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Aggregations of brittle stars can provide similar ecological roles as mussel reefs.

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Running header: Aggregations of mussels and brittle stars

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1 Abstract

2 Biogenic habitats, such as coral reefs, facilitate diverse communities. In aquatic systems,
3 aggregations of mobile benthic species may play a similar ecological role to that of typically
4 sessile biogenic habitats, however, this has rarely been considered. We quantified the abundance
5 of sessile mussels (*Modiolus modiolus*) and aggregating brittle stars (*Ophiotrix fragilis*) and
6 tested for correlations between the density of mussels (live and dead) and brittle stars each with:
7 1) abundance, biomass, diversity and assemblage structure of associated benthic macrofauna;
8 and 2) percent organic matter of the sediment. We found that the abundance of live *M. modiolus*
9 was positively associated with the abundance and biomass of macrofauna. The positive
10 association between *M. modiolus* and macrofauna abundance was further amplified with an
11 increase in brittle stars and a decrease in dead mussel shells. Macrofauna biomass was lower
12 with more dead mussel shells and macrofauna diversity increased with more live *M. modiolus*
13 and brittle stars. Sediment organic matter was positively related with brittle star density, but not
14 with the abundance of live or dead mussels. The positive relationship between brittle stars and
15 sediment organic matter suggests that brittle stars could enhance rates of benthic-pelagic
16 coupling. Given the importance of understanding the functional role of threatened habitats, it is
17 essential that the underlying community patterns be understood through robust observational
18 studies to then derive testable hypotheses to determine drivers. These findings provide novel
19 insight into the ecological role of aggregations of mobile species, which is essential to prioritize
20 conservation and restoration strategies.

21 **Introduction**

22 Habitat-forming species, such as corals or trees, are widespread in terrestrial and aquatic
23 ecosystems worldwide. These species create complex biogenic habitats, which are the foundation
24 of communities that do not exist in their absence (Bertness & Callaway 1994, Stachowicz 2001).
25 Increased habitat complexity facilitates increased species richness by reducing predation,
26 competition and disturbance pressure (Stachowicz 2001, Bruno et al. 2003). In aquatic
27 ecosystems, biogenic habitats such as seagrass, saltmarsh, mangroves, and bivalve reefs provide
28 multiple ecosystem services including, the enhanced production of economically important
29 species, reduced erosion rates, and nutrient removal (Costanza et al. 1997, Grabowski &
30 Peterson 2007, Barbier et al. 2011). Unfortunately, these marine species are impacted heavily by
31 human activities and most are reduced to a fraction of their historical abundance globally
32 (Waycott et al. 2009, Beck et al. 2011, De'ath et al. 2012). The loss of biogenic habitats has in
33 some instances negated their ecological roles and severely diminished the benefits they provide
34 to society (Waycott et al. 2009, Ermgassen et al. 2012, 2013).

35 Aggregations of mobile fauna are generally considered deleterious to ecosystems; for
36 example outbreaks of urchins can denude large areas once covered with macroalgae (Steneck et
37 al. 2004). However, aggregations of mobile species can potentially provide similar functions as
38 sedentary, foundation species and create biogenic habitats. For instance, aggregations of urchins
39 can increase biodiversity and provide shelter for prey (Altieri & Witman 2014).

40 Mussel reefs, similar to oyster reefs, have been depleted worldwide (Lotze et al. 2006).
41 The horse mussel, *Modiolus modiolus*, forms reefs in the North Atlantic ocean (Sanderson et al.
42 2008, Wildish et al. 2009) and its abundance has declined most likely from habitat destruction
43 following fishing practices (Magorrian & Service 1998, Strain et al. 2012, Cook et al. 2013) and

44 global warming (Gormley et al. 2013). The complex habitat that *M. modiolus* reefs create is
45 known for its high diversity of organisms (Rees et al. 2008, Ragnarsson & Burgos 2012, Fariñas-
46 Franco et al. 2013). The loss of live mussels can result in a matrix of dead shell. Experimental
47 studies that held shell structure constant found that loss of live blue mussels reduce abundance
48 but not diversity of macrofauna (Norling & Kautsky 2007, Norling et al. 2015), but the loss of
49 live *M. modiolus* on reef ecology in natural settings is unknown.

50 Brittle stars occur in dense aggregations throughout the globe (Fedra 1977, Fratt &
51 Dearborn 1984). In particular, the suspension feeding brittle star, *Ophiothrix fragilis*, exists in
52 dense beds of more than 1,000 individuals m⁻² around Britain and Ireland (Warner 1971,
53 Aronson 1989, Dauvin et al. 2013). *Ophiothrix fragilis* beds exist in similar environments as *M.*
54 *modiolus* reefs and co-occur in some areas (Sanderson et al. 2008, Ragnarsson & Burgos 2012).
55 Although it was thought that aggregations of *O. fragilis*, which often overlay *M. modiolus* reefs,
56 may have negative effects on benthic macrofauna from smothering and competition effects,
57 many species of macrofauna were recorded beneath brittle star beds (Warner 1971). Brittle stars
58 could facilitate benthic fauna by enhancing deposition of organic material (Warner 1971, Murat
59 et al. 2016) and provide refuge from predators. Dense aggregations of this mobile species could
60 provide ecological benefits similar to sessile reef forming species. In addition, there could be
61 emergent properties when dense aggregations of both sessile and mobile species exist together
62 (Angelini et al. 2011). However, our understanding of ecological roles associated with these two
63 habitats, beds of mobile species and reefs of sessile species, is limited.

64 *Modiolus modiolus* is protected in Europe under Annex I of the EU Habitats Directive
65 (Directive 93/43/EEC) and the OSPAR convention (Rees et al. 2008). Therefore a survey based
66 on benthic grabs was designed to determine the extent and condition of *M. modiolus* reefs in

67 Northern Ireland to inform management decisions about their conservation. This intensive survey
68 allowed an unprecedented opportunity to study this heavily protected biogenic habitat and
69 determine ecological patterns which are a necessary precursor to conducting manipulative
70 experiments to pinpoint underlying mechanisms (Underwood et al. 2000). Our aim was to
71 characterize the ecological roles of a declining biogenic-reef forming species, *M. modiolus*, and a
72 common co-occurring benthic species, *O. fragilis*. We quantified how the abundance of live *M.*
73 *modiolus*, *M. modiolus* shell (dead *M. modiolus*) and *O. fragilis* was related with: (i) macro-
74 benthic species abundance, biomass, richness, diversity and assemblage structure; and (ii)
75 sedimentary organic matter. We hypothesized that the abundance of live *M. modiolus* and *O.*
76 *fragilis* would have a similar positive relationship with the abundance, biomass, richness and
77 diversity of the benthic macrofauna, and the sediment organic matter. Moreover, the abundance
78 of *M. modiolus* shell would not have a correlation with the abundance, biomass, richness, and
79 diversity of the benthic macrofauna, and the sediment organic matter because shell has less
80 structural complexity compared to live mussels and does not produce fecal matter (as would be
81 expected in reefs comprised of mainly live animals). Finally, prevalence of live *M. modiolus*, *M.*
82 *modiolus* shell and *O. fragilis* will explain a similar amount of variation in the macrofauna
83 assemblage.

84

85 **Materials and Methods**

86 **Data collection**

87 To quantify the variation in benthic fauna and sediment organic matter related with the
88 abundance of *M. modiolus*, *M. modiolus* shell, and *O. fragilis*, a grab sampler was used to sample
89 53 sites at two locations off the east coast of Northern Ireland at depths of 20-30 m (Fig. 1).

