

Deposition of Micro-elements through Leaf Fallen from Different Types of Vegetation, North-eastern Mexico

Tilo G. Dominguez-Gomez¹, Humberto Gonzalez-Rodriguez¹, Israel Cantu-Silva¹, Roque G. Ramirez-Lozano^{2*}, Marco V. Gomez-Meza³, Maria Del Socorro Alvarado¹

¹Universidad Autonoma de Nuevo Leon, Facultad de Ciencias Forestales, Carr. Nac. No 85, km 145. Linares, Nuevo Leon 67700, Mexico

²Universidad Autonoma de Nuevo Leon, Facultad de Ciencias Biologicas, Av. Pedro de Alba s/n. San Nicolas de los Garza, Nuevo Leon 66400, Mexico

³Universidad Autonoma de Nuevo Leon, Facultad de Economia, Av. Lazaro Cardenas 4600 Ote, Frac. Residencial Las Torres, Monterrey, Nuevo Leon 64930, Mexico

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Correspondence to

*E-mail: roque.ramirezlz@uanl.edu.mx

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Abstract

The aim of the study was to determine monthly (2007) the content and deposition of Cu, Fe, Mn and Zn in leaf fallen litterfall samples collected in four sites, northeastern Mexico. Site 1 was located at 1600 m of altitude in a pine forest, mixed with deciduous trees; site 2 was at 550 m in the ecotone of a *Quercus* spp. forest and submontane scrub; sites 3 and 4 (at 350 m and at 300 m, respectively) were located in the Tamaulipan Thornscrub vegetation. Leaf fallen litterfall samples at each site (2,500 m²) were obtained from ten canisters of 1.0² m that were randomly situated at each site. The Cu annual deposition (g ha⁻¹ yr⁻¹) was significantly different among sites being higher in site 4 (23.2) and lower in site 1 (4.1). Fe deposition was also significantly higher in site 4 (522.2) and lower in site 1 (120.0). Mn was higher in site 2 (479.4) and lower in site 1 (64.6). Zn was significantly higher in site 1 (62.8) and lower in site 1 (24.3). Micronutrient annual order deposition was as follows: Fe>Mn>Zn>Cu. Differences in deposition may be attributable to environmental conditions and plant species composition at each studied plant community.

1. Introduction

In northeastern Mexico there are different types of vegetation such as pine forest, oak forest, the pine-oak forest, and the subtropical woodlands, typical of the semi-arid plains (Vargas, 1999). These ecosystems include a variety of deciduous species used as forage to domestic and wild ruminants; in addition, as forestry resources to human beings, reforestation practices, timber and charcoal production. Therefore, this region provides an opportunity to investigate the nutrient return and biogeochemical cycles as a measure of ecosystem productivity through litterfall deposition (leaves, twigs, inflorescences, fruits, seeds and others) and to characterize their contribution of nutrients to the forest soil.

Litterfall is an important process in the nutrient cycle, which determines the renewal and input of organic matter to the soil (Cantu and Gonzalez, 2001). More than half of the annual absorption of nutrients in the woods is due to the return of litter to soil (Del Valle, 2003). The production and litter

decomposition are important mechanisms in the nutrient cycling in vegetation areas (Steubing et al., 2001). Nevertheless, Palma et al. (2000) argue that seasonal differences in production and nutritional quality of litter could be related to extreme climatic fluctuations and/or phenological changes such as abscission and bud initiation, flowering and fruiting.

Deposition of micronutrients from leaves plays an important role through soil covering and modifying its environment. In addition, litterfall decomposition becomes a source of organic matter enhancing the geochemical cycles of elements (Gliessman, 2002). Studies conducted by Del Valle (2003), Vasconcelos and Luizao (2004) and Gonzalez et al. (2008) agreed about the quantities of annual inputs of nutrients through litterfall, which are cycled by the leaf fallen litterfall as a main source of organic matter to the forest soil. The leaf component is capable to stock between 30 to 70% of nutrients since in leaves occur most of the physiological processes, specifically photosynthesis (Piatek and Allen, 2000).



