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1 A spatially-resolved pressure-based approach to evaluate combined effects of human activities
2 and management in marine ecosystems

3
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14
15 **Abstract**

16 Our oceans are heavily utilised by a wide variety of human activities that exert pressures which
17 negatively impact marine ecosystems, occasionally leading to unsustainable rates of exploitation.
18 A linkage framework approach can be used to make independent associations between sectors,
19 activities and the pressures they introduce. However in reality, many different sectors and their
20 associated activities overlap in time and space, potentially changing the severity of their impact
21 as pressures combine, and undermine the efforts of environmental managers to mitigate the
22 harmful effects of those activities. Here we present a spatially resolved approach to assess the
23 potential for combined effects using a linkage framework assessment. Using illustrative
24 examples from the North East Atlantic, we show the likelihood of changes in pressure severity as
25 a result of multiple overlapping activities. Management options to limit pressure introduction are

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4 26 explored and their benefit - measured as a reduction in area of seafloor impacted - assessed. In its
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6 27 simplest form, the approach can be used to develop potential precautionary management options
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8 28 in areas where data availability is poor and more comprehensive management measures where
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10 29 data is more widely available.
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15 31 **Key words:** pressure; human activities; management; combined effects; maritime spatial
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17 32 planning
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22 34 **Introduction**

23
24 35 The exploitation of marine resources is causing widespread changes to the structure and state of
25
26 36 marine ecosystems, and at the broad scale, there are few areas of the ocean that are devoid of any
27
28 37 human activity (Halpern *et al.*, 2008). Managing the complex interactions of overlapping
29
30 38 multiple activities that compete for space remains a challenging task, but one which has been
31
32 39 recognised as essential if the sustainable provision of ecosystem goods and services from our
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34 40 oceans is to be achieved (McLeod *et al.*, 2005).
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42 42 In 2007, the European Union (EU) proposed an Integrated Maritime Policy (IMP) (EC, 2007) as
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44 43 the mechanism for providing a long-term (sustainable) improvement in quality of life. A key
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46 44 instrument of the IMP is Maritime Spatial Planning (MSP), which provides a framework for
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48 45 arbitrating between competing sectors and managing their impact on the marine environment and
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50 46 progression towards sustainably exploited ecosystems (Reid *et al.*, 2005; Halpern *et al.*, 2009).
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52 47 The success of MSP is challenged by the need to manage numerous human activities that vary in
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54 48 spatial and temporal footprint (Eastwood *et al.*, 2007), the pressures from which impact the
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4 49 marine environment (Crain, 2008; Knights *et al.*, 2013 and 2015) and overlap in many areas
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6 50 (Figure 1). As marine activities continue to expand in spatial and temporal footprint, competition
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8 51 for maritime space will undoubtedly increase such that there is greater conflict between sectors.
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10 52 MSP has the potential to play an important management role by resolving many of these spatial
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12 53 conflicts (Reid *et al.*, 2005; Halpern *et al.*, 2009), however management success is dependent on
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14 54 realistic assessments of impacts that could be undermined if the overlap of activities (combined
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16 55 effects) is not taken into consideration.
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22 57 Comparing the impact of human activities on the marine ecosystem requires both a method of
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24 58 linking human activities to specific pressures, (NB pressures are also referred to as stressors in
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26 59 the literature, e.g. Halpern *et al.*, 2008, Brown *et al.*, 2014), and a measure of sensitivity of the
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28 60 ecosystem and its components to those pressures (Stelzenmuller *et al.*, 2010a; Foden *et al.*,
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30 61 2011). Rogers (2005) defined a pressure as an anthropogenic factor that induces environmental
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32 62 change and which is generally viewed negatively i.e. a detrimental effect (Gabrielsen and Bosch
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34 63 2003). The types (e.g. Knights *et al.*, 2013) and distribution of pressures (e.g. abrasion, substrate
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36 64 loss, contamination) have been described individually for a number of major sectors and their
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38 65 activities that operate in marine habitats, such as fishing, artificial structures, dredging, shipping
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40 66 (e.g. Eastwood *et al.*, 2007; Foden *et al.*, 2011). Linkages between sectors, their activities and
41
42 67 ecological components of the ecosystem through pressures have also been described using
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44 68 linkage frameworks (e.g. Driver-Pressure-State-Impact-Response (DPSIR); White *et al.*, 2013),
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46 69 however linkages have tended to be viewed independently of one another, such that the full
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48 70 range of pressures affecting a characteristic may not be identified or managed effectively
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50 71 (Knights *et al.*, 2013).
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73 Management measures to limit the harmful effects of activities have thus far generally attempted
74 to protect a specific habitat, species or other feature of interest through management of a single
75 sector/activity and consideration of their pressures (Commission of the European Communities
76 (EC) 2009; Khalilian *et al.*, 2010). Implementation of management measures on a sector-by-
77 sector basis has the potential to miss impacts that only occur when pressures from multiple
78 sectors and their activities act in combination. In such cases, individual sectors would continue to
79 contribute to negative effects on the ecosystem.

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81 Several studies have shown how the severity of a pressure type can change amongst activities
82 depending on the frequency and intensity of introduction i.e. when activities overlap (Eastwood
83 *et al.*, 2007; Evans and Klinger, 2008; Halpern *et al.*, 2008). However, assessing severity is not
84 straightforward. It is often assumed that pressures act in an additive way i.e. the impact of two of
85 the same or different pressure types are simply added together (e.g. Halpern *et al.*, 2007, 2008;
86 Ban and Alder, 2008; Stelzenmuller *et al.*, 2010b). However, pressures may also interact in non-
87 linear, synergistic (Folt *et al.*, 1999; Folke *et al.*, 2004; Christensen *et al.*, 2006; Crain *et al.*,
88 2009) or antagonistic ways (Folt *et al.*, 1999) depending on the pressure type and ecological
89 component impacted such that the outcome of interactions of multiple pressures is hard to predict
90 (see Brown *et al.*, 2014 for an overview of considerations).

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92 Here, for simplicity, we consider only additive interactions, which we define as when the same
93 pressure type occurs in the same place. We refer to this as a *combined* effect. However, it should
94 be noted that this term (as well as *cumulative impact*; e.g. Crain *et al.*, 2009) has been used to

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4 95 describe situations when pressures of different type interact additively (Folt *et al.*, 1999;
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6 96 Breitburg & Riedel, 2005; Darling *et al.*, 2010). We do not consider other relationships (e.g.
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8 97 synergistic or antagonistic) on the basis that it is often unclear, in part due to a paucity of direct
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10 98 manipulative evidence of multiple stressor effects, how the severity of interacting pressures can
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12 99 change beyond a simple additive process of the same pressure type.
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17 101 Combined effects present a considerable challenge to resource managers such that the
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19 102 implementation of a management measure, which was perhaps designed to mitigate the impacts
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21 103 of a single activity, may be undermined by a failure to account for pressures arising from a non-
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23 104 target sector or activity which continue to impact on the ecosystem (Crain *et al.*, 2009) and lead
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25 105 to further or continued departure away from the environmental objective (Breen *et al.*, 2012).
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29 107 The consequences of combined impacts on ecosystems as a result of multiple pressures are
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31 108 discussed in a number of theoretical and empirical studies (Folt *et al.*, 1999; Folke *et al.*, 2004;
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33 109 Vinebrooke *et al.*, 2004; Christensen *et al.*, 2006; Halpern *et al.*, 2008) but quantifying their
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35 110 impact remains a challenge (see Darling and Côté 2008 and references therein). Recent advances
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37 111 have used spatial information to describe the occurrence and intensity of human activities
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39 112 coupled with the sensitivity of the ecosystem to that activity and its pressures (Halpern *et al.*,
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41 113 2007, 2008; Robinson *et al.*, 2013). These studies have been valuable in identifying features of
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43 114 the ecosystem at greatest risk from on-going human activities (e.g. Halpern *et al.*, 2007, 2008),
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45 115 however, the activities contributing to that risk are rarely specified, nor is the potential for
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47 116 management measures to mitigate that risk fully explored (but see Stelzenmüller *et al.*, 2010b
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49 117 and Knights *et al.*, 2015).
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119 Identifying threats to marine ecosystems, the areas at greatest risk of damage or loss from those
120 threats, and strategies to mitigate those risks is at the forefront of recent environmental policy
121 (e.g. the Marine Strategy Framework Directive (MSFD) (EC, 2008). With this in mind and
122 building on previous work, we have developed a geospatial assessment approach with three main
123 objectives. First, the mapping of major marine activities and associated pressures to identify
124 overlap, so that areas with the potential for combined effects can be identified. Second, the
125 outputs may be used in the development of an approach that enables different spatial
126 management options to be assessed and prioritised in terms of reductions in risk. And third, to
127 develop an approach without the need for costly and quantitative spatial monitoring to be
128 undertaken. Here, we demonstrate how our spatially-resolved combined effects framework,
129 when applied in its simplest form, can be used in the determination of management measures,
130 and explore the practical applications of how composite sector-pressure data layers could be used
131 to support the development and evaluation of those measures.

