

1 **Formation of Halogenated C-, N-DBPs from Chlor(am)ination**

2 **and UV Irradiation of Tyrosine in Drinking Water**

3 Wenhai Chu ^a, Naiyun Gao^{a,*}, Stuart W. Krasner ^b, Michael R Templeton ^c, Daqiang Yin^a

4 ^aState Key Laboratory of Pollution Control and Resources Reuse, College of
5 Environmental Science and Engineering, Tongji University, Shanghai, 200092, China

6 ^bMetropolitan Water District of Southern California, 700 Moreno Avenue, La Verne,
7 California 91750-3399, USA

8 ^cDepartment of Civil and Environmental Engineering, Imperial College London, London,
9 SW7 2AZ, UK

10
11
12
13 *Corresponding author.

14 Address: State Key Laboratory of Pollution Control and Resources Reuse, Tongji
15 University, Mingjing Building No. 601, 1239 Siping Road, Yangpu District, Shanghai,
16 200092, China

17 Tel: +86 21 65982691

18 Fax: +86 21 65986839

19 E-mail: feedwater@yahoo.cn, feedwater@126.com.

20 21 22 **Abbreviations:**

23 Alanine –Ala; aspartic acid –Asp; carbonaceous disinfection by-products –C-DBPs;
24 chloroform –CF; chloral hydrate –CH; 4-chlorophenol –4-CP; cyanogen chloride –CNCl;
25 1,1-dichloropropanone –1,1-DCP; disinfection by-products –DBPs; dichloroacetic acid
26 –DCAA; dichloroacetamide –DCAcAm; dichloroacetonitrile –DCAN; dichloromethane
27 –DCM; 2,4-dichlorophenol –2,4-DCP; dissolved organic carbon – DOC; dissolved organic
28 nitrogen – DON; gas chromatography/mass spectrometry -GC/MS; haloacetic acids –
29 HAAs; haloacetamides –HAcAms; haloacetonitriles –HANs; halonitromethanes –HNMs;
30 4-hydroxy-benzyl-cyanide –4-HBC; nitrogenous disinfection by-products- N-DBPs; natural
31 organic matter – NOM; purge & trap- P&T; trichloroacetic acid –TCAA; trichloroacetyl
32 chloride –TCAC; trichloroacetamide –TCAcAm; trichloroacetonitrile –TCAN;
33 trichloronitromethane –TCNM; 2,4,6-trichlorophenol –2,4,6-TCP; 1,1,1-trichloropropanone
34 –1,1,1-TCP; trihalomethanes –THMs; tyrosine –Tyr; U.S. Environmental Protection
35 Agency –USEPA; water treatment plants–WTPs.

39 **Abstract**

40 The formation of regulated and emerging halogenated carbonaceous (C-) and
41 nitrogenous disinfection by-products (N-DBPs) from the chlor(am)ination, UV irradiation of
42 tyrosine (Tyr) was investigated. Increased chlorine contact time and/or Cl₂/Tyr ratio
43 increased the formation of most C-DBPs. In contrast, 4-chlorophenol, dichloroacetonitrile,
44 and dichloroacetamide had their greatest yields at particular conditions. Chloroform and
45 dichloroacetic acid increased with increasing pH, dichloroacetonitrile first increased and
46 then decreased, and other DBPs had maximum yields at pH 7 or 8. The addition of
47 ammonia significantly reduced most C-DBPs formation but increased 4-chlorophenol,
48 dichloroacetonitrile, dichloroacetamide, and trichloroacetonitrile yields for short
49 pre-chlorination contact times before dosing ammonia. When UV irradiation and
50 chlorination were performed simultaneously, the relatively stable C-DBPs increased, and
51 dichloroacetonitrile, dichloroacetamide, and 4-chlorophenol decreased with increasing UV
52 dose. This information was used to develop a mechanistic model for the formation of
53 intermediate DBPs and endproducts from the interaction of disinfectants with tyrosine.

54

55 **Capsule abstract**

56 Exploring the integrated formation mechanism of regulated and emerging highly toxic
57 DBPs, which is expected to preferably reduce their occurrence in drinking water.

58 **Keywords:** Drinking water; Nitrogenous disinfection by-products; Tyrosine; Haloacetamides;
59 Halonitromethanes; Integrated formation pathway

60

61 **1. Introduction**

62 Currently, water resource shortages and growing water demands have spurred utilities
63 to exploit source waters impaired by treated wastewater effluents and/or algal blooms.
64 Such source waters are typically characterized by higher dissolved organic nitrogen (DON)
65 levels, where amino acids constitute an important class of the DON pool and may account
66 for 15% and 35% of the DON (Westerhoff et al., 2002). During chlorination or
67 chloramination, components of the DON in water can react with the disinfectant to form
68 halogenated nitrogenous disinfection by-products (N-DBPs), such as haloacetamides
69 (HAcAms), halonitromethanes (HNMs) and haloacetonitriles (HANs) (Richardson et al.,
70 2007), which represent an emerging concern due to their cytotoxicity and genotoxicity
71 (Plewa et al., 2004; Muellner et al., 2007; Richardson et al., 2011). In particular,
72 haloacetamides (HAcAms), an emerging class of halogenated N-DBPs that have been
73 measured in tap waters (Krasner et al., 2006), exhibited much higher genotoxicity and
74 cytotoxicity than many C-DBPs (e.g., trihalomethanes [THMs] and haloacetic acids
75 [HAAs]) (Plewa et al., 2008).

76 Most formation studies, based on model compounds such as free amino acids, focus
77 on C-DBPs (e.g., THMs and HAAs) and certain N-DBPs (HANs), relatively little involved
78 HAcAms and HNMs. Some laboratory studies reported that aromatic amino acids (e.g.,
79 tyrosine [Tyr]) generally produced more chloroform (CF) and other THMs than
80 non-aromatic ones, and amino acids with a ring structure (e.g., Tyr) resulted in higher
81 yields of dichloro- (DCAA) and trichloroacetic acid (TCAA) during chlorination (Hong et al.,
82 2009). Chlorination (oxidation) of certain amino acids (e.g., Tyr, tryptophan, aspartic acid

83 [Asp]) can result in the formation of aldehydes and nitriles, with subsequent or
84 concomitant chlorine substitution to form trichloroacetaldehyde (chloral hydrate [CH]) and
85 dichloroacetonitrile (DCAN), respectively (Trehy et al., 1986), whereas other amino acids
86 (e.g., glycine, alanine, serine) did not (Trehy and Bieber, 1981). Recently, selected amino
87 acids were shown to react with chlorine to form other N-DBPs, such as cyanogen chloride
88 (CNCl) in the case of glycine (Na and Olson, 2006) and trichloronitromethane (TCNM) in
89 the case of Asp (Hu et al., 2010). For HAcAms, there is only one study that assessed the
90 role of amino acids in the formation of dichloroacetamide (DCAcAm) (Chu et al., 2010b),
91 the most common HAcAm formed in chlorinated and chloraminated drinking water
92 (Krasner et al., 2006), and found that Asp, histidine, and Tyr had the highest DCAcAm
93 yields among 20 free amino acids.

