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## Scaling from single-point sap velocity measurements to stand transpiration in a multi-species deciduous forest: Uncertainty sources, stand structure effect, and future scenarios

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## 1 ABSTRACT

2	A major challenge in studies estimating stand water use in mixed-species forests is how to
3	effectively scale data from individual trees to the stand. This is the case for forest ecosystems in the
4	northeastern USA where differences in water use among species and across different size classes have
5	not been extensively studied, despite their relevance for a wide range of ecosystem services. Our
6	objectives were to assess the importance of different sources of variability on transpiration upscaling
7	and explore the potential impacts of future shifts in species composition on forest water budget. We
8	measured sap velocity in five tree species (Fagus grandiflora, Acer rubrum, A. saccharum, Betula
9	alleghaniensis, B. papyrifera) in a mature and young stand in NH (USA). Our results showed that the
10	greatest potential source of error was radial variability and that tree size was more important than
11	species in determining sap velocity. Total sapwood area was demonstrated to exert a strong controlling
12	influence on transpiration, varying depending on tree size and species. We conclude that the effect of
13	potential species shifts on transpiration will depend on the sap velocity, determined mainly by radial
14	variation and tree size, but also on the sapwood area distribution in the stand.

## 15 KEY WORDS

Sap velocity, stand transpiration, sapwood area, northern hardwood forest, diffuse-porous species,
Heat Ratio Method

#### 18 INTRODUCTION

Forest ecosystems in the northeastern USA provide important ecosystem services to dense population centers in the region, including regulation of water quality and quantity (Campbell et al. 2011), carbon sequestration and storage (Raciti et al. 2012), diverse wood products for timber, energy, and pulp, as well as recreational opportunities. The biogeochemical cycles of nutrients and carbon have been well studied in the region's forests (Likens 2013), as have the hydrological effects of forest management (Hornbeck et al. 1993). However, much less attention has focused on understanding the patterns of water use in these forests.

26 Relative to arid regions and intensive forestry plantations, tree water relations have received less 27 attention in humid temperate forests, where precipitation is plentiful, forests are slow-growing, 28 evapotranspiration is strongly energy-limited, and water stress is relatively infrequent. The diffuse-29 porous hardwoods that dominate the region's forests are generally considered to be relatively poorly 30 adapted to moisture stress (Pederson et al. 2014). Despite the general perception that water is abundant 31 in the region and thus transpiration rates approximate potential evaporation (Zhang et al. 2004), 32 emerging evidence suggests greater limitation on vegetation water use in humid regions than previously 33 thought (Brzostek et al. 2014). Together with the forecasted increases in extended dry periods in the 34 Northeastern USA (Hayhoe et al. 2007), this underscores the need for more in-depth analysis of stand 35 level water use patterns by temperate mesic forests. Over longer time scales, changes in climate 36 variables may produce shifts towards species more adapted to hot, dry summers (Mohan et al. 2009). In 37 turn, changes in forest species composition, structure, and development stage have been widely shown 38 to significantly alter stand water use patterns and, hence, streamflow response at watershed scales 39 (Hornbeck et al. 1993). Thus, information about water use patterns by diverse tree species and the 40 consequences for stand transpiration is critical to assessing potential impacts of environmental change 41 drivers on watershed hydrology and guiding forest management and climate change adaptation 42 practices (Grant et al. 2013).

43 A challenge to estimating stand level water use in highly heterogeneous forests such as the 44 northern hardwoods is how to effectively scale data from individual trees to the stand (Wullschleger et 45 al. 2001). Scaling water fluxes requires an appropriate sampling design to effectively capture stand 46 structure and spatial distribution associated with trees of differing size, dimension, canopy position, leaf 47 area and species (Köstner et al. 1998). In particular, much scientific debate has focused on understanding the relative importance of species identity versus stand structure in controlling stand transpiration. 48 49 While some studies suggest that tree size rather than species is the main determinant of stand 50 transpiration (Meinzer et al. 2001, Wullschleger et al. 2001, McJannet et al. 2007), other studies point to 51 substantial interspecific variability in water use (Granier et al. 1996, Dierick and Hölscher 2009, Cavaleri 52 and Sack 2010). As size is often directly correlated with canopy position, tree size and canopy dominance 53 may be confounded when analyzing different scaling approaches (McGill et al. 2006). Specifically, tree 54 size (e.g., diameter at breast height -DBH, sapwood depth) is typically used as a scaling parameter as it is 55 usually positively related to sap velocity (Meinzer et al. 2001, Jung et al. 2011). Understanding the 56 relative importance of species differences and size is crucial to improving scaling approaches and 57 predictive understanding of water fluxes in structurally complex and species diverse forest ecosystems. 58 While work conducted in a range of forests world-wide generally show that sap velocity varies little with 59 stand age and that scaling parameters such as SAI (sapwood area index) and LAI (leaf area index) often 60 explain changes in water use patterns over time (Vertessy et al. 2001), much of this research has been 61 conducted in relatively species-poor forests and less is known about water use patterns and scaling 62 relationships in mixed-species mesic temperate forests (Grossiord et al. 2013, Kallarackal et al. 2013). 63 We measured sap velocity and estimate tree and stand water use dynamics in five species across 64 two northern hardwood stands in New Hampshire (USA). Our specific objectives were to: (i) Assess the importance of different sources of variability (radial, azimuthal and height 65 66 variation, species identity and size class) when scaling from single-point measurements of 67 sap velocity to tree- and stand scale estimates of transpiration.

68 (ii) Evaluate how species composition and canopy structure affect stand transpiration.

69 (iii) Explore the potential impacts of future shifts in species composition in the region on forest
70 water budget.

71 METHODS

#### 72 Study site

This study was conducted in two forest stands located in the Bartlett Experimental Forest (BEF), in the White Mountain National Forest, New Hampshire, USA (44.05°N, 71.28°W). These two stands (described in Table 1) have been intensively studied since 2004 (e.g. Fatemi et al. 2011, Vadeboncoeur et al. 2012) and have been previously referred to as "C8" (mature; >130 years-old) and "C2" (young; 25 years-old).

78 The climate is humid continental, with warm summers and cold winters; mean annual 79 temperature is 7°C. Annual precipitation averages 1,270 mm and is evenly distributed throughout the 80 year. Soils are predominantly frigid Haplorthods developed on granitic glacial till. At approximately 300 81 m elevation, where our study plots are located, the forest is dominated by late-successional northern 82 hardwood species including American beech, sugar maple, and yellow birch, in some stands mixed with 83 eastern hemlock (Tsuga canadensis) (Leak, 1991). The region was cutover in the late 1800s and early 84 1900s. Since acquisition by the USA Forest Service circa 1915, some areas have been actively managed 85 and others left unmanaged, providing an opportunity to compare forests of different successional stages in close proximity (Leak, 1996). In stands clearcut within the previous 40 years, early-successional species 86 87 including pin cherry (Prunus pensylvanica), paper birch (Betula papyrifera), and red maple (A. rubrum) occur in combination with some of the later-successional species. 88 We measured sap velocity ( $v_s$ , cm<sup>3</sup> cm<sup>-2</sup> h<sup>-1</sup>) during two growing seasons, 2011 and 2013. In May 89 90 2011, fertilization treatments were initiated in our study stands as part of a larger study (MELNHE) aimed

91 at understanding nutrient limitations on various ecosystem processes in northern hardwood forests.

