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Development of a Framework for the Evaluation of the Environmental Benefits of Controlled Traffic Farming

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Abstract: Although controlled traffic farming (CTF) is an environmentally friendly soil management system, no quantitative evaluation of environmental benefits is available. This paper aims at establishing a framework for quantitative evaluation of the environmental benefits of CTF, considering a list of environmental benefits, namely, reducing soil compaction, runoff/erosion, energy requirement and greenhouse gas emission (GHG), conserving organic matter, enhancing soil biodiversity and fertiliser use efficiency. Based on a comprehensive literature review and the European Commission Soil Framework Directive, the choice of and the weighting of the impact of each of the environmental benefits were made. The framework was validated using data from three selected farms. For Colworth farm (Unilever, UK), the framework predicted the largest overall environmental benefit of 59.3% of the theoretically maximum achievable benefits (100%), as compared to the other two farms in Scotland (52%) and Australia (47.3%). This overall benefit could be broken down into: reducing soil compaction (24%), tillage energy requirement (10%) and GHG emissions (3%), enhancing soil biodiversity (7%) and erosion control (6%), conserving organic matter (6%), and improving fertiliser use efficiency (3%). Similar evaluation can be performed for any farm worldwide, providing that data on soil properties, topography, machinery, and weather are available.

Keywords: controlled traffic farming; environmental benefits; prediction framework

1. Introduction

The high demand on food has resulted in increasing the size and weight of agricultural machinery, which has led to higher risk to soil damage and more energy has to be dedicated to cure this damage. The soil damage by agricultural machinery contributes considerably to deepen soil environmental threats, namely, compaction, erosion, increase greenhouse gas (GHG) emission and others. Therefore, soil has to be managed properly to ameliorate the damage, either by mechanical forces e.g., tillage, optimising machine parameters, e.g., reduced tyre inflation pressure or by appropriate management of traffic over the soil. The latter has recently attracted the attention of researchers and farmers. An underestimation of the field area trafficked during one crop cycle is common due to the eradication of wheel tracks by the following tillage. Typically, under conventional tillage with random traffic (RT), almost all field area is trafficked by wheels at least once every year [1]. However, in the last decades, farmers have attempted to concentrate field traffic on temporary or permanent tramlines [2,3].

Controlled farming system (CTF) ensures the crop zone is permanently separated from the traffic zone year after year [4]. It aims at keeping field traffic in the same lanes on the field every year [5]. Thus, the larger bearing capacity of the compacted traffic lanes improves the ability to drive on and the crop zone tends to stay in favourable conditions for growth without the need for deep tillage [4]. There are different ways to implement CTF, consistent for all is the simple principle to not drive at random on the soil [6]. Different guidance systems have been used in CTF to keep the wheels exactly in the same positions from year to year, ranging from concrete tracks [7] to electrical wires and physical markers [8]. However, the recent development of cost effective global positioning systems (GPS) has made it easy to adopt a precise CTF system on farm scale [9,10]. Although CTF has been advocated a long time by scientists in a number of countries, adoption of the new system on large scale has been rare, with one possible exception in Central Queensland, Australia, where over 100,000 ha was converted to CTF over a five year period [10]. This might be attributed to lack of studies, which provide quantitative evaluation of economic and environmental benefits of CTF.

Literature confirms the environmental benefits of CTF, including reduced soil compaction, control of erosion, enhanced soil biodiversity, conserving organic matter, reduced tillage energy requirement, increased fertiliser use efficiency and reduced greenhouse gas emission from the soil [11]. Soil compaction caused by anthropogenic activities such as heavy farm machinery or the result of cyclic tillage is a big concern for farmers as it is directly related to crop growth and potentially to yield. Other factors can also lead to soil compaction occurrence e.g., natural conditions. According to the natural soil susceptibility to compaction, Houšková and Montanarella [12] divided soils in Europe into four categories of low, medium, high, and very high susceptibility to compaction. Soil compaction is associated with increase in bulk density (BD) and penetration resistance (PR), while significant reduction in porosity and pore space may be expected [13]. Therefore, soil compaction also affects the hydraulic properties of the soil. The decrease in infiltration rate leads to surface run off, which enhances soil erosion particularly in areas with intensive rainfall [14]. This also increases the risk of flooding, particularly in areas with steep slopes that experience intensive rainfall [15]. The increase of soil resistance to penetration affects not only plant growth but also leads to increased energy requirement for tillage. Therefore, the occurrence of soil compaction should be avoided otherwise a proper management of tillage should be utilised. Chamen and Longstaff 1995 and Sedaghatpour *et al.*, 1995 [16,17] noted a

decrease in soil BD and PR when adopting CTF. Botta *et al.*, 2007 [18] observed peak PR and BD under the third year of RT, reaching levels up to 3890 kPa and 1.72 Mg m⁻³, respectively, as compared to corresponding CTF peaks of 2556 kPa and 1.53 Mg m⁻³. As a result of compaction amelioration by the adoption of CTF, Meek *et al.*, 1989 and 1990 [19,20] reported improvement of the infiltration rate in cotton production. No tillage combined with CTF retained the highest infiltration rates during the following cotton season, which was probably due to better preservation of macro pores [20]. Similarly, McHugh *et al.*, 2003 [21] experienced a four-fold increase of hydraulic conductivity at 10 cm depth after 22 months of CTF practice in broad acre arable land in Australia. The effect of CTF on water erosion was examined in a number of studies mainly in Australia. Results ranged from a reduction in runoff reaching to 36.3%, which increased to 47.2% when CTF was combined with no tillage (Li *et al.*, 2007), to almost no change in runoff at all [22]. Tullberg *et al.*, 2001 [23] found 44% larger runoff from trafficked plots than from CTF plots. Concerning biodiversity, reports confirmed CTF to increase the number of macrofouna. Pangnakorn *et al.*, 2003 [24] found the number of earthworms to be larger under CTF than under RT, with the highest number tending to occur under no tillage combined with CTF. Literature has not proved that CTF can alleviate soil organic matter (SOM) decline. Potter and Chichester (1993) [25] concluded that no tillage combined with CTF appears to be a longterm sustainable solution. They showed an increase in soil organic carbon when soil was not disturbed by primary tillage. Steadily greater nitrogen content in the CTF swards has resulted in more efficient use of fertiliser than in conventionally farmed plots. As proportion of the nitrogen fertiliser applied, the nitrogen content of the yield was 86% and 64%, for CTF and RT, respectively, nonetheless a reduced ground pressure treatment was almost as good as CTF with 81% of the nitrogen applied being harvested [26]. Soil compaction has a large influence on N₂O emission, mainly due to increased anaerobic conditions [27,28]. However, the quantity of released gas is also related to crop and environmental conditions [29]. Vermeulen *et al.*, 2007 [30] evaluated the impact of seasonal controlled traffic farming (SCTF) on GHG emission, adopting RT during harvest and primary tillage followed by all the other traffic in controlled traffic lanes. They concluded that SCTF resulted in reduction of both N₂O and CH₄ emission as compared to total RT farming.