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133 **Methods**

134 A spatially-resolved approach was developed to assess the combined impact of multiple
135 activities. Building on a linkage framework describing the interactions between sector activities
136 and the ecosystem (White *et al.*, 2013) and an independent assessment of threat to the ecosystem
137 from each of those linkages (a 'pressure assessment', *sensu* Robinson *et al.*, 2013), we consider
138 the potential for combined effects to occur and the potential for their management.

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140 ***Linking Sectors, Activities, Pressures and Ecological components: Creating a Linkage Matrix***

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4 141 Links between sectors, activities, pressures and the ecosystem were identified and compiled to
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6 142 create three linkage matrices, each describing a series of links i.e. (1) sector → activity; (2)
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8 143 activity → pressure; and (3) pressure → ecological component. Each cell in a matrix describes
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10 144 the potential interaction between the two components, for example, indicating which pressures
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12 145 impact which ecological components (White *et al.*, 2013). All 3 matrices can be combined into a
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14 146 single matrix to describing the pathways through which sectors impact the environment. We
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16 147 refer to each pathway i.e. sector → activity → pressure → ecological component combination as
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18 148 an “impact chain” herein (Knights *et al.*, 2013). Impact chains were defined following an
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20 149 extensive review of the peer-reviewed scientific literature and published reports resulting in a
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22 150 combined matrix of 5,515 potential impact chains. In total, we considered 18 sectors, 98
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24 151 activities, 24 pressure types and 11 ecological components in the development of our linkage
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26 152 matrices.
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34 154 ***Assessing the severity of impact chains using a pressure assessment approach***

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36 155 We used a pressure assessment approach to qualitatively assess each impact chain using a
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38 156 categorical assessment of exposure and sensitivity criteria (see Robinson *et al.*, 2013 for details).
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40 157 Impact chains were assessed by expert judgment using five criteria: Exposure of the ecological
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42 158 component to a sector-pressure combination (1) spatial and (2) temporal overlap; (3) Severity of
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44 159 the interaction (i.e. *degree of impact*, *DoI* herein) which incorporates weighting on high versus
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46 160 low severity impacts and *chronic* versus *acute* effects); and Recovery from impact composed of
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48 161 (4) component resilience, and (5) pressure persistence. Each impact chain was evaluated
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50 162 considering prevailing conditions and applied here at a European regional sea scale. The expert
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52 163 group was comprised of >40 academics and researchers who were part of the ODEMM project
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4 164 (www.liv.ac.uk/odemmm) and originating from 17 partner institutions from 13 different countries.

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6 165 Experts undertook an extensive review of both primary and grey literature prior to their
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8 166 evaluations (e.g. Knights *et al.*, 2011) to ensure most up-to-date understanding and knowledge
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10 167 was used in making their judgement.
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14 169 *Spatial Analysis*

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17 170 To illustrate the application of our combined effects assessment methodology, we truncated the
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19 171 outcomes of the pressure assessment (Robinson *et al.*, 2013) to a sub-set of sectors namely:
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21 172 aggregates, fishing, oil and gas, renewable energy (specifically offshore wind farms) and
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23 173 telecommunications, and considered the impact of these sectors on sublittoral sediment habitats
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25 174 (EUNIS Level 2, Class A5.5, Connor *et al.*, 2004) of the North Sea region of the North East
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27 175 Atlantic (Figure 1). These sectors were chosen because of their current prevalence throughout
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29 176 much of the region, the fact that they all introduce multiple types of pressure that can impact on
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31 177 sublittoral habitats and species, and the public availability of spatial (georeferenced) datasets that
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33 178 describe the distribution of those sectors throughout the region (geo-data sources are given in full
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35 179 in Table S1).
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43 181 The spatial data we used was unmodified from its raw form with the exception of the fishing
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45 182 layer. In this case, data were derived from the Vessel Monitoring System (VMS) describing the
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47 183 speed and location of any individual vessels >15 m long operating in the North Sea. These data
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49 184 were filtered to include only information of vessels “at sea” i.e. in operation (after Lee *et al.*,
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51 185 2010) and included all gear types (mobile e.g. benthic and demersal trawls and static e.g. pots
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53 186 and nets). We further truncated the data to include only those vessels at sea for 24 h or more. We
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3 187 did not use VMS data grouped by annum, as the resolution is too coarse to enable spatial and
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6 188 temporal overlap with other sectors to be resolved. Similarly, data for vessels at sea for < 24 h
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8 189 were excluded on the basis that it can include erroneous or ‘non-fishing’ data points and lead to
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10 190 over-estimates of fishing pressure in some instances. By excluding these vessels, our analysis
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12 191 may alternatively under-estimate fishing pressure in some areas, but in the majority of cases,
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14 192 vessels classified as “< 24 h” tend to occur in the same areas as those vessels operating for > 24
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17 193 h, such that underestimates of seafloor area impacted by fishing are unlikely. For all other
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20 194 sectors, the licensed area was used to describe the sector distribution (see Table S1 for full
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22 195 details). Habitat maps were available as georeferenced data layers from EUSeaMap (EMODnet).
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24 196 All shapefiles (georeferenced polygons describing the spatial distribution of sectors and habitat)
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26 197 were compiled within a Geographic Information System (GIS) (ArcInfo 9.3, ESRI).

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32 199 *Mapping the spatial distribution of sector-pressures*

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34 200 Each sector generates several different pressure types: the fewest were generated by the
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36 201 telecommunication sector and the most by the fishing sector (Figure 2). The distribution of
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38 202 pressures can be described in one of two ways: (1) dispersive; or (2) non-dispersive. A dispersive
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40 203 pressure (e.g. marine litter) can spread beyond the operational spatial footprint of the sector,
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42 204 whereas a non-dispersive pressure can only occur where the sector is in operation (e.g. abrasion).
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44 205 Dispersive pressures therefore have the potential to result in combined or cumulative effects at
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46 206 locations different to their source of introduction. Predicting where those locations are, could be
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48 207 achieved using approaches such as hydrodynamic modelling (e.g. Mead and Rodger, 1991;
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50 208 Mead, 2004) but this extends beyond the proof-of-concept objective of this study. As such, here
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52 209 we only consider non-dispersive pressures to illustrate our framework, as the spatial footprint of
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4 210 the sector can simply be used as a proxy for pressure distribution. We recognise that in areas
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6 211 where dispersive pressures occur, we may underestimate combined or cumulative effects as a
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8 212 result of their exclusion from the analysis.
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12 214 *Pressure criterion: Degree of Impact (DoI)*

