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Optimal allocation of effluent charges in regional water resource systems.

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LA THÈSE A ÉTÉ
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OPTIMAL ALLOCATION OF EFFLUENT CHARGES IN REGIONAL
WATER RESOURCE SYSTEMS

A THESIS

SUBMITTED TO THE FACULTY OF GRADUATE STUDIES
THROUGH THE DEPARTMENT OF INDUSTRIAL ENGINEERING
IN PARTIAL FULFILMENT OF THE REQUIREMENTS
FOR THE DEGREE OF MASTER OF APPLIED
SCIENCE AT THE UNIVERSITY OF WINDSOR

BY

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1979

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ABSTRACT

Environmental resources play a major role in economic systems through the service they provide to assimilate or carry away the residuals of production processes. As water resources have become scarce to offer their service to the growing number of population, urbanization, and economic activities, the optimal allocation of their assimilative capacities among the users can be achieved through the forces of the market system. The effluent charges strategy attains this objective. The strategy, in effect, imposes a price on the use of the water resource and thus provides an incentive to the polluters to seek the most efficient waste treatment methods.

A mathematical model is developed to determine, within the framework of a regional water quality management system, the optimal effluent charge rates to be levied against all polluters in a river basin. In addition, the model determines the optimal treatment policy for each polluter such that the required water quality standards in the basin are satisfied at a minimum total cost.

The formulation of the model reflects the presence of a large number of factors which account for the variations in stream characteristics as well as in plant performance at each pollution source. The factors interact with each other through complex, nonlinear transfer functions. Consequently, the objective function and constraints are highly nonlinear.

in nature, which calls for a nonlinear optimization method to find the solution. The Sequential Unconstraint Minimization Technique (SUMT) has been used for this purpose.

Two main features distinguish the present model from previous works. First, the attainment of water quality standards throughout the basin is ensured by explicitly specifying the appropriate constraints in the model. Second, the model determines the effluent charges as a set of nonlinear rates as opposed to a set of uniform rates obtained by the existing models.

The model is applied to two different cases, and the results are interpreted in the context of the complex interactions among all the factors involved. The results point out the importance of developing complex mathematical models in order to deal with complicated problems of environmental pollution.

I wish to dedicate this work to
my parents

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LIST OF NOTATIONS

- BOD - Biochemical oxygen demand
- C - Concentration level of dissolved oxygen, mg/l
- CS - Saturation concentration level of DO, mg/l
- CU - Upstream concentration level of DO, mg/l
- DO - Dissolved oxygen
- EP - Phosphorus concentration level, mg/l
- F - Total value of the objective function
- FC - Charge cost function
- FT - Waste treatment cost function
- Fl - A conversion factor; $\frac{\text{day} \cdot \text{ft}^3 \cdot \text{mg}}{\text{sec} \cdot \text{lb} \cdot \text{l}}$
- G - Constraint function
- K_1 - Biological oxidation rate coefficient; day^{-1}
- K_2 - Rearation rate coefficient; day^{-1}
- K_3 - Sedimentation or absorption rate coefficient of BOD; day^{-1}
- L - Concentration level of biochemical oxygen demand, mg/l
- La - Local BOD runoff rate; mg/l/day
- LB - Maximum BOD content at the beginning of the zone, mg/l
- LE - Stream BOD content at the end of the zone, mg/l
- LU - Upstream concentration of BOD, mg/l
- MAXBOD - Maximum allowable BOD level in the river, mg/l
- MAXPHR - Maximum allowable phosphorus in the river, mg/l
- MDO - Minimum DO concentration level in the river, mg/l
- MGD - Million gallons per day
- MINDO - Minimum allowable DO level in the river, mg/l

- P - Modified objective function in SUMT formulation
- PH - Maximum phosphorus content in the river, mg/l
- <P-R> - Rate of photosynthesis minus the rate of respiration, mg/l/day
- PU - Upstream phosphorus concentration, mg/l
- Q - Stream flow rate; cfs
- \bar{Q} - Average daily flow of the plant, MGD
- QU - Upstream flow rate, cfs
- r - Constant value in SUMT formula
- T - Stream temperature, °K
- WB - Amount of organic waste produced at pollution source, lb/day
- WP - Amount of phosphorus waste produced at pollution source, lb/day
- X - BOD treatment level applied in the theoretical functions
- Y - Effluent charge rate, \$/lb of BOD
- Z - Phosphorus treatment level applied to the empirical functions
- θ_1, θ_2 - Parameter coefficient for effluent charge rate
- τ - Residence time, in days from source of pollution

CHAPTER I
INTRODUCTION

1.1 Economic Aspects of Environmental Pollution

The quality of water, air, and soil, is affected by the economic activities of human beings to the extent that most researchers consider the pollution problem as a phenomenon related to production and consumption processes, (Dales, 1968). This relationship can be clearly shown by tracing the pollution to its primary sources.

Men, through their daily activities, take from the environment a wide variety of materials, agricultural products, and animal materials. They transform them into many varieties of economic goods. During the transformation processes, many unwanted materials - solid, liquid or gaseous - are generated and discarded into the environment as residuals. Some of these residuals are harmful (toxic chemicals, for example) and could damage the environment; some others are harmless, but they could react chemically with other substances or with each other in the air or water in a damaging way (as in the sunlight). On the other hand, the final economic goods are eventually consumed, a process during which they undergo physical or chemical transformation and become unwanted materials. Sooner or later they will be discarded again.

If the environmental capacity to absorb or to carry away all the residuals was unlimited, there would be no pollution problem and residuals could be discharged into

water, air, or soil without limit. In earlier days, a smaller, more dispersed population, and a much lower level of economic activity did not generally impose a real load on the capacity of the environment to assimilate residuals. Environmental resources were in such abundance that they seemed to have an unlimited capacity to assimilate the wastes. Most economists at that time were considering these common resources as perfect examples of free goods. But, as population, urbanization, and economic activity have grown, what were once plentiful environmental resources have become increasingly scarce, and can no longer be fully available to the increasing number of users.

It is clear that the amount of discarded waste depends on two factors: number of people in the economy (population density), and per capita consumptions of materials. Environmental degradation, therefore, will increase with increasing population and economic growth. Assuming the society has no intention to halt or decrease the economy's growth, it has only one alternative left to control pollution, that is to seek new technical methods of pollution abatement to decrease the rate of waste residuals per throughput of raw materials.

On the other hand, eliminating the discarded wastes, which are unwanted residuals with no value, will cost the society at large. The dischargers, however, are under no obligation to incur this cost or even economize in their use of natural resources as waste receptors. This has created a situation in which the costs of production and the prices

of goods and services diverge, in different degrees, from the true costs of production. Such divergence not only leads to a misallocation of the employed resources, but can also result in a deteriorating environmental quality (Ralph, 1963).

Two points emerge from this analysis: First, the environmental resources play an important role in the economic system through the service they offer to the production processes to carry away or assimilate residuals. The concept of the economic system should be broadened to include these resources as a counterpart to such other factors in the production processes as capital, labour, and materials, (Ingemar et al, 1977). Figure 1.1 diagrammatically depicts this concept. Second, any real effort to control pollution would have to provide a set of strong economic incentives to persuade dischargers to seek these new, technical methods of pollution abatement.

In order to alleviate this situation, environmental quality management should be considered as a resource allocation problem; that is, environmental resources, like all other scarce resources, should be optimally allocated among the different competing users so as to maintain a desired level of environmental quality. Although the free market system provides the mechanism for optimal allocation of resources, in the case of environmental resources it could fail. The failure may be due to either the lack of property rights, or the public good nature of these resources.

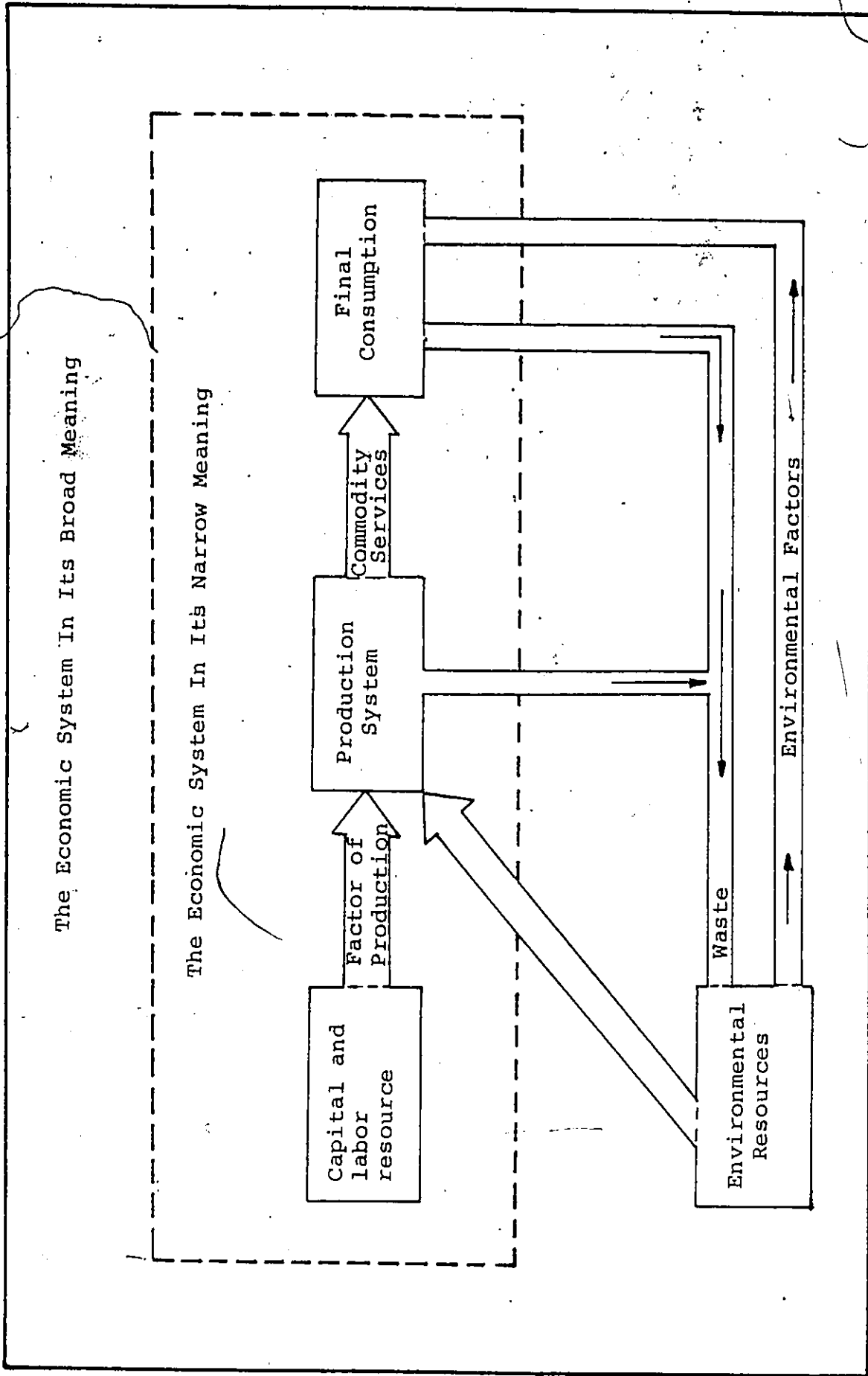


Figure 1.1 Economic system including environmental resources as a factor in the production processes (Ingemar et al, 1977).

To overcome the market failure for the environmental resources, some form of governmental intervention is suggested for maintaining the desired environmental quality levels, and achieving the optimal allocation of resources among different users (Kneese, 1964). The government agency may use the strategy it deems best to fulfill its responsibility. Such methods as the regulation enforcement and subsidies have been widely used in the past one-and-a-half decades. The regulation enforcement strategy imposes discharge levels on firms and municipalities, whereas the subsidies method provides monetary grants for constructing waste treatment facilities. In spite of these efforts, both strategies have failed to bring about the desired levels of environmental quality (Alexander, 1976). The reason lies in the fact that both strategies have two characteristics which render them costly and ineffective. First they rely greatly on detailed information and procedures which are difficult to obtain and apply. Second, they lack economic incentives to realize the social cost of degrading the environment.

The effluent charge strategy is an alternative control policy, and it has the potentiality to overcome the drawbacks inherent in the other two methods. The strategy requires that those who use the environmental resources pay for them at a price that reflects their scarcity. Pricing the environmental resource tends to establish, among the different users, an optimal allocation pattern of these resources. The strategy is simple and applicable; besides,

it places greater reliance on the self-motivation of dischargers to reduce their waste.

1.2 The Purpose and Plan of Study

The primary purpose of this investigation is to develop a mathematical model of water pollution control, which would determine the optimal effluent charge rates to be levied against all the polluters in a river basin system, such that the desired water quality levels in the basin are maintained at a minimum total cost.

In Chapter I, the pollution problem has been embedded in an economic framework; since the problem has arisen in the context of economic activities, its solution may be found in economic theory.

Chapter II presents the technical aspects of the water pollution problem. The competing uses of water resources, are discussed and the sources of pollution and pollutant types are presented. The measures of water quality used in the study are discussed next. Finally, the technical options for water pollution control are exhibited, emphasizing conventional waste treatment methods.

Chapter III is devoted to an evaluation of the effluent charges strategy as an alternative control method. The role played by the price mechanism, through the market function, to solve the problem of resource allocation is discussed. The reasons for market failure in the case of the environmental resources, and the use of effluent charges to overcome this failure are explained. A comparison

is made between the effluent charge strategy and the other two strategies with respect to three important criteria: effectiveness, flexibility, and efficiency. Finally, the criticisms of, and the misunderstanding about, the strategy are evaluated.

In Chapter IV, the historical evolution of the effluent charge strategy is presented in two parts. In the first part, works carried out to date on the development of theoretical bases for the strategy are discussed. The second part offers a survey of the studies dealing with applications of the strategy within framework of water quality management models.

In Chapter V, the mathematical model of the river basin system is developed. The river is divided into a number of zones, each receiving wastes from municipal and/or industrial sources located along the river. Water quality measures are defined and a modified form of the Streeter-Phelps equations, as suggested by Dobbins (1964) are employed to predict the behaviour of these measures at various points of a zone. The model's objective function consists of the treatment costs as well as the effluent charges levied against the untreated portion of the wastes. The constraints of the model define the allowable levels of water quality measures in each zone. The model is highly nonlinear in nature, calling for a nonlinear optimization method to find the solution.

Chapter VI considers the application of the optimi-

zation model in two different cases. The first assumes that the river receives waste from municipal sources. The cost function employed in this example are derived from actual data collected, during the period 1975-1977, from municipal wastewater treatment plants in Southern Ontario. The second example assumes that the river receives waste from different municipal and industrial sources. The cost functions employed are based on the current information on the cost of industrial and municipal waste treatment.

Chapter VII is devoted to discuss the results of the optimization model as it is applied in the two cases. The discussion is followed by drawing a number of conclusions for the present investigation. Proposals are made, at the end of the report, for further research and studies.

CHAPTER II

TECHNICAL ASPECTS OF THE WATER POLLUTION PROBLEM

2.1 Introduction

For the water to be useable, it should possess certain quality levels. Over the years, highly technological treatment methods have been developed to improve the quality of water in rivers and streams. Yet, there are some questions regarding the effectiveness of these methods to cope with increasing problem of water resource pollution.

The chapter discusses, first, the various uses of water resources; second, it identifies the different categories of waste discharges, and explains the measures of water quality. Some technological options for reducing the amount of waste discharge without a decrease in economic growth are presented. Finally, the conventional waste treatment methods - primary, secondary and tertiary - are briefly discussed.

2.2 Importance of Water Resources

The importance of water resources (e.g., streams, rivers, and lakes) can be assessed through the values of their three main services:

- 1) Value in Use: Water is withdrawn from the stream to satisfy domestic, industrial, and agricultural needs.

- 2) Value in Commerce and Recreation: The stream

serves as a means of transportation, a source of power, or a place to fish, boat or swim; it contributes to the environment, which makes a community a pleasant place to live in.

3) Value as a Waste Carrier: Stream flow may be used to carry waste materials away from their sources.

The first two functions compete with the third. If the amount of waste discharged into the stream is greater than the capacity of the stream to assimilate it, pollution problem will arise, and the value of the water course for its other two uses will reduce (Lee, 1970).

The threat to water quality arises mainly from discharging wastes from domestic, industrial, agricultural, and recreational activities into the water. These wastes can be grouped in various ways. One common classification differentiates between non-degradable and degradable wastes:

i) Non-Degradable Wastes: The sources of non-degradable substances are mainly inorganic chemicals such as chlorides, synthetic organic chemicals and inorganic suspended solids. Some industrial wastes contain inorganic or metallic salts and synthetic organic chemicals. Domestic water use results in a small increase in the content of chlorides and other dissolved salts. The return flow from irrigation is generally significantly higher in dissolved salts than the inflow water. Also, some other discharges from mines may contain copper, zinc, and uranium. Many of these non-degradable wastes may be toxic or corrosive. Some of them

may impart color or taste to the water. Suspended materials and colloidal matter cause turbidity in surface waters. This not only makes water less attractive, but can also damage fish life, and inhibit the growth of some aquatic plants such as algae which may or may not be desirable. Non-degradable wastes are usually diluted and may be changed in form, but they are not appreciably reduced in weight in the receiving water (Kneese, 1964).

ii) Degradable Wastes: Degradable wastes include thermal wastes and organic wastes which come mainly from domestic, industrial and agricultural operations. These types of waste are subsequently decomposed into harmless substances by the stream's natural biological, chemical, or physical processes, which result in a significant weight reduction of the waste load. The concentration of a degradable waste, however, is much more difficult to be determined than a non-degradable waste because the capacity of the stream to decompose (or to assimilate) the waste is determined by many factors such as stream flow, stream temperature, and the physical and biological properties of the stream that affect the rates of settling and reaeration (Lee, 1970). Thermal loads are classified as degradable waste since the heat is dissipated in receiving waters, primarily by evaporation and conduction in ground waters. The primary sources of thermal wastes are the generation of electrical energy and the cooling operations in many industries (Fan, et al, 1973).

2.3 Water Quality Measures

A measure of organic waste load is Biochemical Oxygen Demand (BOD), which is a measure of the amount of oxygen needed to oxidize a unit of organic waste material into relatively harmless components. The amount of oxygen demanded and the rate at which it is drawn upon are mainly functions of the quality of the waste, its chemical characteristics, as well as the temperature of the receiving water. For example, toxic substances may appreciably reduce the rate of decomposition by inhibiting bacterial action. At higher temperatures, where the oxygen saturation level of water is relatively low, bacterial action is accelerated, and dissolved oxygen in the water is drawn upon more rapidly (Davidson and Bradshaw, 1967). The imbalance between available oxygen and oxygen demand may proceed to the point of depleting the oxygen level in the stream completely (anaerobic conditions). Such conditions are most likely to occur in the summer, when streamflows tend to be low and temperatures high. All the efforts to control water quality have to avoid such conditions. In fact, the rate at which "BOD" is depleted (which depends on the amount of waste at the point of discharge) combined with the rate at which oxygen is restored, determines the level of Dissolved Oxygen (DO) in the stream.

The combined effect of organic waste discharged at a specific location and reaeration in the stream results first in a decrease and then in an increase in DO as the

waste is carried or moved downstream. This phenomenon is illustrated by a characteristic curve known as the "oxygen sag" curve. In Figure 2.1, the variations in BOD concentration and the associated oxygen sag curves for two different levels of waste load discharge are presented. In Figure 2.1(a) the waste load is below the river's assimilative capacity, and thus a minimum level of "DO" is maintained throughout the stream. However, if the receiving waters are loaded excessively such as in Figure 2.1(b), the process of degradation will proceed anaerobically by the action of bacteria not utilizing free oxygen. Anaerobic processes produce hydrogen sulfide and other gases which have an annoying odor. These, combined with floating materials and sewage solids, are likely to be unpleasant from aesthetic and recreational view points. The bacteria content also makes the stream unsafe for water sports. However, the most direct impact of an excessive organic waste load in the stream is its effect of reducing the amount of "DO" to the point where most species of fish are unable to survive (Kneese and Bower, 1968).

Among the factors contributing to the degradation of water quality are plant nutrients, primarily phosphorus. Although nutrients themselves are not harmful to the water (except when they are present in excessive amounts) they can lead to intense algae growth, or "eutrophication", which, in general, is detrimental to aquatic life (Bishop, 1975). This is of particular importance in quiet waters

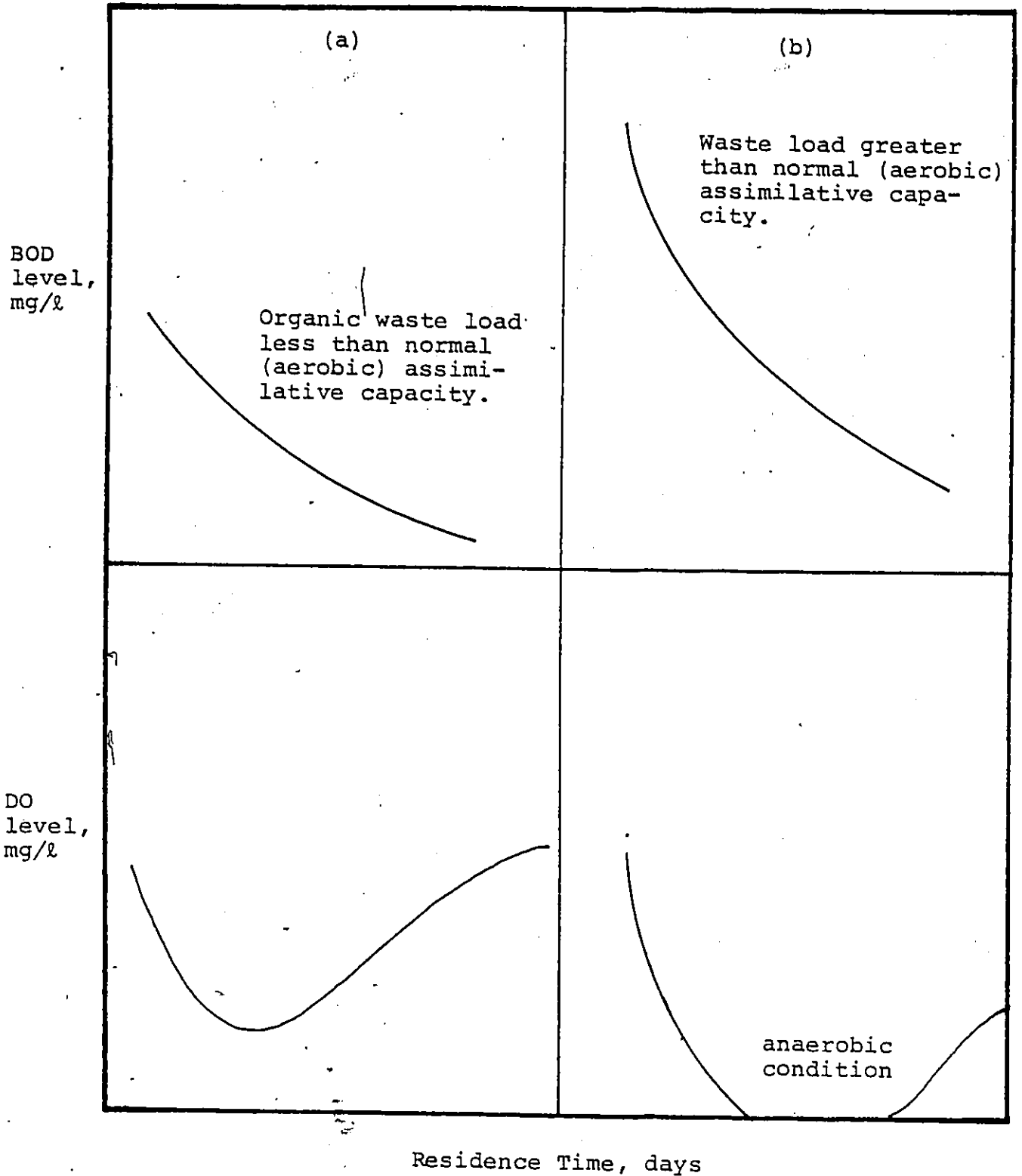


Figure 2.1. Schematic diagram of oxygen sag for two different levels of organic waste load, [Kneese, 1964].

such as lakes, bays, and estuaries. Since most lakes are considered as sinks for rivers, the maximum level of phosphorus substances in rivers should be controlled to prevent eutrophication processes, not only in rivers, but also in lakes.

2.4 Technological Options for Water Pollution Control

The quality of water resources can be improved in two ways (separately, or in combination): first, by making better use of the stream's assimilative capacity, and second, by reducing waste load discharges (Kneese and Bower, 1968).

As far as the first option is concerned, assimilative capacity can be enhanced in various ways. For instance, since the effect of waste discharge depends on the time and/or place of discharge, any change in one or both factors could affect pollution concentration. Some locations along the stream are better suited for carrying away and assimilating the waste than others. Alternatively, discharges can be timed to coincide with periods of maximum stream flow rate, and to avoid periods of high temperature. The assimilative capacity can also be increased by means of artificial reaeration, or by constructing dams to store and release water to dilute wastes during periods of low flow (low flow augmentation).

As for the second option, the reduction of waste load discharges can be accomplished in two broad ways:

(i) by reducing the generation of wastes per unit output,

(ii) by modifying the residual waste after generation.

Within these two broad categories a variety of techniques are available. In the first category, waste generation per unit output could be reduced by changing the types of raw material inputs or by modifying production processes in such a way as to increase the technical efficiency of materials used, which eventually results in less waste. Another possibility is to shift the economic activities toward more durable products. In the second category, the waste reduction could be accomplished either by material recovery or by waste treatment. Material recovery is now technologically possible for most types of material, such as paper, metal, and glass.

Water pollution can also be reduced by various techniques of waste treatment. There are a number of advanced waste treatment methods, capable of significantly reducing the harmful effects of waste discharge. However, the costs entailed in achieving such treatment levels are usually high, compared to other methods, which in many cases embody also a more effective means of waste reduction (Kneese & Schultze, 1975). Nevertheless, most policies have rested almost exclusively on the waste treatment approach to water pollution control. This may be due to the fact that waste treatment is the only general method for reducing waste discharges from different sources, e.g., households, industries, and mines. On the other hand, the absence of an

adequate incentive to motivate the firms to search for alternative methods of waste reduction (such as recycling or material recovery) may also have contributed to this exclusive dependence on waste treatment techniques. In the next section, a brief discussion regarding the conventional treatment methods of domestic and industrial wastes and their economies will be given.

2.5 Conventional Methods of Waste Treatment

Conventional treatment of domestic and industrial wastes is a comparatively standardized process that can greatly reduce, but not eliminate, BOD level. The process involves the following steps:

i) Primary Treatment: In this stage the large solids are removed by screening and sedimentation processes. This step usually produces a wet, difficult-to-handle sludge which is usually digested in heated anaerobic tanks before final disposal. When waste treatment plants are operated adequately, the primary treatment of an influent containing organic wastes can reduce the first stage BOD by 35% to 45% (Kneese, 1964).

ii) Secondary Treatment: This stage is biological in character and essentially controls and accelerates the oxidation processes. Two major techniques are presently used, with various modifications. One is the trickling filter in which a biological film is grown on rocks or some type of plastic medium, and waste is applied intermittently, allowing air contact with the surface of the film. The

other technique is the activated sludge process, in which air or oxygen is forced into a tank containing a mixture of waste and actively feeding biota. Part of the settled sludge containing the biota is recirculated and mixed with the entering waste in the aeration chamber.

Primary and secondary treatments combined usually can reduce BOD by between 80% and 95% (Kneese, 1964).

iii) Tertiary Treatment: Secondary treatment plant effluents can create some undesirable conditions in receiving streams through adverse effects on the dissolved oxygen levels, which result in unwanted changes in the living organisms of the stream and the possible excess growth of algae and other aquatic plants. To arrest these conditions, an advanced type of "tertiary" treatment is being developed as a third stage to follow conventional secondary treatment processes for further reduction of BOD as well as nutrients (such as phosphorus substances) which algae utilize for growth. This third stage can usually reduce the BOD level up to 99%; however, there is no typical tertiary treatment form. This third stage must be adapted to the particular needs of the receiving water concerned. Figure 2.2 illustrates the three conventional treatment stages.

Treatment costs in any plant increase rapidly as the removal of BOD is pushed upward, (Fig. 2.3). This has been confirmed by studies of many industries and different kind of pollutants. Kneese and Schultz (1975) demonstrated the relationship between the percentage of pollutants already

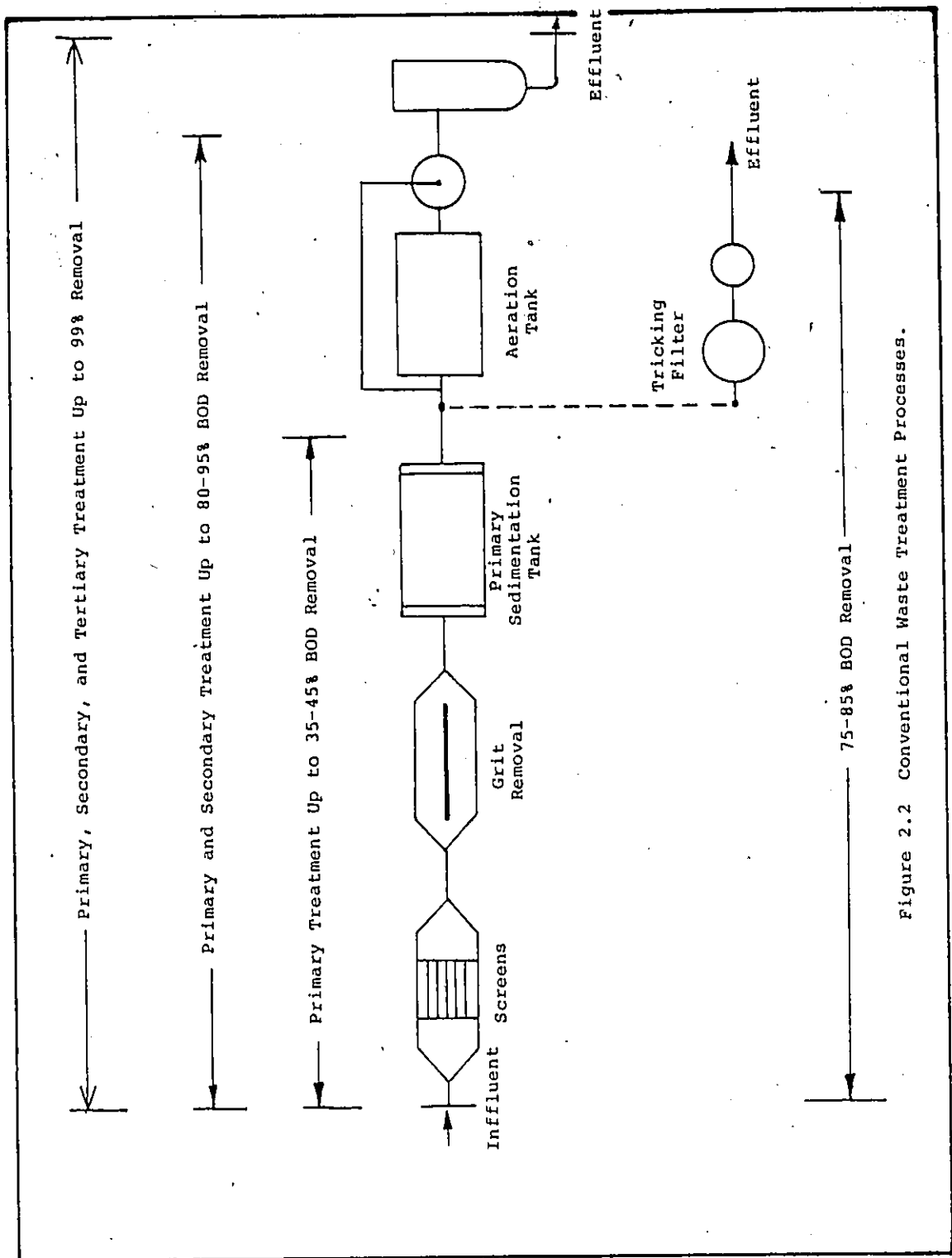


Figure 2.2 Conventional Waste Treatment Processes.

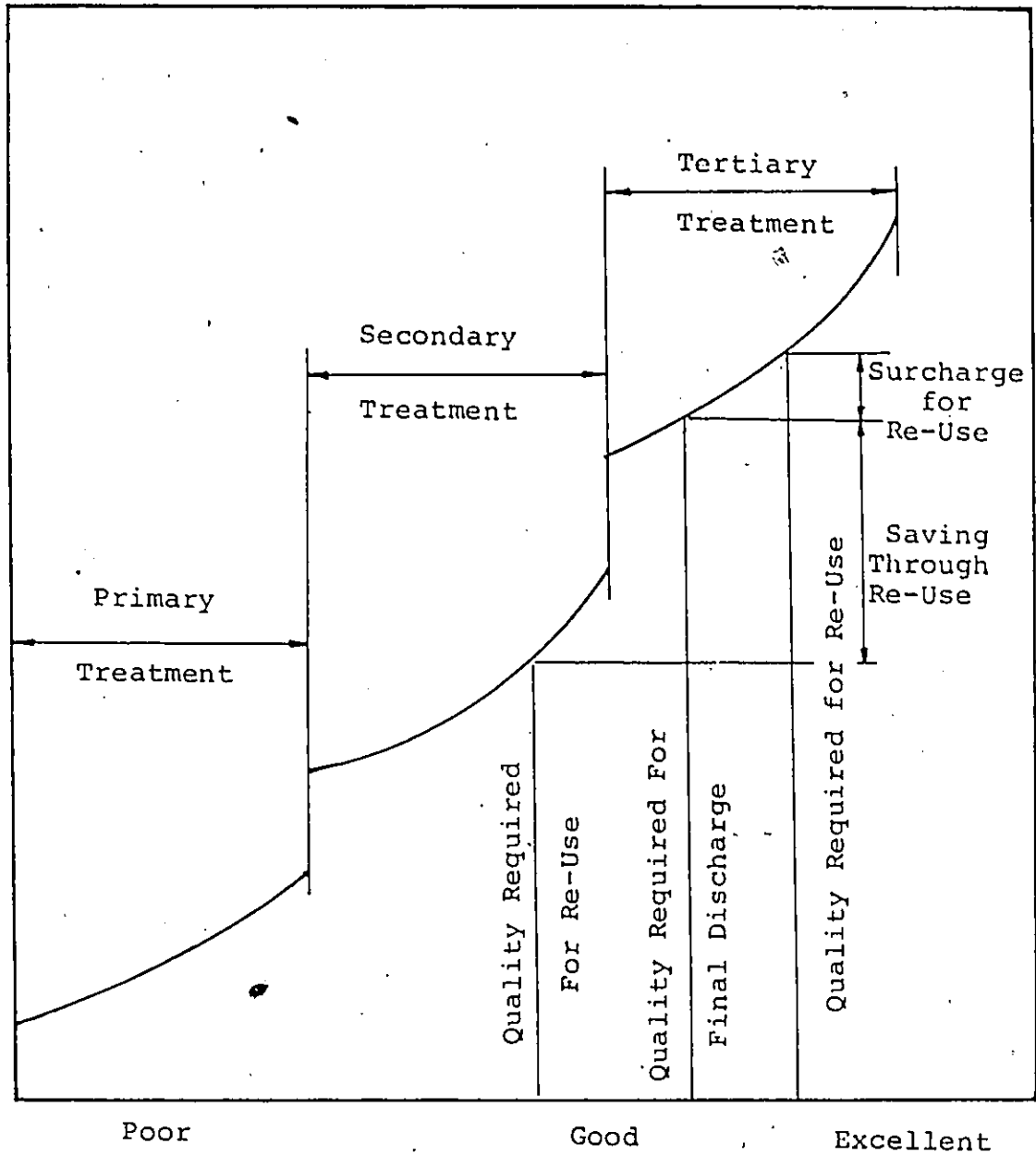


Figure 2.3 Quality of Effluent from Wastewater Treatment Plant (Eckenfelder and Barnard (1971)).

removed and the cost of removing an additional unit. The examples included beet sugar industry, meat-processing plants, and the petroleum refining industry, see Figure 2.4. These examples also illustrate another aspect of the pollution control costs, namely, that they differ widely from industry to industry; the incremental cost of removing pollutants above the 90 percent level is \$0.05 per pound of BOD in beet sugar industry, \$0.60 per pound for meat processing plants, and \$0.22 per pound for petroleum refining industry.

On the other hand, within the same industry, individual firms may incur quite different treatment costs, depending on the magnitude of the waste flow they are handling. The studies given by Shah (1970), and Michel (1970) indicate that as the amount of waste flow increases, the per unit costs of capital invested plus operation and maintenance decreases, see Figure 2.4. This result is valid for both municipal and industrial waste treatment plants.

