



# 

**Citation:** Carrillo P, Medina-Sánchez JM, Herrera G, Durán C, Segovia M, Cortés D, et al. (2015) Interactive Effect of UVR and Phosphorus on the Coastal Phytoplankton Community of the Western Mediterranean Sea: Unravelling Eco-Physiological Mechanisms. PLoS ONE 10(11): e0142987. doi:10.1371/journal.pone.0142987

Editor: Christopher J. Gobler, Stony Brook University, UNITED STATES

Received: May 27, 2015

Accepted: October 29, 2015

Published: November 23, 2015

**Copyright:** © 2015 Carrillo et al. This is an open access article distributed under the terms of the <u>Creative Commons Attribution License</u>, which permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.

**Data Availability Statement:** All relevant data are within the paper and its Supporting Information files.

Funding: This research was supported by the Spanish Ministries of Science and Innovation (CGL2011-23681/BOS), Consejería de Innovación, Ciencia y Empresa – Junta de Andalucía' (Excelencia CVI-02598) and Spanish government 'Formación de Profesorado Universitario fellowship to G. Herrera and C Durán. The funders had no role in study design,data collection and analysis, decision to publish, or preparation of the manuscript. RESEARCH ARTICLE

# Interactive Effect of UVR and Phosphorus on the Coastal Phytoplankton Community of the Western Mediterranean Sea: Unravelling Eco-Physiological Mechanisms

Presentación Carrillo<sup>1,2</sup>\*, Juan M. Medina-Sánchez<sup>1,2</sup>, Guillermo Herrera<sup>1</sup>, Cristina Durán<sup>1</sup>, María Segovia<sup>3</sup>, Dolores Cortés<sup>4</sup>, Soluna Salles<sup>4</sup>, Nathalie Korbee<sup>3</sup>, Félix L. Figueroa<sup>3</sup>, Jesús M. Mercado<sup>4</sup>

1 Instituto del Agua, Universidad de Granada, Granada, Spain, 2 Departamento de Ecología, Facultad de Ciencias, Universidad de Granada, Granada, Spain, 3 Departamento de Ecología, Facultad de Ciencias, Universidad de Málaga, Málaga, Spain, 4 Centro Oceanográfico de Málaga, Instituto Español de Oceanográfía Fuengirola, Málaga, Spain

\* pcl@ugr.es

# Abstract

Some of the most important effects of global change on coastal marine systems include increasing nutrient inputs and higher levels of ultraviolet radiation (UVR, 280-400 nm), which could affect primary producers, a key trophic link to the functioning of marine food webs. However, interactive effects of both factors on the phytoplankton community have not been assessed for the Mediterranean Sea. An in situ factorial experiment, with two levels of ultraviolet solar radiation (UVR+PAR vs. PAR) and nutrients (control vs. P-enriched), was performed to evaluate single and UVR×P effects on metabolic, enzymatic, stoichiometric and structural phytoplanktonic variables. While most phytoplankton variables were not affected by UVR, dissolved phosphatase (APA<sub>EX</sub>) and algal P content increased in the presence of UVR, which was interpreted as an acclimation mechanism of algae to oligotrophic marine waters. Synergistic UVR×P interactive effects were positive on photosynthetic variables (i.e., maximal electron transport rate, ETR<sub>max</sub>), but negative on primary production and phytoplankton biomass because the pulse of P unmasked the inhibitory effect of UVR. This unmasking effect might be related to greater photodamage caused by an excess of electron flux after a P pulse (higher ETR<sub>max</sub>) without an efficient release of carbon as the mechanism to dissipate the reducing power of photosynthetic electron transport.

# Introduction

Coastal marine ecosystems contribute about 30–35% of the global production of phytoplankton in oceanic waters [1] and play a key ecological and economic role for human populations [2]. Due to the pivotal role of the marine primary productivity, any change caused by **Competing Interests:** The authors have declared that no competing interests exist.

PLOS ONE

alterations in abiotic limiting factors such as light or nutrients can have consequences for the functioning of complex marine food webs [2,3]. Some of the most important effects of global change on coastal marine systems include increasing input of nitrate and phosphate, brought about by both climate change (e.g. atmospheric deposition, wind, precipitation) and increased land use (e.g. agricultural activities [4]) causing eutrophication [5].

In addition, phytoplankton might be exposed to higher levels of solar ultraviolet radiation (UVR, 280-400 nm) because of both the greater stratification of the water column due to global warming [6,7,8] as well as the depletion of the ozone layer [9] or low ozone episodes (ozone mini-hole), which imply UVR short and variable intensity pulses, (i.e. 1-5 days [10], as occurs in Mediterranean coastal areas [11]. It has also been reported that at some sites of the Northern Hemisphere, UVB irradiance may continue increasing because of continuous reduction in aerosol extinction since 1990 [12]. Reductions in aerosols and clouds are expected to overcompensate for the effect of ozone recovery UVR after the Montreal protocol. Therefore, UVR still remains as a world-wide stressor with far-reaching implications for ecological interactions [13]. UVR causes damage to the cells directly, via the photochemical degradation of biomolecules (e.g. nucleic acids, lipids, proteins) [14], or indirectly, via the production of reactive oxygen species (ROS) generated by chemical, photolytic reactions inside and outside the cells [15,16]. The negative effects of UVR at the sub-cellular level could lead to reductions in productivity and growth rates at the phytoplankton community level [17,18,19]. However, previous studies also have evidenced a null [20] or positive [21] effect of UVR on phytoplankton. Different protective strategies such as improved efficiency of repair processes at the sub-cellular level of algae exposed to high UVR [22], synthesis of photoprotective compounds [23,24] or changes in taxonomic composition towards communities better adapted to the new conditions [25] are among the mechanisms evolved to allow the acclimation of algae to UVR in marine waters.

The mechanisms that are best suited for adaptation to UVR depend on cell-nutrient content [13]. The direct effects of UVR-inorganic nutrient interactions on phytoplankton can be twofold: firstly, UVR affects nutrient uptake and/or assimilation capacities of the phytoplankton, which is often a UVR species-specific response [26]. Nevertheless, the interaction of UVR with P acquisition mechanism is likely to reflect additional complexities related to biogeochemical processes and/or biotic interactions. Access to dissolved organic phosphorus (DOP) can be increased by alkaline phosphatase (APA), which catalyses the hydrolysis of phosphate esters into orthophosphate [27]. This extracellular enzyme may play a major role in P supply in Pdeficient ecosystems, in which phosphate concentrations are usually negligible and most of the P is bound to organic matter [28]. There is evidence for both inhibition [29] and stimulation [30,31] of APA by UVR. Secondly, the nutrient status of phytoplankton may determine the susceptibility of these organisms to UVR [32]. Despite that nutrient-deficient phytoplankton, especially those limited by N [6,25,33], are generally considered to be more sensitive to UVR than are nutrient-replete phytoplankton, some authors have reported a higher sensitivity of phytoplankton to UVR under nutrient-replete conditions [34]. Variations in the type of nutrient that limits phytoplanktonic growth or the severity of this limitation, as well as differential tolerance/ response to UVR by the phytoplanktonic community [26], are some of the potential interpretations proposed to explain the wide variability of algal responses to UVR and nutrients.

Moreover, it has been shown that, together with the nutrient content in the phytoplanktonic cells, their elemental composition (C:P and N:P ratios) plays a key role in determining the algal response to UVR. Along this line, several studies have demonstrated a direct effect of UVR, decreasing elemental C:P in marine and freshwater phytoplankton [20,31,35,36]. Due to the relevance for C and P cycling in the marine ecosystems, it is of interest whether C:P under UVR exposure relate primarily to metabolic features such as lower C-fixation, greater C-losses

by respiration or organic carbon release, and/or whether the exposed cells also have the capability of improving their P-content via a stimulatory UVR effect on APA activity [31].

The literature addressing UVR effects on inorganic nutrient cycling has focused mainly on N [37,38,24], as this element has generally been reported to limit primary production in marine environments. Less is known about interactions between UVR and P utilization and the phytoplanktonic metabolism in marine ecosystems. The Mediterranean Sea occupies an oligotrophic basin where primary production is recognized to be limited by P [39,40,41], although more-recent reports have modified this paradigm [42]. Several studies on nutrient enrichment have shown that plankton growth is often stimulated after P addition in surface waters during the stratification period [43,44,45,46].

Information about the interactive effect of UVR and P in oligotrophic marine ecosystems is noticeably lacking. This topic is raising great interest within the current global-change scenario, where inorganic inputs from terrestrial origin linked to human activity or due to increasing intensity and frequency in atmospheric dust inputs rich in P from the Sahara Desert is reaching the Mediterranean Sea [47,48] (http://ozoneaq.gsfc.nasa.gov/).

The aim of the present study was to quantify single and combined effects of UVR flux and changes in the P supply on primary producers, paying special attention to the physiological mechanisms involved in the responses of the phytoplankton communities. Our hypothesis is that single effects of UVR and P-addition will be inhibitory and stimulatory, respectively, on marine phytoplankton, thereby producing an antagonistic interactive effect consisting of the attenuation of inhibitory UVR effects by P addition. For this purpose, an *in situ* experiment was conducted in oligotrophic clear waters at Cabo de Gata, on the Alborán Sea, Spain. Light quality and nutrient (P) availability were manipulated at a medium-term scale since the phytoplankton response to these nutrient-enrichment episodes is quite rapid (24–48 h) and phytoplankton bloom decline within 5–6 days after reaching the biomass peak [49,50].