Models to evaluate the impact of tillage and traffic on crop production are available. For instance the advanced PERFECT model was used by Li *et al.*, 2008 [31,32] to predict runoff, plant available water, infiltration rate and yield. Using PERFECT, Li *et al.*, 2008 [31,32] ranked the benefits of tillage systems to be in the following descending order: controlled traffic with zero tillage, controlled traffic with stubble mulch, wheeled with zero tillage, and wheeled with stubble mulch. However, this model does not include all environmental benefits and does not rank them among each other. Furthermore, no overall scoring model for the entire environmental benefits of CTF can be found in the literature. Hamza and Anderson (2005) [33] only listed a range of environmental benefits, without ranking them according to their weight. It can be concluded from the brief literature review that although reports support the environmental benefits of CTF, a quantification of the overall environmental benefits of CTF has not been reported so far.

This study aims to develop a framework that predicts the overall environmental benefits of CTF for farms planning to convert to CTF. This framework considers soil and climate conditions, topography and the current machinery system. The framework is validated using three example farms, whose data were obtained from Unilever (UK) (one farm) and from the literature (two farms).

2. Materials and Methods

2.1. Framework Weighting

When establishing the prediction framework to rank the environmental benefits of CTF, several environmental threats were identified as key threats needed for the successful development of the matrix. The development of the prediction framework was based on three weighting systems, namely, environmental, scientific, and farm specific parameters (Figure 1). The framework was established based on a number of assumptions:

- Environmental weight: Here 75% of weighting was assigned to soil threats and 25% for secondary issues. The importance of each of the soil threats was obtained from answers to a survey questionnaire carried out by the EU.
- Scientific weight: Here also 75% of weighting was assigned to soil threats and 25% for secondary issues. The importance of each soil threat was evaluated based on the number of studies found in the literature confirming the positive or negative influence of the CTF relative to the random traffic. The magnitude of CTF benefit was calculated based on the % of benefit (e.g., in energy saving), as compared to random traffic.
- Probability of occurrence and scale of problem is estimated by ten parameters, which are given two levels of internal weight/influence.

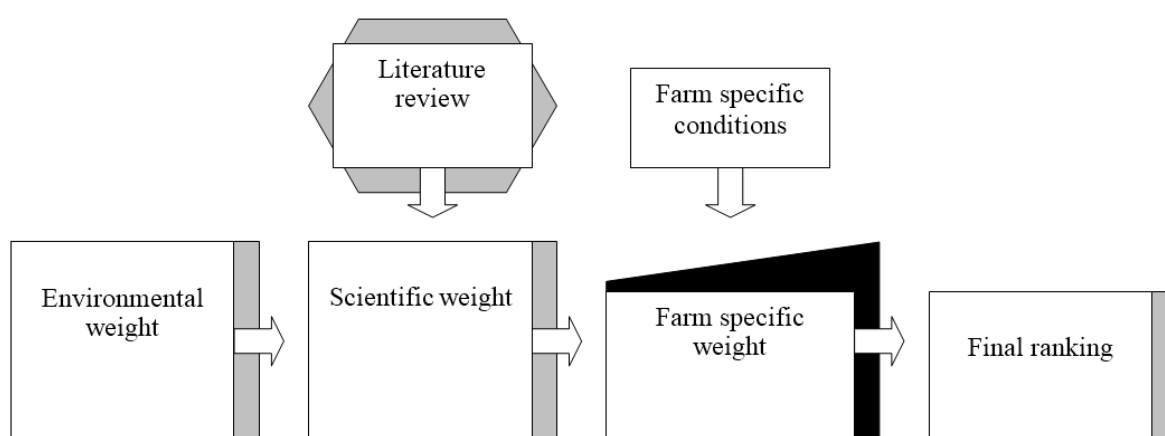


Figure 1. The general structure of the framework for the calculation of the environmental benefits of controlled traffic farming (CTF).

The framework was built using Microsoft Excel, to ensure that the framework was flexible and easily amendable, as a complex model is harder to develop, understand, and validate [34].

Furthermore, a critical requirement when developing scoring functions in functional models is to document why a particular value was chosen as baseline or threshold value. Therefore, a thorough search and documentation of farm specific parameters and their ranges to be considered in this framework was undertaken.

Environmental Weighting. Under the Thematic Strategy for Soil Protection, a soil framework directive prepared by the European Commission (EU Framework Directive 232, 2006) [35] identified nine soil threats, namely, soil compaction, soil erosion, organic matter decline, contamination,

salinisation, soil biodiversity loss, sealing, landslides, and flooding. It was prepared after a web survey provided by soil scientists and research organisations. In this study, only environmental threats that are directly affected by implementing CTF were considered. These were ranked in the Framework Directive in dissenting order as; soil compaction, soil erosion, soil biodiversity loss, and soil organic matter decline. This ranking system was incorporated into the framework, and these soil threats were considered as the main environmental parameters.

Other parameters with indirect environmental benefits of CTF were identified as secondary environmental parameters, such as reduced GHG emissions from soil, reduced energy requirement of tillage, and increased fertiliser efficiency. However, these secondary parameters were not included in the EC Soil Framework Directive. The secondary parameters were assigned a weighting of 25% of the total environmental weighting to be divided equally among them, whereas the remaining 75% weighting was assigned to the main environmental parameters (soil threats). The structure of the environmental weighting part of the framework is illustrated in Figure 2.

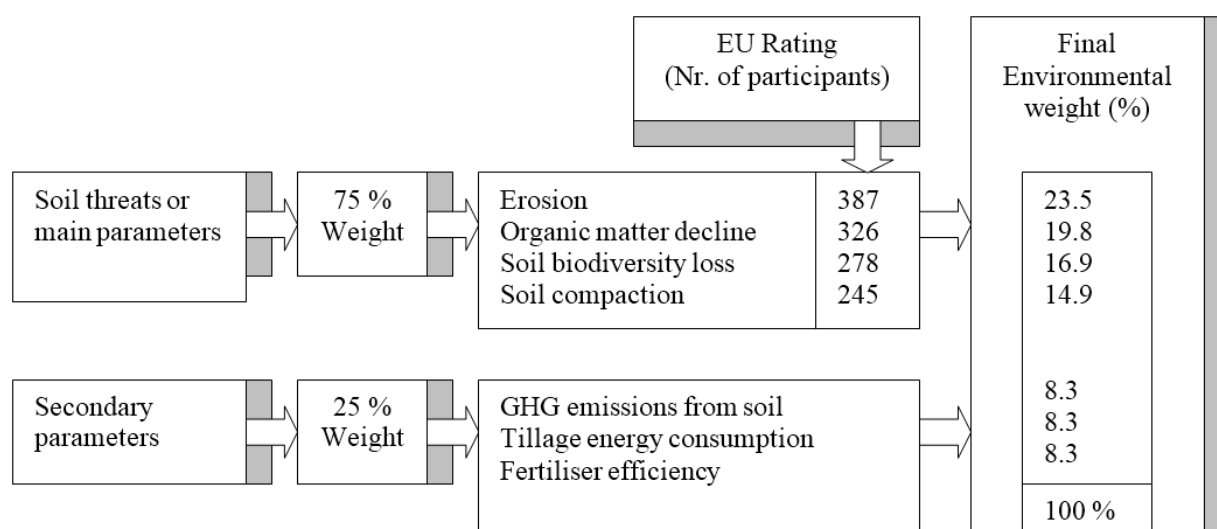


Figure 2. Environmental weighing system of prediction framework of environmental benefits of controlled traffic farming (CTF).

