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```
H. C. Andersens Boulevard 44-46
DK-1553 Copenhagen V
Denmark
Telephone (+45) 33 38 6700
Telefax (+45) 339342 15
www.ices.dk
info@ices.dk
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## Executive summary

## Faroe Bank Cod

The total reported landings in 2014 were the lowest recorded since 1965 ( 30 tonnes).
The spring index suggests that the stock increased from 2012 to 2014 and declined substantially again in 2015. Nevertheless both the summer and spring index suggest that the stock is well below average and there is no indication of strong incoming year classes.

The results of an exploratory production model based on both surveys indicate a good agreement in the stock biomass index in recent years whereas the observed surveybased exploitation rates correlates reasonably well with estimated fishing mortalities. However the model failed to pick up the large increases in stock biomass observed in the 1996-2003 period. Correlation between modelled F's and summer survey based exploitation rates is $\mathrm{R}=0.90$. The exploitation ratio increased in 2011 as a consequence of the increase in landings and it decreased afterwards reflecting the fall of catches observed since 2012.

## Faroe Plateau cod

The input data consisted of the catch-at-age matrix (ages $2-10+$ years) for the period 1959-2014 and two age-disaggregated abundance indices obtained from the two Faroese groundfish surveys: the spring survey 1994-2015 (shifted back to the previous year) and the summer survey 1996-2014. The maturities were obtained from the spring survey 1983-2015.

The assessment settings were the same as in the 2014 assessment. An XSA was tuned with the two survey indices. The fishing mortality in 2014 (average of ages $3-7$ years) was estimated at 0.41 , which was higher than the Fmsy of 0.32 . The total stock size (age $2+$ ) in the beginning of 2014 was estimated at 27700 tonnes and the spawning stock biomass at 21100 tonnes, which was slightly above the limit biomass of 21000 tonnes.

The short term prediction until year 2017 showed a slightly decreasing total stock biomass to 24200 tonnes and a spawning stock biomass to 19500 tonnes. It is advised to reduce the fishing mortality substantially to rebuild the stock

## Faroe haddock

Being an update assessment, the changes compared to last year are additions of new data from 2014 and 2015 and some minor revisions of recent landings data with corresponding revisions of the catch at age data. The main assessment tool is an XSA tuned with two research vessel bottom trawl surveys. The results are in line with those from 2014, showing a very low SSB mainly due to poor recruitment but also due to higher than recommended fishing mortalities in recent years. SSB is now estimated well below Blim and is predicted to stay below Blim in 2016-2017 with status quo fishing mortality. Fishing mortality in 2014 is estimated at 0.29 and the average fishing mortality from 2012-2014 at 0.28 ( $\mathrm{F}_{\mathrm{mSY}}$ and $\mathrm{F}_{\mathrm{pa}}=0.25$ ). Landings in 2014 were 3200 t , which is slightly higher than in 2012 and 2013. This years assessment indicates that the 2014 assessment underestimated the 2013 recruitment by $23 \%$ ( 2 million versus 2.6 million, which still is the lowest on record), overestimated the fishing mortality in 2013 by $6 \%$ ( 0.28 versus 0.26 ) and underestimated the 2013 total- and spawning stock biomasses by $3 \%$ and $6 \%$, respectively ( 20 and 19 thous. t versus 19.6 and 18 thous. t).

## Faroe Saith

The most recent benchmark assessment was completed in 2010.
Nominal landings decreased by more than $25 \%$ from 35 kt. in 2012 to 24 kt. in 2014. The corresponding estimate of fishing mortality in 2014 (average of ages $4-8$ years) decreased to $\mathrm{F}=0.31$ which is lower than the historical average ( $\mathrm{F}=0.36$ ) and very close to $\mathrm{F}_{\mathrm{msy}}=0.30$ and $\mathrm{F}_{\mathrm{pa}}=0.28$. Due to high fishing mortality SSB decreased substantially from 127000 t . in 2005 to 48000 t . in 2013, i.e., below $\mathrm{B}_{\text {trigger }}=55 \mathrm{kt}$. but it increased again to 70 000 t . in 2014 as a consequence of improved weights and maturity ogives.
Numbers of the most recent year-class (2011, age 3 in 2014) has increased substantially from 36 mill. in 2013 to 62 million in 2014. However a statistical separable model suggests that the 2011 year-class is not as strong as the spaly assessment estimate and it predicts recruitment for 2014 at 20 mill.
At status-quo $\operatorname{Fbar}^{\text {(2015 }}=0.31$ and recruitment $\operatorname{Rec}(2015)=27$ mill. the SSB is predicted to increase to 97 kt . in 2016.
Predicted landings for 2014 in the last year assessment were around 38 kt while the actual measurement was 24 kt . The estimate of $\mathrm{F}_{\text {bar }}$ in 2014 was $\mathrm{F}_{\text {bar }}=0.53$ in last year's assessment and $F_{b a r}=0.32$ in the 2015 assessment. Recruitment strength for 2014 was predicted at 28 million while the estimate for that year in the present assessment reached 62 million. SSB was predicted exactly in $2014 \mathrm{SSB}(2014)=70000 \mathrm{t}$.

## Icelandic saithe

The 2015 reference biomass (B4+) is estimated as 255 kt , around the average in the assessment period (1980 to the present). Spawning biomass is estimated as 139 kt , above the average in the assessment period and well above $B_{\text {trigger }}=65 \mathrm{kt}$ and $B_{\lim }=61 \mathrm{kt}$.

Harvest rate has been around the HCR target of $20 \%$ since 2011, with fishing mortality rate between 0.19 and 0.25 . Year classes 2008 and 2009 are above average, but recruitment has declined below average since then.

Weights of ages 3-6 have been low in recent years, but older ages are close to average weight. Maturity at ages 4-9 has decreased in recent years and is currently around average.

The assessment model is a separable statistical catch-at-age model implemented in AD Model Builder. Selectivity is age-specific and varies between three periods: 1980-1996, 1997-2003, and 2004 onwards.

The default separable model (ADSEP) estimates a slightly larger stock size than alternative diagnostic models (ADAPT, TSA, SAM). The estimates of this year's B4+ range from 209 (TSA) to 255 kt (ADSEP).

In 2013, the Icelandic government adopted a harvest control rule for managing the Icelandic saithe fishery, evaluated by ICES (2013). It is similar to the $20 \%$ rule used for the Icelandic cod fishery. When the population is above $B_{\text {trigger, }}$ the TAC set in year t equals the average of $0.2 \mathrm{~B} 4+$ in year t and last year's TAC.

According to the adopted harvest control rule, the TAC will be 55 kt in the next fishing year.

## Icelandic cod

The spawning stock (SSB2015) is estimated to be 547 kt and is higher than has been observed over the last five decades. The reference biomass $(B 4+; 2015)$ is estimated to
be 1302 kt , the highest observed since the late 1970's. Fishing mortality, being 0.3 in 2014, has declined significantly in recent years and is presently the lowest observed in last 6 decades. Year classes since the mid-1980s are estimated to be relatively stable but with the mean around the lower values observed in the period 1955 to 1985. According to the adopted management harvest rule the TAC will be 239 kt in the next fishing season. ICES has evaluated the plan and concludes that it is in accordance with the precautionary approach and the ICES MSY framework.

Mean weight at age in the stock and the catches that were record low in 2006-8 have been increasing in recent years and are now around the long term mean. The input in the analytical assessments are catch at age 1955-2014 and spring groundfish survey (SMB) indices at age from 1985-2015 and fall survey groundfish survey (SMH) indices at age from 1996-2014. The results from the AD-Model builder statistical Catch at Age Model (ADCAM) as was used as the final run. This framework has been the basis for the advice since 2002.

The reference stock (B4+) in 2014 is now estimated to be 1181 kt compared to 1106 kt last year. The SSB in 2014 is now estimated to be 425 kt compared to 427 kt estimated last year. Fishing mortality in 2013 is now estimated 0.29 compared to 0.3 estimated last year. Year classes 2011-2013 were estimated to be 181, 160 and 109 million in last years assessment and are now estimated to be 181, 161 and 115 millions.

## Icelandic haddock

The 2014 yearclass is estimated to be large, after 6 consecutive small yearclass from 2008-2013. The Current assessment shows some upward revision of the stock compared to last years assessment, mostly caused by more growth than predicted. The main features are though the same that recruitment is poor and the stock is predicted to decrease somewhat in next two years before the 2014 yearclass recruits to the stock.

Growth in 2014 was above average since 1985 and more than predicted and the mean weight of young fish is above average while old fish are close to average. The assessment procedure was the same as last year (SPALY), an Adapt type model tuned with both the surveys.

There are differences in the perception of the state of stock in assessment based on either the spring or autumn survey with autumn survey indicating a larger stock. It has been like that since 2009. Different models using the same tuning data show similar results.

Advice is given according to the adopted Harvest Control Rule, and the advice for the fishing year 2014/2015 (September 1st 2015 - August 31st 2016) is 36400 tonnes. The advice for the following fishing year is predicted to be approximately 31000 tonnes but increasing after that when the 2014 yearclass comes in.

No environmental drivers or ecosystem effects are known that can help in prediction of the development of the haddock stock. Some effect of the environment on the stock can though not be excluded.

## Icelandic summer spawning herring

The total reported landings in 2014/15 fishing season were 95.5 kt but the TAC was set at 83 kt -the difference being caused by a transfer of quota between years. The fishable stock (age $4+$ ) in the winter surveys 2014/15 was estimated at 433 kt , compared to 410 kt in the winter 2013/14. The 2013 year class (age 1 in 2014) appears small.

This is an update assessment where the 2014 data have been added to the input data and no revisions of last year's data. The analytical assessment model, NFT-Adapt, indicates that the biomass of age $3+$ is 438 kt and SSB is 342 kt in the beginning of 2015. Record small year class from 2011 entering the spawning stock in 2015, causes a decline in SSB but it is still above BPA. Fishing at F0.1= 0.22 in the fishing season 2015/16 will give a catch of 71 thousand tons. SSB in 2016 is expected to be 327 kt .

Changes in the predictions approach, where the geometric mean for number at age-3 in the assessment year (2015) was replaced by a projection of number at age 3 from a survey estimate at age- 1 of the year class, has decreased the uncertainty and has minor impacts on the advice ( $3 \%$ lower). This year's results support that additional natural mortality in the stock due to Ichthyophonus infection should only be applied for the first two years of the outburst."

## Capelin in the Iceland-East Greenland-Jan Mayen area

In May 2014 ICES advised on the basis of precautionary considerations that the initial quota be set at the $50 \%$ of the predicted quota, implying an initial quota of 225000 t . The quota was revised to 260000 t after an acoustic survey in October. The final TAC of $580000 t$ for the fishing season 2014/2015 was set after an acoustic survey in January 2015.

The total landings in the fishing season 2014/2015 amounted 517 thous. t (preliminary data). Around 40000 t were taken in July 2014, 5000 t in November 2014 and the rest during the winter months January-March 2015.

The stock was benchmarked in January 2015. New methods for setting a final TAC and an initial quota were established. They will both apply for the next fishing season. $\mathrm{B}_{\mathrm{lim}}$ was defined during the meeting.

The acoustic index of 1 year old capelin from the acoustic survey is of an average size. On the basis of the new prediction model the initial quota in 2015/16 is 54 thous. t .

As the capelin increases its weight rapidly over the summer it is recommended that the fishery doesn't start until late autumn.

## Offshore West Greenland Cod

From 2015 the advice for cod in Greenland offshore waters has been split in two stock components (advice year 2016). The West Greenland offshore stock component is now comprised of the NAFO subdivisions 1A-E in West Greenland. The East Greenland stock component is comprised of the area NAFO subdivision 1F in South Greenland and ICES subarea XIV in East Greenland.

Some mixing occurs between the two stocks in West Greenland which at present is considered to act as a nursing area for juveniles of the East Greenland stock component. The offshore fishery in West Greenland was closed in accordance with an implemented management plan in 2014. However, a dispensation was given to a small trawler that fished 116 tons and the 2009 YC dominated the catches.

Survey indices show that the biomass and abundance has increased due to the 2009 YC which is present in considerable numbers. This YC is distributed further south in 2012-2014 than in 2011.

The spatial distribution of the 2009 YC is different than previous year classes that usually migrate out of the area at age 4 , but a large part of the 2009 YC still remains in the southern area (NAFO 1E) at age 5 in 2014.

No formal assessment is conducted and there are no biological reference points for the stock. Information from survey indices (German Groundfish survey and Greenland Shrimp and Fish survey) are used as basis for advice.

No significant spawning has been observed in the area, and fish older than 6 yrs are lacking in the area.

## Inshore Greenland cod

Total catches from the inshore fishery were 18331 t in 2014 which is the highest since early 1990'ies. Several year-classes were caught in the fishery but catches were dominated by the 2009 YC.

The mean length in the fisheries has increases from 44 cm in 2006 to 58 cm in 2014. Survey recruitment indices from the inshore area show that incoming year classes (2011 and 2012) are below average.

The stock was benchmarked in 2015 and a new procedure for making catch advice was adopted. The procedure is based on a linear regression on pairs of survey values (ages $3-8)$ and catches in the following year. The advice is based on the average of the 2013 and 2014 survey values for ages 3-8 multiplied by a scaling factor.

## Cod in East Greenland, South Greenland

From 2014 the management for cod in Greenland offshore waters has been split in two stock components according to areas: NAFO subdivisions 1A-E in West Greenland and NAFO subdivision 1F in South Greenland combined with ICES subarea XIV in East Greenland. The ICES advice for 2016 has for the first time been given according to these two areas.

The offshore fishery in East and South Greenland in 2014 was conducted as an experimental fishery with a TAC of 10000 tons. Total catches were 7893 tons. The year-class dominating the catches was the 2007 YC, which it has done since 2012. The largest cod (mean length of 83 cm ) were caught by trawlers on Dohrn Bank close to the Iceland EEZ.

Available survey biomass indices from the Greenland and German surveys show that the biomass has increased due to the growth of the 2009 YC and in part the 2007 YC. Abundance has however decreased as fewer young fish are observed.

The 2009 YC followed by the 2007 YC has dominated the survey since 2012. The 2009 YC is primarily distributed in South Greenland, whereas the 2007 YC is distributed more to the north in East Greenland. Spawning offshore cod are only found in East Greenland in local high densities.

The procedure suggested as basis for advice at the Benchmark in 2015 was not implemented by NWWG due to shortcomings. Instead, advice was based on an Fproxy multiplier generated from the relationship between the catches and survey index in a period with a considered sustainable fishery, multiplied by the latest year's smoothed survey index (Greenland Shrimp and Fish survey).

## Greenland Halibut in Subareas V, VI, XII, and XIV

Input data to the assessment: current surveys have continued and sampling intensity and coverage remains also unchanged. Logbooks from the fishery are available as haul by haul data. Since 2001 no age readings of otoliths were available from the main fishing areas which impede age based assessment.

A logistic production model in a Bayesian framework has been used to assess stock status and for making predictions. The model includes an extended catch series going back to the assumed virgin status of the stock at the beginning of the fishery in 1961. Estimated stock biomass showed an overall decline along with the high catches in the late 1980s and early 1990s. Since 2004/2005 the stock has increased slowly and is now at $70 \%$ BMSY and fishing mortality has decreased to FMSY. Although the indices that are used for input to the assessment model (combined survey index at Greenland and Iceland) and logbook information from Iceland trawler fishery all show a slight decrease in 2014, the remaining available indices from East Greenland (logbook from trawl fishery) and from Faroese waters (logbooks from trawl fishery and a survey) all suggest high and stable biomass in recent years.

## Golden redfish (Sebastes norvegicus) in Subareas V, VI and XIV

Total landings in 2014 were about 50800 t , which is about 2500 t less than in 2013. About $94 \%$ of the catches were taken in Division Va. A substantial increase in landings from XIVb since 2010, the highest since early 1990s, and is in relation to re-established redfish fishery in 2010. Very little redfish is now taken in Vb .

Catch-at-age data from Va show that the catch was dominated by two strong year classes from 1985 and 1990. From 2008-2011 year-classes 1996-1999 were the most important in the fisheries. Their share has reduced relatively fast and the 2000-2005 year classes are now most important contributing about $60 \%$ of the total catch.

Recruitment seems to be low in all areas, both according to the Icelandic groundfish surveys, and the German survey and the Greenland shrimp and fish survey in EastGreenland. Recruitment is likely to be underestimated as the surveys do not adequately cover nursery areas of the stock.

The stock was benchmarked in 2014 and a management plan evaluated and adopted. The Gadget model was used as basis for advice but the main difference in settings from earlier years was inclusion data from the German survey in East-Greenland and changes in growth rate.

The management plan was based on F9-19=0.097 reducing linearly if the spawning stock is estimated below $220000 \mathrm{t}\left(\mathrm{B}_{\text {trigger }}\right)$. Blim was proposed as 160000 t , lowest SSB in the 2012 run.

According to the management plan the TAC for 2016 will be 51000 t .

## Icelandic slope Sebastes mentella in Va and XIV

ICES concluded in February 2009 that S. mentella is to be divided to three biological stocks and that the S. mentella on the continental shelf and slope of Iceland should be treated as a separate biological stock and management unit.

Total landings of demersal S. mentella in Icelandic waters in 2014 were about 9500 t , 750 t more than in 2013.

No analytical assessment was conducted and there are no biological reference points for the species. Survey indices from the annual autumn survey since 2000 are used as basis for advice.

Available survey biomass indices show that in Division Va the biomass has gradually decreased from 2006 to 2013, but increased in 2014.

The East-Greenland shelf is most likely a nursery area for the stock. No new recruits $(>18 \mathrm{~cm})$ are seen in the survey catches of the German survey and the Greenland
shrimp and fish shallow water survey conducted in the area and no juveniles are present ( $<18 \mathrm{~cm}$ ) recent years.

Icelandic slope S. mentella is considered a data limited stock (DLS) and follows the ICES framework for such (Category 3.2). When the precautionary approach is applied, catches in 2016 should be no more than 9954 t . All catch are assumed to be landed.

## Shallow Pelagic Sebastes mentel/a

ICES concluded in February 2009 that S. mentella is to be divided into three biological stocks and that the shallow pelagic S. mentella in the Irminger Sea and adjacent areas should be treated as separate biological stock and management unit.

Total landings of shallow pelagic S. mentella in 2014 were 6423 t , a significant increase compared to 1527 t in 2013. The catches were almost entirely taken in ICES XII.

No analytical assessment was conducted and there are no biological reference points for the species. Survey indices from the biennial international acoustic redfish survey conducted in the Irminger Sea and adjacent waters since 1991 are used as basis for advice.

The last survey was conducted in June/July 2013. Since 1994, the results of the acoustic survey show a drastic decreasing trend within the deep scattering layer (DSL) from 2.2 million $t$ to 91000 t in 2013. With the trawl method within the DSL $(350-500 \mathrm{~m})$ the biomass was estimated 200000 t , significantly below the 361000 t of 2011. The next international acoustic redfish survey was scheduled to be conducted in June/July 2015.

No signs of recruitment have been observed in the latest German survey on the EastGreenland shelf.

## Deep Pelagic Sebastes mente/la

ICES concluded in February 2009 that S. mentella is to be divided to three biological stocks and that the deep pelagic S. mentella in the Irminger Sea and adjacent areas should be treated as separate biological stock and management unit.

Total landings of deep pelagic S. mentella s in 2014 were 23755 t , half of the 2013 total catch.

No analytical assessment was conducted and there are no biological reference points for the species. Survey indices from the biennial international trawl-acoustic redfish survey conducted in the Irminger Sea and adjacent waters since 1999 are used as basis for advice.

The survey was conducted in June/July 2013. A total biomass of 280900 t was estimated, a 41\% less than in 2011 (474 000 t).Trawl survey estimates in 2011 and 2013 are lower than the average for 1999-2009 and the estimate for 2013 is the lowest observed. The next international trawl-acoustic redfish survey in the Irminger Sea will be conducted in June/July 2015.

No recruitment has been observed on the East-Greenland shelf during 2013 and 2014, which is a concern because it is assumed to contribute to the three $S$. mentella stocks at unknown shares.

## Greenlandic slope Sebastes mente/la in XIVb

ICES concluded in February 2009 that demersal S. mentella is to be divided into three biological stocks and that the $S$. mentella on the continental shelf and slope should be
treated as a separate biological stock and management unit. This separation of the stocks did not include the adult S. mentella on the Greenlandic slopes. ICES therefore decided that NWWG will conduct a separate assessment of S. mentella in subarea XIVb until further information is available to assign stock origin. This chapter therefore deals only with the S. mentella on the Greenlandic Slope.
Total landings of demersal S. mentella in East Greenland waters in 2014 were about 4600 tons, which is less than 2010-2012 landings. The lower catches are partly due to a lower presence of $S$. mentella in the mixed stock fishery and partly due to a lower total landing of demersal redfish.

In the decade before 2009 S. mentella was mainly a valuable by-catch in the fishery for Greenland halibut. However, since 2009 a fishery directed towards demersal redfish has taken place.

Available survey biomass indices show that in Division XIVb the biomass decreased further in 2014. No new recruits ( $>18 \mathrm{~cm}$ ) are seen in the surveys since 2012, and no juveniles are present ( $<18 \mathrm{~cm}$ ) in both 2013 and 2014 surveys

The advice is based on the DLS approach (3.2) using the Greenland shallow water survey as basis for advice. The ratio is applied to the 2014 advice as catches are well above the current advice. The advice for 2016 is 2240 t .

## 1 Introduction

### 1.1 Terms of Reference (ToR)

### 1.1.1 Specific ToR

2014/2/ACOM08 The North-Western Working Group (NWWG), chaired by Rasmus Hedeholm, Greenland, will meet at ICES Headquarters, 28 April - 5 May, 2015 to:
a) Address generic ToRs for Regional and Species Working Groups.

The assessments will be carried out on the basis of the stock annex in National Laboratories, prior to the meeting. This will be coordinated as indicated in the table below.

For capelin in Iceland-East Greenland-Jan Mayen area, Iceland will provide a WG type report and a draft advice sheet on 5 May. NWWG will agree any changes to the WG type report and the Advice sheet no later than 7 May. An ADG will work by correspondence 11 May. The WEBEX will be 15 May, and the Advice Release date 19 May.

Other material and data relevant for the meeting must be available to the group no later than 14 days prior to the starting date.

NWWG will report by 19 May 2015 for the attention of ACOM. For capelin in IcelandEast Greenland-Jan Mayen area NWWG will report by 1 February 2015 for the attention of ACOM.

| Fish Stock | Stock Name | Stock Coord. | Assess. Coord. 1 | Assess. Coord. 2 | Advice |
| :---: | :---: | :---: | :---: | :---: | :---: |
| cod-farp | Cod in Subdivision Vb2 (Faroe Bank) | Faroe Islands | Faroe Islands | Faroe Islands | Update |
| cod-farb | Cod in Subdivision Vb2 (Faroe Bank) | Faroe Islands | Faroe Islands | Faroe Islands | Multiyear |
| had-faro | Haddock in Division Vb | Faroe Islands | Faroe Islands | Faroe Islands | Update |
| sai-faro | Saithe in Division Vb | Faroe Islands | Faroe Islands | Faroe Islands | Update |
| cod-iceg | Cod in Division Va (Icelandic cod) | Iceland | Iceland | Iceland | Update |
| had-iceg | Haddock in Division Va (Icelandic haddock) | Iceland | Iceland | Iceland | Update |
| sai-icel | Saithe in Division Va (Icelandic saithe) | Iceland | Iceland | Iceland | Update |
| her-vasu | Herring in Division Va (Icelandic summer-spawners) | Iceland | Iceland | Iceland | Update |
| cap-icel | Capelin in Subareas V, XIV and Division Ila west of $5^{\circ} \mathrm{W}$ (Iceland-East Greenland-Jan Mayen area | Iceland | Iceland | Iceland | Update |
| cod-ingr | Cod (Gadus morhua) in NAFO Subarea 1, inshore (Inshore West Greenland) | Greenland | Greenland | Greenland | Update |
| cod-segr | Cod (Gadus morhua) in ICES <br> Subarea XIV and NAFO <br> Subdivision 1F (East Greenland, <br> South Greenland) | Greenland | Greenland | Germany | Update |
| cod-wgr | Cod (Gadus morhua) in NAFO Subdivision 1A-E (Offshore West Greenland) | Greenland | Greenland | Germany | Update |
| ghl-grn | Greenland halibut in Subareas V, VI, XII and XIV | Greenland | Greenland | Iceland | Update |
| smr-5614 | Redfish (Sebastes marinus) in Subareas V, VI, XII and XIV | Iceland | Iceland | Faroe Islands | Update |
| smn-con | Beaked redfish (Sebastes mentella) in Division Va and Subarea XIV (Icelandic slope stock). | Iceland | Iceland | Germany | Multiyear |
| smn-sp | Beaked Redfish (Sebastes mentella) in Subareas V, XII, XIV and NAFO Subareas $1+2$ <br> (Shallow Pelagic stock < 500 m deep) | Iceland | Germany | Spain | Update |
| smn-dp | Beaked Redfish (Sebastes mentella) in Subareas V, XII, XIV and NAFO Subareas $1+2$ (Deep Pelagic stock > 500 m deep) | Iceland | Germany | Spain | Update |
| smn-grl | Beaked Redfish (Sebastes mentella) in Subarea XIV (East Greenland Slope) | Greenland | Greenland | Germany | Update |

### 1.1.2 Generic ToRs for Regional and Species Working Groups

The working group should focus on:
a) Consider and comment on ecosystem overviews where available
b) For the fisheries considered by the working group consider and comment on:
i. descriptions of ecosystem impacts of fisheries where available
ii. descriptions of developments and recent changes to the fisheries
iii. Mixed fisheries overview, and
iv. emerging issues of relevance for the management of the fisheries.
c) Conduct an assessment to update advice on the stock(s) using the method (analytical, forecast or trends indicators) as described in the stock annex and produce a brief report of the work carried out regarding the stock, summarising where the item is relevant:
i. Input data (including information from the fishing industry and NGO that is pertinent to the assessments and projections);
ii. Where misreporting of catches is significant, provide qualitative and where possible quantitative information and describe the methods used to obtain the information;
iii. For relevant stocks estimate the percentage of the total catch that has been taken in the NEAFC regulatory area by year in the recent three years.
iv. The developments in spawning stock biomass, total stock biomass, fishing mortality, catches (wanted and unwanted landings and discards) using the method described in the stock annex;
v. The state of the stocks against relevant reference points;
vi. Catch options for next year;
vii. Historical performance of the assessment and catch options and brief description of quality issues with these;
d) Produce a first draft of the advice on the fish stocks and fisheries under considerations according to ACOM guidelines.
The working group is furthermore requested to:
e) Propose specific actions to be taken to improve the quality and transmission of the data (including improvements in data collection);
f) Prepare the data calls for the next year update assessment and for the planned data compilation workshops
g) Update, quality check and report relevant data for the stock:
i. Load fisheries data on effort and catches into the INTERCATCH database by fisheries/fleets;
ii. Abundance survey results;
iii. Environmental drivers.
h) Produce an overview of the sampling activities on a national basis based on the INTERCATCH database or, where relevant, the regional database.
i. Identify research needs of relevance for the working group.

### 1.2 NWWG 2015 work in relation to the ToR

The ToRs were not addressed systematically for all stocks. The main focus was on the adoption of assessments that were the basis for stock status and the premise for the forecasts. This was done to ensure that the basis for the advice was agreed upon. This year, individual report stock sections were not reviewed in plenary due to time constraints, but relevant issues were discussed in plenary and each section was reviewed by other WG members. The summary sheets were all reviewed in plenary.
ad a) Ecosystem overviews were available for the Faroe Islands, Iceland and Greenland ecoregions. In the Icelandic ecosystem the increased temperature/salinity since mid 1990s is regarded as a major factor, which has shifted the distribution of many fish species northwards. The biomass of capelin (an important forage fish) is the only ecosystem driver, which has been used directly in the assessment (predicting individual weights of cod). This relationship became less clear in recent years and the weights of cod were therefore estimated in other ways. No new ecosystem driver was proposed this year. It was, however, remarked that the effects of important ecosystem drivers was expected to be expressed in the input data of the stock assessments and therefore taken into account in an indirect way. In the Greenland ecoregion the effect of temperature and wind is outlined since these measures are good indicators of the recruitment of cod. These measures are, however, not used directly in the assessments or the advice. In the Faroe ecoregion there has been shown a positive relationship between primary production and the production of demersal fish (cod, haddock and saithe). Primary production is, however, not used directly in the assessments or advice.
$a d b)$ In the overview sections there is a description of the fisheries, including mixed fisheries. In the Icelandic ecoregion the ecosystem effect of fisheries is briefly mentioned with corals being destroyed by fishing gears.
ad e-h) INTERCATCH is not used extensively for any stocks considered in this WG as they are mostly national non-EU stocks, and no data was loaded into INTERCATCH that was not already uploaded. The members of the group were however encouraged to use it and to provide the status on the use of INTERCATCH. The 2015 data call was evaluated for all stocks and updated to facilitate the 2016 data call.

### 1.3 MSY/HCR approaches for individual stocks

See Introduction in the 2013 NWWG report.

### 1.4 Assessment methods applied to NWWG stocks

The methods applied to assess the stock status of the NWWG stocks covers a wide range from descriptive to age based analytical assessments as follows:

| Stock | Assessment model | Input* |
| :--- | :--- | :--- |
| Faroe Bank cod | Trend based assessment | Survey |
| Faroe Plateau cod | XSA | Survey |
| Faroe haddock | XSA | Survey |
| Faroe saithe | XSA | CPUE** |
| Iceland saithe | ADCAM (statistical catch-at-age) | Survey |
| Iceland cod | ADCAM (statistical catch-at-age) | Survey |
| Iceland haddock | Adapt type model | Survey |
| Iceland herring | NFT-Adapt | Survey |
| Capelin | Linear regression | Survey |
| Inshore West Greenland cod | Linear regression | Survey |
| East Greenland, South | Fproxy multiplier | Survey |
| Greenland cod | Descriptive | Survey |
| Offshore West Greenland cod | Survey + CPUE |  |
| Greenland halibut | Stock production model (Bayesian) | Survey |
| S. norvegicus | GADGET (age-length based cohort | Survey |
| S. mentella Iceland slope | Descriptive | Survey + CPUE |
| Deep pelagic S. mentella | Descriptive | Survey + CPUE |
| Shallow pelagic S. mentella | Descriptive |  |
| S. mentella Greenland Slope | Descriptive | Sodel) |

* landings or landings by age are input to all assessments
** The CPUE is adjusted by survey information about distribution width.


### 1.5 Benchmarks and workshops

At WKICE cod from Iceland, Greenland and the IGJM capelin stock were benchmarked. For the Iceland cod no significant changes were suggested. For Greenland cod the benchmark concluded that advice should be given for three separate stocks: Inshore West Greenland cod, East Greenland/South Greenland cod and Offshore West Greenland cod. For capelin a new advice procedure was proposed and accepted.

No stocks are scheduled to be benchmarked in 2016 and no new stocks were proposed for a 2017 benchmark. Faroe plateau and bank cod are scheduled to be benchmarked in 2017 and issue lists have previously been prepared for both stocks.

### 1.5.1 Inshore West Greenland cod and the 2015 NWWG assessment

There are age disaggregated data on both survey and landings for this stock. However, the data are highly variable and the internal consistency is low. As a consequence, none of the proposed analytical assessment approaches presented at the benchmark produced convincing results and were not considered appropriate to generate advice. However, the previously used approach (DLS on the survey) was also rejected. Instead an alternative regression approach was approved. It is based on a simple linear regression between survey indices (year y) and subsequent catches (year $y+1$ ). The NWWG concluded that the average of the last two years survey should be used (and not only
the latest as concluded by the benchmark) to smooth the otherwise noisy survey biomass estimate. Both the benchmark and NWWG concluded that this approach should be an interim solution and that a full stock assessment is developed.

### 1.5.2 East Greenland/South Greenland cod and the 2015 NWWG assessment

The benchmark suggested that the advice should be based on the DLS approach since no satisfactory analytical assessment models could be developed in time for the benchmark to properly assess them. The NWWG concluded that this method was inappropriate for this stock but insufficient exploratory analysis at the benchmark failed to realize this. NWWG instead suggested that the advice is based on an estimate of fishing mortality ( $\mathrm{F}_{\text {proxy }}$ ) and the Greenland survey index for the area. To incorporate uncertainty and adopt a precautionary approach, the $\mathrm{F}_{\text {proxy }}$ was based on a period of stock increase and the survey was smoothed using a random effects smoother to minimize year-to-year variation.

### 1.5.3 Offshore West Greenland cod and the 2015 NWWG assessment

The benchmark and the NWWG both concluded that this stock is at a low level although recent survey results show an increase in biomass caused mainly by the 2009 year class. Consequently, the advice continues to be that no fishery should take place on this stock.

### 1.5.4 Iceland-Greenland- Jan Mayen capelin (IGJM) and the 2015 NWWG assessment

The NWWG produced advice in accordance with the benchmark conclusions. Hence, for IGJM capelin the advice will be based on a stochastic projection of the stock that will be conducted starting from acoustic measurements, aiming at a TAC that is associated with less than $5 \%$ probability of the stock being below $\operatorname{Blim}(150.000 \mathrm{t}$ ). An initial TAC advice given for the 2015/2016 fishing season will be subject to revision following a 2015 autumn survey and again following a 2016 winter survey. A special request was received from the coastal states involved in the fishery on IGJM capelin, and they wished to be informed about the advice had it been based on the management in place for this stock. This was provided by the NWWG.

### 1.6 Chairman

This is the first year for the new chairman, Rasmus Hedeholm, Greenland, who is scheduled to chair the group from 2015-2017.

## 2 Demersal Stocks in the Faroe Area (Division Vb and Subdivision Ila4)

### 2.1 Overview Fisheries

The main fisheries in Faroese waters are mixed-species, demersal fisheries and single species pelagic fisheries (Figure 2.1). The demersal fisheries are mainly conducted by Faroese vessels, whereas the pelagic fisheries are conducted both by Faroese vessels and by foreign vessels licensed through bilateral and multilateral fisheries agreements. The usual picture changed in 2011, however, since no mutual agreement could be reached between the Faroe Islands and the EU and Norway, respectively, due to the dispute regarding the share of mackerel. From 2013, the agreement has been re-established.

## Pelagic Fisheries

Three main species of pelagic fish are fished in Faroese waters: blue whiting, herring and mackerel; several nations participate. The Faroese pelagic fisheries are conducted by purse seiners, larger purse seiners also equipped for pelagic trawling and trawlers otherwise performing demersal fisheries. The pelagic fishery by Russian vessels is conducted by large factory trawlers. Other countries use purse seiners and factory trawlers.

## Demersal Fisheries

Although they are conducted by a variety of vessels, the demersal fisheries can be grouped into fleets of vessels operating in a similar manner. Some vessels change between longlining, jigging and trawling, and they therefore can appear in different fleets. The number of licenses can be found in Table 2.3. The grouping of the vessels under the management scheme can be seen in section 2.1.3.

### 2.1.1 Fisheries and management measures

The fishery around the Faroe Islands has for centuries been an almost free international fishery involving several countries. Apart from a local fishery with small wooden boats, the Faroese offshore fishery started in the late $19^{\text {th }}$ century. The Faroese fleet had to compete with other fleets, especially from the United Kingdom with the result that a large part of the Faroese fishing fleet became specialised in fishing in other areas. So except for a small local fleet most of the Faroese fleet were fishing around Iceland, at Rockall, in the North Sea and in more distant waters like the Grand Bank, Flemish Cap, Greenland, the Barents Sea and Svalbard.

Up to 1959, all vessels were allowed to fish around the Faroes outside the 3 nm zone. During the 1960s, the fisheries zone was gradually expanded, and in 1977 an EEZ of 200 nm was introduced in the Faroe area. The demersal fishery by foreign nations has since decreased and Faroese vessels now take most of the catches. The fishery may be considered a multi-fleet and multi-species fishery as described below.

During the 1980s and 1990s the Faroese authorities have regulated the fishery and the investment in fishing vessels. In 1987 a system of fishing licenses was introduced. The demersal fishery at the Faroe Islands has been regulated by technical measures (minimum mesh sizes and closed areas). In order to protect juveniles and young fish, fishing is temporarily prohibited in areas where the number of small cod, haddock and saithe exceeds $30 \%$ (in numbers) of the catches; after $1-2$ weeks, sometimes longer, the areas
are again opened for fishing. A reduction of effort has been attempted through banning of new licenses and buy-back of old licenses.

A quota system, based on individual quotas, was introduced in 1994. The fishing year started on 1 September and ended on 31 August the following year. The aim of the quota system was, through restrictive TACs for the period 1994-1998, to increase the SSBs of Faroe Plateau cod and haddock to 52000 t and 40000 t , respectively. The TAC for saithe was set higher than recommended scientifically. It should be noted that especially cod and haddock but also saithe are caught in a mixed fishery and any management measure should account for this. Species under the quota system were Faroe Plateau cod, haddock, saithe, redfish and Faroe Bank cod.

The catch quota management system introduced in the Faroese fisheries in 1994 was met with considerable criticism and resulted in discarding and in misreporting of substantial portions of the catches. Reorganisation of enforcement and control did not solve the problems. As a result of the dissatisfaction with the catch quota management system, the Faroese Parliament discontinued the system as from 31 May 1996. In close cooperation with the fishing industry, the Faroese government has developed a new system based on individual transferable effort quotas in days within fleet categories. The new system entered into force on 1 June 1996. The fishing year from 1 September to 31 August, as introduced under the catch quota system, has been maintained.

The individual transferable effort quotas apply to 1) the longliners less than 110 GRT, the jiggers, and the single trawlers less than 400 HP (Groups 4,5), 2) the pair trawlers (Group 2) and 3) the longliners greater than 110 GRT (Group 3). The single trawlers greater than 400 HP were in 2011 included into the fishing days system and were allocated a number of fishing days (Tables 1 and 2). They are not allowed to fish within the 12 nautical mile limit and the areas closed to them, as well as to the pair trawlers, have increased in area and time. Their catch of cod and haddock was before 2011 limited by maximum by-catch allocation. This fleet has now started to pair-trawl, and since the fiscal year 2011/12, merged with the pair-trawlers group. The single trawlers less than 400 HP are given special licenses to target flatfishes inside 12 nautical miles with a bycatch allocation of $30 \%$ cod and $10 \%$ haddock. In addition, they are obliged to use sorting devices in their trawls in order to minimize their by-catches. One fishing day by longliners less than 110 GRT is considered equivalent to two fishing days for jiggers in the same gear category. Longliners less than 110 GRT could therefore double their allocation by converting to jigging. Table 2.1 shows the allocated number of fishing days by fleet group since the fiscal year 1996/1997 and in Table 2.2 is a comparison between number of allocated days and number of actually used fishing days. From Table 1 it can been seen that since 1996/1997, the number of days allocated has been reduced considerable and is now $50 \%$ of the originally allocated days. Despite this, there still are many unused days in the system (Table 2.2).

Holders of individual transferable effort quotas who fish outside the thick line on Figure 2.2 can fish for 3 days for each day allocated inside the line. Trawlers are generally not allowed to fish inside the 12 nautical mile limit. Inside the innermost thick line only longliners less than 110 GRT and jiggers less than 110 GRT are allowed to fish. The Faroe Bank shallower than 200 m is closed to trawling. Due to the serious decline of the Faroe Bank cod, the Bank has been closed since 1 January 2009 for all gears except for a minor jigging fishery during summer time.

The fleet segmentation used to regulate the demersal fisheries in the Faroe Islands and the regulations applied are summarized in Table 2.3.

The effort quotas are transferable within gear categories. The allocations of number of fishing days by fleet categories was made such that together with other regulations of the fishery they should result in average fishing mortalities on each of the 3 stocks of 0.45 , corresponding to average annual catches of $33 \%$ of the exploitable stocks in numbers. Built into the system is also an assumption that the day system is self-regulatory, because the fishery will move between stocks according to the relative availability of each of them and no stock will be overexploited. These target fishing mortalities have been evaluated during the 2005 and 2006 NWWG meetings. The realized fishing mortalities have been substantially higher than the target for cod, appear to have exceeded the target for saithe in recent years, while for haddock, fishing mortality remains below the target.

In addition to the number of days allocated in the law, it is also stated in the law what percentage of total catches of cod, haddock, saithe and redfish, each fleet category on average is expected to fish. These percentages are as follows:

| Fleet category | Cod | Haddock | Saithe | Redfish |
| :--- | :--- | :--- | :--- | :--- |
| Longliners < 110GRT, |  |  |  |  |
| jiggers, single trawl. < 400HP | $51 \%$ | $58 \%$ | $17.5 \%$ | $1 \%$ |
| Longliners > 110GRT | $23 \%$ | $28 \%$ |  |  |
| Pairtrawlers | $21 \%$ | $10.25 \%$ | $69 \%$ | $8.5 \%$ |
| Single trawlers > 400 HP | $4 \%$ | $1.75 \%$ | $13 \%$ | $90.5 \%$ |
| Others | $1 \%$ | $2 \%$ | $0.5 \%$ | $0.5 \%$ |

The technical measures as mentioned above are still in effect. An additional measure to reduce the fishing mortality on cod and haddock and to especially reduce the mortality on the youngest age groups has been introduced (See the 2013 NWWG report) in July 2011, but was terminated in August 2013.

### 2.1.2 The marine environment and potential indicators

The waters around the Faroe Islands are in the upper 500 m dominated by the North Atlantic current, which to the north of the islands meets the East Icelandic current. Clockwise current systems create retention areas on the Faroe Plateau (Faroe shelf) and on the Faroe Bank. In deeper waters to the north and east and in the Faroe Bank channel there is deep Norwegian Sea water, and to the south and west is Atlantic water. From the late 1980s the intensity of the North Atlantic current passing the Faroe area decreased, but it has increased again in the most recent years. The productivity of the Faroese waters was very low in the late 1980s and early 1990s. This applies also to the recruitment of many fish stocks, and the growth of the fish was poor as well. Since then, there have been several periods with high or low productivity, which has been reflected in the fish landings a couple of years afterwards.

There has been observed a clear relationship, from primary production to the higher trophic levels (including fish and seabirds), in the Faroe shelf ecosystem, and all trophic levels seem to respond quickly to variability in primary production in the ecosystem (Gaard, E. et al. 2002). There is a positive relationship between primary production and the cod and haddock individual fish growth and recruitment $1 / 2-2$ years later. The primary production index has been below average since 2002 except for 2004 and 20082010 when it was above average (Figure 2.3). The estimate of primary production in 2014 will not be available until July. The primary production index could therefore be
a candidate ecosystem and stock indicator. Another potential indicator candidate is the so-called Sub-polar Gyre Index, which is an index for the primary production in the outer areas (Figure 2.3).

Recent work (Steingrund et al., 2012) shows that there is a moderate positive correlation between primary production on the Faroe Shelf and the subsequent production of cod (Steingrund and Gaard, 2005). There is also a moderate positive correlation for haddock and saithe. However, if all three species are combined, the positive correlation becomes very strong (Figure 2.4). This indicates that a nearly fixed portion of the energy produced by the primary production goes to predatory demersal fish on the Faroe Plateau, but that the portion to each of the fish stocks (to cod, haddock or saithe) may vary much between years. As an example, the last period of high productivity (20082010) did not lead to any marked increase in the stock size of cod/haddock, but only in saithe.

### 2.1.3 Summary of the 2015 assessment of Faroe Plateau cod, haddock and saithe

As mentioned in previous reports of this WG, landings of cod, haddock and saithe on the Faroes appear to be closely linked with the total biomass of the stocks. For cod, the exploitation ratio and fishing mortality have remained relatively stable over time, although they have been more fluctuating in recent years. For haddock, the exploitation rate was decreasing from the 1950s and 1960s, while it have been fluctuating since the mid 1970s. For saithe, there is a suggestion that the exploitation rate was increasing at the beginning of the period, it decreased from the early 1990s to 1998 and has increased close to the highest values observed in 2009. It has since declined again.

Another main feature of the plots of landings, biomasses, mortalities and recruitment is the apparent periodicity during the time series with cod and haddock showing almost the same fluctuations and time-trends.

### 2.1.4 Reference points for Faroese stocks

As explained elsewhere in this report, MSY reference points have recently been estimated for cod, haddock and saithe in addition to the already existing PA reference points. These reference points are all estimated based on single-species models. Multispecies models may give very different perception of $\mathrm{F}_{\text {MSY }}$ reference points than singlespecies models, and for the Faroe area this could be extra true, since there is a close relationship between the environment and the fish stocks and between fish stocks (see section 2.1.3). Adding the recruitment of cod and haddock and relating them to zooplankton concentration shows a strong negative correlation (Figure 2.5), but a potential causal relationship is unknown.

### 2.1.5 Management plan

In 2011 the Faroese minister of fisheries established a group of experts to formulate a management plan for cod, haddock and saithe including a harvest control rule and a recovery plan. The group consisted of scientists from the Faroe Marine Research Institute of one representative from the industry and 1 from the Ministry of Fisheries. The results of this work was delivered to the Minister of Fisheries in the autumn 2011 but the outcome has not been approved by the authorities so far and not been implemented. Basically, the plan builds on the MSY framework developed by ICES.

### 2.1.6 References:

Gaard. E., Hansen, B., Olsen, B and Reinert, J. 2001. Ecological features and recent trends in physical environment, plankton, fish stocks and sea birds in the Faroe plateau ecosystem. In: KSherman and H-R Skjoldal (eds). Changing states of the Large Marine Ecosystems of the North Atlantic.

Steingrund, P., and Gaard, E. 2005. Relationship between phytoplankton production and cod production on the Faroe Shelf. ICES Journal of Marine Science, 62: 163-176.

Steingrund, P., and Hátún, H. 2008. Relationship between the North Atlantic subpolar gyre and fluctuations of the saithe stock in Faroese waters. NWWG 2008 Working Document 20.

Steingrund, P., Gaard, E., Reinert, J., Olsen, B., Homrum, E., and Eliassen, K. 2012. Trophic relationships on the Faroe Shelf ecosystem and potential ecosystem states. In: Homrum, E., 2012. The effects of climate and ocean currents on Faroe Saithe. PhD-thesis, 2012.

Table 2.1. Number of allocated days since the fiscal year 1996/97.

| Fleeet | Smb. LI.: | Serlig viðm. | 1 ytri | 1 innaru | 2 ytri | 2 innari | 3 | 4 A | 4 B | 4 D | 4 T | 5 | \|at rida y yir) | Dagar tils. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1996/97 | (50 20/5-96) | ( $12 / 15 \mathrm{mdr}$ !) |  |  |  | 8225 | 3040 | 4700 | 3080 | 1540 |  | 22000 | 1000 | 43585 |
| 1996/97 | (84 6/6-97) | (12/15mdr!) |  |  |  | 8225 | 3040 | 5600 | 3410 | 1650 |  | 27000 | 660 | 49585 |
| 1997/98 | (133 9/8-97) | 12 mdr ! |  |  |  | 7199 | 2660 | 4696 | 4632 |  |  | 23625 | 577 | 43389 |
| 1998/99 | (69 18/8-98) |  |  |  |  | 6839 | 2527 | 4461 | 4400 |  |  | 22444 | 548 | 41219 |
| 1999/2000 | (8017/8-99) |  |  |  |  | 6839 | 2527 | 4461 | 4400 |  |  | 22444 | 548 | 41219 |
| 2000/2001 | (104 17/8-00) |  |  |  |  | 6839 | 2527 | 4461 | 4400 |  |  | 22,444 | 548 | 41219 |
| 2001/2002 | (115 15/8-01) |  |  |  |  | 6839 | 2527 | 4461 | 4400 |  |  | 22444 | 0 | 40671 |
| 2002/2003 | (76 13/8-02) |  |  |  |  | 6771 | 2502 | 4416 | 4356 |  |  | 22220 | 0 | 40265 |
| 2003/2004 | (100 8/8-03) |  |  |  |  | 6636 | 2452 | 4328 | 4269 |  |  | 21776 | 0 | 39461 |
| 2004/2005 | (49 18/8-04) |  |  |  |  | 6536 | 2415 | 4263 | 4205 |  |  | 21449 | 0 | 38868 |
| 2005/2006 | (98 19/8-05) |  |  |  |  | 5752 | 3578 | 1770 | 2067 |  | 1766 | 21235 | 0 | 36168 |
| 2006/2007 | (81 17/8-06) |  |  |  |  | 5752 | 3471 | 1717 | 2005 |  | 1713 | 20598 | 0 | 35256 |
| 2007/2008 | (80 20/8-07) |  |  |  |  | 5637 | 3402 | 1683 | 1965 |  | 1679 | 20186 | 0 | 34552 |
| 2008/2009 | (76 15/8-08) |  |  |  |  | 5073 | 3062 | 1515 | 1769 |  | 1511 | 18167 | 0 | 31097 |
| 2008/2009 | (62 25/5-09) |  |  |  |  | 4638 | 3095 | 1393 | 1848 |  | 1621 | 18167 | 0 | 30762 |
| 2009/2010 | (106 17/8-09 |  |  |  |  | 4406 | 2940 | 1323 | 1756 |  | 1540 | 17259 | 0 | 29224 |
| 2010/2011 | (87 18/8-10) |  | 1700 | 900 |  | 4274 | 2852 | 1323 | 1756 |  | 1540 | 13259 | 0 | 25004 |
| 2010/2011 | sama - |  | 1700 | 900 |  | 4274 | 2852 | 1323 | 1756 |  | 1540 | 13259 | 0 | 27604 |
| 2011/12 | $\begin{array}{\|l\|} \hline(105 ~ 18 / 8-11) \\ (112 \\ \hline \end{array}$ |  |  |  | 1530 | 4657 | 2567 | 1058 | 1405 |  | 1386 | 10607 |  | 23210 |
| 2012/13 | (89 17/8-12) |  |  |  | 1530 | 4626 | 2567 | 1011 | 1533 |  | 1386 | 10607 |  | 23260 |
| 2013/14 | (109 16/8-13) |  |  |  | 1530 | 4441 | 2387 | 1011 | 1533 |  | 1386 | 9865 |  | 22153 |
| 2014/15 | (L89-18/8-14) |  |  |  | 1530 | 4455 | 2387 | 1029 | 1530 |  | 1386 | 9865 |  | 22182 |

Table 2.2. Number of days allocated and the number actually used since the fiscal year 2010/2011

| Fleet segment | Allocated | Used | \% used | Allocated | Used | \%used | Allocated | Used | \% used | Allocated | Used | \% used | Allocated | Used | \% used |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | days | days | days | days | days | days | days | days | days | days | days | days | days | days | days |
|  | 2010/11 | 2010/11 | 2010/11 | 2011/12 | 2011/12 | 2011/12 | 2012/13 | 2012/13 | 012/13 | 2013/14 | pr. Dato |  | 2014/15 | pr. Dato |  |
| Reference: | L187 18/8-10(JV) |  |  | U105 18/8-11 og L112 2/9-11(JD) |  |  | (89 17/8-12) |  |  | L105 18/8-111 og L1112 2/9-11(JD) |  |  | (L89-18/8-14) |  |  |
| Group 1 - innaru leióir | 900 | 552.39 | 61\% |  |  |  |  |  |  |  |  |  |  |  |  |
| Group 1 - ytri leióir | 1700 | 785.3 | 46\% |  |  |  |  |  |  |  |  |  |  |  |  |
| Group 2 - (innaru leiói | 4274 | 3883.23 | 91\% | 4657 | 4758.02 | 102\% | 4626 | 3952.52 | 85\% | 4441 | 3915.82 | 88\% | 4455 | 1915.88 | 43\% |
| Group 2 - ytri leiôir |  |  |  | 1530 | 894.94 | 58\% | 1530 | 878.57 | 57\% | 1530 | 796.53 | 52\% | 1530 | 367.74 | 24\% |
| Group 3 | 2852 | 2071.16 | 73\% | 2567 | 1985.90 | 77\% | 2567 | 1205.23 | 47\% | 2387 | 1119.66 | 47\% | 2387 | 749.11 | 31\% |
| Group 4A | 1323 | 405.36 | 31\% | 1058 | 259.5 | 25\% | 1011 | 270.72 | 27\% | 1011 | 272.34 | 27\% | 1029 | 118.5 | 12\% |
| Group 4B | 1756 | 1015.65 | 58\% | 1405 | 656.61 | 47\% | 1533 | 687.73 | 45\% | 1533 | 518.77 | 34\% | 1530 | 230.77 | 15\% |
| Group 4T | 1540 | 1411.98 | 92\% | 1386 | 1313.14 | 95\% | 1386 | 1165.71 | 84\% | 1386 | 895.41 | 65\% | 1386 | 243.92 | 18\% |
| Group 5A | 5304 | 2856 | 54\% | 5060 | 1834 | 36\% | 4730 | 1410 | 30\% | 4311 | 998 | 23\% | 2640 | 1000 | 38\% |
| Group 5B | 7955 | 4525 | 57\% | 5547 | 3160 | 57\% | 5877 | 2845 | 48\% | 5554 | 2842 | 51\% | 7225 | 1000 | 14\% |
| Total | 27604 | 17506.07 | 63\% | 23210 | 14862.11 | 64\% | 23260 | 12415.48 | 53\% | 22153 | 11358.53 | 51\% | 22182 | 5625.92 | 25\% |

Table 2.3. Main regulatory measures by fleet in the Faroese fisheries in Vb. The fleet capacity is fixed, based on among other things no. of licenses. Number of licenses within each group (by May 2006) are as follows: 1 : 12; 2:29; 3:25; 4A: 25; 4B: 21; 4T: 19; 5A:140; 5B: 453; 6: 8 . These licenses have been fixed in 1997, but in group 5B a large number of additional licenses can be issued upon request.

| Fleet segment |  | Sub groups |  | Main regulation tools |
| :---: | :---: | :---: | :---: | :---: |
| 1 | $\begin{aligned} & \text { Single trawlers > } 400 \\ & \text { HP } \end{aligned}$ | $\begin{aligned} & \text { non } \\ & \mathrm{e} \end{aligned}$ |  | Fishing days, have from 2011/12 been merged with the <br> pair trawlers, area closures |
| 2 | Pair trawlers > 400 HP | $\begin{aligned} & \text { non } \\ & \mathrm{e} \end{aligned}$ |  | Fishing days, area closures |
| 3 | Longliners > 110 GRT | $\begin{aligned} & \text { non } \\ & \mathrm{e} \end{aligned}$ |  | Fishing days, area closures |
| 4 | Coastal vessels>15 GRT | 4A | Trawlers 15-40 GRT | Fishing days |
|  |  | 4A | Longliners 15-40 GRT | Fishing days |
|  |  | 4B | Longliners>40 GRT | Fishing days |
|  |  | 4 T | Trawlers>40 GRT | Fishing days |
| 5 | Coastal vessels <15 GRT | 5A | Full-time fishers | Fishing days |
|  |  | 5B | Part-time fishers | Fishing days |
| 6 | Others |  | Gillnetters | Bycatch limitations, fishing depth, no. of nets |
|  |  |  | Others | Bycatch limitations |



Figure 2.1. The 2012 distribution of fishing activities by some major fleets. The longline fleet below 15 GRT is not shown here since they are not obliged to keep logbooks.


Figure 2.2. Fishing area regulations in Division Vb . Allocation of fishing days applies to the area inside the outer thick line on the Faroe Plateau. Holders of effort quotas who fish outside this line can triple their numbers of days. Longliners larger than 110 GRT are not allowed to fish inside the inner thick line on the Faroe Plateau. If longliners change from longline to jigging, they can double their number of days. The Faroe Bank shallower than 200 m depths (a, aa) is regulated separate from the Faroe Plateau. It is closed to trawling and the longline fishery is regulated by individual day quotas.


Figure 2.3. Temporal development of the phytoplankton index over the Faroe Shelf area (< $\mathbf{1 3 0} \mathbf{~ m}$ ) and the subpolar gyre index which indicates productivity in deeper waters.


Figure 2.4. Relationship between primary production and production of cod, haddock and saithe.


Figure 2.5. Relationship between zooplankton concentration and recruitment of cod+haddock on the Faroe Plateau.

## 3 Faroe Bank Cod

## Summary

The total reported landings in 2014 were the lowest recorded since 1965 ( 30 tonnes).
The spring index suggests that the stock increased from 2012 to 2014 and declined substantially again in 2015. Nevertheless both the summer and spring index suggest that the stock is well below average and there is no indication of strong incoming year classes.

The results of an exploratory production model based on both surveys indicate a good agreement in the stock biomass index in recent years whereas the observed surveybased exploitation rates correlates reasonably well with estimated fishing mortalities. However the model failed to pick up the large increases in stock biomass observed in the 1996-2003 period. Correlation between modelled F's and summer survey based exploitation rates is $\mathrm{R}=0.90$. The exploitation ratio increased in 2011 as a consequence of the increase in landings and it decreased afterwards reflecting the fall of catches observed since 2012.

### 3.1 State of the stock

Total nominal catches of the Faroe Bank cod from 1987 to 2014 as officially reported to ICES are given in Table 3.1 and since 1965 in Figure 3.1. UK catches reported to be taken on the Faroe Bank are all assumed to be taken on the Faroe Plateau and are therefore not used in the assessment. Landings have been highly variable from 1965 to the mid-1980s, reflecting the opportunistic nature of the cod fishery on the Bank, with peak landings slightly exceeding 5 000t in 1973 and 2003. The trend of landings has been smoother since 1987, declining from about $3500 t$ in 1987 to only 330 t in 1992 before increasing to $3600 t$ in 1997. In 2013 landings were estimated at $36 t$ which is the lowest ever recorded since 1965 (Figure 3.1). Longline fishing effort increased substantially in 2003 and although it decreased in 2004 and 2005 the latter remains the second highest fishing effort observed since 1988 (Figure 3.1). From 2005 to 2007 the effort has been reduced substantially. In the 2010/2011 and 2011/2012 fishing years a total of 61 and 100 fishing-days were allocated to the Bank. No days have been allocated since 2012.

The Faroese groundfish surveys (spring and summer) cover the Faroe Bank and cod is mainly taken within the 200 m depth contour. The catches of cod per trawl hour in depths shallower than 200 meter are shown in Figure 3.2.

The spring survey was initiated in 1983 and discontinued in 1996, 2004 and 2005. The summer survey has been carried out since 1996. The CPUE of the spring survey was low during 1988 to 1995 varying between 73 and 95 kg per tow. Although noisy, the survey suggests higher, possibly increasing biomass during 1995-2003. The 2013 and the 2014 spring point estimates suggest that the stock increased and decreased sharply again in 2015. it is however well below the average of that of the period 1996-2002. The 2014 summer index is estimated at 25 kg per tow, which is the second lowest value in the series. There are conflicting signals between both indices from 2012 to 2014.. The agreement between the summer and spring index is good during 1996 to 2001 and since 2006, but they diverged in the 2002-2003 and 2012-2014 periods.. Both indexes have remained well below average since 2004.

The figure of length distributions (figures 3.3 and 3.5) show in general good recruitment of 1 year old in the summer survey from 2000-2002 (lengths $26-45 \mathrm{~cm}$ ), corresponding to good recruitment of 2 years old in the spring surveys from 2001 to 2003 $(40-60 \mathrm{~cm})$. The spring index shows poor recruitment from 2006 to 2015 reflecting the weak year classes observed in the summer survey since 2004. Age-disaggregated indices confirm the pattern observed in the length composition (figure 3.4 and figure 3.6)

A way to estimate recruitment strength is by simply counting the number of fish in length groups in the surveys. In the spring index, recruitment was estimated as total number of fish below 60 cm (2-year old) and in the summer index as number of fish below 45 cm (1-year old). According to the summer index the recruitment of 1 year old was good from 2000 to 2003, while the recruitment has been relatively poor since 2004 (Figure 3.7) The spring recruitment index in 2014 shows no sign of incoming year classes. Correlation between the spring and summer survey recruitment indices is fairly good ( $\mathrm{r}=0.85$ ). Correlation between numbers of 1-year and 2-years old cod in the agedisaggregated summer and spring surveys respectively is estimated at $\mathrm{r}=0.79$.

