 507-538.

747 Boyle, R. A., Dahl, T. W., Dale, A. W., Shields-Zhou, G. A., Zhu, M., Brasier, M. D., Canfield, D. E. and
748 Lenton, T. M. (2014) Stabilization of the coupled oxygen and phosphorus cycles by the evolution of
749 bioturbation. *Nat. Geosci.* **7**, 671-676.

750 Brock, J. and Schulz-Vogt, H. N. (2010) Sulfide induces phosphate release from polyphosphate in cultures of a
751 marine *Beggiatoa* strain. *ISME J.* **5**, 497-506.

752 Burdige, D. J. (2006) *Geochemistry of Marine Sediments*, Princeton Univ. Press, Princeton, N. J.

753 Burdige, D. J. (2007) Preservation of organic matter in marine sediments: controls, mechanisms, and an
754 imbalance in sediment organic carbon budgets? *Chem. Rev.* **107**, 467-485.

755 Butterfield, N. J. (2009) Oxygen, animals and oceanic ventilation: an alternative view. *Geobiology* **7**, 1-7.

756 Canfield, D. E. and Farquhar, J. (2009) Animal evolution, bioturbation, and the sulfate concentration of the
757 oceans. *Proc. Natl. Acad. Sci. U.S.A.* **106**, 8123-8127.

758 Canfield, D. E., Poulton, S. W., Knoll, A. H., Narbonne, G. M., Ross, G., Goldberg, T. and Strauss, H. (2008)
759 Ferruginous conditions dominated later Neoproterozoic deep-water chemistry. *Science* **321**, 949-952.

760 Canfield, D. E., Poulton, S. W. and Narbonne, G. M. (2007) Late-Neoproterozoic deep-ocean oxygenation and
761 the rise of animal life. *Science* **315**, 92-95.

762 Canfield, D. E., Raiswell, R. and Bottrell, S. (1992) The reactivity of sedimentary iron minerals toward sulfide.
763 *Am. J. Sci.* **292**, 659-683.

764 Clark, L. L., Ingall, E. D. and Benner, R. (1998) Marine phosphorus is selectively remineralized. *Nature* **393**,
765 426.

766 Colman, A. S. And Holland, H. D. (2000) The global diagenetic flux of phosphorus from marine sediments to
767 the oceans: redox sensitivity and the control of atmospheric oxygen levels. In *Marine Authigenesis: From*
768 *Global to Microbial*, SEPM Special Publication No. 66. 53-75.

769 Colman, A. S., Mackenzie, F. T. and Holland, H. D. (1997) Redox stabilization of the atmosphere and oceans
770 and marine productivity. *Science* **275**, 406-408.

771 Dale, A. W., Sommer, S., Bohlen, L., Treude, T., Bertics, V. J., Bange, H. W., Pfannkuche, O., Schorp, T.,
772 Mattsdotter, M. and Wallmann, K. (2011) Rates and regulation of nitrogen cycling in seasonally hypoxic
773 sediments during winter (Boknis Eck, SW Baltic Sea): sensitivity to environmental variables. *Est. Coast.*
774 *Shelf Sci.* **95**, 14-28.

775 Dale, A. W., Bertics, V. J., Treude, T., Sommer, S. and Wallmann, K. (2013) Modeling benthic–pelagic nutrient
776 exchange processes and porewater distributions in a seasonally hypoxic sediment: evidence for massive
777 phosphate release by *Beggiatoa*? *Biogeosciences* **10**, 629-651, doi:10.5194/bg-10-629-2013.

778 Dale, A. W., Nickelsen L., Scholz, F., Hensen, C., Oeschlies, A. and Wallmann, K. (2015a) A revised global
779 estimate of dissolved iron fluxes from marine sediments. *Glob. Biogeochem. Cy.* **29**, doi:
780 10.1002/2014GB005017.

781 Dale, A. W., Sommer, S., Lomnitz, U., Montes, I., Treude, T., Liebetrau, V., Gier, J., Hensen, C., Dengler, M.,
782 Stolpovsky, K., Bryant, L. D. and Wallmann, K. (2015b) Organic carbon production, mineralisation and
783 preservation on the Peruvian margin. *Biogeosciences* **12**, 1537-1559.

784 Davelaar, D. (1993) Ecological significance of bacterial polyphosphate metabolism in sediments. *Hydrobiologia*
785 **253**, 179-192.

786 De Deckere, E., Tolhurst, T. J. and de Brouwer, J. F. C. (2001) Destabilization of cohesive intertidal sediments
787 by infauna. *Estuar. Coast. Shelf Sci.* **53**, 665-669.

788 Delaney, M. L. (1998) Phosphorus accumulation in marine sediments and the oceanic phosphorus cycle. *Glob.*
789 *Biogeochem. Cy.* **12**, 563–572, doi:10.1029/98GB02263.

790 Demaison, G. J. and Moore, G. T. (1980) Anoxic environments and oil source bed genesis. *Am. Assoc. Pet.*
791 *Geol. Bull.* **64**, 1179-1209.

792 Devol, A. H. and Christensen, J. P. (1993) Benthic fluxes and nitrogen cycling in sediments of the continental
793 margin of the eastern North Pacific. *J. Mar. Res.* **51**, 345-372.

794 Diaz, J., Ingall, E., Benitez-Nelson, C., Paterson, D., de Jonge, M. D., McNulty, I. and Brandes, J. A. (2008)
795 Marine polyphosphate: a key player in geologic phosphorus sequestration. *Science* **320**, 652-655.

796 Diaz, J. M. and Ingall, E. D. (2010) Fluorometric quantification of natural inorganic polyphosphate. *Environ.*
797 *Sci. Technol.* **44**, 4665-4671.

798 Diaz, J. M., Ingall, E. D., Snow, S. D., Benitez-Nelson, C. R., Taillefert, M. and Brandes, J. A. (2012) Potential
799 role of inorganic polyphosphate in the cycling of phosphorus within the hypoxic water column of Effingham
800 Inlet, British Columbia. *Glob. Biogeochem. Cy.* **26**, GB2040, doi:10.1029/2011GB004226.

801 Falkowski, P. G., Katz, M. E., Knoll, A. K., Quigg, A., Raven, J. A., Schofield, O. and Taylor, F. J. R. (2004)
802 The evolution of modern eukaryotic phytoplankton. *Science* **305**, 354-360.

803 Faul, K. L., Paytan, A. and Delaney, M. L. (2005) Phosphorus distribution in sinking oceanic particulate matter.
804 *Mar. Chem.* **97**, 307-333.

805 Filippelli, G. M. (2001) Carbon and phosphorus cycling in anoxic sediments of the Saanich Inlet, British
806 Columbia. *Mar. Geol.* **174**, 307-321.