The differential between the pollution removal costs of municipal and industrial plants underlines a significant feature that a flexible pollution control policy should possess. That is, these policies must recognize and exploit these cost differentials so as to encourage the largest reduction in pollution from the firms with the smallest removal costs. This tends to minimize the total cost of pollution control by a greater amount than if uniform regulations of reduction were imposed on discharger, Kneese and Schultze (1975).

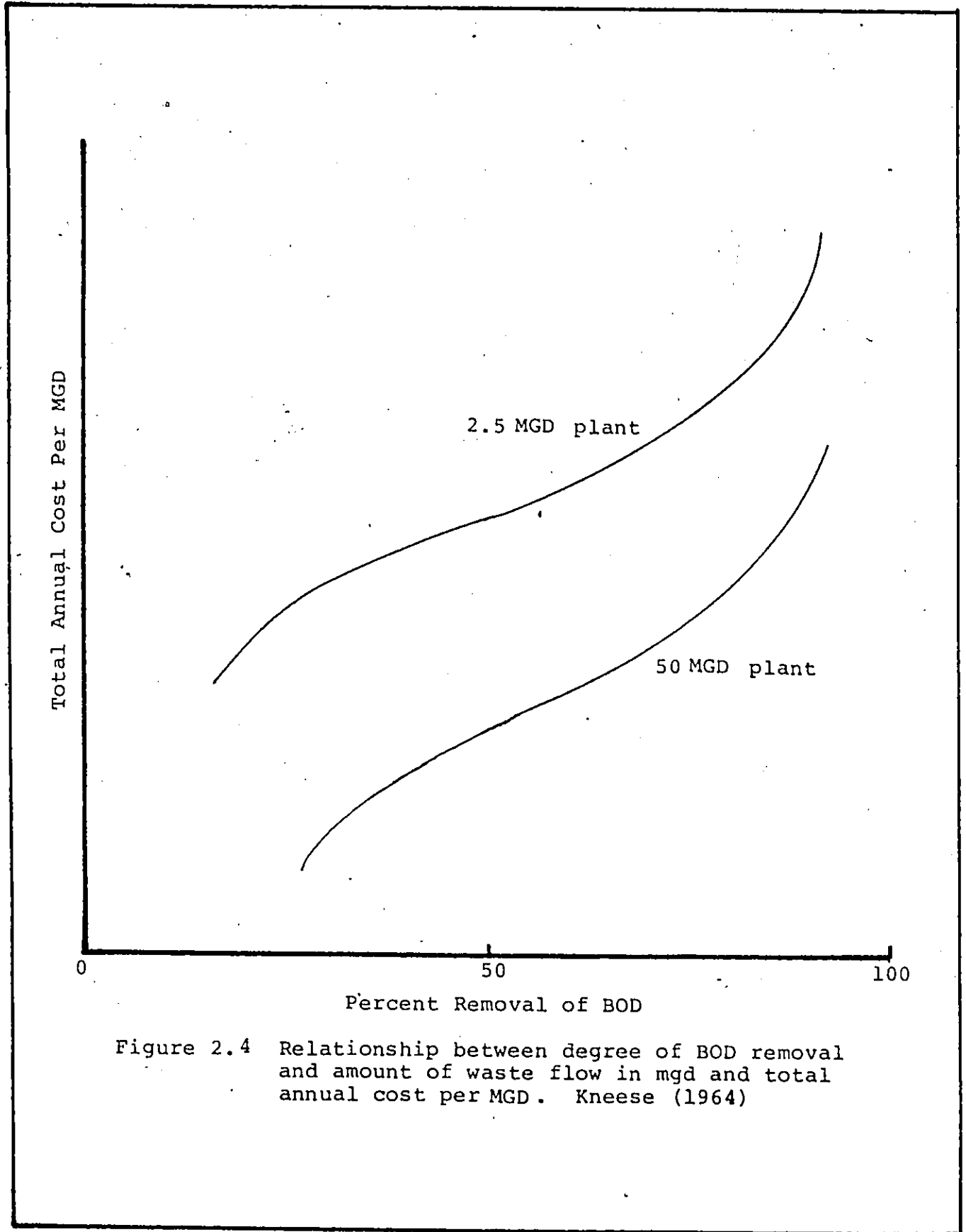


Figure 2.4 Relationship between degree of BOD removal and amount of waste flow in mgd and total annual cost per MGD. Kneese (1964)

CHAPTER III

THE EFFLUENT CHARGE AND WATER POLLUTION CONTROL

3.1 Introduction

Water resources have served as free goods and been considered as infinite sinks for the purpose of waste disposal. In an economic sense, goods or services are not free (scarce) when all users cannot be fully satisfied at zero price (Beckerman, 1975). Nevertheless, when a firm, for example, discharges its pollutants into a stream, it is depriving the community, surround the waterway, of the benefits of clean water. If the community could always provide itself, free of charge, with unlimited amounts of clean water, the use of the stream by the polluters would not be an issue. However, the fact is that there is always a value attached to enjoying clean water.

Accordingly, water resources should no longer be considered as free, and their scarce services should be optimally allocated among different users in order to maintain the quality of the environment.

Economic theory states that the market system, when functioning properly, can solve the problem of scarcity by allocating resources in an efficient way through price mechanism (Stigler, 1960). The market system, however, could fail in its function for many reasons. The phenomenon of market failure is particularly pertinent to the environmental resources due to the nature of these resources as "public goods". To overcome this failure a measure of

public intervention has been suggested. There are many different forms of intervention, one of which is the effluent charge strategy.

The objective of this chapter is two-fold. First, to present the idea of the market system and to discuss the reasons that have led to market failure in water resources planning; second, to evaluate various forms of public intervention, emphasizing the effluent charge strategy.

3.2 Market System and Resource Allocation

According to the theory of welfare economics, a resource is optimally allocated among all users when the welfare of society would decrease more by taking a unit of resource away from one user than it would gain by giving it to some other (Mishan, 1967). The value to society of employing the last unit of a resource in a certain way (i.e., marginal social value) is usually taken as the price. Therefore, if the output of all goods and services was pushed to the point where the benefit loss to society of producing the last unit (marginal social cost) was just equal to its price (marginal social value), resources would be optimally distributed among different uses and the pattern of output would have maximum value to society.

Analogously, the resource is optimally allocated among different uses when the marginal cost of producing the last unit of goods (i.e., marginal private cost) equals its price. Therefore, when the marginal social cost equals

the marginal private cost, and they both equal the unit price of the resource, then the resource will be optimally allocated.

Price in this sense reflects the relative importance of goods and resources in different uses. Usually, prices are determined in a free market through the forces of supply and demand. The perfect market should provide decision makers (producers and consumers) with information (in the form of prices) covering the relative profits and costs of using the resources at hand in different ways. Where all goods and resources pass through free market, and where there is perfect knowledge and resource mobility, prices serve to guide resources to their most beneficial use.

If there is no market, however, for some valuable resources, or if the market does not function properly, the resulting resource allocation will not be optimal. In this case, it is said that market failure has occurred.

The phenomenon of market failure is particularly pertinent to the environmental resources for three main reasons. First, there is lack of property rights. For the market to function properly, the ownership of the resource must be clearly definable and enforceable so that the owner can prevent others from using, benefiting from, or damaging his resource. Since environmental ownership is not established, polluters do not have to either pay the owner for using it or compensate him for the damages

caused.

The second reason relates to the "public goods" nature of the environmental resources. Public goods, once supplied to an individual, are by their very nature freely available to all; it is not practical to exclude some people from using them. Since other users cannot be excluded because of non-payment, producers of public goods are unable to collect revenues from beneficiaries. This situation could induce the producers to provide such goods in less-than-optimal amounts*. Clean water and air are typical examples of public goods. Suppose a single firm is granted exclusive property rights to waterways in a region. The firm has the right to sell clean water (that is, water reduced in pollution) to buyers willing to pay. But if pollution is reduced for one user, it will be reduced for all the users in the region, whether or not they have to pay. Since the firm cannot collect from all beneficiaries the total cost of the clean water, it will have little or no incentive to curtail pollution.

The third reason for the failure of the market system to reach an optimal pattern of resource allocation has to do with the externality effects.

* The optimal quantity of a good is that amount at which marginal value (social cost) equals marginal cost (private cost). In a case where the marginal value of a good is greater than the marginal cost, production of goods should be expanded as long as there is an opportunity to capture such gain. But where a characteristic of the good is its public nature, the producer may not be able to capture the full marginal value and hence lacks the incentive to expand output to the optimal level.

There is a class of economic activities characterized by what is called "externalities", of which pollution output is a prime example. In these activities, optimal output does not result through market function because some of the costs generated by these activities are not borne by those responsible for them. Externalities in using water resources, for example, can occur when upstream users discharge their waste into the stream at no cost to themselves, inflicting damage and costs on downstream users. These costs will not normally enter into the calculations or accounts of the upstream users; they are "external" to such calculations. At the same time, downstream users have no power to ask upstream users for compensation.

The externality effects of water pollution will result in the misallocation of water resources as well as the production of socially undesirable patterns of output. In the absence of any obligation to properly dispose of their wastes, polluters will be encouraged to enhance the amount of their production, which is artificially cheap. This tends to induce the over-production and over-consumption of some products which may have the socially undesirable result of under production and under consumption of other products.

Kneese & Schultze (1975) brought attention to the fact that a society that allows waste dischargers to neglect the compensation costs of their waste disposal, is

encouraging them to produce too much waste without any concern for social needs. This results in a misallocation of water resource services and a considerable deterioration of water quality.

3.3 Economic Incentives For Water Pollution Control

As discussed before, water resources are broadly regarded as public goods. Managing these resources, therefore, is largely a matter of applying the general principles of public goods management. This is achieved through public intervention in order to account for externalities (Kneese, 1964). The public authority responsible for this intervention is required, at first, to establish an acceptable policy regarding water quality standards. These standards are not rigid, and they can be changed according to any circumstances. Once the quality standards are established, the public authority need only seek the best method of implementing the policy. Although a number of methods have been suggested over the years, the following three have received the most attention, both theoretical and practical:

i) Regulation Enforcement Strategy - This is the most widely used control strategy, and it can be viewed as a two-step process. First, the authority establishes allowable quantities and compositions of discharges for each polluter, as well as the time and place of discharge. Second, the authority legally enforces such regulations, i.e., it can impose sanctions when violations are detected.

The strategy could succeed in optimally determining

the amount of waste discharge at each pollution source, provided that the public authority has all the information regarding the time of discharge, the amount of waste, the type of waste treatment technology used, the treatment cost, etc., for every polluter in the region (Dales, 1968; Freeman and Haveman, 1972).

ii) Subsidies Strategy - The public authority may decide to subsidize municipalities or industries on a portion of their expenditures on pollution treatment so as to keep the amount of waste down to the level that it has established (Dales, 1968). The subsidy may take different forms, such as payments for building municipal treatment plants and allowing industries to divert their wastes to these plants at costs below normal, or giving firms grants and loans to set up their own pollution control facilities (Dales, 1968). The subsidy may also be offered in the form of a bonus of a certain number of dollars per unit of treatment achieved. For example, a certain sum may be offered for each percentage reduction of BOD level in a river. In such a case, assuming that the authority has set the tolerable pollution limit, by one way or another, at a certain level "Q", the subsidy should be set equal to the marginal benefit of the unit increase associated with a unit decrease in pollution at the "Q" level. The polluter therefore will have an incentive to increase treatment up to the "Q" level as long as his marginal cost is less than the per unit subsidy, but not beyond that level.

(iii) Effluent Charge Strategy - This strategy requires that anyone discharging waste into a waterway has to pay the public authority a certain sum (effluent charge) per unit of discharged waste. The method was originally suggested by Kneese (1964), then developed and rationalized by many others such as Dales (1968), Baumal and Oates (1971) and Tietenberg (1973 (a), (b)). The method proposes to bring the limited assimilative capacities of watercourses under the effect of the market function by charging those who would use these common resources a certain price. Each user is led to compare the cost of using the resources for discharging waste with the cost of handling his waste disposal problem in some other ways. His only guide is the relative cost of alternative procedures, one of which is discharging the waste into the environment untreated, and paying the charge.

This method provides incentive for the efficient use of the environment, similar to the incentives which induce efficiency in the use of labour, capital, and land (Freeman, 1971). The waste dischargers will realize the cost of degrading the environment through paying effluent charges. This could bring the firm's private costs (including the charges) into equality with the social costs, resulting in an optimal pattern of water resources allocation.

3.4 Evaluation of Effluent Charges Strategy

Choosing appropriately among the above three strategies is not a straightforward task, since the pollution

problem itself has many socio-economic effects on such things as industrial location preferences, income distributions, desired production patterns, population distribution, and standards of living (Solow, 1971). Kneese and Schultze (1975), however, have mentioned three criteria that should at least be met for the chosen strategy to be successful; the strategy should be: (i) effective enough to cope with the complex relationships of the pollution problem (ii) flexible enough to achieve the social need of pollution reduction, and (iii) efficient enough to achieve the least costly means of pollution reduction.

Apparently, the regulation enforcement strategy, which is the most widely used method, has failed to achieve the required reduction at reasonable cost to society. A study issued by the Environmental Conservation Agency of Vermont (1972) gave two reasons for this failure. First is the little attention that has been given, in all legislations concerning water quality management, to the issue of economic incentives. Second is the neglect of the economic value of the waste assimilative capacity of watercourses. Since the early 1970's there has been a growing interest in investigating alternative policies. The effluent charges was one of these alternatives that has received great attention and interest from environmentalists, and from most economists as well. This interest apparently stems from the common view that establishing a price for the water resources would work as an economic incentive to correct and adjust the malprac-

tices of dischargers. The strategy, however, is facing some criticisms and misconceptions. In what follows, the effluent charges strategy is critically evaluated in the light of the above three criteria.

Effectiveness - It is most likely that a monetary incentive will be more effective for reducing water pollution than any legal penalty. The monetary incentives are the essence of the effluent charge strategy. The regulation-enforcement strategy, however, does not provide such incentives; rather, it enforces the regulation through legal penalties which are difficult to administer because of the problems involved in detecting violations (Freeman, 1971). On the other hand, under ideal conditions the regulations could provide enough incentives for the dischargers to just meet the standards, while in the effluent charges strategy, the dischargers would have a continuing incentive to devote research and engineering talent to finding less costly ways of achieving further reductions from the standard requirements, since a charge must be paid for every unit of pollution released.

As for the effectiveness of the subsidies strategy it could generate results entirely opposite to those it is expected to achieve. For example, a large waste-producing industry may be able to reduce its waste at a cost per ton that is less than the subsidy it earns per ton by doing so. The industry may expand its production in order to engage in the profitable business of producing waste and then

treating it.

Flexibility - Since the techniques and costs of removing pollution vary widely from case to case, control policies that emphasize one particular approach, or impose specific limitations on all polluters are bound to be rigid, and to result in inefficiency (Kneese and Schultze, 1975). The effluent charges strategy seems to offer the greatest flexibility, in the sense that it does not specify how to deal with pollution; the choice is left to the dischargers who have a wide range of options. They may be able to undertake treatment of part or all of the residuals, they may change the process to reduce residuals per unit of output, or they may find that recycling and material recovery is less expensive than either treating the waste or paying the charge. In any case the dischargers are led to compare the cost of handling their waste disposal problem with the cost of using the river as a waste receptor. The strategy, moreover, does not need detailed and accurate information, especially regarding the treatment cost for every discharger in the region (Freeman & Haveman, 1972). It is sufficient to start with a little information; then, once charge value is known each polluter makes his own cost calculations and responds to the public authority by reporting the amount of pollution he will discharge. Using this information, the authority can calculate a new schedule of effluent charges. The procedure continues in this way until it forms an optimum schedule of charges for the attainment of the given

quality standards.

On the contrary, in the case of the regulation-enforcement and subsidies strategies, detailed information about the treatment costs and discharge points must be available to the authorities (Baumol & Oates, 1971). These represent a massive administrative problem, and it would be difficult and expensive to collect such information.

Efficiency - Due to the flexibility of the effluent charge strategy's management of pollution control, dischargers, are more likely to seek efficient and least costly methods of pollution reduction. An efficient method requires that different firms reduce pollution by different amounts, depending on the cost of reduction. Moreover, each firm - in addition to applying conventional waste treatment techniques - should take advantage of a wide range of control alternatives: modifying its production processes, recycling its byproduct wastes, and using the varieties of its products which cause less pollution. Efficiency, in this sense, is more inherent in the effluent charges strategy than in the other two.

In spite of the clear advantages of the effluent charges approach, the strategy has been misunderstood and criticized by various groups of environmentalists, businessmen, and economists. It is important, in the course of evaluation of the strategy, to put the more substantive of these criticisms into proper perspective, and to correct any misconceptions:

The main criticism, raised by some environmental groups, is that the strategy is an open invitation for the polluter to purchase the right to degrade the environment. It is important, in evaluating this argument, to recognize the natural capacity of watercourses to assimilate waste loads. It would be unreasonable to neglect the economic value of the waste assimilative service. The effluent charge strategy, like any other economic strategy, makes proper allowances for this within the limits of water quality established by the authorities. The argument, however, could make sense if the required payment of the charge fails to maintain the desired quality limits or to induce dischargers to search for new treatment methods in order to reduce liability to the charge. Nevertheless, these situations can be managed if the charge is set above the costs of the highest treatment level, in order to induce the profit-maximizing firms to search for cheaper methods to reduce their pollution, at least to the specified limits (Kneese & Schultze, 1975).

Another criticism has been advanced by business groups, who argue that the effluent charges would hinder industries' ability to reduce pollution and finance pollution control equipment. They maintain that firms will find it more expensive to reduce pollution by a given amount, because in addition to the real cost of waste treatment they will also have to pay the effluent charge on their residual wastes. By contrast, it is argued that if they were obliged to reduce pollution by direct regulation, for example, they would incur only the real cost of the treatment.

In order to counter the above criticisms, one might put forward two economically-based arguments in favour of the original proposition. First, in the ideal situation, authorities would choose the method which will be cheapest for the economy as a whole, not for the firms only. Beckerman (1975) has argued that, on the national economic scale, the only real cost of pollution treatment is the cost of real resources used by the firms to reduce pollution, that is, the cost of labor and raw materials employed in the treatment processes. Effluent charges, or the effects of the charges, must be excluded in calculating the actual treatment costs, since the charges are considered as transfer payments rather than resource costs. It is assumed that the collected charges must be transferred to other public expenditures or be offset by reduction in other governmental tax revenues. Therefore, the only real cost of pollution control is the total cost of the resources used by the firms.

Second, the proposition of the business groups suggests that the main effect of any decrease in profits would be to reduce expenditures for the treatment of residues. Without appropriate incentives, however, it is probable that businesses do give low priority to investment in pollution control. It is the purpose of the effluent charge strategy to alter these priorities and to put more pressure on firms to seek efficient methods of pollution treatment.

The third criticism came from some economists, primarily from Coase (1960). Coase argued that any form of

taxation such as effluent charges is likely to lead to a misallocation of resources. He argued that private negotiation between the two parties may work as an optimum method. There are, however, three main barriers which prevent greater reliance on this approach. First is the size of the environmental units in which property rights must be vested. A major river system, for example, is inherently indivisible. To secure the economic benefits, based on creation of property rights, control over the whole system must be granted to a single entity (Freeman III, 1971). Yet Coase's approach would require the establishment of a large number of entities to own the system, a case which is quite difficult to arrange.

The second barrier is the presence of public good attributes, in many environmental services, which could not exclude users who have not paid for the services provided.

The third barrier is the cost associated with the bargaining processes. The processes are costly in the sense that it takes time to gather information, evaluate other alternatives, and execute the transaction. Because many parties are involved in the environmental negotiations, this cost tends to be high. Freeman (1971) realized that the cost of government action for administration and enforcement of effluent charge strategy is smaller in comparison to the private transaction costs when the bargaining processes are taking place.

In conclusion, the effluent charge strategy seems to be promising as an economic instrument for controlling water pollution. This stand relies on three main considerations. First, imposition of effluent charges encourages a pattern of waste management among different firms and municipalities that tends to minimize the cost of control for the river system as a whole. Second, the charges provide a continuing incentive to adopt improved technology as it becomes available. Third, this strategy is much more likely to be enforceable than the other alternatives.

CHAPTER IV
EVOLUTION OF EFFLUENT CHARGE STRATEGY -
A LITERATURE SURVEY

4.1 Introduction

Early in the 1960's, research began on the conceptual design of, and rationalization for effluent charges as a strategy for pollution control. Unfortunately, the research progress in this area has been relatively slow. The strategy, at that time, faced opposition and misunderstanding from different industrial, political, and environmental groups. However, the growing awareness, in the early 1970's, of the failure of present control strategies to cope effectively with the pollution problem, has contributed significantly to the willingness of examining other alternatives, particularly the effluent charge strategy.

During the last few years there has been some work to develop both the theoretical basis of effluent charges, and to pave the road toward their practical application. These efforts have successfully produced two theories for establishing charging methods by means of the market system. Yet, little has been done to put these methods into practice.

This chapter presents first, a survey of the works which have been carried out to date on the development of the theory of effluent charges; second, a summary of the studies which have used the strategy within a framework of a water quality management model.

4.2 Theoretical Development of Effluent Charge Strategy

The theoretical foundations of the strategy have been set forth by Kneese (1964). He suggested that a uniform charge rate be levied against all the polluters in a region equal to the dollar value of the marginal damage caused by waste discharge. Kneese realized that a charge equal to the marginal pollution damages will lead to efficient allocation of the water resources. People discharging waste will try to minimize their own costs by equating their marginal costs of waste reduction with the effluent charge.

At that time, the main difficulty with Kneese's proposal lay in obtaining a reasonable estimate of the monetary value of the marginal damage cost of pollution. This difficulty was due to the fact that the number of activities involved and the number of persons affected by the damage were so large that it would be difficult to collect adequate information regarding the damage cost with a reasonable degree of accuracy. In addition, the intangible nature of many of the consequences - damage to health, and the aesthetic costs - added to these difficulties. However, later on Kneese & Bower (1968) reported some promising work constituting a first step towards a monetary estimation of the pollution damages. But the study included the costs due to the loss in recreational benefits only.

To overcome the difficulty entailed in Kneese's proposition, two variant methods for determining the

appropriate charge level have been devised. In the first method, Dales (1968) suggested that the government can control the demand on water resources when it is used as a waste receptor. According to Dales, the government then issues a number of "pollution Rights", each Right giving whoever buys it, the right to discharge one equivalent ton of waste into the river. The Right holder can also sell this Right or buy it from any other Right holder. Because transferable and full property rights always command an explicit price, the establishment of such Rights makes it easy to establish a free market for them. In turn, the buying and selling of Rights in an open market results, theoretically, in an efficient allocation of the water resources among different users. On the other hand, as time goes on, we could expect the growth of population and industry to result in an increase in the demand for Rights. Since the number of Rights issued cannot be increased, the price of the Rights will move upward, thereby stimulating the incentive of waste dischargers to treat or reduce their waste in order to decrease the number of Rights they must purchase.

In the second method, Baumol and Oates (1971) argued that the call for imposing a charge on polluters equal to the marginal damage costs has rarely proved feasible. They suggested, instead, that the government could establish, at first, a set of somewhat arbitrary standards of environmental quality levels. Then they could impose

a set of uniform charge rates on waste emissions rather than attempt to base them on the unknown values of marginal damage. They recommended that the charge should be high enough to attain the environmental quality standards. The appropriate charge rates that yield the desirable level of pollution control at total minimum cost, can be reached by the process of trial and error. If after the first trial, the level of pollution is found to be higher than the desired level, the charge rates can be increased to force dischargers to reduce their pollution level, and vice versa. This method is known as "The Pricing and Standard System".

Since the effluent charge strategy has been set up, some decision and law makers, especially the economists, have expressed their acknowledgement of the strategy. Many papers have been published to clarify the rationality of using it, to defend its theoretical and practical basis, and to explain the merits of the strategy. Boyd (1971) supported the strategy on the basis, that the charges collected would be used to reimburse the public authorities for the cost of constructing and operating dams, aeration devices, etc., in order to expand the waste-receiving capacity of the river. At the same time the water quality requirement would be attained. Solow (1971) appreciated the strategy for three main reasons. First, the fact that the strategy attains the desired abatement level of pollution; second, that it provides enough incentive for the

polluters to search for cheaper methods of waste abatement, and finally, the authorities need little information regarding the quantity, cost, and timing of pollution.

Freeman & Haveman (1972), in their defence against the strategy, made a comparison between Regulation Enforcement and effluent charge strategies. The comparison was based on the effects of each strategy with respect to factors of: inflation, resource misallocation in the presence of monopoly, administrative problems, hindering of industrial abatement, and the cost of attaining environmental quality levels. They concluded that the effluent charges is far more advantageous than the other strategy with respect to the above factors.

Ruff (1970) in his comparison between pollution Rights and The Pricing and the Standard System noted that the two methods will lead to the same results, as long as the authorities succeed in establishing correctly either the amount of pollution Rights, or the starting charge level. Tietenberg (1973,a), on the other hand, realized that the Pricing and Standards System is easier and least costly to administer than the pollution Right method.

Tietenberg, however, disagreed with the notion of proposing a uniform charge rate for all the polluters in the region. To demonstrate this, he employed an abstract model for controlling pollution, using effluent charges as a basis for addressing the issue of whether or not the pricing and standard system is a sufficient means for a government to

efficiently allocate environmental resources, assuming that consumers, producers, and the government are acting non-cooperatively in their own self interest. The model described the relationship of pollution with consumers through their utility functions, with producers through the production functions, and with society through a social welfare function. The analysis showed that the superiority of this control system depended on imposing a set of non-uniform charge rates on the polluters in order to attain the minimum cost advantage. The charge rates should differ according to geographic, temporal and spatial factors. Similar conclusions have been reached by Ackerman (1973), who concluded that the charge rates should be based on the effect of an emission on the community, and not necessarily on the amount generated.

Baumol & Oates agreed with Tietenberg's proposition in their recent work (1975) which dealt with different economic policies for environmental resources control. Brumm and Dick (1976) also concluded that, from a practical standpoint, the Pricing and Standard System has considerable appeal in terms of overcoming the problem of lack of information regarding the treatment cost functions in order to set up optimal charge rates. Orr (1976) realized that the greatest advantage of the effluent charges was in their provision of decentralized incentives for technological change in wastewater treatment methods. Magat (1978) also appraised the effect, in a dynamic world, of the effluent charges and regulation - enforcement strategies on both the rate and the direction of

technical advance for effluent abatement technology and output technology. The results demonstrated that, under certain conditions (easy labor substitutability), constant effluent charge induced the firm to allocate effective research and development efforts to effluent abatement technology. On the other hand, using constant standard policy induced the firm to allocate more efforts to the output production technology.

In addition to technical advance, the constant level of effluent charge was found to fail in limiting the increases in the effluent discharge rates. In other words, for a typical polluting firm, a rising charge rate is needed to halt growth in the effluent level with advancing technology. Regarding this result, Magat realized that, in order to reach a given goal for environmental quality, the iterative process suggested by Baumol & Oates (1971) would never converge to a fixed charge level. Magat therefore suggested that, in order to maintain the environmental quality goal, government would have to either inform polluters that the initial charge level would rise over time at some unspecified rate, or initially announce a rising charge function. The author, however, is in favor of the former approach.

4.3 Practical Application of Effluent Charges

In the realm of practical applications of the strategy, the study given by Johnson (1967) is considered to be one of the pioneer studies in this respect. The study

aimed, in general, at gaining some practical experience in the estimation of effluent charge levels in the Delaware Estuary. It was found that the cost of waste treatment, induced by a charge level, will approach the least costly treatment plan. The study concluded that effluent charges should be seriously considered as a method of attaining water quality management. The study, however, simply formulated the computational model as a linear one.

Upton (1968) presented a linear model of water quality management in order to find the optimal amounts of waste discharge that minimized the total treatment cost incurred by all the polluters and the cost of low flow augmentation, such that the water quality levels should be maintained. The control of waste load discharges by the charging method has been shown to be preferable to the regulation control method under some conditions. The study showed that an optimal set of charges existed which would induce the polluters to discharge optimal amounts of waste.

Hass (1970) suggested a decentralized, decision-making scheme, for obtaining the optimal treatment configuration for meeting water quality standards along a river basin, simultaneously determining optimal pollution taxes to achieve this configuration. It was assumed that the authority knew the quality standards desired and the cost of low flow augmentation, and that polluters knew the individual treatment cost functions. The problem confronted by the authority was to find the combination of

each polluter's treatment plans and the level of flow augmentation, which minimizes the cost of achieving the standard imposed upon the river at the end of each reach. The authority wished to have each polluter treat waste to the point at which the marginal cost to society of disposing waste was equal to the marginal cost to the polluter. Hass formulated the objective function and constraints of his problem in the form of a linear programming problem.

The implications of Hass' study is that authority can, through a planning process involving proposed charge rates and polluter responses, determine a set of near-optimal pollution charges without complete knowledge of treatment cost functions. Disregarding the linearity assumptions in Hass' model, there is no guarantee that the solution is optimal since the polluters have no incentive to be honest in their responses. In fact, it is in their interest to understate their discharges to lower the resulting effluent charges.

In 1972, Agency of Environmental Conservation in the State of Vermont conducted a comprehensive study regarding various methods of establishing effluent charges and the feasibility of their application. Six methods were reported: (1) charges based upon the monetary estimate of the magnitude of pollution damages, (2) a uniform charge to all polluters, (3) uniform charge rate weighted by stream classification or zone, (4) a charge rate based on dilution factors or population equivalents (a composite

index derived from the concentrations of several residual pollutants), (5) charges based upon annual cost of treatment, and finally, (6) charges based upon stream quality degradation.

Each of these methods was evaluated according to six criteria: (1) Efficiency of resource allocation, (2) Relation of charges to instream economic damages, (3) Equity, i.e., whether an effluent charge system would result in any impairment of existing "rights" or "privileges" without some form of compensation, (4) Incentive effects on dischargers, (5) Administrative and technical feasibility and (6) Income potential generated by effluent charges.

The main purpose of the study was to discuss the problems of selecting and implementing the best method, among the six alternatives, which would suit the state of Vermont's water pollution control objectives. In view of the above assessment, the charge method based on discharger's annualized cost of treatment was found to be the most appropriate to the pattern of water quality management set by the state.

Elliot & Seagraves (1972) studied the impact of surcharges on industrial waste, using a sample of 198 observations from 34 cities. The study showed a substantial impact following the imposition of a moderate waste charge. It was indicated that a 45% reduction in waste load would follow from the imposition of a modest surcharge of 2.7¢/lb. of BOD.

Taylor (1973) presented a model for planning an efficient water quality program within a river basin. The control

method, called "rent allotment control", combined the effluent charge with the waste allotment or "Pollution Right" suggested by Dales (1968). It was implemented through a bargaining process between the authority and the waste dischargers that may be characterized as an "N-person prisoner's dilemma". The polluter was asked to respond with an allotment to which he would be willing to adhere and with which he would be willing to pay the effluent charge at the end of a prescribed planning period. If the noted charge was levied, violations of this agreed allotment would not be permitted. Furthermore, the polluter was told, if the allotment response was not low enough, although no indication was given of what "low enough" meant, the effluent charge would be raised and a new request for bids sent out.

The bargaining process, therefore, provides the authority with information about individual waste treatment costs, and terminates in a set of agreements on rents and allotments which depend upon the bidding strategies adopted by each polluter. Through successive iterations of bidding for waste allotment, the authority would reach an optimal situation where each polluter bids for an allotment such that the marginal cost of additional withheld waste is equal to the effluent charge. In fact, the main merit of this method is that it would overcome the dishonest reporting of information by some polluters. The method, however, might fail when all the N polluters

collude and underbid for the allotment.

Ferrari (1973) realized that it was not consistent to establish a static effluent charge rate to account for a situation of dynamic demand on water resources as a waste receptor. He therefore suggested a non-linear, progressive effluent charge scheme to account for the dynamic phenomenon of growing demand in a region. To establish his effluent charge scheme for meeting a specified level of regional environmental quality, he applied the "internal point penalty function" method, suggested by Fiacco & McCormick (1968), such that at any time the polluter will pay a charge directly proportional to the rate of increase in his demand. The scheme, however, requires a public authority to request from polluters the total and marginal costs of present treatment activity. Therefore, such an approach could fail to deal adequately with the problem caused by firms supplying misleading information, particularly cost information, to the public authority.

O'Sullivan (1974), in his survey of the measures taken by common market countries to curb discharge of industrial effluents to waterways, stated that the basis of charging for both community and industrial discharges is the pollution equivalent (P.E.) oxygen demand. This is a measure of the average daily discharge of oxygen-demand substances per head of population. The charge per P.E. of pollution discharge varies according to the area of the country, the total load burden of the receiving

water, and whether the discharge enters large or small rivers. In general, the charge, as O'Sullivan mentioned, varies from \$2.00 per P.E. to about \$7.00. However, the charge scheme of toxic wastes has not been developed yet.

The study given by Wenders (1975) analyzed the impact of improvement in pollution abatement technology on the cost of a firm working under three alternative methods of pollution control. The methods were: effluent charge, subsidy, and regulation-enforcement strategy. It was shown that the cost reduction would be greatest if the firm was operating under the effluent charge strategy. In addition, the incentive to improve pollution abatement technology would be greater if effluent charges were used to control pollution.

However, Nielson and Hwang (1975) reached a different conclusion when they studied the interaction between the incentives offered by the charge, and the individual polluters choice of waste water treatment technology. They concluded that increasing the charge might both fail to improve environmental quality, and increase production costs.

Marin (1978), in his search for the best control method of pollution, concluded that effluent charges seemed to be suitable for controlling river pollution more than air pollution (controlling sulphur dioxide) due to the advantages of: governing the number of polluters involved, monitoring pollution emissions, and the existence, somehow of a relationship between inputs and emissions.

Oates and Strassmann (1978) realized that the compelling case for the use of effluent charges to control polluting activities, as it seems to most of the researchers, proceeds from the assumption of profit maximizing behaviour by polluters. Since profit maximization is not presumably the primary objective of public organization, Oates and Strassmann studied the effectiveness of using the effluent charges to control the public sector sources of pollution. He employed the Niskanen model of bureaucratic behaviour (Niskanen, 1971) to investigate this issue. The Niskanen model operates on two basic premises: first, when the objective of the bureaucrat is to maximize the size of his budget, in order to serve as a proxy variable for the prestige, power, prerequisites, etc. of the bureau. This indicates the presence of "fat" in the bureau's budget, and the bureau is said to be demand constrained; second, when the bureaucrat works under shortage, i.e., the cost of providing output exceeds the bureau's budget. The objective of the bureaucrat is cost minimization, and the bureau is said to be budget constrained.

The study considered the two cases. It was found that the effluent charges can induce significant reduction in pollution activities when the bureau is budget constrained. This is because where there exists a less polluting technology such that the abatement costs are less than the effluent charges, the bureaucrat will shift to alternative technology, since the bureaucrat is a cost minimizer. The study found, however, that the response of the demand constrained bureaucrat to the effluent charges is not clear. This is because

most likely the bureau can absorb the charges with some of its surplus budget with need to search for less costly methods for abatement, unless the charges are sufficiently high to push the bureau from a demand constrained to a budget constrained situation.

In fact, this study is important, since public sector sources of pollution represent a major source of waste emissions (approximately 57% of BOD accounts for the emissions coming from municipal discharges). However, more practical experimentation for the use of the charges is clearly needed.

Rinaldi, et al (1979) studied the possibility of achieving a stable and efficient taxation scheme in regional environmental management systems. Motivated by the concepts of the game theory, the scheme comprises a set of rules which, given a set of polluters, their profits and costs, and a central authority for environmental control, generates taxes to be levied on the waste emissions. The most important aspect of the scheme is its stability which refers to the possibility of overall cooperation between all the polluters. The study demonstrates that, if the damages to the environment are not negligible and if these damages must be refunded by the users of the resource, it is very unlikely that efficient and stable taxation schemes can be found if the regional authority acts as a non-profit corporation. The results also state that if the damages to the environment are neglected, or in other words, if the total benefits of the firms instead of the social benefit is maximized, the efficiency and stability are easily obtained. Increased

public awareness, however, makes this solution no longer feasible. The alternative is to achieve stability and efficiency by letting the regional authority get a profit by the sale of emission rights. The higher the environmental congestion, the greater must be this profit. But if ethical or political attitudes are against this kind of solution, there is no way to maximize the social benefit without generating friction among the polluters.

CHAPTER V

DEVELOPMENT OF AN EFFLUENT CHARGE OPTIMIZATION MODEL

5.1 Introduction

The mathematical model presented in this chapter has been developed to obtain the optimal effluent charge rates to be levied against polluters using the river system as a waste receptor. The model is formulated within the framework of a water quality management system. The modified transfer equations of Streeter-Phelps (1925), as suggested by Dobbins (1964), are employed to predict the values of BOD and DQ along the river. The model's assumptions, and the limitations of its applicability, are discussed first. A description of the model's structure, including the objective function and the constraints, is next exhibited. Finally, the formulation of the model is presented.

