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Effect of the addition of different waste carbonaceous materials 1 on coal gasification in CO₂ atmosphere 2 Ashak M. Parvez^a, Iqbal M. Mujtaba^b, Chengheng Pang^a, Tao Wu^{a,*} 3 ^a Department of Chemical and Environmental Engineering, The University of Nottingham Ningbo China, Ningbo 315100, China 4 b Chemical Engineering Division, School of Engineering , University of Bradford, Bradford BD7 1DP, UK 5 6 Corresponding author: Tao.Wu@nottingham.edu.cn 7 Abstract 8 In order to assess the utilisation of waste carbonaceous materials and the feasibility of using 9 CO₂ as a gasifying agent, the gasification characteristics of coal, a suite of waste carbonaceous materials, and their blends were investigated by using a thermogravimetric 10 analyser (TGA). The results showed that the CO₂ gasification process of polystyrene 11 12 completed at 470 °C which was lower than other carbonaceous materials. This behaviour was attributed due to the high content of volatile coupled with its unique thermal degradation 13 properties. It is found that the initial decomposition temperature decreased with the increasing 14 amount of waste carbonaceous materials in the blends. Overall, CO₂ co-gasification process 15 16 was enhanced as a direct consequence of interactions whose intensity and temperature of occurrence were influenced by the chemical and compositional properties of carbonaceous 17 materials. The strongest interactions were noticed in coal/polystyrene blend at the 18 19 devolatilisation stage as indicated by its highest value of Root Mean Square Interaction Index 20 (RMSII), caused by highly reactive characteristic of polystyrene. On the other hand, coal/oat straw blend revealed the highest interactions at char gasification stage. The catalytic effect of 21 22 alkali metals content in oat straw, particularly CaO and K₂O, was believed to be the reasons

for these strong interactions. Furthermore, co-pyrolysis was compared with CO_2 gasification through RMSII value, while the effect of CO_2 as a gasifying agent on interactions during cogasification was also evaluated. It is clear from this study that CO_2 gasification of coal can be enhanced significantly via the addition of polystyrene and oat straw.

27 Keywords: CO₂ gasification; Carbonaceous materials; Interactions; Catalytic effect

28 **1** Introduction

In order to minimise CO_2 emissions, different waste carbonaceous materials are used as energy source because of their carbon neutral nature. In recent years, co-utilisation of waste carbonaceous materials has become increasingly popular to replace a portion of coal in existing coal-fired boilers for power generation [1, 2].

Over the years, many techniques have been developed to utilise coal with biomass and other carbonaceous materials. Co-gasification is one of the promising methods because it has the potential to improve the gas yield and corresponding heating value of the product when low quality coal is gasified [3, 4]. Moreover, improved overall carbon conversion and cold gas efficiency can also be achieved by increasing the proportion of biomass in the feed.

For gasification process, CO_2 is a potential gasification medium [5-7]. It is less corrosive than steam and can be used to adjust syngas composition for different applications [8]. In the past decades, a lot of research has been carried out on biomass gasification using CO_2 as a

41	gasifying agent [5, 7, 9-12]. These researches were mainly focused on the kinetics [5, 9, 11,
42	12], gasification reactivity [7, 10, 13, 14] and gasification characteristics [14-16] in general.
43	In spite of huge potential of CO ₂ in coal and biomass gasification, limited effort have been
44	made to understand gasification behaviours, particularly on interactions during co-gasification.
45	Apart from these, limited study has been conducted understanding how to enhance co-
46	gasification via choosing suitable waste carbonaceous materials for the co-gasification of
47	certain coal. In addition, information on waste carbonaceous materials such as non-metallic
48	part of printed circuit boards, tyre scraps and polystyrene are very limited compared with the
49	conventional fuel and biomass. Moreover, not much work has been performed to understand
50	the effects of CO_2 on the gasification process of these carbonaceous materials. Thus, there are
51	still needs to achieve in depth understanding of gasification characteristics of coal, waste
52	carbonaceous materials and their blends.
53	In this paper, gasification characteristics of coal, a suite of waste carbonaceous materials and
54	their blends were investigated in a TGA using CO_2 as the only gasifying agent. In addition,
55	the influence of mass fractions of carbonaceous materials on gasification characteristics of the

56 blends was analysed. The interactions during co-gasification were also studied and the reasons57 for these interactions were explored.

58 2 Experimental

59 2.1 Samples preparation

60 In this study, an Australian bituminous coal was selected as the coal sample. Oat straw (OS), 61 Non-metallic part of printed circuit boards (NMPCBs), tyre scraps (tyre) and polystyrene (PS) 62 were chosen as the waste carbonaceous materials samples. All samples were air-dried at 63 ambient temperature prior to further processing. Size reduction was carried out by using an 64 industrial hammer crusher (CSF570, Fengli Pulverization Ltd., China) and ball mill (SM2000, MM400, Retsch, Germany), respectively. All samples were ground to a size smaller than 106 65 microns (Endecott 106 microns aperture sieve) following the standard milling procedure 66 (CEN/TS 15443:2006). Before characterization and testing, all samples were treated 67 following international accepted standard procedure to ensure representativeness of the 68 69 samples [17]. The blend samples were prepared from these materials by mixing manually at 70 (coal to waste carbonaceous material) mass ratios of 90:10 and 70:30 (with a maximum 71 deviation of ± 0.1 mg), which is denoted as high and low bending ratios, respectively. These ratios are representative of typical co-gasification parameters, corresponding to 5%-35% on 72 thermal basis. 73

74 2.2 Sample characterisation

Proximate analysis of the samples was conducted based on the standard practice for
Proximate Analysis using a TGA following procedures described elsewhere [18]. All the

77 experiments were repeated three times and average value was used for the study. Carbon (C), hydrogen (H), nitrogen (N) and sulphur (S) contents were determined using a CHNS/O 78 79 Element Analyser (PE2400, PerkinElmer, USA) following the standard testing procedure described in the manual. The standard sample was firstly applied to calibrate the Elemental 80 81 Analyser. Approximately 1.5 mg of dried fine powder was used for each test, which was 82 repeated at least three times. Oxygen content was calculated by difference on dry and ash free 83 basis. The tests were carried out in helium carrier gas with an accuracy of $\leq 0.3\%$ and a 84 precision of $\leq 0.2\%$. Calorific value of all the samples was determined using a calorimeter 85 (IKA C 200, Germany), which was calibrated prior to testing using benzoic acid to achieve a 86 relative standard deviation less than 0.2%. As for ash composition, each sample was analysed 87 twice using an X-Ray Diffraction (XRD) (Bruker D8 Advance, Germany) to minimize 88 experimental error to be within $\pm 1.5\%$.

