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Abstract: Conservation actions generally benefit some groups more than others, and this inequity is thought to affect the probability of achieving conservation objectives. This has led to the common assumption that triple bottom line solutions -- those that are effective, efficient, and equitable -- are best and most likely to achieve each individual objective. Although this may be true, it has been little tested, and importantly lacks a conceptual foundation for understanding, predicting and evaluating how equity affects conservation outcomes. We describe types of equity relevant to conservation and explore how they may affect the probability of successfully achieving conservation outcomes. Depending on the equity type and context, the relationship between equity and conservation success varies. We find that the best conservation outcome is often achieved without perfect equity; highlighting the risk of ignoring the relationship between equity and success. We offer a conceptual foundation for better addressing this important issue in future research and application.

2 September 2015

Dear Editors,

Re: Submission to Global Environmental Change

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We are pleased to share a revised manuscript that addresses minor edits from two reviewers. We are grateful for their further refinements. We have attached a separate detailed response to reviewers.

For a long time it has been recognized that conservation actions that disproportionately impact the disenfranchised few –termed environmental injustice in the resource management literature – are both morally wrong and less likely to achieve ultimate desired outcomes. The interest in these topics has recently grown dramatically with the heightened attention given to global inequity across many aspects of society, including wealth distribution, climate change impacts, and others. What has been missing from this research and debate is a more complete assessment of how equity in general, across the entire spectrum of equity that ranges from severe injustice to perfect equity, affects the probability of success in achieving the desired environmental management outcome.

Here we tackle this issue within the realm of conservation science by developing and evaluating a formal conceptual foundation for assessing social equity. In particular, we highlight several key lessons and guidelines about how best to address the issue of equity in conservation planning:

- Equity is an increasingly important issue to address in conservation, yet a poorly articulated concept in most of the literature. We developed a conceptual foundation for understanding and evaluating equity within conservation that will help make this science and its application much more rigorous.
- In most cases there is a tradeoff between achieving conservation outcomes and producing equitable solutions – sometimes a strong tradeoff. We offer a formal way of calculating and addressing this tradeoff.
- Triple-bottom line solutions – those that achieve conservation outcomes effectively, efficiently, and equitably – may be quite rare. Acknowledging (and further testing) this result could profoundly change the nature of conservation actions.

Formal evaluation of equity in conservation has only just begun. Our work poses as many questions as it tries to resolve. We anticipate it serving a foundational role in guiding future research addressing this globally important management topic.

If additional information is needed, please do not hesitate to contact me.

Yours sincerely, on behalf of all co-authors,



Dr Carissa Klein

# **Social equity and the probability of success of biodiversity conservation**

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Title: Social equity and the probability of success of biodiversity conservation  
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September 1, 2015

Dear Editors,

We are pleased that our manuscript has been accepted following further revision. We have slightly revised the final manuscript to address the two minor comments from the reviewers. Below we detail our responses to each comment, with the original comment in **bold** and our response in regular font.

**Reviewer #1:**

**The final minor amendment I would recommend relates to their response comment that they have identified whether they are discussing input or output equity in each examples in Section 3.**

Response: We have clarified in the subsections (3.1-3.4) whether the examples described relate to input or output equity. (Edits are within lines 179-182; 193; and 235).

**Reviewer #2:**

**There is one point that I think they have not got yet. It does not preclude publication - for as the authors observe, many people do not get it But they could draw attention to this more explicitly, the text they wrote in their response to me would suffice.**

Response: We have added further justification for our belief that different values of equity might be comparable. See Line 311:

“Not all perceived values of conservation (associated with either costs or benefits) will be tangible or easily quantifiable; yet assessing their relative importance has merit. Any type of equity in principle could be measured subjectively on a unitless scale of low to high. Formalization of problems that involve values can be an anathema to some, but the benefits of explicating integrating these issues into formal conservation planning are greater than ignoring perceived values altogether.”

**Highlights:**

- Social equity, economic efficiency and environmental effectiveness are often sought
- Social equity can be necessary for success, but can compromise other goals
- We enhance our understanding of the social equity-conservation success relationship
- The best conservation outcome is often achieved without perfect social equity



4 **Abstract**

5 Conservation actions generally benefit some groups more than others, and this inequity is  
6 thought to affect the probability of achieving conservation objectives. This has led to the  
7 common assumption that triple bottom line solutions -- those that are effective, efficient, and  
8 equitable -- are best and most likely to achieve each individual objective. Although this may be  
9 true, it has been little tested, and importantly lacks a conceptual foundation for understanding,  
10 predicting and evaluating how equity affects conservation outcomes. We describe types of equity  
11 relevant to conservation and explore how they may affect the probability of successfully  
12 achieving conservation outcomes. Depending on the equity type and context, the relationship  
13 between equity and conservation success varies. We find that the best conservation outcome is  
14 often achieved without perfect equity; highlighting the risk of ignoring the relationship between  
15 equity and success. We offer a conceptual foundation for better addressing this important issue in  
16 future research and application.

17 **Keywords**

18 Biodiversity, benefits, conservation planning, costs, environment, equity, triple bottom line

19



## 20 **1 Introduction**

21 Social equity - the equitable distribution of costs or benefits between individuals or groups of  
22 people - is a highly sought after ideal in many aspects of society. Whether related to education,  
23 employment, or healthcare, equitable outcomes or opportunities can influence the creation,  
24 durability, and success of local, national, and international policies (Solar and Irwin 2007). The  
25 conservation of biodiversity is no exception (Halpern *et al.* 2013). In contrast to health and  
26 education, however, relatively little work has been done to understand how, and in what cases,  
27 explicit consideration of equity influences effectiveness of a conservation plan or policy  
28 (henceforth 'conservation intervention', which can include, but is not limited to: protected area  
29 plans/policies, payments for ecosystem services plans/policies, etc.). Here we aim to enhance  
30 our understanding of the relationship between different types of social equity and success in  
31 biodiversity conservation interventions, with the goal of improving conservation outcomes. A  
32 rich body of literature exists on measuring the effectiveness of conservation interventions, and  
33 understanding factors affecting the probability of their success (Bottrill and Pressey 2012;  
34 Ferraro and Hanauer 2014; Mascia *et al.* 2014). Success in conservation is broadly defined by  
35 achievement of stated goals, which vary according to different values and beliefs. For example,  
36 a successful protected area plan could be measured by ecological representation, biodiversity  
37 persistence, or economic impact (Parrish *et al.* 2003; Klein *et al.* 2010), whereas a successful  
38 conservation policy could be measured by improved strength of legislation governing the use of  
39 natural resources (Gleason *et al.* 2010) or community support (Russ and Alcala 1999). Other  
40 conservation outcomes might be measured by changes in social, institutional or human capital  
41 (Bottrill and Pressey 2012; Ban *et al.* 2013). Ultimately, the success of conservation  
42 interventions is often evaluated on the basis of conservation benefit, social equity, and economic

43 return, the three components to triple bottom line conservation outcomes (Halpern *et al.* 2013).  
44 Yet the feasibility of achieving such triple bottom line solutions, and the potential interactions  
45 and tradeoffs among the three components, remains largely untested. Halpern *et al.* (2013) found  
46 that social equity can compromise achieving efficient conservation outcomes, but highlighted the  
47 importance of further research focused on exploring how the relationship between social equity  
48 and conservation success might influence these trade-offs, in particular with respect to the many  
49 different types of equity. Here, we explore this relationship to provide insight to outstanding  
50 questions in conservation, including: Is probability of conservation success actually optimized  
51 when all three components are maximized? Or, does conservation success require approaches  
52 that deviate from the triple bottom line?

