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1	Historical nectar assessment reveals the fall and rise of Britain in bloom
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# 12 Summary

There is considerable concern over declines in insect pollinator communities and potential impacts on the pollination of crops and wildflowers<sup>1-4</sup>. Among the multiple pressures facing pollinators<sup>2-4</sup>, decreasing floral resources due to habitat loss and degradation has been suggested as a key contributing factor<sup>2-8</sup>. However, a lack of quantitative data has hampered testing for historical changes in floral resources. Here we show that overall floral rewards can be estimated at a national scale by combining vegetation surveys and direct nectar measurements. We find evidence for substantial losses in nectar resources in

England and Wales between the 1930s and 1970s; however, total nectar provision in Great 20 21 Britain as a whole had stabilised by 1978, and increased from 1998 to 2007. These findings concur with trends in pollinator diversity, which declined in the mid-20th century<sup>9</sup> but 22 stabilised more recently<sup>10</sup>. The diversity of nectar sources declined from 1978 to 1990 but 23 stabilised thereafter at low levels, with four plant species accounting for over 50% of 24 25 national nectar provision in 2007. Calcareous grassland, broadleaved woodland and neutral 26 grassland are the habitats that produce the greatest amount of nectar per unit area from the most diverse sources, whereas arable land is the poorest in both respects. While agri-27 environment schemes add resources to arable landscapes, their national contribution is low. 28 29 Due to their large area, improved grasslands could add substantially to national nectar provision if they were managed to increase floral resource provision. This national-scale 30 assessment of floral resource provision brings new insights into the links between plant and 31 pollinator declines, and offers considerable opportunities for conservation. 32

#### 33 Main text

Concerns have been raised about declines in both wild and managed insect pollinators<sup>1–4</sup>. While 34 several potential drivers have been cited $^{2-4}$ , one important factor in pollinator declines may be the 35 loss of floral resources due to changes in land-use and management 5-8. Several factors may have 36 caused decreased floral resources in Great Britain and other developed countries, including 37 increased use of herbicides<sup>11</sup>, destruction of traditional landscape features such as hedgerows<sup>12</sup> 38 and loss and degradation of wildflower-rich natural habitats<sup>13–15</sup>. Current strategies to mitigate 39 pollinator declines focus primarily on enhancing floral resources<sup>4</sup>, including agri-environmental 40 scheme options such as sowing nectar flower mixtures<sup>16,17</sup>. There is evidence for declines in some 41

key pollinator forage plants in Great Britain<sup>5</sup> and the Netherlands<sup>7</sup>, but the notion that the overall
availability of floral resources has declined is largely based on subjective assessments. Floral
resources have never been quantified at national or even landscape scales.

45 While both nectar and pollen are important floral resources, we focus on nectar because of its 46 importance as an energy source in the diets of adult bees, and because it provides a common 47 currency (total sugars) in which we can express the nutritional contribution of all plant species<sup>18</sup>. 48 We quantified the nectar resources in Great Britain by combining directly measured and 49 modelled nectar productivity data per unit cover for 260 common plant species (Supplementary Table 1) with historical vegetative cover estimates from the British Countryside Survey<sup>19</sup>, a 50 representative national-scale survey of plant community composition. Together, the 260 species 51 52 comprise the vast majority of British nectar sources as they include virtually all nectar-producing plants from the set of species covering 99% of the British land area. Using vegetation data from 53 the latest Countryside Survey (2007), we quantified recent nectar productivity of habitats (nectar 54 sugar per unit area and time) and the diversity of their nectar sources (considering nectar 55 production both by species and by floral morphology groups, referred to as "species nectar 56 diversity" and "functional nectar diversity" respectively). Production was scaled up to estimate 57 national nectar provision using the estimated area of habitats<sup>19</sup>, allowing the contributions of 58 59 species, habitats and agri-environment schemes to national nectar provision to be assessed. We 60 estimated historical shifts in nectar provision over recent decades using data from earlier Countryside Survey rounds (1978, 1990, 1998 and 2007), considering both changes in nectar 61 productivity within habitats and changes in habitat area. We also investigated floral resource 62 changes from the 1930s onward for England and Wales, based solely on changes in habitat 63 coverage. 64

Considering the most recent Countryside Survey (2007), there are significant differences in 65 66 annual nectar productivity, species nectar diversity and functional nectar diversity among habitats (Extended Data Table 1). Calcareous grassland, broadleaved woodland and neutral grassland are 67 the best in all three respects (as well as shrub heathland for nectar productivity only) whereas 68 69 arable land is consistently the poorest habitat (Supplementary Table 2). These habitat differences 70 in nectar value create geographical variation in nectar productivity and diversity across Great 71 Britain (Figure 1). After taking into account the national land cover of habitats, improved grassland contributed most (29%) to potential national nectar supply in 2007. Four species of 72 plant, Trifolium repens, Calluna vulgaris, Cirsium palustre and Erica cinerea together produce 73 74 over 50% of nectar nationally (see Extended Data Table 2 and Supplementary Result 1 for further 75 information about these species and their pollinators), and 22 species produce over 90% (Figure 76 2). Other species may of course be important for pollen provision. Considering flowering phenology reveals seasonal variation nationally (Figure 3): 60% of nectar is provided in 77 July/August when the flower density of British dominant species peaks. Because heathland 78 species are unlikely to contribute as much in other European countries, this seasonal pattern may 79 differ. The relative nectar value of linear features (hedgerows, watersides and road verges) 80 depends on habitat. With the exception of those in shrub heathland and bog, linear features 81 82 produce more nectar per unit area (and the contrast is particularly high in landscapes dominated by arable land, improved grassland and conifer woodland; Extended Data Figure 1). Of the five 83 types of agri-environment scheme options we investigated, nectar flower mixtures have the 84 85 highest nectar productivity value, followed by enhanced margins (Extended Data Table 3). Nectar flower mixture options are similar to hedgerows in term of annual nectar productivity per 86 87 unit area, but they cover a much smaller area, and consequently contribute far less to the national

nectar resources (0.1% of nectar supply comes from nectar flower mixtures compared to 3% from
hedgerows in England, Extended Data Table 3).