94 In these studies on DBP formation during chlorination of amino acids, Tyr was a typical
95 precursor for certain C- and N-DBPs. Moreover, Tyr is a naturally occurring amino acid
96 present in many peptides, proteins, and algae (Ram, 1985; Szajdak and Österberg, 1996).
97 Mitch et al (2009) investigated the occurrence of amino acids in some US water treatment
98 plants (WTPs) influent samples, and found the maximum concentration of hydrolyzable Tyr
99 reached 27.4 µg/L (average=9.0 µg/L). In practical WTPs, pre- or post-chlor(am)ination
100 may caused the formation of multiple DBPs from Tyr and other compounds (e.g., proteins,
101 algae, humics) containing Tyr. Pre-chlor(am)ination is a common practice in China now to
102 reduce tastes, odors, and algal growths. Although the above-mentioned studies reported
103 the formation of individual DBP species from the chlorination of Tyr, a comprehensive
104 reaction scheme, which includes key intermediates resulting in multiple DBP formation, is

105 lacking to date. Also, a novel possible pathway for the emerging HAcAms formation was
106 proposed in the present study, which updated the previous speculation that HAcAms was
107 from the hydrolysis of HANs. Besides, chloramination in practice often does not involve
108 adding preformed chloramines, but rather adding ammonia and chlorine. The order of
109 addition and/or free chlorine contact time can impact DBP formation during
110 post-chloramination (Schreiber and Mitch 2009). Especially, the results for the formation
111 of some emerging N-DBPs (e.g., DCAcAm, trichloroacetamide [TCAcAm]) during certain
112 free chlorine contact times before adding ammonia were still lacking. Moreover, UV
113 treatment is an emerging disinfection process for drinking water, and transformation of the
114 NOM by UV irradiation could change the formation of some DBPs during post-chlorination
115 (Liu et al., 2006; Dotson et al., 2010; Reckhow et al., 2010). At present, the contribution of
116 UV irradiation on multiple DBP (esp. HAcAms) formation from selected individual amino
117 acids is unknown.

118 A better understanding of the comprehensive formation mechanism of C- and N-DBPs
119 during disinfection may improve the accuracy of predicting DBP occurrence and make it
120 possible to optimize disinfection practices that minimize the formation of C- and N-DBPs.
121 The objective of this study was to assess the role of Tyr in the formation of multiple
122 chlorinated C- and N-DBPs: CF, DCAA, TCAA, 1,1-dichloro-2-propanone (1,1-DCP),
123 1,1,1-trichloropropanone (1,1,1-TCP), CH, DCAN, trichloroacetonitrile (TCAN), TCNM,
124 DCAcAm, TCAcAm, CNCl, and 4-chlorophenol (4-CP). This was evaluated for
125 chlorination with and without UV irradiation and for chloramination under various
126 conditions. Moreover, an important goal was to explore the formation pathways of these

127 C- and N-DBPs, which is expected to contribute to a better understanding and prediction
128 of the control and formation of these DBPs.

129

130 **2. Materials and methods**

131 **2.1. Materials**

132 THMs (Supelco 47904), HAAs (Supelco 49107-U), 4-CP (Supelco 48689), and CH
133 (47335-U) chemical standards, and EPA 551B standard mixture (Supelco 48046)
134 containing DCAN, TCAN, 1,1-DCP, 1,1,1-TCP, and TCNM were purchased from
135 Sigma-Aldrich (St. Louis, Missouri, USA). Tyr (98.5%) was obtained from Wako (Osaka,
136 Japan). DCACAm (98.5%) and TCACAm (99%) were obtained from Alfa Aesar (Karlsruhe,
137 Germany). The extraction solvent ethyl acetate was obtained from Fisher Scientific
138 (Waltham, Massachusetts, USA). Guaranteed reagent (GR) grade reagents—sodium
139 hypochlorite (NaOCl), sodium hydroxide (NaOH), hydrochloric acid (HCl), ammonium
140 chloride, sodium nitrite, sodium nitrate, buffer salts, glacial acetic acid, ascorbic acid, and
141 anhydrous sodium sulfate—were purchased from Sinopharm Chemical Reagent Co., Ltd.
142 (Shanghai, China). All solutions were prepared using ultrapure water produced with a
143 Millipore Milli-Q Gradient water purification system (Billerica, Massachusetts, USA). All
144 bottles were prewashed with phosphate-free detergent, rinsed with ultrapure water, and
145 dried in an oven at 105°C for 24 h. Chlorine solutions were prepared by diluting a 6%
146 NaOCl solution with ultrapure water, which was standardized daily prior to use.

147 **2.2. Disinfection experiments**

148 Chlorination and chloramination experiments were conducted at a controlled room

149 temperature ($23.0 \pm 0.2^\circ\text{C}$) and under headspace-free conditions in 200-mL brown glass
150 volumetric flasks, which were kept in the dark. During chlorination experiments, a typical
151 run involved applying a certain chlorine dose (0.5, 1.5, 2.5 mM) to a Tyr solution (0.1 mM)
152 for the designated reaction time. To stop the chlorination reaction, the disinfectant residual
153 was quenched with ascorbic acid with a normality twice as high as the initial normality of
154 the chlorine added, because ascorbic acid has little effect in the stability and analysis of
155 these investigated halogenated N-DBPs and C-DBPs (Joo and Mitch, 2007; Chu et al.,
156 2009a). NaOH and HCl were used to adjust the solution pH value (pH=5~9). Buffer
157 solutions were prepared from phosphate and carbonate salts. For comparing the effect of
158 chloramination on DBP formation, certain amounts of ammonium chloride were added to
159 provide ammonia at the desired level. By adjusting the free chlorine contact time before
160 dosing ammonia, three chloramination schemes were examined: I. Chloramination with
161 preformed monochloramine, where ammonia and chlorine with the same molar
162 concentration were dosed to the Tyr solution. II. Prechlorination (1-h), where ammonium
163 chloride with the same molar concentration as the residual chlorine after 1 h was added to
164 a Tyr solution pre-chlorinated for 1 h. III. Prechlorination (6-h), where ammonium chloride
165 with the same molar concentration as the residual chlorine after 6 h was added to a Tyr
166 solution pre-chlorinated for 6 h. In order to examine the effect of UV irradiation on the
167 formation of DBPs during chlorination of Tyr, before chlorination disinfection or at the
168 same time, some water samples were irradiated in a low pressure UV reactor with
169 different UV doses from 19.5 to 585 mJ/cm^2 , as shown in Supplementary Information (SI)
170 (Figure SI and Table S1).