90 Sampling sites were dispersed evenly within different acoustic signatures of a single beam sonar
91 survey conducted by Northern Ireland Agri-Food and Biosciences Institute in 2014 (unpub.
92 data). Samples were collected over 5 days between the 9th of September and the 16th of
93 December 2014. The Day grab used in this study removed approximately 0.1 m² area of
94 substratum (approximately 2 l of sediment; Appendix S1). Grabs with minimal sediment (< 1 L
95 of sediment) were recorded as misfires and the grab was deployed again. Three replicate grabs
96 were taken at each site unless 3 successive misfires were recorded and no more samples were
97 taken at that site. The environment of these locations made diver surveys impractical (depth and
98 currents) and although larger grab samples would have been optimal, a day grab was used to
99 remove a minimal amount of the protected *M. modiolus* reef. Each sample was photographed and
100 the percentage cover of *M. modiolus* shell, mud, and sand were estimated visually based on the
101 grab surface (Appendix 2). Sediment samples were taken from the top 2 cm of 2 haphazardly
102 chosen grabs from each site. However, sediment was collected from all three samples from 9
103 sites because live *M. modiolus* was present in the grabs. Sediment organic matter was collected
104 from 101 grabs at 46 sites. Sediment samples were freeze dried, sieved through 1 mm mesh, and
105 placed in a combustion oven at 500°C for 6 hours (Dean 1974). The percent organic matter was
106 determined by dividing the difference in mass of the sediment before and after combustion by
107 the mass of the sediment before combustion.

108 Conspicuous macrofauna were quantified from 140 grab samples at 53 sites, which were
109 searched thoroughly on the boat after the grab sample was sieved through 1 mm mesh and all
110 bivalves, crustaceans, echinoderms, fish, gastropod, and polychaetes were collected and frozen.
111 Macrofauna were identified to the lowest practical taxon, counted and wet weight recorded for a
112 total taxon biomass. Macrofauna identification was based on morphological characteristics

113 following guides (Hayward & Ryland 1995). A subset of grab samples (approximately 1 grab
114 from each site) were returned to the laboratory and cryptic species, which were not attached to
115 the substratum, were quantified in addition to the conspicuous macrofauna (Appendix S3).
116 Results of analyses from these samples were consistent with those based on conspicuous fauna
117 thus for clarity we are presenting only the findings for the conspicuous fauna. Animal handling
118 protocols followed the ethical guidelines of Queen's University Belfast.

119 **Data Analysis**

120 We tested effect of substrate type (live *M. modiolus*, *M. modiolus* shell, or *O. fragilis*) on benthic
121 macrofauna abundance, biomass, taxon richness, diversity and assemblage structure.
122 The abundance of *M. modiolus* and *O. fragilis* in grab samples were *a priori* determined as
123 predictor variables and not included as benthic fauna in the analyses. Generalized linear models
124 were used to test for variations in total faunal abundance and taxon richness associated with
125 changes of the three habitat types (*M. modiolus*, *M. modiolus* shell, and *O. fragilis*) with Poisson
126 distributions because data were skewed towards zero. General linear models were used to test for
127 variations in non-integer dependent variables (biomass and diversity) associated with changes in
128 the three habitat types. Diversity for each sample was calculated with the Shannon-Weaver
129 index. Multi-collinearity between predictor variables was tested using the variance inflation
130 factor (VIF) and < 10 indicated minimal multi-collinearity (Hair 2006). Site within location
131 (North or South as shown in Fig. 1) was included as random variables in the model to account
132 for the nested sampling design. All interactions were included in the models and the predictor
133 variables were centred (the mean was subtracted from each value) to reduce multi-collinearity
134 between predictor variables and interactions (Quinn & Keough 2002) and scaled (divided by the
135 standard deviation) to reduce the difference in magnitude among the predictor variables (Bates et

136 al. 2014). All analyses were conducted in R (R Development Core Team 2012). Diversity was
137 calculated using the `div` function within the *vegan* package (Oksanen et al.). The `glmer` function
138 within the *lme4* package (Bates et al. 2014 4) was used for both abundance and richness of
139 benthic fauna. Biomass, diversity and organic matter were analysed using `lmer` function within
140 the *lme4* package to calculate *t* values, while the Analysis of Variance (ANOVA) function within
141 the *car* package was used to generate *p* values and test for significance (Fox & Weisberg 2011).
142 Interactions and random variables (site nested in location) were included in the organic matter
143 model. Models were checked to ensure an adequate fit by visually inspecting residuals vs fitted
144 (randomly distributed points) and Q-Q (points were near 1:1 ratio) plots (Crawley 2007).
145 Biomass of macrofauna was log transformed to improve model fit.

146 To quantify the amount of variation in benthic fauna assemblage explained by the
147 abundance of *M. modiolus*, *M. modiolus* shell, and *O. fragilis*, which were all continuous
148 variables, we used Permutational Analysis of Variance (PERMANOVA, McArdle & Anderson
149 2001, Anderson 2001) and redundancy analysis (RDA, Legendre & Anderson 1999).
150 PERMANOVA partitions the variation of a resemblance matrix among sources of variation and
151 fits linear models to test hypotheses and build models without ordination. RDA performs
152 ordination of fitted values to test hypotheses, build models, and create visualizations of the data
153 (Legendre & Anderson 1999). RDA reduces the variance into dimensions, which makes
154 visualizations of the data possible but may reduce the amount of variance explained by predictor
155 variables. However, PERMANOVA analyzes the data without constraining the variance into
156 dimensions so that the relationship between community structure and predictor variables is
157 probably closer to what exists naturally. Histograms of each predictor variable and scatter plots
158 of all combinations of predictor variables were examined to ensure there were no extreme

159 outliers. Linear-based analyses can be biased by multi-collinearity (Legendre & Anderson 1999),
160 we tested for multi-collinearity as previously described using VIF. Multivariate analyses used the
161 *vegan* package version 2.2-0 (Oksanen et al. 2010).

162 Taxon-specific abundances were log-transformed to reduce the influence of abundant
163 taxa in the analyses, and a Bray-Curtis dissimilarity matrix was created (Anderson et al. 2008,
164 Legendre & De Cáceres 2013). The mean of taxon abundance per site was used to remove the
165 possibility of non-independent samples within each site. Interactions were included in the
166 sequential PERMANOVA and predictor variables were centred and scaled to reduce
167 multicollinearity between variables and interactions. A second PERMANOVA was run with
168 taxon-specific biomass following the same procedure described for abundance.

169

170 **Results**

171 *Modiolus modiolus* was present in 45 of the samples and ranged from 1 to 65 individuals per
172 grab (~10 to 650 m⁻²). *Ophiotrix fragilis* was present in 81 (out of 140) of the samples which
173 ranged from 1-203 individuals per grab (~10 to 2,030 m⁻²). Fifty-seven different taxa were
174 quantified in the samples (Appendix S3). The substratum of the grabs was primarily *M. modiolus*
175 shell and mud, with sand being less prevalent (Appendix 4). The abundance of live *M. modiolus*
176 and *O. fragilis* had similar patterns with changes in the different types of substratum. The
177 abundance of benthic fauna increased with the number of live *M. modiolus* and there were
178 interactions between *M. modiolus* and *O. fragilis*, and *M. modiolus* and *M. modiolus* shell (Table
179 1; Fig. 2A, D and G; Fig. 3). The interaction between *M. modiolus* and *O. fragilis* resulted from a
180 greater increase in fauna abundance as *M. modiolus* increased when there were more *O. fragilis*
181 (Fig 3A). The opposite trend existed for the interaction between *M. modiolus* and *M. modiolus*

182 shell; there was a greater increase in fauna abundance as *M. modiolus* increased when there was
183 less shell (Fig 3B). Biomass of macrofauna increased with *M. modiolus*, but decreased with *M.*
184 *modiolus* shell (Table 1; Fig 2B and E). The interaction between all predictor variables was also
185 significant for the biomass of macrofauna (Table 1).