Historical shifts in nectar productivity, species nectar diversity and functional nectar diversity 90 91 over recent decades depended on the habitat type and time period considered (Extended Data 92 Table 1). From 1978 to 1990, annual nectar productivity decreased significantly in arable land 93 and conifer woodland, but from 1990 to 1998, none of the habitats showed significant changes in 94 nectar productivity. From 1998 to 2007, nectar productivity increased significantly in arable land 95 and neutral grassland (Extended Data Figure 2). Nectar diversity, both at the level of plant species and functional groups decreased significantly in arable land and improved grassland from 96 1978 to 2007. Species nectar diversity also significantly decreased in conifer woodland and 97 98 broadleaved woodland during that period. From 1978 to 1990, species nectar diversity declined in 99 all habitats (except bog), significantly so in arable land and conifer woodland; thereafter it remained roughly constant, except in arable land where it rebounded somewhat from 1998 to 100 2007 (see Extended Data Figure 2 and Supplementary Results 2 for details on functional nectar 101 102 diversity). For the 1930s we have information only on shifts in land cover (but not floral abundances within them), and only for England and Wales<sup>20</sup>. Assuming no change in floral 103 104 composition within habitats, we found a strong decline in national nectar provision from 1930s to 1978 (-32%) followed by a period of stagnation from 1978 to 2007 (Figure 4, Supplementary 105 106 Table 3). Incorporating shifts in nectar productivity within habitats for recent decades showed an increase in national nectar provision from 1998 to 2007 (+51% in England & Wales and +25% 107 for Great Britain as a whole, Figure 4, Supplementary Table 4). While shifts in vegetation 108 109 composition within dominant habitats predominate as causes of recent increases, no quantitative data are available before 1978. This recent upturn could be caused by decreased acidification<sup>21</sup>, 110

decreased nitrogen deposition<sup>22</sup> and agricultural set-asides<sup>23</sup> during this period (Supplementary 111 112 Table 5). However, post-war changes in habitat management (e.g. herbicide use in arable land, cessation of woodland coppicing, nitrogen deposition in grasslands; Supplementary Table 5) 113 almost certainly resulted in lower nectar per unit area, suggesting that our estimates of losses 114 115 based on land use change alone are conservative; actual resource declines may have been much larger than the recent increases (see Supplementary Discussion). Due to their large area, 116 117 improved grassland provided the greatest contribution to the increase in national nectar provision from 1998 to 2007 (Extended Data Figure 3). After discounting the contribution of Trifolium 118 repens in improved grasslands, as it may not flower in heavily grazed fields, the increase in 119 nectar provision from 1998 to 2007 remained (Supplementary Result 3 and Extended Data Figure 120 4). 121

The historical pattern of change in nectar resources closely parallels documented shifts in pollinator communities (Extended Data Figure 5). Substantial declines in floral resources and their diversity in the mid to late 20th century, when agricultural intensification peaked, coincide with a period of heightened pollinator extinctions<sup>9</sup>. The stabilization and partial recovery of resources in recent decades corresponds to concomitant periods of decelerated declines and partial recovery in some pollinator groups<sup>10</sup>.

Our findings provide new evidence based on floral resources to support habitat conservation and restoration. First, we provide evidence of the high nectar value of calcareous grassland for pollinating insects. Calcareous grassland area has declined drastically in Great Britain and only a small fraction of the historical national cover remained by 2007<sup>13,14</sup>. Second, the low availability and diversity of nectar sources in arable habitats highlights the need to provide supplementary resources to support pollination services in farmlands, especially as the use of insect-pollinated

crops has increased nationally<sup>24</sup> and globally<sup>25</sup>. The conservation and restoration of broadleaf 134 135 woodland and neutral grassland as components of the farmland matrix could help to support diverse flower-visiting insect communities in arable land. The contrast in nectar productivity 136 between linear features and the surrounding vegetation is particularly high in arable land, 137 138 suggesting that linear features, especially hedgerows, provide an efficient means to enhance floral resources in farmlands if they are managed appropriately to allow flowering<sup>26</sup>. While agri-139 140 environment options such as nectar flower mixtures can also enhance the supply of floral resources locally, their contribution to nectar provision nationally remains low. The higher profile 141 given to floral resource provision in the revised Countryside Stewardship guidelines for 142 England<sup>16</sup> may substantially enhance resources in future. Finally, our results indicate that 143 improved grassland has the potential to contribute massively to the nectar available nationally. 144 Small adjustments to the management cycle in improved grasslands, allowing white clover, the 145 146 dominant resource species, to flower, would help realize this potential, although its utility might be restricted to a limited number of pollinator species (Extended Data Table 2). Together, our 147 results on the nectar values of the commonest British plants and the historical changes in plant 148 149 communities provide the evidence base needed to understand recent national changes in nectar provision and identify the management options needed to restore national nectar supplies. 150

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### 207 End notes

208 **Supplementary Information** is available in the online version of the paper.

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- 217 (www.countrysidesurvey.org.uk).

#### 218 Author Contributions

- The study was conceived by W.E.K. and J.M. The field survey was carried out by M.B. and N.D.
- with the help of J.M. The data were compiled and analysed by M.B. with suggestions from
- 221 W.E.K., J.M., S.S., R.D.M. and M.G. Vegetation data from the Countryside Survey database
- were extracted by S.S. Agri-environment scheme data were provided and analysed by N.D.B. and
- 223 S.C. The national maps were generated by R.D.M. All authors discussed the results and
- 224 contributed during manuscript writing.

#### 225 Author Information

- 226 The floral resource database will be made available from the NERC Environmental Information
- 227 Data Centre (doi:10.5285/69402002-1676-4de9-a04e-d17e827db93c and doi:10.5285/6c6d3844-
- e95a-4f84-a12e-65be4731e934). Reprints and permissions information is available at
- 229 www.nature.com/reprints. The authors declare no competing financial interests. Correspondence
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# 231 Figure legends

- Figure 1. Nectar productivity and diversity in Great Britain in 2007. a, Box plots of log10
- 233 (x+1) nectar productivity (kg of sugars/ha/year) per habitat. **b**, Box plots of species nectar
- 234 diversity (Shannon index of nectar species) per habitat. c, Box plots of functional nectar diversity
- 235 (Shannon index of nectar flower types) per habitat. Box plots are based on 2007 vegetation data

(see Supplementary Table 2 for sample sizes). Habitat types (AR=Arable land, IG=Improved
grassland, AG=Acid grassland, NG=Neutral grassland, CG=Calcareous grassland, CON=Conifer
woodland, BRO=Broadleaf woodland, BOG=Bog, FEN=Fen, BRA=Bracken, SH=Shrub
heathland) significantly different from one another are indicated by different letters. d, Map of
nectar productivity. e, Map of species nectar diversity. f, Map of functional nectar diversity.
Maps are based on 2007 land cover and vegetation data.

#### 242 Figure 2. Plant species' contributions to Great Britain nectar provision and to habitat

nectar provision, based on 2007 land cover and vegetation data. The dotted line represents the
cumulative contribution of plant species to the national nectar provision in 2007 (only species
that contribute to the first 95% are shown). The pie charts represent the contribution of plant
species towards nectar production in each habitat (only the species that contribute to the first 90%
are shown) in 2007. The size of each pie chart is proportional to the contribution of each habitat
to national nectar provision in 2007.