171 **2.3. Analytical Methods**

172 A UV/Vis double-beam spectrophotometer (Unico4802, Dayton, New Jersey, USA)
173 was used to scan the spectrum of different samples from 220 to 500 nm at 10-nm intervals.
174 Residual and total chlorine were detected by a portable spectrophotometer (HACH
175 DR2800) based on HACH method 8021. Prior to DBP analysis, glacial acetic acid was
176 used to lower the pH to 4.8-5.5 for the THM, HAN, and TCNM samples, and to 5.0 ± 0.2
177 for the HAcAm samples to prevent base-catalyzed hydrolysis of HANs or HAcAms (Chu et
178 al., 2009a). CF, CH, DCAN, TCAN, TCNM, 1,1-DCP, and 1,1,1-TCP were measured using
179 purge & trap (P&T) (OI Analytical, Eclipse 4660, College Station, Texas, USA) and gas
180 chromatography/mass spectrometry (GC/MS) (Shimadzu-QP2010, Kyoto, Japan), based
181 on the U.S. Environmental Protection Agency (USEPA) Method 524.2. Two HAAs (DCAA
182 and TCAA) were measured by GC (Shimadzu-QP2010) with an electron capture detector,
183 based on USEPA Method 552.2 (Gao et al., 2009). DCACAm and TCACAm were analyzed
184 using liquid-liquid extraction and GC/MS (Shimadzu-QP2010). The analysis details of
185 DCACAm and TCACAm are available elsewhere (Chu et al., 2010b). Additionally, the
186 details of the 4-CP analysis are briefly described in the SI. This study did not quantify the
187 concentration of CNCl or dichloromethane (DCM), but they were examined qualitatively by
188 P&T and GC/MS (Figure S3). The same was done for trichloroacetyl chloride (TCAC). The
189 yield of each DBP was calculated by the molar ratio of the formed DBP to the initial
190 concentration of Tyr (Eq. S1).

191 **3. Results and Discussion**

192 3.1. Time-dependent formation of DBPs from Tyr

193 [Fig.1]

194 Figures 1A and B show the time-dependent formation of C- and N-DBPs during
195 chlorination of tyrosine at a Cl_2/Tyr molar ratio of 15 at pH 7. There was a free chlorine
196 residual at all times tested at $\text{Cl}_2/\text{Tyr} = 15$ (Figure S4) because the chlorine demand was
197 approximately 13 mol $\text{Cl}_2/\text{mol Tyr}$. As shown in Table S2, at $\text{Cl}_2/\text{Tyr} = 15$ (as well as at 5 or
198 25), 1,1-DCP, 1,1,1-TCP, and TCACAm were not detected, and the yields of TCAN and
199 TCNM were lower than 0.02%, which were too low to distinguish a notable difference.
200 From Figures 1A and B ($\text{Cl}_2/\text{Tyr} = 15$), the yields of 4-CP (2.8%) and DCAN (1.6%) were
201 much higher than other tested DBPs at the first 1 h, and they gradually decreased by 168
202 h to 0.05% and 0.11%, respectively. (Their half-lives were >6 and <6 h, respectively.) The
203 concentrations of four detected C-DBPs (CF, DCAA, TCAA, and CH) increased with
204 increasing contact time and maximized at 4.0, 2.4, 2.9 and 0.46% at 168 h. The yield of
205 the relatively unstable N-DBP DCACAm (Chu et al., 2009a) peaked at 0.16% at 6 h, and
206 then declined at higher contact times (down to 0.05% at 168 h).

207 CF, DCAA, and TCAA were relatively stable in the presence of chlorine and generally
208 were the endproducts of chlorination of Tyr in the absence of bromide. CH can form CF by
209 base-catalyzed hydrolysis under alkaline conditions (Chu et al., 2009b), but the hydrolysis
210 rate of CH is lower than the corresponding formation rate in neutral solution. The trends of
211 DCACAm formation and degradation can be explained by the hydrolysis of DCAN and
212 DCACAm. DCAN can hydrolyze to form DCACAm (e.g., from 1 to 6 h), and DCACAm can
213 further hydrolyze to produce DCAA (e.g., after 6 h) (Table S3, Eq. S2) (Glezer et al., 1999;

214 Reckhow et al., 2001; Chu et al., 2009b; Chu et al., 2010a). Note that the loss of DCAN
215 cannot be fully accounted for by the increases in formation of DCACAm or DCAA.

216 **3.2. Effect of chlorine dosage**

217 The yields of C- and N-DBPs after a 24-h chlorination of Tyr at different Cl_2/Tyr ratios
218 are shown in Figures 1C and D. Four relatively stable C-DBPs (CF, DCAA, TCAA, CH)
219 increased with increasing Cl_2/Tyr , whereas 4-CP decreased with increasing Cl_2/Tyr . The
220 yields of some N-DBPs (DCAN and DCACAm) reached their highest levels at $\text{Cl}_2/\text{Tyr} = 15$.
221 At $\text{Cl}_2/\text{Tyr} = 5$, residual chlorine was barely detected at 0.02 mg/L detection limit (Figure
222 S4), which caused all DBPs except 4-CP to have lower yields (Table S2). At $\text{Cl}_2/\text{Tyr} = 25$,
223 the excess chlorine accelerated the decomposition rate of DCAN (Reckhow et al., 2001),
224 whereas DCACAm was similar to what was present at $\text{Cl}_2/\text{Tyr} = 15$. CNCl and DCM were
225 detected at $\text{Cl}_2/\text{Tyr} = 5$. There was no significant change in CNCl level formed at different
226 chlorine contact times, whereas the DCM peak area gradually decreased with increasing
227 contact time (Figure S3). Na and Olson (2006) demonstrated that the chlorination of
228 glycine could form CNCl, where higher chlorine concentrations promoted its hydrolysis.
229 The absence of CNCl during chlorination of Tyr at $\text{Cl}_2/\text{Tyr} = 15$ or 25 was most likely due to
230 hydrolytic degradation of the CNCl by the free chlorine.