53 Equity is increasingly recognized as a component of conservation success (Ban *et al.* 2013;  
54 Campese *et al.* 2009). However, there are multiple types of equity (Figure 1), and being clear  
55 about what type of equity is important and being measured is critical for understanding the  
56 relationship between conservation success and equity. Equity concerns can arise from both  
57 internal factors (e.g., composition of the project team), which tend to be within the control of the  
58 planning team, and external contextual factors (e.g., social, geographic or economic conditions  
59 of the planning region), which are generally beyond the control of the project. For example, the  
60 design of a stakeholder engagement strategy might consider equal participation of different  
61 groups in a consultation process designed to ensure representation from all affected stakeholders,  
62 an internal factor. Alternatively, the variation and spatial distribution of existing income levels in  
63 the planning region might determine which populations or communities are affected by  
64 restrictions on resource use recommended by a conservation plan, an external contextual factor.  
65 While external factors can rarely be controlled, understanding, anticipating and managing their

66 influence on the design and implementation of a conservation intervention is likely to increase its  
67 probability of success (Berkes 2004; Solar and Irwin 2007). Internal factors can be inputs into,  
68 and/or outcomes of, a conservation intervention, and can influence its success (Figure 1). We  
69 believe that consideration of different types of equity improves the chance of achieving  
70 conservation success.

71 The focus of this manuscript is on how social equity, one of many potential conservation  
72 objectives and factors affecting conservation success, influences the probability a conservation  
73 intervention succeeds in meeting its stated goal. We acknowledge that cases exist where equity  
74 plays little to no role in conservation interventions and their success, for example when  
75 governments impose protected areas despite local protests (Brockington 2004), but our emphasis  
76 here is on cases where equity matters. We identify different types of input and outcome equity  
77 and discuss their possible relationships with conservation success. Finally, we simulate how  
78 understanding these relationships can help us evaluate the feasibility of triple bottom line  
79 solutions, where social equity, environmental benefit, and economic return are maximized.

## 80 **2 Social equity in conservation**

81 A complex collection of social structures, economic systems, and policy frameworks determine  
82 the relevance of equity to conservation outcomes, and thus conservation success. These social  
83 determinants of conservation equity reflect the distribution of wealth, power, and access to  
84 resources within a society, and can in turn have different consequences for different types of  
85 conservation equity. We identified many types of conservation equity, and divided them into two  
86 main categories, input and outcome, that influence conservation success (Figure 1), all of which  
87 can be influenced by socioeconomic and political context (described below in section 2.1).

88 Several types of equity can be either input or outcome equity, or both, depending on the decision  
89 process and goals of the conservation action. The primary distinction is whether the type of  
90 equity is a dimension of the social context that influences the process of making a conservation  
91 decision, i.e., input equity, or is something affected by the conservation action, i.e., outcome  
92 equity. As such, potential metrics of these types of equity are often the same (Fig. 1b), but how  
93 they are used and interpreted will differ. Differences between input and outcome equity are  
94 further explained and illustrated below.

### 95 ***2.1 Socioeconomic and political context***

96 Context variables encompass a broad set of structural, cultural, and functional aspects of a social  
97 system that exert a powerful formative influence on patterns of social stratification and, thus,  
98 influence conservation equity (Ostrom 1990; Solar and Irwin 2007). Fully characterizing all  
99 components of context is beyond the scope of this paper. Context determinants are often beyond  
100 the control of a conservation intervention, representing external factors influencing conservation  
101 success, except when the goal of the intervention is to change existing governance structures or  
102 policies. We highlight context here because it influences equity and thus affects conservation  
103 success. Examples of determinants related to context affecting conservation success include  
104 governance, cultural and societal values, and social/economic/public policies (Figure 1).

### 105 ***2.2 Input Equity***

106 The socioeconomic and political context within a planning region gives rise to different forms of  
107 social position and hierarchy within groups of individuals. Populations can be stratified by  
108 socioeconomic position according to education, occupation, gender/age, race/ethnicity,

109 generational, financial status and other factors (Figure 1). In some cases, these different groups  
110 participate in the conservation intervention through a participatory process, and help guide  
111 decisions about what and where to protect; we classify this as a form of input equity. For  
112 example, a decision process that includes only men or only wealthy people would be inequitable  
113 for those two types of input equity, and this may ultimately affect the ability to achieve the  
114 conservation outcome. In particular, the existence and equitability of the participatory process  
115 can directly influence conservation success by slowing or stopping the decision process, where in  
116 extreme cases the lack of a participatory process is responsible for failure of the intervention  
117 (Gleason *et al.* 2010). In other cases, the participatory process can influence the outcome of the  
118 intervention (e.g., the size or location or regulations of a protected area plan), which can in turn  
119 indirectly influence conservation success.

### 120 **2.3 Outcome Equity**

121 Outcome equity refers to the distribution of costs and benefits of the final outcome of the  
122 conservation intervention (e.g., a protected area plan) to different socio-economic groups and/or  
123 across space (Figure 1). For example, a protected area plan can disproportionately impact  
124 different socioeconomic groups, such as different industry sectors (Adams *et al.* 2010;  
125 occupation equity), by restricting access to a natural resource (access or spatial equity). In many  
126 cases input equity can influence outcome equity, as those involved in the decision process may  
127 design a conservation intervention that favors themselves and thus leads to outcome inequity,  
128 often for the same type of equity (e.g., if men dominate the decision process, they may produce  
129 outcomes that produce greater benefits for men). Outcome equity can be independent of input  
130 equity when conservation interventions do not involve a participatory process.

### 131 **3 Equity and probability of conservation success**

132 Once the types of equity relevant to a conservation intervention have been identified,  
133 conservation success requires understanding how these types of equity affect the probability of  
134 success. Increased social equity is often assumed to improve the probability of conservation  
135 success (Brown 2002; Halpern *et al.* 2013). In some cases, this assumption may be true; for  
136 example, in the implementation of locally managed marine areas, where self enforcement of new  
137 regulations is more likely to occur when local people perceive the regulations as equitable  
138 (Hatcher *et al.* 2000). However, it is also likely that conservation will fail if vocal or powerful  
139 individuals or groups are not satisfied with the outcome, in other words, if the outcomes of  
140 conservation planning and actions do not match the (often inequitable) local context. The  
141 relationship between equity and probability of conservation success is presumed to be positive  
142 (Brown 2002) yet is poorly understood, and further complicated when values and perceptions  
143 among and between different groups are taken into account too (Ravallion 2014; Figure 2).  
144 Recognizing the difference between absolute, relative and perceived is critical for objective  
145 setting and evaluation of intervention outcomes. Absolute equity refers to every participant  
146 experiencing the same, or equal, outcome. For example, regardless of size, every boat is allowed  
147 to catch the same number of fish (Figure 2). Relative equity refers to participants experiencing a  
148 proportional outcome related to a stated variable, e.g., boats receiving fish catch in proportion to  
149 their boat size as compared to other boats. Perceived equity is how those involved in the process  
150 perceive of their allotted outcome compared others, e.g., the size of fish catch relative to other  
151 fishers.