The group tried the ASPIC (Prager 1992) stock production model for the stock. The model requires catch data and corresponding effort or CPUE data that are reasonable indices of the stock biomass.

ASPIC requires starting guesses for $r$, the intrinsic rate of increase, MSY, B1/Bmsy ratio and $q$, catchability coefficients. No sensitivity analysis was performed to explore the stability of parameter estimation.

The program was run with the time-series from 1983-2014 including spring survey and 1996-2014 summer CPUE's separately. The result of the runs are presented in tables 3.2 and 3.3 For both runs the model seemed to follow reasonably well survey trends in periods of low stock abundances but it failed to pick up the large increases observed in the 1996-2003 period (figures 3.8 and 3.9).

However estimates of $r=0.07$ and $F m s y=0.035$ (using the fall survey series) seem spurious given that the Faroe Bank cod is the fastest growing cod stock in the Atlantic.

The ratio of landings to the survey indices provides an exploitation ratio, which can be used as a proxy to relative changes in fishing mortality. For the summer survey, the results suggest that fishing mortality has been reasonably stable during 1996 to 2002, but that it increased steeply in 2003, consistent with the $160 \%$ increase in longline fishing days in that year (Figure 3.1). The exploitation ratio has decreased since 2006 but increased in 2011 due to the increase in catches and decreased again afterwards reflecting the fall of catches observed since 2011.

### 3.2 Comparison with previous assessment and forecast

The status of the stock remains almost unchanged with respect to last year's assessment. Both the spring and the summer indexes suggest the stock is well below average while there are no indications of incoming recruitment. The spring index suggests an increasing stock biomass from 2012 to 2014 which it is however not picked up by the summer survey. The exploratory production model performed since 2013 confirms the poor status of the stock.

### 3.3 Management plans and evaluations

None

### 3.4 Management considerations

The landing estimates are uncertain because since 1996 vessels are allowed to fish both on the Plateau and on Faroe Bank during the same trip, rendering landings from both areas uncertain. Given the relative size of the two fisheries, this is a bigger problem for Faroe Bank cod than for Faroe Plateau cod, but the magnitude remains unquantified for both. The ability to provide advice depends on the reliability of input data. If the cod landings from Faroe Bank are not known, it is difficult to provide advice. If the fishery management agency intends to manage the two fisheries to protect the productive capacity of each individual unit, then it is necessary to identify the catch removed from each stock. Simple measures should make it possible to identify if the catch is originating from the Bank or from the Plateau e.g. by storing in different section of the hold and/or by tagging of the different boxes.

Consistent with the advice given in 2014 the WG suggests the closure of the fishery until the recovery of the stock is confirmed. The reopening of the fishery should not be considered until both surveys indicate a biomass at or above the average that of the period 1996-2002.

### 3.5 Regulations and their effects

In 1990, the decreasing trends in cod landings from Faroe Bank lead ACFM to advise the Faroese authorities to close the bank to all fishing. This advice was followed for depths shallower than 200 meters. In 1992 and 1993 longliners and jiggers were allowed to participate in an experimental fishery inside the 200 meters depth contour. For the quota year 1 September 1995 to 31 August 1996 a fixed quota of 1050 t was set. The new management regime with fishing days was introduced on 1 June 1996 allowing longliners and jiggers to fish inside the 200 m contour. The trawlers are allowed to fish outside the 200 m contour.

A total fishing ban during the spawning period ( 1 March to 1 May) has been enforced since 2005. In 2009, fishing was restricted to all fishing gears from 1 January to 31 August. However, in the 2010/2011 and 2011/2012 fishing years a total of 61 and 100 fish-ing-days were allocated to the Bank to jiggers in the shallow waters of the Bank. No days have been allocated since 2012.

Table 3.1. Faroe Bank (sub-division Vb2) cod. Nominal catches (tonnes) by countries 1986-2014 as officially reported to ICES. From 1992 the catches by Faroe Islands and Norway are used in the assessment.

|  | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | 1836 | 3409 | 2966 | 1270 | 289 | 297 | 122 | 264 | 717 | 561 | 2051 | 3459 | 3092 | 1001 |  |
| Norway | 6 | 23 | 94 | 128 | 72 | 38 | 32 | 2 | 8 | 40 | 55 | 135 | 147 | 88 |  |
| UK (E/N/NI) | - | - | - | - | $2^{2}$ | $1^{2}$ | $74^{2}$ | $186^{2}$ | $56{ }^{2}$ | $43^{2}$ | $126{ }^{3}$ | $61{ }^{3}$ | $27^{3}$ | - |  |
| UK (Scotland) | $63{ }^{3}$ | $47{ }^{3}$ | $37{ }^{3}$ | $14^{3}$ | $205{ }^{3}$ | $90{ }^{3}$ | $176{ }^{3}$ | $118{ }^{3}$ | $227{ }^{3}$ | $551{ }^{3}$ | $382{ }^{3}$ | $277{ }^{3}$ | $265{ }^{3}$ | $51{ }^{3}$ |  |
| Total | 1905 | 3479 | 3097 | 1412 | 568 | 426 | 404 | 570 | 1008 | 1195 | 2614 | 3932 | 3531 | $210{ }^{3}$ |  |
| Used in assessment |  |  |  |  | 289 | 297 | 154 | 266 | 725 | 601 | 2106 | 3594 | 32390 | 1350 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1089 |  |
|  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |
| Faroe Islands |  | 1094 | 1840 | 5957 | 3607 | 1270 | 1005 | 471 | 231 | 81 | 111 | 393 | 115 | 40 | 32 |
| Norway | 49 | 51 | 25 | 72 | 18 | 37 | 10 | 7 | 1 | 4 | 1 |  | 0 |  |  |
| Greenland | - | - | - | - | - | - | - | - | - | - | 5 |  | 1 |  |  |
| UK (E/W/NI) | $18{ }^{3}$ | $50^{3}$ | $42{ }^{3}$ | $15^{3}$ | $15^{3}$ | $24^{3}$ | $1{ }^{3}$ |  |  |  |  |  |  |  |  |
| UK (Scotland) | $245{ }^{3}$ | $288{ }^{3}$ | $218{ }^{3}$ | $254{ }^{3}$ | $244{ }^{3}$ | $1129^{3}$ | $278{ }^{3}$ | 53 | 32 | 38 | 54 |  |  |  | 270 |
| Total | 312 | 1483 | 2125 | 6298 | 3884 | 2460 | 1294 | 531 | 264 | 123 | 171 | 393 | 116 | 40 | 302 |
| Correction of Faroese catches in Vb 2 |  | -65 | -109 | -353 | -214 | -75 | -60 | -28 | -14 | -5 | -7 | -23 | -7 | -2 | -2 |
| Used in assessment | 1194 | 1080 | 1756 | 5676 | 3411 | 1232 | 955 | 450 | 218 | 80 | 105 | 370 | 108 | 38 | 30 |

Table 3.2. Faroe Bank (sub-division Vb2) cod. Surplus production model output using the summer index.

Faroe Bank Cod RV
Page 1
14 Apr 2015 at 12:00.44
ASPIC -- A Surplus-Production Model Including Covariates (Ver. 3.82)
FIT Mode

Author: Michael H. Prager; NOAA/NMFS/S.E. Fisheries Science Center 101 Pivers Island Road; Beaufort, North Carolina 28516 USA

## ASPIC User's Manual

 is available gratisfrom the author.
Ref: Prager, M. H. 1994. A suite of extensions to a nonequilibrium surplus-production model. Fishery Bulletin 92: 374-389.

## CONTROL PARAMETERS USED (FROM INPUT FILE)

| Number of years analyzed: | 50 | Number of bootstrap trials: | 0 |
| :---: | :---: | :---: | :---: |
| Number of data series: | 1 , | Lower bound on MSY: | $5.000 \mathrm{E}+02$ |
| Objective function computed: | in effort | Upper bound on MSY: | $1.000 \mathrm{E}+09$ |
| Relative conv. criterion (simplex): | $1.000 \mathrm{E}-08$ | Lower bound on r : | $7.000 \mathrm{E}-02$ |
| Relative conv. criterion (restart): | $3.000 \mathrm{E}-08$ | Upper bound on r : | $2.500 \mathrm{E}+00$ |
| Relative conv. criterion (effort): | $1.000 \mathrm{E}-04$ | Random number seed: | 2010417 |
| Maximum F allowed in fitting: | 8.000 | Monte Carlo search mode, trials: | 110000 |

## PROGRAM STATUS INFORMATION (NON-BOOTSTRAPPED ANALYSIS)

 20ERROR: Estimate of $r$ is at or near minimum constraint, 7.000E-02
Solution may be trivial--examine carefully.


## MANAGEMENT PARAMETER ESTIMATES (NON-BOOTSTRAPPED)


........ Fishing effort at MSY in units of each fishery:
fmsy( 1) Survey CPUE Summer $\quad 2.045 \mathrm{E}+00 \quad \mathrm{r} / 2 q(1) \quad \mathrm{f}(0.1)=1.840 \mathrm{E}+00$

## ESTIMATED POPULATION TRAJECTORY (NON-BOOTSTRAPPED)



| 38 | 2002 | 0.139 | $1.316 \mathrm{E}+04$ | $1.267 \mathrm{E}+04$ | $1.756 \mathrm{E}+03$ | $1.756 \mathrm{E}+03$ | $7.759 \mathrm{E}+02$ | $3.961 \mathrm{E}+00$ | $2.595 \mathrm{E}-01$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 39 | 2003 | 0.603 | $1.218 \mathrm{E}+04$ | $9.414 \mathrm{E}+03$ | $5.676 \mathrm{E}+03$ | $5.676 \mathrm{E}+03$ | $5.964 \mathrm{E}+02$ | $1.723 \mathrm{E}+01$ | $2.402 \mathrm{E}-01$ |
| 40 | 2004 | 0.628 | $7.104 \mathrm{E}+03$ | $5.435 \mathrm{E}+03$ | $3.411 \mathrm{E}+03$ | $3.411 \mathrm{E}+03$ | $3.595 \mathrm{E}+02$ | $1.793 \mathrm{E}+01$ | $1.401 \mathrm{E}-01$ |
| 41 | 2005 | 0.349 | $4.052 \mathrm{E}+03$ | $3.532 \mathrm{E}+03$ | $1.232 \mathrm{E}+03$ | $1.232 \mathrm{E}+03$ | $2.386 \mathrm{E}+02$ | $9.965 \mathrm{E}+00$ | $7.990 \mathrm{E}-02$ |
| 42 | 2006 | 0.360 | $3.059 \mathrm{E}+03$ | $2.653 \mathrm{E}+03$ | $9.550 \mathrm{E}+02$ | $9.550 \mathrm{E}+02$ | $1.808 \mathrm{E}+02$ | $1.029 \mathrm{E}+01$ | $6.031 \mathrm{E}-02$ |
| 43 | 2007 | 0.211 | $2.285 \mathrm{E}+03$ | $2.129 \mathrm{E}+03$ | $4.500 \mathrm{E}+02$ | $4.500 \mathrm{E}+02$ | $1.459 \mathrm{E}+02$ | $6.039 \mathrm{E}+00$ | $4.505 \mathrm{E}-02$ |
| 44 | 2008 | 0.112 | $1.981 \mathrm{E}+03$ | $1.938 \mathrm{E}+03$ | $2.180 \mathrm{E}+02$ | $2.180 \mathrm{E}+02$ | $1.331 \mathrm{E}+02$ | $3.214 \mathrm{E}+00$ | $3.905 \mathrm{E}-02$ |
| 45 | 2009 | 0.042 | $1.896 \mathrm{E}+03$ | $1.922 \mathrm{E}+03$ | $8.000 \mathrm{E}+01$ | $8.000 \mathrm{E}+01$ | $1.320 \mathrm{E}+02$ | $1.189 \mathrm{E}+00$ | $3.738 \mathrm{E}-02$ |
| 46 | 2010 | 0.054 | $1.948 \mathrm{E}+03$ | $1.963 \mathrm{E}+03$ | $1.050 \mathrm{E}+02$ | $1.050 \mathrm{E}+02$ | $1.347 \mathrm{E}+02$ | $1.529 \mathrm{E}+00$ | $3.840 \mathrm{E}-02$ |
| 47 | 2011 | 0.200 | $1.977 \mathrm{E}+03$ | $1.853 \mathrm{E}+03$ | $3.700 \mathrm{E}+02$ | $3.700 \mathrm{E}+02$ | $1.274 \mathrm{E}+02$ | $5.704 \mathrm{E}+00$ | $3.899 \mathrm{E}-02$ |
| 48 | 2012 | 0.062 | $1.735 \mathrm{E}+03$ | $1.741 \mathrm{E}+03$ | $1.080 \mathrm{E}+02$ | $1.080 \mathrm{E}+02$ | $1.198 \mathrm{E}+02$ | $1.773 \mathrm{E}+00$ | $3.420 \mathrm{E}-02$ |
| 49 | 2013 | 0.021 | $1.747 \mathrm{E}+03$ | $1.789 \mathrm{E}+03$ | $3.800 \mathrm{E}+01$ | $3.800 \mathrm{E}+01$ | $1.230 \mathrm{E}+02$ | $6.070 \mathrm{E}-01$ | $3.444 \mathrm{E}-02$ |
| 50 | 2014 | 0.016 | $1.832 \mathrm{E}+03$ | $1.881 \mathrm{E}+03$ | $3.000 \mathrm{E}+01$ | $3.000 \mathrm{E}+01$ | $1.292 \mathrm{E}+02$ | $4.558 \mathrm{E}-01$ | $3.611 \mathrm{E}-02$ |
| 51 | 2015 | $1.931 \mathrm{E}+03$ |  |  |  | $3.807 \mathrm{E}-02$ |  |  |  |



| 38 | 2002 | $3.472 \mathrm{E}+02$ | $2.168 \mathrm{E}+02$ | 0.1386 | $1.756 \mathrm{E}+03$ | $1.756 \mathrm{E}+03$ | -0.47080 | $0.000 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 39 | 2003 | $1.618 \mathrm{E}+02$ | $1.611 \mathrm{E}+02$ | 0.6029 | $5.676 \mathrm{E}+03$ | $5.676 \mathrm{E}+03$ | -0.00396 | $0.000 \mathrm{E}+00$ |
| 40 | 2004 | $7.304 \mathrm{E}+01$ | $9.303 \mathrm{E}+01$ | 0.6276 | $3.411 \mathrm{E}+03$ | $3.411 \mathrm{E}+03$ | 0.24190 | $0.000 \mathrm{E}+00$ |
| 41 | 2005 | $6.188 \mathrm{E}+01$ | $6.046 \mathrm{E}+01$ | 0.3488 | $1.232 \mathrm{E}+03$ | $1.232 \mathrm{E}+03$ | -0.02321 | $0.000 \mathrm{E}+00$ |
| 42 | 2006 | $2.927 \mathrm{E}+01$ | $4.541 \mathrm{E}+01$ | 0.3600 | $9.550 \mathrm{E}+02$ | $9.550 \mathrm{E}+02$ | 0.43905 | $0.000 \mathrm{E}+00$ |
| 43 | 2007 | $3.331 \mathrm{E}+01$ | $3.644 \mathrm{E}+01$ | 0.2114 | $4.500 \mathrm{E}+02$ | $4.500 \mathrm{E}+02$ | 0.08977 | $0.000 \mathrm{E}+00$ |
| 44 | 2008 | $3.117 \mathrm{E}+01$ | $3.317 \mathrm{E}+01$ | 0.1125 | $2.180 \mathrm{E}+02$ | $2.180 \mathrm{E}+02$ | 0.06204 | $0.000 \mathrm{E}+00$ |
| 45 | 2009 | $4.927 \mathrm{E}+01$ | $3.289 \mathrm{E}+01$ | 0.0416 | $8.000 \mathrm{E}+01$ | $8.000 \mathrm{E}+01$ | -0.40409 | $0.000 \mathrm{E}+00$ |
| 46 | 2010 | $4.164 \mathrm{E}+01$ | $3.359 \mathrm{E}+01$ | 0.0535 | $1.050 \mathrm{E}+02$ | $1.050 \mathrm{E}+02$ | -0.21484 | $0.000 \mathrm{E}+00$ |
| 47 | 2011 | $5.854 \mathrm{E}+01$ | $3.172 \mathrm{E}+01$ | 0.1996 | $3.700 \mathrm{E}+02$ | $3.700 \mathrm{E}+02$ | -0.61266 | $0.000 \mathrm{E}+00$ |
| 48 | 2012 | $3.425 \mathrm{E}+01$ | $2.979 \mathrm{E}+01$ | 0.0620 | $1.080 \mathrm{E}+02$ | $1.080 \mathrm{E}+02$ | -0.13949 | $0.000 \mathrm{E}+00$ |
| 49 | 2013 | $1.737 \mathrm{E}+01$ | $3.062 \mathrm{E}+01$ | 0.0212 | $3.800 \mathrm{E}+01$ | $3.800 \mathrm{E}+01$ | 0.56678 | $0.000 \mathrm{E}+00$ |
| 50 | 2014 | $2.575 \mathrm{E}+01$ | $3.219 \mathrm{E}+01$ | 0.0160 | $3.000 \mathrm{E}+01$ | $3.000 \mathrm{E}+01$ | 0.22324 | $0.000 \mathrm{E}+00$ |

[^0]Page 4

UNWEIGHTED LOG RESIDUAL PLOT FOR DATA SERIES \# 1

| -1 | -0.75 | -0.5 | -0.25 | 0 | 0.25 | 0.5 | 0.75 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |




Observed (O) and Estimated (*) CPUE for Data Series \# 1 -- Survey CPUE Summer
900.
.

600. -: **

```
\begin{tabular}{cc}
\(:\) & \(* *\) \\
\(:\) & \(* *\)
\end{tabular}
450. -: \(\quad * * * *\)
```








```
    150. -:
                            O 2
                                    O
                                    *
                                    O 2* O O
                                    O 22*2*222
0. -:
1959. 1965. 1971. 1977. 1983. 1989. 1995. 2001. 2007. 2013. 2019.
Time Plot of Estimated F-Ratio and B-Ratio
```

18. -:
F
F
```
15. -:
:
:
:
12. -:
: FF
9. -:
    :
    :
6. -: FF F F
    : F F FF
    : F
F
3. -: F F F
: F F FFFF F FF
    : F FFFFF F F F
    : -- --- - 22 - -- --- -- 2-2 2 2- - -- --- --- --2 --- --- -- --- - -2 - -- - -- --- --
        B BB B BB B BB B BB B BB B BB B BB B BB B 22 B 2B 2 BB B
        F F
0. -:
    F BB B BB B BB B BB B BB B BB
1959. 1965. 1971. 1977. 1983. 1989. 1995. 2001. 2007. 2013. 2019.
```

Table 3.3. Faroe Bank (sub-division Vb2) cod. Surplus production model output using the spring index.


Normal convergence.

GOODNESS-OF-FIT AND WEIGHTING FOR NON-BOOTSTRAPPED ANALYSIS

| Weighted | Weighted | Current | Suggested | R -squar |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Loss component number and title | SSE N | MSE | weight | weight | in CPUE |

Loss(-1) SSE in yield
Loss( 0) Penalty for B1R > 2
Loss( 1) Survey CPUE Spring
TOTAL OBJECTIVE FUNCTION:
$0.000 \mathrm{E}+00$
$0.000 \mathrm{E}+00 \quad 1 \quad$ N/A $1.000 \mathrm{E}-01 \quad$ N/A
$1.895 \mathrm{E}+01 \quad 29 \quad 7.020 \mathrm{E}-01 \quad 1.000 \mathrm{E}+00 \quad 1.000 \mathrm{E}+00 \quad 0.131$ $1.89548543 \mathrm{E}+01$

Number of restarts required for convergence: 18
Est. B-ratio coverage index ( 0 worst, 2 best): 0.6336 < These two measures are defined in Prager
Est. B-ratio nearness index ( 0 worst, 1 best): $0.7091<$ et al. (1996), Trans. A.F.S. 125:729

MODEL PARAMETER ESTIMATES (NON-BOOTSTRAPPED)


| MSY | Maximum sustainable yield | $2.931 \mathrm{E}+03$ |  | $3.000 \mathrm{E}+03$ | 1 | 1 |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| r | Intrinsic rate of increase | $3.993 \mathrm{E}-01$ | $8.000 \mathrm{E}-01$ | 1 | 1 |  |  |
| ....... | Catchability coefficients by fishery: |  |  |  |  |  |  |
| q (1) | Survey CPUE Spring | $3.065 \mathrm{E}-02$ | $1.000 \mathrm{E}-02$ | 1 | 1 |  |  |

MANAGEMENT PARAMETER ESTIMATES (NON-BOOTSTRAPPED)

........ Fishing effort at MSY in units of each fishery:
fmsy( 1) Survey CPUE Spring
$6.514 \mathrm{E}+00$
r/2q(1)
$\mathrm{f}(0.1)=5.863 \mathrm{E}+00$

Faroe Bank Cod RV
Page 2 ESTIMATED POPULATION TRAJECTORY (NON-BOOTSTRAPPED)

|  | Estin Year | mated E total | Estimated <br> starting | Estimated average tot | Observed total | Model Estimated surplus F mort |  | Ratio of biomass |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Obs | or ID | F mort | t biomass | biomass | yield yi | yield produc | , | to Bmsy |  |
| 1 | 1965 | 0.269 | $8.662 \mathrm{E}+03$ | $8.716 \mathrm{E}+03$ | $2.341 \mathrm{E}+03$ | $2.341 \mathrm{E}+03$ | $2.447 \mathrm{E}+03$ | $1.345 \mathrm{E}+00$ | $5.900 \mathrm{E}-01$ |
| 2 | 1966 | 0.211 | 8.7 | 9.0 | $1.909 \mathrm{E}+03$ | 1. | $2.502 \mathrm{E}+03$ | $1.054 \mathrm{E}+00$ | $5.972 \mathrm{E}-01$ |
| 3 | 1967 | 0.159 | 9.36 | 9.8 | $1.569 \mathrm{E}+03$ | 1.5 | $2.618 \mathrm{E}+03$ | 7.947E-01 | $6.376 \mathrm{E}-01$ |
| 4 | 1968 | 0.397 | 1.041 | $9.746 \mathrm{E}+03$ | $3.871 \mathrm{E}+03$ | $3.871 \mathrm{E}+03$ | $2.598 \mathrm{E}+03$ | $1.989 \mathrm{E}+00$ | 7.091E-01 |
| 5 | 1969 | 0.268 | $9.138 \mathrm{E}+03$ | 9. | 2.4 | $2.457 \mathrm{E}+03$ | $2.518 \mathrm{E}+03$ | 0 | $6.224 \mathrm{E}-01$ |
| 6 | 1970 | 0.336 | $9.199 \mathrm{E}+03$ | $8.931 \mathrm{E}+03$ | $3.002 \mathrm{E}+03$ | $3.002 \mathrm{E}+03$ | $2.481 \mathrm{E}+03$ | $684 \mathrm{E}+00$ | $6.266 \mathrm{E}-01$ |
| 7 | 1971 | 0.234 | $8.678 \mathrm{E}+03$ | $8.877 \mathrm{E}+03$ | $2.079 \mathrm{E}+03$ | $2.079 \mathrm{E}+03$ | $2.473 \mathrm{E}+03$ | 0 | $5.911 \mathrm{E}-01$ |
| 8 | 1972 | 0.234 | 9.0 | 9. | $2.168 \mathrm{E}+03$ | $2.168 \mathrm{E}+03$ | $2.531 \mathrm{E}+03$ | 0 | $6.179 \mathrm{E}-01$ |
| 9 | 197 | 0.643 | 9. | 7.9 | 5. | $5.101 \mathrm{E}+03$ | $2.303 \mathrm{E}+03$ | 0 | 1 |
| 10 | 1974 | 0.312 | $6.636 \mathrm{E}+03$ | 6. | $2.068 \mathrm{E}+03$ | 2.068E+03 | 3 | 0 | 1 |
| 11 | 1975 | 0.307 | $6.618 \mathrm{E}+03$ | 6.6 | $2.036 \mathrm{E}+03$ | - $2.036 \mathrm{E}+03$ | $2.048 \mathrm{E}+03$ | 0 | 01 |
| 12 | 1976 | 0.347 | 6.630 E | 6.5 | $2.258 \mathrm{E}+03$ | - $2.258 \mathrm{E}+03$ | $2.023 \mathrm{E}+03$ | 0 | 01 |
| 13 | 1977 | 0.138 | $6.395 \mathrm{E}+03$ | $6.969 \mathrm{E}+03$ | $9.590 \mathrm{E}+02$ | 9.590E+02 | $2.121 \mathrm{E}+03$ | $6.892 \mathrm{E}-01$ | $4.356 \mathrm{E}-01$ |
| 14 | 1978 | 0.700 | $7.557 \mathrm{E}+03$ | $6.252 \mathrm{E}+03$ | $4.379 \mathrm{E}+03$ | 4.379E+03 | $1.959 \mathrm{E}+03$ | $3.508 \mathrm{E}+00$ | $5.148 \mathrm{E}-01$ |
| 15 | 1979 | 0.244 | $5.137 \mathrm{E}+03$ | 5.358 | $1.306 \mathrm{E}+03$ | 1.306E+03 | $1.749 \mathrm{E}+03$ | $221 \mathrm{E}+00$ | 01 |
| 16 | 1980 | 0.203 | $5.580 \mathrm{E}+03$ | $5.920 \mathrm{E}+03$ | $1.203 \mathrm{E}+03$ | 1.203E+03 | $1.887 \mathrm{E}+03$ | $1.018 \mathrm{E}+00$ | $3.801 \mathrm{E}-01$ |
| 17 | 1981 | 0.184 | $6.264 \mathrm{E}+03$ | $6.676 \mathrm{E}+03$ | $1.229 \mathrm{E}+03$ | 1.229E+03 | $2.059 \mathrm{E}+03$ | $9.219 \mathrm{E}-01$ | $4.266 \mathrm{E}-01$ |
| 18 | 1982 | 0.309 | $7.094 \mathrm{E}+03$ | $7.074 \mathrm{E}+03$ | $2.184 \mathrm{E}+03$ | 2.184E+03 | $2.144 \mathrm{E}+03$ | $1.546 \mathrm{E}+00$ | $4.832 \mathrm{E}-01$ |
| 19 | 1983 | 0.328 | $7.054 \mathrm{E}+03$ | $6.972 \mathrm{E}+03$ | $2.284 \mathrm{E}+03$ | -2.284E+03 | $2.123 \mathrm{E}+03$ | $1.641 \mathrm{E}+00$ | $4.805 \mathrm{E}-01$ |
| 20 | 1984 | 0.320 | $6.893 \mathrm{E}+03$ | $6.846 \mathrm{E}+03$ | $2.189 \mathrm{E}+03$ | 2.189E+03 | $2.096 \mathrm{E}+03$ | $1.601 \mathrm{E}+00$ | $4.695 \mathrm{E}-01$ |
| 21 | 1985 | 0.461 | $6.801 \mathrm{E}+03$ | $6.315 \mathrm{E}+03$ | $2.913 \mathrm{E}+03$ | 2.913E+03 | $1.978 \mathrm{E}+03$ | $2.310 \mathrm{E}+00$ | $4.632 \mathrm{E}-01$ |
| 22 | 1986 | 0.312 | $5.866 \mathrm{E}+03$ | $5.888 \mathrm{E}+03$ | $1.836 \mathrm{E}+03$ | 1.836E+03 | $1.880 \mathrm{E}+03$ | $1.562 \mathrm{E}+00$ | $3.996 \mathrm{E}-01$ |
| 23 | 1987 | 0.687 | $5.910 \mathrm{E}+03$ | $4.965 \mathrm{E}+03$ | $3.409 \mathrm{E}+03$ | $3.409 \mathrm{E}+03$ | $1.644 \mathrm{E}+03$ | $3.438 \mathrm{E}+00$ | $4.025 \mathrm{E}-01$ |
| 24 | 1988 | 0.951 | $4.145 \mathrm{E}+03$ | $3.118 \mathrm{E}+03$ | $2.966 \mathrm{E}+03$ | 2.966E+03 | $1.109 \mathrm{E}+03$ | $4.764 \mathrm{E}+00$ | $2.823 \mathrm{E}-01$ |
| 25 | 1989 | 0.630 | $2.288 \mathrm{E}+03$ | $2.015 \mathrm{E}+03$ | $1.270 \mathrm{E}+03$ | 1.270E+03 | $7.492 \mathrm{E}+02$ | $3.156 \mathrm{E}+00$ | $1.558 \mathrm{E}-01$ |
| 26 | 1990 | 0.146 | $1.767 \mathrm{E}+03$ | $1.984 \mathrm{E}+03$ | $2.890 \mathrm{E}+02$ | $2.890 \mathrm{E}+02$ | $7.387 \mathrm{E}+02$ | 7.294E-01 | $1.204 \mathrm{E}-01$ |
| 27 | 1991 | 0.118 | $2.217 \mathrm{E}+03$ | $2.516 \mathrm{E}+03$ | $2.970 \mathrm{E}+02$ | $2.970 \mathrm{E}+02$ | $9.183 \mathrm{E}+02$ | $5.911 \mathrm{E}-01$ | $1.510 \mathrm{E}-01$ |
| 28 | 1992 | 0.046 | $2.838 \mathrm{E}+03$ | $3.327 \mathrm{E}+03$ | $1.540 \mathrm{E}+02$ | $1.540 \mathrm{E}+02$ | $1.177 \mathrm{E}+03$ | $2.318 \mathrm{E}-01$ | $1.933 \mathrm{E}-01$ |
| 29 | 1993 | 0.060 | $3.861 \mathrm{E}+03$ | $4.460 \mathrm{E}+03$ | $2.660 \mathrm{E}+02$ | $2.660 \mathrm{E}+02$ | $1.509 \mathrm{E}+03$ | $2.987 \mathrm{E}-01$ | $2.630 \mathrm{E}-01$ |
| 30 | 1994 | 0.129 | $5.104 \mathrm{E}+03$ | $5.640 \mathrm{E}+03$ | $7.250 \mathrm{E}+02$ | 7.250E+02 | $1.818 \mathrm{E}+03$ | $6.438 \mathrm{E}-01$ | $3.476 \mathrm{E}-01$ |
| 31 | 1995 | 0.087 | $6.197 \mathrm{E}+03$ | $6.937 \mathrm{E}+03$ | $6.010 \mathrm{E}+02$ | 6.010E+02 | $2.113 \mathrm{E}+03$ | $4.339 \mathrm{E}-01$ | $4.221 \mathrm{E}-01$ |
| 32 | 1996 | 0.270 | $7.709 \mathrm{E}+03$ | $7.801 \mathrm{E}+03$ | $2.106 \mathrm{E}+03$ | 2.106E+03 | $2.288 \mathrm{E}+03$ | $1.352 \mathrm{E}+00$ | $5.251 \mathrm{E}-01$ |
| 33 | 1997 | 0.504 | $7.891 \mathrm{E}+03$ | $7.135 \mathrm{E}+03$ | $3.594 \mathrm{E}+03$ | 3.594E+03 | $2.155 \mathrm{E}+03$ | $2.523 \mathrm{E}+00$ | $5.375 \mathrm{E}-01$ |
| 34 | 1998 | 0.567 | $6.451 \mathrm{E}+03$ | $5.712 \mathrm{E}+03$ | $3.239 \mathrm{E}+03$ | 3.239E+03 | $1.835 \mathrm{E}+03$ | $2.840 \mathrm{E}+00$ | 4.394E-01 |
| 35 | 1999 | 0.185 | $5.047 \mathrm{E}+03$ | $5.425 \mathrm{E}+03$ | $1.001 \mathrm{E}+03$ | 1.001E+03 | $1.766 \mathrm{E}+03$ | $9.240 \mathrm{E}-01$ | $3.438 \mathrm{E}-01$ |
| 36 | 2000 | 0.193 | $5.812 \mathrm{E}+03$ | $6.187 \mathrm{E}+03$ | $1.194 \mathrm{E}+03$ | 1.194E+03 | $1.950 \mathrm{E}+03$ | $9.665 \mathrm{E}-01$ | $3.959 \mathrm{E}-01$ |
| 37 | 2001 | 0.152 | $6.568 \mathrm{E}+03$ | $7.097 \mathrm{E}+03$ | $1.080 \mathrm{E}+03$ | 1.080E+03 | $2.148 \mathrm{E}+03$ | $7.622 \mathrm{E}-01$ | $4.473 \mathrm{E}-01$ |
| 38 | 2002 | 0.222 | $7.635 \mathrm{E}+03$ | $7.913 \mathrm{E}+03$ | $1.756 \mathrm{E}+03$ | 1.756E+03 | $2.308 \mathrm{E}+03$ | $1.111 \mathrm{E}+00$ | $5.201 \mathrm{E}-01$ |


| 39 | 2003 | 0.932 | $8.187 \mathrm{E}+03$ | $6.087 \mathrm{E}+03$ | $5.676 \mathrm{E}+03$ | $5.676 \mathrm{E}+03$ | $1.911 \mathrm{E}+03$ | $4.670 \mathrm{E}+00$ | $5.577 \mathrm{E}-01$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 40 | 2004 | 1.092 | $4.422 \mathrm{E}+03$ | $3.123 \mathrm{E}+03$ | $3.411 \mathrm{E}+03$ | $3.411 \mathrm{E}+03$ | $1.109 \mathrm{E}+03$ | $5.469 \mathrm{E}+00$ | $3.012 \mathrm{E}-01$ |
| 41 | 2005 | 0.672 | $2.120 \mathrm{E}+03$ | $1.832 \mathrm{E}+03$ | $1.232 \mathrm{E}+03$ | $1.232 \mathrm{E}+03$ | $6.857 \mathrm{E}+02$ | $3.367 \mathrm{E}+00$ | $1.444 \mathrm{E}-01$ |
| 42 | 2006 | 0.714 | $1.574 \mathrm{E}+03$ | $1.338 \mathrm{E}+03$ | $9.550 \mathrm{E}+02$ | $9.550 \mathrm{E}+02$ | $5.099 \mathrm{E}+02$ | $3.574 \mathrm{E}+00$ | $1.072 \mathrm{E}-01$ |
| 43 | 2007 | 0.402 | $1.129 \mathrm{E}+03$ | $1.118 \mathrm{E}+03$ | $4.500 \mathrm{E}+02$ | $4.500 \mathrm{E}+02$ | $4.296 \mathrm{E}+02$ | $2.015 \mathrm{E}+00$ | $7.688 \mathrm{E}-02$ |
| 44 | 2008 | 0.177 | $1.108 \mathrm{E}+03$ | $1.231 \mathrm{E}+03$ | $2.180 \mathrm{E}+02$ | $2.180 \mathrm{E}+02$ | $4.708 \mathrm{E}+02$ | $8.871 \mathrm{E}-01$ | $7.549 \mathrm{E}-02$ |
| 45 | 2009 | 0.050 | $1.361 \mathrm{E}+03$ | $1.612 \mathrm{E}+03$ | $8.000 \mathrm{E}+01$ | $8.000 \mathrm{E}+01$ | $6.079 \mathrm{E}+02$ | $2.486 \mathrm{E}-01$ | $9.271 \mathrm{E}-02$ |
| 46 | 2010 | 0.047 | $1.889 \mathrm{E}+03$ | $2.230 \mathrm{E}+03$ | $1.050 \mathrm{E}+02$ | $1.050 \mathrm{E}+02$ | $8.224 \mathrm{E}+02$ | $2.358 \mathrm{E}-01$ | $1.287 \mathrm{E}-01$ |
| 47 | 2011 | 0.126 | $2.606 \mathrm{E}+03$ | $2.938 \mathrm{E}+03$ | $3.700 \mathrm{E}+02$ | $3.700 \mathrm{E}+02$ | $1.055 \mathrm{E}+03$ | $6.307 \mathrm{E}-01$ | $1.775 \mathrm{E}-01$ |
| 48 | 2012 | 0.028 | $3.292 \mathrm{E}+03$ | $3.882 \mathrm{E}+03$ | $1.080 \mathrm{E}+02$ | $1.080 \mathrm{E}+02$ | $1.344 \mathrm{E}+03$ | $1.393 \mathrm{E}-01$ | $2.242 \mathrm{E}-01$ |
| 49 | 2013 | 0.007 | $4.527 \mathrm{E}+03$ | $5.345 \mathrm{E}+03$ | $3.800 \mathrm{E}+01$ | $3.800 \mathrm{E}+01$ | $1.743 \mathrm{E}+03$ | $3.561 \mathrm{E}-02$ | $3.084 \mathrm{E}-01$ |
| 50 | 2014 | 0.004 | $6.232 \mathrm{E}+03$ | $7.271 \mathrm{E}+03$ | $3.000 \mathrm{E}+01$ | $3.000 \mathrm{E}+01$ | $2.179 \mathrm{E}+03$ | $2.066 \mathrm{E}-02$ | $4.245 \mathrm{E}-01$ |
| 51 | 2015 | $8.381 \mathrm{E}+03$ |  |  | $5.709 \mathrm{E}-01$ |  |  |  |  |



| 38 | 2002 | $4.439 \mathrm{E}+02$ | $2.425 \mathrm{E}+02$ | 0.2219 | $1.756 \mathrm{E}+03$ | $1.756 \mathrm{E}+03$ | -0.60439 | $0.000 \mathrm{E}+00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 39 | 2003 | $8.671 \mathrm{E}+02$ | $1.866 \mathrm{E}+02$ | 0.9325 | $5.676 \mathrm{E}+03$ | $5.676 \mathrm{E}+03$ | -1.53632 | $0.000 \mathrm{E}+00$ |
| 40 | 2004 | $*$ | $9.574 \mathrm{E}+01$ | 1.0921 | $3.411 \mathrm{E}+03$ | $3.411 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 41 | 2005 | $*$ | $5.616 \mathrm{E}+01$ | 0.6724 | $1.232 \mathrm{E}+03$ | $1.232 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 42 | 2006 | $6.051 \mathrm{E}+01$ | $4.102 \mathrm{E}+01$ | 0.7136 | $9.550 \mathrm{E}+02$ | $9.550 \mathrm{E}+02$ | -0.38875 | $0.000 \mathrm{E}+00$ |
| 43 | 2007 | $5.206 \mathrm{E}+01$ | $3.428 \mathrm{E}+01$ | 0.4023 | $4.500 \mathrm{E}+02$ | $4.500 \mathrm{E}+02$ | -0.41779 | $0.000 \mathrm{E}+00$ |
| 44 | 2008 | $6.402 \mathrm{E}+01$ | $3.772 \mathrm{E}+01$ | 0.1771 | $2.180 \mathrm{E}+02$ | $2.180 \mathrm{E}+02$ | -0.52893 | $0.000 \mathrm{E}+00$ |
| 45 | 2009 | $5.550 \mathrm{E}+01$ | $4.940 \mathrm{E}+01$ | 0.0496 | $8.000 \mathrm{E}+01$ | $8.000 \mathrm{E}+01$ | -0.11647 | $0.000 \mathrm{E}+00$ |
| 46 | 2010 | $5.808 \mathrm{E}+01$ | $6.836 \mathrm{E}+01$ | 0.0471 | $1.050 \mathrm{E}+02$ | $1.050 \mathrm{E}+02$ | 0.16300 | $0.000 \mathrm{E}+00$ |
| 47 | 2011 | $1.224 \mathrm{E}+02$ | $9.006 \mathrm{E}+01$ | 0.1259 | $3.700 \mathrm{E}+02$ | $3.700 \mathrm{E}+02$ | -0.30687 | $0.000 \mathrm{E}+00$ |
| 48 | 2012 | $4.454 \mathrm{E}+01$ | $1.190 \mathrm{E}+02$ | 0.0278 | $1.080 \mathrm{E}+02$ | $1.080 \mathrm{E}+02$ | 0.98270 | $0.000 \mathrm{E}+00$ |
| 49 | 2013 | $1.390 \mathrm{E}+02$ | $1.638 \mathrm{E}+02$ | 0.0071 | $3.800 \mathrm{E}+01$ | $3.800 \mathrm{E}+01$ | 0.16444 | $0.000 \mathrm{E}+00$ |
| 50 | 2014 | $2.092 \mathrm{E}+02$ | $2.229 \mathrm{E}+02$ | 0.0041 | $3.000 \mathrm{E}+01$ | $3.000 \mathrm{E}+01$ | 0.06331 | $0.000 \mathrm{E}+00$ |

[^1]
## UNWEIGHTED LOG RESIDUAL PLOT FOR DATA SERIES \# 1



| 2005 | 0.0000 | $=======\mid$ |
| :--- | :--- | :--- |
| 2006 | -0.3887 | $========\mid$ |
| 2007 | -0.4178 | $==========\mid$ |
| 2008 | -0.5289 | $==\mid$ |
| 2009 | -0.1165 | $\mid===$ |
| 2010 | 0.1630 | $=====\mid$ |
| 2011 | -0.3069 | $\mid===================$ |
| 2012 | 0.9827 | $\mid===$ |
| 2013 | 0.1644 | $\mid=$ |
| 2014 | 0.0633 |  |

```
Faroe Bank Cod RV
                                    Page 5
        Observed (O) and Estimated (*) CPUE for Data Series # 1 -- Survey CPUE Spring
    1200. -
    O
    1000. -:
        :
        800. -:
        :
        :
        600. -:
        OO
                            O O
        400. -:
            : O
        ******* O
        * * * *
        200. -: ************* * * *** O
                        * OO 2
                            O 2*** OO * O O **
                                    O OO O O **2*2O O
        0. -:
            :.......................................................................................
        1959. 1965. 1971. 1977. 1983. 1989. 1995. 2001. 2007. 2013. 2019.
```

```
    F
5. -:
                                F
                                    F
4. -:
    : F F
        F
    F
3. -:
: F
F
2. -: F F
F
    F FF FFFF
    F F F
        FF F
1. -: -- - -- - 2- - -- - -- - -- - -- 2 2- - -- - -- - -- - -- - -- - 22
        FB F F F
    B BB BB B BB F B F F F BB BB F
        B BB B B B BB B BB B B B 2 B BB B B B B
            B B BB 22 BB F 2 B2
            B BB FF
1959. 1965. 1971. 1977. 1983. 1989. 1995. 2001. 2007. 2013. 2019.
```



Figure 3.1. Faroe Bank (sub-division Vb2) cod. Reported landings 1965-2014. Since 1992 only catches from Faroese and Norwegian vessels are considered to be taken on Faroe Bank. Lower plot: fishing days (fishing year) 1997-2015 for long line gear type in the Faroe Bank.


Figure 3.2. Faroe Bank (subdivision Vb2) cod. Catch per unit of effort in the spring groundfish survey (1983-2015)(red line) and summer survey (1996-2014)(black line). Vertical bars and shaded areas show the standard error in the estimation of indexes.


Figure 3.3. Faroe Bank (sub-division Vb2) cod. Length distributions in summer survey (1996-2014)


Figure 3.4. Faroe Bank (sub-division Vb2) cod. Age-disaggregated indices in the summer survey (ages 1-11)(1996-2014)


Figure 3.5. Faroe Bank (sub-division Vb2) cod. Length distributions in spring survey (1994-2015). No surveys were conducted in 1996, 2004 and 2005.


Figure 3.6. Faroe Bank (sub-division Vb2) cod. Age-disaggregated indices in the spring survey (ages 1-11) (1994-2015). No surveys were conducted in 1996, 2004 and 2005.

Recruitment yearclasses of Faroe Bank cod
(correlation from 1995 to 2013 equals 0.85)


Figure 3.7. Faroe Bank (sub-division Vb2) cod. Correlation between recruitment year classes in both survey indices.


Figure 3.8. Results from the surplus production model using the summer index. Observed (points) and expected catch rates (kg/hour) (top panel). Estimated fishing mortality (black line) and exploitation ratios (ratio of spring index to landings)(green line) (ratio of summer index to landings)(red line)(middle panel). Model residuals in log scale (bottom panel)

Catch rates

F
แ


Log residuals


Figure 3.9. Results from the surplus production model using the spring index. Observed (points) and expected catch rates (kg/hour) (top panel). Estimated fishing mortality (black line) and exploitation ratios (ratio of spring index to landings)(green line) (ratio of summer index to landings)(red line)(middle panel). Model residuals in log scale (bottom panel)

## 4 Faroe Plateau cod

## Summary

The input data consisted of the catch-at-age matrix (ages 2-10+ years) for the period 1959-2014 and two age-disaggregated abundance indices obtained from the two Faroese groundfish surveys: the spring survey 1994-2015 (shifted back to the previous year) and the summer survey 1996-2014. The maturities were obtained from the spring survey 1983-2015.

The assessment settings were the same as in the 2014 assessment. An XSA was tuned with the two survey indices. The fishing mortality in 2014 (average of ages 3-7 years) was estimated at 0.41 , which was higher than the Fmsy of 0.32 . The total stock size (age $2+$ ) in the beginning of 2014 was estimated at 27700 tonnes and the spawning stock biomass at 21100 tonnes, which was slightly above the limit biomass of 21000 tonnes.

The short term prediction until year 2017 showed a slightly decreasing total stock biomass to 24200 tonnes and a spawning stock biomass to 19500 tonnes.

It is adviced to reduce the fishing mortality substantially to rebuild the stock

### 4.1 Stock description and management units

Both genetic and tagging data suggest that there are three cod stocks present in Faroese waters: on the Faroe Bank (Division Vb2), on the Faroe Plateau (Division Vb1) and on the Faroe-Iceland Ridge. Cod on the Faroe-Iceland Ridge seem to belong to the cod stock at Iceland, and the WG in 2005 decided to exclude these catches from the catch-at-age calculations. The annex provides more information.

### 4.2 Scientific data

### 4.2.1 Trends in landings and fisheries

The landings were obtained from the Fisheries Ministry and Statistics Faroe Islands. The landings are presented in Table 4.2.1 and the working group estimates are presented in Table 4.2.2. The catches on the Faroe-Iceland Ridge, i.e. for the large single trawlers and the large longliners were not included in the catch-at-age calculations. In recent years the longliners have taken the majority of the cod catches (Table 4.2.3).

### 4.2.2 Catch-at-age

Landings-at-age for 2014 are provided for the Faroese fishery in Table 4.2.4. Faroese landings from most of the fleet categories were sampled (Table 4.2.5). The catch-at-age is shown in Table 4.2.6. Catch curves are shown in Fig. 4.2.1. They show atypical patterns in 1996 and to some extent in 2001-2002 when there appears to be an increase over the previous year for ages where a decrease would normally have been expected. This could be due to catchability for longliners depending on fish growth, causing atypical catch curves for longliners.

### 4.2.3 Weight-at-age

Mean weight-at-age data are provided for the Faroese fishery in Table 4.2.7. These were calculated using the length/weight relationship based on individual length/weight measurements of samples from the landings. The sum-of-products-check for 2014 showed a discrepancy of $0 \%$. The weights have increased in recent years (Figure 4.2.2).

### 4.2.4 Maturity-at-age

The proportion of mature cod by age during the Faroese groundfish surveys carried out during the spawning period (March) is given in Table 4.2.8 and in Figure 4.2.3. Full maturity is generally reached at age 5 or 6 , but considerable changes have been observed in the proportion mature for younger ages between years.

### 4.2.5 Catch, effort and research vessel data

The spring groundfish surveys in Faroese waters with the research vessel Magnus Heinason is used as a tuning series. The catch curves showed a normal pattern (Figure 4.2.4), i.e., a decreasing trend after age 5 . The stratified mean catch of cod per unit effort (Figure 4.2.5) has been low in the recent years.
The other tuning series used is the Summer Groundfish Survey. The stratified mean catch of cod per unit effort has been low in recent years (Figure 4.2.5). The catch curves (Figure 4.2.6) show that the fish are fully recruited to the survey gear at an age of 4 or 5 years. Both tuning series are presented in Table 4.2.9 and they show that there are few small cod in the stock.

Three commercial cpue series (longliners and pairtrawlers) are also presented (Tables 4.2.10, 4.2.11, and 4.2.12 as well as Figure 4.2.7), although they are not used as tuning series. All these series show that the incoming year classes are small. Note that the small boats (0-25 GRT) operating with longlines and jigging reels close to land have had a relatively higher cpue in recent years compared with the other cpue series and the two tuning series (Figure 4.2.8 and Figure 4.2.9), although the larger longliners also have had a high catchability in recent years. When that happens, the recruitment of 2year old cod tends to be low.

### 4.3 Information from the fishing industry

The sampling of the catches is included in the 'scientific data'. The fishing industry has since 1996 gathered data on the size composition of the landings but this information has not been used in this assessment.

### 4.4 Methods

This is an update assessment using XSA and the procedure is described in stock annex and the results of the assessment is mostly data-driven implying that there may be little difference in the assessment results by using another method.

### 4.5 Reference points

The reference points are dealt with in the general section of Faroese stocks. The PA reference points for Faroe Plateau cod are the following: Bpa $=40 \mathrm{kt}$, Blim $=21 \mathrm{kt}$, Fpa $=0.35$ and Flim $=0.68$.

The reference points based on the yield-per-recruit curve are the following: $\mathrm{F}_{\max }=0.25$, $\mathrm{F}_{0.1}=0.11, \mathrm{~F} 35 \% \mathrm{SPR}=0.17, \mathrm{~F}_{\text {med }}=0.41, \mathrm{~F}_{\text {low }}=0.10, \mathrm{~F}_{\text {high }}=0.97$.

The group adopted in 2011 following preliminary MSY reference points: $\mathrm{F}_{\text {msy }}=0.32$, see section 4.8. The $\mathrm{B}_{\text {trigger }}$ was set at $\mathrm{B}_{\mathrm{pa}}=40 \mathrm{kt}$.

### 4.6 State of the stock

Since the current assessment is an update assessment, the same procedure is followed as last year: to use the two surveys for tuning. The commercial series showed a similar
overall tendency as the surveys (Figure 4.2.7) but were not used in the tuning. The XSA-run (Table 4.6.1) showed that the fit between the model and the tuning series (logQ residuals, Figure 4.6.1) was rather poor for the young ages and there seemed to be both year class effects and year effects.

The results from the XSA-run shows that fishing mortality (F3-7) has fluctuated in recent years without a trend (Table 4.6.2, Figure 4.6.2), and other measures of fishing mortality have done so as well (Table 4.6.4, Figure 4.6.3). The population numbers, total biomass and spawning stock biomass have been low compared with other years in the series (Table 4.6.3, Table 4.6.4, Figure 4.6.2). The poor state of the stock since 2005 has been due to poor recruitment (not poor individual growth). Prior to that time, extremely weak year classes ( $<5$ million individuals) were only observed two times, whereas it has happened four times since 2005 (in 2011-2014). In the past there has been a poor relationship between the size of the spawning stock and subsequent recruitment (Figure 4.6.4), but the increasing number of low data points in recent years have strengthened the stock-recruitment relationship. The spawning stock biomass in the terminal year was close to Blim and the fishing mortality above Fmsy (Figure 4.6.5).

During the years 1938-55 a large work was undertaken in ICES ("The North-Western Area Committee", which established the "Sub-Committee on the Faroe Question", sometimes referred to as "The Sub-Committee on a proposed Closure of Certain Extraterritorial Waters off the Faroes") to investigate whether certain areas around the Faroe Islands should be closed to fishing. Although no areas were closed as a result of this work a large amount of data became available. These data, together with other data, are now used to estimate the stock size of Faroe Plateau cod back to 1906, which puts the present stock size into a wider perspective (Working Document no. 32). A cpue series (tonnes per million tonn-hours) for British steam trawlers 1924-1972 was available from the data presented in Jákupsstovu and Reinert (1994). The cpue series was also used, and explained, in Jones (1966). There was an overlap between the cpue series and the stock assessment for the years 1959-1972. Another cpue series (cwts per day of absence from port, $1 \mathrm{cwt}=50.8 \mathrm{~kg}$ ) was available for English steam trawlers 1906-1954 (with gaps). In addition there was a record of Faroese boat catches that extended into the war periods. In WD 32 the biomass back in time is estimated in four steps: 1) Extending the British cpue back to 1906 by the use of English steam trawlers. 2-3) Extending the British cpue to the World War 1 period and World War 2 period (with gaps) by the Faroe boat catches. 4) Extending the age 2+ biomass from the age-based assessment back to 1906 by using the raw or constructed British cpue series. The result depended upon whether a regression line (biomass versus cpue) was used or a scaling factor (sum of biomass divided by the sum of cpue), the latter giving a higher biomass estimate back in time. The resulting exploitation ratio of the higher biomass was in better correspondence with tagging returns and a Faroese longliner series (see WD 32) and is probably more reliable. The results are presented in Table 4.6.5 and Figure 4.6.6. The biomass in 2005-2014 was very low compared with the entire period, but it is worth noting that the fishing mortality (exploitation ratio) was high already in the 1930s. The extension of biomass back in time can likely be improved in the future by including the Faroe longliner CPUE series mentioned above and also to include age data prior to 1959.

### 4.7 Short term forecast

### 4.7.1 Input data

The input data for the short term prediction are given in Table 4.7.1. Note the extremely weak YC2010, YC2011, YC2012 and YC2013, which were set to the face value from the XSA-run, i.e., according to the Annex. Estimates of stock size (ages 3+) were taken directly from the XSA stock numbers. The exploitation pattern was estimated as the average fishing mortality for 2012-2014. The weights at age in the catches in 2015 were estimated from the spring survey (ages 2 and 6-8 years) whereas the other ages were estimated from the catch weights in January-February 2015. The weights in the catches in 2016 were set to the values in 2015 and the average of 2013-2015 was expected for 2017. The proportion mature in 2014 was set to the 2014 values from the spring groundfish survey, and for 2015-2016 to the average values for 2012-2014.

### 4.7.2 Results

The landings in 2015 are expected to be 6600 tonnes (Table 4.7.2) (the landings from the Faroe-Icelandic ridge should be added to this figure in order to get the total Faroese landings within the Vb 1 area). The spawning stock biomass is expected to be 18900 tonnes in 2015, 19700 tonnes in 2016 and eventually 19500 tonnes in 2017. The "old" year classes (YC 2008 and YC2009) are still important for the SSB in 2016 and 2017 (Figure 4.7.1).

### 4.8 Long term forecast

The input to the traditional long term forecast (yield per recruit) is presented in Table 4.8.1 and the result is presented in Table 4.8.2 and Figure 4.8.1.

Single species long term forecasts for Faroe Plateau cod indicated Fmsy values lower than Fpa. An FLR procedure (MSE, Management strategy evaluations using FLR standard packages; a simulation of management and stock response over a 20 yr period) for Faroe Plateau cod indicates that Fmsy is 0.32 . This value ( 0.32 ) was adopted by the NWWG 2011 as a preliminary Fmsy.

### 4.9 Uncertainties in assessment and forecast

Since there is no incentive to discard fish or misreport catches under the effort management system, the catch figures are considered adequate, as well as the catch-at-age, although the number of otoliths should have been higher.

There was a clear retrospective pattern (Figure 4.9.1), indicating uncertainties in the assessment.

Steingrund et al. (2010) found that the recruitment of Faroe Plateau cod (age 2) could be rather precisely estimated as there is a relationship between cod biomass (age 3+) and the amount of cannibalistic cod in nearshore waters in June-October the previous year. This approach showed that the recent year classes were extremely weak (Figure 4.9.2).

### 4.10 Comparison with previous assessment and forecast

The assessment settings were according to the Stock Annex. The 2015 assessment was much in line with the 2014 assessment and forecast (Figure 4.10.1).

### 4.11 Management plans and evaluations

There is no explicit management plan for this stock. A management system based on number of fishing days, closed areas and other technical measures was introduced in 1996 with the purpose to ensuring sustainable demersal fisheries in Vb . This was before ICES introduced PA and MSY reference values and at the time it was believed that the purpose was achieved, if the total allowable number of fishing days was set such, that on average $33 \%$ of the cod exploitable stock in numbers would be harvested annually. This translates into an average F of 0.45 , above the Fpa of 0.35 . ICES considers this to be inconsistent with the PA and MSY approaches. Some work has been done in the Faroes to move away from the Ftarget of 0.45 to be more consistent with the ICES advice.

### 4.12 Management considerations

The cod stock is assessed to be in a very poor state and is predicted to remain so for the next two years due to poor recruitment. Although the environmental conditions have been rather special since 2007 (lots of mackerel) and may partly be responsible for the poor state of the cod stock, it is certainly necessary to protect the cod stock as much as possible. The reason is not only that it may prevent a total collapse of the stock but also that the stock may recover faster in the future.

Hence, the number of fishing days should be considered and further area closures might be necessary.

The managers should consider changing the management system, or changing the implementation of it, in order to rebuild the cod stock.

### 4.13 Ecosystem considerations

Regarding the ecosystem effects on fishing, this issue is partly addressed in the ecological modelling work presented in the overview section for Faroese stocks.

### 4.14 Regulations and their effects

There seems to be a poor relationship between the number of fishing days and the fishing mortality because of large fluctuations in catchability. Area restrictions may help to reduce fishing mortality, but they cause practical problems for the fishing fleets (e.g. high concentrations of vessels in certain areas). Area restrictions may be best suited to protect certain fish species/sizes in certain areas, whereas the number of fishing days remains the only tool to reduce the overall fishing mortality, given the effort management system.

The area closure (for commercial longliners close to land) introduced in July 2011 and ending in August 2013 to protect young fish has not yet resulted in strong recruitment, since the 2008 year class is below average size, and the 2009-2011 year classes either poor or exceptionally poor.

### 4.15 Changes in fishing technology and fishing patterns

Fishing effort per fishing day may have increased gradually since the effort management system was introduced in 1996, although little direct quantitative information exists. There also seems to have been substantial increases in fishing power when new vessels are replacing old vessels.

The fishing pattern in recent years has changed in comparison to previous years. The large longliners seem to have exploited the deep areas (> 200 m ) to a larger extent (ling and tusk) because the catches in shallower waters of cod and haddock have been so poor - which was also observed in the beginning of the 1990s. This could reduce the fishing mortality on cod and haddock, but the small longliners and jiggers still exploit the shallow areas.

### 4.16 Changes in the environment

The primary production has been low for a number of years, albeit high in 2008 to 2010, but it is not believed that this has any relationship with a change in the environment. The temperature has been high in recent years, which may have a negative effect on cod recruitment (Planque and Fredou, 1999).