#### 249 Figure 3. Seasonal nectar productivity in Great Britain, based on 2007 land cover and

vegetation data. Maps of nectar productivity in kg of sugars/ha from March to October (panels a
to h). Hot colours correspond to high nectar productivity while cold colours correspond to low
nectar productivity (see colours scale). Note that urban areas are assigned with nectar
productivity values equal to zero, hence the blue colours in cities. Nectar productivity values for
mapping correspond to back-transformed estimates of the linear mixed model fitted on log10
(x+1) nectar productivity of 2007 Countryside Survey non-linear plots with habitat, month and
their interaction as fixed effects and plots nested within squares as random effects.

Figure 4. Historical changes in nectar provision (in kg of sugars/year) at the national scale 257 258 in England & Wales (1930-2007) and in Great Britain (1978-2007): Nectar provision partitioned by habitat, based on land cover for 1930 (England & Wales only), 1978, 1990, 1998 259 and 2007, using vegetation data from 1978 for all years (assuming unchanged nectar productivity 260 261 within habitats across time) in **a**, England & Wales and **b**, Great Britain. Nectar provision 262 partitioned by habitat, based on land cover and vegetation data for 1978, 1990, 1998 and 2007 in 263 c, England & Wales and d, Great Britain. See Figure 1 for habitat type codes and Supplementary Table 6 for habitat land cover values. 264

265 Methods

# Stage 1: Constructing the nectar database by scaling up nectar resources from the flower to the vegetative scale

#### 268 Identifying the key plant species to be sampled

While there are >2800 plant species in Great Britain<sup>27</sup>, only 1341 of them are common enough to 269 270 have been encountered in the Countryside Survey. Of these, the 454 commonest species accounted for 99% of national plant cover in 2007. More than half of these 454 species are 271 unrewarding to pollinators (mainly bryophytes, pteridophytes, gymnosperms and wind-pollinated 272 angiosperms<sup>28</sup>), leaving 220 species that are likely to contribute substantially to floral resources 273 274 at a national scale. We focus here on these 220 species, along with an additional 50 species that 275 we believe to be locally important floral sources (e.g. Buddleja davidii, Impatiens glandulifera, Knautia arvensis). Together, these 270 plant species provide a focal set of potential importance in 276 277 national nectar provision (Supplementary Table 1).

279 Of the 270 species, 175 were surveyed in the field from February 2011 to October 2012, mainly 280 in the South of England. When possible (112 species), nectar was collected from plants in at least 281 two populations in two locations. For three species (Caltha palustris, Lamium purpureum, and 282 Sinapis arvensis), half the nectar samples, and for Viola arvensis all the samples were collected 283 from pot-grown plants, because insufficient flowering field populations were found. For the 284 remaining species, nectar was collected from plants in one field population. When possible, the 285 different populations were sampled on different dates, thus providing some measure of variation due to differences in location and weather. Note that nectar was collected in only 1-2 sites per 286 species, and so intraspecific variation in production per flower was not assessed (but see 287 288 Supplementary Result 4).

Nectar was collected from ten single flowers in each population between 0900-1600 hours 289 (median: 20 and range: 5-30 flowers collected per species in total; see Extended Data Figure 6 290 and Supplementary Result 4 for site correlation); these had been bagged (using 1.4 x 1.7mm 291 292 fabric mesh) for 24h to prevent depletion by nectar-feeding insects. When possible (76 species), 293 glass microcapillaries (1 and 5µL Minicaps, Hirshmann, Eberstadt, Germany) were used directly 294 to collect the nectar, otherwise single flowers were rinsed twice with 1-5  $\mu$ L of distilled water added to the nectaries with a pipette for one minute, and the diluted nectar solution was collected. 295 296 The sugar concentration of nectar (%; g sucrose/100 g solution) was measured by using a hand held refractometer modified for small volumes (Eclipse, Bellingham and Stanley, Tunbridge 297 Wells, UK). The amount of sugar produced per flower basis over 24h (s; µg of 298 sugars/flower/24h) was calculated using the formula<sup>29</sup> 299

s = 10 dvC

where v is the volume collected ( $\mu$ L), and d is the density of a sucrose solution at a concentration C (g sucrose/100 g solution) as read on the refractometer. The density of the sucrose solution was calculated by the formula<sup>29</sup>

304 d=0.0037921C+0.0000178C<sup>2</sup>+0.9988603

305 The number of open flowers per unit area of vegetative cover (flower density) was estimated for 179 species by placing five quadrats  $(0.5m \times 0.5m)$  haphazardly on each flowering population 306 (median: 10 quadrats, range: 1-20 quadrats; see Extended Data Figure 6 and Supplementary 307 308 Result 4 for site correlation). In each quadrat, we counted the number of open floral units of the focal species (a "floral unit" is one or multiple flowers that can be visited by insects without 309 310 flying<sup>30</sup>; for example a composite flowerhead of daisy, *Bellis perennis*). We also counted the number of open flowers present in one typical open floral unit in each quadrat. Vegetative cover 311 for each plant species was estimated using a point-quadrat approach with the cross-strings of the 312 313 quadrat: cover was expressed as proportional to the number of the 36 cross-points covered by the 314 foliage of the species of interest in each quadrat. For trees, instead of using quadrats, we counted 315 the number of floral units in a 3D cube  $(0.5 \times 0.5 \times 0.5m)$  that was placed in the outer areas of foliage. This was extrapolated to the whole column situated above the unit of vegetative cover by 316 317 measuring the height of tree foliage with an inclinometer (PM-5/360 PC Suunto) and by estimating the distribution of the flowers within the tree foliage (subjectively assessed scores: 318 319 from 1 for a strongly biased flower distribution on the outer edges of the foliage to 5 for a 320 homogeneous full flower distribution). Given that flower density is not constant throughout the flowering season, we estimated variations in flower density according to a triangular function 321

from the estimated peak of flowering through the flowering season which was documented from recorded phenologies<sup>28,31,32</sup> (see Supplementary Method 1 and Extended Data Figure 6 for phenology parameter relationships). An alternative nectar rectangular phenology productivity database was also generated by keeping nectar productivity of each species constant throughout the flowering season; this was used to perform sensitivity analyses.