152 Here, we describe four general relationships that have been observed between equity and  
153 probability of conservation success ( $P(x)$ ; Figure 3): A) Linear, where  $P(x)$  increases  
154 proportionally with increasing equity; B) Asymptotic, where  $P(x)$  increases rapidly with initial  
155 increases in equity and then plateaus; C) Humped, where  $P(x)$  rises initially and then drops off  
156 with higher levels of equity, and D) Sigmoidal, where  $P(x)$  responds slowly at first to increases  
157 in equity and then rises quickly. For nonlinear shapes, the location of inflection points (i.e.,  
158 change in slope) is likely connected to a contextual determinant, such as governance or cultural  
159 value. For each relationship, we describe it in the terms of individual types of equity and support  
160 it using empirical evidence, where possible. These four relationships are hypotheses; their  
161 frequency of occurrence and impact on overall conservation outcomes are still to be fully tested.  
162 We hope the conceptual foundation described here helps make such testing more rigorous. For  
163 any equity type, its relationship with conservation success will likely vary from case to case  
164 depending on how equity is considered in the process (as an input or an outcome), how equity is  
165 measured (as quantitative or qualitative values, e.g., dollars versus participation effort), and how  
166 equity is defined (as absolute, relative or perceived) (McClanahan *et al.* 2008) (Figure 2).

### 167 3.1. *Linear*

168 Occupational and spatial equity are two of several types of equity that may relate linearly with  
169 conservation success (Figure 3a). For example, it seems reasonable to expect conservation plans  
170 that produce more equitable relative impact to each key occupational sector, would be more  
171 successful. In California, the Marine Life Protection Act Initiative is an example of a successful  
172 conservation plan that made considerable effort to equitably impact commercial fishery sectors  
173 in each major region (Klein *et al.* 2010; White *et al.* 2013).

174 With spatial equity, a linear relationship between equity and conservation success has been  
175 observed with a type of spatial fisheries management, Territorial User Rights in Fisheries  
176 (TURFs), which allow individuals or a set group of people to fish in a particular area. TURFs  
177 have demonstrated increasingly positive outcomes with increasing levels of both input and  
178 output equity. For example, Chilean TURF cooperatives allocate effort temporally and spatially  
179 via a pooling scheme (input equity), to equalize the work burden and spread effort in a more  
180 efficient manner (Cancino *et al.* 2007), and this program has successfully met conservation goals  
181 (by not exceeding the total allowable catch) and social goals (by equally distributing the  
182 transaction costs and benefits of the TURF) – an example of output equity.

### 183 3.2 *Asymptotic*

184 Financial and participation equity are two of several types of equity that could relate to  
185 conservation success asymptotically (Figure 3b), where conservation success increases with  
186 increasing levels of equity to a point, after which equity does not influence success. With  
187 financial equity, conservation success is assumed to increase with increasing financial equity  
188 (i.e., distribution between groups regardless of financial status or profitability). However, in  
189 some cases conservation success is likely to peak, and remain constant, when more powerful or  
190 vocal stakeholders receive the greatest benefit. For example, when the Great Barrier Reef was  
191 rezoned, the government provided monetary compensation to commercial fishermen but not to  
192 other, more profitable industries (Macintosh *et al.* 2010). As fishermen were the most vocal  
193 stakeholder group, allocation of additional money to other groups, an example of output equity,  
194 may not have impacted conservation success, resulting in an asymptotic relationship.



195 Similarly to financial equity, the probability of success of conservation interventions could  
196 increase, to a point, with increasing participation from stakeholder groups (participation equity).  
197 An example of how stakeholder participation can lead to successful conservation was  
198 demonstrated using data from 84 forest management cases around the world (Persha *et al.* 2011);  
199 whereas, lack of stakeholder participation lead to an unsuccessful conservation was shown in the  
200 first attempt to implement the California's Marine Life Protection Act (Gleason *et al.* 2010).  
201 Similarly, in Alaska where all federal fisheries are managed by annual catch limits and some  
202 type of limited access program, stakeholders and the public have several opportunities for  
203 participation input during the development phase, which is recognized as critical for building  
204 stakeholder acceptance of the program and balancing divergent interests (Fina 2011). However,  
205 this relationship is unlikely to be linear, as conservation success likely stabilizes once the most  
206 vocal or influential stakeholders are included in the process (i.e., engaging additional, less  
207 influential stakeholders in the decision process might increase equity but likely have little effect  
208 on conservation success).

### 209 3.3 *Humped*

210 Generational, gender, social, ethnicity, and financial (described above) are types of equity that  
211 could affect conservation success in a humped fashion, where the peak of the hump reflects the  
212 point in which conservation success is maximized. For example, some conservation initiatives  
213 favor current generations and disproportionately impose costs on future generations, indicating a  
214 humped shaped relationship that peaks early to reflect the bias towards current generations  
215 (Figure 3c) (Dobbs 1982). Generational equity would be difficult to achieve as a type of input  
216 equity given timeframes involved in most decision processes. In many societies, conservation

217 success is generally assumed to increase linearly with increased gender equity inputs and outputs  
218 (Agarwal 2009; Figure 3a). However, conservation success probably peaks at a point that  
219 matches the power structure of a society. In many places, decisions are often made by, or favor, a  
220 single gender (Martin and Lemon 2001; Agarwal 2009; Tsikata and Golah 2010), thus  
221 conservation success would peak at the point that reflects this power structure. Other types of  
222 equity, in particular social class and ethnicity, often reflect different power and influence among  
223 groups within regions. In community forestry programs in Nepal, while socially dominant  
224 (higher caste) individuals make management decisions affecting all groups, lower caste social  
225 classes harvest a majority of the forest resources, and therefore conservation success is unlikely  
226 to occur until they are involved, even if at a minimal level. Yet, higher caste groups might not  
227 tolerate a substantial redistribution of decision-making rights among other social classes  
228 reflecting a humped relationship (Nightingale 2002).

### 229 3.4 *Sigmoidal*

230 Types of equity that potentially have an asymptotic relationship to conservation success would  
231 exhibit a sigmoidal relationship in cases where some minimum threshold level of equity exists  
232 that is needed to achieve success. For example, in fisheries management based on individual  
233 transferable quotas, each fisher (or fisher group) is allowed a 'catch share' (i.e., access equity)  
234 that can be used, sold, or leased. This form of regulation is only likely to be successful if some  
235 minimum threshold of output equity is achieved, or in other words, fishermen are not entirely  
236 excluded from the process. If access equity increases, more people are given access to a smaller  
237 portion of the fishery, assuming a total allowable catch has been set and remains constant, and  
238 thus individual catch would decrease. In this case, probability of success likely plateaus at some

239 intermediate level of equity (sigmoidal relationship). For example, the halibut and sablefish  
240 fisheries have historically supported a large number of small vessels (Fina 2011). Both set  
241 individual fishing quotas (IQFs) to reflect historic fisheries access, but entry into the fishery is  
242 limited. Thus, probability of success increases to a point where enough of the fishers buy into the  
243 program, but probably plateaus at a point where entry (access equity) is limited and total  
244 allowable catch and catch shares remain steady.