186 The three predictor variables did not explain a significant amount of variation in the
187 number of taxon (richness) in a sample and there were no significant interactions (Table 1). *M.*
188 *modiolus* and *O. fragilis* were associated with an increase in the macrofauna diversity (Table 1,
189 Fig 2C and I). *O. fragilis* was positively related with organic matter, while the prevalence of *M.*
190 *modiolus* or *M. modiolus* shell did not explain a significant amount of variation in organic matter
191 (Table 1; Fig. 4A-C). The three-way interaction was significant for organic matter and resulted
192 from a positive relationship between *O. fragilis* and organic matter, which was greatly reduced
193 with an increase in *M. modiolus* abundance and reduced with an increase in *M. modiolus* shell
194 cover (Fig. 4D).

195 The amount of variation in faunal assemblage using abundance explained by *M.*
196 *modiolus*, *O. fragilis*, and *M. modiolus* shell was quantified using a PERMANOVA. *Modiolus*
197 *modiolus* ($F_{1,52} = 3.16$; $P < 0.001$), *O. fragilis* ($F_{1,52} = 5.45$; $P < 0.001$), and *M. modiolus* shell ($F_{1,52}$
198 $= 5.11$; $P < 0.001$) were significant and explained more variation in fauna assemblage than would
199 be expected by random chance. No interactions were significant. *Modiolus modiolus* shell
200 explained the most variation in macrofauna assemblage of the 3 continuous predictor variables
201 ($R^2 = 8.4\%$), followed by *O. fragilis* ($R^2 = 5.7\%$), and *M. modiolus* ($R^2 = 5.2\%$). The RDA
202 represents the relationship between predictor variables and individual taxon. RDA explained
203 10.7% of the variation in fauna assemblage. The first and second axes explained 6.5 and 2.9% of
204 the variation respectively. *Modiolus modiolus* was positively related with axis 1 and *M. modiolus*

205 shell was positively related with axis 2 (Appendix S5). *Ophiocomina nigra* (a brittle star) was
206 positively related with the second axis, and *Ophiura* spp. (a brittle star) and *Timoclea ovata* (a
207 bivalve) were negatively related with the second axis (Appendix S5). The fauna assemblage
208 based on biomass had similar findings as the assemblage using abundance with all three
209 predictor variables explaining a significant amount of variation. *Modiolus modiolus* shell
210 explained the most variation in macrofauna assemblage of the 3 continuous predictor variables
211 ($R^2 = 8.7\%$), followed by *O. fragilis* ($R^2 = 6.3\%$), and *M. modiolus* ($R^2 = 4.1\%$).

212

213 **Discussion**

214 Biogenic habitats composed of aggregations of sessile species, often referred to as
215 meadows or reefs, are touted for their ecological and economic benefits (Anton et al. 2011,
216 Barbier et al. 2011, Firth et al. 2015). On the other hand, dense aggregations of mobile species
217 are generally viewed negatively; however, this study indicates that increasing densities of *O.*
218 *fragilis* were associated with greater macrofauna diversity and organic matter, and had a positive
219 emergent effect on the total abundance of fauna within *M. modiolus* reefs. Although these results
220 are correlations, they could suggest that aggregates of brittle stars enhance diversity of
221 macrofauna and increase sediment organic matter similar to or more than filter feeding bivalves.

222 The ecological effect of a single foundation species on the local community has been
223 extensively studied (Grabowski et al. 2005, Geraldi et al. 2009), however, multiple species often
224 coexist together and little is known about potential interactions among different species
225 (Angelini et al. 2011, Donadi et al. 2015). One study that included multiple ecosystem engineers
226 found that the presence of *Caulerpa taxifolia*, a macroalgae, near *Anadora trapezia*, a clam,
227 increased diversity and abundance of epibiota on the bivalve (Gribben et al. 2009). Most of these

228 studies focused on relatively sessile species and there is the potential for mobile species to also
229 enhance both density and diversity of associated fauna (Altieri & Witman 2014). The abundance
230 of *O. fragilis* was related with enhanced diversity of macrofauna and had a positive emergent
231 effect with *M. modiolus* reefs on the abundance of macrofauna. In addition, minimal
232 multicollinearity among predictor variables indicates that there was no facilitation between *O.*
233 *fragilis* and *M. modiolus*, and that abundances of live and dead *M. modiolus* were independent.
234 Finally, all three habitat types measured had similar influence on the macrofauna assemblage
235 (explained between 5 and 8% of the variation in assemblage).

236 Our conclusions are based on a robust survey, which aimed to identify ecological patterns
237 associated with different dominant species (mussels and/or brittle stars). Experimental
238 manipulation is required to determine the mechanisms driving these differences, which is
239 difficult given the ethical and logistical limitations of manipulating a rare species that primarily
240 exist in areas with high currents and deeper than 20 m. Given our existing knowledge,
241 aggregations of brittle stars and other mobile species appear to share similar roles as some well-
242 described sessile foundation species. For example, positive effects on the macrofauna
243 community associated with aggregations of mobile fauna could result from reduced predation
244 from provision of shelter (Bruno et al. 2003) or from increased food provision via biodeposition
245 (Norling & Kautsky 2007).

246 Understanding how the loss of individual bivalves from reefs affects ecological
247 functioning is important given the prevalence of reef degradation (Beck et al. 2011, Ermgassen et
248 al. 2012). Teasing apart the provision of habitat by the physical structure from the biotic function
249 of bivalve reefs has been studied using experimental reefs. For example, the diversity of
250 macrofauna was similar on blue mussel (*Mytilus edulis*) reefs compared to reefs made of intact

251 shells, while the abundance was greater on live reefs possibly because of resources supplied by
252 biodeposition (Norling & Kautsky 2007). Similarly, another study found that diversity of
253 epibenthic fauna was similar among live and dead experimental reefs of oysters or mussels,
254 while abundance of epibenthic fauna was greatest on oyster shell, moderate on live bivalve reefs
255 and lowest on mussel shell reefs (Norling et al. 2015). We found that the amount of naturally
256 occurring dead shell was not related with the abundance, richness and diversity of benthic
257 macrofauna, and that dead shell cover was negatively related with the biomass of macrofauna.
258 Separating the role of the physical reef structure from associated biotic functioning is necessary
259 to identify ecological mechanisms, and also to predict changes in ecosystem functioning
260 associated to bivalve mortality from direct or indirect anthropogenic impacts.