89 **2.3 CO₂ gasification**

 CO_2 gasification of the coal, waste carbonaceous materials and their blends was carried out using a STA 449/F3 Jupiter thermal analyser (NETZSCH Geraetebau GmbH, Germany) with a weighing precision of $\pm 0.01\%$ and TGA resolution of 0.1 µg. In each test, the sample was heated to 50 °C in N₂ (30 ml/min) from room temperature and kept isothermal for 10 min. It was then heated from 50 to 1200 °C at a constant heating rate of 20 K/min in the presence of pure CO₂ (40 ml/min). Prior to each test, the sample was further manually grounded to 96 smaller than 90 microns to minimize diffusion effects. About 5 mg of sample (accuracy up to 97 0.01 mg via PerkinElmer AD 6 autobalance) was loaded into a ceramic crucible for each test. 98 This slow heating rate and small amount of sample used would minimise mass transfer effects 99 and eliminate the heat transfer limitations in the CO_2 gasification process. Each test was 100 repeated three times to minimize the experimental error.

101 **2.4 Interactions indexes**

102 Thermal characteristics of coal, waste carbonaceous materials and their blends in CO₂ 103 atmosphere were extracted from the TG/DTG profiles. The theoretical TG/DTG curves of the 104 blends were calculated by Eq. (1) based on the mass loss rates of each individual sample 105 assuming additive property applies. The deviation between experimental and calculated 106 curves was used to show the degree of interactions between samples [19, 20]:

107
$$\frac{dm}{dt} = x_{SP1} \left(\frac{dm}{dt}\right)_{SP1} + x_{SP2} \left(\frac{dm}{dt}\right)_{SP2}$$
(1)

108 where $\left(\frac{dm}{dt}\right)_{SP1}$ and $\left(\frac{dm}{dt}\right)_{SP2}$ are the mass loss rates (%/min) of individual samples while 109 x_{SP1} and x_{SP2} are the corresponding mass fractions in the blends, respectively. The Root 110 Mean Square Interactions Index (RMSII) was applied to quantify the interactions between 111 components in the blend which compares the deviation of calculated value with experimental 112 value. Normally, a higher value of RMSII indicates a larger interaction taking place in the 113 process. The RMSII can be calculated using the Eq. (2) [21].

114
$$RMSII = \sqrt{\frac{\sum_{i=1}^{N} \left(\frac{(dm/dt)_{EXP}^{i} - (dm/dt)_{CAL}^{i}}{(dm/dt)_{CAL}^{i}} \right)^{2}}{N}}$$
 (2)

115 where $(dm/dt)_{EXP}$ and $(dm/dt)_{CAL}$ denote experimental and calculated mass loss rates values, 116 respectively. *N* represents the number of points undertaken.

117 **3 Results and discussion**

118 **3.1** Properties of the samples

119 Table 1 shows proximate, ultimate and ash analysis results of the samples studied in present 120 work. For the proximate analysis, the relative standard deviation was ± 3 wt%, while for 121 ultimate analysis, it was ± 2 wt%. As can be seen, volatile content and oxygen content varied 122 significantly. Among all the tested samples, NMPCBs had the highest percentage of ash, 123 which is mainly due to the presence of glass fiber. Hence, it would have some influence on 124 gasification process. It is clear that tyre showed very similar properties in term of proximate 125 and ultimate analysis data as coal. Its heating value is also very close to that of coal. Therefore, it is expected that the co-gasification of tyre with coal would have very little influence on the 126 127 overall gasification process. On the other hand, negligible amount of ash was detected in 128 polystyrene, which was completely decomposed, resulting in no fixed carbon formed after 129 pyrolysis. Besides, polystyrene contained no sulphur and low percentage of nitrogen 130 compared with the other samples.

131	Higher heating value (HHV) of individual samples were also included in Table 1 with relative
132	standard deviation controlled within $\pm 4.5\%$. It is clear that polystyrene (41.7 MJ/kg, daf)
133	exhibited the highest heating value, greater than that of coal (35.5 MJ/kg, daf). Meanwhile,
134	the heating values of tyre (33.2 MJ/kg, daf) and NMPCBs (31.5 MJ/kg, daf) were all slightly
135	lower than coal. Hence, blend with polystyrene, tyre or NMPCBs might not have a negative
136	impact on co-gasification process. In contrast, oat straw (19.5 MJ/kg, daf) had the lowest
137	heating value, which subsequently would result in a lower gasification temperature. As
138	moderate low gasification temperature is one of the desired properties of the fuel, thus, it
139	would enhance gasification performance in co-gasification process.

1	4	0
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Table 1: Proximate, ultimate analyses and calorific values of the samples.

