

1 **Structural attributes of individual trees for identifying homogeneous patches**

2 **in a tropical rainforest**

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6 **Abstract**

7 Mapping and monitoring tropical rainforests and quantifying their carbon stocks are important, both
8 for devising strategies for their conservation and mitigating the effects of climate change. Airborne
9 Laser Scanning (ALS) has advantages over other remote sensing techniques for describing the three-
10 dimensional structure of forests. This study identifies forest patches using ALS-based structural
11 attributes in a tropical rainforest in Sumatra, Indonesia. A method to group trees with similar
12 attributes into forest patches based on Thiessen polygons and k-medoids clustering is developed,
13 combining the advantages of both raster and individual tree-based methods. The structural
14 composition of the patches could be an indicator of habitat type and quality. The patches could also
15 be a basis for developing allometric models for more accurate estimation of carbon stock than is
16 currently possible with generalised models.

17 **1. Introduction**

18 Tropical forests play a major role in regulating the Earth's climate, being a large sink for carbon
19 dioxide, and storing much of the terrestrial carbon pool (Dixon et al. 1994). An accurate estimation
20 of carbon components within a forest is a first step in the United Nations initiative for Reducing
21 carbon Emissions from Deforestation and forest Degradation (REDD). However, limited knowledge
22 about the quantity and spatial distribution of biomass at the landscape level has led to considerable
23 uncertainties in the estimation of carbon stocks. Human activities such as logging and clearing of
24 forests for agriculture and agro-forestry continue to alter the extent and composition of tropical
25 rainforests. Natural causes such as death of large trees, and subsequent regrowth in the gaps, also
26 contribute to the generation of patches in the landscape. This increases complexity in carbon
27 estimation and causes fragmentation of habitats. Mapping and monitoring these structural changes
28 are pre-requisites for devising strategies for conservation of many endangered species.

29 Airborne Laser Scanning (ALS), an active remote sensing technique based on the technique of Light
30 Detection and Ranging (LiDAR), is now extensively used for describing the three-dimensional
31 structure of forests to understand the habitat requirements of species and to quantify above-ground
32 biomass (AGB), and thereby carbon stocks (Asner and Mascaro 2014). A standard approach to area-

33 based AGB estimation with ALS data uses grid cells, which has limitations given that ALS datasets are
34 generally obtained as point clouds. LiDAR metrics aggregated from the attributes of points within
35 grid cells are highly scale-dependent, and in forests, a grid cell could include part of a large tree, or
36 many small trees, depending on the cell size. Thus, Ferraz et al. (2016) noted that the predictive
37 power of ALS-based AGB models decreased with increasing spatial resolution due to edge effects
38 associated with tree crowns.

39 Patches with different canopy structure and composition can be distinguished in Canopy Height
40 Models (CHMs) derived from ALS data, which could correspond to different habitat types and
41 quality. These could also form the basis for carbon stock estimation which is mid-way between plot-
42 based and individual tree-based approaches, in terms of accuracy, computational time and
43 complexity. The aim of this study is to identify forest patches based on the structural composition of
44 individual trees using ALS data in a tropical rainforest to facilitate estimations of habitat
45 fragmentation and carbon stock. The objectives are: (i) to estimate the locations and attributes of
46 single trees based on a Canopy Height Model; (ii) to group the single trees based on their structural
47 attributes into homogeneous forest patches; and (iii) to analyse the attributes of trees within
48 clusters of similar patches.

49 **2. Study Area and Dataset**

50 The study area (centre: 99.00°E; 1.89°N), with an area of 400 ha, is in Batang Toru in the province of
51 North Sumatra, Indonesia. A history of logging and clearing of land for agro-forestry, selective
52 logging to establish “forest gardens” and natural dynamics have created a mosaic of forest patches.
53 The forests are home to a number of unique plant and animal species (Fredriksson et al. 2014),
54 including the critically endangered Sumatran orang-utans (*Pongo abelii*).