5.2 Model's Assumptions

The assumptions involved in the development of the model are as follows:

- 1) The model is limited to stream and river basins as main receptors of waste. The model establishes different zones along the river at points of discontinuity. A point of discontinuity arises whenever the existing conditions violate the assumptions of the transfer equations, or changes in water quality standards, or of boundaries determined by law that place a region under separate jurisdiction. Examples of discontinuity that require establishing a new zone are:

- i) organic and/or thermal inputs,
- ii) a change in the flow regime such as changes in stream width or depth,
- iii) increase in the flow due to tributary input,
- iv) changes in water quality standards as the river crosses a provincial line, (Shojalashkari, 1974).

Figure 5.1 schematically depicts such points.

The establishment of zones along the river facilitates the modelling process. Specifically, the complex nature of pollution interaction at various discharge points causes difficulty in the application of the transformation equations along the entire river if it is considered as one zone.

However, by dividing the river into a number of segments or reaches, adequate results would be obtained by applying the equations, independently, to each segment.

2) Since municipal and industrial waste treatment plants are the major polluters of waterways, the model considers them as the main sources of pollution. The largest portion of waste released from these sources is degradable (Kneese, 1964); hence, the model is developed to deal mainly with this type of waste.

3) The most common measures of pollution intensity are the dissolved oxygen (DO) content of water, the biological oxygen demand (BOD), and the phosphorus concentration (PO_4). These measures are used as pollution indices in the development of the model.

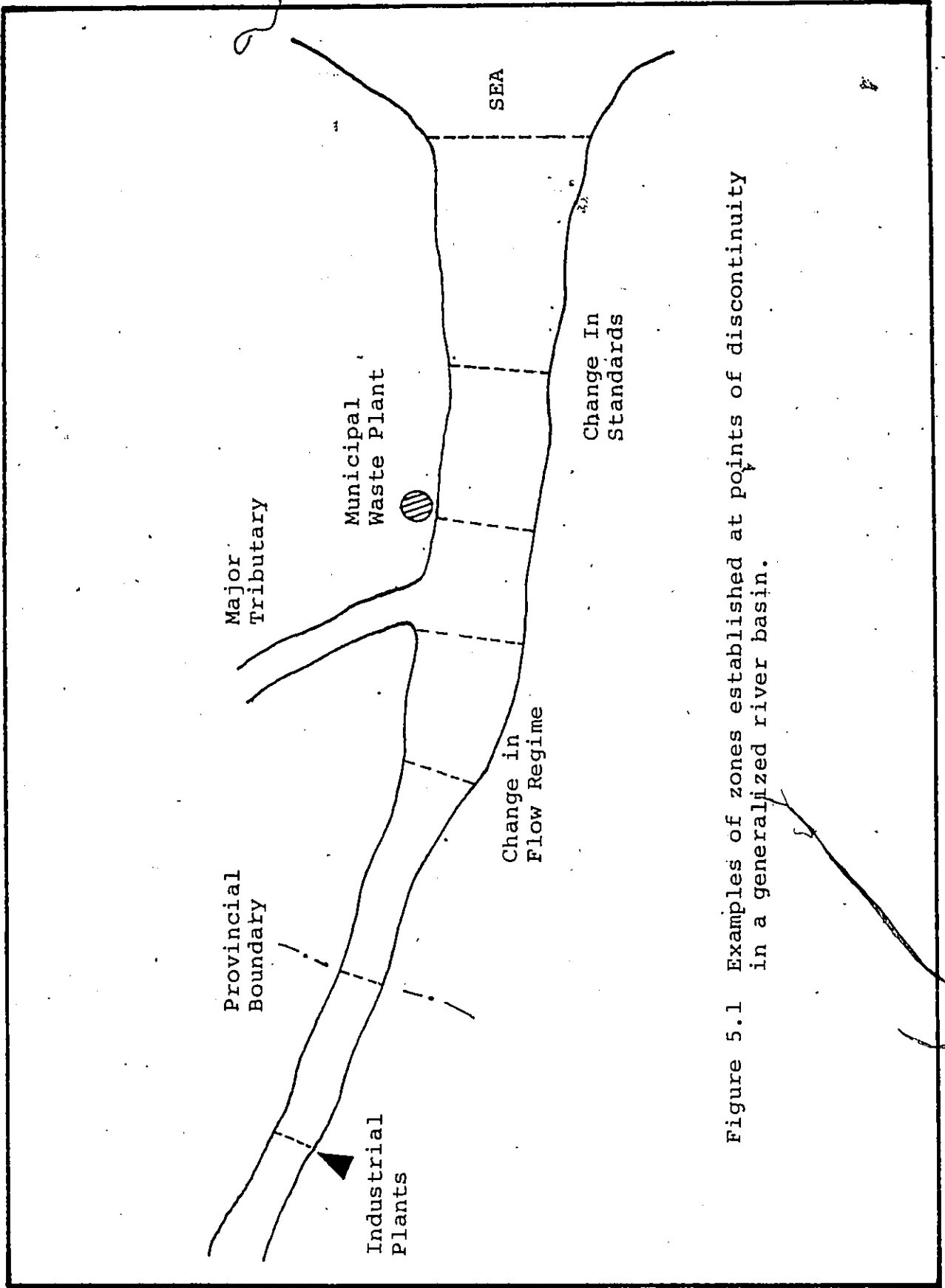


Figure 5.1 Examples of zones established at points of discontinuity in a generalized river basin.

4) The model assumes that the phosphorus has no chemical or biological interaction with the organic load and that its concentration level remains constant along the river as long as there is no change in the flow rate.

5) There are two major one-dimensional models which describe the concentrations of BOD and DO with respect to the distance from discharge source. The first is the mixed model, which predicts the BOD and DO levels along the river by assuming a complete mixing of pollutants with the water body. This mixing becomes appreciable if the stream velocity is low, as in estuaries or streams with dams. The second, or plug flow model is applicable to streams with a high flow rate and comparatively negligible mixing effects. This situation is more likely to occur in most rivers and streams, especially in upstream sections (Fan, et al, 1973).

Since the study is limited to streams and river basins, it is appropriate to employ the plug flow model, represented by a modified version of the Streeter-Phelps equations (1925), as suggested by Dobbins (1964). The modified model considers the following:

- i) the sedimentation effect to remove BOD,
- ii) local runoff effect to add BOD,
- iii) photosynthetic action of plankton and plants to add oxygen,
- iv) the respiration of plankton and plants to remove oxygen from the river.

According to the modified Streeter-Phelps equations, the BOD concentration L , and DO concentration C vary with

time according to the following relationships:

$$\frac{dL}{d\tau} = -(K_1 + K_3) * L + L_a \quad (5.1)$$

$$\frac{dC}{d\tau} = -K_1 L + K_2 (C_s - C) + \langle P - R \rangle \quad (5.2)$$

where:

τ = residence time, in days, from source of pollution

K_1 = the biological oxidation rate coefficient, day⁻¹

K_2 = the reparation rate coefficient, day⁻¹

K_3 = the sedimentation or absorption rate coefficient for BOD, day⁻¹

L_a = local BOD runoff rate, mg/l/day

C_s = the saturation DO concentration, mg/l

$\langle P - R \rangle$ = the rate of photosynthesis minus the rate of respiration, mg/l/day.

The values of the parameters K_1 , K_2 , C_s , and $\langle P - R \rangle$, which are temperature-dependent, may be given as follows, (Davidson and Bradshaw, 1967):

$$K_1 = 2.35 * 10^{-7} \text{ Exp}(0.0464T) \quad (5.3)$$

$$K_2 = 0.43 * \text{Exp}(0.025(T - 273)) \quad (5.4)$$

$$C_s = 4000.0 * \text{Exp}(-0.021T) \quad (5.5)$$

$$\langle P - R \rangle = \frac{\pi - \alpha}{\pi(1-\alpha)} * (25 - 0.028 (T - 303)^2) \quad (5.6)$$

where:

α = a constant between 1 and 0

T = stream temperature in degree Kelvin

π = 3.14

6) The present study assumes that the temperature of waterways may change from zone to zone, but is constant within the same zone. This assumption is applicable in some situations. For example, when the river is long enough to flow through different climatic zones, (e.g., the river Nile in Africa where it starts at the middle of the continent, ends at far north in the Mediterranean, and passes through different climatic zones). This causes a temperature rise at some zones and a drop at some others.

7) The model considers the amount of pollutants, measured in pounds of BOD per day, as the basis for calculating the effluent charge. It seems that this method creates more incentive for dischargers to reduce the strength of their waste than if it was based on measurement of volume, for example (Maystre and Geyer, 1970).

The effluent charge, y , itself is set to be inversely proportional to the percentage BOD treatment, x , so that the dischargers are motivated to increase their treatment level and avoid the high charge. The inverse relationship may take several forms, such as:

$$\text{Linear Form: } y = \theta_1 - \theta_2 * x \quad (5.7.1)$$

$$\text{Nonlinear Form: } y = \theta_1 * \left(\frac{1}{x} - 1\right)^{\theta_2} \quad (5.7.2)$$

$$\text{Exponential Form: } y = \theta_1 * \exp\left(\frac{-\theta_2}{1-x}\right) \quad (5.7.3)$$

where θ_1 and θ_2 are parameters, whose optimal values are to be determined.

The first form is simple and straightforward. The second requires an upper bound constraint to limit the charge rate from approaching infinity when x approaches zero. The third, which is used in this study is a self-bounded form, see Figure 5.2.

5.3 Development of the Model

In the development of the water quality management model, a river basin which is divided into a number of zones according to the criteria discussed before is considered. Each zone receives pollutants from different municipal and/or industrial sources, located along the river. The objective of the model is to determine the optimal values of the effluent charge rates, as well as the treatment configuration for each source, such that the desired water quality levels are maintained at a minimum total cost to all the polluters.

The objective function of the model describes the total costs which consist of the waste treatment cost incurred by all dischargers and the effluent charges paid on the untreated portion of their waste discharges. The constraints of the model represent, for each zone, the water quality constraints, and structural and non-negativity constraints for the variables involved.

In the following sections, the objective function and the constraints are developed in detail.

5.3.1 The Objective Function

The costs incurred by each polluter includes one or both of the following components: (1) the cost of cleaning

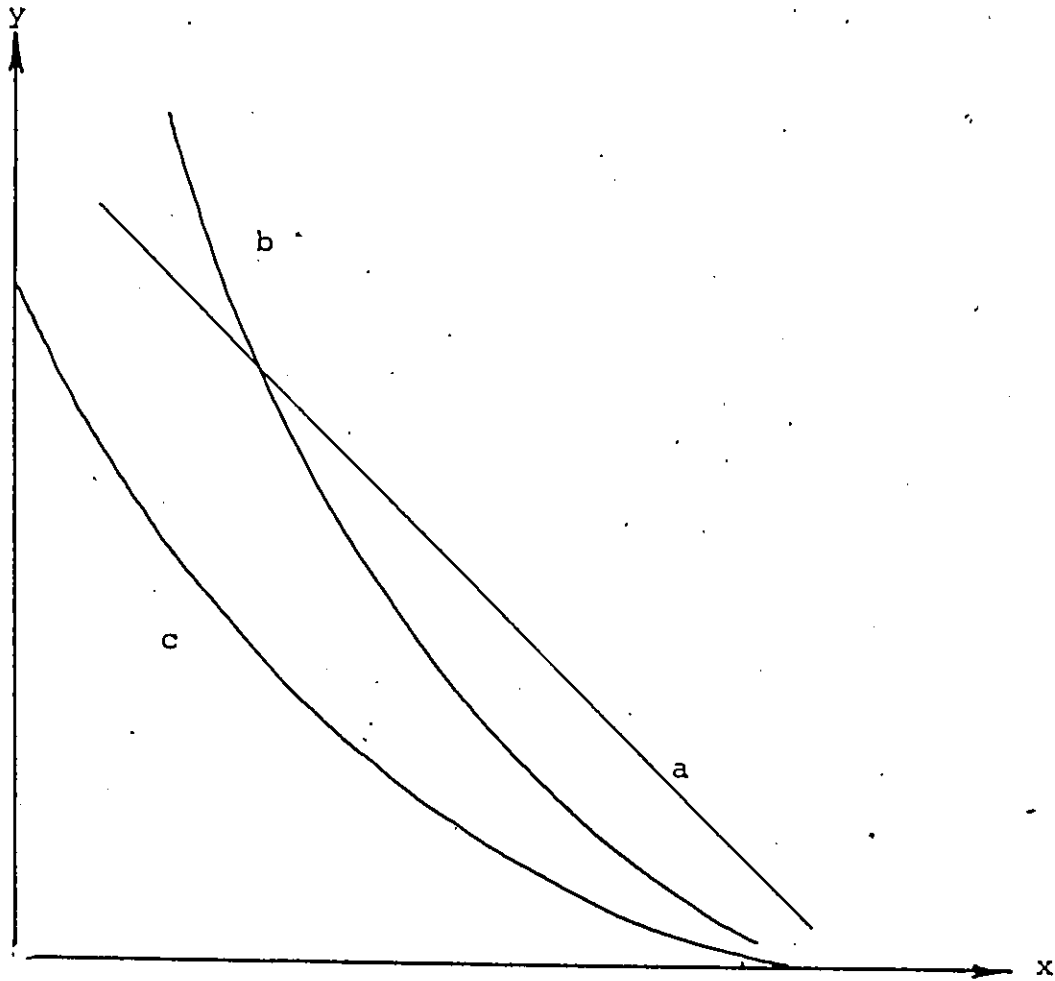


Figure 5.2 Relationship Between Effluent Charge Rate y , and Waste Treatment Level, x .

- (a) Linear Relationship
- (b) Nonlinear Relationship (with upper bound)
- (c) Exponential Relationship (bounded).

up that portion of the waste decided to be treated, and (2) effluent charges on the remaining part of the waste that is discharged untreated. The objective function includes all such costs for the N polluters in the region.

Consider zone i , the treatment cost in this zone, FT_i , which covers only the operating and maintenance cost of the waste treatment plants, is a function of:

- (i) Waste loads, measured in lb /day, of BOD, WB_i , and of phosphorus, WP_i .
- (ii) Percentage treatment of BOD, x_i , and of phosphorus, z_i .
- (iii) Average waste flow, \bar{Q}_i , measured in million gallons per day (MGD).

The effluent charge cost, FC_i , is directly related to the amount of waste load discharged, WB_i , and the effluent charge rate, y_i , which is inversely related to the treatment level of BOD, x_i .

The objective function can thus be mathematically expressed as follows:

$$\begin{aligned} \text{Minimize} \quad & \sum_{i=1}^N \{ FT_i(WB_i, WP_i, x_i, z_i, \bar{Q}_i) \\ & + FC_i(WB_i, y_i, x_i) \} \end{aligned} \quad (5.8)$$

5.3.2 The Constraints

The model is subject to two sets of constraints; the first accounts for the desired water quality limits, defined by the maximum concentration of BOD, the minimum concentration of DO, and the maximum concentration of phosphorus. The

second set is of a structural type, accounting for the maximum and minimum limits of effluent charge rates y_i , BOD treatment level x_i , and phosphorus treatment level, z_i . The following are descriptions of these constraints.

1) Water Quality Constraints

(i) Maximum Concentration of BOD:

Assuming that LB_i is the maximum BOD content in zone i due to waste discharge, this can be represented as follows:

$$LB_i < \text{MAXBOD}_i, \quad i = 1, 2, \dots, N \quad (5.9)$$

where MAXBOD_i is the maximum allowable level of BOD in zone i .

To explain how this is related to waste discharge level, the following analysis can be considered:

The amount of organic load, WB_i , of BOD in lbs/day is discharged at the beginning of zone i , causing a rise in BOD level at the discharge point, and at subsequent points downstream, as shown in Figure 5.3.

Assuming that x_i percent of the waste load is treated, then the increased amount of BOD, in mg/l, at the discharge point can be calculated as:

$$\Delta L_i = \left(\frac{1.0 - x_i}{1.0} \right) * \frac{F1 * WB_i}{Q_i} \quad (5.10)$$

where:

$F1$ = a conversion factor = $0.185405 \frac{\text{day} \cdot \text{ft}^3 \cdot \text{mg}}{\text{sec} \cdot \text{lb} \cdot \text{l}}$

Q_i = stream flow rate in zone i , cfs

ΔL_i = increase in BOD level at the beginning of zone i , mg/l

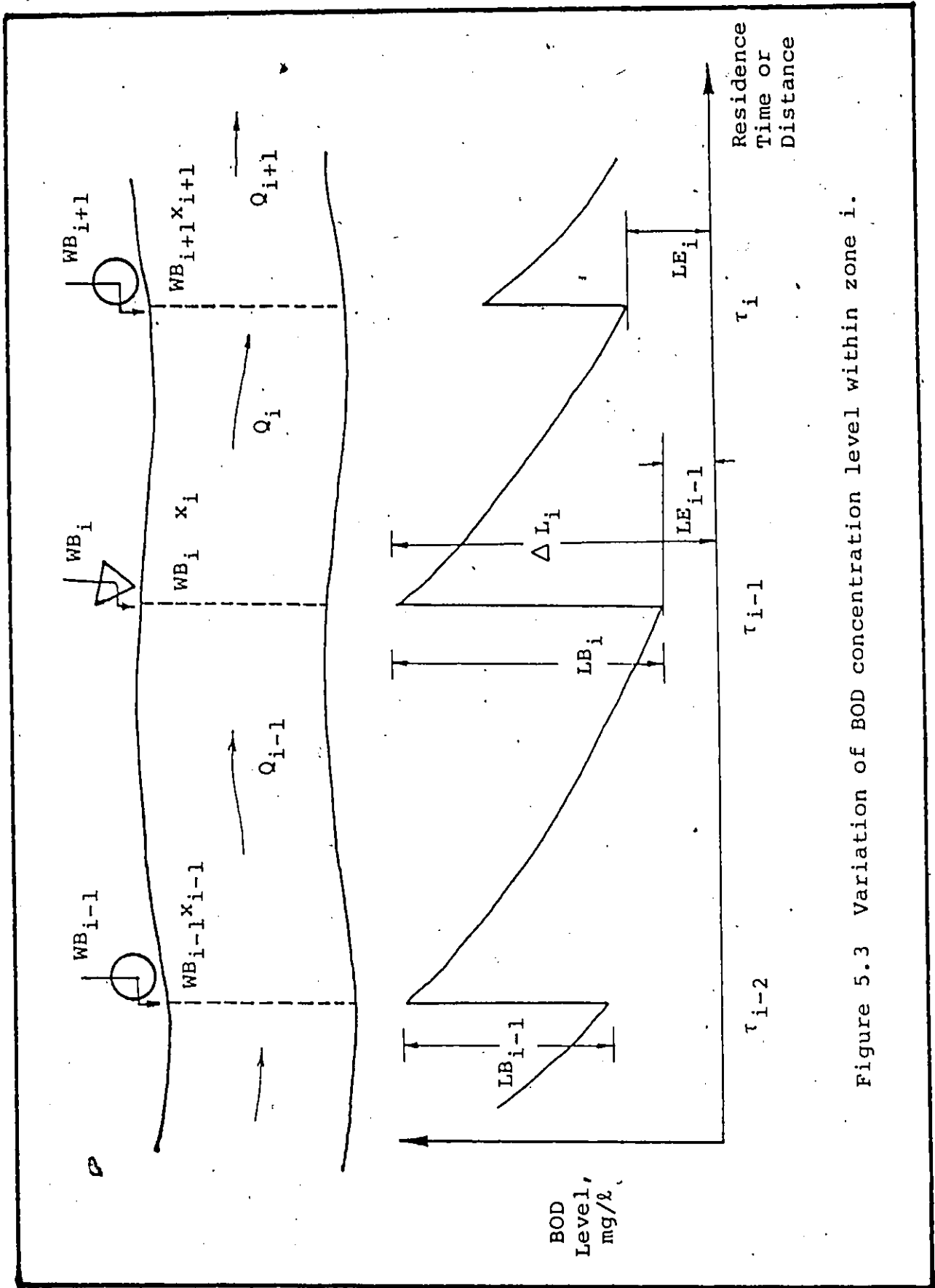


Figure 5.3 Variation of BOD concentration level within zone i .

As is shown in Figure 5.3, the maximum BOD content, LB_i , occurs at the discharge point. Therefore, we have:

$$LB_i = LE_{i-1} + \Delta L_i \quad (5.11)$$

where:

LB_i = stream BOD content at the beginning of zone i , mg/l

LE_{i-1} = stream BOD content at the end of zone $i-1$, mg/l

For determining LE_{i-1} , note that the BOD concentration at the end of a zone is equal to the BOD content at the beginning of that zone plus the variation of the BOD along that zone. The variation of BOD can be calculated according to Equation (5.1). LE_i can be determined as follows:

$$LE_i = LB_i + \int_{\tau_{i-1}}^{\tau_i} [-(K_{1,i} + K_{3,i}) * LB_i + L_{a,i}] dt ;$$

$$i = 1, 2, \dots, N \quad (5.12)$$

For the first zone, $i = 1$, $LE_0 = LU$ and thus:

$$LB_1 = LU + \Delta L_1 \quad (5.13)$$

where:

LU = upstream concentration of BOD, mg/l

(ii) Minimum Concentration of DO

Assuming that MDO_i is the minimum DO concentration level at zone i , the constraint states that:

$$MDO_i > MINDO_i, \quad i = 1, 2, \dots, N \quad (5.14)$$

where $MINDO_i$ is the minimum level of DO allowed in zone i .

The minimum concentration of DO within the zone can occur anywhere within that zone, and not necessarily at the boundaries, Figure 5.4. Therefore a search process is required to locate the point corresponding to the minimum DO at each zone.

Assuming C_i as the DO concentration at the end of zone i , this is equal to the DO concentration at the end of the previous zone, C_{i-1} , plus the variations of the DO content (according to Equation(5.2)). This can be expressed as follows:

$$C_i = C_{i-1} + \int_{\tau_{i-1}}^{\tau_i} \{-K_{1,i} LB_i + K_{2,i}(C_{s,i} - C_{i-1}(\tau)) + \langle P - R \rangle_i\} d\tau \quad (5.15)$$

To determine the distribution of DO concentration in the i^{th} zone, the above expression is evaluated through numerical integration. The minimum DO concentration and its location within the zone then are determined during the course of integration, (See Appendix A). This constraint can now be formulated as:

$$MDO = \min\{C_{i-1} + \int_{\tau_{i-1}}^{\tau_i} -K_{1,i} LB_i + K_{2,i}(C_{s,i} - C_{i-1}(\tau)) + \langle P - R \rangle_i\} d\tau \quad MINDO_i, \quad i=1,2,\dots,N \quad (5.16)$$

(iii) Maximum Concentration of Phosphorus

Assuming PH_i is the maximum phosphorus content at

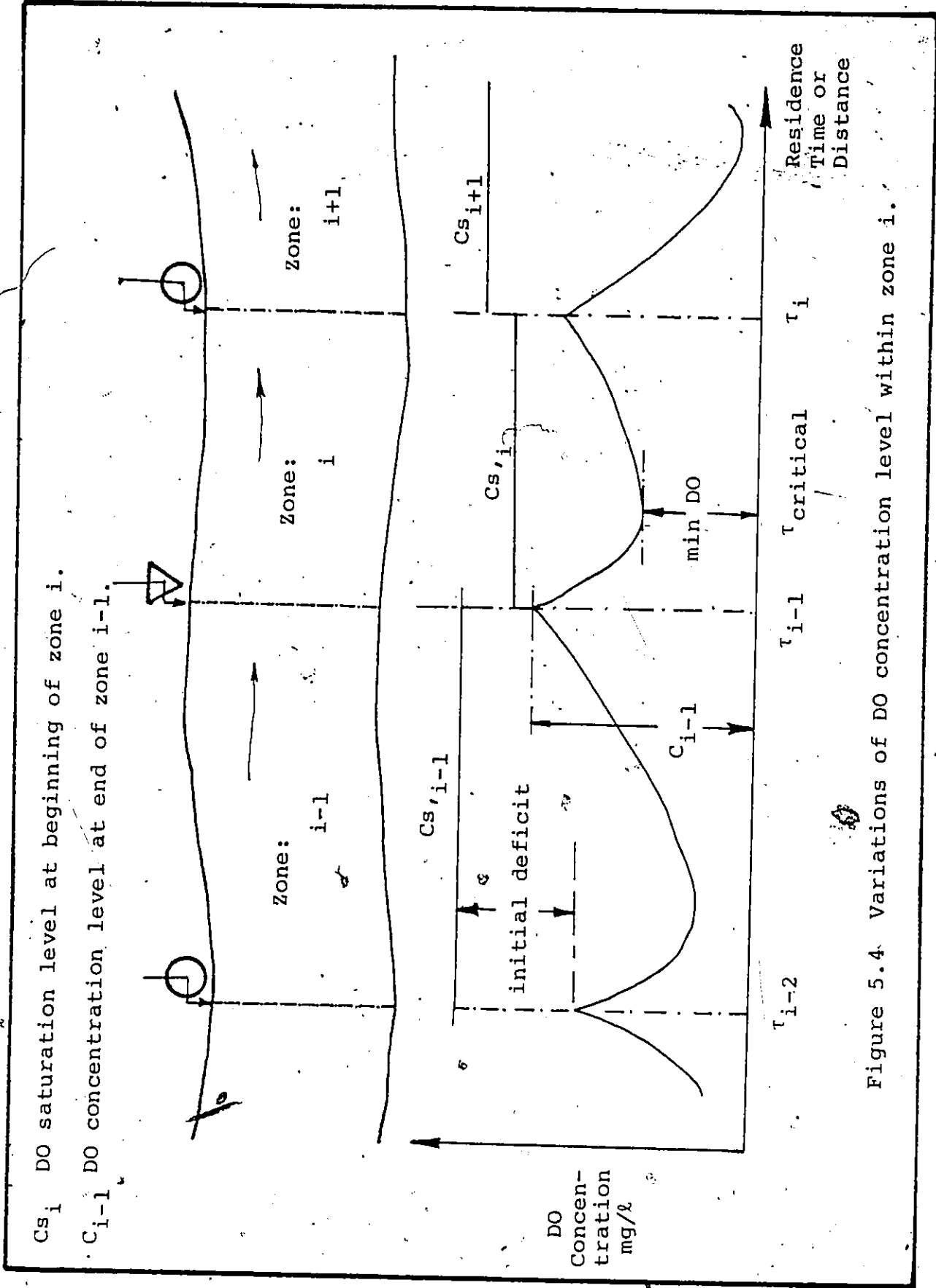


Figure 5.4 Variations of DO concentration level within zone i.

zone i due to waste discharge, this can be represented as:

$$PH_i \leq MAXPHR_i, \quad i = 1, 2, \dots, N \quad (5.17)$$

where $MAXPHR_i$ is the maximum allowable level of phosphorus in zone i .

The maximum concentration occurs at the point of discharge, and remains at a constant level throughout the zone as long as there is no change in flow rate, since the variation in phosphorus concentration level is due to dilution processes. This phenomenon is shown in Figure 5.5.

The maximum concentration level at the discharge point can be calculated as follows: Assume the produced amount of phosphorus waste load is WP_i in lbs/day. If z_i percent is the treatment level to reduce the phosphorus pollutant level in the influents, then the phosphorus concentration, EP_i , in the effluents, measured in mg/l, can be given as:

$$EP_i = \left(\frac{1.0 - z_i}{1.0} \right) * \frac{Fl * WP_i}{Q_{e_i}} \quad (5.18)$$

where:

Q_{e_i} = the effluent rate, cfs

Fl = the conversion factor, $\frac{\text{day} \cdot \text{ft}^3 \cdot \text{mg}}{\text{sec} \cdot \text{lb} \cdot \text{l}}$

If it is assumed that the flow of the stream at the end of zone $i-1$ has a phosphorus concentration level PH_{i-1} , in mg/l, due to the accumulated waste effluent discharged at previous zones, then the total phosphorus concentration

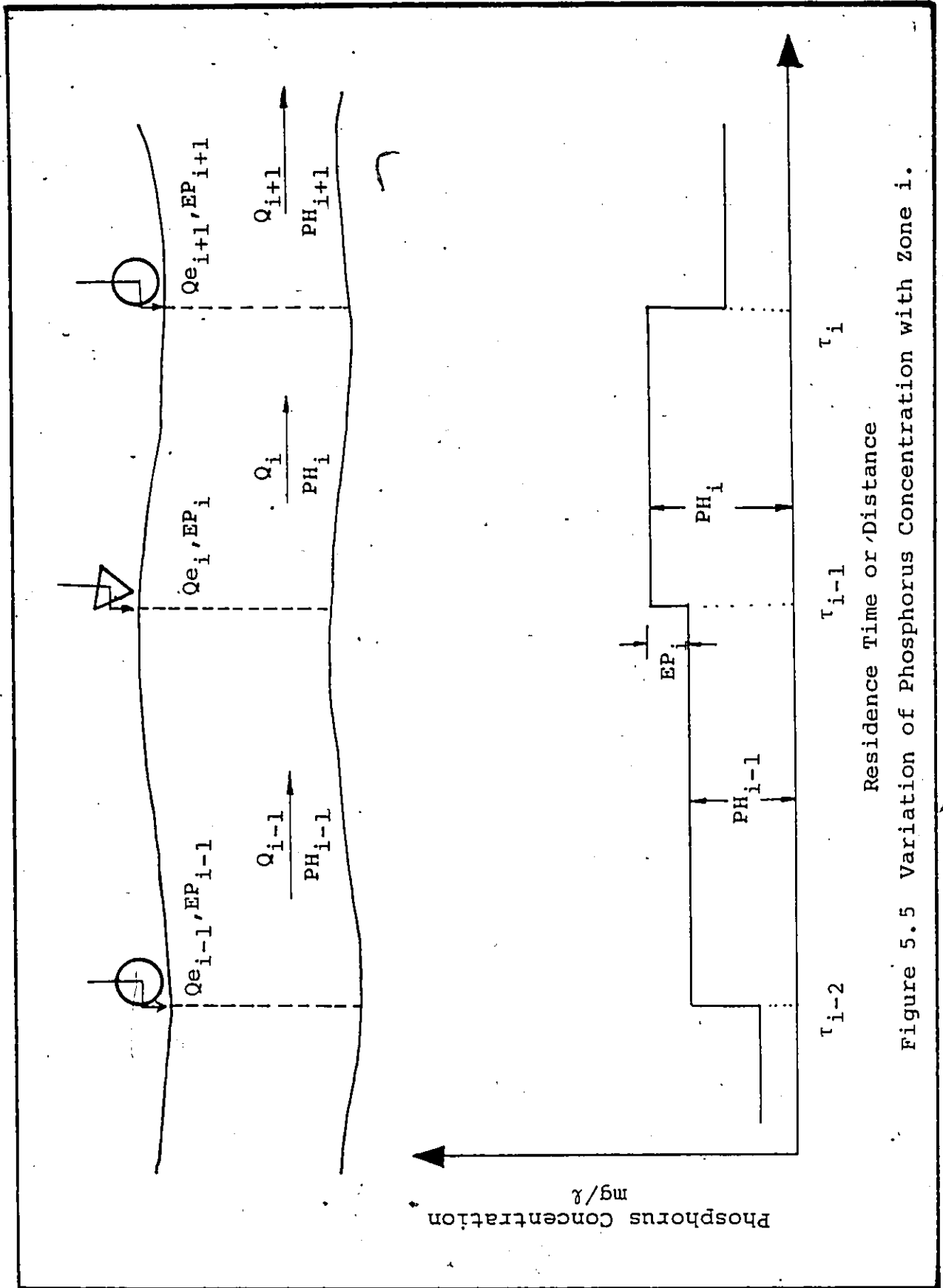


Figure 5.5 Variation of Phosphorus Concentration with Zone i.

PH_i , in mg/l, at the beginning of zone i would be:

$$PH_i = \frac{EP_i \cdot Qe_i + Q_{i-1} \cdot PH_{i-1}}{Qe_i + Q_{i-1}}, \quad i = 1, 2, \dots, N \quad (5.19)$$

where for the first zone:

$PH_0 = PU$, the upstream phosphorus concentration in mg/l.

$Q_0 = QU$, the upstream flow rate, cfs.

2) Structural Constraints

(i) The parameters θ_1 , and θ_2 of the effluent charge rate (Equations (5.7.1) to (5.7.3)) are non-negative, thus:

$$\theta_{1,i} > 0$$

$$\theta_{2,i} > 0;$$

$$i = 1, 2, \dots, N \quad (5.20)$$

(ii) The treatment levels of BOD and phosphorus, i.e., x_i , and z_i respectively, take values between zero and 100 percent, thus:

$$0 < x_i < 1.$$

$$0 < z_i < 1. ;$$

$$i = 1, 2, \dots, N \quad (5.21)$$

The model can now be comprehensively formulated to consider simultaneously municipal and/or industrial pollution sources at each zone as follows:

$$\text{Minimize } \sum_{\text{zone (i)}} \sum_{\text{source (j)}} \left[FT_{ij}(WB_{ij}, WP_{ij}, x_{ij}, z_{ij}, \bar{Q}_{ij}) \right. \\ \left. + FC_{ij}(WB_{ij}, y_{ij}, x_{ij}) \right]$$

with respect to x_{ij}, z_{ij} , and y_{ij} ; $i=1,2,\dots,N, j=1,2$
such that:

(1) To account for maximum level of BOD:

$$LB_1 = LU + \frac{F1}{Q_1} \sum_{j=1}^2 WB_{1,j} \frac{(1.0-x_{1,j})}{100} \ll \text{MAXBOD}_1$$

$$LB_2 = LE_1 + \frac{F1}{Q_2} \sum_{j=1}^2 WB_{2,j} \frac{(1.0-x_{2,j})}{1.0} \ll \text{MAXBOD}_2$$

$$LB_N = LE_{N-1} + \frac{F1}{Q_N} \sum_{j=1}^2 WB_{N,j} \frac{(1.0-x_{N,j})}{1.0} \ll \text{MAXBOD}_N$$

(2) To account for minimum DO level:

$$MDO_1 = \min \left\{ CU + \int_{\tau_0}^{\tau_1} \left[-K_{1,1} \cdot LB_1 + K_{2,1} (C_{s_1} - C_0(\tau)) \right. \right. \\ \left. \left. + \langle P - R \rangle_1 \right] d\tau \right\} \gg \text{MINDO}_1$$

$$MDO_2 = \min \left\{ C_1 + \int_{\tau_1}^{\tau_2} \left[-K_{1,2} \cdot LB_2 + K_{2,1} (C_{s_2} - C_1(\tau)) \right. \right. \\ \left. \left. + \langle P - R \rangle_2 \right] d\tau \right\} \gg \text{MINDO}_2$$

$$MDO_N = \min \left\{ C_{N-1} + \int_{\tau_{N-1}}^{\tau_N} \left[-K_{1,N} \cdot LB_N + K_{2,1} (C_{s_N} - C_{N-1}(\tau)) \right. \right. \\ \left. \left. + \langle P - R \rangle_N \right] d\tau \right\} \gg \text{MINDO}_N$$

(3) To account for maximum level of phosphorus:

$$PH_1 = \frac{EP_{1,1} \cdot Qe_{1,1} + EP_{1,2} \cdot Qe_{1,2} + PU \cdot QU}{Qe_{1,1} + Qe_{1,2} + QU} \leqslant MAXPHR_1$$

$$PH_2 = \frac{EP_{2,1} \cdot Qe_{2,1} + EP_{2,2} \cdot Qe_{2,2} + PH_2 \cdot Q_1}{Qe_{1,1} + Qe_{1,2} + Q_1} \leqslant MAXPHR_2$$

$$PH_N = \frac{EP_{N,1} \cdot Qe_{N,1} + EP_{N,2} \cdot Qe_{N,2} + PH_{N-1} \cdot Q_{N-1}}{Qe_{N,2} + Qe_{N,2} + Q_{N-1}} \leqslant MAXPHR_N$$

(4) Constraints on parameters of effluent charge rate equations:

$$\theta_{1,1,i} ; \theta_{1,2,i} \geqslant 0$$

$$\theta_{2,1,i} ; \theta_{2,2,i} \geqslant 0; \quad i = 1, 2, \dots, N$$

(5) Boundary conditions on treatment levels:

$$1 \gg x_{i,1} \gg 0$$

$$1 \gg x_{i,2} \gg 0 ; \quad i = 1, 2, \dots, N$$

$$1 \gg z_{i,1} \gg 0$$

$$1 \gg z_{i,2} \gg 0 ; \quad i = 1, 2, \dots, N$$

5.4 Solution Approach

The objective function and the constraints of the model are highly nonlinear; therefore, a nonlinear optimization technique, is required to solve the problem. In this study, the sequential unconstrained Minimization Technique (SUMT) of Fiacco and McCormick (1968) which is one of the simplest and most efficient nonlinear optimization techniques, is employed. The idea of the technique is to use the original objective function and the constraints to form a sequence of modified unconstrained objective functions which could be minimized by using any appropriate unconstrained multivariate optimization technique. In the context of the model developed in Section 5.3, the problem is first stated as follows:

Minimize:

$$F = \sum_{i=1}^N \sum_{j=1}^2 \left[FT_{ij}(WB_{ij}, WP_{ij}, x_{ij}, z_{ij}, Q_{ij}) + FC_{ij}(WB_{ij}, Y_{ij}, x_{ij}) \right]$$

Subject to:

$$G_i = LE_i + \frac{F_1}{Q_i} \sum_{j=1}^2 \left(WB_{ij} \frac{(1.0 - x_{ij})}{1.0} \right) + \text{MAXBOD}_i \geq 0;$$

$$i = 1, 2, \dots, N$$

$$G_k = \min \left\{ C_{i-1} + \int_{\tau_{i-1}}^{\tau_i} -K_{1,i} \cdot LB_i + K_{2,i} (C_{s_i} - C_{i-1}(\tau)) + \langle P - R \rangle_i \right\} d\tau - \text{MINDO}_i \geq 0;$$

$$k = N+1, \dots, 2N$$

$$G_l = \frac{EP_{i,1} \cdot Q_{e_{r,1}} + EP_{i,2} \cdot Q_{e_{i,2}} + PH_{i-1} \cdot Q_{i-1}}{Q_{e_{i,1}} + Q_{e_{i,2}} + Q_{i-1}}$$

$$+ \text{MAXPH}_i \geq 0; \quad l = 2N+1, \dots, 3N$$

$$G_r = x_{ij} \geq 0; \quad r = 3N+1, \dots, 5N$$

$$G_s = 1 - x_{ij} \geq 0; \quad s = 5N+1, \dots, 7N$$

$$G_t = z_{ij} \geq 0; \quad t = 7N+1, \dots, 9N$$

$$G_u = 1 - z_{ij} \geq 0; \quad u = 9N+1, \dots, 11N$$

$$G_v = \theta_{1,i j} \geq 0; \quad v = 11N+1, \dots, 13N$$

$$G_q = \theta_{2,i j} \geq 0; \quad q = 13N+1, \dots, 15N$$

where $j=1,2$ indicates number of pollution source at each zone, and $i=1,2,\dots,N$, where N is the number of zones in the river.