	AC	OS	NMPCBs	Tyre	PS
HHV (MJ/kg dry ash	35.5	19.5	31.5	33.2	41.7
free)			0110		
	Prox	imate analysis (wt %)		
Moisture	0.7	4.2	0.8	0.4	0.9
Volatile matter	34.6	70.7	25.1	62.9	99.1
Fixed carbon	48.2	18.2	6.9	19.6	-
Ash	16.5	6.9	67.2	17.1	-
	Ultin	nate analysis ^{a,b} (wt %	%)		
С	81.3	37.4	39.8	80.9	92.1
Н	4.9	5.3	3.6	6.4	7.7
O^C	9.7	47.9	55.7	10.7	-
Ν	1.9	2.3	0.6	0.5	0.2
S	2.2	0.2	0.3	1.5	-
Ash composition (wt %, standard deviation $\pm 5\%$)					

SiO ₂	42.6	43.3	29.7	38.5	-
SO ₃	1.4	3.4	-	6.2	-
CaO	4.2	10.6	5.3	3.5	-
Na ₂ O	2.3	2.6	1.4	2.7	-
Fe ₂ O ₃	7.4	0.9	4.6	45.1	-
MgO	2.2	4.2	-	0.7	-
Al_2O_3	39.1	2.2	5.3	2.4	-
K ₂ O	0.8	32.8	-	0.9	-
CuO	-	-	38.4	-	-
ZnO	-	_	5.7	-	-
Ni ₂ O ₃	_	-	9.6	-	-

^{*a*}Dry basis. ^{*b*}Ash free basis. ^{*c*}By difference.

141 **3.2 Thermal behaviours of the individual samples**

Fig. 1 shows the TGA and DTG curves of coal and different waste carbonaceous materials 142 under CO₂ gasification. In this study, the instrument was calibrated to ensure that the 143 144 precision of $\leq 0.2\%$ in TGA analysis was achieved. In Fig. 1, it is apparent that the gasification 145 processes involved a multi-stage thermal degradation where the number of stages and the unique features of individual stage varied. The main mass losses in coal gasification took 146 147 place at 400-570 °C and 935-1190 °C, respectively (Fig. 1a). The first mass loss was due to 148 the release of volatiles while char gasification and thermal decomposition of minerals 149 contributed to the second mass loss. As expected in CO₂ gasification, the heterogeneous reaction between coal char and CO₂ was much slower compared with combustion process. 150 151 This is because CO₂ is a weak oxidizing agent and the CO₂ gasification via Boudouard 152 reaction (C (S) + CO₂ \rightarrow 2CO) is slow [22, 23]. Normally, at atmospheric pressure, the

153	Boudouard reaction is thermodynamically favourable at temperatures above 900 °C (initiated
154	at 800 °C) while the combustion of char can occur at significantly lower temperatures (below
155	630 °C). Two stages of mass losses were observed in oat straw as seen in Fig. 1b. The first
156	loss (240-375 °C) was attributed to the complete decomposition of hemi-cellulose and
157	cellulose and the partial decomposition of lignin [20, 24, 25]. The decomposition of
158	remaining lignin, char gasification and minerals decomposition contributed to the second
159	mass loss (820-950 °C). Unlike coal, CO ₂ gasification of char derived from oat straw took
160	place at a lower temperature range. In general, biomass has a higher amount of volatiles
161	which generate a highly porous char compared with char derived from coal [26]. The porous
162	structure of char facilitates CO ₂ diffusion, which leads to a better gasification performance.
163	The gasification of oat straw therefore completed at lower temperatures. It was reported that
164	the reactivity of biomass char is usually higher than the chars produced from coal [27, 28].
165	This implies that compared with coal char, biomass char can be gasified at much lower
166	temperature which is the case for oat straw.

167 Fig. 1c shows the main mass loss experienced by NMPCBs in the second stage which is
168 illustrated at temperature from 295 to 410 °C. In spite of NMPCBs having the lowest
169 volatiles, its maximum mass loss was occurred at low temperature region.



171 Figure 1: CO₂ gasification distribution of the TGA and DTG curves of (a) coal, (b) oat straw, (c)
172 NMPCBs, (d) tyre and (e) polystyrene.

173 Typically, epoxy resin is the major component of the NMPCBs which is mixed with 174 brominated compounds [29, 30]. The decomposition of epoxy resin, degradable part of 175 NMPCBs, started at around 300 °C. Consequently, thermal decomposition of NMPCBs

176	consists of two stages, the decomposition of brominated compounds and decomposition of
177	non-brominated compounds. The possible reason for the existence of small peak before the
178	main peak was the partial overlapping of these two reactions [30]. As seen in Fig. 1c, a
179	smaller DTG peak which was situated between the temperatures of 465 to 510 °C, denoted the
180	second main mass loss due to the release of volatiles due to the decomposition of brominated
181	compounds in NMPCBs. It is important to notice that the value of this mass loss rate was
182	smaller (0.85 wt%/min) in comparison with mass loss rates noticed in other samples. As seen
183	in Fig. 1d, tyre also exhibited the first mass loss at low temperature region (300-490 °C)
184	similar to NMPCBs while the second mass loss happened at relatively high temperature
185	region (875-1110 °C). Generally, tyre contains small portion of additives (less than 30 wt%)
186	and the rest is rubber (natural and synthetic) [31, 32]. The first mass loss was caused by the
187	decomposition of the additives and both types of rubber which occurred at lower temperature.
188	Afterwards, the second mass loss depicted CO ₂ char gasification. Higher percentage of carbon
189	in tyre might be the reason for this slow gasification reaction. On the other hand, polystyrene
190	experienced a single stage degradation at a higher temperature region (350-455 °C) compared
191	with other waste carbonaceous materials. From Fig. 1e, it can be seen that polystyrene had a
192	negligible mass loss up to 250 °C, whereas a slight mass loss was detected at temperature
193	between 250 to 350 °C. This confirmed that polystyrene was thermally stable up to 350 °C.
194	Degradation completed at 455 °C and during this stage the mass loss observed was 98.5 wt%
195	due to the thermal cracking of light organics and other carbonaceous heavy organics to low

196 molecular weight compounds. The overall mechanism is believed to be dictated by the 197 combination of end-chain initiation, depolymerisation, intramolecular hydrogen transfer, and 198 bimolecular termination processes. The changes in mass loss are caused by inter-molecular 199 transfer reactions while devolatilisation is governed by intramolecular transfer reactions. 200 Above 455 °C, no residue was observed, which confirmed the complete degradation of polystyrene in CO₂ gasification. In general, gasification characteristics such as initial 201 202 decomposition temperature and final mass loss temperature are the crucial parameters of the 203 fuel performance. As can be seen in Fig. 1, oat straw had the lowest initial decomposition 204 temperature (281 °C). In contrast, coal and polystyrene experienced the two highest initial 205 decomposition temperatures (415 and 407 °C). However, the final maximum mass loss temperatures was in the order of polystyrene (428 °C), NMPCBs (479 °C) and oat straw (931 206 207 °C), respectively which suggested that polystyrene and NMPCBs were more reactive 208 compared with oat straw. Moreover, the highest maximum mass loss rate (49.7 wt%/min) and the lowest gasification time (18.9 min) were observed for polystyrene. Therefore, the use of 209 210 polystyrene as a fuel for co-gasification application would be more beneficial.