261 Biodeposition by filter feeding bivalves is an important process in coastal ecosystems
262 because it couples pelagic and benthic communities. Benthic-pelagic coupling may reduce
263 occurrences of hypoxia by directly reducing phytoplankton abundance (Dame & Olenin 2005,
264 Grizzle et al. 2008) and indirectly through nitrogen removal by enhancing denitrification on the
265 sediment because of the high quality resources provided by biodeposits (Kellogg et al. 2013,
266 Smyth et al. 2013, 2015). *Modiolus modiolus* produce nutrient rich biodeposits (Navarro &
267 Thompson 1997), however, we did not identify a relationship between *M. modiolus* density and
268 sediment organic matter content. We did find a positive relationship between brittle star density
269 and organic matter content. This relationship could have resulted from brittle stars preferring
270 benthos with greater organic matter. However, our results indicate that *O. fragilis* and *M.*
271 *modiolus* have similar abundance patterns in shell, mud and sandy substratum suggesting that
272 brittle stars are not preferentially selecting one type of substratum that could be causing this
273 relationship, which is likely driven by benthic-pelagic coupling. A positive relationship between

274 total organic carbon and brittle star density was found in stable environments (Murat et al. 2016)
275 and benthic-pelagic coupling associated with brittle star beds was suggested to reduce
276 eutrophication in coastal bays (Hily 1991). Both *M. modiolus* and *O. fragilis* are suspension
277 feeders but use entirely different mechanisms to collect suspended particles. *M. modiolus* is an
278 active filter feeder while *O. fragilis* passively feeds on phytoplankton (Roushdy & Hansen 1960,
279 Migne et al. 2012, BlanchetAurigny et al. 2015). The stronger association between brittle stars
280 and sediment organic matter compared to *M. modiolus* could result from *O. fragilis* having a low
281 absorption efficiency (Migné & Davoult 1998) or that aggregations reduce water motion and the
282 erosion of biodeposits (Warner 1971) more than mussel reefs. Our findings, that organic matter
283 was positively related with brittle star abundance and not *M. modiolus* density, may suggest that
284 benthic-pelagic coupling in brittle star beds is potentially greater than in bivalve reefs and this
285 should be investigated further.

286 The ecological and economic benefits of marine biogenic habitats, such as coral reefs,
287 salt marshes and bivalve reefs, are well known and are the impetus for their conservation and
288 restoration (Brumbaugh & Coen 2009, Barbier et al. 2011, Geraldi et al. 2013, La Peyre et al.
289 2014). Beds of brittle stars may enhance the diversity of macrofauna and increase benthic-
290 pelagic coupling equal to or greater than bivalve reefs. The carbon budgets associated with
291 biomass production and calcification has been quantified for brittle star beds (Migne et al. 1998,
292 Davoult et al. 2009, Lebrato et al. 2010), however, their potential importance for other rates of
293 ecosystem functioning and associated services is relatively unknown. The ecosystem functions
294 provided by brittle stars are probably context-dependent, but the global functional role of these
295 taxa may be equal to or greater than other sessile foundation species for multiple reasons. First,
296 brittle star beds are prevalent around the globe given that they have been documented from the

297 Arctic (Piepenburg & Schmid 1996, Blicher & Sejr 2011) to the Antarctic (Fratt & Dearborn
298 1984) and throughout the mid-latitudes (Haedrich et al. 1980, Fujita 1990). They are also present
299 at broad depth ranges (Lebrato et al. 2010) and not restricted to estuaries and coasts like
300 traditional biogenic habitats.

301 Determining the ecological functions provided by aggregations of mobile species and
302 comparing these to functions provided by traditional biogenic habits, as well as potential
303 emergent effects between the two are needed to understand the relative importance of these
304 species to broader ecosystem processes and functions. This is of utmost importance as humans
305 are constantly altering the abundance and extent of both sessile and mobile species. The applied
306 implications of these results, if confirmed by manipulative experiments, include assigning
307 aggregates of mobile species similar conservation status as sessile foundation species (Peterson
308 & Lipcius 2003, Byers et al. 2006, Lampert & Hastings 2014).

309

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- 514

515 Table 1. Summary of statistical models to assess the relationship between the predictor variables
 516 (abundance of *M. modiolus*, *M. modiolus* shell, and *O fragilis*) and the abundance, biomass,
 517 richness, and diversity of benthic fauna, and organic matter collected in samples (significant
 518 predictor variables and interactions are bold).

Dependent variable	Independent variable	Estimate	Std. Error	z or t value	p value
# of individuals	<i>M. Modiolus</i>	0.981	0.111	8.81	<0.001
	<i>M. Modiolus</i> shell	0.004	0.066	0.06	0.957
	<i>O. fragilis</i>	-0.083	0.061	-1.35	0.178
	<i>M. Modiolus</i>: <i>M. Modiolus</i> shell	-0.678	0.106	-6.38	<0.001
	<i>M. Modiolus</i>: <i>O. fragilis</i>	0.162	0.078	2.07	0.038
	<i>M. Modiolus</i> shell: <i>O. fragilis</i>	0.012	0.045	0.27	0.789
	<i>M. Modiolus</i> : <i>M. Modiolus</i> shell: <i>O. fragilis</i>	-0.004	0.075	-0.06	0.954
Biomass	<i>M. Modiolus</i>	0.009	0.028	0.32	<0.001
	<i>M. Modiolus</i> shell	-0.008	0.003	-2.39	<0.001
	<i>O. fragilis</i>	0.002	0.001	1.52	0.214
	<i>M. Modiolus</i> : <i>M. Modiolus</i> shell	0.001	0.001	1.18	0.744
	<i>M. Modiolus</i> : <i>O. fragilis</i>	0.001	0.000	1.67	0.920
	<i>M. Modiolus</i> shell: <i>O. fragilis</i>	0.000	0.000	0.90	0.230
	<i>M. Modiolus</i>: <i>M. Modiolus</i> shell: <i>O. fragilis</i>	0.000	0.000	-2.09	0.037
# of species	<i>M. Modiolus</i>	0.090	0.095	0.95	0.344
	<i>M. Modiolus</i> shell	-0.022	0.050	-0.44	0.662
	<i>O. fragilis</i>	0.016	0.040	0.40	0.691
	<i>M. Modiolus</i> : <i>M. Modiolus</i> shell	-0.062	0.093	-0.66	0.508
	<i>M. Modiolus</i> : <i>O. fragilis</i>	0.027	0.084	0.33	0.744
	<i>M. Modiolus</i> shell: <i>O. fragilis</i>	0.003	0.038	0.09	0.928
	<i>M. Modiolus</i> : <i>M. Modiolus</i> shell: <i>O. fragilis</i>	0.029	0.087	0.33	0.739
Diversity	<i>M. Modiolus</i>	0.256	0.170	1.51	0.032
	<i>M. Modiolus</i> shell	-0.116	0.086	-1.35	0.069
	<i>O. fragilis</i>	0.129	0.068	1.90	0.021
	<i>M. Modiolus</i> : <i>M. Modiolus</i> shell	-0.127	0.166	-0.77	0.335
	<i>M. Modiolus</i> : <i>O. fragilis</i>	0.042	0.145	0.29	0.963
	<i>M. Modiolus</i> shell: <i>O. fragilis</i>	0.081	0.066	1.23	0.235
	<i>M. Modiolus</i> : <i>M. Modiolus</i> shell: <i>O. fragilis</i>	-0.055	0.152	-0.36	0.719
Organic matter	<i>M. Modiolus</i>	0.001	0.002	0.61	0.605
	<i>M. Modiolus</i> shell	-0.002	0.001	-1.16	0.636
	<i>O. fragilis</i>	0.002	0.001	1.72	0.048
	<i>M. Modiolus</i> : <i>M. Modiolus</i> shell	-0.001	0.002	-0.69	0.928
	<i>M. Modiolus</i> : <i>O. fragilis</i>	-0.005	0.002	-2.95	0.388
	<i>M. Modiolus</i> shell: <i>O. fragilis</i>	-0.002	0.001	-1.36	0.677
	<i>M. Modiolus</i>: <i>M. Modiolus</i> shell: <i>O. fragilis</i>	0.006	0.002	3.79	<0.001