#### **Material and Methods**

#### Study site

Water samples for the experiments were collected on board B/O F. de P.Navarro (Spanish Oceanography Institute) offshore of Cabo de Gata Natural Park (36° 33'N, 2° 16'W), located on the eastern end of the Alborán Sea. The sampling was performed during summer, when the water column was characterized by a strong thermal stratification, leading to an impoverishment of nutrients in the euphotic layer. Cabo de Gata is bio-optically characterized by very clear waters, resulting in very low values of diffuse attenuation coefficients for downward irradiance and high surface irradiances (Fig 1A). No specific permissions were required to conduct our experiments on microplanktonic communities due to the low impact on area of study. Our research involve no endangered or protected species.

#### Experimental set-up

The experiment consisted of a factorial design with two light treatments [full sunlight (UVR +PAR; 280–700 nm) and only photosynthetic active radiation (PAR; PAR;400-700nm)] and two phosphorus (P) treatments [increased P (P-enriched treatment) and ambient nutrient conditions (control treatment)]. Water samples were collected with Niskin bottles at different depths within the upper 10-m layer at the sampling station (photic layer receiving > 1% surface UVR<sub>305</sub>). The bottle water was pooled and mixed in order to fill a 240-L tank. Water was pre-screened through a 200- $\mu$ m nylon mesh to remove mesozooplankton. Mesozooplankton was excluded because the volume of the microcosms was not suitable to maintain a grazer community for a week, and therefore we avoided the lack of replicability in the microcosms due to





**Fig 1. Solar radiation during the experiments in Cabo de Gata.** (A) Depth profiles of the irradiance at 305, 320, and 380 nm, and PAR (400–700 nm) and values of vertical attenuation coefficients (m<sup>-1</sup>); (B) Solar radiation reaching the deck of the boat during the experiments. Solar radiation was continuously monitored using a broad-band filter radiometer instrument (ELDONET, Real Time Computer, Möhrendorf, Germany) placed on the deck.

the uneven effect of herbivores on phytoplankton. Tank water was used to fill twelve 20-L lowdensity polyethylene (LPDE) (Plásticos Andalucía, Spain) microcosms in triplicate. LPDE transmits 90% of photosynthetic active radiation, 75% of UVA and 60% of UVB).

Microcosms were incubated from 21 to 26 September 2009 in two tanks on deck of the boat to avoid possible damage underwater. To ensure natural light conditions, we painted the two tanks black, thereby avoiding background reflection. Temperature was kept constant by continuously pumping surface seawater. For the UVR-exclusion treatment, the tank was covered

with a polycarbonate filter of UF3 plexiglass (Atohaas, USA) which screens out UVR < 390nm and transmits 90% of the PAR. For the P-enrichment treatments,  $KH_2PO_4$  was added to half of the microcosms to a final concentration of  $30\mu g P L^{-1}$ , which are the estimated P inputs from Saharan dust to the Mediterranean area [51]. Different samples of each microcosm were collected on the sixth incubation day in order to measure different chemical, biological, and physiological variables.

### Physical measurements

Downwelling irradiances for the bands 305, 320, and 380 nm, and PAR as well as temperature were measured at different depths within the water column at noon by using a submersible BIC Compact 4-Channel Radiometer (Biospherical Instruments Inc. CA, USA). Diffuse attenuation coefficients for downward irradiance ( $k_d$ ) were calculated from the slope of the linear regression of the natural logarithm of downwelling irradiance vs. depth for the different radiation bands. Solar radiation received by the microcosms at the surface was continuously monitored using a broad-band filter radiometer instrument (ELDONET, Real Time Computer, Möhrendorf, Germany) placed on the boat deck.

#### Chemical and stoichiometric analyses

Samples for total and dissolved inorganic nutrient (N and P) were frozen at -30°C until analysed. Nitrate, nitrite, phosphate, and silicate were measured by a Technicon Auto Analyzer (TrAAcs 800, Bran-Leubbe) following [52] Ramírez et al. (2005). For determination of total nitrogen (TN) and total phosphorus (TP), the samples were treated with an oxidizing agent (a mixture of peroxysulphate and boric acid in NaOH) and autoclaved for 30–40 min. Nitrate and phosphate concentrations in these samples were determined by following the methods described by Grasshoff et al. [53].

Sestonic elemental composition and stoichiometric ratios were determined from 500-mL aliquots of sample water filtered through pre-combusted (1h at 550°C) 0.7-µm filters (What-man GF/F) at low pressure (< 100mmHg). Filters for P, N, and C were immediately frozen at -20°C. C and N analyses were performed using a Perkin-Elmer 2400 elemental analyser. P filters were introduced in acid-washed vials and digested with potassium persulfate and boric acid at 120°C for 30 min. Particulate P was then determined as SRP in 10-cm quartz cuvettes by the acid molybdate technique [53]. C:P, N:P and C:N ratios were calculated on a molar basis. Dissolved organic carbon (DOC) was measured using a total organic carbon analyser (TOC-V CSH/CSN, Shimadzu) in samples pre-filtered through pre-combusted (2 h at 500°C) glass-fiber filters (Whatman GF/F and acidified with HCl 1N (2%).

#### Analyses of functional and structural variables

Alkaline phosphatase activity. APA was determined following [54] total and extracellular APA (APA<sub>T</sub> and PA<sub>EX</sub>) were determined from unfiltered water sample from the microcosms (< 200 µm mesh) and samples filtered through 0.2-µm filters (Anodisc, Whatman) respectively. APA was calculated in nMP h<sup>-1</sup> from a reference curve constructed using different concentrations of 4-methylumbelliferyl phosphate hydrolysed to methylumbelliferone (MUF, Sigma-Aldrich) and measured fluorometrically ( $\lambda_{em}$ 365nm and  $\lambda_{ex}$ 440nm)(Perkin-Elmer LS 55 luminescence-spectrometer) after 45 min. Particulate APA was calculated as the difference between APA<sub>T</sub> and APA<sub>EX</sub>.

**Respiration rates.** Two 25-mL samples from each microcosm were used to measure the total microplanktonic respiration (TMR) and respiration of the fraction  $< 0.7 \ \mu m \ (R_{< 0.7 \mu m})$ . It was estimated from oxygen depletion in darkness measured with sensor-spot optodes

(SP-PSt3-NAU-D5-YOP and Fibox3; PreSens GmbH, Germany). Optodes followed a twopoint calibration (0 and 100% oxygen saturation). Zero % point calibration was performed by adding sodium sulphite (Na<sub>2</sub>SO<sub>3</sub>) to a final concentration exceeding 0.1mg mL<sup>-1</sup>, and 100% by inserting wet cotton wool into the closed flask to ensure 100% O<sub>2</sub>-saturated water-vapour air. Flasks were incubated under dark conditions and constant ambient temperature (20°C) in culture chambers for 24 h and measured every 2h after 30 min of acclimation. Oxygen data were then adjusted to a linear model via least-squares regression. The slope of the regressions provided the oxygen-consumption rates. Respiration of phytoplankton was calculated as the difference between TMR and R  $_{< 0.7 \mu m}$ .

**Reactive oxygen species.** The intracellular accumulation of reactive oxygen species was estimated by using a modification of the method used by Segovia and Berges [55] described elsewhere [15]. Briefly,ROS was detected using carboxy-H<sub>2</sub>DFFDA (Invitrogen, Oregon, USA). Cells were incubated with  $5\mu$ M c-H<sub>2</sub>DFFDA (final concentration) at 16°C for 90 min in darkness. Fluorescence was detected using a microplate fluorescence reader (FL-600, BIO-TEK) at an excitation of 490 nm and emission of 525 nm. The concentration of ROS in the cells was then expressed as relative fluorescence units per cell.

**Photosynthetic rate as** *in vivo* **Chlorophyll Fluorescence**. *In vivo* chlorophyll fluorescence was determined by using a Water-PAM pulse amplitude modulated fluorometer (Walz GmbH, Effeltrich, Germany). First, 1.5 mL of water from each microcosm was placed into the 15-mm-diameter quartz cuvette inserted in the measuring chamber. After a 15-min dark-adaptation period followed by a 5-s far-red (FR) light pulse to ensure full oxidation of  $Q_A$ , a rapid light curve (RLC) program (WINCONTROL software, Walz GmbH) was initiated. RLCs were constructed from a 20-s exposure to each of 11 incremental irradiances (30, 46, 69, 104,158, 230, 350, 520,737, 1017, and 1642, µmol photons·m<sup>-2</sup>·s<sup>-1</sup>).