55 ALS data were collected by PT McElhanney (Indonesia) between 23rd March and 4th April, 2015, using
56 a Leica ALS-70 HP LiDAR system from a fixed wing aircraft. The flying height was between 900 m and
57 1350 m above ground level, and the scan half angle was 22.5°. This generated an ALS point cloud
58 with an average density of 23.63 returns m⁻². The returns were classified into ground (0.97%) and
59 non-ground (99.03%) using an algorithm based on adaptive TIN filtering implemented in Terrasolid
60 software (Axelsson 2000; PT McElhanney 2015).

61 **3. Methods**

62 **3.1 Attributes of individual trees**

63 The ground returns, with an average density of 0.23 returns m⁻², were used to generate a Digital
64 Terrain Model (DTM) using FUSION v3.50 (McGaughey 2009). The ground and non-ground returns

65 were merged, and the 95th percentile height of returns above the DTM was used to generate a CHM
66 with a cell size of 1 m; the 95th percentile height was used instead of the maximum to exclude
67 outliers. Individual tree locations, and their heights and crown radii were estimated from the CHM,
68 using the *CanopyMaxima* function in FUSION. This algorithm identifies local maxima using a variable
69 sized filtering window based on canopy height variances (Popescu et al. 2002). The number, mean
70 height and mean canopy radius of all trees within a 25 m radius of each tree were derived, using
71 *Generate Near Table* in ArcGISTM (v10.1), with those summary attributes assigned to each individual
72 tree. A 25 m buffer radius was selected because less than 1% of the trees had a crown radius larger
73 than 12.5 m.