327 The mean nectar sugar content from a single flower (produced over a 24h period) was multiplied 328 up to the nectar content of a single floral unit (number of flowers in a floral unit), then to the amount of nectar per unit area (number of flowers per m<sup>2</sup>), to the amount of nectar per unit area 329 330 for each month (variation in flower density over the flowering season) and finally to the amount of nectar per unit area per year. Despite relatively low sample sizes per species compared to 331 332 species-specific studies, our estimates of sugar production were well correlated with published 333 values both per flower/day and per area/year (Extended Data Figure 6 and Supplementary Result 4). This empirical method provided the nectar productivity values for 161 plant species amongst 334 the 175 initially surveyed (nectar productivity could not be scaled up for some species due to 335 mismatches with phenological data, see Supplementary Method 1). 336

#### 337 Modelling nectar productivity: the 'unsurveyed species'

To model the nectar productivity of the plant species that could not be surveyed in the field, we used a predictive modelling approach. We first analysed variation in the nectar values from the surveyed species. A linear model was fitted to annual nectar sugar productivity (log10 (x+1) transformed) as a function of plant traits. Plants traits were mainly collected from the BiolFlor database<sup>33</sup>, and included: "flower shape", "breeding system", "life span", the degree of "dicliny", the maximum "height", the "flowering period" and "family" (see Supplementary Method 2 for

definitions). The estimates from the most parsimonious statistical model based on AIC criterion 344 345 (Supplementary Table 7, N=153; Adjusted  $r^2=0.55$ ) were used to predict the annual nectar sugar productivity for the initial list of surveyed and unsurveyed species on the basis of their traits. To 346 check the validity of the predicted values, we adopted a repeated "leave-one-out" approach to 347 348 model successively all the excluded values from the empirically derived datasets. Then, we applied a standardized major axis regression on the  $\log 10$  (x+1) transformed empirically derived 349 350 and modelled nectar values of the surveyed species (Extended Data Figure 6). We predicted the nectar values for 252 species; and giving priority to empirical and default values, we included 94 351 352 of them in our database. An alternative nectar productivity database was also generated by 353 considering only the species with empirical nectar values; this was used to perform sensitivity 354 testing.

#### 355 Ascribing default values for nectar productivity

356 For four crop species harvested before flowering; onion (Allium cepa), cabbage (Brassica oleracea cultivated), turnip (Brassica rapa) and radish (Raphanus sativus) we assigned a value of 357 zero for nectar productivity. A zero-value was also assigned to *Helianthemum nummularium*, 358 359 despite the missing flower density data, given that we collected no nectar in flowers. In the 360 Countryside Survey vegetation dataset, some taxa are only identified at the genus level; we 361 interpreted these taxa to represent the commonest species in the genus (e.g. *Centaurea* sp. was 362 interpreted as Centaurea nigra). For 10 species out of the initial list of 270 it was not possible to quantify nectar production, leading to a total of 260 species with quantified annual and monthly 363 364 nectar productivity values (161 values from empirical research, 94 modelled values, and 5 default 365 values, Supplementary Table 1). All the above steps of scaling-up process are summarized in 366 Supplementary Table 8.

# 367 Stage 2: Using the Countryside Survey vegetation database to scale up nectar resources from 368 plant species to communities at the habitat and national scales

Spatio-temporal variations in nectar provision at the national scale were calculated by combining 369 370 our nectar productivity dataset with vegetation and land cover data already recorded during the Countryside Survey<sup>19</sup>. The Countryside Survey is a national survey of plant communities 371 372 conducted in 1978, 1990, 1998 and 2007 in Great Britain (England, Wales and Scotland). The survey was conducted by selecting 1-km sample squares at random from 32 Land Classes<sup>19</sup> 373 374 representing physiographically similar sampling domains throughout Great Britain, ensuring an 375 unbiased representation of the British non-urban landscape. Within each square, a random, 376 stratified sample of five areal (non-linear) square plots (200 m<sup>2</sup>) was established and the presence 377 and the percentage cover of all vascular plant species were recorded. These plots were classified 378 to 17 habitat classes, but we only used data from 11 habitats: acid grassland, arable land, bog, bracken, broadleaf woodland, calcareous grassland, conifer, fen, improved grassland, neutral 379 grassland and shrub heath (Supplementary Table 9 for habitat description). The habitats not used 380 were inland rock, littoral rock/supralittoral rock, littoral sediment/supralittoral sediment, montane 381 382 and urban habitats; these were excluded due to low sample sizes. Even though urban habitats 383 probably contribute to the national nectar provision, we were unable to include this habitat in this study because the Countryside Survey was not designed to survey urban areas. In 1.14% of 384 385 Countryside Survey plots, two or more habitats were attributed to the same plot; these were excluded for this study. Additional plots were used to sample linear features in each 1km square, 386 covering hedgerows, streamsides and road verges (1x10m and oriented along the linear feature). 387 388 Each linear plot was also attributed to its nearest adjacent habitat.

389 To investigate the most recent nectar patterns, we used the most comprehensive vegetation 390 dataset from the Countryside Survey 2007 that encompasses all non-linear plots (2576 plots in 2007). To focus on linear features, we included vegetation data from linear features plots (1951 391 plots in 2007). To test for historical changes from 1978 to 2007, we used vegetation data from 392 393 non-linear plots shared between the 1978, 1990, 1998 and 2007 Countryside Surveys (529 shared 394 plots in England & Wales and 768 in Great Britain; Supplementary Table 10). We focussed on 395 the shared plots across years because the Countryside Survey sampling design was modified over time (e.g., from fixed to proportional plot number per Land Class from 1978 to 1990). 396

397 The annual nectar productivity within each plot (kg/ha/year) is the sum of the nectar productivity of each species (kg/ha cover/year) weighted by their vegetative cover in the plot (%), assuming 398 399 that the vegetative cover is representative of floral abundance (see Extended Data Figure 7 and 400 Supplementary Results 4 for details). Nectar productivity values of plots were used to statistically estimate the annual nectar productivity for each habitat (kg/ha/year). The annual nectar provision 401 of each habitat (kg/year) was computed from their annual habitat nectar productivity (kg/ha/year) 402 403 multiplied by their respective national land covers for each survey (areas of habitats in ha from Countryside Surveys<sup>19,34,35</sup>; Supplementary Table 6). These were summed to estimate the annual 404 405 national nectar provision in 1978, 1990, 1998 and 2007. For the 1930s period, areas of habitats (only available for England and Wales) were derived from the digitalised Dudley Stamp land 406 utilisation survey maps<sup>20</sup>; see Supplementary Method 3 and Supplementary Table 6). Because 407 nectar productivity can't be assessed for this period, we quantified nectar provision in 1930, 408 1978, 1990, 1998 and 2007 assuming unchanged nectar productivity within habitats but using 409 410 observed shifts in land cover among habitats across time. The national nectar provision of hedgerows was calculated from their mean nectar productivity (kg/ha/year) multiplied by their 411

estimated area in England (length of hedgerows from Countryside Survey 2007 for England<sup>35</sup>,
assuming a 1m width).