807 Filippelli, G. M. and Delaney, M. L. (1996) Phosphorus geochemistry of equatorial Pacific sediments. *Geochim.*
808 *Cosmochim. Acta* **60**, 1479-1495.

809 Fossing, H., Ferdelman, T. G., and Berg, P. (2000) Sulfate reduction and methane oxidation in continental
810 margin sediments influenced by irrigation (South-East Atlantic off Namibia). *Geochim. Cosmochim. Acta*
811 **64**, 897-910.

812 Froelich, P., Arthur, M. A., Burnett, W. C., Deakin, M., Hensley, V., Jahnke, R., Kaul, N., Kim, K. H., Roe, K.,
813 Soutar, A. and Vathakanon, C. (1988) Early diagenesis of organic matter in Peru continental margin
814 sediments: Phosphorite precipitation. *Mar. Geol.* **80**, 309-343.

815 Froelich, P. N., Bender, M. L., Luedtke, N. A., Heath, G. R. and DeVries, T. (1982) The marine phosphorus
816 cycle. *Am. J. Sci.* **282**, 474-511.

817 Froelich, P. N., Kim, K. H., Jahnke, R., Burnett, W. C., Soutar, A. and Deakin, M. (1983) Pore water fluoride in
818 Peru continental margin sediments: Uptake from seawater. *Geochim. Cosmochim. Acta* **47**, 1605-1612.

819 Froelich, P. N., Klinkhammer, G. P., Bender, M. L., Luedtke, N. A., Heath, G. R., Cullen, D. and Dauphin, P.
820 (1979) Early oxidation of organic matter in pelagic sediments of the eastern equatorial Atlantic: suboxic
821 diagenesis. *Geochim. Cosmochim. Acta* **43**, 1075-1090.

822 Gachter, R. and Meyer, J. S. (1993) The role of microorganisms in mobilization and fixation of phosphorus in
823 sediments. *Hydrobiologia* **253**, 103-121.

824 Gachter, R., Meyer, J. S. and Mares, A. (1988) Contribution of bacteria to release and fixation of phosphorus in
825 lake sediments. *Limnol. Oceanogr.* **33**, 1542-1558.

826 Glud, R. N. (2008) Oxygen dynamics of marine sediments. *Mar. Biol. Res.* **4**, 243-289.

827 Hartnett H. E., Keil R. G., Hedges J. I. and Devol A. H. (1998) Influence of oxygen exposure time on organic
828 carbon preservation in continental margin sediments. *Nature* **391**, 572-574.

829 Hedges, J. I., Hu, F. S., Devol, A. H., Hartnett H. E., Tsamakis, E. and Keil, R. G. (1999) Sedimentary organic
830 matter preservation: A test for selective degradation under oxic conditions. *Am. J. Sci.* **299**, 529-555.

831 Houben, G. J. (2003) Iron oxide incrustations in wells. Part 1: genesis, mineralogy and geochemistry. *App.*
832 *Geochem.* **18**, 927-939.

833 Hupfer, M., Rube, B. and Schmieder, P. (2004) Origin and diagenesis of polyphosphate in lake sediments: A P-
834 31-NMR study. *Limnol. Oceanogr.* **49**, 1-10.

835 Ingall, E. and Jahnke, R. (1994) Evidence for enhanced phosphorus regeneration from marine sediments
836 overlain by oxygen depleted waters. *Geochim. Cosmochim. Acta* **58**, 2571-2575. 1994.

837 Ingall, E. and Jahnke, R. (1997) Influence of water-column anoxia on the elemental fractionation of carbon and
838 phosphorus during sediment diagenesis. *Mar. Geol.* **139**, 219-229.

839 Ingall, E., Kolowith, L., Lyons, T. and Hurtgen, M. (2005) Sediment carbon, nitrogen and phosphorus cycling in
840 an anoxic fjord, Effingham Inlet, British Columbia. *Am. J. Sci.* **305**, 240-258.

841 Ingall, E. D., Bustin, R. M. and Van Cappellen, P. (1993) Influence of water column anoxia on the burial and
842 preservation of carbon and phosphorus in marine shales. *Geochim. Cosmochim. Acta* **57**, 303-316.

843 Ingall, E. D. and Van Cappellen, P. (1990) Relation between sedimentation rate and burial of organic
844 phosphorus and organic carbon in marine sediments. *Geochim. Cosmochim. Acta* **54**, 373-386.

845 Jahnke, R. A., Emerson, S. R., Roe, K. K., and Burnett, W. C. (1983) The present day formation of apatite in
846 Mexican continental margin sediments. *Geochim. Cosmochim. Acta* **47**, 259-266.

