



Improved representation of plant functional types and physiology in the Joint UK Land Environment Simulator (JULES v4.2) using plant trait information

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Abstract. Dynamic global vegetation models are used to predict the response of vegetation to climate change. They are essential for planning ecosystem management, understanding carbon cycle–climate feedbacks, and evaluating the potential impacts of climate change on global ecosystems. JULES (the Joint UK Land Environment Simulator) represents terrestrial processes in the UK Hadley Centre family of models and in the first generation UK Earth System Model. Previously, JULES represented five plant functional types (PFTs): broadleaf trees, needle-leaf trees, C₃ and C₄ grasses, and shrubs. This study addresses three developments

in JULES. First, trees and shrubs were split into deciduous and evergreen PFTs to better represent the range of leaf life spans and metabolic capacities that exists in nature. Second, we distinguished between temperate and tropical broadleaf evergreen trees. These first two changes result in a new set of nine PFTs: tropical and temperate broadleaf evergreen trees, broadleaf deciduous trees, needle-leaf evergreen and deciduous trees, C₃ and C₄ grasses, and evergreen and deciduous shrubs. Third, using data from the TRY database, we updated the relationship between leaf nitrogen and the maximum rate of carboxylation of Rubisco (V_{cmax}), and updated the leaf

turnover and growth rates to include a trade-off between leaf life span and leaf mass per unit area.

Overall, the simulation of gross and net primary productivity (GPP and NPP, respectively) is improved with the nine PFTs when compared to FLUXNET sites, a global GPP data set based on FLUXNET, and MODIS NPP. Compared to the standard five PFTs, the new nine PFTs simulate a higher GPP and NPP, with the exception of C₃ grasses in cold environments and C₄ grasses that were previously over-productive. On a biome scale, GPP is improved for all eight biomes evaluated and NPP is improved for most biomes – the exceptions being the tropical forests, savannahs, and extratropical mixed forests where simulated NPP is too high. With the new PFTs, the global present-day GPP and NPP are 128 and 62 Pg C year⁻¹, respectively. We conclude that the inclusion of trait-based data and the evergreen/deciduous distinction has substantially improved productivity fluxes in JULES, in particular the representation of GPP. These developments increase the realism of JULES, enabling higher confidence in simulations of vegetation dynamics and carbon storage.

1 Introduction

The net exchange of carbon dioxide between the vegetated land and the atmosphere is predominantly the result of two large and opposing fluxes: uptake by photosynthesis and efflux by respiration from soils and vegetation. CO₂ can also be released by land ecosystems due to vegetation mortality resulting from human and natural disturbances, such as changes in land use practices, insect outbreaks, and fires. Vegetation models are used to quantify many of these fluxes, and the evolution of the terrestrial carbon sink strongly affects future greenhouse gas concentrations in the atmosphere (Friedlingstein et al., 2006; Friedlingstein, 2015; Arora et al., 2013). A subset of vegetation models also predicts both compositional and biogeochemical responses of vegetation to climate change (dynamic global vegetation models, DGVMs), one of these being the Joint UK Land Environment Simulator (JULES). (Best et al., 2011; Clark et al., 2011). JULES predecessor, the Met Office Surface Exchange Scheme (MOSES: Cox et al., 1998, 1999; Essery et al., 2001, 2003) was the land component of the Hadley Centre Global Environmental Model (HadGEM2), and JULES will represent the land surface in the next generation UK Earth System Model (UKESM). Within JULES, the TRIFFID model (Top-down Representation of Foliage and Flora Including Dynamics; Cox, 2001) predicts changes in biomass and the fractional coverage of five plant functional types (PFTs; broadleaf trees, needle-leaf trees, C₃ grass, C₄ grass, and shrubs) based on cumulative carbon fluxes and a predetermined dominance hierarchy. DGVMs such as JULES are essential for planning ecosystem management, understanding carbon cycle–climate feedbacks, and evaluating the potential

Table 1. Parameters used for the five PFT experiment (JULES5). The standard PFTs are broadleaf trees (BT), needle-leaf trees (NT), C₃ grass, C₄ grass, and shrubs (SH). N_m was calculated by dividing the default N_{10} by C_{mass} (0.5 in this study), LMA was calculated as $\sigma_L \times C_{mass}$, and s_v was calculated to yield the same $V_{cmax,25}$ as with the default five PFTs. All other parameters were taken from Clark et al. (2011). Parameters are defined in Table A1.

	BT	NT	C ₃	C ₄	SH
a_{wl}	0.65	0.65	0.005	0.005	0.10
D_{crit}	0.09	0.06	0.10	0.075	0.10
d_T	9	9	9	9	9
f_0	0.875	0.875	0.900	0.800	0.900
f_d	0.010	0.015	0.015	0.025	0.015
i_v	0	0	0	0	0
L_{max}	9	6	4	4	4
L_{min}	1	1	1	1	1
LMA	0.075	0.200	0.050	0.100	0.100
N_a^*	1.73	3.30	1.83	3.00	3.00
N_m	0.023	0.0165	0.0365	0.030	0.030
rootd	3	1	0.5	0.5	0.5
s_v	21.33	8.00	32.00	8.00	16.00
T_{low}	0	−10	0	13	0
T_{off}	5	−40	5	5	5
T_{opt}	32	22	32	41	32
T_{upp}	36	26	36	45	36
$V_{cmax,25}$	36.8	26.4	58.4	24.0	48.0
α	0.08	0.08	0.12	0.06	0.08
γ_0	0.25	0.25	0.25	0.25	0.25
γ_p	20	15	20	20	15
μ_{rl}	1.0	1.0	1.0	1.0	1.0
μ_{sl}	0.10	0.10	1.00	1.00	0.10

* These are derived from other parameters. Here N_a is $g N m^{-2}$.

impacts of climate change on global ecosystems. However, the use of DGVMs in ESMs is relatively rare. For example, of the nine coupled carbon cycle–climate models evaluated by Arora et al. (2013), only three distinct DGVMs interactively simulated changes in the spatial distribution of PFTs (the spatially explicit individual-based (SEIB)-DGVM, JS-BACH (the Jena Scheme for Biosphere–Atmosphere Coupling in Hamburg), and JULES/TRIFFID).

Previous benchmarking studies of JULES and MOSES identified certain areas needing improvement, such as the seasonal cycle of evaporation, gross primary productivity (GPP), and total respiration in regions with seasonally frozen soils and in the tropics; too high growing season respiration; and too low GPP in temperate forests (Blyth et al., 2011); and too high GPP in the tropics (20° S–20° N) (Blyth et al., 2011; Anav et al., 2013). In 21st century simulations, JULES vegetation carbon was sensitive to climate change. In particular, the tropics were very sensitive to warming, with large simulated losses of carbon stored in the Amazon forest when the climate became very dry and hot (Cox et al., 2000, 2004, 2013; Galbraith et al., 2010; Huntingford et al., 2013).

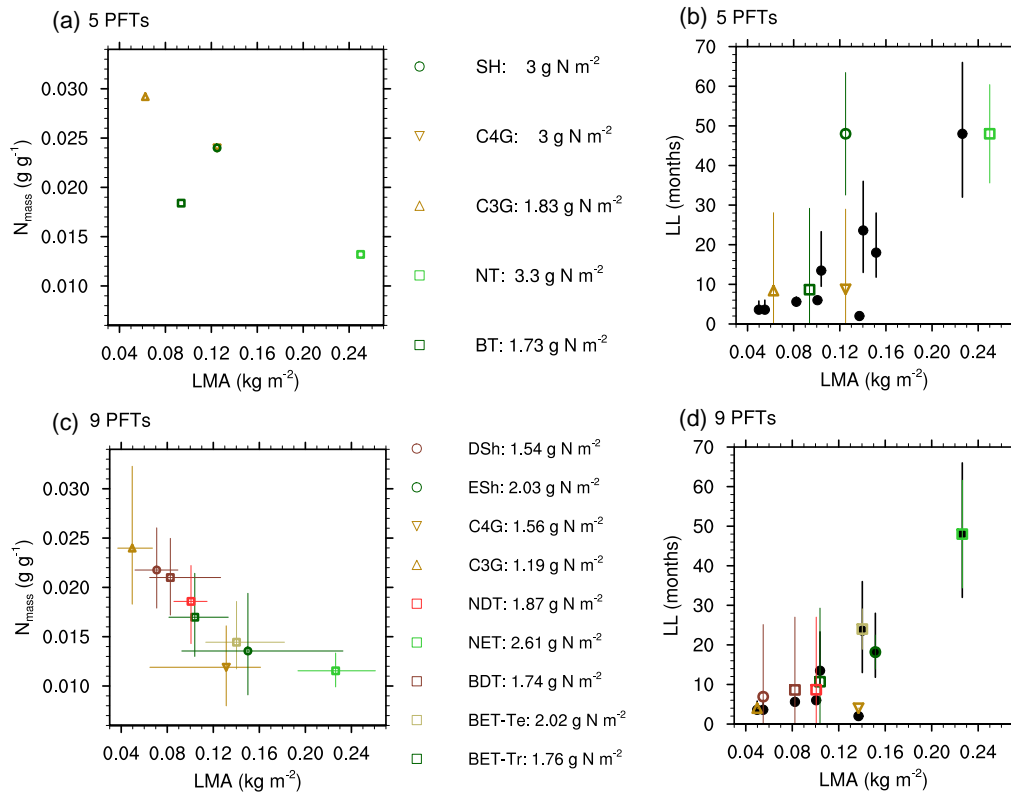


Figure 1. Trade-offs between leaf mass per unit area (LMA; kg m⁻²) and (a, c) leaf nitrogen (g g⁻¹), and between LMA and (b, d) leaf life span (LL). (a, b) Parameters in the standard JULES, converted from N_{l0} and σ_l based on 0.4 kg C per kg dry mass (assumed parameter in JULES from Clark et al., 2011). (c, d) Median values from the TRY database for the new nine PFTs. In (b) and (d), the filled circles show the observed data and the open shapes show the median values from global simulations of JULES from 1982 to 2012. Vertical and horizontal lines show the range of values between the lower and upper quartile of data.

Based on these previous results, our study addressed three potential improvements in the parameterization and representation of PFTs in JULES. First, the original five PFTs (Table 1) did not represent the range of leaf life spans and metabolic capacities that exists in nature, and so trees and shrubs were split into deciduous and evergreen PFTs. In a broad sense, the differences between evergreen and deciduous strategies can be summarized in a leaf economics spectrum, where leaves employ trade-offs in their nitrogen use (Reich et al., 1997; Wright et al., 2004; Fig. 1). When photosynthesis is limited by CO₂, the photosynthetic capacity of a leaf is dependent on the maximum rate of carboxylation of Rubisco (V_{cmax}). Plants allocate about 10–30 % of their nitrogen into synthesis and maintenance of Rubisco (Evans, 1989), while a portion of the remaining nitrogen is put toward leaf structural components, and hence the strong relationship between photosynthetic capacity and leaf nitrogen concentration (e.g., Meir et al., 2002; Reich et al., 1998; Wright et al., 2004) and leaf structure (Niinemets, 1999). On average, evergreen species have a lower photosynthetic capacity and respiration per unit leaf mass (Reich et al., 1997; Wright et al., 2004; Takashima et al., 2004), higher leaf mass

per unit area (LMA) (Takashima et al., 2004; Poorter et al., 2009), allocate a lower fraction of leaf N to photosynthesis (Takashima et al., 2004), and exhibit lower N loss at senescence (Aerts, 1995; Silla and Escudero, 2003; Kobe et al., 2005) than deciduous species. There is also a positive relationship between LMA and leaf life span (Reich et al., 1992, 1997; Wright et al., 2004). Leaves with high nutrient concentration tend to have a short life span and low LMA. They are able to allocate more nutrients to photosynthetic machinery to rapidly assimilate carbon at a relatively high rate (but they also have high respiration rates). Conversely, leaves with less access to nutrients use a longer-term investment strategy, allocating nutrients to structure, defense, and tolerance mechanisms. They tend to have longer life spans, low assimilation and respiration rates, but high LMAs.

Second, we distinguished between tropical broadleaf evergreen trees and broadleaf evergreen trees from warm-temperate and Mediterranean climates, based on fundamental differences in leaf traits, chemistry, and metabolism (Niinemets et al., 2007, 2015; Xiang et al., 2013). For example, measured V_{cmax} for a given leaf N per unit area (N_A) can be lower in tropical evergreen trees than in temperate broadleaf

Table 2. Updated parameters used in JULES_{9ALL}. The new PFTs are tropical broadleaf evergreen trees (BET-Tr), temperate broadleaf evergreen trees (BET-Te), needle-leaf evergreen trees (NET), needle-leaf deciduous trees (NDT), C₃ grass, C₄ grass, evergreen shrubs (ESh), and deciduous shrubs (DSh).