231 **3.3. Effect of pH**

232 **[Fig.2]**

233 The yields of C- and N-DBPs upon 24-h chlorination of Tyr at different pH levels are
234 summarized in Figures 2A and B, respectively. As shown, CF and DCAA yields kept
235 growing with increasing pH from 5 to 9. However, during the chlorination of humic acid,

236 increase in pH increases CF formation, but decreases DCAA formation (Babcock and
237 Singer, 1979). The impact of pH (e.g., from 7 to 9) was higher for CF (yield at pH 9 was
238 147% higher than yield at pH 7) than for DCAA (yield at pH 9 was 124% higher than yield
239 at pH 7). DCAN reached a maximum yield of 2.4% at pH = 6 and dropped substantially (to
240 a 0.35% yield) at pH 7. The highest formation of TCAA, CH, and DCACAm occurred at pH
241 7 to 8, which had maximum yields of 0.76-0.78, 0.18-0.24, and 0.12-0.14%, respectively.
242 However, 4-CP yields (0.59 to 0.74%) were relatively insensitive to pH. At different pH
243 levels, chlorination of Tyr was not able to produce sufficient TCNM to make notable
244 differences. TCAN gave low yields at pH 5 to 8 and was not detected at pH 9, and
245 TCACAm was not detected at all.

246 The effect of pH on the DBP yields is mainly attributable to their stability and formation
247 pathway from Tyr. Many of these results are consistent with previous research in bulk
248 water (e.g., conducted at pH levels of 5, 7, and 9.4) that has shown CF formation
249 increasing with pH; whereas CH formation increased over time at pH 5 and 7, and CH that
250 had formed within 4 h at pH 9.4 decayed over time at the elevated pH; and DCAN
251 formation only increased over time at pH 5, while DCAN formed within 4 h at pH 7
252 decayed over time at the neutral pH, and DCAN formed to a low extent at all reaction
253 times at pH 9.4 (Stevens et al., 1989). Relatively stable CH and TCAA, and relatively
254 unstable TCAN, DCAN, and DCACAm are also all easily hydrolyzed to CF or DCAA under
255 alkaline conditions (Yang et al., 2007; Chu et al., 2009b). The incremental increase in
256 DCAA under alkaline conditions was due in part to the hydrolysis of DCAN and DCACAm.
257 As shown in Table S4, with pH increasing from 7 to 9, the sum of the losses of DCAN and

258 DCACAm (0.41%) was greater than the increase in DCAA (0.17%). Alternatively, the sum
259 of the losses of TCAA, CH, and TCAN (0.089%) was far lower than the increase in CF
260 (0.52%). Therefore, there are other reasons to cause the incremental increase in CF
261 rather than only the hydrolysis of TCAA, CH, and TCAN. For example, the alkaline
262 environment can facilitate the rapid opening of the benzene ring, which can promote the
263 formation of CF (Chu et al., 2009c)

264 Apparent first-order rate constants for the decomposition of DCAN in the presence or
265 absence of chlorine were $1.8 \times 10^{-5} \text{ s}^{-1}$ (initial free chlorine was 10 mg/L, pH = 7.0) and 7.5
266 $\times 10^{-7} \text{ s}^{-1}$ (pH = 7.5) (Reckhow et al., 2001; Yang et al., 2007). These were higher than the
267 apparent first-order rate constants for the decomposition of DCACAm, which were $4.55 \times$
268 10^{-6} s^{-1} (initial free chlorine was 10 mg/L, pH = 7.0) and $2.37 \times 10^{-7} \text{ s}^{-1}$ (pH = 7.5) (Chu et
269 al., 2009a). At pH 7 to 8, the difference between the formation rate of DCACAm from
270 DCAN hydrolysis and the hydrolysis rate of DCACAm was probably higher than that at
271 other pH levels, as the net yield of DCACAm was highest at pH 7 to 8. At pH 8 the
272 maximum formation of CH, similar to that of DCACAm, was probably due to a larger
273 difference between CH formation and decomposition at this pH level. TCNM was relatively
274 stable at pH 5 to 9, in agreement with earlier studies (Joo and Mitch, 2007), and there was
275 no observed effect of pH on the TCNM yield.

276 **3.4. Effect of chloramination**

277 The yields of C- and N-DBPs after a 24-h disinfection of Tyr with different
278 chloramination schemes are summarized in Figures 2C and D, respectively. The use of
279 chloramines with less free chlorine contact time reduced the yields of the C-DBPs, except

280 for 4-CP, which is in agreement with studies with NOM or algae as the precursor (Zhang et
281 al., 2000; Fang et al., 2010). With increasing free chlorine contact time, relatively stable
282 C-DBPs (CF, DCAA, TCAA, and CH) increased, whereas 4-CP and N-DBP yields
283 increased and then decreased. Yang et al. (2007) found that a short period (1 to 30 min) of
284 chlorination of NOM solutions before switching to chloramination did not form substantial
285 quantities of these DBPs, except for DCaAm and TCaAm. A similar phenomenon was
286 also found in this study using Tyr as a precursor, whereas some significant differences in
287 some DBP yields were found when the prechlorination time was increased to 1 or 6 h
288 before dosing ammonia. However, the typical prechlorination time is less than 6 h in water
289 treatment plants.

290 Of note, the DCaAm yield was higher during 6-h prechlorination and 18-h
291 chloramination (0.17%) than 24-h chlorination alone (0.12%), whereas its yield was similar
292 to 6-h chlorination alone (0.16%). This was probably due to the stability and formation
293 mechanism of DCaAm (Chu et al., 2010a). Although DCAN was relatively more stable in
294 monochloramine solutions than in free chlorine solutions (Yang et al., 2007), some DCAN
295 can still hydrolyze to DCaAm during chloramination. Moreover, the hydrolysis rate of
296 DCaAm in monochloramine solutions was much lower than that in free chlorine solutions
297 (Chu et al., 2009a), thus this probably resulted in the net yield of DCaAm to reach its
298 highest level after 6 h of prechlorination and 18 h of chloramination.

299 The last detail of note is that TCaAm was first detected in this study during two
300 chloramination schemes, which were chloramination with preformed monochloramine and
301 1-h prechlorination before dosing ammonia. For these two chloramination schemes, the

302 formation of TCAC was tentatively identified (Figure S5). TCAC can react with ammonia to
303 form TCACAm (Montalbetti and Falque, 2005) (Table S3, Eq. S3). No TCAC was detected
304 during 24-h chlorination or 6-h prechlorination, probably because the TCAC formed was
305 further oxidized by chlorine or hydrolyzed to TCAA (Table S3, Eq. S4).