- 847 Jennings, S., Melin, F., Blanchard, J. L., Forster, R. M., Dulvy, N. K. and Wilson, R. W. (2008) Global-scale
848 predictions of community and ecosystem properties from simple ecological theory. *Proc. R. Soc. B.* **275**,
849 1375-1383.
- 850 Jensen, H. S., Mortensen P. B., Andersen F. O. and Rasmussen E. (1995) Phosphorus cycling in a coastal
851 marine sediment, Aarhus Bay, Denmark. *Limnol. Oceanogr.* **40**, 908-917.
- 852 Jensen, H. S. and Thamdrup, B. (1993) Iron-bound phosphorus in marine sediments as measured by
853 bicarbonate-dithionite extraction. *Hydrobiologia* **253**, 47-59.
- 854 Jilbert, T., Slomp, C. P., Gustafsson, B. G. and Boer, W. (2011) Beyond the Fe-P-redox connection: preferential
855 regeneration of phosphorus from organic matter as a key control on Baltic Sea nutrient cycles.
856 *Biogeosciences* **8**, 1699-1720.
- 857 Johnston, D. T., Poulton, S. W., Goldberg, T., Sergeev, V. A., Podkovyrov, V., Vorob'eva, N. G. , Bekker, A.
858 and Knoll, A. H. (2012) Late Ediacaran redox stability and metazoan evolution. *Earth Planet. Sci. Lett.* **335-336**,
859 25-35. 2012.
- 860 Jordan, T. E., Cornwell, J. C., Boynton, W. R. and Anderson, J. T. (2008) Changes in phosphorus
861 biogeochemistry along an estuarine salinity gradient: The iron conveyor belt. *Limnol. Oceanogr.* **53**, 172-
862 184.
- 863 Kendall, B., Anbar, A. D., Kappler, A. and Konhauser, K. O. (2012) The global iron cycle. In *Fundamentals of*
864 *Geobiology* (eds A. H. Knoll, D. E. Canfield and K. O. Konhauser), John Wiley and Sons, Ltd, Chichester,
865 UK. doi: 10.1002/9781118280874.ch6.
- 866 Kraal, P., Slomp, C. P., Reed, D. C., Reichart, G.-J. and Poulton, S. W. (2012) Sedimentary phosphorus and iron
867 cycling in and below the oxygen minimum zone of the northern Arabian Sea. *Biogeosciences* **9**, 2603-2624.
- 868 Kraal, P., Bostick, B. C., Reichart, G.-J. and Slomp, C. P. (2015) Characterization of phosphorus species in
869 sediments from the Arabian Sea oxygen minimum zone: Combining sequential extractions and X-ray
870 spectroscopy. *Mar. Chem.* **168**, 1-8.
- 871 Krajewski K. P., Van Cappellen P., Trichet J., Kuhn O., Lucas J., Martin-Algarra A., Prévot L., Tewari V. C.,
872 Gaspar L., Knight R. I. and Lamboy M. (1994) Biological processes and apatite formation in sedimentary
873 environments. *Eclogae Geol. Helv.* **87**, 701-745.
- 874 Krom, M. D. and Berner, R. A. (1981) The diagenesis of phosphorus in a nearshore marine sediment. *Geochim.*
875 *Cosmochim. Acta* **45**, 207-216.
- 876 Kulaev, I. and Kulakovskaya, T. (2000) Polyphosphate and phosphate pump. *Ann. Rev. Microbiol.* **54**, 709-734.
- 877 Lebo, M. E. Particle-bound phosphorus along an urbanized coastal plain estuary. *Mar. Chem.* **34**, 225-246.
878 1991.
- 879 Lenton, T. M., Boyle, R. A., Poulton, S. W., Shields-Zhou, G. A. and Butterfield, N. J. (2014) Co-evolution of
880 eukaryotes and ocean oxygenation in the Neoproterozoic era. *Nat. Geosci.* **7**, 257-265.
- 881 Lenton, T. M. and Watson, A. J. (2000) Redfield revisited 2. What regulates the oxygen content of the
882 atmosphere? *Glob. Biogeochem. Cy.* **14**, 249-268.
- 883 Levin, L. A. and Gage, J. D. (1998) Relationships between oxygen, organic matter and the diversity of bathyal
884 macrofauna. *Deep-Sea Res. Part II* **45**, 129-163.
- 885 Li, C., Love, G. D., Lyons, T. W., Fike, D. A., Sessions, A. L. and Chu, X. (2010) A stratified redox model for
886 the Ediacaran ocean. *Science* **328**, 80-83.
- 887 Lijklema, L. (1980) Interaction of orthophosphate with iron(III) and aluminum hydroxides. *Env. Sci. Technol.*
888 **14**, 537-541.
- 889 Lucotte, M. and D'Anglejan, B. (1983) Forms of phosphorus and phosphorus iron relationships in the suspended
890 matter of the St. Lawrence estuary. *Canad. J. Earth Sci.* **20**, 1880-1890.
- 891 Lyons, T. W., Reinhard, C. T., Love, G. D. and Xiao, S. (2012) Geobiology of the Proterozoic Eon. In
892 *Fundamentals of Geobiology* (eds A. H. Knoll, D. E. Canfield and K. O. Konhauser), John Wiley and Sons,
893 Ltd, Chichester, UK. doi: 10.1002/9781118280874.ch20.
- 894 Lyons, T. W., Reinhard, C. T. and Planavsky, N. J. (2014) The rise of oxygen in Earth's early ocean and
895 atmosphere. *Nature* **506**, 307-315.
- 896 Maloof, A. C., Porter, S. M., Moore, J. L., Dudas, F. O., Bowring, S. A., Higgins, J. A., Fike, D. A. and Eddy,
897 M. P. (2010) The earliest Cambrian record of animals and ocean geochemical change. *Geolog. Soc. Am.*
898 *Bull.* **122**, 1731-1774.
- 899 Meile, C. and Van Cappellen, P. (2005) Particle age distributions and O₂ exposure times: Timescales in
900 bioturbated sediments. *Glob. Biogeochem. Cy.* **19**, GB3013, doi:10.1029/2004GB002371.
- 901 Mermillod-Blondin, F. and Rosenberg, R. (2006) Ecosystem engineering: the impact of bioturbation on
902 biogeochemical processes in marine and freshwater benthic habitats. *Aquat. Sci.* **68**, 434-442.
- 903 Meysman, F. J. R., Boudreau, B. P., and Middelburg, J. J. (2003) Relations between local, nonlocal, discrete and
904 continuous models of bioturbation. *J. Mar. Res.* **61**, 391-410.
- 905 Meysman, F. J. R., Middelburg, J. J. and Heip, C. H. R. (2006) Bioturbation: a fresh look at Darwin's last idea.
906 *Trends Ecol. Evol.* **21**, 688-695.