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15 215 The DoI criterion (see Robinson *et al.*, 2013), which describes the likely severity of the impact,
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17 216 was chosen as a primary mechanism to assess the combined impact of sector-pressures. The DoI
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19 217 of each pressure type can differ in classification between either a **low severity** impact pressure, a
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21 218 **chronic** severe impact pressure, or an **acute** severe impact pressure (see definitions in Table 1
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23 219 and distributions amongst sectors in Figure 2). Low severity pressures were removed from the
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25 220 analysis based on the definition that, irrespective of the frequency or magnitude of introduction
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27 221 (or overlap) of that pressure, no significant adverse effects on the ecosystem would occur
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29 222 (Robinson *et al.*, 2013). Here, a significant adverse effect is seen when a pressure causes changes
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31 223 in characteristic structure or functioning of a habitat. Furthermore, at this stage we only
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33 224 considered additions of the same pressure type (e.g. abrasion from fishing combining with
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35 225 abrasion from aggregate dredging) rather than combinations of different pressure types (e.g.
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37 226 abrasion from fishing with changes in siltation from aggregates), although we recognise the
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39 227 potential for such interactions to occur (see introduction above for references).
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48 229 The distribution of each pressure was mapped in GIS (ArcInfo 9.3, ESRI) based on the spatial
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50 230 distribution of the sector and the number of pressure polygons determined from the linkage
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52 231 tables (after White *et al.*, 2013). Attributes of each polygon included the DoI which was assigned
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54 232 to the pressure type from the Robinson *et al.*, (2013) pressure assessment database (Figure 2).
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4 233 The extent of each sector and its overlap with the seafloor was calculated using join and union
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6 234 tools from the analysis toolbox (ArcInfo 9.3, ESRI) and the number of overlapping pressures was
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8 235 calculated using merge (data management) and intersect (analysis) tools, allowing the area of
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10 236 pressure overlap(s) with the seafloor (km²) to be estimated.

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15 238 *An independent assessment of acute pressure distribution*

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17 239 Based on the spatial distribution of pressures and following an independent assessment of degree
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19 240 of impact for those pressures, the distribution and number of acute pressures per unit area of
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21 241 seafloor was determined. The number of acute pressures was calculated by overlaying all
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23 242 pressure shapefiles and the number of acute pressures present summed with no assumption of
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25 243 combined effects or pressure interaction, e.g. the potential for chronic severity pressures to
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27 244 become acute in severity was ignored (Figure 5a).

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34 246 *A combined assessment of acute pressure distribution*

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36 247 Where pressures overlap in space or time, there is potential for combined effects to occur which
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38 248 may alter the severity of the impact on sublittoral habitats. A set of rules was developed to
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40 249 determine if and when the severity of a pressure could change (e.g. from chronic to acute
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42 250 severity) following overlap of the same pressure type from different sectors.

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48 252 Spatial overlap is a relatively straightforward concept, although the resolution of data describing
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50 253 the spatial extent of sector pressures is generally coarse such that methods are used (e.g.
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52 254 buffering, Lee *et al.*, 2010) when plotting spatial extents to allow for uncertainty. Here, we
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54 255 adopted a precautionary approach whereby we consider the spatial extent of pressures as the

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4 256 scale of the licenced area of the sector rather than considering smaller spatial scales (e.g. the

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6 257 foundation of a single wind turbine).

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10 259 Temporal overlap is perhaps a more challenging concept. Here, ‘time’ refers to not only the

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12 260 persistence of the pressure (see Robinson *et al.*, 2013 for definitions) but also its intensity. This

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14 261 interaction determines how frequently a pressure type needs to be introduced to cause a

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16 262 combined effect. However, the persistence and intensity thresholds required for combined effects

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18 263 to occur are unknown and remain un-quantified.

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24 265 On that basis, we assumed that when two or more chronic pressures of the same type (but from

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26 266 different sectors) overlapped, all contributing impact chains in that area would become acute in

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28 267 severity (Figure 3). For example, fishing and aggregate dredging both introduce a chronic

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30 268 pressure (e.g. sedimentation) and where the footprint of these sectors overlaps in space, the

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32 269 pressure combines and becomes acute i.e. causing instantaneous mortality due to the volume of

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34 270 sediment introduced to an area (Figure 3). We recognise that both the spatial and temporal

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36 271 overlap assumptions made may lead to overestimates of combined effects in some cases and

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38 272 therefore the outcomes are precautionary.

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45 274 In areas where an acute pressure is present, the area of seafloor underlying that pressure was

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47 275 assumed to be immediately impacted (i.e. organisms are instantly killed). Given that “acute” was

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49 276 the most severe degree of impact category, this severity category cannot become more severe

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51 277 irrespective of whether or not there is overlap. As such, DoI assessment of each impact chain had

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53 278 one of three possible states: (1) DoI (stays) chronic, (2) DoI (stays) acute, or (3) DoI changes

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4 279 state from chronic to acute. For each linkage, all possible DoI states were calculated should
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6 280 pressure overlap occur (i.e. DoI stays the same or changes from the initial assessment). All
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8 281 possible DoI states were determined using a chain of conditions and functions in *R* (*R*
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10 282 Development Core Team, 2011) (Figure 4). The resulting DoI state data table was then joined to
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12 283 data describing the spatial extent of the targeted sectors in ArcGIS and a Python script used to
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14 284 determine which of the DoI states was appropriate in a given unit area km². The spatial extent of
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16 285 sector-pressures (km²) of combined pressures was estimated in the same way as the independent
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18 286 assessment above. Difference plots (e.g. Figure 5c) were generated by subtracting the number of
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20 287 pressures km⁻² in the combined assessment from the number of pressures km⁻² in the independent
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22 288 assessment. For brevity, only the spatial extent of acute pressures (km²) is shown although
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24 289 similar figures can be produced for areas of chronic DoI where the intensity of pressure
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26 290 introduction may be an issue.
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34 292 *Assessing the performance of management measures in light of overlapping sector activities*

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36 293 To assess the potential for single-sector management to alleviate pressure and reduce impact on
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38 294 the seafloor, we identified a case study area of 67,500 km² (Figure 1 inset) and evaluated the
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40 295 change in spatial extent of the seafloor (km²) impacted by sector-pressures if a single sector was
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42 296 excluded entirely. However we recognise that it may not be feasible to remove some sectors
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44 297 entirely (e.g. permanent structures such as oil and gas and renewables). Changes in impacted
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46 298 seafloor area (km²) were calculated for five scenarios with management targeted at one of five
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48 299 sectors in each scenario. Given that each sector introduces at least one acute pressure (Figure 2),
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50 300 the spatial extent of the sector footprints could be used without the need to differentiate between
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52 301 pressure types.
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Results

When considering all five sectors in combination, there was no change in the total spatial extent of acute pressure between the independent assessment and the combined assessment (Figures 5a and b). This was because all sectors included in the assessment introduce at least one pressure type that is acute in its severity (Figure 2). When considering each sector and its pressures separately (e.g. fishing-abrasion or aggregate extraction-abrasion as a separate un-combined footprint), a large area of the seafloor was impacted by at least one acute sector-pressure, and in worst cases, up to 30 acute sector-pressures per km² (Figure 5b) where there was considerable overlap between multiple sectors (Figure 1).

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The change in the number of acute sector-pressures occurring per km² of seafloor between the independent and combined assessment increased indicating the presence of overlap between two or more chronic pressures of the same type (Figure 5c). However, the area of overlap and a change in pressure severity (i.e. chronic to acute) was limited to relatively small areas of the North East Atlantic and there was no change in the total number of acute sector-pressures km² across large areas of the region, irrespective of whether an independent or combined assessment methodology was used (Figure 6). Both methodologies estimated ~89 % of the North East Atlantic seafloor area was impacted by fewer than five acute sector pressures per km², indicating that, of the sectors included in this assessment, large areas of the seafloor are impacted by a single sector operating in isolation (Figure 6).