RLCs were constructed by calculating the electron-transport rate (ETR) through PSII for each level of actinic light:

$$ETR = \left(\frac{F'_m - F}{F'_m}\right) * AQ \quad \cdot \quad \text{FII} \tag{1}$$

where (F'm-F)/F'm estimates the effective quantum yield of PSII, and AQ is the absorbed quantum expressed as  $\mu$ mol m<sup>-2</sup> s<sup>-1</sup>, calculated as follows:

$$AQ = \sum_{400}^{700} a_{\rm ph}(\lambda) \times E(\lambda) \tag{2}$$

The specific absorption coefficient  $(a_{ph} (\lambda), expressed as m^{-1})$  was estimated by calculating the light absorption by phytoplankton concentrated on glass-fiber filters corrected as described by Korbee [56].  $E(\lambda)$  is the spectral irradiance of the LED lamp of the Water-PAM and  $F_{II}$  is a multiplication factor (adimensional) that expresses the fraction of absorbed quanta to PS<sub>II</sub>, between 400–700 nm. This value was calculated for each treatment as a function of the percentages of bacillariophyceae, dinophyceae I, and flagellates, taking into account the  $F_{II}$  values proposed by Johnsen and Sakshaug [57] for the different groups. In all cases, we used the values for the high-light-acclimated cultures, as these organisms, being cultured at the surface, were exposed to high irradiances.RLC data were fitted to the model of Eilers and Peeters [58] to calculate values for the initial slope ( $\alpha_{ETR}$ ) and maximal electron transport rate (ETR<sub>max</sub>). The light-saturation parameter ( $E_k$ ) was derived from ETR<sub>max</sub> and  $\alpha_{ETR}$ :

$$E_k = \frac{ETR_{max}}{\alpha} \tag{3}$$

**Photosynthetic pigments and xanthophyll cycle.** A 300-mL sample of water from each microcosm was filtered through glass fibre (Whatman GF/F filters) at < 100mmHg. Filters were preserved at -80°C until analysed. The filters were immersed in N, N-dimethylformamide and extracted overnight at 4°C. Pigment concentration in the extract was determined by HPLC after filtration through 0.2  $\mu$ m following Lubian and Montero [59]. Identification and quantification were made using commercial standard (DHI LAB products). Concentrations of chlorophyll a (Chl *a*), chlorophyll c2 (Chl c2), diadianoxanthin (Dd), diatoxanthin (Dt), fucoxanthin, zeaxanthin, and traces of other xanthophyll pigments were detected. The Dd de-epoxidation degree (in %) was calculated as follows:

$$Dd \ de - epoxidation = \frac{Dt}{Dd + Dt}$$
(4)

Primary production and excreted organic carbon. From each microcosm, 200 mL were used for primary production (PP) measurements following the <sup>14</sup>C method proposed by Steeman-Nielsen [60]. For this, 9.25 MBq of NaH<sup>14</sup>CO<sub>3</sub> (specific activity: 310.8 MBq mmol<sup>-1</sup>, DHI Water and Environment, Germany) were added to sets of four 50-mL flasks (three clear and one dark as control). The flasks were incubated in the tanks at the same depth as the microcosms, for 4 h symmetrically distributed around noon. The total organic carbon (TOC) produced was measured on 4-mL aliquots before filtration. Microplankton primary production  $(PP_M)$  was determined from particles retained by filtering the sample through 3-µm glass-fiber filters (Whatman GF/D). The filtered water was used for picoplanktonic primary production  $(PP_P)$  in a subsequent filtration and estimated from particles retained in 0.7  $\mu$ m (Whatman GF/F). Filtrations at low pressure (< 100 mmHg) were used to minimize cell breakage. The excreted organic carbon (EOC) was measured from 4-mL aliquots of the < 0.7-µm filtrate. Filters and filtrate were placed in 20-mL scintillation vials and acidified with 100  $\mu$ L of 1N HCl to remove DI14C. Vials were then kept open for 24 h in an aeration hood following the recommendations of Lignell [61]. Sixteen mL of scintillation cocktail (Ecoscint A) were added for scintillating the vials and counted using a scintillation counter (Beckman LS 6000TA) equipped with autocalibration. The %EOC was estimated as the percentage of EOC to TOC:

$$\% EOC = \frac{EOC}{TOC} \times 100 \tag{5}$$

Abundance/biomass and taxonomical composition of phytoplankton. Water samples of 200 mL were taken for taxonomical identification and quantification of autotrophic microplankton > 5 µm) by means of inverted microscope. The sample was preserved in acetic lugol solution and cell abundance was determined by the Utermöhl technique. Also, 5 mL of sample were fixed with glutaraldehyde in order to quantify cell abundance of autotrophic picoplankton (AP; *Prochlorococcus, Synechococcus* and pico-eukaryotes with a Becton Dickinson FACScan flow cytometer; more details in [62]). Biovolumes of the three pico-phytoplankton cell groups analysedwere calculated following Ribes et al. [63] for samples collected in the north-western Mediterranean Sea. Biovolume values were converted into biomass by using the formulae proposed by Morel et al. [64] for *Prochlorococcus* and Kana and Glibert [65] for *Synechococcus*. Cell biovolumes corresponding to the most abundant species, genus, and other taxa of microphytoplankton identified by inverted microscope were calculated by using the appropriate formula according to their geometric shape [66, 67, 68]. Biovolumes were converted into biomass using the formulae proposed by Verity et al. [69] for phytoplankton < 15 µm and Menden-Deuder and Lessard [70] for >15 µm dinophycecae and bacillariophyceae.

# Statistical analysis

A full factorial two-way ANOVA test was used to evaluate the interactive effects of light and nutrients on all variables after normality (Kolmogorov-Smirnov test) and homoscedasticity (Cochran C test) were verified. When interactive effects of both factors were found, a *post hoc* Tukey HSD test was used to account for significant differences between treatments. All tests were performed using R2.15 software (R Development Core Team).

# Results

#### Starting conditions of the experiments

Changes with depth of the downwelling irradiance of the different spectral bands as well as surface solar radiation during the experimental period are shown in <u>Fig 1A and 1B</u>. There was a weak at tenuation of UVR in the water column, since UVB and UVA spectral bands reached 13 and 25 m in depth, respectively (Fig 1A). The penetration depth of PAR was greater, although the attenuation coefficient ( $k_{PAR}$ ) was also relatively low (0.075 m<sup>-1</sup>). Surface UVR and PAR irradiance (Fig 1B) proved similar during the first five days of the experiment, reaching maximal values of ca. 70 and 450 W m<sup>-2</sup> for UVR and PAR, respectively. By contrast, the last day of experimentation was cloudy and consequently the maximal irradiance was comparatively lower. The daily mean irradiance values during the experimental period (i.e. total duration of exposure of microcosms) for UVB, UVA, and PAR were 0.14, 18.72 and 135.88 W m<sup>-2</sup>, respectively. The concentrations of total and inorganic forms of N and P indicate that most of the nutrients were in particulate form incorporated into the organisms. The DIN:TP ratio was low, implying likelihood of N limitation for the pelagic community, although the sestonic N:P ratio was close to the Redfield ratio (Table 1). Chl *a* concentrations were < 1 µg L<sup>-1</sup>, this being characteristic of oligotrophic ecosystems.

Total initial phytoplankton was ~7.8  $10^3$  cells mL<sup>-1</sup> and 172.14 µgC L<sup>-1</sup>, in term of abundance and biomass, respectively (<u>Table 1</u>). Diatoms represented 58% total biomass, followed by flagellates (26%). Picoplankton made up 14% of the total biomass (<u>Table 2</u>), although the dominant group in terms of abundance was Synechococcus sp, which accounted for 33% and picoeukaryotes 44% of total abundance

# UVR and P effect on ROS, R, APA, DOC, and C:N:P ratio

Under non-enriched conditions, UVR significantly augmented R and ROS values (Fig 2A and 2B). P enrichment eliminated this increasing effect of UVR, generating a significant UVR×P effect (Table A in S1 File), which, according to our hypothesis, reduced the UVR-stress effect on both variables. Similarly, UVR alone had a significant effect on APA<sub>T</sub> and APA<sub>EX</sub> (Fig 2C and 2D), increasing APA activities by 25 and 265% respectively, but not on particulate APA (data not shown). As expected, P-enrichment significantly reduced APA activities in both fractions and cancelled the stimulatory UVR effect on APA<sub>T</sub> and APA<sub>EX</sub>, although no significant UVR×P effect on them was found (Table A in S1 File). DOC, in non-enriched treatments, decreased (60%) under UVR compared to PAR (Fig 2E). After P enrichment, DOC decreased in relation to PAR control treatment in both light treatments that generated a significant interactive UVR×P effect on this variable (Fig 2E, Table A in S1 File).

Noticeably, under non-enriched conditions, C, N, and P cell-quota and POC,PON, and POP increased significantly under UVR (Fig 3A–3F). The relative increase of POP was higher in the UVR (60%) than in the PAR treatment, resulting in a significant decline in C:P molar ratio (25%) (Fig 3G). P enrichment significantly diminished values of C-, N-but not P cell-quota (Fig 3C). By contrast, P enrichment resulted in a significant increasing effect only on

 Table 1. Physical, chemical, and biological variables at initial conditions of the experiment in Cabo de Gata.

 Values are mean and standard deviation (sd)Variables.