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524 Figure legends

525 Fig. 1. Sampling sites near the Outer Ards Peninsula, east coast of Northern Ireland. Three
526 replicate grab samples were taken at each site

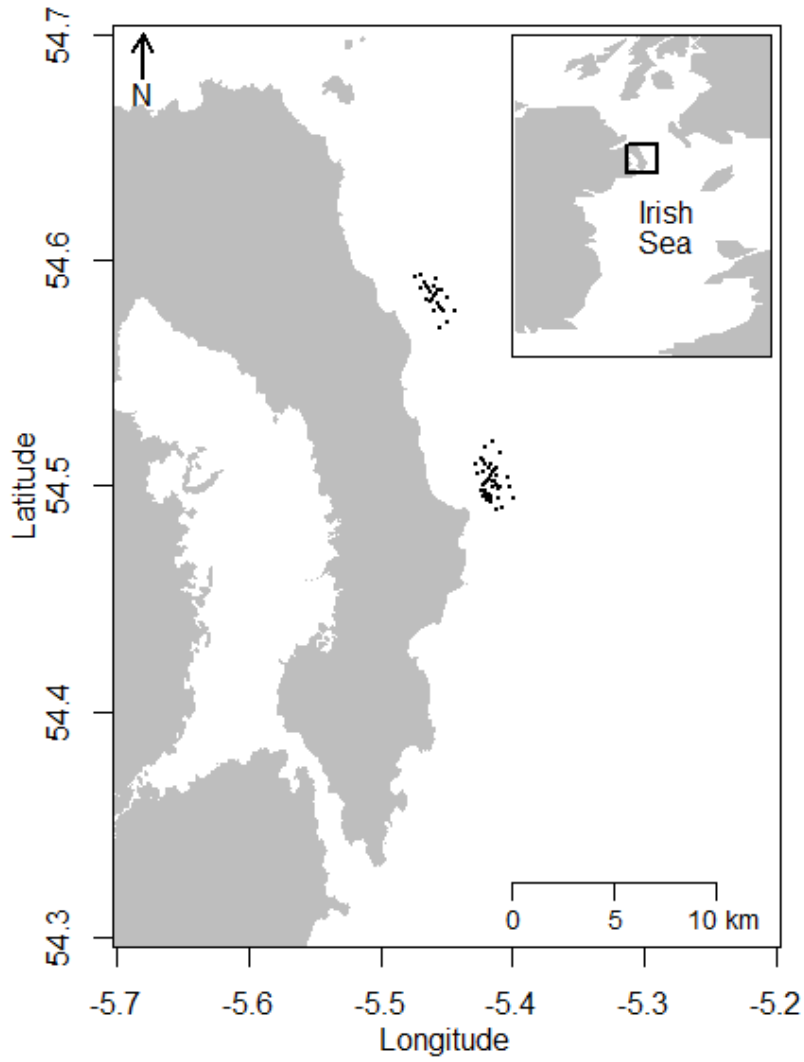
527 Fig. 2. The relationship between the abundance (left column), biomass (middle column), and
528 diversity (right column) of benthic fauna quantified in day grabs, and *M. modiolus*, *M. modiolus*
529 shell or *O. fragilis*. Predictor variables that were significant are indicated by black lines (See
530 Table 1 for statistical summary).

531 Fig. 3. The interactions between the abundance of *M. modiolus* and *O. fragilis* (A) and between
532 *M. modiolus* and *M. modiolus* shell (B) in explaining variation in the abundance of benthic fauna.
533 The data points and trend lines were categorized based on abundance of *O. fragilis* (A) or
534 percent shell cover (B).

535 Fig. 4. The relationship between the abundance of *M. modiolus* (A), *M. Modiolus* shell (B), *O.*
536 *fragilis* (C), and the percent organic matter in sediment collected in day grabs. The interaction
537 between all three predictor variables and the percent organic matter (D). *O. fragilis* and the
538 interaction between the three predictor variables explained a significant amount of variation in
539 percent organic matter. Predictor variables that were significant are indicated by black lines (See
540 Table 1 for statistical summary).