	BET-Tr	BET-Te	BDT	NET	NDT	C ₃	C ₄	ESh	DSh
a_{wl}	0.65	0.65	0.65	0.65	0.75	0.005	0.005	0.10	0.10
D_{crit}	0.090	0.090	0.090	0.060	0.041	0.051	0.075	0.037	0.030
d_T	9	9	9	9	9	0	0	9	9
f_0	0.875	0.892	0.875	0.875	0.936	0.931	0.800	0.950	0.950
f_d	0.010	0.010	0.010	0.015	0.015	0.019	0.019	0.015	0.015
i_v	7.21	3.90	5.73	6.32	6.32	6.42	0.00	14.71	14.71
L_{max}	9	7	7	7	6	3	3	4	4
L_{min}	1	1	1	1	1	1	1	1	1
LMA	0.1039	0.1403	0.0823	0.2263	0.1006	0.0495	0.1370	0.1515	0.0709
N_a^*	1.76	2.02	1.74	2.61	1.87	1.19	1.55	2.04	1.54
N_m	0.017	0.0144	0.021	0.0115	0.0186	0.0240	0.0113	0.0136	0.0218
rootd	3	2	2	1.8	2	0.5	0.5	1	1
s_v	19.22	28.40	29.81	18.15	23.79	40.96	20.48	23.15	23.15
T_{low}	13	13	5	5	-5	10	13	10	0
T_{off}	0	-40	5	-40	5	5	5	-40	5
T_{opt}	39	39	39	33	34	28	41	32	32
T_{upp}	43	43	43	37	36	32	45	36	36
$V_{cmax,25}$	41.16	61.28	57.25	53.55	50.83	51.09	31.71	62.41	50.40
α	0.08	0.06	0.08	0.08	0.10	0.06	0.04	0.06	0.08
γ_0	0.25	0.50	0.25	0.25	0.25	3.0	3.0	0.66	0.25
γ_p	15	15	20	15	20	20	20	15	30
μ_{rl}	0.67	0.67	0.67	0.67	0.67	0.72	0.72	0.67	0.67
μ_{sl}	0.10	0.10	0.10	0.10	0.10	1.00	1.00	0.10	0.10

* These are derived from other parameters. Here N_a is $g\ N\ m^{-2}$.

evergreen trees (Kattge et al., 2011), resulting in lower V_{cmax} and maximum assimilation rates for tropical forests (Carswell et al., 2000; Meir et al., 2002, 2007; Domingues et al., 2007, 2010; Kattge et al., 2011). Collectively, the evergreen/deciduous and tropical/temperate distinctions resulted in a new set of nine PFTs for JULES: tropical broadleaf evergreen trees (BET-Tr), temperate broadleaf evergreen trees (BET-Te), broadleaf deciduous trees (BDT), needle-leaf evergreen trees (NET), needle-leaf deciduous trees (NDT), C₃ grasses, C₄ grasses, evergreen shrubs (ESh), and deciduous shrubs (DSh) (Table 2).

Lastly, several parameters relating to variation in photosynthesis and respiration have not been updated since MOSES was developed in the late 1990s. We used data on LMA ($kg\ m^{-2}$), leaf N per unit mass, N_m ($kg\ N\ kg^{-1}$), and leaf life span from the TRY database (Kattge et al., 2011; accessed November 2012). The new parameters for leaf nitrogen and LMA were used to calculate a new V_{cmax} at 25 °C, and to update phenological parameters that determine leaf life span. Other parameters related to leaf dark respiration, canopy radiation, canopy nitrogen, stomatal conductance, root depth, and temperature sensitivities of V_{cmax} were revised based on a review of recently available observed values, which are described in Sect. 2.

The purpose of our paper is to document these changes, and to evaluate their impacts on the ability of JULES to model CO₂ exchange for selected sites and globally on the scale of biomes, with a focus on the gross and net primary productivity. Specifically, we explore the consequences for carbon fluxes on seasonal and annual timescales of switching from the current five PFTs to a greater number of PFTs (nine) that account for growth habit (evergreen versus deciduous) and temperate/tropical plant types.

2 Model description

Full descriptions of the model equations are in Clark et al. (2011) and Best et al. (2011). Here we briefly describe relevant current equations in JULES, associated changes in terms of updated parameter values, and document new equations and parameters. The revisions discussed in our study fall into three categories: (1) changes to model physiology based on leaf trait data from TRY; (2) adjustment of parameters to account for the properties of the new PFTs (evergreen/deciduous, tropical/temperate); and (3) calibration of parameters based on known biases in the model and a review of the literature. Parameters for the standard five PFTs and for the new nine PFTs are given in Tables 1 and 2, re-

spectively, and a summary of all parameters are in Table A1 in Appendix A. For the site-level simulations, we incrementally made changes to the model to determine whether or not changes improved the simulations. This resulted in a total of eight experiments (Table 3). The version of JULES with five PFTs (Experiment 0) is kept as similar as possible to the configuration used in the TRENDY experiments, which are a set of historical simulations to quantify the global carbon cycle (e.g., Le Quéré et al., 2014; Sitch et al., 2015) that have been included in several recent publications. In the supplement, we provide a set of recommended parameters and guidance for users who wish to run JULES with the original five PFTs (Table A2).

2.1 JULES model

In JULES, leaf-level photosynthesis for C_3 and C_4 plants (Collatz et al., 1991, 1992) is calculated based on the limiting factor of three potential photosynthesis rates: W_1 (light limited rate), W_e (transport of photosynthetic products for C_3 and PEPCarboxylase limitation for C_4 plants), and W_c (Rubisco limited rate) (see Supplement). W_e and W_c depend on V_{cmax} , the maximum rate of carboxylation of Rubisco, which is a function of the V_{cmax} at 25 °C ($V_{cmax,25}$):

$$V_{cmax} = \frac{V_{max,25} f_T(T_C)}{[1 + \exp(0.3(T_C - T_{upp}))][1 + \exp(0.3(T_{low} - T_C))]}, \quad (1)$$

where T_C is the canopy temperature in Celcius, and

$$f_T(T_C) = Q_{10,leaf}^{0.1(T_C-25)}, \quad (2)$$

where T_{upp} and T_{low} are PFT-dependent parameters. $Q_{10,leaf}$ is 2.0.

JULES has several options for representing canopy radiation. Option 5, as described in Clark et al. (2011), includes a multi-layer canopy with sunlit and shaded leaves in each layer, two-stream radiation with sunflecks penetrating below the top layer, and light-inhibition of leaf respiration. Additionally, N is assumed to decay exponentially through the canopy with an extinction coefficient, k_n , of 0.78 (Mercado et al., 2007). $V_{cmax,25}$ is calculated in each canopy layer (i) as

$$V_{max,25,i} = n_{eff} N_{10} e^{-k_n(i-1)/10}, \quad (3)$$

assuming a 10-layer canopy. The parameter N_{10} is the top-leaf nitrogen content (kg N kg C^{-1}), and n_{eff} linearly relates leaf N concentration to $V_{cmax,25}$.

Leaf dark respiration is assumed to be proportional to the V_{cmax} calculated in Eq. (1):

$$R_d = f_d V_{cmax} \quad (4)$$

with a 30 % inhibition of leaf respiration when irradiance is $> 10 \mu\text{mol quanta m}^{-2} \text{s}^{-1}$ (Atkin et al., 2000; Mercado et

al., 2007; Clark et al., 2011). Plant net primary productivity (NPP) is very sensitive to f_d , and since the vegetation fraction depends on NPP when the TRIFFID competition is turned on, the distribution of PFTs can also be sensitive to f_d . The parameter was modified from 0.015 (Clark et al., 2011) to 0.010 for all broadleaf tree PFTs in this study, based on underestimated coverage of broadleaf trees in previous versions of JULES. Leaf photosynthesis is calculated as

$$A_1 = (W - R_d)\beta, \quad (5)$$

where W is the smoothed minimum of the three limiting rates (W_1 , W_e , W_c), and β is a soil moisture stress factor. The factor β is 1 when soil moisture content of the root zone (θ : $\text{m}^3 \text{m}^{-3}$) is at or above a critical threshold (θ_{crit}), which depends on the soil texture. When soil water content drops below θ_{crit} , β decreases linearly until θ reaches the wilting point (where $\beta = 0$) (Cox et al., 1998).

Stomatal conductance (g_s) is linked to leaf photosynthesis:

$$A = \frac{g_s(C_s - C_i)}{1.6}, \quad (6)$$

where C_s and C_i are the leaf surface and internal CO_2 concentrations, respectively. The gradient in CO_2 between the internal and external environments is related to leaf humidity deficit at the leaf surface (D) following Jacobs (1994):

$$\frac{C_i - \Gamma^*}{C_s - \Gamma^*} = f_0 \left(1 - \frac{D}{D_{crit}}\right). \quad (7)$$

Here, Γ^* is the CO_2 compensation point – or the internal partial pressure of CO_2 at which photosynthesis and respiration balance, and D_{crit} is the critical humidity deficit (f_0 and D_{crit} are PFT-dependent parameters). In JULES, the surface latent heat flux (LE) is due to evaporation from water stored on the canopy, evaporation of water from the top layer of soil, transpiration through the stomata, and sublimation of snow. Any change to LE will also impact the sensible heat and ground heat fluxes, since these are linked to the total surface energy balance (Best et al., 2011).

Total plant (autotrophic) respiration, R_a , is the sum of maintenance and growth respiration (R_{pm} and R_{pg} , respectively):

$$R_{pm} = 0.012 R_d \left(\beta + \frac{N_r + N_s}{N_l}\right) \quad (8)$$

and

$$R_{pg} = r_g (GPP - R_{pm}), \quad (9)$$

where r_g is a parameter set to 0.25 (Cox et al., 1998, 1999), and the nitrogen concentration of roots, stem, and leaves are given by N_r , N_s , and N_l , respectively. When using canopy

Table 3. Experiments for the FLUXNET site-level evaluation.

Experiment number	Description
0: JULES5	Five PFTs (Table 1)
1	Nine PFTs with N_m , LMA, and $V_{cmax,25}$ from TRY
2: JULES9-TRY	Exp. 1 + parameters affecting leaf life span
3	Exp. 2 + f_0 and D_{crit}
4	Exp. 2 + α
5	Exp. 2 + adjusted f_d , T_{upp} , T_{low} , and s_v
6	Exp. 2 + rootd, a_{wl}
7: JULES9	All new PFT parameters (Table 2)

radiation model 5 in JULES, these are calculated as

$$N_l = N_{l0} \sigma_1 \cdot LAI, \quad (10)$$

$$N_r = N_{l0} \sigma_1 \mu_{rl} \cdot L_{bal}, \quad (11)$$

$$N_s = N_{l0} \mu_{sl} \eta_{sl} h \cdot L_{bal}, \quad (12)$$

where σ_1 is specific leaf density ($\text{kg C m}^{-2} \text{LAI}^{-1}$), h is the vegetation height in meters, L_{bal} is the balanced leaf area index (LAI) (the seasonal maximum of LAI based on allometric relationships, Cox, 2001), μ_{rl} and μ_{sl} relate N in roots and stems to top-leaf N, and η_{sl} is $0.01 \text{ kg C m}^{-1} \text{LAI}^{-1}$. In Eqs. (10)–(12), N_{l0} , σ_1 , μ_{rl} , and μ_{sl} are PFT-dependent parameters.

The NPP is

$$\text{NPP} = \text{GPP} - R_a. \quad (13)$$

For each PFT in JULES, the NPP determines the carbon available for spreading (expanding fractional coverage in the grid cell, only relevant when the TRIFFID competition is turned on) or for growth (growing leaves or height). The net ecosystem exchange (NEE; positive flux from the land to the atmosphere) is

$$\text{NEE} = R_{eco} - \text{GPP}, \quad (14)$$

where R_{eco} is the total ecosystem respiration.

Phenology in JULES affects leaf growth rates and timing of leaf growth/senescence based on temperature alone (Cox et al., 1999; Clark et al., 2011). When canopy temperature (T_c) is greater than a temperature threshold (T_{off}), the leaf turnover rate (γ_m) is equal to γ_0 . When $T_c < T_{off}$, the turnover rate is modified as in Eq. (15a) (where T_{off} , γ_0 , and d_T are PFT-dependent parameters):

$$\gamma_m = \gamma_0 (1 + d_T (T_{off} - T_c)) \quad \text{for } T_c \leq T_{off}, \quad (15a)$$

$$\gamma_m = \gamma_0 \quad \text{for } T_c > T_{off}. \quad (15b)$$

The leaf turnover rate affects phenology ($p = \frac{LAI}{L_{bal}}$) by triggering a loss of leaf area for $\gamma_m > 2\gamma_0$, and a growth of leaf

area when $\gamma_m \leq 2\gamma_0$:

$$\frac{dp}{dt} = \gamma_p (1 - p) \quad \text{for } \gamma_m \leq 2\gamma_0, \quad (16a)$$

$$\frac{dp}{dt} = -\gamma_p \quad \text{for } \gamma_m > 2\gamma_0, \quad (16b)$$

where γ_p is the leaf growth rate.

2.2 Updated leaf N, $V_{cmax,25}$, and leaf life span (Experiments 1–2)

Essentially, with the revised trait-based physiology, the parameter σ_1 (Eqs. 10–11) and N_{l0} (Eqs. 3, 10–12) were replaced with LMA and N_m , respectively, from the TRY database. N_{l0} and N_m both describe the nitrogen content at the top of the canopy, but the former is N per unit carbon, while the latter is the more commonly observed N per unit dry mass. N_m can be converted to N_{l0} using leaf carbon content per dry mass (C_m). Historically, C_m was 0.4 in JULES (Schulze et al., 1994), but we updated it to 0.5 in all versions of JULES evaluated in this study (Reich et al., 1997; White et al., 2000; Zaehle and Friend, 2010).