- 907 Middelburg, J. J. (1989) A simple rate model for organic matter decomposition in marine sediments. *Geochim.*
908 *Cosmochim. Acta* **53**, 1577-1581.
- 909 Middelburg, J. J. and Levin, L. A. (2009) Coastal hypoxia and sediment biogeochemistry. *Biogeosciences* **6**,
910 1273-1293.
- 911 Mángano, M. G. and Buatois, L. A. (2015) Decoupling of body-plan diversification and ecological structuring
912 during the Ediacaran–Cambrian transition: evolutionary and geobiological feedbacks. *Proc. R. Soc. B.* **281**,
913 20140038.
- 914 Noffke, A., Hensen, C., Sommer, S., Scholz, F., Bohlen, L., Mosch, T., Graco, M. and Wallmann, K. (2012)
915 Benthic iron and phosphorus fluxes across the Peruvian oxygen minimum zone. *Limnol. Oceanogr.* **57**, 851-
916 867.
- 917 Och, L. M. and Shields-Zhou, G. A. (2012) The Neoproterozoic oxygenation event: Environmental
918 perturbations and biogeochemical cycling. *Earth-Sci. Rev.* **110**, 26-57.
- 919 Papineau, D. (2010) Global biogeochemical changes at both ends of the Proterozoic: insights from phosphorites.
920 *Astrobiology* **10**, 165-181.
- 921 Paytan, A., Cade-Menun, B. J., McLaughlin, K. and Faul, K. L. (2003) Selective phosphorus regeneration of
922 sinking marine particles: evidence from ³¹P-NMR. *Mar. Chem.* **82**, 55-70.
- 923 Paytan, A. and McLaughlin, K. (2007) The oceanic phosphorus cycle. *Chemical Reviews* **107**, 563-576.
- 924 Peters, S. E. and Gaines, R. R. (2012) Formation of the ‘Great Unconformity’ as a trigger for the Cambrian
925 explosion. *Nature* **484**, 363-366.
- 926 Poulton, S. W. and Canfield, D. E. (2006) Co-diagenesis of iron and phosphorus in hydrothermal sediments
927 from the southern East Pacific Rise: Implications for the evaluation of paleoseawater phosphate
928 concentrations. *Geochim. Cosmochim. Acta* **70**, 5883-5898.
- 929 Poulton, S. W., Krom, M. D. and Raiswell, R. (2004) A revised scheme for the reactivity of iron (oxyhydr)oxide
930 minerals towards dissolved sulfide. *Geochim. Cosmochim. Acta* **68**, 3703-3715.
- 931 Reed, D. C., Slomp, C. P. and Gustafsson, B. G. (2011) Sedimentary phosphorus dynamics and the evolution of
932 bottom-water hypoxia: A coupled benthic–pelagic model of a coastal system. *Limnol. Oceanogr.* **56**, 1075-
933 1092.
- 934 Renz, J. R. and Forster, S. (2014) Effects of bioirrigation by the three sibling species of *Marenzelleria* spp. on
935 solute fluxes and porewater nutrient profiles. *Mar. Ecol. Prog. Ser.* **505**, 145-159.
- 936 Robbins, J. A. (1986) A model for particle-selective transport of tracers in sediments with conveyor belt deposit
937 feeders. *J. Geophys. Res.* **91**, 8542-8558.
- 938 Ruttenger, K. C. (2014) The global phosphorus cycle. In *Treatise on Geochemistry Volume 10*, (eds. H. D.
939 Holland and K. K. Turekian), Elsevier, p 499-558.
- 940 Ruttenger, K. C. and Berner, R. A. (1993) Authigenic apatite formation and burial in sediments from non-
941 upwelling, continental margin environments. *Geochim. Cosmochim. Acta* **57**, 991-1007.
- 942 Ruttenger, K. C. and Goni, M. A. (1997) Phosphorus distribution, C:N:P ratios, and $\delta^{13}\text{C}_{\text{OC}}$ in arctic, temperate,
943 and tropical coastal sediments: Tools for characterizing bulk sedimentary organic matter. *Mar. Geol.* **139**,
944 123-145.
- 945 Saltzman, M. R., Edwards, E. T., Adrain, J. M. And Westrop, S. R. (2015) Persistent oceanic anoxia and
946 elevated extinction rates separate the Cambrian and Ordovician radiations. *Geology*, **43**, 807-810.
- 947 Sannigrahi, P. and Ingall, E. (2005) Polyphosphates as a source of enhanced P fluxes in marine sediments
948 overlain by anoxic waters: Evidence from ³¹P NMR. *Geochem. Trans.* **6**, 52.
- 949 Sarmiento, J. L. and Gruber, N. (2006) *Ocean Biogeochemical Dynamics*. Princeton University Press, Princeton,
950 526 pp.
- 951 Schenau, S. J. and de Lange, G. J. (2001) Phosphorus regeneration vs. burial in sediments of the Arabian Sea.
952 *Mar. Chem.* **75**, 201-217.
- 953 Schenau, S. J., Reichart, G. J. and de Lange, G. J. (2005) Phosphorus burial as a function of paleoproductivity
954 and redox conditions in Arabian Sea sediments. *Geochim. Cosmochim. Acta* **69**, 919-931.
- 955 Schuffert, J. D., Jahnke, R. A., Kastner, M., Leather, J., Sturz, A. and Wing, M. R. (1994) Rates of formation of
956 modern phosphorite off western Mexico. *Geochim. Cosmochim. Acta* **58**, 5001-5010.
- 957 Schulz, H. N. and Schulz, H. D. (2005) Large sulfur bacteria and the formation of phosphorite. *Science* **307**,
958 416-418.
- 959 Shapiro, J. (1967) Induced rapid release and uptake of phosphate by microorganisms. *Science* **155**, 1269-1271.
- 960 Shen, Y. N., Zhang, T. G. and Hoffman, P. F. (2008) On the coevolution of Ediacaran oceans and animals. *Proc.*
961 *Natl. Acad. Sci. U.S.A.* **105**, 7376-7381.
- 962 Slomp, C. P. (2011) Phosphorus cycling in the estuarine and coastal zones: sources, sinks, and transformations.
963 In *Treatise on Estuarine and Coastal Science 5* (eds. E. Wolanski and D. McLusky), 201-229.
- 964 Slomp, C. P., Epping, E. H. G., Helder, W. and Van Raaphorst, W. (1996) A key role for iron-bound
965 phosphorus in authigenic apatite formation in North Atlantic continental platform sediments. *J. Mar. Res.*
966 **54**, 1179-1205.