Treatments included 30 kg N ha<sup>-1</sup> y<sup>-1</sup>, 10 kg P ha<sup>-1</sup> y<sup>-1</sup>, both N and P combined, as well as a control; these 92 93 were assigned randomly to the 4 plots within each stand (Fisk et al. 2013). These relatively low 94 fertilization rates are intended to show effects of nutrient enrichment over many years. Our data showed no significant treatment effects on  $v_s$  in either 2011 or 2013, with the exception of red maple in 95 96 the N plot in the young stand (Hernandez-Hernandez 2014). Thus, the data for red maple in this plot 97 were excluded from the analyses presented here, and all other trees were pooled by species and stand 98 for analysis of species and tree size differences in v<sub>s</sub>, as well as sources of error in estimating water flux 99 at the stand scale.

### 100 **Observation conditions**

During the 10-day preliminary observation period in 2011 (22 August to 31 August), maximum temperature averaged 18.2°C and ranged from 15.6 to 20.3, and minimum temperature averaged 12.5°C, ranging from 9.7 to 17.3. Daily maximum vapor pressure deficit ranged from 0.4 to 1.6 kPa, and averaged 1.3 kPa. Daily incoming global radiation averaged 5000 W h<sup>-1</sup> m<sup>-2</sup>, with a range of 300-6800 W h<sup>-1</sup> m<sup>-2</sup>.

106 During the 41-day primary observation period in 2013 (5 July to 14 August), daily maximum temperature averaged 25.2 +/- 3.8 °C, with a range of 17.8-33.0, and minimum temperature averaged 107 14.8 +/- 3.4°C, with a range of 8.9 - 20.8. Daily maximum vapor pressure deficit ranged from 0.0 - 2.1 108 kPa, and averaged 1.1 kPa. Daily incoming global radiation averaged 5500 +/- 2060 W  $h^{-1}$  m<sup>-2</sup>, with a 109 range of 640-8210 W h<sup>-1</sup> m<sup>-2</sup>. Soil moisture sensors were installed during this period as well. Soil 110 111 moisture at 15 cm averaged about 30% at both sites, and the mean of 4 sensors did not decline below 112 27% at the young stand or 23% at the old stand. All the data were collected at the nearby Ameriflux 113 tower (http://ameriflux.lbl.gov/).

## 114 Sap velocity measurements

Sap velocity was measured using the Heat Ratio Method (HRM) (Burgess et al. 2001). This
 method uses three probes (a heater and two temperature probes) arranged vertically with the heater

117 located between the temperature probes. Following the release of a pulse of heat from the middle 118 probe, the HRM estimates sap velocity from the ratio of the increase in temperature, at points 119 equidistant downstream and upstream. Heat pulse velocity ( $v_h$ ) is calculated as (Marshall, 1958):

120 
$$v_h = \frac{k}{x} Ln\left(\frac{v_1}{v_2}\right) 3600$$
 (1)

Where k is thermal diffusivity of a green fresh wood ( $\text{cm}^2 \text{ s}^{-1}$ ), x is the distance between the heater probe 121 122 to either temperature probes (cm),  $v_1$  is upstream increase in temperature and  $v_2$  is downstream 123 increase in temperature from initial temperature ( $^{\circ}$ C). Sap velocity is then calculated after v<sub>h</sub> is corrected 124 for wounding and misalignment following Burgess et al. (2001). Each temperature probe consisted of three thermocouples located at depths of 1.0 (outer), 2.3 (middle) and 3.5 cm (inner) from the bark in 125 126 order to measure three different depths across the radial sapwood profile (Gebauer et al. 2008). 127 In the mature stand, two probe sets were installed in each sample tree in 2013 at N and S 128 orientations (in 2011 only N orientation) whereas in the young stand only the N orientation was used 129 due to the small DBH of these trees. To install each probe set, we carefully removed the bark from a 130 small area before drilling three vertically-aligned holes 1.3 mm in diameter, 38 mm deep, and 6 mm 131 apart. Petroleum jelly was used to ease probe insertion and maintain thermal contact between the 132 probe and xylem tissue (Burgess et al. 2001). The whole area was then wrapped with reflective 133 polyethylene air-cell insulation to minimize external heat exchange via radiation and conduction. Each 134 probe set was connected via a 10 m cable and a multiplexer (AM 16/32B, Campbell Scientific Inc., Logan, 135 UT, USA) to a datalogger (CR1000, Campbell Scientific Inc.). Temperature from each thermocouple was 136 logged following a heat pulse every 15 min, and v<sub>s</sub> calculated according to Burgess et al. (2001). 137 In order to assess in greater detail the radial, azimuthal, and height variation in v<sub>s</sub>, we conducted 138 a more intensive measurement campaign on American beech trees in the mature stand from 22-31 August 2011. Twelve trees of different sizes (DBH 5.4-45.1 cm) were instrumented with one probe set 139

140 each installed with N orientation and at 1.3 m to evaluate changes in the radial profile of v<sub>s</sub> with tree

size. Two additional American beech trees of DBH of 7.6 cm and 6.9 cm were instrumented with eight

probe sets each to assess the radial, azimuthal and height variability of v<sub>s</sub>. Probe sets were installed at

143 two azimuthal orientations (N and S) and four different heights starting 25 cm above the ground surface.

144 The vertical distance between probe pairs with the same orientation was 86-89 cm.

For our primary data set, v<sub>s</sub> was measured from 5 July to 14 August 2013 in three dominant species in each stand (Table 1). The species examined for v<sub>s</sub> were sugar maple, American beech, and yellow birch in the mature stand, and red maple, American beech, and white birch in the young stand.

148 Pin cherry, despite being a dominant species, was not measured in the young stand due to unsatisfactory

data quality during a preliminary data collection period, likely due to its very narrow sapwood. After

150 removing trees with large data gaps or poor-quality data, our final data set v<sub>s</sub> for the entire

151 measurement period included 7-8 trees per species in the mature stand and 9-12 trees per species in the

152 young stand. All selected trees were healthy, dominant or co-dominant in canopy position, and had no

153 major injuries or defects apparent on the lower few meters of the bole. The DBH range of sample trees

154 was similar among the species in each stand: 29.4-58.4 for American beech, 28.9-55.4 for yellow birch,

and 32.0-50.2 for sugar maple in the mature stand, and 9.3-16.9 for American beech, 9.3-16.7 for white

birch, and 10.2-13.7 for red maple in the young stand.

To determine whether it was appropriate to compare the data from the two years, we compared the v<sub>s</sub> of the outer and middle sensors in the 8 American beech trees in which v<sub>s</sub> was measured in both years. A mean v<sub>s</sub> value for each tree was used as the dependent variable in a linear model, where "year" was the fixed factor. We observed no significant difference between years (P=0.61 and P=0.17, for outer and middle sensor, respectively), and thus, v<sub>s</sub> data from both years were pooled for analyses of height and azimuthal variability and canopy position effect.

163 Sapwood area estimates

164 Twenty representative (non-sample) trees of each dominant species in the young and old stand 165 were surveyed in September 2012 and 2010, respectively, to derive allometric relationships between 166 DBH and sapwood area (A<sub>s</sub>), so we could estimate A<sub>s</sub> for the trees monitored with sap flow sensors. For

each tree we recorded the DBH and extracted at least one increment core with a 5 mm diameter borer,

and sapwood depth was determined on each core by measuring the translucent section between the

169 bark and heartwood using a Vernier caliper (Table 2). Species-specific allometric equations were

170 developed to describe the relationship between sapwood depth and tree DBH data for the stand. For

each species, the relationship between A<sub>s</sub> and DBH was best explained by a power equation of the form

172  $A_s = a * DBH^b$ .