306 **3.5. Effect of UV irradiation**

307 **[Fig.3]**

308 A significant increase of the studied DBPs during chlorination after UV exposure was
309 not detected by GC/MS. Moreover, as shown in Figure S6, the optical spectra of Tyr were
310 similar or slightly larger with UV doses ranging from 19.5 to 585 mJ/cm², suggesting that
311 Tyr was relatively unchanged during UV irradiation. Previous studies have shown some
312 structural changes of NOM during UV irradiation, which resulted in an increase in some
313 DBPs (e.g., CF) after UV exposure and subsequent chlorination (Liu et al., 2006; Dotson
314 et al., 2010; Reckhow et al., 2010). However, in another recent study, there was an impact
315 from medium pressure UV whereas there was no impact from low pressure UV (Reckhow
316 et al., 2010), and it was a low pressure system that was used in this study.

317 The formation of various C- and N-DBPs was evaluated when both UV irradiation and
318 chlorination were conducted at the same time. This resulted in an increase in CF,
319 somewhat of an increase in DCAA and TCAA, and decreases in 4-CP, DCAN, and
320 DCAcAm.

321 As shown in Table S3, Eq. S5, the DCAN and DCAcAm formation pathway during
322 chlorination was proposed to include substitution, elimination, and decarboxylation
323 reactions, and a further substitution reaction to the methylene group (-CH₂-) in the main

324 chain of an amino acid (Reckhow et al.,2001; Chu et al., 2010b), where -CH₂- is an
325 electron-donating group. The chlorine substitution reaction rate in -CH₂- of Tyr was faster
326 than many other amino acids, probably because of the R- group including an aromatic
327 ring in the side chain of Tyr, which enhanced the electron-donating ability of -CH₂- in the
328 main chain of Tyr (Chu et al., 2010b). The aromatic ring in the R- group was probably
329 broken by a strong oxidizer (e.g., hydroxyl radical, Table S3, Eq. S6) formed under the
330 combined action of UV irradiation and chlorine (Nowell and Hoigne, 1992; Feng et al.,
331 2007), and the extent of the damage to the R- group in the side chain of Tyr increased with
332 increasing UV dose. This probably lessened the electron-donating ability of -CH₂- in the
333 main chain of Tyr and caused the decrease in DCAN and DCAcAm with increasing UV
334 doses. The decrease in 4-CP was also likely caused by the damage of the aromatic ring in
335 the side chain of Tyr. The increase in CF, DCAA, and TCAA could have been caused by
336 the strong oxidizer transforming the Tyr structure to a form that was more reactive with
337 chlorine to form CF, DCAA, and TCAA.

338 **3.6. Preliminary hypothesis of C- and N-DBPs formation pathways from Tyr**

339 **[Scheme 1]**

340 A possible integrated pathway of C- and N-DBP formation during Tyr chlorination is
341 proposed in Scheme1. For the reactions A1 to A7, B6, C1 to C6, D1 to D2, and E1 to E4 in
342 Scheme1, the possible formation pathways of DCAN, DCAcAm, DCAA, CH, and TCAN
343 from chlorination of Tyr was modified from previous studies (Reckhow et al., 2001; Joo
344 and Mitch, 2007; Hong et al., 2009; Chu et al., 2010b). From reaction C1 to C10
345 (secondary reaction), relatively small amounts of CF was formed from further chlorination

346 of Ala (Chu et al., 2009b). From reaction B1 to B5 (main reaction), the formation of CF
347 mainly goes through 4-CP, 2,4-dichlorophenol (2,4-DCP), 2,4,6-trichlorophenol
348 (2,4,6-TCP), and a ring opening reaction, which has been confirmed in an earlier study
349 (Chu et al., 2009c). As shown in Figure S7 and Table S5, 4-CP, 2,4-CP, and 2, 4, 6-TCP
350 were formed to much greater extents during chlorination, which was in agreement with
351 above-mentioned results of DBP yields.

352 Additionally, 4-hydroxyl-benzyl cyanide (4-HBC) during chloramination of Tyr had a
353 high peak (Figure S7), it was likely the main intermediate to form DCACAm during
354 chloramination by the reactions A1 to A4 (Scheme 1). During chlorination, the formation of
355 DCACAm was probably from the reactions E1, E2, E3, and A4, where benzyl cyanide
356 (Figure S7) was an important intermediate. TCACAm was less likely to occur from the
357 hydrolysis of TCAN, because TCAN yields were relatively low at all contact times tested.
358 As involved in "3.4. Effect of chloramination", TCACAm was detected, and TCAC was also
359 identified tentatively, during the same chloramination schemes. A novel possible formation
360 pathway for TCACAm was proposed; TCAC (Figure S5) could be produced by the ring
361 opening (main reaction B7) and was also probably from chlorination of formaldehyde
362 (secondary reaction C7). Also, it has been well known that TCAC could react quickly with
363 ammonia to yield TCACAm (reaction C8) (Montalbetti and Falque, 2005). We will confirm
364 the novel possible formation pathway by quantifying TCAC and other intermediates at
365 different experimental conditions in further studies.

366 **4. Conclusion**

367 Amino acids are an important component of the DON in water. Algal activity or treated
368 wastewater discharges are important sources of DON in watersheds. Amino acids are
369 precursors to certain C- and N-DBPs. Recent research has indicated that certain
370 emerging DBPs (e.g., HANs, HAcAms, HNMs, haloacetaldehydes) are more toxic than
371 regulated C-DBPs (THMs, HAAs). Thus, it is important to better understand the formation
372 and control of regulated and emerging DBPs. It was found that factors that increased the
373 formation of some by-products of Tyr decreased the formation of others. Moreover, the
374 degradation of some DBPs resulted in the formation of other DBP. The information in this
375 study was used to augment the scheme proposed for the formation of DBPs from Tyr.

376 Increased chlorine dose and pH could decrease the production of some N-DBPs, but
377 increase the formation of the relatively stable C-DBPs. If a utility had low C-DBP formation
378 potential but was experiencing halogenated N-DBPs (e.g., HAcAms, HANs), it is
379 recommended to try to switch chlorination to chloramination, and chlorine should be
380 dosed after ammonia, based on the results of this study. Meanwhile, it is necessary to
381 monitor the concentration of non-halogenated N-DBPs (e.g., nitrosamines) because
382 chloramines may increase their formation. Low-pressure UV before chlorination did not
383 impact the formation of DBPs from Tyr, whereas with post-chlorination, increased levels of
384 some DBPs were found with UV was used together with chlorine for disinfection. In the
385 future study on health effects and risk assessment in drinking water, the effect of some
386 emerging DBPs may be considered more important than regulated DBPs because their
387 concentration and toxic potency could be much greater than these regulated DBPs in
388 certain disinfection conditions

389 **Acknowledgments**

390 This project was supported by the national major science and technology project of
391 China (2008ZX07421-002), International Science & Technology Cooperation Program of
392 China (2010DFA91800), Postdoctoral Science Foundation of China (20110490073) and
393 Shanghai Postdoctoral Sustentation Fund of China (11R21415800). The authors sincerely
394 appreciate comments and revision suggestions from associate editor Prof. Kevin C. Jones
395 (Lancaster University) which improved this manuscript substantially.