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4 324 In areas where sector-pressure overlap did occur, the combined assessment revealed relatively
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6 325 small (~140 km²) areas of seafloor where the number of acute sector-pressures km² increased
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8 326 beyond levels predicted by the independent assessment (Figure 6). In the independent
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10 327 assessment, a maximum of 18 acute pressures (km⁻²) were identified, but this increased to a
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12 328 maximum of 30 acute sector-pressures (km⁻²) in some areas in the combined assessment (Figures
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15 329 5a, 5b and 6). The increase in pressure density was primarily due to the overlap of fishing with
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17 330 the non-renewable energy (oil and gas) and offshore renewable energy sectors. There was an
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19 331 especially large increase in the areas with 16-20 pressures per km² (equivalent to 4,153km²)
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21 332 attributed to the overlap between offshore renewable energy, aggregate extraction and fishing in
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23 333 the southern North Sea area (Figure 1 inset area).
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28 29 335 *Single Sector Management*

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31 336 Overall 63% of the seafloor (42,839km²) within the management area explored was impacted by
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33 337 one or more sectors (Figure 7 and Table 2). The spatial distribution of each sector varied
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35 338 between sectors, ranging from 20 km² in the case of telecommunications to in excess of 30,000
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37 339 km² for fishing (Table 2). Estimates of the overlap of a single sector with any other sector ranged
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39 340 from as little as 45% of the total sector extent in the case of telecommunications, up to a
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41 341 maximum of 90% for offshore wind farms (Table 2). The potential for single-sector management
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43 342 to reduce combined effects on the seafloor was evaluated by determining the extent to which
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45 343 sector-pressures occurs in isolation or overlap with one another. Reductions would only be seen
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47 344 if sector-pressures occur in isolation. The greatest reduction in impacted seafloor area under this
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49 345 scenario/ approach would be achieved following the exclusion of fishing (37 %) (not least due to
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51 346 the greater spatial extent in comparison to other sectors), followed by oil & gas (22 %), offshore
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4 347 wind (2.4 %), aggregates (1.5 %) and telecommunications (< 0.1 %) respectively (Figure 7,
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6 348 Table 2).

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10 350 **Discussion**

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12 351 We developed a methodology to assess the potential for combined effects from multi-sector use
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14 352 of marine ecosystems in the North East Atlantic, which if unaccounted for, can lead to
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16 353 underestimates of threat to marine ecosystems and undermine management objectives and
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18 354 measures. We applied a simple set of rules to available spatial datasets describing the distribution
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20 355 of marine sectors coupled to an expert judgement assessment of threats to marine ecosystems
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22 356 from sector activities. Similar data have previously been used to indicate where sectors can be in
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24 357 competition for maritime space (Ban and Alder, 2008; Stelzenmüller *et al.*, 2010b). Where
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26 358 overlap occurs, pressures have the potential to become more severe in their impact. Our approach
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28 359 evaluates how and where overlapping pressures might occur and represent a greater risk to the
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30 360 marine environment. This approach will assist resource managers to identify areas that are not
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32 361 only potentially more difficult to manage but also areas where improvements could be achieved
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34 362 more simply, either by way of cross-sectoral or pressure management (e.g. Knights *et al.*, 2013;
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36 363 Piet *et al.*, *in press*), or by maritime spatial planning to reduce overlap between sectors in space
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38 364 and time. Coupling these outcomes with more detailed maps of sensitive ecosystem
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40 365 characteristics or protected features should further support managers in achieving environmental
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42 366 objectives. The approach was designed to make use of existing data and expert knowledge,
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44 367 making it a cost-effective mechanism for prioritising and streamlining management activities.
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46 368 Minimising costs is desirable for policy makers in the current austere economic climate,
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4 369 especially where there is a desire to meet environment targets towards sustainable use of marine
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6 370 space.
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10 372 Comparison of the independent and combined effects assessment outputs demonstrated the
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12 373 potential for independent assessments to underestimate threats to the ecosystem by failing to
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14 374 consider potential changes in pressure severity where sectors overlap. This was demonstrated
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16 375 considering a simple additive interaction between the same pressure type and non-dispersive
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18 376 pressures. If cumulative impacts are as pervasive in the environment as proposed (e.g. Darling &
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20 377 Côté, 2008; Crain *et al.*, 2009), it can be hypothesised that our combined estimate is in fact a
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22 378 ‘best-worse’ case scenario rather than a worst-case scenario. Inclusion of other interactions e.g.
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24 379 between different pressure types as well as synergistic and antagonistic relationships (see Brown
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26 380 *et al.*, 2014 and references therein), can be expected to lead to further increases in the threat to
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28 381 ecosystem habitats and ideally should be included in the future, although it remains to be seen
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30 382 the extent and change in severity that will occur as a result of such interactions.
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34 384 Considering only non-dispersive pressures is also likely to underestimate the spatial extent of
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36 385 pressure overlap and thus combined and cumulative effects, as numerous pressures are dispersive
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38 386 and may well move beyond the physical footprint of a sector. Modelling approaches (e.g.
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40 387 hydrodynamic modelling, Mead and Rodger, 1991) or the use of buffers around known pressure
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42 388 distributions (e.g. Eastwood *et al.*, 2007; Halpern *et al.*, 2008) could shed further light on areas
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44 389 that may be more susceptible to combined effects from dispersive pressures, although in the
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46 390 latter case, such an approach should be driven by some underlying understanding of possible
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48 391 dispersal distance as otherwise, applying a fixed buffer distance (i.e. assuming symmetrical
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4 392 dispersal) could misrepresent locations of combined effects as dispersal distances (kernel) are

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6 393 often asymmetrical (e.g. Nickols *et al.*, 2015).

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10 395 While the total spatial extent of sector activities were the same in both the independent and

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12 396 combined assessments, the independent assessment did not identify areas where multiple sectors

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14 397 overlap (sectoral conflict and management complexity) or areas that may be under greater threat

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16 398 from pressures as a result of combined effects. Current policy drivers such as the Marine

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18 399 Strategy Framework Directive (EC, 2008) promote the use of an ecosystem-based approach to

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20 400 management to improve ecosystem health and (any) improvement will be reliant upon

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22 401 appropriate interventions to reduce or remove pressures that are causing negative effects on the

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24 402 functioning of the ecosystem. An independent assessment could mislead managers into

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26 403 developing management strategies that are unable to reduce the magnitude of threat (Breen *et al.*,

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28 404 2012) and/or lack the complexity to manage all relevant combining pressures, resulting in some

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30 405 activities or pressures being unregulated (Smith *et al.*, 2007).

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34 407 It is common that the same pressure type is introduced by different activities, which overlap in

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36 408 time and/or space (Stelzenmuller *et al.*, 2010b; Knights *et al.*, 2013). Here, using relatively

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38 409 simple scenarios that excluded a single sector from an area where considerable sector conflicts

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40 410 occur, we have shown the limited environmental benefits that might be achieved (i.e. a reduction

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42 411 in impacted seafloor area) versus the cost (a 'cost' being a reduction in the operational area of a

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44 412 sector) of doing so. When the cost of management exceeds the perceived benefits and assuming

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46 413 decisions are economically biased, it is considered unlikely that a management measure will be

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48 414 considered viable (Baral *et al.*, 2004). Here, we have not considered the value of an area of

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4 415 seafloor to a particular industry in terms of its 'quality' (Klein *et al.*, 2013) and used a broad
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6 416 habitat classification (i.e. EUNIS Level 2, sublittoral sediment) in our analysis rather than
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8 417 adopting more specific habitat classifications. Clearly spatial management decisions are likely to
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10 418 be driven by the value of the resource to society versus the cost of management and value to the
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12 419 environment (Klein, 2013; Auerbach *et al.*, 2014; Knights *et al.*, 2014) and an analysis using
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14 420 more highly resolved habitat classifications could easily be undertaken to determine if certain
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16 421 habitats that are identified as being particularly valuable to the region from an environmental,
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18 422 societal or economic perspective are impacted by combined effects.
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24 424 We considered only five of the many sectors that operate in the North East Atlantic (Halpern *et*
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26 425 *al.*, 2008; Knights *et al.*, 2013), yet large areas of seafloor were impacted by multiple
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28 426 overlapping acute pressures. The inclusion of additional sectors is likely to further compound the
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30 427 scale of effects from multiple pressures on seafloor integrity. Managing sectors in isolation is
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32 428 therefore only likely to partially address problems of environmental degradation and in the worst
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34 429 cases, management objectives may have no net benefit due to other acute pressures remaining
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36 430 unmanaged. This emphasises the need for a multi-sectoral approach to management of combined
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38 431 pressures if improvements in seafloor state are to be achieved.
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46 433 Overlap between sector activities presents several challenges to managers implementing an
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48 434 ecosystem approach. Firstly, it is unlikely that any one management measure will control all
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50 435 drivers that influence a policy objective/target such as good environmental status (GES). Rather,
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52 436 it is more likely that a suite of measures (i.e. a strategy) will be required to control the threats
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54 437 from pressures arising from multiple sectors and activities (Knights *et al.*, 2013) such that
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4 438 ecosystem state improves. This alternative approach to management has been shown to be
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6 439 successful in several cases where it has been applied (e.g. the Nitrates Directive)
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8 440 (see OSPAR, 2010 for further details). The implementation of pressure-targeted management
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10 441 programmes, does however, have its own challenges. These include the considerable time and
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12 442 resource commitments required for participating and building capacity among stakeholders
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14 443 (often arising from the limited trust and sectoral protectionism between sectors) as well as the
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16 444 capacity and/or willingness to participate in long-term integrated management programmes
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18 445 (Rutherford *et al.*, 2005; Knights *et al.*, 2014). However, these challenges can be overcome, for
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20 446 example, by developing measures in conjunction with stakeholders that are easily associated
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22 447 with a specific policy objective and display tangible targets and benefits (Watson, 2005;
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24 448 Carwardine *et al.*, 2009).
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31 450 **Improvements to future assessments**