### 4.17 References

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Table 4.2.1. Faroe Plateau cod (sub-division Vb1). Nominal catch (t) by countries, as officially reported to ICES.

|  | Denmark | Faroe Islands | France | Germany | Iceland | Norway | Greenland | Portugal | UK (E/W/NI) | UK (Scotland) | United Kingdom | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 8 | 34,492 | 4 | 8 |  | 83 | - |  | - | - | - | 34,595 |
| 1987 | 30 | 21,303 | 17 | 12 |  | 21 | - |  | 8 | - | - | 21,391 |
| 1988 | 10 | 22,272 | 17 | 5 |  | 163 | - |  | - | - | - | 22,467 |
| 1989 | $\cdots$ | 20,535 | - | 7 |  | 285 | - |  | - | - | - | 20,827 |
| 1990 | - | 12,232 | - | 24 |  | 124 | - |  | - | - | - | 12,380 |
| 1991 | - | 8,203 | $-1$ | 16 |  | 89 | - |  | 1 | - | - | 8,309 |
| 1992 | - | 5,938 | $3^{2}$ | 12 |  | 39 | - |  | 74 | - | - | 6,066 |
| 1993 | - | 5,744 | $1^{2}$ | + |  | 57 | - |  | 186 | - | - | 5,988 |
| 1994 | - | 8,724 | - | 2 |  | 36 | - |  | 56 | - | - | 8,818 |
| 1995 | - | 19,079 | $2^{2}$ | 2 |  | 38 | - |  | 43 | - | - | 19,164 |
| 1996 | - | 39,406 | $1^{2}$ | + |  | 507 | - |  | 126 | - | - | 40,040 |
| 1997 | - | 33,556 | - | + |  | 410 | - |  | $61^{2}$ | - | - | 34,027 |
| 1998 | - | 23,308 | - | - |  | 405 | - |  | $27^{2}$ | - | - | 23,740 |
| 1999 | - | 19,156 | - | 39 | - | 450 | - |  | 51 | - |  | 19,696 |
| 2000 |  | 0 | 1 | 2 | - | 374 | - |  | 18 | - |  | 395 |
| 2001 |  | 29,762 | $9^{2}$ | 9 | - | 531. | - |  | 50 | - |  | 30,361 |
| 2002 |  | 40,602 | 20 | 6 | 5 | 573 |  |  | 42 | - |  | 41,248 |
| 2003 |  | 30,259 | 14 | 7 | - | 447 | - |  | 15 | - |  | 30,742 |
| 2004 |  | 17,540 | 2 | $3^{2}$ |  | 414 |  | 1 | 15 | \% |  | 17,975 |
| 2005 |  | 13,556 | - |  |  | 201 |  |  | 24 | - |  | 13,781 |
| 2006 |  | 11,629 | 7 | $1^{2}$ |  | 49 | 5 |  | 1 | - |  | 11,691 |
| 2007 |  | 9,905 | $1^{2}$ |  |  | 71 | 7 |  | 3 | 358 |  | 10,344 |
| 2008 |  | 9,394 | 1 |  |  | 40 |  |  |  | 383 |  | 9,818 |
| 2009 |  | 10,736 | 1 |  |  | 14 | 7 |  |  | 300 |  | 11,058 |
| 2010 |  | 13,878 | 1 |  |  | 10 |  |  |  | 312 |  | 14,201 |
| 2011 |  | 11,348 | - |  |  |  |  |  |  |  |  | 11,348 |
| 2012 |  | 8,437 | 0 |  | 28 |  |  |  |  |  |  | 8,465 |
| 2013 . |  | 5,331 | 0 |  | 20 |  | 2 |  |  |  |  | 5,333 |
| 2014 |  | 7,037 * |  |  |  | 6 |  |  |  | 270 |  | 7,314 |

Table 4.2.2. Faroe Plateau cod (sub-division Vb1). Nominal catch (t) used in the assessment.


Table 4.2.3. Faroe Plateau cod (sub-division Vb1). The landings of Faroese fleets (in percents) of total catch ( $\mathbf{t}$ ). Note that the catches on the Faroe-Iceland ridge (mainly belonging to single trawlers $>\mathbf{1 0 0 0} \mathbf{H P}$ ) are included in this table, but excluded in the XSA-run.


Table 4.2.4. Faroe Plateau cod (sub-division Vb1). Catch in numbers at age per fleet in terminal year. Numbers are in thousands and the catch is in tonnes, gutted weight.

| AgelFleet | Open boats | Longliners Jiggers$<100 \text { GRT }$ |  | Single trwl $0-399 \mathrm{HP}$ |  | Single trwl $400-1000 \mathrm{H}$ | Single trwl $H>1000 \mathrm{HP}$ | $\begin{aligned} & \text { Pair trwl } \\ & 700-999 \end{aligned}$ |  |  | trwl $000 \mathrm{HP}$ | Longliners $>100 \text { GRT }$ | Gillnetters | Others (scaling) |  | Catch-at -age |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 0 | 346 | 29 | 0 | O | 8 | 3 |  | 1 |  | 11 | 48 | 0 |  | -16 | 430 |
| 3 | 0 | 342 | 44 | 0 | 0 | 16 | 14 |  | 2 |  | 50 | 129 | 0 |  | -22 | 575 |
| 4 | 0 | 160 | 24 | 0 | O | 16 | 10 |  | 2 |  | 37 | 81 | 0 |  | -12 | 318 |
| 5 | 0 | 234 | 40 | 0 | O | 38 | 20 |  | 4 |  | 87 | 153 | 0 |  | -22 | 554 |
| 6 | 0 | 111 | 22 | 0 | 0 | 30 | 19 |  | 4 |  | 88 | 130 | 0 |  | -15 | 389 |
| 7 | 0 | 23 | 6 | 0 | O | 8 | 4 |  | 1 |  | 21 | 36 | 0 |  | -4 | 95 |
| 8 | 0 | 5 | 1 | 0 | 0 | 1 | 1 |  | 0 |  | 4 | 4 | 0 |  | -1 | 15 |
| 9 | 0 | 3 | 1 | 0 | 0 | 0 | 1 |  | 0 |  | 2 | 4 | 0 |  | -1 | 10 |
| 10+ | 0 | 0 | 0 | 0 |  | 0 | 0 |  | 0 |  | 0 | 0 | 0 |  | 1 | 1 |
| Sum | 0 | 1224 | 167 | 0 | O | 117 | 72 |  | 14 |  | 300 | 585 | 0 |  | -92 | 2387 |
| G.weight | 0 | 2242 | 368 | 0 | 0 | 354 | 258 |  | 49 |  | 1105 | 1565 | 0 |  | -221 | 5720 |

Gutted total catch is calculated as round weight divided by 1.11.

Table 4.2.5. Faroe Plateau cod (sub-division Vb1). Number of samples, lengths, otoliths, and individual weights in terminal year.

| Fleet | Size | Samples | Lengths | Otoliths | Weights |
| :--- | :--- | ---: | ---: | ---: | ---: |
| Open boats |  | 4 | 677 | 20 | 677 |
| Longliners | $<100$ GRT | 15 | 2,985 | 460 | 2,580 |
| Longliners | $>100$ GRT | 14 | 2,926 | 317 | 2,926 |
| Jiggers |  | 0 | 0 | 0 | 0 |
| Gillnetters |  | 0 | 0 | 0 | 0 |
| Sing. trawlers | $<400 \mathrm{HP}$ | 0 | 0 | 0 | 0 |
| Sing. trawlers | $400-1000 \mathrm{HP}$ | 17 | 3,560 | 319 | 3,560 |
| Sing. trawlers | $>1000 \mathrm{HP}$ | 0 | 0 | 0 | 0 |
| Pair trawlers | $<1000 \mathrm{HP}$ | 0 | 0 | 0 | 0 |
| Pair trawlers | $>1000 \mathrm{HP}$ | 30 | 5,718 | 479 | 4,514 |
| Total |  | 80 | 15,866 | 1,595 | 14,257 |

Table 4.2.6. Faroe Plateau cod (sub-division Vb1). Catch in numbers at age used in the XSA model.

|  | Ag |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 1959 | 0 | 2002 | 4239 | 858 | 1731 | 200 | 207 | 50 | 10 | 0 |
| 1960 | 0 | 4728 | 4027 | 2574 | 513 | 876 | 171 | 131 | 61 | 0 |
| 1961 | 0 | 3093 | 2686 | 1331 | 1066 | 232 | 372 | 78 | 29 | 0 |
| 1962 | 0 | 4424 | 2500 | 1255 | 855 | 481 | 93 | 94 | 22 | 0 |
| 1963 | 0 | 4110 | 3958 | 1280 | 662 | 284 | 204 | 48 | 30 | 0 |
| 1964 | 0 | 2033 | 3021 | 2300 | 630 | 350 | 158 | 79 | 41 | 0 |
| 1965 | 0 | 852 | 3230 | 2564 | 1416 | 363 | 155 | 48 | 63 | 0 |
| 1966 | 0 | 1337 | 970 | 2080 | 1339 | 606 | 197 | 104 | 33 | 0 |
| 1967 | 0 | 1609 | 2690 | 860 | 1706 | 847 | 309 | 64 | 27 | 0 |
| 1968 | 0 | 1529 | 3322 | 2663 | 945 | 1226 | 452 | 105 | 11 | 0 |
| 1969 | 0 | 878 | 3106 | 3300 | 1538 | 477 | 713 | 203 | 92 | 0 |
| 1970 | 0 | 402 | 1163 | 2172 | 1685 | 752 | 244 | 300 | 44 | 0 |
| 1971 | 0 | 328 | 757 | 821 | 1287 | 1451 | 510 | 114 | 179 | 0 |
| 1972 | 0 | 875 | 1176 | 810 | 596 | 1021 | 596 | 154 | 25 | 0 |
| 1973 | 0 | 723 | 3124 | 1590 | 707 | 384 | 312 | 227 | 120 | 97 |
| 1974 | 0 | 2161 | 1266 | 1811 | 934 | 563 | 452 | 149 | 141 | 91 |
| 1975 | 0 | 2584 | 5689 | 2157 | 2211 | 813 | 295 | 190 | 118 | 150 |
| 1976 | 0 | 1497 | 4158 | 3799 | 1380 | 1427 | 617 | 273 | 120 | 186 |
| 1977 | 0 | 425 | 3282 | 6844 | 3718 | 788 | 1160 | 239 | 134 | 9 |
| 1978 | 0 | 555 | 1219 | 2643 | 3216 | 1041 | 268 | 201 | 66 | 56 |
| 1979 | 0 | 575 | 1732 | 1673 | 1601 | 1906 | 493 | 134 | 87 | 38 |
| 1980 | 0 | 1129 | 2263 | 1461 | 895 | 807 | 832 | 339 | 42 | 18 |
| 1981 | 0 | 646 | 4137 | 1981 | 947 | 582 | 487 | 527 | 123 | 55 |
| 1982 | 0 | 1139 | 1965 | 3073 | 1286 | 471 | 314 | 169 | 254 | 122 |
| 1983 | 0 | 2149 | 5771 | 2760 | 2746 | 1204 | 510 | 157 | 104 | 102 |
| 1984 | 0 | 4396 | 5234 | 3487 | 1461 | 912 | 314 | 82 | 34 | 66 |
| 1985 | 0 | 998 | 9484 | 3795 | 1669 | 770 | 872 | 309 | 65 | 80 |
| 1986 | 0 | 210 | 3586 | 8462 | 2373 | 907 | 236 | 147 | 47 | 38 |
| 1987 | 0 | 257 | 1362 | 2611 | 3083 | 812 | 224 | 68 | 69 | 26 |
| 1988 | 0 | 509 | 2122 | 1945 | 1484 | 2178 | 492 | 168 | 33 | 25 |
| 1989 | 0 | 2237 | 2151 | 2187 | 1121 | 1026 | 997 | 220 | 61 | 9 |
| 1990 | 0 | 243 | 2849 | 1481 | 852 | 404 | 294 | 291 | 50 | 26 |
| 1991 | 0 | 192 | 451 | 2152 | 622 | 303 | 142 | 93 | 53 | 24 |
| 1992 | 0 | 205 | 455 | 466 | 911 | 293 | 132 | 53 | 30 | 34 |
| 1993 | 0 | 120 | 802 | 603 | 222 | 329 | 96 | 33 | 22 | 25 |
| 1994 | 0 | 573 | 788 | 1062 | 532 | 125 | 176 | 39 | 23 | 16 |
| 1995 | 0 | 2615 | 2716 | 2008 | 1012 | 465 | 118 | 175 | 44 | 49 |
| 1996 | 0 | 351 | 5164 | 4608 | 1542 | 1526 | 596 | 147 | 347 | 47 |
| 1997 | 0 | 200 | 1278 | 6710 | 3731 | 657 | 639 | 170 | 51 | 120 |
| 1998 | 0 | 455 | 745 | 1558 | 5140 | 1529 | 159 | 118 | 28 | 25 |
| 1999 | 0 | 1185 | 993 | 799 | 1107 | 2225 | 439 | 59 | 17 | 7 |
| 2000 | 0 | 2091 | 2637 | 782 | 426 | 674 | 809 | 104 | 7 | 1 |
| 2001 | 0 | 3912 | 3759 | 2101 | 367 | 367 | 718 | 437 | 36 | 6 |


|  | Age |  |  |  | $\mathbf{4}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |  |  |  |  |  |
| 2002 | 0 | 2079 | 7283 | 3372 | 1671 | 470 | 533 | 413 | 290 | 7 |
| 2003 | 0 | 678 | 2128 | 4572 | 1927 | 640 | 177 | 91 | 115 | 20 |
| 2004 | 0 | 100 | 691 | 1263 | 2105 | 736 | 240 | 65 | 42 | 37 |
| 2005 | 0 | 494 | 592 | 877 | 1122 | 823 | 204 | 41 | 19 | 30 |
| 2006 | 0 | 1182 | 1167 | 499 | 706 | 852 | 355 | 81 | 11 | 3 |
| 2007 | 0 | 540 | 1308 | 771 | 336 | 308 | 273 | 91 | 21 | 3 |
| 2008 | 0 | 293 | 776 | 799 | 439 | 191 | 160 | 159 | 58 | 20 |
| 2009 | 0 | 875 | 2267 | 863 | 619 | 297 | 85 | 55 | 43 | 17 |
| 2010 | 0 | 2113 | 2034 | 861 | 468 | 481 | 178 | 58 | 33 | 38 |
| 2011 | 0 | 330 | 2360 | 1242 | 367 | 189 | 127 | 50 | 19 | 2 |
| 2012 | 0 | 49 | 518 | 1348 | 556 | 201 | 99 | 69 | 25 | 22 |
| 2013 | 0 | 57 | 179 | 344 | 608 | 182 | 40 | 26 | 15 | 6 |
| 2014 | 0 | 430 | 575 | 318 | 554 | 389 | 95 | 15 | 10 | 1 |

Table 4.2.7. Faroe Plateau cod (sub-division Vb1). Mean weight at age (kg) in the catches.

|  | A |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 1959 | 0 | 0.850 | 1.730 | 3.230 | 4.400 | 5.800 | 6.370 | 7.340 | 7.880 | 10.270 |
| 1960 | 0 | 1.000 | 2.030 | 3.370 | 4.420 | 6.020 | 6.650 | 8.120 | 11.000 | 10.270 |
| 1961 | 0 | 1.080 | 2.220 | 3.450 | 4.690 | 5.520 | 7.090 | 9.910 | 8.030 | 10.270 |
| 1962 | 0 | 1.000 | 2.270 | 3.350 | 4.580 | 4.930 | 9.080 | 6.590 | 6.660 | 10.270 |
| 1963 | 0 | 1.040 | 1.940 | 3.510 | 4.600 | 5.500 | 6.780 | 8.710 | 11.720 | 10.820 |
| 1964 | 0 | 0.970 | 1.830 | 3.150 | 4.330 | 6.080 | 7.000 | 6.250 | 6.190 | 14.390 |
| 1965 | 0 | 0.920 | 1.450 | 2.570 | 3.780 | 5.690 | 7.310 | 7.930 | 8.090 | 11.110 |
| 1966 | 0 | 0.980 | 1.770 | 2.750 | 3.510 | 4.800 | 6.320 | 7.510 | 10.340 | 11.650 |
| 1967 | 0 | 0.960 | 1.930 | 3.130 | 4.040 | 4.780 | 6.250 | 7.000 | 11.010 | 10.690 |
| 1968 | 0 | 0.880 | 1.720 | 3.070 | 4.120 | 4.650 | 5.500 | 7.670 | 10.950 | 9.280 |
| 1969 | 0 | 1.090 | 1.800 | 2.850 | 3.670 | 4.890 | 5.050 | 7.410 | 8.660 | 14.390 |
| 1970 | 0 | 0.960 | 2.230 | 2.690 | 3.940 | 5.140 | 6.460 | 10.310 | 7.390 | 9.340 |
| 1971 | 0 | 0.810 | 1.800 | 2.980 | 3.580 | 3.940 | 4.870 | 6.480 | 6.370 | 10.220 |
| 1972 | 0 | 0.660 | 1.610 | 2.580 | 3.260 | 4.290 | 4.950 | 6.480 | 6.900 | 11.550 |
| 1973 | 0 | 1.110 | 2.000 | 3.410 | 3.890 | 5.100 | 5.100 | 6.120 | 8.660 | 7.570 |
| 1974 | 0 | 1.080 | 2.220 | 3.440 | 4.800 | 5.180 | 5.880 | 6.140 | 8.630 | 7.620 |
| 1975 | 0 | 0.790 | 1.790 | 2.980 | 4.260 | 5.460 | 6.250 | 7.510 | 7.390 | 8.170 |
| 1976 | 0 | 0.940 | 1.720 | 2.840 | 3.700 | 5.260 | 6.430 | 6.390 | 8.550 | 13.620 |
| 1977 | 0 | 0.870 | 1.790 | 2.530 | 3.680 | 4.650 | 5.340 | 6.230 | 8.380 | 10.720 |
| 1978 | 0 | 1.112 | 1.385 | 2.140 | 3.125 | 4.363 | 5.927 | 6.348 | 8.715 | 12.229 |
| 1979 | 0 | 0.897 | 1.682 | 2.211 | 3.052 | 3.642 | 4.719 | 7.272 | 8.368 | 13.042 |
| 1980 | 0 | 0.927 | 1.432 | 2.220 | 3.105 | 3.539 | 4.392 | 6.100 | 7.603 | 9.668 |
| 1981 | 0 | 1.080 | 1.470 | 2.180 | 3.210 | 3.700 | 4.240 | 4.430 | 6.690 | 10.000 |
| 1982 | 0 | 1.230 | 1.413 | 2.138 | 3.107 | 4.012 | 5.442 | 5.563 | 5.216 | 6.707 |
| 1983 | 0 | 1.338 | 1.950 | 2.403 | 3.107 | 4.110 | 5.020 | 5.601 | 8.013 | 8.031 |
| 1984 | 0 | 1.195 | 1.888 | 2.980 | 3.679 | 4.470 | 5.488 | 6.466 | 6.628 | 10.981 |
| 1985 | 0 | 0.905 | 1.658 | 2.626 | 3.400 | 3.752 | 4.220 | 4.739 | 6.511 | 10.981 |
| 1986 | 0 | 1.099 | 1.459 | 2.046 | 2.936 | 3.786 | 4.699 | 5.893 | 9.700 | 8.815 |
| 1987 | 0 | 1.093 | 1.517 | 2.160 | 2.766 | 3.908 | 5.461 | 6.341 | 8.509 | 9.811 |
| 1988 | 0 | 1.061 | 1.749 | 2.300 | 2.914 | 3.109 | 3.976 | 4.896 | 7.087 | 8.287 |
| 1989 | 0 | 1.010 | 1.597 | 2.200 | 2.934 | 3.468 | 3.750 | 4.682 | 6.140 | 9.156 |
| 1990 | 0 | 0.945 | 1.300 | 1.959 | 2.531 | 3.273 | 4.652 | 4.758 | 6.704 | 8.689 |
| 1991 | 0 | 0.779 | 1.271 | 1.570 | 2.524 | 3.185 | 4.086 | 5.656 | 5.973 | 8.147 |
| 1992 | 0 | 0.989 | 1.364 | 1.779 | 2.312 | 3.477 | 4.545 | 6.275 | 7.619 | 9.725 |
| 1993 | 0 | 1.155 | 1.704 | 2.421 | 3.132 | 3.723 | 4.971 | 6.159 | 7.614 | 9.587 |
| 1994 | 0 | 1.194 | 1.843 | 2.613 | 3.654 | 4.584 | 4.976 | 7.146 | 8.564 | 8.796 |
| 1995 | 0 | 1.218 | 1.986 | 2.622 | 3.925 | 5.180 | 6.079 | 6.241 | 7.782 | 8.627 |
| 1996 | 0 | 1.016 | 1.737 | 2.745 | 3.800 | 4.455 | 4.978 | 5.270 | 5.593 | 7.482 |
| 1997 | 0 | 0.901 | 1.341 | 1.958 | 3.012 | 4.158 | 4.491 | 5.312 | 6.172 | 7.056 |
| 1998 | 0 | 1.004 | 1.417 | 1.802 | 2.280 | 3.478 | 5.433 | 5.851 | 7.970 | 8.802 |
| 1999 | 0 | 1.050 | 1.586 | 2.350 | 2.774 | 3.214 | 5.496 | 8.276 | 9.129 | 10.652 |
| 2000 | 0 | 1.416 | 2.170 | 3.187 | 3.795 | 4.048 | 4.577 | 8.182 | 11.895 | 13.009 |


|  | Age |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| 2001 | 0 | 1.164 | 2.076 | 3.053 | 3.976 | 4.394 | 4.871 | 5.563 | 7.277 | 12.394 |
| 2002 | 0 | 1.017 | 1.768 | 2.805 | 3.529 | 4.095 | 4.475 | 4.650 | 6.244 | 7.457 |
| 2003 | 0 | 0.820 | 1.362 | 2.127 | 3.329 | 4.092 | 4.670 | 6.000 | 6.727 | 6.810 |
| 2004 | 0 | 1.037 | 1.154 | 1.693 | 2.363 | 3.830 | 5.191 | 6.326 | 7.656 | 9.573 |
| 2005 | 0 | 0.986 | 1.373 | 1.760 | 2.293 | 3.138 | 5.287 | 8.285 | 8.703 | 9.517 |
| 2006 | 0 | 0.839 | 1.304 | 1.988 | 2.386 | 3.330 | 4.691 | 7.635 | 9.524 | 11.990 |
| 2007 | 0 | 0.937 | 1.324 | 1.970 | 3.076 | 3.529 | 4.710 | 6.464 | 9.461 | 9.509 |
| 2008 | 0 | 1.209 | 1.478 | 2.104 | 2.714 | 3.804 | 4.669 | 5.915 | 7.233 | 9.559 |
| 2009 | 0 | 0.805 | 1.431 | 2.287 | 2.723 | 3.435 | 5.081 | 6.281 | 8.312 | 9.959 |
| 2010 | 0 | 1.049 | 1.642 | 2.400 | 3.212 | 3.678 | 4.774 | 5.973 | 7.094 | 9.800 |
| 2011 | 0 | 0.815 | 1.367 | 2.413 | 3.493 | 4.525 | 5.076 | 6.631 | 6.863 | 10.089 |
| 2012 | 0 | 1.007 | 1.315 | 1.893 | 3.102 | 4.279 | 5.573 | 5.871 | 7.482 | 9.206 |
| 2013 | 0 | 1.011 | 1.527 | 2.528 | 3.180 | 4.672 | 6.776 | 6.966 | 9.028 | 10.324 |
| 2014 | 0 | 1.099 | 1.653 | 2.466 | 3.000 | 4.148 | 6.489 | 9.394 | 9.236 | 12.120 |

Table 4.2.8. Faroe Plateau cod (sub-division Vb1). Proportion mature at age. From 1961-1982 the average from 1983-1996 is used (as it was used in the 1990s). In 2002, the high maturities for age 2 in 1983 (0.63), 1984 (0.4) and in 1993 ( 0.25 ) were revised, but not the maturities back in time.

|  | Age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 1959 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1960 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1961 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1962 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1963 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1964 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1965 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1966 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1967 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1968 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1969 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1970 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1971 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1972 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1973 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1974 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1975 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1976 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1977 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1978 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1979 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1980 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1981 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1982 | 0.00 | 0.17 | 0.64 | 0.87 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1983 | 0.00 | 0.03 | 0.71 | 0.93 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1984 | 0.00 | 0.07 | 0.96 | 0.98 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1985 | 0.00 | 0.00 | 0.50 | 0.96 | 0.96 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1986 | 0.00 | 0.00 | 0.38 | 0.93 | 1.00 | 1.00 | 0.96 | 0.94 | 1.00 | 1.00 |
| 1987 | 0.00 | 0.00 | 0.67 | 0.91 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1988 | 0.00 | 0.06 | 0.72 | 0.90 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1989 | 0.00 | 0.05 | 0.54 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1990 | 0.00 | 0.00 | 0.68 | 0.90 | 0.99 | 0.96 | 0.98 | 1.00 | 1.00 | 1.00 |
| 1991 | 0.00 | 0.00 | 0.72 | 0.86 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1992 | 0.00 | 0.06 | 0.50 | 0.82 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1993 | 0.00 | 0.03 | 0.73 | 0.78 | 0.91 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1994 | 0.00 | 0.05 | 0.33 | 0.88 | 0.96 | 1.00 | 0.96 | 1.00 | 1.00 | 1.00 |
| 1995 | 0.00 | 0.09 | 0.35 | 0.33 | 0.66 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1996 | 0.00 | 0.04 | 0.43 | 0.74 | 0.85 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1997 | 0.00 | 0.00 | 0.64 | 0.91 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1998 | 0.00 | 0.00 | 0.62 | 0.90 | 0.99 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1999 | 0.00 | 0.02 | 0.43 | 0.88 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |


| 2000 | 0.00 | 0.02 | 0.39 | 0.69 | 0.92 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2001 | 0.00 | 0.07 | 0.47 | 0.86 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2002 | 0.00 | 0.04 | 0.37 | 0.76 | 0.97 | 0.93 | 0.97 | 1.00 | 1.00 | 1.00 |
| 2003 | 0.00 | 0.00 | 0.29 | 0.79 | 0.88 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2004 | 0.00 | 0.00 | 0.51 | 0.78 | 0.92 | 0.89 | 0.87 | 1.00 | 1.00 | 1.00 |
| 2005 | 0.00 | 0.05 | 0.66 | 0.90 | 0.93 | 0.98 | 0.92 | 1.00 | 1.00 | 1.00 |
| 2006 | 0.00 | 0.04 | 0.59 | 0.80 | 0.99 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2007 | 0.00 | 0.00 | 0.47 | 0.78 | 0.91 | 0.99 | 0.97 | 1.00 | 1.00 | 1.00 |
| 2008 | 0.00 | 0.10 | 0.78 | 0.91 | 0.90 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2009 | 0.00 | 0.09 | 0.61 | 0.81 | 0.96 | 0.94 | 0.96 | 1.00 | 1.00 | 1.00 |
| 2010 | 0.00 | 0.08 | 0.61 | 0.77 | 0.94 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2011 | 0.00 | 0.06 | 0.51 | 0.69 | 0.84 | 0.93 | 0.98 | 1.00 | 1.00 | 1.00 |
| 2012 | 0.00 | 0.00 | 0.63 | 0.85 | 0.94 | 0.97 | 1.00 | 1.00 | 1.00 | 0.83 |
| 2013 | 0.00 | 0.24 | 0.82 | 0.95 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2014 | 0.00 | 0.24 | 0.73 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |

Table 4.2.9. Faroe Plateau cod (sub-division Vb1). Summer survey tuning series (number of individuals per 200 stations) and spring survey tuning series (number of individuals per 100 stations) used as tuning series in the XSA model.


SPRING SURVEY (shifted back to december)

19932014
110.91 .0

18

| 100 | 612.5 | 336.9 | 912.8 | 508.5 | 129.7 | 187.2 | 28.6 | 0.1 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 100 | 623.2 | 845.7 | 1528.4 | 1525.2 | 1191.4 | 285.6 | 350.8 | 48.9 |
| 100 | 215.5 | 4043.9 | 3984.4 | 1892.1 | 1372 | 420.8 | 82.8 | 169.7 |
| 100 | 72.5 | 834.4 | 5398.3 | 2359.5 | 333.9 | 227 | 58.8 | 5.3 |


| 100 | 69.7 | 425.2 | 1572.1 | 4919.3 | 1136 | 82.3 | 40.7 | 35.2 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 100 | 704.7 | 674.9 | 991.3 | 1225.2 | 2079.2 | 252.1 | 25.2 | 13.4 |
| 100 | 316 | 1432.4 | 746.1 | 441 | 506.7 | 836.7 | 63.8 | 3.1 |
| 100 | 938.4 | 2387.8 | 1993.8 | 456.2 | 324.4 | 578.6 | 128.6 | 3.9 |
| 100 | 383 | 4564.1 | 2892.1 | 1579.7 | 331.9 | 231.8 | 178.9 | 131.9 |
| 100 | 90.2 | 719 | 3915 | 1260.4 | 528.7 | 67.4 | 51.7 | 39.7 |
| 100 | 609.5 | 575.8 | 844.6 | 1175.1 | 292.9 | 66 | 22.2 | 11.9 |
| 100 | 383.1 | 438.2 | 1151.7 | 1440.2 | 844.5 | 140.6 | 14 | 3.8 |
| 100 | 167.5 | 156.7 | 177.3 | 360.1 | 292 | 95 | 15.5 | 4 |
| 100 | 41.1 | 270.9 | 286.6 | 155.2 | 170.4 | 105.1 | 37.8 | 14.4 |
| 100 | 176.6 | 474.5 | 851.9 | 479.2 | 151.5 | 83.9 | 39.4 | 13.3 |
| 100 | 307.8 | 475.5 | 977.7 | 1159.1 | 427.3 | 73.7 | 31.6 | 24.9 |
| 100 | 697.6 | 1318.8 | 745.6 | 538.1 | 381 | 98.9 | 41 | 17.2 |
| 100 | 148.4 | 1319 | 1240.3 | 562.4 | 300.2 | 237.8 | 85.2 | 21.9 |
| 100 | 41.1 | 273.8 | 1303.8 | 326.7 | 73.6 | 27 | 23.7 | 6.2 |
| 100 | 68 | 377.6 | 1699.8 | 2053.2 | 295.6 | 32.6 | 22.4 | 17.7 |
| 100 | 130.9 | 113.4 | 159.6 | 419.7 | 333 | 74.8 | 22 | 13.6 |
| 100 | 22.4 | 533.3 | 225.6 | 193.9 | 305.2 | 138.9 | 32.6 | 8 |

Table 4.2.10. Faroe Plateau cod (sub-division Vb1). Pair trawler abundance index (number of individuals per 1000 fishing hours). This series was not used in the tuning of the XSA. The season is June - December. The otoliths are selected from deep ( $>150 \mathrm{~m}$ ) locations.

|  | Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| 1989 | 1200 | 1638 | 1783 | 1381 | 928 | 719 | 297 | 194 |
| 1990 | 116 | 2856 | 2057 | 834 | 465 | 419 | 200 | 0 |
| 1991 | 8 | 148 | 1401 | 869 | 329 | 225 | 65 | 93 |
| 1992 | 84 | 487 | 696 | 1234 | 760 | 353 | 129 | 62 |
| 1993 | 51 | 1081 | 2192 | 746 | 1062 | 398 | 67 | 107 |
| 1994 | 1314 | 2129 | 1457 | 2208 | 697 | 1241 | 461 | 53 |
| 1995 | 577 | 3645 | 5178 | 4199 | 2769 | 543 | 539 | 106 |
| 1996 | 242 | 10608 | 16683 | 7985 | 4410 | 194 | 0 | 723 |
| 1997 | 28 | 674 | 6038 | 9375 | 2413 | 944 | 113 | 0 |
| 1998 | 80 | 731 | 1805 | 5941 | 4904 | 801 | 286 | 0 |
| 1999 | 444 | 2082 | 1933 | 3008 | 5136 | 2220 | 218 | 4 |
| 2000 | 3478 | 3956 | 1737 | 956 | 1003 | 1694 | 382 | 0 |
| 2001 | 3385 | 6700 | 3009 | 555 | 415 | 797 | 862 | 25 |
| 2002 | 571 | 6409 | 5019 | 1235 | 432 | 400 | 41 | 228 |
| 2003 | 63 | 1341 | 4450 | 3630 | 870 | 270 | 152 | 145 |
| 2004 | 23 | 0 | 278 | 2534 | 2831 | 1733 | 274 | 184 |
| 2005 | 42 | 399 | 655 | 1766 | 2171 | 860 | 148 | 70 |
| 2006 | 93 | 135 | 699 | 755 | 1580 | 612 | 787 | 71 |
| 2007 | 64 | 916 | 1767 | 1392 | 802 | 656 | 206 | 46 |
| 2008 | 54 | 295 | 418 | 573 | 387 | 456 | 487 | 182 |
| 2009 | 11 | 734 | 801 | 756 | 448 | 247 | 147 | 105 |
| 2010 | 1578 | 2917 | 1787 | 543 | 603 | 190 | 0 | 81 |
| 2011 | 22 | 1487 | 4078 | 1967 | 622 | 441 | 95 | 25 |
| 2012 | 0 | 95 | 1531 | 1789 | 950 | 223 | 40 | 107 |
| 2013 | 35 | 102 | 761 | 1583 | 670 | 103 | 57 | 36 |
| 2014 | 292 | 1631 | 1006 | 1690 | 1812 | 477 | 94 | 101 |

Table 4.2.11. Faroe Plateau cod (sub-division Vb1). Longliner abundance index (number of individuals per 100000 hooks). This series was not used in the tuning of the XSA. The age composition was obtained from all longliners > $\mathbf{1 0 0}$ GRT. The area was restricted to the area west of Faroe Islands at depths between 100 and 200 m .

|  | Age |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ |
| $\mathbf{1 9 9 3}$ | 405 | 2610 | 9306 | 3330 | 806 | 2754 | 847 | 258 |
| $\mathbf{1 9 9 4}$ | 101 | 8105 | 14105 | 7863 | 4659 | 962 | 1187 | 71 |
| 1995 | 0 | 15249 | 23062 | 2895 | 2505 | 1568 | 708 | 1073 |
| 1996 | 0 | 2269 | 18658 | 13265 | 4153 | 8435 | 4513 | 1147 |
| 1997 | 0 | 1738 | 5837 | 26368 | 18089 | 2805 | 2807 | 402 |
| 1998 | 1892 | 4490 | 2025 | 2565 | 11738 | 2732 | 131 | 19 |
| 1999 | 849 | 10968 | 3811 | 985 | 1891 | 3759 | 548 | 109 |
| 2000 | 2695 | 10983 | 6710 | 998 | 780 | 1473 | 2136 | 109 |
| 2001 | 287 | 12999 | 7409 | 2660 | 515 | 1135 | 1808 | 2545 |
| 2002 | 105 | 6862 | 20902 | 10819 | 7759 | 1561 | 1945 | 1265 |
| 2003 | 16 | 2099 | 6057 | 15910 | 7778 | 1830 | 708 | 650 |
| 2004 | 59 | 510 | 1773 | 2438 | 3214 | 1059 | 293 | 71 |
| 2005 | 297 | 2169 | 1543 | 2313 | 2327 | 1360 | 170 | 13 |
| 2006 | 151 | 5813 | 5319 | 674 | 2205 | 2352 | 1148 | 56 |
| 2007 | 274 | 3578 | 6383 | 2778 | 1927 | 1159 | 1118 | 134 |
| 2008 | 1270 | 2243 | 4449 | 4773 | 2564 | 1133 | 816 | 716 |
| 2009 | 294 | 2670 | 15107 | 6308 | 3028 | 2491 | 683 | 132 |
| 2010 | 23 | 20287 | 16914 | 8733 | 2595 | 4780 | 1878 | 864 |
| 2011 | 160 | 2817 | 28218 | 14391 | 4295 | 2207 | 1252 | 195 |
| 2012 | 0 | 1833 | 9562 | 8309 | 2364 | 1296 | 403 | 197 |
| 2013 | 0 | 52 | 209 | 2887 | 5132 | 2654 | 1222 | 359 |
| 2014 | 93 | 5898 | 9602 | 4695 | 4398 | 3475 | 1289 | 116 |
|  |  |  |  |  |  |  |  |  |

Table 4.2.12. Longliner abundance index (number of individuals per day) for longliners < 25 GRT operating mainly near shore. This series was not used in the tuning of the XSA. The age composition was obtained from all longliners.

|  | Age |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ |
| $\mathbf{1 9 8 3}$ | 0.9 | 7.5 | 4.7 | 3.8 | 1.6 | 0.9 | 0.5 | 0.2 |
| $\mathbf{1 9 8 4}$ | 0.0 | 33.3 | 32.1 | 13.2 | 5.8 | 6.3 | 1.0 | 0.7 |
| 1985 | 0.0 | 3.7 | 50.1 | 35.0 | 25.3 | 14.1 | 19.6 | 5.8 |
| 1986 | 0.0 | 5.6 | 41.6 | 24.0 | 15.3 | 6.8 | 6.2 | 2.2 |
| 1987 | 0.0 | 6.8 | 11.3 | 16.6 | 27.5 | 12.4 | 5.3 | 0.9 |
| 1988 | 0.0 | 3.1 | 6.4 | 13.0 | 8.5 | 19.1 | 6.5 | 2.6 |
| 1989 | 0.1 | 43.7 | 21.3 | 20.5 | 13.9 | 7.5 | 16.1 | 2.2 |
| 1990 | 0.0 | 7.9 | 40.3 | 8.6 | 12.2 | 6.5 | 7.7 | 4.2 |
| 1991 | 0.0 | 0.0 | 5.2 | 27.0 | 8.7 | 3.9 | 2.4 | 0.7 |
| 1992 | 0.0 | 6.2 | 17.1 | 6.9 | 3.9 | 3.6 | 1.8 | 1.4 |
| 1993 | 0.4 | 4.6 | 19.2 | 7.3 | 1.4 | 1.3 | 0.3 | 1.3 |
| 1994 | 0.1 | 14.9 | 18.4 | 15.4 | 6.6 | 2.1 | 2.6 | 0.5 |
| 1995 | 0.0 | 53.6 | 47.8 | 12.2 | 8.4 | 5.1 | 2.0 | 3.1 |
| 1996 | 0.0 | 5.9 | 76.2 | 52.1 | 13.1 | 28.8 | 14.3 | 4.2 |
| 1997 | 0.0 | 4.6 | 16.6 | 71.8 | 54.5 | 7.9 | 7.6 | 0.9 |
| 1998 | 5.8 | 12.1 | 5.6 | 8.2 | 33.1 | 9.9 | 0.4 | 0.4 |
| 1999 | 0.3 | 29.2 | 10.0 | 4.7 | 7.0 | 15.9 | 2.5 | 0.1 |
| 2000 | 9.6 | 40.4 | 23.5 | 1.3 | 1.3 | 2.4 | 4.2 | 0.5 |
| 2001 | 0.6 | 96.6 | 48.7 | 17.1 | 3.0 | 5.7 | 12.6 | 12.9 |
| 2002 | 0.1 | 47.6 | 97.2 | 43.4 | 30.0 | 7.3 | 11.5 | 6.8 |
| 2003 | 0.0 | 17.5 | 37.4 | 106.4 | 59.1 | 12.9 | 4.1 | 1.5 |
| 2004 | 0.0 | 7.0 | 21.5 | 21.0 | 31.1 | 8.2 | 0.3 | 0.0 |
| 2005 | 0.6 | 14.7 | 20.5 | 18.5 | 32.9 | 15.6 | 1.5 | 0.0 |
| 2006 | 2.0 | 58.7 | 47.0 | 9.1 | 10.6 | 13.6 | 4.1 | 0.4 |
| 2007 | 0.2 | 11.2 | 23.2 | 8.9 | 4.2 | 4.9 | 3.5 | 0.6 |
| 2008 | 0.3 | 3.4 | 16.2 | 21.1 | 14.4 | 3.3 | 1.5 | 2.1 |
| 2009 | 3.1 | 33.3 | 154.6 | 57.5 | 33.9 | 23.5 | 9.6 | 5.9 |
| 2010 | 2.6 | 135.7 | 147.1 | 62.4 | 27.3 | 28.5 | 8.5 | 1.8 |
| 2011 | 0.0 | 19.7 | 156.5 | 65.0 | 25.2 | 15.6 | 8.5 | 1.9 |
| 2012 | 0.3 | 4.6 | 39.3 | 59.0 | 15.1 | 5.2 | 2.6 | 1.3 |
| 2013 | 1.2 | 16.6 | 23.8 | 63.6 | 58.0 | 7.8 | 2.9 | 0.0 |
| 2014 | 2.1 | 103.4 | 102.0 | 46.9 | 27.3 | 17.1 | 1.4 | 0.0 |
|  |  |  |  |  |  |  |  |  |

Table 4.6.1. Faroe Plateau cod (sub-division Vb1). The XSA-run.

Lowestoft VPA Version 3.1


XSA population numbers (Thousands)

$2005,9.31 \mathrm{E}+03,6.10 \mathrm{E}+03,2.88 \mathrm{E}+03,3.06 \mathrm{E}+03,3.30 \mathrm{E}+03,1.69 \mathrm{E}+03,3.95 \mathrm{E}+02,1.04 \mathrm{E}+02,3.06 \mathrm{E}+01$, 2006 , $\quad 6.25 \mathrm{E}+03,7.62 \mathrm{E}+03,4.54 \mathrm{E}+03,1.82 \mathrm{E}+03,1.71 \mathrm{E}+03,1.68 \mathrm{E}+03,6.37 \mathrm{E}+02,1.39 \mathrm{E}+02,4.79 \mathrm{E}+01$, $2007,7.95 \mathrm{E}+03,5.12 \mathrm{E}+03,5.17 \mathrm{E}+03,2.66 \mathrm{E}+03,1.04 \mathrm{E}+03,7.60 \mathrm{E}+02,6.07 \mathrm{E}+02,2.00 \mathrm{E}+02,4.07 \mathrm{E}+01$ $2008,1.03 \mathrm{E}+04,6.51 \mathrm{E}+03,3.70 \mathrm{E}+03,3.05 \mathrm{E}+03,1.48 \mathrm{E}+03,5.49 \mathrm{E}+02,3.43 \mathrm{E}+02,2.50 \mathrm{E}+02,8.15 \mathrm{E}+01$ $2009,1.50 \mathrm{E}+04,8.43 \mathrm{E}+03,5.06 \mathrm{E}+03,2.33 \mathrm{E}+03,1.77 \mathrm{E}+03,8.17 \mathrm{E}+02,2.77 \mathrm{E}+02,1.36 \mathrm{E}+02,6.09 \mathrm{E}+01$, $2010, \quad 5.06 \mathrm{E}+03,1.23 \mathrm{E}+04,6.11 \mathrm{E}+03,2.09 \mathrm{E}+03,1.13 \mathrm{E}+03,8.91 \mathrm{E}+02,4.00 \mathrm{E}+02,1.50 \mathrm{E}+02,6.17 \mathrm{E}+01$ $2011,2.17 \mathrm{E}+03,4.14 \mathrm{E}+03,8.16 \mathrm{E}+03,3.16 \mathrm{E}+03,9.34 \mathrm{E}+02,4.99 \mathrm{E}+02,2.95 \mathrm{E}+02,1.67 \mathrm{E}+02,7.00 \mathrm{E}+01$,
$2012, \quad 3.00 \mathrm{E}+03,1.77 \mathrm{E}+03,3.09 \mathrm{E}+03,4.54 \mathrm{E}+03,1.46 \mathrm{E}+03,4.33 \mathrm{E}+02,2.38 \mathrm{E}+02,1.26 \mathrm{E}+02,9.13 \mathrm{E}+01$,
2013 , $\quad 8.27 \mathrm{E}+03,2.45 \mathrm{E}+03,1.41 \mathrm{E}+03,2.07 \mathrm{E}+03,2.50 \mathrm{E}+03,6.94 \mathrm{E}+02,1.72 \mathrm{E}+02,1.05 \mathrm{E}+02,4.09 \mathrm{E}+01$,,
$\begin{array}{ll}2013, & 8.27 \mathrm{E}+03,2.45 \mathrm{E}+03,1.41 \mathrm{E}+03,2.07 \mathrm{E}+03,2.50 \mathrm{E}+03,6.94 \mathrm{E}+02,1.72 \mathrm{E}+02,1.05 \mathrm{E}+02,4.09 \mathrm{E}+01, \\ 2014, & 1.07 \mathrm{E}+03,6.77 \mathrm{E}+03,1.96 \mathrm{E}+03,9.90 \mathrm{E}+02,1.38 \mathrm{E}+03,1.50 \mathrm{E}+03,4.04 \mathrm{E}+02,1.05 \mathrm{E}+02,6.23 \mathrm{E}+01,\end{array}$

Estimated population abundance at 1st Jan 2015
$0.00 \mathrm{E}+00,8.74 \mathrm{E}+02,5.16 \mathrm{E}+03,1.08 \mathrm{E}+03,5.23 \mathrm{E}+02,6.28 \mathrm{E}+02, \quad 8.73 \mathrm{E}+02,2.45 \mathrm{E}+02,7.24 \mathrm{E}+01$,
Taper weighted geometric mean of the VPA populations:
$1.39 \mathrm{E}+04,1.19 \mathrm{E}+04,8.96 \mathrm{E}+03,5.56 \mathrm{E}+03,3.10 \mathrm{E}+03,1.52 \mathrm{E}+03,6.74 \mathrm{E}+02,2.75 \mathrm{E}+02,1.11 \mathrm{E}+02$, Standard error of the weighted Log (VPA populations) :


| Age | , | 1995, | 1996, | 1997, | 1998, | 1999, | 2000, | 2001, | 2002, | 2003, | 2004 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 |  | No data | for t | is fle | et at t | his age |  |  |  |  |  |
| 2 | , | 99.99, | -. 16, | . 21 , | . 35 , | -.87, | .13, | . 66, | 1.10, | -. 07, | . 62 |
| 3 | , | 99.99, | .10, | -.25, | -. 63, | .49, | -.45, | . 04 , | . 57 , | -.39, | . 01 |
| 4 | , | 99.99, | .19, | . 31 , | -. 60, | -.13, | .07, | .10, | .09, | .10, | -. 20 |
| 5 | , | 99.99, | . 65 , | -.08, | . 23 , | -.71, | -. 80, | -. 12, | .11, | -. 35, | . 43 |
| 6 | , | 99.99, | . 14 , | -. 21 , | . 59, | .11, | -. 65, | -.59, | -.35, | -. 73, | . 26 |
| 7 | , | 99.99, | . 27 , | -.05, | -.39, | . 53, | . 04 , | -. 32 , | -.42, | -1.40, | . 09 |
| 8 | , | 99.99, | -. 15, | -.29, | .10, | . 40 , | -. 24 , | -.05, | -.47, | -1.07, | . 19 |
| Age | , | 2005, | 2006, | 2007, | 2008, | 2009, | 2010, | 2011, | 2012, | 2013, | 2014 |
| 1 |  | No data | for t | s fle | et at t | his age |  |  |  |  |  |
| 2 | , | . 49, | . 79, | -.29, | -1.80, | -.18, | . 56, | .16, | -1.65, | -. 45, | . 41 |
| 3 | , | . 37 , | -.11, | -.66, | -.56, | 1.08, | .49, | . 04 , | -.89, | -. 25 , | 1.00 |
| 4 | , | . 21 , | -. 20 , | -.67, | -.79, | . 51, | . 89 , | . 52, | . 04 , | -.94, | . 49 |
| 5 | , | . 25 , | -. 32, | -.49, | -.07, | . 13, | . 25 , | .45, | . 28 , | .15, | . 01 |
| 6 | , | . 65 , | -. 40 , | -.40, | .03, | .55, | . 24 , | .16, | -. 17, | .92, | -. 15 |
| 7 | , | .47, | -.08, | -.69, | -.46, | . 48 , | . 40 , | . 33, | -.55, | .63, | -. 04 |
| 8 | , | . 48 , | .02, | -.47, | -.47, | . 28 , | .03, | -.29, | -.32, | . 28 , | . 00 |

Mean log catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

| Age, | 2, | 3, | 4, | 5, | 6, | 7, |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log q, | -7.8857, | -6.7379, | -6.3878, | -6.1480, | -6.1144, | -6.1144, |
| S.E (Log q), | .7684, | .5560, | .4813, | .3902, | .4676, | .5196, |

Regression statistics :

Ages with $q$ independent of year class strength and constant w.r.t. time.
Age, Slope, t-value , Intercept, RSquare, No Pts, Reg s.e, Mean Q

| 2, | .70, | 1.742, | 8.19, | .66, | 19, | .51, | -7.89, |
| :--- | ---: | ---: | ---: | :--- | :--- | :--- | :--- |
| 3, | .99, | .059, | 6.76, | .67, | 19, | .57, | -6.74, |
| 4, | .99, | .067, | 6.41, | .73, | 19, | .49, | -6.39, |
| 5, | .98, | .111, | 6.17, | .76, | 19, | .40, | -6.15, |
| 6, | 1.04, | -.209, | 6.08, | .64, | 19, | .50, | -6.11, |
| 7, | 1.04, | -.209, | 6.17, | .60, | 19, | .55, | -6.17, |
| 8, | 1.31, | -1.697, | 6.53, | .64, | 19, | .47, | -6.22, |

Fleet : SPRING SURVEY (shift

| Age, | 1993, | 1994 |
| ---: | ---: | ---: |
| 1, | -.05, | -.56 |
| 2, | -.91, | -.89 |
| 3, | -.65, | -.01 |
| 4, | -.58, | -.02 |
| 5, | -.57, | .76 |
| 6, | -.63, | .92 |
| 7, | -.32, | .35 |
| 8, | -4.57, | .74 |


| Age |  | 1995, | 1996, | 1997, | 1998, | 1999, | 2000, | 2001, | 2002, | 2003, | 2004 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | , | -. 42 , | -.82, | -.78, | . 65 , | -. 47 , | .21, | .12, | -. 57, | 1.88, | . 90 |
| 2 | , | .19, | -. 23, | -.21, | . 39 , | .27, | .49, | . 76 , | -. 26 , | . 22 , | 39 |
| 3 | , | .10, | -.05, | -.17, | .09, | . 05 , | .19, | .29, | . 36, | -. 51, | . 38 |
| 4 | , | .58, | -.06, | .19, | -.23, | -. 51, | -.14, | . 33, | -.03, | -.27, | . 24 |
| 5 | , | . 39, | -. 12, | . 26 , | . 21, | -. 55, | -.33, | .09, | . 28 , | -.41, | . 39 |
| 6 | , | . 55, | -.06, | -.01, | . 28 , | . 44 , | . 38, | .15, | -.23, | -.43, | . 32 |
| 7 | , | .21, | -.11, | -. 20, | -. 18, | . 20 , | -.68, | .09, | .17, | -. 24, | -. 66 |
| 8 |  |  |  |  |  |  |  |  |  |  |  |

Age , 2005, 2006, 2007, 2008, 2009, 2010, 2011, 2012, 2013, 2014
$1,-.15,-1.16,-.06, \quad .35, \quad .79, \quad .33,-.10, \quad .08,-.28, \quad .00$

| 2, | -.15, | -1.16, | .06, | .35, | .79, | .33, |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| .- .10, | .08, | -.28, | .00 |  |  |  |


| 2, | -1.09, | -.68, | .22, | -.09, | .74, | .45, | -.15, |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 3, | -.98, | -.89, | .07, | .48, | .29, | .40, | .09, |
| 1.16, | -.47, | -.23 |  |  |  |  |  |


| 4, | -.49, | -.84, | -.06, | .64, | .32, | .55, | -.44, |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 5, | -.65, | -.40, | -.18, | .46, | .25, | -.11, | .07 |


| 5, | -.65, | -.40, | -.18, | .46, | .25, | .59, | -.68, | .24, |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 6, | -.54, | -.40, | -.04, | .05, | -.02, | 1.14, | -.80, | -.30, |
| 7 | -.30, | -.46 |  |  |  |  |  |  |
| 7 |  |  |  |  |  |  |  |  |


| -.84, | -.32, | -.48, | -.10, | .08, | .69, | -.31, | -.18, |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| , | -.18, | -.63 |  |  |  |  |  |
| .31, | -.45, | .44, | .09, | .21, | -1.31, | .51, | -.14, |
| -.81 |  |  |  |  |  |  |  |


| Age, | 1, | 2, | 3, | 4, | 5, | 6, | 7, |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log q, | -8.2792, | -6.8938, | -5.9763, | -5.7006, | -5.7402, | -6.0145, | -6.0145, |
| S.E (Log q), | .6722, | .5682, | .4868, | .4368, | .4385, | .4933, | .4069, |

Regression statistics :

Ages with $q$ independent of year class strength and constant w.r.t. time
Age, Slope, t-value, Intercept, RSquare, No Pts, Reg s.e, Mean Q

| 1, | 1.08, | -.456, | 8.21, | .62, | 22, | .74, | -8.28, |
| :--- | ---: | ---: | :--- | :--- | :--- | :--- | :--- |
| 2, | 1.02, | -.092, | 6.86, | .65, | 22, | .59, | -6.89, |
| 3, | .92, | .582, | 6.19, | .75, | 22, | .46, | -5.98, |
| 4, | .94, | .481, | 5.86, | .76, | 22, | .42, | -5.70, |
| 5, | .92, | .584, | 5.90, | .74, | 22, | .41, | -5.74, |
| 6, | .90, | .662, | 6.13, | .67, | 22, | .45, | -6.01, |
| 7, | .95, | .420, | 6.18, | .78, | 22, | .36, | -6.18, |
| 8, | .63, | 1.649, | 6.02, | .50, | 22, | .71, | -6.47, |

Terminal year survivor and $F$ summaries :
Age 1 Catchability constant w.r.t. time and dependent on age
Year class $=2013$

| Fleet, | Estimated, | Int, | Ext, | Var, | N, Scaled, | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors, | s.e, | s.e, | Ratio, | , Weights, |  |
| SUMMER SURVEY , | 1., | . 000 , | .000, | .00, | 0, .000, | . 000 |
| SPRING SURVEY (shift, | 874., | .687, | .000, | .00, | 1, 1.000, | . 000 |
| F shrinkage mean | 0., | 2.00, |  |  | .000, | . 000 |

Weighted prediction :
Survivors, Int, Ext, N, Var, F
$\begin{array}{ccccc}\text { at end of year, s.e, s.e, } & \text { Ratio, } & \\ 874 ., & .69, & .00, & 1, & .000,\end{array}$

Age 2 Catchability constant w.r.t. time and dependent on age
Year class = 2012


Weighted prediction :
Survivors, Int, Ext, N, Var, F
$\begin{array}{cccc}\text { at end of year, s.e, s.e, } & \text { Ratio, } & \\ 5155 ., & .38, & .15, & 4, \\ .392, & .073\end{array}$

Age 3 Catchability constant w.r.t. time and dependent on age
Year class $=2011$


Age 4 Catchability constant w.r.t. time and dependent on age
Year class $=2010$

| Fleet, | Estimated, Survivors, | $\begin{aligned} & \text { Int, } \\ & \text { s.e, } \end{aligned}$ | $\begin{aligned} & \text { Ext, } \\ & \text { s.e, } \end{aligned}$ | Var, Ratio, |  | Scaled, Weights, | $\underset{F}{\text { Estimate }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SUMMER SURVEY | 469., | . 339 , | . 538, | 1.59, | 3, | . 382 , | . 479 |
| SPRING SURVEY (shift, | 562., | . 267 , | .282, | 1.06, | 4, | .600, | . 413 |
| F shrinkage mean | 487., | 2.00, |  |  |  | .018, | . 464 |

Weighted prediction :
Survivors, Int, Ext, N, Var, F

```
at end of year, s.e, S.e, %, Ratio, 
Age 5 Catchability constant w.r.t. time and dependent on age
Year class = 2009
\begin{tabular}{|c|c|c|c|c|c|c|c|}
\hline Fleet, & Estimated, Survivors, & \[
\begin{aligned}
& \text { Int, } \\
& \text { s.e, }
\end{aligned}
\] & \[
\begin{aligned}
& \text { Ext, } \\
& \text { s.e, }
\end{aligned}
\] & Var, Ratio, & & Scaled, Weights, & \[
\underset{F}{\text { Estimate }}
\] \\
\hline SUMMER SURVEY & 432., & . 262 , & .273, & 1.04, & 4, & . 452 , & . 771 \\
\hline SPRING SURVEY (shift, & 860., & .233, & .227, & . 97 , & 5, & .532, & . 459 \\
\hline F shrinkage mean & 754., & 2.00, & & & & .016, & . 510 \\
\hline
\end{tabular}
Weighted prediction :
Survivors, Int, Ext, N, Var, F
at end of year, s.e, S.e, 年, Ratio, 
```

Age 6 Catchability constant w.r.t. time and dependent on age
Year class $=2008$

Age 7 Catchability constant w.r.t. time and age (fixed at the value for age) 6
Year class $=2007$


Age 8 Catchability constant w.r.t. time and age (fixed at the value for age) 6
Year class $=2006$


Age 9 Catchability constant w.r.t. time and age (fixed at the value for age) 6
Year class $=2005$


Table 4.6.2. Faroe Plateau cod (sub-division Vb1). Fishing mortality at age from the XSA model.