396 **References**

- 397 Babcock DB, Singer PC. Chlorination and coagulation of humic and fulvic acids. *J Am Water*
398 *Works Assoc* 1979; 71: 149–58.
- 399 Chu WH, Gao NY, Deng Y. Stability of newfound nitrogenous disinfection by-products:
400 Haloacetamides in drinking water. *Chinese J Org Chem* 2009a; 29: 1569-74.
- 401 Chu WH, Gao NY, Deng Y., Dong BZ. Formation of chloroform during chlorination of alanine in
402 drinking water. *Chemosphere* 2009b; 77: 1346-51.
- 403 Chu WH, Gao NY, Zhao SJ, Deng HP. The mechanism analysis of formation of chloroform
404 during typical dissolved organic nitrogen tyrosine chlorination in drinking water. *Acta Chimica*
405 *Sinica* 2009c; 67: 2505-10.
- 406 Chu WH, Gao NY, Deng Y. Formation of haloacetamides during chlorination of dissolved
407 organic nitrogen aspartic acid. *J Hazard Mater* , 2010a; 173: 82-6.
- 408 Chu WH, Gao NY, Deng Y, Krasner SW. Precursors of dichloroacetamide, an emerging
409 nitrogenous DBP formed during chlorination or chloramination. *Environ Sci Technol* 2010b; 44:
410 3908-12.
- 411 Dotson AD, Keen VS, Metz D, Linden KG. UV/H₂O₂ treatment of drinking water increases
412 post-chlorination DBP formation. *Water Res* 2010; 44: 3703-13.
- 413 Fang J, Ma J, Yang X, Shang C. Formation of carbonaceous and nitrogenous disinfection
414 by-products from the chlorination of *Microcystis aeruginosa*. *Water Res* 2010; 44: 1934-40.
- 415 Feng YG, Smith DW, Bolton JR. Photolysis of aqueous free chlorine species (HOCl and OCl⁻)
416 with 254 nm ultraviolet light. *J Environ Eng Sci* 2007; 6: 277-84.
- 417 Gao NY, Chu WH, Deng Y, Xu B. TCAA degradation in ultraviolet (UV) irradiation/hydrogen
418 peroxide (H₂O₂)/micro-aeration (MCA) combination process. *J Water Supply Res*
419 *Technol–Aqua* 2009; 58: 510-8.
- 420 Glezer V, Harris B, Tal N, Iosefzon B, Lev O. Hydrolysis of haloacetonitriles: Linear free energy
421 relationship, kinetics and products. *Water Res* 1999; 33: 1939-48.

422 Hong HC, Wong MH, Liang Y. Amino acids as precursors of trihalomethane and haloacetic
423 acid formation during chlorination. *Arch Environ Contam Toxicol* 2009; 56: 638-45.

424 Hu J, Song H, Addison JW, Karanfil T. Halonitromethane formation potentials in drinking
425 waters. *Water Res* 2010; 44: 105-14.

426 Joo SH, Mitch WA. Nitrile, aldehyde, and halonitroalkane formation during
427 chlorination/chloramination of primary amines. *Environ Sci Technol* 2007; 41: 1288-96.

428 Krasner SW, Weinberg HS, Richardson SD, Pastor SJ, Chinn R, Scilimenti MJ, Onstad GD,
429 Thruston AD. Occurrence of a new generation of disinfection byproducts. *Environ Sci Technol*
430 2006; 40: 7175-85.

431 Liu W, Cheung LM, Yang X, Shang C. THM, HAA and CNCI formation from UV irradiation and
432 chlor(am)ination of selected organic waters. *Water Res.* 2006; 40: 2033-43.

433 Mitch WA, Krasner SW, Westerhoff P, Dotson A. Occurrence and Formation of Nitrogenous
434 Disinfection By-Products. Denver, Colo: American Water Works Association Research
435 Foundation, 2009.

436 Montalbetti C, Falque V. Amide bond formation and peptide coupling. *Tetrahedron* 2005; 61:
437 10827-52.

438 Muellner MG, Wagner ED, McCalla K, Richardson SD, Woo YT, Plewa MJ. Haloacetonitriles vs.
439 regulated haloacetic acids: Are nitrogen-containing DBPs more toxic? *Environ Sci Technol*
440 2007; 41: 645-51.

441 Na CZ, Olson TM. Mechanism and kinetics of cyanogen chloride formation from the
442 chlorination of glycine. *Environ Sci Technol* 2006; 40: 1469-77.

443 Nowell LH, Hoigne J. Photolysis of aqueous chlorine at sunlight and ultraviolet wavelengths. I.
444 Degradation rates. *Water Res* 1992; 26: 593-8.

445 Plewa MJ, Wagner ED, Jazwierska P, Richardson SD, Chen PH, McKague AB.
446 Halonitromethane drinking water disinfection byproducts: Chemical characterization and
447 mammalian cell cytotoxicity and genotoxicity. *Environ Sci Technol* 2004; 38: 62-8.

448 Plewa MJ, Muellner MG, Richardson SD, Fasano F, Buettner KM, Woo YT, McKague AB,
449 Wagner ED. Occurrence, synthesis, and mammalian cell cytotoxicity and genotoxicity of
450 haloacetamides: An emerging class of nitrogenous drinking water disinfection byproducts.
451 *Environ Sci Technol* 2008; 42: 955-61.

452 Ram NM. A review of the significance and formation of chlorinated N-organic compounds in
453 water supplies including preliminary studies on the chlorination of alanine, tryptophan, tyrosine,
454 cytosine, and syringic acid. *Environ Int* 1985; 11: 441-51.

455 Reckhow DA, Linden KG, Kim J, Shemer H, Makdissy G. Effect of UV treatment on DBP
456 formation. *J Am Water Works Assoc* 2010; 102: 100-13.

457 Reckhow DA, Platt TL, MacNeill AL, McClellan JN. Formation and degradation of
458 dichloroacetonitrile in drinking waters. *J Water Supply Res Technol-Aqua* 2001; 50: 1-13.