- 967 Slomp, C. P., Thomson, J. and de Lange, G. J. (2004) Controls on phosphorus regeneration and burial during
968 formation of eastern Mediterranean sapropels. *Mar. Geol.* **203**, 141-159.
- 969 Slomp, C. P. and Van Cappellen, P. (2007) The global marine phosphorus cycle: sensitivity to oceanic
970 circulation. *Biogeosciences* **4**, 155-171.
- 971 Smith, J. N., Boudreau, B. P. And Noshkin, V. (1986) Plutonium and ²¹⁰Pb distributions in northeast Atlantic
972 sediments: subsurface anomalies caused by nonlocal mixing. *Earth Planet. Sci. Lett.*, **81**, 15-28.
- 973 Spiteri, C., Van Cappellen, P. and Regnier, P. (2008) Surface complexation effects on phosphate adsorption to
974 ferric iron oxyhydroxides along pH and salinity gradients in estuaries and coastal aquifers. *Geochim.*
975 *Cosmochim. Acta* **72**, 3431-3445.
- 976 Steenbergh, A. K., Bodelier, P. L. E., Heldal, M., Slomp, C. P., Laanbroek, H. J. (2012) Does microbial
977 stoichiometry modulate eutrophication of aquatic ecosystems? *Env. Microbiol.* **15**, 1572-1579.
- 978 Suess, E. (1981) Phosphate regeneration from sediments of the Peru continental margin by dissolution of fish
979 debris. *Geochim. Cosmochim. Acta* **45**, 577-588.
- 980 Sun, M. Y., Aller, R. C., Lee, C., & Wakeham, S. G. (2002) Effects of oxygen and redox oscillation on
981 degradation of cell-associated lipids in surficial marine sediments. *Geochim. Cosmochim. Acta* **66**, 2003-
982 2012.
- 983 Sundby, B., Gobeil, C., Silverberg, N. and Mucci, A. (1992) The phosphorus cycle in coastal marine sediments.
984 *Limnol. Oceanogr.* **37**, 1129-1145.
- 985 Tarhan, L. G. and Droser, M. L. (2014) Widespread delayed mixing in early to middle Cambrian marine shelfal
986 settings. *Palaeogeogr. Palaeoclimatol. Palaeoecol.* **399**, 310-322.
- 987 Tarhan, L. G., Droser, M. L., Planavsky, N. J. and Johnston, D. T. (2015) Protracted development of
988 bioturbation through the early Palaeozoic Era. *Nat. Geosci.* **8**, 865-869.
- 989 Thamdrup, B., Dalsgaard, T., and Revsbech, N. P. (2012) Widespread functional anoxia in the oxygen minimum
990 zone of the Eastern South Pacific. *Deep Sea Research Part I: Oceanographic Research Papers* **65**, 36-45.
- 991 Van Cappellen, P. and Berner, R. A. (1988) A mathematical model for the early diagenesis of phosphorus and
992 fluorine in marine sediments; apatite precipitation. *Am. J. Sci.* **288**, 289-333.
- 993 Van Cappellen, P. and Ingall, E. D. (1994) Benthic phosphorus regeneration, net primary production, and ocean
994 anoxia: A model of the coupled marine biogeochemical cycles of carbon and phosphorus. *Paleoceanography*
995 **9**, 677-692.
- 996 Van Cappellen P. and Ingall, E. D. (1997) Redox stabilization of the atmosphere and oceans and marine
997 productivity - Response. *Science* **275**, 407-408.
- 998 Van Cappellen, P. and Ingall, E. D. (1996) Redox stabilization of the atmosphere and oceans by phosphorus-
999 limited marine productivity. *Science* **271**.
- 1000 van der Zee, C., Roberts, D. R., Rancourt, D. G. and Slomp, C. P. (2003) Nanogoethite is the dominant reactive
1001 oxyhydroxide phase in lake and marine sediments. *Geology* **31**, 993-996.
- 1002 Wallmann, K. (2014) Is late Quaternary climate change governed by self-sustained oscillations in atmospheric
1003 CO₂? *Geochim. Cosmochim. Acta* **132**, 413-439.
- 1004 Wallmann, K. (2010) Phosphorus imbalance in the global ocean? *Glob. Biogeochem. Cy.* **24**, GB4030,
1005 doi:10.1029/2009GB003643.
- 1006 Westrich, J. T. and Berner, R. A. (1984). The role of sedimentary organic matter in bacterial sulfate reduction:
1007 the G-model tested. *Limnol. Oceanogr.* **29**, 236-249.
- 1008

1009 **Figure captions**

1010

1011 Fig. 1. Conceptual P cycle in fine-grained, oxic (bioturbated and bioirrigated) continental margin
1012 sediments outside of oxygen minimum zones. White squares and circles denote solids and solutes,
1013 respectively. Particulate P is added to the sediment as P_{org} and FeP. PO_4^{3-} released to the porewater
1014 from these phases is free to be transported out of the sediment by diffusion and bioirrigation. PO_4^{3-}
1015 can also be re-precipitated into authigenic FeP and CFA. The extended model considers P cycling by
1016 microorganisms (blue arrows). Dissolved PO_4^{3-} can be assimilated into an intracellular polyphosphate
1017 pool (P_{poly}). A fraction (f_{poly}) of P_{poly} is transformed into an unreactive particulate organic P phase (P_{org-U})
1018 with the remainder ($1-f_{poly}$) shunted back to the bulk P_{org} pool. If P_{poly} is mixed or buried below the
1019 aerobic sediment layers it is rapidly hydrolyzed back to PO_4^{3-} . The black bars on the right
1020 schematically indicate the spatial separation of organic matter respiration by electron acceptors with
1021 depth in the bioturbated layer. See Table S4 for stoichiometries of reactions, R.

1022

1023 Fig. 2. Simulated rates of P turnover in (a) oxic and (b) anoxic sediments in modern shelf sediments
1024 ($\mu\text{mol m}^{-2} \text{d}^{-1}$). P_{reac} burial fluxes are calculated at 100 cm. The relative magnitude of the fluxes is
1025 indicated schematically by the arrow thicknesses.

1026

1027 Fig. 3. Simulated concentrations in modern shelf sediments. (a) Dissolved oxygen, (b) nitrate + nitrite,
1028 (c) phosphate, (d) organic C, (e) organic P (including P_{org-U}), (f) iron-associated P, (g) carbonate
1029 fluorapatite, and (h) total reactive P. Green and black curves represent scenarios with oxic and anoxic
1030 bottom waters, respectively. Dashed red curves show simulation results for oxic sediments without the
1031 microbial P pump and thus the effect of animals only (red star in Fig. 4a). The dashed blue curves
1032 examine the effect of microbial P without bioturbation and bioirrigation (blue star in Fig. 4a). Despite
1033 preferential mineralization of P_{org} in oxic sediment layers, P_{org} concentrations are highest in the oxic
1034 scenario due to synthesis of microbial P_{org} . C_{org} concentrations are, in contrast, lower in oxic
1035 sediments. P_{org} is absent below the bioturbation zone in the simulation without microbial P synthesis.
1036 All geochemical profiles are shown in the Electronic Annex. Note different depth scales.

1037

1038 Fig. 4. Simulated steady state P_{reac} burial flux for sediments over a range of bottom water O_2
1039 concentrations. (a) Modern sediments. Fluxes using the baseline model parameters (Table 2) are
1040 shown as black curves with circles, and the grey shaded area shows P burial for the full range of f_{poly}
1041 (0 to 1). Results are also shown for a deeper sediment mixing and burrowing (doubling of parameters
1042 z_{bt} and z_{bio}), and for more intense mixing and burrowing (doubling of parameters $D_b(0)$ and $\alpha(0)$). The
1043 dashed red curve shows P burial without the microbial P pump, whereas the dashed blue curve shows
1044 results where bioturbation and bioirrigation are turned off. The stars correspond to the respective
1045 sediment profiles in Fig. 3. The green circle shows the anoxic baseline simulation with a lower

1046 porosity (see text). (b) Early Paleozoic sediments. P burial fluxes using the parameters in Table 2 are
1047 shown as the black curve and the dashed blue curve is the result without bioturbation and
1048 bioirrigation. The red curve is P burial with a deeper 2 cm sediment mixed depth, tentatively
1049 applicable to the late Cambrian / Ordovician shelf ($z_{br} = 1$ cm, $z_{bio} = 1$ cm, $D_b(0) = 5$ cm² yr⁻¹, $\alpha(0) = 50$
1050 y⁻¹). The arrows schematically indicate the expected trajectory of P burial following the radiation of
1051 deposit feeders and burrowing fauna with bottom water O₂ concentrations believed to characterize the
1052 early Paleozoic (Canfield et al., 2007).