### 245 *3.5. Additional aspects of the curves*

246 Some types of equity may express different relationship curves depending on the context. For  
247 example, with the catch allocation example in section 3.4, if individuals become less satisfied  
248 with their shrinking allocation of catch with increasing equity, they may begin violating  
249 regulations, in turn decreasing conservation success at higher levels of equity (humped shape  
250 curve instead of sigmoidal). Similarly, the relationship between financial equity and conservation  
251 success may be humped if groups without much power or voice receive money that could have  
252 gone to groups that feel they deserve more, causing those groups to perceive the allocation as  
253 inequitable and unacceptable for success.

254 A key unknown about any of the potential relationships between equity and conservation success  
255 is where the curve crosses an axis (Fig. 1a, inset). It is often assumed that conservation  
256 interventions will fail without some minimum level of equity (Borrini *et al.* 2004), such that the  
257 curves would intersect the x-axis at some value greater than zero. Yet there are other examples  
258 where conservation has been successful despite highly inequitable outcomes, for example where  
259 top-down management displaces local communities (Brockington 2004, de Santo *et al.* 2011). In  
260 these cases, the curves would intersect the y-axis at a value greater than zero.

261 Additionally, different types of equity, each with its own curve, may be relevant and important  
262 within the same management plan. Such differences further challenge incorporating equity into  
263 conservation planning, but can be resolved at least partially by efforts to elicit the relative  
264 importance of each type of equity to stakeholder groups and then incorporate those weights into  
265 formal multi-criteria decision making (Kittinger *et al.* 2014).

266

## 267 **4 Discussion**

268 We need a better understanding of the relationship between equity and conservation success,  
269 including when and how much social equity contributes to conservation success, to achieve  
270 conservation goals. We provide a conceptual foundation for understanding how and when  
271 different types of equity can influence conservation success relative to how equity is measured  
272 and perceived. Understanding the nature of these interactions between equity, conservation  
273 success, and economic return is fundamental for determining the feasibility of triple bottom line  
274 solutions. In conservation planning, expected conservation benefit is typically calculated as the  
275 product of probability of success and conservation benefit. In general, conservation benefit  
276 reflects both biodiversity conservation and economic efficiency objectives, addressing two  
277 pillars of the triple bottom line (Halpern *et al.* 2013). Here we demonstrate, in theory, how a  
278 third pillar, equity, potentially affects probability of conservation success (shown in Figure 3),  
279 and how this in turn interacts with the way equity can limit potential conservation benefits  
280 (Figure 4). The implication of these results is that equity can either exacerbate (Fig. 4b-d) or  
281 mitigate (Fig. 4a) the ability to achieve biodiversity and economic conservation objectives. In  
282 most cases, the optimal conservation outcome is achieved without perfect equity. In fact, high

283 levels of equity could severely compromise conservation outcomes (e.g. Figure 4c) if, for  
284 example, existing power structures are themselves inequitable, which highlights the risk of not  
285 considering the relationship between equity and probability of success.

286

287 We simplified the problem by considering each type of equity separately, but acknowledge that  
288 complex relationships exist among specific types of equity and between context determinants  
289 (Adelman and Morris 1973), and that these interactions influence the degree of success. Further,  
290 we acknowledge that additional relationships are likely to exist (e.g., nonlinear shapes with  
291 multiple inflection points, flat lines where equity has no bearing on conservation success), and  
292 that the relationships may change through time, as people learn and adapt, and among  
293 communities that have different contexts. Similarly, different groups within a planning process  
294 may value different types of equity, and if those types influence the process (input equity) or  
295 respond differently to the conservation intervention (outcome equity), then overall conservation  
296 success could be compromised. A more indepth understanding of these relationships and  
297 interactions is important and will require empirical research focused on determining or  
298 evaluating specific relationships between the probability of success and equity, as well as how  
299 different types of equity are valued by stakeholders within a planning process (i.e., how much  
300 weight to give each one in planning decisions). Embarking on this substantial research agenda  
301 requires a conceptual foundation, which is the crux of this manuscript.

302

303 Complicating matters further, the actual relationship between equity and conservation success  
304 may differ from the perceived relationship of equity for different individuals or groups (Webb *et*  
305 *al.* 2004; McClanahan *et al.* 2008). Perceptions of equity and conservation success reflect the

306 values of those involved in, or affected by, a program or strategy, their expectations, and whether  
307 goals are achieved (Axford *et al.* 2008). Perceptions are important as they lead people to change  
308 their behavior (*e.g.*, whether or not to comply to new regulations) and/or lead to new  
309 conservation actions (Claus *et al.* 2010). As with absolute equity, perceptions of equity will  
310 likely change through time and vary among individuals and communities, creating an additional  
311 challenge for understanding the relationship between equity and conservation success. Not all  
312 perceived values of conservation (associated with either costs or benefits) will be tangible or  
313 easily quantifiable; yet assessing their relative importance has merit. Any type of equity in  
314 principle could be measured subjectively on a unitless scale of low to high. Formalization of  
315 problems that involve values can be an anathema to some, but the benefits of explicating  
316 integrating these issues into formal conservation planning are greater than ignoring perceived  
317 values altogether.

318

319 Social equity in conservation has emerged from concern for environmental justice and fairness,  
320 in particularly, for those groups most affected by conservation interventions or most dependent  
321 on natural resources for their livelihoods. These issues reflect two important key ethical  
322 considerations. The first, which has been the primary focus of this paper, relates to how social  
323 equity among and between different groups might be represented in the process or outcomes of  
324 conservation planning. The second relates more specifically to how different types of equity are  
325 defined, by whom and for which groups. Goals reflect the values and beliefs of those individuals  
326 or groups that set them. We have suggested several key types of equity, but these are by no  
327 means exhaustive or prescriptive. Rather we provide a conceptual basis for articulating types of  
328 equity, the possible relationships between equity and conservation outcomes, and ways to

329 interpret trade-offs among types of equity and between equity and conservation outcomes. Such  
330 a framework has the potential to inform and support rights-based approaches to conservation. It  
331 would be nearly impossible to consider all types of equity at once, thus conservation planners  
332 have to make some decision as to which types of equity to consider. Similar decisions are made  
333 when considering economic and ecological objectives, e.g., which actions to take to conserve  
334 which species (Bottrill *et al.*, 2008). How these decisions are made will depend on the local  
335 context in which the conservation intervention occurs, but we recommend an explicit conceptual  
336 framework to promote transparency and balance different perspectives.

337

338 Our conceptual foundation provides a lens through which issues of equity and conservation  
339 success can be viewed and studied using empirical data. This foundation informs further  
340 research required to resolve outstanding issues, including: 1) empirical evidence to document  
341 and measure the frequency of occurrence and effect of different types of social equity on the  
342 probability of conservation success; 2) information on whether minimum thresholds of equity are  
343 required to achieve conservation success (Figure 3a); 3) data on the contribution of equity versus  
344 other factors in affecting conservation success among different interventions, and potential  
345 tradeoffs among these factors; 4) a systematic review to synthesize existing evidence on which  
346 types of interventions, and their relative conservation success, are most influenced by which  
347 types of social equity; and 5) definitions and perceptions of conservation success among and  
348 across different groups and contexts. Greater knowledge of these issues will improve our  
349 understanding of how and when to consider equity in conservation decisions making.