414 The contribution of habitat or species to the national nectar provision in 2007 is the fraction of 415 nectar provided by these entities (in %). The amount of nectar offered by each habitat in 2007 is 416 calculated from habitat nectar productivity (estimated value of habitat productivity) multiplied by 417 its national area. The amount of nectar offered by each species in 2007 is calculated from the sum 418 of its average nectar productivity stratified by habitat and multiplied by habitat national area. The 419 contribution of habitat or species to the historical changes in national nectar provision is 420 expressed by the absolute change (in kg of sugars), which is the difference in the amount of nectar produced by the entity during the time period considered. Relative change (in %) which is 421 422 the absolute change multiplied by 100 and divided by the amount of nectar produced at the initial 423 date, refers to the magnitude of change for each entity.

Nectar diversity was estimated through two Shannon indexes (using 'vegan' package in R<sup>36</sup>) that
encompass both the richness and the evenness of nectar producing sources (see Supplementary
Method 4). The species nectar diversity index, based on the proportion of nectar produced by
each species, was calculated as follows:

428  $H_{sp}' = -\sum_{i=1}^{S} p_i \times \ln(p_i)$ 

429 where  $p_i$  is the proportional nectar contribution of plant species *i* and *S* is the total number of 430 plant species in each plot.

The functional nectar diversity index, based on the proportion of nectar produced by each floralmorphology group, reflects the diversity of nectar sources in terms of resource accessibility for

flower-visiting insects. Flower types were derived from Müller flower classification system
recorded from the BiolFlor database<sup>33</sup> which was condensed into five classes: pollen rewarding
flowers, open, partly-hidden, hidden, and bee flowers (see Supplementary Method 4). The
functional nectar diversity index was computed as follows:

437 
$$H_{fun}' = -\sum_{i=1}^{S} p_i \times \ln(p_i)$$

438 where  $p_i$  is the proportional nectar contribution of flower type *i* and *S* is the total number of 439 flower types in each plot.

The annual nectar productivity (kg of sugars/ha/year), species nectar diversity (Shannon index of
nectar contribution of plant species) and functional nectar diversity (Shannon index of nectar
contribution of floral morphology groups) in 2007 were mapped at the British national scale
using the Great Britain Land Cover Maps of 2007<sup>37</sup>.

# 444 Stage 3: Using Agri-environment scheme flower abundance data to estimate nectar provision 445 within agri-environment scheme options at the national scale

Various options are available for managing habitats to provide floral resources for pollinators, 446 some of which are eligible for grant aid under European Union funded agri-environment 447 448 schemes. Agri-environment options within the English 'Environmental Stewardship' scheme included sowing nectar flower mixtures (EF4/HF4), sowing wild bird seed mixtures (EF2/HF2), 449 creation or enhancement of floristically-enhanced buffer strips (HE10), re-introduction or 450 451 continuation of haymaking (haymaking supplement HK18) and creation, restoration and maintenance of species-rich semi-natural grassland (HK6/7/8). These five options were selected 452 as the most likely to provide floral resources for pollinators. 453

Field study sites were located on farmland and nature reserves in which the following replicates 454 455 of the pollinator habitats were present: nectar flower mixtures (n=32), wild bird seed mixtures (n=4), enhanced field margins/road verges (n=7), hay meadows (n=5) and species-rich grasslands 456 (n=7). These were existing habitats representing ongoing management by the land owners or land 457 458 managers concerned. Transects 100m long x 6m wide were established in each habitat. The number of floral units of each flowering species was recorded on 1 to 3 occasions, in 20 x 1m<sup>2</sup> 459 460 quadrats per transect. Annual nectar productivity (kg of sugars/ha/year) was calculated for each species at each site from the average estimated nectar productivity at the peak of the flowering 461 season derived from the several counts of floral units across the flowering period (analogous to 462 463 Supplementary Method 1). The values for the species present in each habitat were then summed 464 to estimate productivity for each habitat.

465 National areas of options providing floral resources in the English agri-environment scheme "Environmental Stewardship" were extracted for 2007 for England (data for Great Britain was 466 unavailable) from data supplied by Natural England<sup>38,39</sup>. Mean nectar productivity per unit area 467 was multiplied by the national area of each option to give nectar provision by that option (kg of 468 469 sugars/year). The total contribution of nectar provision provided by Environmental Stewardship 470 in England is a minimum value, as it has been compared to national provision estimated from vegetative cover rather than direct flower counts and we did not take into account the more 471 472 limited floral resources potentially provided by other options.

#### 473 Stage 4: Statistical analyses

474 Statistical analyses were carried out with Linear Mixed-Effect Models (lme function from 'nlme'
475 package) in R 3.0.1<sup>(36)</sup>. To investigate the most recent nectar variations (2007), we analysed the

log10(x+1) annual nectar productivity, species nectar diversity and functional nectar diversity 476 477 according to the type of habitat ("HABITAT"; 11 habitats) of the non-linear plots. The differences in log10(x+1) nectar productivity, species nectar diversity and functional nectar 478 diversity between non-linear and linear features were analysed according to the type of habitat 479 480 ("HABITAT"; 11 habitats), the type of vegetation surveyed ("TYPE"; non-linear vs linear 481 features) and the interaction between these two terms. Countryside Survey square ("SQUARE") 482 was included as a random term in these models in order to account for the spatial auto-correlation of plots nested into 1km squares. In order to investigate historical changes over recent decades 483 484 (1978-2007), we analysed the log10(x+1) annual nectar productivity, species nectar diversity and 485 functional nectar diversity computed from the shared non-linear plots in 1978, 1990, 1998 and 486 2007 according to the type of habitat ("HABITAT"), the year ("YEAR") considered as a 487 categorical factor, and the interaction between these two terms. We included plots nested within square ("SQUARE/PLOTS") as random terms to account for the spatial and temporal auto-488 correlation of the data in this latter model. This latter statistical test was repeated considering all 489 shared plots in Great Britain or only those in England & Wales to provide estimates of habitat 490 491 nectar productivity across time for distinct areas, allowing comparisons with earlier (1930s) habitat information only available for that latter area. Significant differences among modalities 492 493 were analysed with multiple comparisons (single-step method adjusted p-values from glht function in "multcomp" package in R<sup>36</sup>). Model residuals were plotted to visually check that 494 normality and homoscedasticity assumptions were satisfied. We re-ran the same analyses with the 495 496 Countryside Survey vegetation data combined with (i) the alternative nectar rectangular phenology productivity database (created by keeping constant nectar productivity of each species 497 during the flowering season); and (ii) using only the empirical nectar productivity database, as 498 499 sensitivity tests (Extended Data Figure 4, Supplementary Result 3). Plots were performed with

500 ggplot2 package in  $R^{36}$ . All box plots show the median,  $25^{th}$  and  $75^{th}$  percentiles (lower and upper

501 hinges), trimmed ranges that extend from the hinges to the lowest and highest values within 1.5 x

502 inter-quartile range of the hinge (lower and upper whiskers) plus outliers (filled circles). Notches

- that extend 1.58 x inter-quartile range / square root of the number of observations were
- represented to give a roughly 95 interval for comparing medians.