The decomposition and mineralization of litterfall are given mainly by a limited number of elements, being of great importance some essential nutrients and some heavy metals potentially toxic (Guerrero et al., 1998). Studies on nutrient cycling were conducted in forest areas to gather information about the contribution of litterfall and nutrients (Mlambo and Nyathi, 2007). These studies are limited to the availability of mineral elements such as Ca, K, Mg, Si, P, Fe, Mn, B, Cl, Mo, Zn, Na and Co. However, few studies are focused to quantitate the amount of micronutrient deposition in different plant communities of the northeastern region of Mexico. Therefore, the aim of this research was to assess the content and deposition of micronutrients through leaf fallen in diverse types of vegetation.

2. Materials and Methods

2.1. Sites of study

This study was carried out at four undisturbed grazing sites located in the state of Nuevo Leon, northeastern Mexico. Site 1 was located in the pine (*Pinus pseudostrubus* Lindl.) forest mixed with deciduous trees (*Quercus* spp) located at the Experimental Forest Research Station of the Universidad Autonoma de Nuevo Leon in the Sierra Madre Oriental Mountain Range, Iturbide county (24° 43' N; 99° 52' W; 1600 m of altitude). Annual mean air temperature is about 14 °C, average annual rainfall is approximately 639 mm (INEGI, 2002). The soils are mainly rocky and comprise upper cretaceous lutite or siltstone. Site 2 was located in the ecotone of a *Quercus* spp. forest and the Piedmont scrub vegetation (24° 46' N; 99° 41' W; 550 m of altitude) in the “Ejido Crucitas” in Linares county. Average total annual rainfall is about 755 mm (INEGI, 2002). Site 3 was located at the Experimental Research Station of the Faculty of Forest Sciences of the Universidad Autonoma de Nuevo Leon (24° 47' N; 99° 32' W; 350 m of altitude) in Linares county. Average total annual rainfall is about 805 mm (INEGI, 2002). Peak rainfall months are May, June, and September (Gonzalez-Rodriguez et al., 2004). Site 4 was located in the “Ejido Hacienda de Guadalupe” in Linares county (24° 54' N; 99° 32' W; elevation, 300 m of altitude). Total cumulative annual rainfall is about 672 mm. Vegetation of sites 2 and 3 is known as the Tamaulipan Thornscrub or Subtropical Thornscrub Woodlands. In the last three sites, the climate is subtropical and semiarid with warm summer, monthly mean air temperature ranges from 14.7 °C in January to 22.3 °C in August, although daily high temperatures of 45 °C are common during summer. The dominant soils are deep, dark-gray, lime-gray, lime-clay vertisols, with montmorillonite, which shrink and swell noticeably in response to changes in soil moisture content (Gonzalez et al., 2011).

2.2. Litterfall sampling

An experimental plot (2,500 m²) was delimited at each studied site. At each site, ten canisters of 1.0×1.0 m were randomly scattered inside the plot. Canisters were made of a wooden beveled frame (10 cm) with a plastic mesh (1.0 mm) on the bottom. The canisters were placed at 0.30 m above soil surface, in this way throughfall is allowed to pass thru. In this study, the sampling area, number of canisters, and height above soil level are within the range of previous studies (Fang et al., 2007; Zhou et al., 2007). The frequency of litter sampling was at intervals of 30 days between December 21, 2006 and December 20, 2007. Dry weight of samples was determined by drying samples at 65°C for 72 hours until constant weight using a forced air oven. The weight of each sample was determined using a digital scale (Sartorius Brand, Model C1), with a resolution of a thousandth of a gram. Dry samples were ground in a Wiley mill (Thomas Scientific) to pass 1.0 mm mesh sieve and were kept in closed paper envelopes.

2.3. Chemical analyses

Since both, macro- (Ca, K, Mg, P, and N) and micro-nutrients (Cu, Fe, Mn and Zn) play essential roles in plant cell metabolism and plant life cycle, most of the total litterfall is constituted mainly by leaves (from 74% to 86%, on a yearly basis), and because the contribution of each litterfall component deposition and macronutrient deposition through leaves have been previously documented at research sites by Gonzalez et al. (2011), it was widely considered by the authors of this research paper to know and to document the content and deposition pattern of micronutrients on a monthly as well as on a yearly basis, since very little information is available of this kind.