173 Selecting sensor depths for comparisons among trees

174 As sensor depths for all sample trees were the same, sapwood depths at which v<sub>s</sub> was measured

175 (1.0, 2.2 and 3.5 cm) were expressed relative to the maximum sapwood depth of each sample tree

176 ranging from 0 (at the cambium) to 100% (at the estimated transition from sapwood to heartwood). This

177 information was used to select the most comparable sensor depths when comparing trees that varied in

178 DBH and sapwood thickness (Alvarado-Barrientos et al. 2013). For comparisons across size classes, all

trees across both stands were binned into DBH categories with break-points at 11.1, 19.1, 29.1, and 38.1

180 cm, in order to distribute the sample trees approximately evenly among size classes.

#### 181 Canopy position effect

182 To explore the influence of canopy position on v<sub>s</sub>, we analyzed data from both 2011 and 2013

183 collected in trees of similar DBH range (9.3-13.0 cm in young stand and 8.6-11.7 cm in mature stand) but

184 varying in canopy position (dominant or co-dominant in the young stand and sub-dominant,

185 intermediate in the mature stand). Using trees of similar size but differing in canopy position (dominant

in the young stand or subdominant in the mature stand) allowed us to separate the effect of tree size

187 from the effect of canopy position.

### 188 Scaling sap velocity to transpiration estimates for trees and stands

189 In contrast to the selection of one depth for comparisons of v<sub>s</sub> among trees (described above),

190 we used data from all depths to calculate whole tree sap flow  $(Q_s, cm^3 h^{-1})$  for each day. For each sample

191 tree, we divided the estimated sapwood area into three concentric rings, each corresponding to a sensor

depth using the weighted-average method (Hatton et al. 1990). The area of each cylinder was multiplied

193 by  $v_s$  for that depth. The total sap flow of each effective area was summed to determine  $Q_s$ .

194 Stand-scale transpiration (T, mm h<sup>-1</sup>) was then calculated for both the mature and young stands

using stand inventories from 2011. Within the four 30 x 30 m plots in each stand, all trees >10 cm DBH

196 were identified and measured (0.36 ha total). Trees between 2-10 cm DBH were measured in twenty

subplots, each 5x5 m in size (0.05 ha total). Unsampled species accounted for 33% of basal area in the

198 young stand but only 10% in the mature stand (Table 1). In order to estimate stand-scale T, the sapwood

area and v<sub>s</sub> of subdominant tree species were estimated using the allometric equation of the study

200 species within the same stand having the most similar DBH range.

201 Using the mean species  $Q_s$  of every tree of each stand and the tree density of each stand 202 (considering separately the trees with DBH less and greater than 10 cm), we observed no significant 203 relationship between DBH and  $v_s$  for each stand.

We further calculated stand T for different scenarios of shifts in species composition to heuristically assess boundaries to the potential impacts of species change on stand T, with theoretical mono-species stands of each study species, assuming the same tree Q<sub>s</sub> as obtained from the study.

#### 207 Statistical analysis

208 We used linear mixed models (LMM) to analyze the effects of radial variation, azimuthal, height, 209 species, tree size (here as DBH) and canopy position (fixed factors) on  $v_s$  (dependent variable). We 210 generally obtained normal and homoscedastic residuals and thus no transformations of the variables 211 were necessary. The protocol described in Zuur et al. (2009) was followed to determine the optimal 212 random structure in each case, using Akaike's information criterion (AIC) to select the best models. The 213 random structures we compared were ~1|stand/plot/,~1|stand, ~1|plot and a simple linear model with 214 no random structure, where stand is young or mature stand and plot indicates in which treatment plot 215 the tree was located. The first two random structures were only used when we pooled together the data

of the two stands for the statistical analysis (tree size and canopy position). The fixed effects were
analyzed in simple models, each of them at a time, comparing each model with the fixed effect with a
null model without the factor we are interested. The likelihood ratio was used to test for the significance
of each fixed factor (Zuur et al. 2009). All analyses were conducted with the R package 'nlme R' (Pinheiro
et al. 2011). When multilevel analyses were necessary, "pair.wise.test" analysis was conducted and P
values were Bonferroni adjusted. For mixed models, the package "multcomp" was used (Hothorn et al.
2014).

223 **RESULTS** 

## 224 Within-tree sap velocity variability

225 Analysis of the radial variation of  $v_s$  within sapwood in the young stand showed no consistent 226 relationship for the different sensor depths (1.0, 2.2 and 3.5 cm for the inner, middle and outer sensor in 227 each studied individual) across the three species (Fig. 1). On the contrary, the radial profile was similar in 228 the three species monitored in the mature stand (no significant interaction between radial profile and 229 species). In the mature stand,  $v_s$  measured at the middle and inner sensor positions were similar to each 230 other and greater than the outer sensor position. This difference was significant in yellow birch and 231 sugar maple but not in American beech, although the mean values in American beech followed the same 232 general pattern. No interaction between tree size and radial profile was found in either stand probably 233 due to the small DBH range analyzed. 234 For the intensive measurements on American beech, no significant differences were observed 235 for v<sub>s</sub> measurements between the N and S orientation in either 2011 or 2013, nor was there an effect of

height or the interaction between radial and azimuthal measurements on  $v_s$  of American beech in 2011

237 (P>0.05, data not shown).

#### 238 Species differences

239 To examine the differences among species and remove the size effect from the analysis, we 240 calculated the relative depth of each sensor within the sapwood as described previously, i.e., each 241 sensor depth was normalized by the maximum sapwood depth of each individual. The sensors used for 242 the species comparison were in a relative position range of 30%-60% of total sapwood depth. In general, 243 for mature stand trees, the middle sensor (and in a few individuals, the inner sensor) was located within 244 this range, whereas for young stand trees, it was the outer sensor. In general, the species in the young 245 stand had greater  $v_s$  (Fig. 2). White birch had significantly higher  $v_s$  than the other species in both stands, 246 except in the young stand that showed a statistically similar  $v_s$  to American beech. Comparing between congeneric species in the mature and young stand, we observed no significant site differences in 247 248 American beech or between the two maple species, although there was a trend of greater  $v_s$  for the American beech and maple in the young stand relative to the old stand (26.2 $\pm$ 0.7 and 20.3 $\pm$ 1.6 cm<sup>3</sup> cm<sup>-2</sup> 249  $h^{-1}$  in American beech in young and old stand, respectively and 17.2 ± 0.5 and 15.1 ± 1.0 cm<sup>3</sup> cm<sup>-2</sup>  $h^{-1}$  in 250 251 red maple in young stand and sugar maple in old stand, respectively; all results hereafter reported as 252 mean ± SE). However, a significant difference was observed for the two birch species, white birch in the young stand having significantly higher  $v_s$  (31.8 ± 0.6 cm<sup>3</sup> cm<sup>-2</sup> h<sup>-1</sup>) than yellow birch in the mature stand 253  $(15.8 \pm 1.3 \text{ cm}^3 \text{ cm}^{-2} \text{ h}^{-1})$ . Comparing across all species and sites, the maple species had the lowest v<sub>s</sub>, but 254 255 these values were only significantly different from white birch.

## 256

## Tree size and canopy position effect

Using the same relative position approach as for the species comparison (above) and all sample trees in both stands for 2013, the results showed significant differences in  $v_s$  due to tree size (Fig. 3), with higher values (25.1 cm<sup>3</sup> cm<sup>-2</sup> h<sup>-1</sup> on average) associated with smaller DBH size classes (< 29.1 cm) and significantly smaller values (17.6 cm<sup>3</sup> cm<sup>-2</sup> h<sup>-1</sup> on average) recorded for trees with DBH > 29 cm. This DBH division corresponds to the DBH ranges in the young and old stands.