|  | Age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | FBAR 3- |
| 1959 | 0.1829 | 0.4853 | 0.4463 | 0.6303 | 0.3909 | 0.6060 | 0.3005 | 0.4784 | 0.4784 | 0.5117 |
| 1960 | 0.4570 | 0.6793 | 0.6222 | 0.5290 | 0.7826 | 0.6920 | 1.0328 | 0.7389 | 0.7389 | 0.6610 |
| 1961 | 0.3346 | 0.5141 | 0.4986 | 0.5737 | 0.4863 | 0.9566 | 0.8116 | 0.6715 | 0.6715 | 0.6059 |
| 1962 | 0.2701 | 0.4982 | 0.4838 | 0.7076 | 0.5569 | 0.3662 | 0.6826 | 0.5641 | 0.5641 | 0.5226 |
| 1963 | 0.2534 | 0.4138 | 0.5172 | 0.5124 | 0.5405 | 0.4879 | 0.3269 | 0.4806 | 0.4806 | 0.4944 |
| 1964 | 0.1086 | 0.2997 | 0.4523 | 0.5229 | 0.5659 | 0.6677 | 0.3531 | 0.5164 | 0.5164 | 0.5017 |
| 1965 | 0.1209 | 0.2518 | 0.4498 | 0.5622 | 0.6604 | 0.5305 | 0.4345 | 0.5318 | 0.5318 | 0.4909 |
| 1966 | 0.0829 | 0.1969 | 0.2552 | 0.4499 | 0.5016 | 0.9680 | 0.8520 | 0.6106 | 0.6106 | 0.4743 |
| 1967 | 0.0789 | 0.2389 | 0.2687 | 0.3442 | 0.5779 | 0.5203 | 1.0438 | 0.5556 | 0.5556 | 0.3900 |
| 1968 | 0.1010 | 0.2318 | 0.3949 | 0.5339 | 0.4472 | 0.7132 | 0.3331 | 0.4882 | 0.4882 | 0.4642 |
| 1969 | 0.1099 | 0.3063 | 0.3806 | 0.4180 | 0.5709 | 0.5118 | 0.8457 | 0.5499 | 0.5499 | 0.4375 |
| 1970 | 0.0530 | 0.2081 | 0.3654 | 0.3409 | 0.3709 | 0.6559 | 0.4208 | 0.4339 | 0.4339 | 0.3882 |
| 1971 | 0.0309 | 0.1337 | 0.2225 | 0.3845 | 0.5572 | 0.4651 | 0.7528 | 0.4800 | 0.4800 | 0.3526 |
| 1972 | 0.0464 | 0.1476 | 0.2070 | 0.2497 | 0.6058 | 0.4686 | 0.2464 | 0.3578 | 0.3578 | 0.3358 |
| 1973 | 0.0657 | 0.2322 | 0.3048 | 0.2813 | 0.2526 | 0.3722 | 0.3259 | 0.3091 | 0.3091 | 0.2886 |
| 1974 | 0.0816 | 0.1568 | 0.2046 | 0.2953 | 0.3797 | 0.5330 | 0.3052 | 0.3457 | 0.3457 | 0.3139 |
| 1975 | 0.0774 | 0.3193 | 0.4359 | 0.4134 | 0.4544 | 0.3504 | 0.4485 | 0.4235 | 0.4235 | 0.3947 |
| 1976 | 0.0933 | 0.1723 | 0.3665 | 0.5568 | 0.5167 | 0.7619 | 0.6429 | 0.5738 | 0.5738 | 0.4749 |
| 19 | 0.0481 | 0.303 | 0.4748 | 0.7532 | 0.7333 | 1.1138 | 0.7776 | 0.7783 | 0.7783 | 0.6757 |
| 1978 | 0.0588 | 0.1896 | 0.4291 | 0.4289 | 0.4850 | 0.5968 | 0.5674 | 0.5054 | 0.5054 | 0.4259 |
| 1979 | 0.0433 | 0.2623 | 0.4309 | 0.5049 | 0.4906 | 0.4480 | 0.6903 | 0.5170 | 0.5170 | 0.4273 |
| 1980 | 0.0544 | 0.2391 | 0.3695 | 0.4337 | 0.5182 | 0.4119 | 0.6437 | 0.4790 | 0.4790 | 0.3945 |
| 1981 | 0.0523 | 0.2877 | 0.3409 | 0.4369 | 0.5644 | 0.6940 | 0.5015 | 0.5115 | 0.5115 | 0.4648 |
| 1982 | 0.0586 | 0.2227 | 0.3602 | 0.3887 | 0.4047 | 0.6926 | 0.5526 | 0.4834 | 0.4834 | 0.4138 |
| 1983 | 0.0991 | 0.4672 | 0.5585 | 0.6411 | 0.7835 | 1.0779 | 0.9416 | 0.8087 | 0.8087 | 0.7056 |
| 1984 | 0.1073 | 0.3711 | 0.5790 | 0.6609 | 0.4533 | 0.4761 | 0.4791 | 0.5340 | 0.5340 | 0.5081 |
| 1985 | 0.0658 | 0.3543 | 0.5075 | 0.6134 | 0.9234 | 1.1081 | 1.3203 | 0.9042 | 0.9042 | 0.7013 |
| 1986 | 0.0247 | 0.3544 | 0.6225 | 0.7030 | 0.8256 | 0.8399 | 0.5407 | 0.7131 | 0.7131 | 0.6691 |
| 1987 | 0.0291 | 0.2208 | 0.4753 | 0.4849 | 0.5555 | 0.4895 | 0.6221 | 0.5297 | 0.5297 | 0.4452 |
| 1988 | 0.0666 | 0.3530 | 0.5637 | 0.5489 | 0.7732 | 0.7979 | 0.8639 | 0.7163 | 0.7163 | 0.6073 |
| 1989 | 0.1633 | 0.4395 | 0.7614 | 0.7614 | 0.9611 | 1.0566 | 1.0988 | 0.9381 | 0.9381 | 0.7960 |
| 1990 | 0.0778 | 0.3287 | 0.6376 | 0.8014 | 0.7129 | 0.8504 | 1.1337 | 0.8358 | 0.8358 | 0.6662 |
| 1991 | 0.0324 | 0.1990 | 0.4365 | 0.5987 | 0.7459 | 0.5797 | 0.7153 | 0.6207 | 0.6207 | 0.5120 |
| 1992 | 0.0201 | 0.1001 | 0.3256 | 0.3326 | 0.6381 | 0.8909 | 0.4433 | 0.5304 | 0.5304 | 0.4575 |
| 1993 | 0.0132 | 0.1020 | 0.1868 | 0.2535 | 0.1912 | 0.4421 | 0.5779 | 0.3325 | 0.3325 | 0.2351 |
| 1994 | 0.0255 | 0.1129 | 0.1907 | 0.2501 | 0.2212 | 0.1481 | 0.3228 | 1.0965 | 1.0965 | 0.1846 |
| 1995 | 0.0704 | 0.1619 | 0.4651 | 0.2805 | 0.3615 | 0.3361 | 0.2156 | 0.7443 | 0.7443 | 0.3210 |
| 1996 | 0.0306 | 0.1935 | 0.4530 | 0.8107 | 0.9071 | 1.1451 | 0.9361 | 0.8738 | 0.8738 | 0.7019 |
| 1997 | 0.0348 | 0.1489 | 0.4138 | 0.8362 | 1.0504 | 1.4087 | 1.3725 | 1.0695 | 1.0695 | 0.7716 |
| 1998 | 0.0887 | 0.1760 | 0.2732 | 0.6530 | 1.0628 | 0.7973 | 1.1943 | 0.8977 | 0.8977 | 0.5925 |
| 1999 | 0.0958 | 0.2841 | 0.2904 | 0.3183 | 0.6678 | 1.0916 | 0.8046 | 0.5191 | 0.5191 | 0.5304 |
| 2000 | 0.1247 | 0.3191 | 0.3799 | 0.2477 | 0.3268 | 0.5479 | 0.8501 | 0.1974 | 0.1974 | 0.3643 |
| 2001 | 0.1574 | 0.3448 | 0.4554 | 0.3078 | 0.3506 | 0.6988 | 0.6566 | 0.8359 | 0.8359 | 0.4315 |


| 2002 | 0.1903 | 0.4904 | 0.5998 | 0.8219 | 0.8296 | 1.3662 | 1.2399 | 1.3959 | 1.3959 | 0.8216 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2003 | 0.1279 | 0.3039 | 0.6642 | 0.8523 | 0.9072 | 0.9026 | 0.9391 | 1.8019 | 1.8019 | 0.7260 |
| 2004 | 0.0309 | 0.1862 | 0.2977 | 0.7556 | 0.9873 | 1.1291 | 1.0716 | 2.1045 | 2.1045 | 0.6712 |
| 2005 | 0.0938 | 0.2575 | 0.3816 | 0.4720 | 0.7745 | 0.8443 | 0.5739 | 1.1569 | 1.1569 | 0.5459 |
| 2006 | 0.1881 | 0.3339 | 0.3600 | 0.6103 | 0.8196 | 0.9574 | 1.0306 | 0.2931 | 0.2931 | 0.6163 |
| 2007 | 0.1239 | 0.3281 | 0.3855 | 0.4407 | 0.5945 | 0.6869 | 0.6982 | 0.8459 | 0.8459 | 0.4871 |
| 2008 | 0.0511 | 0.2634 | 0.3420 | 0.3961 | 0.4853 | 0.7241 | 1.2123 | 1.5432 | 1.5432 | 0.4422 |
| 2009 | 0.1219 | 0.6832 | 0.5265 | 0.4876 | 0.5136 | 0.4148 | 0.5911 | 1.5140 | 1.5140 | 0.5252 |
| 2010 | 0.2106 | 0.4591 | 0.6065 | 0.6145 | 0.9074 | 0.6761 | 0.5594 | 0.8932 | 0.8932 | 0.6527 |
| 2011 | 0.0921 | 0.3853 | 0.5701 | 0.5695 | 0.5423 | 0.6474 | 0.4027 | 0.3566 | 0.3566 | 0.5429 |
| 2012 | 0.0310 | 0.2045 | 0.3975 | 0.5450 | 0.7201 | 0.6174 | 0.9268 | 0.3607 | 0.3607 | 0.4969 |
| 2013 | 0.0260 | 0.1515 | 0.2035 | 0.3132 | 0.3421 | 0.2962 | 0.3202 | 0.5197 | 0.5197 | 0.2613 |
| 2014 | 0.0728 | 0.3927 | 0.4384 | 0.5867 | 0.3388 | 0.3012 | 0.1718 | 0.1951 | 0.1951 | 0.4115 |

Table 4.6.3. Faroe Plateau cod (sub-division Vb1). Stock number at age from the XSA model.

| Age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | $10+$ | Total |
| 1959 | 13238 | 12185 | 2634 | 4092 | 683 | 503 | 213 | 29 | 0 | 50976 |
| 1960 | 14245 | 9027 | 6141 | 1380 | 1784 | 378 | 225 | 129 | 0 | 47989 |
| 1961 | 12019 | 7385 | 3747 | 2699 | 666 | 668 | 155 | 66 | 0 | 52630 |
| 1962 | 20654 | 7042 | 3616 | 1863 | 1245 | 335 | 210 | 56 | 0 | 59804 |
| 1963 | 20290 | 12907 | 3503 | 1825 | 752 | 584 | 190 | 87 | 0 | 66807 |
| 1964 | 21834 | 12893 | 6986 | 1710 | 895 | 358 | 294 | 112 | 0 | 55183 |
| 1965 | 8269 | 16037 | 7823 | 3639 | 830 | 416 | 151 | 169 | 0 | 60009 |
| 1966 | 18566 | 5999 | 10207 | 4085 | 1698 | 351 | 200 | 80 | 0 | 69829 |
| 1967 | 23451 | 13990 | 4034 | 6475 | 2133 | 842 | 109 | 70 | 0 | 72579 |
| 1968 | 17582 | 17744 | 9020 | 2525 | 3757 | 980 | 410 | 31 | 0 | 63439 |
| 1969 | 9325 | 13012 | 11522 | 4976 | 1212 | 1967 | 393 | 240 | 0 | 53161 |
| 1970 | 8608 | 6840 | 7843 | 6447 | 2682 | 561 | 965 | 138 | 0 | 48654 |
| 1971 | 11928 | 6684 | 4548 | 4456 | 3754 | 1516 | 238 | 519 | 0 | 59683 |
| 1972 | 21320 | 9469 | 4788 | 2981 | 2483 | 1760 | 779 | 92 | 0 | 59029 |
| 1973 | 12573 | 16664 | 6689 | 3187 | 1901 | 1109 | 902 | 499 | 400 | 81153 |
| 1974 | 30480 | 9639 | 10816 | 4037 | 1969 | 1209 | 626 | 533 | 342 | 106456 |
| 1975 | 38319 | 23000 | 6747 | 7217 | 2460 | 1103 | 581 | 378 | 476 | 102968 |
| 1976 | 18575 | 29035 | 13683 | 3572 | 3908 | 1279 | 636 | 304 | 466 | 83665 |
| 1977 | 9995 | 13853 | 20010 | 7765 | 1676 | 1909 | 489 | 274 | 18 | 69116 |
| 1978 | 10748 | 7799 | 8372 | 10190 | 2993 | 659 | 513 | 184 | 154 | 59931 |
| 1979 | 14998 | 8298 | 5282 | 4463 | 5433 | 1509 | 297 | 238 | 103 | 69424 |
| 1980 | 23583 | 11759 | 5226 | 2811 | 2206 | 2723 | 789 | 122 | 52 | 66371 |
| 1981 | 14001 | 18286 | 7580 | 2957 | 1491 | 1076 | 1477 | 339 | 150 | 74384 |
| 1982 | 22128 | 10878 | 11228 | 4413 | 1564 | 694 | 440 | 732 | 348 | 83159 |
| 1983 | 25162 | 17087 | 7128 | 6412 | 2450 | 854 | 284 | 207 | 200 | 118129 |
| 1984 | 47769 | 18656 | 8767 | 3339 | 2765 | 916 | 238 | 91 | 174 | 103874 |
| 1985 | 17323 | 35132 | 10538 | 4023 | 1412 | 1439 | 466 | 121 | 146 | 82219 |
| 1986 | 9513 | 13280 | 20182 | 5194 | 1784 | 459 | 389 | 102 | 81 | 63093 |
| 1987 | 9914 | 7598 | 7628 | 8867 | 2106 | 640 | 162 | 185 | 69 | 47827 |
| 1988 | 8726 | 7884 | 4989 | 3883 | 4470 | 989 | 321 | 71 | 53 | 51427 |
| 1989 | 16408 | 6684 | 4535 | 2324 | 1836 | 1689 | 365 | 111 | 16 | 38422 |
| 1990 | 3646 | 11410 | 3526 | 1734 | 889 | 575 | 481 | 99 | 50 | 30547 |
| 1991 | 6662 | 2762 | 6725 | 1526 | 637 | 357 | 201 | 127 | 57 | 32967 |
| 1992 | 11392 | 5280 | 1853 | 3558 | 687 | 247 | 164 | 81 | 90 | 35685 |
| 1993 | 10097 | 9142 | 3911 | 1096 | 2089 | 297 | 83 | 86 | 97 | 57624 |
| 1994 | 25156 | 8158 | 6759 | 2657 | 696 | 1413 | 156 | 38 | 26 | 96979 |
| 1995 | 42508 | 20078 | 5967 | 4573 | 1694 | 457 | 997 | 93 | 102 | 92173 |
| 1996 | 12858 | 32437 | 13981 | 3068 | 2828 | 966 | 267 | 658 | 88 | 75034 |
| 1997 | 6454 | 10210 | 21884 | 7277 | 1117 | 935 | 252 | 86 | 198 | 55645 |
| 1998 | 5922 | 5103 | 7203 | 11846 | 2582 | 320 | 187 | 52 | 46 | 50773 |
| 1999 | 14338 | 4437 | 3504 | 4487 | 5048 | 730 | 118 | 46 | 19 | 56802 |
| 2000 | 19710 | 10667 | 2734 | 2146 | 2672 | 2119 | 201 | 43 | 6 | 76558 |
| 2001 | 29687 | 14245 | 6347 | 1531 | 1371 | 1578 | 1003 | 70 | 12 | 72038 |


| 2002 | 13258 | 20766 | 8262 | 3295 | 921 | 791 | 642 | 426 | 10 | 55993 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2003 | 6240 | 8974 | 10411 | 3713 | 1186 | 329 | 165 | 152 | 26 | 35631 |
| 2004 | 3631 | 4496 | 5422 | 4387 | 1296 | 392 | 109 | 53 | 45 | 27275 |
| 2005 | 6095 | 2882 | 3055 | 3296 | 1687 | 395 | 104 | 31 | 47 | 26900 |
| 2006 | 7619 | 4543 | 1824 | 1708 | 1683 | 637 | 139 | 48 | 13 | 24469 |
| 2007 | 5120 | 5169 | 2664 | 1042 | 760 | 607 | 200 | 41 | 6 | 23554 |
| 2008 | 6506 | 3703 | 3048 | 1483 | 549 | 343 | 250 | 82 | 27 | 26282 |
| 2009 | 8425 | 5061 | 2330 | 1773 | 817 | 277 | 136 | 61 | 23 | 33923 |
| 2010 | 12297 | 6106 | 2093 | 1127 | 891 | 400 | 150 | 62 | 70 | 28258 |
| 2011 | 4145 | 8156 | 3159 | 934 | 499 | 295 | 167 | 70 | 7 | 19597 |
| 2012 | 1773 | 3095 | 4542 | 1462 | 433 | 238 | 126 | 91 | 80 | 14836 |
| 2013 | 2453 | 1407 | 2065 | 2499 | 694 | 172 | 105 | 41 | 16 | 17724 |
| 2014 | 6772 | 1957 | 990 | 1380 | 1496 | 404 | 105 | 62 | 6 | 14239 |

Table 4.6.4. Faroe Plateau cod (sub-division Vb1). Summary table from the XSA model. The results from the short term prediction are shown in bold.

|  | RECRUITS | TOTALBIO | TOTSPBIO | LANDINGS | YIELD/SSB | FBAR 3-7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age 2 |  |  |  |  |  |
| 1959 | 13238 | 67803 | 48869 | 22415 | 0.4587 | 0.5117 |
| 1960 | 14245 | 75862 | 54447 | 32255 | 0.5924 | 0.661 |
| 1961 | 12019 | 65428 | 46439 | 21598 | 0.4651 | 0.6059 |
| 1962 | 20654 | 68225 | 43326 | 20967 | 0.4839 | 0.5226 |
| 1963 | 20290 | 77602 | 49054 | 22215 | 0.4529 | 0.4944 |
| 1964 | 21834 | 84666 | 55362 | 21078 | 0.3807 | 0.5017 |
| 1965 | 8269 | 75043 | 57057 | 24212 | 0.4244 | 0.4909 |
| 1966 | 18566 | 83919 | 60629 | 20418 | 0.3368 | 0.4743 |
| 1967 | 23451 | 105289 | 73934 | 23562 | 0.3187 | 0.39 |
| 1968 | 17582 | 110433 | 82484 | 29930 | 0.3629 | 0.4642 |
| 1969 | 9325 | 105537 | 83487 | 32371 | 0.3877 | 0.4375 |
| 1970 | 8608 | 98398 | 82035 | 24183 | 0.2948 | 0.3882 |
| 1971 | 11928 | 78218 | 63308 | 23010 | 0.3635 | 0.3526 |
| 1972 | 21320 | 76439 | 57180 | 18727 | 0.3275 | 0.3358 |
| 1973 | 12573 | 110713 | 83547 | 22228 | 0.2661 | 0.2886 |
| 1974 | 30480 | 139266 | 98434 | 24581 | 0.2497 | 0.3139 |
| 1975 | 38319 | 153664 | 109566 | 36775 | 0.3356 | 0.3947 |
| 1976 | 18575 | 161260 | 123077 | 39799 | 0.3234 | 0.4749 |
| 1977 | 9995 | 136212 | 112057 | 34927 | 0.3117 | 0.6757 |
| 1978 | 10748 | 96227 | 78497 | 26585 | 0.3387 | 0.4259 |
| 1979 | 14998 | 85112 | 66723 | 23112 | 0.3464 | 0.4273 |
| 1980 | 23583 | 85038 | 58887 | 20513 | 0.3483 | 0.3945 |
| 1981 | 14001 | 88411 | 63562 | 22963 | 0.3613 | 0.4648 |
| 1982 | 22128 | 98964 | 67033 | 21489 | 0.3206 | 0.4138 |
| 1983 | 25162 | 123256 | 78543 | 38133 | 0.4855 | 0.7056 |
| 1984 | 47769 | 152162 | 96774 | 36979 | 0.3821 | 0.5081 |
| 1985 | 17323 | 131245 | 84789 | 39484 | 0.4657 | 0.7013 |
| 1986 | 9513 | 99280 | 73698 | 34595 | 0.4694 | 0.6691 |
| 1987 | 9914 | 78372 | 62249 | 21391 | 0.3436 | 0.4452 |
| 1988 | 8726 | 66185 | 52134 | 23182 | 0.4447 | 0.6073 |
| 1989 | 16408 | 59280 | 38427 | 22068 | 0.5743 | 0.796 |
| 1990 | 3646 | 38547 | 29450 | 13692 | 0.4649 | 0.6662 |
| 1991 | 6662 | 28951 | 21301 | 8750 | 0.4108 | 0.512 |
| 1992 | 11392 | 36023 | 21073 | 6396 | 0.3035 | 0.4575 |
| 1993 | 10097 | 51491 | 33502 | 6107 | 0.1823 | 0.2351 |
| 1994 | 25156 | 84335 | 42937 | 9046 | 0.2107 | 0.1846 |
| 1995 | 42508 | 144616 | 54735 | 23045 | 0.421 | 0.321 |
| 1996 | 12858 | 142597 | 85457 | 40422 | 0.473 | 0.7019 |
| 1997 | 6454 | 96379 | 81121 | 34304 | 0.4229 | 0.7716 |
| 1998 | 5922 | 65797 | 55445 | 24005 | 0.4329 | 0.5925 |
| 1999 | 14338 | 64613 | 44611 | 18306 | 0.4103 | 0.5304 |
| 2000 | 19710 | 90668 | 45736 | 21033 | 0.4599 | 0.3643 |


| 2001 | 29687 | 109541 | 58652 | 28183 | 0.4805 | 0.4315 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2002 | 13258 | 98034 | 55679 | 38457 | 0.6907 | 0.8216 |
| 2003 | 6240 | 60425 | 40399 | 24501 | 0.6065 | 0.726 |
| 2004 | 3631 | 37025 | 27059 | 13178 | 0.487 | 0.6712 |
| 2005 | 6095 | 31865 | 23470 | 9906 | 0.4221 | 0.5459 |
| 2006 | 7619 | 30285 | 20897 | 10479 | 0.5015 | 0.6163 |
| 2007 | 5120 | 27367 | 17387 | 8015 | 0.461 | 0.4871 |
| 2008 | 6506 | 29800 | 20433 | 7465 | 0.3653 | 0.4422 |
| 2009 | 8425 | 29989 | 19563 | 10002 | 0.5113 | 0.5252 |
| 2010 | 12297 | 38774 | 21525 | 12757 | 0.5927 | 0.6527 |
| 2011 | 4145 | 30826 | 19114 | 9760 | 0.5106 | 0.5429 |
| 2012 | 1773 | 24323 | 19290 | 7210 | 0.3738 | 0.4969 |
| 2013 | 2453 | 23476 | 20785 | 4630 | 0.2228 | 0.2613 |
| 2014 | 6772 | 27720 | 21142 | 6349 | 0.3003 | 0.4115 |
| 2015 | 874 | 26110 | 18781 | 6648 | 0.3540 | 0.3899 |
| 2016 | 3666 | 24201 | 19687 | 6037 | 0.3066 | 0.3899 |
| 2017 | 3666 | 24179 | 19472 |  |  |  |
| Avg.59-14 | 14720 | 80017 | 55471 | 21817 | 0.41 | 0.51 |

Table 4.6.5. Faroe Plateau cod (sub-division Vb1). Results from the back-calculation of the age2+ biomass back to 1906 (in tonnes). The exploitation ratio (catch/biomass, C/B) is also provided. The higher biomass estimate is obtained by using a scaling factor between age $2+$ biomass from the agebased assessment and the cpue of British trawlers. The lower estimate is obtained by using a regression line.

|  |  | Factor | Regression |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Catch (t) | Biomass (t) | Biomass (t) | C/B | C/B |
| 1906 | 18510 | 125162 | 108644 | 0.148 | 0.170 |
| 1907 | 19802 | 148793 | 122844 | 0.133 | 0.161 |
| 1908 | 11609 | 108532 | 98651 | 0.107 | 0.118 |
| 1909 | 19825 | 175051 | 138622 | 0.113 | 0.143 |
| 1910 | 21682 | 149669 | 123370 | 0.145 | 0.176 |
| 1911 | 31406 | 175051 | 138622 | 0.179 | 0.227 |
| 1912 | 38718 | 161922 | 130733 | 0.239 | 0.296 |
| 1913 | 33228 | 137415 | 116007 | 0.242 | 0.286 |
| 1914 | 30580 | 112908 | 101281 | 0.271 | 0.302 |
| 1915 | 19810 | 122391 | 106979 | 0.162 | 0.185 |
| 1916 | 17785 | 143731 | 119802 | 0.124 | 0.148 |
| 1917 | 18155 | 171332 | 136387 | 0.106 | 0.133 |
| 1918 | 23160 | 213691 | 161841 | 0.108 | 0.143 |
| 1919 | 43468 | 205685 | 157030 | 0.211 | 0.277 |
| 1920 | 17726 | 98904 | 92866 | 0.179 | 0.191 |
| 1921 | 12088 | 117284 | 103911 | 0.103 | 0.116 |
| 1922 | 19315 | 160172 | 129681 | 0.121 | 0.149 |
| 1923 | 25553 | 133039 | 113377 | 0.192 | 0.225 |
| 1924 | 45197 | 136895 | 115694 | 0.330 | 0.391 |
| 1925 | 38296 | 129353 | 111163 | 0.296 | 0.345 |
| 1926 | 44066 | 185574 | 144945 | 0.237 | 0.304 |
| 1927 | 45172 | 162034 | 130800 | 0.279 | 0.345 |
| 1928 | 30303 | 126611 | 109515 | 0.239 | 0.277 |
| 1929 | 26506 | 135524 | 114871 | 0.196 | 0.231 |
| 1930 | 33022 | 142608 | 119128 | 0.232 | 0.277 |
| 1931 | 45418 | 139409 | 117205 | 0.326 | 0.388 |
| 1932 | 44646 | 121354 | 106356 | 0.368 | 0.420 |
| 1933 | 37087 | 108327 | 98529 | 0.342 | 0.376 |
| 1934 | 35495 | 107870 | 98254 | 0.329 | 0.361 |
| 1935 | 32125 | 91187 | 88229 | 0.352 | 0.364 |
| 1936 | 34758 | 102385 | 94958 | 0.339 | 0.366 |
| 1937 | 26639 | 95758 | 90976 | 0.278 | 0.293 |
| 1938 | 23755 | 93244 | 89465 | 0.255 | 0.266 |
| 1939 | 6399 | 143439 | 119627 | 0.045 | 0.053 |
| 1940 | 8113 | 193635 | 149789 | 0.042 | 0.054 |
| 1941 | 6559 | 216611 | 163595 | 0.030 | 0.040 |
| 1942 | 6791 | 188465 | 146682 | 0.036 | 0.046 |
| 1943 | 9850 | 196270 | 151372 | 0.050 | 0.065 |
| 1944 | 7847 | 210683 | 160033 | 0.037 | 0.049 |


|  | Factor |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Year | Catch (t) | Biomass $(\mathbf{t})$ | Biomass $(\mathbf{t})$ | C/B | C/B |
| 1945 | 8646 | 225096 | 168693 | 0.038 | 0.051 |
| 1946 | 30485 | 239509 | 177354 | 0.127 | 0.172 |
| 1947 | 30993 | 177346 | 140001 | 0.175 | 0.221 |
| 1948 | 20712 | 122497 | 107043 | 0.169 | 0.193 |
| 1949 | 28134 | 164777 | 132448 | 0.171 | 0.212 |
| 1950 | 35973 | 152207 | 124895 | 0.236 | 0.288 |
| 1951 | 35076 | 124325 | 108142 | 0.282 | 0.324 |
| 1952 | 30259 | 116783 | 103610 | 0.259 | 0.292 |
| 1953 | 27055 | 116783 | 103610 | 0.232 | 0.261 |
| 1954 | 36170 | 146493 | 121462 | 0.247 | 0.298 |
| 1955 | 38583 | 149464 | 123247 | 0.258 | 0.313 |
| 1956 | 27628 | 108327 | 98529 | 0.255 | 0.280 |
| 1957 | 31393 | 112898 | 101275 | 0.278 | 0.310 |
| 1958 | 27807 | 84102 | 83972 | 0.331 | 0.331 |

Table 4.7.1. Faroe Plateau cod (sub-division Vb1). Input to management option table.

|  | Recr. |  | Source | Stock size |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Age | 2015 Source |
|  |  |  | 2 | 874 XSA-output |
|  |  |  | 3 | 6772 XSA-output |
|  |  |  | 4 | 1082 XSA-output |
|  |  |  | 5 | 523 XSA-output |
| 2014 | YC2012 | 6772 |  | XSA-output | 6 | 628 XSA-output |
| 2015 | YC2013 | 874 |  | XSA-output | 7 | 873 XSA-output |
| 2016 | YC2014 | 3666 |  | Average R 2012-14 | 8 | 245 XSA-output |
| 2017 | YC2015 | 3666 |  | Average R 2012-14 | 9 | 72 XSA-output |
|  |  |  |  |  | 10+ | 46 XSA-output |


| Maturity |  |  |  | Exploitation pattern (not rescaled) |  |  | eights |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Observed | Av. 13-15 | Av. 13-15 | Av. 12-14 | Av. 12-14 | Av. 12-14 |  | As 2015 | Av.13-15 |
| Age | 2015 | 2016 | 2017 | 2015 | 2016 | 2017 | 2015 | 2016 | 2017 |
| 2 | 0.28 | 0.25 | 0.25 | 0.0433 | 0.0433 | 0.0433 | 1.098 | 1.098 | 1.069 |
| 3 | 0.48 | 0.68 | 0.68 | 0.2496 | 0.2496 | 0.2496 | 1.648 | 1.648 | 1.609 |
| 4 | 0.70 | 0.88 | 0.88 | 0.3465 | 0.3465 | 0.3465 | 2.098 | 2.098 | 2.364 |
| 5 | 0.95 | 0.98 | 0.98 | 0.4816 | 0.4816 | 0.4816 | 2.82 | 2.82 | 3 |
| 6 | 0.97 | 0.99 | 0.99 | 0.4670 | 0.4670 | 0.4670 | 4.241 | 4.241 | 4.354 |
| 7 | 1.00 | 1.00 | 1.00 | 0.4049 | 0.4049 | 0.4049 | 5.269 | 5.269 | 6.178 |
| 8 | 1.00 | 1.00 | 1.00 | 0.4729 | 0.4729 | 0.4729 | 7.182 | 7.182 | 7.847 |
| 9 | 1.00 | 1.00 | 1.00 | 0.3585 | 0.3585 | 0.3585 | 9.236 | 9.236 | 9.167 |
| $10+$ | 1.00 | 1.00 | 1.00 | 0.3585 | 0.3585 | 0.3585 | 12.12 | 12.12 | 11.521 |
|  |  |  | Fbar: | 0.3899 | 0.3899 | 0.3899 |  |  |  |

Table 4.7.2. Faroe Plateau cod (sub-division Vb1). Management option table.

| 2015 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar | Landings |  |  |
| 26110 | 18781 | 1.0000 | 0.3899 | 6648 |  |  |
| 2016 |  |  |  |  | 2017 |  |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 24201 | 19687 | 0.0000 | 0.0000 | 0 | 31564 | 26685 |
| . | 19687 | 0.1000 | 0.0390 | 710 | 30677 | 25817 |
| . | 19687 | 0.2000 | 0.0780 | 1394 | 29824 | 24983 |
| . | 19687 | 0.3000 | 0.1170 | 2053 | 29003 | 24181 |
| . | 19687 | 0.4000 | 0.1560 | 2688 | 28214 | 23409 |
| . | 19687 | 0.5000 | 0.1950 | 3299 | 27455 | 22667 |
| . | 19687 | 0.6000 | 0.2340 | 3888 | 26725 | 21954 |
| . | 19687 | 0.7000 | 0.2729 | 4455 | 26022 | 21267 |
| . | 19687 | 0.8000 | 0.3119 | 5002 | 25346 | 20607 |
| . | 19687 | 0.9000 | 0.3509 | 5529 | 24695 | 19973 |
| . | 19687 | 1.0000 | 0.3899 | 6037 | 24069 | 19362 |
| . | 19687 | 1.1000 | 0.4289 | 6526 | 23466 | 18774 |
| . | 19687 | 1.2000 | 0.4679 | 6998 | 22886 | 18209 |
| . | 19687 | 1.3000 | 0.5069 | 7453 | 22328 | 17665 |
| . | 19687 | 1.4000 | 0.5459 | 7891 | 21790 | 17142 |
| - | 19687 | 1.5000 | 0.5849 | 8314 | 21272 | 16638 |
| . | 19687 | 1.6000 | 0.6239 | 8722 | 20774 | 16154 |
| - | 19687 | 1.7000 | 0.6629 | 9115 | 20294 | 15687 |
| . | 19687 | 1.8000 | 0.7019 | 9495 | 19832 | 15238 |
| . | 19687 | 1.9000 | 0.7408 | 9861 | 19387 | 14806 |
| . | 19687 | 2.0000 | 0.7798 | 10214 | 18958 | 14390 |

Table 4.8.1. Faroe Plateau cod (sub-division Vb1). Input to yield per recruit calculations (long term prediction).

|  | Expl. pattern | Weight at age | Prop mature |
| :--- | :--- | :--- | :--- |
|  | Average | Average | Average |
| Age | $2002-2014$ | $1978-2014$ | $1983-2015$ |
|  | Not rescaled |  |  |
| 2 | 0.105 | 1.040 | 0.06 |
| 3 | 0.342 | 1.559 | 0.57 |
| 4 | 0.444 | 2.272 | 0.83 |
| 5 | 0.574 | 3.066 | 0.94 |
| 6 | 0.674 | 3.881 | 0.98 |
| 7 | 0.736 | 4.977 | 0.99 |
| 8 | 0.749 | 6.200 | 1.00 |
| 9 | 0.999 | 7.703 | 1.00 |
| $10+$ | 0.999 | 9.638 | 0.99 |

Table 4.8.2. Faroe Plateau cod (sub-division Vb1). Output from yield per recruit calculations (long term prediction).

| Reference point | F multiplier | Absolute $F$ |
| :--- | :--- | :--- |
| Fbar(3-7) | 1.0000 | 0.554 |
| FMax | 0.4513 | 0.25 |
| F0.1 | 0.2087 | 0.1156 |
| F35\%SPR | 0.3147 | 0.1743 |
| Flow | 0.16 | 0.0886 |
| Fmed | 0.6973 | 0.3863 |
| Fhigh | 1.6044 | 0.8888 |

Weights in kilograms


Figure 4.2.1. Faroe Plateau cod (sub-division Vb1). Catch in numbers at age shown as catch curves.


Figure 4.2.2. Faroe Plateau cod (sub-division Vb1). Mean weight at age. The predicted weights are also shown.


Figure 4.2.3. Faroe Plateau cod (sub-division Vb1). Proportion mature at age as observed in the spring groundfish survey. The predicted values are shown in grey.


Figure 4.2.4. Faroe Plateau cod (sub-division Vb 1 ). Catch curves from the spring groundfish survey.

Faroe Plateau cod


Figure 4.2.5. Faroe Plateau cod (sub-division Vb1). Stratified $\mathrm{kg} /$ hour in the spring and summer surveys (upper figure). The age $3+$ biomass obtained from the assessment is also included as an index.


Figure 4.2.6. Faroe Plateau cod (sub-division Vb1). Catch curves from the summer groundfish survey.


Figure 4.2.7. Faroe Plateau cod (sub-division Vb1). Standardised catch per unit effort for pair trawlers and longliners. The two surveys are shown as well.


Figure 4.2.8. Faroe Plateau cod (sub-division Vb1). Catch per unit effort for small and large longliners compared with the fishable (age 3+) biomass.


Figure 4.2.9. Faroe Plateau cod (sub-division Vb1). Catchability (cpue divided by age 3+ biomass) for small and large longliners and pair trawlers.

Spring survey (shifted back to December)


## Summer survey



Figure 4.6.1. Faroe Plateau cod (sub-division Vb1). Log catchability residuals for age 2 to 7 for the spring (upper figure) and summer survey. The residuals for age 8 are not presented because some values were off scale. White bubbles indicate negative residuals.

## Spawning stock and recruitment



Yield and fishing mortality


Figure 4.6.2. Faroe Plateau cod (sub-division Vb1). Spawning stock biomass (SSB) and recruitment (year class) versus year (upper figure) and yield and fishing mortality versus year. Points (white and grey) are taken from the short term projections.


Figure 4.6.3. Faroe Plateau cod (sub-division Vb1). Different measures of fishing mortality: straight arithmetic average (Avg F), weighted by stock numbers (Nwtd), weighted by stock biomass (Bwtd) or weighted by catch (Cwtd).


Figure 4.6.4. Faroe Plateau cod (sub-division Vb1). Spawning stock - recruitment relationship. Years are shown at each data point.


Figure 4.6.5. Faroe Plateau cod (sub-division Vb1). Spawning stock biomass versus fishing mortality.


Figure 4.6.6. Faroe Plateau cod (sub-division Vb1). Biomass obtained from the age-based assessment as well as from cpue of British trawlers back in time. There was an overlap between cpue and the age-based assessment in the period 1959-72 and the two versions of the biomass prior to 1959 was whether a regression line was used or a scaling factor. During the wars (grey symbols) catch data from Faroe boats were used as indicative of stock biomass and regressed against cpue of British trawlers for a period prior to the wars. The missing years of data were estimated by linear interpolation (open symbols).

Faroe Plateau cod, biomass 1906-58 estimated by CPUE (a factor)


Figure 4.6.7. Faroe Plateau cod (sub-division Vb1). Exploitation ratio (based on the higher biomass) compared with tag returns. The taggings in 1909-13 were on small cod close to land, in 1930s on large spawning cod, in 1950s-60s and in 1997-2013 on cod on the feeding grounds.

SSB 2016


Figure 4.7.1. Faroe Plateau cod (sub-division Vb1). Predictions of the contribution of various year classes to the spawning stock biomass in terminal year +1 (upper figure) and terminal year +2 (lower figure).


Figure 4.8.1. Faroe Plateau cod (sub-division Vb1). Yield per recruit and spawning stock biomass (SSB) per recruit versus fishing mortality (left figure). Landings and SSB versus Fbar (3-7) (right figure).


Figure 4.9.1. Faroe Plateau cod (sub-division Vb 1 ). Results from the XSA retrospective analysis of fishing mortality (ages 3-7).


Figure 4.9.1. Faroe Plateau cod (sub-division Vb1). Results from the XSA retrospective analysis (continued). Recruitment at age 1 (upper figure) and at age 2.


Figure 4.9.1. Faroe Plateau cod (sub-division Vb 1 ). Results from the XSA retrospective analysis (continued). Spawning stock biomass (upper figure) and total stock biomass.


Figure 4.9.2. Faroe Plateau cod (sub-division Vb1). Modelling cod recruitment in three steps. First, the catch-per-unit -effort of cod (C) for small boats operating close to land, as being indicative of the amount of cannibalistic cod. Second, the amount of cod (older than the recruiting cod) (B), as being indicative of e.g. the amount of schools to which recruiting cod can join and hide in. Third, the ratio between B and C, as indicative of recruitment success. Fourth and fifth, a comparison with observed recruitment. Note that the model predicts that the recruitment in recent years is very poor.


Figure 4.10.1. Faroe Plateau cod (sub-division Vb1). Comparison between the results from the current assessment (Assm. 2015) and the assessment last year (Assm. 2014) for recruitment (upper left), fishing mortality (upper right), stock biomass (lower left) and spawning stock biomass (lower right).

## 5 Faroe haddock

## Executive summary

Being an update assessment, the changes compared to last year are additions of new data from 2014 and 2015 and some minor revisions of recent landings data with corresponding revisions of the catch at age data. The main assessment tool is an XSA tuned with two research vessel bottom trawl surveys. The results are in line with those from 2014, showing a very low SSB mainly due to poor recruitment but also due to higher than recommended fishing mortalities in recent years. SSB is now estimated well below Blim and is predicted to stay below Blim in 2016-2017 with status quo fishing mortality. Fishing mortality in 2014 is estimated at 0.29 and the average fishing mortality from 2012-2014 at 0.28 ( $\mathrm{F}_{\mathrm{MSY}}$ and $\mathrm{F}_{\mathrm{pa}}=0.25$ ). Landings in 2014 were 3200 t , which is slightly higher than in 2012 and 2013. This years assessment indicates that the 2014 assessment underestimated the 2013 recruitment by $23 \%$ ( 2 million versus 2.6 million, which still is the lowest on record), overestimated the fishing mortality in 2013 by $6 \%$ ( 0.28 versus 0.26 ) and underestimated the 2013 total- and spawning stock biomasses by $3 \%$ and $6 \%$, respectively ( 20 and 19 thous. t versus 19.6 and 18 thous. t).

### 5.1 Stock description and management units

Haddock in Faroese Waters, i.e. ICES Sub-Divisions Vb1 and Vb2 and in the southern part of ICES Division IIa, close to the border of Sub-Division Vb1, are generally believed to belong to the same stock and are treated as one management unit named Faroe haddock. Haddock is distributed all over the Faroe Plateau and the Faroe Bank from shallow water down to more than 450 m . A more detailed description of haddock in Farose waters is given in the stock annex. The spatial distribution of the haddock in the summer survey and in the spring survey is shown in figure 5.9. The figure do clearly illustrate the drastic decrease in the stock biomass in recent years.

### 5.2 Scientific data

### 5.2.1 Trends in landings and fisheries

Nominal landings of Faroe haddock increased very rapidly from only 4000 t in 1993 to 27000 t in 2003, but have declined drastically since and amounted in 2014 to only about 3200 t . Most of the landings are taken from the Faroe Plateau; the 2014 landings from the Faroe Bank (Sub-Division Vb2), where the area shallower than 200 m depths has been closed to almost all fishing since the fiscal year 2008-2009, amounted to only about 64 t (Tables 5.1 and 5.2). The cumulative landings by month are shown in Figure 5.2.

Faroese vessels have taken almost the entire catch since the late 1970s (Figure 5.1). Due to the dispute on mackerel quota share, there has been no agreement on mutual fishery rights between the Faroe Islands and Norway and EU, respectively, since 2011 and therefore there was no fishery by those parties in Vb in 2012 and 2013; in 2014 the parties happened to make an agreement again. The proportion of the Faroese landings taken by each fleet category since 1985 are shown in the annex. The longliners have taken most of the catches in recent years followed by the trawlers. This was also the case in 2014, where the share by longliners was $82 \%$ and that by trawlers $18 \%$ (Figure 5.3).

### 5.2.2 Catch-at-age

Catch-at-age data were provided for fish taken by the Faroese fleets from Vb 1 and Vb 2 . The sampling intensity in 2014 is shown in Table 5.4 showing some decrease in intensity as compared to 2013. There is a need to increase the sampling level. Reasons for the inadequate sampling level are shortage of resources (people, money) but also that the total catches (and stock) are so small that it is difficult to obtain enough samples. From late 2011, a landing site has been established in Tórshavn close to the Marine Research Institute and it is the intention that technicians from the Institute will regularly be sampling the landings there; this will increase the sampling level in coming years. This has also turned out to be difficult of the above mentioned reasons but the outlook is very positive regarding raising enough money to hire a new technician to among other things do the sampling.

The normal procedure has been to disaggregate samples from each fleet category by season (Jan-Apr, May-Aug and Sep-Dec) and then raise them by the corresponding catch proportions to give the annual catch-at-age in numbers for each fleet This year, all longliners were grouped into 2 fleets (larger and smaller than 100 GRT, respectively), and all trawlers were also grouped into 2 fleets (larger and smaller than 1000 Hp, respectively) The longliner samples had to be treated by using 2 seasons only (JanJun, Jul-Dec. The results are given in Table 5.3. No catch-at-age data were available from other nations (Norwegian longliners and British trawlers) and they were assumed to have the same age composition as the Faroese corresponding fleets. The most recent data were revised according to the final catch figures. The resulting total catch-at-age in numbers are given in Tables 5.4 and 5.5, and in Figure 5.4 the LN(catch-at-age in numbers) is shown since 1957.

In general the catch-at-age matrix in recent years appears consistent although from time to time a few very small year classes are disturbing this consistency, both in numbers and mean weights at age. The recent very small year classes need to be very carefully inspected when the FBAR is calculated. Also there are some problems with what ages should be included in the plus group; there are some periods where only a few fishes are older than 9 years, and other periods with a quite substantial plus group $(10+)$. These problems have been addressed in former reports of this WG and will not be further dealt with here (See the 2005 NWWG report). No estimates of discards of haddock are available. However, since almost no quotas are used in the management of the fisheries on this stock, the incentive to discard in order to high-grade the catches should be low. The landings statistics is therefore regarded as being adequate for assessment purposes. The ban on discarding as stated in the law on fisheries should also - in theory - keep the discarding at a low level.

### 5.2.3 Weight-at-age

Mean weight-at-age data are provided for the Faroese fishery (Table 5.5). Figure 5.5 shows the mean weights-at-age in the landings for age groups $2-7$ since 1976. During this period, weights have shown cyclical changes. They were at a minimum in 20072009, but have increased again since then In the 3 latest years the weights have been fluctuated without a clear trend and a simple average of these years will be used in the short term predictions (Figure 5.5). The mean weights at age in the stock are assumed equal to those in the landings.

### 5.2.4 Maturity-at-age

Maturity-at-age data is available from the Faroese Spring Groundfish Surveys 1982 2014. The survey is carried out in February-March, so the maturity-at-age is determined just prior to the spawning of haddock in Faroese waters and the determinations of the different maturity stages is relatively easy.

In order to reduce year-to-year effects due to possible inadequate sampling and at the same time allow for trends in the series, the routine by the WG has been to use a 3-year running average in the assessment. For the years prior to 1982, average maturity-atage from the surveys 1982-1995 was adopted (Table 5.6 and Figure 5.6).

### 5.3 Information from the fishing industry

There exists a considerable amount of data on fish size in the fishing industry. No such information was used directly in the 2015 assessment but catch per unit effort for some selected fleets (logbook data) is used as an additional information on the status of the stock (see section 5.4.1.1).

### 5.4 Methods

This assessment is an update of the 2014 assessment, with exactly the same settings of the XSA. The only changes are minor revisions of recent landings according to revised data and corresponding revisions of the catch-at-age input. All other input files (VPA) are the same except for the addition of the 2014 data.

### 5.4.1 Tuning and estimates of fishing mortality

## Commercial cpue series

Several commercial catch per unit effort series are updated every year, but as discussed in previous reports of this WG they are not used directly for tuning of the VPA but as additional information on stock trends (for details see the stock annex). The age-aggregated cpue series for longliners and pair trawlers are presented in Figure 5.7. In general the two series show the same trends although in some periods the two series are conflicting; this has been explained by variations in catchability of the longlines due to changes in productivity of the ecosystem (see chapter 2). Both series, however, indicate that the stock is very low. The longliner cpue's do not decrease as much as the trawler cpue's which in addition to the explanation given above may be attributed to the fact that in the management of the demersal Faroese stocks, large areas have been closed to trawling with the effect that when the haddock stock is small, the distribution of it is mainly outside the "trawl areas".

## Fisheries independent cpue series

Two annual groundfish surveys are available, one carried out in February-March since 1982 ( 100 stations per year down to 500 m depth), and the other in August-September since 1996 ( 200 stations per year down to 500 m depth). The spatial distribution of haddock catches in the surveys Biomass estimates (kg/hour) are available for both series since they were initiated (Figure 5.8). The main trends from the surveys are the same but the summer survey indicates a considerably more depleted stock in recent years than the summer survey. Age disaggregated data are available for the whole summer series, but due to problems with the database (see earlier reports), age disaggregated data for the spring survey are only available since 1994. The calculation of indices at age is based on age-length keys with a smoother applied. This is a useful method but,
some artefacts may be introduced because the smoothing can assign wrong ages to some lengths, especially for the youngest and oldest specimen. As in recent years, the length distributions have been used more directly for calculation of indices at age (ages $0-2$ ), since these ages have length distributions almost without overlap. LN(numbers at age) for the surveys are presented in Figures $5.10-5.11$. Further analyses of the performances of the two series are shown in the stock annex. In general there is a good relationship between the indices for one year class in two successive years. The same applies when comparing the corresponding indices at age from both surveys .

A SPALY (same procedure as last year) run, with the same settings of the XSA as in 2014 (tuned with the two surveys combined, Table 5.8), with 2015 data included and some minor revisions of recent catch figures, gave in general similar results as last year (Table 5.9), although this years assessment indicates that the 2014 assessment underestimated the 2013 recruitment by $23 \%$ ( 2 million versus 2.6 million, which still is the lowest on record), overestimated the fishing mortality in 2032 by $6 \%$ ( 0.28 versus 0.26 ) and overestimated the 2013 total- and spawning stock biomasses by $3 \%$ and $6 \%$, respectively ( 20 and 19 thous. t versus 20 and 18 thous. t ). The $\log \mathrm{q}$ residuals for the two surveys are shown in Figure 5.12.

The retrospective analysis of fishing mortality, recruitment and spawning stock biomass of this XSA is shown in Figure 5.13. The retrospective pattern of the fishing mortality is hampered by strange values of some small poorly sampled year classes which in some years are included in the FBAR reference ages and consequently they will create problems for estimation of the stock (see the 2005 NWWG report); this is not a problem for the time being but the development of recent small year classes should be carefully inspected.

It has been questioned if a rather heavy shrinkage of 0.5 is the most appropriate for a stock like Faroe haddock where biological parameters and fishing mortality (catchability) are closely linked to productivity changes in the ecosystem. In order to investigate the possible effect of the shrinkage, the 2010 NWWG carried out an exploratory XSA without shrinkage (Shr. 2.0). Based on that it was concluded to continue with a shrinkage of 0.5 and this shrinkage was also applied this year.

## Results

The fishing mortalities from the final XSA run are given in Table 5.9 and in Figure 5.14. The fishing mortality was high (around 0.6) in the 1950s and early 1960s but declined to around 0.2 from 1965-1975. Since then, fishing mortality has usually been low, the exceptions are peaks in 1977, 1982, 1997-1999 and 2003-2006. They occur near the end of relatively high catch periods and some of the highest values $(0.32-0.45)$ are nearly certainly an artefact of the unweighted fishing mortality. Exploitation ratio (Yield/Biomass) is a bit more stable and may be used to indicate the level of fishing mortality.

### 5.5 Reference points

The yield- and spawning stock biomass per recruit (age 2) based on the long-term data are shown in Table 5.16 and Figure 5.16. Fmed, and Fhigh were calculated at 0.23 and 0.89 , respectively. The $\mathrm{F}_{\max }$ of 0.89 should not be used since it is very poorly determined due to the flat YPR curve. $\mathrm{F}_{0.1}$ is estimated at 0.18 . The F35\%SPR was estimated at 0.24 .

The precautionary reference fishing mortalities were set in 1998 by ACFM with $\mathrm{F}_{\mathrm{pa}}$ as the $\mathbf{F}_{\text {med }}$ value of 0.25 and Flim two standard deviations above $\mathbf{F}_{\mathrm{pa}}$ equal to 0.40 . The precautionary reference spawning stock biomass levels were changed by ACFM in
2007. Blim was set at $22000 t\left(B_{\text {loss }}\right)$ and $B_{p a}$ at $35000 t$ based on the formula $B_{p a}=B_{\text {lime }}{ }^{1.645 \sigma}$, assuming a $\sigma$ of about 0.3 to account for the uncertainties in the assessment.

The working group in 2012 investigated possible candidates for $\mathrm{F}_{\text {msy. }}$. Based on Medium -term projections, Medium-term projections the NWWG suggested, that FMSY preliminary could be set at 0.25 and the MSY $B_{\text {trigger }}$ at 35 thous. $t$ (same as $B_{p a}$ ) These values were accepted by ACOM. Some further analyses have indicated that these values are acceptable, but it is anticipated that further work will be untertaken in connection with the next benchmark assessment. See the stock annex for more details.

### 5.6 State of the stock

The stock size in numbers is given in Table 5.11 and a summary of the VPA with the biomass estimates is given in Table 5.12 and in Figure 5.14. According to this assessment, the period up to the mid 1970s was characterized by relative high and stable landings, recruitment and spawning stock biomass and the stock was able to withstand relatively high fishing mortalities. Since then the spawning stock biomass has shown large fluctuations due to cyclical changes in recruitment, growth and maturity (Figures 5.5 and 5.6). The fishing mortality does not seem to be the decisive factor in this development since it most of the period has fluctuated around the $\mathrm{F}_{\text {msy }}$ and $\mathrm{F}_{\mathrm{pa}}$. It must though be remembered that the characteristics of the stock in recent decades with long periods of poor recruitment make it less resilient to high fishing mortality.

The most recent increase in the spawning stock is due to new strong year classes entering the stock of which the 1999 year class is the highest on record (103 million at age 2). Also the YC's from 2000 and 2001 are estimated well above average and the 2002 YC above average, but the more recent YC's are all estimated to be very small except the 2009 YC, which is estimated to be slightly above the half of the average for the whole series back to 1957 and the 2012 and 2013 YC's, which are estimated somewhat higher than the other small year-classes. Fishing mortality has been relatively high since 2003, highest whent the stock was large leading to large variability in catches. Currently fishing mortality is estimated close to $\mathrm{F}_{\text {MSY }}$ (0.25).

### 5.7 Short term forecast

### 5.7.1 Input data

The input data for the short-term predictions are estimated in accordance with the procedures last year and explained in Tables 5.12-13. The YC 2015 at age 2 in 2017 is estimated as the geometric mean of the 2-year-olds since 2005. This procedure was introduced in 2011. All available information suggests that using the recent short series with poor recruitment is more appropriate than the longer period used in the past. However, the choice of recruitment in 2017 has little effect on the short term prediction.

### 5.7.2 Results

Although the allocated number of fishing days for the fishing year 2014 - 2015was reduced for some fleets as compared to the year before (see section 2), it should not be unrealistic to assume fishing mortalities in 2015 as the average of some recent years, here the average of $\mathrm{F}(2012-2014)$, since not all allocated days were actually used; however, possible changes in the catchability of the fleets (which seems to be linked to productivity changes in the environment) could undermine this assumption; price differences between cod and haddock may also influence this assumption. The landings in 2015 are then predicted to be about 3800 t , and continuing with this fishing mortality
will result in 2016 landings of about 4700 t . The SSB will decline to 19000 t in 2015, will be 19000 t in 2016 and decrease to 18000 t in 2017 i.e. will be below $\operatorname{Blim}(22000 \mathrm{t})$ in the next years. The results of the short-term prediction are shown in Table 5.16 and in Figure 5.14. The contribution (\%) by year-classes to the age composition of the predicted 2016 and 2017 SSB's is shown in Figure 5.17. It should be noted that young YC's which not have really entered the fishery in 2014/15, will contribute by a heavy proportion of the SSB in 2016/17.

### 5.8 Medium term forecasts and yield per recruit

No medium term projections were made this year; however, the 2013 projections, which were the basis for suggested MSY reference points, are presented in the stock annex.

The input data for the long-term yield and spawning stock biomass (yield-per-recruit calculations) are listed in Table 5.15. Mean weights-at-age (stock and catch) are averages for the 1977-2014 period. The maturity o-gives are averages for the years 1982 2014. The exploitation pattern is the same as in the short term prediction.

The results are given in Table 5.16, in Figure 5.16 and under Reference points (section 5.5).

### 5.9 Uncertainties in assessment and forecast

Retrospective analyses indicate periods with tendencies to overestimate spawning stock biomass and underestimate fishing mortality and vice versa. Similar things can be seen with the recruitment. This years assessment indicates that the 2014 assessment underestimated the 2013 recruitment $23 \%$ ( 20 millions versus 26 million, which still is the lowest on record), overestimated the fishing mortality in 2013 by $6 \%$ ( 0.28 versus 0.26 ) and underestimated the 2013 total- and spawning stock biomasses by $3 \%$ and $6 \%$, respectively ( 20 and 19 thous. t versus 19.6 and 18 thous. t ), see text table below..

Recruitment estimates from surveys are not very consistent for small cohorts.
The sampling of the catches for length measurements, otolith readings and lengthweight relationships has decreased somewhat compared to 2014. Although it is regarded to be adequate for the assessment, there is a need to improve it again (see 5.2).

### 5.10 Comparison with previous assessment and forecast

As explained previously in the report, this assessment is an update of the 2014 assessment. The only changes are minor revisions of recent landings according to revised data and corresponding revisions of the catch-at-age input. All other input files (VPA and tuning fleets) are the same except for the addition of the 2014 data.
Following differences in the 2013 estimates were observed as compared to last year (see text above):

Comparisons between 2014 and 2015 assessment of 2013 data
The year of comparison is 2013

|  | R at age 2 <br> (thousands) | Total B <br> (tonnes) | SSB <br> (tonnes) | Landings <br> (tonnes) | F (3-7) |
| :--- | ---: | ---: | :---: | :---: | :---: |
| 2014 spaly | 1992 | 20183 | 19017 | 3105 | 0.2753 |
| 2015 spaly | 2596 | 19643 | 17931 | 2950 | 0.2595 |
| \%-change | 23 | -3 | -6 | -5 | -6 |

### 5.11 Management plans and evaluations

There is no explicit management plan for this stock. A management system based on number of fishing days, closed areas and other technical measures was introduced in 1996 with the purpose of ensuring sustainable fisheries. There has been some work with establishing a management plan with a harvest control rule for cod, haddock and saithe including a recovery plan, but the proposal has not yet been officially accepted. See overview in section 2 for details.

### 5.12 Management considerations

Management of fisheries on haddock also needs to take into account measures for cod and saithe.

### 5.13 Ecosystem considerations

Since on average about $80 \%$ of the catches are taken by longlines and the remaining by trawls, the effects of the haddock fishery on the bottom is moderate.

### 5.14 Regulations and their effects

As explained in the overview (section 2), the fishery for haddock in Vb is regulated through a maximum number of allocated fishing days, gear specifications, closed areas during spawning times, closed areas for longlining close to land and large areas closed to trawling. As a consequence, around $80 \%$ of the haddock landings derive from long line fisheries. Since the minimum mesh size in the trawls (codend) is 145 mm , the trawl catches consist of fewer small fish than the long line fisheries. Other nations fishing in Faroese waters are regulated by TAC's obtained during bilateral negotiations; their total landings are minimal, however, and in 2011-2013 no agreement could be made between the Faroe Islands and EU and Norway, respectively, due to the dispute on mackerel quota sharing. In 2014, however, the parties managed to get an agreement in place again. Discarding of haddock is considered minimal and there is a ban to discarding.

### 5.15 Changes in fishing technology and fishing patterns

See section 2.

### 5.16 Changes in the environment

See section 2.