We also changed the equation for $V_{cmax,25}$ from a function of N_{l0} (Eq. 3) to a function of leaf N per unit area, N_a , a more commonly observed leaf trait, calculated as the product of the observed leaf traits LMA (kg m^{-2}) and N_m (kg N kg^{-1}):

$$N_a = N_m \cdot \text{LMA} \quad (17)$$

and $V_{cmax,25}$ ($\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) is

$$V_{cmax,25} = i_v + s_v N_a, \quad (18)$$

where parameters i_v ($\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) and s_v ($\mu\text{mol CO}_2 \text{ gN}^{-1} \text{ s}^{-1}$) were taken directly from Kattge et al. (2009; hereafter K09) (see also Medlyn et al., 1999), with two exceptions. First, the V_{cmax} parameterization from K09 was based on the leaf C_3 photosynthesis model. C_4 plants have high CO_2 concentration at the site of Rubisco, and therefore require less Rubisco than C_3 plants (von Caemmerer and Furbank, 2003). C_4 species typically have 30–50% as much Rubisco per unit N as C_3 species (Sage and Percy, 1987; Makino et al., 2003; Houborg et al.,

2013). We chose a slope (s_v) for C_4 to give a $V_{\text{cmax},25}$ that is half of that for C_3 grass, and set the intercept (i_v) to 0. This resulted in a $V_{\text{cmax},25}$ of $32 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ for C_4 grass, which is similar to observed values in natural grasses (Kubien and Sage, 2004; Domingues et al., 2007) and $V_{\text{cmax},25}$ in seven other ESMs ($13\text{--}38 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$; Rogers, 2013). Second, K09 reported a separate $V_{\text{cmax},25}$ for tropical trees growing on oxisols (old tropical soils with low phosphorous availability) and non-oxisols. For the BET-Tr PFT, we calculated a weighted mean slope and intercept from their Table 2 to represent an “average” tropical soil.

The new $V_{\text{cmax},25}$ for canopy level i is calculated as (replacing Eq. 3)

$$V_{\text{max},25_i} = i_v + s_v N_a e^{-K_n(i-1)/10}. \quad (19)$$

The leaf, root, and stem nitrogen contents are (replacing Eqs. 10–12)

$$N_l = N_m \text{LMA} \cdot \text{LAI}, \quad (20)$$

$$N_r = N_m \text{LMA} \mu_{rl} \cdot L_{\text{bal}}, \quad (21)$$

$$N_s = \frac{N_m}{C_m} \mu_{sl} \eta_{sl} \cdot h \cdot L_{\text{bal}}, \quad (22)$$

Four phenological parameters (T_{off} , d_T , γ_0 , and γ_p , Eqs. 15–16) were adjusted to capture the trade-off between leaf life span and LMA. We set T_{off} to 5°C for deciduous trees and shrubs, to -40°C for BET-Te, NET, and ESh, and to 0°C for BET-Tr. The latter reflects the fact that many tropical evergreen tree species cannot tolerate frost (Woodward and Williams, 1987; Prentice et al., 1992). For the other evergreen PFTs, the value of -40°C ensured that plants only lose their leaves in extremely cold environments. Second, we changed d_T to 0 for grasses to attain constant leaf turnover rates (Eq. 15). This fixed an unrealistic seasonal cycle in LAI of grasses and makes grasses more competitive in very cold environments (Hopcroft and Valdes, 2015). Third, we adjusted γ_0 for grasses and evergreen species to reflect the median observed leaf life span in the TRY database. Last, we changed γ_p from its default value of 20 to 15 year^{-1} for the PFTs with the thickest leaves (NET, ESh, BET-Temp, BET-Trop) and to 30 year^{-1} for the PFT with the thinnest leaves (DSh). The parameter γ_p controls the rate of leaf growth in the spring and senescence at the end of the growing season (Eq. 16b). To reduce an overestimation of uptake during the spring with the new phenology for grass, the maximum LAI for grasses was reduced from 4 to 3.

2.3 Other updates to JULES parameters with new PFTs (Experiments 3–6)

Additional changes to JULES were made to account for the properties of the new PFTs, to incorporate recent observations, and to correct known biases in the model. These fall into four categories: radiation, stomatal conductance, photosynthesis and respiration, and plant structure. For the site-

level evaluation of JULES, we incrementally added these changes (Table 3).

2.3.1 Stomatal conductance (Experiment 3)

JULES stomatal conductance is related to the leaf internal CO_2 , where C_i/C_s is proportional to the parameters f_0 and $1/D_{\text{crit}}$ (Eq. 7). For vapor pressure deficits (D) greater than D_{crit} , the stomata close. For $D < D_{\text{crit}}$, stomata gradually open in response to a reducing evaporative demand. Needle-leaf species in JULES have a lower D_{crit} than other trees, grasses, and shrubs. The lower D_{crit} increases the likelihood of the stomata being closed – similar to Mediterranean conifers that tend to close their stomata earlier than angiosperms (Carnicer et al., 2013) – and it tightly regulates the stomatal aperture, making plants more sensitive to increasing D . This is analogous to plants conserving water at the expense of assimilation. We use updated f_0 and D_{crit} from a synthesis of water use efficiency at the FLUXNET sites (Dekker et al., 2016). Compared to the standard five PFT parameters, the D_{crit} was decreased for BET-Te, NDT, C_3 grass, and shrubs. The parameter f_0 was increased for these PFTs, which increased C_i for all $D < D_{\text{crit}}$.

2.3.2 Radiation (Experiment 4)

The light-limited photosynthesis rate (W_l) is proportional to $\alpha \times [\text{absorbed PAR}]$, where α is the quantum efficiency of photosynthesis ($\text{mol CO}_2 [\text{mol quanta}]^{-1}$). We reduced α from 0.08 to 0.06 for C_3 grass and evergreen PFTs typical of semi-arid and arid environments, and from 0.06 to 0.04 for C_4 grass, where previously the model over-predicted GPP for a given PAR. Quantum efficiency was set at 0.10 for NDT. These values are still within the range reported in Skillman (2008). An example of the changes is shown in the Supplement, Fig. S1. Decreasing the α for BET-Te and ESh PFTs helped reduce a high bias in the GPP at low irradiances at Las Majadas (Spain – a savannah site), while increasing α for NDT improved the light response of GPP at Tomakai (Japan – a larch site).

2.3.3 Photosynthesis and respiration parameters (Experiment 5)

The leaf dark respiration is calculated as a fraction, f_d , of V_{cmax} (Eq. 4). In testing JULES, we found that C_3 grasses were overly productive and tended to be the dominant grass type even in tropical ecosystems where we expected C_4 dominance. Therefore, we increased the f_d for C_3 (from 0.015 to 0.019) and decreased the f_d for C_4 (from 0.025 to 0.019) so the two grass PFTs would have similar R_d rates for a given V_{cmax} .

Preliminary evaluation of JULES GPP at the FLUXNET sites in Table 4 revealed the need for a higher (lower) $V_{\text{cmax},25}$ for the BET-Tr and NDT (BET-Te) PFTs than the mean value reported in K09. For these PFTs, the slope parameter (s_v)

Table 4. Sites used in the site simulations. Land cover is according to site PI.

Site name	Location	Simulated years	Land cover	Dominant PFT(s)
BR-Ma2	Manaus, Brazil	2002–2005	Evergreen broadleaf forest	100 % BET
BR-Sa1	Santarem (Tapajós Forest, KM67), Brazil	2002–2004	Evergreen broadleaf forest	100 % BET
BR-Sa3	Santarem (Tapajós Forest, KM77), Brazil	2001–2005	Pasture	20 % BET, 75 % C ₄ , 5 % soil
DE-Tha	Tharandt, Germany	1998–2006	Needle-leaf evergreen forest	100 % NET
ES-ES1	El Saler, Spain	1999–2006	Needle-leaf evergreen forest	100 % NET
ES-LMa	Las Majadas, Spain	2004–2006	Closed shrub	33 % Temp-BET, 33 % C ₃ , 33 % ESh
FI-Hyy	Hyytiälä, Finland	1998–2002	Needle-leaf evergreen forest	100 % NET
FI-Kaa	Kaamanen, Finland	2000–2005	Wetland (simulated as C ₃ grass)	80 % C ₃ grass, 20 % bare soil
JP-Tom	Tomakai, Japan	2001–2003	Needle-leaf deciduous plantation	10 % BDT, 10 % NET, 80 % NDT
US-Bo1	Bondville, IL, US	1997–2006	Crop (rotating C ₃ /C ₄)	40 % C ₃ , 40 % C ₄ , 20 % soil
US-FPe	Fort Peck, MT, US	2000–2006	Grassland (C ₃)	80 % C ₃ grass, 20 % bare soil
US-Ha1	Harvard, MA, US	1995–2001	Broadleaf deciduous forest	100 % BDT
US-MMS	Morgan Monroe Forest, US	2000–2004	Broadleaf deciduous forest	100 % BDT
US-Ton	Tonzi, CA, US	2001–2006	Woody savannah	33 % BDT, 33 % C ₃ , 33 % DS

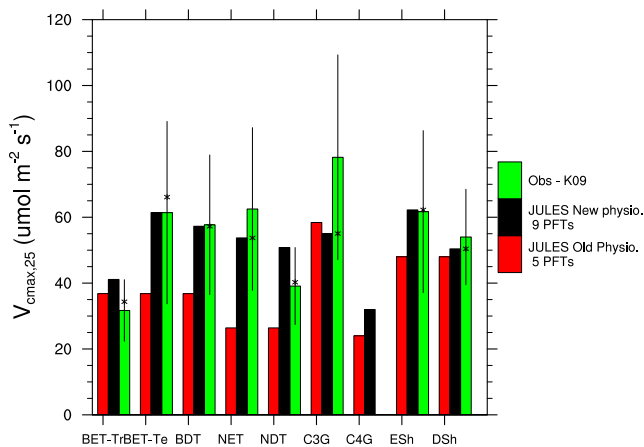


Figure 2. $V_{cmax,25}$ for the new nine PFTs (black), from the comparable PFT from the TRY data (Kattge et al., 2009) (green), and from the standard five PFTs (red). Asterisks indicate the $V_{cmax,25}$ for JULES9 prior to calibration based on the FLUXNET sites. The standard deviation reported in Kattge et al. (2009) are also shown for the observations with the vertical lines. BET-Tr – Tropical broadleaf evergreen trees, BET-Te – Temperate broadleaf evergreen trees, BDT – Broadleaf deciduous trees, NET – Needle-leaf evergreen trees, NDT – Needle-leaf deciduous trees, C3G – C₃ grass, C4G – C₄ grass, ESh – Evergreen shrubs, DSh – Deciduous shrubs.

was adjusted to result in the final $V_{cmax,25}$ for each PFT (black bars, Fig. 2), using the mean ± 1 standard deviation of $V_{cmax,25}$ from K09 as an upper limit.

T_{upp} and T_{low} were also modified, as optimal V_{cmax} can occur at temperatures near 40 °C (Medlyn et al., 2002), and the previous optimal temperature for V_{cmax} was 32 °C for broadleaf trees (BT) and 22 °C for NT. A study of seven broadleaf deciduous tree species found T_{opt} for V_{cmax} ranging from 35.9 °C to > 45 °C (Dreyer et al., 2001), and maximum V_{cmax} can occur at temperatures of at least 38 °C in the Amazon forest (B. Kruijt, personal communication, 2015). There-

fore, we changed T_{opt} from 32 to 39 °C for all broadleaf trees and from 22 to 33 and 32 °C for NET and NDT, respectively. C₃ grass T_{opt} was decreased from 32 to 28 °C to help reduce the high productivity bias in grasses.

Additionally, the ratio of nitrogen in roots to leaves (μ_{rl}) was updated following the relationships in Table 1 of Kerkhoff et al. (2006). However, instead of assigning a separate μ_{rl} for each PFT, we assigned the mean values for trees/shrubs and grasses (0.67 and 0.72, respectively).

2.3.4 Plant structure (Experiment 6)

There is evidence that larch trees (NDT) can be tall with a relatively low LAI compared to needle-leaf evergreen trees (Ohta et al., 2001; Hirano et al., 2003) and compared to broadleaf deciduous trees (Gower and Richards, 1990). In JULES, canopy height (h) is proportional to the balanced LAI, L_b :

$$h = \frac{a_{wl}}{a_{ws} \cdot \eta_{sl}} L_b^{b_{wl}-1}. \quad (23)$$

The parameter a_{wl} relates the LAI to total stem biomass, and for trees it is 0.65. Hirano et al. (2003) found $h = 15$ m and maximum LAI = 2.1, which would imply $a_{wl} = 0.91$, and Ohta et al. (2001) found $h = 18$ m and LAI = 3.7, implying $a_{wl} = 0.75$. Therefore, we adjusted a_{wl} for NDT to 0.75, which was an important change for allowing NDT to out-compete BDT in high latitudes.

We also changed the root depths, although these changes were constrained by the 3 m deep soil in the standard JULES setup. Previously, root depths were 3 m for broadleaf trees, 1 m for needle-leaf trees, and 0.5 m for grasses and shrubs (Best et al., 2011). With the new PFTs, roots are shallower for BET-Te and BDT (2 m), and deeper for NET (1.8 m), NDT (2 m), and shrubs (1 m) (Zeng, 2001).