- 262 Regarding the effect of canopy position on v<sub>s</sub>, the results (Fig. 4) showed that the trees in the
- 263 young stand had significantly higher v<sub>s</sub> than trees of similar size in the mature stand. This difference can
- be seen in both the outer sensor (26.0 $\pm$ 2.5 in dominant and 13.3 $\pm$ 4.1 cm<sup>3</sup> cm<sup>-2</sup> h<sup>-1</sup> in sub-dominant trees)
- and middle sensor (26.4 $\pm$ 3.5 in dominant and 4.1 $\pm$ 2.4 cm<sup>3</sup> cm<sup>-2</sup> h<sup>-1</sup> in sub-dominant trees).
- 266

## Stand-scale transpiration estimates

- Although  $v_s$  was in general higher in the species of the young stand than in the mature stand, the greater total sapwood area of the mature stand (27.2 m<sup>2</sup> ha<sup>-1</sup>) compared to the young stand (18.8 m<sup>2</sup> ha<sup>-1</sup>) resulted in a greater total daily stand T in the mature stand (5.3 mm day<sup>-1</sup>) than in the young stand (4.9 mm h<sup>-1</sup>) (Fig. 5).
- 271 Species change scenarios

272 Under a scenario whereby future management leads to dominance by sugar maple, stand T 273 would increase substantially (Fig. 5) (9.2 mm day<sup>-1</sup>), while dominance by American beech would have the 274 lowest stand T in both the mature and young stands (3.2 and 3.9 mm day<sup>-1</sup> in mature and young stand, 275 respectively) (Fig. 5). A future scenario where the young stand is dominated by white birch would have 276 the greatest transpiration (7.4 mm day<sup>-1</sup>) of the species considered in the young plot.

## 277 **DISCUSSION**

278 Our analysis of the different sources of uncertainty in the v<sub>s</sub> measurements (radial, azimuthal 279 and height heterogeneity) showed that only radial variation in v<sub>s</sub> had a significant effect on v<sub>s</sub> 280 determination and thus should be considered in sample design and scaling (Fig. 1). Previous works have 281 attributed radial variability in v<sub>s</sub> to the type and arrangement of sap conducting tissue (e.g. vessel 282 diameter, spatial distribution of the vessels within a given growth ring; Swanson 1994), the vertical 283 distribution of foliage in the crown (Fiora and Cescatti 2008), the effect of differential water stored along 284 the sapwood (Ford et al. 2004), and outer rings compensating for embolization in older sapwood 285 (Granier et al. 1994). The azimuthal variation observed in other works (Lu et al. 2000) was not found in

286 our study probably due to the closed canopy of the study stand, and consequently, the lack of strong 287 azimuthal variation in solar radiation within individual crowns. The lack of an effect of height variation on 288  $v_s$  in our study may be at least in part due to the small sample size (n = 2); nevertheless, some researchers have previously reported a correlation between  $v_s$  and height (Loustau et al. 1998). 289 290 The sources of variability in plant water use patterns in structurally complex and species diverse 291 forests are often much greater compared to species-poor forests or timber plantations, and this 292 heterogeneity has important implications for estimating stand water use (Wullschleger et al. 2001). No 293 major differences in v<sub>s</sub> were observed among the studied species other than white birch presenting 294 higher  $v_s$  than the other species (Fig. 2). The differences between species may be attributed to several 295 factors, including leaf-level properties, growth strategy, site characteristics, and canopy position. White 296 birch is a shade-intolerant early successional species, and therefore has fast growth rates and quickly 297 establishes canopy dominance in young stands. To achieve higher v<sub>s</sub> than other co-existing species, white 298 birch should have higher stomatal conductance, a hydraulic system allowing high water use rates, and 299 the capacity to tolerate a higher water potential gradient. Although studies comparing these 300 characteristics for our study species are lacking, our results showing that white birch had a higher v<sub>s</sub> than 301 red maple is consistent with reported lower stomatal densities and longer guard cells in white birch 302 compared to red maple (Abrams and Kubiske 1990). While previous studies have assessed water use 303 patterns in some of these species (e.g., sugar maple, yellow birch: Tang et al. 2006; red maple and white 304 birch: Bovard et al. 2005), these studies were conducted in different forest types (northern central 305 hardwoods, mixed-hardwood forest in northern Lower Michigan, respectively), and we are not aware of 306 any simultaneous comparisons within forests where these species co-exist. Although it is difficult to 307 compare the absolute numbers of other studies with our study due to differences in tree age, stand 308 structure, and climate, the trends found in our study are consistent with the findings of Tang et al. 309 (2006), which showed a similar sap flux per unit of sapwood area for both yellow birch and sugar maple. 310 However, contrary to our findings, Bovard et al. (2005) found that red maple and white birch presented

similar stand transpiration for both species ( $0.1 \text{ kg m}^{-2} \text{ day}^{-1}$ ). Federer and Gee (1976) predicted that yellow birch should have higher transpiration rates than American beech and sugar maple due to the differences in diffusion resistances of abaxial leaf surfaces. However, we did not observe significant differences in v<sub>s</sub> among these species (Fig. 2).

315 Nevertheless, we did observe a clear effect of tree size on  $v_s$ , with the normalized sensor in the 316 sapwood having significantly greater v<sub>s</sub> in trees with DBH < 29 cm relative to larger trees (Fig. 3, this 317 comparison shows data only from dominant trees either in the young or mature stand). Smaller, younger 318 trees generally have faster growth rates (including DBH, height, and foliage) in order to compete with 319 other trees, whereas old trees have reached their maximum growth rate and may allocate resources to 320 other functions (e.g., maintenance respiration). Our findings that size had a direct influence on water use 321 patterns while species did not is consistent with the idea of functional convergence, which suggests that 322 because plants operate within given biophysical limits, they develop common physiologies for water 323 uptake across taxa (Meinzer et al. 2001, Jung et al. 2011). Thus, tree water uptake would be more 324 controlled by biophysical limits such as radiation than by any differences between species. A positive 325 relationship between tree size and v<sub>s</sub> and Q<sub>s</sub> has been widely reported (McJannet et al. 2007, Dierick and 326 Hölscher 2009, Jung et al. 2011), although in some works negative relationships have been found 327 (Meinzer et al. 2001). The lack of a significant statistical correlation between DBH and  $v_s$  in our study for 328 each stand is likely explained by the relatively small range of DBH sizes sampled in each stand.

The intensive study of American beech trees of similar size but in different canopy positions pointed to a clear effect of canopy position on v<sub>s</sub>, as reported previously (Hernandez-Santana et al. 2011). We observed much lower v<sub>s</sub> in the subdominant trees of the mature stand than in the dominant trees of the young stand, likely due to the lower leaf-to-sapwood area ratio and lower solar radiation for mature stand American beech trees. Although these data were collected in different years, which may introduce confounding factors due to differing environmental conditions, performing the statistical

analyses on the entire data set encompassing a large degree of environmental variability providedgreater confidence in our results of a strong significant difference.