Scientific Weighting. A comprehensive literature review on CTF has been carried out to evaluate the scientific weighting based on the number of studies reported with positive effect in reducing the level of any of the main or secondary environmental parameters. Many journal papers, reports, and books on CTF were closely examined. However, a much larger number of studies available on e.g., soil compaction under RT can be found in the literature, which was not included in the scientific weighting of the framework. The reader is referred to the Appendix for more details about the studies used to calculate the weighting factors.

Studies about the environmental benefits of CTF were categorised under one soil threat and/or secondary parameter. The number of studies reported about a parameter was adopted to determine the scientific weight of that parameter. Some studies considered different parameters to represent a threat, e.g., soil compaction was indicated by BD, PR and/or infiltration rate. When the same study considered multiple compaction measures, the study was weighted once in the framework to eliminate double or triple weighting of a given parameter. Again the main environmental parameters were given

75% of the total scientific weighting, whereas the secondary parameters were given a total weighting of 25%. The structure of the scientific weighting part of the framework is illustrated in Figure 3.

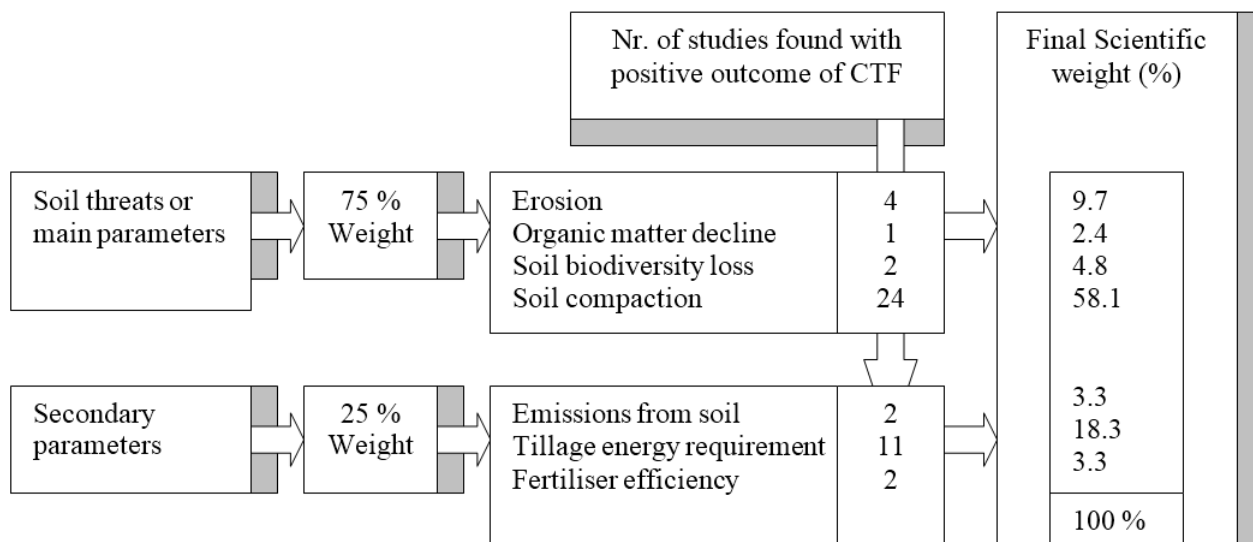


Figure 3. Scientific weighing system of prediction framework of environmental benefits of controlled traffic farming (CTF).

Weighting of Farm Specific Parameters. The parameters considered as farm specific parameters were those related to selected soil properties, climate conditions, topography, and existing machinery system in the farm. These were soil moisture content at traffic and at tillage, annual rainfall, slope length and steepness, machine parameters, present tillage system (no tillage, reduced tillage and conventional tillage), soil texture, BD at traffic and at tillage, and SOM content. Each parameter was given a weighting factor ranged between 0 and 10, which affects the final ranking of the environmental benefits. The structure of the farm specific part of the framework is illustrated in Figure 4. Each farm specific parameter was assigned one or two +, based on its impact on soil threats and on secondary parameters (Table 1).

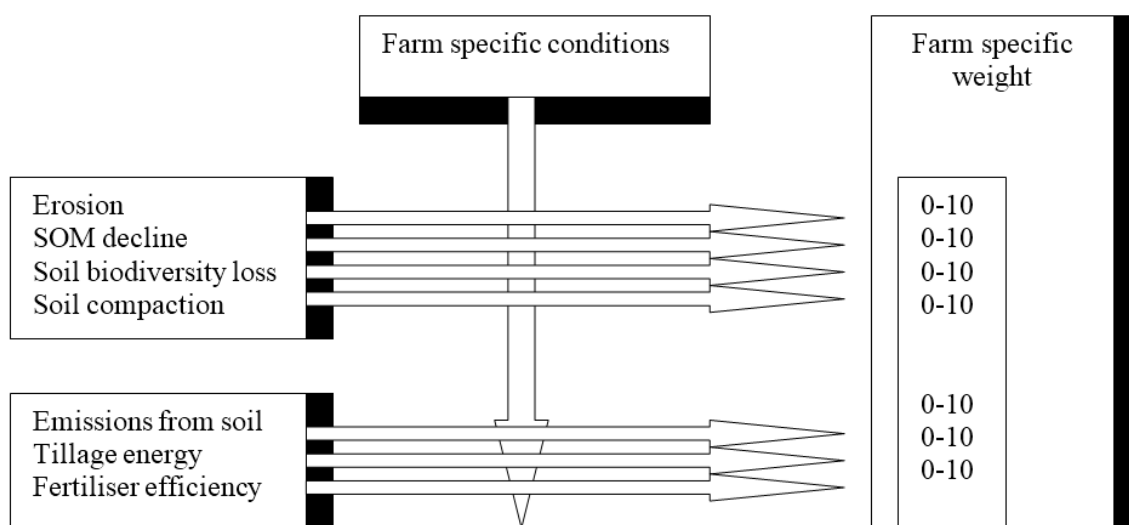


Figure 4. The structure of framework part related to farm specific parameters.