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## 538 Extended Data Legends

539 Extended Data Table 1. ANOVA results for annual nectar productivity, species nectar

540 diversity and functional nectar diversity. a, 2007 values according to habitat. The linear mixed

effect models were performed on data from 2576 non-linear plots surveyed in 2007. b, 2007

values according to habitat and location. The linear mixed effect models were performed on data

543 from 4527 plots (2576 non-linear plots and 1951 linear plots) surveyed in 2007. c, 1978-2007

values according to habitat and year. The linear mixed effect models were performed on data

from 768 shared plots surveyed in 1978, 1990, 1998 and 2007. The annual nectar productivity

546 was systematically log10 (x+1) transformed. See Supplementary Table 2 and Supplementary

547 Table 4 for sample sizes.

#### 548 Extended Data Table 2. Flower morphology and flower-visiting insects of the four main

549 nectar providing species. Flower morphology parameters (mean and standard error for depth

and width of flower tubes) were measured on 20-40 flowers per species in the field. Flower-

visiting insects were listed from published and unpublished plant-insect visiting networks from

Memmott's group to which recorded interactions from a review of literature have been added(see Supplementary Table 12 for reference list).

#### 554 Extended Data Table 3. Agri-environment schemes and linear features: nectar productivity 555 and provision in England in 2007. a, Mean nectar productivity values of agri-environment 556 schemes were estimated from our nectar productivity database combined with flower counts in 557 these options. Areas of options providing floral resources in the English agri-environment scheme 558 "Environmental Stewardship" were extracted for 2007 from data supplied by Natural England<sup>38,39</sup>. **b**, Mean nectar productivity values of linear features correspond to back-559 560 transformed $(10^{x} - 1)$ estimates of the linear mixed model fitted on log10 (x+1) nectar 561 productivity of all Countryside Survey linear plots surveyed in England in 2007. National areas 562 of hedgerows were estimated from the length given in Countryside Survey 2007 for England<sup>35</sup> and assuming a 1m width. 563

#### 564 Extended Data Figure 1. Annual nectar productivity and diversity in linear features in

**2007.** a, Box plots of log10 (x+1) nectar productivity according to the location of the vegetation 565 surveyed (non-linear vs linear features) in each habitat. **b**, Box plots of species nectar diversity 566 567 according to the location of the vegetation surveyed (non-linear vs linear features) in each habitat. c, Box plots of functional nectar diversity according to the location of the vegetation surveyed 568 569 (non-linear vs linear features) in each habitat. Significant differences of locations (linear vs nonlinear) in habitats are indicated by asterisks as follows: \* for  $p \le 0.05$ ; \*\* for  $p \le 0.01$ ; \*\*\* for  $p \le 0.01$ ; 570 571 0.001. Statistical model were re-run without calcareous grassland habitat (to meet residuals 572 homoscedasticity constraint) in order to check that significant effects remained. See Extended 573 Data Table 1 for ANOVA results.

Extended Data Figure 2. Historical changes in nectar productivity and diversity per habitat over recent decades (1978 to 2007). a, Box plots of log 10 (x+1) nectar productivity per habitat, based on vegetation data for 1978, 1990, 1998 and 2007. b, Box plots of species nectar diversity per habitat, based on vegetation data for 1978, 1990, 1998 and 2007. c, Box plots of functional nectar diversity per habitat, based on vegetation data for 1978, 1990, 1998 and 2007. Significant differences of time periods per habitats are indicated by stars (\* for  $p \le 0.05$ ; \*\* for  $p \le 0.01$ ; \*\*\* for  $p \le 0.001$ ). See Extended Data Table 1 for ANOVA results.

581 Extended Data Figure 3. Habitat contributions to the national nectar provision shifts and species contributions to habitats over recent decades (1978 to 2007). Habitat contributions to 582 the national nectar provision changes from **a**, 1978 to 1990 **b**, 1990 to 1998 and **c**, 1998 to 2007. 583 584 All barplots represent the absolute changes (in 000 000 kg of sugars) for each habitat during the 585 time period considered. Numbers in brackets indicate the relative changes (in %). Species contributions to nectar provision in 1978, 1990, 1998 and 2007 per habitat type (panels **d-n**). 586 Only species that contribute to the first 90% are shown. See Supplementary Table 11 for main 587 contributing species to the national changes from 1978 to 2007. 588

589 Extended Data Figure 4. Sensitivity analyses of historical trends from 1978 to 2007 in nectar productivity and species diversity with alternative datasets. a, Box plots of log 10 590 591 (x+1) nectar productivity and **b**, Box plots of species nectar diversity per habitat based on 592 vegetation data for 1978, 1990, 1998 and 2007 discounting the contribution of grazed white clover in improved grassland. c, Box plots of  $\log 10 (x+1)$  nectar productivity and d, Box plots of 593 594 species nectar diversity per habitat, based on vegetation data for 1978, 1990, 1998 and 2007 and 595 computed with the alternative rectangular phenology function. e, Box plots of log 10 (x+1) nectar productivity and **f**, Box plots of species nectar diversity per habitat, based on vegetation data for 596

597 1978, 1990, 1998 and 2007 and computed considering only the species with empirical nectar 598 values. Significant differences of time periods per habitats are indicated by stars (\* for  $p \le 0.05$ ; 599 \*\* for  $p \le 0.01$ ; \*\*\* for  $p \le 0.001$ ). See Supplementary Table 4 for sample sizes and

600 Supplementary Result 3 for details.

#### 601 Extended Data Figure 5. Historical timeline in changes in nectar resources and flower-

visiting insects in Great Britain. Historical periods with the greatest negative changes in nectar
resources and flower-visiting insects are indicated in red, those with intermediate changes are in
orange and those with the lowest (or even reversing) changes are in green. Main historical trends
from this study (Baude et al.) are presented in regard to those described in Carvalheiro et al.
2014<sup>10</sup> and Ollerton et al. 2014<sup>9</sup> studies. The white chevron indicates a provisional extinction rate
that needs to be confirmed on a 20 year period of time (see supplementary materials from
Ollerton et al. 2014<sup>9</sup>).