1053

1054 Fig. 5. Organic phosphorus mineralization rate (sum of R₁ to R₇, Table S5) in oxic sediments
1055 inhabited by fauna (green curve) and oxic sediments without fauna (dashed blue curves). Bioturbation
1056 leads to lower rates at the surface and higher rates towards the base of the bioturbated zone, favoring
1057 the sequestration of P into authigenic phases including microbial P. The bottom water O₂
1058 concentration is 150 μM in both cases.

1059

Table 1. Observations of C and P geochemistry in oxic and anoxic margin sediments.

	Oxic	Anoxic
$C_{org}:P_{org}$	30 – 115 ^a	200 – 700 ^b
$C_{org}:P_{reac}$	21 ^c	54 – 161 ^c
C_{org} burial efficiency (CBE, %)	25 ± 10 ^d	> 40 ^e
P_{org} burial efficiency (PBE, %)	20 – 40 ^f	< 2 – 11 ^g
$(C:P)_{REG}$ ^h	118 ± 24 ⁱ	< 70 ^j

1061
1062
1063
1064
1065
1066
1067
1068
1069
1070
1071
1072
1073
1074
1075
1076
1077
1078
1079
1080
1081
1082
1083
1084
1085
1086
1087
1088
1089
1090
1091
1092
1093
1094

1095

^a From the data corresponding to ‘oxic sites’ in Fig. 3 of Slomp and Van Cappellen (2007).

^b From the data corresponding to ‘low oxygen and anoxic sites’ in Fig. 3 of Slomp and Van Cappellen (2007), as well as data from the anoxic Arabian Sea oxygen minimum zone from Kraal et al. (2012). Values exceeding 1000 observed in the paleo record are rare in modern sediments.

^c Slomp et al. (2004)

^d Literature data show that CBE depends non-linearly on the sediment mass accumulation rate (e.g. Burdige, 2007) and, possibly, the availability of dissolved oxygen. A mass accumulation rate of 0.075 g cm⁻² yr⁻¹ can be calculated for a typical upper slope / shelf setting using the formula $\rho \cdot (1-\phi) \cdot \omega_{acc}$ where ρ (2.5 g cm⁻³) is the density of sediment particles, and ϕ (0.7) and ω_{acc} (0.1 cm yr⁻¹) are the porosity and sedimentation rate of compacted sediments, respectively. For oxygenated bottom waters (> 20 μM O₂), this mass accumulation rate corresponds to a CBE of around 25 % (Dale et al., 2015b).

^e As footnote ^d, for < 20 μM O₂

^f Data on PBE in oxic, bioturbated sediments are scarce and there is high uncertainty in this value. PBE can be estimated using the CBE and $C_{org}:P_{org}$ burial ratios for which more data is available:

$PBE_{OX} = \frac{PBE_{AN} \cdot CBE_{OX} \cdot (C_{org}:P_{org})_{AN}}{CBE_{AN} \cdot (C_{org}:P_{org})_{OX}}$, where subscripts ‘OX’ and ‘AN’ denote oxic and anoxic sediments,

respectively. Assuming average values for CBE_{AN} , PBE_{AN} and $(C_{org}:P_{org})_{AN}$ of 50 %, 7 % and 400, respectively, and CBE_{OX} and $(C_{org}:P_{org})_{OX}$ burial ratios of 25 % and 50 (respectively), PBE_{OX} is 28 ± 14 %, hence the proposed range of 20 – 40. Notwithstanding the uncertainty arising from the input parameters, PBE in oxic sediments is at least a factor of 2 – 4 that in anoxic sediments. For comparison, Jensen et al. (1995) calculated a PBE of 29 – 35 % in oxic Aarhus Bay sediments, whereas Ingall and Jahnke (1994) reported lower PBE in oxic sediments ranging from 6 to 23 % based on a few observations on the continental slope.

^g For oxygen deficient bottom waters (< 20 μM O₂) (Schenau et al., 2000; Slomp et al., 2004; Ingall and Jahnke, 1994)

^h Benthic regeneration ratio = DIC flux / PO₄³⁻ flux at sediment surface determined in situ using benthic chambers. It is important to point out that Colman and Holland (2000) caution that benthic phosphate fluxes, and indeed dissolved inorganic carbon fluxes, may exhibit seasonal variability. For the present study, we note these concerns but lay them to one side, and consider that $(C:P)_{REG}$ ratios are indicative of steady state or seasonally-averaged conditions.

ⁱ Using the empirical transfer function of Wallmann (2010).

^j Values for anoxic bottom waters on the Peruvian margin range from 4 to 68 (Noffke et al., 2012). The empirical function of Wallmann (2010) based on in situ flux measurements predicts a value of 11 ± 24 for near-anoxic conditions, although the full range of values for anoxic waters is very similar to the Peruvian margin. Sannigrahi and Ingall (2005) report a value of 39 for an anoxic site in Effingham Inlet.

1096 Table 2. Key model parameters for modern oxic sediments displaying bioturbation and bioirrigation. Values for
 1097 anoxic and early Paleozoic (Pz) settings that differ from these are also given. Parameter sources are given in
 1098 Table S3.