Table 1. Impact of farm specific parameters.

| Item | Moisture at traffic | Moisture at tillage | Rainfall | Slope | Machine parameter | Present tillage system | Soil texture | Bulk density in traffic zones | Bulk density in tillage zones | Soil organic matter |
|------------------------------------|---------------------|---------------------|----------------|-------|-------------------|------------------------|--------------|-------------------------------|-------------------------------|---------------------|
| Soil erosion | | | + ¹ | + | | + | + | | | + |
| soil organic matter decline | | | | | | + | + | | | |
| Soil biodiversity loss | | | | | | ++ ² | + | | | + |
| Soil compaction | ++ | | | | ++ | | + | + | | + |
| Emissions from soil | | | ++ | | | | | ++ | | |
| Tillage energy | | ++ | | | | ++ | + | | ++ | |
| Fertiliser efficiency ³ | | | | | | | | | | |

¹ Minor influence (+); ² Major influence (++); ³ No specific influence on fertiliser efficiency found. Soil organic matter (SOM)

A parameter having a major influence scored two +, whereas a parameter having minor influence scored one + only. This impact categorisation was based on information available in the literature concerning CTF, as described here below.

Moisture Content at Traffic and at Tillage. The effect of soil moisture content on compaction occurrence is difficult to predict, since this is also dependent on soil texture, BD, and organic matter content. Based on the Proctor test, literature has shown optimal moisture contents for compaction occurrence, which varies with soil texture [36]. For simplicity, gravimetric soil moisture content was classified, with a linear scale into four different categories. In general, the wetter the soil at traffic, the lower is the soil resistance to compaction occurrence. Also, the dryer the soil at tillage, the higher is the draught and energy requirement of tillage. Based on these four categories a simple classification system of the framework impact points for moisture content, similar to that proposed by Spoor *et al.*, 2003 [37] was adopted (Table 2).

Table 2. Impact points of moisture content.

| Wetness Condition | Impact Points |
|---|----------------------|
| Wet (close to field capacity) | 10 |
| Moist | 7 |
| Dry (approaching permanent wilting point) | 4 |
| Very dry | 0 |

If the framework is to be more precise in defining the exact trend of compaction occurrence according to soil moisture content during traffic for different soil textures, a more complicated function that correlates soil compaction occurrence with moisture content has to be adopted based on experimental data. Nonetheless, both moisture contents at traffic and at tillage are classified according to Table 2, with the moisture content at traffic has a negative effect on soil compaction occurrence, whereas soil moisture content at tillage has a positive effect on energy consumption due to tillage operations.

Rainfall. The mean annual precipitation was considered in the framework as the factor affecting erosion. High precipitation can also lead to increased soil moisture content, which in turn leads to increased GHG emission under anaerobic conditions. The scale of conversion proposed was linear with the exception of a smaller step at the lowest precipitation to distinguish the places with very low rainfall (Table 3).

Table 3. Impact points of rainfall.

| Mean annual Precipitation (mm) | Impact Points |
|---------------------------------------|----------------------|
| >2000 | 10 |
| 15,000–2000 | 8 |
| 1000–1500 | 6 |
| 500–1000 | 4 |
| 250–500 | 2 |
| Under 250 | 0 |

Slope. The impact points for slope (SL) designated as slope factor (SF) were calculated in a similar way to that of the universal soil loss equation (USLE) method, by substituting the values of average slope steepness and length into Equation (1).

$$(SF) = (0.065 + 0.045S + 0.0065S^2) \times (L/22)^n \quad (1)$$

where: SF is slope factor, S is slope steepness (%), L is slope length (m) and n is a constant calculated according to the slope value (Table 4)

Table 4. Values of constant n for different values of slope.

| Slope (%) | <1 | 1 < Slope < 3 | 3 < Slope < 5 | >5 |
|-----------|-----|---------------|---------------|-----|
| n | 0.2 | 0.3 | 0.4 | 0.5 |

The final weighting scale was transformed into a framework of impact points using SF values calculated from Equation (1), and reported in Table 5.

Table 5. Impact points of slope.

| Slope Factor | Impact Points |
|--------------|---------------|
| > 1.5 | 10 |
| 1.0–1.5 | 7 |
| 0.5–1.0 | 4 |
| < 0.5 | 0 |

Machine Parameters. No precise relationship between axle load and soil compaction exists, probably because of the fact that there are many other factors affecting soil compaction occurrence. However, practical experience (thanks to the anonymous reviewer) showed a large axle load of 18 Mg to cause severe yield losses (at least 10%), whereas a 9 Mg load caused only short term damage to the soil and minor yield loss. Therefore, farms with light loads are less likely to benefit from CTF and less likely to adopt it regardless of environmental benefits.

The relative importance of tyre inflation pressure and axel load has been disputed for a long time [38]. The combination determines the compaction level, although the severity of compaction declines with depth. However, it is a challenge to create a simple model that can predict the actual level of compaction, as related to axle load and inflation pressure, since there is still an inter-correlation between those factors and other factors [38]. In the framework, the impact of the axle load and tyre inflation pressure during one crop cycle was considered.

The scale for this framework was based on a simple “rule of thumb” that the soil can withstand compaction damage from axle loads up to 6 Mg and 7.5 Mg at inflation pressure values of 150 kPa and 100 KPa, respectively [39]. From those two values, a linear regression was carried out to derive a matrix (not shown) about the impact of both tyre inflation pressure and axle load, which was used to derive the framework impact points.

Present Tillage System. The tillage system affects several soil threats and secondary parameters in the framework. In general, the higher the tillage intensity, the larger is the effect. Table 6 shows the framework impact points considered for different tillage systems. Conventional tillage represents primary tillage by mouldboard or disc plough followed by discs and/or tine harrows *etc.* and drilling

with at least three passages. Reduced tillage involves shallow tillage followed by drilling, whereas no tillage consists of only direct drilling of soil.

Texture. The classification of texture considered in this study was based on FAO standard triangle soil texture classification. A scale of soil vulnerability to compaction proposed by Spoor *et al.*, 2003 [37] formed the basis to establish impact points of different textures (Table 7). For example, clay soils were considered less vulnerable to compaction when compared to sandy soils at a same moisture content level.

Table 6. Impact points of tillage system.

| Tillage System | Impact Points |
|----------------------|---------------|
| Conventional tillage | 10 |
| Reduced tillage | 5 |
| No tillage | 0 |

Table 7. Impact points of soil texture.

| Texture (FAO–UNESCO) | | Impact Points |
|----------------------|-------------------------|---------------|
| | Coarse | |
| Light soils | Medium (<18% clay) | 10 |
| | Medium fine (<18% clay) | |
| Medium soils | Medium (>18% clay) | 5 |
| | Medium fine (>18% clay) | |
| Heavy soils | Fine | 0 |
| | Very fine | |

Bulk Density at Traffic and at Tillage. Soil BD is a measure of soil compaction, which unlike PR is independent on moisture content at sampling [40]. Therefore, BD was adopted, since it is a more reliable measurement for establishing the scale of impact points. Furthermore, soil texture had to be taken into account, since BD is a function of soil texture [41]. The impact points scale, provided in Table 8 was based on a guideline for seedbed compaction [42], where BD has a different impact depending on the soil type considered. The limiting BD for root growth generally ranges from 1.45 Mg m⁻³ for clay soils to 1.85 Mg m⁻³ for loamy sands [41], which proves that Table 8 is of the correct magnitude.