3.3 Co-gasification characteristics of coal and carbonaceous materials

Fig. 2 describes the co-gasification behaviours of coal and carbonaceous materials and thecharacteristics of individual samples and blends are listed in Table 2.



Figure 2: CO₂ gasification distributions of (a) coal/oat straw, (b) coal/NMPCBs, (c) coal/tyre and (d)
 coal/polystyrene blends (standard deviation ± 1.3%).

In general, curves of the blends were found in the location between those of the original samples. From Fig. 2a, it is clear that coal/oat straw blends had three-stage mass loss rather than the two-stage mass loss identified for the original oat straw sample. In the case of blends, the first mass loss was mainly due to the volatiles release from decomposition of hemicellulose and cellulose in oat straw whereas the third one was mostly caused by char gasification. Commonly, the initial devolatilisation temperature of coal is approximately the same as the terminal devolatilisation temperature of biomass. Therefore, the second mass loss

224 was attributed mainly to the devolatilisation of coal, while the gasification of chars derived 225 from both oat straw and coal collectively resulted in the third mass loss. As shown in Table 2, 226 the coal (415 °C) started to devolatilise at a higher temperature than that of the oat straw (281 °C). In addition, the maximum mass loss temperatures of coal and oat straw were 458/1108 227 228 °C and 332/931 °C, respectively. It is interesting to see that the first and the second mass loss 229 temperatures of the blends did not vary significantly. However, the change of the third mass 230 loss temperatures was noticed. This indicated that the blending of oat straw with coal did not 231 have significant influence on the devolatilisation stage but affected the char gasification 232 process. Besides, as oat straw percentage in blend increased, the reduction in the initial 233 decomposition temperature (T_i) of the blends was also observed, which contributed to a lower 234 devolatilisation temperature in co-gasification process. The experimental results also 235 demonstrated that the addition of oat straw shifted the CO₂ gasification process to a low 236 temperature region by reducing the overall maximum mass loss temperatures (T_{max}) . The 237 higher volatiles content in oat straw (shown in Table 1) was the reason for these phenomena 238 as oat straw burns at a much lower temperature compared with coal. Besides, DTG value of the first mass loss increased (1.9 wt%/min to 5.2 wt%/min due to increment of blending 239 240 proportion from 10% to 30%) due to higher volatiles content while it decreased (6 wt%/min 241 to 5.9 wt%/min) in the third mass loss due to increasing amount of the char derived from coal.

Table 2: CO ₂	gasification characteri	stic parameters of bler	nds and individual samples.
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Sample	T _i	T_{\max} (°C)	t _{max} (min)	DTG _{max} (%/min)
Australian coal	415	458/1108	20.5/52.8	4.1/6.1
Blend 1 (10% oat straw)	301	335/460/1091	14.2/20.5/52.2	1.9/3.9/6
Blend 2 (30% oat straw)	294	333/456/1053	14.1/20.3/50.1	5.2/3.2/5.9
Oat straw	281	332/931	14.1/44.2	14.7/4.1
Blend 1 (10% NMPCBs)	407	456/1100	20.6/53.2	3.8/5.9
Blend 2 (30% NMPCBs)	353	344/453/1082	14.7/20.5/51.7	1.7/3.0/4.3
NMPCBs	305	364/479	15.7/21.5	4.5/0.9
Blend 1 (10% tyre)	400	457/1106	20.5/52.7	4.2/5.9
Blend 2 (30% tyre)	365	395/458/1088	17.3/20.3/51.8	3.9/4.2/5.3
Tyre	357	394/1046	17.2/50.1	10.4/3
Blend 1 (10% PS)	412	456/1104	20.2/52.5	7.3/5.6
Blend 2 (30% PS)	410	453/1082	20/51.8	15.1/4.5
Polystyrene (PS)	407	428	18.9	49.7

²⁴⁴ T_i =Initial decomposition temperature, T_{max} = Maximum mass loss temperature, t_{max} =maximum 245 reaction time and DTG_{max} = Maximum mass loss rate.

²⁴⁶ Uncertainty of temperature: ±1°C. Uncertaintty of DTG_{max}: ±0.15 %/min.

²⁴⁷ When NMPCBs was gasified with coal at the high blending ratio, the gasification process 248 featured three main mass loss stages as seen in Fig. 2b. The first mass loss stage occurred at 249 temperature between 330 and 380 °C where the decomposition of NMPCBs took place. The 250 second DTG_{max} (400-485 °C) indicated the release of volatiles from both coal and NMPCBs 251 whereas the third DTG_{max} depicted the stage of CO₂ gasification of NMPCBs and coal 252 derived chars. In contrast, at low blending ratios, such as 10 wt% NMPCBs, only two main 253 mass loss stages were noticed at similar temperature ranges as coal in co-gasification. Here, 254 the value of first DTG_{max} (Table 2) decreased because of the lower amount of volatiles in the