350

351 Multiple objectives are common in conservation, yet there is not always consensus on objectives

352 among individuals and groups. Conservation planning can only strive to achieve the stated  
353 objectives and ensure that the objectives are clear, measurable and identified through a  
354 transparent and participatory process with multiple stakeholders. Governments and organizations  
355 are increasingly moving away from purely biophysical approaches to biodiversity conservation  
356 to more holistic approaches based on sustainable human interactions, which require integration  
357 of environmental, social, and economic demands. Although substantial work has been done to  
358 promote the need for addressing social factors in effective planning design and implementation  
359 (Ban *et al.* 2013), there has been little focus on social equity and its influence on conservation  
360 outcomes, despite the assumption that triple bottom line solutions are commonly held as ideal.  
361 We hope our work here will help improve conservation success by shedding light on how and  
362 why equity influences the probability of success, the consequences of not adequately considering  
363 equity on conservation outcomes, and provide guidance on tradeoffs among social equity,  
364 economic efficiency, and conservation effectiveness for conservation interventions.

## 365 **References**

- 366 Adams V, Pressey R, and Naidoo R. (2010). Opportunity costs: Who really pays for  
367 conservation? *Biol Conserv* **143**: 439–48.
- 368 Adelman I and Morris C. (1973). *Economic growth & social equity in developing countries.*  
369 Stanford: Stanford University Press.
- 370 Agarwal B. (2009). Gender and forest conservation: The impact of women’s participation in  
371 community forest governance. *Ecol Econ* **68**: 2785–99.
- 372 Axford JC, Hockings MT, and Carter RW. (2008). What constitutes success in Pacific island  
373 community conserved areas? *Ecol Soc* **13**: 45–61.
- 374 Ban NC, Mills M, Tam J, *et al.* (2013). A social-ecological approach to conservation planning:  
375 embedding social considerations. *Front Ecol Environ* **11**: 194–202.
- 376 Berkes F. 2004. Rethinking community-based conservation. *Conserv Biol* **18**: 621–30.



- 377 Borrini, Grazia, Ashish Kothari, and Gonzalo Oviedo. (2004) Indigenous and local communities  
 378 and protected areas: Towards equity and enhanced conservation: Guidance on policy and  
 379 practice for co-managed protected areas and community conserved areas. No. 11. IUCN,  
 380 Gland, Switzerland.
- 381 Bottrill, M., Joseph, L. N., Carwardine, J., Bode, M., Cook, C., Game, E. T., Grantham, H.,  
 382 Kark, S., Linke, S., McDonald-Madden, E., Pressey, R. L., Walker, S., Wilson, K. A. &  
 383 Possingham, H. P. (2008). Is conservation triage just smart decision-making? *Trends in*  
 384 *Ecology & Evolution*, 23, 649-654.
- 385 Bottrill MC and Pressey RL. (2012). The effectiveness and evaluation of conservation planning.  
 386 *Conserv Lett* 5: 407–20.
- 387 Brockington, D. (2004). Community conservation, Inequality and Injustice. Myths of Power in  
 388 Protected Area Management. *Conservation and Society* 2 (2): 411-432
- 389 Brown K. (2002). Innovations for conservation and development. *Geogr J* 168: 6–17.
- 390 Cancino JP, Wilen JE, and Uchida H. (2007). TURFs and ITQs : Collective vs . individual  
 391 decision making. University of Rhode Island. *Mar Resour Econ* 22: 391–406.
- 392 Claus C, Chan K, and Satterfield T. (2010). The roles of people in conservation. *Conserv Biol*  
 393 *all*: 262–83.
- 394 De Santo, E. M., P. J. S. Jones, and A. M. M. Miller. (2011) Fortress conservation at sea: a  
 395 commentary on the Chagos marine protected area. *Marine Policy* 35.2: 258-260.
- 396 Dobbs I.M. (1982). Discounting, intergenerational equity and the almost-anywhere dominance  
 397 criterion. *Futures* 14: 307–12.
- 398 Ferraro, P.J and Hanauer, M.M. (2014). Quantifying causal mechanisms to determine how  
 399 protected areas affect poverty through changes in ecosystem services and infrastructure.  
 400 *Proc Natl Acad Sci U S A* 111: 4332–7.
- 401 Fina, M. (2011). Evolution of Catch Share Management: Lessons from Catch Share Management  
 402 in the North Pacific, *Fisheries* 26(4): 164-177
- 403 Gleason M, McCreary S, Miller-Henson M, *et al.* (2010). Science-based and stakeholder-driven  
 404 marine protected area network planning: A successful case study from north central  
 405 California. *Ocean Coast Manag* 53: 52–68.
- 406 Halpern, B.S., Klein, C.J., Brown, C.J., *et al.* (2013). Achieving the triple bottom line in the face  
 407 of inherent tradeoffs among social equity, economic return and conservation. *Proc Natl*  
 408 *Acad Sci* doi:10.107.

- 409 Hatcher A, Jaffry S, Thebaud O, and Bennett E. (2000). Normative and social influences  
410 affecting compliance with fishery regulations. *Land Econ* **76**: 448–61.
- 411 Kittinger, John N., et al. (2014) A practical approach for putting people in ecosystem-based  
412 ocean planning." *Frontiers in Ecology and the Environment* 12.8: 448-456.
- 413 Klein CJ, Steinback C, Watts ME, et al. (2010). Spatial marine zoning for fisheries and  
414 conservation. *Front Ecol Environ* **8**: 349–53.
- 415 Macintosh A, Bonyhady T, and Wilkinson D. (2010). Dealing with interests displaced by marine  
416 protected areas: A case study on the Great Barrier Reef Marine Park Structural Adjustment  
417 Package. *Ocean Coast Manag* **53**: 581–8.
- 418 Martin A and Lemon M. (2001). Challenges for participatory institutions: The Case of Village  
419 Forest Committees in Karnataka, South India. *Soc Nat Resour* **14**: 585–97.
- 420 Mascia MB, Pailler S, Thieme ML, et al. (2014). Commonalities and complementarities among  
421 approaches to conservation monitoring and evaluation. *Biol Conserv* **169**: 258–67.
- 422 McClanahan TR, Cinner J, Kamukuru AT, et al. (2008). Management preferences, perceived  
423 benefits and conflicts among resource users and managers in the Mafia Island Marine Park,  
424 Tanzania. *Environ Conserv* **35**: 340.
- 425 Nightingale AJ. (2002). Participating or just sitting in? The dynamics of gender and caste in  
426 community forestry. *J For livelihood vol* **2**: 1.
- 427 Ostrom E. (1990). *Governing the Commons: The evolution of institutions for collective action.*  
428 Cambridge, United Kingdom: Cambridge University Press.
- 429 Parrish JD, Braun DP, and Unnasch RS. (2003). Are we conserving what we say we are?  
430 Measuring ecological integrity within protected areas. *Bioscience* **53**: 851–60.
- 431 Persha L, Agrawal A, and Chhatre A. (2011). Social and ecological synergy: local rulemaking,  
432 forest livelihoods, and biodiversity conservation. *Science* **331**: 1606–8.
- 433 Ravallion M. (2014). Income inequality in the developing world. *Science* **344**: 851–5.
- 434 Russ GR and Alcala AC. (1999). Management histories of Sumilon and Apo Marine Reserves,  
435 Philippines, and their influence on national marine resource policy. *Coral Reefs* **18**: 307–19.
- 436 Solar O and Irwin A. (2007). *A Conceptual Framework for Action on the Social Determinants of*  
437 *Health.*
- 438 Tsikata D and Golah P (Eds). (2010). *Land tenure, gender, and globalisation: research and*  
439 *analysis from Africa, Asia, and Latin America.* New Delhi and Ottawa: International  
440 *Development Research Centre and Zubaan.*

441 Webb EL, Maliano RJ, and Siar S V. (2004). Using local user perceptions to evaluate outcomes  
442 of protected area management in the Sagay Marine Reserve, Philippines. *Environ Conserv*  
443 **31**: 138–48.