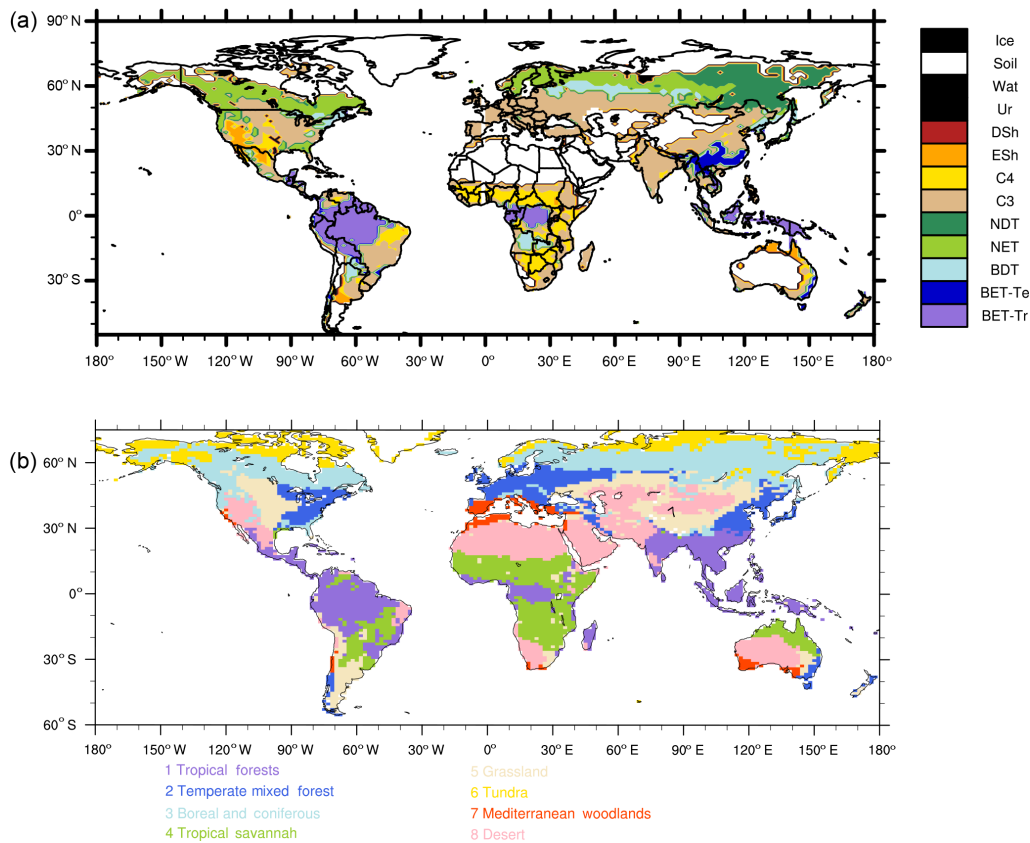


Figure 3. (a) Dominant vegetation type from the ESA LC_CCI data set, aggregated to the new nine PFTs. (b) Color-coded map of global biomes, based on World Wildlife Fund biomes.

3 Methods

3.1 Data

We analyzed leaf N_m , specific leaf area ($= 1/LMA$), and leaf life span from the TRY database (accessed in November 2012). Data were translated from species level to both the standard five and new nine PFTs based on a look-up table provided by TRY, and screened for duplicate entries. We only selected entries with measurements for both LMA and N_m . This resulted in 9372 LMA / N_m pairs and 1176 leaf life span measurements (Supplement).

To evaluate the model performance we used GPP from the model tree ensemble (MTE) of Jung et al. (2011), MODIS NPP from the MOD17 algorithm (Zhao et al., 2005; Zhao and Running, 2010), and GPP and NEE from 13 and 14 FLUXNET sites (Table 4). Using the net exchange of CO_2 observed at the FLUXNET sites, NEE was partitioned into GPP and R_{eco} . Assuming that nighttime $NEE = R_{eco}$, R_{eco} was estimated as a temperature function of nighttime NEE (Reichstein et al., 2005; Groenendijk et al., 2011).

3.2 Model simulations

We performed two sets of simulations to evaluate the impacts of the new PFTs in JULES v4.2. First, site-level simulations used observed meteorology from 14 FLUXNET towers – these include the nine original sites benchmarked in the study of Blyth et al. (2011), plus an additional five to represent more diversity in land cover types and climate. The vegetation cover was prescribed as in Table 4, and vegetation competition was turned off. The changes described in Sect. 2.2 and 2.3 were incrementally added to evaluate the effect of each group of changes (Table 3). Full results are shown in the Supplement, but for the main text we focus the discussion on JULES with five PFTs (JULES5); JULES with nine PFTs and updated N_m , LMA, $V_{cmax,25}$, and leaf life span from the TRY database (JULES9_{TRY}); and JULES with nine PFTs and all updated parameters described in Sect. 2.3 (JULES9_{ALL}). These are, respectively, Experiments 0, 2, and 7 in Table 3.

Soil carbon takes more than 1000 years to equilibrate in JULES, so we used an accelerated method that only requires 200–300 years of spin-up (depending on the site). JULES has four soil pools (decomposable and resistant plant material, long-lived humus, and microbial biomass), and the decom-

possible material pool has the fastest turnover rate (equivalent to $\sim 10 \text{ year}^{-1}$) (Clark et al., 2011). For each experiment, soil carbon was spun-up using accelerated turnover rates in the three slower soil pools for the first 100 years. The rates of the resistant, humus, and biomass pools were increased by a factor of 33, 15, and 500, respectively, so all pools had the same turnover time as the fastest pool. This resulted in unrealistically depleted soil carbon pools. The second step of the spin-up was to multiply the pool sizes by these same factors, and then allow the soil carbon to spin-up under normal conditions for an additional 100–200 years.

Second, global simulations were conducted for JULES5 and JULES9_{ALL}. It could be argued that similar model improvements might be gained with the original five PFTs with improved parameters. We tested this hypothesis with a third global experiment, JULES5_{ALL}, with five PFTs but improved parameters (Table A2). The global simulations followed the protocol for the S2 experiments in TRENDY (Sitch et al., 2015), where the model was forced with observed annual-average CO₂ (Dlugokencky and Tans, 2013), climate from the CRU-NCEP data set (v4, N. Viovy, personal communication, 2013), and time-invariant fraction of agriculture in each grid cell (Hurtt et al., 2011). Vegetation cover was prescribed based on the European Space Agency's Land Cover Climate Change Initiative (ESA LC_CCI) global vegetation distribution (Poulter et al., 2015, processed to the JULES 5 and nine PFTs by A. Hartley) (Fig. 3a). JULES did not predict vegetation coverage in this study, which enabled us to evaluate JULES GPP and NPP given a realistic land cover. The evaluation of vegetation cover and updated competition for nine PFTs will be evaluated in a follow-up paper. Since the land cover was prescribed based on a 2010 map, we also set the agricultural mask based on land use in 2010, and enforced consistency between the two maps such that fraction of agriculture could not exceed the fraction of grass in each grid cell. During the spin-up (300 years with 100 years of accelerated turnover rates as at the sites), we used atmospheric CO₂ concentration from 1860 and recycled climate from 1901–1920. The transient simulation (with time-varying CO₂ and climate) was from 1901–2012. The model spatial resolution was N96 (1.875° longitude \times 1.25° latitude).

3.3 Model evaluation

The model evaluation is presented in two stages. First, using the site-level simulations, we evaluated GPP and NEE with the root mean square error (RMSE) and correlation coefficient, r , based on daily and monthly averaged fluxes, respectively. Site history can result in non-zero annual NEE, but JULES maintains annual carbon balance, so it is not realistic to expect the simulated annual NEE to match the observations. Therefore, we compared anomalies of NEE instead.

We summarized the changes in RMSE and r using relative improvements for each experiment in Table 4, *i*. The statistics were calculated such that positive values denote an

improvement compared to JULES5 (Experiment 0):

$$\text{RMSE_rel}_i = \frac{\text{RMSE}_{5\text{pfts}} - \text{RMSE}_i}{\text{RMSE}_{5\text{pfts}}}, \quad (24)$$

$$r_rel_i = \frac{r_i - r_{5\text{pfts}}}{r_{5\text{pfts}}}. \quad (25)$$

Second, we compared the model from global simulations to biome-averaged fluxes in eight biomes based on 14 World Wildlife Fund terrestrial ecoregions (Olson et al., 2001) (Fig. 3b, Table S3). Fluxes were averaged for the land in each biome in both the model and the observations. We evaluated seasonal cycles of GPP from the MTE (Jung et al., 2011), and annually averaged GPP (from the MTE) and NPP (from MODIS). The tropical forest biome includes regions of tropical grasslands and pasture – in the ESA LC_CCI data set, the BET-Tr PFT is dominant in only 38 % of the biome and grasses occupy 36 %. Therefore, we only included the grid cells where the dominant PFT in the ESA data is BET-Tr. The extratropical mixed forest biome has a large coverage of agricultural land, and as a result 46 % of the biome is C₃ grass, while BDT and NET only cover 14 and 8 % of the biome, respectively. We omitted grid cells with > 20 % agriculture in 2012 to calculate the biome average fluxes.

4 Results

4.1 Data analysis of leaf traits

With the previous five PFTs, only the needle-leaf tree PFT occupied the “slow investment” end of the leaf economics spectrum (high LMA and low N_m) (Fig. 1). The new PFTs were given the median N_m and LMA from the TRY data set (Fig. 1c), and these exhibit a range of deciduous and evergreen strategies, although there is substantial overlap between PFTs. The needle-leaf evergreen trees, evergreen shrubs, and temperate broadleaf evergreen trees have low N_m and thick leaves, but their N_A (shown in the legend of Fig. 1a, c) is relatively high ($> 2 \text{ g m}^{-2}$), which has been long known for species with long leaf life spans ($> 1 \text{ year}$) (Reich et al., 1992). These traits on aggregate indicate that they use the “slow investment” strategy of growing thick leaves with low rates of photosynthesis per unit investment of biomass.

Compared to the evergreen PFTs, the deciduous shrubs and broadleaf deciduous trees have higher N_m , thinner leaves, lower N_A (1.3–1.7 g N m^{-2}), and leaf life spans of less than 6 months. The tropical broadleaf evergreen trees have a moderate N_m and leaf thickness, with an average life span of 11 months, reflecting a mixture of successional stages in the database. The grasses have the shortest leaf life spans. C₄ grasses have high LMA, low N_m , and a high N_A ; while the thinner C₃ grasses have a high N_m and low N_A . Figure 1 also shows the impacts of changing the phenological parameters (T_{off} , d_T , γ_0 , and γ_p , Eqs. 15–16) on median leaf life

Table 5. Comparison of simulated and observed annual GPP and NPP at FLUXNET sites, listed in order from most to least productive. Units: $\text{g C m}^{-2} \text{ year}^{-1}$. Results are color-coded so blue shows when there is an improvement. The GPP and NPP are based on similar data processing between the FLUXNET observations and model. Sources: ¹ Malhi (2009); ² Gower and Richards (1990), assuming 0.5 g C g^{-1} biomass.

Site	GPP JULES5 JULES9 OBS			NPP JULES5 JULES9 OBS		
BR-Sa1	2671	2795	3314 ± 600	850	1048	1440 ± 130 ¹
BR-Ma2	2848	3225	3285 ± 835	867	1198	1011 ± 140 ¹
BR-Sa3	3318	2116		1623	1125	
DE-Tha	1364	1876	1923 ± 547	700	1004	
JP-Tom	1306	1361	1723 ± 641	691	747	1100 ²
ES-ES1	1164	1087	1458 ± 383	513	404	
US-MMS	1135	1234	1445 ± 463	603	693	
US-Ha1	1229	1438	1433 ± 531	686	851	
US-Bo1	896	1006	1233 ± 568	457	591	
ES-LMA	1095	1257	1133 ± 305	500	644	
FI-Hyy	1124	1465	1084 ± 324	605	834	
US-Ton	818	794	924 ± 256	365	405	
US-FPe	238	368	354 ± 185	88	192	
FI-Kaa	633	512	297 ± 126	359	311	

span during a 30-year global simulation, where now JULES captures the observed leaf life spans.

Based on the new N_A , $V_{\text{cmax},25}$ was updated using the new parameters i_v and s_v (Eq. 18; Fig. 2). The values calculated from the TRY data are shown with asterisks, and these were used in the JULES9_{TRY} experiments. The black bars show the final $V_{\text{cmax},25}$ after adjusting s_v for the two broadleaf evergreen tree PFTs and the needle-leaf deciduous trees (see Sect. 2.3.3). Within the trees, the temperate broadleaf evergreen PFT has the highest $V_{\text{cmax},25}$, while the needle-leaf deciduous and tropical broadleaf evergreen PFTs have the lowest. Because the JULES C_3 and C_4 PFTs are assumed representative of natural vegetation, they have relatively low $V_{\text{cmax},25}$ (compared to the range from K09 for C_3). The N_A calculated from median N_m and LMA in this study (1.19 g N m^{-2}) is lower than the average N_A reported in K09 (1.75 g N m^{-2}). However, the $C_3 V_{\text{cmax},25}$ ($51.09 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$) is close to values reported for European grasslands ($41.9 \pm 6.9 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ and $48.6 \pm 3.5 \mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ for graminoids and forbs, respectively, in Wohlfahrt et al., 1999). In comparison to JULES5, the new $V_{\text{cmax},25}$ is higher for all PFTs except for C_3 grass. Previously, the $V_{\text{cmax},25}$ was lower than the observed range for all non-tropical trees, but now the $V_{\text{cmax},25}$ for all PFTs is within the range of observed values.