255	blend contributed by NMPCBs. Meanwhile, the decrement of the second DTG_{max} was as a
256	result of the lower amount of NMPCBs char. From Table 2, it is clear that the initial
257	devolatilisation temperature and the first DTG_{max} temperature of the blend were shifted to a
258	lower temperature region as the proportion of NMPCBs increased while the final DTG_{max}
259	temperature was also shifted to a lower temperature due to the reduced amount of coal derived
260	char. Therefore, it can be concluded that the co-gasification of coal and NMPCBs completed
261	at a relatively lower temperature despite of a slight reduction of the maximum mass loss rate.
262	For coal/tyre blends as presented in Fig. 2c, two-stage mass loss was noticed at the low
263	blending ratio. The results on gasification parameters listed in Table 2 demonstrated that coal
264	had a dominant effect on the blend of 10 wt% tyre since it showed similar behaviours with the
265	characteristics of coal. It's mainly because coal and tyre have similar physical and chemical
266	properties as shown in Table 1. These properties were assumed as a result of the narrow
267	difference of initial decomposition temperatures between coal (415 °C) and tyre (357 °C), and
268	also a very small amount of tyre being blended. In contrast, the blend of 30 wt% tyre
269	experienced three-stage mass loss. From Table 2, it can be found that the first and the second
270	mass loss of the blend occurred at the same temperature as the first mass loss of original tyre
271	and coal, respectively. This means that the decomposition of tyre and coal in the blend was
272	took place sequentially rather than simultaneously. However, it can be seen from Fig. 2c that
273	gasification was shifted toward lower temperature with the increase of tyre fraction in the

blends. The reason behind the shifting was because of the thermal properties of tyre which decomposes relatively lower temperature than coal. Besides, the third mass loss rate which represented the char gasification of both coal and tyre was reduced (about 10% because of increment of blending proportion from 10% to 30%) due to the low amount of char generated from coal in the blend.

279 Among all the blend samples investigated in this study, coal/polystyrene blends exhibited 280 different behaviours as illustrated in Fig. 2d. As polystyrene decomposed through a single-281 stage mass loss without any solid residue and its initial decomposition took place at a slightly 282 slower temperature than coal, a complete coincidence between whole polystyrene mass loss 283 and first mass loss of coal was observed. Hence, devolatilisation stage of the blend samples 284 was represented by a single peak in DTG profile. Compared with coal, polystyrene contained 285 approximately three times higher amount of volatiles, which resulted in the substantial 286 increment of the first mass loss rates in blends. Similar to other blends, a gradual decrement 287 of the second mass loss rate was observed when polystyrene was blended with coal. At the 288 same time, the temperature corresponding to the maximum mass loss was also reduced. This 289 was affected by the combination of the decrease in the percentage of coal and the increment of volatiles from polystyrene in the blends. 290

3.4 Interactions between coal and carbonaceous materials

292 To further understand the interactions in blend, the experimental and calculated DTG curves of the blends under CO₂ atmosphere are compared in Fig. 3. Overall, the curves illustrate that 293 294 the interactions existed as the experimental curves of blends deviated from the calculated 295 curves for the same blends [19, 20]. However, the deviation was not consistent at different 296 stages of the entire process and varied with the blending of different materials, blending ratio 297 and temperature range. From Fig. 3a, deviations were observed for coal/oat straw blends at 298 two different temperature regions. This indicated the existence of significant interactions in the range of 950-1200 °C and weaker interactions between 390-700 °C for both high and low 299 300 blending ratios. Similarly, as seen in Fig. 3d, coal/polystyrene blends also showed significant 301 interactions at different temperature ranges of 400-500 °C and 1000-1200 °C. For 302 coal/NMPCBs blends (Fig. 3b), analogous behaviours were noticed where stronger and weaker interactions occurred in the temperatures ranges of 1050-1200 °C and 600-900 °C, 303 304 respectively. Correspondingly, Fig. 3c depicted that in temperature range of 1160-1200 °C, 305 coal/tyre blends experienced noticeable interactions while insignificant interactions occurred 306 from 400 to 600 °C.

307 In order to determine the intensity of the interactions in the whole process, the root mean 308 square interactions index (RMSII) of the blends (as detailed in Eq. 2) were also evaluated and 309 listed in Table 3, which is of an accumulated uncertainty of ± 0.01 . Overall, the presence of 310 interactions was evident in the blends as demonstrated by RMSII value. From Table 3, it is 311 apparent that the higher the amount of carbonaceous material was added into the blend, the 312 higher the value of RMSII was observed. Moreover, regardless of blending ratio, RMSII 313 values of char gasification stage were higher than those of devolatilisation stage, except for 314 polystyrene, which ratified the strong interactions during char gasification stage. However, 315 significant interactions were noticed in devolatilisation stage for coal/polystyrene blends as 316 demonstrated by high RMSII value. At both blending ratios, the highest RMSII values in 317 devolatilisation and char gasification stages were exhibited by coal/polystyrene blends and 318 coal/oat straw blends, respectively. It is important to point out that the interactions in char 319 gasification stage of coal/NMPCBs blend at high blending ratio were considerably lower than 320 those of other ash containing blends. As seen in Table 1, NMPCBs had the highest ash 321 content (67.2 wt%) among all the carbonaceous materials studied. Thus, the high ash content 322 of NMPCBs which subsequently result in ash inhibition effect contributing to its weakest 323 interactions [33].

Table 4 shows the comparison of RMSII values between co-pyrolysis and CO₂ co-gasification processes which were conducted under the same temperature but different atmosphere. From Table 4, it is apparent that coal/oat straw blend showed higher interactions during gasification process as demonstrated by its RMSII value which was over two times higher than that of 328 pyrolysis process. The reasons behind the aforementioned characteristics can be well329 explained as follows.