Table 5.1 Faroe Plateau (Sub-division Vb1) HADDOCK. Nominal catches (tonnes) by countries
2000-2014 and Working Group estimates in Vb

| Country | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | $2014{ }^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | $13,620^{8}$ | $13,457^{8}$ | 20,776 ${ }^{6}$ | 21,615 | 18,995 | 18,172 | 15,600 | 11,689 | 6,728 | 4,895 | 4,932 | 3,350 | 2,490 | 2,877 | 2,704 |
| France ${ }^{1}$ | 6 | $8^{7}$ | 2 | 4 | $1^{5}$ | + | $12^{5}$ | $4^{5}$ | $3^{5}$ | $2^{5}$ | $1^{7}$ | 3 |  |  |  |
| Germany | 1 | 2 | 6 | 1 | 6 |  | 1 |  |  |  |  |  |  |  |  |
| Greenland | $22^{6}$ | $0^{6}$ | $4^{4}$ |  |  |  | 1 | 94 |  | $6^{4}$ | $12^{6}$ | + | $1{ }^{4}$ |  |  |
| Iceland |  |  | 4 |  |  |  |  |  |  |  |  |  | 2 | $26^{4}$ |  |
| Norway | 355 | $257^{2}$ | 227 | 265 | 229 | 212 | 57 | 61 | 26 | 8 | 5 |  |  |  | 2 |
| Russia |  |  |  |  | 16 |  |  |  | 10 |  |  |  |  |  |  |
| Spain |  |  |  |  | 49 |  |  |  |  |  |  |  |  |  |  |
| UK (Engl. and Wales) | $19^{7}$ | $4^{7}$ | $11^{5}$ | 14 | 8 | 1 | 1 |  |  |  |  |  |  |  |  |
| UK (Scotland) ${ }^{5}$ |  |  |  | 185 | 186 | 126 | 106 | 35 | 60 | 64 |  |  |  |  |  |
| United Kingdom |  |  |  |  |  |  |  |  |  |  | $73{ }^{4}$ |  |  |  | 424 |
| Total | 14,023 | 13,728 \#' | 21,030 | 22,084 ${ }^{\prime}$ | 19,490 | 18,511 ${ }^{\prime}$ | 15,778 | 11,798 | 6,827 | 4,975 | 5,023 | 3,353 ${ }^{\circ}$ | 2,493 | 2,903 | 3,130 |
| Used in the assessmer | 15,821 0 | 15,890 | 24,933 | 27,072 | 23,101 | 20,455 | 17,154 | 12,631 | 7,388 | 5,197 | 5,202 | 3,540 | 2,634 | 2,950 | 3,194 |

1) Including catches from Sub-division Vb2. Quantity unknown 1989-1991, 1993 and 1995-2001.
2) Preliminary data
3)From 1983 to 1996 catches included in Sub-division Vb2.
3) Reported as Division Vb , to the Faroese coastal guard service.
4) Reported as Division Vb.
5) Includes Faroese landings reported to the NWWG by the Faroe Marine Research Institute

Table 5.2 Faroe Bank ( Sub-division Vb2) HADDOCK. Nominal catches (tonnes) by countries, 2000-2014.

| Country | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | $2014{ }^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | $1,565^{\text {s }}$ | 1,948 | 3,698 | 4,934 | 3,594 | 2,444 | 1,375 | 810 | 556 | 192 | 178 | 194 | 141 | 47 | 63 |
| Francel |  |  |  |  |  | + |  |  |  |  |  |  |  |  |  |
| Norway | 48 | 66 | 28 | 54 | 17 | 45 | 1 | 8 |  | 3 | 1 |  |  |  | 1 |
| UK (Engl. and Wales) | : | : | : | : | : | 1 | 4 |  |  |  |  |  |  |  |  |
| UK (Scotland) 3 |  |  | 177 | ${ }^{4}$ | : | ${ }^{1}$ | ${ }^{4}$ | 15 | 5 | 27 | 33 |  |  |  |  |
| Total | 1,798 | 2,162 | 3,903 | 4,988 | 3,611 | 1,944 | 1,376 | $833^{\prime \prime}$ | $561{ }^{\prime}$ | $222^{\prime}$ | $212^{\prime \prime}$ | $194{ }^{\prime \prime}$ | $141^{\prime \prime}$ | $47^{\prime \prime}$ | 64 |

1) Catches included in Sub-division Vbl.
2) Provisional data
3)From 1983 to 1996 includes also catches taken in Sub-division Vbl (see Table 2.4.1)
3) Reported as Division Vb
4) Provided by the NWWG

Table 5.3

Catch at age 2014

| Age | $\begin{gathered} \hline \mathrm{Vb} \\ \text { LLiners } \\ <100 \mathrm{GRT} \end{gathered}$ | $\begin{gathered} \hline \mathrm{Vb} \\ \text { LLiners } \\ >100 \mathrm{GRT} \\ \hline \end{gathered}$ | $\begin{gathered} \hline \mathrm{Vb} \\ \text { Trawl } \\ <1000 \mathrm{HP} \\ \hline \end{gathered}$ | $\begin{gathered} \mathrm{Vb} \\ \text { Trawl } \\ >1000 \mathrm{HP} \end{gathered}$ | $\begin{gathered} \hline \mathrm{Vb} \\ \text { Others } \end{gathered}$ | Vb <br> All Faroese <br> fleets | Vb <br> Foreign <br> Trawlers | $\begin{array}{\|c\|} \hline \mathrm{Vb} \\ \text { Total } \\ \text { All fleets } \\ \hline \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 157 | 35 | 12 | 10 | 0 | 214 | 18 | 232 |
| 3 | 101 | 30 | 145 | 39 | 0 | 316 | 69 | 385 |
| 4 | 235 | 122 | 160 | 39 | 0 | 556 | 70 | 626 |
| 5 | 321 | 251 | 343 | 71 | 1 | 986 | 126 | 1112 |
| 6 | 19 | 27 | 26 | 9 | 0 | 82 | 17 | 99 |
| 7 | 11 | 19 | 16 | 5 | 0 | 51 | 8 | 59 |
| 8 | 5 | 10 | 8 | 3 | 0 | 26 | 5 | 31 |
| 9 | 1 | 1 | 7 | 2 | 0 | 11 | 4 | 5 |
| 10 | 4 | 8 | 3 | 3 | 0 | 18 | 4 | 22 |
| 11 | 3 | 7 | 3 | 2 | 0 | 15 | 4 | 19 |
| 12 | 0 | 0 | 1 | 1 | 0 | 2 | 1 |  |
| 13 | 1 | 1 | 0 | 0 | 0 | 2 | 0 | 3 |
| 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total no. | 858 | 513 | 722 | 184 | 2 | 2279 | 327 | 2606 |
| Catch, t. | 885 | 656 | 738 | 215 | 2 | 2495 | 382 | 2877 |

Notes: $\quad$ Numbers in $1000^{\circ}$
Catch, gutted weight in tonnes
Others includes netters, jiggers, other small categories and catches not otherwise accounted for LLiners = Longliners OB.trawl. = Otterboard traı Pair Trawl $=$ Pair trawlers

| Comm. <br> Sampling | Vb <br> LLiners <br> $<\mathbf{1 0 0 G R T}$ | Vb <br> LLiners <br> $>\mathbf{~ 1 0 0 G R T ~}$ | Vb <br> Trawl <br> $<\mathbf{< 1 0 0 0 H P}$ | Vb <br> Trawl <br> $<1000 \mathrm{HP}$ | Vb <br> Others | Vb <br> All Faroese <br> Fleets | Vb <br> Foreign <br> Trawlers | Vb <br> Total |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| No. samples | 10 | 14 | 8 | 37 | 0 | 73 | 0 | 73 |
| No. lengths | 1918 | 2923 | 1722 | 8182 | 0 | 16942 | 0 | 16942 |
| No. weights | 1718 | 2923 | 1722 | 7951 | 0 | 16942 | 0 | 16942 |
| No. ages | 180 | 360 | 20 | 679 | 0 | 1379 | 0 | 1379 |

As compared to 2013, the sampling in 2014 was:
no samples $-5 \%$, no of lengths $-13 \%$, no of weights $16 \%$, no of otoliths $-10 \%$

## Tabel 5.4 Faroe haddock. Catch number-at-age

Run title : FAROE HADDOCK (ICES DIVISION Vb)

```
At 3/05/2015 14:07
```

| Table 1 | Catch numbers at age |  |  | Numbers*10**-3 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR, | 1957, | 1958, | 1959, | 1960, | 1961, | 1962, | 1963, | 1964, |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 0 , | 0 , | 0 , | 0 , | 0, | 0, | 0, | 0, | 0 , |  |  |
| 1, | 45, | 116, | 525, | 854, | 941, | 784, | 356, | 46, |  |  |
| 2, | 4133, | 6255, | 3971, | 6061, | 7932, | 9631, | 13552, | 2284, |  |  |
| 3, | 7130, | 8021, | 7663, | 10659, | 7330, | 13977, | 8907, | 7457, |  |  |
| 4, | 8442, | 5679, | 4544 , | 6655, | 5134, | 5233, | 7403, | 3899, |  |  |
| 5, | 1615, | 3378, | 2056, | 2482, | 1937, | 2361, | 2242, | 2360, |  |  |
| 6 , | 894, | 1299, | 1844, | 1559, | 1305, | 1407, | 1539, | 1120, |  |  |
| 7, | 585, | 817, | 721, | 1169, | 838, | 868, | 860, | 728, |  |  |
| 8 , | 227, | 294, | 236, | 243, | 236, | 270, | 257, | 198, |  |  |
| 9, | 94, | 125, | 98, | 85, | 59, | 72, | 75, | 49, |  |  |
| +gp, | 58, | 105, | 47, | 28, | 13, | 22, | 23, | 7, |  |  |
| TOTALNUM, | 23223, | 26089, | 21705, | 29795, | 25725, | 34625, | 35214, | 18148, |  |  |
| TONSLAND, | 20995, | 23871, | 20239, | 25727, | 20831, | 27151, | 27571, | 19490, |  |  |
| SOPCOF \%, | 89, | 90, | 90, | 88, | 88, | 89, | 89, | 101, |  |  |
| Table 1 | Catch | numbers at |  |  |  |  | bers*10 |  |  |  |
| YEAR, | 1965, | 1966, | 1967, | 1968, | 1969, | 1970, | 1971, | 1972, | 1973, | 1974, |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 0 , | 0 , | 0 , | 0, | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 0, |
| 1, | 39, | 90, | 70, | 49, | 95, | 57, | 55, | 43, | 665, | 253, |
| 2, | 1368, | 1081, | 1425, | 5881, | 2384, | 1728, | 717, | 750, | 3311, | 5633, |
| 3, | 4286, | 3304, | 2405, | 4097, | 7539, | 4855, | 4393, | 3744, | 8416, | 2899, |
| 4, | 5133, | 4804, | 2599, | 2812, | 4567, | 6581, | 4727, | 4179, | 1240, | 3970, |
| 5, | 1443, | 2710, | 1785, | 1524, | 1565, | 1624, | 3267, | 2706, | 2795, | 451, |
| 6 , | 1209, | 1112, | 1426, | 1526, | 1485, | 1383, | 1292, | 1171, | 919, | 976, |
| 7, | 673, | 740, | 631, | 923, | 1224, | 1099, | 864, | 696, | 1054, | 466, |
| 8 , | 1345, | 180, | 197, | 230, | 378, | 326 , | 222, | 180, | 150, | 535, |
| 9, | 43, | 54, | 52, | 68, | 114, | 68, | 147, | 113, | 68 , | 68, |
| +gp, | 8, | 9, | 13, | 12, | 20, | 10, | 102, | 95, | 11, | 147, |
| TOTALNUM, | 15547, | 14084, | 10603, | 17122, | 19371, | 17731, | 15786, | 13677, | 18629, | 15398, |
| TONSLAND, | 18479, | 18766, | 13381, | 17852, | 23272, | 21361, | 19393, | 16485, | 18035, | 14773, |
| SOPCOF \%, | 94, | 109, | 101, | 102, | 108, | 102, | 97, | 96, | 97, | 97, |


| Table 1 | Catch numbers at age |  |  | Numbers*10**-3 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR, | 1975, | 1976, | 1977, | 1978, | 1979, | 1980, | 1981, | 1982, | 1983, | 1984, |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , |
| 1, | 94, | 40, | 0 , | 0, | 1, | 0 , | 0 , | 0 , | 0 , | 25, |
| 2, | 7337, | 4396, | 255, | 32, | 1, | 143, | 74, | 539, | 441, | 1195, |
| 3, | 7952, | 7858, | 4039, | 1022, | 1162, | 58, | 455, | 934, | 1969, | 1561, |
| 4, | 2097, | 6798, | 5168, | 4248, | 1755, | 3724, | 202, | 784, | 383, | 2462, |
| 5, | 1371, | 1251, | 4918, | 4054, | 3343, | 2583, | 2586, | 298, | 422, | 147, |
| 6 , | 247, | 1189, | 2128, | 1841, | 1851, | 2496, | 1354, | 2182, | 93, | 234, |
| 7, | 352, | 298, | 946, | 717, | 772, | 1568, | 1559, | 973, | 1444, | 42, |
| 8 , | 237, | 720, | 443, | 635, | 212, | 660, | 608, | 1166, | 740, | 861, |
| 9, | 419, | 258, | 731, | 243, | 155, | 99, | 177, | 1283, | 947, | 388, |
| +gp, | 187, | 318, | 855, | 312, | 74, | 86, | 36, | 214, | 795, | 968, |
| TOTALNUM, | 20293, | 23126, | 19483, | 13104, | 9326, | 11417, | 7051, | 8373, | 7234, | 7883, |
| TONSLAND, | 20715, | 26211, | 25555, | 19200, | 12424, | 15016, | 12233, | 11937, | 12894, | 12378, |
| SOPCOF \%, | 117, | 107, | 98, | 99, | 104, | 100, | 109, | 92, | 106, | 106, |

Tabel 5.4 Faroe haddock. Catch number-at-age (cont.)

| Table 1 | Catch numbers at age |  |  | Numbers*10**-3 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR, | 1985, | 1986, | 1987, | 1988, | 1989, | 1990, | 1991, | 1992, | 1993, | 1994, |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 0, | 0 , |
| 1, | 0, | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 43, | 1, |
| 2, | 985, | 230, | 283, | 655, | 63, | 105, | 77, | 40, | 113, | 277, |
| 3, | 4553, | 2549, | 1718, | 444, | 1518, | 1275, | 1044, | 154, | 298, | 191, |
| 4, | 2196, | 4452, | 3565, | 2463, | 658, | 1921, | 1774, | 776, | 274, | 307, |
| 5, | 1242, | 1522, | 2972, | 3036, | 2787, | 768, | 1248, | 1120, | 554, | 153, |
| 6 , | 169, | 738, | 1114, | 2140, | 2554, | 1737, | 651, | 959, | 538, | 423, |
| 7, | 91, | 39, | 529, | 475, | 1976, | 1909, | 1101, | 335, | 474, | 427, |
| 8, | 61, | 130, | 83, | 151, | 541, | 885, | 698, | 373, | 131, | 383, |
| 9, | 503, | 71, | 48, | 18, | 133, | 270, | 317, | 401, | 201, | 125, |
| +gp, | 973, | 712, | 334, | 128, | 81, | 108, | 32, | 162, | 185, | 301, |
| TOTALNUM, | 10773, | 10443, | 10646, | 9510, | 10311, | 8978, | 6942, | 4320, | 2811, | 2588, |
| TONSLAND, | 15143, | 14477, | 14882, | 12178, | 14325, | 11726, | 8429, | 5476, | 4026, | 4252, |
| SOPCOF \%, | 106, | 101, | 102, | 97, | 100, | 102, | 106, | 106, | 103, | 100, |
| Table 1 | Catch | numbers at | age |  |  |  | ers*10 |  |  |  |
| YEAR, | 1995, | 1996, | 1997, | 1998, | 1999, | 2000, | 2001, | 2002, | 2003, | 2004, |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 0 , | 0 , | 0, | 0, | 0 , | 0, | 0, | 0, | 0, | 0, | 0, |
| 1, | 0 , | 1, | 0 , | 0 , | 9, | 73, | 19, | 0 , | 0 , | 3, |
| 2, | 804, | 326, | 77, | 106, | 174, | 1461, | 4380, | 1515, | 133, | 243, |
| 3, | 452, | 5234, | 2913, | 1055, | 1142, | 3061 , | 3128, | 14039, | 3436, | 2007, |
| 4, | 235, | 1019, | 10517, | 5269, | 942, | 210, | 2423, | 2879, | 13551, | 4802, |
| 5, | 226, | 179, | 710, | 9856, | 4677, | 682, | 173, | 1200, | 2224, | 10426, |
| 6 , | 132, | 163, | 116, | 446, | 6619, | 2685, | 451, | 133, | 949, | 1163, |
| 7, | 295, | 161, | 123, | 99, | 226, | 2846, | 1151, | 239, | 163, | 409, |
| 8 , | 290, | 270, | 93, | 87, | 26, | 79, | 1375, | 843, | 334, | 89, |
| 9, | 262, | 234, | 220, | 95, | 20, | 1, | 17, | 1095, | 858, | 166, |
| +gp, | 295, | 394, | 516, | 502, | 192, | 71, | 18, | 33, | 924, | 811, |
| TOTALNUM, | 2991, | 7981, | 15285, | 17515, | 14027, | 11169, | 13135, | 21976, | 22572, | 20119, |
| TONSLAND, | 4948, | 9642, | 17924, | 22210, | 18482, | 15821, | 15890, | 24933, | 27072, | 23101, |
| SOPCOF \%, | 103, | 100, | 103, | 101, | 100, | 103, | 100, | 100, | 100, | 99, |
| Table 1 | Catch | numbers at | age |  |  |  | ers*10 |  |  |  |
| YEAR, | 2005, | 2006, | 2007, | 2008, | 2009, | 2010, | 2011, | 2012, | 2013, | 2014, |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , | 0 , |
| 1, | 0, | 0 , | 0, | 6, | 0 , | 0 , | 0, | 0 , | 0, | 0 , |
| 2, | 85, | 247, | 76, | 66, | 27, | 389, | 170, | 8, | 83, | 232, |
| 3, | 1671, | 446, | 982, | 204, | 329, | 445, | 773, | 960, | 510, | 385, |
| 4, | 3852, | 2566, | 547, | 918, | 402, | 426, | 324, | 513, | 1118, | 626, |
| 5, | 6753, | 3949, | 2732, | 424, | 555, | 279, | 198, | 156, | 219, | 1112, |
| 6 , | 6127, | 5423, | 3309, | 1471, | 514, | 484, | 186, | 114, | 95, | 99, |
| 7, | 542, | 3278, | 2758, | 1706, | 1133, | 553, | 280, | 123, | 78, | 59, |
| 8, | 147, | 136, | 1117, | 1254, | 739, | 718, | 353, | 94, | 88, | 31, |
| 9, | 28, | 63, | 89, | 320, | 285, | 444, | 367, | 171, | 71, | 15, |
| +gp, | 154, | 70, | 9, | 39, | 48, | 159, | 187, | 114, | 119, | 47, |
| TOTALNUM, | 19359, | 16178, | 11619, | 6408, | 4032, | 3897, | 2838, | 2253, | 2381, | 2606, |
| TONSLAND, | 20455, | 17154, | 12631, | 7388, | 5197, | 5202, | 3540, | 2634, | 2950, | 3194, |
| SOPCOF \%, | 100, | 100, | 100, | 101, | 100, | 101, | 101, | 102, | 101, | 101, |

Table 5.5 Faroe haddock. Catch weight-at-age.


| Table 2 | Catch | weights at | age (kg) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR, | 1965, | 1966, | 1967, | 1968, | 1969, | 1970, | 1971, | 1972, | 1973, | 1974, |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 0 , | . 0000 , | . 0000, | . 0000, | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000, | . 0000, | . 0000, |
| 1, | . 2500 , | . 2500 , | . 2500 , | . 2500 , | . 2500, | . 2500, | . 2500 , | . 2500, | . 2500 , | . 2500 , |
| 2, | . 4700 , | . 4700 , | . 4700 , | . 4700, | . 4700 , | . 4700 , | . 4700, | .4700, | . 4700 , | . 4700 , |
| 3, | . 7300 , | . 7300 , | . 7300 , | . 7300 , | . 7300 , | . 7300 , | . 7300 , | . 7300 , | . 7300 , | . 7300 , |
| 4, | 1.1300, | 1.1300, | 1.1300, | 1.1300, | 1.1300, | 1.1300, | 1.1300, | 1.1300, | 1.1300, | 1.1300, |
| 5, | 1.5500, | 1.5500, | 1.5500, | 1.5500, | 1.5500, | 1.5500, | 1.5500, | 1.5500, | 1.5500, | 1.5500, |
| 6 , | 1.9700, | 1.9700, | 1.9700, | 1.9700, | 1.9700, | 1.9700, | 1.9700, | 1.9700, | 1.9700, | 1.9700, |
| 7, | 2.4100, | 2.4100, | 2.4100, | 2.4100, | 2.4100, | 2.4100, | 2.4100, | 2.4100, | 2.4100, | 2.4100, |
| 8 , | 2.7600, | 2.7600, | 2.7600, | 2.7600, | 2.7600, | 2.7600, | 2.7600, | 2.7600, | 2.7600, | 2.7600, |
| 9, | 3.0700 , | 3.0700 , | 3.0700, | 3.0700, | 3.0700 , | 3.0700 , | 3.0700 , | 3.0700 , | 3.0700 , | 3.0700 , |
| +gp, | 3.5500, | 3.5500, | 3.5500, | 3.5500, | 3.5500 , | 3.5500, | 3.5500, | 3.5500, | 3.5500, | 3.5500, |
| SOPCOFAC, | . 9383 , | 1.0885, | 1.0117, | 1.0246, | 1.0787, | 1.0249, | . 9688 , | . 9597 , | . 9690 , | . 9678 , |


| Table 2 | Catch | weights at | age (kg) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR, | 1975, | 1976, | 1977, | 1978, | 1979, | 1980, | 1981, | 1982, | 1983, | 1984, |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 0 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 00000 , | . 0000 , |
| 1, | . 2500, | . 2500 , | . 0000 , | . 0000 , | . 3000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 3590 , |
| 2, | . 4700 , | . 4700 , | . 3110 , | . 3570 , | . 3570 , | .6430, | . 4520 , | . 7000 , | . 4700 , | . 6810, |
| 3, | . 7300 , | . 7300 , | .6330, | . 7900 , | .6720, | . 7130 , | . 7250 , | . 8960 , | . 7400 , | 1.0110, |
| 4, | 1.1300, | 1.1300, | 1.0440, | 1.0350, | . 8940 , | . 9410, | . 9570 , | 1.1500, | 1.0100, | 1.2550, |
| 5, | 1.5500, | 1.5500, | 1.4260, | 1.3980, | 1.1560, | 1.1570, | 1.2370, | 1.4440, | 1.3200, | 1.8120, |
| 6, | 1.9700, | 1.9700, | 1.8250, | 1.8700, | 1.5900, | 1.4930, | 1.6510, | 1.4980, | 1.6600, | 2.0610, |
| 7, | 2.4100, | 2.4100, | 2.2410, | 2.3500, | 2.0700, | 1.7390, | 2.0530, | 1.8290, | 2.0500, | 2.0590, |
| 8 , | 2.7600, | 2.7600, | 2.2050, | 2.5970, | 2.5250, | 2.0950, | 2.4060, | 1.8870, | 2.2600, | 2.1370, |
| 9, | 3.0700 , | 3.0700 , | 2.5700, | 3.0140, | 2.6960, | 2.4650, | 2.7250, | 1.9610, | 2.5400, | 2.3680, |
| +gp, | 3.5500, | 3.5500 , | 2.5910, | 2.9200, | 3.5190, | 3.3100, | 3.2500 , | 2.8560, | 3.0400 , | 2.6860, |
| SOPCOFAC, | 1.1696, | 1.0741, | . 9784 , | . 9947 , | 1.0380, | 1.0017, | 1.0870, | . 9238, | 1.0554, | 1.0593, |

Table 5.5 Faroe haddock. Catch weight-at-age (cont.).

| Table 2 | Catch | weights at | age (kg) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR, | 1985, | 1986, | 1987, | 1988, | 1989, | 1990, | 1991, | 1992, | 1993, | 1994, |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 0 , | . 0000, | . 0000, | . 0000, | . 0000 , | . 0000, | . 0000 , | . 0000, | . 0000 , | . 0000, | . 0000 , |
| 1, | . 0000, | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000, | . 0000 , | . 3600, | . 0000 , |
| 2, | . 5280, | .6080, | .6050, | . 5010, | .5800, | . 4380, | . 5470, | . 5250, | . 7550 , | . 7540 , |
| 3 , | . 8590, | .8870, | . 8310, | . 7810 , | . 7790 , | . 6990, | . 6930, | . 7240 , | . 9820 , | 1.1030, |
| 4, | 1.3910, | 1.1750, | 1.1260, | . 9740 , | . 9230, | . 9390, | . 8840 , | . 8170 , | 1.0270, | 1.2540, |
| 5, | 1.7770, | 1.6310, | 1.4620, | 1.3630, | 1.2070, | 1.2040, | 1.0860, | 1.0380, | 1.1920, | 1.4650, |
| 6 , | 2.3260, | 1.9840, | 1.9410, | 1.6800, | 1.5640, | 1.3840, | 1.2760, | 1.2490, | 1.3780, | 1.5930, |
| 7, | 2.4400, | 2.5190, | 2.1730, | 1.9750, | 1.7460, | 1.5640, | 1.4770, | 1.4300, | 1.6430, | 1.8040, |
| 8, | 2.4010, | 2.5830, | 2.3470, | 2.3440, | 2.0860, | 1.8180, | 1.5740, | 1.5640, | 1.7960, | 2.0490, |
| 9, | 2.5320, | 2.5700, | 3.1180, | 2.2480, | 2.4240, | 2.1680, | 1.9300, | 1.6330, | 1.9710, | 2.2250, |
| +gp, | 2.6860, | 2.9220, | 2.9330, | 3.2950, | 2.5140, | 2.3350, | 2.1530, | 2.1260, | 2.2400, | 2.4230, |
| SOPCOFAC, | 1.0559, | 1.0141, | 1.0197, | . 9695 , | 1.0025, | 1.0195, | 1.0635, | 1.0554, | 1.0320, | . 9969 , |


| Table 2 | Catch weights at age (kg) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR, | 1995, | 1996, | 1997, | 1998, | 1999, | 2000, | 2001, | 2002, | 2003, | 2004, |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 0 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , |
| 1, | . 0000, | . 3600 , | . 0000 , | . 0000 , | . 2780 , | . 2800 , | . 2800 , | . 0000 , | . 0000 , | . 3670 , |
| 2, | . 6660, | . 5340, | . 5190, | . 6220, | . 5040, | .6610, | .6080, | . 5840, | . 5710, | . 5740, |
| 3 , | 1.0540, | . 8580 , | . 7710 , | . 8460 , | . 6240, | . 9360 , | . 9400 , | . 8570, | . 7150 , | . 7700 , |
| 4, | 1.4890, | 1.4590, | 1.0660, | 1.0160, | . 9740 , | 1.1660, | 1.3740, | 1.4050, | 1.0080, | . 8870, |
| 5, | 1.7790, | 1.9930, | 1.7990, | 1.2830, | 1.2200, | 1.4830, | 1.7790, | 1.7990, | 1.5370, | 1.1590, |
| 6, | 1.9400, | 2.3300, | 2.2700, | 2.0800, | 1.4900, | 1.6160, | 1.9710, | 1.9740, | 1.9110, | 1.6380, |
| 7, | 2.1820, | 2.3510, | 2.3400, | 2.5560, | 2.4560, | 1.8930, | 2.1190, | 2.3010, | 2.0910, | 1.8700, |
| 8 , | 2.3570, | 2.4690, | 2.4750, | 2.5720, | 2.6580, | 2.8210, | 2.3730, | 2.3700, | 2.3010, | 2.4380, |
| 9, | 2.4900, | 2.7770, | 2.5010, | 2.4520, | 2.5980, | 3.7490 , | 2.7500, | 2.6260, | 2.4060, | 2.3570, |
| +gp, | 2.6780, | 2.5820, | 2.6760, | 2.7530, | 2.9530, | 3.1960, | 3.9660, | 3.1300, | 2.5350, | 2.4170, |
| SOPCOFAC, | 1.0331, | 1.0043, | 1.0250, | 1.0106, | . 9973, | 1.0349, | . 9960 , | 1.0010, | 1.0049, | . 9929 , |



Table 5.6 Faroe haddock. Proportion mature-at-age.


| Table | 5 | Proportion mature at age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR, |  | 1975, | 1976, | 1977, | 1978, | 1979, | 1980, | 1981, | 1982, | 1983, | 1984, |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
| 0 , |  | . 0000 , | . 0000 , | . 0000, | . 0000 , | . 0000 , | . 0000, | . 0000, | . 0000, | . 0000, | . 0000, |
| 1, |  | . 0000 , | . 0000 , | . 0000, | . 0000 , | . 0000 , | . 0000 , | . 0000, | . 0000, | . 0000 , | . 0000 , |
| 2, |  | . 0600 , | . 0600 , | . 0600, | . 0600 , | . 0600 , | . 0600 , | . 0600, | . 0800 , | . 0800 , | . 0800 , |
| 3 , |  | . 4800 , | . 4800 , | . 4800, | . 4800 , | . 4800 , | . 4800, | . 4800, | .6200, | . 6200, | . 7600 , |
| 4, |  | . 9100, | . 9100, | . 9100, | . 9100, | . 9100, | . 9100, | .9100, | . 8900, | . 8900 , | . 9800 , |
| 5, |  | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |
| 6 , |  | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |
| 7, |  | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000 , | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |
| 8, |  | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |
| 9, |  | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |
| +gp, |  | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, | 1.0000, |

Table 5.6 Faroe haddock. Proportion mature-at-age (cont.).


Table 5.7 Faroe haddock. 2015 tuning file.


## Table 5.8 Faroe haddock 2015 xsa.

Lowestoft VPA Version 3.1
2/05/2015 23:12

Extended Survivors Analysis

| FAROE HA | HADDOCK ( | (ICES DI | VISION | $\mathrm{Vb})$ |  |  | HAD_IND |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CPUE data from file D:\Vpa\vpa2015\input-files ${ }^{\text {comb-survey-spaly-15-jr.txt }}$ |  |  |  |  |  |  |  |  |  |  |
| Catch data for 58 years. 1957 to 2 |  |  |  |  |  |  |  |  |  |  |
| Fleet, |  |  | First, Last, First, Last, Alpha, Betayear, year, age, age |  |  |  |  |  |  |  |
| SUMMER SURVEY , |  |  | 1996, | 2014, | 1, | 8, | . 600, | . 700 |  |  |
| SPRING SURVEY SHIFTE, |  |  | 1993, 2014, 0, 6, |  |  |  | . 950 , | 1.000 |  |  |
| Time series weights : |  |  |  |  |  |  |  |  |  |  |
| Tapered time weighting not applied |  |  |  |  |  |  |  |  |  |  |
| Catchability analysis : |  |  |  |  |  |  |  |  |  |  |
| Catchability independent of stock size for all ages |  |  |  |  |  |  |  |  |  |  |
| Catchability independent of age for ages $>=6$ |  |  |  |  |  |  |  |  |  |  |
| Terminal population estimation : |  |  |  |  |  |  |  |  |  |  |
| Survivor estimates shrunk towards the mean $F$ |  |  |  |  |  |  |  |  |  |  |
| S.E. of the mean to which the estimates are shrunk = . 500 |  |  |  |  |  |  |  |  |  |  |
| Minimum standard error for population estimates derived from each fleet $=$. 300 |  |  |  |  |  |  |  |  |  |  |
| Prior weighting not applied |  |  |  |  |  |  |  |  |  |  |
| Tuning converged after 35 iterations |  |  |  |  |  |  |  |  |  |  |
| Regression weights |  |  |  |  |  |  |  |  |  |  |
| Fishing mortalities |  |  |  |  |  |  |  |  |  |  |
| Age, | , 2005, | 2006, | 2007, | 2008, | 2009, | 2010, | 2011, | 2012, | 2013, | 2014 |
| 0 , | , . 000 , | . 000 , | . 000 , | . 000 , | . 000, | . 000 , | . 000 , | . 000, | . 000 , | . 000 |
| 1, | , .000, | . 000 , | . 000 , | . 002 , | . 000, | .000, | . 000 , | . 000, | . 000 , | . 000 |
| 2, | , .011, | . 036, | . 028, | . 029 , | .013, | .089, | . 012, | . 003 , | . 036 , | . 029 |
| 3, | , .085, | . 074 , | . 198, | . 096 , | .194, | . 301 , | . 258, | . 087, | . 249, | . 233 |
| 4, | , .178, | . 183, | . 123, | . 287 , | . 277 , | . 415, | . 374 , | . 272 , | .139, | . 550 |
| 5, | , . 352 , | . 280 , | . 302 , | . 132, | . 282, | . 316 , | . 345 , | . 310 , | . 178 , | . 200 |
| 6, | , .559, | . 535, | . 402 , | . 263, | . 235 , | . 425 , | . 361 , | . 342 , | . 316 , | . 114 |
| 7, | , .671, | . 672 , | . 579, | . 373, | . 333 , | . 428, | . 469 , | . 432, | . 416 , | . 330 |
| 8 , | , .622, | . 347 , | . 509, | . 571, | . 274 , | . 366 , | . 539, | . 282, | . 638, | . 288 |
| 9, | , .766, | . 600, | . 403, | . 264 , | . 241 , | . 263 , | . 322 , | . 550, | . 357 , | . 206 |

## Table 5.8 Faroe haddock 2015 xsa (cont.)

XSA population numbers (Thousands)



Taper weighted geometric mean of the VPA populations:
$2.31 \mathrm{E}+04,1.96 \mathrm{E}+04,1.64 \mathrm{E}+04,1.29 \mathrm{E}+04,8.88 \mathrm{E}+03,5.45 \mathrm{E}+03,3.20 \mathrm{E}+03,1.79 \mathrm{E}+03,8.97 \mathrm{E}+02,4.33 \mathrm{E}+02$, Standard error of the weighted Log(VPA populations) :

$$
1.1032,1.0995,1.0995,1.0746,1.0521,1.0264,1.0194,1.0309,1.1351,1.3687,
$$

Log catchability residuals.

Fleet : SUMMER SURVEY

| Age | , | 1995, | 1996, | 1997, | 1998, | 1999, | 2000, | 2001, | 2002, | 2003, | 2004 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  | No data | for t | s fle | at | age |  |  |  |  |  |
| 1 | , | 99.99, | 1.21, | . 26 , | -.15, | -. 22 , | .11, | . 15, | . 41, | . 18, | -. 28 |
| 2 | , | 99.99, | .17, | . 67, | . 07 , | -. 14, | . 27 , | . 31 , | . 21, | . 20, | . 53 |
| 3 | , | 99.99, | . 35, | .19, | -.39, | 1.53, | . 22 , | . 40 , | . 36 , | -.14, | -. 23 |
| 4 | , | 99.99, | -. 43, | . 43, | . 04 , | -.51, | -.69, | . 27 , | . 12, | . 34, | -. 17 |
| 5 | , | 99.99, | -. 11, | . 03, | . 11, | . 15, | -.10, | -.92, | . 18, | . 59 , | 31 |
| 6 | , | 99.99, | . 21 , | . 43, | -. 28 , | . 07 , | . 09 , | -.33, | -.51, | -. 14, | -. 09 |
| 7 | , | 99.99, | -.02, | -. 35, | . 97, | . 29, | . 05, | . 00 , | -. 35, | -.28, | -. 44 |
| 8 | , | 99.99, | -. 07 , | . 16, | . 63 , | . 44 , | . 29 , | -.08, | -. 27 , | . 42 , | -. 73 |
| Age | , | 2005, | 2006, | 2007, | 2008, | 2009, | 2010, | 2011, | 2012, | 2013, | 2014 |
| 0 |  | No data | for t | s fle | at | s age |  |  |  |  |  |
| 1 | , | . 29, | -. 23 , | . 06 , | . 34 , | . 39, | . 10, | -1.93, | -. 44, | -.35, | . 10 |
| 2 | , | . 25 , | . 58, | 1.18, | . 08, | -. 17, | . 11, | -. 06 , | -1.71, | -2.13, | -. 43 |
| 3 | , | . 04 , | -. 64, | -.61, | -. 15, | -.94, | . 37, | -. 10, | -.21, | .09, | -. 14 |
| 4 |  | .17, | -. 01, | -. 65, | . 22 , | . 37, | . 53, | -.29, | -.11, | -.04, | . 41 |
| 5 |  | . 09 , | . 12 , | -.19, | -.62, | . 14, | . 15, | .11, | -.31, | . 39, | -. 10 |
| 6 |  | . 73 , | . 27 , | .14, | . 01, | -. 26 , | . 28 , | -. 30, | -. 29 , | . 20 , | -. 23 |
| 7 |  | . 22 , | . 30, | . 00 , | . 25 , | . 18 , | . 10, | -. 35, | -. 04 , | . 01, | -. 19 |
| 8 |  | -1.21, | -. 53, | -. 74, | . 18, | -. 20 , | . 18 , | -. 17, | -. 66 , | . 79, | . 26 |

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age , | 1, | 2, | 3, | 4, | 5, | 6, | 7, |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log q, | -5.0205, | -5.5052, | -5.7113, | -5.6741, | -5.7982, | -5.8181, | -5.8181, |
| S.E (Log q), | .5965, | .7636, | .5269, | .3751, | .3441, | .3130, | .3288, |

## Table 5.8 Faroe haddock 2015 xsa (cont.)

Regression statistics :


Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age , | 0, | 1, | 2, | 3, | 4, | 5, |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log q, | -6.0068, | -5.3462, | -5.8437, | -5.8842, | -6.0745, | -6.3176, |
| S.E (Log q), | .7512, | .5191, | .6440, | .4340, | .7748, | .6068, |

## Table 5.8 Faroe haddock 2015 xsa (cont.)

Regression statistics :


## Table $5.8 \quad$ Faroe haddock 2015 xsa (cont.)



Age 4 Catchability constant w.r.t. time and dependent on age
Year class $=2010$


Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | Ratio, |  |  |
| $773 .$, | .18, | .39, | 10, | 2.101, | .550 |

## Table 5.8 Faroe haddock 2015 xsa (cont.)



## Table 5.8 Faroe haddock 2015 xsa (cont.)



Table 5.9 Faroe haddock. Fishing mortality (F) at age.



Terminal Fs derived using XSA (With F shrinkage)

|  | Table 8 | Fishing mortality (F) at age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAR, | 1975, | 1976, | 1977, | 1978, | 1979, | 1980, | 1981, | 1982, | 1983, | 1984, |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 0 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 |
|  | 1, | . 0015 , | . 0014 , | . 0000 , | . 0000 , | . 0002 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0006 |
|  | 2, | . 1230, | . 0908, | . 0108, | . 0010, | . 0004 , | . 0325 , | . 0237, | . 0383, | . 0252, | . 0329 |
|  | 3, | . 2650 , | . 1878, | . 1128, | . 0547 , | . 0458 , | . 0285 , | .1374, | . 4618, | .1917, | . 1167 |
|  | 4, | . 2412 , | . 3810 , | . 1815, | . 1665 , | . 1255 , | . 2025 , | . 1314 , | . 3709 , | . 3481 , | .3896 |
|  | 5, | . 2116, | . 2216, | . 5273, | . 2115, | .1913, | . 2750 , | . 2112 , | . 2918, | . 3498 , | . 2171 |
|  | 6 , | . 0957 , | . 2871, | . 7246 , | . 3820 , | .1409, | . 2136, | . 2264 , | . 2775, | .1383, | . 3336 |
|  | 7, | .0859, | . 1601 , | . 3904 , | . 5760 , | . 2721, | .1702, | . 2004 , | . 2524, | . 2991, | . 0853 |
|  | 8 , | .1599, | . 2539, | . 3788, | . 4969, | . 3303 , | . 3954 , | . 0920 , | . 2266 , | . 3102 , | . 2929 |
|  | 9, | . 1595, | . 2621, | . 4437, | . 3690 , | . 2130, | . 2526 , | . 1730 , | . 2854 , | . 2907, | . 2651 |
|  | +gp, | . 1595, | . 2621 , | . 4437, | . 3690 , | . 2130, | . 2526 , | .1730, | . 2854 , | . 2907, | . 2651 |
| FBAR | R 3-7, | .1799, | . 2476 , | . 3873, | . 2782 , | . 1551, | . 1780 , | . 1814 , | . 3309 , | . 2654 , | . 2285 |

Table 5.9 Faroe haddock. Fishing mortality (F) at age (cont.).



Terminal Fs derived using XSA (With F shrinkage)

|  | $\begin{aligned} & \text { Table } 8 \\ & \text { YEAR, } \end{aligned}$ | Fishing 1975, | $\begin{aligned} & \text { mortality } \\ & \text { 1976, } \end{aligned}$ | $\begin{aligned} & \text { (F) at } \\ & 1977, \end{aligned}$ | 1978, | 1979, | 1980, | 1981, | 1982, | 1983, | 1984, |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 0 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , |
|  | 1, | . 0015 , | . 0014, | . 0000 , | . 0000 , | . 0002 , | . 0000 , | . 0000 , | . 0000 , | . 0000 , | . 0006 , |
|  | 2, | . 1230, | . 0908 , | . 0108, | . 0010, | . 0004 , | . 0325 , | . 0237, | . 0383, | . 0252, | .0329, |
|  | 3, | . 2650 , | . 1878, | . 1128 , | . 0547 , | . 0458 , | . 0285 , | . 1374 , | . 4618 , | . 1917 , | . 1167 , |
|  | 4, | . 2412 , | . 3810 , | .1815, | . 1665 , | .1255, | . 2025, | .1314, | . 3709 , | . 3481 , | . 3896 , |
|  | 5, | . 2116, | . 2216, | . 5273, | . 2115 , | .1913, | . 2750 , | . 2112 , | . 2918, | . 3498 , | . 2171, |
|  | 6 , | . 0957 , | . 2871, | . 7246 , | . 3820 , | .1409, | . 2136, | . 2264 , | . 2775 , | .1383, | . 3336 , |
|  | 7, | .0859, | .1601, | . 3904 , | . 5760 , | . 2721, | . 1702, | . 2004 , | . 2524 , | . 2991, | . 0853, |
|  | 8, | . 1599, | . 2539, | . 3788 , | . 4969, | . 3303, | . 3954 , | .0920, | . 2266 , | . 3102 , | . 2929, |
|  | 9, | . 1595, | . 2621 , | . 4437, | . 3690 , | . 2130, | . 2526 , | . 1730 , | . 2854 , | . 2907, | . 2651 |
|  | +gp, | .1595, | . 2621, | . 4437 , | . 3690 , | . 2130, | . 2526, | . 1730, | . 2854 , | . 2907, | . 2651 |
| FBAR | 3-7, | .1799, | . 2476 , | . 3873 , | . 2782 , | . 1551, | . 1780 , | . 1814 , | . 3309 , | . 2654 , | . 2285 |

Table 5.10 Faroe haddock. Stock number (N) at age.

| Run title : FAROE HADDOCK (ICES DIVISION Vb) HAD_IND |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| At | 3/05/2015 | 14:07 |  |  |  |  |  |  |  |
|  |  | Terminal Fs derived using XSA (With F shrinkage) |  |  |  |  |  |  |  |
|  | Table 10 | Stock number at age (start of year) |  |  |  |  | Numbers*10**-3 |  |  |
|  | YEAR, | 1957, | 1958, | 1959, | 1960, | 1961, | 1962, | 1963, | 1964, |
|  | AGE |  |  |  |  |  |  |  |  |
|  | 0 , | 64927, | 54061 , | 77651, | 58761, | 71715, | 45399, | 33843, | 30192, |
|  | 1, | 47944, | 53158, | 44261, | 63576, | 48109, | 58715, | 37170, | 27709, |
|  | 2, | 35106, | 39212, | 43417, | 35763, | 51279, | 38537, | 47362, | 30110, |
|  | 3, | 25440, | 25003, | 26445, | 31954, | 23796, | 34806, | 22837, | 26515, |
|  | 4, | 20280, | 14377, | 13213, | 14717, | 16517, | 12850, | 15850, | 10638, |
|  | 5, | 5517, | 8965, | 6632, | 6706, | 6028, | 8877, | 5786, | 6278, |
|  | 6 , | 2786, | 3055, | 4284, | 3570, | 3245, | 3182, | 5132, | 2708, |
|  | 7, | 1377, | 1472, | 1326, | 1839, | 1512, | 1476, | 1332, | 2809, |
|  | 8, | 585, | 598, | 466, | 433, | 448, | 480, | 423, | 313, |
|  | 9, | 252, | 274, | 224, | 168, | 135, | 153, | 148, | 114, |
|  | +gp, | 154, | 227, | 106, | 54, | 29, | 46, | 45, | 16, |
|  | TOTAL, | 204367, | 200401, | 218024, | 217540, | 222811, | 204522, | 169929, | 137402, |


| Table 10 | Stock number at age (start of year) |  |  |  |  | Numbers*10**-3 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR, | 1965, | 1966, | 1967, | 1968, | 1969, | 1970, | 1971, | 1972, | 1973, | 1974, |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 0 , | 37948, | 81923, | 47768, | 53237, | 23136, | 49622, | 35418, | 78970, | 104848, | 83625, |
| 1, | 24719, | 31069, | 67073, | 39109, | 43587, | 18942, | 40627 , | 28998, | 64655 , | 85842, |
| 2, | 22644, | 20203, | 25356, | 54851, | 31975, | 35600, | 15457, | 33213, | 23702, | 52333, |
| 3 , | 22585, | 17302, | 15563, | 19470, | 39587, | 24022, | 27583, | 12006, | 26514, | 16410, |
| 4, | 14961, | 14613, | 11176, | 10566, | 12234, | 25590, | 15275, | 18608, | 6442, | 14092, |
| 5, | 5182, | 7604, | 7617, | 6798, | 6106, | 5884, | 14996, | 8229, | 11454, | 4152, |
| 6 , | 3005, | 2937, | 3774, | 4622, | 4187, | 3583, | 3348, | 9322, | 4288, | 6849, |
| 7, | 1204, | 1366, | 1398, | 1800, | 2403, | 2084, | 1682, | 1572, | 6573, | 2680, |
| 8 , | 1641, | 377, | 449, | 574, | 638, | 860, | 712, | 595, | 657, | 4427, |
| 9, | 77, | 127, | 146, | 189, | 262, | 180, | 409, | 382, | 325, | 402, |
| +gp, | 14, | 21, | 36, | 33, | 45, | 26, | 281, | 319, | 52, | 865, |
| TOTAL, | 133981, | 177542, | 180355, | 191249, | 164160, | 166393, | 155787, | 192213, | 249510, | 271679, |


| Table 10 | Stock number at age (start of year) |  |  |  |  | Numbers*10**-3 |  |  | 1983, | 1984, |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR, | 1975, | 1976, | 1977, | 1978, | 1979, | 1980, | 1981, | 1982, |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 0 , | 39127, | 52360, | 4153, | 7376, | 5208, | 23621, | 29256, | 60793, | 58813, | 39477, |
| 1, | 68467 , | 32035, | 42869, | 3400, | 6039, | 4264, | 19339, | 23953, | 49773, | 48152, |
| 2, | 70053, | 55971, | 26192, | 35098, | 2784, | 4944, | 3491, | 15834, | 19611, | 40751, |
| 3 , | 37750, | 50715, | 41847, | 21213, | 28707, | 2278, | 3918, | 2791, | 12476, | 15657, |
| 4, | 10812, | 23712, | 34412, | 30607 , | 16443, | 22452, | 1813, | 2796, | 1440, | 8433, |
| 5, | 7946, | 6955, | 13262, | 23498, | 21215, | 11875, | 15012, | 1301, | 1580, | 832, |
| 6 , | 2992, | 5265, | 4562, | 6408, | 15570, | 14345, | 7385, | 9951, | 796, | 912, |
| 7, | 4724, | 2226 , | 3235, | 1810, | 3581, | 11073, | 9486, | 4821, | 6173, | 567, |
| 8 , | 1772, | 3549, | 1553, | 1792, | 833, | 2233, | 7647, | 6356, | 3067, | 3747, |
| 9, | 3141, | 1237, | 2254, | 870, | 893, | 490, | 1231, | 5711, | 4149, | 1841, |
| +gp, | 1396, | 1515, | 2613, | 1109, | 424, | 423, | 249, | 946, | 3460, | 4566, |
| TOTAL, | 248179, | 235539, | 176952, | 133182, | 101697, | 97997, | 98827, | 135253, | 161337, | 164936, |

Table 5.10 Faroe haddock. Stock number (N) at age (cont.).


Table 5.11. Faroe haddock. Stock summary of the 2015 VPA.

> Run title : FAROE HADDOCK (ICES DIVISION Vb) HAD_IND

At 15/04/2014 20:12


Table 5.12. Management options table INPUT DATA descriptions

## Stock size

The stock in numbers 2015 is taken directly from the 2015 XSA. The yearclass 2014 at age 2 (in 2016) is estimated from the 2015 XSA age 1 applying a natural mortality of 0.2 in foreward calculation of the number using the standard VPA equation. The yearclass 2015 at age 2 (in 2017) is estimated as the geomean of the numbers at age 2 since 2005.

| Age | 2015 | 2016 | 2017 |
| :--- | :--- | :--- | :--- |
| 2 | 8558 | 5078 | 5089 |
| 3 | 7054 |  |  |
| 4 | 1330 |  |  |
| 5 | 773 |  |  |
| 6 | 4548 |  |  |
| 7 | 743 |  |  |
| 8 | 136 |  |  |
| $10+$ | 84 |  |  |

Numbers in thousands (predicted values rounded).

## Proportion mature at age

The proportion mature at age in 2015 is estimated as the average of the observed data in 2014 and 2015. For 2016 and 2017, the average of 2013 to 2015 is used.

| Age | 2015 | $\mathbf{2 0 1 6}$ | 2017 |
| :--- | :--- | :--- | :--- |
| 2 | 0.17 | 0.16 | 0.16 |
| 3 | 0.83 | 0.83 | 0.83 |
| 4 | 0.99 | 0.99 | 0.99 |
| 5 | 1.00 | 1.00 | 1.00 |
| 6 | 1.00 | 1.00 | 1.00 |
| 7 | 1.00 | 1.00 | 1.00 |
| 8 | 1.00 | 1.00 | 1.00 |
| 9 | 1.00 | 1.00 | 1.00 |
| $10+$ | 1.00 | 1.00 | 1.00 |

Table 5.12. Management options table INPUT DATA descriptions (cont.).

## Catch\&Stock weights at age

Catch and stock weights at age for all ages and for each of the years 2015-2017 are simply the average of the estimated point-values for 2012-2014 not re-scaled to 2014 since weights have been fluctuating without any trend during the last 3 years (no model was available to predict future mean weights at age).

| Age | 2015 | 2016 | 2017 |
| :--- | :--- | :--- | :--- |
| 2 | 0.581 | 0.581 | 0.581 |
| 3 | 0.839 | 0.839 | 0.839 |
| 4 | 1.128 | 1.128 | 1.128 |
| 5 | 1.408 | 1.408 | 1.408 |
| 6 | 1.673 | 1.673 | 1.673 |
| 7 | 1.775 | 1.775 | 1.775 |
| 8 | 1.811 | 1.811 | 1.811 |
| 9 | 1.813 | 1.813 | 1.813 |
| $10+$ | 1.938 | 1.938 | 1.938 |

## Exploitation pattern

The exploitation pattern 2015 is estimated like last year as the average fishing mortality matrix in the 3 preceding years (2012-2014) from the final VPA in 2015, without re-scaling to the terminal year (2014) since fishing mortalities have been fluctuating without any general trend during the last 3 years; the same exploitation pattern was used for all 3 years.

| Age | 2015 | 2016 | 2017 |
| :--- | :--- | :--- | :--- |
| 2 | 0.0227 | 0.0227 | 0.0227 |
| 3 | 0.1895 | 0.1895 | 0.1895 |
| 4 | 0.3203 | 0.3203 | 0.3203 |
| 5 | 0.2294 | 0.2294 | 0.2294 |
| 6 | 0.2570 | 0.2570 | 0.2570 |
| 7 | 0.3928 | 0.3928 | 0.3928 |
| 8 | 0.4028 | 0.4028 | 0.4028 |
| 9 | 0.3709 | 0.3709 | 0.3709 |
| $10+$ | 0.3709 | 0.3709 | 0.3709 |

Table 5.13
Faroe haddock. Management option table - Input data

MFDP version 1
Run: jak
Time and date: 18:12 24/04/2015
Fbar age range: 3-7


| 2017 |  |  | Mat |  | PF |  | PM |  | SWt |  | Sel | CWt |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | N | M |  |  |  |  |  |  |  |  |  |  |  |
|  | 2 | 5089 | 0.2 |  | 0.17 |  | 0 |  | 0 | 0.581 |  | 0.023 | 0.581 |
|  | 3 |  | 0.2 |  | 0.83 |  | 0 |  | 0 | 0.839 |  | 0.189 | 0.839 |
|  | 4 |  | 0.2 |  | 0.99 |  | 0 |  | 0 | 1.128 |  | 0.320 | 1.128 |
|  | 5 |  | 0.2 |  | 1 |  | 0 |  | 0 | 1.408 |  | 0.229 | 1.408 |
|  | 6 |  | 0.2 |  | 1 |  | 0 |  | 0 | 1.673 |  | 0.257 | 1.673 |
|  | 7 |  | 0.2 |  | 1 |  | 0 |  | 0 | 1.775 |  | 0.393 | 1.775 |
|  | 8 |  | 0.2 |  | 1 |  | 0 |  | 0 | 1.811 |  | 0.403 | 1.811 |
|  | 9 |  | 0.2 |  | 1 |  | 0 |  | 0 | 1.813 |  | 0.371 | 1.813 |
|  | 10 |  | 0.2 |  | 1 |  | 0 |  | 0 | 1.938 |  | 0.371 | 1.938 |

Input units are thousands and kg - output in tonnes

Table 5.14
Faroe haddock. Management option table - Results
MFDP version 1
Run: jak
Index file 24/04/2015
Time and date: 18:12 24/04/2015
Fbar age range: 3-7

| 2015 |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
| Biomass SSB | FMult | FBar | Landings |  |
| 23279 | 18133 | 1 | 0.2778 | 3820 |


| 2016 |  | FMult | FBar | 2017 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB |  |  | Landings | iomass | SSB |
| 22390 | 18912 | 0 | 0 | 0 | 25383 | 22274 |
|  | 18912 | 0.1 | 0.0278 | 522 | 24836 | 21729 |
| . | 18912 | 0.2 | 0.0556 | 1028 | 24306 | 21202 |
| . | 18912 | 0.3 | 0.0833 | 1518 | 23791 | 20690 |
| . | 18912 | 0.4 | 0.1111 | 1995 | 23293 | 20193 |
| . | 18912 | 0.5 | 0.1389 | 2457 | 22809 | 19712 |
|  | 18912 | 0.6 | 0.1667 | 2905 | 22340 | 19245 |
|  | 18912 | 0.7 | 0.1945 | 3340 | 21884 | 18792 |
|  | 18912 | 0.8 | 0.2222 | 3762 | 21443 | 18353 |
|  | 18912 | 0.9 | 0.25 | 4171 | 21014 | 17927 |
|  | 18912 | 1 | 0.2778 | 4569 | 20599 | 17514 |
|  | 18912 | 1.1 | 0.3056 | 4954 | 20195 | 17113 |
| . | 18912 | 1.2 | 0.3334 | 5329 | 19804 | 16723 |
|  | 18912 | 1.3 | 0.3611 | 5692 | 19424 | 16346 |
|  | 18912 | 1.4 | 0.3889 | 6045 | 19055 | 15979 |
|  | 18912 | 1.5 | 0.4167 | 6388 | 18697 | 15624 |
|  | 18912 | 1.6 | 0.4445 | 6721 | 18350 | 15278 |
|  | 18912 | 1.7 | 0.4723 | 7044 | 18013 | 14943 |
|  | 18912 | 1.8 | 0.5001 | 7358 | 17685 | 14618 |
| . | 18912 | 1.9 | 0.5278 | 7663 | 17367 | 14302 |
|  | 18912 | 2 | 0.5556 | 7960 | 17058 | 13996 |

Input units are thousands and kg - output in tonnes

Table $5.15 \quad$ Faroe haddock. Long-term Prediction - Input data
MFYPR version 1
Run: rei
Index file 24/04/2015
Time and date: 19:23 24/04/2015
Fbar age range: 3-7

| Age | M |  | Mat |  | PF | PM |  | SWt |  | Sel |  | CWt |  |
| ---: | ---: | ---: | :--- | :--- | :--- | :--- | :--- | :--- | :---: | :---: | :---: | :---: | :---: |
| 2 | 0.2 | 0.060 | 0 | 0 | 0.563 | 0.023 | 0.563 |  |  |  |  |  |  |
| 3 | 0.2 | 0.516 | 0 | 0 | 0.803 | 0.189 | 0.803 |  |  |  |  |  |  |
| 4 | 0.2 | 0.923 | 0 | 0 | 1.067 | 0.320 | 1.067 |  |  |  |  |  |  |
|  | 0.2 | 0.992 | 0 | 0 | 1.370 | 0.229 | 1.370 |  |  |  |  |  |  |
|  | 0.2 | 0.999 | 0 | 0 | 1.653 | 0.257 | 1.653 |  |  |  |  |  |  |
|  | 7 | 0.2 | 1.000 | 0 | 0 | 1.908 | 0.393 | 1.908 |  |  |  |  |  |
|  | 0 | 0.2 | 1.000 | 0 | 0 | 2.123 | 0.403 | 2.123 |  |  |  |  |  |
|  | 0.2 | 1.000 | 0 | 0 | 2.342 | 0.371 | 2.342 |  |  |  |  |  |  |
|  | 10 | 0.2 | 1.000 | 0 | 0 | 2.637 | 0.371 | 2.637 |  |  |  |  |  |

Weights in kilograms

Table 5.16
Faroe haddock. Long-term Prediction - Results
MFYPR version 1
Run: rei
Time and date: 19:23 24/04/2015
Yield per results

| FMult | Fbar |  | CatchNos | Yield | StockNos | Biomass | SpwnNosJan | SSBJan | SpwnNosSpwn | SSBSpwn |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 | 0 | 0 | 5.5167 | 8.2855 | 4.1248 | 7.3771 | 4.1248 | 7.3771 |
|  | 0.1 | 0.0278 | 0.1081 | 0.1877 | 4.978 | 7.0177 | 3.5884 | 6.1115 | 3.5884 | 6.1115 |
|  | 0.2 | 0.0556 | 0.1883 | 0.3116 | 4.5792 | 6.1066 | 3.1918 | 5.2025 | 3.1918 | 5.2025 |
|  | 0.3 | 0.0833 | 0.2504 | 0.3967 | 4.2706 | 5.4225 | 2.8853 | 4.5206 | 2.8853 | 4.5206 |
|  | 0.4 | 0.1111 | 0.3002 | 0.4569 | 4.0236 | 4.8914 | 2.6405 | 3.9916 | 2.6405 | 3.9916 |
|  | 0.5 | 0.1389 | 0.3411 | 0.5003 | 3.8208 | 4.468 | 2.4397 | 3.5703 | 2.4397 | 3.5703 |
|  | 0.6 | 0.1667 | 0.3755 | 0.5323 | 3.6507 | 4.1231 | 2.2717 | 3.2273 | 2.2717 | 3.2273 |
|  | 0.7 | 0.1945 | 0.4049 | 0.5561 | 3.5056 | 3.8369 | 2.1286 | 2.9431 | 2.1286 | 2.9431 |
|  | 0.8 | 0.2222 | 0.4303 | 0.5741 | 3.3801 | 3.5959 | 2.005 | 2.7041 | 2.005 | 2.7041 |
|  | 0.9 | 0.25 | 0.4526 | 0.5878 | 3.2702 | 3.3903 | 1.8971 | 2.5004 | 1.8971 | 2.5004 |
|  | 1 | 0.2778 | 0.4724 | 0.5982 | 3.1731 | 3.2129 | 1.8019 | 2.3248 | 1.8019 | 2.3248 |
|  | 1.1 | 0.3056 | 0.49 | 0.6063 | 3.0864 | 3.0583 | 1.7172 | 2.172 | 1.7172 | 2.172 |
|  | 1.2 | 0.3334 | 0.5059 | 0.6125 | 3.0086 | 2.9224 | 1.6412 | 2.0379 | 1.6412 | 2.0379 |
|  | 1.3 | 0.3611 | 0.5203 | 0.6173 | 2.9382 | 2.802 | 1.5727 | 1.9194 | 1.5727 | 1.9194 |
|  | 1.4 | 0.3889 | 0.5334 | 0.6211 | 2.8742 | 2.6947 | 1.5105 | 1.8138 | 1.5105 | 1.8138 |
|  | 1.5 | 0.4167 | 0.5454 | 0.6239 | 2.8157 | 2.5984 | 1.4538 | 1.7193 | 1.4538 | 1.7193 |
|  | 1.6 | 0.4445 | 0.5564 | 0.626 | 2.762 | 2.5116 | 1.4018 | 1.6341 | 1.4018 | 1.6341 |
|  | 1.7 | 0.4723 | 0.5666 | 0.6276 | 2.7125 | 2.4329 | 1.354 | 1.5571 | 1.354 | 1.5571 |
|  | 1.8 | 0.5001 | 0.5761 | 0.6288 | 2.6667 | 2.3613 | 1.3099 | 1.487 | 1.3099 | 1.487 |
|  | 1.9 | 0.5278 | 0.5849 | 0.6296 | 2.6242 | 2.2958 | 1.2691 | 1.4232 | 1.2691 | 1.4232 |
|  | 2 | 0.5556 | 0.5931 | 0.6302 | 2.5846 | 2.2357 | 1.2312 | 1.3646 | 1.2312 | 1.3646 |


| Reference point | F multiplier | Absolute $F$ |
| :--- | ---: | ---: |
| Fbar(3-7) | 1 | 0.2778 |
| FMax | 2.2382 | 0.6218 |
| F0.1 | 0.6571 | 0.1826 |
| F35\%SPR | 0.8581 | 0.2384 |
| Fhigh | 3.1887 | 0.8859 |
| Fmed | 0.8397 | 0.2333 |
| Flow | -99 |  |

Weights in kilograms


Figure 5.1. Haddock in ICES Division Vb. Landings by all nations 1904-2014. Horisontal line average for the whole period.