4.2 Site-level simulations

In most cases, the higher V_{cmax} from trait data increased the GPP and NPP, and resulted in higher respiration fluxes due to both autotrophic (responding to higher GPP) and heterotrophic (responding to higher litterfall due to higher NPP) respiration. First, we compared JULES with five PFTs (JULES5) to JULES with nine PFTs and the TRY data

(JULES9_{TRY}) (Experiments 1 and 2, respectively, in Table 3) at the sites listed in Table 4. The results are summarized in Fig. 4, where yellows and reds indicate increased correlation (Fig. 4a, b) or reduced RMSE (Fig. 4c, d) in each experiment compared to JULES5. Using the N_m , LMA, and $V_{\text{cmax},25}$ data from TRY improved the seasonal cycle of GPP at the two tropical forest sites, the evergreen savannah, and the crop site, and decreased the daily RMSE at one NET site (Tharandt), all grass sites, and the NDT site (Tomakai) (Experiment 1, Fig. 4). Enforcing the LMA–leaf life span relationship further improved the seasonal cycle at both savannah sites, the two natural C_3 grass sites (the seasonal cycle was worse at the crop site), and the NDT site, and further reduced RMSE at the deciduous savannah site and one BDT site (Harvard) (Experiment 2, a.k.a. JULES9_{TRY}). In comparison, applying all parameter changes summarized in Table 3 further reduced the RMSE at every site except the two tropical forests and further increased r at every site except the tropical forests and the evergreen savannah (Experiment 7, a.k.a. JULES9_{ALL}).

Overall, the carbon and energy exchanges were best captured with JULES9_{ALL}. Compared to JULES5, the RMSE for GPP in JULES9_{ALL} decreased by more than 40 % at Kaamanen (C_3 grass), Tharandt (NET), and Tomakai (NDT); the daily RMSE of NEE decreased at eight sites; and r increased for NEE at 11 sites. The only sites without an improvement in either metric for NEE were Manaus (BET-Tr) and Bondville (Crop). The improvements to NEE were large at Tharandt (r from 0.61 to 0.76), Fort Peck C_3 grass (0.05 to 0.38), and Tomakai (0.09 to 0.93), and RMSE for NEE decreased by more than 35 % at Kaamanen and Tomakai. Respiration and latent heat fluxes are discussed in the supplemental material.

On an annual basis, GPP was higher in JULES9_{ALL} than in JULES5 at every site except for the Tapajós K77 pasture,

Table 6. (a) Area-weighted GPP from each biome ($\text{g C m}^{-2} \text{ year}^{-1}$). The biome total GPP from MTE is given in Pg C year^{-1} to give perspective of each biome's role in the global total. (b) Area-weighted NPP from each biome ($\text{g C m}^{-2} \text{ yr}^{-1}$).

(a) Biome	JULES5	JULES9	JULES5-ALL	MTE	MTE total
Tropical forest	2403 ± 217	2295 ± 191	2505 ± 217	2244 ± 297	49.9
Tropical forest: only BET-Tr.	2924 ± 144	2955 ± 147	3279 ± 178	2790 ± 273	
Tropical savannah	1355 ± 244	1268 ± 223	1320 ± 237	1111 ± 257	21.9
Extratropical mixed forests	947 ± 147	1082 ± 158	1119 ± 167	1119 ± 212	2.9 (13.4*)
Boreal and coniferous forests	514 ± 99	597 ± 118	645 ± 122	650 ± 203	12.1
Temperate grasslands	420 ± 145	465 ± 138	477 ± 140	509 ± 184	8.1
Deserts and shrublands	82 ± 48	91 ± 46	91 ± 47	283 ± 200	4.9
Tundra	86 ± 20	94 ± 20	101 ± 20	279 ± 233	1.9
Mediterranean woodlands	324 ± 147	407 ± 136	405 ± 140	510 ± 190	1.5

(b) Biome	JULES5	JULES9	JULES5-ALL	MODIS17
Tropical forest	956 ± 144	1007 ± 125	951 ± 143	786 ± 352
Only BET-Tr.	1141 ± 101	1233 ± 103	1109 ± 126	929 ± 315
Tropical savannah	527 ± 158	591 ± 143	584 ± 152	451 ± 319
Extratropical mixed forests	586 ± 93	631 ± 104	640 ± 110	563 ± 231
Boreal and coniferous forests	307 ± 65	358 ± 77	385 ± 80	350 ± 155
Temperate grasslands	180 ± 94	243 ± 89	242 ± 90	304 ± 247
Deserts and shrublands	16 ± 29	35 ± 29	33 ± 29	111 ± 133
Tundra	52 ± 14	61 ± 13	65 ± 13	136 ± 94
Mediterranean Woodlands	118 ± 94	201 ± 89	195 ± 89	324 ± 184

* Value for EMF (extra-tropical mixed forest) biome when agricultural mask is not applied.

El Saler (NET), Tonzi (savannah), and Kaamanen, and NPP was higher at every site except for Tapajós K77, El Saler, and Kaamanen (Table 5). Total GPP was improved at every site except for Hyytiälä (NET) and Las Majadas (savannah), where annual GPP was too high in JULES5, and at El Saler and Tonzi, where the modeled GPP was too low. However, for every site except Hyytiälä, JULES9_{ALL} was within the range of observed annual GPP. We now explore some site-specific aspects of the carbon cycle results.

4.2.1 Broadleaf forests

Both GPP and NPP were higher in JULES9_{ALL} than JULES5 for broadleaf forests due to a higher T_{opt} of V_{cmax} and a higher $V_{\text{cmax},25}$. Simulated GPP was similar to observations in the absence of soil moisture stress. The increase in GPP occurred year-round at Manaus, but only during the wet season at Tapajós K67 (Fig. 5). GPP was similar in all JULES simulations during the dry season (October–December), when soil moisture deficits limited photosynthesis. The soil moisture stress factor, β , was < 0.7 during these months, while it was > 0.87 all year at Manaus (recall that a higher β indicates less stress). The reduction in GPP during the dry season at both sites is in contrast to the observations, which show an increase from August–December. As a result, the simulated seasonal cycle of GPP was incorrect at both sites, and although the annual total GPP was closer to observations, the monthly RMSE was higher in JULES9_{ALL}

compared to JULES5. The simulated NPP was too low in JULES5 at both sites. In JULES9_{ALL}, the NPP was too high at Manaus (by $187 \text{ g C m}^{-2} \text{ year}^{-1}$) and too low at Tapajós (by $396 \text{ g C m}^{-2} \text{ year}^{-1}$).

At the two BDT sites (Harvard and Morgan Monroe), the peak summer GPP was closer to observations in JULES9_{ALL}. GPP was very well reproduced at Harvard (BDT), where the average JJA temperature was 4°C cooler than at Morgan Monroe (29°C compared to 33°C), and, due to differences in the soil parameters, the soil moisture stress factor was higher ($\beta > 0.8$ at Harvard compared to $0.5 < \beta < 0.7$ at Morgan Monroe). At Morgan Monroe, the observed GPP was nearly zero from November–March, but all versions of JULES simulated uptake during November–December, when the average temperatures were still above freezing, possibly due to leaves staying on the trees for too long in the model. The RMSE of NEE decreased (Fig. 5b), but the amplitude of the seasonal cycle was too small at both BDT sites.

4.2.2 Needle-leaf forests

The seasonal cycle of GPP improved at the needle-leaf forests, but JULES9_{ALL} underestimated GPP during mid-summer at the larch site (Tomakai) and during the summer at a Mediterranean site (El Saler), and overestimated summertime GPP at a cold conifer site (Hyytiälä). Although there was a large improvement in the seasonal cycle at El Saler in JULES9_{ALL}, the GPP was still underestimated during the dry

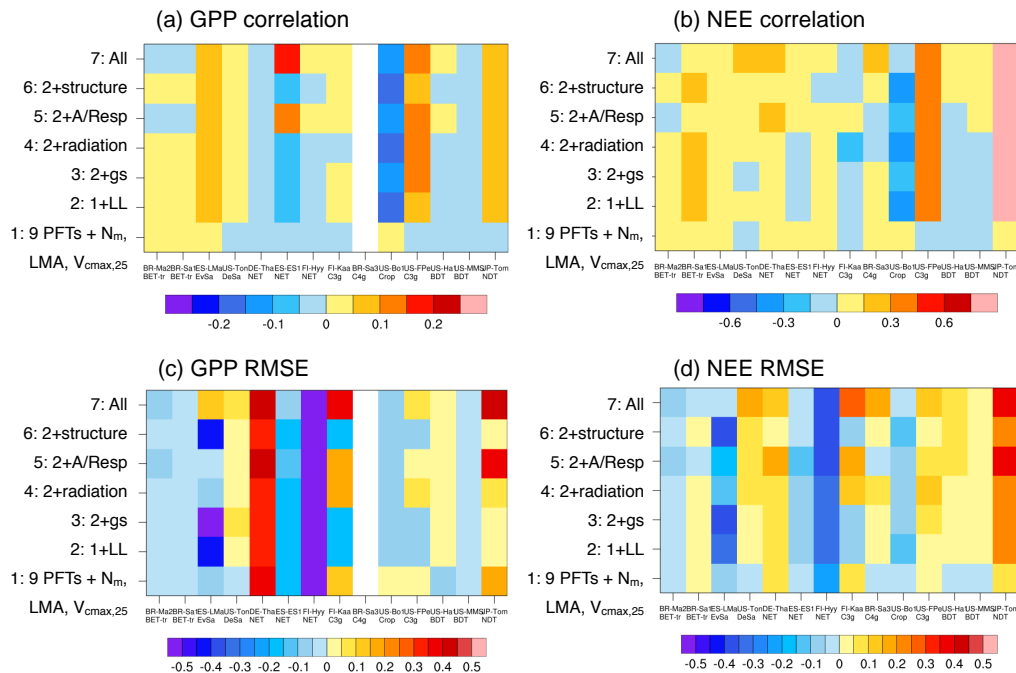


Figure 4. Relative changes in daily RMSE (Eq. 24) and monthly correlation coefficients (Eq. 25) for the JULES experiments in Table 4 compared to JULES5. Yellows and reds indicate an improvement in JULES compared to the FLUXNET observations.

months of June–October. During this period, β reduced to a minimum of 0.17 in August, and the GPP was too low by an average $1.83 \text{ g C m}^{-2} \text{ d}^{-1}$. At all sites there was shift toward stronger net carbon uptake during the summer months with the new PFTs, which increased the correlation with observed NEE. At El Saler, the RMSE of NEE increased due to a change in the seasonal cycle of leaf dark respiration (R_d , Eq. 8) resulting from the higher T_{opt} . At Hyytiälä, the RMSE of NEE increased due to higher rates of soil respiration during the winter months (Fig. S3; where soil respiration is the difference between total and autotrophic respiration).

Compared to JULES5 (with a needle-leaf PFT), both GPP and respiration were improved with the new NDT PFT at Tomakai, primarily due to an improved seasonal cycle of GPP with the deciduous phenology (Experiment 2). In JULES5, the LAI at the site was $6.0 \text{ m}^2 \text{ m}^{-2}$, compared to a summer maximum of $\sim 3.5 \text{ m}^2 \text{ m}^{-2}$ with the deciduous phenology and to a reported average LAI of larch of $3.8 \text{ m}^2 \text{ m}^{-2}$ (Gower and Richards, 1990). The new deciduous PFT also improved the seasonal cycle of NEE, and reduced errors in LE and SH (Fig. S4). The magnitude of maximum summertime GPP was still underestimated, but this could be because the site is a plantation, where trees are evenly planted to optimize the incoming radiation, rather than a natural larch forest.

4.2.3 Grasses

GPP and NEE were improved for temperate grasslands (Kaamanen and Fort Peck) and NEE was improved at a tropical pasture (Tapajós K77). Compared to JULES5, productivity in JULES9_{ALL} was higher at a temperate C₃ site (Fort Peck), and lower at a cold C₃ site (Kaamanen) and the tropical C₄ site. In terms of GPP, these changes brought JULES9_{ALL} closer to the observations (Table 5). With the new PFT parameters, grasses had higher year-round LAI due to the removal of phenology, and GPP increased earlier in the year at Kaamanen, Bondville, and Fort Peck in JULES9_{ALL} compared to JULES5. Net uptake also occurred 1–2 months earlier in JULES9_{ALL} (compared to JULES5), which decreased RMSE and increased r for NEE at the three natural grassland sites. JULES9_{ALL} underestimated productivity at Bondville (crop site), but this is not surprising given that the PFT is meant to represent natural grasses. There is a separate crop model available for JULES (Osborne et al., 2015).