337 Although trees of the younger stand had higher  $v_s$  compared to the mature stand, stand 338 transpiration was still greater in the mature stand, primarily attributed to differences in basal area and, 339 thus, sapwood area. Our results are contrary to studies showing that water use by younger stands is 340 greater than for older stands (Vertessy et al. 2001, Moore et al. 2004, Buckley et al. 2012). However, in 341 these studies the greater water use observed in the younger stand was mainly driven by greater 342 sapwood area instead of higher v<sub>s</sub>, as observed in our study. In the pure stands examined in these 343 studies, greater tree density per surface area was associated with greater A<sub>s</sub> and LAI. In our study, 344 however, the young stand has a higher tree density than mature stand (Table 1), but the mature stand 345 has a higher tree density of bigger trees (>10 cm). That difference could explain why we found a similar 346 LAI but a higher  $A_s$  in mature stand than in young stand. Therefore, the greater T observed in our 347 northern hardwood forest mature stand may be explained by the greater complexity in tree species 348 composition and shade tolerances, resulting in the higher total stand A<sub>s</sub> we observed (Table 1) Shade 349 tolerant tree species such as American beech and sugar maple of DBH ≥ 10 cm were the main 350 contributors to mature stand T. The young stand on the other hand, had lower basal area,  $A_s$  (Table 1) 351 and consequently lower T, which was mainly a result of a larger number of trees with DBH between 2-10 352 cm, particularly American beech, red maple and pin cherry. Total leaf area and mass were similar 353 between the two stands (Table 1), as is common following the earliest successional stages in northern 354 hardwood forests (Covington and Aber 1980). Our study species differed in A<sub>s</sub>, particularly in the young 355 stand, where red maple and white birch had little or no heartwood up to 5 cm DBH, while American 356 beech had a substantial amount of heartwood (e.g. heartwood area was ~30% of basal area in 10 cm 357 trees). This difference in A<sub>s</sub> partly explains why American beech T was lower than red maple (Fig 5), 358 despite v<sub>s</sub> in American beech being higher than in maple. Our scaled estimates of water use in the young 359 stand were in fact considerably lower than in the mature stand, though some caution in interpreting

360 these results is warranted due to uncertainty resulting from extrapolating sap velocities to trees of sizes 361 and species (including pin cherry) for which we did not measure sap velocities. Thus, in our study, 362 species-specific sapwood area per ground area and the spatial distribution of trees of different species 363 were more important than interspecific differences in  $v_s$  in determining water use at the tree and stand 364 scales as recognized in other studies (Wullschleger et al. 2001, Bovard et al. 2005, Mackay et al. 2010). 365 For instance, Wullschleger et al. (2001) working in a forest composed of both ring-porous and diffuse-366 porous species, concluded that transpiration is likely dominated by the species that dominates total 367 sapwood area. Similarly, Bovard et al. (2005) reported that much of the variation in transpiration among 368 different stands within a forest composed by bigtooth aspen (Populus grandidentata Michx.), white 369 birch, red maple, and red oak (Quercus rubra L.) was due to large differences in stand level sapwood 370 area.

371 A clear effect of the differences in species' sapwood area and  $v_s$  on stand T was also observed 372 when the hypothetical scenarios of shifts in species composition were assessed. The transpiration results 373 calculated using the sap flow data by species and tree density for each scenarios of 100% dominance by 374 each species (Fig. 5) showed that a shift towards dominance by sugar maple would result in the highest T  $(9.2 \text{ mm day}^{-1})$  despite having the lowest v<sub>s</sub> of the mature stand species (Fig. 2). Sugar maple has the 375 376 highest sapwood area at any DBH of the three species studied and American beech the lowest. For 377 example, for a tree of a DBH of 30 cm, the sapwood area is around 33% higher in sugar maple than in 378 American beech. Management efforts to increase the dominance of sugar maple would therefore 379 increase T and decrease water yield. We did not study mature red maple, but it also had quite high 380 sapwood area as a fraction of total basal area, and increases in red maple may be more likely a result of 381 climate change, as it is a highly plastic species with a wide geographic and edaphic range (Abrams 1998). 382 On the other hand, dominance by American beech (as a consequence of disturbance or management 383 that inadvertently encourages the success of root-sprouting species), would lead to the lowest stand transpiration at both stand ages (3.9 and 3.2 mm day<sup>-1</sup> in young and mature stands, respectively) despite 384

being the species presenting the highest v<sub>s</sub> in the mature stand and the second highest v<sub>s</sub> in the young
stand. Shifts in forest composition from mixed mature northern hardwoods to sprout-regenerated
American beech, as has occurred to some extent in other stands with the Hubbard Brook Experimental
Forest (Hane 2003; Weeks et al. 2009) might therefore be expected to decrease evapotranspiration,
which has indeed been observed (Hamburg et al. 2013).

390 A hypothetical young stand with 100% white birch would have the greatest transpiration (7.4 391 mm day<sup>-1</sup>). These bounding cases are heuristic tools and do not necessarily represent realistic 392 trajectories. For example, relative shade intolerance of yellow birch would limit its dominance in a 393 mature stand, though it can be quite important in post-disturbance forests even into maturity (Nowacki 394 and Abrams 2014). With climate warming over the next century, most of the northern hardwood species 395 we studied are projected to decline in dominance in the region, though red maple may increase along 396 with species with more southern distributions such as oaks (Mohan et al. 2009). Red oak may be poised 397 to do particularly well, as it is near the northern limit of its range in northern New Hampshire, and is 398 already present in low abundance near these study plots. As a ring-porous species, it is functionally 399 quite distinct from the diffuse-porous hardwoods we studied; sapwood tends to be narrow but highly 400 conductive, and may use considerably more water than the diffuse-porous species it replaces (Catovsky 401 et al. 2002), in which case we would not expect the rough scaling of transpiration with sapwood area to 402 hold as does within the diffuse-porous hardwoods.

Long-term monitoring of species composition and regeneration dynamics together with modeling simulations may enhance capacity to accurately predict likely future scenarios of species shifts and their implications for stand water use patterns and responses to extreme climate events. Within the forecasted future scenario of high intensity precipitation and extended dry periods (Swain and Hayhoe 2014), forest management efforts might focus not only on favoring species with efficient water use (Grant et al. 2013), which in this landscape seems to mean species with less sapwood area. Stand

409 transpiration is mainly dominated by the species dominating the total A<sub>s</sub>, and thus, favoring diffuse-

410 porous species with less A<sub>s</sub> may better maintain water yields.

#### 411 **CONCLUSIONS**

412 In this study, we found that tree size is more important than species in determining sap velocity 413 in this mesic temperate forest of the northeastern USA. Tree size also had an important effect on 414 estimates of total sapwood area and canopy structure. However, total sapwood area, which has been 415 demonstrated to exert a strong controlling influence on stand transpiration, also varies depending on 416 species composition. The effects of radial variability on sap velocity should be considered as a major 417 potential source of error when scaling sap velocity to tree water use, and was found to differ by both 418 tree size and species in the young and small trees. Thus, reliable measurements of stand water use in 419 multi-species diffuse-porous deciduous forests could be achieved considering the variability of sap 420 velocity in size class and canopy position, rather than species and the sapwood area dependent on both 421 the species and the size. Furthermore, the effect of potential species shifts on stand transpiration will 422 depend on the sap velocity, determined mainly by radial variation and tree size, but also on the sapwood 423 area distribution in the stand.

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#### REFERENCES

Abrams, M.D. 1998. The red maple paradox. Bioscience **48**:355–364.