Figure 5.2. Faroe haddock. Cumulative Faroese landings from Vb.


Figure 5.3. Faroe haddock. Contribution (\%) by fleet to the total Faroese landings 2014.

Faroe Haddock LN(catch at age in numbers) for YC's 1948 onwards


Figure 5.4.


Figure 5.5. Faroe haddock. Mean weight at age (2-7). 2015-2017 are predicted values used in the short term prediction (open symbols).

Faroe Haddock - Maturity at age 1982-2014


Figure 5.6. Faroe haddock. Maturity at age since 1982. Running 3-years average of survey observations.


Figure 5.7. Commercial CPUE's for Pairtrawlers > 1000 HP and longliners > $\mathbf{1 0 0} \mathrm{HP}$.


Figure 5.8. Faroe haddock. CPUE (kg/trawlhour) in the spring and summer surveys.



Figure 5.9. Distribution of Faroe haddock catches in the summer survey (upper page) and in the spring survey (this page).


Figure 5.10. Faroe haddock. LN (catch-at-age in numbers) in the spring survey.


Figure 5.11. Faroe haddock. LN (catch-at-age in numbers) in the summer survey.

Faroe haddock. Spring survey log q residuals.


Faroe haddock. Summer survey log q residuals.


Figure 5.12. Faroe haddock survey $\log q$ residuals.


Figure 5.13. Faroe haddock. Retrospective analysis on the 2015 XSA.


Figure 5.14. Faroe haddock (Division Vb) standard graphs from the 2015 assessment.



Figure 5.14 (cont.). Faroe haddock (Division Vb) standard graphs from the 2013 assessment.


MFYPR version 1
Run: rei
Time and date: 19:23 24/04/2015

| Reference point | F multiplier | Absolute F |
| :---: | ---: | ---: |
| Fbar(3-7) | 1 | 0.2778 |
| FMax | 2.2382 | 0.6218 |
| F0.1 | 0.6571 | 0.1826 |
| F35\%SPR | 0.8581 | 0.2384 |
| Flow | -99 |  |
| Fmed | 0.8397 | 0.2333 |
| Fhigh | 3.1887 | 0.8859 |
| Weights in kilograms |  |  |

MFDP version
Run: jak
Index file 24/04/2015
Time and date: 18:12 24/04/201 Fbar age range: 3-7

Figure 5.16. Faroe haddock. Prediction output.

## SSB composition in 2016



SSB composition in 2017


Figure 5.17. Faroe haddock. Projected composition of the number by year-classes in the SSB's in 2016 and 2017.

## 6 Faroe Saithe

## Summary

The most recent benchmark assessment was completed in 2010.
Nominal landings decreased by more than $25 \%$ from 35 kt . in 2012 to 24 kt . in 2014. The corresponding estimate of fishing mortality in 2014 (average of ages 4-8 years) decreased to $\mathrm{F}=0.31$ which is lower than the historical average ( $\mathrm{F}=0.36$ ) and very close to $F_{m s y}=0.30$ and $\mathrm{F}_{\mathrm{pa}}=0.28$. Due to high fishing mortality SSB decreased substantially from 127000 t . in 2005 to 48000 t . in 2013, i.e., below $\mathrm{B}_{\text {trigger }}=55 \mathrm{kt}$. but it increased again to 70 000 t . in 2014 as a consequence of improved weights and maturity ogives.

Numbers of the most recent year-class (2011, age 3 in 2014) has increased substantially from 36 mill. in 2013 to 62 million in 2014. However a statistical separable model suggests that the 2011 year-class is not as strong as the spaly assessment estimate and it predicts recruitment for 2014 at 20 mill.

At status-quo $F_{b a r}(2015)=0.31$ and recruitment $\operatorname{Rec}(2015)=27$ mill. the SSB is predicted to increase to 97 kt . in 2016.

Predicted landings for 2014 in the last year assessment were around 38 kt while the actual measurement was 24 kt . The estimate of $\mathrm{F}_{\text {bar }}$ in 2014 was $\mathrm{F}_{\mathrm{bar}}=0.53$ in last year's assessment and $\mathrm{F}_{\mathrm{bar}}=0.32$ in the 2015 assessment. Recruitment strength for 2014 was predicted at 28 million while the estimate for that year in the present assessment reached 62 million. SSB was predicted exactly in 2014 SSB(2014)=70 000 t .

### 6.1 Stock description and management units.

See the stock annex.

### 6.2 Scientific data

### 6.2.1 Trends in landings and fisheries

Nominal landings of saithe from Faroese grounds (Division Vb) have varied cyclically between 10000 t and 68000 t since 1961. After a third high of about 60000 t in 1990, landings declined steadily to 20000 t in 1996. Since then landings have increased to 68 000 tonnes in 2005 (Table 6.2.1.1, Figure 6.2.1.1) but has declined to 57000 tonnes in 2008 and 2009. After a substantial drop in landings in 2011 which was the lowest observed since 1999 ( 33000 t ) landings increased by $20 \%$ in 2012 up to 35000 t . The total tonnage in 2014 is the lowest observed since 1997. The historical average landings for saithe since 1961 is 37000 t .

Since the introduction of the 200 miles EEZ in 1977, the saithe fishery has been prosecuted mostly by Faroese vessels. The principal fleet consists of large pair trawlers ( $>1000 \mathrm{HP}$ ), which have a directed fishery for saithe, about $50-77 \%$ of the reported landings in 1992-2011 (Table 6.2.1.2). The smaller pair trawlers ( $<1000 \mathrm{HP}$ ) and single trawlers $(400-1000 \mathrm{HP})$ have a more mixed fishery and they have accounted for about $10-20 \%$ of the total landings of saithe in the 1997-2011 period while the percentage of total landings by large single trawlers $(>1000 \mathrm{HP})$ has declined drastically to just $1 \%$. Historically the catch composition by the pair-trawler fleet has accounted for about $75 \%$ of the total tonnage for saithe but since 2007 it has increased gradually up to $96 \%$ in 2014 due mainly to the gear-shifting of single-trawlers to pair-trawling. The share of catches by the jigger fleet was about $8 \%$ in the 1985-1998 period but has decreased to
less than $0.5 \%$ since 2000 and it now accounts for only $2 \%$ of the total domestic landings for saithe in 2013. Foreign catches that have been reported to the Faroese Authorities but not officially reported to ICES are also included in the Working Group estimates. Catches in Subdivision IIa, which lies immediately north of the Faroes, have also been included. Little or no discarding is thought to occur in this fishery. Effort (measured as the ratio of nominal to used fishing days by the pair-trawl fleet segment) has diminished considerably in recent years. In the 2012/2013 fishing year only $85 \%$ and $57 \%$ of fishing days were utilized in the inner and outer areas respectively while in the 2013/2014 fishing year these ratios went down to $58 \%$ and $41 \%$.

Cumulative landings of saithe for the domestic fleets since 2000 are shown in Figure 6.2.1.2. The period from 2011 to 2014 are among the poorest in the time series. The progression of landings in the first two months of 2015 is below monthly averages and suggest a poor fishing year.

### 6.2.2 Catch at age

Catch at age is based on length, weight and otoliths samples from Faroese landings of small and large single and pair trawlers, and landing statistics by fleet provided by the Faroese Authorities. Catch at age is calculated for each fleet by four-month periods and the total is raised by the foreign catches. Minor adjustments were made to the catch-atage matrix for 2013 due to revised final catch statistics (Tables 6.2.2.1 and 6.2.2.2). Most of the age-disaggregated catch matrix is comprised of catches of the pair-trawl fleet. Since 2010 catch numbers is mostly comprised of age-groups 4 to 6 whereas in the period from 2005 to 2009 it is mainly composed of age-groups 4 to 8 . Numbers of 4 to 6 years old were higher in 2014 than in 2013. while catches of 3-year old saithe in increased from 721 thous. in 2013 to 878 thous. in 2014.

The sampling program and sampling intensity in 2014 as well as the approach used in compiling catch numbers is the same as in preceding years. Sampling levels of catches in both 2012 and 2013 are quite similar ( $5.6 \%$ and $5.4 \%$ respectively) going up to $8.9 \%$ in 2014 (Table 6.2.2.3.) The average amount sampled per tonnes landed since 2000 is $5.9 \%$.

### 6.2.3 Weight at age

Mean weights at age have varied by a factor of about 2 during the 1961-2013 period. Mean weights at age were generally high during the early 1980s and they subsequently decreased from the mid 1980s to the early 1990s (Table 6.2.3.1 and Figure 6.2.3.1). Mean weights increased again in the period 1992-96 but have shown a general decrease thereafter. With the exception of 3-years old saithe all age groups were showing signs of increasing size since 2006. By 2011 age-classes 4 to 8 were approaching or at long term average. This trend seemed to continue for older age groups ( 7 and older) whereas weight of 4 to 6 years old individuals appeared to decrease again in 2012 and 2013. Mean weight of the 2011 year-class (age 3 in 2014) is estimated at 1.37 kg . which is an increase with respect to that in 2013 ( 1.21 kg .). Since 2001 all age groups have remained below the historical average with the only exception of 7 -years old saithe, which reached the long-term mean value ( 3.785 kg .) in 2012 and 3-years old with size above average in 2009. In 2014 all age classes are above or just above the historical average. Mean weights at age in the stock are assumed equal to those in the catch.

### 6.2.4 Maturity at age

Maturity at age data from the spring survey is available from 1983 onward (Steingrund, 2003.) Due to poor sampling in 1988 the proportion mature for that year was calculated as the average of the two adjacent years. At the 2012 working group a model using maturity at age from the Faroese groundfish spring survey was implemented to derive smoothed trends in maturity by age and year. The fitting was done locally and the smoothing level was chosen as a trade-off between retaining the trend in maturities and reducing the data noise. For 1962 to 1982 the average maturity of predicted ogives of the 1983-2011 period was used (Table 6.2.4.1 and Figure 6.2.4.1.) Maturity ogives were low from the early and mid-1990s up to 2001 where they began to rise considerably and are above historical average since 2012.

Faroe saithe begins to mature at 3 years old, approximately $20 \%$ are mature at age 4, $50 \%$ at 5 years old and $100 \%$ are mature at age 9 and onwards.

### 6.2.5 Indices of stock size

### 6.2.5.1 Surveys

There are two annual groundfish surveys conducted in Faroese waters. The spring survey series (FGFS1) are available since 1994, while the summer survey (FGFS2) was initiated in 1996. The design for both bottom-trawl surveys is depth stratified with randomised stations covering the Faroe Plateau area. The total number of stations in the summer and spring is 100 and 200 respectively. Effort is recorded in terms of minutes towed approximately 60 min . Large proportion of saithe is caught in relatively few hauls and the inter-annual variability of these hauls is considerable.

Survey catch rates (kg per hour), length composition and age-disaggregated indices are presented in figures 6.2.5.1.1 to 6.2.5.1.5. Both surveys suggest low abundances of saithe in mid- and late 1990's and increasing numbers from 2001 to 2005 although they differ in the order of this magnitude. Since 2006 the indexes show that the saithe stock is at low levels while there are indications of a slight upward trend since 2011. Both surveys agreed not only in the direction but also in the magnitude of this positive trend. Since 2011 the most recent estimate of the spring survey suggest a slightly decrease in stock biomass for 2015 but given the uncertainty associated with the index the point estimate ought to be taken with caution. Both survey at age numbers agreed in the lack of year classes present in the stock since 2007. The spring index suggest that the 2002 year class (age 3 in 2015) may be relatively strong, which is confirmed by more abundant individuals in the $35-45 \mathrm{~cm}$ size range from length distribution data.

Given the extreme schooling behaviour of saithe the internal consistency in the spring survey measured by the correlation of numbers in the data matrix for the same year class is reasonably good, with $\mathrm{R}^{2}$ close to 0.85 for the best defined age groups and below $\mathrm{R}^{2}=0.3$ for other age classes (Figure 6.2.5.1.6). Internal consistency in the age-disaggregated fall survey is displayed in figure 6.2.5.1.7. In terms of internal consistency the spring survey outperforms the fall survey.

### 6.2.5.2 Commercial CPUE

The CPUE series that has been used in the assessment since 2000 was introduced in 1998 (ICES C.M. 1998/ACFM:19), and consists of saithe catch at age and effort in hours, referred to as the pair trawler series. A GLM model and a survey spatial scaling factor is used to standardised the CPUE series (Stock Annex B.4., Benchmark report, WKROUND 2010.) The benchmark working group regarded this novel approach to
developing the commercial series as reasonable (Benchmark report, WKROUND 2010.) Predicted annual CPUEs derived from this approach suggests that stock abundance was low in the 1990s and increased subsequently in the 2000s. and a sharp downward trend from 2006 to 2011. Since 2012 the predicted CPUE has remained remarkably stable at approximately $375 \mathrm{~kg} /$ hour (Figure 6.2.5.1.1)

The correlation between predicted CPUE and the spring and summer surveys is $\mathrm{R}^{2}=0.56$ and $\mathrm{R}^{2}=0.68$ respectively. The agreement between the survey indices measured by their correlation is estimated at $\mathrm{R}^{2}=0.36$.
. The age composition indicates that the pair-trawl fleet targets mostly age groups 4 to 6. (Figure 6.2.5.2.1) There is a good agreement between age-disaggregated indices in the commercial index and indices of the same year class one year later (Figure 6.2.5.2.2) as measured by $\mathrm{R}^{2}>0.35$ for all age-classes.

### 6.2.5.3 Information from the fishing industry

No additional information beyond the landings from the commercial fleet was presented for incorporation in the assessment.

### 6.3 Methods

The assessment model adopted at the benchmark assessment in 2010 is described in the Stock annex (Sec. C) and in the benchmark report (WKROUND 2010.) The 2010 XSA was calibrated with the standardized pair trawlers with catchability independent of stock size for all ages, catchability independent of age for ages $\geq 8$, the shrinkage of the SE of the mean $=2.0$, and no time tapered weighting. The tunings series used are shown in Table 6.3.1. Commercial catch-at age data (ages 3-14+, years 1961-2013) were calibrated in the XSA model using the commercial pair-trawl fleet (ages 3-11, years 1995-2013). XSA model diagnostics of the spaly run is presented in Table 6.3.2. Patterns in log-catchability residuals from the XSA model are relatively random but with large positive blocks in 2006-2010 for 3 to 4 age-classes (Figure 6.3.1.). Residuals from a separable statistical model predicting catch numbers at age and survey data and modelling selectivity over 3 distinct periods are also presented (Figure 6.3.3)

### 6.4 Reference points

### 6.4.1 Biological reference points and MSY framework

In 2014 at the WKMSYREF2 workshop the EqSim simulation framework was used to explore candidates to Fmsy. The work was presented at the NWWG meeting in 2014 and the results agree with the previous simulations (see above) in that estimates of Fmsy are in the range of Fmsy= 0.30 and Fmsy=0.34 and not as the present level of Fmsy $=0.28$. In the 2014 meeting ACOM adopted the EqSim framework and agreed to set Fmsy=0.30, which agrees with the estimation of Fmed=0.31. Below it is an excerpt from the WKMSYREF2 report:

The EqSim framework fits three stock-recruit functions (Ricker, Beverton-Holt and Hockey-stick) on the bootstrap samples of the stock and recruit pairs from which approximate joint distributions of the model parameters can be made. The result of this is projected forward for a range of F's values and the last 50 years are retained to calculate summaries. Each simulation is run independently from the distribution of model and parameters. Error is introduced within the simulations by randomly generating process error about the constant stock recruit fit, and by using historical variation in maturity, natural mortality, weight at age, etc.

In the EqSim simulations the Hockey-Stick stock-recruit function were used assuming assessment and autocorrelation errors. Figures 6.4.1.1 and 6.4.1.2 illustrate the results of these simulations which suggest that candidates for FMSY are FMSY $=0.34$ (median yield) and FMSY $=0.30$ ( F that gives the maximum mean yield in the long term) lie above the current FMSY $=\mathrm{Fpa}=0.28$ if autocorrelation and assessment errors are included in the simulation framework. If errors are ignored then estimates for FMSY are predicted to FMSY $=0.38$ (median yield), FMSY $=0.35$ (maximum mean yield). No Blim is defined for faroe saithe but for the purposes of the analysis a value of Blim=Bpa/1.4 was set for the simulations. A more detailed information of the simulations are available under http://www.ices.dk/community/groups/Pages/WKMSYREF2.aspx A summary is given in the table below.

|  | F | SSB | Catch | option |
| :--- | :--- | :--- | :--- | :--- |
| Flim | 0.34 | 87327.43 | 36479.8 | ass. Error |
| Flim | 0.37 | 79116.87 | 35447.45 | ass. Error |
| Flim | 0.46 | 38905.3 | 22023.28 | ass. Error |
| MSY:median | 0.34 | 88565.78 | 36665.24 | ass. Error |
| Maxmeanland | 0.30 | 101372.9 | 37109.88 | ass. Error |
| FCrash5 | 0.41 | 63312 | 31637.31 | ass. Error |
| FCrash50 | 0.52 | 855.73 | 550.19 | ass. Error |
| Flim | 0.40 | 78435.72 | 38526.07 | No ass. Error |
| Flim | 0.42 | 73052.08 | 37660.27 | No ass. Error |
| Flim | 0.50 | 38910.57 | 24279.75 | No ass. Error |
| MSY:median | 0.38 | 82329.53 | 38694.43 | No ass. Error |
| Maxmeanland | 0.35 | 90688.34 | 39167.13 | No ass. Error |
| FCrash5 | 0.43 | 69750.99 | 37114.99 | No ass. Error |
| FCrash50 | 0.54 | 2847.53 | 1910.51 | No ass. Error |

MSY and revised precautionary reference points (Section 2. Demersal stocks in the Faroe Area, Subsection 2.1.7 Faroe saithe) for faroe saithe are listed below:

| Biological reference points | NWWG 2012 | NWWG2014 |
| :--- | :--- | :--- |
| Btrigger | 55000 t. | 55000 t. |
| Blim | not defined. |  |
| Bpa | 60000 t. |  |
| Flim | not defined |  |
| Fpa | 0.28 | 0.30 |
| Fmsy | 0.32 |  |

The Yield/R and SSB-R calculations with respect to reference fishing mortalities (Fmax, Fmed and F0.1) is presented in the table below. The SSB-R plot in relation to Fhigh, Fmed and Flow is shown in Figure 6.4.1.3.

|  | Fish Mort <br> Ages 4-8 | Yield/R | SSB/R |
| :--- | :--- | :--- | :--- |
| Average last 3 years | 0.44 | 1.29 | 2.23 |
| Fmax | 0.42 | 1.29 | 2.36 |
| F0.1 | 0.15 | 1.15 | 6.10 |
| Fmed | 0.31 | 1.28 | 3.27 |

### 6.5 State of the stock

Recruitment in the 1980s was close to the historical average ( 32 millions). The strongest year class since 1986 was produced in the 1990s and the average for that decade was about 28 millions (Figures 6.5.1 to 6.5.4. and Tables 6.5.1 to 6.5.3). The 1998 ( 88 millions) and 1999 ( 106 millions) are the largest observed in the time series. Since 2006 estimated recruitment has remained at low levels in comparison with the exceptionally high recruitment pulses observed from 2001 to 2005. However the 2011 year-class (numbers of age-3 saithe in 2014) is estimated at 62 million and therefore far above the historical average of 32 million. Nevertheless the most recent recruitment estimate is highly unreliable and it contradicts with the estimate from a more sophisticated statistical model, which predicts recruitment at $\mathrm{N}_{3}(2014)=20$ million and thus in line with the present low productivity period.

Relatively low Fs during the 1960s and recruitment above average in early-1970s caused an increase in SSB well above the historical average around the mid-1970s while landings peaked to almost 58000 t . in 1973. Increasing Fs since 1980 lead to a decrease in the spawning stock biomass of saithe throughout the mid-1980s although recruitment of the 1983 year class rose to 662000 millions, i.e. double the average from 1961 to 2014. The historically low SSB persisted in 1992-1998 and this along with low Fs caused landings to steeply decline to around 20000 tonnes in 1996. The SSB increased since 1999 to above 128000 t in 2005 with the maturation of the 1995, 1996, 1997 and 1999 year classes and decreased to 93000 t in 2009. The 2014 spaly assessment indicates that the point estimator of $\operatorname{SSB}(2013)$ is approximately 70000 t . Since 2005 SSB has been declining sharply and at present is above $B_{\text {trigger }}=55000 \mathrm{t}$. Figure 6.5.6 illustrates the numbers of mature fish in the stock forage-groups from 3 to 9 in 2006, 2013 and 2014. It is quite clear that there has been a substantial increase in the numbers of mature fish over the age groups 3 to 6 a phenomenon supported by increased maturity ogives in recent years The separable catch-at-age model predicts $\operatorname{SSB}(2014)=94000 \mathrm{t}$. and is thus at historical average.

In 2014 average fishing mortality over age groups 4 to 8 (Fbar) is estimated at $\mathrm{F}(2014)=0.32$ and therefore very closed to $\mathrm{F}_{\text {msy }}=0.30$ and below average for the first time since 2005. On the other hand the statistical model framework suggests that $F(2014)=0.23$ is even lower than that of the spaly assessment. The assessment model suggests a drop in fishing mortality from 2013 to 2014 reflecting the abrupt decline in landings from 26 kt . to 24 kt . Estimated $\mathrm{F}^{\prime}$ s have been above $\mathrm{F}_{\mathrm{msy}}=0.30$ and $\mathrm{F}_{\mathrm{pa}}=0.28$ since 1998.

The relation between stock and recruitment is presented in figure 6.5.7.

### 6.6 Short term forecast

### 6.6.1 Input data

Population numbers at age 3 for the base short term prediction is calculated as the geometric mean of estimated recruitment strength from 2008 to 2012. Natural mortality is set to constant 0.2. Weight-at-age for 3-years old saithe is predicted by the year class strength (number of 3 -years old in the stock) with a 3 year time lag (Eq. 1) whereas weight for ages 4 to 8 is estimated by weight-at-age the previous year from the same year class (Eq. 2) Weight for ages 9 to $14+$ is an average of the most 3 recent years. Diagnostics and results of the model are shown in Figures 6.6.1.1 and 6.6.1.2. For older age groups ( 9 to $14+$ ) a 3-year average is used.
$\mathrm{W} 3, \mathrm{y}=\alpha \mathrm{N} 3, \mathrm{y}-3+\beta$
for $\mathrm{a}=3$ (Eq. 1)
$\mathrm{Wa}+1, \mathrm{y}+1=\alpha \mathrm{Wa}, \mathrm{y}+\beta$
$W a, y=(W a-3, y W a-2, y W a-1, y) / 3$
for $4 \leq \mathrm{a} \leq 8 \quad$ (Eq. 2)

Proportion mature for 2015-2017 is taken as the average of predicted maturity ogives from 2013 and 2015. The exploitation pattern used is a 3 year average rescaled to last year as specified in the stock annex.

Input data for the prediction with management options for the spaly scenario are presented in Table 6.6.1.1.

### 6.6.2 Projection of catch and biomass

Results from predictions with management option is presented in Table 6.6.2.1 and Figure 6.6.2.1.
At status quo $\mathrm{F}=0.32$ landings would increase to 35 kt . in 2015 and 37 kt . in 2016 while spawning stock biomass is expected to around 82 kt . in 2015 and increase to 96 kt . tonnes in 2016. Landings in 2015 are predicted to rely on the 2009, 2010 and 2011 year classes (79\%) while in the SSB these year-classes will contribute to around $73 \%$ of the spawning biomass in 2015 (Figure 6.6.2.2.)

### 6.7 Yield per recruit and medium term forecasts

No medium term projections were performed for faroe saithe.

## Input data to yield per recruit

The input data to long-term prediction are shown in Table 6.7.1.1.
Mean weights-at-age for 1981-2013 were used for the long term projection. Natural mortality is set to constant 0.2 . Proportion mature-at-age is taken as the average from 1983-2014.

The exploitation pattern was set equal to the average of the last five years (2005-2013) (as suggested from ACFM, 2004). Results from the yield per recruit analysis is shown in Figure 6.7.1.1.

### 6.8 Uncertainties in assessment and forecast

In 2014 the amount of catch sampled was $8.9 \%$, which is regarded as adequate.
The assessment of Faroe saithe is relatively uncertain due to lack of good tuning data although the internal consistency in the commercial fleets used to calibrate the XSA
model is reasonable considering the nature of the species that is highly schooling, and widely migrating. The retrospective pattern (Figure 6.8.1) reveals some of the assessment uncertainty. It shows periods of over- and underestimation in average fishing mortality and consequently under- and overestimation in spawning stock biomass. Over- and underestimation seem to occur in periods of poor and high abundances respectively. Various factors could explain this phenomenon, e.g., by changes in the vertical distribution of the stock or changes in the selection pattern that have been observed in recent years. With respect to recruitment the retrospective trend suggests an overestimation of incoming year-classes. To avoid large year-to-year fluctuations in the spawning stock biomass (also dependent on age structure) a locally fitting model was implemented in 2012 to reduce variability in maturities.

### 6.9 Comparison with previous assessment and forecast

The 2014 assessment predicted recruitment for 2014 to around 28 million while the observed year-class strength was 62 million (Table 6.9.1). Fishing mortality was overestimated from $\mathrm{F}=0.53$ to $\mathrm{F}=0.32$. The spawning stock biomass was predicted exatly. Landings for 2014 were predicted at $\operatorname{Land}(2014)=38 \mathrm{kt}$. while actual observed catches in that year reached $\operatorname{Land}(2014)=24 \mathrm{kt}$ an overestimation of $40 \%$. Landings and F estimates from the statistical model were however closer to the actual measurements $F(2014)=0.23$, $\operatorname{Land}(2014)=27 \mathrm{kt}$. while recruitment $\operatorname{Rec}(2014)=20$ mill. was three times lower than that of the spaly run.

### 6.10 Management plans and evaluations

No management plan exists for saithe in Division Vb

### 6.11 Management considerations

Management consideration for saithe is under the general section for Faroese stocks.
In 2014 ACOM adopted $\mathrm{F}_{\text {msy }}=0.30$ presented at the NWWG meeting for the same year and produced in the WKMSYREF2 workshop on reference points. Btrigger is set at Bloss $=55 \mathrm{kt}$. ( $\left.\mathrm{B}_{\text {trigger }}=55 \mathrm{kt}\right)$.

### 6.12 Ecosystem considerations

No evidence is available to indicate that the fishery is impacting the marine environment. A Ph.D. project was initiated in 2008, with the aim of investigate the role of environmental indicators in the dynamics of Faroe saithe. The results and conclusions of the PhD will be available to the working group in future meetings.

### 6.13 Regulations and their effects

It seems to be no relationship between number of fishing days and fishing mortality, probably because of large fluctuations in catchability. Area restriction is an alternative to reduce fishing mortality- and this is used to protect small saithe in Faroese area.

### 6.14 Changes in fishing technology and fishing patterns

See section 6.2.

### 6.15 Changes in the environment

According to existing literature the productivity of the ecosystem clearly affects both cod and haddock recruitment and growth (Gaard et al., 2002), a feature outlined in Steingrund and Gaard (2005). The primary production on the Faroe Shelf (< 130 m depth), over the period May through June, varied interannually by a factor of five, giving rise to low- or high-productive periods of $2-5$ years duration (Steingrund and Gaard, 2005). The productivity over the outer areas seems to be negatively correlated with the strength of the Subpolar Gyre (Hátún et al., 2005; Hátún et al., 2009; Steingrund et al., 2010), which may regulate the abundance of saithe in Faroese waters (Steingrund and Hátún, 2008). When comparing a gyre index (GI) to saithe in Faroese waters there was a marked positive relationship between annual variations in GI and the total biomass of saithe lagged 4 years (Figure 6.15.1.)

There is a negative relationship between mean weight-at-age and the stock size of saithe in Faroese waters. This could be due to simple density-dependence, where there is a competition for limited food resources. Stomach content data show that the food of saithe is dominated by blue whiting, Norway pout, and krill, and the annual variations in the stomach fullness are mainly attributable to variations in the feeding on blue whiting. There seems to be no relationship between stomach fullness and weights-atage for saithe (í Homrum et al. WD 2009).

### 6.16 References

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Table 6.2.1.1. Faroe saithe (Division Vb). Nominal catches (tonnes round weight) by countries 19882014 as officially reported to ICES.

| Country | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 94 | - | 2 | - | - | - | - | - | - | - | - | - | - | - |
| Estonia | - | - | - | - | - | - | - | - | - | 16 | - | - | - | - |
| Faroe Islands | 44402 | 43624 | 59821 | 53321 | 35979 | 32719 | 32406 | 26918 | 19267 | 21721 | 25995 | 32439 |  | 49676 |
| France 3 | 313 | - | - | - | 120 | 75 | 19 | 10 | 12 | 9 | 17 | - | 273 | 934 |
| Germany | - | - | - | 32 | 5 | 2 | 1 | 41 | 3 | 5 | - | 100 | 230 | 667 |
| German Dem.Rep. | - | 9 | - | - | - | - | - | - | - | - | - | - | - | - |
| German Fed. Rep. | 74 | 20 | 15 | - | - | - | - | - | - | - | - | - | - | 5 |
| Greenland | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Ireland | - | - | - | - | - | - | - | - | - | - | - | 0 | 0 | 0 |
| Netherlands | - | 22 | 67 | 65 | - | - | - | - | - |  | - | 160 | 72 | 60 |
| Norway | 52 | 51 | 46 | 103 | 85 | 32 | 156 | 10 | 16 | 67 | 53 | - | - | - |
| Portugal | - | - | - | - | - | - | - | - | - | - | - | - | 20 | 1 |
| UK (Eng. \& W.) | - | - | - | 5 | 74 | 279 | 151 | 21 | 53 | - | 19 | 67 | 32 | 80 |
| UK (Scotland) | 92 | 9 | 33 | 79 | 98 | 425 | 438 | 200 | 580 | 460 | 337 | 441 | 534 | 708 |
| USSR/Russia 2 | - | - | 30 | - | 12 | - | - | - | 18 | 28 | - | - | - | - |
| Total | 45027 | 43735 | 60014 | 53605 | 36373 | 33532 | 33171 | 27200 | 19949 | 22306 | 26065 | 33207 | 1161 | 52131 |
| Working Group estimate 45 | 45285 | 44477 | 61628 | 54858 | 36487 | 33543 | 33182 | 27209 | 20029 | 22306 | 26421 | 33207 | 39020 | 51786 |
| Country | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |  |
| Denmark | - | - | - | - | 34 | - | - | - | - | - | - | - |  |  |
| Estonia | - | - | - | - | - | - | - | - | - | - | - | - |  |  |
| Faroe Islands | 55165 | 47933 | 48222 | 71496 | 70696 | 64552 | 61117 | 61889 | 46686 | 32056 | 38175 | 28609 | 25440 |  |
| France | 607 | 370 | 147 | 123 | 315 | 108 | 97 | 68 | 46 | 135 | 40 | 31 |  |  |
| Germany | 422 | 281 | 186 | 1 | 49 | 3 | 3 | 0 |  |  |  |  |  |  |
| Greenland | 125 | - |  |  | 73 | 239 | 0 | 1 |  |  | 1 |  |  |  |
| Irland | - | - | - | - | - | - | - | - |  |  |  |  |  |  |
| Iceland | - | - | - | - | - | - | - | 148 | - |  |  |  |  |  |
| Netherlands | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 |  |  |  |  |  |
| Norway | 77 | 62 | 82 | 82 | 35 | 81 | 38 | 23 | 28 |  |  |  | 165 |  |
| Portugal | - | - | 5 | - | - | - | - | - |  |  |  |  |  |  |
| Russia | 10 | 32 | 71 | 210 | 104 | 159 | 38 | 44 | 3 |  |  | 1 |  |  |
| UK (E/W/NI) | 58 | 89 | 85 | 32 | 88 | 4 | - | - |  |  |  |  |  |  |
| UK (Scotland) | 540 | 610 | 748 | 4322 | 1011 | 408 | 400 | 685 |  |  |  |  |  |  |
| United Kingdom | - | - | - | - | - | - | - | - | 706 | 19 |  | 1 | 340 |  |
| Total | 57004 | 49377 | 49546 | 76266 | 72405 | 65557 | 61693 | 62858 | 47469 | 32210 | 38216 | 28642 | 25945 |  |
| Working Group estimate 4567 | 53546 | 46555 | 46355 | 67967 | 66902 | 60785 | 57044 | 57949 | 43885 | 29658 | 35314 | 26463 | 23854 |  |

Table 6.2.1.2. Faroe saithe (Division Vb). Total Faroese landings (rightmost column) and the contribution (\%) by each fleet category (1985-2014). Averages for 1985-2014 are given at the bottom.

|  |  |  |  | $\frac{\stackrel{ \pm}{ \pm}}{\overline{=}}$ | ¿ 응 응 |  |  |  |  |  | Industrial trawl |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 0.2 | 0.1 | 0.1 | 0.0 | 2.6 | 6.6 | 33.7 | 28.2 | 28.2 | 0.1 | 0.2 | 0.2 | 42598 |
| 1986 | 0.3 | 0.2 | 0.1 | 0.1 | 3.6 | 2.8 | 27.3 | 27.5 | 36.5 | 0.1 | 0.7 | 0.9 | 40107 |
| 1987 | 0.7 | 0.1 | 0.3 | 0.4 | 5.6 | 4.1 | 20.4 | 22.8 | 44.2 | 0.1 | 1.1 | 0.0 | 39627 |
| 1988 | 0.4 | 0.3 | 0.1 | 0.3 | 6.5 | 6.8 | 20.8 | 19.6 | 43.6 | 0.1 | 1.3 | 0.1 | 43940 |
| 1989 | 0.9 | 0.1 | 0.3 | 0.2 | 9.3 | 5.4 | 17.7 | 23.5 | 41.1 | 0.1 | 1.3 | 0.0 | 43624 |
| 1990 | 0.6 | 0.2 | 0.2 | 0.2 | 7.4 | 3.9 | 19.6 | 24.0 | 42.8 | 0.2 | 0.9 | 0.0 | 59821 |
| 1991 | 0.6 | 0.1 | 0.1 | 0.6 | 9.8 | 1.3 | 13.9 | 26.5 | 46.2 | 0.1 | 0.8 | 0.0 | 53321 |
| 1992 | 0.4 | 0.4 | 0.0 | 0.0 | 10.5 | 0.5 | 7.1 | 24.4 | 55.6 | 0.1 | 1.0 | 0.0 | 35979 |
| 1993 | 0.6 | 0.2 | 0.1 | 0.0 | 9.3 | 0.6 | 6.5 | 21.4 | 60.6 | 0.1 | 0.7 | 0.0 | 32719 |
| 1994 | 0.4 | 0.4 | 0.1 | 0.0 | 12.6 | 1.1 | 6.8 | 18.5 | 59.1 | 0.2 | 0.7 | 0.0 | 32406 |
| 1995 | 0.2 | 0.1 | 0.4 | 0.0 | 9.6 | 0.9 | 9.9 | 17.7 | 60.9 | 0.3 | 0.0 | 0.0 | 26918 |
| 1996 | 0.0 | 0.0 | 0.1 | 0.0 | 9.2 | 1.2 | 6.8 | 23.7 | 58.6 | 0.2 | 0.0 | 0.0 | 19267 |
| 1997 | 0.0 | 0.1 | 0.1 | 0.0 | 8.9 | 2.5 | 10.7 | 17.8 | 58.9 | 0.4 | 0.4 | 0.0 | 21721 |
| 1998 | 0.1 | 0.4 | 0.1 | 0.0 | 8.1 | 2.8 | 13.8 | 16.5 | 57.6 | 0.3 | 0.4 | 0.0 | 25995 |
| 1999 | 0.0 | 0.1 | 0.1 | 0.0 | 5.7 | 1.2 | 12.6 | 18.5 | 60.0 | 0.2 | 1.6 | 0.0 | 32439 |
| 2000 | 0.1 | 0.1 | 0.2 | 0.0 | 3.7 | 0.3 | 15.0 | 17.5 | 62.3 | 0.1 | 0.7 | 0.0 | 39020 |
| 2001 | 0.1 | 0.1 | 0.1 | 0.0 | 2.8 | 0.3 | 20.2 | 16.5 | 58.8 | 0.2 | 0.8 | 0.1 | 51786 |
| 2002 | 0.1 | 0.2 | 0.1 | 0.0 | 1.6 | 0.1 | 26.5 | 10.5 | 60.8 | 0.1 | 0.0 | 0.0 | 53546 |
| 2003 | 0.0 | 0.0 | 1.9 | 0.0 | 0.9 | 0.4 | 17.4 | 14.7 | 64.7 | 0.1 | 0.0 | 0.0 | 46555 |
| 2004 | 0.1 | 0.2 | 3.7 | 0.0 | 1.9 | 0.4 | 15.1 | 14.4 | 63.8 | 0.2 | 0.0 | 0.0 | 44605 |
| 2005 | 0.2 | 0.1 | 4.4 | 0.0 | 2.4 | 0.2 | 12.7 | 20.6 | 59.2 | 0.2 | 0.0 | 0.0 | 66394 |
| 2006 | 0.2 | 0.4 | 0.3 | 0.0 | 3.9 | 0.1 | 19.8 | 20.6 | 54.1 | 0.6 | 0.0 | 0.0 | 65394 |
| 2007 | 0.2 | 0.2 | 0.2 | 0.0 | 2.0 | 0.1 | 30.4 | 16.0 | 50.6 | 0.3 | 0.0 | 0.0 | 41341 |
| 2008 | 0.2 | 0.3 | 1.5 | 0.0 | 3.2 | 0.2 | 20.4 | 16.0 | 57.7 | 0.5 | 0.0 | 0.0 | 27475 |
| 2009 | 0.4 | 0.2 | 3.3 | 0.0 | 4.3 | 0.1 | 9.6 | 15.1 | 66.8 | 0.2 | 0.0 | 0.0 | 47122 |
| 2010 | 0.1 | 0.1 | 1.2 | 0.0 | 3.9 | 2.4 | 8.3 | 15.1 | 68.3 | 0.6 | 0.0 | 0.0 | 38293 |
| 2011 | 0.1 | 0.1 | 0.5 | 0.0 | 3.6 | 1.3 | 2.6 | 14.1 | 77.1 | 0.5 | 0.0 | 0.0 | 26854 |
| 2012 | 0.2 | 0.1 | 1.9 | 0.0 | 2.4 | 0.1 | 2.2 | 18.6 | 73.5 | 1.0 | 0.0 | 0.0 | 31633 |
| 2013 | 0.1 | 0.3 | 1.0 | 0.0 | 3.2 | 0.2 | 0.6 | 24.9 | 69.0 | 0.5 | 0.0 | 0.1 | 22339 |
| 2014 | 0.2 | 0.3 | 0.5 | 0.0 | 1.9 | 0.2 | 0.2 | 15.6 | 80.7 | 0.3 | 0.0 | 0.1 | 20793 |
| Avg. | 0.3 | 0.2 | 0.8 | 0.1 | 5.3 | 1.6 | 14.3 | 19.4 | 57.4 | 0.3 | 0.4 | 0.0 | 39121 |

Table 6.2.2.1. Faroe saithe (Division Vb). Catch number at age by fleet categories in 2014 (calculated from gutted weights).

| Age | Jiggers | Single <br> trawlers $>1000 \mathrm{HP}$ | Pair trawlers $<1000$ HP | Pair trawlers $>1000 \mathrm{HP}$ | Others | Total Division Vb |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0 | 0 | 2 | 6 | 0 | 8 |
| 3 | 9 | 8 | 135 | 625 | 10 | 788 |
| 4 | 37 | 55 | 334 | 1624 | 30 | 2081 |
| 5 | 55 | 90 | 404 | 2226 | 40 | 2815 |
| 6 | 26 | 41 | 218 | 1200 | 19 | 1505 |
| 7 | 13 | 20 | 120 | 613 | 10 | 775 |
| 8 | 12 | 23 | 40 | 216 | 4 | 295 |
| 9 | 1 | 1 | 13 | 72 | 2 | 89 |
| 10 | 0 | 0 | 11 | 70 | 2 | 82 |
| 11 | 1 | 2 | 8 | 50 | 1 | 63 |
| 12 | 0 | 0 | 8 | 39 | 1 | 49 |
| 13 | 0 | 0 | 1 | 13 | 0 | 14 |
| 14 | 0 | 0 | 0 | 1 | 0 | 1 |
| 15 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total No. | 155 | 241 | 1294 | 6755 | 119 | 8564 |
| Catch t. | 411 | 654 | 3251 | 16774 | 304 | 21394 |

Table 6.2.2.2. Faroe saithe (Division Vb). Catch number at age (thousands) from the commercial fleet (1961-2014)

| CN | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1961 | 183 | 379 | 483 | 403 | 216 | 129 | 116 | 82 | 45 | 27 | 6 | 49 |
| 1962 | 562 | 542 | 617 | 495 | 286 | 131 | 129 | 113 | 71 | 29 | 13 | 63 |
| 1963 | 614 | 340 | 340 | 415 | 406 | 202 | 174 | 158 | 94 | 169 | 61 | 44 |
| 1964 | 684 | 1908 | 1506 | 617 | 572 | 424 | 179 | 150 | 100 | 83 | 47 | 44 |
| 1965 | 996 | 850 | 1708 | 965 | 510 | 407 | 306 | 201 | 156 | 120 | 89 | 76 |
| 1966 | 488 | 1540 | 1201 | 1686 | 806 | 377 | 294 | 205 | 156 | 94 | 52 | 79 |
| 1967 | 595 | 796 | 1364 | 792 | 1192 | 473 | 217 | 190 | 97 | 75 | 38 | 27 |
| 1968 | 614 | 1689 | 1116 | 1095 | 548 | 655 | 254 | 128 | 89 | 59 | 40 | 88 |
| 1969 | 1191 | 2086 | 2294 | 1414 | 1118 | 589 | 580 | 239 | 115 | 100 | 36 | 54 |
| 1970 | 1445 | 6577 | 1558 | 1478 | 899 | 730 | 316 | 241 | 86 | 48 | 46 | 38 |
| 1971 | 2857 | 3316 | 5585 | 1005 | 828 | 469 | 326 | 164 | 100 | 54 | 13 | 33 |
| 1972 | 2714 | 1774 | 2588 | 2742 | 1529 | 1305 | 1017 | 743 | 330 | 133 | 28 | 49 |
| 1973 | 2515 | 6253 | 7075 | 3478 | 1634 | 693 | 550 | 403 | 215 | 103 | 25 | 58 |
| 1974 | 3504 | 4126 | 4011 | 2784 | 1401 | 640 | 368 | 340 | 197 | 124 | 45 | 96 |
| 1975 | 2062 | 3361 | 3801 | 1939 | 1045 | 714 | 302 | 192 | 193 | 126 | 64 | 108 |
| 1976 | 3178 | 3217 | 1720 | 1250 | 877 | 641 | 468 | 223 | 141 | 96 | 60 | 131 |
| 1977 | 1609 | 2937 | 2034 | 1288 | 767 | 708 | 498 | 338 | 272 | 129 | 80 | 121 |
| 1978 | 611 | 1743 | 1736 | 548 | 373 | 479 | 466 | 473 | 407 | 211 | 146 | 178 |
| 1979 | 287 | 933 | 1341 | 1033 | 584 | 414 | 247 | 473 | 368 | 206 | 136 | 349 |
| 1980 | 996 | 877 | 720 | 673 | 726 | 284 | 212 | 171 | 196 | 156 | 261 | 369 |
| 1981 | 411 | 1804 | 769 | 932 | 908 | 734 | 343 | 192 | 92 | 128 | 176 | 717 |
| 1982 | 387 | 4076 | 994 | 1114 | 380 | 417 | 296 | 105 | 88 | 56 | 49 | 797 |
| 1983 | 2483 | 1103 | 5052 | 1343 | 575 | 339 | 273 | 98 | 98 | 99 | 25 | 416 |
| 1984 | 368 | 11067 | 2359 | 4093 | 875 | 273 | 161 | 52 | 65 | 59 | 18 | 176 |
| 1985 | 1224 | 3990 | 5583 | 1182 | 1898 | 273 | 103 | 38 | 26 | 72 | 41 | 162 |
| 1986 | 1167 | 1997 | 4473 | 3730 | 953 | 1077 | 245 | 104 | 67 | 33 | 56 | 69 |
| 1987 | 1581 | 5793 | 3827 | 2785 | 990 | 532 | 333 | 81 | 43 | 5 | 11 | 81 |
| 1988 | 866 | 2950 | 9555 | 2784 | 1300 | 621 | 363 | 159 | 27 | 43 | 15 | 2 |
| 1989 | 451 | 5981 | 5300 | 7136 | 793 | 546 | 185 | 83 | 55 | 10 | 2 | 27 |
| 1990 | 294 | 3833 | 10120 | 9219 | 5070 | 477 | 123 | 61 | 60 | 18 | 19 | 42 |
| 1991 | 1030 | 5125 | 7452 | 5544 | 3487 | 1630 | 405 | 238 | 128 | 77 | 22 | 19 |
| 1992 | 521 | 4067 | 3667 | 2679 | 1373 | 894 | 613 | 123 | 63 | 37 | 52 | 19 |
| 1993 | 1316 | 2611 | 4689 | 1665 | 858 | 492 | 448 | 245 | 54 | 34 | 10 | 8 |
| 1994 | 690 | 3961 | 2663 | 2368 | 746 | 500 | 307 | 303 | 150 | 28 | 19 | 2 |
| 1995 | 398 | 1019 | 3468 | 1836 | 1177 | 345 | 241 | 192 | 104 | 73 | 25 | 19 |
| 1996 | 297 | 1087 | 1146 | 1449 | 1156 | 521 | 132 | 77 | 64 | 45 | 29 | 8 |
| 1997 | 344 | 832 | 2440 | 1767 | 1335 | 624 | 165 | 71 | 29 | 48 | 29 | 23 |
| 1998 | 163 | 1689 | 1934 | 3475 | 1379 | 683 | 368 | 77 | 32 | 28 | 24 | 21 |
| 1999 | 322 | 655 | 3096 | 2551 | 4113 | 915 | 380 | 147 | 24 | 27 | 5 | 37 |
| 2000 | 811 | 2830 | 1484 | 4369 | 2226 | 2725 | 348 | 186 | 56 | 18 | 2 | 5 |
| 2001 | 1125 | 2452 | 8437 | 2155 | 3680 | 1539 | 1334 | 293 | 90 | 24 | 19 | 13 |
| 2002 | 302 | 8399 | 5962 | 9786 | 862 | 1280 | 465 | 362 | 33 | 36 | 8 | 1 |


| CN | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2003 | 330 | 2432 | 11152 | 3994 | 4287 | 417 | 419 | 304 | 91 | 40 | 3 | 0 |
| 2004 | 76 | 2011 | 8544 | 8762 | 2125 | 1807 | 265 | 293 | 146 | 100 | 10 | 2 |
| 2005 | 454 | 2948 | 9486 | 16606 | 7099 | 843 | 810 | 32 | 102 | 27 | 3 | 0 |
| 2006 | 1475 | 5045 | 7781 | 7712 | 10296 | 3760 | 640 | 282 | 32 | 12 | 12 | 5 |
| 2007 | 831 | 3320 | 11305 | 6473 | 3781 | 4294 | 1538 | 406 | 81 | 11 | 9 | 3 |
| 2008 | 4784 | 3108 | 3598 | 9370 | 3594 | 2223 | 2048 | 444 | 159 | 12 | 6 | 0 |
| 2009 | 459 | 7412 | 4978 | 1842 | 5167 | 2009 | 1696 | 1069 | 292 | 41 | 3 | 1 |
| 2010 | 2324 | 2916 | 5298 | 1125 | 1009 | 2098 | 1248 | 832 | 376 | 51 | 22 | 0 |
| 2011 | 1897 | 2744 | 1940 | 1804 | 477 | 530 | 704 | 521 | 439 | 138 | 34 | 4 |
| 2012 | 859 | 9833 | 4142 | 1252 | 901 | 304 | 307 | 399 | 229 | 136 | 91 | 21 |
| 2013 | 721 | 5172 | 4219 | 2242 | 511 | 209 | 122 | 96 | 146 | 85 | 39 | 36 |
| 2014 | 878 | 2320 | 3139 | 1679 | 864 | 329 | 99 | 92 | 70 | 55 | 16 | 1 |

Table 6.2.2.3. Faroe saithe (Division Vb). Sampling intensity in 2001-2013.

| Year |  | Jiggers | Single trawlers $>1000$ HP | Pair trawlers <1000 HP | Pair trawlers $>1000$ HP | Others | Total | Amount sampled pr tons landed (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | Lengths | 1788 | 4388 | 5613 | 30341 | 0 | 42130 | 7.7 |
|  | Otoliths | 180 | 450 | 480 | 3237 | 0 | 4347 |  |
|  | Weights | 180 | 420 | 420 | 3177 | 0 | 4197 |  |
| 2002 | Lengths | 1197 | 9235 | 5049 | 30761 | 0 | 46242 | 5.8 |
|  | Otoliths | 120 | 1291 | 422 | 3001 | 0 | 4834 |  |
|  | Weights | 120 | 420 | 240 | 2760 | 0 | 3540 |  |
| 2003 | Lengths | 0 | 4959 | 6393 | 34812 | 1388 | 47552 | 7.0 |
|  | Otoliths | 0 | 719 | 960 | 3719 | 180 | 5578 |  |
|  | Weights | 0 | 420 | 239 | 2999 |  | 3658 |  |
| 2004 | Lengths | 916 | 2665 | 3455 | 35609 | 1781 | 44426 | 5.9 |
|  | Otoliths | 180 | 180 | 240 | 3537 | 240 | 4377 |  |
|  | Weights | 180 | 120 | 120 | 3357 | 1364 | 5141 |  |
| 2005 | Lengths | 1048 | 4266 | 6183 | 32046 | 1564 | 45107 | 3.6 |
|  | Otoliths | 120 | 413 | 690 | 2760 | 240 | 4223 |  |
|  | Weights | 340 | 385 | 791 | 3533 | 1564 | 6613 |  |
| 2006 | Lengths | 1059 | 7979 | 8115 | 23082 | 1139 | 41374 | 3.5 |
|  | Otoliths | 180 | 598 | 1138 | 2096 | 60 | 4072 |  |
|  | Weights | 180 | 60 | 1620 | 5678 | 812 | 8350 |  |
| 2007 | Lengths | 683 | 10525 | 10593 | 18045 | 381 | 40227 | 4.1 |
|  | Otoliths | 120 | 748 | 960 | 1977 | 0 | 3805 |  |
|  | Weights | 120 | 697 | 5603 | 9884 | 120 | 16424 |  |
| $2008$ | Lengths | 0 | 6892 | 3694 | 13995 | 234 | 24815 | 2.5 |
|  | Otoliths | 0 | 690 | 600 | 1500 | 0 | 2790 |  |
|  | Weights | 0 | 0 | 2517 | 12914 | 234 | 15665 |  |
| $2009$ | Lengths | 511 | 5273 | 3695 | 23352 | 0 | 32831 | 4.1 |
|  | Otoliths | 97 | 301 | 599 | 2519 | 0 | 3516 |  |
|  | Weights | 511 | 0 | 3494 | 19060 | 0 | 23065 |  |
| $2010$ | Lengths | 209 | 1442 | 3663 | 25793 | 151 | 31258 | 6.0 |
|  | Otoliths | 5 | 119 | 480 | 2459 | 0 | 3063 |  |
|  | Weights | 5 | 0 | 3060 | 18749 | 151 | 21965 |  |
| 2011 | Lengths | 583 | 18 | 1874 | 19990 | 753 | 23218 | 8.5 |
|  | Otoliths | 60 | 0 | 300 | 2459 | 60 | 2879 |  |
|  | Weights | 583 | 18 | 1458 | 14256 | 753 | 17068 |  |
| 2012 | Lengths | 6 | 0 | 1060 | 24924 | 211 | 26201 | 5.6 |
|  | Otoliths | 6 | 0 | 120 | 2516 | 0 | 2642 |  |
|  | Weights | 6 | 0 | 1060 | 17593 | 211 | 18870 |  |
| 2013 | Lengths | 0 | 0 | 1465 | 18015 | 920 | 20400 | 5.2 |
|  | Otoliths | 0 | 0 | 360 | 1979 | 120 | 2459 |  |
|  | Weights | 0 | 0 | 1465 | 13544 | 1325 | 16334 |  |
| 2014 | Lengths | 0 | 201 | 0 | 22131 | 920 | 23252 | 8.9 |
|  | Otoliths | 0 | 0 | 0 | 2542 | 120 | 2662 |  |
|  | Weights | 0 | 0 | 0 | 15448 | 920 | 16368 |  |

Table 6.2.3.1. Faroe saithe (Division Vb ). Catch weights at age (kg)(equal to stock-weights) from the commercial fleet (1961-2014). The value for 2015 is used for short-term projections.

| CW | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | $14+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1961 | 1.43 | 2.302 | 3.348 | 4.287 | 5.128 | 6.155 | 7.06 | 7.265 | 7.497 | 8.198 | 9.154 | 9.992 |
| 1962 | 1.273 | 2.045 | 3.293 | 4.191 | 5.146 | 5.655 | 6.469 | 6.706 | 7.15 | 7.903 | 8.449 | 9.658 |
| 1963 | 1.28 | 2.197 | 3.212 | 4.568 | 5.056 | 5.932 | 6.259 | 8 | 7.265 | 8.551 | 9.02 | 9.818 |
| 1964 | 1.175 | 2.055 | 3.266 | 4.255 | 5.038 | 5.694 | 6.662 | 6.837 | 7.686 | 8.348 | 8.123 | 9.423 |
| 1965 | 1.181 | 2.125 | 2.941 | 4.096 | 4.878 | 5.932 | 6.321 | 7.288 | 8.074 | 7.878 | 9.479 | 9.849 |
| 1966 | 1.361 | 2.026 | 3.055 | 3.658 | 4.585 | 5.52 | 6.837 | 7.265 | 7.662 | 8.123 | 10.21 | 9.883 |
| 1967 | 1.273 | 1.78 | 2.534 | 3.572 | 4.368 | 5.313 | 5.812 | 6.554 | 7.806 | 7.591 | 8.551 | 9.135 |
| 1968 | 1.302 | 1.737 | 2.036 | 3.12 | 4.049 | 5.183 | 6.238 | 7.52 | 8.049 | 8.654 | 8.298 | 9.748 |
| 1969 | 1.188 | 1.667 | 2.302 | 2.853 | 3.673 | 5.002 | 5.714 | 6.405 | 6.554 | 7.591 | 7.951 | 9.096 |
| 1970 | 1.244 | 1.445 | 2.249 | 2.853 | 3.515 | 4.418 | 5.444 | 5.733 | 6.662 | 7.31 | 9.047 | 9.634 |
| 1971 | 1.101 | 1.316 | 1.818 | 2.978 | 3.702 | 4.271 | 5.388 | 5.972 | 6.49 | 7.173 | 7.38 | 9.612 |
| 1972 | 1.043 | 1.485 | 2.055 | 2.829 | 3.791 | 4.175 | 4.808 | 5.294 | 6.948 | 6.727 | 7.591 | 9.609 |
| 1973 | 1.306 | 1.754 | 1.899 | 2.7 | 4.426 | 5.264 | 6.156 | 6.334 | 8.076 | 8.777 | 9.782 | 11.115 |
| 1974 | 1.615 | 1.723 | 2.493 | 2.824 | 3.524 | 5.197 | 6.279 | 6.454 | 7.07 | 7.773 | 8.763 | 10.83 |
| 1975 | 1.293 | 1.924 | 2.623 | 3.621 | 4.128 | 4.754 | 5.952 | 7.073 | 8.352 | 9.032 | 9.984 | 11.082 |
| 1976 | 1.162 | 1.79 | 3.074 | 3.291 | 4.579 | 4.648 | 5.116 | 6.314 | 7.069 | 7.069 | 7.808 | 9.714 |
| 1977 | 1.223 | 1.641 | 2.66 | 3.79 | 4.239 | 5.597 | 5.35 | 5.912 | 6.837 | 6.727 | 6.948 | 9.258 |
| 1978 | 1.493 | 2.324 | 3.068 | 3.746 | 4.913 | 4.368 | 5.276 | 5.832 | 6.053 | 6.706 | 7.686 | 8.516 |
| 1979 | 1.22 | 1.88 | 2.62 | 3.4 | 4.18 | 4.95 | 5.69 | 6.38 | 7.02 | 7.26 | 8.15 | 9.618 |
| 1980 | 1.23 | 2.12 | 3.32 | 4.28 | 5.16 | 6.42 | 6.87 | 7.09 | 7.93 | 8.07 | 8.59 | 10.142 |
| 1981 | 1.31 | 2.13 | 3 | 3.81 | 4.75 | 5.25 | 5.95 | 6.43 | 7 | 7.47 | 8.14 | 9.43 |
| 1982 | 1.337 | 1.851 | 2.951 | 3.577 | 4.927 | 6.243 | 7.232 | 7.239 | 8.346 | 8.345 | 8.956 | 10.227 |
| 1983 | 1.208 | 2.029 | 2.965 | 4.143 | 4.724 | 5.901 | 6.811 | 7.051 | 7.248 | 8.292 | 9.478 | 10.509 |
| 1984 | 1.431 | 1.953 | 2.47 | 3.85 | 5.177 | 6.347 | 7.825 | 6.746 | 8.636 | 8.467 | 8.556 | 10.802 |
| 1985 | 1.401 | 2.032 | 2.965 | 3.596 | 5.336 | 7.202 | 6.966 | 9.862 | 10.67 | 10.46 | 10.202 | 13.055 |
| 1986 | 1.718 | 1.986 | 2.618 | 3.277 | 4.186 | 5.589 | 6.05 | 6.15 | 9.536 | 9.823 | 7.303 | 12.773 |
| 1987 | 1.609 | 1.835 | 2.395 | 3.182 | 4.067 | 5.149 | 5.501 | 6.626 | 6.343 | 10.245 | 8.491 | 10.482 |
| 1988 | 1.5 | 1.975 | 1.978 | 2.937 | 3.798 | 4.419 | 5.115 | 6.712 | 9.04 | 9.364 | 9.142 | 10.216 |
| 1989 | 1.309 | 1.735 | 1.907 | 2.373 | 3.81 | 4.667 | 5.509 | 5.972 | 6.939 | 8.543 | 9.514 | 10.484 |
| 1990 | 1.223 | 1.633 | 1.83 | 2.052 | 2.866 | 4.474 | 5.424 | 6.469 | 6.343 | 8.418 | 7.383 | 8.64 |
| 1991 | 1.24 | 1.568 | 1.864 | 2.211 | 2.648 | 3.38 | 4.816 | 5.516 | 6.407 | 7.395 | 8.079 | 8.674 |
| 1992 | 1.264 | 1.602 | 2.069 | 2.554 | 3.057 | 4.078 | 5.012 | 6.768 | 7.754 | 8.303 | 7.786 | 9.301 |
| 1993 | 1.408 | 1.86 | 2.323 | 3.131 | 3.73 | 4.394 | 5.209 | 6.54 | 8.403 | 7.275 | 9.414 | 9.64 |
| 1994 | 1.503 | 1.951 | 2.267 | 2.936 | 4.214 | 4.971 | 5.657 | 5.95 | 6.891 | 8.752 | 9.752 | 7.989 |
| 1995 | 1.456 | 2.177 | 2.42 | 2.895 | 3.651 | 5.064 | 5.44 | 6.167 | 7.08 | 7.736 | 7.295 | 7.104 |
| 1996 | 1.432 | 1.875 | 2.496 | 3.229 | 3.744 | 4.964 | 6.375 | 6.745 | 7.466 | 7.284 | 8.47 | 10.125 |
| 1997 | 1.476 | 1.783 | 2.032 | 2.778 | 3.598 | 4.766 | 5.982 | 7.658 | 7.882 | 8.539 | 9.488 | 10.413 |
| 1998 | 1.388 | 1.711 | 1.954 | 2.405 | 3.3 | 4.22 | 4.999 | 6.391 | 6.665 | 8.214 | 8.485 | 8.845 |
| 1999 | 1.374 | 1.712 | 1.905 | 2.396 | 2.845 | 4.124 | 5.256 | 5.526 | 6.956 | 8.03 | 8.349 | 8.907 |
| 2000 | 1.477 | 1.606 | 2.077 | 2.36 | 2.977 | 3.48 | 4.851 | 5.268 | 6.523 | 4.727 | 8.807 | 8.972 |
| 2001 | 1.33 | 1.59 | 1.785 | 2.586 | 3.059 | 3.871 | 4.374 | 5.565 | 6.703 | 5.776 | 7.745 | 7.773 |
| 2002 | 1.142 | 1.46 | 1.652 | 1.969 | 3.13 | 3.589 | 4.513 | 5.138 | 6.422 | 8.026 | 4.759 | 11.357 |
| 2003 | 1.123 | 1.304 | 1.614 | 1.977 | 2.532 | 3.97 | 4.834 | 5.499 | 6.099 | 6.987 | 5.961 | 10 |
| 2004 | 1.143 | 1.333 | 1.45 | 1.789 | 2.56 | 3.159 | 4.154 | 5.167 | 6.015 | 6.186 | 7.056 | 9.391 |
| 2005 | 1.148 | 1.325 | 1.516 | 1.672 | 2.087 | 2.975 | 3.79 | 6.087 | 6.134 | 6.651 | 7.424 | 10 |
| 2006 | 1.126 | 1.218 | 1.462 | 1.79 | 2.035 | 2.436 | 3.861 | 4.222 | 5.149 | 6.437 | 6.905 | 5.365 |
| 2007 | 1.058 | 1.391 | 1.413 | 1.824 | 2.361 | 2.682 | 3.278 | 4.104 | 4.998 | 6.331 | 7.844 | 7.971 |
| 2008 | 1.146 | 1.312 | 1.672 | 1.816 | 2.395 | 2.902 | 3.1 | 3.728 | 4.769 | 6.072 | 6.451 | 10 |
| 2009 | 0.938 | 1.485 | 1.893 | 2.411 | 2.601 | 3.147 | 3.634 | 4.024 | 5.014 | 5.828 | 6.308 | 9.011 |


| CW | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2010 | 1.429 | 1.706 | 2.166 | 2.551 | 3.172 | 3.411 | 3.972 | 4.352 | 5.083 | 4.941 | 5.305 | 10 |
| 2011 | 1.111 | 1.693 | 2.253 | 2.918 | 3.609 | 4.204 | 4.531 | 5.087 | 5.416 | 6.087 | 6.763 | 7.916 |
| 2012 | 1.029 | 1.334 | 1.626 | 2.709 | 3.785 | 4.448 | 4.799 | 5.207 | 5.562 | 6.018 | 7.143 | 6.247 |
| 2013 | 1.208 | 1.466 | 1.778 | 2.069 | 3.553 | 4.292 | 5.191 | 5.742 | 5.919 | 6.417 | 7.941 | 7.138 |
| 2014 | 1.369 | 1.724 | 2.163 | 2.868 | 3.325 | 5.903 | 5.899 | 6.877 | 6.784 | 7.467 | 7.121 | 11.31 |
| 2015 | 1.299 | 1.528 | 1.850 | 2.239 | 2.602 | 4.451 | 5.296 | 5.942 | 6.088 | 6.634 | 7.402 | 8.232 |