By quintuplicate, 2.0 g of leaf litter samples collected from each canister were subjected to micronutrients analysis. Firstly, leaf samples were incinerated during 5 h in a furnace at 550°C. Ashes were digested by the wet digestion technique using a solution containing HCl and HNO₃ (Cherney, 2000). Contents (µg g⁻¹ dry weight) of Cu, Fe, Mn and Zn (air flame/acetylene) were determined by atomic absorption spectrophotometry using a Varian spectrophotometer (model SpectrAA-200). Micronutrient deposition (mg m⁻²) at each site was calculated by multiplying leaf fallen production of each sampling date by the mineral content of the corresponding date of sampling, site, replication, and adding them over the entire year. The cumulative monthly values at each site were used as an estimate of the annual micronutrient deposition (mg ha⁻¹ yr⁻¹).

2.4. Statistical analyses

Data of micronutrient content and deposition were subjected to one-way analysis of variance (Steel and Torrie, 1980). Normal distribution and homogeneity of variances of deposition and micronutrient content data were analyzed by using the



Kolmogorov-Smirnov, Shapiro-Wilk, and Levene tests (Brown and Forsythe, 1974; Steel and Torrie, 1980) indicating that deposition and micronutrient content data were non-normally distributed. Since for most sampling dates, ANOVA did not show the assumption of equality of variances, the Kruskal-Wallis nonparametric test was employed (Ott, 1993) to detect significant differences among sites at each sampling date. Hence, differences in micronutrient content and deposition among sites were validated using the Mann-Whitney U nonparametric test with the Bonferroni's correction method at $p=0.05$ (Wackerly et al., 2002). All applied statistical methods were according to the SPSS 2000; (Statistical Package for the Social Sciences) software package (standard released version 13.0 for Windows, SPSS Inc., Chicago, IL).

3. Results and Discussion

3.1. Monthly microelement content of fallen leaves

The Cu content ($\mu\text{g g}^{-1}$ of dry weight; Figure 1a) in site 1 ranged from 1.5 (May) to 5.3 (February); in site 2, from 4.2 (October) to 8.3 (April); in site 3, from 3.3 (January) to 6.8 (April); and in site 4, from 6.0 (July) to 11.2 (March). Leaf litter Fe content ($\mu\text{g g}^{-1}$ of dry weight; Figure 1b) in site 1 varied from 57.8 (March) to 101.8 (April); for site 2, from 44.7 (March) to 262.4 (December); for site 3, from 81.7 (February) to 208.7 (July); and for site 4, from 114.7 (January) to 325.9 (June). The Mn content ($\mu\text{g g}^{-1}$ of dry weight; Figure 1c) in site 1 varied from 16.1 (May) to 177.3 (March); for site 2, from 87.79 (June) to 263.0 (October); for site 3, from 26.83 (April) to 36.0 (July); and for site 4, from 56.2 (November) to 110.6 (January). The Zn content ($\mu\text{g g}^{-1}$ of dry weight; Figure 1d) of site 1 ranged from 10.7 (December) to 30.0 (September); in site 2, from 7.5 (March) to 26.4 (June); in site 3, from 12.2 (June) to 25.0 (September); and in site 4, from 15.0 (July) to 29.3 (February).

3.2. Monthly deposition of microelements

The Cu deposition (mg m^{-2} ; Figure 2a) in site 1, ranged from 0.01 (January) to 0.10 (November); in site 2, from 0.05 (May) to 0.18 (September); in site 3, from 0.06 (July) to 0.28 (December); and in site 4, from 0.03 (May) to 0.46 (November). Leaf litter Fe deposition (mg m^{-2} ; Figure 2b) in site 1 varied from 0.21 (February) to 1.88 (November); for site 2, from 0.50 (May) to 4.72 (December); for site 3, from 1.35 (May) to 7.80 (September); and site 4, from 0.71 (May) to 13.5 (August). The Mn deposition (mg m^{-2} ; Figure 2c) in site 1 varied from 0.15 (May and July) to 1.14 (November); for site 2, from 0.97 (June) to 10.4 (September); for site 3, from 0.34 (June) to 1.37 (December); and for site 4, from 0.34 (May) to 3.02 (February). The Zn deposition (mg m^{-2} ; Figure 2d) for site 1 ranged from 0.3 (February) to 0.40 (November); in site