- Alvarado-Barrientos, M.S., Hernandez-Santana, V., and Asbjornsen, H. 2013. Variability of the radial profile of sap velocity in *Pinus patula* from contrasting stands within the seasonal cloud forest zone of Veracruz, Mexico. Agr. Forest Meteorol. **168**: 108-119
- Bovard, B.D., Curtis, P.S., Vogel, C.S., Su, C.S., and Schmid, H.P. 2005. Environmental controls on sap flow in a northern hardwood forest. Tree Physiol. **25**: 31-38
- Brzostek, E. R., D. Dragoni, H. P. Schmid, A. F. Rahman, D. Sims, C. a Wayson, D. J. Johnson, and Phillips,
   R.P. 2014. Chronic water stress reduces tree growth and the carbon sink of deciduous hardwood forests. Global Change Biol. 20:2531–2539.
- Buckley, T.N., Turnbull, T.L., Pfautsch, S., Gharun, M., Adams, M.A., 2012. Differences in water use between mature and post-fire regrowth stands of subalpine *Eucalyptus delegatensis* R. Baker. Forest Ecol. Manag. **270**: 1-10.
- Burgess, S.S.O, Adams, M.A., Turner, N.C., Beverly, C.R., Ong, C.K., Khan, A.A.H., and Bleby, T.M. 2001. An improved heat pulse method to measure low and reverse rates of sap flow in woody plants. Tree Physiol. **21**: 589-598.
- Campbell, J.K., Driscoll, C.T., Pourmokhtarian, A., and Hayhoe, K. 2011. Streamflow responses to past and projected future changes in climate at the Hubbard Brook Experimental Forest, New Hampshire, United States. Water Resour. Res. 47, DOI: 10.1029/2010WR009438
- Catovsky, S., Holbrook, N.M., Bazzaz, F.A. 2002. Coupling whole-tree transpiration and canopy photosynthesis in coniferous and broad-leaved tree species. Can. J. For. Res. **32**:295–309.
- Cavaleri, M.A., and Sack, L. 2010. Comparative water use of native and invasive plants at multiple scales: a global meta-analysis. Ecology **91**: 2705-2715.
- Covington, W.W., Aber, J.D. 1980. Leaf production during secondary succession in northern hardwoods. Ecology 61:200–204.
- Dierick, D., and Hoelscher, D. 2009. Species-specific tree water use characteristics in reforestation stands in the Philippines. Agr. Forest Meteorol. **149**: 1317-1326.
- Fatemi, F. R., R. D. Yanai, S. P. Hamburg, M. A. Vadeboncoeur, M. A. Arthur, R. D. Briggs, and Levine, C.R. 2011. Allometric equations for young northern hardwoods: the importance of age-specific equations for estimating aboveground biomass. Can. J. For. Res. **41**:881–891.
- Federer, C.A. and Gee, G.W. 1976. Diffusion resistance and xylem potential in stressed and unstressed northern hardwood trees. Ecology **57**: 975-984.
- Fiora, A., and Cescatti, A. 2008. Vertical foliage distribution determines the radial pattern of sap flux density in *Picea abies*. Tree physiol. **28**: 1317-1323
- Fisk, M.C., Ratliff, T.J., Goswami, S., and Yanai, R.D., 2013. Synergistic soil response to nitrogen plus phosphorus fertilization in hardwood forests. Biogeochemistry.
- Ford, C.R., McGuire, M.A., Mitchell, R.J., and Teskey, R.O. 2004. Assessing variation in the radial profile of sap flux density in *Pinus* species and its effect on daily water use. Tree Physiol. **24**: 241–249

- Gebauer, T., Horna, V., and Leuschner, C. 2008. Variability in radial sap flux density patterns and sapwood area among seven co-occurring temperate broad-leaved tree species. Tree Physiol. . 28: 1821-1830
- Granier, A., Anfodillo, T., Sabatti, M., Cochard, H., Dreyer, E., Tomasi, M., Valentini, R., and Bréda, N.
  1994. Axial and radial water flow in the trunks of oak trees: a quantitative and qualitative analysis.
  Tree Physiol. 14: 1383-1396
- Granier, A., Biron, P., Breda, N., Pontailler, J.Y., and Saugier, B. 1996. Transpiration of trees and forest stands: Short and longterm monitoring using sapflow methods. Global Change Biol. **2**: 265-274.
- Grant, G. E., C. L. Tague, and Allen, C.D. 2013. Watering the forest for the trees: an emerging priority for managing water in forest landscapes. Front. Ecol. Environ. **11**:314–321.
- Grossiord, C., Granier, A., Gessler, A., Pollastrini, M., and Bonal, D. 2013. The influence of tree species mixture on ecosystem-level carbon accumulation and water use in a mixed boreal plantation. Forest Ecol. Manag. **298**: 82-92
- Hamburg, S.P., Vadeboncoeur, M.A., Richardson, A.D., Bailey, A.S. 2013. Climate change at the ecosystem scale: a 50-year record in New Hampshire. Clim. Change **116**:457–477.
- Hane, E.N. 2003. Indirect effects of beech bark disease on sugar maple seedling survival. Can J For Res **33**:807–813.
- Hayhoe, K., Wake, C.P., Huntington, and Luo, T.G. 2007. Past and future changes in climate and hydrological indicators in the USA Northeast. Clim. Dyn. **28**: 381-407.
- Hatton, T.J., Catchpole, E.A., and Vertessy, R.A. 1990. Integration of sapflow velocity to estimate plant water-use. Tree Physiol. **6**: 201-209.
- Hernandez-Hernandez, A. 2014. Effects of nutrient amendments on water use and water use efficiency in a Northeastern forest ecosystem. Master thesis, University of New Hampshire.
- Hernandez-Santana, V., Asbjornsen, H., Sauer, T., Isenhart, T., Schilling, K., Schultz, D. 2011. Enhanced transpiration by riparian buffer trees in response to advection in a humid temperate agricultural landscape. Forest Ecol. Manag. **261**: 1415–1427
- Hornbeck, J.W., Adams, M.B., Corbett, E.S., Verry, E.S., and Lynch, J.A. 1993. Long-term impacts of forest treatments on water yield- a summary for northeastern USA. J. Hydrol. **150**: 323-344
- Hothorn, T., Bretz, F., Westfall, P., Heiberger, R.M., and Schuetzenmeister, A. 2014. Simultaneous Inference in General Parametric Models. Package Multcomp. http://multcomp.r-forge.r-project.org/
- Jung, E.Y., Otieno, D., Lee, B., Lim, J.H., Kang, S.K., Schmidt, M.W.T., and Tenhunen, J. 2011. Up-scaling to stand transpiration of an Asian temperate mixed-deciduous forest from single tree sapflow measurements. Plant Ecol. 212: 383-395.
- Köstner, B., Falge, E.M., Alsheimer, M., Geyer, R., and Tenhunen, J.D. 1998. Estimating tree canopy water use via xylem sapflow in an old Norway spruce forest and a comparison with simulation-based canopy transpiration estimates. Ann. For. Sci. **55**: 125-139
- Kallarackal, J., Otieno, D.O., Reineking, B., Jung, E.Y., Schmidt, M.W.T., Granier, A., and Tenhunen, J.D.
   2013. Functional convergence in water use of trees from different geographical regions: a metaanalysis. Trees 27: 787-799
- Leak, W. 1991. Secondary forest succession in New Hampshire, USA. Forest Ecol. Manag. 43: 69-86
- Leak W.B., Smith, M.-L. 1996. Sixty years of management and natural disturbance in a New England forested landscape. Forest Ecol. Manag. **81**:63–73.