The modified objective function, P , is next formulated in terms of the original objective function, F , and a penalty term. This is expressed as follows:

$$P = F + r \sum_{t=1}^{15N} \frac{1}{G_t} \quad (5.22)$$

where r is a positive real number which is re-evaluated to form a monotonically decreasing sequence $r_1 > r_2 > \dots > 0$. The penalty term is designed so as to insure that the solution is always located in the interior of the constrained region by avoiding to cross the boundaries of the

feasible region.

The study has employed the Hooke and Jeeves technique (1961) to solve the modified unconstrained problem. The technique is based on direct search and requires no derivatives. Since the technique assumes a unimodal function, several sets of starting values had to be tried to find the global minimum. For a detailed description of SUMT, and its computational procedures, the interested reader is referred to Appendix B at the end of this report.

CHAPTER VI

APPLICATION OF THE OPTIMIZATION MODEL

6.1 Introduction

The validity of the developed model has been verified by applying it to a hypothetical river basin in two different cases. In the first case (Example 1), three municipal waste treatment plants are considered, which discharge different waste loads into the river. Accordingly, the river is divided into three zones. It is assumed that three plants have the same treatment cost function.

In the second case (Example 2), five pollution sources (two municipal sources, and three industrial sources) are assumed to discharge different amounts of waste in the river which is divided into three zones. The treatment cost function of the industrial source is different from that of the municipal source.

An empirical treatment cost function for municipal sources has been derived and employed in the study. The function is based on the data collected, for the period 1975 - 1977, on municipal wastewater treatment plants in Southern Ontario. The cost function employed for industrial sources is based on the current information for the treatment cost of particular chemical industrial wastes, adopted from literature (e.g., Cost Engineering Journal, 1971).

This chapter is divided into two main parts; in the first part, the treatment cost functions for both municipal and industrial sources are presented. In the case of

municipal waste treatment cost function, the data are examined and the development of the cost model is discussed. In the case of industrial waste treatment cost function, the cost models for primary, secondary and tertiary treatment are given.

The second part of this chapter is devoted to a discussion of the input data used in each example (i.e., the river characteristics, types and amounts of waste, water quality standards in each zone, etc.), followed by a detailed presentation of the optimization models for both examples.

It is assumed that municipal sources produce organic and phosphorus waste loads, while industrial sources produce organic loads only. This assumption is based on the fact that most of the phosphorus wastes discharged into rivers originate from domestic activities, (Donald , and Bishop, 1975).

The water quality limits in the first example are held constant throughout the river, whereas they are set differently at each zone in the second example. The water quality limits, in both examples, have been chosen to fall within the acceptable ranges of water quality for public service supplies, (e.g., Guides and Criteria for Water Quality Management in Ontario, Ministry of Environment, June, 1973).

6.2 Development of Waste Treatment Cost Functions

The costs of waste water treatment plants may generally be divided into two main categories: i) capital costs, and ii) operating and maintenance costs (O & M). The capital costs include the cost of land purchases, equipment or

structure, installation, and engineering contracting fees. Such costs are generally paid in installments over the useful life of the plant or over an accelerated depreciation period. This cost is usually represented as a fixed part of the cost accounts, (Shah and Reid, 1971). The O & M costs, on the other hand, are the expenses incurred in the day-to-day operation and upkeep of the treatment facilities. This part of the costs is not fixed, but varies with the amount of waste flow, percentage of the treatment required, characteristics of the waste, and other factors. The O & M costs can be divided into: (i) direct costs which include the costs of maintenance, plant chemical supplies, labour and supervision, and utilities; (ii) indirect costs which include the costs of depreciation, insurance, and other general overhead.

Tihansky (1974) indicated that most of the derived functions for operating and maintenance costs pertain only to direct costs, since indirect costs, in most cases, do not constitute a large share of the total operating and maintenance costs. The treatment cost functions, related to municipal and industrial wastes, employed in the present study, are concerned only with the direct portion of operating and maintenance costs.

6.2.1 Cost Function for Municipal Waste Treatment Plants

The derived function of operating and maintenance costs for municipal wastewater treatment plants are based on data collected for 33 municipal wastewater treatment plants in Southern Ontario for the period of 1975-1977. The data

include, for each plant, two sets of information. The cost data, such as total salaries, wages, employee benefits, repair and maintenance costs, chemicals, utilities, and other miscellaneous expenses. The details of cost information for each plant are shown in Appendix C1. The technical information regarding the hydraulic capacity, the produced amount of organic and phosphorus waste loads, and the performance of each plant is shown in Appendix C2.

Analysis of Data The collected data presented some diversity regarding the percentages of BOD reduction. The percentages were generally higher than normally expected, particularly at primary and secondary levels. In the course of examining the data, it was found that the estimation for the reduction percentages of BOD are based on calculating the average differences of BOD in samples withdrawn, randomly, from the plant's influent and effluent respectively. Improper sampling methods, or bias regarding the method and/or time of taking the observations might be the reason for such diversity. It was decided, however, to omit such extreme values from the analysis.

Since the data are extended over a four-year period it is important to keep the observations unaffected from the influence of factors extraneous to the cost relationship itself, (Johnston, 1960). For example, observations of the cost values should not be influenced by the variations, from year-to-year, in the prices paid for production factors such

as labour and raw materials. A price index is used to diminish such influences. There are many up-to-date price indices available for most production factors of many typical industries (Statistics of Canada, Cat.#62). It is difficult, however, to find an aggregated measure to represent all the variations in prices paid for the factors constituting the operating and maintenance costs of wastewater treatment plants. Therefore, a price index has been established based on the collected cost data using the year 1977 as the base year. Only those plants that did not undergo any changes in design capacity or in the method of treatment during the four-year period are considered in the price indices calculation. This is necessary to insure that the variations in costs are only due to variations in prices paid for the operating and maintenance factors. The details of the derivation of the price indices are shown in Appendix C3.

Cost Model The most important factors used to establish operating and maintenance cost functions for municipal wastewater treatment plants have been reported in many studies. (Shah 1970, Michel 1970, and Marsden, et al, 1973). According to these studies, three independent variables are considered in the derivation of the present models: (i) the hydraulic capacity of the plant, \bar{Q} , measured in terms of the average daily flow of influents in millions of gallons per day (MGD). (ii) the amount of reduction in BOD, which is a function of the percentage treatment level of BOD, X , and the organic waste load, WB , measured in lbs/day; (iii) the amount

of reduction in phosphorus load which, in turn, is a function of the percentage treatment level, Z , and the phosphorus waste load, WP , measured in lb /day.

The three independent variables were plotted against the annual operating and maintenance cost, the dependent variable, in order to determine the type of the relationship between the independent and dependent variables. Identifying these relationships helps in proposing the form of the cost model. Figures 6.1, 6.2 and 6.3 depict these relationships; in each case an exponential relationship may be inferred.

Marsden, et.al (1973) suggested three functional forms that may be used in characterizing cost models:

- (i) General Exponential (GE): This form is expressed as follows:

$$c = f_1(x_1, x_2, \dots, x_n) = e^{\beta_0} \prod_{j=1}^k x_j^{\beta_j}; x_j > 0 \quad (6.1)$$

- (ii) Generalized Quadratic (GQ): This form is expressed as follows:

$$c = f_2(x_1, x_2, \dots, x_n) = A \left(\prod_{j=1}^k x_j^2 \right) \sum_{j=1}^k \frac{\beta_j}{x_j} / \sum_{j=1}^k x_j^4 \quad (6.2)$$

$$x_i > 0, A > 0$$

- (iii) Constant Cost Elasticity of Substitution (CCES): This form is expressed as follows:

$$c = f_3(x_1, x_2, \dots, x_n) = A \left[\sum_{i=1}^n \alpha_i x_i^{-B} \right]^{-1/B} \quad (6.3)$$

$$x_i > 0, A > 0, B > -1$$

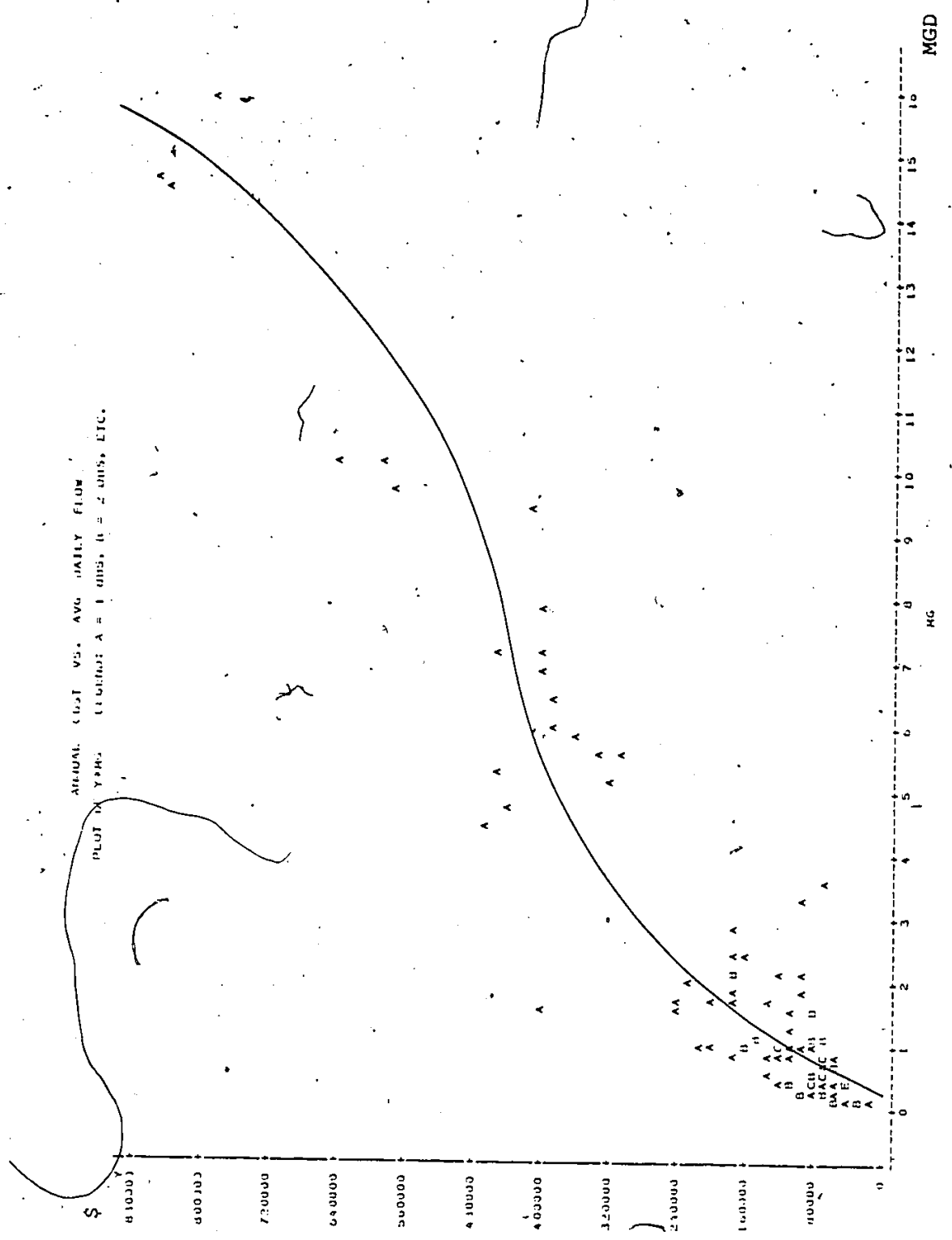


Figure 6.1

ANNUAL COST VS. BASIC REMOVAL
PLANT CAPACITY - 1 GPD, 10 - 2 GPD, ETC.

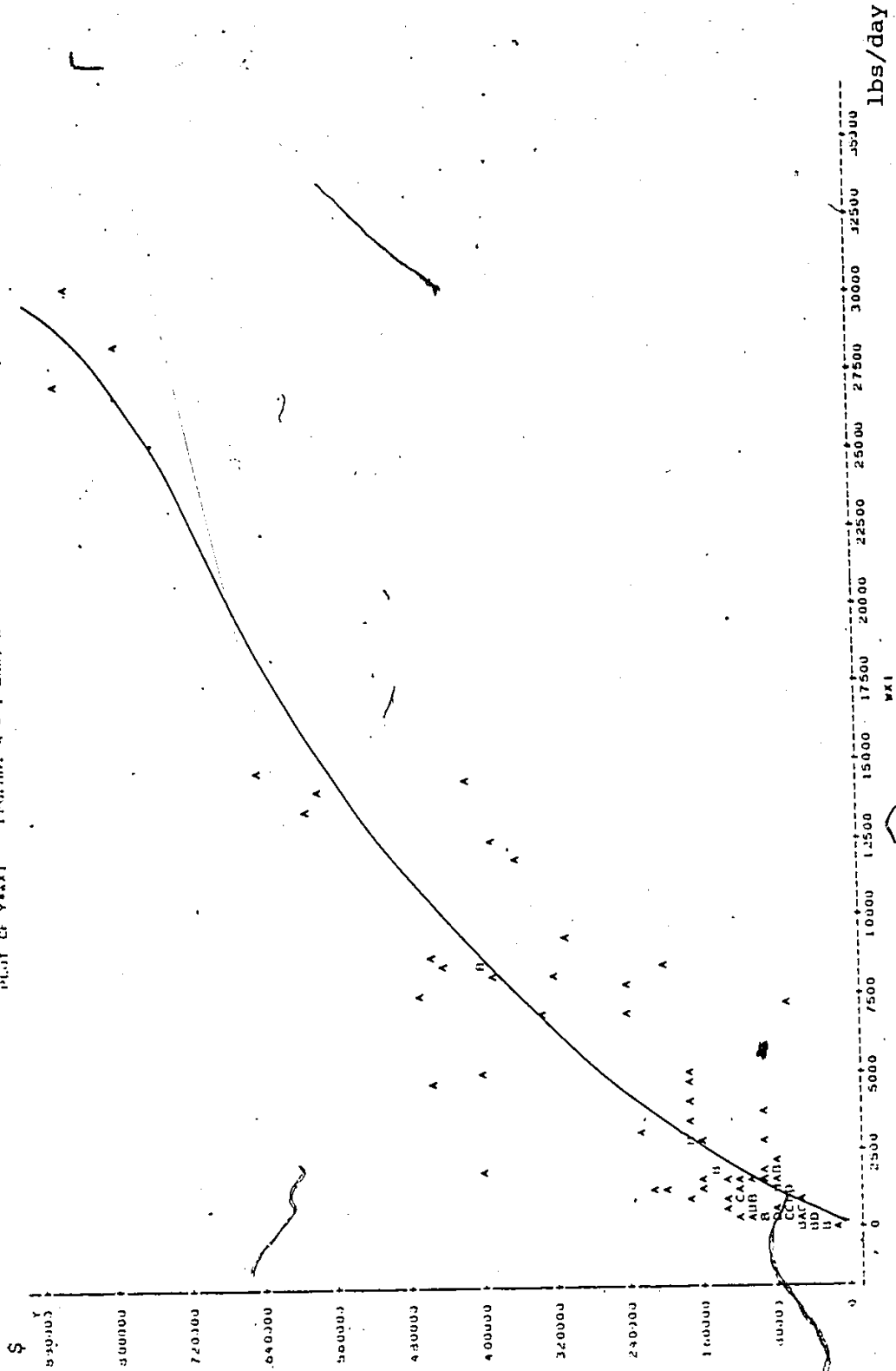


Figure 6.2

ANNUAL COST VS. PRODU. BASIC MODEL
PLOT OF YEARLY COST VS. PRODUCTION

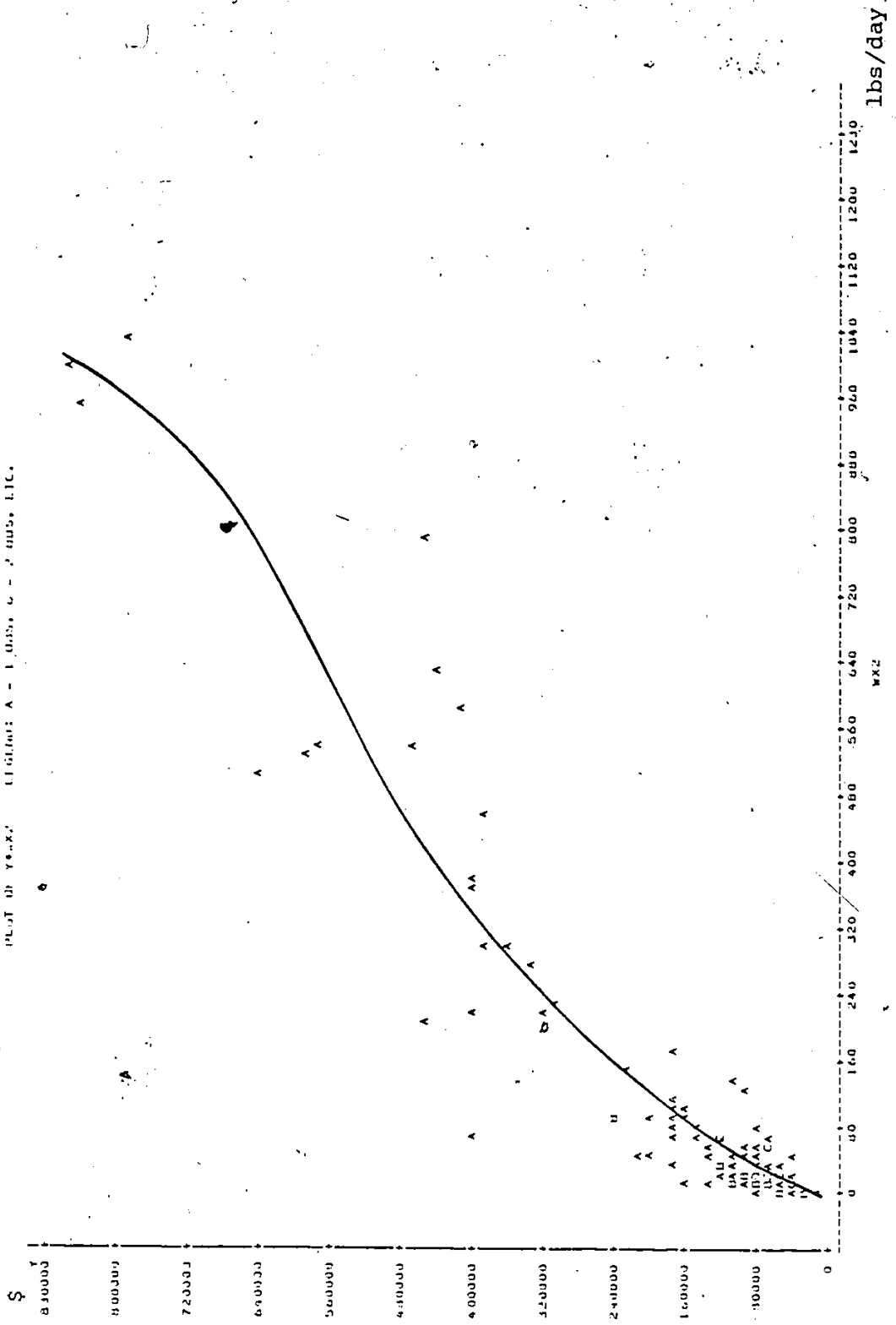


Figure 6.3

and, $\sum_{i=1}^n \alpha_i = 1, \alpha_i > 0 (i=1, \dots, n)$

where,

c = a cost measure;

x_i = cost factor measures;

e = exponential constant;

α_i, B, A, β_j = parameters of the various functions.

The General Exponential form is easier than the other two forms to deal with for estimation purposes. It is easily transformed into a log-linear form to estimate the parameters involved. Direct, ordinary least-squares method can be applied in this case. Functions in the other two forms may be estimated using nonlinear least-square techniques, or one of several step-wise procedures. These techniques require, however, the evaluation of the derivatives of the dependent variable with respect to function's parameters, which, most of the time, are difficult to obtain.

The General Exponential form has been chosen in the present study to represent the cost function. The function contains the three independent variables given above, and takes the following formulation:

$$c = e^{10.1125} \cdot Q^{0.3495} \cdot (WB \cdot X)^{0.1778} \cdot (WP \cdot Z)^{0.0685} \quad (6.4)$$

The computer program, results, and statistical analysis for the derivation of the cost model are exhibited in Appendix D.

6.2.2 Cost Function for a Typical Chemical Industrial Plant

At present, the available literature on industrial waste treatment costs is ~~not complete~~ because of the diversity of industries in general, and of technological abatement methods, in particular. The Cost Engineering Journal is the main publication devoted exclusively to present cost estimations and cost functions for industrial wastes. The function presented here pertains to chemical industrial plants as reported by Eckenfelder and Barnard (1971). It represents the biological treatment costs at the primary, secondary and tertiary stages as follows:

Primary Treatment Cost Function

$$FT = 909 + 2273 (1.1/\text{MGD})^{0.5} \quad (6.5)$$

Secondary Treatment Cost Function

$$FT = 2700 + 2500/(\text{MGD})^{0.67} \\ + \text{UNITS} (0.02 - 0.0001 * \text{MGD}) \quad (6.6)$$

Tertiary Treatment Cost Function

$$FT = 1500 + 6450 (1/\text{MGD})^{0.63} \quad (6.7)$$

where,

FT = Annual operating and maintenance costs in \$/MGD.

MGD = Average flow rate of influents in millions of gallons per day

UNITS = Amount of power consumption in kw.hr/day.

The three cost relationships are functions of the average daily flow of the plant. In addition, the secondary treatment cost function is formulated in terms of the amount of power consumption in kw.hr. The cost functions for the three stages, however, are not formulated so as to include other relevant factors which reflect the strength of the waste. Nevertheless, these functions are satisfactory for the purpose of demonstrating how the model can handle situations involving wastes generated at industrial and municipal sources.

6.3 Model Description for Example 1

6.3.1 Input Data Description

The river basin, in this example, is assumed to receive both organic and phosphorus wastes from three municipalities located along the river. The river basin is thus divided into three zones, as depicted in Figure 6.4. The hydraulic characteristics of the river and parameter values used in the transfer equations (5.1) and (5.2) are given in Table 6.1. It is assumed that the water at the upstream portion is relatively clean. The upstream concentration levels of DO, BOD, and phosphorus are 6, 5, and 0.05 mg/l respectively. The local BOD flow rate as well as the stream temperature are assumed constant throughout the segment of the river under consideration. The stream flow rate is 500 cfs, while the stream temperature is 290°K. The residence time at the first, second, and third zone are 1, 1.2, and 1.5 days, respec-

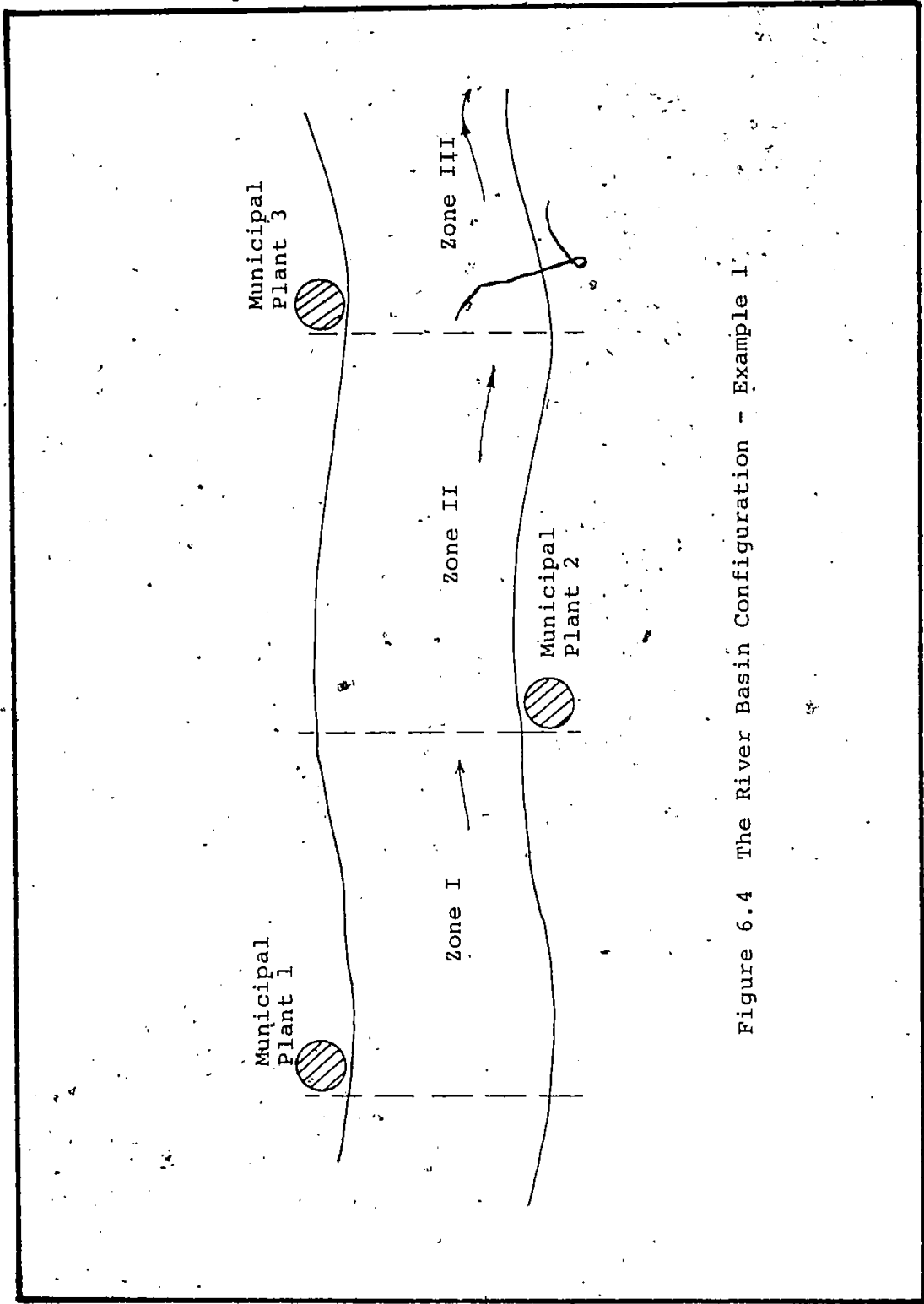


Figure 6.4 The River Basin Configuration - Example 1

tively.

The operating conditions of the waste treatment plants in terms of the average daily flow, and the organic and phosphorus waste loads are shown in Table 6.2. The plant at zone 1 has an average daily flow of 4 MGD, an organic waste load of about 0.2×10^5 lb /day, a phosphorus waste load of 0.8×10^3 lb /day. The second plant has an average daily flow of 5 MGD, an organic waste load of 0.22×10^5 lb /day, a phosphorus waste load of 0.9×10^3 lb /day. The corresponding figures for the third plant are 5 MGD, 0.1×10^5 lb /day, and 0.7×10^3 lb/day.

The water quality standards are given in Table 6.3. The maximum allowable BOD concentration in all three zones is 15 mg/l. The minimum allowable DO concentration in all three zones, is 5 mg/l, and the maximum allowable concentration of phosphorus in all the three zones is 0.1 mg/l.

6.3.2 Model Description

In this section, the SUMT formulation of Example 1 is described. The objective function is to minimize:

$$f(XB_1, XB_2, XB_3; XP_1, XP_2, XP_3; \theta_{11}, \theta_{21}; \theta_{12}, \theta_{22}; \theta_{13}, \theta_{23}) =$$

$$\sum_{\text{source } i=1}^3 \left[e^{10.1125 \cdot \bar{Q}_i} \cdot 0.3495 \cdot (WB_i \cdot X_i)^{0.1778} \cdot (WP_i \cdot Z_i)^{0.0685} \right]$$

$$+ \sum_{i=1}^3 \left[(1 - X_i) \cdot WB_i \cdot \theta_{1i} \cdot e^{-\theta_{2i}/(1 - X_i)} \right] \quad (6.8)$$

Subject to the following constraints:

(a) BOD Constraints:

$$g_i = -LB_i + 15.0 \geq 0; \quad i=1, 2, 3$$

Table 6.1

Characteristics of the River and the Parameter Values of the Model for Example 1

Parameter	Description	Numerical Value
LU	Upstream BOD Concentration	5.0 mg/%
CU	Upstream DO Concentration	6.0 mg/%
PU	Upstream Phosphorus Concentration	0.05 mg/%
La	Local BOD Runoff Rate	0.20 mg/%
Q	Stream Flow Rate	500 CFS for Zone 1 500 CFS for Zone 2 500 CFS for Zone 3
α	The Constant in Photosynthesis Respiration Expression	3.46
τ	Residence Time	1 day for Zone 1 1.2 day for Zone 2 1.5 day for Zone 3
T	Stream Temperature	290° K for Zone 1 290° K for Zone 2 290° K for Zone 3
K ₃	BOD Adsorption and Sedimentation Coefficient	0.1 day ⁻¹

Table 6.2

Operating Conditions of Wastewater Treatment Plants - Example 1

Plant #	Average Daily Flow MGD	Waste Load in lbs/day	
		Organic Waste	Phosphorus Waste
1	4	0.20×10^5	0.8×10^3
2	5	0.22×10^5	0.7×10^3
3	2.5	0.10×10^5	0.6×10^3

Table 6.3

Water Quality Standards - Example 1

Zone	The Desired Water Quality Limits mg/l		
	BOD	DO	Phosphorus
1	15.0	5.0	0.1
2	15.0	5.0	0.1
3	15.0	5.0	0.1

where LB_i is given by Equation (5.11).

(b) DO Constraints:

$$g_k = MDO_i - 5.0 \geq 0 ; k=4,5,6$$

where MDO_i is given by Equation (5.16).

(c) Phosphorus Constraints:

$$g_l = -PH_i + 0.1 \geq 0 ; l=7,8,9$$

where HP_i is given by Equation (5.19).

(d) Boundary Conditions on X_i :

$$(1) \quad g_r = X_i \geq 0 ; r=10,11,12$$

$$(2) \quad g_s = 1 - X_i \geq 0 ; s=13,14,15$$

(e) Boundary Conditions on Z_i :

$$(1) \quad g_t = Z_i \geq 0 ; t=16,17,18$$

$$(2) \quad g_u = 1 - Z_i \geq 0 ; u=19,20,21$$

(f) Non-Negativity Conditions on θ_{1j} and θ_{2j} :

$$g_v = \theta_{1i} \geq 0 ; v=22,23,24$$

$$g_q = \theta_{2i} \geq 0 ; q=25,26,27$$

and $i = 1,2,3$

The above optimization model consists of twelve decision variables and twenty-seven constraints. The decision variables represent the BOD treatment levels at the three plants (X_1, X_2, X_3), the phosphorus treatment levels (Z_1, Z_2, Z_3), and the coefficients of the effluent charge

formula $(\theta_{11}, \theta_{21}; \theta_{12}, \theta_{22}; \theta_{13}, \theta_{23})$. The constraints represent, for each zone, the maximum allowable BOD concentration (3 constraints), the minimum DO concentration (3 constraints), and the maximum limits for phosphorus concentration (3 constraints). The next twelve constraints account for the boundary conditions on the variables representing the organic and phosphorus waste treatment levels, and the remaining six constraints account for the non-negativity of the effluent charge rate coefficients.

The problem has been solved on an IBM 370/3031 computer in 3.31 minutes of CPU time, after 13 iterations of the SUMT procedure. The last iteration represents the final optimum obtained at the required accuracy limit of 10^{-6} , see Appendix . The optimal policy corresponds to a total annual cost of \$1,223,334, which includes the effluent charges as well as the treatment costs for all polluters in the basin. The optimal effluent charges are 7.17, 5.37 and 3.18 ¢/lb of BOD discharged in the first, second and third zone, respectively.

The optimal treatment levels of organic waste in the three zones are 58.07%, 68.0% and 78.0%, respectively. The corresponding levels of phosphorus waste treatment in the three zones are 88.9%, 98.9%, and 91%.

6.4 Model Description for Example 2

6.4.1 Input Data Description

This example considers a river basin which receives wastes from both municipal and industrial sources located along the river. It is assumed that the industrial sources generate only organic wastes, while the municipal sources generate not only organic wastes, but also a considerable amount of phosphorus waste, originating mainly from residential areas. Two industrial, and three municipal waste sources are considered, which according to their relative locations, necessitate the division of the river into three zones. Figure 6.5 shows the geographical locations of the plants along the river. The first zone receives wastes from a municipal and an industrial source located across the river from one another. The second zone receives waste from only one municipal source, whereas, the third zone receives wastes from an industrial and a municipal source located across the river from one another.