Physical, chemical, and biological variables at initial conditions of the experiment in Cabo de Gata	Mean values ± sd
T (°C)	21.1±1.20
SRP (µM)	0.10 ±0.06
ΤΝ (μΜ)	4.12 ±0.60
ΤΡ (μΜ)	1.50 ±0.45
DIN (NO <sub>3</sub> +NO <sub>2</sub> <sup>-</sup> )(µM)	0.41±0.01
DIN/TP (molar)	0.27±0.05
Si (µM)	0.24± 0.012
Chl a (μg L⁻¹)	0.99±0.23
Total APA (nM P h <sup>-1</sup> )	8.57±1.27
PP <sub>M</sub> (µg C·L <sup>-1</sup> )	2.37±0.88
PP <sub>P</sub> (μg C·L <sup>-1</sup> )	0.62±0.25
POC (µM)	26.43±3.04
PON (µM)	3.09±0.39
ΡΟΡ (μΜ)	0.15±0.02
C:N ratio	8.57±0.17
N:P ratio	20.30± 4.79
C:P ratio	173.49 ±38.62
DOC (mgL <sup>-1</sup> )	1.72±0.23
PA (cell·mL <sup>-1</sup> )	7,895±1,489
PB (µg C·L <sup>-1</sup> )	172.14±32.20

Abbreviations: T = temperature; SRP = soluble reactive phosphorus; TN = total nitrogen; TP = total phosphorus; DIN (NO<sub>3</sub><sup>-</sup>+NO<sub>2</sub><sup>-</sup>) = dissolved inorganic nitrogen (nitrate + nitrite); Si = silicates; Chl a = chlorophyll a; APA = total alkaline phosphatase activity; PP<sub>M</sub> = microplanktonic primary production; PP<sub>P</sub> = picoplanktonic primary production; POC = particulate organic carbon; PON = particulate organic nitrogen; POP = particulate organic phosphorus; DOC = dissolved organic carbon; PA = phytoplanktonic abundance; PB = phytoplanktonic biomass. vAll measurements were made on the mixed water simple, taken from the photic layer receiving>1% surface UVR<sub>305</sub>

doi:10.1371/journal.pone.0142987.t001

Table 2. Changes in the taxonomical composition of phytoplankton communities during the experiment. Phytoplanktonic groups or species biomass ( $\mu$ g C L<sup>-1</sup>) and percentage of biomass (%) under initial conditions and after each treatment. Values are mean and standard deviation whereas the letters indicate differences among treatments for the different phytoplanktonic groups and total biomass.

			CONTROL				P-ENRICHED			
	INITIAL		UVR+PAR		PAR		UVR+PAR		PAR+P	
	Mean ±sd	%	Mean ±sd	%	Mean ± sd	%	Mean ± sd	%	Mean ± sd	%
Bacillariophyceae	100.46±16.13	58.36	205.74±23.27 <sup>a</sup>	57.73	175.42±62.81 <sup>a</sup>	52.36	136.24±26.13 <sup>a</sup>	68.95	157.98±35.95 <sup>a</sup>	47.90
Dinophyceae	1.65±0.96	0.01	2.94±1.09 <sup>a</sup>	0.01	2.02±1.00 <sup> a</sup>	0.01	2.32±2.16 <sup> a</sup>	0.01	1.32±0.46 <sup>a</sup>	0.00
Flagellates	44.89±0.87	26.07	146.45±34.39 <sup>a</sup>	41.10	156.15±36.66 <sup>a</sup>	46.61	52.47±13.54 <sup>b</sup>	26.56	156.54±26.84 <sup>a</sup>	47.47
Picoeukaryotes	22.31±0.45	12.96	0.85±0.19 <sup>a</sup>	0.24	0.74 ± 0.07 <sup>a</sup>	0.22	5.22±0.25 <sup>b</sup>	2.64	11.94±0.50 <sup>c</sup>	3.62
Prochlorococcus sp.	0.004±0.001	0.00	0.003±0.002 <sup>a</sup>	0.00	0.001 ± 0.0003 <sup>a</sup>	0.00	0.002±0.0003 <sup>a</sup>	0.00	0.002±0.001 <sup>a</sup>	0.00
Synechococcus sp.	2.83±0.08	1.64	0.37±0.02 <sup>a</sup>	0.11	0.68 ±0.02 <sup>b</sup>	0.20	1.34±0.12 <sup>c</sup>	0.68	2.02±0.12 <sup>d</sup>	0.61
Total biomass	172.14±32.39		356.37±5 9.31 <sup>a</sup>		335.02 ±99.45 <sup>a</sup>		197.61±41.24 <sup>b</sup>		329.80±62.43 <sup>a</sup>	

doi:10.1371/journal.pone.0142987.t002



**Fig 2.** Response of metabolic, enzymatic variables and dissolved organic carbon to experimental conditions. (A) Respiration rates (R, in  $\mu$ MO<sub>2</sub> h<sup>-1</sup>); (B) Reactive oxygen species (ROS, in RFUs); (C) total alkaline phosphatase (APA<sub>T</sub> in nMP h<sup>-1</sup>); (D) Dissolved alkaline phosphatase (APA<sub>EX</sub> in nMP h<sup>-1</sup>); (E) Dissolved organic carbon (DOC in mgC L<sup>-1</sup>) under photosynthetically active radiation (PAR), and full sunlight (PAR+UVR) in nutrient-enriched (P-enriched) and non-enriched (control) treatments. Horizontal dashed lines indicate value for initial day. Data are expressed as mean values ± sd (n = 3). Significant differences among treatments are denoted by lower-case letters.

POP, and therefore C:P (~140) and N:P (~20) decreased regardless of light treatments (<u>Fig 3G</u> and 3I; Table B in <u>S1 File</u>). The C:N ratio (mean values 7.40) did not change among treatments.

# UVR and P effect on photosynthetic activity, pigments, and primary production

Under non-enriched conditions, UVR did not affect  $ETR_{max}$  or  $F_v/F_m$  (data not shown). However,  $ETR_{max}$  and photosynthetic efficiency ( $\alpha_{ETR}$ ) were higher in the UVR and P-enriched treatment (Fig 4A, 4B and 4C). Hence, the UVR×P interaction was synergistically positive on  $ETR_{max}$ and ( $\alpha_{ETR}$ ) (Table C in <u>S1 File</u>). In general, chlorophyll and carotenoid pigments (fucoxanthin) did not respond significantly to any treatment assayed (data not shown), and only the D<sub>t</sub>: (D<sub>d</sub>+D<sub>t</sub>) ratio decreased under UVR at ambient nutrient conditions (Fig 4D; Table C in <u>S1 File</u>).





**Fig 3. Elemental content of the sestonic fraction under experimental conditions.** (A) C cell-quota (pgC cell<sup>-1</sup>); (B) N cell-quota (pgN cell<sup>-1</sup>); (C) P cellquota (pgP cell<sup>-1</sup>); (D)particulate organic carbon (POC;  $\mu$ gC L<sup>-1</sup>); (E) Particulate organic nitrogen (PON;  $\mu$ gN L<sup>-1</sup>); (F) Particulate organic phosphorus (POP;  $\mu$ gP L<sup>-1</sup>); (G) C:P; (H) C:N and (I) N:P ratios under photosynthetically active radiation (PAR) and full sunlight (PAR+UVR) in nutrient-enriched (P-enriched) and non-enriched (control) treatments. Horizontal dashed lines indicate value for initial day. Data are expressed as mean values ± sd (n = 3). Significant differences among treatments are denoted by lower-case letters.

 $PP_M$  represented more than 88% of the total PP. The effect of UVR on PP varied according to the phytoplankton size fraction being considered (Fig 5A and 5B). Thus, under nonenriched conditions  $PP_M$  was not affected by UVR but PPP decreased (up to 68%; Fig 5B). Penrichment had a stimulatory effect on both size fractions, unmasking a negative UVR effect on  $PP_M$  and productivity, but suppressing the inhibitory UVR effect on  $PP_P$  (antagonistic effect). EOC rates (data not shown) and %EOC had significantly the lowest values in samples under UVR under non-enriched conditions whereas P-enrichment increased both variables, generating an antagonistic UVR×P effect (i.e. joint UVR and P eliminated the decreasing UVR effect on the EOC rate (data not shown) and %EOC (Fig 5C).

# UVR and P effect on abundance/biomass and taxonomic composition of the phytoplankton

Total phytoplankton abundance and biomass, in non-enriched treatments, showed no significant differences between light treatments (Fig 6A; Table 2). Total phytoplankton abundance increased in P-enriched treatments, reaching 20.1 and  $36.6 \times 10^3$  cell mL<sup>-1</sup>, for



Fig 4. Response of photosynthetic rate and xanthophyll cycle pigments to experimental conditions. (A) The electron-transport rate (ETR) vs. irradiance; (B) ETR<sub>max</sub>; (C) Initial slopes ( $\alpha_{ETR}$ ) of curves ETR; (D) Dd de-epoxidation state of the xanthophyll cycle pigments (D<sub>t</sub>:(D<sub>t</sub>+D<sub>d</sub>) (diadinoxanthin, D<sub>d</sub>, and diatoxanthin, D<sub>t</sub>) under photosynthetically active radiation (PAR) and full sunlight (PAR+UVR) in nutrient enriched (P-enriched) and non-enriched (control) treatments. Data are expressed as mean values ± sd (n = 3). Significant differences among treatments are denoted by lower-case letters.

UVR+PAR and PAR treatments, respectively (Fig 6A). By contrast, total biomass, significantly decreased under UVR+PAR treatments compared to PAR after P-enrichment. Consequently, a negative synergistic UVR×P effect was exerted on total abundance and biomass, since the P-enrichment unmasked the inhibitory UVR effect on these variables (Table C in <u>S1 File</u>). The picoplanktonic community played the main role in the magnitude of this response of total phytoplanktonic abundance to UVR×P (Fig 6B). Among microphytoplankton groups, the flagellates underwent a negative synergetic effect UVR×P both in abundance (Fig 6C) and in biomass (Table 2).