Table 8. Impact points of bulk density (BD).

| Bulk Density (Mg m ⁻³) | Impact Points | | |
|------------------------------------|---------------------|----------------|--------------------|
| | Light (Sandy) Soils | Medium (Loams) | Heavy (Clay) Soils |
| > 1.8 | 10 | - | - |
| 1.6–1.8 | 8 | 10 | 10 |
| 1.4–1.6 | 5 | 5 | 8 |
| 1.2–1.4 | 1 | 3 | 5 |
| < 1.2 | 0 | 1 | 3 |

Soil Organic Matter. The “base limit” of SOM was correlated with clay content, so that the minimal SOM content tended to rise with clay content [43]. However, very limited data are available to make a texture based scale for SOM, and hence a simple linear scale was proposed (Table 9).

Table 9. Impact points of soil organic matter (SOM).

| Soil Organic Matter Content (%) | <0.5 | 0.5–1.5 | 1.5–2.5 | 2.5–3.5 | 3.5–4.5 | 4.5< |
|---------------------------------|------|---------|---------|---------|---------|------|
| Impact points | 10 | 8 | 6 | 4 | 2 | 0 |

Erosion Factor Using USLE. The USLE is an empirical model to calculate soil erodability as a function of dimension and dimensionless factors. However, if metric units are used to obtain the rainfall erosivity factor, a metrical value of soil loss will be derived according to the following equation [44].

$$E = R \times K \times L \times S \times C \times P \quad (2)$$

where: E is the mean annual soil loss according to USLE (Mg ha^{-1}), R is the rainfall erosivity factor ($\text{MJ ha}^{-1} \text{mm h}^{-1}$), K is the soil erodibility factor, $L \times S$ is the slope factor, C is the crop management factor and P is the erosion control practice factor.

A USLE value estimating soil loss for a site or a field can be transformed into a framework point, as shown in Table 10, which was constructed after a classification proposed by Morgan (2005) [44]. When USLE is to be considered, the individual weight of the other factors involved in the calculation of the erosion factor has to be omitted from the matrix. This included rainfall, slope, soil organic matter and texture. In this case, the impact of moisture content at traffic and at tillage, machine parameters, present tillage system, bulk density at traffic and at tillage, and erosion factor using USLE were only considered.

Table 10. Impact points of universal soil loss equation (USLE).

| Soil Loss According to USLE (t ha^{-1}) | Impact Points |
|--|---------------|
| >10 High | 10 |
| 5–10 Moderate | 7 |
| 2–5 Slight | 4 |
| <2 Very slight | 0 |

2.2. Effect of Farm Specific Parameters on Environmental Threats and Secondary Parameters

Soil Erosion. Erosion is a complicated environmental threat, whose occurrence is affected by many parameters. Morgan (2005) [44] identified a number of parameters affecting soil erosion, namely, rainfall amount and intensity, soil structure and texture, SOM, permeability, slope length and steepness, land use, and land cover. In the framework presented in this paper, all farm specific parameters, were considered equally to have a minor influence (one +) on erosion. However, since there already existed models on erosion prediction e.g., USLE [44] it would probably be more scientifically sound to use the USLE model, expressed as an erosion factor. This is particularly because of the fact that the internal weighting of parameters in USLE is more advanced and precise than that proposed for the specific farm parameters considered in the framework. Therefore, the calculation of erosion occurrence in the excel sheet matrix allowed for the use of both either the farm specific parameters or the erosion factor based on the USLE model. However, due to its relative simplicity the former option might be a better option for farmers.

Soil Organic Matter Decline. It is a challenge to find consistent data on factors affecting SOM decline. This is because some researchers suggested that no tillage systems might increase SOM content [45], whereas others stated that SOM is just relocated with depth [46,47]. However, the predominant result after a comprehensive evaluation of a large number of studies worldwide suggested that no tillage and reduced tillage systems, in descending order, increase SOM [48]. In conclusion, the present tillage system was regarded in the present study to have a minor influence (one +) on SOM decline. Even though, no literature about a high declining rate of SOM was found for any soil texture, texture was regarded as having a minor effect on SOM decline in the framework. This is because of the important water holding capacity and aggregate building capability and stability of SOM, which is particularly important for sandy soils. In fact, SOM decline is a larger problem on sandy than on clayey soils.

Soil Biodiversity. The incidence of occurrence of earthworms was studied by Pangnakorn *et al.*, 2003 [24], who found that absence of tillage had a large positive effect on the occurrence. Also, fungal and bacterial biomasses were higher in surface soil when less disturbed by tillage [49]. Thus, the present tillage system was regarded as being a major influencing factor on soil biodiversity. In a model used by Fox *et al.*, 2004 [50], the number of earthworms was positively correlated with increasing SOM and clay content. Consequently, soil biodiversity decline can be regarded as being a bigger problem in sandy soils with low organic matter. Therefore, texture and SOM were considered to have a minor influence on soil biodiversity in the framework.

Soil Compaction. The fact that soil compaction is, to a large extent, affected by the moisture content of the soil during traffic and by machine parameters (size, tyre inflation pressure, *etc.*) is well accepted and also supported by researchers [33]. Therefore, these parameters were regarded to have the major effect (two +) on soil compaction. Since SOM has marginal influence on compaction [33] it was considered to have only a minor effect in the framework. The vulnerability to soil compaction is also, to some extent, affected by soil texture [37,51]. However, texture was considered to have a minor effect in the framework. Initial BD at traffic also affects the degree of soil compaction occurrence. Furthermore, a low BD at traffic that occurs after tillage makes soil more vulnerable to compaction [52]. Therefore, BD at traffic was considered to have a minor influence (one +) in the framework. The same arguments apply to the present tillage system. The higher the intensity of tillage the higher is the soil vulnerability to compaction. However, it is not common to use these three tillage systems worldwide. For instance, Munkholm *et al.*, 2003 [53] indicated the low adoption of reduced tillage systems and no tillage in Scandinavia, which was partly attributed to major problems associated with compaction of the topsoil. Therefore, no influence of the tillage system on soil compaction occurrence was assigned in the framework.

Greenhouse Gas (GHG) Emission from Soil. Since soil compaction and soil moisture content have a large effect on N₂O emissions from soil [29], soil BD at traffic and annual rainfall were regarded as having a major influence (two +) on GHG emissions from soil. The relationship between emissions and tillage practices is not clear and rather complicated. The common idea is that a no tillage system sequesters carbon but this might be offset by a higher level of N₂O emissions [29,54]. Therefore, no influence of the tillage system on GHG emission was assigned in the framework.