541

542 Fig. 1

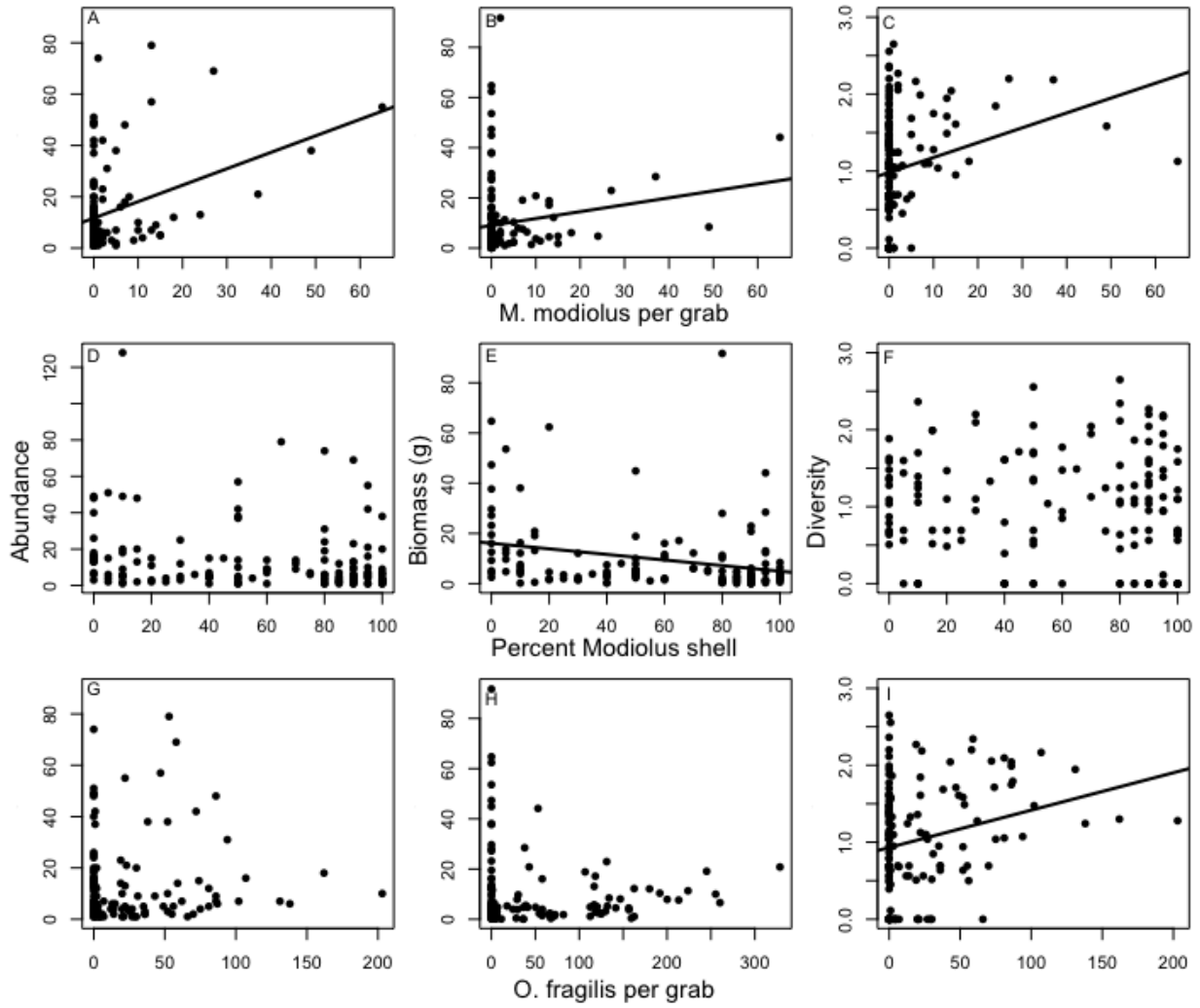


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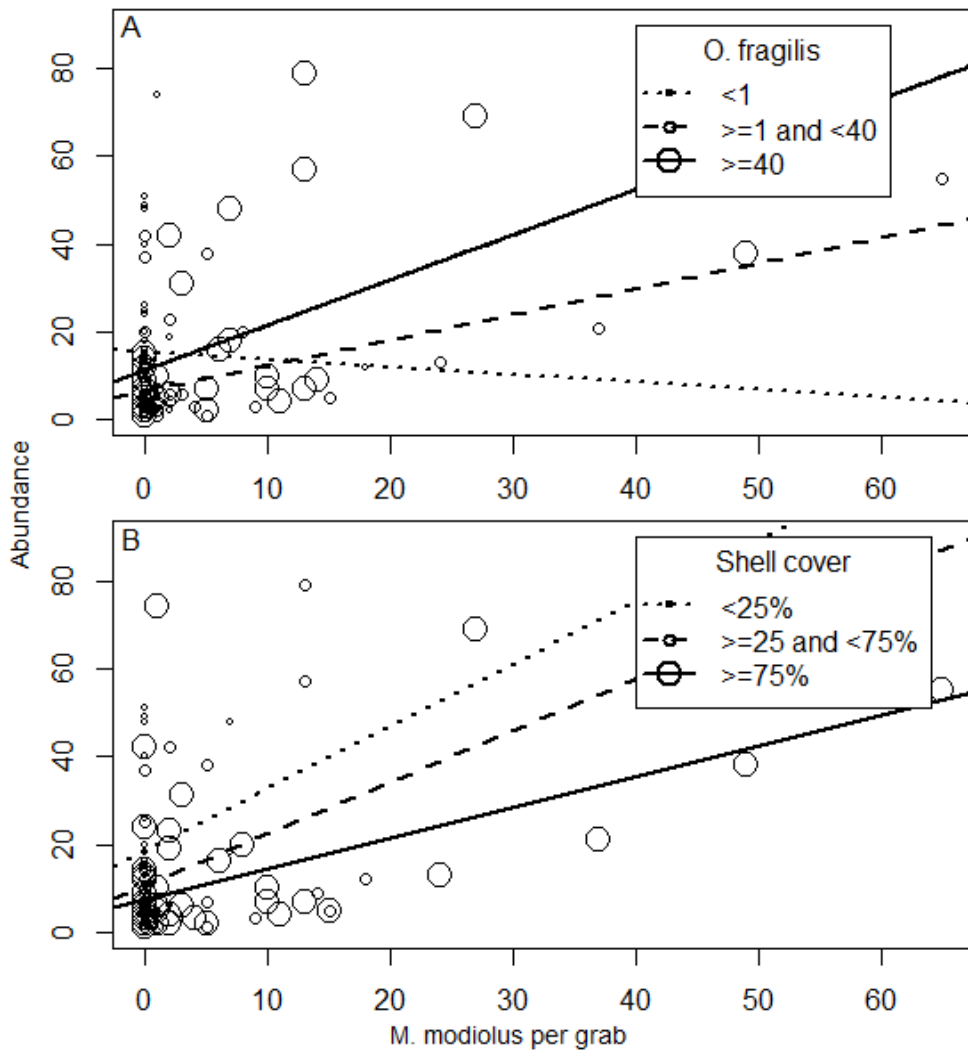
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546 Fig. 2



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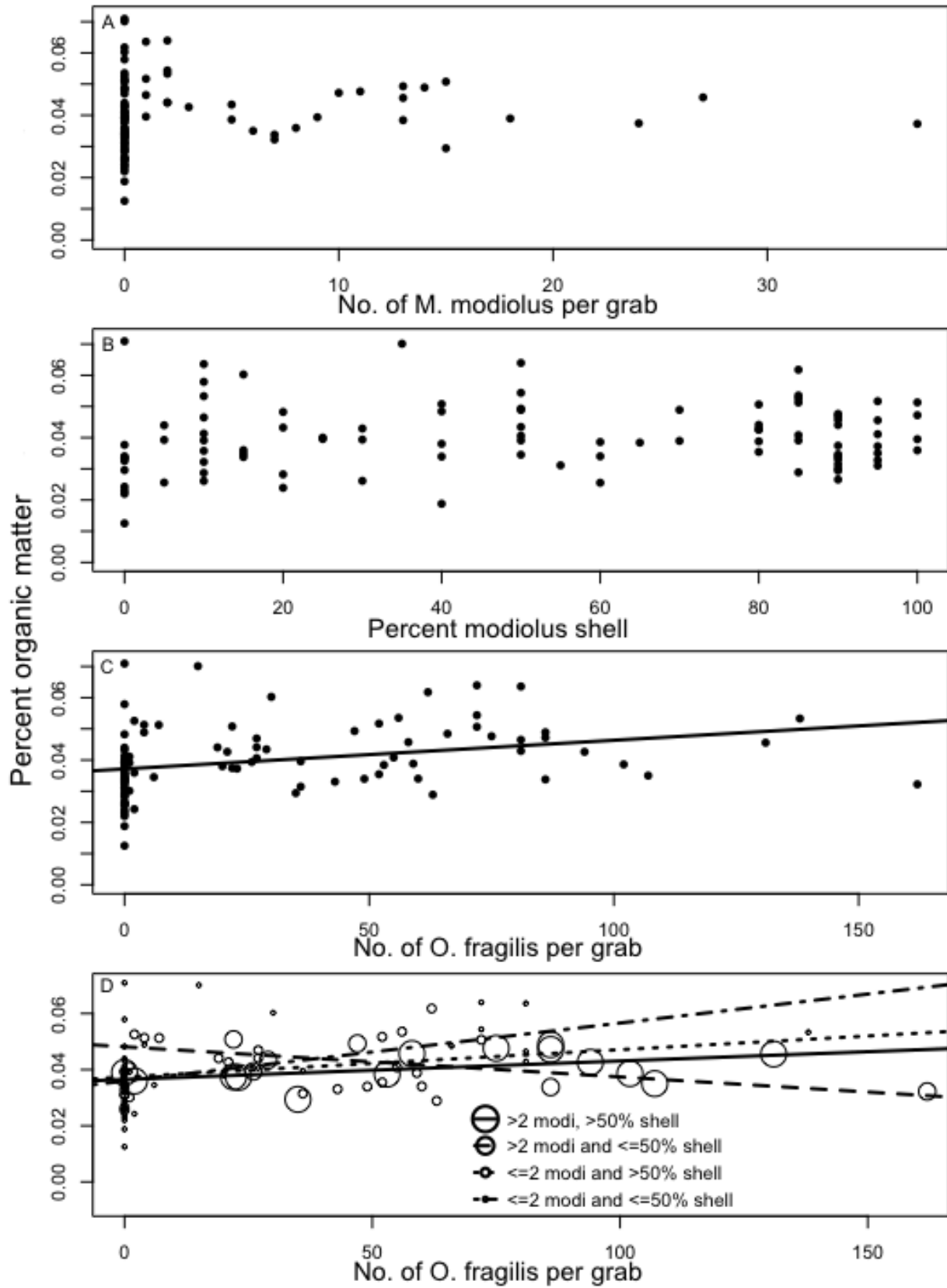
549 Fig. 3



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552 Fig. 4



Running header: Aggregations of mussels and brittle stars

554 Appendices

555 Appendix S1. Picture of the day grab used to sample fauna. Photo credit: C. Bertolini



556

557 Appendix S2. Day grabs illustrating typical samples with *M. modiolus* and *O. fragilis* (A), *M.*
558 *modiolus* shell (B), *O. fragilis* (C), and mud (D). The grab sampled 0.1m². Photo credits: C.
559 Bertolini and N. Galdi.