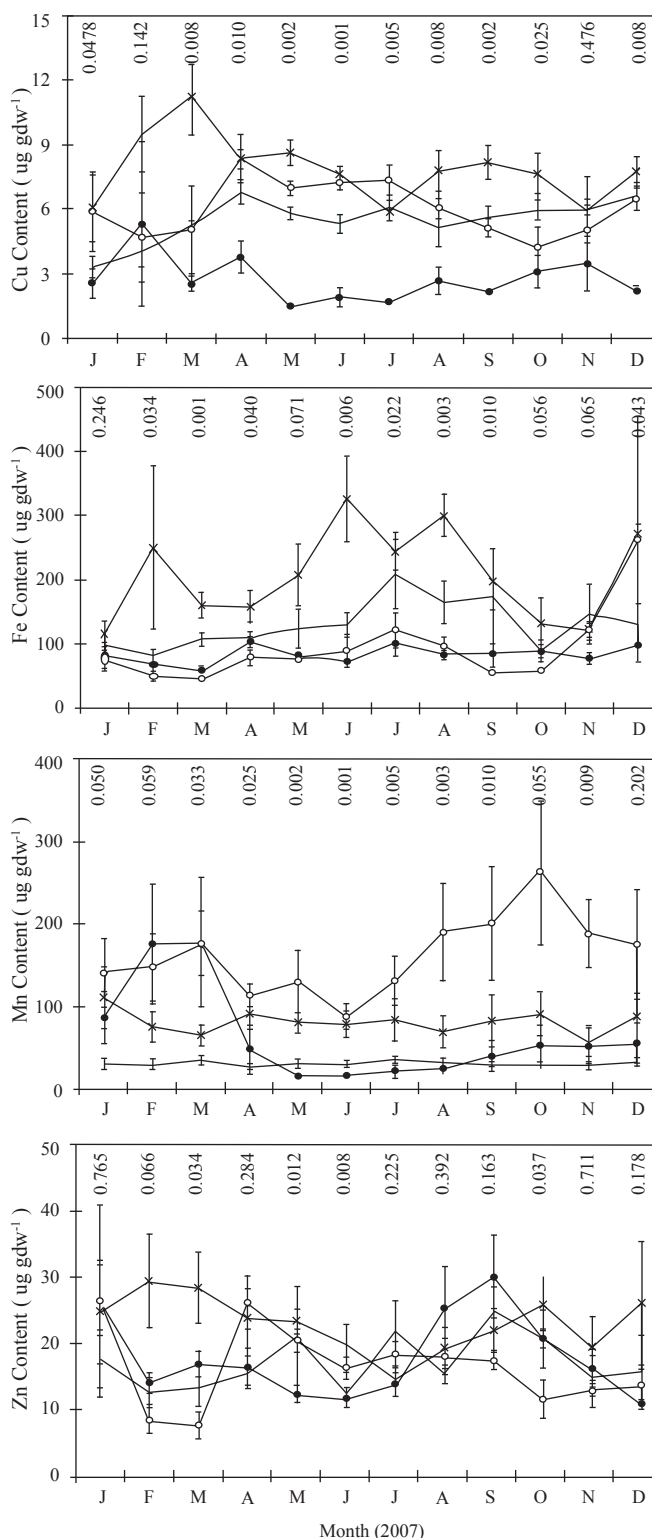


Figure 1: Monthly variations of Cu, Fe, Mn, and Zn contents in leaf fallen litter at research sites. *P*-values of the Kruskal-Wallis test to detect significant differences among sites are shown at each sampling date within the graph. *P*-values ≤ 0.001 are denoted as 0.001. Statistically significant probabilities ($P \leq 0.05$) are shown in boldface. ●=site 1, ○=site 2, □=site 3, ×=site 4

2, from 0.13 (May) to 0.64 (September); in site 3, from 0.17 (March) to 0.89 (September); and in site 4, from 0.09 (July) to 1.37 (February).

3.3. Annual deposition of microelements

Annual deposition of Cu ranged from 4.1 (Site 1; Table 1) to 23.2 g ha⁻¹ yr⁻¹ (site 4); Fe from 120 (site 1) to 522.1 g ha⁻¹ yr⁻¹ (site 4); Mn from 64.4 (site 3) to 479.4 g ha⁻¹ yr⁻¹ (site 2); and Zn from 24.3 (site 1) to 62.8 g ha⁻¹ yr⁻¹ (site 4). The total annual contribution (Cu+Fe+Mn+Zn) in the site 1, 2, 3 and 4 was of 213.0, 738.1, 522.2 and 822.9 g ha⁻¹ yr⁻¹, respectively. In general, the order of deposition of trace elements was Fe>Mn>Zn>Cu.