Description	Value		
	Oxic	Anoxic	Pz
Sediment accumulation rate, ω_{acc} (cm yr ⁻¹)	0.1		
Surface sediment porosity, $\phi(0)$ (-)	0.9		
Bioturbation coefficient at sediment surface, $D_b(0)$ (cm ² yr ⁻¹) ^a	28	0.2	5
Bioturbation halving depth, z_{bt} (cm)	3		0.25
Bioirrigation coefficient at sediment surface, $\alpha(0)$ (yr ⁻¹) ^a	465	4	50
Bioirrigation attenuation coefficient, z_{bio} (cm)	2		1
Average lifetime of the reactive C _{org} components, a (yr)	3×10 ⁻⁴		
Shape of gamma distribution for C _{org} mineralization, ν (-)	0.05		
Rate constant for precipitation of apatite, k_{29} (yr ⁻¹)	1		
Rate constant for polyphosphate synthesis, k_{30} (M ⁻¹ yr ⁻¹)	3×10 ⁷		
Rate constant for polyphosphate hydrolysis, k_{31} (yr ⁻¹)	50		
Rate constant for transformation of P _{poly} to P _{org} , k_{32} (yr ⁻¹)	50		
Rate constant for transformation of P _{poly} to P _{org-U} , k_{33} (yr ⁻¹)	50		
Fraction of P _{poly} converted to P _{org-U} , f_{poly} (-)	0.25		
Fe:P ratio in allochthonous Fe _{HR} , ϵ_{all} (mol Fe (mol P) ⁻¹)	25		
Fe:P ratio in authigenic Fe _{HR} , ϵ_{aut} (mol Fe (mol P) ⁻¹)	10		
Fe:P ratio in Fe _{MR} recrystallized from Fe _{HR} , ϵ_{age} (mol Fe (mol P) ⁻¹)	40		
Equilibrium PO ₄ ³⁻ concentration for CFA precipitation, C_{eq} (μM)	10		
Atomic P-C ratio in deposited organic matter, r_{PC} (mol P (mol C) ⁻¹)	1/106		
Acceleration factor for aerobic (relative to anaerobic) C _{org} degradation, $foxC$ (-)	2		
Acceleration factor for aerobic P _{org} degradation relative to C _{org} , $foxP$ (-)	2		
Acceleration factor for anaerobic P _{org} degradation relative to C _{org} , $fanoxP$ (-)	2		
Bottom water concentration of O ₂ (μM)	150	0.1	Variable ^b
Bottom water concentration of NO ₃ ⁻ (μM)	35		
Bottom water concentration of SO ₄ ²⁻ (μM)	28000		3000
Bottom water concentration of Fe ²⁺ (μM)	0		
Bottom water concentration of PO ₄ ³⁻ (μM)	0		
Bottom water concentration of H ₂ S (μM)	0		
Seafloor flux of C _{org} (mmol cm ⁻² yr ⁻¹)	0.34		
Seafloor flux of P _{org} (mmol cm ⁻² yr ⁻¹)	0.34/ r_{CP}		
Seafloor flux of Fe _{HR} (mmol cm ⁻² yr ⁻¹)	0.011		
Seafloor flux of P associated with Fe _{HR} (mmol cm ⁻² yr ⁻¹)	0.011/ ϵ_{all}		
Seafloor flux of CFA (mmol cm ⁻² yr ⁻¹)	0		
Seafloor flux of P _{poly} (mmol cm ⁻² yr ⁻¹)	0		
Seafloor flux of P _{org-U} (mmol cm ⁻² yr ⁻¹)	0		

1099 ^a Bioturbation ($D_b(0)$) and bioirrigation ($\alpha(0)$) are scaled to bottom water O₂ concentration (see Table S1) which
 1100 causes a reduction in the mixing and burrowing intensities as O₂ falls below ca. 20 μM.

1101 ^b For the Paleozoic simulations, a ranges of O₂ from 0.1 to 150 μM is used (Fig. 4).

1102

1103

1104 Table 3. Modelled C and P geochemistry in modern oxic and anoxic shallow margin sediments from
 1105 the simulations that include the microbial P loop.

	Oxic	Anoxic
C_{org}/P_{org} burial ratio ^{a, b}	73	497
C_{org}/P_{reac} burial ratio ^a	44	196
C_{org} burial efficiency (CBE, %) ^a	16	47
P_{org} burial efficiency (PBE, %) ^{a, b}	24	10
Inorganic carbon-to-phosphorus regeneration ratio, $(C:P)_{REG}$	116	63