Table 6.2.4.1. Faroe saithe (Division Vb). Proportion mature at age (1982-2014). Maturities-at-age from 1961 to 1981 are fixed and equal to those in 1982. The value for 2015 is used for short-term prognosis.

| Mat | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 0.03 | 0.22 | 0.52 | 0.79 | 0.92 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1983 | 0.03 | 0.27 | 0.61 | 0.91 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1984 | 0.04 | 0.28 | 0.60 | 0.88 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1985 | 0.05 | 0.29 | 0.59 | 0.85 | 0.97 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1986 | 0.05 | 0.28 | 0.57 | 0.82 | 0.94 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1987 | 0.05 | 0.27 | 0.55 | 0.79 | 0.92 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1988 | 0.05 | 0.26 | 0.53 | 0.77 | 0.90 | 0.96 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1989 | 0.04 | 0.23 | 0.51 | 0.76 | 0.89 | 0.96 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1990 | 0.03 | 0.19 | 0.49 | 0.75 | 0.89 | 0.96 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1991 | 0.03 | 0.17 | 0.48 | 0.75 | 0.88 | 0.96 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1992 | 0.02 | 0.17 | 0.48 | 0.75 | 0.89 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1993 | 0.02 | 0.17 | 0.49 | 0.77 | 0.91 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1994 | 0.01 | 0.17 | 0.49 | 0.78 | 0.93 | 1.00 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1995 | 0.01 | 0.17 | 0.49 | 0.78 | 0.93 | 1.00 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1996 | 0.01 | 0.17 | 0.47 | 0.75 | 0.90 | 1.00 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1997 | 0.01 | 0.16 | 0.44 | 0.70 | 0.87 | 0.98 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1998 | 0.02 | 0.16 | 0.41 | 0.64 | 0.83 | 0.96 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1999 | 0.02 | 0.16 | 0.38 | 0.60 | 0.79 | 0.94 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2000 | 0.02 | 0.16 | 0.37 | 0.58 | 0.77 | 0.92 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2001 | 0.01 | 0.17 | 0.37 | 0.56 | 0.75 | 0.91 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2002 | 0.01 | 0.17 | 0.37 | 0.56 | 0.74 | 0.89 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2003 | 0.01 | 0.18 | 0.37 | 0.56 | 0.74 | 0.88 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2004 | 0.01 | 0.18 | 0.38 | 0.57 | 0.74 | 0.88 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2005 | 0.00 | 0.18 | 0.39 | 0.59 | 0.76 | 0.89 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2006 | 0.00 | 0.18 | 0.40 | 0.62 | 0.78 | 0.90 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2007 | 0.00 | 0.19 | 0.42 | 0.64 | 0.80 | 0.91 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2008 | 0.01 | 0.20 | 0.43 | 0.66 | 0.82 | 0.92 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2009 | 0.01 | 0.21 | 0.45 | 0.68 | 0.84 | 0.94 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2010 | 0.02 | 0.23 | 0.47 | 0.71 | 0.87 | 0.95 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2011 | 0.03 | 0.24 | 0.49 | 0.72 | 0.88 | 0.96 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2012 | 0.03 | 0.25 | 0.50 | 0.73 | 0.89 | 0.97 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2013 | 0.04 | 0.25 | 0.50 | 0.74 | 0.90 | 0.97 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2014 | 0.04 | 0.26 | 0.51 | 0.74 | 0.90 | 0.98 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2015 | 0.04 | 0.26 | 0.51 | 0.74 | 0.90 | 0.98 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |

Table 6.3.1. Faroe saithe (Division Vb ). Effort (hours) and catch in number at age for the commercial pair trawlers (1995-2013)

| year | effort | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1995 | 11016 | 47 | 180 | 577 | 236 | 146 | 49 | 24 | 19 | 14 |
| 1996 | 48205 | 310 | 958 | 821 | 1119 | 503 | 282 | 133 | 127 | 70 |
| 1997 | 34828 | 199 | 533 | 1488 | 1013 | 768 | 333 | 73 | 33 | 10 |
| 1998 | 34422 | 107 | 656 | 1148 | 1486 | 730 | 325 | 170 | 40 | 13 |
| 1999 | 43528 | 174 | 487 | 1554 | 2016 | 2024 | 817 | 190 | 83 | 12 |
| 2000 | 44280 | 434 | 1566 | 913 | 2700 | 1333 | 1604 | 192 | 106 | 31 |
| 2001 | 41860 | 611 | 1438 | 4946 | 1165 | 1855 | 748 | 618 | 127 | 29 |
| 2002 | 41914 | 133 | 3976 | 3964 | 6888 | 520 | 682 | 246 | 177 | 25 |
| 2003 | 38489 | 141 | 1494 | 6560 | 2373 | 2263 | 197 | 212 | 124 | 35 |
| 2004 | 35525 | 43 | 1200 | 5089 | 5116 | 1035 | 762 | 113 | 116 | 53 |
| 2005 | 32860 | 188 | 1189 | 4039 | 7266 | 3130 | 320 | 291 | 7 | 43 |
| 2006 | 25334 | 140 | 1176 | 2410 | 2584 | 3700 | 1376 | 268 | 85 | 14 |
| 2007 | 25218 | 204 | 879 | 2913 | 1815 | 1034 | 1215 | 435 | 110 | 19 |
| 2008 | 25259 | 796 | 762 | 947 | 2641 | 1063 | 726 | 611 | 156 | 51 |
| 2009 | 68408 | 154 | 4082 | 3377 | 1283 | 3612 | 1402 | 1153 | 751 | 195 |
| 2010 | 61563 | 459 | 2019 | 3586 | 737 | 657 | 1325 | 814 | 518 | 245 |
| 2011 | 64272 | 397 | 1936 | 1367 | 1257 | 323 | 356 | 488 | 366 | 310 |
| 2012 | 57749 | 366 | 5652 | 2332 | 756 | 554 | 187 | 189 | 252 | 143 |
| 2013 | 43325 | 424 | 3047 | 2462 | 1295 | 293 | 122 | 71 | 56 | 83 |
| 2014 | 48205 | 625 | 1624 | 2226 | 1200 | 613 | 216 | 72 | 70 | 50 |

Table 6.3.2. Faroe saithe (Division Vb). Diagnostics from XSA with commercial pair trawler tuning series (spaly)

## FLR XSA Diagnostics 2015-04-15 15:45:12

CPUE data from indices
Catch data for 54 years 1961 to 2014. Ages 3 to 14.
fleet first age last age first year last year alpha beta
1 PairTrawlers_GLM_SD 31119952014 <NA> <NA>
Time series weights :
Tapered time weighting not applied

Catchability analysis :
Catchability independent of size for all ages
Catchability independent of age for ages > 8

Terminal population estimation :
Survivor estimates shrunk towards the mean F
of the final 5 years or the 3 oldest ages.
S.E. of the mean to which the estimates are shrunk $=2$

Minimum standard error for population
estimates derived from each fleet $=0.3$
prior weighting not applied

| Regression weights year |
| :---: |
| age 2005200620072008200920102011201220132014 |
| $\begin{array}{llllllllll}\text { all } & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1\end{array}$ |
| Fishing mortalities year |
| age 2005200620072008200920102011201220132014 |
| 30.0070 .0760 .0500 .1840 .0380 .1120 .0620 .0280 .0220 .016 |
| 40.0770 .1030 .2450 .2660 .4810 .3540 .1870 .5220 .2340 .093 |
| 50.2950 .2970 .3530 .4590 .9070 .7740 .4230 .4760 .4450 .217 |
| 60.4770 .4170 .4330 .5600 .4530 .5230 .6650 .5350 .5170 .319 |
| 70.5850 .6210 .3710 .4590 .7030 .4840 .4400 .8580 .4350 .383 |
| 80.3520 .7220 .5770 .3890 .5070 .7050 .5100 .5630 .4860 .560 |
| 90.8890 .4960 .7530 .6060 .5870 .6940 .5440 .6350 .4630 .450 |
| 100.2960 .9390 .6890 .5040 .7570 .6500 .7150 .6950 .4140 .783 |
| 110.6210 .5450 .7900 .6430 .7470 .6670 .8920 .8230 .5950 .610 |
| 120.3480 .1320 .3620 .2460 .3350 .2700 .5530 .7870 .8650 .469 |
| 130.5410 .2570 .1380 .3440 .0890 .3020 .2910 .9050 .5440 .380 |
| 40.5410 .2570 .1380 .3440 .0890 .3020 .2910 .9050 .5440 .380 |

```
XSA population number (Thousand)
    age
year 
20056998444103410054840517716 31401520 138244101 8 0
20062222256888 334412498824605 80811808512841075924
2007188801686042011203391348110828 3214901164407726
200831507 1470610799241661079576164980124037061 23 0
200913724214679228558611307558642242224613159 3913
2010242911082110869 30512907 45822756 1924 854 238 93 0
201134720 17785622141051480146718531127 822 35914917
201234440 2671012078 3338 1729 780 721 880 451 27616938
2013359512742012971 6141 1600 600 364 313 36016210394
2014616192878217770 68022999848 302 18716916256 3
Estimated population abundance at 1st Jan 2015
    age
year 
2015 049655214651170840501674396158 70 75 83 31
Fleet: PairTrawlers_GLM_SD
Log catchability residuals.
year
age 1995 1996 19971998199920002001200220032004200520062007200820092010201120122013 2014
\(3-0.3890 .4920 .0540 .413-0.868 \quad 0.5330 .027-1.689-1.044-1.979-0.6830 .4630 .9941 .9050 .0280 .6900 .1210 .1380 .527\) 0.267
\(4-0.033-0.733-0.518-0.609-0.172-0.560-0.059 \quad 0.066-1.075-0.706-0.451-0.4440 .5520 .5540 .9550 .9850 .3261 .2500 .763\) -0.088
\(50.449-0.659-0.672-0.422-0.637-0.190 \quad 0.041 \quad 0.400 \quad 0.078-0.461-0.032-0.083-0.0920 .1890 .8130 .7590 .1570 .1590 .415\) \(-0.211\)
\(6-0.194-0.178-0.074-0.663-0.047 \quad 0.014 \quad 0.3490 .6540 .2100 .056\) 0.097-0.042-0.178 \(0.079-0.221-0.0340 .221-0.030 \quad 0.178\) -0.196
\(7 \quad 0.152-0.407 \quad 0.2250 .058-0.171-0.0340 .328 \quad 0.2100 .367-0.006 \quad 0.1730 .287-0.492-0.2040 .083-0.253-0.3500 .320-0.134\) -0.154
\(80.0960 .1500 .113-0.0130 .5670 .2720 .1190 .1490 .0050 .171-0.580 \quad 0.356-0.119-0.367-0.343-0.010-0.313-0.194-0.106\) 0.046
\(9-0.0390 .3950 .0030 .260-0.020-0.118 \quad 0.411-0.185-0.151 \quad 0.4980 .2820 .1190 .144-0.019-0.2240 .007-0.216-0.074-0.156\) -0.070
10-0.362 1.060 \(0.0750 .1930 .2100 .2470 .5290 .295-0.020 \quad 0.115-1.3050 .4210 .014-0.038 \quad 0.062-0.1040 .068 \quad 0.040-0.265\) 0.525
\(11-0.0580 .144-0.393-0.060-0.560 \quad 0.074 \quad 0.046-0.034-0.3340 .1520 .088 \quad 0.2550 .0040 .113-0.002-0.0340 .2910 .1950 .068\) 0.215
```

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Mean_Logq -15.5347-13.4272 -12.4482-12.0741-11.9394-11.8402-11.8402 -11.8402 -11.8402 $\begin{array}{llllllllll}\text { S.E_Logq } & 0.4575 & 0.4575 & 0.4575 & 0.4575 & 0.4575 & 0.4575 & 0.4575 & 0.4575 & 0.4575\end{array}$

Terminal year survivor and F summaries:

Age 3 Year class =2011

```
source
        scaledWts survivors yrcls
PairTrawlers_GLM_SD 0.82 64821 2011
fshk 0.18 14726 2011
    Age 4 Year class=2010
source
    scaledWts survivors yrcls
PairTrawlers_GLM_SD 0.888 19660 2010
fshk 0.112 4881 2010
    Age 5 Year class =2009
source
        scaledWts survivors yrcls
PairTrawlers_GLM_SD 0.941 9482 2009
fshk 0.059 3377 2009
    Age 6 Year class =2008
source
        scaledWts survivors yrcls
PairTrawlers_GLM_SD 0.97 3329 2008
fshk 0.03 2106 2008
    Age 7 Year class =2007
source
        scaledWts survivors yrcls
PairTrawlers_GLM_SD 0.968 1435 2007
fshk 0.032 974 2007
```

    Age 8 Year class \(=2006\)
    source
scaledWts survivors yrcls
PairTrawlers_GLM_SD 0.962 4152006
fshk $\quad 0.038 \quad 3982006$
Age 9 Year class $=2005$
source
scaledWts survivors yrcls
PairTrawlers_GLM_SD 0.966 1472005
fshk $\quad 0.034 \quad 1112005$

Age 10 Year class =2004
source
scaledWts survivors yrcls
PairTrawlers_GLM_SD 0.891 1182004
fshk $\quad 0.109 \quad 912004$

Age 11 Year class $=2003$
source
scaledWts survivors yrcls
PairTrawlers_GLM_SD 0.96 932003
fshk $0.04 \quad 562003$

Age 12 Year class $=2002$
source
scaledWts survivors yrcls
fshk $1 \quad 652002$

Age 13 Year class =2001
source
scaledWts survivors yrcls
fshk $1 \quad 172001$

Table 6.5.1. Faroe saithe (Division Vb). Fishing mortality at age (1961-2013). The value for 2015 is used for short-term prognosis.

| F | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | $14+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1961 | 0.026 | 0.058 | 0.109 | 0.143 | 0.12 | 0.1 | 0.11 | 0.106 | 0.112 | 0.181 | 0.134 | 0.134 |
| 1962 | 0.052 | 0.101 | 0.127 | 0.156 | 0.143 | 0.099 | 0.138 | 0.149 | 0.125 | 0.098 | 0.124 | 0.124 |
| 1963 | 0.035 | 0.04 | 0.085 | 0.118 | 0.185 | 0.142 | 0.185 | 0.25 | 0.178 | 0.491 | 0.308 | 0.308 |
| 1964 | 0.052 | 0.144 | 0.251 | 0.218 | 0.236 | 0.301 | 0.18 | 0.241 | 0.248 | 0.235 | 0.243 | 0.243 |
| 1965 | 0.05 | 0.085 | 0.186 | 0.253 | 0.283 | 0.263 | 0.37 | 0.316 | 0.424 | 0.532 | 0.427 | 0.427 |
| 1966 | 0.026 | 0.103 | 0.167 | 0.283 | 0.348 | 0.35 | 0.308 | 0.456 | 0.433 | 0.493 | 0.464 | 0.464 |
| 1967 | 0.027 | 0.053 | 0.125 | 0.158 | 0.332 | 0.354 | 0.349 | 0.335 | 0.407 | 0.384 | 0.378 | 0.378 |
| 1968 | 0.03 | 0.099 | 0.098 | 0.14 | 0.156 | 0.307 | 0.326 | 0.358 | 0.258 | 0.467 | 0.363 | 0.363 |
| 1969 | 0.034 | 0.136 | 0.189 | 0.175 | 0.207 | 0.25 | 0.493 | 0.586 | 0.639 | 0.518 | 0.586 | 0.586 |
| 1970 | 0.044 | 0.262 | 0.142 | 0.179 | 0.16 | 0.202 | 0.206 | 0.39 | 0.431 | 0.609 | 0.48 | 0.48 |
| 1971 | 0.086 | 0.135 | 0.373 | 0.128 | 0.144 | 0.117 | 0.13 | 0.157 | 0.277 | 0.534 | 0.325 | 0.325 |
| 1972 | 0.094 | 0.07 | 0.148 | 0.316 | 0.293 | 0.354 | 0.4 | 0.49 | 0.541 | 0.73 | 0.592 | 0.592 |
| 1973 | 0.125 | 0.325 | 0.438 | 0.304 | 0.315 | 0.209 | 0.246 | 0.272 | 0.253 | 0.32 | 0.283 | 0.283 |
| 1974 | 0.222 | 0.311 | 0.358 | 0.307 | 0.192 | 0.195 | 0.164 | 0.237 | 0.207 | 0.227 | 0.225 | 0.225 |
| 1975 | 0.141 | 0.345 | 0.528 | 0.293 | 0.18 | 0.141 | 0.132 | 0.12 | 0.205 | 0.198 | 0.175 | 0.175 |
| 1976 | 0.196 | 0.34 | 0.298 | 0.328 | 0.208 | 0.16 | 0.129 | 0.137 | 0.122 | 0.149 | 0.136 | 0.136 |
| 1977 | 0.146 | 0.281 | 0.376 | 0.382 | 0.344 | 0.259 | 0.179 | 0.13 | 0.246 | 0.156 | 0.178 | 0.178 |
| 1978 | 0.085 | 0.233 | 0.267 | 0.163 | 0.18 | 0.375 | 0.272 | 0.259 | 0.228 | 0.307 | 0.266 | 0.266 |
| 1979 | 0.037 | 0.18 | 0.283 | 0.251 | 0.261 | 0.31 | 0.338 | 0.49 | 0.329 | 0.172 | 0.333 | 0.333 |
| 1980 | 0.088 | 0.153 | 0.205 | 0.224 | 0.281 | 0.195 | 0.258 | 0.415 | 0.386 | 0.226 | 0.344 | 0.344 |
| 1981 | 0.014 | 0.227 | 0.194 | 0.447 | 0.533 | 0.512 | 0.383 | 0.394 | 0.412 | 0.471 | 0.429 | 0.429 |
| 1982 | 0.028 | 0.184 | 0.188 | 0.477 | 0.329 | 0.502 | 0.399 | 0.191 | 0.315 | 0.477 | 0.33 | 0.33 |
| 1983 | 0.07 | 0.103 | 0.366 | 0.419 | 0.486 | 0.552 | 0.736 | 0.221 | 0.275 | 0.711 | 0.405 | 0.405 |
| 1984 | 0.016 | 0.498 | 0.332 | 0.575 | 0.535 | 0.451 | 0.558 | 0.292 | 0.224 | 0.265 | 0.262 | 0.262 |
| 1985 | 0.062 | 0.236 | 0.507 | 0.276 | 0.579 | 0.314 | 0.304 | 0.243 | 0.232 | 0.415 | 0.298 | 0.298 |
| 1986 | 0.021 | 0.138 | 0.452 | 0.774 | 0.375 | 0.785 | 0.518 | 0.578 | 0.895 | 0.518 | 0.67 | 0.67 |
| 1987 | 0.037 | 0.138 | 0.423 | 0.57 | 0.476 | 0.372 | 0.598 | 0.32 | 0.503 | 0.141 | 0.323 | 0.323 |
| 1988 | 0.022 | 0.089 | 0.355 | 0.631 | 0.576 | 0.629 | 0.471 | 0.65 | 0.167 | 1.599 | 0.813 | 0.813 |
| 1989 | 0.018 | 0.203 | 0.228 | 0.492 | 0.366 | 0.511 | 0.384 | 0.184 | 0.489 | 0.086 | 0.254 | 0.254 |
| 1990 | 0.016 | 0.203 | 0.627 | 0.784 | 0.801 | 0.392 | 0.203 | 0.209 | 0.196 | 0.29 | 0.233 | 0.233 |
| 1991 | 0.047 | 0.414 | 0.768 | 0.875 | 0.799 | 0.658 | 0.689 | 0.756 | 0.903 | 0.415 | 0.698 | 0.698 |
| 1992 | 0.03 | 0.262 | 0.596 | 0.707 | 0.551 | 0.483 | 0.558 | 0.459 | 0.455 | 0.73 | 0.552 | 0.552 |
| 1993 | 0.063 | 0.205 | 0.547 | 0.601 | 0.514 | 0.388 | 0.478 | 0.454 | 0.374 | 0.477 | 0.438 | 0.438 |
| 1994 | 0.046 | 0.274 | 0.334 | 0.597 | 0.599 | 0.652 | 0.448 | 0.706 | 0.562 | 0.339 | 0.54 | 0.54 |
| 1995 | 0.011 | 0.089 | 0.411 | 0.406 | 0.684 | 0.623 | 0.778 | 0.565 | 0.563 | 0.595 | 0.579 | 0.579 |
| 1996 | 0.014 | 0.039 | 0.137 | 0.3 | 0.486 | 0.757 | 0.517 | 0.616 | 0.37 | 0.51 | 0.502 | 0.502 |
| 1997 | 0.011 | 0.048 | 0.115 | 0.324 | 0.5 | 0.532 | 0.575 | 0.589 | 0.496 | 0.528 | 0.741 | 0.741 |
| 1998 | 0.014 | 0.071 | 0.15 | 0.238 | 0.454 | 0.52 | 0.706 | 0.586 | 0.582 | 1.419 | 0.552 | 0.552 |
| 1999 | 0.006 | 0.073 | 0.181 | 0.302 | 0.492 | 0.627 | 0.623 | 0.694 | 0.361 | 1.681 | 1.151 | 1.151 |
| 2000 | 0.025 | 0.068 | 0.235 | 0.418 | 0.471 | 0.721 | 0.52 | 0.727 | 0.628 | 0.508 | 0.503 | 0.503 |
| 2001 | 0.014 | 0.1 | 0.294 | 0.634 | 0.763 | 0.709 | 1 | 1.206 | 0.998 | 0.612 | 1.927 | 1.927 |
| 2002 | 0.003 | 0.14 | 0.372 | 0.661 | 0.566 | 0.667 | 0.479 | 0.843 | 0.389 | 1.815 | 0.421 | 0.421 |
| 2003 | 0.006 | 0.032 | 0.279 | 0.46 | 0.696 | 0.597 | 0.476 | 0.675 | 0.522 | 1.216 | 0.741 | 0.741 |
| 2004 | 0.002 | 0.043 | 0.148 | 0.37 | 0.478 | 0.73 | 1.006 | 0.736 | 0.834 | 2.518 | 1.289 | 1.289 |
| 2005 | 0.007 | 0.077 | 0.295 | 0.477 | 0.585 | 0.352 | 0.889 | 0.296 | 0.621 | 0.348 | 0.541 | 0.541 |
| 2006 | 0.076 | 0.103 | 0.297 | 0.417 | 0.621 | 0.722 | 0.496 | 0.939 | 0.545 | 0.132 | 0.257 | 0.257 |
| 2007 | 0.05 | 0.245 | 0.353 | 0.433 | 0.371 | 0.577 | 0.753 | 0.689 | 0.79 | 0.362 | 0.138 | 0.138 |
| 2008 | 0.184 | 0.266 | 0.459 | 0.56 | 0.459 | 0.389 | 0.606 | 0.504 | 0.643 | 0.246 | 0.344 | 0.344 |
| 2009 | 0.038 | 0.481 | 0.907 | 0.453 | 0.703 | 0.507 | 0.587 | 0.757 | 0.747 | 0.335 | 0.089 | 0.089 |


| $\mathbf{F}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2010 | 0.112 | 0.354 | 0.774 | 0.523 | 0.484 | 0.705 | 0.694 | 0.65 | 0.667 | 0.27 | 0.302 | 0.302 |
| 2011 | 0.062 | 0.187 | 0.423 | 0.665 | 0.44 | 0.51 | 0.544 | 0.715 | 0.892 | 0.553 | 0.291 | 0.291 |
| 2012 | 0.028 | 0.522 | 0.476 | 0.535 | 0.858 | 0.563 | 0.635 | 0.695 | 0.823 | 0.787 | 0.905 | 0.905 |
| 2013 | 0.022 | 0.234 | 0.445 | 0.517 | 0.435 | 0.486 | 0.463 | 0.414 | 0.595 | 0.865 | 0.544 | 0.544 |
| 2014 | 0.016 | 0.093 | 0.217 | 0.319 | 0.383 | 0.56 | 0.45 | 0.783 | 0.61 | 0.469 | 0.38 | 0.38 |
| 2015 | 0.016 | 0.201 | 0.269 | 0.324 | 0.397 | 0.381 | 0.366 | 0.448 | 0.480 | 1.00 | 1.00 | 1.00 |

Table 6.3.2. Faroe saithe (Division Vb). Stock number at age (start of year) (Thousands)(1961-2013). The value for 2015 is used for short-term prognosis.

| year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1961 | 7827.25 | 7421.86 | 5158.38 | 3351.65 | 2113.91 | 1494.26 | 1232.82 | 904.51 | 468.22 | 179.78 | 53.02 | 431.33 |
| 1962 | 12256.25 | 6242.83 | 5733.57 | 3786.29 | 2379.45 | 1535.28 | 1106.68 | 904.39 | 666.35 | 342.63 | 122.76 | 592.7 |
| 1963 | 19837.07 | 9526.05 | 4620.77 | 4135.96 | 2652.05 | 1689.34 | 1138.44 | 789.35 | 638.21 | 481.32 | 254.28 | 182.18 |
| 1964 | 14811.79 | 15685.65 | 7491.63 | 3475.53 | 3010.73 | 1803.95 | 1200.34 | 774.64 | 503.3 | 437.46 | 241.15 | 224.48 |
| 1965 | 22362.92 | 11507.96 | 11115.89 | 4770.94 | 2287.23 | 1947.41 | 1093.3 | 820.79 | 498.49 | 321.58 | 283.06 | 239.61 |
| 1966 | 21229.27 | 17407.99 | 8652.81 | 7555.46 | 3032.95 | 1411.16 | 1226.14 | 618.24 | 490.13 | 266.98 | 154.71 | 232.85 |
| 1967 | 24897.65 | 16939.49 | 12859.01 | 5997.61 | 4660.33 | 1753.87 | 814.24 | 737.85 | 320.68 | 260.13 | 133.53 | 94.13 |
| 1968 | 22879.37 | 19846.09 | 13148.63 | 9293.87 | 4193.8 | 2736.99 | 1007.96 | 470.29 | 432.19 | 174.78 | 145.12 | 316.81 |
| 1969 | 39798.56 | 18176.48 | 14720.33 | 9755.39 | 6618.38 | 2937.74 | 1648.19 | 595.42 | 269.22 | 273.31 | 89.71 | 133.05 |
| 1970 | 37092.13 | 31506.65 | 12994.15 | 9976.29 | 6707.6 | 4407.06 | 1872.27 | 824.62 | 271.23 | 116.37 | 133.29 | 109.05 |
| 1971 | 38446.65 | 29060.97 | 19844.34 | 9228.97 | 6830.55 | 4678.27 | 2947.67 | 1246.96 | 457.08 | 144.25 | 51.84 | 130.67 |
| 1972 | 33424.45 | 28892.33 | 20792.67 | 11193.66 | 6646.68 | 4843.17 | 3405.87 | 2118.37 | 872.53 | 283.74 | 69.24 | 119.79 |
| 1973 | 23621.85 | 24909.9 | 22049.86 | 14681.88 | 6683.53 | 4058.35 | 2784.44 | 1868.27 | 1062.08 | 415.77 | 111.96 | 258.1 |
| 1974 | 19420.6 | 17064.27 | 14736.55 | 11651.17 | 8873.48 | 3993.51 | 2695.64 | 1782.05 | 1164.96 | 675.02 | 247.2 | 524.53 |
| 1975 | 17327.15 | 12729.69 | 10237.68 | 8435.96 | 7020.11 | 5997.32 | 2690.51 | 1874.02 | 1151.37 | 775.54 | 440.46 | 739.88 |
| 1976 | 19709.19 | 12320.5 | 7381.03 | 4942.62 | 5152.3 | 4802.02 | 4264.13 | 1929.54 | 1360.59 | 768.03 | 520.95 | 1132.94 |
| 1977 | 13106.08 | 13260.95 | 7176.31 | 4486.76 | 2915.63 | 3424.81 | 3351.56 | 3067.71 | 1378 | 986.38 | 541.95 | 815.91 |
| 1978 | 8332.93 | 9274.47 | 8199.64 | 4035.03 | 2508.02 | 1693.1 | 2163.37 | 2293.42 | 2205.8 | 882.09 | 690.85 | 837.16 |
| 1979 | 8686.33 | 6269.57 | 6016.16 | 5142.5 | 2807.75 | 1715.89 | 952.78 | 1349.56 | 1449.7 | 1437.68 | 531.28 | 1353.59 |
| 1980 | 13075.22 | 6852.07 | 4288.88 | 3712.23 | 3275.62 | 1770.37 | 1030.25 | 556.58 | 676.94 | 853.94 | 990.68 | 1390.32 |
| 1981 | 33145.15 | 9803.87 | 4816.46 | 2859.95 | 2430.36 | 2024.94 | 1192.48 | 651.67 | 300.96 | 376.88 | 557.99 | 2253.33 |
| 1982 | 15676.15 | 26765.06 | 6394.4 | 3247.57 | 1498.22 | 1168.22 | 993.73 | 665.96 | 359.81 | 163.16 | 192.75 | 3112.79 |
| 1983 | 40830.06 | 12484.37 | 18225.26 | 4335.88 | 1650.89 | 882.8 | 579.14 | 545.77 | 450.23 | 214.96 | 82.91 | 1368.13 |
| 1984 | 26075.32 | 31182.12 | 9223.3 | 10350.34 | 2334.72 | 831.36 | 416.04 | 227.14 | 358.16 | 279.95 | 86.42 | 839.91 |
| 1985 | 22332.2 | 21015.69 | 15515.92 | 5416.89 | 4770.65 | 1119.78 | 433.64 | 194.94 | 138.91 | 234.42 | 175.82 | 690.1 |
| 1986 | 61856.33 | 17176.53 | 13595.89 | 7651.66 | 3365.46 | 2188.49 | 669.78 | 261.83 | 125.22 | 90.21 | 126.78 | 154.24 |
| 1987 | 48619.31 | 49587.74 | 12256 | 7084.03 | 2889.6 | 1893.09 | 817.28 | 326.68 | 120.27 | 41.9 | 44 | 321.7 |
| 1988 | 44855 | 38375.57 | 35357.28 | 6571.55 | 3279.94 | 1470.02 | 1068.56 | 367.82 | 194.17 | 59.56 | 29.78 | 3.91 |
| 1989 | 28601.04 | 35940.58 | 28749.99 | 20302.37 | 2861.26 | 1509.1 | 641.64 | 546.41 | 157.28 | 134.54 | 9.85 | 132.25 |
| 1990 | 20712.55 | 23008.47 | 24013.82 | 18742.86 | 10165.26 | 1625.07 | 741.51 | 357.94 | 372.26 | 79 | 101.11 | 222.27 |
| 1991 | 24971.59 | 16691.98 | 15369.5 | 10503.9 | 7003.66 | 3735.08 | 898.88 | 495.8 | 237.86 | 250.49 | 48.39 | 41.25 |
| 1992 | 19572.3 | 19513.03 | 9028.95 | 5840.63 | 3583.45 | 2578.95 | 1583.14 | 369.49 | 190.57 | 78.92 | 135.41 | 48.95 |
| 1993 | 23780.38 | 15553.02 | 12295.94 | 4074.24 | 2357.85 | 1691.54 | 1302.54 | 741.5 | 191.21 | 99.02 | 31.14 | 24.69 |
| 1994 | 16877.27 | 18278.96 | 10371.21 | 5824.28 | 1829.15 | 1154.09 | 939.73 | 661.06 | 385.4 | 107.69 | 50.31 | 5.24 |
| 1995 | 38973.1 | 13193.6 | 11381.49 | 6081.64 | 2625.86 | 822.57 | 492.47 | 491.6 | 267.07 | 179.82 | 62.83 | 47.22 |
| 1996 | 24356.89 | 31548.35 | 9879.98 | 6180.4 | 3317.95 | 1084.88 | 361.3 | 185.14 | 228.76 | 124.55 | 81.17 | 22.17 |
| 1997 | 33517.36 | 19673 | 24846.05 | 7052.1 | 3748.97 | 1670.51 | 416.81 | 176.37 | 81.9 | 129.38 | 61.26 | 47.91 |
| 1998 | 12756.42 | 27130.43 | 15354.06 | 18134.42 | 4174.92 | 1861.44 | 803.08 | 191.95 | 80.15 | 40.82 | 62.5 | 54.1 |
| 1999 | 58813.51 | 10296.59 | 20684.24 | 10820.89 | 11702.9 | 2170.37 | 906.01 | 324.53 | 87.49 | 36.67 | 8.08 | 58.61 |
| 2000 | 35840.25 | 47861.07 | 7837.47 | 14133.45 | 6551.15 | 5859.93 | 949.02 | 397.94 | 132.69 | 49.91 | 5.59 | 13.84 |
| 2001 | 88038.88 | 28609.69 | 36624.64 | 5074 | 7618.26 | 3349.46 | 2332.02 | 462.11 | 157.51 | 57.97 | 24.58 | 16.29 |
| 2002 | 105902.44 | 71062.19 | 21204.97 | 22351.6 | 2204.31 | 2907.5 | 1349.76 | 702.24 | 113.23 | 47.52 | 25.74 | 3.19 |


| year | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2003 | 64250.85 | 86432.32 | 50581.07 | 11966.52 | 9445.21 | 1024.77 | 1222.27 | 684.34 | 247.4 | 62.84 | 6.33 | 0 |
| 2004 | 53951.7 | 52305.55 | 68564.24 | 31321.53 | 6183.44 | 3854.04 | 461.69 | 621.58 | 285.22 | 120.21 | 15.26 | 2.98 |
| 2005 | 69984.38 | 44103.14 | 41004.54 | 48404.72 | 17715.72 | 3139.79 | 1520.38 | 138.22 | 243.79 | 101.41 | 7.94 | 0 |
| 2006 | 22222.42 | 56887.57 | 33441.14 | 24988.39 | 24604.7 | 8080.96 | 1807.87 | 511.87 | 84.21 | 107.3 | 58.6 | 24.27 |
| 2007 | 18880.03 | 16859.55 | 42010.7 | 20338.75 | 13480.66 | 10828.42 | 3213.94 | 901.06 | 163.92 | 39.99 | 77 | 25.56 |
| 2008 | 31507.33 | 14705.74 | 10799.37 | 24166.26 | 10794.95 | 7615.84 | 4980.19 | 1239.71 | 370.36 | 60.91 | 22.79 | 0 |
| 2009 | 13723.87 | 21467.28 | 9227.81 | 5586.17 | 11307.34 | 5586.17 | 4223.87 | 2224.33 | 613.24 | 159.36 | 39.01 | 12.96 |
| 2010 | 24290.81 | 10820.83 | 10869.27 | 3050.81 | 2906.86 | 4582.37 | 2755.75 | 1923.6 | 853.85 | 237.87 | 93.37 | 0 |
| 2011 | 34720.15 | 17784.79 | 6220.84 | 4105.17 | 1479.85 | 1466.95 | 1853.38 | 1126.98 | 822.09 | 358.86 | 148.6 | 17.37 |
| 2012 | 34439.89 | 26709.98 | 12078.08 | 3337.81 | 1728.71 | 779.99 | 721.48 | 880.41 | 451.27 | 275.85 | 168.94 | 38.35 |
| 2013 | 35951.06 | 27419.74 | 12971.02 | 6140.86 | 1599.91 | 600.09 | 363.53 | 312.91 | 359.79 | 162.26 | 102.79 | 93.87 |
| 2014 | 61619.17 | 28781.85 | 17769.57 | 6802.26 | 2999.07 | 847.52 | 302.2 | 187.24 | 169.32 | 162.46 | 55.94 | 3.47 |
| 2015 | 26993.00 | 49649.00 | 21472.00 | 11711.00 | 4048.00 | 1674.00 | 396.00 | 158.00 | 70.00 | 75.00 | 83.00 | 33.00 |

Table 6.3.3. Faroe saithe (Division Vb). Summary table (1961-2014). Values for 2015-2017 are estimates.

| year | Recruits (age 3) | SSB (tonnes) | Yield (tonnes) | Yield/SSB | Fbar(4-8) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1961 | 7827 | 68467 | 9592 | 0.13 | 0.106 |
| 1962 | 12256 | 72862 | 10454 | 0.154 | 0.125 |
| 1963 | 19837 | 76441 | 12693 | 0.173 | 0.114 |
| 1964 | 14811 | 80928 | 21893 | 0.272 | 0.23 |
| 1965 | 22362 | 84690 | 22181 | 0.284 | 0.214 |
| 1966 | 21229 | 87313 | 25563 | 0.3 | 0.25 |
| 1967 | 24897 | 85361 | 21319 | 0.241 | 0.204 |
| 1968 | 22879 | 93938 | 20387 | 0.213 | 0.16 |
| 1969 | 39798 | 103452 | 27437 | 0.274 | 0.191 |
| 1970 | 37092 | 109688 | 29110 | 0.275 | 0.189 |
| 1971 | 38446 | 121969 | 32706 | 0.245 | 0.179 |
| 1972 | 33424 | 137956 | 42663 | 0.308 | 0.236 |
| 1973 | 23621 | 130735 | 57431 | 0.439 | 0.318 |
| 1974 | 19420 | 134009 | 47188 | 0.352 | 0.272 |
| 1975 | 17327 | 135484 | 41576 | 0.307 | 0.297 |
| 1976 | 19709 | 129099 | 33065 | 0.256 | 0.267 |
| 1977 | 13106 | 122227 | 34835 | 0.273 | 0.328 |
| 1978 | 8332 | 105216 | 28138 | 0.266 | 0.243 |
| 1979 | 8686 | 96036 | 27246 | 0.277 | 0.257 |
| 1980 | 13075 | 96216 | 25230 | 0.264 | 0.211 |
| 1981 | 33145 | 85056 | 30103 | 0.37 | 0.382 |
| 1982 | 15676 | 94389 | 30964 | 0.341 | 0.336 |
| 1983 | 40830 | 98639 | 39176 | 0.397 | 0.385 |
| 1984 | 26075 | 104707 | 54665 | 0.523 | 0.478 |
| 1985 | 22332 | 110005 | 44605 | 0.431 | 0.382 |
| 1986 | 61856 | 91583 | 41716 | 0.484 | 0.505 |
| 1987 | 48619 | 94297 | 40020 | 0.441 | 0.396 |
| 1988 | 44854 | 103005 | 45285 | 0.443 | 0.456 |
| 1989 | 28601 | 107398 | 44477 | 0.427 | 0.36 |
| 1990 | 20712 | 103216 | 61628 | 0.609 | 0.562 |
| 1991 | 24971 | 76177 | 54858 | 0.725 | 0.703 |
| 1992 | 19572 | 59993 | 36487 | 0.579 | 0.52 |
| 1993 | 23780 | 59260 | 33543 | 0.557 | 0.451 |
| 1994 | 16877 | 57407 | 33182 | 0.564 | 0.491 |
| 1995 | 38973 | 55521 | 27209 | 0.48 | 0.442 |
| 1996 | 24356 | 60584 | 20029 | 0.32 | 0.344 |
| 1997 | 33517 | 68222 | 22306 | 0.327 | 0.304 |
| 1998 | 12756 | 74050 | 26421 | 0.349 | 0.287 |
| 1999 | 58813 | 77635 | 33207 | 0.42 | 0.335 |
| 2000 | 35840 | 80387 | 39020 | 0.478 | 0.383 |
| 2001 | 88038 | 83993 | 51786 | 0.616 | 0.5 |
| 2002 | 105902 | 81692 | 53546 | 0.655 | 0.481 |


| year | Recruits (age <br> 3) | SSB (tonnes) | Yield (tonnes) | Yield/SSB | Fbar(4-8) |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 2003 | 64250 | 97221 | 46555 | 0.478 | 0.413 |
| 2004 | 53951 | 112980 | 46355 | 0.409 | 0.354 |
| 2005 | 69984 | 127585 | 67967 | 0.533 | 0.357 |
| 2006 | 22222 | 127123 | 66902 | 0.528 | 0.432 |
| 2007 | 18880 | 120818 | 60785 | 0.505 | 0.396 |
| 2008 | 31507 | 104362 | 57044 | 0.542 | 0.427 |
| 2009 | 13723 | 93278 | 57949 | 0.614 | 0.61 |
| 2010 | 24290 | 69401 | 43885 | 0.632 | 0.568 |
| 2011 | 34720 | 56238 | 29658 | 0.527 | 0.445 |
| 2012 | 34439 | 49174 | 35314 | 0.718 | 0.591 |
| 2013 | 35951 | 48637 | 26463 | 0.544 | 0.423 |
| 2014 | 61619 | 70026 | 23854 | 0.341 | 0.315 |
| 2015 | 26993 | 82089 | 35361 |  | 0.314 |
| 2016 | 26993 | 96782 | 37467 |  | 0.314 |
| 2017 | 26993 | 104194 |  |  |  |
| Avg. | 31662 | 92151 | 36994 | 0.41 | 0.36 |

Table 6.6.1.1. Faroe saithe (Division Vb). Input data for prediction with management options for the SPALY assessment .

| 2015 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 3 | 26993 | 0.2 | 0.04 | 0 | 0 | 1.299 | 0.016 | 1.299 |
| 4 | 49649 | 0.2 | 0.26 | 0 | 0 | 1.528 | 0.201 | 1.528 |
| 5 | 21472 | 0.2 | 0.51 | 0 | 0 | 1.850 | 0.269 | 1.850 |
| 6 | 11711 | 0.2 | 0.74 | 0 | 0 | 2.239 | 0.324 | 2.239 |
| 7 | 4048 | 0.2 | 0.90 | 0 | 0 | 2.602 | 0.397 | 2.602 |
| 8 | 1674 | 0.2 | 0.98 | 0 | 0 | 4.451 | 0.381 | 4.451 |
| 9 | 396 | 0.2 | 0.98 | 0 | 0 | 5.296 | 0.366 | 5.296 |
| 10 | 158 | 0.2 | 1.00 | 0 | 0 | 5.942 | 0.448 | 5.942 |
| 11 | 70 | 0.2 | 1.00 | 0 | 0 | 6.088 | 0.480 | 6.088 |
| 12 | 75 | 0.2 | 1.00 | 0 | 0 | 6.634 | 1.000 | 6.634 |
| 13 | 83 | 0.2 | 1.00 | 0 | 0 | 7.402 | 1.000 | 7.402 |
| 14 | 33 | 0.2 | 1.00 | 0 | 0 | 8.232 | 1.000 | 8.232 |
| $2016$ |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 3 | 26993 | 0.2 | 0.04 | 0 | 0 | 1.299 | 0.016 | 1.299 |
| 4 | - | 0.2 | 0.26 | 0 | 0 | 1.528 | 0.201 | 1.528 |
| 5 | - | 0.2 | 0.51 | 0 | 0 | 1.850 | 0.269 | 1.850 |
| 6 | - | 0.2 | 0.74 | 0 | 0 | 2.239 | 0.324 | 2.239 |
| 7 | - | 0.2 | 0.90 | 0 | 0 | 2.602 | 0.397 | 2.602 |
| 8 | - | 0.2 | 0.98 | 0 | 0 | 4.451 | 0.381 | 4.451 |
| 9 | - | 0.2 | 0.98 | 0 | 0 | 5.296 | 0.366 | 5.296 |
| 10 | - | 0.2 | 1.00 | 0 | 0 | 5.942 | 0.448 | 5.942 |
| 11 | - | 0.2 | 1.00 | 0 | 0 | 6.088 | 0.480 | 6.088 |
| 12 | - | 0.2 | 1.00 | 0 | 0 | 6.634 | 1.000 | 6.634 |
| 13 | - | 0.2 | 1.00 | 0 | 0 | 7.402 | 1.000 | 7.402 |
| 14 | - | 0.2 | 1.00 | 0 | 0 | 8.232 | 1.000 | 8.232 |
| 2017 |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 3 | 26993 | 0.2 | 0.04 | 0 | 0 | 1.299 | 0.016 | 1.299 |
| 4 | - | 0.2 | 0.26 | 0 | 0 | 1.528 | 0.201 | 1.528 |
| 5 | - | 0.2 | 0.51 | 0 | 0 | 1.850 | 0.269 | 1.850 |
| 6 | - | 0.2 | 0.74 | 0 | 0 | 2.239 | 0.324 | 2.239 |
| 7 | - | 0.2 | 0.90 | 0 | 0 | 2.602 | 0.397 | 2.602 |
| 8 | - | 0.2 | 0.98 | 0 | 0 | 4.451 | 0.381 | 4.451 |
| 9 | - | 0.2 | 0.98 | 0 | 0 | 5.296 | 0.366 | 5.296 |
| 10 | - | 0.2 | 1.00 | 0 | 0 | 5.942 | 0.448 | 5.942 |
| 11 | - | 0.2 | 1.00 | 0 | 0 | 6.088 | 0.480 | 6.088 |
| 12 | - | 0.2 | 1.00 | 0 | 0 | 6.634 | 1.000 | 6.634 |
| 13 | - | 0.2 | 1.00 | 0 | 0 | 7.402 | 1.000 | 7.402 |
| 14 | - | 0.2 | 1.00 | 0 | 0 | 8.232 | 1.000 | 8.232 |

Input units are thousands and kg - output in tonnes

Table 6.6.2.1. Faroe saithe (Division Vb). Prediction with management option for SPALY assessment.

| 2015 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar | Landings |  |  |
| 199713 | 82089 | 1.000 | 0.314 | 35361 |  |  |
| 2016 |  |  |  |  | 2017 |  |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 195333 | 96782 | 0.0000 | 0.0000 | 0 | 232926 | 138536 |
| . | 96782 | 0.1000 | 0.0314 | 4263 | 228057 | 134572 |
| . | 96782 | 0.2000 | 0.0629 | 8399 | 223339 | 130737 |
| . | 96782 | 0.3000 | 0.0943 | 12414 | 218766 | 127027 |
| . | 96782 | 0.4000 | 0.1258 | 16310 | 214335 | 123437 |
| - | 96782 | 0.5000 | 0.1572 | 20092 | 210039 | 119962 |
| . | 96782 | 0.6000 | 0.1886 | 23764 | 205875 | 116600 |
| . | 96782 | 0.7000 | 0.2201 | 27329 | 201838 | 113345 |
| . | 96782 | 0.8000 | 0.2515 | 30791 | 197924 | 110195 |
| . | 96782 | 0.9000 | 0.2830 | 34153 | 194128 | 107146 |
| . | 96782 | 1.0000 | 0.3144 | 37418 | 190448 | 104194 |
| - | 96782 | 1.1000 | 0.3458 | 40590 | 186878 | 101335 |
| . | 96782 | 1.2000 | 0.3773 | 43671 | 183416 | 98567 |
| . | 96782 | 1.3000 | 0.4087 | 46664 | 180058 | 95887 |
| . | 96782 | 1.4000 | 0.4402 | 49572 | 176800 | 93292 |
| . | 96782 | 1.5000 | 0.4716 | 52398 | 173640 | 90778 |
| . | 96782 | 1.6000 | 0.5030 | 55144 | 170574 | 88344 |
| . | 96782 | 1.7000 | 0.5345 | 57813 | 167599 | 85986 |
| . | 96782 | 1.8000 | 0.5659 | 60407 | 164713 | 83701 |
| . | 96782 | 1.9000 | 0.5974 | 62929 | 161911 | 81488 |
| . | 96782 | 2.0000 | 0.6288 | 65380 | 159192 | 79345 |

Input units are thousands and kg - output in tonnes

Table 6.7.1.1. Faroe saithe (Division Vb). Yield per recruit input data.

| Age | $\mathbf{M}$ | Mat | PF | PM | WeSt | Sel | WeCa |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 0.2 | 0.02 | 0 | 0 | 1.304 | 0.048 | 1.304 |
| 4 | 0.2 | 0.21 | 0 | 0 | 1.668 | 0.278 | 1.668 |
| 5 | 0.2 | 0.47 | 0 | 0 | 2.031 | 0.467 | 2.031 |
| 6 | 0.2 | 0.71 | 0 | 0 | 2.602 | 0.5118 | 2.602 |
| 7 | 0.2 | 0.86 | 0 | 0 | 3.373 | 0.52 | 3.373 |
| 8 | 0.2 | 0.95 | 0 | 0 | 4.318 | 0.5648 | 4.318 |
| 9 | 0.2 | 0.99 | 0 | 0 | 5.085 | 0.5572 | 5.085 |
| 10 | 0.2 | 1 | 0 | 0 | 5.904 | 0.6514 | 5.904 |
| 11 | 0.2 | 1 | 0 | 0 | 6.777 | 0.7174 | 6.777 |
| 12 | 0.2 | 1 | 0 | 0 | 7.472 | 0.5888 | 7.472 |
| 13 | 0.2 | 1 | 0 | 0 | 7.835 | 0.4844 | 7.835 |
| 14 | 0.2 | 1 | 0 | 0 | 9.388 | 0.4844 | 9.388 |

Table 6.9.1. Faroe saithe (Division Vb). Comparison between the current assessment (NWWG2015 SPALY) statistical assessment (NWWG2015 ADMB) and predictions from last year in the terminal year (2014).

|  | NWWG2014 <br> prediction | NWWG2015 <br> (SPALY) | NWWG2015 (ADMB) |
| :--- | :--- | :--- | :--- |
| Recruitment | 28 mill. | 62 mill. | 20 mill. |
| SSB | 70000 t. | 70000 t. | 94000 t. |
| Fbar(4-8) | 0.53 | 0.32 | 0.23 |
| Landings | 38000 t. | 24000 t. | 27000 t. |



Figure 6.2.1.1. Faroe saithe (Division Vb). Landings in 1000 tonnes (1961-2014). Horizontal red line represents historical average landings.


Figure 6.2.1.2. Saithe in the Faroes (Division Vb). Cumulative domestic landings (2000-2015).


Figure 6.2.3.1. Faroe saithe (Division Vb). Mean weight at age (kg) in commercial catches (ages 3-9) (1961-2017). Weights from 2015 to 2017 are estimates. Horizontal lines show historical average.


Figure 6.2.4.1. Faroe saithe (Division Vb). Smoothed maturity ogives (ages 3-8)(1983-2015) from FGFS1 (spring survey). Horizontal lines show historical average.


Figure 6.2.5.1.1. Faroe saithe (Division Vb). Predicted catch rates from the commercial fleet (pairtrawlers) used for tuning the assessment (black line). Catch rates (kg/hour) from the Faroese bot-tom-trawl fall FGFS2 (1996-2014)(red line) and spring survey FGFS1 (1994-2015)(blue line). Dotted lines and shade areas show standard errors in the estimation of indices.


Figure 6.2.5.1.2. Faroe saithe (Division Vb). Length composition from the Faroese bottom-trawl spring survey FGFS1 (1994-2015)


Figure 6.2.5.1.3. Faroe saithe (Division Vb). Length composition from the Faroese bottom-trawl summer survey FGFS2 (1996-2014)


Figure 6.2.5.1.4. Faroe saithe (Division Vb). Age-disaggregated indices in the Faroese bottom-trawl spring survey FGFS1 (ages 3-10, years 1994-2015)


Figure 6.2.5.1.5. Faroe saithe (Division Vb). Age-disaggregated indices in the Faroese bottom-trawl fall survey FGFS2 (ages 3-10, years 1996-2014)


Figure 6.2.5.1.6. Faroe saithe (Division Vb). Numbers from spring survey (FGFS1) plotted against numbers of the same year class one year later. Letters in the figures represent year classes.


Figure 6.2.5.1.7. Faroe saithe (Division Vb). Numbers from summer survey (FGFS2) plotted against numbers of the same year class one year later. Letters in the figures represent year classes.


Figure 6.2.5.2.1. Faroe saithe (Division Vb). Age-disaggregated indices in the commercial pair-trawl fleet (ages 3-10, years 1995-2014)


Figure 6.2.5.2.2. Faroe saithe (Division Vb). Indices from in the commercial pair-trawl plotted against indices of the same year class one year later. Letters in the figures represent year classes.


Figure 6.3.1. Faroe saithe (Division Vb). Log-catchability residuals of the spaly assessment calibrated with the commercial series (ages 3-11, years 1995-2014). Blue and red bubbles represent positive and negative residuals respectively.



Fig-
ure 6.3.3. Faroe saithe (Division Vb). Catch- (ages 3-14+, years 1961-2014)(top plot) and survey-atage (ages 3-11, years 1995-2014)(bottom plot) residuals from a statistical catch-at-age model. Red and white bubbles represent positive and negative residuals respectively.


Figure 6.4.1.1. Faroe saithe (Division Vb). EqSim simulation. Stock-recruitment function used in the simulations (Hockey-stick).


Figure 6.4.1.2. Faroe saithe (Division Vb). EqSim simulation outputs with assessment errors and Hockey-stick function from WKMSYREF2 report. Blim is undefined but was set as Blim=Bpa/1.4.


Figure 6.4.1.3. Faroe saithe (Division Vb). Stock-Recruitment plot in relation to Flow=0.13 (lowest regression line), Fmed=0.31 (middle regression line) and Fhigh= 0.80 (top regression line). Vertical red line represents Btrigger= 55000 t .


Figure 6.5.1. Faroe saithe (Division Vb). Recruitment (age 3) in millions (top-left), total stock biomass (thousand tonnes)(top-middle), spawning stock biomass (thousand tonnes) (bottom-left), landings (thousand tonnes)(middle-left), landings SSB ratio (middle-middle), Fbar (ages 4 to 8)(middle-right), reference biomass (B4+) (thousand tonnes) (bottom-left) and landings B4+ ratio (bottom-right). Black line represents the spaly run. Green lines show estimates from a catch-at-age statistical model implemented in ADMB. Red lines show a 'a4a' statistical model implemented in R. Horizontal blue lines represent historical averages.


Figure 6.5.2. Faroe saithe (Division Vb). Fishing mortality (average over ages 4-8)(1961-2014)


Figure 6.5.3. Faroe saithe (Division Vb). Recruitment at age 3 (millions)(1961-2015). The 2015 recruitment estimate is used in the short-term forecast.


Figure 6.5.4. Faroe saithe (Division Vb). Spawning stock biomass (' 000 tonnes)(1961-2015). The 2015 SSB estimate is used in the short-term forecast. Horizontal lines represent Btrigger=Bpa=55 000 t .


Figure 6.5.6. Faroe saithe (Division Vb). Numbers of mature fish in the stock (ages 3-9) for 2006, 2013 and 2014.


Figure 6.5.7. Faroe saithe (Division Vb). SSB - Recruitment (age 3) plot. Btrigger=Bpa=55 000 t .


Figure 6.6.1.1. Faroe saithe (Division Vb). Residual plots from a 3-year running average weight model and the model in which weights are predicted from the previous year in the same year class. Red and white bubbles represent positive and negative residuals respectively.


Figure 6.6.1.2. Faroe saithe (Division Vb). Observed (stapled lines) and predicted weights (solid lines)(ages 4-8, years 1985-2014)


Figure 6.6.2.1. Faroe saithe (Division Vb). Short-term prediction output (spaly assessment). Solid and broken lines represent landings ( t ) and spawning stock biomass ( t ) respectively.


Figure 6.6.2.2 Faroe saithe (Division Vb). Composition of landings (upper figure) and SSB (lower figure) by year classes in 2015.


Figure 6.7.1.1. Faroe saithe (Division Vb). Yield and spawning per-recruit calculations. Dashed and solid lines represent Yield/R and SSB/R respectively.


Figure 6.8.1. Faroe saithe (Division Vb). Retrospective analysis of recruitment at age 3 (millions)(top figure), spawning stock biomass (' 000 tonnes)(middle figure) and average fishing mortality over age groups 4-8 (bottom figure) from the spaly assessment.


Figure 6.15.1. Faroe saithe (Division Vb). Relationship between the Gyre index (4 years shifted) and saithe biomasse (age 3+) in Faroese waters.

## 7 Overview on ecosystem, fisheries and their management in Icelandic waters

This section gives a very broad and general overview of the marine ecosystem, fishery, fleet, species composition and some bycatch analysis of the commercially landed species as well as management measures in the Icelandic Exclusive Economic Zone. The Icelandic EEZ covers partly the IIa2, Va1, Va2, Vb1b, XIIa4, XIVa and XIVb2 ICES statistical regions. In practice however, the Icelandic landings of different species are generally reported as catches/landings in Va.
The information on the ecosystem of Icelandic waters is brief but a more detailed description is available in the WGRED report (ICES 2008).