74 The tree location points (X, Y) were converted to Thiessen polygons, with the attributes of the
75 enclosed tree assigned to the polygons. In fitting the Thiessen polygons the area of the polygon was
76 determined by the spacing between adjacent points, with adjacency based on a Triangulated
77 Irregular Network (TIN) generated from the points. The line connecting two points in the TIN was
78 bisected, and these bisectors formed the edges of the Thiessen polygons.

79 **3.2 Delineation of patches and analysis of clusters**

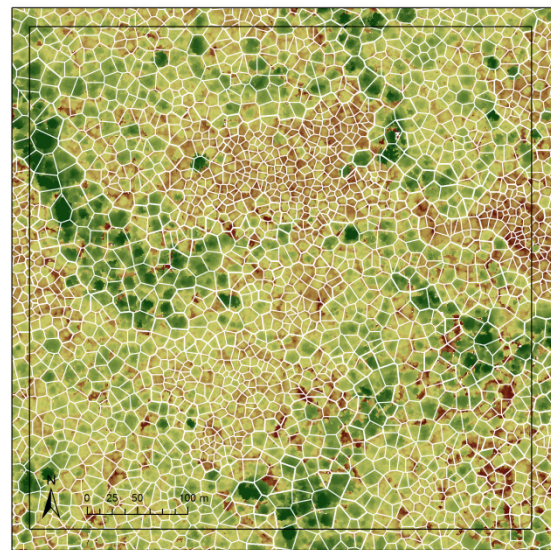
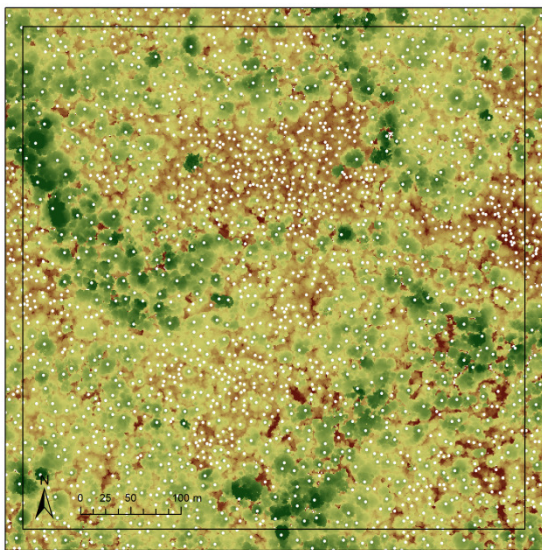
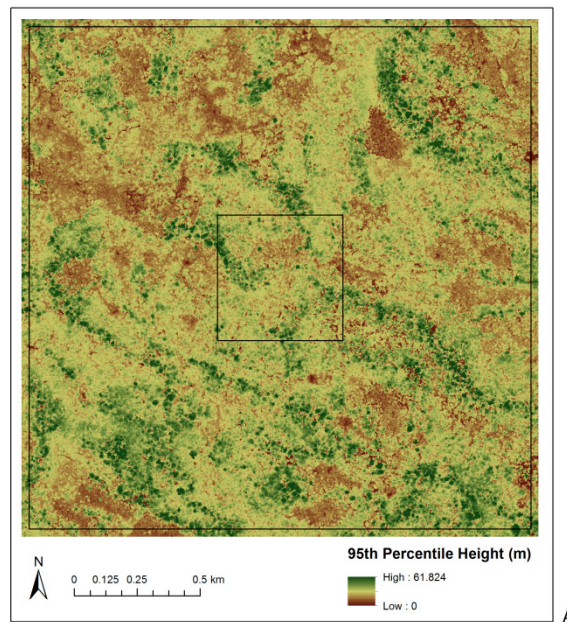
80 The individual Thiessen polygons were clustered into patches using the five attributes (Height and
81 Crown Radius of each tree, and the Count, Mean Height, and Mean Crown Radius of trees within a
82 25 m radius) in a k-medoids algorithm implemented in MATLAB R2015. Silhouette values, a measure
83 of the separability of clusters, were used to determine the number of clusters; the one with the
84 lowest number of negative Silhouette values was selected as the optimum. All adjacent polygons
85 belonging to the same cluster were merged to generate patches in ArcGISTM. All the patches with an
86 area less than 0.25 ha were merged based on the longest shared border. Statistical analyses were
87 performed in MATLAB with α set to 0.001. Crown areas and Thiessen polygons were compared using
88 a Pearson correlation. An ANOVA (one-way analysis of variance) was used to test for differences
89 between clusters, using Scheffe's procedure for post hoc pair-wise comparisons.