444 White JW, Scholz AJ, Rassweiler A, *et al.* (2013). A comparison of approaches used for  
445 economic analysis in marine protected area network planning in California. *Ocean Coast*  
446 *Manag* **74**: 77–89.

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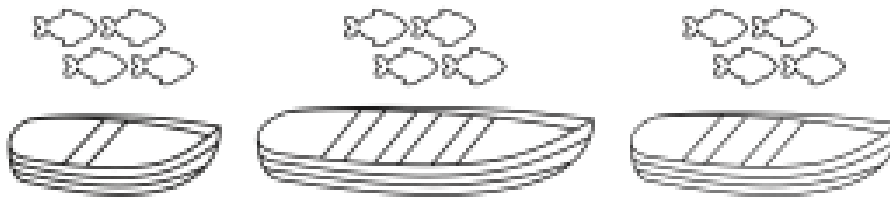
448

449 **Figures**

450

451 **Figure 1.**

(a) Absolute equity



(b) Relative equity

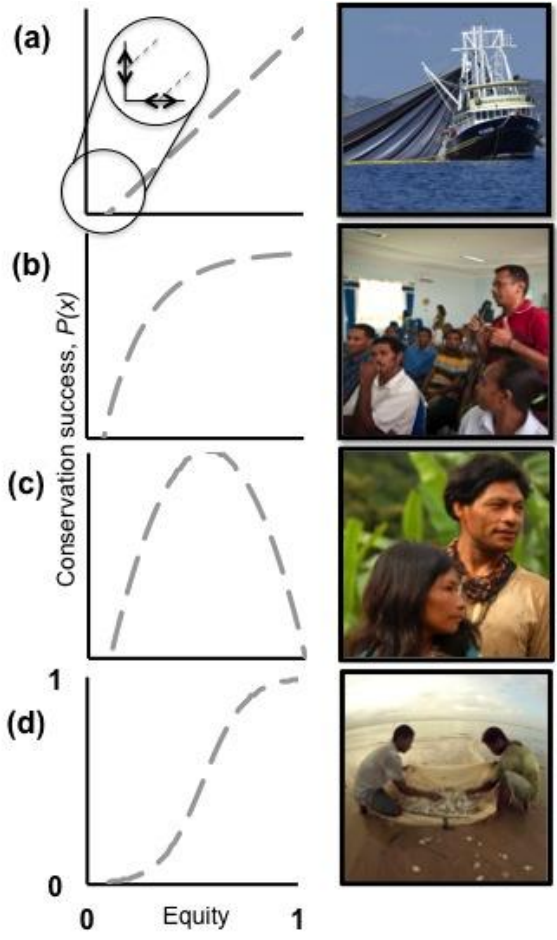


(c) Perceived equity



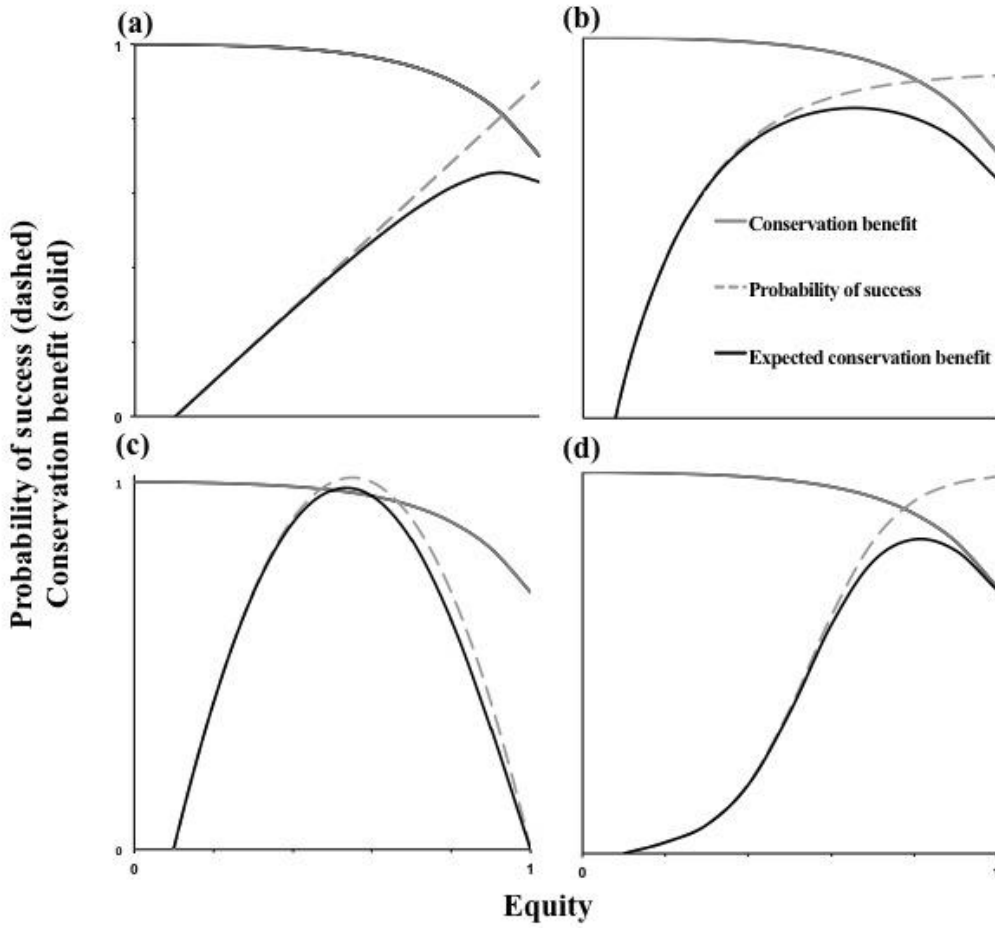
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453 **Figure 2**



454

455 **Figure 3**



456

457 **Figure 4**

458 **Figure Legends**

459 Figure 1. (a) Conservation success can be influenced by several different types of equity  
460 (described in b), both as an input into (e.g., participation by stakeholder groups) and/or an  
461 outcome of the conservation intervention (e.g., access to natural resources by individuals or  
462 groups). Each type of equity can be influenced by a variety of socioeconomic and political  
463 context determinants.