The hydraulic characteristics of the river, as well as the parameters values used in the transfer equations are given in Table 6.4. The upstream water quality levels are assumed to be 5 mg/l of BOD, 6 mg/l of DO, and 0.05 mg/l of phosphorus. The stream flow rate is 1500 cfs throughout the segment of the river under consideration. The stream temperature, however, has been assumed to change from zone to zone; it is 290°K in zone 1, 295°K in zone 2, and 300°K in zone 3. The residence times are assumed to be 2.5, 2,

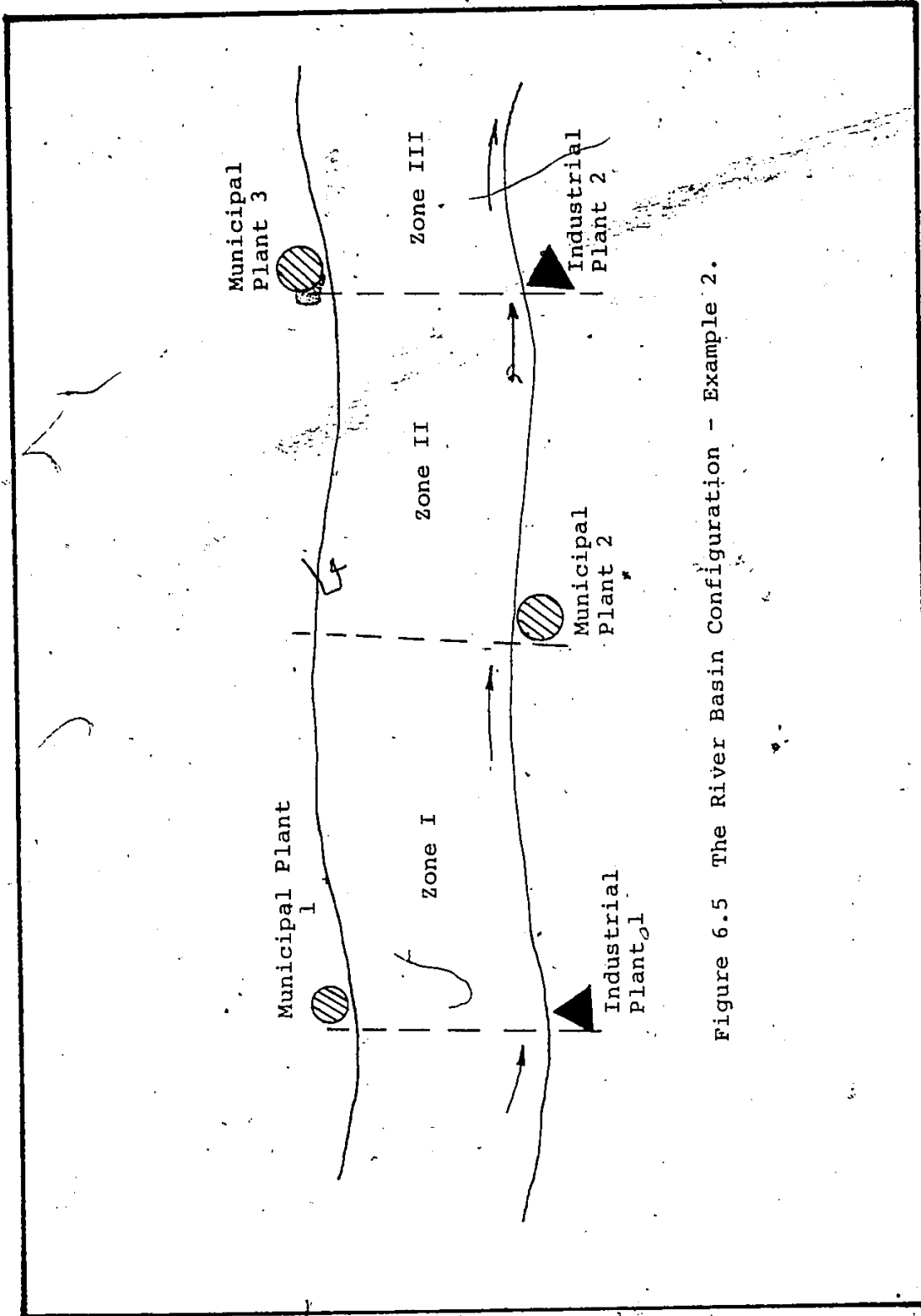


Figure 6.5 The River Basin Configuration - Example 2.

Table 6.4
 Characteristics of the River and the Parameter Values of the Model for Example 2

Parameter	Description	Numerical Value
IU	Upstream BOD Concentration	5.0 mg/l
CU	Upstream DO Concentration	6.0 mg/l
PU	Upstream Phosphorus Concentration	0.05 mg/l
Ia	Local BOD Runoff Rate	0.20 mg/l
Q	Stream Flow Rate	1500 cfs for Zone 1 1500 cfs for Zone 2 1500 cfs for Zone 3
α	The Constant in Photosynthesis Respiration Expression	3.46
τ	Residence Time	2.5 day for zone 1 2.0 day for zone 2 3.0 day for zone 3
T	Stream Temperature	290° K for Zone 1 295° K for Zone 2 300° K for Zone 3
k_3	BOD Adsorption and Sedimentation Coefficient	0.1 day ⁻¹

and 3 days at the corresponding zones.

The operating conditions of the municipal and industrial sources, as well as their waste loads, are given in Table 6.5. At the first zone, the municipal source has an average daily flow of 10 MGD, an organic waste load of 0.2×10^5 lb/day, and a phosphorus waste load of 8.7×10^3 lb/day. The industrial source has a daily flow of 6 MGD, and about 0.1×10^5 lb/day of organic waste. At the second zone, the municipal source has an average daily flow of 3 MGD, an organic waste load of 0.4×10^4 lb/day, and a phosphorus waste load of 0.6×10^3 lb/day. At the third zone, the municipal source has an average daily flow of 4 MGD, an organic waste load of 0.75×10^4 lb/day, and a phosphorus waste load of 0.65×10^3 lb/day. The industrial source in the same zone has 8 MGD average daily flow, and 0.3×10^5 lb/day of organic waste. The water quality standards are given in Table 6.6. The maximum allowable BOD level are 15, 12 and 15 mg/l at the first, second and third zone, respectively. The corresponding figures for DO concentration are 4.5, 5 and 4 mg/l, and for phosphorus is 0.1 mg/l at the three zones.

6.4.2 Model Description

In this section, the SUMT formulation of the objective function and constraints, for the second example is described. The objective function is to minimize:

$$f(X_1, X_2, X_3, X_4, X_5; Z_1, Z_2, Z_3; \theta_{11}, \theta_{21}; \theta_{12}, \theta_{22}; \theta_{13}, \theta_{23}; \theta_{14}, \theta_{24}; \theta_{15}, \theta_{25}) =$$

$$\sum_{\text{municipal source } i=1}^3 \left[e^{10.1125 * Q_i} * 0.3495 * (WB_i * X_i)^{0.1778} * (WP_i * Z_i)^{0.0685} \right]$$

Table 6.5

Operating Conditions of Municipal and Industrial Wastewater Treatment Plants - Example 2

Zone	Source	Average Daily Flow MGD	Waste Load in lb /day	
			Organic Waste	Phosphorus Waste
1	Municipal	10	$0.2 * 10^5$	$0.70 * 10^3$
	Industrial	6	$0.10 * 10^5$	--
2	Municipal	3	$0.40 * 10^4$	$0.60 * 10^3$
3	Municipal	4	$0.75 * 10^4$	$0.65 * 10^3$
	Industrial	8	$0.3 * 10^5$	--

Table 6.6

Water Quality Standards - Example 2

Zone	The Desired Water Quality Limits mg/l		
	BOD	DO	Phosphorus
1	15.0	4.5	0.1
2	12.0	5.0	0.1
3	15.0	4.0	0.1

$$\begin{aligned}
 & + \sum_{j=4}^5 \text{ industrial source } j=4 \left\{ \begin{aligned} & \left[909.0 + 2273(1.1/\bar{Q}_j)^{0.5} \right]; 0 \leq x_j < 0.4 \\ & \text{or} \\ & \left[2700 + (2500.0/\bar{Q}_j)^{0.67} \right. \\ & \quad \left. + 1000.0(0.02 - 0.0001 * \bar{Q}_j) \right]; \\ & \quad \quad \quad .4 \leq x_j < 0.9 \\ & \text{or} \\ & \left[1500 + 6450(1/\bar{Q})^{0.63} \right]; .9 \leq x_j < 1.0 \end{aligned} \right\} \\
 & + \sum_{k=1}^5 \text{ municipal } + \text{ industrial source } k=1 \left\{ (1-x_k) * WB_k * \theta_{1k} * e^{-\theta_{2k}/(1-x_k)} \right\}
 \end{aligned}
 \tag{6.9}$$

Subject to the following constraints:

(a) BOD constraints:

$$g_1 = -LB_1 + 15.0 \geq 0$$

$$g_2 = -LB_2 + 12.0 \geq 0$$

$$g_3 = -LB_3 + 15.0 \geq 0$$

(b) DO Constraints:

$$g_4 = MDO_1 - 4.5 \geq 0$$

$$g_5 = MDO_2 - 5.0 \geq 0$$

$$g_6 = MDO_3 - 4.0 \geq 0$$

where LB_i and MDO_i ; $i = 1, 2, 3$ are given by Equation (5.13) and (5.16) respectively.

(c) Phosphorus Constraints:

$$g_l = -PH_i + 0.1 \geq 0$$

where $i=1,2,3$; $l=7,8,9$; PH_i is given by Equation (5.19).

(d) Boundary Conditions on X_j :

$$(1) g_r = X_j \geq 0$$

$$(2) g_s = 1 - X_j \geq 0$$

where $j=1,2,\dots,5$; $r=10,11,\dots,15$; $s=16,17,\dots,19$.

(e) Boundary Conditions on Z_j :

$$(1) g_t = Z_i \geq 0$$

$$(2) g_u = 1 - Z_i \geq 0$$

where $i=1,2,3$; $t=20,21,22$; $u=23,24,25$

(f) Non-Negativity Constraints on θ_{1j} and θ_{2j} :

$$g_v = \theta_{1j} \geq 0$$

$$g_q = \theta_{2j} \geq 0$$

where $j=1,2,\dots,5$; $v=26,27,\dots,30$; $q=31,32,\dots,35$.

The optimization model of Example 2 consists of eighteen decision variables, and thirty-five constraints. The decision variables represent the BOD treatment levels at all municipal and industrial sources (X_1 through X_5), the phosphorus treatment levels at the three municipal sources (Z_1 through Z_3), and the coefficients of the effluent charge formula (θ_{11} through θ_{25}). The constraints represent for each zone, the maximum allowable BOD concentration (3 constraints),

the minimum allowable concentration of DO (3 constraints), and the maximum limits for phosphorus constraints (3 constraints). The next sixteen constraints account for the boundary conditions on the variables, representing the organic and phosphorus treatment levels at the three zones, and the remaining ten constraints account for the non-negativity of the charge rate coefficients.

The problem has been solved on the same IBM 370/3031 computer, in 6.5 minutes of CPU time. The SUMT procedure converged to the optimal solution in twelve iterations at the specified accuracy limits of 10^{-6} . The computer output illustrating the final results of the SUMT procedure is given in Appendix E

The optimal treatment policy has been obtained at a total annual cost of \$2,685,790, of which \$2,452,890 is the total treatment costs, and \$232,900 is the total effluent charges paid to the river basin authority. The optimal rates of the effluent charge levied against the municipal sources in the first, second, and third zone are 2.98, 4.29, and 6.06 ¢/lb of BOD, respectively. The charge rates levied against the industrial sources in the first, and third zones are 2.95, and 2.98 ¢/lb of BOD, respectively.

The optimal treatment levels of organic waste loads in the first zone are 77.7% for the municipal source, and 64.2% for the industrial source. The municipal source in zone 2 is required to treat 60.7% of its waste. Finally, the required treatment levels in zone 3 are 62.7% for the

municipal source, and 82.7% for the industrial source. As for the phosphorus wastes, the optimal treatment levels for the three municipal sources at the first zone are 98.5%, 90.5% at the second, and 92.0% at the third zones.

The description of the computer program and the results for the two examples are presented in Appendix E.

CHAPTER VIIDISCUSSION OF RESULTS AND SUGGESTIONS FOR FURTHER STUDIES7.1 Introduction

This chapter discusses the results of the two applications of the optimization model. The discussion emphasizes the relationships between the effluent charge rates, on one hand, and waste treatment levels, amount and intensity of the discharged wastes, treatment cost functions, and water quality standards; on the other hand. On the basis of the results of the present investigation, a number of conclusions are drawn, and proposals are made, at the end, for further researchs and studies.

7.2 Discussion of Results - Example (1)

This example deals with three municipal waste sources located at three different zones in a river basin. The sources discharge various amounts of organic and phosphorus waste into the river. However, it is assumed that the same treatment cost function applies at each source. The water quality standards are also assumed to be uniform in all three zones.

In Table 7.1, the results of the optimization of the model in Example 1 are presented. In column (1), the total waste load at each zone is given. The optimal treatment level at each source is given in column (2), and the amount of waste treated at each source is calculated in column (3). The cost of treatment, which is based on the amount indicated in column (3) is computed in column (4). The effluent

Table 7.1
Results of the Optimization of Model for Example 1

Zone No.	ORGANIC WASTE LOAD						PHOSPHORUS WASTE LOAD				(11) Grand Total Cost (4)+(7)
	(1) Total Load WBi lb /day	(2) Optimal Treatment Level X 1 %	(3) Amount of Waste Treated (1) x (2) lb /day	(4) Annual Cost of Treatment \$/year	(5) Amount of Waste Untreated (1) - (3) lb /day	(6) Effluent Charge Rate \$/lb	(7) Total Effluent Charge Cost (5) · (6) \$/year	(8) Total Load WPI lb /day	(9) Optimal Treatment Level Z-1 %	(10) Amount of Waste Treated (8) · (9) lb /day	
1	0.20*10 ⁵	58.07	0.116*10 ⁵	281246	0.084*10 ⁵	7.17	219604	0.8*10 ³	88.9	0.77*10 ³	500850
2	0.22*10 ⁵	68.00	0.149*10 ⁵	318414	0.071*10 ⁵	5.37	137973	0.9*10 ³	98.9	0.88*10 ³	456387
3	0.10*10 ⁵	78.00	0.01*10 ⁵	240532	0.03*10 ⁵	3.18	25565	0.7*10 ³	90.9	0.63*10 ³	266097

charges apply to the untreated portion of waste (column (5)) and are computed in column (7). It is noted that the highest treatment cost, for example, corresponds to the second source with the highest amount of waste treatment, whereas the lowest level of effluent charges corresponds to the third source, which discharges the least amount of untreated waste. This result is mainly due to the fact that the three sources have the same treatment cost function. The results do not indicate that the concerned relationship follows a general pattern, since the operating conditions at each source, such as the average daily flow, the amount of waste, etc., are different.

The optimal treatment configurations for the organic and phosphorus waste loads are given in columns (2), and (9), respectively. About 58% of the organic load generated at the first source should be treated in order to satisfy the waste quality standards in the first and subsequent zones. At the second source, which has the highest amount of organic waste, the required treatment level is 68%. The corresponding figure for the third source, which generates the least amount of organic waste, is 78%.

The behaviour of the BOD, DO and phosphorus concentrations along the river basin is shown in Figures 7.1(a), (b), and (c), respectively. It is interesting to note that, although the BOD levels in all three zones are below the stipulated standard of 15 mg/l (Figure 7.1(a)), treatment of the organic waste is still necessary due to the stringent standard on DO (5 mg/l) as it is depicted in Figure 7.1(b).

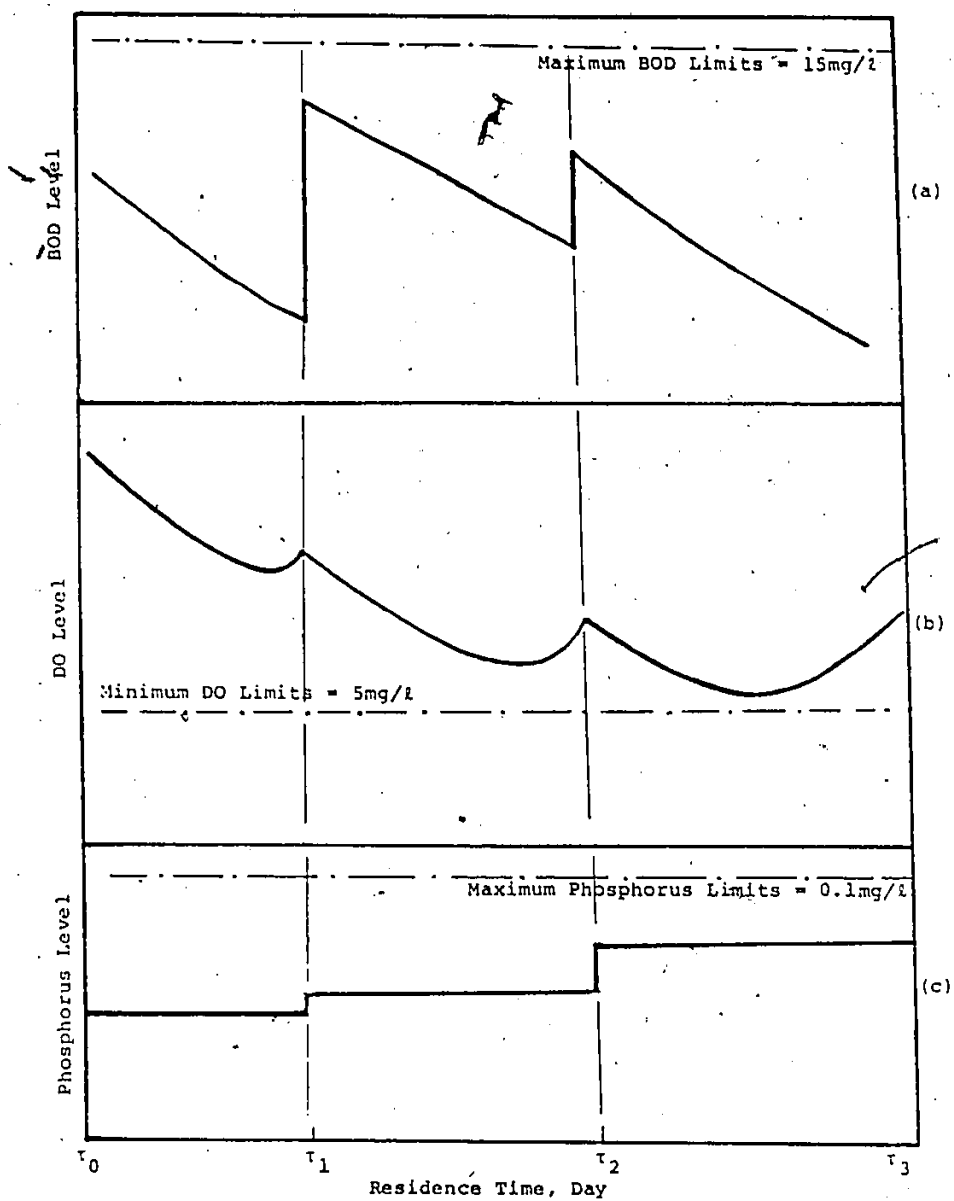


Figure 7.1 (a) BOD Concentration Level
 (b) DO Concentration Level
 (c) Phosphorus Concentration Level

On the other hand, the results indicate that the phosphorus treatment levels are generally high in the three zones. The treatment levels are 88%, 98%, and 91% at the first, second and third source, respectively. This is expected due to the strict concentrations in the river (0.1 mg/l). The results are in general agreement with current governmental regulations for pollution control, which require at least 80% treatment level of the phosphorus wastes produced at any pollution source. (Environment Canada, Ontario - Conference Proceedings, No. 1, May, 1973).

The behaviour of the phosphorus concentration along the river basin is depicted in Figure 7.1(c), which shows that the phosphorus concentration level changes in step-wise manner.

The effluent charge rates are found to be consistent with the above results. They are also in agreement with the nature of the conditions imposed on the basin system, such as the quality requirements at each zone, the nature of the cost function for each source, and the amount of waste generated at each source. The optimal effluent charge rates for each source are given in column (6) of Table 7.1. The results show that the effluent charge rate for the first zone is 7.17 ¢/lb of BOD discharged. This represents the highest charge rate among the three zones. For the second and third zones, the charges are 5.37 and 3.18 ¢/lb of BOD, respectively. The effluent charge rates are inversely related to the BOD treatment level. The highest charge rate corresponds to the lowest BOD treatment level at the first zone, whereas the lowest charge rate corresponds to the

highest treatment level at the third zone.

7.3 Discussion of Results - Example (2)

This example demonstrates the application of the model to a more general case. At each zone in the river, two pollution sources, industrial and/or municipal, are considered. Furthermore, the waste treatment cost functions at various sources are assumed to be different. One or both sources simultaneously discharge their wastes into each zone in the river. Each zone requires different quality limits.

The results of the optimization of the model are presented in Table 7.2. The optimal policy is obtained at the minimum annual cost of \$2,685,790. This includes the effluent charge and the treatment cost for each discharger in the basin. In column (11) of Table 7.2, the breakdown of the total cost among different dischargers is given. It is noticed, in this example, that the treatment costs for industrial sources are higher than those for municipal sources. Also, as can be seen from column (7), the effluent charges for industrial sources are less than those for municipal sources. These results will be explained in detail during the course of the discussion.

The optimal treatment configurations for organic and phosphorus wastes at the five pollution sources are given in columns (2), and (9), respectively. The results indicate that the municipal and industrial sources in the first zone have to treat 77.7% and 64.2% of their organic wastes, respec-

Table 7.2
Results of the Optimization of Model for Example 2

Zone No.	ORGANIC WASTE LOAD					PHOSPHORUS WASTE LOAD					
	(1) Total Load WB lb/day	(2) Optimal Treatment Level X_i	(3) Amount of Waste Treated (1) x (2) lb/day	(4) Annual Cost of Treatment \$/year	(5) Amount of Waste Untreated (1) - (3) lb/day	(6) Effluent Charge Rate \$/lb	(7) Total Effluent Charge Cost (5) x (6) \$/year	(8) Total Load WPI lb/day	(9) Optimal Treatment Level Z _i	(10) Amount of Waste Treated (8) x (9) lb/day	(11) Grand Total Cost (4) + (7)
1 (M) †	$0.2 \cdot 10^5$	77.7	$0.155 \cdot 10^5$	479884.5	$0.045 \cdot 10^5$	2.986	48443.4	$0.70 \cdot 10^3$	98.5	$0.689 \cdot 10^3$	528327.9
1 (I) x	$0.1 \cdot 10^5$	64.2	$0.064 \cdot 10^5$	641432.4	$0.036 \cdot 10^5$	2.958	38576.8	---	---	---	680009.2
2 (M) †	$0.4 \cdot 10^4$	60.7	$0.024 \cdot 10^5$	227533.7	$0.016 \cdot 10^5$	4.293	27665.9	$0.60 \cdot 10^3$	90.5	$0.543 \cdot 10^3$	255199.6
3 (M) †	$0.75 \cdot 10^4$	62.7	$0.047 \cdot 10^5$	278950.6	$0.028 \cdot 10^5$	6.067	61839.7	$0.65 \cdot 10^3$	92.0	$0.598 \cdot 10^3$	340790.3
3 (I) x	$0.3 \cdot 10^5$	82.7	$0.248 \cdot 10^5$	825088.7	$0.052 \cdot 10^5$	2.987	56374.3	---	---	---	881463.0

N.B. † Indicates Municipal (M)
x Indicates Industrial (I)

tively. The treatment level for the municipal source is higher than that for industrial source, because the former has a more concentrated waste (199.8 mg/l) than the latter (166.5 mg/l). In order to satisfy the water quality standards in the first and subsequent zones, the municipal source, therefore has to treat its waste at a higher level than the industrial source.

The municipal source in the second zone, is the sole pollution source in the zone, and is required to treat about 60.7% of its organic waste.

The optimal organic waste treatment levels for the municipal and industrial sources in the third zone are 62.7% and 82.7%, respectively. The high concentration of BOD at the industrial source (374.7 mg/l) is the reason for the correspondingly high treatment level at this source, relative to the municipal source which has a waste concentration of 187.3 mg/l only.

The behaviours of BOD and DO along the river basin are depicted in Figure 7.2(a) and (b). Again, it is noticed that, although the BOD concentration in all three zones is below the maximum allowable level (15 mg/l), the strict DO concentration requirements (4.5 mg/l in zone 1, 5 mg/l in zone 2, and 4 mg/l in zone 3) necessitate the treatment of the organic waste in all three zones.

As for the phosphorus treatment levels (column (9)), the results indicate that the phosphorus concentration standard in the first zone is attained at a high treatment level of 98.5%. The corresponding standards in zones 2 and 3

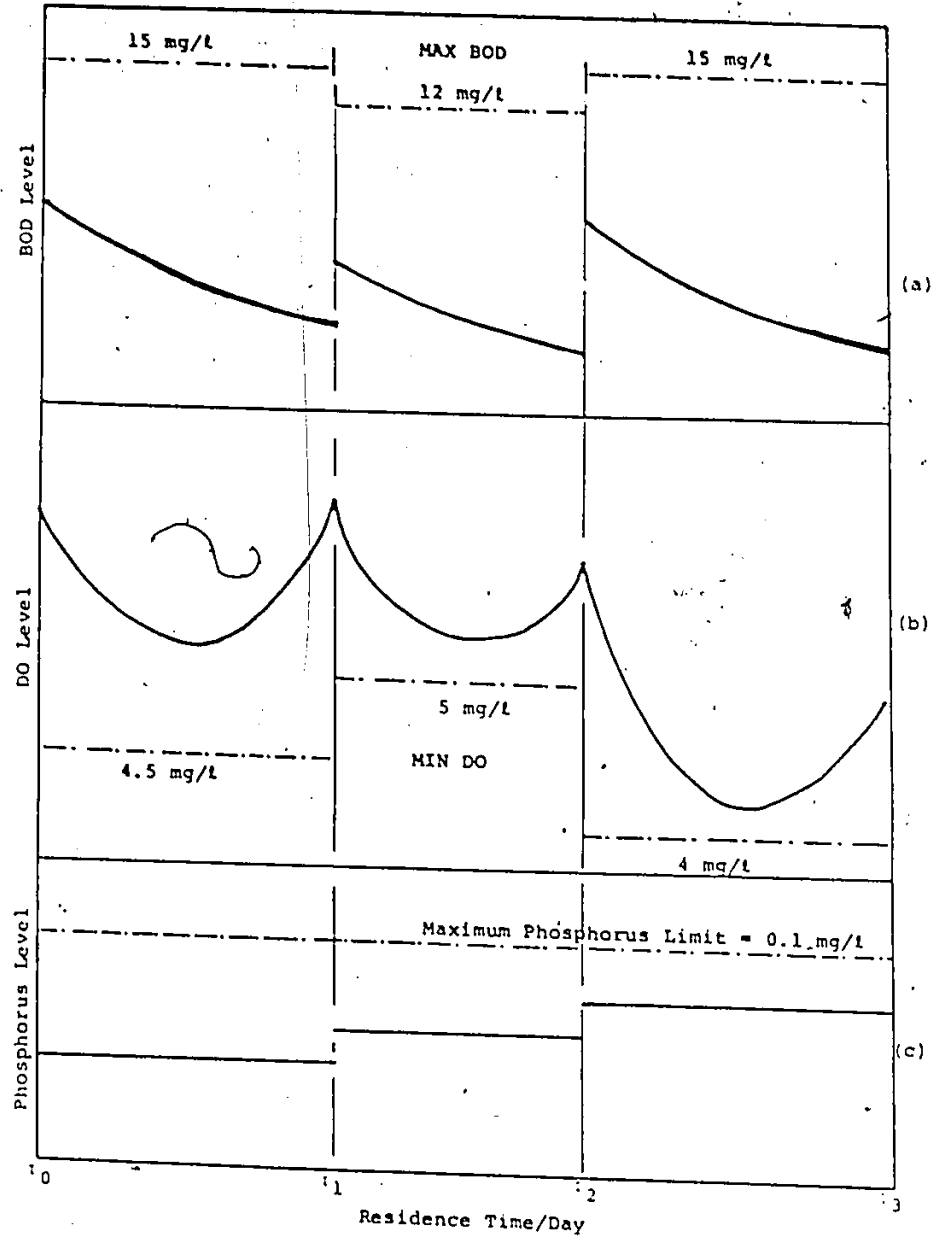


Figure 7.2 (a) BOD Concentration Level
 (b) DO Concentration Level.
 (c) Phosphorus Concentration Level

are attained at treatment levels of 90.5%, and 92.0% for both municipal sources located at the second and third zones, respectively. The high level of phosphorus treatment at the first source is due to the upstream conditions, i.e., the phosphorus concentration of 0.05 mg/l upstream. The behaviour of phosphorus concentration level in the river is shown in Figure 7.2(c), indicating the familiar step-wise variation.

The effluent charge rates to be levied against all the dischargers in the basin are given in column (6). As is evident from these results, dischargers in the same zone may be subject to different charge rates; the municipal source in the first zone, for example, is required to pay 2.98¢/lb of BOD discharged into the river, whereas the industrial source in the same zone is required to pay 2.95¢/lb. The variation in the charge rates, within the same zone, is due to differences in the marginal treatment costs incurred by each source to treat its wastes. To illustrate, while the industrial source at the first zone incurs \$641,432.4 annually to treat about 0.1×10^5 lb/day of BOD, the municipal source incurs less cost (\$479,884.5 per year) to treat more waste (0.2×10^5 lb/day). The allocation of effluent charges between the two sources, therefore, is in agreement with the idea of imposing a higher rate on the source which incurs smaller treatment cost (Kneese & Schultze, 1975). This tends to encourage such dischargers to treat a larger portion of their wastes than they normally do in the absence of effluent charges.

The municipal source in the second zone is required to pay 4.29 ¢/lb of BOD. In the third zone, the effluent charge rate for the municipal source is 6.06¢/lb of BOD, which is higher than the rate for industrial source which is 2.95 ¢/lb. The difference between the two effluent charge rates may be explained in the same way as was given for the first zone.

7.4 Conclusions

The evaluation of the available literature on water pollution control methods, presented in this study, indicates that the effluent charges strategy is an effective, flexible, and efficient approach to deal with the complicated phenomenon of water pollution. In the present work, the strategy has been incorporated within the framework of a water quality management model, which aims at allocating, optimally, the effluent charge rates among all the polluters using the river basin as waste receptor.

There are two features in the present model which distinguish it from other works seeking the same objective (e.g., Hass (1970) and Taylor (1973)). First, the attainment of the water quality standards throughout the river basin is ensured by explicitly specifying the appropriate constraints in the model's formulation. Moreover, these water quality constraints represent non-linear relations. In previous works, either the water quality constraints have not been explicitly considered, or they have been approximated into linear forms. In both cases, however,

there is no guarantee that the required water quality standards are met. This is because the estimation of the optimal effluent charge rates is carried out through an iterative procedure. If water quality standards are not achieved under a given set of effluent charges, the basin authority revises them accordingly. The revised values of effluent charges are transmitted to the polluters in the basin, and then the water quality measures are monitored. The process, which depends entirely on the accuracy of the polluters' response to the charge rates, may take several iterations before it converges on the optimal set of effluent charges. In contrast, the present model determines the effluent charges as well as the levels of waste treatment directly.

Second, the present model determines the effluent charges as a set of non-linear rates as opposed to a set of uniform rates obtained by existing models.

Imposing uniform charge rates on all polluters does not lead to optimal allocation of water resources. For example, a firm which has decided to treat its effluent wastes and/or the incoming water supply for its production processes should, in an efficient resource allocation, face a lower charge rate than a firm which has chosen not to do so. As Tietenberg (1973(a), (b)) has pointed out, theoretically, non-uniform charge rates, rather than uniform rates, constitute the least costly approach to water pollution control.

It is noted that the nonlinearity of the effluent charges is a consequence of the presence of a large number

factors which have to be considered in the development of the water of water quality management. These factors may be classified into two categories: (1) factors which describe the variations in stream characteristics (e.g., stream flow rate, stream width and depth, water quality standards, etc.) at different points, and (2) factors which describe the variations in plant performance (e.g., waste treatment costs, amount and intensity of the discharged wastes, levels of waste treatments, etc.) at each pollution source. The consideration of the factors in the first category necessitates the establishment of different zones in the river, so that the model can deal with each zone independently, and determine the optimal charge rate for that zone. On the other hand, consideration of the factors in the second category is reflected in the formulation of the model's objective function and constraints.

The present model, therefore, is developed to account for the complex interactions among the factors mentioned above, and for the nonlinearity of the effluent charges. Accordingly, the results obtained from the applications of the model are to be interpreted in accordance with these interactions. In the first application, for example, the marginal treatment costs for all the polluters as well as the prevailing stream characteristics in all three zones were the same, but the charge rates (7.17, 5.37, and 3.18 ¢/lb of BOD, at the first, second, and third zone, respectively) were different. To interpret these results, one must focus on the overall picture representing the problem,

and considering such factors as the intensity of the pollutants generated at each source, the treatment levels at each source, etc. Also, it is noted that in the second applications, not only are the effluent charge rates different in different zones, there may even be different rates applied to a number of polluters in the same zone. These results, which are not obvious beforehand, should be interpreted in the light of the complex interactions among the factors involved. This points to the importance of developing complex mathematical models in order to deal with such complicated problems as environmental pollution control.

The model developed in this study provides a vehicle for the basin authority to determine, for any given set of conditions, the optimal effluent charges. The model is flexible in the sense that any adjustment to input data can easily be implemented. In addition, the data preparations and set up are kept to a minimum, so as to make the computer program easy to use.

7.5 Suggestions for Further Studies

(1) The present model is static in nature, in the sense that it does not consider the effect of time on various parameters under consideration. For instance, it is assumed that the waste discharges at various sources do not vary with time; or, that the flow rate in the river basin

is constant. While such assumptions may not represent the situation realistically, they could be conceived of as representing either the average conditions, or the extreme conditions existing in the river basin.

If the model is based on the average conditions prevailing in the river basin, then the results would be adequate in most cases except when extreme conditions exist. It is, however, possible to run the model under these conditions and determine the corresponding optimal policy.

If, on the other hand, the model is based on the extreme conditions that may prevail in the basin, then the results are applicable in all cases. Nevertheless, the obtained results would place a burden, in different degrees, on the dischargers according to the prevailing feasible conditions. The burden is represented in terms of paying more effluent charges or increasing the levels of the treatment.

To alleviate these shortcomings, a dynamic version of the model may be considered which takes into the account the time-dependency of the model's parameters, hydraulic conditions of stream, and waste discharges. However, the set of partial differential equations that result, are too complicated to handle. Some work for the single-stage model have been developed using the concepts of the maximum principal and optimal control theory (Fan, et al 1973; Fan and Hong, 1972). It is suggested that the application of these concepts to multi-stage models be investigated.

(2) The degradation of water resources is caused, not only by the discharge of organic and inorganic waste, but also by the discharge of "waste heat" in the form of heated water discharges from the cooling processes of various industries, notably the power generation stations. Although the present model does not deal with thermal effects of waste discharges, however, it could be expanded to include the effect of varying temperature on BOD and DO along the river.

(3) It is understood that the phosphorus pollutants have a direct, adverse effect on lakes rather than on rivers. Since lakes are the natural destination of many rivers, in order to halt or decrease their enrichment with plant nutrients (eutrophication process), which accelerates in the presence of phosphorus wastes, the concentration levels of these pollutants in the rivers must be kept within acceptable limits. The present model has recognized this process by imposing a set of control limits on the phosphorus concentration in the river. The model, however, could be extended to include effluent charges to be levied on each unit of the phosphorus waste discharged into the river.

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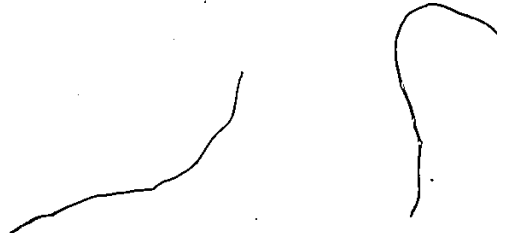
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APPENDICES



APPENDIX A

NUMERICAL SOLUTION OF DIFFERENTIAL EQUATIONS
(RUNGE-KUTTA METHOD)

NUMERICAL SOLUTION OF DIFFERENTIAL EQUATIONS USING THE
RUNGE-KUTTA METHOD

An approximate solution could be obtained, for the two differential equations (5.1), and (5.2) given in Chapter V, through a numerical integration method. The Runge-Kutta method is considered one of the most accurate procedures used in such cases. The solution is obtained in the form of a set of values for the dependent variables corresponding to given values of the independent variables (N.G. Bakhoom, 1962). The method can be described as follows:

Assume two simultaneous differential equations:

$$\frac{dy}{dx} = f(x, y, z) ; \quad \frac{dz}{dx} = g(x, y, z) \quad (1)$$

with initial conditions:

$$x = x_0, \quad y = y_0, \quad z = z_0$$

where it is required to find y and z for a given value of x .

The Algorithm

- (1) Let h : denote the increment in x
 k : denote the increment in y
 ℓ : denote the increment in z

- (2) Compute K and ℓ according to the following expressions:

$$k = \frac{1}{6}k_1 + \frac{1}{3}k_2 + \frac{1}{3}k_3 + \frac{1}{6}k_4;$$

$$\ell = \frac{1}{6}\ell_1 + \frac{1}{3}\ell_2 + \frac{1}{3}\ell_3 + \frac{1}{6}\ell_4$$

(3) Compute $k_1, k_2, k_3,$ and k_4 as follows:

$$k_1 = h * f(x_0, y_0, z_0)$$

$$k_2 = h * f(x_0 + \frac{1}{2}h, y_0 + \frac{1}{2}k_1, z_0 + \frac{1}{2}l_1)$$

$$k_3 = h * f(x_0 + \frac{1}{2}h, y_0 + \frac{1}{2}k_2, z_0 + \frac{1}{2}l_2)$$

$$k_4 = h * f(x_0 + h, y_0 + k_3, z_0 + l_3)$$

(4) $l_1, l_2, l_3,$ and l_4 are sequentially computed as follows:

$$l_1 = h * g(x_0, y_0, z_0)$$

$$l_2 = h * g(x_0 + \frac{1}{2}h, y_0 + \frac{1}{2}k_1, z_0 + \frac{1}{2}l_1)$$

$$l_3 = h * g(x_0 + \frac{1}{2}h, y_0 + \frac{1}{2}k_2, z_0 + \frac{1}{2}l_2)$$

$$l_4 = h * g(x_0 + h, y_0 + k_3, z_0 + l_3)$$

(5) The solution at the new point $x_1 = x_0 + h$ is:

$$y_1 = y_0 + k$$

$$z_1 = z_0 + l$$

(6) Set the new value of x such that:

$$x_2 = x_1 + h_1$$

and repeat the procedure from step (2) to step (6) for all points in the domain of x variable.