Taxonomic composition of the phytoplankton communities at the end of the incubation time are shown in <u>Table 2</u>. In the non-enriched treatment, microphytoplankton biomass groups increased whereas picoeukaryotes and *Synechococcus* sp. biomass declined regardless of the light treatments when compared with initial conditions. P-enrichment did not promote an increase in microphytoplankton biomass, and flagellate biomass significantly decreased in UVR and P-enriched treatment (UVR×P p-value < 0.01) (<u>Table 2</u>). However, the response of picoplankton to P-enrichment was the opposite. Picoeukaryotes and *Synechococcus* sp. increased in both light treatments after P-enrichment. UVR×P exerted a significantly negative synergistic effect (UVR×P p-value < 0.01) on both groups since P enrichment triggered the inhibitory UVR effect.





**Fig 5.** Primary production in different size fractions, excreted organic carbon (%) and productivity under experimental conditions. (A) Microplanktonic primary production ( $PP_{M}$ , in µgC L<sup>-1</sup> h<sup>-1</sup>); (B) Picoplanktonic primary production ( $PP_{P}$ , in µgC L<sup>-1</sup> h<sup>-1</sup>); (C) Excreted organic carbon (% EOC); (D) Productivity (µgC µgChl a<sup>-1</sup>) under photosynthetically active radiation (PAR) and full sunlight (PAR+ UVR) in nutrient-enriched (P-enriched) and nonenriched (control) treatments. Horizontal dashed lines indicate value for the initial day. Data are expressed as mean values ± sd (n = 3). Significant differences among treatments are denoted by lower-case letters.

doi:10.1371/journal.pone.0142987.g005



**Fig 6. Phytoplanktonic abundance under experimental conditions.** (A) Total phytoplankton abundance (AT) (cells  $mL^{-1}$ ); (B) Picoplanktonic abundance ( $A_P$ ) in cells  $mL^{-1}$ ); (C) Microplanktonic abundance ( $A_M$  in cells  $mL^{-1}$ ) under photosynthetically active radiation (PAR) and full sunlight (PAR+UVR) in nutrient-enriched (P-enriched) and non-enriched (control) treatments. Data are expressed as mean values ± sd (n = 3). Significant differences among treatments are denoted by lower-case (total abundance and *Synechococcus* sp. abundance) and capital letters (picoeukaryotes abundance).

doi:10.1371/journal.pone.0142987.g006

### Discussion

In this study, we evaluate, for the first time, the phytoplanktonic responses to the interactive effects of UVR and P enrichment at different biological organization levels (from the physiological to the community level), in an oligotrophic P-limited Mediterranean area. Furthermore, in our experimental approach, we simulated the P enrichment provided by current Saharan dust loads in the western Mediterranean Sea [47, 48], which constitute one of the consequences of global change at the regional scale [3]. Under these conditions, we measured the phytoplank-tonic responses to P enrichment over a realistic time scale similar to previously reported phytoplankton blooms in this area of the Mediterranean Sea [71, 72]. This approach provides a framework for unravelling the mechanisms that enable algae to tolerate UVR stress. (attenuation) and at the community level (unmasking) to UVR effect after P-enrichment.

#### Tolerance to UVR stress

The C expenditure by respiration, which has been proposed as an indicator of physiological stress, results from catabolic pathways developed to maintain the cell functionality of primary producers under UVR [7]. Our results evidence that UVR increased the phytoplanktonic respiration rate under nutrient ambient conditions. The respiratory electron-transfer chain generates oxygen-free radicals, which increase under photoinhibitory conditions, resulting in the accumulation of ROS [73, 15, 16], an indicator of oxidative stress. Although enzymatic antioxidant activities were not measured in our experiment, the absence of violaxanthin or the decrease in the  $D_t:(D_t+D_d)$  ratio, pigments related to the xanthophyll cycle, involved in the thermal dissipation of excess light [74], could partly support the idea of the existence of metabolic stress. Nevertheless, depressed Dd de-epoxidation might also imply an active photoprotective response to UVR exposure via enhanced diadinoxanthin synthesis [75], helping to alleviate the UVR stress, as found in species of bacillariophyceae, haptophyceae or dinophyceae over the time scale of our experiment. In addition, the lack of negative UVR effect on Chl pigments, PP<sub>M</sub>, microphytoplanktonic or picoeukaryotic abundance and biomass could be determined by the start-up of repair mechanisms. Cells reportedly boost the RNA (a P-rich biomolecule) content under UVR to activate the expression of genes related to repair proteins or to provide metabolites needed for cell repair (e.g. ATP) [76, 77, 16]. The start-up of repair mechanisms could be guaranteed by the increase in sestonic P found here (further discussion below). Consequently, the increase in respiration rates could reflect energy costs related to the repair of cellular components damaged by UVR [78, 79]. These above-described mechanisms could operate together, helping to explain that increased oxidative stress did not transfer as damage to the microphytoplankton variables studied. Nevertheless, the metabolic stress generated (increased ROS) was transferred to picoplanktonic C-incorporation and Synechococcus sp. abundance and biomass. This result agrees with previous findings showing that eukaryotic phytoplankton have a higher photoacclimation potential than do picoprokaryotic species when both undergo identical experimental light conditions in oligotrophic marine areas [80, 81].

Another less specific mechanism that underlies the eukaryotic phytoplankton acclimation may be the improvement in the cell-quota nutrient shown by the sestonic fraction under UVR, and the decrease in their stoichiometric ratios. The lower C:P ratio under UVR could be attributed either to C-losses (greater respiration or C excretion), or to increased sestonic P content. Our findings imply that the C:P (and N:P) ratios declined under UVR due mainly to greater sestonic P content, since the sestonic C content increased under UVR.

Therefore, the next key question is to ascertain which processes can be related to the increase in the sestonic P content. On the one hand, our results showed a decrease in DOC concentration under UVR in the non-enriched treatment. The lack of increase in bacterial

production in our experiment [82] precludes considering the bacterial consumption of DOC as a major mechanism responsible for the observed DOC decrease. Such a drop in the DOC concentration might indicate the UVR-induced photolysis of organic matter, which could actually increase P availability for phytoplankton [83]. On the other hand, our results show that the greater sestonic P per cell was related to surge in enzymatic hydrolysis due to APA, because of the significant increase of APA<sub>EX</sub> under UVR. Similar UVR effects increasing dissolved APA activity values have been described in freshwater ecosystems [30, 31], where the APA values for different fractions were in the same range as in the present study. Thus, cells exposed to UVR may stimulate P uptake, boosting the sestonic P content. Our results suggest that microplankton phytoplankton is acclimated to UVR, due to induced extracellular APA by UVR. We propose that this mechanism could guarantee P acquisition in P-limited marine ecosystems, where the autotrophic fractions could have a competitive advantage over bacteria to take up inorganic P [84].

# Interactive UVR x P effect

The growth of phytoplankton was stimulated by P addition. Thus, UVR-acclimated phytoplankton (see above) that can use UVA radiation through carotenoids (accessory pigments) as an energy source for photosynthesis [21] might be expected to enhance their growth after a P pulse, and even a positive synergistic interactive effect. This response was found only in ETR<sub>max</sub> and photosynthetic efficiency ( $\alpha_{ETR}$ ). By contrast, the UVR×P interactive effect was antagonistic on picoplanktonic C-incorporation (PP<sub>P</sub>) and negatively synergetic on other variables such as PP<sub>M</sub>, productivity, and abundance of both size fractions. Therefore, an inhibitory effect of UVR was unmasked after P enrichment.

The finding that UVR×P on picoplanktonic production was antagonistic could be the result, described for these organisms, of an alternative electron flow to  $O_2$ , which extracts electrons from the intersystem electron-transport chain, prior to photosystem I [85]. This pathway alleviates excessive photosystem II excitation pressure that could occur after P enrichment.

The negative synergistic UVR×P effect reported for most of the variables suggests that colimitation is exerted by high UVR and low nutrient levels on the C metabolism and growth of phytoplankton, because only the removal of both stress factors led to the highest stimulation [35, 20]. The harmful UVR effect after P-enrichment has been widely reported in oligotrophic freshwater ecosystems on phytoplankton abundance [35, 86, 87, 20] and for C fixation (PP) and productivity [20,31]. There are several mechanisms that could alter the coupling of the PSII function, estimated using chlorophyll fluorescence, with photosynthetic carbon fixation, such as photorespiration [88], chlororespiration via a plastid terminal oxidase (PTOX) [85, 89] and the Mehler reaction [90]. The lack of increased ROS precludes considering a direct photoreduction of  $O_2$  by thylakoids, known as the Mehler reaction [91, 92], as a mechanism to explain the decoupling between electron transfer and carbon fixation under UVR and Penriched conditions. Based on our findings of non-increase in R under UVR after P enrichment, we can rule out a higher carbon loss by respiration, as has been recently proposed [93, 94]. Unfortunately, we did not measure photorespiration, a key process which could account for the mismatch between ETR<sub>max</sub> and C incorporation into biomass, because the photorespiration can imply a loss of up to 25% of the C fixed in photosynthetic processes [95]. Nevertheless, we may speculate that the excreted C (not increased under UVR after P-enrichment) failed to eliminate phosophoglycolate (a by-product of photorespiration), which inflicts damage similar to that of ROS [96, 97]. Furthermore, this unmasking effect on phytoplanktonic abundance could be the result of growth stimulated by enrichment with limiting nutrients, thereby inducing higher rates of DNA synthesis. This may result in a greater propensity to

UVR-induced DNA damage, making the effects of UVR on cell division variables more evident after P enrichment [98].