According to micronutrient contents, the order of magnitude was as follows: Fe>Mn>Zn>Cu, and by site was: site 4>site 2>site 3>site 1. The annual Cu deposition via leaf litter was in the following decreasing order: site 4>site 3>site 2>site 1. The extreme low and high of total annual Cu depositions were in site 1, with 4.1 and in site 4 with 23.2 g ha⁻¹ yr⁻¹. Higher Cu deposition values, in different plant communities, that varied from 40 to 90 g ha⁻¹ yr⁻¹ were reported by Del Valle (2003), Gonzalez et al. (2006) and Ramirez et al. (2007). With respect to Fe, the decreasing order of total annual deposition was: site 4>site 3>site 2>site 1 with a range from 120 (site 1) to 522 (site 4) g ha⁻¹ yr⁻¹. Gonzalez et al. (2006) found a Fe deposition range between 607 and 1,965 g ha⁻¹ yr⁻¹ when they studied micronutrient deposition in the Tamaulipan thornscrub in the northeastern region of Mexico. In another study, Ramirez et al. (2007) reported Fe depositions of 700, 630 and 420 g ha⁻¹ yr⁻¹ in an oak, *Pinus patula* and cypress forests, respectively. Del Valle (2003) observed a Fe deposition value of 730 g ha⁻¹ yr⁻¹. Xiaoni et al. (2004) documented values of 860 g ha⁻¹ yr⁻¹ in a subtropical forest in Okinawa Island, Japan. The annual deposition of Mn among sites was in the following order: site 2>site 4>site 3>site 1. Mn annual deposition ranged from 64.6 (site 1) to 479.4 (site 2) g ha⁻¹ yr⁻¹. Gonzalez et al. (2006) studied the micronutrient deposition in the Tamaulipan Thornscrub and reported a Mn deposition from 131 to 275 g ha⁻¹ yr⁻¹ while Hagen et al. (2006) found a range of deposition between 270 and 458 g ha⁻¹ yr⁻¹ in a *Quercus robur* forest. Similarly, Xiaoni

Table 1: Annual micronutrient deposition (g ha⁻¹ yr⁻¹) through leaf litterfall at each research site.

Deposition	Sites			
	1	2	3	4
Cu	4.1	14.1	15.4	23.2
Fe	120.0	206.7	378.1	522.1
Mn	64.6	479.4	82.3	214.8
Zn	24.3	37.6	46.4	62.8
Total	213.0	738.1	522.2	822.9