Application of the Runge-Kutta Method to Solve the Two Simultaneous Differential Equations of the Transformation Equations

The transform equations are expressed as follows:

$$\frac{dC}{d\tau} = g(C, L, T) \quad (2)$$

$$\frac{dL}{d\tau} = f(L, T) \quad (3)$$

where C = DO concentration level; mg/l

L = BOD concentration level; mg/l

T = Temperature level; °K

τ = Residence time; days

Assume: 1) T is constant throughout the zone

2) The initial conditions are $L = L_0$, $C = C_0$, at

$$\tau = \tau_0.$$

3) ΔC is the increment in C;

ΔL is the increment in L;

$\Delta \tau$ is the increment in τ .

Then:

$$\Delta C = \frac{1}{6}D_1 + \frac{1}{3}D_2 + \frac{1}{3}D_3 + \frac{1}{6}D_4$$

$$\Delta L = \frac{1}{6}B_1 + \frac{1}{3}B_2 + \frac{1}{3}B_3 + \frac{1}{6}B_4$$

where:

$$D_1 = \Delta \tau * g(L_0, C_0, T_0)$$

$$D_2 = \Delta \tau * g(L_0 + \frac{B_1}{2}, C_0 + \frac{D_1}{2}, T_0)$$

$$D_3 = \Delta \tau * g(L_0 + \frac{B_2}{2}, C_0 + \frac{D_2}{2}, T_0)$$

$$D_4 = \Delta \tau * g(L_0 + B_3, C_0 + D_3, T_0)$$

and

$$B_1 = \Delta\tau * f(L_0, T_0)$$

$$B_2 = \Delta\tau * f(L_0 + \frac{B_1}{2}, T_0)$$

$$B_3 = \Delta\tau * f(L_0 + \frac{B_2}{2}, T_0)$$

$$B_4 = \Delta\tau * f(L_0 + \frac{B_3}{2}, T_0)$$

The solution at the new point $\tau = \tau_0 + \Delta\tau$

$$L_1 = L_0 + \Delta L$$

$$C_1 = C_0 + \Delta C$$

APPENDIX B

SEQUENTIAL UNCONSTRAINED MINIMIZATION TECHNIQUE (SUMT)

SEQUENTIAL UNCONSTRAINED MINIMIZATION
TECHNIQUE (SUMT)

The sequential unconstrained minimization technique (SUMT) solves the constrained minimization problem with non-linear or linear objective function. The technique was originally proposed by Carroll (1959, 1961), and developed further by Fiacco and McCormick (1964 (a), (b), 1965, 1966, 1968).

The general nonlinear programming problem with non-linear inequality constraints is to choose x to:

$$\begin{aligned} &\text{minimize } f(x), \\ &\text{subject to:} \qquad \qquad \qquad (1) \\ &\qquad \qquad \qquad g_i(x) \geq 0, \quad i = 1, 2, \dots, m \end{aligned}$$

where x is an n -dimensional column vector $(x_1, x_2, \dots, x_n)^T$. If the variables are required to be non-negative, each constraint is included in the g_i 's.

To solve problem (1), function P can be defined as follows:

$$P(x, r_K) = f(x) + r_K \sum_{i=1}^m \frac{1}{g_i(x)} \qquad (2)$$

where r_K is a positive constant. Subscript K indicates the number of times P function has been set up to solve the problem given by Equation (1). The conditions imposed on the P function are as follows:

- 1) r_K , $K = 1, 2, \dots$, is a positive real number and $r_1 > r_2 > \dots > r_K \dots > 0$. This indicates that $\{r_K\}$ is

a strictly monotonic decreasing sequence and $r_K \rightarrow 0$ as $K \rightarrow \infty$.

2) $R^0 = \{x \mid g_i(x) \geq 0, i = 1, 2, \dots, m\}$ is a non-empty set. This condition indicates that at least one point must exist within the interior of the feasible region.

3) The functions $f(x), g_1(x), \dots, g_m(x)$ are twice continuously differentiable.

4) The function $f(x)$ is convex.

5) The functions $g_1(x), \dots, g_m(x)$ are concave.

6) For every finite $K, \{x \mid f(x) \leq K; x \in R\}$ is a bounded set, where $R = \{x \mid g_i(x) \geq 0; i = 1, 2, \dots, m\}$.

7) The function $P(x, r_K) = f(x) + r_K \sum_{i=1}^m \frac{1}{g_i(x)}$, is, for each $r > 0$, strictly convex for $x \in R^0$. This also indicates that either $f(x)$ is strictly convex or one of g_1, \dots, g_m is strictly concave.

Practical experience indicates that the problem given by Equation (1) can be solved even when these conditions are not met. The three conditions which are absolutely required to obtain any useful results are conditions (1), (2), and (6). Condition (1) guarantees that the sequential minimization of the P function will eventually lead to the solution of minimization of function $f(x)$. Condition (2) eliminates problems with equality constraints. Condition (6) eliminates problems having local minimum at infinite points.

The characteristics of the P function are as follows:

$$1) \lim_{K \rightarrow \infty} \left[r_K \sum_{i=1}^m \frac{1}{g_i(x)} \right] = 0;$$

$$2) \lim_{K \rightarrow \infty} f[x(r_K)] = u^*; \text{ where } x(r_K) \text{ is the value of vector } x \text{ obtained at iteration } K, \text{ and } u^* \text{ is a}$$

finite value.

- 3) $\text{Lim } P[x(r_K), r_K] = u^*$;
- 4) $\{f(x(r_K))\}$ is a monotonically decreasing sequence.
- 5) $\sum_{i=1}^m \frac{1}{g_i(x)}$ is a monotonically increasing sequence.

The proofs of these characteristics are presented in detail in Fiacco and McCormick (1968).

Intuitive Concept of P Function

The problem is now transformed from a constrained one to an unconstrained problem which can be solved by any of the search techniques. This can be explained as follows:

The term $r_K \sum_{i=1}^m \frac{1}{g_i(x)}$ in P function of Equation (2) can be considered as a penalty factor attached to the objective function $f(x)$. By adding the penalty term, the minimization of P function will assure a minimum to be in the interior of the inequality constrained region by avoiding crossing the boundaries of the feasible region. Since the feasible boundary is defined by one or more of the $g_i(x)=0, i=1, \dots, m$, the value of $r_K \sum_{i=1}^m \frac{1}{g_i(x)}$ will approach infinity as the value of x approaches one of the boundary lines. Hence the value of x will tend to remain inside the inequality-constrained region.

Computational Procedure

- (1) Select the initial value of r_0 arbitrarily.
- (2) Select a feasible starting point $x^0 = (x_1^0, x_2^0, \dots, x_n^0)$. If the feasible point cannot be easily obtained, select x^0 arbitrarily. The computer program can search for a feasible one.

- (3) Minimize the P function for the current value of r_K by using the second order optimum gradient method.
- (4) Check if a stopping criterion such as:

$$\frac{f[x(r_K)]}{G[x(r_K)]} - 1 < \varepsilon \quad (3)$$

is satisfied. The solution is the optimal one if the criterion is satisfied; otherwise, go to step 5.

The dual value, $G[x(r_K)]$, is defined as:

$$G[x(r_K)] = f[x(r_K)] - r_K \sum_{i=1}^m \frac{1}{g_i[x(r_K)]} \quad (4)$$

- (5) Set $K = K + 1$ and $r_{K+1} = r_K/c$ where $c > 1$. Repeat the iteration from step (3).

APPENDIX C

COLLECTED DATA

- C1 - Operating and Maintenance Cost
- C2 - Performance of the Plants
- C3 - Calculation of Index Numbers

TABLE C1(1)
Total Operating and Maintenance Costs For Municipal Wastewater Treatment Plants - 1975.

Plant Site	Grand Total	Total Salaries And Employee Benefits	Total Transportation And Communication	Total Repair And Maintenance And Other Services	Total Chemicals And Utilities And Other Supplies	Other Transactions
Dunnville	80432	36969	1591	10020	31852	—
Fergus	42710	23808	1439	5432	12031	—
Kitchener	613929	285624	3111	114538	201873	8783
Simcoe	125984	64076	1455	7733	52720	—
Waterloo	299103	107730	3828	91020	81027	15501
Elmira	46412	33377	1080	151	10124	1680
Burlington	67240	33461	511	14444	18824	—
Campbellford	51867	31629	1396	4528	11510	2804
Huntsville	46799	30592	365	4958	10615	269
Markham	56127	17422	762	23537	14406	—
Newmarket	162219	79152	1109	40130	39790	2038
Richmond Hill	290781	67897	1079	171233	33386	17186
Belleville	329804	159263	2926	58381	76474	32760
Kingston Twp.	123384	64727	2175	11197	45285	—
Sidney Twp.	22319	16283	600	1027	4404	5
Nickel Center	38859	22455	1141	5107	9805	351
North Bay	310622	129528	13725	41927	117133	2315

N.B. All units are dollars

Table CI(1) Continued

Plant Site	Grand Total	Total Salaries And Employee Benefits	Total Transportation And Communication	Total Repair And Maintenance And Other Services	Total Chemicals And Utilities And Other Supplies	Other Transactions
Sturgeon Falls	70539	29388	1339	8243	28469	3105
Valley East	127558	72903	1260	5416	47144	835
Point Edward	45780	26540	1079	2874	15287	—
Port Dover	50306	29982	177	2225	17922	—
Espanola	25266	16718	583	2109	5856	—
Prescott	74592	30369	1089	8020	32069	3045
Trenton	66300	29233	797	14452	21828	—
Chatham	341959	187256	5483	31385	117098	737
Ingersoll	118902	58245	3177	17532	39948	—
St. Mary's	61733	32442	1472	10348	14615	2856
Tillsonburg	88630	48476	1889	13943	24322	—
Van Astra	38132	17665	1508	975	17500	484
Wallaceburg	122443	64476	1801	15605	29304	—
Brantford	465803	261629	3613	66853	133707	11257
Galt	243628	107495	1646	51571	74491	8425
Preston	177549	95612	2459	32436	41690	5352
Delhi	74422	29014	1649	8020	35739	—

TABLE C1(2)

Total Operating and Maintenance Costs for Municipal Wastewater Treatment Plants - 1976

Plant Site	Grand Total	Total Salaries And Employee Benefits	Total Transportation And Communication	Total Repair And Maintenance And Other Services	Total Chemicals And Utilities And Other Supplies	Other Transactions
Point Edward	53267	29756	1101	4652	16898	860
Port Dover	52669	31023	730	2302	20074	—
Prescott	79405	36364	1152	5215	34840	1834
Espanola	33928	24200	987	2216	6525	—
Chatham	376756	195218	5880	37520	137313	825
Ingersoll	121647	56772	2883	13681	43729	4582
St. Mary's	64405	36985	1145	7372	15815	3088
Tillsonberg	110400	54422	2409	19768	33801	—
Van Astra	40260	27475	1724	2122	8759	179
Wallaceburg	172722	72456	4374	39504	52074	4314
Brantford	500570	278562	3372	30855	156330	31451
Galt	254934	119454	1466	50513	78768	4733
Preston	200602	113189	1900	36839	45627	3047
Delhi	62314	32002	1059	10191	19063	—
Dunnville	104355	46591	615	21749	35400	—
Fergus	60146	27564	1990	10064	20528	—

N.B. All units are dollars.

TABLE C1(2) Continued

Plant Site	Grand Total	Total Salaries And Employee Benefits	Total Transportation And Communication	Total Repair And Maintenance And Other Services	Total Chemicals And Utilities And Other Supplies	Other Transactions
Caledonia	90328	34122	1082	29989	25635	—
Hagersville	50515	24670	431	6918	19357	—
Kitchener	668682	308387	3942	154037	192926	9390
Simcoe	151755	76059	869	16514	58313	—
Waterloo	325801	121217	4889	87071	101532	9092
Elmira	55170	34863	1265	4577	14465	—
Burlington DL	72308	36357	416	7449	28086	—
Burlington EG	42277	22130	259	10309	9579	—
Campbellford	72837	49637	3035	4941	13372	2852
Halton Hills	100494	50703	1933	21570	26288	—
Huntsville	77726	43478	2425	9399	18474	3950
Belleville	333530	166023	3345	73966	89969	227
Kingston Twp.	151942	76443	2300	20935	52264	—
Sidney Twp.	15561	9038	592	912	5019	—
North Bay	355286	157006	7475	77720	96801	16284
Sturgeon Falls	89505	47053	1149	9273	26602	4428
Valley East	108517	56069	15107	5124	44366	2853

TABLE CI(3)
Total Operating And Maintenance Costs For Municipal Wastewater Treatment Plants - 1977

Plant Site	Grand Total	Total Salaries And Employee Benefits	Total Transportation And Communication	Total Repair And Maintenance And Other Services	Total Chemicals And Utilities And Other Supplies	Other Transactions
Point Edward	54322	32558	728	3014	16815	1207
Port Dover	67546	33025	—	2860	31719	—
Prescott	79513	39458	1722	6887	29341	2105
Espanola	42861	60332	849	1772	10408	1700
Chatham	447057	213979	7739	76146	149193	—
Ingersoll	149479	75481	8638	16309	44199	4852
St. Mary's	78664	41190	1027	10472	22886	—
Tillsonburg	116697	62374	1919	12416	40056	68
Van Astra	39739	27366	2244	1782	8516	2027
Wallaceburg	216880	81719	5085	58720	59104	12252
Brantford	573841	282967	2918	36826	199658	51472
Galt	314736	135612	2193	67884	103745	5329
Preston	203572	95427	2395	42405	60178	3166
Delhi	59910	28976	525	7961	22448	—

N.B. All units are dollars.

Table Cl(3) Continued

Plant Site	Grand Total	Total Salaries And Employee Benefits	Total Transportation And Communication	Total Repair And Maintenance And Other Services	Total Chemicals And Utilities And Other Supplies	Other Transactions
Dunnville	117062	55624	1196	17401	42841	—
Fergus	59481	31164	1221	8883	18213	—
Caledonia	75616	41476	2606	13390	24144	—
Hagetsville	44603	23525	561	5130	16509	—
Kitchener	836386	337149	5025	219238	268944	6030
Simcoe	175343	76241	182	38963	60321	—
Waterloo	362745	123927	10519	91253	130688	6358
Elmira	65635	36051	6746	6744	14239	1856
Burlington DL	81233	43650	481	12356	24806	—
Campbellford	85546	50648	2132	9619	20305	2841
Halton Hills	166407	76037	3342	25660	49657	11711
Huntsville	116687	53077	6029	10410	43619	3552
Belleville	400354	164649	1092	73793	137775	23045
Kingston Twp.	167599	79409	2258	27531	58401	—
Sidney Twp.	25110	15517	1040	473	8080	—
Trenton	87297	43243	1124	6321	36609	—
North Bay	398466	181948	7575	75336	115260	18347
Sturgeon Falls	101191	41355	1000	15943	39940	2953
Valley East	156464	73445	1741	18439	53612	3227

TABLE C2(1)
 Plant Loading and Performance For Municipal Wastewater Treatment Plants - 1975

Plant Site	Hydraulic Load		Organic Load		Phosphorus Load		
	Average Daily Flow MGD	BOD lb/day	Reduction Percentage	Po4 lb/day	Influent mg/l	Effluent mg/l	
Point Edward	0.27	380	0.68	767.96	22.3	3.8	
Port Dover	0.84	1300	0.64	46.08	5.5	1.4	
Espanola	0.47	550	0.57	27.66	5.9	4.0	
Prescott	0.90	1100	0.57	35.01	3.9	2.1	
Trenton	2.2	4300	0.37	107.51	4.9	4.1	
Chatham	4.4	7800	0.96	570.46	13.0	0.80	
Ingersoll	1.0	1800	0.95	112.7	11.3	2.2	
St. Mary's	0.45	1287	0.98	33.66	7.5	1.2	
Tillsonburg	0.87	930	0.95	72.88	8.4	0.9	
Van Astra	0.17	68	0.95	4.75	2.8	1.4	
Wallaceburg	0.82	928	0.93	39.25	4.8	0.7	
Brantford	10.1	17000	0.86	644.66	6.4	1.4	
Galt	5.5	7500	0.93	362.02	6.6	1.6	
Preston	1.5	7200	0.97	104.72	7.0	1.0	
Delhi	0.39	630	0.94	32.28	8.3	2.2	
Dunnville	0.93	1700	0.98	43.59	4.7	0.6	
Fergus	0.67	800	0.97	25.85	7.4	1.3	
Kitchener	14.6	30000	0.91	1135.74	7.8	1.0	
Simcoe	1.9	5300	0.96	121.27	6.4	0.8	
Waterloo	5.8	15000	0.95	665.61	7.1	0.9	
Elmira	0.52	8200	0.87	30.60	5.2	2.2	
Burlington	1.80	2200	0.89	122.07	6.8	3.5	
Campbellford	1.1	550	0.94	15.36	1.4	0.7	

TABLE C2(1) Continued

Plant Site	Hydraulic Load		Organic Load		Phosphorus Load		
	Average Daily Flow MGD	BOD lb/day	Reduction Percentage	PO ₄ lb/day	Influent mg/l	Effluent mg/l	
Huntsville	0.35	320	0.91		7.4	1.2	
Markham	1.0	1300	0.85	73.80	7.4	4.9	
Newmarket	2.0	3500	0.86	203.45	10.2	2.8	
Richmond Hill	1.6	2000	0.95	122.87	7.7	3.7	
Belleville	7.1	6000	0.80	644.36	9.1	5.2	
Kingston Twp.	1.76	3400	0.86	136.91	7.8	3.9	
Sidney Twp.	0.14	151	0.86	4.33	3.1	1.4	
Nickel Center	0.34	400	0.88				
North Bay	6.6	8300	0.89		7.6	1.4	
Sturgeon Falls	1.2	790	0.91	112.7	11.2	2.2	
Valley East	0.38	360	0.86		6.7	5.6	

TABLE C2 (2)
 Plant Loading and Performance for Municipal Wastewater Treatment Plants - 1976

Plant Site	Hydraulic Load		Organic Load		Phosphorus Load		
	Average Daily Flow MGD	BOD lb/day	Reduction Percentage	PO ₄ lb/day	Influent mg/l	Effluent mg/l	
Point Edward	0.42	600	0.64	68.69	16.4	2.7	
Port Dover	1.1	2700	0.75	74.6	6.8	1.0	
Prescott	0.94	6000	0.62	38.44	4.1	1.7	
Espanola	0.48	600	0.65	29.68	6.2	3.8	
Chatham	4.7	9000	0.94	674.98	14.4	1.0	
Ingersoll	1.2	2000	0.95	87.36	7.3	1.6	
St. Mary's	0.55	2000	0.97	51.56	9.4	1.3	
Tillsonburg	0.91	800	0.96	66.25	7.3	0.7	
Van Astra	0.16	70	0.89	4.79	3.0	0.9	
Wallaceburg	1.0	1300	0.93	51.86	5.2	0.9	
Brantford	10.1	15000	0.90		6.8	1.6	
Galt	5.6	10000	0.95	296.0	5.3	1.2	
Preston	1.7	8000	0.96	106.81	6.3	0.9	
Delhi	0.44	500	0.92	37.74	8.6	0.7	
Dunnville	.96	1100	0.95	42.13	4.4	0.5	
Fergus	.87	1000	0.93	68.55	7.9	1.3	
Caledonia	.36	400	0.85	22.62	6.3	3.7	
Hagersville	.35	200	0.92	14.31	4.1	0.9	
Kitchener	15.9	33000	0.86	236.87	7.8	1.3	
Simcoe	2.1	4200	0.96	117.28	5.8	0.4	
Waterloo	6.0	13000	0.95	371.0	6.5	1.2	

TABLE C2(2) Continued

Plant Site	Hydraulic Load		Organic Load		Phosphorus Load		
	Average Daily Flow MGD	BOD lb/day	Reduction Percentage	PO4 lb/day	Influent mg/l	Effluent mg/l	
Elmira	0.57	9000	0.91	33.54	5.9	2.5	
Burlington DL	1.50	2000	0.92	106.21	7.1	3.3	
Burlington EG	0.86	900	0.88	55.75	6.5	3.9	
Campbellford	1.1	400	0.65	16.46	1.1	0.9	
Halton Hills	2.1	2000	0.85	123.57	5.9	2.5	
Huntsville	0.32	300	0.88	44.64	14.3	1.3	
Belleville	7.9	6000	0.84	323.03	4.1	1.3	
Kingston	2.12	3000	0.93	103.6	4.9	1.0	
Sidney	0.13	80	0.91	5.19	4.0	0.9	
North Bay	6.4	9000	0.89	527.7	8.3	1.2	
Sturgeon Falls	1.3	1000	0.90	171.14	13.2	2.9	
Valley East	0.52	900	0.90	48.75	9.4	6.1	

TABLE C2(3)
 Plant Loading and Performance for Municipal Wastewater Treatment Plants - 1977

Plant Site	Hydraulic Load	Organic Load		Phosphorus Load		
	Average Daily Flow MGD	BOD lb/day	Reduction Percentage	PO ₄ lb/day	Influent mg/l	Effluent mg/l
Point Edward	0.41	652	0.6	50.29	12.3	2.5
Port Dover	0.66	1900	0.78	59.90	9.1	0.9
Prescott	0.96	540	0.38	16.75	3.5	1.3
Espanola	0.48	530	0.57	30.16	6.3	4.2
Chatham	5.30	9500	0.93	845.72	16.0	1.0
Ingersoll	1.12	1900	0.96	99.41	8.9	1.7
St. Mary's	0.54	1500	0.97	32.85	6.1	1.2
Tillsonburg	1.01	1100	0.97	70.51	7.0	0.7
Van Astra	0.17	88	0.96	5.09	3.0	0.4
Wallaceburg	0.99	1200	0.96	55.29	5.6	0.9
Brantford	9.7	15000	0.94	667.50	6.9	1.4
Galt	5.2	8800	0.92	326.72	6.3	2.1
Preston	1.7	8800	0.96	101.73	6.0	0.9
Delhi	0.43	540	0.92	28.73	6.7	0.3
Dunnville	0.98	1400	0.95	40.07	4.1	0.8
Fergus	0.67	770	0.93	43.43	6.7	1.1
Caledonia	0.29	250	0.75	9.54	3.3	1.7
Hagersville	0.37	380	0.96	19.93	5.4	0.3
Kitchener	14.4	33000	0.92	1105.82	7.7	1.1
Simcoe	2.3	5000	0.94	189.89	6.8	0.8
Waterloo	5.9	13000	0.92	353.05	6.0	0.9

TABLE C2(3) Continued

Plant Site	Hydraulic Load		Organic Load		Phosphorus Load		
	Average Daily Flow MGD	BOD lb/day	Reduction Percentage	PO4 lb/day	Influent mg/l	Effluent mg/l	
Elmira	0.55	1200	0.93	35.65	6.5	2.5	
Burlington DL	1.6	2300	0.91	94.15	5.9	1.0	
Campbellford	1.21	834	0.88	35.0	2.9	1.1	
Halton Hills	2.4	2900	0.93	148.4	6.2	1.7	
Huntsville	0.36	380	0.89	25.49	7.1	1.3	
Belleville	7.1	9500	0.89	424.85	6.0	0.7	
Kingston Twp.	2.4	3800	0.87	134.04	5.6	1.7	
Sidney	0.12	64	0.75	5.74	4.8	1.0	
Trenton	3.28	3500	0.84	440.41	6.4	1.1	
North Bay	6.90	10000	0.84	49.37	3.3	0.6	
Sturgeon Falls	1.50	1100	0.90	88.5	8.7	7.3	
Valley East	1.02	1300	0.87				

TABLE (C.3)
THE DATA USED FOR CALCULATING THE INDEX NUMBERS FOR MAINTENANCE AND OPERATING COSTS OF THE
MUNICIPAL WASTEWATER TREATMENT PLANTS

Plant Site	1975		1976		1977		Index Number			
	Annual Cost	Design Capacity MGD	Annual Cost	Design Capacity MGD	Annual Cost	Design Capacity MGD	75	76	77	
Chatham	341959	4.5	376756	4.5	447057	4.5	1.3	1.18	1	
Ingersoll	118902	2.25	121647	2.25	149479	2.25	1.25	1.22	1	
St. Mary's	61733	0.85	64405	0.85	78664	0.85	1.27	1.22	1	
Tillsonburg	88630	1.80	110400	1.8	116697	1.8	1.31	1.05	1	
Wallaceburg	122443	1.5	172722	1.5	216880	1.5	1.77	1.25	1	
Brantford	465803	12.5	500570	12.5	573841	12.5	1.23	1.14	1	
Galt	243628	8.5	254934	8.5	314763	8.5	1.29	1.23	1	
Preston	177549	3.71	200602	3.71	203572	3.71	1.14	1.01	1	
Delhi	74422	0.7	62314	0.7	59910	0.7	1.45	1.12	1	
Dunnville	80432	1.7	104355	1.7	117062	1.7	1.39	.98	1	
Fergus	42710	1.1	60146	1.1	59481	1.1	1.15	.83	1	
Caledonia	65666	.5	90328	0.5	75616	.5	1.01	.88	1	
Hagersville	43807	0.2	50515	0.2	44603	0.7	1.36	1.25	1	
Kitchener	613929	13.5	668682	13.5	836386	13.5	1.39	1.15	1	
Simcoe	125984	2.0	151755	2.0	175343	2.0	1.20	1.15	1	
Burlington	67040	2.5	72308	2.5	81299	2.5	1.21	1.12	1	
Campbellford	51867	1.0	72837	1.0	85545	1.0	1.64	1.17	1	
Halton Hills	92523	1.5	100494	1.5	166407	1.5	1.79	1.65	1	
Huntsville	46799	0.25	77726	0.25	116687	0.25	2.49	1.5	1	
South Peel	2321849	50	2971284	50	4001712	50	1.72	1.34	1	
Belleville	329804	8	333530	8	400354	8	1.21	1.01	1	
Kingston	129384	2.42	151942	2.42	167599	2.24	1.29	1.1	1	
Sidney	22319	0.12	15561	0.12	25110	0.12	1.12	1.61	1	
North Bay	310622	8	355285	8	398466	8	1.28	1.12	1	
Sturgeon Falls	70539	1	88505	1	101191	1	1.43	1.14	1	
Valley East	127558	2.5	108517	2.5	156464	2.5	1.22	1.44	1	
AVERAGE VALUE OF PRICE INDEX								1.38	1.18	1

APPENDIX D

STATISTICAL ANALYSIS OF THE COST MODEL

STATISTICAL ANALYSIS OF THE COST MODEL

The proposed cost model has been transformed into a Log-Linear form before the direct ordinary least-square method was applied to estimate the parameters of the model β_0 , β_1 , β_2 and β_3 . The estimation process was carried out using the General Linear Model (GLM) computer routine, which was developed at North Carolina State University (1976), and which is part of the SAS (Statistical Analysis System) Library. The procedure uses the principle of the least squares to fit a fixed effect linear model to any type of data.

The statistical results of the GLM procedure are exhibited in the following three tables. Table D.1 exhibits the parameter values of the model and the statistical 'T' test for the estimated mean value of each parameter. The null hypothesis of the test is:

$$H_0 : \beta_j = 0 ; j = 0,1,2,3$$

that is, the estimated value of the parameter, β_j does not differ significantly from zero. The test rejects the null hypothesis for β_0 , β_1 , and β_2 at a high level of significance ($\alpha = 0.01$), but accepts the hypothesis for the third parameter β_3 at $\alpha = 0.1$

Table D.2 indicates the results of the 'F' test for the individual effects of the variables in explaining the model. The three variables \bar{Q} , $(X * WB)$, $(Z * WB)$ are shown to be significant in explaining the model.

Table D.3 indicates the combined effect of the three independent variables as a whole is shown to be highly sig-

nificant (at $\alpha = 0.001$) to explain the model. The third variable, the amount of treated phosphorus ($Z * WP$), has a small coefficient value, β_3 , as indicated by the 'T' test. However, the variable itself is highly significant in explaining the model, as is indicated in Tables D.2 and D.3. It was decided, therefore, to keep the third variable in the model.

Table D.1
Parameter Estimates and the Statistical Significance Model For
 $C = e^{\beta_0 + Q\beta_1 + (XB*WB)_j \beta_2 + (XP*WP) \beta_3}$

Parameters	Estimate	T Value For $H_0: \beta_i = 0$	$Pr\left\{\frac{\beta - 0}{\frac{\sigma}{n}} > T \right\}$	STD Error of Estimate
β_0	10.1125	22.30	0.0001	0.4535
β_1	0.3495	3.66	0.0004	0.0668
β_2	0.1778	2.66	0.0091	0.0668
β_3	0.0685	1.03	0.3078	0.0955

Table D.2
The ANOVA Table for Individual Effects of Model's Parameters

Source Of Variation	D.F.	Sum Of Squares	Mean Square	F Value	PR	F
\bar{Q}	1	52.39122	52.39122	367.57	0.0001	
XB*WB	1	1.7216	1.7216	12.08	0.0008	
XP*WP	1	1.9085	1.9085	13.39	0.0004	
ERROR	96	13.6831	0.142533			
CORRECTED TOTAL		69.7046				

Table D.3

The ANOVA Table for Combined Effect of Model's Parameters

Source of Variation	D.F.	Sum of Squares	Mean Square	F Value	PR F
Model's Parameters	3	56.02140	18.6780	131.01	0.0001
Error	96	13.6831	0.1425		
Corrected Total	99	69.7046			

S T A T I S T I C A L A N A L Y S I S S Y S T E M

NOTE: THE JOB COST HAS BEEN RUN UNDER RELEASE 70.00 OF SAS AT THE UNIVERSITY OF MINDOUR.

```

1 DATA ; INPUT YEAR IN ET Y WD PIRG MG PD ;
2 IF YEAR=3 THEN V=Y+1.25 ;
3 IF YEAR=2 THEN V=Y+1.10 ;
4 IF YEAR=1 THEN V=Y+1.10 ;
5 W=(INR)/IN ;
6 WNGAI=25 ;
7 W=(0+IN)/20.186 ;
8 X1=N0*PD ;
9 X2=WP*PIP ;
10 LY=LUG(Y) ; X1=LUG(MG) ; X2=LUG(WX1) ; X3=LUG(WX2) ;
11 CAPDS ;

```

NOTE: DATA SET WORK-DATA1 HAS 100 OBSERVATIONS AND 17 VARIABLES. 52 OBS/TK. NOTE: THE DATA STATEMENT USED 0.59 SECONDS AND 102K.

```

112 PROC GLM ;
113 TITLE 1 DERIVED COST MODEL FOR MUNICIPAL ;
114 TITLE 2 WASTE WATER TREATMENT PLANT IN ;
115 TITLE 3 SOUTHERN ONTARIO ;
116 TITLE 4 WASTE WATER TREATMENT COST ;
117 TITLE 5 V= AVERAGE DAILY COSTING COST ;
118 TITLE 6 MG= AVERAGE DAILY FLOW IN MILLIO ;
119 TITLE 7 X1= PERCENTAGE REMOVAL OF BOD ;
120 TITLE 8 X2= PERCENTAGE REMOVAL OF P04 ;
121 TITLE 9 W1= AMOUNT OF ORGANIC WASTE LOAD LB/DAY ;
122 TITLE 10 W2= AMOUNT OF PHOSPHORUS WASTE LOAD ;
123 MODEL LY= X1 X2 X3 /P CLM ;

```

NOTE: THE PROCEDURE GLM USED 1.22 SECONDS AND 158K AND PRINTED PAGES 1 TO 2.

```

125 PROC PLOT ; PLOT Y=MG ;
126 TITLE ANNUAL COST VS. AVG DAILY FLOW ;

```

NOTE: THE PROCEDURE PLOT USED 0.42 SECONDS AND 120K AND PRINTED PAGE 3.

```

127 PROC PLOT ; PLOT Y=WX1 ;
128 TITLE ANNUAL COST VS. WASTE REMOVED ;

```

NOTE: THE PROCEDURE PLOT USED 0.40 SECONDS AND 120K AND PRINTED PAGE 4.

```

129 PROC PLOT ; PLOT Y=WX2 ;
130 TITLE ANNUAL COST VS. PHOSPH. WASTE REMOVED ;

```

NOTE: THE PROCEDURE PLOT USED 0.40 SECONDS AND 120K AND PRINTED PAGE 5.

NOTE: SAS USED 150K MEMORY.

NOTE: SAHR, GOODRICH, SALL AND HELWIG
SAS INSTITUTE INC.
P.O. BOX 10060
RALEIGH, N.C. 27605

DEPENDENT VARIABLE: Y
 OBSERVED VALUE
 PREDICTED VALUE
 RESIDUAL
 LURK 95% CL FOR MEAN
 UPPER 95% CL FOR MEAN

INDEPENDENT VARIABLE	OBSERVED VALUE	PREDICTED VALUE	RESIDUAL	LURK 95% CL FOR MEAN	UPPER 95% CL FOR MEAN
45	11.1680466	11.5507771	-0.3827305	11.45215614	11.69820323
46	10.9555408	12.8305666	-1.8750258	10.78761071	10.94861194
47	15.0232749	12.4047902	0.0474847	12.07114562	12.26756019
48	11.0136809	11.3110578	0.0447231	11.20612947	11.41759031
49	12.0367605	12.0891290	0.0004495	12.00196339	12.17479685
50	9.8180375	12.7788631	-0.4391875	10.09871149	9.91527912
51	12.0579714	11.7424106	0.0791796	12.57314268	11.91513353
52	11.5675621	12.7686503	-0.4417018	11.5318303	11.85753374
53	11.9149114	11.7880175	0.2479045	12.58703969	12.95026218
54	11.2294090	11.4164823	0.1269268	11.28367994	11.89235509
55	10.5908835	11.6671361	-0.1435472	11.26024748	11.57573177
56	12.2709948	10.3037713	0.2063112	10.22391453	11.73574875
57	11.2681076	11.6854934	-0.2612015	10.22391453	10.70343116
58	12.9595725	13.0359873	-0.0764148	13.08140483	13.18938876
59	12.2837593	12.6572176	0.0265417	12.51388907	12.78196048
60	11.9202596	11.1484613	0.0123526	12.01300720	12.40903759
61	9.9314321	11.6825715	-0.1478856	11.02577795	11.27118460
62	12.0744933	11.3497407	0.0478870	11.53282738	11.71231443
63	12.6368455	11.3770776	0.0478870	11.28008870	11.47406499
64	10.7056091	11.3770776	-0.3036653	10.89867884	11.53169830
65	12.8145319	13.3974079	-0.2071042	13.12203225	12.42677806
66	11.9518637	12.7341504	-0.2523867	11.23260312	12.92338101
67	10.9214187	11.9523131	-0.2712105	11.23260312	11.49356772
68	12.9810443	12.1439636	0.6301347	11.85683961	12.01358701
69	11.1310215	12.8117759	-0.1217710	12.05567389	12.23225447
70	10.1636490	10.1636490	0.0000000	9.96604190	10.35815897
71	11.3707138	12.2818181	-0.9047473	12.17319158	12.56580604
72	12.3537745	12.8007751	-0.2050054	11.22010666	12.36681699
73	11.52476510	11.5247651	0.0000000	10.84409570	11.84251839
74	11.66662193	10.5316673	0.6554830	10.84409570	11.10387917
75	11.0291950	11.5214492	-0.4381363	11.50488465	11.731800520
76	11.36679783	12.1716307	-0.1410335	12.07912726	12.26159888
77	11.23342318	10.7155489	0.5178742	11.42459490	11.70229968
78	12.88300253	12.7200879	0.1629146	10.56132358	10.86377429
79	11.57857833	13.37812560	-0.2003473	12.53137445	12.80377429
80	11.6333674	11.9868679	0.2003473	11.48232810	12.56041917
81	11.42445954	10.2013726	0.9277086	10.74927336	11.87222009
82	11.8175392	11.45245186	0.0419389	10.98156068	11.54302518
83	11.5767121	10.94504161	0.6316705	11.36348746	11.55541625
84	10.74862814	10.28992311	0.4587050	10.82740628	11.00259665
85	12.04514022	12.01897224	0.0261679	10.121427	10.46714295
86	11.4512823	11.86189828	-0.1046024	11.92350836	12.61232911
87	11.41442007	11.94755255	-0.1460103	11.66011893	12.03392316
88	11.08733642	10.75164186	0.3356946	10.81173423	11.03574394
89	11.08733642	0.27799094	0.5567855	10.59098933	10.80027753
90	13.28901717	11.66418231	0.5567855	11.30091347	12.02745103
91	11.35420449	13.04101393	-0.2440037	12.80268815	13.19915905
92	11.71430918	11.55031789	0.1575912	11.79000329	11.94680370
93			0.0000000	11.44138046	11.67125932
94			13.64318781		
95			-0.0000000		
96			0.10870445		
97			1.77548176		

SUM OF RESIDUALS
 SUM OF SQUARED RESIDUALS - L INQUIRY 55
 SUM OF OBSERVED VALUES
 CURRIM-WAT SUN D

VERY POOR COPY

TO W-2 AMOUNT OF PENSIONERS WAST LEAD
ALLEGED LINEAR MODEL'S PRO. CLASH

DEPENDENT VARIABLE: LY

SOURCE DF J SUM OF SQUARES MEAN SQUARE F VALUE PR > F TYPE IV SS F VALUE PR > F R-SQUARE C.V.