Extended Data Figure 6. Validity of the datasets. a, Major axis linear regression of log10 609 (x+1) nectar values per flower obtained in the second location against those obtained in the first 610 one. **b**, Major axis linear regression of log10 (x+1) flower density values obtained in the second 611 612 location against those obtained in the first one. c, Major axis linear regression of log10 (x+1)peak flower density values obtained in the second location against those obtained in the first one. 613 614 **d**, Standardized major axis regression of the log(x+1) length of the flowering period used for 615 analyses with those derived from IPI AgriLand floral transects. e, Standardized major axis 616 regression of peak date of flowering season used for analyses with those derived from IPI 617 AgriLand floral transects. **f**, Major axis linear regression performed on the log10 (x+1) empirical (empirical dataset) and published nectar values (literature dataset from Raine & Chittka 2007<sup>40</sup>) 618 at the flower scale. **g**, Standardized major axis linear regression performed on the  $\log 10 (x+1)$ 619

empirical (empirical dataset) and published nectar values (literature dataset, see Supplementary Table 13 for references) at the vegetative scale. **h**, Standardized major axis linear regression performed on the log10 (x+1) empirical and modelled nectar values generated by a leave-one-out approach. Estimates of all equations are derived from (standardized) major axis regression (ma and sma function from 'smatr' package in  $\mathbb{R}^{36}$ ; see Supplementary Result 4 for details).

#### 625 Extended Data Figure 7. Flower number and vegetative cover relationships. Linear

626 regressions between the number of open flowers counted in a quadrat of 0.5m<sup>2</sup> according to the 627 vegetative cover of the focus species in the quadrat (in %). Data are extracted from IPI AgriLand 628 floral transects survey in 2012 for 23 (panels **a-w**) out of the 35 main nectar contributing species. 629 The number of flowers was analyzed according to the vegetative cover ("Cover"), the month of 630 the survey ("Month") and the interaction between these two terms ("Cover:Month") using 631 negative binomial generalized linear models (see Supplementary Result 4 for details). Colored lines represent the linear regression between flower abundance and vegetative cover for each 632 month of the survey. Black lines represent the overall linear regression between flower 633 634 abundance and vegetative cover when the "Month" covariate cannot be included in the model. 635 Line equations were derived from statistical intercept and slope estimates.

# Figures

# Figure 1















# **Extended Data Tables**

# **Extended Data Table 1.**

а									
		nectar prod (kg of suga cover/year)	uctivity rs/ha	nectar o (Shanno	liversity on index)	land cove (000s ha)	r	nectar pro (000 000s sugars/ye	ovision kg of ar)
Habitat	plot number	mean	se	mean	se	England	Great Britain	England	Great Britain
Acid Grass	195	29.38	0.12	0.78	0.04	396	1589	11.64	46.69
Arable	531	6.90	0.08	0.61	0.03	4002	4608	27.61	31.79
Bog	271	39.53	0.11	0.84	0.04	140	2232	5.53	88.23
Bracken	42	36.98	0.25	0.89	0.08	91	260	3.37	9.61
Broadleaf	177	70.04	0.12	1.03	0.04	981	1406	68.71	98.47
Calcareous Grass	9	97.48	0.68	1.54	0.18	30	57	2.92	5.56
Conifer	185	14.49	0.13	0.71	0.04	257	1319	3.72	19.11
Fen	59	39.22	0.21	0.95	0.07	117	392	4.59	15.37
Improved Grass	623	51.73	0.07	0.73	0.02	2856	4494	147.74	232.47
Neutral Grass	317	64.84	0.09	1.03	0.03	1453	2176	94.21	141.09
Shrub Heath	167	82.43	0.13	0.72	0.04	331	1343	27.28	110.70

		nectar prod (kg of suga cover/year)	uctivity rs/ha	nectar o (Shanno	liversity on index)	land cove (000s ha)	land cover (000s ha)		vision kg of ar) <u>Great Britair</u> 53.39 29.39 73.76	
Habitat	plot number	mean	se	mean	se	England	Great Britain	England	Great Britain	
Acid Grass	59	33.60	0.23	0.73	0.08	396	1589	13.31	53.39	
Arable	199	6.38	0.13	0.58	0.04	4002	4608	25.53	29.39	
Bog	37	33.05	0.30	0.84	0.10	140	2232	4.63	73.76	
Bracken	5	55.75	0.98	1.23	0.23	91	260	5.07	14.49	
Broadleaf	46	78.24	0.25	0.97	0.08	981	1406	76.75	110.00	
Calcareous Grass	5	212.54	1.02	1.91	0.24	30	57	6.38	12.11	
Conifer	62	14.01	0.23	0.69	0.08	257	1319	3.60	18.47	
Fen	14	45.18	0.50	0.77	0.14	117	392	5.29	17.71	
Improved Grass	196	51.39	0.12	0.74	0.04	2856	4494	146.77	230.94	
Neutral Grass	109	71.33	0.16	1.03	0.05	1453	2176	103.64	155.21	
Shrub Heath	36	91.03	0.31	0.60	0.09	331	1343	30.13	122.25	

# **Extended Data Table 2.**

Response variable	Effect	df	F value	P-value
Nectar productivity	Habitat	10	75.081	<.0001
	Location	1	0.560	0.455
	Habitat:Location	10	63.519	<.0001
Nectar diversity	Habitat	10	22.061	<.0001
	Location	1	0.147	0.701
	Habitat:Location	10	10.396	<.0001
	Habitat:Location	10	10.396	<.0001

# **Extended Data Table 3.**

		Mean nectar productivity	England land cover	England nectar provision
Option	Option code	(kg of sugars/ha/year)	(ha)	(000s kg of sugars/year)
Wild bird seed mixture	EF2/HF2	56	2966.447	166.12
Enhanced margin	HE10	166.8	617.27	102.96
Nectar strip	EF4/HF4	244	1611.146	393.12
Haymaking supplement	HK18	18.6	1122.83	20.88
Maintenance spp rich grassland	HK6	31.9	2769.69	88.35

# **Extended Data Table 4.**

Response variable	Effect	df	F value	P-value
Nectar productivity	Habitat	10	26.860	<.0001
	Year	3	1.473	0.220
	Habitat:Year	30	1.793	0.005
Nectar diversity	Habitat	10	5.137	<.0001
	Year	3	2.600	0.050
	Habitat:Year	30	2.523	<.0001

# **Extended Data Table 5.**

				Nectar pr	oductivity			
	1978-1990		1990-2000		2000-2007		1978-2007	,
Habitat	est (diff)	p-value	est (diff)	p-value	est (diff)	p-value	est (diff)	p-value
Acid Grass	0.031	0.747	0.074	0.472	0.087	0.394	0.192	0.047
Arable	-0.207	<.0001	0.053	0.331	0.347	<.0001	0.192	<.0001
Bog	-0.009	0.948	0.063	0.645	-0.119	0.363	-0.065	0.638
Bracken	0.045	0.898	0.057	0.854	-0.036	0.918	0.067	0.868
Broadleaf	0.113	0.407	0.026	0.829	0.084	0.473	0.224	0.093
Calcareous Grass	0.153	0.717	0.235	0.504	-0.032	0.927	0.356	0.399
Conifer	-0.274	0.009	0.138	0.169	-0.045	0.655	-0.181	0.084
Fen	-0.383	0.122	-0.097	0.623	0.086	0.675	-0.393	0.131
Improved Grass	0.046	0.41	-0.045	0.422	0.125	0.029	0.125	0.028
Neutral Grass	-0.095	0.319	-0.058	0.514	0.209	0.007	0.055	0.52
Shrub Heath	0.06	0.608	0.012	0.927	0.021	0.874	0.093	0.452