90 **4. Results**

91 **4.1 Identification of single trees**

92 The mean height of the CHM (Figure 1) was 20.37 ± 7.31 m. There were 34,484 trees identified with
93 heights ≥ 5 m within the study area, with an overall tree density of 86.21 trees ha^{-1} . The mean tree
94 height was 21.26 ± 6.98 m, and the mean crown radius was 6.39 ± 2.08 m. The mean number of trees
95 within a radius of 25 m for all trees was 22.35 ± 12.37 , and their mean crown radius was 6.35 ± 1.01 m.
96 The mean crown area calculated from the crown radii was 141.77 ± 93.87 m^2 , whereas the mean area

97 of Thiessen polygons was $115.99 \pm 84.13 \text{ m}^2$. The areas of Thiessen polygons correlated only
98 moderately with the crown areas ($r=0.4$; $n=34484$; $p<0.001$).



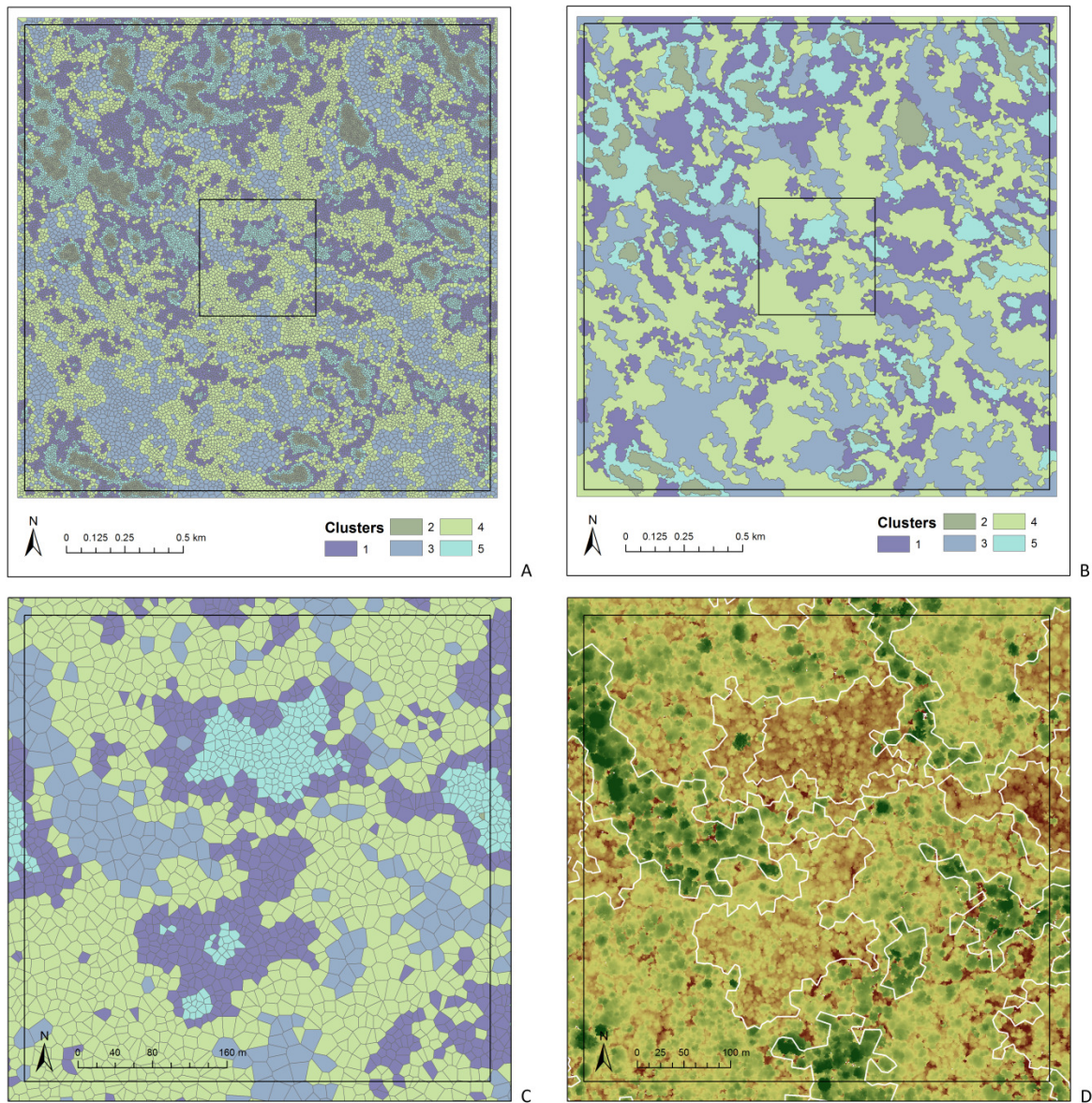
99

100 Figure 1: Canopy Height Model generated from the ALS dataset (A); locations of individual trees in a subset of the study area (B); and the
101 Thiessen polygons generated from the locations of individual trees (C)

102 4.2 Delineation of forest patches

103 The tree clustering process identified an optimum number of five cluster types based on the five
104 input structural variables. The shortest trees (Cluster 2) occupied only 4.58% of the area, while
105 accounting for 13.38% of the tree count, while the tallest trees (Cluster 3) occupied 21.97% of the
106 area, with only 8.86% of the tree count. Cluster 4 (mean tree height: 25.58 m), covered the largest
107 area (37.26%), based on the clustered Thiessen polygons (Table 1). There were 1082 patches when
108 the Thiessen polygons were merged based on clusters, with a mean area of $0.37 \pm 2.94 \text{ ha}$. These

109 were merged into 189 patches with a mean area of 2.11 ± 8.71 ha, by iterative merging of patches
110 with an area less than 0.25 ha (Figure 2).



111
112 Figure 2: Thiessen polygons grouped into clusters based on the attributes of individual trees for the whole study area (A) and for a subset
113 (C); the merged patches after dissolving the patches with area less than 0.25 ha for the whole study area (B), and boundaries of patches
114 overlaid on the Canopy Height Model for a subset (D)