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34 451 We have presented an approach that can be used to assess and identify areas of habitat that are
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36 452 greatest risk from human activities, providing a mechanism for targeted decision-making by
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38 453 resource managers that best supports current environmental objectives. Here, our objective was
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40 454 to demonstrate proof-of-concept, doing so using broad-scale distribution maps of sector activities
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42 455 and non-dispersive 'static' pressures to illustrate key considerations. However, the approach can
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44 456 be applied to address impacts at other, more refined, spatial scales. For example, the
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46 457 incorporation of high-resolution GIS data layers describing the location of wind turbines or oil
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48 458 well heads could be used to address small-scale impacts at a site level, rather than the broad scale
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50 459 approach we have implemented here which adopted licenced areas as its basis. In this instance,
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52 460 we may well overestimate combined impacts.
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6 462 In developing the underlying rules of the methodology, we made some general assumptions.
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8 463 First, we assumed that the recovery of the habitat to a pre-impacted condition would be
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10 464 instantaneous once the pressure was removed, but in reality, sublittoral habitats can take days,
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12 465 weeks or years to recover from an impacted to a pre-impacted state (Thrush and Dayton, 2002)
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14 466 depending on local conditions. This limitation could be addressed by the inclusion of habitat
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16 467 resilience data to improve estimates of the relationship between impact and recovery (see Eno *et*
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18 468 *al.*, 2013).
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22 469 Second, we assumed that when a chronic pressure of the same type (e.g. sedimentation) was
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24 470 introduced by two or more sectors (e.g. fishing and aggregate dredging) that co-occurred, then
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26 471 both pressure impact chains changed from chronic to acute in their severity (see Table 1). We
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28 472 used this simplified assumption for demonstration purposes only, but recognise that the intensity
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30 473 of pressure introduction may vary both within and among sectors depending on the specific
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32 474 activity being undertaken, such that the number of overlaps that are required to impart a change
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34 475 in degree of impact from chronic to an acute severity may vary. In many cases, the intensity of
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36 476 introduction of a specific pressure type by a specific sector activity is unknown, as is the critical
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38 477 intensity threshold that is required to impart a detrimental impact on the ecosystem or any of its
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40 478 components.
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45 479 Third, we assumed that where acute interactions occur, and irrespective of the number of
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47 480 overlapping acute pressures, the severity of interaction would not increase. This assumption was
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49 481 based on our definition that an acute pressure would lead to instantaneous mortality, which we
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51 482 view as the worst-case scenario, and further additions could not lead to further reductions in
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4 483 ecosystem state (Table 1). An improved understanding of interaction of pressures where they
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6 484 overlap will shed light on whether this is an appropriate assumption moving forward.
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10 486 Ecosystem-based management is continually challenged by the complexity of interactions
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12 487 between on-going sector activities and the environment. It is clear that combined effects are
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14 488 prevalent in many areas of the ocean as a result of competition for resources such that the
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16 489 pressures generated by those activities overlap, which can lead to more severe impacts than if
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18 490 those sectors occurred in isolation, presenting complex spatial management issues if
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20 491 conservation is to be achieved. Our approach utilises broad-scale geographic data coupled with
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22 492 an expert judgement assessment of threat that does not require exhaustive and prohibitively
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24 493 expensive studies to underpin it. As such, we suggest that the approach can be applied to assess
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26 494 and identify areas of habitat that are at greatest risk from human activities and provide a
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28 495 mechanism for targeted decision-making by resource managers that best support current
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30 496 environmental objectives.
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36 653 EU sea map disclaimer
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38 654 *Information contained here has been derived from EUSeaMap Consortium webGIS data*
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40 655 (www.jncc.gov.uk/page-5040) which is made available under the pilot project for the European
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42 656 *Marine Observation Data Network (EMODnet), funded by the European Commission’s*
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44 657 *Directorate-General for Maritime Affairs and Fisheries (DG MARE).*
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49 659 *Fishing VMS data*
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51 660 *VMS data were provided by the UK’s Department for Environment, Food and Rural Affairs*
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53 661 *(Defra) in raw, uninterpreted form. The Secretary of State for the Environment, Food and Rural*
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662 *Affairs does not accept any liability whatsoever for the interpretation of the data or any reliance*
663 *placed thereon.*

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Table 1. (a) Degree of Impact definitions used to classify a sector-pressure combination in the pressure assessment, and (b) the rules to determine possible Degree of Impact changes should pressure overlap occur. **NB** Degree of Impact classification can vary between pressure types but not within pressure types (i.e. among sectors) (after Robinson *et al.* 2013)

a). Degree of Impact classification definitions		
<p>A Low severity (L) DoI is an interaction that, irrespective of the frequency and magnitude of the events, never causes a noticeable effect for the ecological component of interest in the area of interaction. There are never high levels of mortality, sustained and noticeable reductions in breeding/recruitment success, or loss of habitat or change in its typical species or functioning, at the spatial scale of the interaction, i.e. proximate ecological responses (<i>sensu</i> Harley <i>et al.</i>, 2006).</p> <p>NB These sector-pressure combinations were removed from the analysis.</p>	<p>A Severe-Chronic (C) interaction is described as an impact that will eventually have severe consequences at the spatial scale of the interaction, if it occurs often enough and/or at sufficiently high levels e.g. where disease levels might build up over time, eventually leading to levels where a high number of individuals would be killed or habitat features would change. No inference is made as to when the pressure impact becomes severe; simply that at some frequency and intensity, a pressure can lead to severe impacts on that ecological component.</p>	<p>A Severe-Acute (A) interaction is described as a severe impact over a short duration e.g. for species, a high proportion of individuals are killed immediately where there is an interaction of the pressure and the component. In the case of habitats, such interactions cause an immediate change in habitat type, i.e. change or loss of characteristic features and/or species in the area of interaction. An Acute (A) interaction can occur after just one event.</p>
b). Rules for changes in DoI classification		
DoI remains Chronic	DoI becomes Acute	DoI remains Acute
<ul style="list-style-type: none"> • There is no overlap between a pressure of the same type (e.g. abrasion) arising from different sectors. • Pressure of the same type (e.g. abrasion) from two or more sectors overlap in space but not time. 	<p>Pressure of the same type (e.g. abrasion) from two or more sectors overlap in space and time.</p>	<p>The pressure DoI is already acute.</p>

Table 2. The extent of sublittoral seafloor impacted by acute sector-pressures when a sector is managed in isolation using a hypothetical management box (Figure. 1 inset and Figure. 7).