459 Richardson SD, Plewa MJ, Wagner ED, Schoeny R, DeMarini DM. Occurrence, genotoxicity,
460 and carcinogenicity of regulated and emerging disinfection by-products in drinking water: A
461 review and roadmap for research. *Mutation Res* 2007; 636: 178-242.

462 Richardson SD, Ternes TA. Water analysis: Emerging contaminants and current issues. *Anal*
463 *Chem* 2011; 83 4614-48.

464 Schreiber IM, Mitch WA. Influence of the order of reagent addition on NDMA formation during
465 chloramination. *Environ Sci Technol* 2009; 39: 3811-8.

466 Stevens AA, Moore LA, Miltner RJ. Formation and control of non-trihalomethane disinfection
467 by-products. *J Am Water Works Assoc* 1989; 81: 54-60.

468 Szajdak L, Österberg R. Amino acids present in humic acids from soils under different
469 cultivations *Environ Int* 1996; 22: 331-4.

470 Trehy ML, Bieber TI, 1981. Detection, identification and quantitative analysis of
471 dihaloacetonitriles in chlorinated natural waters, in: Keith LH (Ed.), *Advances in the*
472 *Identification & Analysis of Organic Pollutants in Water, Vol. 2.* Ann Arbor Sci. Publ., Inc., Ann
473 Arbor, Mich., pp. 941-75.

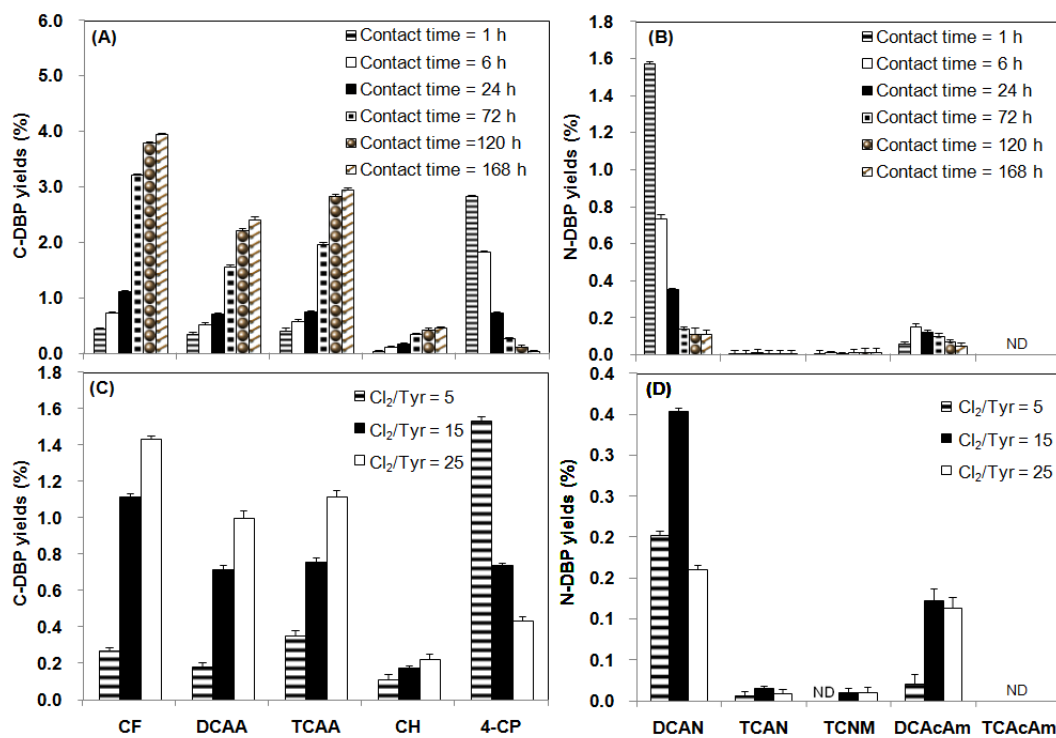
474 Trehy ML, Yost RA, Miles CJ. Chlorination byproducts of amino acids in natural waters,
475 *Environ Sci Technol* 1986; 20: 1117-22.

476 Westerhoff P, Mash H. Dissolved organic nitrogen in drinking water supplies: a review. *J Water*
477 *Supply Res Technol-Aqua* 2002; 51: 415-48.

478 Yang X, Shang C, Westerhoff P. Factors affecting formation of haloacetonitriles, haloketones,
479 chloropicrin and cyanogen halides during chloramination. *Water Res* 2007; 41: 1193-200.

480 Zhang XR, Echigo S, Minear RA, Plewa MJ, 2000. Characterization and comparison of
481 disinfection by-products of four major disinfectants, in: Barrett SE, Krasner SW, Amy GL (Eds),
482 *Natural Organic Matter and Disinfection By-products: Characterization and Control in Drinking*
483 *Water.* American Chemical Society: Washington, DC, pp. 299-314.

484



486

487

Fig.1 – Formation of C- and N-DBPs during chlorination of Tyr at different contact times (A and B)

488

and different Cl₂/Tyr molar ratios (C and D). Tyr concentration = 0.1 mM, pH = 7.0 ± 0.3. Cl₂/Tyr = 15

489

and contact time = 24 h, except as noted. The bars represent the standard deviation of replicate measurements (n = 3).