115 4.3 Analysis of clusters

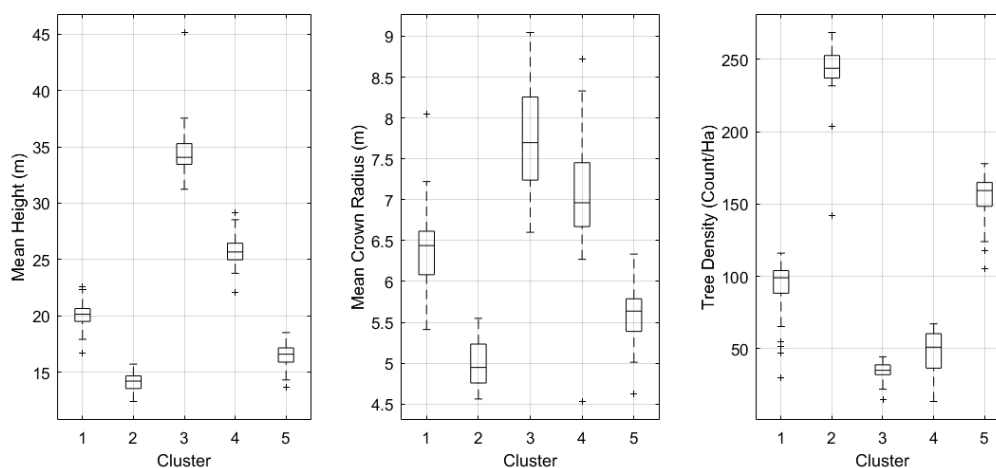
116 The mean height, mean crown radius and density of trees in each patch (Table 1; Figure 3) were
117 significantly different between the clusters (all $p < 0.001$; $F_{4,183} = 1032.41$; $F_{4,183} = 132.9$; $F_{4,183} = 679.3$
118 respectively). When the clusters were compared pairwise, all differences were significant except for
119 the crown radii for clusters 2 and 5 ($p = 0.002$), and the density of trees for clusters 3 and 4 ($p = 0.042$).

120

121 Table 1: Attributes of patches within the five clusters before (above) and after (below) merging

Cluster ID	1	2	3	4	5
Number of Patches	414	62	181	268	157
	62	20	39	33	35
Total Number of trees	9832	4615	3054	9219	7764
	9573	4256	3057	9693	7905
Total Number of trees (%)	28.51	13.38	8.86	26.73	22.51
	27.76	12.34	8.86	28.11	22.92
Total Surface Area (%)	24.20	4.58	21.97	37.26	11.99
	23.31	4.28	21.52	38.55	12.34
Overall Density (Trees ha ⁻¹)	101.57	251.99	34.75	61.85	161.90
	102.68	248.60	35.51	62.87	160.13
Mean Height of Trees (m)	19.89±3.64	13.93±2.57	35.68±5.05	25.58±3.55	16.54±3.18
	19.99±3.90	13.99±2.61	34.49±6.46	25.32±4.37	16.60±3.29
Mean Crown Radius of Trees (m)	6.33±1.87	4.91±1.40	7.82±2.37	7.33±2.04	5.65±1.69
	6.39±1.89	4.93±1.42	7.62±2.37	7.24±2.08	5.65±1.70

122



123

124 Figure 3: Box Plots of mean height, mean crown radius and overall density of trees in the five clusters (after merging)

125 5. Discussion and Conclusion

126 Identification of homogeneous patches in tropical forests based on tree heights, crown radii and
 127 density could have relevance for estimating habitat fragmentation and biomass. The method
 128 developed in this study, based on Thiessen polygons and k-medoids clustering, groups trees of
 129 similar structural attributes combining the advantages of raster and individual tree-based methods.
 130 The structural composition of the patches could be an indicator of habitat type and quality for
 131 species which are increasingly under threat from anthropogenic and natural disturbances. Distances
 132 between suitable habitats, in the case of fragmentation, could potentially be more accurately
 133 estimated using these tree crown-following tessellations rather than grid cells, especially if they are
 134 at a low resolution.

135 Natural and anthropogenic factors have contributed to the generation of a mosaic of forest patches
 136 in the study area, which were clearly visible in the CHM. The tallest trees with the largest crown radii
 137 (Cluster 3) occupied a large percentage of the area but had relatively low tree density. Mapping the

138 extent of these tall patches is important, even if the accuracy of estimated tree density is low, since
139 the large trees account for most of the biomass in tropical forests. They also serve as a focal point
140 for biological activity and create large gaps at death, altering the forest structure dynamics in
141 addition to releasing the sequestered carbon (Chambers et al. 2007; Ferraz et al. 2016). Note that
142 the average estimated tree density in the study area (86.21 trees ha⁻¹) could be lower than the
143 actual density since the algorithm identifies only the dominant and co-dominant trees as
144 represented in the CHM (McGaughey 2009).