Energy Requirement of Tillage. Since a positive correlation between draught requirement and BD exists [55], BD at traffic was regarded as a major factor affecting energy consumption of tillage in the framework. The tillage system has also a major influence on the energy required [56], since more intensive

tillage requires higher energy. It has been demonstrated that a higher moisture content resulted in a lower draught [55,57]. Thus, soil moisture content at tillage was considered to be a major influencing factor. Since clay soils (heavy soils) require higher energy to break up during tillage than sandy soils (light soils) [43], soil texture was considered to have a minor effect on energy consumption by tillage.

Fertiliser Use Efficiency. The benefit of fertiliser use efficiency was given an average value of five, which means that the benefit of CTF is regarded equally high regardless of how the user specifies the other parameters.

Construction of the Framework. The two different weighting parts, namely, environmental and scientific were combined in the excel-sheet framework, as shown in Figure 5. The percentage for each individual benefit was then multiplied by the weighting of the farm specific parameters (0–10). The sum of the individual values after dividing by 10 becomes 92.3%.

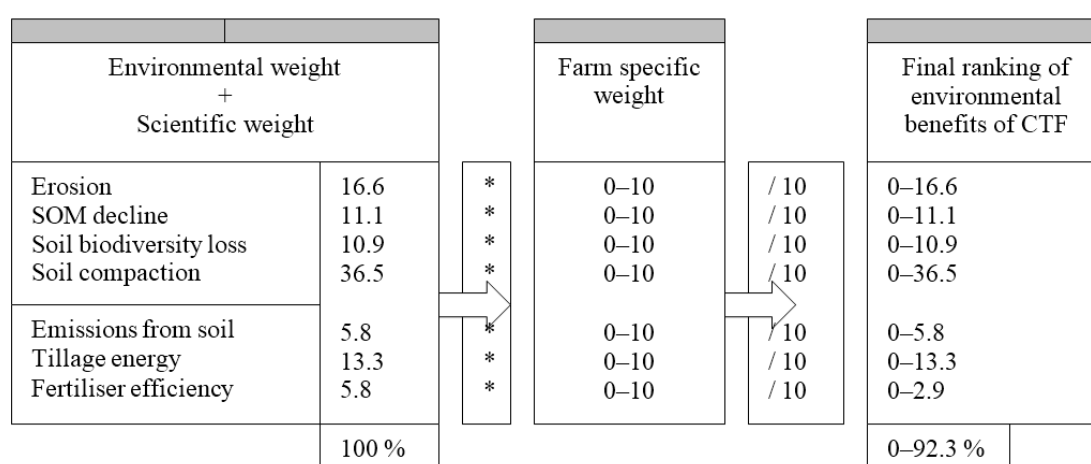


Figure 5. Calculation of the overall ranking of the framework.

This percentage was considered to be the maximum achievable environmental benefit at optimal soil and climate conditions, for topography and machinery systems for eliminating or reducing probability of occurrence of soil threats, and for secondary environmental parameters. For example, light rain, small slopes, small size agricultural machinery, stable aggregate soils and texture, *etc.* would result in reducing the probability of the occurrence of soil threats (e.g., soil erosion or compaction). However, the prevention of occurrence of soil threats is not practically possible and thus the maximum benefit level of 92.3% is not achievable. Therefore, this maximum benefit level was designated as “the theoretical maximum environmental benefit of CTF”.

The constant impact points assigned to the fertiliser efficiency (5), and the contradictory positive and negative impacts of soil moisture content during tillage and compaction is the reason for having a theoretical maximum of 92.3% instead of 100%. To overcome this issue, each individual benefit was rescaled in order to have the final evaluation presented to 100%, allowing the evaluation to range between 0 and 100%.

3. Sensitivity Analysis

When all farm specific parameters in the framework were assumed to be of an average impact of five (Figure 5), the overall environmental benefit of CTF becomes 54.2% of the theoretical maximum

(100%). Relatively large differences of contribution from each benefit were revealed e.g., impact of reduced soil compaction was 19.8%, controlling erosion was 9.0%, reduced tillage energy consumption was 7.2%, declining impede organic matter was 6.0%, enhanced soil biodiversity was 5.9%, improved fertiliser efficiency was 3.2%, and reduced GHG emissions was 3.2%. The sensitivity analysis revealed that the magnitude of weighting seems to have been correctly assumed. This is because of the fact that the results of the sensitivity analysis were in line with those reported in the literature. Generally, reduced soil compaction, reduced tillage energy requirement, and controlling soil erosion are, according to the literature, the top three environmental benefits expected from CTF.

4. Validation of the Framework

Validation of the framework was conducted using two data sets obtained from the literature. The first data set was reported by Dickson *et al.*, 1992 [58] and Dickson and Ritchie (1996 and 1996) [59,60], investigating the effect of no-traffic and reduced ground pressure traffic systems in an arable rotation in Scotland, whereas the second data set was reported by Li *et al.*, 2007 [61], investigating the effect of wheel traffic and tillage on runoff and crop yield in Australia. A third validation data set was obtained from Colworth farm (Unilever, UK), where CTF trial fields to investigate the environmental and economic benefits of CTF were established.

4.1. Validation Study 1

When the farm specific parameters of Dickson *et al.*, 1992 [58] and Dickson and Ritchie (1996 and 1996) [59,60] (Table 11), were substituted into the framework as shown in Table 12, the framework predicted an overall environmental benefit from CTF of 52.0% of the theoretical maximum, which is below the average environmental benefit of CTF of 54.2%.

Table 11. Farm specific parameters from Dickson *et al.*, 1992 [58] and Dickson and Ritchie (1996 and 1996) [59,60].

| Parameter | Impact Point |
|--|------------------|
| Moisture content at traffic | Moist (assumed) |
| Moisture content at tillage | Moist (assumed) |
| Annual rainfall (mm) | 845 |
| Average slope steepness (%) | 6.7 |
| Slope length (m) | 100 (assumed) |
| Highest axle load (Mg) | 4.59 |
| Tyre inflation pressure (kPa) | 160 |
| Present tillage system | Conventional |
| Soil type | Clay loam (fine) |
| Bulk density at traffic (Mg m^{-3}) | ~1.5 |
| Bulk density at tillage (Mg m^{-3}) | ~1.5 |
| Soil organic matter (%) | 4.1 |

Table 12. Framework impact points for validation study 1.

| Parameter | Impact Point |
|--|--------------|
| Moisture content at traffic | 7 |
| Moisture content at tillage | 7 |
| Annual rainfall (mm) | 4 |
| Slope factor | 7 |
| Machine parameters | 3 |
| Present tillage system | 10 |
| Soil type | 0 |
| Bulk density at traffic (Mg m^{-3}) | 8 |
| Bulk density at tillage (Mg m^{-3}) | 8 |
| Soil organic matter content (%) | 2 |

This overall benefit is broken down into percentage units in descending order as: reduced soil compaction (13.6%), reduced tillage energy consumption (10.7%), controlling erosion (8.3%), enhanced soil biodiversity (6.5%), declining impeded organic matter (6.0%), reduced GHG emissions (3.8%) and improved fertiliser efficiency (3.2%).