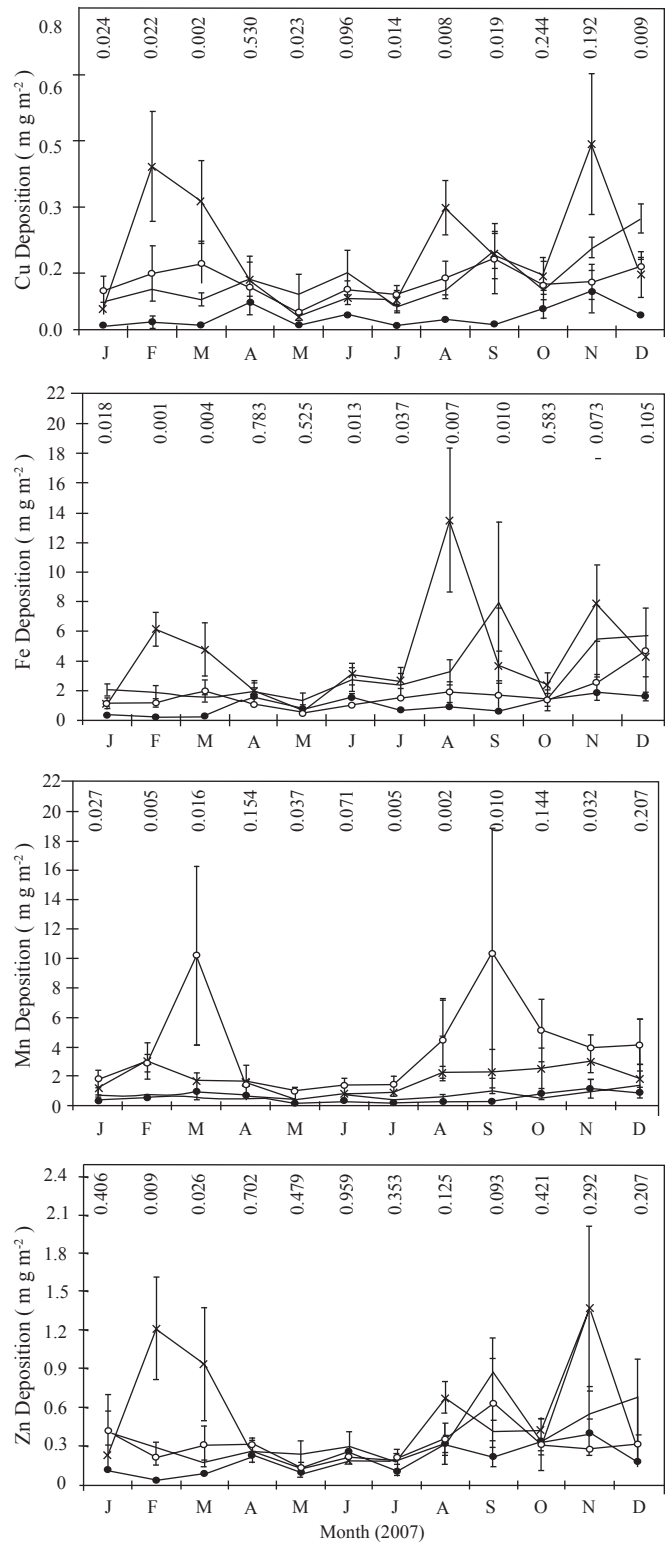


Figure 2: Monthly variations of Cu, Fe, Mn, and Zn deposition in leaf fallen litter at research sites. P-values of the Kruskal-Wallis test to detect significant differences among sites are shown at each sampling date within the graph. p-values ≤ 0.001 are denoted as 0.001. Statistically significant probabilities (P ≤ 0.05) are shown in boldface. ●=site 1, ○=site 2, =site 3, ×=site 4.

et al. (2004) observed Mn deposition values of about 391 g ha⁻¹ yr⁻¹ in a subtropical forest. In contrast, Del Valle (2003) reported values of 930 g ha⁻¹ yr⁻¹ in Mn deposition in South of Colombia. The annual deposition of Zn decreased among sites as follows: site 4>site 3>site 2>site 1 with values that varied from 24.3 (site 1) to 62.8 (site 4) g ha⁻¹ yr⁻¹. Del Valle (2003) found deposition values of 230 g ha⁻¹ yr⁻¹ and Ramirez et al. (2007) reported 75, 105, and 70 g ha⁻¹ yr⁻¹ in Oak, *Pinus patula* and cypress, forest ecosystems, respectively, in Colombia. Studies undertaken in the Tamaulipan thornscrub vegetation, northeast Mexico, by Gonzalez et al. (2006) documented a deposition of Zn from 28.2 to 540 g ha⁻¹ yr⁻¹. In present study, for a given site, Cu and Zn deposition values showed a similar trend during the experimental period; however, Fe and Mn exhibited major peaks and a higher variability in figures. These results may allow to hypothesize that high Fe and Mn content may be associated with translocation processes through leaves just before foliar abscission takes place or the promotion of phenological events. In addition, higher content and deposition values of Fe and Mn could be associated to resource availability, which lead to high absorption of these micronutrients, particularly in sites 2 and 4. This rationale could be used in further research objectives, although nutrient content was not estimated in other litterfall components such as twigs and reproductive structures to find out nutrient allocation and compartmentalization. It has been demonstrated that nutrient concentration is higher in leaves than other litterfall constituents (Yang et al., 2006). Nevertheless, nutrient deposition depends on tissue age, type of structure being studied, season, herbivory, edaphic features and prevailing environmental conditions at research sites (Gliessman, 2002).

4. Conclusions

Results of present study showed a spatial and temporal variation in the content and deposition of micronutrients in the litterfall in the four studied sites and the potential contribution of nutrients varied among plant communities. In addition, the findings of this research confirm the importance of deposition of microelements through leaves, to maintain forest ecosystem productivity. However, further nutrient deposition studies are required to quantify the contribution of other litterfall components in order to understand the dynamic of biogeochemical cycles in forest ecosystems.

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