It was observed that DTG curves for the devolatilisation stage in pyrolysis were similar to those in CO₂ gasification so that the main DTG_{max} was almost coincided each other at the same temperature. With the increase of temperature (above 900 °C), considerable additional weight loss was noticed in CO₂ gasification process. This was because of the reaction of CO₂ with the remaining char (mostly) (Eq. 3) as CO₂ has reaction activity at high temperature condition [34, 35]. Thus, it is confirmed that char gasification process played a key role for these interactions.

$$337 \qquad C + CO_2 \to 2CO \tag{3}$$

338 Similarly, enhanced interactions were noticed for the blends of coal with each NMPCB and 339 tyre in CO₂ co-gasification. However, the enhancements in these blends were considerably 340 lower than that of coal/oat straw blend. This was due to the fact that char derived from 341 biomass are normally more reactive compared with those derived from other carbonaceous 342 materials such as NMPCBs. Furthermore, this phenomenon might be originated from mineral 343 content in these samples which will be discussed in the following section. Exceptionally, 344 coal/polystyrene blend had a slightly lower RMSII value in CO₂ gasification process which 345 revealed weaker interactions than in its pyrolysis process. This indicated the influence of CO₂ in polystyrene devolatilisation process as the contribution of C + CO₂ reaction is extremely 346

347	small since char content was literally zero. It is reported that the devolatilisation of
348	polystyrene mainly occurred because of the breaking of C-C and C-H bonds [36]. During CO ₂
349	gasification, CO ₂ reacts randomly with these bonds [34]. Hence, coal/polystyrene blend is
350	likely to demonstrate less interaction in this process as also presented in Table 4. Thus, it can
351	be summarised that char gasification process increased the interactions on blends of oat straw,
352	NMPCBs and tyre with coal while the devolatilisation process had a significant role on
353	coal/polystyrene blend. Nonetheless, it is interesting to note that coal/polystyrene blend
354	revealed the highest interactions in both processes as shown by its highest RMSII values
355	compared to the other blends. This was due to the high volatiles content in polystyrene sample
356	which will be detailed in section 3.5.

The aforementioned discussion has suggested that the mechanism of interactions between coal and carbonaceous materials were highly influenced by the properties and percentage of carbonaceous materials being added into the blends. Furthermore, the effect of CO_2 as gasification agent on interactions was also noticed.

Table 3: RMSII value of experimental and calculated DTG curves.

	RMSII (SD±0.01)				
Carbonaceous materials	Devolat	ilisation	Char gasification		
muter funs	10%	30%	10%	30%	
Oat straw	0.10	0.09	0.26	0.71	
NMPCBs	0.13	0.25	0.20	0.24	
Tyre	0.06	0.10	0.18	0.60	
Polystyrene	0.32	0.92	0.13	0.17	

 Blends
 CO₂ co-gasification
 Co-pyrolysis

 AC/OS
 0.43
 0.20

 AC/NMPCBs
 0.25
 0.22

 AC/tyre
 0.29
 0.25

 AC/PS
 0.56
 0.60

362 Table 4: Comparison of RMSII values between co-gasification and co-pyrolysis processes (70:30 blends).



Figure 3: Comparison of experimental and calculated DTG curves of (a) coal/oat straw, (b) coal/NMPCBs,
 (c) coal/tyre and (d) coal/polystyrene blends.

368 3.5 Mechanism of interactions

In spite of blending equal proportions of different carbonaceous materials with coal, intensityand temperature at which interactions took place varied considerably. High level of volatile

371 content and high reactivity of carbonaceous materials were the reasons for the existence of 372 interactions in the blends [19, 37]. As previously explained, at both low and high blending 373 ratios, coal/polystyrene blends exhibited the highest interactions during devolatilisation stage. 374 This was due to the generation of large amount of reactive free radicals from polystyrene 375 cracking of as a result of containing a relatively higher amount of volatiles. The free radicals 376 were released through thermal decomposition of polystyrene, which react with coal and 377 enhance the devolatilisation and coal char gasification. In addition, the volatiles generated by 378 coal can react with the gases resulted from the cracking of the heavy volatiles as well as light 379 molecules (mostly hydrocarbon) derived from polystyrene. Therefore, a sluggish increment of 380 the secondary char generation is observed because of the inhibition of reverse Boudouard 381 reaction. Furthermore, the maximum weight loss rate also influences the generation of 382 different kinds of radicals. The rate of radicals formation is increased by their stability; hence 383 the conversion rate is greater when more stable radicals formed during devolatilisation 384 process [38]. Applying this concept to carbonaceous materials, the maximum weight loss rate 385 obtained by polystyrene is very high and this suggests the generation of stable radicals. This 386 phenomenon also assisted to the significant interactions during devolatilisation stage. On the 387 other side, even though NMPCBs had lower amount of volatiles, stronger interactions were 388 observed at devolatilisation stage. This behaviour was due to epoxy resin, a main part of 389 NMPCBs, which decomposed at relatively low temperature because of its high reactivity.

390	The interactions can also be caused by the catalytic effect of alkali and alkaline earth metals
391	(AAEM) contained in ash of the original samples [37, 39, 40]. It is reported that AAEM,
392	particularly, Ca and K species are active catalyst for gasification process [41, 42].
393	In the beginning of decomposition process, calcium, potassium and the other AAEM species
394	are combined with the carboxyl and phenolic groups, forming organic molecules in coal or
395	carbonaceous sample [43]. At the same time, the organic substances in the solid surface are
396	transformed to a micro-crystalline char, favouring the structure of char to become more
397	porous which enhances the reactivity in gasification [42].
398	Afterwards, this porous char adsorbs the AAEM onto its surface during the volatile-char
399	interactions stage [44]. Moreover, the presence of intermediate alkali-surface compounds
400	creates a bigger gap between AAEM and carbon which result in significant volume
401	expansion. These behaviours therefore contribute to the promotion of the gasification reaction
402	through weakening the carbon-carbon bonds present between layers [41].
403	As seen in Table 1, oat straw ash has demonstrated mostly the highest amount of these AAEM
404	among the others. In particular, CaO and K_2O in oat straw were found to be 10.6 wt% and
405	32.8 wt%, respectively, which were essential components for catalytic effect. Therefore, these
406	mainly contributed to the highest interactions in char gasification stage shown by coal/oat
407	straw blends compared with other blends. Meanwhile, the ash from NMPCBs, tyre and coal

408 samples contain a comparable percentage of AAEM, much lower than those of oat straw.