4.2. Validation Study 2

When the farm specific parameters of Li *et al.*, 2007 [61] of Table 13 were substituted into the framework as shown in Table 14, the framework predicted an overall environmental benefit from CTF of 47.3% of the theoretical maximum, which is also below the average environmental benefit of CTF of 54.2%. This overall benefit is broken down into percentage units in descending order as: reduced soil compaction (10.2%), reduced tillage energy consumption (9.5%), controlling erosion (8.3%), enhanced soil biodiversity (7.4%), declining impeded organic matter (6.0%), improved fertiliser efficiency (3.2%), and reduced GHG emissions (2.9%). Li *et al.*, 2007 [61] did not examine all environmental benefits of CTF but only runoff, soil water, and crop production. They measured a reduced mean annual runoff of 36.3% with CTF, which increased to 47.2% when combined with no tillage practices. Therefore, the fact that the framework rates control erosion as the third most important environmental benefit may be argued. However, the benefit of controlling erosion contributed to a larger part of the overall benefit (8.3% of the total benefits of 47.3%), as compared to validation Study 1 (8.3% of the total benefits of 52.0%).

Table 13. Farm specific parameters [61].

| Parameter | Impact Point |
|-------------------------------|-----------------|
| Moisture content at traffic | Dry (assumed) |
| Moisture content at tillage | Moist (assumed) |
| Annual rainfall (mm) | 789 |
| Average slope steepness (%) | 7 |
| Slope length (m) | 30 |
| Highest axle load (Mg) | 4.5 |
| Tyre inflation pressure (kPa) | 100 |

Table 13. *Cont.*

| Parameter | Impact Point |
|--|-------------------------------------|
| Present tillage system | Conventional |
| Soil type | Vertisol (fine) |
| Bulk density at traffic (Mg m^{-3}) | Assumed average due to lack of data |
| Bulk density at tillage (Mg m^{-3}) | Assumed average due to lack of data |
| Soil organic matter (%) | Assumed average due to lack of data |

Table 14. Framework impact points for validation study 2.

| Parameter | Impact Point |
|--|--------------|
| Moisture content at traffic | 4 |
| Moisture content at tillage | 7 |
| Annual rainfall (mm) | 4 |
| Slope factor | 4 |
| Machine parameters | 0 |
| Present tillage system | 10 |
| Soil type | 0 |
| Bulk density at traffic (Mg m^{-3}) | 5 |
| Bulk density at tillage (Mg m^{-3}) | 5 |
| Soil organic matter content (%) | 5 |

4.3. Validation Study 3

When the farm specific parameters of Colworth farm (Unilever, UK) shown in (Table 15) were substituted into the framework as described in Table 16, the framework predicted an overall environmental benefit from CTF of 59.3% of the theoretical maximum, which was higher than the average environmental benefit of CTF of 54.2%. This overall benefit is broken down into percentage units in descending order as: reduced soil compaction (24.3%), reduced tillage energy consumption (9.5%), enhanced soil biodiversity (7.1%), controlling erosion (6.5%), declining impeded organic matter (6.0%), improved fertiliser efficiency (3.2%), and reduced GHG emissions (2.9%). Alternatively, if the erosion factor of USLE soil loss value is used for weighting controlling erosion instead, an increase of benefit of control erosion by 0.7% is expected. This would raise the overall score of controlling erosion to 7.2%, overtaking soil biodiversity but leaving the remaining ranking of all other benefits unchanged.

Table 15. Farm specific parameters of Colworth farm (Unilever, UK).

| Parameter | Impact Point |
|-------------------------------|--------------|
| Moisture content at traffic | Moist |
| Moisture content at tillage | Moist |
| Annual rainfall (mm) | 611 |
| Average slope steepness (%) | 2 |
| Slope length (m) | 270 |
| Highest axle load (Mg) | 11 |
| tyre inflation pressure (kPa) | 160 |

Table 15. *Cont.*

| Parameter | Impact Point |
|--|-----------------|
| Present tillage system | Conventional |
| Soil type | Fine, very fine |
| Bulk density at traffic (Mg m^{-3}) | 1.2–1.4 |
| Bulk density at tillage (Mg m^{-3}) | 1.2–1.4 |
| Soil organic matter (%) | 2.5–3.5 |

Soil loss according to USLE is $2\text{--}5 \text{ Mg ha}^{-1}$, which represents an erosion factor 4.

Table 16. Framework impact points for Colworth farm.

| Moisture Content at Traffic | Impact Point |
|--|--------------|
| Moisture content at traffic | 7 |
| Moisture content at tillage | 7 |
| Annual rainfall (mm) | 4 |
| Slope factor | 0 |
| Machine parameters | 10 |
| Present tillage system | 10 |
| Soil type | 0 |
| Bulk density at traffic (Mg m^{-3}) | 5 |
| Bulk density at tillage (Mg m^{-3}) | 5 |
| Soil organic matter content (%) | 4 |

The two largest benefits of CTF at Colworth farm (Unilever, UK), namely, reduced soil compaction and tillage energy consumption were also the highest ranked benefits in the framework due to their higher environmental and scientific weightings. Furthermore, the current research available in the literature confirms that these two benefits are of the highest importance. In fact, the reduction of soil compaction is interlinked with all other benefits, which also highlights the complexity of ranking the framework.

Among the three data sets, it is clear that the highest environmental benefit is expected at Colworth farm. The datasets used for the validation, based on Dickson *et al.*, 1992 [58] and Dickson and Ritchie (1996) [59,60] and Li *et al.*, 2007 [61] were generated in Scotland and Australia, respectively, representing two different farming, weather, topography, and soil conditions. This makes direct comparison between the two data sets very difficult. However, one factor that clearly differentiates these two conditions from those of Colworth farm was the high maximum axle load of machinery used under random traffic at Colworth farm. Consequently, the higher benefit of reduced compaction by CTF at Colworth farm as compared to the other two data sets with smaller axial load machinery is justified.

Although this study ranked environmental benefits in a prediction framework, it is recognised that many benefits of CTF tend to be relatively small, which corresponds well with a report from Spoor (1997) [5]. However, when all these small benefits are combined, they might add up to a significant overall benefit. The adoption of CTF depends on the overall benefits (including ecological and economic benefits) and not only on the environmental benefits. Ecological and economical evaluations of CTF have to be combined with environmental evaluation to achieve an overall evaluation system, which would encourage farmers to convert into CTF.