The Tapajós K77 pasture was not included in the set of sites with GPP/ R_{eco} partitioning. The simulated GPP was lower in JULES9_{ALL} than in JULES5 due to the lower quantum efficiency (Fig. S3c). The seasonal cycle of NEE was close to that observed during most months (Fig. 5b), and in terms of r and RMSE JULES9_{ALL} were better than JULES5. In JULES5, the GPP and NPP were higher at the Tapajós K77 pasture than at the Tapajós K67 forest site despite being driven by the same meteorology (Table 5). In JULES9_{ALL},

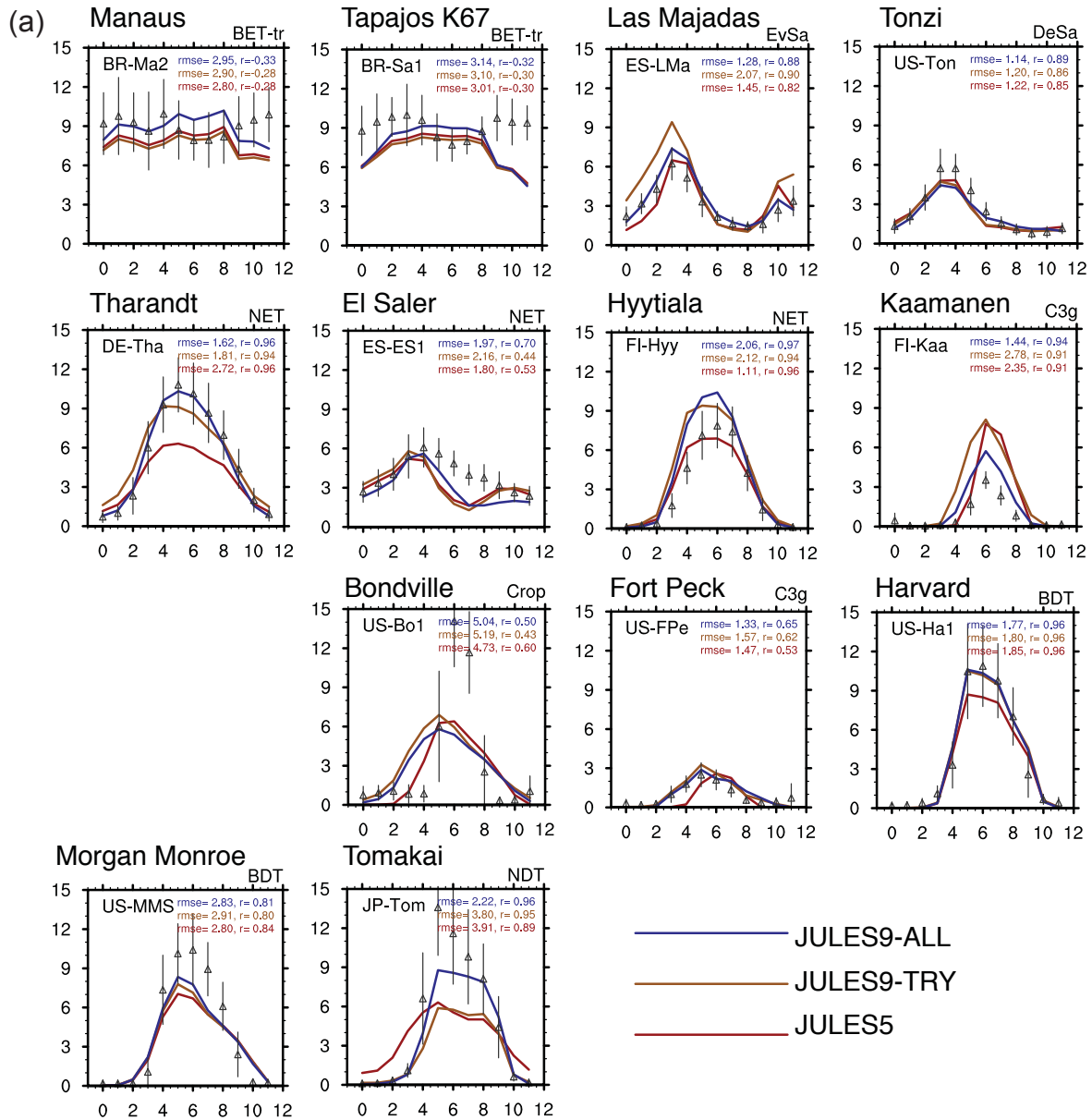


Figure 5.

GPP was higher at the forest site than at the pasture, and the NPP was similar.

4.2.4 Mixed vegetation sites

Las Majadas and Tonzi are savannah sites dominated by evergreen and deciduous plants, respectively (assumed in the simulations to be an equal mix of trees, shrubs, and C₃ grass, Table 4). Both GPP and NPP were better simulated with JULES9_{ALL} at both sites, and the annual GPP was within the range of the observations (although it was too high at Las Majadas and too low at Tonzi).

At Las Majadas, the GPP increased in JULES9_{ALL} (compared to JULES5) during the wet spring (January–April) due to high GPP from the BET-Te and C₃ grass PFTs. The former had a higher year-round LAI (~4.6 m² m⁻²), V_{cmax,25}, and T_{opt} for V_{cmax} compared to the BT from the five PFTs (which simulated maximum summer LAI of 3.8 m² m⁻²). For C₃ grass, the new V_{cmax,25} and T_{opt} were lower in JULES9_{ALL}, but the removal of phenology (setting d_T to 0) increased the LAI during the cool, mild winter months when photosynthesis could still occur. Grid-cell mean GPP was also slightly higher during the hot, dry summer, again owing to the BET-Te PFT. The simulated seasonality NEE was similar to obser-

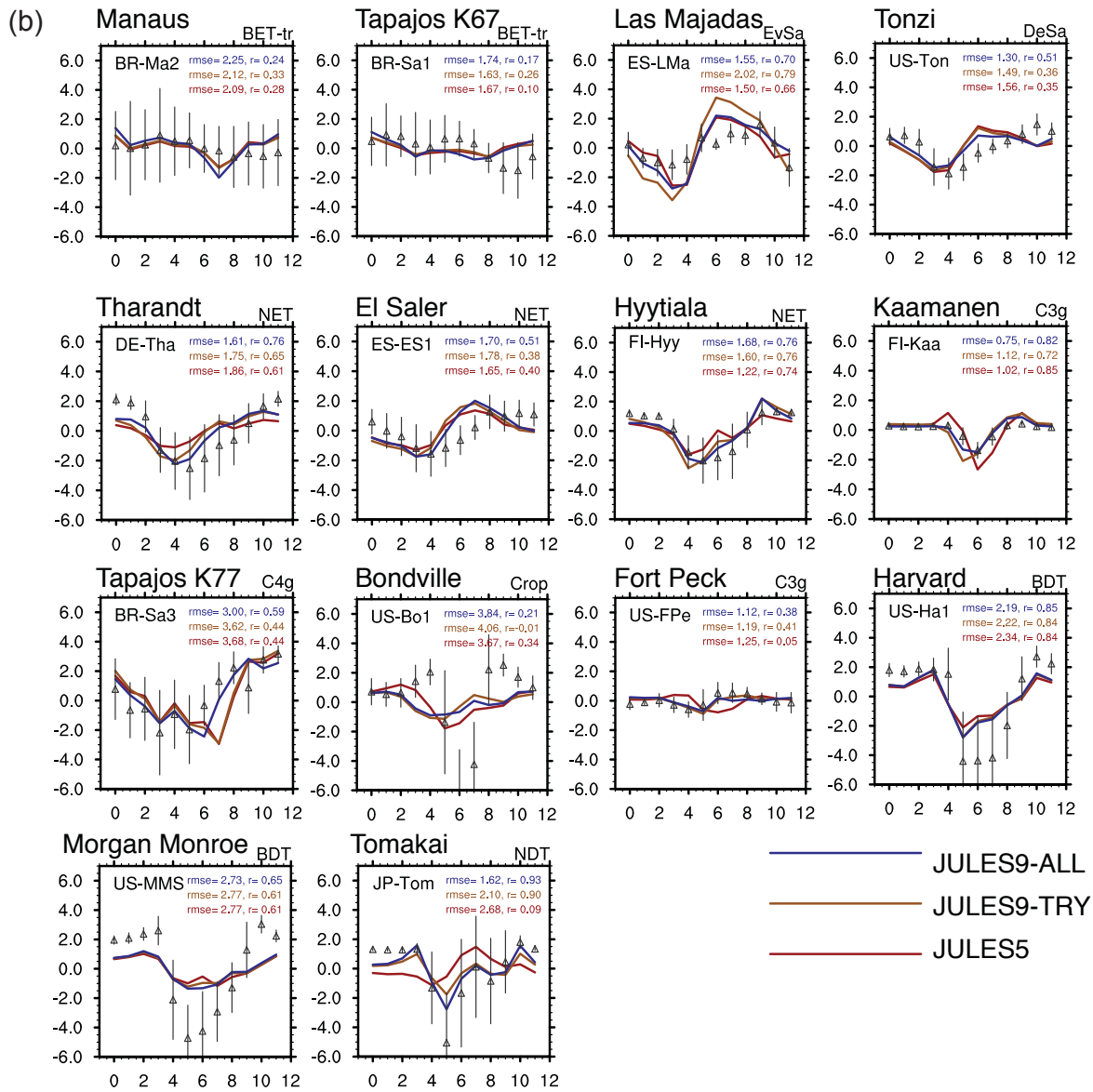


Figure 5. (a) Monthly mean fluxes of GPP. Observations \pm standard deviation from FLUXNET are shown with triangles and vertical lines. The three JULES simulations are JULES5 with standard five PFTs (JULES5, red); JULES with nine PFTs and new LMA, N_m , and $V_{cmax,25}$ from TRY (JULES9_{TRY}, orange); JULES9-TRY plus new parameters for the PFTs as discussed in Sect. 2.3 (JULES9_{ALL}, blue). Also shown are the daily root mean square error (RMSE) based on daily fluxes and the correlation coefficient (r) based on monthly mean fluxes for all years of the simulations. Site information is given in Table 3. All units are in $\text{g C m}^{-2} \text{ d}^{-1}$. (b) As in (a) but for monthly anomalies of NEE.

variations ($r = 0.70$), but the April–May uptake was too strong and contributed to an overestimation of the annual GPP.

At Tonzi, GPP was similar to observations except during April–July, when it was too low. The modeled photosynthesis began to decline after March, coinciding with a rapid increase in simulated soil moisture stress and stomatal resistance. Moving from a generic to a deciduous shrub resulted in a large decrease in simulated GPP at this site. The shrub LAI decreased from $\sim 3.3 \text{ m}^2 \text{ m}^{-2}$ to a maximum of $1.5 \text{ m}^2 \text{ m}^{-2}$, and the $V_{cmax,25}$ for the DSh was slightly lower

than the $V_{cmax,25}$ for the generic shrub. Slightly compensating for the lower shrub GPP was a higher broadleaf tree GPP, with a higher $V_{cmax,25}$ and T_{opt} compared to the previous values in JULES5.

4.3 Global results

In this section, we analyze the impact of the PFT-specific biases and improvements on biome-scale GPP and NPP fluxes in global simulations. The area-weighted fluxes are displayed in Table 6 and Fig. 6 for the biomes shown in

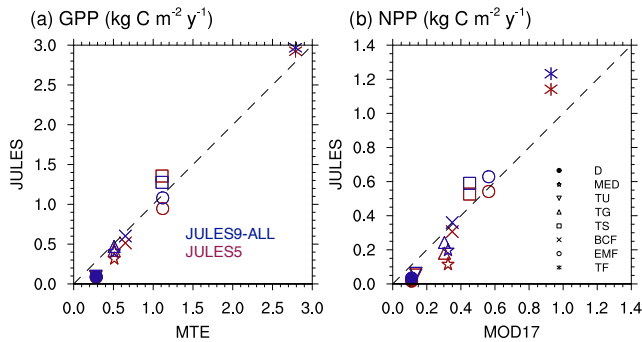


Figure 6. Annual GPP and NPP for the eight biomes shown in Fig. 3b. Biome abbreviations are D – deserts, M – Mediterranean woodlands, TU – tundra, TG – temperate grasslands, TS – tropical savannahs, BCF – boreal and coniferous forests, EMFs – extratropical mixed forests, TF – tropical forests.

Fig. 3, and seasonal cycles are shown in Fig. 7. GPP increased in JULES9_{ALL} compared to JULES5 in all extratropical biomes, but it decreased in the two biomes with significant coverage by C₄ grass. For all biomes, the representation of GPP in JULES9_{ALL} was closer to the observed (MTE) value. NPP increased in every biome, and this was an improvement (relative to MOD17) in five biomes (boreal and coniferous forests, temperate grasslands, deserts/shrublands, tundra, and Mediterranean woodlands), but NPP was too high in tropical biomes and extratropical mixed forests.

In the tropical forests, the biome-averaged GPP and NPP increased in JULES9_{ALL} compared to JULES5, and both fluxes were $\sim 200 \text{ g C m}^{-2} \text{ yr}^{-1}$ higher than their respective observational value. The seasonality of rainfall in the tropics has a hemispheric dependence. Splitting the biome into the Northern and Southern hemispheres revealed that the seasonal cycle in Fig. 7a was most similar to the Southern Hemisphere in terms of the climate and fluxes. In both hemispheres, the JULES GPP was higher than the MTE GPP during the transition period from the wet to the dry season and the early dry season. This is in contrast to the results at the two Brazilian FLUXNET sites, where JULES GPP was lower than that observed during the dry season.

Most of the differences between JULES5_{ALL} and JULES9_{ALL} were in the tropics (Fig. 9, Table 6). The global GPP was relatively high ($135 \text{ Pg C year}^{-1}$) in JULES5_{ALL} (compared to $127 \text{ Pg C year}^{-1}$ for JULES9_{ALL}), primarily because V_{cmax} for the generic broadleaf tree was much higher than for the tropical broadleaf evergreen PFT, based on the data from K09. Although tropical GPP was higher in JULES5_{ALL} compared to JULES9_{ALL}, the NPP in tropical forests was lower and closer to the values from MODIS NPP. The reason was the differences in leaf nitrogen, which increased respiratory costs in JULES5_{ALL} compared to JULES9_{ALL}. Both N_A and N_m were higher for the broadleaf tree PFT than for the tropical evergreen broadleaf tree PFT.

Over the tropical savannah biome, the GPP decreased in JULES9_{ALL} compared to JULES5 due to lower productivity from C₄ grasses, and GPP was within the uncertainty range of the MTE GPP, although slightly higher. The overestimation occurred during most of the year (Fig. 7b), except during the late dry season/early wet season (October–December). Although C₄ grasses had a lower NPP in JULES9_{ALL}, a significant fraction of the biome is composed of C₃ grass, BDT, ESh, and DSh in the ESA data, which all had higher NPP in JULES9_{ALL}. For this reason, biome-scale NPP was higher in JULES9_{ALL} than in JULES5, and simulated NPP was $140 \text{ g C m}^{-2} \text{ year}^{-1}$ higher than the MOD17 value. In the temperate grasslands biome, both GPP and NPP were higher in JULES9_{ALL} compared to JULES5, and closer to the MTE and MOD17 values. However, compared to the MTE, the JULES9 GPP increased 1 month early, it was too low in the mid-summer, and it declined too slowly in the autumn.