409 Moreover, ash compositional data showed that oat straw had high silica content which has 410 been demonstrated to reduce the reactivity of char [45]. However, the proportion of silica in 411 the other samples were relatively similar to each other, thus, the overall effect from silica was 412 assumed to be similar regardless of types of samples. Similarly, alumina also acts as a 413 deactivator for the catalytic effect [46]. As the lowest percentage of Al₂O₃ (2.2 wt%) was 414 observed in oat straw ash, so that oat straw had the least obstacle to undergo the catalytic 415 activities. It is worth noting that coal/tyre blend exhibited higher interactions than NMPCBs at 416 high blending ratio despite both had almost equal percentage of AAEM. The reason behind 417 these characteristics can be explained by Fe₂O₃ in the ash. Previous study revealed that metal 418 oxide such as Fe₂O₃ increased the gasification activity of coal due to the catalytic effect in 419 CO_2 gasification [47]. As tyre ash contain higher percentage of Fe₂O₃ (45.1 wt%), hence, it 420 could act as a catalyst for char-CO₂ reaction.

421 **4** Conclusion

422 CO_2 co-gasification of coal with four different waste carbonaceous materials was investigated 423 in a thermogravimetric analyser. The influence of waste origin on the mechanism of 424 interactions was also discussed. The main findings are summarised as follows: (i) significant 425 high volatiles content of carbonaceous materials could plays a vital role in enhancing the 426 degree of interactions during devolatilisation stage; (ii) AAEM metals content such as CaO 427 and K₂O had a key influence on strong interactions in char gasification stage due to their 428 catalytic effect; (iii) the improvement of coal gasification performance is noticed as 429 demonstrated by its gasification characteristics such as initial decomposition temperature and 430 final mass loss temperature with the addition of carbonaceous materials, irrespective of stage 431 at which the interactions occurred. The best performances of gasification process in CO₂ 432 atmosphere were achieved by using polystyrene and oat straw as carbonaceous materials to be 433 blended with coal. These phenomena were as expected mainly due to polystyrene having the 434 highest volatiles content whereas oat straw presented the catalytic effect from its AAEM 435 metals. Furthermore, the comparison between co-pyrolysis and CO₂ gasification processes by 436 using RMSII values confirmed the influence of CO₂ as a gasifying agent on the presence of 437 interactions in the blends. Hence, these experimental findings have confirmed that CO₂ 438 gasification of coal can be considerably improved by adding polystyrene and oat straw into 439 the process.

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447 **5 References**

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449 450 1. Wu, T., et al., Characteristics and synergistic effects of co-firing of coal and 451 carbonaceous wastes. Fuel, 2013. 104(0): p. 194-200. 452 2. Sami, M., K. Annamalai, and M. Wooldridge, Co-firing of coal and biomass fuel 453 blends. Progress in Energy and Combustion Science, 2001. 27(2): p. 171-214. 454 3. Pan, Y.G., et al., Fluidized-bed co-gasification of residual biomass/poor coal blends 455 for fuel gas production. Fuel, 2000. 79(11): p. 1317-1326. 456 4. Seo, M.W., et al., Gasification characteristics of coal/biomass blend in a dual 457 circulating fluidized bed reactor. Energy & Fuels, 2010. 24(5): p. 3108-3118. 458 5. Wang, L., et al., CO2 Gasification of Chars Prepared from Wood and Forest Residue: 459 A Kinetic Study. Energy & Fuels, 2013. 27(10): p. 6098-6107. 460 Skodras, G., G. Nenes, and N. Zafeiriou, Low rank coal - CO2 gasification: 6. 461 Experimental study, analysis of the kinetic parameters by Weibull distribution and 462 compensation effect. Applied Thermal Engineering, 2015. 74(0): p. 111-118. 463 7. Lahijani, P., et al., CO2 gasification reactivity of biomass char: Catalytic influence of 464 alkali, alkaline earth and transition metal salts. Bioresource Technology, 2013. 465 144(0): p. 288-295. 466 8. Butterman HC, C.M., CO2 as a Carbon Neutral Fuel Source via Enhanced Biomass 467 Gasification. Environ. Sci. Technol., 2009(43): p. 9030-9037. 468 9. Edreis, E.M.A., G. Luo, and H. Yao, Investigations of the structure and thermal 469 kinetic analysis of sugarcane bagasse char during non-isothermal CO2 gasification. 470 Journal of Analytical and Applied Pyrolysis, 2014. 107(0): p. 107-115. 471 10. Kirtania, K., et al., Comparison of CO2 and steam gasification reactivity of algal and 472 woody biomass chars. Fuel Processing Technology, 2014. 117(0): p. 44-52. 473 11. Khalil, R., et al., CO2 gasification of biomass chars: a kinetic study. Energy & Fuels, 474 2008. **23**(1): p. 94-100. 475 Sircar, I., et al., Experimental and modeling study of pinewood char gasification with 12. 476 CO2. Fuel, 2014. 119(0): p. 38-46. 477 13. Lewis, A.D., E.G. Fletcher, and T.H. Fletcher, Pyrolysis and CO2 Gasification Rates 478 of Biomass at High Heating Rate Conditions. 2013. 479 14. Huo, W., et al., Study on CO2 gasification reactivity and physical characteristics of 480 biomass, petroleum coke and coal chars. Bioresource Technology, 2014. 159(0): p. 481 143-149. 482 15. Pohořelý, M., et al., CO2 as moderator for biomass gasification. Fuel, 2014. 117, Part 483 A(0): p. 198-205.