464

465 Figure 2. Equity influences conservation success in different ways, depending on how it is  
466 measured and perceived. Potential measures and perceptions are illustrated for access equity,  
467 where a management plan limits fishing access to different fisher groups (each with a different  
468 size boat). When measured in absolute terms (a), each group benefits equally, represented by  
469 catching the same number of fish; when measured in relative terms (b), the benefit is distributed  
470 proportionally to the size of the boat. (c) the group with the largest boat has a positive  
471 perception of the relative benefits, whereas groups with smaller boats have a negative  
472 perception.

473

474 Figure 3. Four broad classes of relationship between equity and the probability of conservation  
475 success,  $P(x)$ : (a) Linear; (b) Asymptotic; (c) Humped; and (d) Sigmoidal. A value of 1 indicates  
476 perfect equity and conservation success. For each relationship, we do not know where they cross  
477 an axis (shown in (a)). If there is a minimum threshold of equity, below which there is zero  
478 chance of success, then the lines would cross the x-axis; whereas if success is possible in

479 inequitable situations, the lines would intercept the y-axis. Photos represent equity types that can  
480 exhibit the associated relationship, occupational, participation, gender, and access, respectively.

481 Photos courtesy of (a) Ulrich Karlowski; (b) World Wildlife Fund, Inc. Tory Read; (c) Trond  
482 Larsen; (d) Cristina Mittermeier.

483 Figure 4. The relationship between equity and conservation benefit (i.e., success), and how  
484 different relationships between probability of success ( $P(x)$ , from Fig. 3), given different levels  
485 of equity modifies the ability of the conservation intervention to achieve biodiversity  
486 conservation outcomes. The solid gray line shows a general possible trade-off between  
487 conservation benefit and equity (taken from Halpern et al. 2013). The dashed gray lines show  
488 four possible relationships between equity and probability of success, described in Figure 3. The  
489 solid black lines are the resulting consequence of these probability relationships on the degree to  
490 which conservation success is achieved (expected conservation benefit).

491

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495



Figure 1a  
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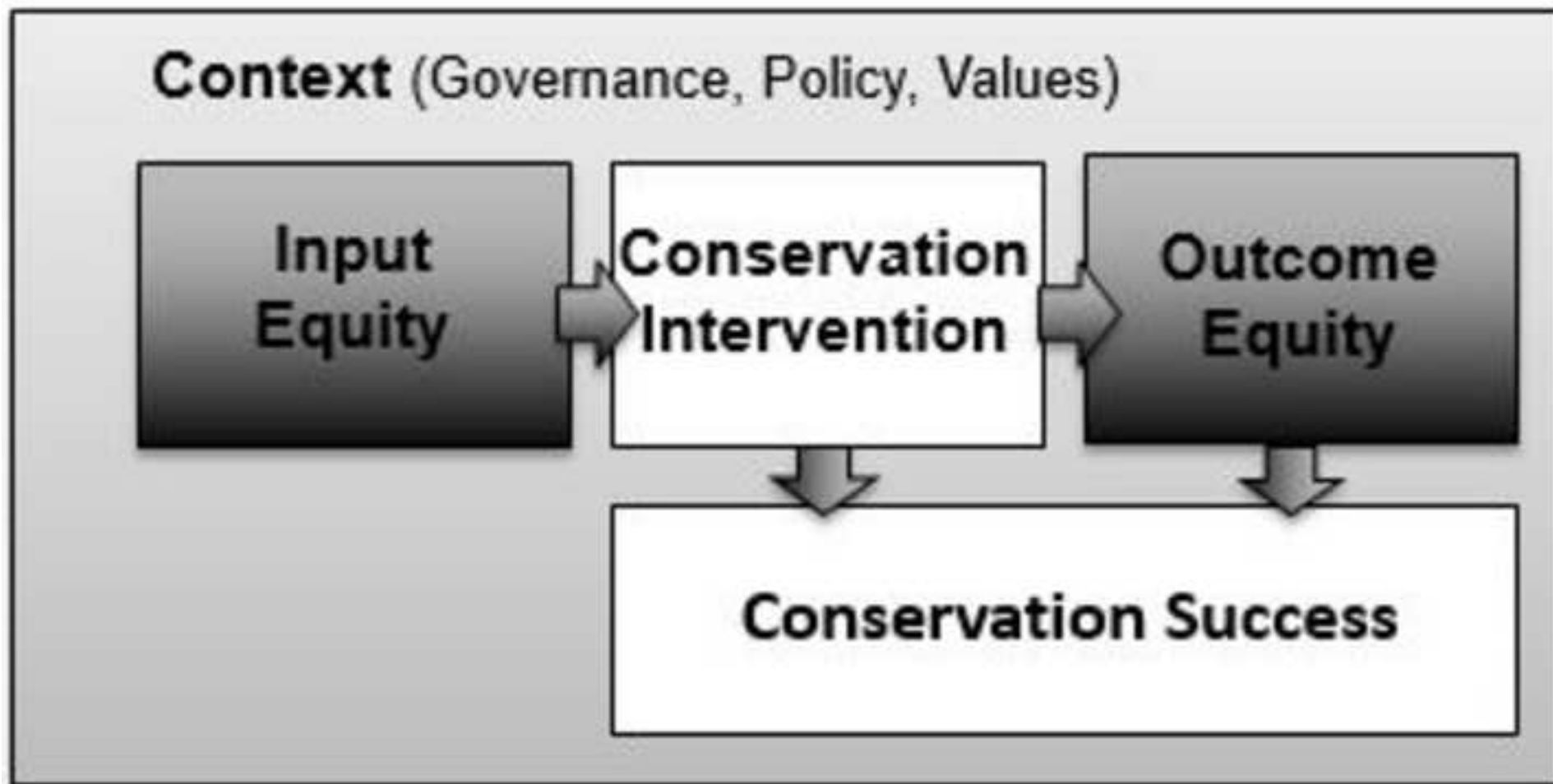


Figure 1b  
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Equity Type	Description
<b>Social class</b>	Distribution of benefits or costs to each class, or people, according to set of hierarchical social categories
<b>Gender</b>	Distribution of benefits or costs to each gender group, where 50:50 ratio is perfect equity
<b>Ethnicity</b>	Distribution of benefits or costs to individuals or groups by cultural or indigenous status
<b>Generational</b>	Distribution of costs or benefits to each generation
<b>Educational</b>	Distribution of benefits or costs based upon level of education attainment (e.g., primary, secondary, tertiary)
<b>Occupation</b>	Distribution of benefits or costs to each occupational sector, where a sector could be individual types of fisheries (e.g., crab, tuna), different fishery groups (e.g., commercial, recreational), or different industries (e.g., wind farming, tourism)