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Running header: Aggregations of mussels and brittle stars

562 Appendix S3. Taxon and their respective groups quantified in the grab samples. The proportion
 563 for each taxon of the total number of individuals for the grabs that had all taxon quantified and
 564 for the grabs that only conspicuous taxon were quantified. Only data from conspicuous taxon
 565 were used for all analyses.
 566

Taxa	Taxa group	Full	Conspicuous
<i>Abra alba</i>	Bivalve	0.248	0.254
<i>Amphipholis squamata</i>	Echinoderm		0.002
<i>Amphiura chiajei</i>	Echinoderm		0.008
<i>Amphiura filiformis</i>	Echinoderm		0.010
Anomiidae	Bivalve	0.018	
<i>Aphrodita aculeata</i>	Polychaete	0.001	0.001
<i>Astarte sulcata</i>	Bivalve	0.033	0.036
<i>Atelecyclus rotundatus</i>	Crustacean		0.001
<i>Buccinum undatum</i>	Gatropod	0.005	0.012
Capitellidae	Polychaete	0.046	
<i>Caprella acanthifera</i>	Crustacean		0.001
<i>Clausinella fasciata</i>	Bivalve	0.009	0.015
<i>Crossaster papposus</i>	Echinoderm		0.003
<i>Diodera graeca</i>	Gatropod		0.001
<i>Ebalia tuberosa</i>	Crustacean		0.003
<i>Echinocardium cordatum</i>	Echinoderm		0.001
<i>Emarginula fissura</i>	Gatropod	0.003	
<i>Eschinus esculentus</i>	Echinoderm	0.003	0.003
<i>Eteone longa</i>	Polychaete	0.004	
<i>Eunereis longissima</i>	Polychaete	0.004	
Eunicidae	Polychaete	0.005	
<i>Euspira nitida</i>	Gatropod	0.002	
<i>Galathea</i>	Crustacean		0.017
<i>Galathowenia oculata</i>	Polychaete	0.002	
Gammaridae	Crustacean		0.008
<i>Gari depressa</i>	Bivalve	0.024	0.021
<i>Gari tellinella</i>	Bivalve	0.027	0.030
<i>Gattyana cirrhosa</i>	Polychaete	0.002	
<i>Gibbula cineraria</i>	Gatropod	0.002	
<i>Glycera spp.</i>	Polychaete	0.004	
<i>Glycimeris glycimeris</i>	Bivalve	0.004	0.003
Gobiesocidae	Fish		0.004
Golfingiidae	Sipunucla	0.005	0.005
Harmothoe	Polychaete	0.014	0.009
Hesionidae	Polychaete	0.002	
<i>Hiatella arctica</i>	Bivalve	0.013	0.010
Hippolytidae	Crustacean		0.017

Running header: Aggregations of mussels and brittle stars

<i>Lepidonotus squamatus</i>	Polychaete	0.013	0.010
<i>Leptochiton asellus</i>	Chiton	0.029	0.018
<i>Limaria sp.</i>	Bivalve	0.010	0.007
<i>Liocarcinus spp.</i>	Crustacean		0.004
<i>Lumbrineridae</i>	Polychaete	0.015	
<i>Marthasterias glacialis</i>	Echinoderm		0.008
<i>Mediomastus fragilis</i>	Polychaete	0.002	
<i>Mimachlamys varia</i>	Bivalve	0.018	0.017
<i>Mya arenaria</i>	Bivalve	0.009	0.009
<i>Mya truncata</i>	Bivalve	0.009	0.007
<i>Myrtea spinifera</i>	Bivalve	0.003	
<i>Mytilus edulis</i>	Bivalve		0.006
Nemertea	Nematode	0.001	0.003
Nephtheidae	Polychaete	0.027	0.018
Nereis spp.	Polychaete	0.001	0.001
<i>Nucula nucleus</i>	Bivalve	0.079	0.069
Nuculanidae	Bivalve	0.002	
Oeononidae	Polychaete	0.003	
<i>Onchidoris spp.</i>	Nudibranch	0.001	
<i>Onoba semicostata</i>	Gastropod	0.002	
<i>Ophelina acuminata</i>	Polychaete	0.009	
<i>Ophiocomina nigra</i>	Echinoderm		0.057
<i>Ophiura spp.</i>	Echinoderm		0.062
Orbiniidae	Polychaete	0.003	
<i>Owenia fusiformis</i>	Polychaete	0.013	0.008
Paguridae	Crustacean		0.024
<i>Parvicardium pinnulatum</i>	Bivalve	0.004	
<i>Pecten maximus</i>	Bivalve	0.003	
<i>Pectinariidae</i>	Polychaete	0.004	
<i>Pherusa plumosa</i>	Polychaete	0.058	0.038
<i>Pholas dactylus</i>	Bivalve		0.014
<i>Pilumnus hirtellus</i>	Crustacean		0.003
<i>Pisa spp.</i>	Crustacean		0.009
<i>Pisidia longicornis</i>	Crustacean		0.006
<i>Platyhelminthes</i>	Platyhelminthes	0.001	
Polynoidae	Polychaete	0.016	0.013
<i>Psammechinus miliaris</i>	Echinoderm	0.001	0.002
Sabellidae	Polychaete	0.045	
<i>Scalibregma inflatum</i>	Polychaete	0.011	
<i>Scoloplos armiger</i>	Polychaete	0.004	
<i>Sepiola spp.</i>	Cephalopod	0.001	
Serpulidae	Polychaete	0.005	0.006
<i>Spatangus purpureus</i>	Echinoderm		0.001

Running header: Aggregations of mussels and brittle stars

Spionidae	Polychaete	0.001	0.001
<i>Talochlamys pusio</i>	Bivalve		0.001
<i>Tapes aureus</i>	Bivalve	0.002	
<i>Tapes rhomboides</i>	Bivalve	0.005	
Terebellidae	Polychaete	0.020	0.017
<i>Timoclea ovata</i>	Bivalve	0.087	0.085
<i>Tritia incrassata</i>	Gatropod	0.001	0.005
<i>Trivia arctica</i>	Gatropod	0.002	
<i>Tubificoides spp.</i>	Polychaete	0.001	
<i>Velutina velutina</i>	Gatropod	0.002	