484 16. Hanaoka, T., S. Hiasa, and Y. Edashige, Syngas production by CO2/O2 gasification of 485 aquatic biomass. Fuel Processing Technology, 2013. 116(0): p. 9-15. 486 BS/EN15443:2011, Solid recovered fuels. Methods for the preparation of the 17. 487 laboratory sample. 2011. 488 ISO17246:2010, Coal - Proximate analysis. 2010. 18. 489 19. Edreis, E.M.A., et al., Synergistic effects and kinetics thermal behaviour of petroleum 490 coke/biomass blends during H2O co-gasification. Energy Conversion and 491 Management, 2014. 79(0): p. 355-366. 492 20. Gil, M.V., et al., Thermal behaviour and kinetics of coal/biomass blends during co-493 combustion. Bioresource Technology, 2010. 101(14): p. 5601-5608. 494 Wang, Q., et al., Interactions and kinetic analysis of oil shale semi-coke with cornstalk 21. 495 during co-combustion. Applied Energy, 2011. 88(6): p. 2080-2087. 496 22. Butterman, H.C. and M.J. Castaldi. CO2 enhanced steam gasification of biomass 497 fuels. in 16th Annual North American Waste-to-Energy Conference. 2008. American 498 Society of Mechanical Engineers. 499 Basu, P., Combustion and gasification in fluidized beds. 2006: CRC press. 23. 500 24. Sahu, S.G., N. Chakraborty, and P. Sarkar, Coal-biomass co-combustion: An 501 overview. Renewable and Sustainable Energy Reviews, 2014. 39(0): p. 575-586. 502 Xie, Z. and X. Ma, The thermal behaviour of the co-combustion between paper sludge 25. 503 and rice straw. Bioresource Technology, 2013. 146(0): p. 611-618. 504 26. Basu, P., Biomass gasification and pyrolysis: practical design and theory. 2010: 505 Academic press. 506 27. Backreedy, R.I., et al., Modeling the reaction of oxygen with coal and biomass chars. 507 Proceedings of the Combustion Institute, 2002. 29(1): p. 415-421. 508 Wang, X., et al., Kinetics investigation on the combustion of biochar in O2/CO2 28. 509 atmosphere. Environmental Progress & Sustainable Energy, 2015. 34(3): p. 923-932. 510 Wu, W. and K. Qiu, Vacuum co-pyrolysis of Chinese fir sawdust and waste printed 29. 511 circuit boards. Part I: Influence of mass ratio of reactants. Journal of Analytical and 512 Applied Pyrolysis, 2014. 105(0): p. 252-261. 513 Quan, C., A. Li, and N. Gao, Thermogravimetric analysis and kinetic study on large 30. 514 particles of printed circuit board wastes. Waste Management, 2009. 29(8): p. 2353-515 2360. 516 31. Martínez, J.D., et al., Waste tyre pyrolysis – A review. Renewable and Sustainable 517 Energy Reviews, 2013. 23: p. 179-213. 518 32. Chen, F. and J. Qian, Studies of the thermal degradation of waste rubber. Waste 519 Management, 2003. 23(6): p. 463-467. 520 33. Zhang, Y., et al., Effect of fuel origin on synergy during co-gasification of biomass 521 and coal in CO2. Bioresource Technology, 2016. 200: p. 789-794.

522 34. Chen, T., et al., Key thermal events during pyrolysis and CO2-gasification of selected 523 combustible solid wastes in a thermogravimetric analyser. Fuel, 2014. 137: p. 77-84. 524 Lai, Z., et al., Thermogravimetric analysis of the thermal decomposition of MSW in 35. 525 N2, CO2 and CO2/N2 atmospheres. Fuel Processing Technology, 2012. 102: p. 18-23. 526 36. Grammelis, P., et al., Pyrolysis kinetics and combustion characteristics of waste 527 recovered fuels. Fuel, 2009. 88(1): p. 195-205. 528 37. Tchapda, A.H. and S.V. Pisupati, A review of thermal co-conversion of coal and 529 biomass/waste. Energies, 2014. 7(3): p. 1098-1148. 530 Kiran, N., E. Ekinci, and C. Snape, Recyling of plastic wastes via pyrolysis. 38. 531 Resources, Conservation and Recycling, 2000. 29(4): p. 273-283. 532 39. Edreis, E.M.A., et al., CO2 co-gasification of lower sulphur petroleum coke and sugar 533 cane bagasse via TG-FTIR analysis technique. Bioresource Technology, 2013. 534 **136**(0): p. 595-603. 535 40. Nemanova, V., et al., Co-gasification of petroleum coke and biomass. Fuel, 2014. 117, 536 **Part A**(0): p. 870-875. 537 Huang, Y., et al., Effects of metal catalysts on CO2 gasification reactivity of biomass 41. 538 char. Biotechnology Advances, 2009. 27(5): p. 568-572. 539 42. Mitsuoka, K., et al., Gasification of woody biomass char with CO2: The catalytic 540 effects of K and Ca species on char gasification reactivity. Fuel Processing 541 Technology, 2011. 92(1): p. 26-31. 542 43. Quyn, D.M., et al., Volatilisation and catalytic effects of alkali and alkaline earth 543 metallic species during the pyrolysis and gasification of Victorian brown coal. Part 544 IV. Catalytic effects of NaCl and ion-exchangeable Na in coal on char reactivity x. 545 Fuel, 2003. 82(5): p. 587-593. 546 Li, X. and C.-Z. Li, Volatilisation and catalytic effects of alkali and alkaline earth 44. 547 metallic species during the pyrolysis and gasification of Victorian brown coal. Part 548 VIII. Catalysis and changes in char structure during gasification in steam. Fuel, 2006. 549 **85**(10–11): p. 1518-1525. 550 45. Kannan, M.P. and G.N. Richards, Gasification of biomass chars in carbon dioxide: 551 dependence of gasification rate on the indigenous metal content. Fuel, 1990. 69(6): p. 552 747-753. 553 46. Kajita, M., et al., Catalytic and Noncatalytic Mechanisms in Steam Gasification of 554 Char from the Pyrolysis of Biomass[†]. Energy & Fuels, 2009. **24**(1): p. 108-116. 555 47. He, P., et al., Simultaneous Low-Cost Carbon Sources and CO2 Valorizations through 556 Catalytic Gasification. Energy & Fuels, 2015. 29(11): p. 7497-7507. 557 558