The biome-scale GPP in the extratropical mixed forests improved in JULES9_{ALL} compared to JULES5, and was very close to the MTE estimate. The simulated GPP was overestimated during the autumn (September–October) and underestimated during the winter. Simulated NPP was very close to the MOD17 NPP in JULES5, but it is too high by $\sim 100 \text{ g C m}^{-2} \text{ year}^{-1}$ in JULES9_{ALL}. The predominant vegetation types in the “boreal and coniferous forests” biome are NET (26% coverage), C₃ grass (20%), and NDT (14%). Shrubs, deciduous broadleaf trees, and bare soil cover the remaining 40% of the biome. There was a large increase in summertime GPP in this biome, bringing JULES9_{ALL} closer to the MTE GPP than JULES5. The NPP increased in JULES9, compared to JULES5, and was within $10 \text{ g C m}^{-2} \text{ year}^{-1}$ of the MOD17 NPP.

Deserts/shrublands and tundra are both dry environments with annual-average GPP of $\sim 280 \text{ g C m}^{-2} \text{ year}^{-1}$ according to the MTE data set. Although GPP increased in both biomes in JULES9_{ALL} relative to JULES5, it was much lower than the MTE value. In the tundra biome, GPP was underestimated during the entire growing season, and it was underestimated all year in the desert biome. The simulated NPP was also significantly lower than MOD17 in these two biomes, although it was slightly improved in JULES9_{ALL}. These results indicate that the JULES plants struggle in extremely cold and arid environments.

In the Mediterranean woodlands, GPP increased by $90 \text{ g C m}^{-2} \text{ year}^{-1}$ and NPP increased by $80 \text{ g C m}^{-2} \text{ year}^{-1}$ in JULES9_{ALL} compared to JULES5, but both fluxes were still $\sim 100 \text{ g C m}^{-2} \text{ year}^{-1}$ lower than the MTE GPP and MOD17 NPP. The simulated GPP (in JULES9_{ALL}) was close to the MTE value during most of the year except the dry season, when it declined more in the model than in the MTE estimate.

On a global scale, JULES9_{ALL} had a similar GPP but higher NPP compared to JULES5 (Fig. 8). In both simulations, the global GPP was $128\text{--}129 \text{ Pg C year}^{-1}$ (average from 2000–2012), compared to the MTE average of

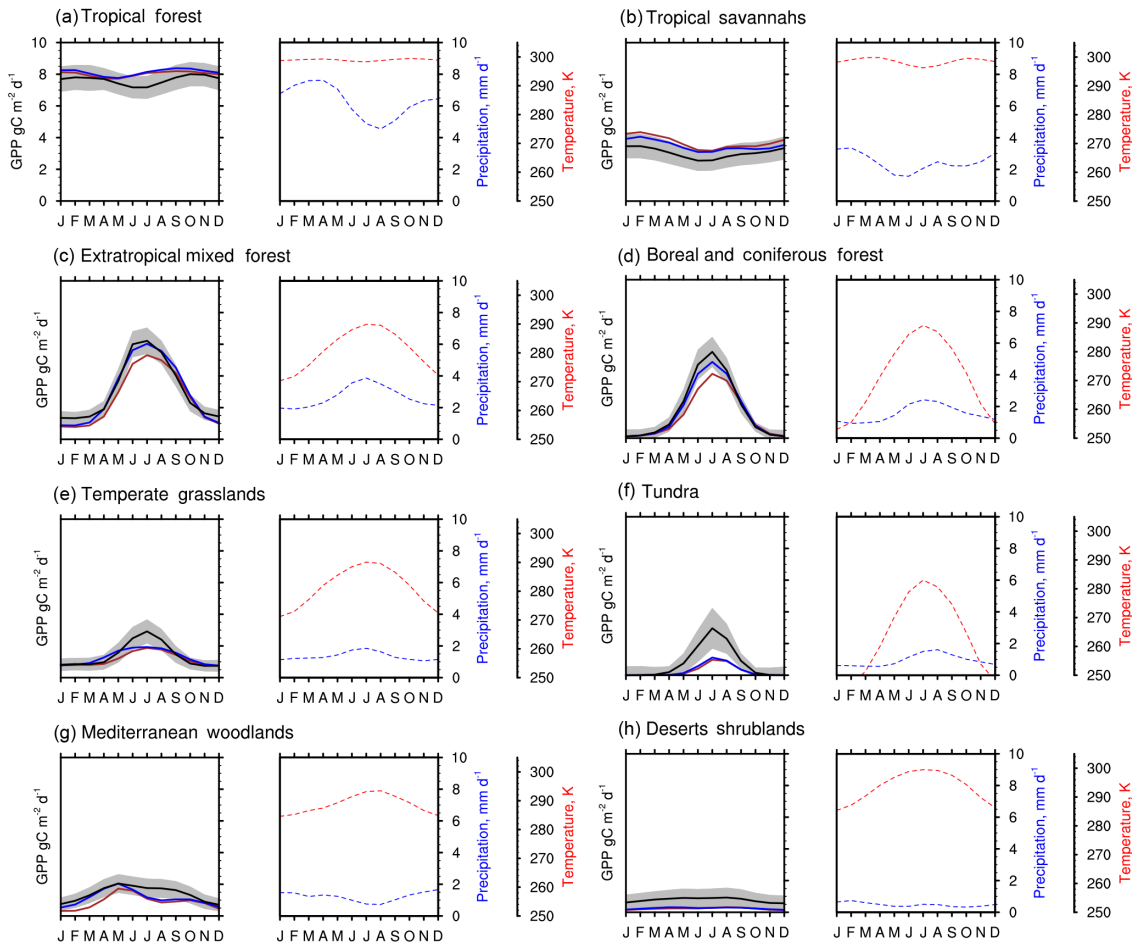


Figure 7. Area-averaged seasonal cycles of GPP from the biomes shown in Fig. 3b, comparing JULES5, JULES9, and the Jung et al. (2011) MTE. Also shown are the temperature and precipitation from the CRU-NCEP data set used to force the JULES simulations. The gray shading in the GPP plots shows the MTE GPP ± 1 standard deviation based on the area-averaged standard deviations of monthly fluxes for each grid cell.

$122 \pm 8 \text{ Pg C year}^{-1}$. GPP was higher in JULES9_{ALL} compared to JULES5 in the core of the tropical forests, but lower in tropical/subtropical South America, Africa, and Asia. These are regions with significant grass coverage (Fig. 3a), especially C₄ grasses. Poleward of 30°, GPP was higher in JULES9_{ALL} due to higher productivity in trees. In JULES5, the global NPP ($55 \text{ Pg C year}^{-1}$) was close to the value from MODIS NPP ($54 \text{ Pg C year}^{-1}$). In JULES9_{ALL}, the NPP was higher than JULES5 almost everywhere (except for southern Brazil where C₄ grasses are dominant), and the global NPP was $62 \text{ Pg C year}^{-1}$.

5 Discussion

5.1 Impacts of trait-based parameters and new PFTs

Including trait-based data on leaf N, $V_{\text{cmax},25}$, and leaf life span improved the seasonal cycle of GPP at seven sites, es-

pecially sites with C₃ grass and NDT. Parameterizing leaf life span correctly has been shown to be important, even within biomes (Reich et al., 2014). Our study confirms this, as the simulation of GPP improved at fewer sites in the simulations without the improved leaf life span. However, compared to the standard five PFTs, the RMSE of GPP was only improved at four sites in JULES9_{TRY}. Despite this, the new PFTs with the new trait data include observed trade-offs between leaf structure and life span. These trade-offs are important for enabling JULES to represent observed vegetation distribution and for predictions of future fluxes.

Incorporating more data and accounting for evergreen and deciduous habits further improved the model, as indicated by the closer model-data comparison obtained with JULES9_{ALL} at both the site and global level. The distinction between the tropical and temperate broadleaf evergreen trees provided mixed results. While there was no improvement in the seasonal cycles at the two tropical forest sites, both GPP and the

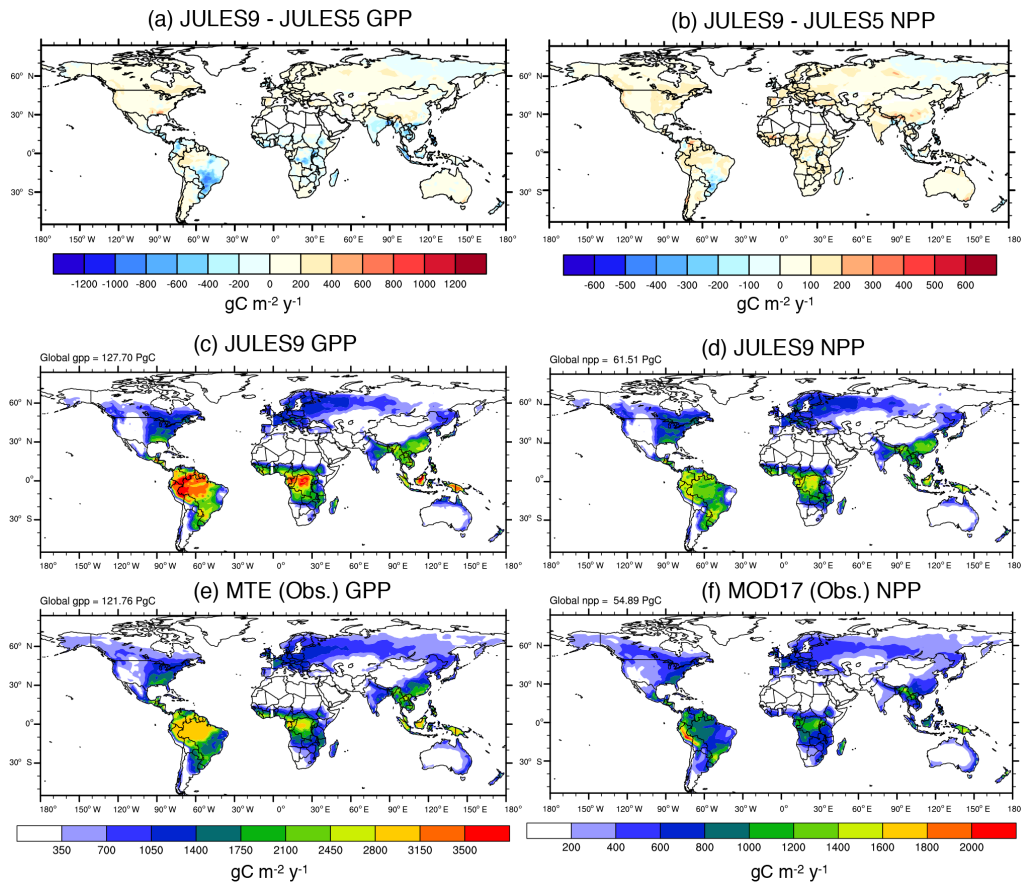


Figure 8. Global maps of carbon cycle fluxes from 2000 to 2012. The observation sources are MTE (GPP) and MODIS MOD17 (NPP, 2000–2013).

seasonal cycle of NEE were improved at the warm-temperate evergreen savannah site (Las Majadas). This study has laid the groundwork for further improvements to JULES GPP and plant respiration by incorporating trait-based physiological relationships and allowing for a flexible number of PFTs. Future development can focus on more biome-specific data-model mismatches than was possible with the generic set of five PFTs.

The nine PFTs were chosen as they represent the range of deciduous and evergreen plant types with minimal externally determined bioclimatic limits. The distinction between tropical and temperate broadleaf evergreen trees account for the important differences between these types of trees (e.g., a lower V_{cmax} for a given N_A in tropical broadleaf evergreen trees: Kattge et al., 2009). The comparison of JULES5_{ALL} and JULES9_{ALL} indicates that even using improved parameters with five PFTs based on the TRY data and the literature reviewed in this study will give improved productivity fluxes in JULES. However, an important caveat is JULES was not run with dynamic vegetation for this analysis. The additional PFTs enable more diverse and specific dynamic responses to climate change.

5.2 Future development priorities

The biome-level evaluation of GPP and NPP provides insight into potential areas for improvement in JULES: in particular boreal forests, tundra, Mediterranean woodlands and desert/xeric shrublands (Fig. S6). GPP was systematically underestimated in regions experiencing seasonal soil moisture stress, such as the tropical forests, summer at Morgan Monroe, and the dry season at El Saler. A similar result was seen with the arid biomes and in the Mediterranean biome during summer. The fact that the model did not match the seasonal cycle of GPP at the two tropical forest sites with improved parameters indicates that processes such as the representation of plant water access and/or soil hydraulic properties need to be addressed in JULES. However, the dry season bias was not present when JULES was compared to the biome-scale MTE GPP. This underscores the complexity of modeling tropical forest productivity and the need to evaluate multiple data sources. High latitude grasses were underproductive, which also contributed to an underestimation of soil carbon (not shown). Further development of a tundra-specific PFT(s) could improve the carbon cycle in these regions.