5. Conclusions

The evaluation of the overall environmental benefit of adopting CTF was successfully carried out based on a framework developed. The study confirms the following conclusions:

- (1) There is a substantial evidence in the literature of environmental benefits associated with CTF, which justifies the adoption of CTF, recognizing that even small benefits might add up to a significant sum of environmental benefits.
- (2) A framework for predicting the environmental benefits for farm converting to CTF is feasible to construct and implement. The framework is flexible and can be further modified responding to changes in local and/or national agricultural practices.
- (3) Literature is in line with the frameworks result, confirming that, reducing compaction, reducing tillage energy consumption and controlling soil erosion are the most pronounced environmental benefits of CTF.
- (4) The new framework introduced in this work can be implemented in practice today. Farmers wishing to implement CTF can in advance evaluate the environmental benefits of adopting CTF in their farms with specific practical and environmental conditions. Policy makers can also benefit from this framework by convincing farmers of the environmental benefits that they would expect if they were to adopt CTF.

Further research might be needed to determine the exact impact range of different environmental benefits of CTF subjected to availability of experimental data. This would make it possible to refine the framework, and convince farmers to convert to CTF, when economical benefits are also proved.

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Author Contributions

The research in the current work was carried out by master student Martin Palmqvist with supervision of Abdul Mouazen. Palmqvist performed the literature review and data analysis, while the paper was written by Mouazen, based on the master thesis of Palmqvist. Both authors have read and approved the manuscript.

Conflicts of Interest

The authors declare no conflict of interest.

Appendix

Table A1. Weighting factor calculated for soil compaction comparing trafficked with un-trafficked practices. Compaction indicators included bulk density, penetration resistance and infiltration rate.

| Measurement | Results | % of Relative Improvement | Study |
|---------------------|------------------------|---------------------------|--|
| Bulk density | CTF positive | 90 | Botta <i>et al.</i> , 2007 [18] |
| | CTF positive | 83–91 | Campbell <i>et al.</i> , 1986 [62] |
| | CTF positive | 81–100 | Chamen and Longstaff, 1995 [16] |
| | CTF positive | 93 * | Unger, 1996 [63] |
| | CTF positive | 90 | Dickson and Ritchie, 1996 [60] |
| | CTF positive | 91 | McHugh <i>et al.</i> , 2003 [21] |
| | CTF positive | 83 * | Raper and Reeves, 2007 [64] |
| | CTF positive | 96 | Dickson <i>et al.</i> , 1992 [58] |
| | CTF positive | 85 * | Bauder <i>et al.</i> , 1985 [65] |
| | CTF positive | 84 * | Gerik <i>et al.</i> , 1987 [66] |
| | CTF positive | 88 | Voorhees <i>et al.</i> , 1984 [67] |
| | CTF positive | 87 * | Wagger and Denton, 1989 [68] |
| | CTF positive | 97 * | Liebig <i>et al.</i> , 1993 [69] |
| | CTF positive | 85 | Sedaghatpour <i>et al.</i> , 1995 [17] |
| | CTF positive | 90 | Douglas <i>et al.</i> , 1992 [26] |
| | CTF positive | 80 | Braunack and McGarry, 2006 [70] |
| | Penetration resistance | No difference | 100 |
| <u>CTF negative</u> | | - | Potter and Chichester, 1993 [25] |
| CTF positive | | - | Sommer and Zach, 1992 [72] |
| CTF positive | | 50 * | Radcliffe <i>et al.</i> , 1989 [73] |
| CTF positive | | 57 | Carter <i>et al.</i> , 1991 [74] |
| CTF positive | | - | Willcocks, 1981 [75] |
| Infiltration rate | CTF positive | - | Raper <i>et al.</i> , 1994 [76] |
| | CTF positive | 170 | Meek <i>et al.</i> , 1989 [19] |
| | CTF positive | 128 | Meek <i>et al.</i> , 1990 [20] |

* Studies comparing trafficked and un-trafficked row under ridge and furrow farming.

Table A2. Weighting factor calculated for soil erosion, comparing trafficked with un-trafficked practices.

| Measurement | Results | % of Relative Improvement | Study |
|-------------|---------------|---------------------------|--|
| Runoff | CTF positive | 64 | Li <i>et al.</i> , 2007 [61] |
| | CTF positive | 50 | Li <i>et al.</i> , 2001 [77] |
| | CTF positive | - | Sedaghatpour <i>et al.</i> , 1995 [17] |
| | CTF positive | 70 | Tullberg <i>et al.</i> , 2001 [23] |
| | No difference | 0 | Reyes <i>et al.</i> , 2005 [22] |
| | No difference | 0 | Rohde and Yule, 2003 [78] |

Table A3. Weighting factor calculated for soil biodiversity, comparing trafficked with un-trafficked practices.

| Measurement | Results | % of Relative Improvement | Study |
|---------------|--------------|---------------------------|--------------------------------------|
| Earthworms | CTF positive | 125 | Pangnakorn <i>et al.</i> , 2003 [24] |
| Herbicide use | CTF positive | - | Kurstjens, 2007 [79] |

Table A4. Weighting factor calculated for soil organic matter content, comparing trafficked with un-trafficked practices.

| Measurement | Results | % of Relative Improvement | Study |
|----------------|--------------|---------------------------|----------------------------------|
| Organic carbon | CTF positive | - | Potter and Chichester, 1993 [25] |

Table A5. Weighting factor calculated for energy requirement, comparing trafficked with un-trafficked practices.

| Measurement | Results | % of Relative Improvement | Study |
|---------------------|--------------|---------------------------|-----------------------------------|
| Energy requirements | CTF positive | 50–80 | Chamen <i>et al.</i> , 1992 [80] |
| | CTF positive | - | McPhee <i>et al.</i> , 1995 [56] |
| | CTF positive | 75 | Lamers <i>et al.</i> , 1986 [80] |
| | CTF positive | 85 | Dickson and Campbell, 1990 [81] |
| | CTF positive | 85 | Vermeulen and Klooster, 1992 [82] |
| | CTF positive | - | Chamen and Cavalli, 1994 [83] |
| | CTF positive | 70 | Chamen and Longstaff, 1995 [16] |
| | CTF positive | 57 | Dickson and Ritchie, 1996 [59] |
| | CTF positive | - | Willcocks, 1981 [75] |
| | CTF positive | 50–68 | Dickson <i>et al.</i> , 1992 [58] |
| | CTF positive | - | Chamen <i>et al.</i> , 1994 [84] |

Table A6. Weighting factor calculated for fertiliser use efficiency, comparing trafficked with un-trafficked practices.

| Measurement | Result | % of Relative Improvement | Study |
|----------------|--------------|---------------------------|-----------------------------------|
| Fertiliser use | CTF positive | 70–80 | Dickson and Ritchie, 1996b [60] |
| | CTF positive | - | Douglas <i>et al.</i> , 1992 [26] |

Table A.7 Weighting factor calculated for GHG emission from soils, comparing trafficked with un-trafficked practices.

| Measurement | Result | % of Relative Improvement | Study |
|--|---------------|---------------------------|-------------------------------------|
| N ₂ O-emissions | CTF positive | - | Ball <i>et al.</i> , 1999 [85] |
| | CTF positive | - | Vermeulen <i>et al.</i> , 2007 [30] |
| CO ₂ - and H ₂ O-emissions | No difference | - | Reicosky <i>et al.</i> , 1999 [54] |

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