145 The areas of Thiessen polygons did not have a high correlation with crown areas since the polygons
146 were constructed based only on the distances between adjacent points. The associated Thiessen
147 polygon would be larger than the estimated crown area if the distance from a tree to adjacent trees
148 is greater than its crown radius. However, this would not pose a problem in the delineation of
149 patches, and the estimation of patch areas could in fact, be more accurate than one based on crown
150 areas if tree crowns overlap. The distance for locating neighbouring trees could be modified based
151 on the study area, especially if there are a high proportion of very large or very small trees.

152 Although generalised allometric models have been developed for tropical forests (Chave et al. 2005),
153 species-specific and site-specific models are considered to be more accurate. Even within a single
154 species, the biomass may vary depending on environmental factors such as rainfall, soil or
155 topography (Litton and Kauffman 2008). The species diversity in tropical forests is extremely high,
156 and AGB estimations based on individual tree species may be difficult with the current knowledge of
157 taxonomy and tree species distribution. Allometric models based on identified clusters within a
158 landscape could be developed based on systematic field surveys, rather than random sampling of
159 stands or limited species-specific models, leading to more accurate estimation of carbon stock in
160 tropical forests.

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168 **References**

169 Asner, G.P., & Mascaro, J. (2014). Mapping tropical forest carbon: Calibrating plot estimates to a simple LiDAR
170 metric. *Remote Sensing of Environment*, 140, 614-624

171 Axelsson, P. (2000). DEM generation from laser scanner data using adaptive tin models. *International Archives*
172 *of Photogrammetry and Remote Sensing, XXXIII, Part B4*, 110-117

173 Chambers, J.Q., Asner, G.P., Morton, D.C., Anderson, L.O., Saatchi, S.S., Espírito-Santo, F.D.B., Palace, M., &
174 Souza Jr, C. (2007). Regional ecosystem structure and function: ecological insights from remote sensing of
175 tropical forests. *Trends in Ecology & Evolution, 22*, 414-423

176 Chave, J., Andalo, C., Brown, S., Cairns, M.A., Chambers, J.Q., Eamus, D., Fölster, H., Fromard, F., Higuchi, N.,
177 Kira, T., Lescure, J.-P., Nelson, B.W., Ogawa, H., Puig, H., Riéra, B., & Yamakura, T. (2005). Tree allometry and
178 improved estimation of carbon stocks and balance in tropical forests. *Oecologia, 145*, 87-99

179 Dixon, R.K., Solomon, A.M., Brown, S., Houghton, R.A., Trexler, M.C., & Wisniewski, J. (1994). Carbon pools and
180 flux of global forest ecosystems. *Science, 263*, 185-190

181 Ferraz, A., Saatchi, S., Mallet, C., & Meyer, V. (2016). Lidar detection of individual tree size in tropical forests.
182 *Remote Sensing of Environment, 183*, 318-333

183 Fredriksson, G., Nowak, M., & Usher, G. (2014). Biodiversity Monitoring Stage 1: Camera trapping and Wildlife
184 Surveys. In PanEco / Yayasan Ekosistem Lestari (Ed.), *Biodiversity Monitoring for Sarulla Operations Ltd.*
185 Indonesia

186 Litton, C.M., & Kauffman, J.B. (2008). Allometric Models for Predicting Aboveground Biomass in Two
187 Widespread Woody Plants in Hawaii. *Biotropica, 40*, 313-320

188 McGaughey, R.J. (2009). FUSION/LDV: Software for LIDAR data analysis and visualization. *US Department of*
189 *Agriculture, Forest Service, Pacific Northwest Research Station: Seattle, WA, USA, 123*

190 Popescu, S.C., Wynne, R.H., & Nelson, R.F. (2002). Estimating plot-level tree heights with lidar: local filtering
191 with a canopy-height based variable window size. *Computers and Electronics in Agriculture, 37*, 71-95

192 PT McElhanney (2015). LiDAR & DAP Survey. In PT McElhanney (Ed.), *Project Report for BUT Sarulla Operations*
193 *Ltd.* Indonesia