Sectors and Management Scenarios (a-e)	Sector spatial extent (km²)	Area not overlapping with another sector (km²)	Area overlapping with one or more sectors (km²)	Remaining area of seafloor impacted after sector removal (km²)
(a) Aggregates	4,243	660	3,583	42,179
(b) Fishing	30,185	16,050	14,135	26,789
(c) Oil & Gas	17,152	9,383	7,769	33,456
(d) Renewables	10,463	1,017	9,446	41,822
(e) Telecoms	20	8	12	42,831

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3 Figure 1. The location and overlap of sectoral activities (top) and the number of sector-
4 pressures impacting the seafloor (bottom) for a discrete area of the North East Atlantic. Inset:
5 a hypothetical management area (67,500 km²) for evaluating the performance of management
6 areas in light of combined pressure effects.
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10 Figure 2. The number of different pressure types introduced by five different sectors and
11 classified by their degree of impact (low; chronic; acute) that operate in a discrete area of the
12 North East Atlantic. The Degree of Impact categories Low, Chronic, Acute are defined in
13 Table 2.
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17 Figure 3. Illustration of combined effects where sector-pressures overlap in the same
18 geographic space. Each cell represents a unit area of seafloor. Chronic pressures from sectors
19 A and B overlap in space/time resulting in combined effects and an increase in severity (e.g.
20 cell 1a). Where an acute sector-pressure occurs, there is not further increase in pressure
21 severity (e.g. cells 1b and 3c). Where there is no pressure overlap, the pressure severity does
22 not change (e.g. cells 1c, 2b and c, and 3b).
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27 Figure 4. A schematic of the combined effects framework. Spatial data were compiled within
28 ArcGIS 9.3. Changes in pressure DoI were predicted if overlap were to occur between a
29 sector-pressure type using the conditions described in Table 2 resulting in multiple outcomes
30 for a given sector-pressure combination being generated. A Python script in ArcGIS was used
31 to select the appropriate DoI combination per unit area of space (km²) i.e. if overlap occurred
32 given the spatial extent of sector-pressures.
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36 Figure 5. The spatial extent of acute sector-pressures derived from (a) independent
37 assessment, (b) Combined assessment and (c) difference plot indicating an increase in the
38 number of acute sector-pressures per unit area following the combined assessment.
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42 Figure 6. The number of acute-sector pressures per km² identified by the independent (white
43 bars) and combined (grey bars) assessments. The independent assessment identified no areas
44 where acute sector-pressure density exceeded 21 km⁻².
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48 Figure 7. Change in the spatial extent of sublittoral seafloor that is impacted when a sector is
49 removed from the area. Estimated changes in the impacted extent (km²) are given in Table 2.
50 Scenarios tested were the removal of: a) Aggregates; b) Demersal fishing; c) Oil and gas
51 infrastructure; d) Renewable energy installations; and e) Telecommunications. Box area =
52 67,500km².
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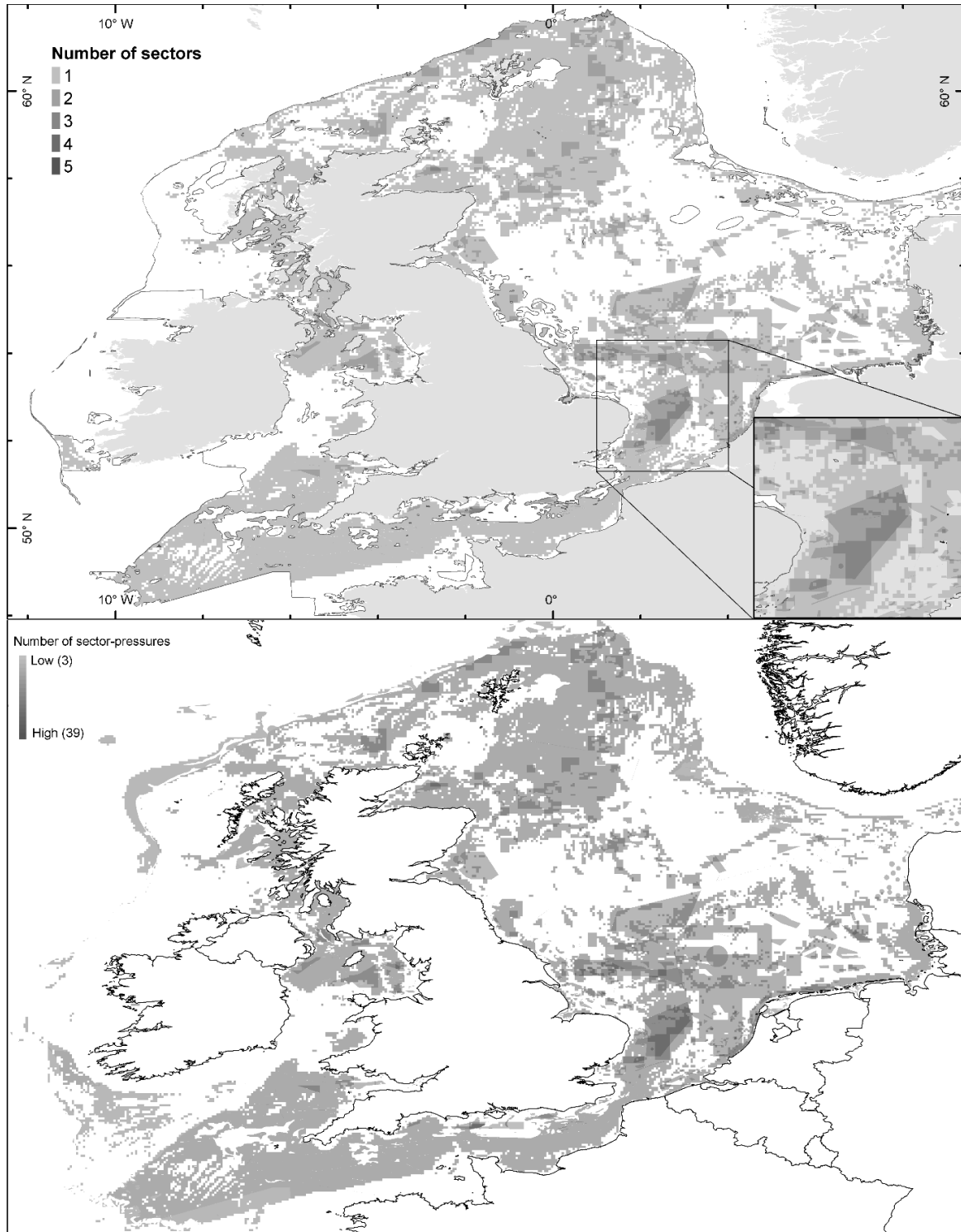


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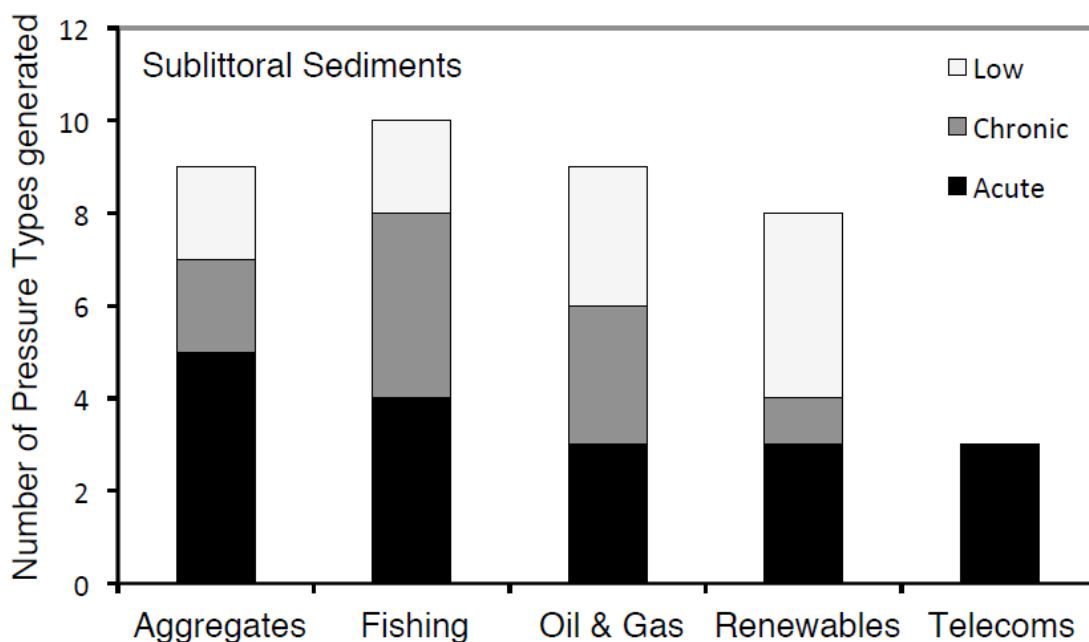