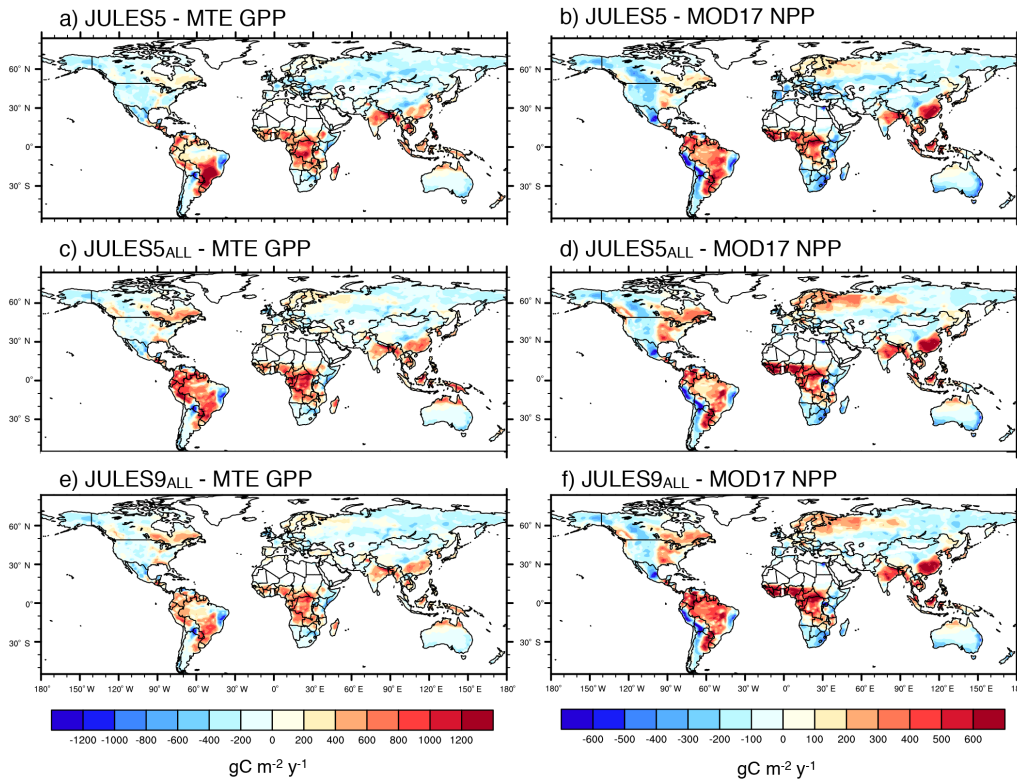


Figure 9. Differences between modeled and observed GPP (observed – MTE) and NPP (observed – MOD17). (a, b) JULES with the standard five PFTs and default parameters; (c, d) JULES with five PFTs and improved parameters; (e, f) JULES with nine PFTs and improved parameters.

A side effect of the trait-based parameters was increased respiration, and comparison to both FLUXNET sites and the MTE suggest it is now too high for most biomes. Total ecosystem respiration was higher than that observed at Manaus, Harvard, Morgan Monroe, Tharandt, Hyytiälä, Kaamanen, Las Majadas, and Tonzi (75 % of the sites with respiration data) (Fig. S3). As this study has focused primarily on improving the GPP, the next step should be to include a more mechanistic representation of growth and maintenance respiration in JULES to improve the net productivity (e.g., using data from Atkin et al., 2015). Comparison to the MTE respiration also suggests that JULES soil respiration is too high during the winter in the temperate and boreal biomes. In the latter, both versions of JULES predicted positive respiration flux during the winter, while the MTE product showed negligible fluxes (Fig. S5). The average winter temperatures in the biome were $< -13^{\circ}\text{C}$, yet soil respiration continued during these months because the Q_{10} soil respiration scheme has a very slow decay of soil respiration flux at sub-zero temperatures (see Fig. 2 of Clark et al., 2011). A similar result was seen at Hyytiälä (Fig. S3b), which further indicates that wintertime respiration might be too high.

Last, the simulation of GPP could be further improved by replacing the static $V_{\text{cmax},25}$ per PFT. Simultaneous with this

study, there is work to include temperature acclimation for photosynthesis JULES, which is more realistic than a set T_{opt} for each PFT. Also, the data exhibit large within-PFT variation in $V_{\text{cmax},25}$ (Fig. 2) and photosynthetic capacity can depend on the time of year. Recent work relating photosynthetic capacity to climate variables, environmental factors, and soil conditions shows promise for better capturing the dynamic nature of this parameter (e.g., Verheijen et al., 2013; Ali et al., 2015; Maire et al., 2015).

6 Conclusions

We evaluated the impacts on GPP, NEE, and NPP of new plant functional types in JULES. All changes were evaluated in version 4.2 with the canopy radiation model 5 option (Clark et al., 2011). At the base of the new PFTs was inclusion of new data from the TRY database. N_m and LMA replaced the parameters N_{10} and σ_1 . These were used to calculate new $V_{\text{cmax},25}$, which was higher for all of the new PFTs compared to the original five, except for C_3 grasses. The higher $V_{\text{cmax},25}$ resulted in higher GPP. The GPP did not increase for C_4 grasses due to a lower quantum efficiency, or for cold grasslands due to a lower optimal temperature for

V_{cmax} . Increases in NPP generally followed on from the increases in GPP.

A trade-off between LMA and leaf life span was enforced by changing parameters relating to leaf phenology, growth and senescence. The new parameter values changed the turnover rate of leaves on trees in the spring and fall, therefore altering the leaf life span in JULES in a manner consistent with observations. In JULES9_{TRY}, the median leaf life span of grasses and shrubs were reduced, which improved the seasonal cycle at the relevant sites (Las Majadas, Tonzi, Fort Peck, Kaamanen, and Tomakai). The exception was the Bondville crop site.

Including the full range of updated parameters (in JULES9_{ALL}) resulted in an improved seasonal cycle of GPP at 10 sites and reductions to daily RMSE at 11 sites (out of 13 sites with GPP data) compared to JULES9_{TRY}. The annual GPP was within the range of the FLUXNET observations at every site except for one (Hyytiälä). On a biome scale, we compared GPP to the MTE product of Jung et al. (2011) and NPP to the MODIS17 product. GPP was improved in JULES9 for all eight biomes evaluated, although for the tundra and desert/shrubland biome the GPP was much lower than the MTE value. The global NPP was slightly higher than that observed, but JULES9 was closer to MOD17 in most biomes – the exceptions being the tropical forests, savannahs, and extratropical mixed forests where JULES9 was too high. The biome-averaged NPP from JULES9 was within the range of MOD17 NPP for all biomes.

Overall, the simulation of gross and net productivity was improved with the nine PFTs. The present study can be thought of as a “bottom-up” approach to improving JULES fluxes, with new parameters being based on large observationally based data sets. The next step for improving PFTs in JULES is to evaluate the nine PFTs when the dynamic vegetation is turned on. This will be addressed in a follow-up paper. A complimentary, “top-down” method for reducing uncertainty in JULES is to optimize PFT parameters based on minimizing errors between simulated and observed fluxes. This is currently being done with adJULES, an adjoint version of JULES (Raoult et al., 2016). Future model development within JULES will have more flexibility for improving the model with more PFTs, and the improvements presented in this study increase our confidence in using JULES in carbon cycle studies.

7 Code availability

The simulations discussed in this manuscript were done using JULES version 4.2. This can be accessed through the JULES FCM repository: <https://code.metoffice.gov.uk/trac/jules> (registration required). For further details, see <https://code.metoffice.gov.uk/trac/jules/wiki/9PFTs>. An example with the nine PFTs and parameters in this paper is provided for Loobos in the documentation directory of the JULES trunk. Summary tables of the traits LMA, N_m , and leaf life span are included in the Supplement.

Appendix A

Table A1. List of parameters and symbols in the text.

Symbol	Units	Equation	Description	Default value*
A_l	$\text{kg C m}^{-2} \text{s}^{-1}$	5	Leaf-level photosynthesis	
a_{wl}	kg C m^{-2}	24	Allometric coefficient	
a_{ws}	–	24	Ratio of total to respiring stem carbon	
b_{wl}	–	24	Allometric exponent	1.667
C_i	Pa	6	Internal leaf CO_2 concentration	
C_{mass}	$\text{kg C} [\text{kg biomass}]^{-1}$	23	Leaf carbon concentration per unit mass	0.5 for this study
C_s	Pa	6	Leaf surface CO_2 concentration	
D_{crit}	kg kg^{-1}	7	Critical humidity deficit	
d_T	–	16	Rate of change of leaf turnover with temperature	
f_0	–	7	Stomatal conductance parameter	
f_d	–	4	Leaf dark respiration coefficient	
g_s	m s^{-1}	6	Leaf-level stomatal conductance	
i_v	$\mu\text{mol CO}_2 \text{m}^{-2} \text{s}^{-1}$	19	Intercept for relationship between N_A and $V_{\text{cmax},25}$	
k_n	–	3, 20	Extinction coefficient for nitrogen	0.78
h	m	13, 23, 24	Canopy height	
L_{bal}	$\text{m}^2 \text{m}^{-2}$	12, 13, 22–24	Balanced leaf area index (maximum LAI given the plant's height)	
L_{max}	$\text{m}^2 \text{m}^{-2}$		Maximum LAI	
L_{min}	$\text{m}^2 \text{m}^{-2}$		Minimum LAI	
LMA	kg m^{-2}	18, 21, 22	Leaf mass per unit area (new parameter)	
N_a	kg N m^{-2}	18	Leaf nitrogen per unit area	
n_{eff}	$\text{mol CO}_2 \text{m}^{-2} \text{s}^{-1} \text{kg C} [\text{kg N}]^{-1}$	3	Constant relating leaf nitrogen to Rubisco carboxylation capacity	
N_{l0}	$\text{kg N} [\text{kg C}]^{-1}$	3	Top-leaf nitrogen concentration (old parameter, mass basis)	
N_m	kg N kg^{-1}	18, 21–23	Top-leaf nitrogen concentration (new parameter)	
N_l	kg N m^{-2}	11, 21	Total leaf nitrogen concentration	
N_r	kg N m^{-2}	12, 22	Total root nitrogen concentration	
N_s	kg N m^{-2}	13, 23	Total stem nitrogen concentration	
p	–	17	Phenological state ($\text{LAI}/L_{\text{bal}}$)	
$Q_{10,\text{leaf}}$	–	2	Constant for exponential term in temperature function of V_{cmax}	2
R_a	$\text{kg C m}^{-2} \text{s}^{-1}$	8	Total plant autotrophic respiration	
R_d	$\text{kg C m}^{-2} \text{s}^{-1}$	4, 5	Leaf dark respiration	
r_g	–	10	Growth respiration coefficient	0.25
rootd	m		e -folding root depth	
s_v	$\mu\text{mol CO}_2 \text{g N}^{-1} \text{s}^{-1}$	19	Slope between N_A and $V_{\text{cmax},25}$	
T_{low}	$^{\circ}\text{C}$	1	Upper temperature parameter for V_{cmax}	
T_{off}	$^{\circ}\text{C}$	16	Threshold temperature for phenology	
T_{opt}^b	$^{\circ}\text{C}$		Optimal temperature for V_{cmax}	
T_{upp}	$^{\circ}\text{C}$	1	Upper temperature parameter for V_{cmax}	
$V_{\text{cmax},25}$	$\mu\text{mol m}^{-2} \text{s}^{-1}$	1, 9	The maximum rate of carboxylation of Rubisco at 25°C	
W	$\text{kg C m}^{-2} \text{s}^{-1}$	5	Smoothed minimum of the potential limiting rates of photosynthesis	
α	$\text{mol CO}_2 [\text{mol PAR photons}]^{-1}$		Quantum efficiency	
β	–	5	Soil moisture stress factor	
α^*	Pa	7	CO_2 compensation point	
γ_0	$[\text{360 days}]^{-1}$	16	Minimum leaf turnover rate	
γ_{lm}	$[\text{360 days}]^{-1}$	16	Leaf turnover rate	
γ_p	$[\text{360 days}]^{-1}$	17	Leaf growth rate	20
μ_{rl}	–	12, 22	Ratio of nitrogen concentration in roots and leaves	
μ_{sl}	–	13, 23	Ratio of nitrogen concentration in stems and leaves	
η_{sl}	$\text{kg C m}^{-2} \text{LAI}^{-1}$	13, 23	Live stemwood coefficient	0.01
σ_L	$\text{kg C m}^{-2} \text{LAI}^{-1}$	11, 12	Specific leaf density (old parameter)	

* Default values only provided for non-PFT-dependent parameters.

Table A2. New trait-based parameters for five PFTs that are consistent with the data used in this study. Used in the JULES_{5ALL} experiments.

	BT	NT	C3	C4	SH
N_m	0.0185	0.0117	0.0240	0.0113	0.0175
LMA	0.1012	0.2240	0.0495	0.1370	0.1023
s_v	25.48	18.15	40.96	20.48	23.15
i_v	6.12	6.32	6.42	0.00	14.71
$V_{cmax,25}$	53.84	53.88	55.08	31.71	56.15
T_{off}	5	-40	5	5	-40
d_T	9	9	0	0	9
γ_0	0.25	0.25	3.0	3.0	0.66
γ_p	20	15	20	20	15
L_{min}	1	1	1	1	1
L_{max}	9	7	3	3	4
D_{crit}	0.09	0.06	0.051	0.075	0.037
f_0	0.875	0.875	0.931	0.800	0.950
f_d	0.010	0.015	0.019	0.019	0.015
rootd	3	2	0.5	0.5	1
T_{low}	5	0	10	13	0
T_{opt}	39	32	28	41	32
T_{upp}	43	36	32	45	36
α	0.08	0.08	0.06	0.04	0.08
μ_{rl}	0.67	0.67	0.72	0.72	0.67

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