On the other hand, our results demonstrate that picoprokaryotes in the near-surface layers of the Alborán Sea may be severely affected by exposure to ambient levels of UVR, limiting their growth despite their competitive advantage under P-enriched conditions. The *Synechococcus* sp. populations underwent higher inhibitory UVR effects than did *Prochlorococcus* sp.. These results are contrary to those of Sommaruga et al. [99] and LLabré et al. [18] for the north-western Mediterranean Sea (NWS). It is remarkable that the abundance of picoprokaryotes was lower (by one order of magnitude) in Alborán Sea than in NWS [99,100]. The higher sensitivity of *Synechococcus* sp. even after a P pulse suggests that the levels of photoprotection and repair systems could be insufficient to repair the cell damage induced by solar radiation.

## **Concluding Remarks**

Based on our findings, we propose that (1) the main mechanism of autotrophic eukaryote tolerance to UVR is the increase in P content mediated by (i) a direct stimulatory effect of UVR on dissolved fraction of extracellular APA reinforced by greater DOC-photolysis, (ii) an increase in P uptake under UVR and an improvement of the P cell-quota, allowing the repair of cell components damaged by sublethal levels of UVR; (2) the mechanism involved in the unmasking effect of UVR after P enrichment may be the photodamage caused by excessive electron flux with the activation of photosynthetic electron transport, in the absence of an efficient C-release mechanism(by eliminating the phosophoglycolate) to dissipate the reducing power of photosynthetic electron transport. This photodamage would explain the mismatch between the synergistic positive UVR×P effect on photosynthetic variables and the synergistic negative UVR×P effect on C incorporation and productivity that constrains the growth of the phytoplankton community.

## **Supporting Information**

S1 File. Results of the two-way ANOVA of the interaction of UVR and P-enrichment. Interactive effect on respiration, reactive oxygen species, total and dissolved alkaline phosphatase and dissolved organic carbon (Table A). Interactive effect on C, P and N cell-quota, particulate organic carbon, particulate organic nitrogen, particulate organic phosphorus, C:P, C:N and N: P ratios (Table B). Interactive effect on maximal electron transport rate, photosynthetic efficiency, Dd de-epoxidation; microphytoplanktonic and, picophytoplanktonic carbon incorporation, percentage excreted organic carbon, productivity and total abundance (Table C). F values and significance levels (p) are shown, numbers in bold indicate, p-value < 0.05. R: respiration; ROS: reactive oxygen species; APA<sub>T</sub>:t otal alkaline phosphatase activity; APA<sub>EX</sub>: dissolved alkaline phosphatase activity; DOC: dissolved organic carbon. POC: particulate organic carbon; PON: particulate organic nitrogen; POP: particulate organic phosphorus ETR<sub>max</sub>: maximal electron transport rate;  $\alpha$ ETR: photosynthetic efficiency; Dt:Dt+Dd: Dd de-epoxidation; PP<sub>M</sub>: microphytoplanktonic carbon incorporation; PP<sub>P</sub>: picophytoplanktonic carbon incorporation;%EOC: Percentage excreted organic carbon.

## Acknowledgments

This study was supported by Ministerio Ciencia e Innovación (CGL2011-23681/BOS), Consejería de Innovación, Ciencia y Empresa—Junta de Andalucía (Excelencia CVI-02598). GH and CD were supported by the Spanish Government -Formación de Profesorado Universitario Grant. We thank the crew of the B/O F. de P.

#### **Author Contributions**

Conceived and designed the experiments: PC JMMS. Performed the experiments: PC JMMS GH CD MS NK FLF JMM. Analyzed the data: PC JMMS GH CD MS SS DC NK FLF JMM. Contributed reagents/materials/analysis tools: PC JMM. Wrote the paper: PC JMMS.

#### References

- 1. Longhurst AR, Sathyendranath S, Platt T. Caverhill CM. An estimate of global primary production in the ocean from satellite data. J Plankton Res. 1995; 17:1245–1271.
- 2. Cloern JE, Foster SQ, Kleckner AE. Phytoplankton primary production in the world's estuarine-coastal ecosystems. Biogeosciences 2014; 11: 2477–2501, doi: 10.5194/bg-11-2477-2014
- Stocker TFD, Qin G-K, Plattner M, Tignor SK, Allen J, Boschung A, et al. IPCC,2013: Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change, ed. Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 2013.
- Philippart CJM, Anadón R, Danovaro R, Dippnerd JW, Drinkwater KF, Hawkins SJ, et al. Impacts of climate change on European marine ecosystems: observations, expectations and indicators. J. Exp Mar Biol Ecol. 2011; 400:52–69.
- Cloern JE, Jassby AD. Drivers of change in estuarine-coastal ecosystems: discoveries from four decades of study in San Francisco Bay. Rev Geophys. 2012; 50: RG4001, doi: <u>10.1029/2012RG000397</u>
- Beardall J, Sobrino C, Stojkovic S. Interactions between the impacts of ultraviolet radiation, elevated CO2 and nutrient limitation on marine primary producers. Photochem Photobiol Sci. 2009; 8: 1257– 1265. doi: 10.1039/b9pp00034h PMID: 19707614
- 7. Gao K, Helbling EW, Häder DP, Hutchins DA. Responses of marine primary producers to interactions between ocean acidification, solar radiation, and warming. Mar Ecol Prog Ser. 2012; 470: 167–189.
- Häder DP, Williamson CE, Wängberg SA, Rautio M, Rose KC, Gao K, et al. Effects of UV radiation on aquatic ecosystems and interactions with other environmental factors. Photochem Photobiol Sci. 2014; doi: <u>10.1039/c4pp90035a</u>
- Shanklin J. Reflections on the ozone hole. Nature. 2010; 465: 34–35. doi: <u>10.1038/465034a</u> PMID: <u>20445611</u>
- Sola Y, Lorente J. Impact of two low ozone events on surface solar UV radiation over the northeast of Spain. Int J Climatol. 2011; 31:1724–1734.
- Martínez-Lozano JA, Utrillas MP, Núñez JA, Tamayo J, Marín MJ, Esteve AR, et al. Ozone mini-holes over Valencia (Spain) and their influence on the UV erythemal radiation. Int J Climatol. 2011; 31: 1554–1566.
- McKenzie RL, Aucamp PJ, Bais AF, Björn LO, Ilyas M, Madronich S. Ozone depletion and climate change: impacts on UV radiation. Photochem Photobiol.2011; 10, 182–198.
- Häder DP, Helbling EW, Williamson CE, Worrest RC. Effects of UV radiation on aquatic ecosystems and interactions with climate change: 2010 assessment.Photochem Photobiol Sci. 2011; 10(2):242– 260. doi: <u>10.1039/c0pp90036b</u> PMID: <u>21253662</u>
- Leu E, Falk-Petersen S, Hessen DO. Ultraviolet radiation negatively affects growth but not food quality of arctic diatoms. Limnol Oceanogr. 2007; 52:787–797.
- Bouchard JN, García-Gómez C, Lorenzo MR, Segovia M. Differential effect of ultraviolet exposure (UVR) in the stress response of the Dinophycea Gymnodinium sp. and the Chlorophyta Dunaliella tertiolecta: mortality versus survival. Mar Biol. 2013; 160:2547–2560.
- García-Gómez C, Gordillo FJL, Palma A, Lorenzo MR, Segovia M. Elevated CO2 alleviates high PAR and UV stress in the unicellular chlorophyte Dunaliella tertiolecta. Photochem Photobiol Sci. 2014; 13:1347–1358. doi: <u>10.1039/c4pp00044g</u> PMID: <u>25043601</u>
- Villafañe V, Sundback K, Figueroa F, Helbling E. Photosynthesis in the aquatic environment as affected by UVR. In: Helbling EW, Zagarese H, editors. UVeffects in aquatic organisms and ecosystems. Royal Society Chemistry, Cambridge; 2003. pp. 357–397.
- Llabrés M, Agustí S, Alonso-Laita P, Herndl GJ. Synechococcus and Prochlorococcus cell death induced by UV radiation and the penetration of lethal UVR in the Mediterranean Sea. Mar Ecol Prog Ser. 2010; 399: 27–37, doi: <u>10.3354/meps08332</u>