Likens G.E. 2013. Biogeochemistry of a Forested Ecosystem, 3rd ed. Springer

- Loustau, D., Domec, J. C., and Bosc, A. 1998. Interpreting the variations in xylem sap flux density within the trunk of maritime pine (*Pinus pinaster* Ait.): application of a model for calculating water flows at tree and stand levels. Ann. For. Sci. **55**: 29-46.
- Lu, P., Müller, W. J., and Chacko, E. K. 2000. Spatial variations in xylem sap flux density in the trunk of orchard-grown, mature mango trees under changing soil water conditions. Tree Physiol. 20: 683-692.
- Mackay, D.S., Ewers, B.E., Loranty, M.M., and Kruger, E.K. 2010. On the representativeness of plot size and location for scaling transpiration from trees to a stand. J. Geophys. Res. 115

Marshall, D.C. 1958. Measurement of sap flow in conifers by heat transport. Plant Physiol. 33: 385–396

- McGill, B.J., Enquist, B.J., Weiher, E., and Westoby, M. 2006. Rebuilding community ecology from functional traits. Trends Ecol. Evol. **21**: 178-185.
- McJannet, D., Fitch, P., Disher, M., and Wallace, J. 2007. Measurements of transpiration in four tropical rainforest types of north Queensland, Australia. Hydrol. Process. **21**: 3549-3564.
- Meinzer, F.C., Goldstein, G., and Andrade, J.K. 2001. Regulation of water flux through tropical forest canopy trees: Do universal rules apply? Tree Physiol. **21**: 19-26.
- Mohan, J.E., Cox, R.M., and Iverson, L.R. 2009. Composition and carbon dynamics of forests in northeastern North America in a future, warmer world. Can. J. For. Res. **39**: 213-230.
- Moore, G.W., Bond, B.J., Jones, J.A., Phillips, N., Meinzer, F.C., 2004. Structural and compositional controls on transpiration in 40-and 450-year-old riparian forests in western Oregon, USA. Tree Physiol. **24**: 481-491.
- Nowacki, G. J., and Abrams, M.D. 2014. Is climate an important driver of post-European vegetation change in the Eastern United States? Global Change Biol.: in press. http://dx.doi.org/10.1111/gcb.12663
- Pederson, N., J. M. Dyer, R. W. McEwan, A. E. Hessl, C. J. Mock, D. A. Orwig, H. E. Rieder, and Cook, B.I.
  2014. The legacy of episodic climatic events in shaping temperate, broadleaf forests. Ecological
  Monographs:in press. doi: 10.1890/13-1025.1
- Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D. and Team, R.D.C. 2011. nlme: Linear and Nonlinear Mixed Effects Models.
- Raciti, S. M., T. J. Fahey, R. Q. Thomas, P. B. Woodbury, C. T. Driscoll, F. J. Carranti, D. R. Foster, P. S.
  Gwyther, B. R. Hall, S. P. Hamburg, J. C. Jenkins, C. Neill, B. W. Peery, E. E. Quigley, R. Sherman, M. A.
  Vadeboncoeur, D. A. Weinstein, and Wilson G. 2012. Local-scale carbon budgets and mitigation opportunities for the northeastern United States. BioScience 62:23–38.
- Swain, S., and Hayhoe K. 2014. CMIP5 projected changes in spring and summer drought and wet conditions over North America. Clim. Dyn.
- Tang, J., Bolstad, P.V., Ewers, B.E., Desai, A.R., Davis, K.J., and Carey, E.V. 2006. Sap flux-upscaled canopy transpiration, stomatal conductance, and water use efficiency in an old growth forest in the Great Lakes region of the United States. J. Geophys. Res. 111, DOI: 10.1029/2005JG000083
- Vadeboncoeur, M. A., Hamburg, S. P., Blum, J. D., Pennino, M. J., Yanai, R. D. and Johnson, C. E. 2012. The quantitative soil pit method for measuring belowground carbon and nitrogen stocks. S.S.S.A.J. 76: 2241–2255.

- Vertessy, R.A., Watson, F.G.R., and O'Sullivan, S.K. 2001. Factors determining relations between stand age and catchment water balance in mountain ash forests. Forest Ecol. Manag. **146**: 13-26
- Weeks, B.C., Hamburg, S.P., Vadeboncoeur, M.A. 2009. Ice storm effects on the canopy structure of a northern hardwood forest after 8 years. Can. J. For. Res. **39**:1475–1483.
- Wullschleger, S.D., Hanson, P.J., and Todd, D.E. 2001. Transpiration from a multi-species deciduous forest as estimated by xylem sap flow techniques. Forest Ecol. Manag. **143**: 205-213.
- Yanai, R. D., M. A. Vadeboncoeur, S. P. Hamburg, M. A. Arthur, C. B. Fuss, P. M. Groffman, T. G. Siccama, and Driscoll, C. T. 2013. From missing source to missing sink: Long-term changes in the nitrogen budget of a northern hardwood forest. Environ. Sci. Technol. 47: 11440–11448.
- Zhang, L., Hickel, K., Dawes, W.R., Chiew, F.H.S., Western, A.W., and Briggs, P.R. 2004. A rational function approach for estimating mean annual evapotranspiration. Water Resour. Res. 40, W02502, doi:10.1029/2003WR002710
- Zuur, A.F., Ieno, E.N., Walker, N.J., Saveliev, A.A. and Smith, G.M. 2009. Mixed effects models and extensions in ecology with R. Springer, New York.

	young stand	mature stand	
Year cut	1988 <sup>a</sup>	1883 <sup>b</sup>	
Elevation (m)	340	330	
Slope (%)	15-30	5-35	
Mean canopy height (m) <sup>c</sup>	27 ± 3	12 ± 1	
Mean canopy tree crown area (m <sup>2</sup> ) <sup>d</sup>	45	11	
LAI $(m^2 m^{-2})^{e}$	$4.1 \pm 0.6$	$4.4 \pm 0.2$	
Foliar litterfall (g m <sup>-2</sup> y <sup>-1</sup> ) $f$	261 ± 25	286 ± 24	
Sapwood water content (cm <sup>3</sup> cm <sup>-3</sup> )	0.456 ± 0.025	$0.436 \pm 0.008$	
Sapwood area (m <sup>2</sup> ha <sup>-1</sup> )	19.6	28.4	
Overall stem density in 2011 (ha <sup>-1</sup> )			
DBH > 10 cm	319	496	
DBH 2-10 cm	11,800	960	
Basal area by species in 2011 (m <sup>2</sup> ha <sup>-1</sup> )			
American beech (Fagus grandifolia Ehrh.)			
	7.8	16.4	
sugar maple (Acer saccharum Marsh.)	-	16.4 12.8	
sugar maple ( <i>Acer saccharum</i> Marsh.) red maple ( <i>Acer rubrum</i> L.)	7.8 - 4.9	-	
	-	12.8	
red maple (Acer rubrum L.)	4.9	<b>12.8</b> 0.4	
red maple ( <i>Acer rubrum</i> L.) yellow birch ( <i>Betula alleghaniensis</i> Britton)	<b>4.9</b> 1.1	<b>12.8</b> 0.4	
red maple ( <i>Acer rubrum</i> L.) yellow birch ( <i>Betula alleghaniensis</i> Britton) white birch ( <i>Betula papyrifera</i> Marsh.)	<b>4.9</b> 1.1 <b>2.9</b>	<b>12.8</b> 0.4	
red maple ( <i>Acer rubrum</i> L.) yellow birch ( <i>Betula alleghaniensis</i> Britton) white birch ( <i>Betula papyrifera</i> Marsh.) pin cherry ( <i>Prunus pensylvanica</i> L.)	<b>4.9</b> 1.1 <b>2.9</b> 4.7	<b>12.8</b> 0.4 <b>2.5</b> -	
red maple ( <i>Acer rubrum</i> L.) yellow birch ( <i>Betula alleghaniensis</i> Britton) white birch ( <i>Betula papyrifera</i> Marsh.) pin cherry ( <i>Prunus pensylvanica</i> L.) white ash ( <i>Fraxinus americana</i> L.)	<b>4.9</b> 1.1 <b>2.9</b> 4.7 0.1	<b>12.8</b> 0.4 <b>2.5</b> - - 2.4	
red maple ( <i>Acer rubrum</i> L.) yellow birch ( <i>Betula alleghaniensis</i> Britton) white birch ( <i>Betula papyrifera</i> Marsh.) pin cherry ( <i>Prunus pensylvanica</i> L.) white ash ( <i>Fraxinus americana</i> L.) striped maple ( <i>Acer pensylvanicum</i> L.)	<b>4.9</b> 1.1 <b>2.9</b> 4.7 0.1 1.5	<b>12.8</b> 0.4 <b>2.5</b> - 2.4 -	
red maple ( <i>Acer rubrum</i> L.) yellow birch ( <i>Betula alleghaniensis</i> Britton) white birch ( <i>Betula papyrifera</i> Marsh.) pin cherry ( <i>Prunus pensylvanica</i> L.) white ash ( <i>Fraxinus americana</i> L.) striped maple ( <i>Acer pensylvanicum</i> L.) eastern hemlock ( <i>Tsuga canadensis</i> L.)	<b>4.9</b> 1.1 <b>2.9</b> 4.7 0.1 1.5	<b>12.8</b> 0.4 <b>2.5</b> - 2.4 - 0.2	