491

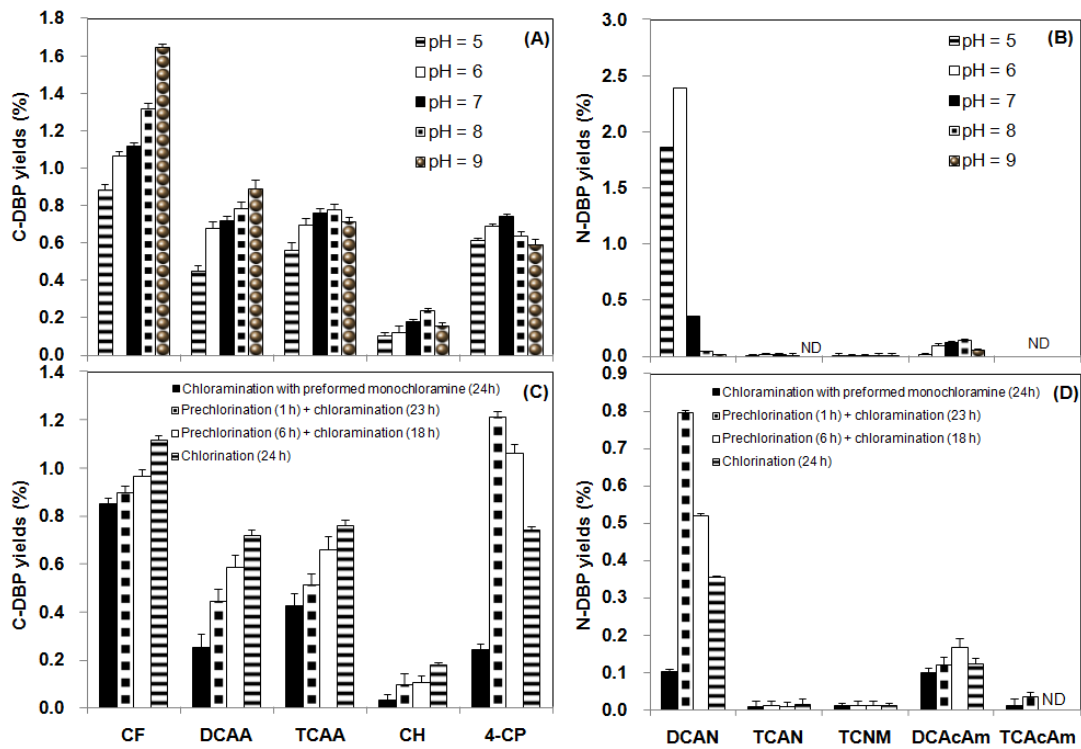
492

493

494

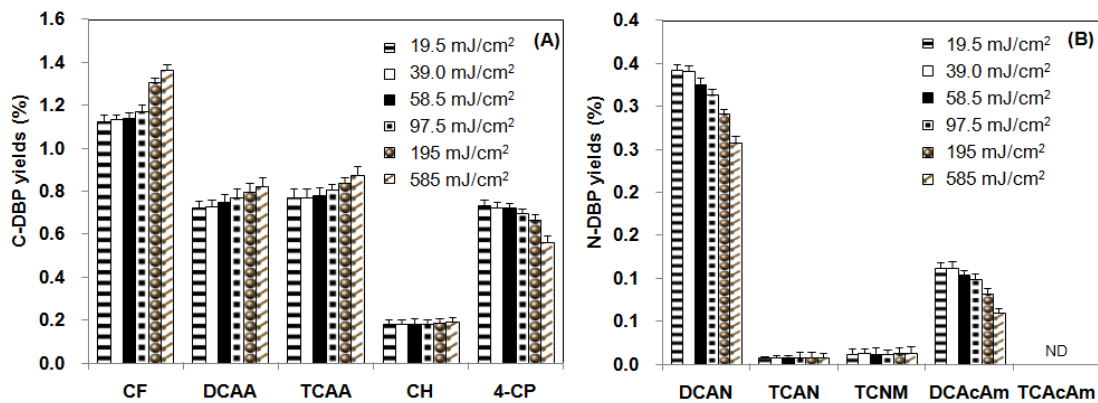
495

496



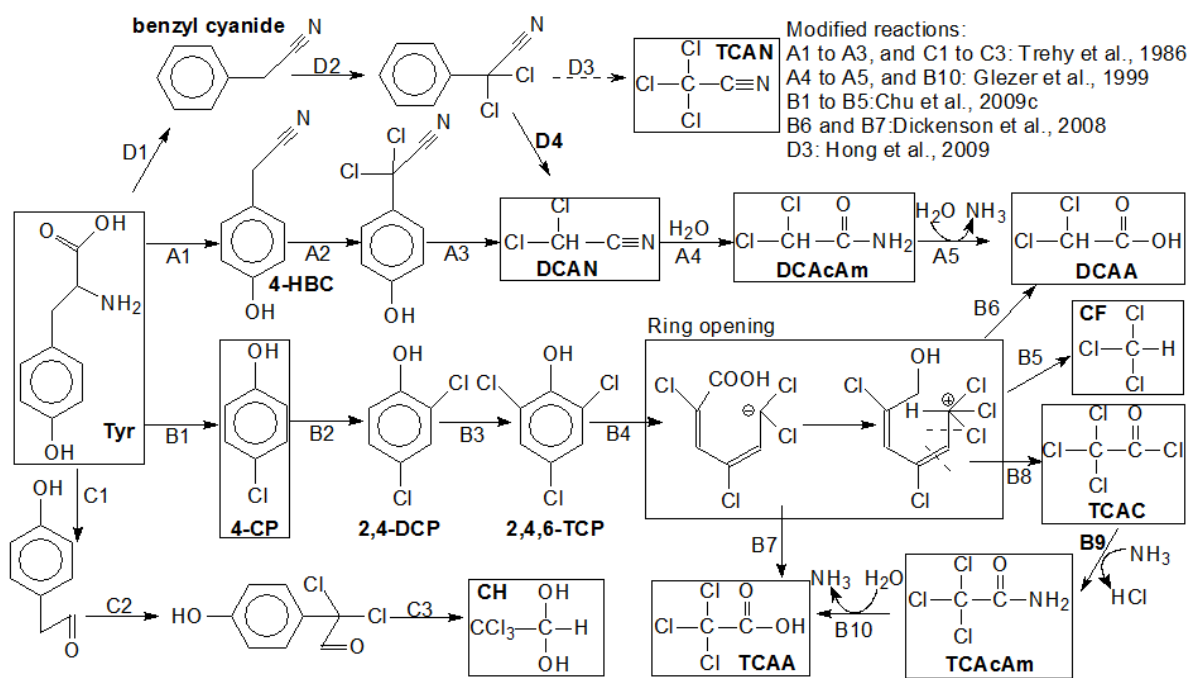
497
498
499
500
501
502
503
504
505
506
507
508
509
510
511
512
513
514
515
516

Fig.2 – Formation of C- and N-DBPs during chlorination of Tyr at different pH levels (A and B) and chloramination schemes (C and D). Tyr concentration = 0.1 mM, total disinfectant contact time = 24 h, pH = 7.0, Cl₂/Tyr = 15, and Cl₂/ammonia = 1, except as noted. The bars represent the standard deviation of replicate measurements (n = 3).



517
518
519
520
521
522
523

Fig.3 – Formation of C- and N-DBPs at different UV dosages (UV irradiation and chlorination were conducted at the same time). Tyr concentration = 0.1 mM, chlorination contact time = 24 h, pH = 7.0, and Cl₂/Tyr = 15. The bars represent the standard deviation of replicate measurements (n = 3).



524
525
526
527

Scheme 1 – Proposed formation pathway of C- and N-DBPs from Tyr (Ala = alanine).