- Agustí S. Regaudie-de-Gioux A, Arrieta JM, Duarte CM. Consequence of UV-enhanced community respiration for plankton metabolic balance Limnol. Oceanogr. 2014; 59(1): 223–232.
- Carrillo P, Delgado-Molina JA, Medina-Sánchez JM, Bullejos FJ, Villar-Argaiz M. Phosphorus inputs unmask negative effects of ultraviolet radiation on algae in a high mountain lake, Glob Chang Biol. 2008; 14: 423–439.
- Gao K, Wu Y, Li G, Wu H, Villafañe VE, Helbling EW. Solar UV radiation drives CO2 fixation in marine phytoplankton: a double-edged sword. Plant Physiol. 2007; 144:54–59. PMID: <u>17494919</u>
- Helbling EW, Villafañe VE, Ferrario M, Holm-Hansen O. Impact of natural ultraviolet radiation on rates of photosynthesis and on specific marine phytoplankton species, Mar Ecol Progr Ser. 1992; 80: 89– 100.
- Zudaire L, Roy S. Photoprotection and long-term acclimation to UV radiation in the marine diatom Thalassiosira weissflogii. J. Photochem Photobiol B. 2001; 62:26–34. PMID: <u>11693364</u>
- Korbee N, Huovinen P, Figueroa FL, Aguilera J, Karsten U. Availability of ammonium influences the photosynthesis and the accumulation of MAAs in two Porphyra species (Bangiales, Rhodophyta) from different latitudes. Mar Biol. 2005; 146: 645–654.
- **25.** Villafañe VE, Erzinger GS, Strauch SM, Helbling EW. Photochemical activity of PSII of tropical phytoplankton communities of Southern Brazil exposed to solar radiation and nutrient addition. J Exp Mar Biol Ecol. 2014; 459: 199–207, doi: 10.1016/j.jembe.2014.05.027
- Hessen DO, Frigstad H, Færøvig PJ, Wojewodzic MW, Leu E. UV radiation and its effects on Puptake in arctic diatoms J Exp Mar Biol Ecol. 2012; 411:45–51.
- Jansson M, Olsson H, Petterson K. Phosphatase: origin, characteristics and function in lakes. Hydrobiologia. 1988; 170: 157–175.
- Hudson JJ, Taylor WD, Schindler DW. Phosphate concentrations in lakes. Nature. 2000; 406:54–56. PMID: <u>10894537</u>
- Tank SE, Xenopoulos MA, Hendzel LL. Effect of ultraviolet radiation on alkaline phosphatase activity and planktonic phosphorus acquisition in Canadian boreal shield lakes, Limnol Oceanogr. 2005; 50: 1345–1352.
- **30.** Sereda JM, Vandergucht DM, Hudson JJ. Disruption of planktonic phosphorus cycling by ultraviolet radiation, Hydrobiologia. 2011; 665: 205–217.
- Korbee N, Carrillo P, Mata M, Rosillo S, Medina-Sánchez JM, Figueroa LF. Effects of ultraviolet radiation and nutrients on the structure-function of phytoplankton in a high mountain lake, Photochem Photobiol Sci. 2012; 11(6):1087–98. doi: 10.1039/C2PP05336E PMID: 22544332
- **32.** Medina-Sánchez JM, Villar-Argaiz M, Carrillo P. Solar radiation–nutrient interaction enhances the resource and predation algal control on bacterioplankton: a short-term experimental study, Limnol Oceanogr. 2006; 51: 913–924.
- Marcoval MA, Villafañe VE, Helbling EW. Combined effects of solar ultraviolet radiation and nutrients addition on growth, biomass and taxonomic composition of coastal phytoplankton communities of Patagonia. J Photochem Photobiol B. 2008; 91:157–166. doi: <u>10.1016/j.jphotobiol.2008.03.002</u> PMID: <u>18456506</u>
- 34. Veen A, Reuvers M, Roncak P. Effects of acute and chronic UVB exposure on a green alga: a continuous culture study using a computer-control dynamic light regime. Plant Ecol. 1997; 128: 29–40.
- Xenopoulos MA, Frost PC, Elser JJ. Joint effects of UV radiation and phosphorus supply on Igal growth rate and elemental composition. Ecology.2002; 83: 423–435.
- **36.** Hessen DO, Leu E, Færøvig PJ, Petersen SF. Light and spectral properties as determinants of C:N: P-ratios in phytoplankton. Deep-Sea Res PT II. 2008; 55: 2169–2175.
- Korbee N, Abdala R, Figueroa FL, Helbling WE. Ammonium and UVR stimulate the accumulation of mycosporine-like amino acids in <u>Porphyra columbina</u> (Rhodophyta) from Patagonia, Argentina. J Phycol. 2004; 40:248–259.
- Villafañe VE, Marcoval MA. Helbling EW. Photosynthesis vs. irradiance characteristics in phytoplankton assemblages off Patagonia (Argentina): Temporal variability and solar UVR effects. Mar Ecol Progr Ser. 2004, 284: 23–34.
- Krom MD, Kress N, Brenner S. Phosphorous limitation of primary productivity in the eastern Mediterranean Sea. Limnol Oceanogr. 1991; 36:424–432.
- Krom MD, Herut B, Mantoura RFC. Nutrient budget for the Eastern Mediterranean: Implication for phosphorous limitation. Limnol Oceanogr. 2004; 49:1582–1592.
- Siokou-Frangou I, Christaki U, Mazzocchi MG, Montresor M, Ribera d'Alcala M, Vaqué D, et al. Plankton in the open Mediterranean Sea: a review. Biogeosciences. 2010; 7:1543–1586.

- 42. Tanaka T, Thingstad TF, Christaki U, Colombet J, Cornet-Barthaux V, Courties C, et al. Lack of P-limitation of phytoplankton and heterotrophic prokaryotes in surface waters of three anticyclonic eddies in the stratified Mediterranean Sea, Biogeosciences. 2011; 8: 525–538.
- **43.** Pinhassi J, Gómez-Consarnau L, Alonso-Sáez L, Sala MM, Vidal M, Pedrós-Alió C, et al. Seasonal changes in bacterioplankton nutrient limitation and their effects on bacterial community composition in the NW Mediterranean Sea. Aquat Microb Ecol. 2006; 44: 241–252.
- Tanaka T, Thingstad TF, Gasol JM, Cardelus C, Jezbera J, Sala MM, et al. Determining the availability of phosphate and glucose for bacteria in P-limited mesocosms of NW Mediterranean surface waters. Aquat Microb Ecol. 2009; 56(1):81–91.
- **45.** Tanaka T, Rassoulzadegan F. Orthophosphate uptake by heterotrophic bacteria, cyanobacteria, and autotrophic nanoflagellates in Villefranche Bay, northwestern Mediterranean: Vertical, seasonal, and short-term variations of the competitive relationship for phosphorus. Limnol Oceanogr. 2004; 49(4): 1063–1072.
- **46.** Llebot C, Spitz Yvette H, Solé J, Estrada M. The role of inorganic nutrients and dissolved organic phosphorus in the phytoplankton dynamics of a Mediterranean bay: A modeling study J. Mar. Syst. 2010; 83: 192–209.
- 47. Querol X, Alastuey A, Pey J, Cusack M, Perez N, Mihalopoulos N, et al. Variability in regional background aerosols within the Mediterranean. Atmos Chem Phys. 2009; 9:4575–4591.
- Gallisai R, Peters F, Volpe G, Basart S, Baldasano JM. Saharan dust deposition may affect phytoplankton growth in the Mediterranean Sea at ecological time scales. PLoS ONE.2014; 9(10): e110762. doi: 10.1371/journal.pone.0110762 PMID: 25333783
- Mercado JM, Cortés D, Garcia A, Ramírez T. Seasonal and inter-annual changes in the planktonic communities of the northwest Alborán Sea (Mediterranean Sea). Progr Oceanogr. 2007; doi: <u>10.</u> <u>1016/j.pocean.2007.04.013</u>
- Mercado JM, Ramírez T, Cortés D. Changes in nutrient concentration induced by hydrological variability and its effect on light absorption by phytoplankton in the Alborán Sea (Western Mediterranean Sea). J Mar Syst. 2008; 71:31–45. doi: 10.1016/j.jmarsys.2007.05.009
- Morales-Baquero R, Pulido-Villena E. Reche I. Atmospheric inputs of phosphorus and nitrogen to the southwest Mediterranean region: biogeochemical response of high mountain lakes. Limnol Oceanogr. 2006; 51:830–837.
- Ramírez T, Cortés D, Mercado JM, Vargas-Yañez M, Sebastián M, Liger E. Seasonal dynamics of inorganic nutrients and phytoplankton biomass in the NWAlboran Sea. Estuar Coast Shelf Sci. 2005; 65: 654–670.
- 53. Grasshoff K, Kremling K, Ehrhardt M. Methods of Seawater Analysis, 3rd ed. Wiley-VCH, Weinheim, Germany; 1999.
- Petterson K, Janson M. Determination of phosphatase activity in lake water- a study methods, Verh InternatVerein Limnol. 1978; 20: 226–1330.
- Segovia M, Berges JA. Inhibition of caspase-Like activities prevents the appearance of reactive oxygen species and dark-induced apoptosis in the unicellular Chlorophyte Dunaliella tertiolecta, J. Phycol. 2009; 45: 1116–1126.
- 56. Korbee N, Mata MT, Figueroa FL. Photoprotection mechanisms against ultraviolet radiation in Heterocapsa sp. (Dinophyceae) are influenced by nitrogen availability: Mycosporine-like amino acids vs. xanthophyll cycle, Limnol Oceanogr. 2010; 55: 899–908.
- 57. Johnsen G, Sakshaug E. 2007. Bio-optical characteristics of PSII and PSI in 33 species (13 pigment groups) of marine phytoplankton, and the relevance for pulse-amplitude-modulated and fast-repetition-rate fluorometry. J. Phycol. 43: 1236–1251, doi: 10.1111/j.1529-8817.2007.00422.x
- Eilers PHC, Peeters JCH. A model for the relationship between light intensity and the rate of photosynthesis in phytoplankton. Ecol Modell. 1988; 42: 199–215.
- Lubián LM, Montero O. Excess light-induced violaxanthin cycle activity in Nannochloropsis gaditana (Eustigmatophyceae): Effects of exposure time and temperature. Phycologia. 1998; 37: 16–23.
- Steeman-Nielsen E. The use of radioactive carbon (14C) for measuring organic production in the sea. J Cons Int. Explor Mer. 1952; 18, 117–140.
- Lignell R. Problems in filtration fractionation of 14C primary productivity samples, Limnol Oceanogr. 1992; 37, 172–178.
- Mercado JM, Ramírez T, Cortés D, Sebastián M, Reul A, Aguilera J, et al. Diurnal changes in the biooptical properties of the phytoplankton in the Alborán Sea (Mediterranean Sea). Estuar Coast Shelf Sci. 2006; 69:459–470. doi: <u>10.1016/j.ecss.2006.05.019</u>