	Nectar diversity									
	1978-1990		1990-2000	1990-2000		2000-2007		1978-2007		
Habitat	est (diff)	p-value	est (diff)	p-value	est (diff)	p-value	est (diff)	p-value		
Acid Grass	-0.161	0.031	0.021	0.798	-0.047	0.557	-0.189	0.014		
Arable	-0.313	<.0001	-0.096	0.024	0.278	<.0001	-0.131	0.003		
Bog	0.103	0.351	0.009	0.935	-0.01	0.923	0.102	0.346		
Bracken	-0.006	0.981	0.037	0.883	0.142	0.605	0.171	0.588		
Broadleaf	-0.184	0.087	-0.01	0.92	-0.088	0.339	-0.282	0.007		
Calcareous Grass	-0.56	0.092	0.226	0.414	0.222	0.421	-0.112	0.736		
Conifer	-0.299	<.0001	0.102	0.193	-0.037	0.642	-0.234	0.005		
Fen	-0.181	0.355	0.081	0.598	-0.272	0.093	-0.372	0.07		
Improved Grass	-0.11	0.012	-0.067	0.136	0.022	0.624	-0.154	<.0001		
Neutral Grass	-0.115	0.126	0.007	0.33	0.013	0.831	-0.034	0.621		
Shrub Heath	-0.021	0.82	0.001	0.927	-0.067	0.515	-0.13	0.181		

# **Extended Data Table 6.**

а												
	Me	ean necta	ar product	ivity	N	ational I	and cov	er		Nectar p	rovision	
	(kg of sugars/ha cover/year)				(000s ha)				(000 000s kg of sugars/year)			
Habitat	1978	1990	2000	2007	1978	1990	2000	2007	1978	1990	2000	2007
Acid Grass	28.00	30.13	35.90	44.13	1786	1821	1502	1589	50.01	54.86	53.91	70.13
Arable	4.73	2.56	3.01	7.93	5105	5025	5067	4608	24.16	12.84	15.27	36.52
Bog	38.97	38.15	44.28	33.44	2004	2050	2222	2232	78.10	78.20	98.39	74.64
Bracken	30.39	33.82	38.69	35.55	258	272	315	260	7.84	9.20	12.19	9.24
Broadleaf	39.69	51.78	55.08	67.10	995	1343	1328	1406	39.49	69.54	73.15	94.34
Calcareous Grass	62.33	89.11	153.89	142.76	53	78	61	57	3.30	6.95	9.39	8.14
Conifer	23.93	12.25	17.20	15.41	1413	1239	1386	1319	33.81	15.18	23.84	20.33
Fen	93.89	38.27	30.46	37.39	231	427	425	392	21.69	16.34	12.94	14.66
Improved Grass	34.29	38.24	34.34	46.09	5188	4619	4251	4494	177.88	176.64	145.98	207.13
Neutral Grass	54.20	43.33	37.75	61.70	1442	1669	2007	2176	78.15	72.32	75.77	134.25
Shrub Heath	59.53	68.49	70.42	73.97	1677	1436	1299	1343	99.82	98.35	91.47	99.34
								Sum	614.26	610.42	612.30	768.71

b													
	Me	ean necta	r product	ivity	Ν	ational I	and cov	er		Nectar p	rovision		
	(kg of sugars/ha cover/year)				(000s ha)				(000 000s kg of sugars/year)				
Habitat	1978	1990	2000	2007	1978	1990	2000	2007	1978	1990	2000	2007	
Acid Grass	22.48	21.45	30.59	37.30	1786	1821	1502	1589	40.16	39.07	45.95	59.26	
Arable	3.89	2.21	2.70	7.32	5105	5025	5067	4608	19.84	11.12	13.70	33.71	
Bog	36.76	36.28	40.83	31.67	2004	2050	2222	2232	73.68	74.38	90.72	70.69	
Bracken	26.59	29.12	35.55	32.56	258	272	315	260	6.86	7.92	11.20	8.47	
Broadleaf	33.80	43.01	48.46	54.67	995	1343	1328	1406	33.63	57.76	64.35	76.87	
Calcareous Grass	50.39	75.40	125.71	118.53	53	78	61	57	2.67	5.88	7.67	6.76	
Conifer	23.90	12.62	18.11	16.27	1413	1239	1386	1319	33.76	15.63	25.10	21.45	
Fen	82.57	24.98	24.29	27.59	231	427	425	392	19.07	10.67	10.32	10.81	
Improved Grass	13.66	11.12	10.27	13.12	5188	4619	4251	4494	70.89	51.38	43.67	58.97	
Neutral Grass	27.70	19.18	19.00	30.53	1442	1669	2007	2176	39.94	32.01	38.12	66.43	
Shrub Heath	56.17	64.67	66.04	69.78	1677	1436	1299	1343	94.20	92.87	85.78	93.72	
								Sum	434.71	398.69	436.58	507.15	

# **Extended Data Figures**

# **Extended Data Figure 1**



## **Extended Data Figure 2**



# **Extended Data Figure 3**

а		b		
	40 -		Contribution to 2000-2007 nectar	shift
			Species	%
			Acer pseudoplatanus	-1.33
			Achillea millefolium	1.66
			Ajuga reptans	1.04
-			Brassica sp.	/
7(%)			Calluna vulgaris	-7.75
200			Centaurea nigra	6.26
0 to	20 -		Chamerion angustifolium	1.03
1ges 200			Cirsium arvense	2.19
chai			Cirsium palustre	5.13
n to ision			Cirsium vulgare	-2.28
provi			Crataegus monogyna	1.76
ontril ctar			Erica cinerea	5.64
ŭ e			Erica tetralix	/
tiona			Hederahelix	3.34
na	0 -		Heracleum sphondylium	10.51
			Prunella vulgaris	1.14
			Ranunculus repens	1.19
			Oilseed rape	1.02
			Rubus fruticosus agg.	3.47
			Salix caprea	1.30
			Salix cinerea	-1.44
			Salix repens agg.	3.33
		Grand Land and Charles Charles Charles and Counter Bors	Senecio erucifolius	1.36
	arovec	Bentan , Bio boil Ching , Bone Bi	Senecio jacobaea	2.97
	100.	Calco.	Taraxacum agg.	3.95
			Trifolium pratense	3.25
			Trifolium repens	45.88
			Vaccinium myrtillus	3.02

45.8	Trifolium repens
3.0	Vaccinium myrtillus