MODEL 1 1.42796313 1.42796313 1.42796313 0.0001 1.00000000 13.39 0.0004

ERROR 30 1.42796313 0.04759878 0.15866257 0.0001 1.00000000 17.08 0.0004

CORRECTED TOTAL 31 2.85592626 0.09242626 0.29866257 0.0001 1.00000000 1.05 0.0004

ADJUSTED TOTAL 31 2.85592626 0.09242626 0.29866257 0.0001 1.00000000 1.05 0.0004

PARAMETER ESTIMATE

INTERCEPT 10.1125225

X1 0.17951510

X2 0.1452448

X3 0.06082198

TYPE I SS

PARMETER=0 23.30

1 2.66

2 1.03

3 1.03

STANDARD ERROR OF ESTIMATE

0.45253281

0.05251485

0.06682419

0.05683625

PARAMETER	ESTIMATE	TYPE I SS	PR > F	F VALUE	DF	STD ERROR OF ESTIMATE	LOWER 95% CL FOR MEAN	RESIDUAL	UPPER 95% CL FOR MEAN
1	10.1125225	23.30	0.0001	10.08	1	0.45253281	11.3332720	-0.2744601	11.52208710
2	0.17951510	2.66	0.0001	1.05	1	0.05251485	10.7130460	-0.4284047	11.06236684
3	0.1452448	1.03	0.3078	0.3078	1	0.06682419	11.90350083	-1.2242829	12.57366593
4	0.06082198	1.03	0.3078	0.3078	1	0.06682419	11.28142958	-0.4715496	11.54200459
5							11.60023838	-0.4715496	12.13053200
6							10.86221340	-0.32263163	11.12848500
7							10.81962243	-0.32263163	11.10850178
8							11.35597350	-0.32263163	11.10233274
9							10.79977171	-0.32263163	11.27526732
10							11.66317326	-0.32263163	11.70046378
11							11.54413261	-0.32263163	11.98049255
12							12.49012125	-0.4724068	11.9531937
13							11.61912371	-0.22922567	12.80443064
14							11.16298103	0.06957495	11.87130176
15							11.40694003	0.2574577	10.4954171
16							12.85533431	0.32110238	11.56538616
17							12.5424244	0.05646050	13.20966407
18							11.94794064	0.27249122	12.79121895
19							11.00401553	0.40397917	12.3241472
20							11.55390855	-0.03075736	11.26560688
21							11.25676997	-0.39839589	11.75731394
22							12.07913417	-0.1091062	13.40836653
23							11.34052408	-0.3234542	13.10546550
24							11.15728309	-0.21006308	11.53473384
25							10.82959234	-0.11043062	11.63588791
26							11.47770755	-0.32158085	11.10053364
27							13.16093749	0.31105734	11.68036366
28							12.01960893	0.19743142	12.22302874
29							10.82516724	0.94830336	12.22302874
30							12.66231774	0.35952397	12.44286439
31							12.53256632	-0.43010924	12.44286439
32							11.71885870	-0.06648261	12.86970378
33							11.32075663	-0.27197739	12.6975913
34							11.42955884	0.23594353	11.64875813
35							11.5517853	-0.5927032	11.65331381
36							12.07517853	0.10142119	11.7892082
37							12.07517853	-0.10142119	12.30364163
38							11.10053364	0.07221155	11.29871698
39							11.50563620	0.13055371	11.65959592

APPENDIX E

DESCRIPTION OF THE COMPUTER PROGRAM

- (1) Program Description
- (2) Flow Chart
- (3) Program List
- (4) Program Set-Up and Results For Example (1)
- (5) Program Set-Up and Results For Example (2)

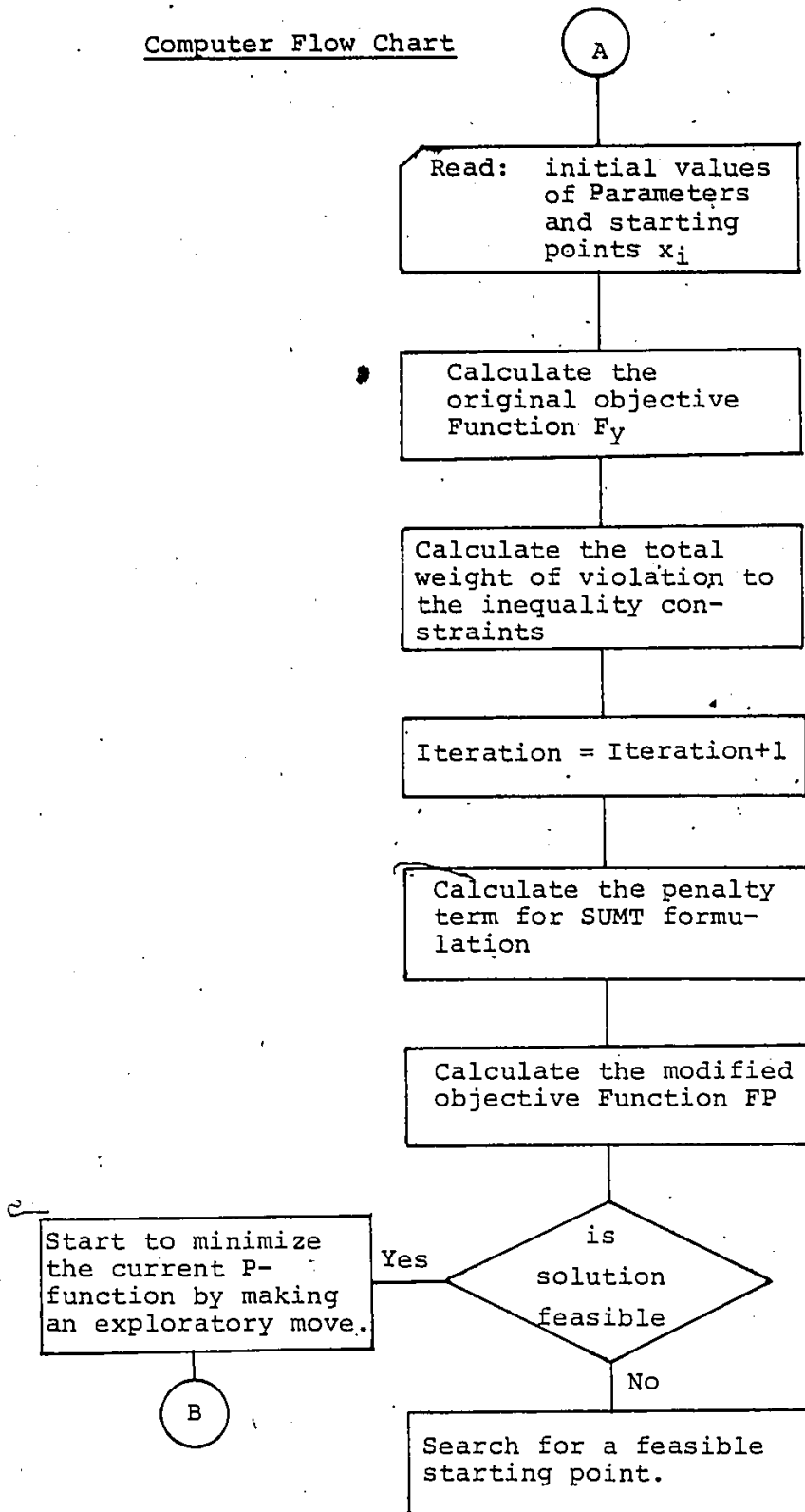
DESCRIPTION OF THE COMPUTER PROGRAM

The computer program used in this study has been developed to determine the optimal charge rates as well as the optimal treatment configuration at each source of pollution. The computer program consists of a main program, three special-purpose subroutines, and five user-supplied subroutines. The main program contains the optimization algorithm, which is based on the search developed by Hooke and Jeeves (1961). It performs the minimization procedure on the unconstrained objective function, supplied to it by the SUMT procedure. The following three special-purpose subroutines transform the constrained optimization problem into an unconstrained one:

- (1) SUBROUTINE WEIGH is used to compute, for the initial solution, the total weigh of violation to the inequality constraints. The weigh is used to adjust the initial solution to be located in the interior of the constrained region. The search procedure for the global minimum, therefore, starts at a feasible point.
- (2) SUBROUTINE PENAT is used to calculate the penalty term for the modified objective function in the SUMT formulation.
- (3) SUBROUTINE BACK is a heuristic program used to pull the current infeasible solution back into the feasible region during the search procedure.

The five user-supplied subroutines are:

- (1) SUBROUTINE READIN which reads any additional data necessary for the computation purpose.
- (2) SUBROUTINE OUTPUT which is used to print out any additional information at any stage of the search procedure.
- (3) SUBROUTINE OBRES which contains the objective function, F , and the problem constraints $G(I)$.
- (4) SUBROUTINE NUMERIC which is developed to numerically compute the variations of the BOD concentration in each zone, the minimum level of DO concentration and its location in each zone, and the maximum concentration of phosphorus level in each zone.
- (5) SUBROUTINE COST which is developed to calculate both the treatment and effluent charge costs for each pollution source at any zone. The output of this subroutine is directly supplied to the objective function, F , in SUBROUTINE OBRES. The following is the flow chart for the computer program used. The program listing follows the flow chart.

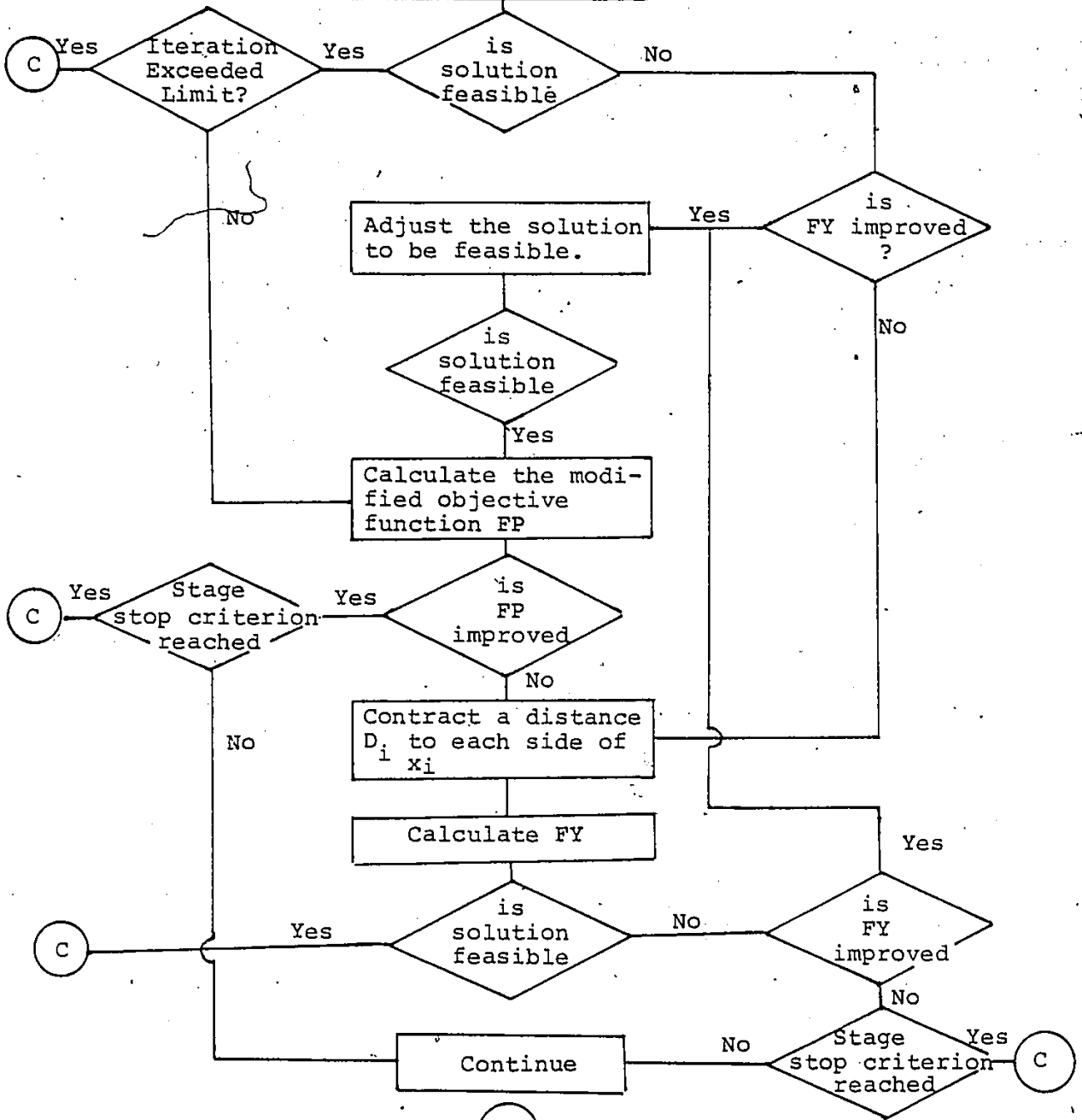
Computer Flow Chart

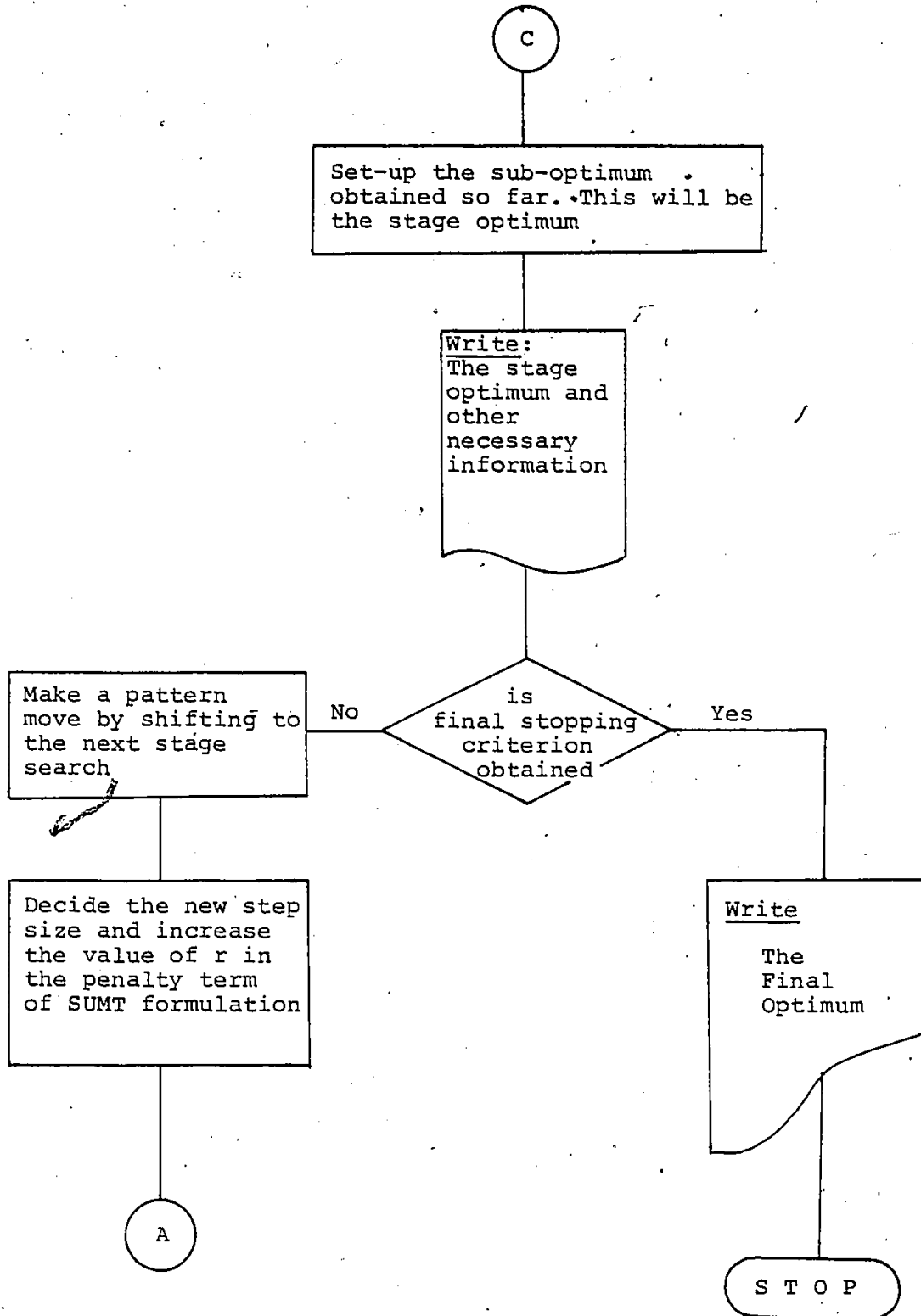
B

Start local search for the optimum.

Expand a distance D_i to each side of x_i

Calculate the original objective Function F_y





PAGE 0002

03/10/17

DATE = 79

MATH

FORTRAN IV G LEVEL 21

```

0004 CALL WEIGHTS(FH, MG, FG, MH, FH)
0005 TIER=0
0006 CALL PENAL(FG, FH, PNA1, PNA2)
C***** COMPUTE AN INITIAL VALUE OF R WHEN INPUT R VALUE IS .LE. 0.
0007 IF(R) 12, 12, 13
0008 R=0.0051*FY/(PNA1+PNA2)
0009 R=R/7A-0.00
C***** USE RATIO =4.0 *RWD INPUT RATIO VALUE IS .LE. 0.
0010 IF(RATIO) 14, 14, 15
0011 RATIO=4.000
0012 R=FY*REPNA1/R*(-0.3)*PLNA2
0013 WRITE(6,333) ITMAX, MAXP, ISIZE, ICUT
0014 JJJ FORMAT(77, 5X, 11MAX=, 15.5X, *MAXP=, 15.5X, *ISIZE=, 15.5X, *ICUT=,
0015 1, 15.77)
0016 WRITE(6, 1005) FY, FP, R, RATIO, R, INCUT, THEA
0017 WRITE(6, 1006) (1, FX(I), I, D(I), I=1, N)
0018 WRITE(6, 1007)
0019 IF(LDST-2) 50, 10, 10
C***** SELECT A FEASIBLE STARTING POINT WHEN INPUT INITIAL POINT IS NOT
C***** FEASIBLE SUBJECT TO IRREGULARITY CONSTRAINTS.
C***** MAKE EXPLORATORY MOVE FOR SELECT A FEASIBLE STARTING POINT.
0020 NUF=0
0021 DO 28 I=1, N
0022 FX(I)=X(I)+2.0000*(I)
0023 CALL WRES(FX, FY, FG, FH)
0024 CALL WEIGHT(TGH, MG, FG, MH, FH)
0025 IF(LDST-2) 44, 10, 10
0026 IF(STGH-TGH) 20, 20, 26
0027 FX(I)=FX(I)-4.0000*(I)
0028 CALL WRES(FX, FY, FG, FH)
0029 CALL WEIGHT(TGH, MG, FG, MH, FH)
0030 IF(LDST-2) 44, 22, 22
0031 IF(STGH-TGH) 24, 24, 26
0032 NUF=NUF+1
0033 GO TO 28
0034 STGH=TGH
0035 X(I)=FX(I)
0036 CONTINUE
0037 IF(NUF-N) 34, 30, 30
C***** CUT STEP SIZE FOR SELECTING A FEASIBLE STARTING POINT.
0038 DO 32 I=1, N
0039 DI(I)=DI(I)+0.570
0040 GO TO 16
C***** MAKE PATTERN MOVE FOR SELECTING A FEASIBLE STARTING POINT
0041 DO 36 I=1, N
0042 PX(I)=FX(I)+FX(I)
0043 CALL WRES(PX, FY, FG, FH)
0044 CALL WEIGHT(TGH, MG, FG, MH, FH)
0045 IF(STGH-TGH) 16, 16, 40
0046 DO 42 I=1, N
0047 X(I)=PX(I)
0048 IF(LDST-2) 44, 43, 43
0049 STGH=TGH
0050 GO TO 16
0051 DO 46 I=1, M
0052 O(I)=O(I)
0053 OX(I)=FX(I)
0054 DO 48 I=1, M
0055 WRITE(6, 1020)
0056 GO TO 11
0057 DO 49 I=1, M
0058 X(I)=OX(I)
0059 LDST=LDST
0060 CALL WRES(PX, FY, FG, FH)
C***** START TO MINIMIZE THE CURRENT P/ FUNCTION.
C***** MAKE EXPLORATORY MOVE FOR MINIMIZING THE P- FUNCTION
0061 NUF=0
0062 MCUT=1
0063 DO 51 I=1, 52, 52, 52, MCUT
0064 DO 52 I=1, 52, 52, 52, MCUT
0065 DO 51 I=1, 52, 52, 52, MCUT
0066 X(I)=X(I)+DI(I)
0067 CALL WRES(PX, FY, FG, FH)
0068 IF(LDST-1) 62, 62, 62

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MAIN

FURKAN IV G LEVEL 21

```

0132 53 (LY-FY) 55.55,68
0133 55 CALL BACK(X,X,Y,G,H)
0134 56 NUTD=NUITD+1
0135 57 NUTD=NUITD+1
C *****CHECK THE ITMAX IS EXCEEDED OR NOT IN (BACK) *****LST=1 MEANS
C THE RETURNED POINT IS INFEASIBLE.
0136 50 LUST=0
0137 51 IF(LST=1) 56,150,56
C *****CHECK THE ITMAX IS EXCEEDED OR NOT IN (BACK) *****LST=# MEANS
C MEANS THE ENTERD POINT NEAR - FEASIBLE.
0138 52 IF(CHECK-1) 54,140,140
0139 53 CALL PENAL(G,H,PENAL,PENAL)
0140 54 P-Y*(PENAL**(-0.5))+PENAL
0141 55 IF(P-FP) 98,68,68
0142 56 X(I)=FX(I)-D(I)
0143 LUST=0
0144 CALL QURES(X,Y,G,H)
0145 70 IF(LST=1) 90,80,70
0146 71 IF(Y-FY) 73,73,92
0147 73 CALL BACK(X,X,Y,G,H)
0148 74 NUTD=NUITD+1
0149 75 NUTD=NUITD+1
C *****CHECK THEC ITMAX IS EXCEEDED OR NOT IN (BACK) *****LST=# MEANS
C THE ENTERED POINT IS NEAR FEASIBLE
0150 74 LUST=0
0151 75 IF(CHECK-1) 82,140,140
0152 76 CALL PENAL(G,H,PENAL,PENAL)
0153 77 P-Y*(PENAL**(-0.5))+PENAL
0154 78 IF(P-FP) 88,88,88
0155 79 X(I)=FX(I)
0156 80 NUF=NUF+1
0157 81 GJ TO 99
0158 82 FY=Y
0159 83 FP=P
0160 84 NUTD=NUITD+1
0161 85 FX(I)=X(I)
0162 86 LUST=LST
0163 87 IF(MG) 94,94,90
0164 88 PJ 92 JJ=1,AG
0165 89 IF(JJ)=G(JJ)
0166 90 IF(MJ) 99,99,96
0167 91 DU 20 LK=LK+1
0168 92 CHECK THE STAGE STOPPING CRITERIUM IS SATISFID OR NOT
0169 93 IF(INCLV) INCLV 102,150,150
0170 94 CHECK THE STAGE STOPPING CRITERIUM IS SATISFID OR NOT
0171 100 IF(CHECK-1) 101,150,150
0172 101 CONTINUE
0173 102 DU 101 LK=LK+1
0174 103 X(I)=FX(I)+D(I)
0175 104 CALL QURES(X,Y,G,H)
0176 105 LUST=LST
0177 106 IF(LY-FY) 107,107,1104
0178 107 CALL BACK(X,X,Y,G,H)
0179 108 NUTD=NUITD+1
0180 109 NUTD=NUITD+1
0181 110 IF(LST=1) 1106,150,1106
0182 111 LUST=0
0183 112 IF(CHECK-1) 1107,140,140
0184 113 CALL PENAL(G,H,PENAL,PENAL)
0185 114 P-Y*(PENAL**(-0.5))+PENAL
0186 115 IF(P-FP) 115,1106,1108
0187 116 DU 1109 LK=LK+1
0188 117 X(I)=FX(I)-D(I)
0189 118 CALL QURES(X,Y,G,H)
0190 119 IF(LST=1) 1113,1113,1110
0191 120 IF(Y-FY) 111,111,1114
0192 121 CALL BACK(X,X,Y,G,H)
0193 122 NUTD=NUITD+1
0194 123 NUTD=NUITD+1
0195 124 IF(LST=1) 1112,150,1112
0196 125 LUST=0
0197 126 IF(CHECK-1) 1113,140,140
0198 127 CALL PENAL(G,H,PENAL,PENAL)
0199 128 P-Y*(PENAL**(-0.5))+PENAL
0200 129 IF(P-FP) 1115,1114,1114
0201 130 MCO1=J
0202

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```

0203 GO TO 51
0204 FY=9
0205 MCUT=1
0206 FX=1
0207 IF (X(1)=1) I=1,N
0208 IF (X(1)=1) I=1,N
0209 IF (X(1)=1) I=1,N
0210 IF (X(1)=1) I=1,N
0211 IF (X(1)=1) I=1,N
0212 IF (X(1)=1) I=1,N
0213 IF (X(1)=1) I=1,N
0214 IF (X(1)=1) I=1,N
0215 IF (X(1)=1) I=1,N
C*****CUT STEP SIZE FOR MINIMIZING THE P- FUNCTION**
105 DO 110 I=1,N
106 DO 110 J=1,N
107 MCUT=MCUT+1
108 CALL PENAT(FG,FH,PENAL,PENAZ)
109 IF (I*J) MCUT=106
110 IF (MCUT=1) 107,107,110
111 MCUT=2
112 R=2
113 CALL PENAT(FG,FH,PENAL,PENAZ)
114 FP=FP+R*PENAL+R*(-0.5)*PENAZ
115 INCU=INCU+1
116 MCUT=0
117 DO 109 I=1,N
118 DO 109 J=1,N
119 PD(I)=PD(I)+.000
120 D(I)=PD(I)
121 WRITE(6,1022) MCUT
122 IF (151Z) 2109,2109,51
2109 DO 2110 I=1,N
2110 D(I)=D(I)/FNDR
123 GO TO 51
124 IF (MCUT=INCU) 111,150,150
125 MCUT=J
C*****MAKE PATTERN MOVE FOR MINIMIZING THE P- FUNCTION.
126 DO 112 I=1,N
127 PX(I)=FX(I)+FX(I)
128 DX(I)=FX(I)
129 LOST=0
130 CALL DBRES(PX,Y,G,H)
131 IF (LST=1) 124,124,113
132 IF (Y-Y) 114,114,51
133 CALL BACK(PX,X,Y,G,H)
134 NUTD=NOITB+1
135 NUTP=NDP+1
136 CHECK THE IFMAX EXCEEDED OR NOT IN (BACK) *****LOST=1 MEANS THE
137 RETURNED POINT IS INFEASIBLE.
138 IF (LOST=1) 115,150,115
139 LOST=0
140 CHECK THE IFMAX IS EXCEEDED OR NOT IN (BACK) *****LOST.NE.1 MEANS
141 THE ENTERED POINT IS NEAR FEASIBLE.
142 IF (CHECK=1) 123,140,140
143 IF (SKIP=1) 124,48,48
144 CALL PENAT(G,H,PENAL,PENAZ)
145 P-Y+R*PENAL+R*(-0.5)*PENAZ
146 IF (FP) 128,48,48
147 NUTP=NDP+1
148 NUTP=NDP+1
149 DO 129 I=1,N
150 FX(I)=PX(I)
151 LOST=LUST
152 IF (X(1)=1) 133,133,131
153 DO 132 J=1,N
154 F(J)=G(J)
155 IF (H) 136,136,134
156 DO 135 K=1,MH
157 F(K)=H(K)
158 FY=9
159 FP=9
C*****CHECK THE STAGE STOPPING CRITERION IS SATISFIED OR NOT.
160 IF (MCUT=INCU) 150,150,150
161 IF (CHECK=1) 50,150,150
C*****CHECK THE IFMAX EXCEEDED PLINT (WHEN IT IS RETURNED FROM BACK)
162 IS BETTER OR NOT AND SET PROPER STAGE - OPTIMUM.
163 CALL DBRES(X,Y,G,H)

```

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```

0274 CALL PENAL(AM,PLM1,PLM2)
0275 IF (NOR) PENAL=0.5*PENAL2
0276 IF (NOR) PENAL=142.156.150
142 DO 114 J=1,N
144 FX(J)=X(J)
145 Y(J)=Y(J)
146 Z(J)=Z(J)
C ***** SET THE SUB OPTIMUM GOT BEFORE ENTERED TO BACK BE THE
C STAGE--OPTIMUM
150 NUPULL=0
151 DO 152 J=1,M
152 IF (NG(J)) 162,162,152
C ***** CHECK THE STAGE OPTIMUM IS FEASIBLE OR NOT
160 IF (LUST=1) 170,162,162
C ***** PULL BACK THE INFESIBLE STAGE--OPTIMUM INTO THE FEASIBLE REGION
163 DO 163 J=1,N
164 FX(J)=PULL*(X(J)-UX(J))+CX(J)
165 NUPULL=NUPULL+1
CALL UDRES(FX,FY,FG,FI)
167 NUPULL=0
168 DO 168 J=1,N
169 CALL UDRES(FX,FY,FG,FI)
170 LOST=0
CALL PENAL(FG,FI,PLM1,PLM2)
FP=FY+R*PENAL+R*(-0.5)*PENAL2
NUPULL=NUPULL+1
171 YSTOP=DABS(YSTOP-1.000)
CALL UDRES(FX,FY,FG,FI)
172 INPAT,NOCUT,YSTOP
WRITE(6,1006) (I,FX(I),I,0(I),J=1,N)
WRITE(6,1011)
173 IF (NG) 216,216,215
174 WRITE(6,1012) (J,FG(J),J=1,NG)
175 IF (MH) 218,218,217
176 WRITE(6,1013) (K,H(K),K=1,MH)
177 CALL OUTPUT(M,NG,MH)
178 WRITE(6,1007)
C ***** CHECK THE FINAL STOPPING CRITERION IS SATISFIED OR NOT
IF (VSTOP-(META) 210,230,220
220 IF (INDH-MAXPI) 221,232,232
C ***** STOP LAST SUB-OPTIMUM
221 DO 222 J=1,N
222 D(I)=PD(I)
C ***** SHUFF TO THE NEXT STAGE SEARCH.
R=0/RATIO
FP=FY+R*PENAL+R*(-0.5)*PENAL2
NUPULL=NUPULL+1
IF (NOR=5*MP) 224,224,223
223 MP=MP+1
224 IF (NDMP) 226,226,225
225 INCUTE=INCUTE+1
226 NUIP=0
MULT=1
NUIB=0
NUIP=0
ICHECK=0
NDEXP=0
NUPAT=0
NOCUT=0
IEND=0
ID=0
FNUN=NUR
B=0.000
MCUI=1
IDIFF=0
C ***** DECIDE THE INITIAL STEP-SIZE AND TOLERANCE LIMIT.

```

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```

0140 IF(1(CUT) 229,229,227
0141 DO 228 I=1,N
0142 D(I)=PD(I)/FNR
0143 248 D=0+0.500*D(I)
0144 B=B/7FN
0145 GU TU 50
0146 229 D=PH
0147 GU TU 50
0148 230 WRITE(6,1015)
0149 CALL OUTPUT( N,MC,MH)
0150 GU TU 234
0151 232 WRITE(6,1016) MAXP
0152 234 IOPM=IDPM+1
0153 IF(10PH-NUPH) 1,1,236
0154 236 STOP
0155 END

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MAIN

FURTHER IV G LEVEL 21

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*PTIONS IN EFFECT*, ID,EBODIC, SOURCE,NU,IST,NUDECK,LOAD,NUMAP
*PTIONS IN EFFECT*, NAME,MAIN, LINECNT = , LINECNT = 80
*STATISTICS*, SOURCE STATEMENTS = 355,PROGRAM SIZE = 13726
*STATISTICS*, NU DIAGNOSTICS GENERATED

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PAGE 0001

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MAIN

FURTHER IV G LEVEL 21

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```

SUBROUTINE WLGHTGH(MG,G,MH,IJ)
THIS SUBROUTINE COMPUTES THE TOTAL WEIGH OF VIOLATION
IN THE INEQUALITY CONSTRAINTS.
IMPLICIT REAL*8(A-H,O-Z)
DIMENSION G(125),MH(20)

```

```

1 DO 3 IK=1,MH
2 WGH=0.0
3 IF(G(IK)) 2,3
4 CONTINUE
5 DO 7 IH=1,MH
6 WGH=WGH+G(IH)*IH**2
7 CONTINUE
8 TWTGH=WGH*0.500
9 RETURN
END

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WEIGH

FURTHER IV G LEVEL 21

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*DEVIATIONS IN EFFECT* ID,ENDIC, SOURCE,NULIST,NUDECK,LOAD,NUMAP
*ADDITIONS IN EFFECT* NAME = WEIGH * LINECNT = 00
*STATISTICS* SOURCE STATEMENTS = 17,PROGRAM STZL = 050
*STATISTICS* NO DIAGNOSTICS GENERATED

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MAIN

FURTHAN 'IV G LEVEL -1

```

C.....
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C
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C
SUBROUTINE PENAL (G,M,ITER,AL,PENAZ)
THIS SUBROUTINE COMPUTES THE PENALTY TERMS FOR SORT FORMULATION
INPUT: REAL*8(A=N, D=2)
DIMENSION G(1:25),I(20)
COMMON /BLDOST/ N,NU,MI,ITER,ITMAX,ICHECK,IB,LLST
PENAL=0.000
PENAZ=0.000
IF(NG) 5,5,1
1 DO 9 I=1,NG
IF(G(I)) 2,2,4
**SET G(I)=0.000
2 G(I)=G(I)+DABS(I-.000/G(I))
4 PENAL=PENAL+DABS(I-.000/G(I))
5 IF(MI)=1,MI,0
6 PLN2=PENAL*ICHECK**2
9 CONTINUE
10 RETURN
END
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PENAL

FURTHAN 'IV G LEVEL -1

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*OPTIMUMS IN EFFECT* ID,EBCDIC, SOURCE, NOLIST, NUDECK, L, DAD, NDMAP
*OPTIONS IN EFFECT* NAME = PENAL , LINECT = 00
*STATISTICS* SOURCE STATEMENTS = 17, PROGRAM SIZE = 618
*STATISTICS* NU DIAGNOSTICS GENERATED

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0001 SUBROUTINE BACK(XD,Y,G,N)
0002 THIS SUBROUTINE PUTS THE FEASIBLE POINTS BACK INTO THE FEASIBLE
0003 OR NEAR FEASIBLE REGION
0004 IMPLICIT REAL*8(A-H,O-Z)
0005 DIMENSION XD(50),X(50),G(125),H(20),DI(50)
0006 COMMON /LOGS/ N,MG,MH,ITER,ITMAX,ICHECK,IB,LUST
0007 ITEMP=1
0008 DO 10 I=1,ITER
0009 FRAC=0.5000
0010 CALL WEIGH(IGH,MG,G,MH,H)
0011 IF (IGH) B,B,4
0012 DECREASE THE VALUE OF B IN RETURN
0013 IF (IB-IGH) 12,12,C
0014 B=0.7500*B
0015 IF (B) 10,10,10
0016 IF (IGH) 12,12,C
0017 MAKE EXPLURATORY MOVE FOR MINIMIZING TGH.
0018 DO 30 N=1,N
0019 XD(N)=XB(N)-FRAC*(ND)
0020 CALL UDRES(XD,Y,G,H)
0021 IF (LUST-2) 24,24,25
0022 IF (LUST-2) 24,24,25
0023 N=NTIP-NDI*PT
0024 GO TO 46
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*OPTIONS IN EFFECT: IN,LEUCDIC,SOURCE,NOLIST,NOCHECK,LOAD,NOHMAP
*OPTIONS IN EFFECT: NOHMAP,NOHMAP,NOHMAP,NOHMAP
*STATISTICS: SOURCE STATEMENTS = 1684
*STATISTICS: NO DIAGNOSTICS GENERATED

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SUBROUTINE CUST
THIS SUBROUTINE CALCULATE FOR EACH ZONE :
(1) TOTAL TREATMENT COST FOR ALL SOURCES
(2) CHANGE RATES FOR ALL SOURCES
(3) INITIAL CHANGE COST FOR ALL SOURCES
.....
IMPLICIT REAL*8(A-H,O-Z)
COMMON/AREA / F1,F2,PAY
COMMON/FRI/SMAXT(10),RHC(10),RHM(10),RPM(10),CF(10),JJ,PP
COMMON/LLST/TC,KC,YC,TRC,CC,TCC,TTC,HR,IL
DIMENSION TRC(10),CC(10),XC(10),YC(10),RH(10)
CUST_FUNCTIONS
C.....
CR(U,V,XI)=UPDEAP(-V*(1.000/(1.000-XI)))
CHMU(V,XI,WI)=U*(XPI-V*(1.000/(1.000-XI)))
SHCO(XI)=(0.3785*0)*((1.2500/(1.1000*(C2/3785-0.00)))+(1.000)/(14500)
+3.000*(U.71+XI-0.75*(XI+2)*0.5*(XI+5))+50000.0*(XI+100)
IC(10)=A1*(2+XI*X2)-DEXP(10.1125000)*(10+90.3475000)*((XI+11))
PIC(10)=90*(0.0+271.30*(XI+29.01)*0.5
SIC(10)=2700*(0.02000-3.000*(101+305.000))*0.6700011+
TIC(10)=300.000*(0.02000-3.000*(101+305.000))
TTC=0.000
DT 7 KL = 1, JJ
XI=AT(KL)
WI=WM(KL)
GU TU (1,2) ,KL
1 O=Q(KL)
V=XC(KL)
Y=YC(KL)
X2=PA(KL)
HM(KL)=CR(U,V,XI)
TRC(KL)=CHMU(V,XI,WI)
CC(KL)=CHMU(V,XI,WI)
TTC=TTC+TRC(KL)
TCC=TCC+CC(KL)
GU TU 7
2 UI=JF(KL)
V=YC(KL)
U=XC(KL)
HM(KL)=CR(U,V,XI)
F=CF(KL)*F1*(101+JF(KL))*XI+0.5-0.000
IF (X1-0.0-90-AND(X1LE,0.40) GU TU 5
TRC(KL)=TTC(KL)
CC(KL)=CHMU(V,XI,WI)
HM(KL)=CR(U,V,XI)
GU TU 6
4 TRC(KL)=PIC(KL)
CC(KL)=CHMU(V,XI,WI)
HM(KL)=CR(U,V,XI)
GU TU 8
5 IC(KL)=SIC(KL)
CL(KL)=CHMU(V,XI,WI)
HM(KL)=SHMU(KL,XI)
6 TTC=TTC+TRC(KL)
TCC=TCC+CC(KL)
7 CONTINUE
IL=JJ
IC=TTC+TCC
RETURN
END

```

OPTIMUMS IN EFFECT ID=CRDTC, SOURCE=NOI1ST,NUDLCK=LOAD,NUMAP
OPTIMUMS IN EFFECT NAME=CUST
STATISTICS SOURCE STATEMENTS = 57, PROGRAM SIZE = 2040
STATISTICS NO DIAGNOSTICS GENERATED

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Program Set-Up and Final Results for Example 1.