- 63. Ribes M, Coma R, Gili JM. Seasonal variation of particulate organic carbon, dissolved organic carbon and the contribution of microbial communities to the live particulate organic carbon in a shallow nearbottom ecosystem in the Northwestern Mediterranean Sea. J Plankton Res. 1999; 21:1077–1100.
- **64.** Morel A, Ahn Y-H, Partensky F, Vaulot D, Claustre H. Prochlorococcus and Synechococcus: a comparative study of their optical properties in relation to their size and pigmentation. J Mar Res. 1993; 51:617–649.
- Kana TM, Glibert PM. Effect of irradiance of up to 2000 mEm-2s-1 on marine Synechococcus WH7803. 1. Growth, pigmentation and cell composition. Deep-Sea Res. 1987; 34:479–495.
- **66.** Sun J, Liu D. Geometric models for calculating cell biovolume and surface area for phytoplankton. J Plankton Res. 2003; 25:1331–1346.
- 67. Olenina I, Hajdu S, Edler L, Andersson A, Wasmund N, Busch S, et al. Biovolumes and size-classes of phytoplankton in the Baltic Sea. HELCOM Balt Sea Environ Proc. 2006. No 106, 144 pp.
- **68.** Vadrucci MR, Cabrini M, Basset A. Biovolume determination of phytoplankton guilds in transitional water ecosystems of Mediterranean Ecoregion. TWB Transit. Waters Bull. 2007; 2:83–102.
- Verity PG, Robertson CY, Tronzo CR, Andrews MG, Nelson JR, Sieracki ME. Relationships between cell volume and the carbon and nitrogen content of marine photosynthetic nanoplankton. Limnol Oceanogr. 1992; 37, 1434–1446.
- **70.** Menden-Deuer S, Lessard EJ. Carbon to volume relationships for dinoflagellates, diatoms and other protest plankton. Limnol Oceanogr. 2000; 45:569–579.
- 71. Mercado JM, Sobrino C, Neale PJ, Segovia M, Reul A, Amorim AL, et al. Effect of CO2, nutrients and light on coastal plankton. II. Metabolic rates. Aquatic Biol. 2014; 22: 43–57, doi: 10.3354/ab00606
- 72. Mercado JM, Ramírez T, Sebastián M, Vargas M. Seasonal and inter-annual variability of the phytoplankton communities in an upwelling area of the Alborán Sea (SW Mediterranean Sea). Scientia Marina. 2005; 69(4):451–465.
- **73.** Rijstenbil JW. Assessment of oxidative stress in the planktonic diatom Thalassiosira pseudonana in response to UVA and UVB radiation, J. Plankton Res. 2002; 24: 1277–1288.
- Demmig-Adams B, Adams WW III. Photoprotection in an ecological context: the remarkable complexity of thermal energy dissipation. New Phytologist. 2006; 172:11–21. PMID: <u>16945085</u>
- 75. van de Poll WH, Buma AGJ. Does ultraviolet radiation affect the xanthophyll cycle in marine phytoplankton? Photochem Photobiol Sci. 2009; 8: 1295–1301. doi: <u>10.1039/b904501e</u> PMID: <u>19707617</u>
- Heraud P, Roberts S, Shelly K, Beardall J. Interactions between UV-B exposure and phosphorus nutrition. II. Effects on rates of damage and repair. J Phycol. 2005; 41: 1212–1218.
- Jahid IK, Silva AJ, Benitez JA. Polyphosphate stores enhance the ability of Vibrio cholerae to overcome environmental stresses in a low-phosphate environment. Appl. Environ Microb. 2006; 72: 7043–7049.
- 78. Raven JA, Beardall J. Respiration and photorespiration. Can Bull Fish Aquat Sci. 1981; 210: 55–82.
- Xu Z, Gao K. Impacts of UV radiation on growth and photosynthetic carbon acquisition in Gracilaria lemaneiformis (Rhodophyta) under phosphorus-limited and replete conditions. Funct Plant Biol. 2009; 36: 1057–1064.
- Kulk G, de Poll WHv, Visser RJW, Buma AGJ. Distinct differences in photoacclimation potential between prokaryotic and eukaryotic oceanic phytoplankton, J. Exp. Mar. Biol Ecol. 2011; 398: 63–72, doi: 10.1016/j.jembe.2010.12.011
- Neale PJ, Pritchard AL, Ihnacik R. UV effects on the primary productivity of picophytoplankton: biological weighting functions and exposure response curves of Synechococcus. Biogeosciences, 2014; 11: 2883–2895, doi: 10.5194/bg-11-2883-2014
- 82. Durán C. Multiple stressors acting on the algal-bacterial interaction in Mediterranean oligotrophic system. PhD. The University of Granada 2014. Available: <u>http://ecologia.ugr.es/pages/banners/2014/tesis-cristina-duran</u>.
- **83.** Findlay S, Sinsabaugh RL. Aquatic ecosystems: Interactivity of dissolved organic matter. Academic Press; 2003.
- 84. Carrillo P, Medina-Sánchez JM, Villar-Argaiz M. Does microorganism stoichiometry predict microbial food web interactions after a phosphorus pulse? Microb Ecol. 56(2):350–363. doi: <u>10.1007/s00248-007-9353-8</u> PMID: <u>18165873</u>
- Bailey S, Mehlis A, Mackey KRM, Cardol P, Finazi G, van Dijken G, et al. Alternative photosynthetic electron flow to oxygen in marine Synechococcus. Biochim Biophys Acta. 2008; 1777: 269–276. doi: 10.1016/j.bbabio.2008.01.002 PMID: 18241667

- Doyle SA, Saros JE. Williamson C. Interactive effects of temperature and nutrient limitation on the response of alpine phytoplankton growth to ultraviolet radiation. Limnol Oceanogr. 2005; 50:1362– 1367.
- Xenopoulos MA, Frost PC. UV radiation, phosphorus, and their combined effects on the taxonomic composition of phytoplankton in a boreal lake. J Phycol. 2003; 39: 291–302.
- Badger MR, von Caemmerer S, Ruuska S, Nakano H. Electron flow to oxygen in higher plants and algae: rates and control of direct photoreduction (Mehler reaction) and rubisco oxygenase. Phil Trans R Soc Lond D. 2000; 355: 1433–1446.
- 89. Mackey KRM, Paytan A, Grossman AR, Bailey S. A photosynthetic strategy for coping in a high-light, low-nutrient environment. Limnol Oceanogr. 2008; 53:900–913.
- Lawrenz E, Silsbe G, Capuzzo E, Ylöstalo P, Forster RM, Simis SGH, et al. Predicting the Electron Requirement for Carbon Fixation in Seas and Oceans. PLoSONE. 2013; 8(3): e58137. doi: <u>10.1371/</u> journal.pone.0058137
- Asada K. The water to water cycle in chloroplasts: scavenging of active oxygens and dissipation of excess photons. A. Rev. Plant Physiol. Plant Mol Biol. 1999; 50: 601–639.
- 92. Badger MR. 1985. Photosynthetic oxygen-exchange. A Rev Plant Physiol. 36, 27–53.
- 93. Gao K, Xu J, Gao G, Li Y, Hutchins DA, Huang B, et al. Rising CO2 and increased light exposure synergistically reduce marine primary productivity. Nat Clim Change 2012; 2:519–23. doi: <u>10.1038/</u> NCLIMATE1507
- 94. Yang G, Gao K. Physiological responses of the marine diatom Thalassiosira pseudonana to increased pCO2 and seawater acidity. Mar Environ Res. 2012; 79: 142–151. doi: <u>10.1016/j.</u> <u>marenvres.2012.06.002</u> PMID: <u>22770534</u>
- Ludwig J, Canvin DT. The rate of photorespiration during photosynthesis and therelationship of the substrate of light respiration to the products of photosynthesis in sunflower leaves. Plant Physiol. 1971; 48: 712–719. PMID: <u>16657866</u>
- **96.** Somerville CR, Ogren WL. Genetic modification of photorespiration. Trends Biochem Sci.1982; 7: 171–174.
- 97. Goyal A, Tolbert NE. Association of glycolate oxidation with photosynthetic electron transport in plant and algal chloroplasts. Proc Natl Aca. Sci.1996; 93:3319–3324.
- Karentz D, Cleaver J, Mitchell DL. Cell survival characteristics and molecular responses of Antarctic phytoplankton to ultraviolet-B radiation, J Phycol. 1991; 27: 326–341.
- Sommaruga R, Hofer JS, Alonso-Sáez L, Gasol JM. Differential sunlight sensitivity of picophytoplankton from surface Mediterranean coastal waters. Appl Environ Microbiol. 2005; 71:2154–2157. PMID: 15812050
- Neale PJ, Sobrino C, Segovia M, Mercado JM, León P, Cortés MD, et al. Effect of CO2 nutrients and light on coastal plankton. I. Abiotic conditions and biological responses. Aquat Biol. 2014; 22: 25–41.