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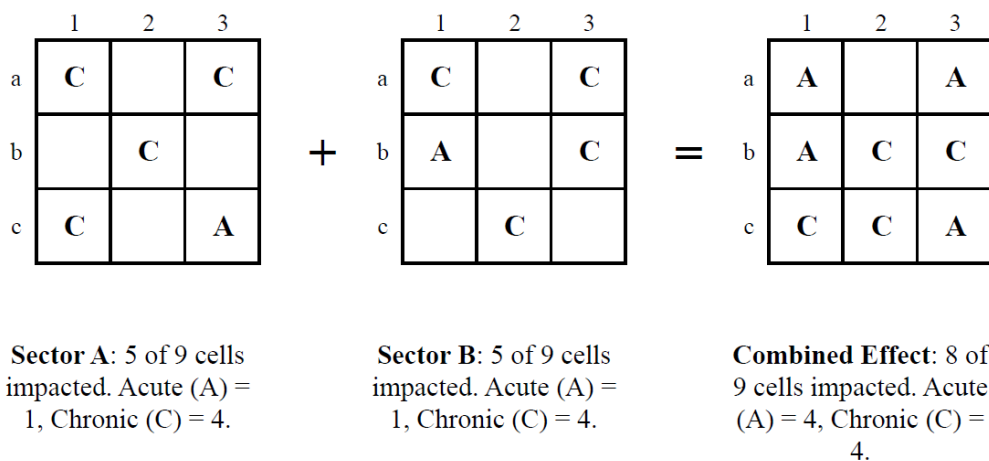


Figure 3. Illustration of combined effects where sector-pressures overlap in the same geographic space. Each cell represents a unit area of seafloor. Chronic pressures from sectors A and B overlap in space/time resulting in combined effects and an increase in severity (e.g. cell 1a). Where an acute sector-pressure occurs, there is not further increase in pressure severity (e.g. cells 1b and 3c). Where there is no pressure overlap, the pressure severity does not change (e.g. cells 1c, 2b and c, and 3b).

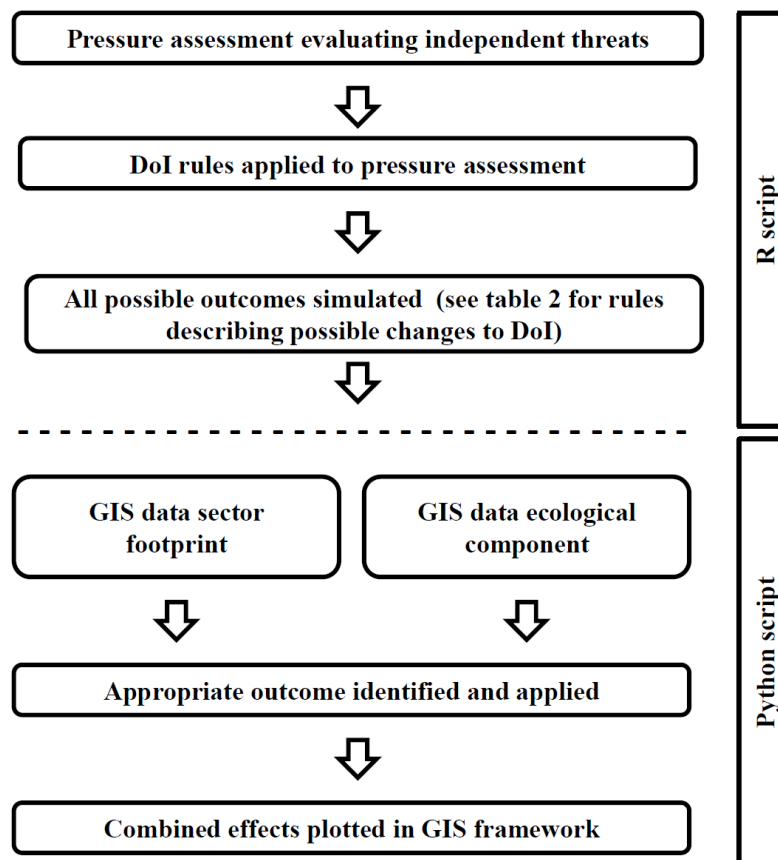


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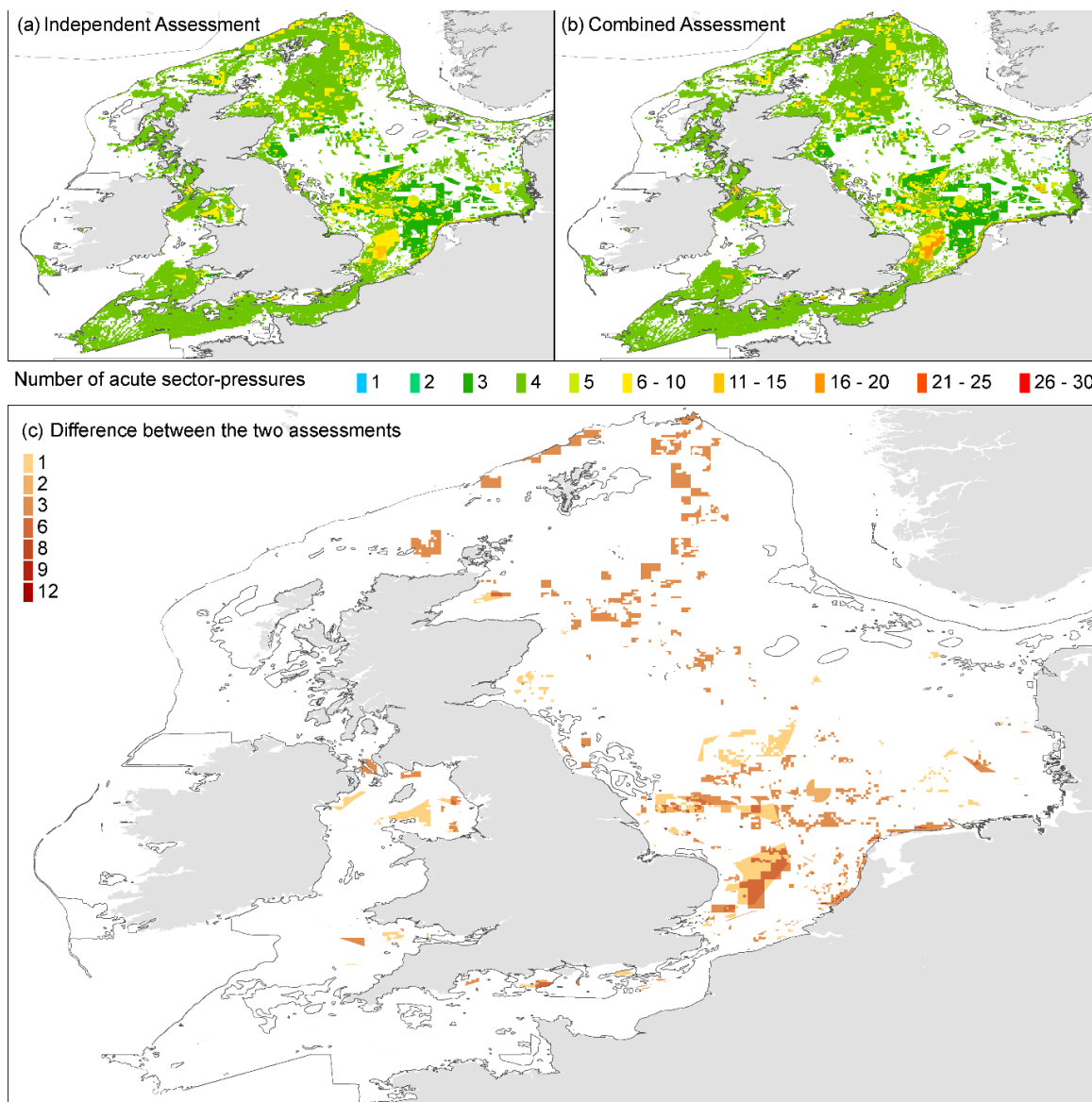