*SUNT * PROBLEMS

NU: UF XIJ) : 12
NO: OF GIJ) : 27
NO: OF HIJ) : 0

NO OF PROBLEMS ... 1

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```

//SUMT JOB (XXXXXXXXX.10.15.8207) 'HUUGY', CLASS=S JUB 117
//EXEC FORTGCLG,REGION=450K
//FORT.SYSIN DD *
//STEP WAS EXECUTED - COND CODE 0000
IEF1421 - STOP /FORT /START 79170.0310 CPU
IEF1421 - STOP /FORT /STOP 79170.0310 CPU
IEF1421 - STEP WAS EXECUTED - COND CODE 0000
IEF1421 - STOP /LKED /START 79170.0310 CPU
IEF1421 - STOP /LKED /STOP 79170.0310 CPU
//JOB.SYSIN DD *
//STEP WAS EXECUTED - COND CODE 0000
IEF1421 - STOP /GO /START 79170.0310 CPU
IEF1421 - STOP /GO /STOP 79170.0316 CPU
IEF1421 - JOB /SUMT /START 79170.0310 CPU
IEF1421 - JOB /SUMT /STOP 79170.0310 CPU
//SUMT JOB (XXXXXXXXX.10.15.8207) 'HUUGY', CLASS=S JUB 117
//EXEC FORTGCLG,REGION=450K
//FORT.SYSIN DD *
//STEP WAS EXECUTED - COND CODE 0000
IEF1421 - STOP /FORT /START 79170.0310 CPU
IEF1421 - STOP /FORT /STOP 79170.0310 CPU
IEF1421 - STEP WAS EXECUTED - COND CODE 0000
IEF1421 - STOP /LKED /START 79170.0310 CPU
IEF1421 - STOP /LKED /STOP 79170.0310 CPU
//JOB.SYSIN DD *
//STEP WAS EXECUTED - COND CODE 0000
IEF1421 - STOP /GO /START 79170.0310 CPU
IEF1421 - STOP /GO /STOP 79170.0316 CPU
IEF1421 - JOB /SUMT /START 79170.0310 CPU
IEF1421 - JOB /SUMT /STOP 79170.0310 CPU

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MATH

FURKAN IV G LEVEL 21

```

C *****
C SUBROUTINE HEADIN(N,MC,MH)
C *****
C TITLES SUBROUTINE IS FOR HEAD IN ADDITIONAL DATA
C USER SUPPLIES HIS OWN HEAD STATEMENT AND FORMAT
C ARGUMENTS N,MC,MH ARE NUMBERS OF VARIABLES OF INEQUALITY CONSTRAINTS
C AND OF EQUALITY CONSTRAINTS
C COMMON/BLDGRT/..... STATEMENT IS FOR TRANSFER DATA USE
C *****

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```

IMPLICIT REAL*8(A-H,O-Z)
REAL*8 LU,LA,MGO
COMMON/AREA 1/ F1,F2,PAY
COMMON/TRNSF /ALPH(10),B(10),BP(10),UR(10),T(10),LA(10),LA(10)
1,MGO(10),AK3(10),MN(10),IH(10),NZ
COMMON/BLDGRZ(10)

```

```

C *****
C *****
C *****

```

```

0007 NZ=J
0008 EPS=0.1000000
0009 PAY=0.3140001
0010 F1=1.055000
0011 F2=0.1854000
0012 IH(1)=0
0013 IH(2)=0
0014 IH(3)=0
0015 MH(1)=1
0016 MH(2)=1
0017 MH(3)=1
0018 ALPH(1)=3000000
0019 ALPH(2)=3000001
0020 ALPH(3)=3000001
0021 B(1)=0.20005
0022 B(2)=0.22005
0023 B(3)=0.24005
0024 B(4)=0.26005
0025 B(5)=0.28005
0026 B(6)=0.30005
0027 T(1)=200.000
0028 T(2)=200.000
0029 T(3)=200.000
0030 T(4)=200.000
0031 T(5)=200.000
0032 T(6)=200.000
0033 T(7)=200.000
0034 T(8)=200.000
0035 T(9)=200.000
0036 T(10)=200.000
0037 LA(1)=2000000
0038 LA(2)=2000000
0039 LA(3)=2000000
0040 DR(1)=500.000
0041 DR(2)=500.000
0042 DR(3)=500.000
0043 MGO(1)=2.5000
0044 MGO(2)=3.0000
0045 MGO(3)=2.0000
0046 RETURN
0047 END

```

03/10/17

DATE = 79170

READIN

FURKAN IV G LEVEL 21

```

*OPTIONS IN EFFECT* ID=ENCDDIC, SOURCE=MULIST, NUDECK=LOAD+NUMAP
*OPTIONS IN EFFECT* NAME=READIN, LINECNT=80
*STATISTICS* SOURCE STATEMENTS = 820
*STATISTICS* AD=PROGRAM SIZE =
*STATISTICS* NJ DIAGNOSTICS GENERATED

```

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PAGE 0001

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DATE = 79170

MAIN

21

FURTRAN IV G LEVEL

```

0001 C*****
      SUBROUTINE OUTPUT(NC,MH)
      C*****
      THIS SUBROUTINE IS FOR USER TO PRINT OUT ADDITIONAL INFORMATION
      C*****
      WANTED. ARGUMENTS N,NC,MH ARE NUMBERS OF VARIABLES, OF INEQUALIT
      C*****
      -Y CONSTRAINTS, AND OF EQUALITY CONSTRAINTS.
      C*****
      THE NEEDED DATA INFORMATION NEEDED DATA IN MAIN TO
      C*****
      CUMMOR /BL000/..... IS FOR TRANSFER NEEDED DATA IN MAIN TO
      C*****
      THE SUBROUTINE OUTPUT
      C*****
      USER SUPPLIES ALL NECESSARY FORMATS
      C*****
      IMPLICIT REAL*8(A-H,O-Z)
      CUMMOR/AREA3/CHG(10),CHR(10),TMC(10),UUD,DD,TIME,LS
      CUMMOR/DL000/G(125)
      REAL *4 BUD(500),DUT(500),TIME(500)
      DO 10 I=1,3
      WRITE(6,30) CHR(I)
      10 WRITE(6,20) I,CHG(I),TMC(I)
      20 FORMAT(5X,'CHARGING COST',I2,')=,DIS,6,10X,'TREATMENT COST (',I
      *2,')=,DIS,6,7)
      30 FORMAT(5X,'CHARGE RATE =,DIS,6,7)
      RETURN
      END

```

PAGE 0002

03/10/17

DATE = 79170

OUTPUT

21

FURTRAN IV G LEVEL

```

*OPTIONS IN EFFECT: IO, ERCDIC, SOURCE, NOLIST, NUDECK, LUAD, HOMAP
*OPTIONS IN EFFECT: NAME = OUTPUT, LUNECNT = 80
*STATISTICS* SOURCE STATEMENTS = 12, PROGRAM SIZE = 504
*STATISTICS* NO DIAGNOSTICS GENERATED

```

VERY POOR COPY

DATE 7/9/70 03/10/77

MATH

FURMAN IV G LEVELL J1

```

0001 ***** SUBROUTINE LURBS(X,Y,Z,H) *****
0002 ***** THIS SUBROUTINE COMPUTES OBJ. AND CONSTRAINT VALUES. *****
0003 ***** USER SHOULD SUPPLY ALL NECESSARY STATEMENTS IN THE FURN ** *****
0004 ***** V. FUNCTION OF X(I), FOR OBJECTIVE FUNCTION. *****
0005 ***** J FROM 1 TO M. FOR CONSTRAINTS G(J),GT, 0.0. *****
0006 ***** K FROM 1 TO M1. FOR CONSTRAINTS H(K),HO,0.0. *****
0007 ***** THESE STATEMENTS IN THE BLOCK BELOW LINED BY *****
0008 ***** APPLICIT REAL*8(A-H,U-Z) *****
0009 DIMENSION X(50),G(125),H(20) *****
0010 REAL*4 W(500),DU(500),TIME(500) *****
0011 REAL*8 MIND(10),LU,LB,MIN,MINM,MAXDU(10),K3,LAI,LA,MCD *****
0012 COMMON/IRNSF /ALPH(10),WP(10),MP(10),OR(10),T(10),LA(10) *****
0013 I,MCD(10),AK(10),MN(10),IN(10),NZ *****
0014 COMMON/ALIST/TC,XC,YC,TRC,CC,TCC,TR,HR,IL *****
0015 COMMON/FRI/SU,XI(10),XII(10),MB(10),APM(10),CP(10),JJ,MM *****
0016 COMMON/AREA2/T1,ALPH1,OR1,TIME,K3,LAI,NT *****
0017 COMMON/AREA1/CHG(10),CHK(10),TMC(10),BUD,DO,TIME,LS *****
0018 COMMON /BLDGY /M,MC,MN1,TER,ITMAX,LCHECK,IB,LCST *****
0019 COMMON /BLDGR /G(10) *****
0020 FURMAT(JK,25H**THE ITERATION EXCEEDED .15,1P*) *****
0021 *****
0022 SUM=0.000 *****
0023 LI=1 *****
0024 KU=0 *****
0025 LS=1 *****
0026 LU=5.000 *****
0027 CU=0.000 *****
0028 PU=0.05000 *****
0029 KK=1 *****
0030 MX=0 *****
0031 NI=0 *****
0032 NC=0 *****
0033 DU 9 J=1,NZ *****
0034 NI=N1+MN(J)+IN(J) *****
0035 NC=NC+2*MN(J)+IN(J) *****
0036 J CUNTINUE *****
0037 NL=N1 *****
0038 NC=NC+J *****
0039 DU 10 J=1,NZ *****
0040 JJ=MN(J)+IN(J) *****
0041 MM=HN(J) *****
0042 LL=LN(J) *****
0043 TT=T(J) *****
0044 ALPHA=ALPH(J) *****
0045 TIME=TIM(J) *****
0046 K=AK(J) *****
0047 LAI=LA(J) *****
0048 NT=TIM(J)+25.0 *****
0049 OR1=OR(J) *****
0050 IF(MN(J).LE.0) GU TU J1 *****
0051 ***** ARRANGE DATA FOR ORGANIC WASTES FROM MUNICIPAL SOURCES *****
0052 MZ=HN(J) *****
0053 NZ=LN(J) *****
0054 NI=MX+I *****
0055 MX1=MX+HN(J) *****
0056 K=1 *****
0057 DO 11 L=NI,MX1 *****
0058 XT(K)=X(L) *****
0059 WJ(K)=W(L) *****
0060 WPM(K)=MP(L) *****
0061 OF(K)=MCD(L) *****
0062 K=K+1 *****
0063 IF(K.GT.M2) GU TU 22 *****
0064 ***** ARRANGE DATA FOR PHOSPHORUS WASTE FROM MUNICIPAL SOURCES *****
0065 11 CONTINUE *****
0066 22 NA=NI+1 *****
0067 MA=MN(J)+NI *****
0068 H=1 *****
0069 DO 12 I=NA,MA *****
0070 XII(I)=X(I) *****
0071 M=MI+1 *****
0072 IF(I4.GT.M2) GU TU J1 *****
0073 *****
0074 12 CONTINUE *****

```

VERY POOR COPY

```

0069 JJ NI=NI+MN(J)
0070 C ARRANGE DATA FOR ORGANIC WASTE FROM INDUSTRIAL SOURCES
0071 1J MA=MX+MN(J)
0072 IF (IN(J),LE,0) GO TO 55
0073 MXI=MX+I
0074 MAXI=MX+IN(J)
0075 MIA=MI(J)+I
0076 MIE=MI(J)+IN(J)
0077 K=MNA
0078 DU 14 L=MXI,MAXI
0079 XI(K)=X(I)
0080 WM(K)=WB(L)
0081 OF(K)=MOD(L)
0082 K=K+1
0083 IF (K,GT,MNI) GO TO 44
0084 C CONTINUE
0085 44 ARRANGE DATA FOR PHOSPHORUS WASTE FROM INDUSTRIAL SOURCES
0086 IF (KK,EQ,1) GO TO 55
0087 MA=NI+1
0088 MAI=NI+IN(J)
0089 M=MNA
0090 DU 16 I= MA,MAI
0091 XH(I)=X(I)
0092 WPM(I)=WP(I)
0093 M=M+1
0094 IF (M,GT,MNI) GO TO 55
0095 C CONTINUE
0096 16 CALCULATE THE CONCENTRATION LEVELS OF DU,DU,PCA AND FND
0097 C THE MINIMUM POINT OF DU AT THE ZONE
0098 C CALL NUMRIC (MINN,LU,PH,LU,CU,PU)
0099 MINDU(J)=MINM
0100 MAXDD(J)=LB
0101 PHIS(J)=PH
0102 DU 18 I=1,JJ
0103 XC(I)=X(INC)
0104 YC(I)=X(NL,INC)
0105 NC=NC+1
0106 C CONTINUE
0107 18 CALCULATE ALL THE COST TERMS
0108 CALL CUST
0109 SWS=SWATC
0110 KU=KOTIL
0111 K=1
0112 DU 20 L=L1,KU
0113 CMI(L)=MI(L)
0114 CMI(L)=CCK(L)
0115 TAC(L)=TRC(K)
0116 K=K+1
0117 IF (K,GT,IL) GO TO 77
0118 C CONTINUE
0119 77 L=LI
0120 10 CONTINUE
0121 C ***** STATEMENT NUMBERS 1,2,3,4,5,6,7,8,100 HVE BEEN USED
0122 C ***** FOR DEFINITION THE DECISION VARIABLES
0123 C ***** THE VALUES ARE DECLARATED BY PUTTING A,C IN COLUMN 1.
0124 C ***** IF THE OUTPUT OF THESE STATEMENTS ARE NEEDED, REWAVE THE YC+.
0125 C ***** (X(1),X(11),X(15),X(16)
0126 C ***** (X(1),X(11),X(15),X(16)
0127 C ***** FOR DEFINITION
0128 C ***** WJ, FUKTIUN
0129 C ***** Y-SUM
0130 C ***** EQUALITY CONSTRAINTS
0131 G(1)=-MAXDU(1)+15.000
0132 G(2)=-MAXDD(2)+15.000
0133 G(3)=-MAXDU(3)+15.000
0134 G(4)=MINDU(1)-5.000
0135 G(5)=MINDU(2)-5.000
0136 G(6)=MINDU(3)-5.000
0137 G(7)=-PHIS(1)+0.1000
0138 G(8)=-PHIS(2)+0.1000
0139
0140
0141
0142

```

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PAGE 0003

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DATE = 79170

DIRMS

FURTRAN IV G LEVEL 21

```

0122      G(10)=PMS(3)*0.1000
0123      G(11)=X(1)
0124      G(12)=X(2)
0125      G(13)=X(3)
0126      G(14)=1.0000-X(1)
0127      G(15)=1.0000-X(2)
0128      G(16)=X(4)
0129      G(17)=X(5)
0130      G(18)=X(6)
0131      G(19)=1.0000-X(4)
0132      G(20)=1.0000-X(5)
0133      G(21)=1.0000-X(6)
0134      G(22)=X(7)
0135      G(23)=X(8)
0136      G(24)=X(9)
0137      G(25)=X(10)
0138      G(26)=X(11)
0139      G(27)=X(12)
0140      LUST=0
0141      ICHK=ILR+1
0142      IF (ITER-ITMAX) J1,2
0143      *****QUIP ***** THE MESSAGE OF ITMAX EXCEEDED.
0144      1 WR LK(4,100) ITMAX
0145      2 CHECK FOR THE VIOLATION TO INEQUALITY CONSTRAINTS
0146      3 IF (MG) 4,4
0147      4 0171=1.4G
0148      5 LG(1)=5.0,7
0149      6 LUST=2
0150      7 GO TO 7
0151      8 IN=1
0152      9 CONTINUE
0153      0 RETURN
0154      END
0155

```

03/10/17

DATE = 79170

DIRMS

FURTRAN IV G LEVEL 21

```

*****OPTIONS IN EFFECT* LD,EDCLOC,SOURCE,NUL,IST,NODECK,LOAD,NUMAP
*****OPTIONS IN EFFECT* NAME,DIRMS
*****STATISTICS* SOURCE STATEMENTS = 133,PROGRAM SIZE = 3100
*****STATISTICS* NO DIAGNOSTICS GENERATED
*****STATISTICS* NO DIAGNOSTICS THIS STEP

```

F08-LEVEL LINKAGE EDITOR OPTIONS SPECIFIED LET,LIST
 DEFAULT OPTION(S) USED - SIZE=(1,3,120,40960)
 *****MAIN DOES NOT EXIST BUT HAS BEEN ADDED TO DATA SET

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```

*****
**THE ITERATION EXCEEDED 50.
**OPTIMUM ( 1 )
FY = 0.122330 07.FP = 0.122330 07. K = 0.2384D-03. ITER = 52.
NUIT = 0.638. NUB = 0. NUP = 0. NUDP = 0
MUEXP = 2. MOPAT = 0. NUCUT = 0.
YSTOP = 0.50316D-07.
X( 1 ) = 0.500700 00. ( 1 ) = 0.7692310-03 .
X( 2 ) = 0.800760 00. ( 2 ) = 0.7692310-03 .
X( 3 ) = 0.780070 00. ( 3 ) = 0.7692310-03 .
X( 4 ) = 0.809920 00. ( 4 ) = 0.7692310-03 .
X( 5 ) = 0.989920 00. ( 5 ) = 0.7692310-03 .
X( 6 ) = 0.909240 00. ( 6 ) = 0.7692310-03 .
X( 7 ) = 0.139240 00. ( 7 ) = 0.7692310-03 .
X( 8 ) = 0.139240 00. ( 8 ) = 0.7692310-03 .
X( 9 ) = 0.139240 00. ( 9 ) = 0.7692310-03 .
X( 10 ) = 0.200760 00. ( 10 ) = 0.7692310-03 .
X( 11 ) = 0.106340 00. ( 11 ) = 0.7692310-03 .
X( 12 ) = 0.1255130 00. ( 12 ) = 0.7692310-03 .

```

```

**CONSTRAINTS **
C( 1 ) = 0.6090510 01 .
C( 2 ) = 0.5866230 01 .
C( 3 ) = 0.7190560 01 .
C( 4 ) = 0.7564310 00 .

```

```

C( 5 ) = 0.4496550 00 .
C( 6 ) = 0.4055790 00 .
C( 7 ) = 0.3273370-01 .
C( 8 ) = 0.3133300-01 .
C( 9 ) = 0.1905300-01 .
C( 10 ) = 0.5007060 00 .
C( 11 ) = 0.6800760 00 .
C( 12 ) = 0.7800760 00 .
C( 13 ) = 0.4192940 00 .
C( 14 ) = 0.3199240 00 .
C( 15 ) = 0.2199240 00 .
C( 16 ) = 0.8899240 00 .
C( 17 ) = 0.9899240 00 .
C( 18 ) = 0.9099240 00 .
C( 19 ) = 0.1100760 00 .
C( 20 ) = 0.1007500-01 .
C( 21 ) = 0.9007560-01 .
C( 22 ) = 0.1399240 00 .
C( 23 ) = 0.1399240 00 .
C( 24 ) = 0.1399240 00 .
C( 25 ) = 0.2800760 00 .
C( 26 ) = 0.1063460 00 .
C( 27 ) = 0.1255130 00 .
CHARGE RATE = 0.7174600-01

```

```

CHARGING COST ( 1 ) = 0.219040 0E TREATMENT COST ( 1 ) = 0.2812460 00
CHARGE RATE = 0.5370700-01
CHARGING COST ( 2 ) = 0.137970 0E TREATMENT COST ( 2 ) = 0.3184140 00
CHARGE RATE = 0.118040-01
CHARGING COST ( 3 ) = 0.255650 0E TREATMENT COST ( 3 ) = 0.2405320 00

```

*****THE ABOVE RESULTS ARE THE FINAL OPTIMUM*****

VERY POOR COPY

Program Set Up and Results for Example 2

***** PROBLEMS *****

NO. OF X(I) : 13
NO. OF C(J) : 15
NO. OF H(K) : 0

NO. OF PARAMETERS : 1

VERY POOR COPY


```

//JOBT  EXEC  (XXXXXXXXXX(10.15.00007) *MUJCV) CLASSES  JOBT 105
//EQUIT  SYSIN  DD
//EQUIT  EXEC  EXECUTED - COBJ CODE 0000
IEF1121  - STOP WAS  / START 79197.1816
IEF3741  STEP /JOB  / STJP 79197.1816 CPU
IEF1121  - STOP WAS  / EXECUTED - COBJ CODE 0000
IEF1121  - STOP /KED  / START 79197.1816
IEF3741  STEP /JOB  / STJP 79197.1816 CPU
//SO.SYSIN  DD
IEF1121  - STOP WAS  EXECUTED - COBJ CODE 0000
IEF3731  STEP /JOB  / STJP 79197.1816 CPU
IEF3741  STEP /JOB  / STJP 79197.1816 CPU
IEF3751  JOB /JOB  /
IEF3761  JOB /JOB  /

```

041 14.2751C MAIN 90K LCS OK
 021 01.0051C MAIN 131K LCS OK
 041M 34.9051C MAIN 60K LCS OK
 041M 50.2251C

VERY POOR COPY

0001 ***** SUBROUTINE DBRES(K,Y,G,H) *****

0002 ***** THIS SUBROUTINE COMPUTES OBJ. AND CONSTRAINT VALUES *****

0003 ***** USER SHOULD SUPPLY ALL NECESSARY DATA IN COMMON BLOCKS *****

0004 ***** Y2... FUNCTION OF X TO BE MINIMIZED *****

0005 ***** G... FUNCTION OF X TO BE MAXIMIZED *****

0006 ***** U... FUNCTION OF X TO BE MINIMIZED *****

0007 ***** U... FUNCTION OF X TO BE MAXIMIZED *****

0008 ***** U... FUNCTION OF X TO BE MINIMIZED *****

0009 ***** U... FUNCTION OF X TO BE MAXIMIZED *****

0010 ***** U... FUNCTION OF X TO BE MINIMIZED *****

0011 ***** U... FUNCTION OF X TO BE MAXIMIZED *****

0012 ***** U... FUNCTION OF X TO BE MINIMIZED *****

0013 ***** U... FUNCTION OF X TO BE MAXIMIZED *****

0014 ***** U... FUNCTION OF X TO BE MINIMIZED *****

0015 ***** U... FUNCTION OF X TO BE MAXIMIZED *****

0016 ***** U... FUNCTION OF X TO BE MINIMIZED *****

0017 ***** U... FUNCTION OF X TO BE MAXIMIZED *****

0018 ***** U... FUNCTION OF X TO BE MINIMIZED *****

0019 ***** U... FUNCTION OF X TO BE MAXIMIZED *****

0020 ***** U... FUNCTION OF X TO BE MINIMIZED *****

0021 ***** U... FUNCTION OF X TO BE MAXIMIZED *****

0022 ***** U... FUNCTION OF X TO BE MINIMIZED *****

0023 ***** U... FUNCTION OF X TO BE MAXIMIZED *****

0024 ***** U... FUNCTION OF X TO BE MINIMIZED *****

0025 ***** U... FUNCTION OF X TO BE MAXIMIZED *****

0026 ***** U... FUNCTION OF X TO BE MINIMIZED *****

0027 ***** U... FUNCTION OF X TO BE MAXIMIZED *****

0028 ***** U... FUNCTION OF X TO BE MINIMIZED *****

0029 ***** U... FUNCTION OF X TO BE MAXIMIZED *****

0030 ***** U... FUNCTION OF X TO BE MINIMIZED *****

0031 ***** U... FUNCTION OF X TO BE MAXIMIZED *****

0032 ***** U... FUNCTION OF X TO BE MINIMIZED *****

0033 ***** U... FUNCTION OF X TO BE MAXIMIZED *****

0034 ***** U... FUNCTION OF X TO BE MINIMIZED *****

0035 ***** U... FUNCTION OF X TO BE MAXIMIZED *****

0036 ***** U... FUNCTION OF X TO BE MINIMIZED *****

0037 ***** U... FUNCTION OF X TO BE MAXIMIZED *****

0038 ***** U... FUNCTION OF X TO BE MINIMIZED *****

0039 ***** U... FUNCTION OF X TO BE MAXIMIZED *****

0040 ***** U... FUNCTION OF X TO BE MINIMIZED *****

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PAUL 0002

01/14/70

DATE = 7/105

DIRLS

FURFRAN IV G LEVEL :1

```

0003 JJ NI=MIN(MI(J))
0006 C ARRANGE DATA FOR ORGANIC WASTE FROM INDUSTRIAL SOURCES
0007 13 MA=MAX(MI(J))
0008 IF (INI(J).LE.0) GJ TU 55
0009 MAXI=MAX(I)
0010 AHA=MIN(J)
0011 PHI=MIN(J)+IN(J)
0012 K=MNA
0013 DU 14 L=MAXI,MAXI
0014 XT(K)=X(L)
0015 MIA(K)=#B(L)
0016 QI(K)=AGD(L)
0017 K=K+1
0018 IF (K.GT.MNI) GG TU 44
0019 C CHANGE DATA FOR PHOSPHORUS WASTE FROM INDUSTRIAL SOURCES
0020 14 AVERAGE DATA FOR PHOSPHORUS WASTE FROM INDUSTRIAL SOURCES
0021 44 MA=MAX(MI(J)) GU TU 55
0022 MI=MIN(MI(J))
0023 M=MNI+IN(J)
0024 DU 15 I=NA,NAI
0025 XI(M)=X(I)
0026 WUM(M)=W(I)
0027 MEM(M)
0028 *IF (M.GT.MNI) GO TU 55
0029 C CONTINUE
0030 C CALCULATE THE CONCENTRATION LEVELS OF BOD,DD,PO4 AND FINE.
0031 C THE MINIMUM POINT OF DO AT THE ZONE
0032 55 CALL NUMRIC (MIH,LU,PH,LU,CU,PU)
0033 MIND(J)=MINH
0034 MAX(DO(J))=PH
0035 PLUS(J)=PH
0036 C ARRANGE DATA FOR THE COEFFICIENTS OF THE CHARG
0037 DU 16 I=1,JJ
0038 XC(I)=X(NC)
0039 YC(I)=X(NL+NC)
0040 NC=NC+1
0041 C CONTINUE
0042 18 CALCULATE ALL THE COST TERMS
0043 C CALL COST
0044 SUM=SUM+TC
0045 KU=XO*IL
0046 K=1
0047 DU 20 L=L1,KU+1
0048 CHG(L)=HR(K)
0049 CHG(L)=CC(K)
0050 TAC(L)=TRC(K)
0051 K=K+1
0052 IF (K.GT.IL) GU TU 77
0053 20 CONTINUE
0054 77 L=L+1
0055 10 CONTINUE
0056 C ***** STATEMENT NUMBERS 1,2,3,4,5,6,7,8,100 HVE BEEN USED *****
0057 C ***** FUNCTION DEFINITION *****
0058 C ***** WRITE STATEMENTS ARE DEACTIVATED BY PUTTING A'C' IN COLUMN 1. *****
0059 C ***** IF THE OUTPUT OF THESE STATEMENTS ARE WANTED, REMOVE THE 'C'. *****
0060 C ***** IF(0,1000) (X(1),I=1,6) *****
0061 C1000 FORMAL(, X(1))=.6015(.7)
0062 C ***** FUNCTION DEFINITION *****
0063 C ***** *****
0064 C ***** G1J. FUNCTION. *****
0065 C ***** *****
0066 C ***** Y-SUM *****
0067 C ***** *****
0068 C ***** EQUALITY CONSTRAINTS *****
0069 C ***** *****
0070 G(1)=-MAX(DO(1))+15.000
0071 G(2)=-MAX(DO(2))+12.000
0072 G(3)=-MAX(DO(3))+15.000
0073 G(4)=-MIND(1)-4.5000
0074 G(5)=-MIND(2)-5.000
0075 G(6)=-MIND(3)-4.000
0076 G(7)=-PH(1)+0.1000
0077 G(8)=-PH(1)+0.1000
0078
0079
0080
0081
0082
0083
0084
0085
0086
0087
0088
0089
0090
0091
0092
0093
0094
0095
0096
0097
0098
0099
0100
0101
0102
0103
0104
0105
0106
0107
0108
0109
0110
0111
0112
0113
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0115
0116
0117
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0120
0121

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PAGE 0003

01/14/50

DATE = 791053

00005

FURTHER IN G LEVEL -1

```

0121 G(10)=-PHS(3)+0.1000
0122 G(11)=X(1)
0123 G(12)=X(2)
0124 G(13)=X(3)
0125 G(14)=X(4)
0126 G(15)=X(5)
0127 G(16)=1.0000-X(1)
0128 G(17)=1.0000-X(2)
0129 G(18)=1.0000-X(3)
0130 G(19)=1.0000-X(4)
0131 G(20)=X(6)
0132 G(21)=X(7)
0133 G(22)=X(8)
0134 G(23)=1.0000-X(6)
0135 G(24)=1.0000-X(7)
0136 G(25)=1.0000-X(8)
0137 G(26)=X(9)
0138 G(27)=X(10)
0139 G(28)=X(11)
0140 G(29)=X(12)
0141 G(30)=X(13)
0142 G(31)=X(14)
0143 G(32)=X(15)
0144 G(33)=X(16)
0145 G(34)=X(17)
0146 G(35)=X(18)
0147 LUST=0
0148 ITER=1
0149 IF(ITER-ITMAX) J,I,2
0150 C*****OUTPUT THE MESSAGE IF ITMAX EXCEEDED.
0151 1 WRITE(6,100) ITMAX
0152 2 ICHECK=1
0153 C*****CHECK FOR THE VIOLATION TO INEQUALITY CONSTRAINTS.
0154 3 ID=0
0155 IF(ING) D,D,4
0156 4 DU 7 I=1,XG
0157 5 IF(G(I)) S,0,7
0158 6 LUST=7
0159 7 CU ID 7
0160 8 IN=1
0161 9 CONTINUE
0162 A RETURN
0163 END

```

VERY POOR COPY

REP OPT 1404... (15)
PRNT = 9.265790 107.53 - 0.2645740 07.44 - 0.95373-0.1. ITER = 0.2,
MUNIT = 0.09, NUNIT = 0.1, NUNIT = 2, NUNIT = 9
MUNIT = 1, NUNIT = 0.1, NUNIT = 0.1
YSTEP = 0.1907222D-07

X(1)	0.1777090	00.	0.1111133	-04
X(2)	0.420113	00.	0.1111133	-04
X(3)	0.0706493	00.	0.1111133	-04
X(4)	1.270592	00.	0.1111133	-04
X(5)	0.1276490	00.	0.1111133	-04
X(6)	0.850363	00.	0.1111133	-04
X(7)	0.430660	00.	0.1111133	-04
X(8)	0.4236883	00.	0.1111133	-04
X(9)	0.406360	-01.	0.1111133	-04
X(10)	0.240813	00.	0.1111133	-04
X(11)	0.056160	-01.	0.1111133	-04
X(12)	0.1440813	00.	0.1111133	-04
X(13)	0.2450813	00.	0.1111133	-04
X(14)	0.549390	00.	0.1111133	-04
X(15)	0.592593	00.	0.1111133	-04
X(16)	0.1344500	00.	0.1111133	-04
X(17)	0.219393	00.	0.1111133	-04
X(18)	0.1676063	00.	0.1111133	-04

CONSTRAINTS

GC(1)	0.027633	01.
GC(2)	0.778963	01.
GC(3)	0.1864129	01.
GC(4)	0.1271280	01.
GC(5)	0.7329315	00.
GC(6)	0.7636220	00.
GC(7)	0.1930583	-01.
GC(8)	0.1805033	-01.
GC(9)	0.181420	-01.
GC(10)	0.7777090	00.
GC(11)	0.0426113	00.
GC(12)	0.0076390	00.
GC(13)	0.0276093	00.
GC(14)	0.1270893	00.
GC(15)	0.222113	00.
GC(16)	0.3571093	00.
GC(17)	0.0924113	00.
GC(18)	0.1724113	00.
GC(19)	0.0826063	00.
GC(20)	0.0826063	00.
GC(21)	0.0826063	00.
GC(22)	0.206063	00.
GC(23)	0.1419190	-01.
GC(24)	0.7939190	-01.
GC(25)	0.7939190	-01.
GC(26)	0.4280813	-01.
GC(27)	0.3280813	00.
GC(28)	0.056063	-01.
GC(29)	0.1410013	00.
GC(30)	0.0430613	00.
GC(31)	0.0430613	00.
GC(32)	0.0430613	00.
GC(33)	0.1349393	00.
GC(34)	0.1219393	00.
GC(35)	0.3626063	00.

TREATMENT COST (1) = 0.4790072D 00

TREATMENT COST (2) = 0.6143174D 00

TREATMENT COST (3) = 0.2275333D 00

TREATMENT COST (4) = 0.2705040D 00

TREATMENT COST (5) = 0.8250027D 00

CHARGE COST (1) = 0.4084341D 05
CHARGE RATE = 0.2958720D -01

CHARGE COST (2) = 0.8876700D 05
CHARGE RATE = 0.1234488D -01

CHARGE COST (3) = 0.6760091D 05
CHARGE RATE = 0.0074748D -01

CHARGE COST (4) = 0.1039721D 05
CHARGE RATE = 0.0976295D -01

CHARGE COST (5) = 0.5017438D 05
CHARGE RATE = 0.1219393D 00

VITA AUCTORIS

- 1946 Born on December 23rd, Mansoura, Egypt.
- 1968 Graduated from Cairo University, Egypt with Bachelor Degree of Science in Mechanical Engineering.
- 1968 Worked as Production Engineer at the Imperial British Tobacco Company, Giza, Egypt.
- 1974 Received Diploma in Mathematical Statistics from Cairo University, Egypt.
- 1975 Completed Degree of Master of Science in Production Engineering from Cairo University, Egypt.
- 1976 Held a teaching and research assistantship at the University of Windsor, Ontario.
- 1978 Employed as an Industrial Engineer, with Pyramid Homes Company, Windsor, Ontario.
- 1979 Candidate for Master of Applied Science Degree in Industrial Engineering.