1106

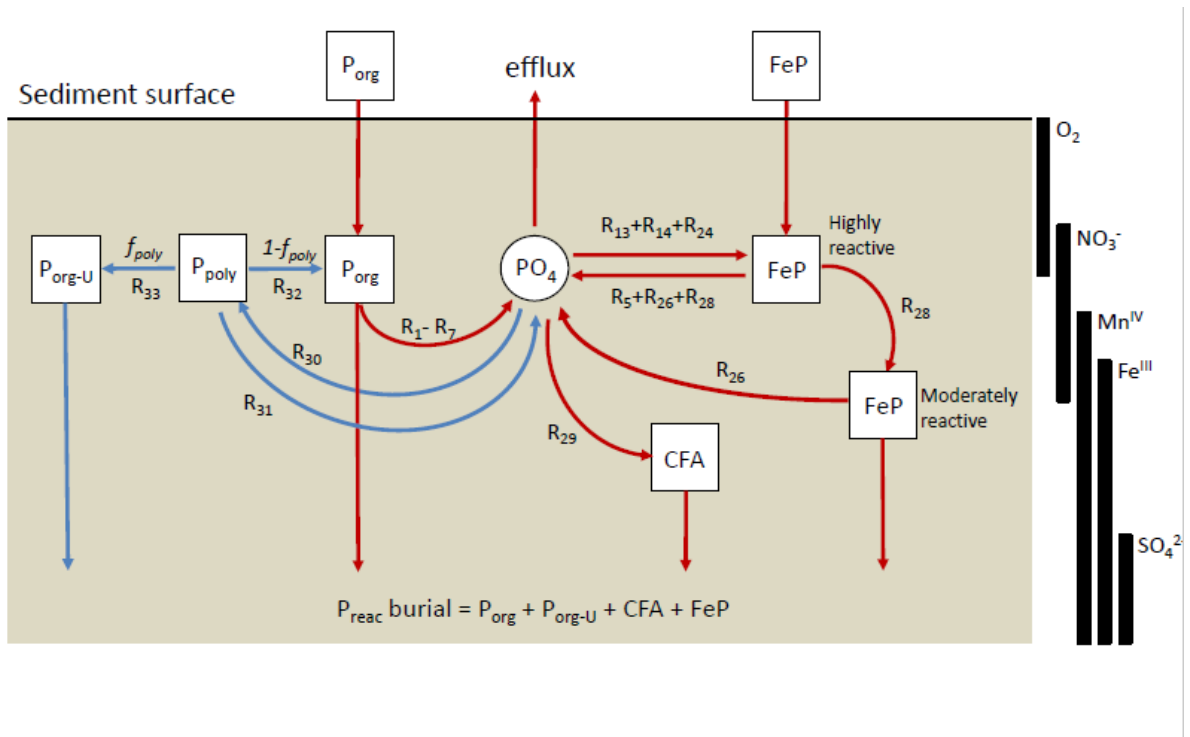
1107 ^a At 100 cm

1108 ^b Includes P_{org-U}

1109

1110

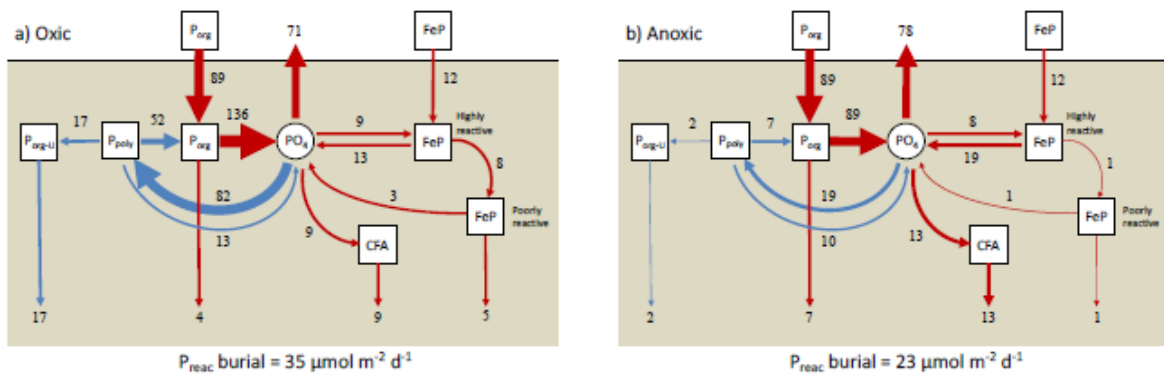
1111 Fig. 1



1112

1113

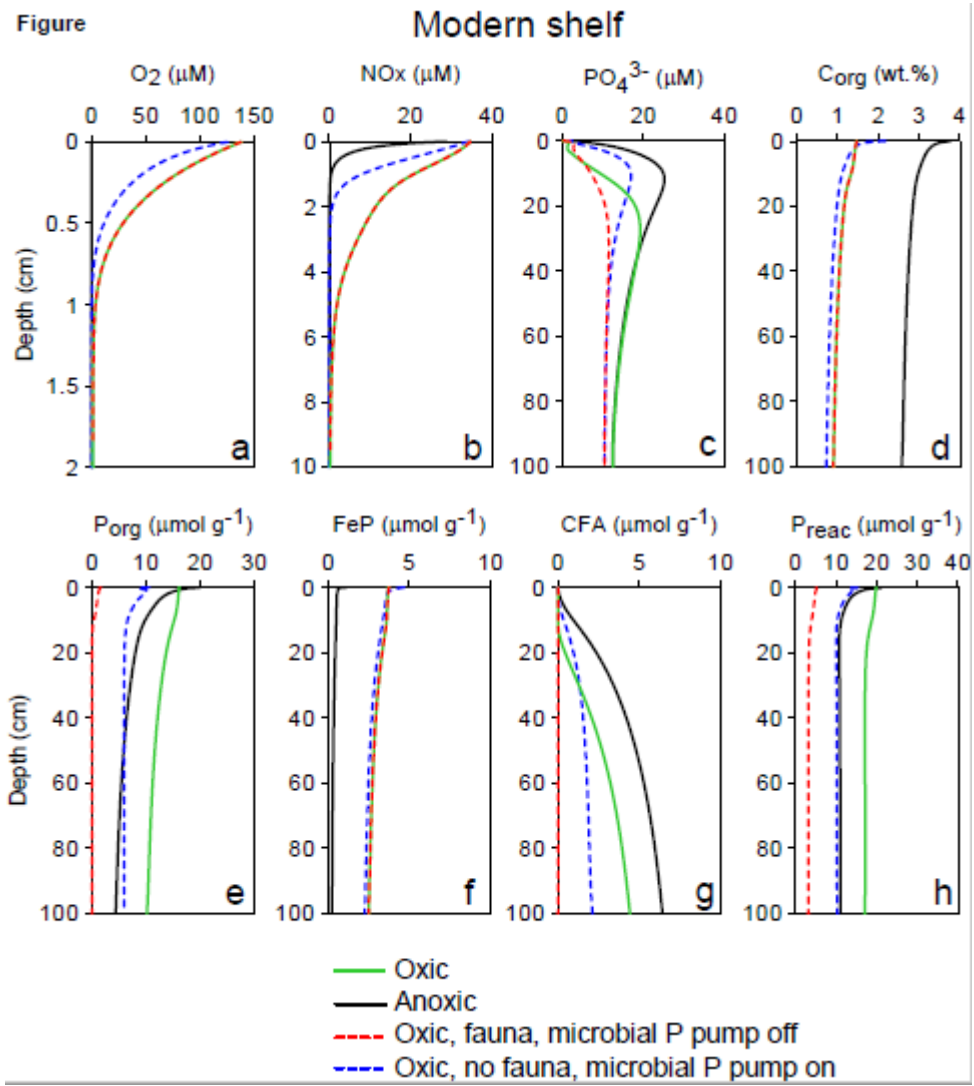
1114 Fig. 2.



1115

1116

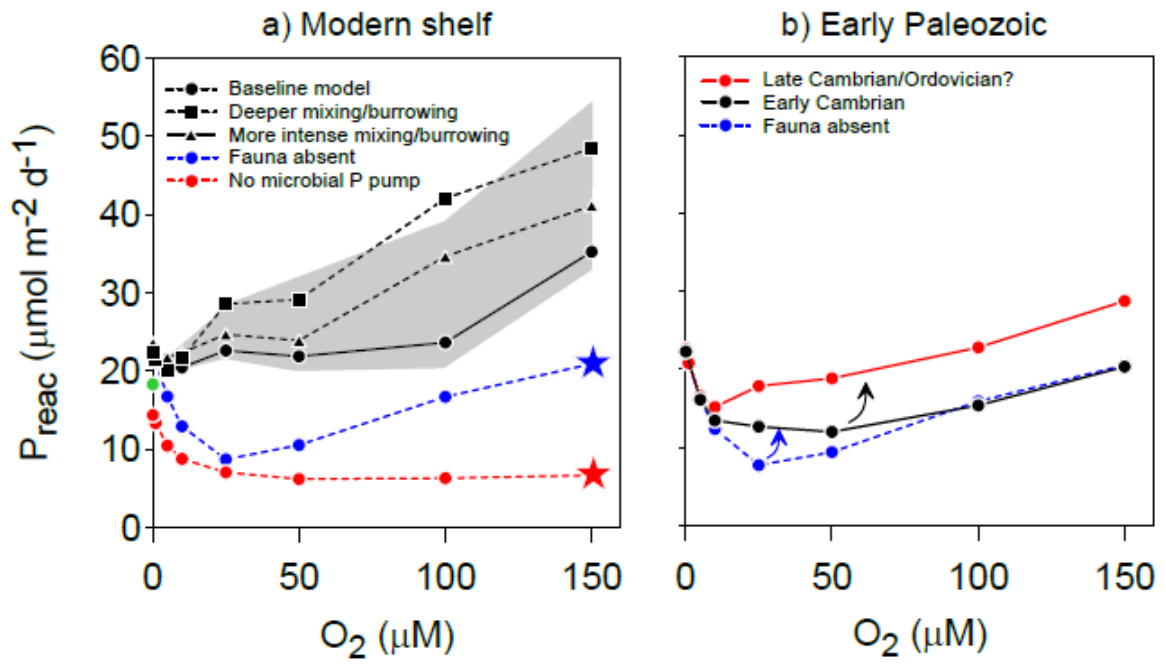
1117 Fig. 3



1118

1119

1120 Fig. 4



1121

1122