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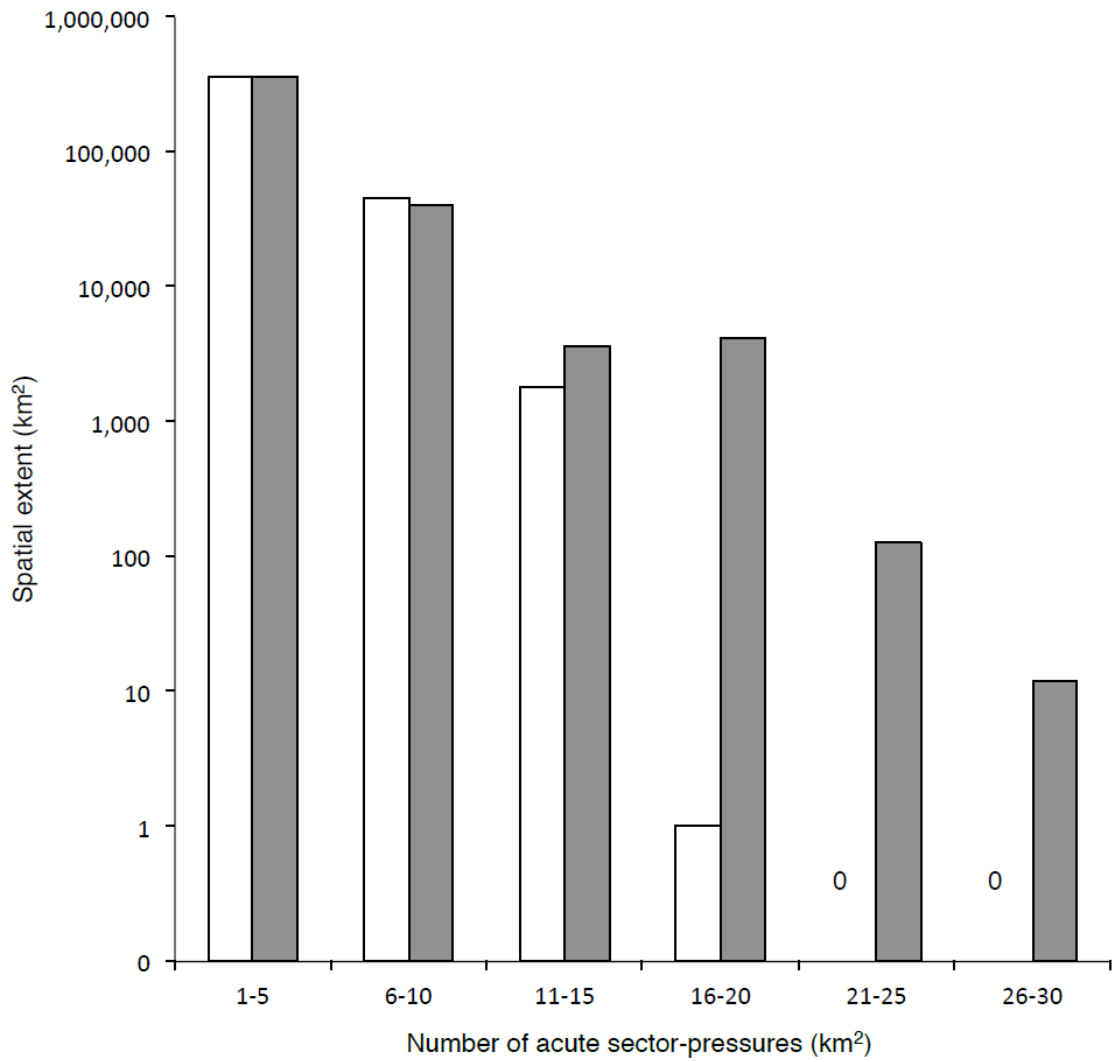
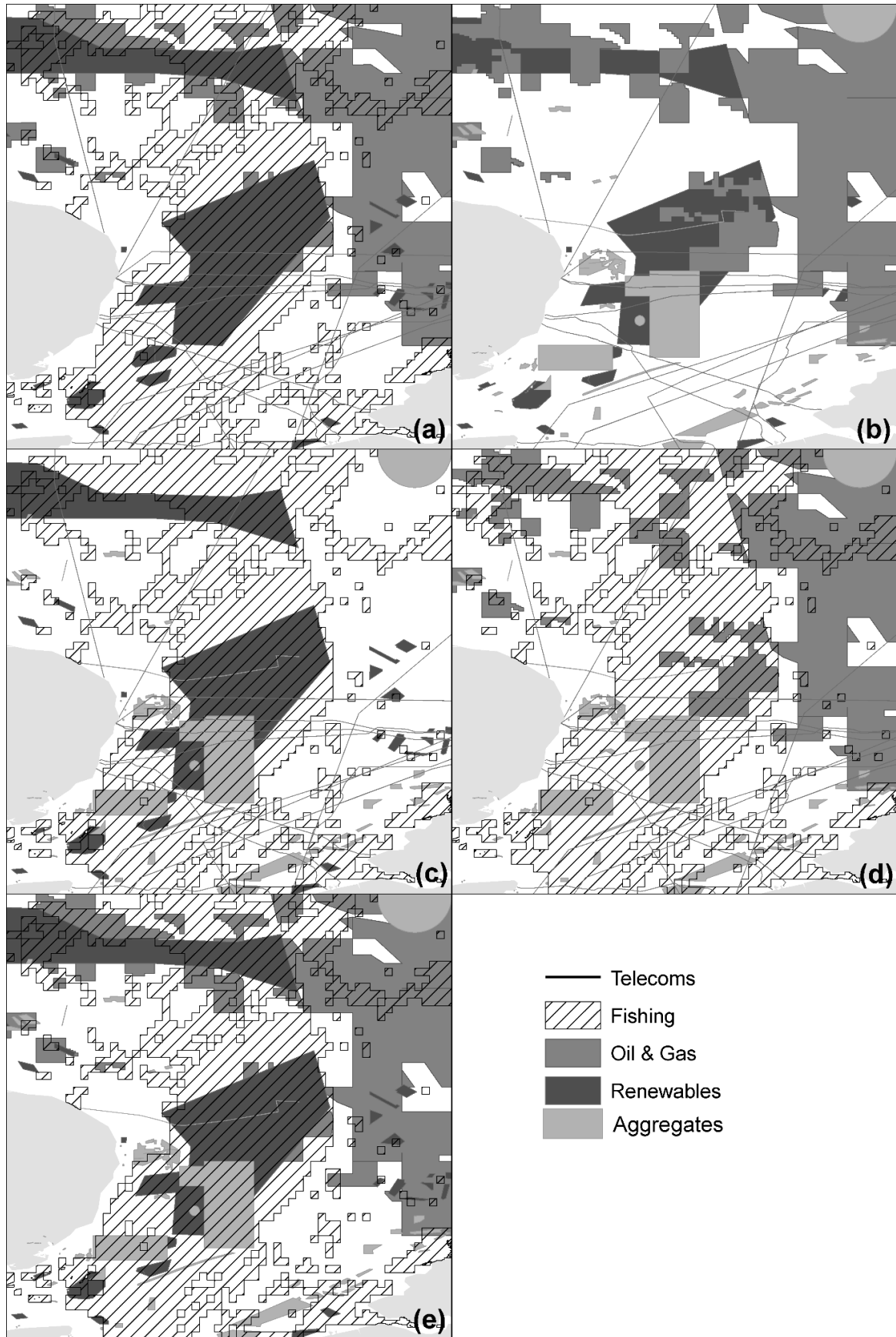


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