567

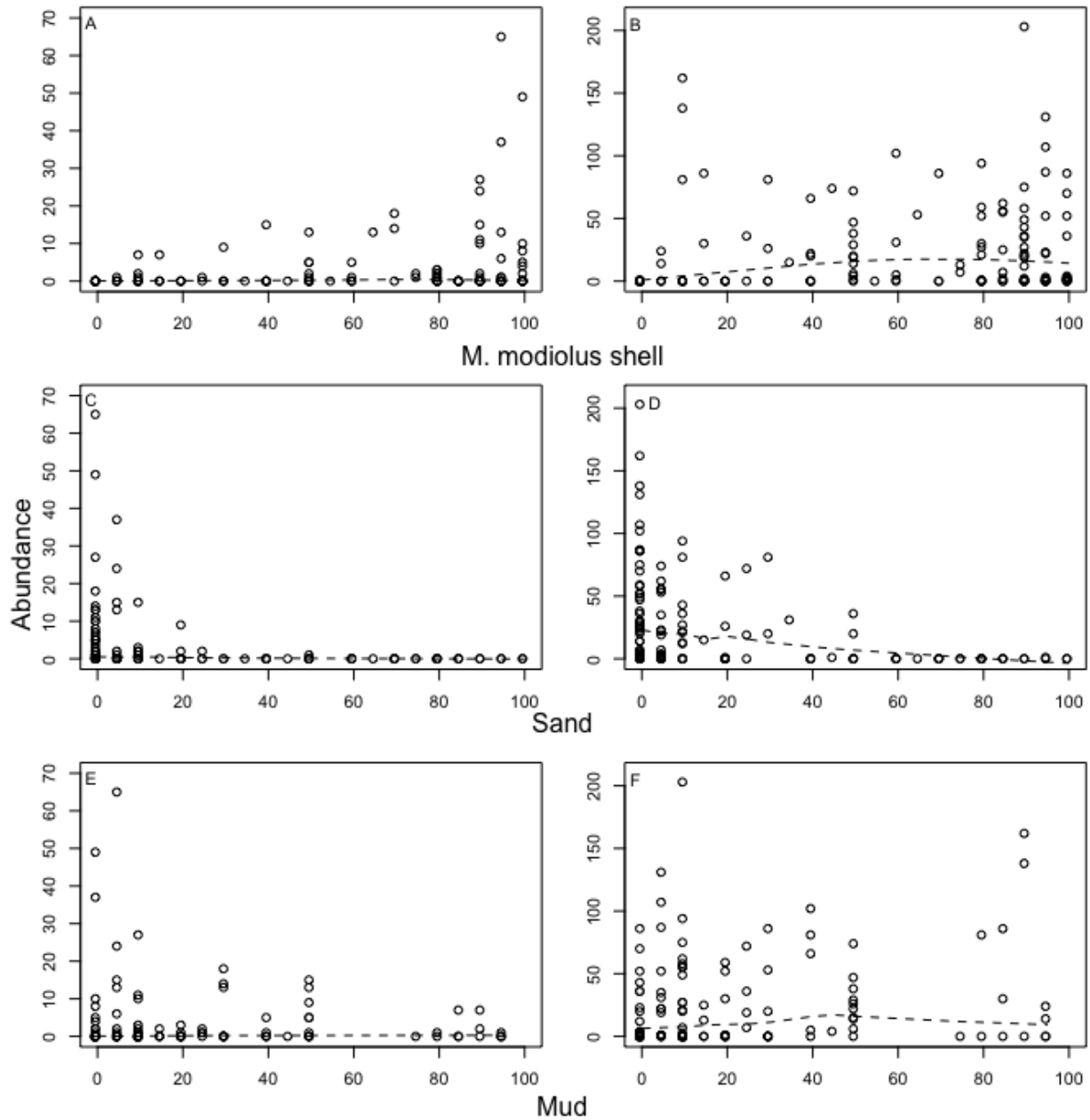
568

569 Appendix S4

570 The abundance of *M. modiolus* (left column) and *O. fragilis* (right column) in grab samples

571 compared to the percent cover of 3 substrate categories. Substrate type was determined from

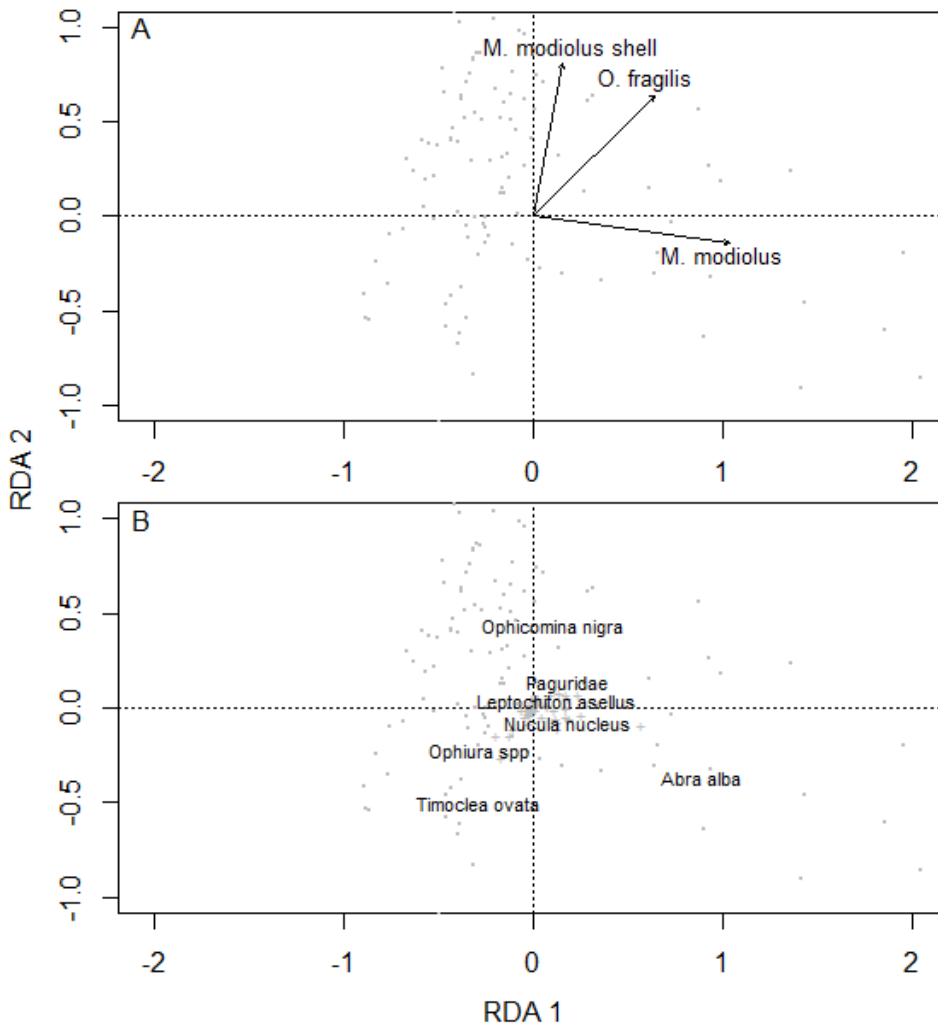
572 photos of grabs. Fitted lines were determined by the lowess function in R.



573

574

575 Appendix S5. The first and second axis of RDA of the fauna assemblage in grab samples
576 overlaid with vectors of predictor variables (A) and taxon centroids (B). The first axis explained
577 6.5% and the second explained 2.9% of the variation respectively. To make taxon labels readable
578 and to reduce clutter in B, only the most abundant taxon that did not overlap with other taxon
579 labels are shown. Grey dots represent individual samples and vector length is relative to the
580 variance explained by the variable.



581