**Table 1.** Stand characteristics. Basal area of the three species for which sapflow was measured in each stand are indicated in bold.

a. Timber sale date from USFS records.

b. Approximate date from local history and hardwood increment cores. Best estimate of cutting date is based on a release recorded in the growth rings of a 200+ year-old hemlock adjacent to the plots.

c. Measured May 2011; mean of randomly chosen trees >10cm DBH. SD is among trees.

d. Measured in trees corresponding with plot corners on a 2013 aerial photograph with 0.3m resolution. Estimates of the mean are rough as individuals vary greatly.

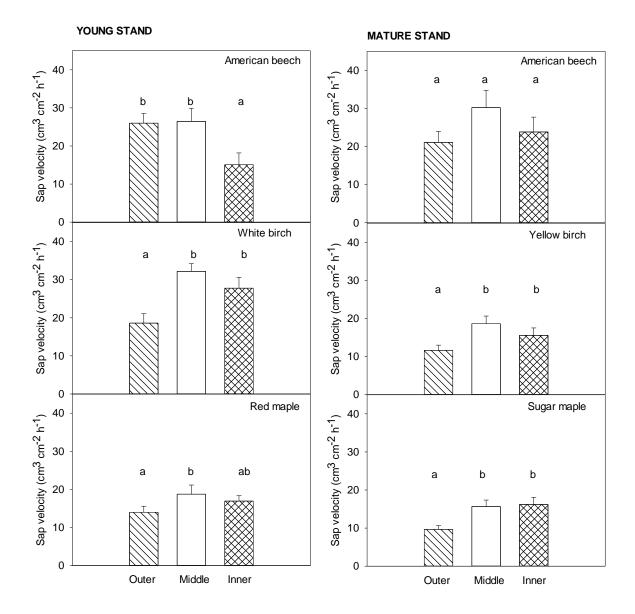
e. Measured in August 2004 with a LICOR LAI-2000; SD is shown for n=15 observations per stand.

f. Litterfall collected in 15-20 baskets per site, deployed from early September to early November in 2004, '05, '09, '11, '12, and '13; SD shown among years. No significant trends were observed in litter mass at either site. R.D. Yanai lab, unpublished.

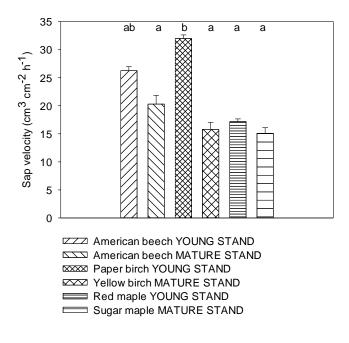
**Table 2.** Allometric equations between DBH (cm) and sapwood area (cm<sup>2</sup>). The equation is in the form of a power function ( $A_s = a * DBH^b$ ).

Tree species	n (trees)	a	h	r <sup>2</sup>	DBH range (cm)
sugar maple	24	0.78	1.97	0.95	22.6-42.7
(Acer saccharum)					
yellow birch	21	1.17	1.79	0.95	12.5-46.5
(Betula alleghaniensis)					
beech	47	0.67	1.92	0.98	2.2-33.4
(Fagus granalfiora)					
white birch	20	0.71	2.04	0.99	5.5-14.5
(Betulu pupyrijeru)					
red maple (Acer rubrum)	20	0.83	1.97	0.99	5.6-12.2
	(Acer saccharum) yellow birch (Betula alleghaniensis) beech (Fagus grandiflora) white birch (Betula papyrifera)	sugar maple 24 (Acer saccharum) yellow birch 21 (Betula alleghaniensis) beech 47 (Fagus grandiflora) white birch 20 (Betula papyrifera) red maple 20	sugar maple240.78(Acer saccharum)91.17yellow birch211.17(Betula alleghaniensis)470.67beech470.67(Fagus grandiflora)0.71white birch200.71(Betula papyrifera)200.83	sugar maple240.781.97(Acer saccharum)211.171.79yellow birch211.171.79(Betula alleghaniensis)470.671.92beech470.671.92(Fagus grandiflora)200.712.04white birch200.712.04(Betula papyrifera)200.831.97	sugar maple         24         0.78         1.97         0.95           (Acer saccharum)         21         1.17         1.79         0.95           yellow birch         21         1.17         1.79         0.95           (Betula alleghaniensis)         47         0.67         1.92         0.98           (Fagus grandiflora)         20         0.71         2.04         0.99           (Betula papyrifera)         20         0.83         1.97         0.99

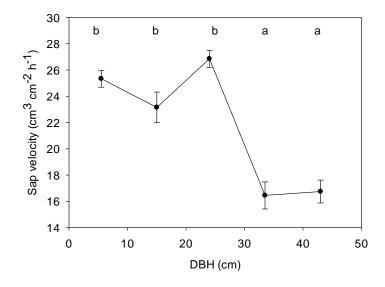
**Figure 1.** Mean sap velocity measured at three radial positions along the sapwood depth (outer, middle and inner for 1.0, 2.2, and 3.5 cm from the cambium), left figures are the species located in the young stand and on the right are the species of the mature stand. Error bars denote 1 SE. Letters indicate significant differences among radial positions.



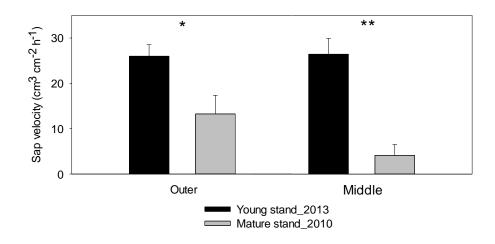
**Figure 2.** Comparison of mean sap velocity measured at 30%-60% of total sapwood depth along the study period for the different studied species both in the mature and young stand. The bars represent mean sap velocity and the error bars are 1 SE. Different letters indicate significant differences across all categories (P<0.05).



**Figure 3.** Mean sap velocity measured at 30%-60% of total sapwood depth estimated allometrically in trees of different sizes of the two stands studied. DBH ranges considered are ≤11.0, 11.1-19.0, 19.1-29.0, 29.1-38.0, ≥38.1 cm. The points plotted at the midpoint of each DBH range represent mean sap velocity. Error bars show 1 SE.



**Figure 4.** Mean sap velocity in American beech trees of similar DBH but with different canopy position. In the young stand, trees of ~10 cm DBH are dominant, while in the mature stand trees of this same size are sub-dominant. Error bars are 1 SE (\* indicates P<0.05, \*\* indicates P<0.01). The inner sensor was not included because it was located in the heartwood of some young trees.



**Figure 5.** Hypothetical transpiration of scenarios with 100% dominance by each of the species for young and mature stand using the data obtained from 5 July to 14 August 2013. Black bar represents transpiration of trees > 10 cm in DBH, while empty bar represent trees 2-10 cm in DBH.