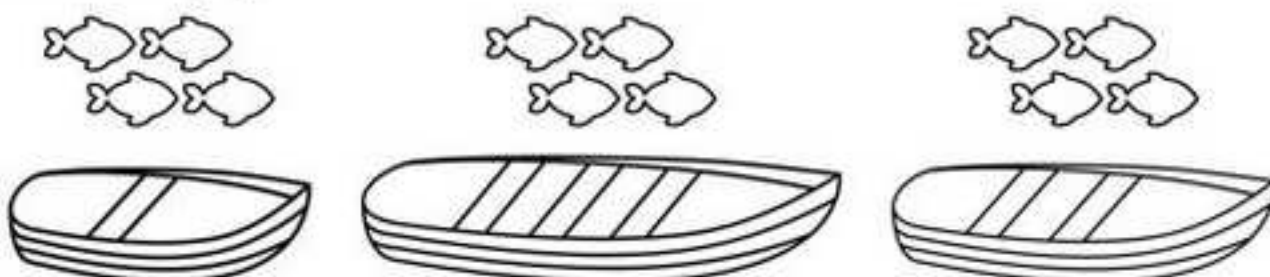
Figure 1c

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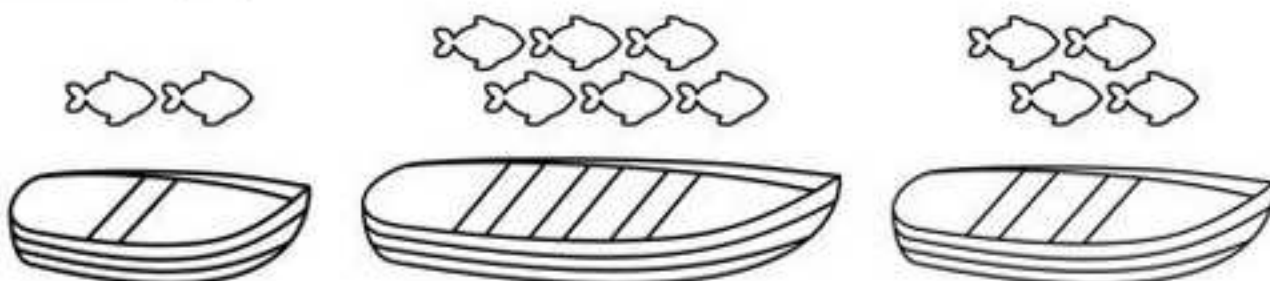
Equity Metric	Description
<b>Participation</b>	Representation by type of stakeholder group in participatory conservation process of decision
<b>Access</b>	Amount of non-spatial access allocated to individuals or groups, such as access (or no access) to natural resources during a particular time period or season.
<b>Spatial</b>	Amount of space or area on the landscape/seascape allocated to individuals or groups
<b>Financial</b>	Amount of income or profitability allocated to individuals or groups

Figure 2  
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(a) Absolute equity



(b) Relative equity



(c) Perceived equity

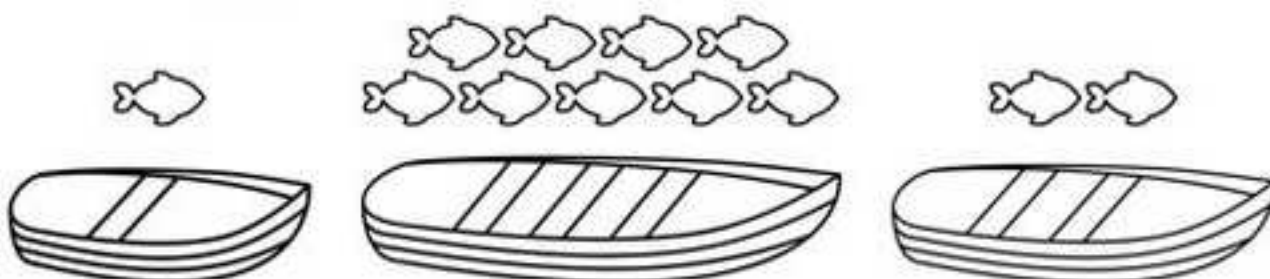


Figure 3  
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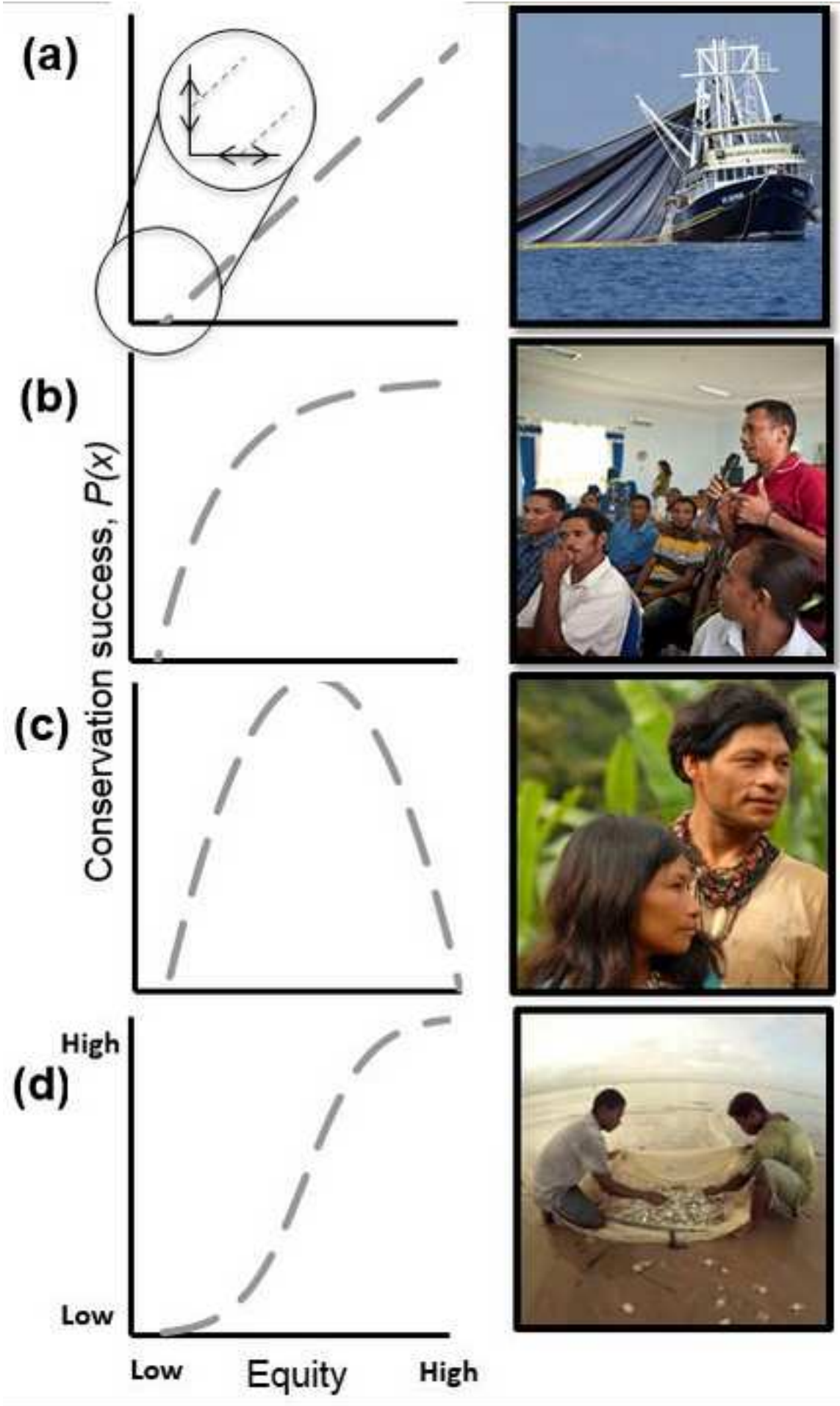


Figure 4  
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