### 7.1 Environmental and ecosystem information

Iceland is located at the junction of the Mid-Atlantic Ridge and the Greenland-Scotland Ridge just south of the Arctic Circle and this is reflected in the topography around the country. Substrate characteristics can be largely influenced by depth. Hard bottom is more often found in shallower waters compared to deep waters. In deeper waters, hard bottom is often confined to abrupt features such as ridges and seamounts. Soft sediments often dominate in the troughs and outside the continental slope. The shelf around Iceland is narrowest off the south coast (Figure 7.3.4) and is cut by submarine canyons around the country (Figure 7.3.4).

The Polar Front lies west and north off Iceland and separates the cold and southward flowing waters of Polar origin from the northward flowing waters of Atlantic origin. South and east off Iceland the North Atlantic Current flows towards the Norwegian Sea. The Irminger Current is a branch of the North Atlantic Current and flows northwards over and along the Reykjanes Ridge and along the western shelf break. In the Denmark Strait it divides into a branch that flows northeastward and eastward to the waters north off Iceland, as the North Icelandic Irminger Current, and another branch that flows south-westward along the East Greenland Current. In the Iceland Sea north off Iceland, a branch originating from the cold East Greenland Current flows over the Kolbeinsey Ridge and continues to the southeast along the northeastern shelf brake as the East Icelandic Current, which is part of a cyclonic gyre in the Iceland Sea. This current subsequently continues into the Norwegian Sea along the Atlantic water flowing eastwards over the Iceland-Faroes Ridge (Stefansson 1962, Valdimarsson and Malmberg 1999).

The Icelandic Shelf is a high ( $150-300 \mathrm{gC} / \mathrm{m} 2-\mathrm{yr}$ ) productivity ecosystem according to SeaWiFS global primary productivity estimates. Productivity is higher in the southwest regions than to the northeast and higher on the shelf areas than in the oceanic regions (Gudmundsson 1998). In terms of abundance, copepods dominate the mesozooplankton within Icelandic waters with Calanus finmarchicus being the most abundant species, often comprising between $60-80 \%$ of net-caught zooplankton in the uppermost 50 m (Astthorsson and Vilhjalmsson 2002, Astthorsson et al. 2007).

The structure of benthic communities in Icelandic waters is likely to be influenced by a large number of factors. Amongst these, water mass characteristics will have profound effects on species composition and spatial distribution patterns at the largest spatial scales (e.g. $>50 \mathrm{~km}$ ) whereas substrate characteristics (e.g. sediment type and rugosity) and topography will have profound effects on smaller scales (e.g. meters to
kilometers), (e.g. Weisshappel and Svavarsson 1998). Shrimp biomass in Icelandic waters, both in inshore and offshore waters, has been declining in recent years. Consequently the fishing effort was reduced and is now banned in most inshore areas. The causes for the decline in the inshore shrimp biomass is in part considered to be environmentally driven, both due to increasing water temperature north of Iceland and due to increasing biomass of younger cod, haddock and whiting.
Based on information from fishermen, eleven cold-water coral areas were known to exist close to the shelf break off the northwest towards southeast Iceland around 1970. During the 70s and 80s, more coral areas were found by fishermen as a direct consequence of the bottom trawling fisheries extending into deeper waters. More recently there has been a considerable effort in mapping cold-water coral habitats in Icelandic waters and to investigate their biology using the state of the art technology such as unmanned submersibles. At present, large cold-water coral areas have been located on the Reykjanes Ridge and on the shelf break south and southeast Iceland (Steingrímsson and Einarsson 2004). Many of the cold-water coral areas that have been surveyed have already been destroyed. Currently, 5 areas with relatively undisturbed cold-water corals have received full protection and several other areas are under consideration for further protection.

The database of the BIOICE programme provides information on the spatial distribution of benthic organisms within the Icelandic territorial waters based on samples collected from 579 locations, including horny corals (Gorgonacea) and seapens (Pennatulacea) that are considered sensitive to fishing. Gorgonian corals occur all around Iceland but these are relatively uncommon on the shelf ( $<500 \mathrm{~m}$ depth) but can be found in relatively high numbers in deep waters ( $>500 \mathrm{~m}$ ) off south, west and north coasts of Iceland, given the right environmental conditions. Similar distribution patterns were observed in the distribution of pennatulaceans, these being common in deeper waters, especially off South Iceland (Guijarro et al. 2007).

About 25 species of stocks of fish and marine invertebrates are exploited commercially on a regular basis in Icelandic waters.
Icelandic waters are comparatively rich in species and contain around 30 commercially exploited stocks of fish and marine invertebrates. The most important commercial species are cod, haddock, saithe, redfish, Greenland halibut and various other flatfish, wolffish, tusk (Brosme brosme), ling (Molva molva), herring, capelin and blue whiting. Most fish species spawn in the warm Atlantic water off the south and southwest coasts. Fish larvae and 0-group subsequently drift west and then north from the spawning grounds to nursery areas on the shelf off northwest, north and east Iceland, where they grow in a mixture of Atlantic and Arctic water.

Capelin is important in the diet of cod as well as a number of other fish stocks, marine mammals and seabirds. Unlike other commercial stocks, adult capelin undertake extensive feeding migrations north into the cold waters of the Denmark Strait and Iceland Sea during summer. Capelin abundance has been oscillating on roughly a decadal period since the 1970s, producing a yield of up to 1600 Kt at the most recent peak. In recent years the stock size of capelin has decreased from about 2000 Kt in 1996/97 to about 900 Kt in 2012/13 (Anon. 2013). Herring were very abundant in the early 1960s until the stock collapsed in the nineteen sixties due to overfishing. From 1970 onwards the stock size has increased until attaining historical high levels in the last decade. Abundance of demersal species have been generally trending downward since the 1950s with total catches dropping from over 800 Kt to less than 500 Kt in the early 2000s.

A number of species of sharks and skates are known to be caught as a by-catch in Icelandic waters, but information on amount of the catches is incomplete, and the status of these species is not known. Information on status and trends of non-commercial species are collected in extensive bottom trawl surveys conducted in early spring and autumn.
The seabird community in Icelandic waters is composed of relatively few but mostly abundant species, accounting for roughly $1 / 4$ of total number and biomass of seabirds within the whole ICES area (ICES 2002). Auks and petrels are the most important groups, comprising almost $3 / 5$ and $1 / 4$ of the total abundance and biomass in the area, respectively. The estimated annual food consumption is on the order of 1.5 million tonnes.

At least 12 species of cetaceans occur regularly in Icelandic waters, and additional 10 species have been recorded more sporadically. In the continental shelf area, the minke whale (Balaenoptera acutorostrata) probably has the largest biomass. Based on the 2001 sightings survey, 67000 minke whales were estimated in the Central North Atlantic stock region, with 44000 animals in Icelandic coastal waters (NAMMCO 2004, Borchers et al. 2003, Gunnlaugsson 2003). In the 2007 aerial survey the abundance of minke whales was estimated at around 21000 animals on the Icelandic shelf. The reasons for this decrease are not known. Two species of seals, common seal (Phoca vitulina) and grey seal (Halichoerus grypus) breed in Icelandic waters, while 5 other species are found as vagrants (Sigurjonsson and Hauksson, 1994; Hauksson, 1993, 2004).

### 7.2 Environmental drivers of productivity

Mean weight at age of Icelandic cod have been shown to correlate well with the size of the capelin stock and therefore the capelin stock was used as a predictor of weights in the landings in 1991-2007. In 1981-1982, cod weights were low following collapse of the capelin stock and were also relatively low in 1990-1991 when the capelin stock was small. In recent years this relationship seems to be much weaker and have not been used for predictions. The reasons for these changes are most likely changes in the spatial distribution of capelin or uncertainties in the estimation of the capelin stock size.
No other ecosystem drivers of productivity that may affect the assessment of the Icelandic stocks assessed in this report were presented to the NWWG in 2013.

### 7.3 Ecosystem considerations (General)

After 1996 a rise in both temperature and salinity were observed in the Atlantic water south and west of Iceland. Temperature and salinity have remained at similar high levels since and west of Iceland amounts to an increase of temperature of about $1^{\circ} \mathrm{C}$ and salinity by one unit on average (Figure 7.3.1.) and these changes can therefore be regarded as conspicuous. Off central N -Iceland, similar trends have been observed although with higher inter-annual variability. This period has been characterized with an increase of temperature and salinity in the winter north of Iceland in the last $12-14$ years which is on average above $1^{\circ} \mathrm{C}$ and 1 salinity units. (Figure 7.3.2)

It appears that these changes in seawater temperature have had considerable effects on the spatial distribution of fish species in Icelandic waters with many species now found further northwards. The most obvious examples of such changes is the increased abundance of haddock, mackerel, whiting, monkfish, lemon sole and witch in the mixed water area north of Iceland.

On the other hand, coldwater species like Greenland halibut and northern shrimp have become scarcer. Capelin have shifted their larval drift and nursing areas westwards to the colder waters off E-Greenland. Furthermore, the arrival of adult capelin to the overwintering grounds on the outer shelf off N -Iceland has been delayed and migration routes to the spawning grounds off S- and W-Iceland are currently located farther off N - and E-Iceland and do not reach as far west along the south coast as was the rule in most earlier years (Figure 7.3.3. and 7.3.4.). These changes in the spatial distribution patterns of capelin may have had an effect on the growth rate of various predators, as is reflected in low weight of cod in recent years.

There is one demersal stock, which apparently has not taken advantage, or not been able to take advantage, of the milder marine climate of Icelandic waters. This is the Icelandic cod, which was very abundant during the last warm epoch, which began around 1920 and lasted until 1965. By the early 1980s the cod stock had been fished down to much lower levels as compared to previous decades and has remained relatively low since. During the last 20 years the Icelandic cod stock has not produced a large year class and the average number of age 3 recruits in the last 20 years is about 150 million fish per annum, as compared to 205-210 recruits in almost any period prior to that, even during the ice years of 1965-1971. Immigrants from Greenland are not included in this comparison. It is not possible to pinpoint exactly what has caused this change, but a very small and young spawning stock is the most obvious common denominator for this protracted period of impaired recruitment to the Icelandic cod stock. Regulations, particularly the implementation of the catch rule in 1993 have resulted in lower fishing mortalities in the last ten years when compared with the years prior to 2000. Further, despite the overall low recruitment, this reduction in fishing mortality has almost resulted in almost doubling of the spawning stock biomass. This increase in the SSB biomass has however not resulted in significant increase in recruitment in recent years, although year classes 2008 and 2009 are now estimated around average size.

Associated with the large warming of the 1920s, was a well documented drift of larval and 0-group cod as well as some other fish species, from Iceland across the northern Irminger Sea to East and then West-Greenland. Although many of these fish apparently returned to Iceland to spawn and did not leave again, there is little doubt that the cod, remaining in West-Greenland waters which also had warmed, were instrumental in establishing a self-sustaining Greenlandic cod stock that eventually became very large. It seems that significant numbers of cod of the 2003 year class have drifted across to Greenland in that year. Tag returns, survey estimates in Greenlandic waters as well as anomalies in the catch-at-age matrix in Iceland indicate that a portion of the moderate 2003 year class that has been observed in Greenlandic waters in recent years may have migrated to Icelandic waters in 2009.

### 7.4 Description of fisheries [Fleets]

Only Icelandic vessels are considered in the following analysis since they constitute the largest operational players in Icelandic waters. Few trawlers and longliners of other nationalities operate in the Icelandic region principally targeting deep-sea redfish, cod, tusk, ling and, with some bycatch of other species. Additionally some limited pelagic fishery of foreign boats on capelin, herring and blue whiting also takes place in Icelandic waters.

The data sources used in this section are landings, boat, log book and discard databases. Landings of species by each boat and gear are effectively available electronically
in real time (end of day of landing). Log-book statistics are generally available in a centralized database about 1 month after the day of fishing operation. Since 2009 increasing proportion of vessels are using electronic logbooks. Fisheries scientists have direct access to the logbook database.
The Icelandic fishing fleet can be characterised by the most sophisticated technological equipment available in this field. This applies to navigational techniques and fish-detection instruments as well as the development of more effective fishing gear. The most significant development in recent years is the increasing size of pelagic trawls and with increasing engine power the ability to catch pelagic fishes at greater depths than previously possible. There have also been substantial improvements in recent decades with respect to technological aspects of other gears such as bottom trawl, longline and handline. Each fishery uses a variety of gears and some vessels frequently shift from one gear to another within each year. The most common demersal fishing gear are otter trawls, longlines, seines, gillnets and jiggers while the pelagic fisheries use pelagic trawls and purse seines. The total recorded landings of the Icelandic fleets in 2010 amounted to around 1 million tonnes where pelagic fishes amounted to 0.5 million tonnes. Spatial distributions of the catches are shown in figure 7.4.1. Detailed information of landings by species and gear type are given in Table 7.1. Spatial overviews of the removal of the some important species by different gear are given in Figures 7.4.2.-7.4.5.

A simple categorization of boats among the different fisheries types is impossible as many change gear depending on fish availability in relation to season, quota status of the individual companies, fish availability both in nature and on the quota exchange market, market price, etc. E.g. larger trawl vessels may operate both on demersal species using bottom trawls as well as using purse seine and pelagic trawls on pelagic species. Total number of vessels within each fleet category in 2010 is thus limited to the broad categories given below:

| Type | No. vessels 1) | Gear type used |
| :--- | :--- | :--- |
| Trawlers | 57 | Pelagic and bottom trawl |
| Vessels $>100 \mathrm{t}$ | 140 | Purse seine, longline, trawl, gillnet |
| Vessels $<100 \mathrm{t}$ | 621 | Gillnet, longline, danish seine, trawl, jiggers |
| Open boats | 807 | Jiggers, longliners (including recreational fishers) |
| Total | 1625 |  |
| 1)Source: Statistic Iceland $-\mathrm{http}: / /$ www.statice.is/ |  |  |

The demersal fisheries take place all around Iceland including variety of gears and boats of all sizes. The most important fleets targeting them are:

Large and small trawlers using demersal trawl. This fleet is the most important one fishing cod, haddock, saithe, redfish as well as a number of other species. This fleet is operating year around; mostly outside 12 nautical miles from the shore.

Boats (<300 GRT) using gillnet. These boats are mostly targeting cod but haddock and a number of other species are also target. This fleet is mostly operating close to the shore.

Boats using longlines. These boats are both small boats (<10GRT) operating in shallow waters as well as much larger vessels operating in deeper waters. Cod and haddock are the main target species of this fleet but a number other species are also caught, some of them in directed fisheries.

Boats using jiggers. These are small boats ( $<10$ GRT). Cod is the most important target species of this fleet with saithe of secondary importance.
Boats using Danish seine. (20-300 GRT) Cod, haddock and variety of flatfishes, e.g. plaice, dab, lemon sole and witch are the target species of this fleet.

Although different fleets may be targeting the main species the spatial distribution of effort may different. In general it can be observed that the bottom trawl fleet is fishing in deeper waters than the long line fleet (Figures 7.4.6. and 7.4.7).
The pelagic fisheries targeting capelin, herring, blue whiting and mackerel is almost exclusively carried out by larger vessels. The fisheries in Icelandic waters for capelin and herring are carried out using both purse seine and pelagic trawl while that of blue whiting and mackerel is exclusively carried out with pelagic trawl. Additionally a significant part of the pelagic fisheries of the Icelandic fleet is caught outside the Icelandic EEZ, both on the Atlanto-Scandian herring and on blue whiting.

### 7.5 Regulations

The Ministry of Fisheries is responsible for management of the Icelandic fisheries and implementation of the legislation. The Ministry issues regulations for commercial fishing for each fishing year, including an allocation of the TAC for each of the stocks subject to such limitations. Below is a short account of the main feature of the management system.

### 7.5.1 The ITQ system

A system of transferable boat quotas was introduced in 1984. The agreed quotas were based on the Marine Research Institute's TAC recommendations, taking some socioeconomic effects into account. Until 1990, the quota year corresponded to the calendar year but since then the quota, or fishing year, starts on September 1 and ends on August 31 the following year. This was done to meet the needs of the fishing industry. In 1990, an individual transferable quota (ITQ) system was established for the fisheries and they were subject to vessel catch quotas. Since 2006/2007 fishing season, all boats operate under the TAC system.

With some minor exceptions it is required by law to land all catches. Consequently, no minimum landing size is in force. To prevent fishing of small fish various measures such as mesh size regulation and closure of fishing areas are in place (see below).

Within this system individual boat owners have substantial flexibility in exchanging quota, both among vessels within individual company as well as among different companies. The latter can be done via temporary or permanent transfer of quota. In addition, some flexibility is allowed by individual boats with regard to transfer allowable catch of one species to another. These measures, which can be acted on more or less instantaneously, are likely to result in lesser initiative to discards and misreporting than can be expected if individual boats are restricted by strict TAC measures alone. They may however result in fishing pressures of individual species to be different than intended under the single species TAC allocation.

### 7.5.2 Mesh size regulations

With the extension of the fisheries jurisdiction to 200 miles in 1975, Iceland introduced new measures to protect juvenile fish. The mesh size in trawls was increased from 120 mm to 155 mm in 1977. Mesh size of 135 mm was only allowed in the fisheries for redfish in certain areas. Since 1998 a minimum mesh size of 135 is allowed in the
codend in all trawl fisheries not using "Polish cover" and in the Danish seine fisheries. For the gillnet fishery both minimum and maximum mesh-sizes are restricted. Since autumn 2004 the maximum allowed mesh-size in the gillnet fishery is 8 inches. The objective of this measure is to decrease the effort directed towards bigger spawners.

### 7.5.3 Area closures

Real time area closure: A quick closure system has been in force since 1976 with the objective to protect juvenile fish. Fishing is prohibited for at least two weeks in areas where the number of small fish in the catches has been observed by inspectors to exceed certain percentage ( $25 \%$ or more of $<55 \mathrm{~cm}$ cod and saithe, $25 \%$ or more of $<45 \mathrm{~cm}$ haddock and $20 \%$ or more of $<33 \mathrm{~cm}$ redfish). If, in a given area, there are several consecutive quick closures the Minister of Fisheries can with regulations close the area for longer time forcing the fleet to operate in other areas. Inspectors from the Directorate of Fisheries supervise these closures in collaboration with the Marine Research Institute. In 2010, 113 such closures took place:

Permanent area closures: In addition to allocating quotas on each species, there are other measures in place to protect fish stocks. Based on knowledge on the biology of various stocks, many areas have been closed temporarily or permanently aiming at protect juveniles. Figure 7.5 .1 shows map of such legislation that was in force in 2004. Some of them are temporarily, but others have been closed for fishery for decades.

Temporary area closures: The major spawning grounds of cod, plaice and wolfish are closed during the main spawning period of these species. The general objectives of these measures, which were in part initiated by the fishermen, are to reduce fishing during the spawning activity of these species.

### 7.5.4 Discards

Discarding measurements have been carried out in Icelandic fisheries since 2001, based on extensive data collection and length based analysis of the data (Pálsson 2003). The data collection is mainly directed towards main fisheries for cod (Gadus morhua) and haddock (Melanogrammus aeglefinus) and towards saithe (Pollachius virens) and golden redfish (Sebastes marinus) fisheries in demersal trawl and plaice in Danish seine. Sampling for other species is not sufficient to warrant a satisfactory estimation of discarding. The discard rate for cod has been in the range of $0.2-2.2 \%$ of the reported landings over the time investigated (Figure 7.5.2.). The discard estimates for haddock are somewhat higher ranging between $0.7-5 \%$ annually. Discarding of saithe and golden redfish has been negligible over time period of investigation. Estimates of discards of cod and haddock in 2010 by individual fleets are given in table 7.2. These relatively low discard rates compared to what is generally assumed to be a side effect of a TAC system may be a result of the various measures, including the flexibility within the Icelandic ITQ system (see above). Since the time series of discards is relatively short it is not included in the assessments.

All catch that is brought ashore must by law be weighted by a licensed body. The monitoring and enforcement is under the realm of the Directorate of Fisheries. Under the TAC system there are known incentives for misreporting, both with regards to the actual landings statistics as well as with regards to the species recorded. This results in bias in the landings data but detailed quantitative estimates of how large the bias may be, is not available to the NWWG. Unpublished report from the Directorate of Fisheries, partly based on investigation comparing export from fish processing plants with
the amount of fish weighted in the landing process indicate that this bias may be of the order of single digit percentages and not in double digits.

### 7.6 Mixed fisheries, capacity and effort

A number of species caught in Icelandic waters are caught in fisheries targeting only one species, with very little bycatch. These include the pelagic fisheries on herring, capelin and blue whiting (see however below), the Greenland halibut fishery in the west and southeast of Iceland and the S. mentella fishery. Advice given for these stocks should thus not influence the advice of other stocks.

Other fisheries, particularly demersal fisheries may be classified as more mixed, where a target species of e.g. cod, haddock, saithe or S. marinus may be caught in a mixture with other species in the same haul/setting (Figure 7.6.1.). Fishermen can however have a relatively good control of the relative catch composition of the different species. E.g. the saithe fishery along the shelf edge is often in the same areas as the redfish fisheries: Fleets are often targeting at redfish during daytime and saithe during nights. Therefore the fishery for one of those species is relatively free of bycatch of the other species even though they take place in the same area. Small differences in the location of setting are also known to affect the catch composition. This has for example been documented in the long line fisheries in Faxabay, where in adjacent areas cod catches and wolfish catches are known to consistently dominate the catches in individual setting. There are however numerous species in Icelandic waters that can be classified as "bycatch species" in some fisheries. E.g. in the bottom trawl fisheries $75 \%$ of the annual plaice yield is caught in hauls where plaice is minority of the catches. In a proper fisheries based advice taking mixed fisheries issues into account, such stocks may have a greater influence on the advice on the main stocks that are currently assessed by ICES than fisheries linkage among the latter.

In the pelagic fisheries catch other than the targeted species is considered rare. In some cases juveniles of other species are caught in significant numbers. When observers are on board or when fishermen themselves provide voluntary information, the fishing areas have in such cases been closed for fishing, temporarily or permanently. By catch of adults of other species in the blue whiting fishery have been estimated (Pálsson 2005).

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Table 7.1 Overview of the 2010 landings of fish and marine invertebrates caught by the Icelandic fleet categorized by gear types. Based on landing statistics from the Directorate of Fisheries. Landings are given in thous. tonnes.

| Species/gear | Long line | Gillnets Jiggers |  | Danish seine | Bottom trawl | Nephros trawl | Pelagic trawl | Purse seine | Shrimp trawl | edge | Other | otal |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Herring | 0.000 | 0.000 | 0.000 | 0.000 | 0.112 | 0.000 | 213.528 | 40.836 | 0.000 | 0.000 | 0.000 | 254.476 |
| Cod | 57.493 | 16.552 | 3.721 | 8.285 | 82.996 | 1.581 | 0.923 | 0.009 | 1.006 | 0.000 | 0.784 | 173.349 |
| Mackerel | 0.000 | 0.001 | 0.180 | 0.000 | 0.164 | 0.000 | 121.680 | 0.001 | 0.000 | 0.000 | 0.000 | 122.028 |
| Capelin | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 3.187 | 112.328 | 0.000 | 0.000 | 0.000 | 115.515 |
| Blue whiting | 0.000 | 0.000 | 0.000 | 0.000 | 0.124 | 0.000 | 87.784 | 0.000 | 0.000 | 0.000 | 0.000 | 87.908 |
| Haddock | 23.916 | 0.380 | 0.012 | 10.137 | 29.481 | 0.212 | 0.630 | 0.000 | 0.041 | 0.000 | 0.028 | 64.836 |
| Saithe | 0.594 | 4.453 | 2.383 | 1.093 | 42.441 | 0.404 | 1.216 | 0.000 | 0.007 | 0.000 | 0.068 | 52.660 |
| Golden redfish | 1.080 | 0.194 | 0.058 | 0.513 | 35.777 | 0.932 | 0.594 | 0.000 | 0.014 | 0.000 | 0.014 | 39.176 |
| Pearlside | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 17.912 | 0.000 | 0.000 | 0.000 | 0.000 | 17.912 |
| Atlantic argentine | 0.000 | 0.000 | 0.000 | 0.000 | 16.321 | 0.001 | 0.256 | 0.000 | 0.000 | 0.000 | 0.000 | 16.579 |
| Golden redfish | 0.000 | 0.000 | 0.000 | 0.000 | 1.921 | 0.000 | 12.872 | 0.000 | 0.000 | 0.000 | 0.000 | 14.794 |
| Deepwater redfish | 0.052 | 0.002 | 0.000 | 0.000 | 14.149 | 0.000 | 0.181 | 0.000 | 0.000 | 0.000 | 0.000 | 14.384 |
| Greenland halibut | 0.033 | 0.000 | 0.000 | 0.000 | 12.147 | 0.000 | 0.263 | 0.000 | 0.861 | 0.000 | 0.001 | 13.305 |
| Atlantic catfish | 6.915 | 0.020 | 0.002 | 1.032 | 4.490 | 0.083 | 0.033 | 0.000 | 0.000 | 0.000 | 0.027 | 12.602 |
| Ling | 6.529 | 0.363 | 0.011 | 0.404 | 1.538 | 0.981 | 0.011 | 0.000 | 0.000 | 0.000 | 0.028 | 9.865 |
| Shrimp | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.155 | 0.000 | 7.607 | 0.000 | 0.000 | 7.762 |
| Tusk | 6.760 | 0.052 | 0.003 | 0.000 | 0.093 | 0.005 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 6.915 |
| Blue Lling | 3.978 | 0.091 | 0.000 | 0.092 | 1.901 | 0.283 | 0.013 | 0.000 | 0.002 | 0.000 | 0.015 | 6.375 |
| Plaice | 0.105 | 0.118 | 0.006 | 3.640 | 2.020 | 0.003 | 0.015 | 0.000 | 0.001 | 0.000 | 0.077 | 5.984 |
| Monkfish | 0.079 | 0.176 | 0.001 | 0.430 | 0.452 | 0.556 | 0.000 | 0.000 | 0.001 | 0.000 | 1.586 | 3.281 |
| Whiting | 0.425 | 0.030 | 0.002 | 0.191 | 2.037 | 0.155 | 0.000 | 0.000 | 0.001 | 0.000 | 0.001 | 2.842 |
| Redfish | 0.001 | 0.000 | 0.000 | 0.000 | 2.446 | 0.000 | 0.154 | 0.000 | 0.000 | 0.000 | 0.000 | 2.601 |
| Nephrops | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 2.541 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 2.541 |
| Sea cucumber | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 2.246 | 0.000 | 2.246 |
| Lumpfish roe | 0.000 | 0.000 | 0.002 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 2.133 | 2.135 |
| Lemon sole | 0.000 | 0.002 | 0.001 | 0.992 | 0.886 | 0.078 | 0.007 | 0.000 | 0.000 | 0.000 | 0.001 | 1.968 |
| Leopardfish | 1.045 | 0.003 | 0.000 | 0.004 | 0.805 | 0.002 | 0.022 | 0.000 | 0.037 | 0.000 | 0.003 | 1.922 |
| Witch | 0.000 | 0.000 | 0.000 | 0.733 | 0.075 | 0.514 | 0.000 | 0.000 | 0.000 | 0.000 | 0.002 | 1.325 |
| Starry ray | 0.776 | 0.005 | 0.000 | 0.188 | 0.057 | 0.001 | 0.000 | 0.000 | 0.001 | 0.000 | 0.001 | 1.029 |
| Common dab | 0.007 | 0.002 | 0.004 | 0.574 | 0.025 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.612 |
| Halibut | 0.377 | 0.004 | 0.000 | 0.034 | 0.114 | 0.014 | 0.001 | 0.000 | 0.000 | 0.000 | 0.008 | 0.552 |
| Lumpfish | 0.000 | 0.017 | 0.001 | 0.002 | 0.002 | 0.000 | 0.037 | 0.000 | 0.000 | 0.000 | 0.333 | 0.391 |
| Megrim | 0.000 | 0.000 | 0.000 | 0.089 | 0.052 | 0.111 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.252 |
| Long rough dab | 0.009 | 0.004 | 0.000 | 0.173 | 0.031 | 0.000 | 0.000 | 0.000 | 0.001 | 0.000 | 0.000 | 0.220 |
| Sea-urchins | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.146 | 0.000 | 0.146 |
| European whelk | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.142 | 0.142 |
| Skate | 0.042 | 0.007 | 0.000 | 0.026 | 0.024 | 0.008 | 0.000 | 0.000 | 0.000 | 0.000 | 0.009 | 0.117 |
| Black scabbard-fish | 0.002 | 0.000 | 0.000 | 0.000 | 0.107 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.109 |
| Boston hake | 0.109 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.109 |
| Blue mussel | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.077 | 0.000 | 0.077 |
| Dogfish | 0.011 | 0.039 | 0.000 | 0.004 | 0.006 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.002 | 0.062 |
| Rat-tail | 0.000 | 0.000 | 0.000 | 0.000 | 0.058 | 0.000 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.059 |
| Squid | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.051 | 0.000 | 0.000 | 0.000 | 0.000 | 0.051 |
| Greenland shark | 0.000 | 0.000 | 0.000 | 0.000 | 0.043 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.043 |
| Norway pout | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.039 | 0.000 | 0.000 | 0.000 | 0.000 | 0.039 |
| onioin eye | 0.000 | 0.000 | 0.000 | 0.000 | 0.023 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.023 |
| Fuller's ray | 0.018 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.019 |
| Arctiv wolffish | 0.000 | 0.000 | 0.000 | 0.000 | 0.017 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.017 |
| sailray | 0.012 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.012 |
| Deal fish | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.011 | 0.000 | 0.000 | 0.000 | 0.000 | 0.011 |
| Gurnard | 0.000 | 0.000 | 0.000 | 0.001 | 0.000 | 0.000 | 0.010 | 0.000 | 0.000 | 0.000 | 0.000 | 0.010 |
| Black dogfish | 0.001 | 0.000 | 0.000 | 0.000 | 0.009 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.010 |
| Total | 110.370 | 22.520 | 6.386 | 28.638 | 252.947 | 8.466 | 461.586 | 153.175 | 9.579 | 2.470 | 5.263 | 1,061.401 |

Table 7.2. Estimates of discard of cod and haddock in the Icelandic fisheries in 2008. Source: Ólafur K. Pálsson, Höskuldur Björnsson, Eypór Björnsson, Guð̈mundur Jóhannesson og Pórhallur Ottesen 2009. Discards in demersal Icelandic fisheries 2009. Marine Research Institute, 2009, report series no. 154 .

|  | Gear | Landings (tonnes) | Discards |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Numbers (thous.) | Weight (tonnes) | \% Weight |
| COD | Longline Gillnet Danish Seine Bottom trawl Total | $\begin{array}{r} \hline 61008 \\ 21859 \\ 10369 \\ 77172 \\ \mathbf{1 7 0 4 0 8} \\ \hline \end{array}$ |  | 308 0 18 635 961 | $\begin{aligned} & \hline 0.51 \\ & 0.00 \\ & 0.18 \\ & 0.82 \\ & 0.56 \end{aligned}$ |
| HADDOCK | Longline <br> Danish Seine <br> Bottom trawl <br> Total | $\begin{array}{r} \hline 26573 \\ 15126 \\ 38822 \\ \mathbf{8 0 8 5 2 1} \end{array}$ | $\begin{array}{r} 155 \\ 36 \\ 1042 \\ 1233 \end{array}$ | 79 9 465 553 | $\begin{aligned} & \hline 0.30 \\ & 0.06 \\ & 1.20 \\ & 0.69 \end{aligned}$ |



Figure 7.3.1. Temperature and salinity in winter west of Iceland 1971-2011. Mean 0-200m


Figure 7.3.2. Temperature and salinity off central North-Iceland 1974-2011.


Figure 7.3.3. Distribution and migrations of capelin in the Iceland/East-Greenland/Jan Mayen area before 2001. Red: Spawning grounds; Green: Adult feeding area; Blue: Distribution and feeding area of juveniles; Green arrows: Adult feeding migrations; Blue arrows: Return migrations; Red arrows: Spawning migrations; Depth contours are 200, 500 and 1000 m (Vilhjalmsson 2002)


Figure 7.3.4. Likely changes of distribution and migration routes of capelin in the Iceland/Greenland/Jan Mayen area in the last 3-4 years. Green: Feeding area; Light blue: Juvenile area; Red area: Main spawning grounds; Lighter red colour: Lesser importance of W-Iceland spawning areas; Light blue arrows: Larval drift; Dark green arrows: Feeding migrations; Dark blue arrows: Return migrations; Red arrows: Spawning migrations. Depth contours are 200, 500 and 1000 m .


Figure 7.4.1. Distribution of total catch of all species by the Icelandic fishing fleet in Icelandic EEZ and adjacent waters in 2012. The EEZs are shown as white lines.


Figure 7.4.2. Location of catches of cod, saithe, haddock, redfish, Greenland halibut and others caught with bottom trawl in 2012.


Figure 7.4.3. Location of catches of cod, saithe, haddock, redfish, Greenland halibut and others caught with long-line in 2012.


Figure 7.4.4. Location of catches of cod, saithe, haddock, redfish, Greenland halibut and others caught with gillnets in 2012.


Figure 7.4.5. Location of catches of capelin, Icelandic summer spawning herring, blue whiting and mackerel with purse seine and pelagic trawls in 2012.


Figure 7.4.6 Spatial distribution of the trawler fleet effort (in hours trawled) in 2000-2012 and as a time-series.


Figure 7.4.7. Spatial distribution of the longlinefleet effort (in number of hooks) in 2000-2012. The main targeted species for longline fishing are cod, haddock, catfish, tusk, ling and blue ling.


Figure 7.5.1. Overview of closed areas around Iceland in 2006 . The boxes are of different nature and can be closed for different time period and gear type.


Figure 7.5.2. Estimates of discard percentage by weight for cod and haddock. Source: Ólafur K. Pálsson, Höskuldur Björnsson, Eypór Björnsson, Guðmundur Jóhannesson , og Pórhallur Ottesen 2009. Discards in demersal Icelandic fisheries 2009. Marine Research Institute, report series Nr. 154. 2010 figures are preliminary .


Figure 7.6.1. Cumulative plot for bottom trawl in 2008. An example describes this probably best. Looking at the figure above it can be seen from the dashed lines that $30 \%$ of the catch of haddock comes from hauls where haddock is less than $60 \%$ of the total catch while only $4 \%$ of the catch of greenland halibut comes from hauls where it is less than $50 \%$ of the total catch. $75 \%$ of the plaice is on the other hand caught in hauls where plaice is minority of the catches. The figures also shows that $70 \%$ of the catch of greenland halibut comes from hauls where nothing else is caught but only $10 \%$ of the haddock. Of the species shown in the figure plaice is the one with largest proportion as bycatch while greenland halibut is the one with largest proportion caught in mixed fisheries.

## Summary

The 2015 reference biomass ( $B_{4+}$ ) is estimated as 255 kt , around the average in the assessment period (1980 to the present). Spawning biomass is estimated as 139 kt , above the average in the assessment period and well above $B_{\text {trigger }}=65 \mathrm{kt}$ and $B_{\lim }=61 \mathrm{kt}$.

Harvest rate has been around the HCR target of $20 \%$ since 2011, with fishing mortality rate between 0.19 and 0.25 . Year classes 2008 and 2009 are above average, but recruitment has declined below average since then.

Weights of ages 3-6 have been low in recent years, but older ages are close to average weight. Maturity at ages 4-9 has decreased in recent years and is currently around average.

The assessment model is a separable statistical catch-at-age model implemented in AD Model Builder. Selectivity is age-specific and varies between three periods: 1980-1996, 1997-2003, and 2004 onwards.

The default separable model (ADSEP) estimates a slightly larger stock size than alternative diagnostic models (ADAPT, TSA, SAM). The estimates of this year's $B_{4+}$ range from 209 (TSA) to 255 kt (ADSEP).

In 2013, the Icelandic government adopted a harvest control rule for managing the Icelandic saithe fishery, evaluated by ICES (2013). It is similar to the $20 \%$ rule used for the Icelandic cod fishery. When the population is above $B_{\text {trigger, }}$ the TAC set in year $t$ equals the average of $0.2 B_{4+}$ in year $t$ and last year's TAC.

According to the adopted harvest control rule, the TAC will be 55 kt in the next fishing year.

### 8.1 Stock description and management units

Description of the stock and management units is provided in the stock annex.

### 8.2 Fisheries-dependent data

### 8.2.1 Landings, advice and TAC

Landings of saithe in Icelandic waters in 2014 are estimated to have been 46500 t (Table 8.1 and Figure 8.1). Of the landings, 38600 t were caught by trawl, 2400 t by gillnets, and the rest caught by other fishing gear. The domestic as well as ICES advice for the fishing year 2014/2015 was based on the $20 \%$ harvest control rule and was 58 kt . The TAC issued was also 58 kt . The trajectory of the landings in the current fishing year and calendar year is shown in Figure 8.2.
Most of the catch is caught in bottom trawl ( $80 \%$ in 2010-2014), with gillnet and jiggers taking the majority of the rest. The share taken by the gillnet fleet was larger in the past, $26 \%$ in 1982-1996 compared to $9 \%$ in 1997-2014 (Figure 8.1).

### 8.2.2 Landings by age

Catch in numbers by age based on landings are listed in Table 8.2. Discarding is not considered to be a problem in the Icelandic saithe fisheries, with an estimated discard proportion of $0.1 \%$ (annual reports by Palsson et al. 2003 and later). Comparison of sea
and harbour samples indicate that discards have been small in most years since 2000. The sea samples constitute about $60-70 \%$ of the length samples used in the calculation of the catch in number. Since the amount of discards is likely to be small, not taking discards into account in the total catches and catch in numbers is not considered to have major effect on the stock assessment.

The sampling program was slightly revised in 2013 and 2014, but the approach used for calculating catch in numbers has not changed. In 2013, the sampling frequency was reduced for bottom trawl, while the sampling frequency was increased for gillnets, jiggers, and demersal seine in 2014. Also in 2014, the number of otoliths from each sample was halved from 50 to 25 for all fishing gears. These revisions in the sampling program were based on the analysis of Thordarson (2012). The age and length sampling in 2014 is indicated in the following table:

|  |  | No. of <br> otolith <br> (t) | No. of <br> samples | No. of <br> length <br> otoliths read | No. of length <br> measurements |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Gillnets | 2355 | 9 | 250 | 10 | 1036 |
| Jiggers | 2115 | 14 | 370 | 15 | 1601 |
| Demersal seine | 1005 | 4 | 150 | 4 | 471 |
| Bottom trawl | 38634 | 52 | 1625 | 224 | 32251 |
| Other gear | 1624 | - | - | 189 | 2354 |
| Foreign landings | 750 | - | - | - | - |
| Total | 46483 | 79 | 2395 | 442 | 37713 |

Two age-length keys are used to calculate catch at age, one key for the gillnet catch and another key for other gears combined. The same length-weight relationship ( $W=0.02498^{*} L^{\wedge} 2.75674$ ) is applied to length distributions from both fleets.

### 8.2.3 Mean weight and maturity at age

Weights of ages 3-6 have been low in recent years, but older ages are close to average weight (Table 8.3 and Figure 8.3). The long-term trend since 1980 has been a gradual decline in the weight of all ages. Weight at age in the landings is also used as weight at age in the stock. Weights for the current calendar year are predicted by applying a linear model using survey weights and the weight of a year class in the previous year as predictors (Magnusson 2012).

Maturity at ages 4-9 has decreased in recent years and is currently around average (Table 8.4 and Figure 8.4). A model using maturity at age from the Icelandic groundfish spring survey is used to derive smoothed trends in maturity by age and year (see stock annex).

### 8.2.4 Logbook data

Commercial CPUE indices are not used for tuning in this assessment. Although these indices have been explored for inclusion in the past, they were not considered for inclusion in the benchmark (ICES 2010), as the trends in CPUE are considered unreliable as an indicator of changes in abundance.

### 8.3 Scientific surveys

In the benchmark, spring survey data were considered superior to the autumn survey for calibrating the assessment. Saithe is among the most difficult demersal fishes to get
reliable information on from bottom trawl surveys. In the spring survey, which has 500-600 stations, a large proportion of the saithe is caught in relatively few hauls and there seems to be considerable inter-annual variability in the number of these hauls.
The survey biomass indices fluctuated greatly in 1985-1995, but were consistently low in 1995-2001, high in the period around 2005, declining to a relatively low level in 20072011. The 2012 and 2013 survey biomass indices were relatively high (Table 8.5 and Figure 8.5).
Internal consistency in the surveys measured by the correlation of the indices for the same year class in 2 adjacent surveys is poor, with $R^{2}$ close to 0.3 for the best-defined age groups, and much lower for some other.

Young saithe tend to live very close to shore, so it is not surprising that survey indices for ages 1 and 2 are poor measures of recruitment, and the number of young saithe caught in the survey is very low.

### 8.4 Assessment method

In accordance with the recommendation from the benchmark (ICES 2010), a separable forward-projecting statistical catch-age model, developed in AD Model Builder, is used to fit commercial catch at age (ages 3-14 from 1980 onwards) and survey catch at age (ages 2-10 from 1985 onwards). The selectivity pattern is constant within each period (Figure 8.6). Natural mortality is set at 0.2 for all ages.

The commercial catch-at-age residuals (Table 8.6 and Figure 8.7) are relatively small in recent years, owing to the model flexibility provided by the two recent selectivity periods 1997-2003 and 2004 onwards. The survey catch-at-age residuals (Table 8.7 and Figure 8.7) have year blocks with all residuals being only negative or only positive in some years. The survey residuals are modelled as multivariate normal distribution with the correlation estimated (one coefficient).

### 8.5 Reference points and HCR

In April 2013, the Icelandic government adopted a management plan for managing the Icelandic saithe fishery (Ministry of Industries and Innovation 2013). ICES evaluated this management plan and concluded that it was in accordance with the precautionary approach and the ICES MSY framework. In the harvest control rule (HCR) evaluation (ICES 2013) $B_{\text {lim }}$ was defined as 61 kt , based on $B_{\text {loss }}$ as estimated in 2010, and $B_{\text {trigger }}$ was defined as 65 kt , based on an estimated hockey-stick recruitment function.
The TAC set in year $t$ is for the upcoming fishing year, from 1 September in year $t$, to 31 August in year $t+1$. The $20 \%$ HCR consists of two equations, as follows.
When $S S B \geq B_{\text {trigger, }}$ the TAC set in year $t$ equals the average of 0.20 times the current biomass and last year's TAC:

$$
\begin{equation*}
T A C_{t}=0.5 \times 0.20 B_{t, 4+}+0.5 T A C_{t-1} \tag{Eq.1}
\end{equation*}
$$

When SSB is below $B_{\text {trigger, }}$ the harvest rate is reduced below 0.20 :

$$
\begin{equation*}
\left.T A C_{t}=S S B_{t} / B_{\text {trigger }}\left[\left(1-0.5 S S B_{t} / B_{\text {trigger }}\right) 0.20 B_{t, 4+}\right)+0.5 T A C_{t-1}\right] \tag{Eq.2}
\end{equation*}
$$

Equation 1 is a plain average of two numbers. Equation 2 is continuous over $S S B_{t} / B_{\text {trigger, }}$, so the rule does not lead to very different TAC when $S S B_{t}$ is slightly below or above $B_{\text {trigger }}$ (Magnusson 2013).

### 8.6 State of the stock

The results of the principal stock quantities (Table 8.8 and Figure 8.8) show that the reference biomass has historically ranged from 410 to 130 kt (in 1988 and 1999), but this range has been narrower since 2003, between 220 and 320 kt . The current stock size of 255 kt is around the average in the assessment period ( 1980 to the present). Spawning biomass is estimated as 139 kt , above the average in the assessment period and well above $B_{\text {trigger }}$ and $B$ lim.
The harvest rate peaked around $30 \%$ in the mid 1990s, but has fluctuated around the HCR target of $20 \%$ since 2011, with fishing mortality rate between 0.19 and 0.25 . SSB has been stable at a relatively high level during the last ten years, having declined to its historical minimum in the mid 1990s.

Year classes 2008 and 2009 are above average, but recruitment has declined below average since then. The details of the fishing mortality and stock in numbers are presented in Tables 8.9 and 8.10.

### 8.7 Short-term forecast

The input for the short-term forecast is shown in Table 8.11. Future weights, maturity, and selectivity are assumed to be the same as in the assessment year, as described in the stock annex. Recruitment predictions are based on the segmented stock-recruitment function estimated in the assessment model.

The landings for the ongoing calendar year are predicted based on the $20 \% \mathrm{HCR}$, with the calendar year landings consisting of $2 / 3$ of the ongoing fishing year's TAC and $1 / 3$ of the next fishing year's TAC.
Following the HCR, the predicted landings in 2016 are 54 kt and the resulting SSB in 2017 is predicted to be 130 kt .

### 8.8 Uncertainties in assessment and forecast

The assessment of Icelandic saithe is relatively uncertain due to fluctuations in the survey data, as well as irregular changes in the fleet selectivity. The internal consistency in the spring bottom trawl survey is very low for saithe. This is not surprising, considering the nature of the species that is partly pelagic, schooling, and relatively widely migrating. There are also indications of time-varying selectivity, so changes in the commercial catch at age may not reflect changes in the age dstribution of the population. The retrospective pattern (Figure 8.9) reveals some of the assessment uncertainty. The harvest control rule evaluation incorporated uncertainties about assessment estimates, among other sources of uncertainty (ICES 2013).

The results from the default separable assessment model (ADSEP) are compared to alternative diagnostic model runs, involving ADAPT, TSA, and SAM, in order to explore the overall uncertainty in the assessment. The comparison involved four models which differ mainly in the way the commercial catch-at-age variability and F-matrix is modelled:

|  | Model | Family | CA variability | F matrix |
| :--- | :--- | :--- | :--- | :--- |
| 1 | ADSEP <br> (default) | separable | observation <br> error | multiplicative <br> in 3 periods |
| 2 | ADAPT | vpa | process error | no constraints |
| 3 | TSA | state-space <br> (kalman filter) | observation \& process <br> error | orthogonal <br> polynomials |


| 4 | SAM | state-space <br> (random effects) | observation \& process <br> error | correlated <br> random walk |
| :--- | :--- | :--- | :--- | :--- |

The results from the model comparison (Figure 8.10) show that the default model estimates a slightly larger stock size than the other models, which has also been the case for saithe assessments in recent years. The estimates of this year's B4+ range from 209 (TSA) to 255 kt (ADSEP).

### 8.9 Comparison with previous assessment and forecast

Compared to last year's assessment the estimated reference biomass B4+ in 2014 has decreased from 296 to 265 kt , SSB 2014 has decreased from 150 to 132 kt , and the harvest rate $\mathrm{u}_{2013}$ has increased from $19 \%$ to $22 \%$ (fishing mortality 0.22 to 0.25 ). Stock numbers at age 5 have increased slightly, while stock numbers at ages 6 and 7 have decreased as shown below.

|  | NWWG 2014 | NWWG 2015 |
| :--- | :--- | :--- |
| B4+(2014) | 296 | 265 |
| SSB(2014) | 150 | 132 |
| $\mathbf{u}(2013)$ | $19 \%$ | $22 \%$ |
| F4-9(2013) | 0.22 | 0.25 |
| N5(2014) | 24 | 26 |
| N6(2014) | 21 | 17 |
| N7(2014) | 11 | 9 |

### 8.10 Ecosystem considerations

Changes in the distribution of large pelagic stocks (blue whiting, mackerel, Norwegian spring-spawning herring, Icelandic summer-spawning herring) may affect the propensity of saithe to migrate off shelf and between management units. Saithe is a migrating species and makes both vertical and long-distance feeding and spawning migrations (Armannsson et al. 2007, Armannsson and Jonsson 2012, i Homrum et al. 2013). The evidence from tagging experiments (ICES 2008) show some migrations along the Faroe-Iceland Ridge, as well as onto the East Greenland shelf.

### 8.1 1 Changes in fishing technology and fishing patterns

According to the stock assessment model fit to the commercial catch-at-age data, the fleet is targeting younger fish since around 2004, compared to earlier years. This can be partly explained by reduced use of gillnets in the saithe fishery.

### 8.12 References

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Table 8.1. Saithe in division Va. Nominal catch ( $t$ ) by countries, as officially reported to ICES.

|  |  |  |  |  |  | UK | UK |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | Belgium | Faroes | France | Germany | Iceland | Norway | (E/W/NI) | (Scot) | UK | Total |
| :--- |
| 1980 |
| 980 |


|  |  |  |  |  | UK |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | Belgium | Faroes | France | Germany | Iceland | Norway | (E/W/NI) <br> (Scot) | UK | Total

Table 8.2. Saithe in division Va. Commercial catch at age (millions).

|  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 198 | 0.27 | 2.540 | 5.21 | 2.596 | 2.169 | 1.34 | 0.38 | 0.26 | 0.15 | 0.11 | 0.06 | 0.03 |
| 0 | 5 |  | 4 |  |  | 1 | 7 | 2 | 5 | 2 | 4 | 3 |
| 198 | 0.20 | 1.325 | 3.50 | 5.404 | 1.457 | 1.41 | 0.57 | 0.24 | 0.06 | 0.15 | 0.13 | 0.12 |
| 1 | 3 |  | 3 |  |  | 5 | 8 | 2 | 1 | 4 | 5 | 8 |
| 198 | 0.50 | 1.092 | 2.80 | 4.845 | 4.293 | 1.21 | 0.97 | 0.30 | 0.05 | 0.03 | 0.04 | 0.04 |
| 2 | 8 |  | 4 |  |  | 5 | 5 | 6 | 9 | 5 | 8 | 6 |
| 198 | 0.10 | 1.750 | 1.06 | 2.455 | 4.454 | 2.31 | 0.50 | 0.25 | 0.03 | 0.01 | 0.00 | 0.00 |
| 3 | 7 |  | 5 |  |  | 1 | 1 | 1 | 8 | 2 | 2 | 4 |
| 198 | 0.05 | 0.657 | 0.80 | 1.825 | 2.184 | 3.61 | 0.84 | 0.37 | 0.29 | 0.13 | 0.18 | 0.22 |
| 4 | 3 |  | 0 |  |  | 0 | 4 | 6 | 1 | 5 | 5 | 6 |
| 198 | 0.37 | 4.014 | 3.36 | 1.958 | 1.536 | 1.17 | 0.74 | 0.47 | 0.07 | 0.02 | 0.07 | 0.07 |
| 5 | 6 |  | 6 |  |  | 2 | 7 | 9 | 4 | 3 | 2 | 1 |
| 198 | 3.10 | 1.400 | 4.17 | 2.665 | 1.550 | 1.11 | 0.62 | 1.54 | 0.21 | 0.05 | 0.03 | 0.01 |
| 6 | 8 |  | 0 |  |  | 6 | 8 | 9 | 6 | 1 | 0 | 4 |
| 198 | 0.95 | 5.135 | 4.42 | 5.409 | 2.915 | 1.34 | 0.66 | 0.49 | 0.49 | 0.05 | 0.02 | 0.04 |
| 7 | 6 |  | 8 |  |  | 8 | 1 | 6 | 8 | 8 | 7 | 8 |
| 198 | 1.31 | 5.067 | 6.61 | 3.678 | 2.859 | 1.77 | 0.84 | 0.22 | 0.27 | 0.10 | 0.02 | 0.00 |
| 8 | 8 |  | 9 |  |  | 5 | 5 | 6 | 0 | 7 | 4 | 1 |
| 198 | 0.31 | 4.313 | 8.47 | 7.309 | 1.794 | 1.92 | 0.84 | 0.27 | 0.19 | 0.13 | 0.07 | 0.01 |
| 9 | 5 |  | 1 |  |  | 8 | 8 | 0 | 1 | 5 | 6 | 0 |
| 199 | 0.14 | 1.692 | 5.47 | 10.11 | 6.174 | 1.81 | 1.08 | 0.38 | 0.15 | 0.05 | 0.07 | 0.03 |
| 0 | 3 |  | 1 | 2 |  | 6 | 7 | 0 | 1 | 5 | 6 | 7 |
| 199 | 0.19 | 0.874 | 3.61 | 6.844 | 10.77 | 3.22 | 0.85 | 0.83 | 0.22 | 0.04 | 0.00 | 0.00 |
| 1 | 8 |  | 3 |  | 2 | 3 | 8 | 8 | 8 | 0 | 6 | 5 |
| 199 | 0.24 | 2.928 | 3.84 | 4.355 | 3.884 | 4.04 | 1.29 | 0.35 | 0.19 | 0.05 | 0.05 | 0.01 |
| 2 | 2 |  | 4 |  |  | 6 | 0 | 0 | 6 | 6 | 4 | 5 |
| 199 | 0.65 | 1.083 | 2.84 | 2.252 | 2.247 | 2.31 | 3.67 | 0.83 | 0.22 | 0.18 | 0.08 | 0.01 |
| 3 | 7 |  | 1 |  |  | 4 | 1 | 0 | 3 | 8 | 1 | 2 |
| 199 | 0.70 | 2.955 | 1.77 | 2.603 | 1.377 | 1.24 | 1.26 | 2.00 | 0.45 | 0.15 | 0.18 | 0.08 |
| 4 | 2 |  | 0 |  |  | 3 | 3 | 9 | 4 | 8 | 8 | 2 |
| 199 | 1.57 | 1.853 | 2.66 | 1.807 | 2.370 | 0.90 | 0.57 | 0.48 | 0.52 | 0.10 | 0.03 | 0.01 |
| 5 | 3 |  | 1 |  |  | 5 | 4 | 2 | 1 | 6 | 5 | 3 |
| 199 | 1.10 | 2.608 | 1.86 | 1.649 | 0.835 | 1.23 | 0.38 | 0.26 | 0.21 | 0.23 | 0.14 | 0.07 |
| 6 | 2 |  | 8 |  |  | 3 | 5 | 7 | 0 | 2 | 1 | 4 |
| 199 | 0.60 | 2.960 | 2.76 | 1.651 | 1.178 | 0.59 | 0.45 | 0.12 | 0.09 | 0.11 | 0.07 | 0.04 |
| 7 | 3 |  | 6 |  |  | 9 | 4 | 5 | 5 | 4 | 7 | 3 |
| 199 | 0.18 | 1.289 | 1.76 | 1.545 | 1.114 | 0.65 | 0.35 | 0.26 | 0.12 | 0.08 | 0.08 | 0.08 |
| 8 | 3 |  | 7 |  |  | 8 | 1 | 5 | 0 | 1 | 5 | 5 |
| 199 | 0.98 | 0.732 | 1.56 | 2.176 | 1.934 | 0.66 | 0.32 | 0.14 | 0.07 | 0.02 | 0.02 | 0.02 |
| 9 | 9 |  | 4 |  |  | 9 | 4 | 0 | 2 | 5 | 8 | 2 |
| 200 | 0.85 | 2.383 | 0.89 | 1.511 | 1.612 | 1.80 | 0.33 | 0.17 | 0.05 | 0.03 | 0.01 | 0.00 |
| 0 | 0 |  | 6 |  |  | 6 | 5 | 3 | 7 | 3 | 7 | 7 |
| 200 | 1.22 | 2.619 | 2.18 | 0.591 | 0.977 | 0.94 | 0.81 | 0.18 | 0.09 | 0.02 | 0.02 | 0.01 |
| 1 | 3 |  | 4 |  |  | 3 | 9 | 6 | 4 | 8 | 8 | 3 |
| 200 | 1.18 | 4.190 | 3.14 | 2.970 | 0.519 | 0.82 | 0.57 | 0.30 | 0.10 | 0.02 | 0.01 | 0.01 |
| 2 | 7 |  | 7 |  |  | 0 | 0 | 9 | 1 | 7 | 5 | 1 |
| 200 | 2.28 | 4.363 | 4.03 | 2.472 | 1.942 | 0.28 | 0.43 | 0.28 | 0.19 | 0.02 | 0.02 | 0.01 |
| 3 | 4 |  | 1 |  |  | 5 | 8 | 9 | 6 | 8 | 9 | 5 |


| 200 | 0.95 | 7.841 | 7.19 | 5.363 | 1.563 | 1.05 | 0.21 | 0.22 | 0.15 | 0.07 | 0.03 | 0.01 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 4 | 2 |  | 5 |  |  | 7 | 1 | 4 | 7 | 4 | 9 | 1 |
| 200 | 2.60 | 3.089 | 7.33 | 6.876 | 3.592 | 0.97 | 0.64 | 0.11 | 0.14 | 0.08 | 0.04 | 0.01 |
| 5 | 7 |  | 3 |  |  | 8 | 2 | 9 | 9 | 9 | 6 | 2 |
| 200 | 1.38 | 10.05 | 2.61 | 5.840 | 4.514 | 1.98 | 0.66 | 0.48 | 0.11 | 0.11 | 0.08 | 0.03 |
| 6 | 0 | 1 | 6 |  |  | 9 | 7 | 5 | 8 | 2 | 6 | 1 |
| 200 | 1.24 | 6.552 | 8.75 | 2.124 | 2.935 | 1.81 | 0.96 | 0.39 | 0.19 | 0.04 | 0.03 | 0.02 |
| 7 | 4 |  | 1 |  |  | 7 | 4 | 5 | 0 | 3 | 6 | 0 |
| 200 | 1.43 | 3.602 | 5.87 | 6.706 | 1.155 | 1.89 | 1.24 | 0.80 | 0.26 | 0.17 | 0.08 | 0.04 |
| 8 | 2 |  | 4 |  |  | 4 | 8 | 3 | 2 | 6 | 7 | 4 |
| 200 | 2.82 | 5.166 | 2.08 | 2.734 | 2.883 | 0.77 | 1.10 | 0.84 | 0.55 | 0.20 | 0.13 | 0.03 |
| 9 | 0 |  | 4 |  |  | 7 | 1 | 7 | 5 | 3 | 4 | 6 |
| 201 | 2.14 | 6.284 | 3.05 | 0.997 | 1.644 | 1.57 | 0.51 | 0.65 | 0.52 | 0.23 | 0.11 | 0.06 |
| 0 | 6 |  | 8 |  |  | 1 | 4 | 6 | 2 | 1 | 4 | 4 |
| 201 | 2.00 | 4.850 | 4.00 | 1.502 | 0.677 | 1.06 | 1.14 | 0.32 | 0.43 | 0.24 | 0.15 | 0.07 |
| 1 | 4 |  | 6 |  |  | 5 | 5 | 3 | 3 | 4 | 0 | 5 |
| 201 | 1.18 | 4.816 | 3.51 | 2.417 | 0.903 | 0.43 | 0.88 | 1.01 | 0.35 | 0.27 | 0.17 | 0.09 |
| 2 | 3 |  | 4 |  |  | 2 | 3 | 5 | 4 | 7 | 3 | 9 |
| 201 | 1.16 | 5.538 | 6.36 | 2.963 | 1.610 | 0.66 | 0.37 | 0.53 | 0.46 | 0.12 | 0.11 | 0.07 |
| 3 | 3 |  | 6 |  |  | 4 | 5 | 7 | 0 | 4 | 8 | 8 |
| 201 | 0.66 | 3.499 | 4.86 | 2.805 | 1.276 | 0.72 | 0.34 | 0.24 | 0.31 | 0.19 | 0.12 | 0.07 |
| 4 | 8 |  | 7 |  |  | 5 | 7 | 1 | 2 | 9 | 8 | 4 |

Table 8.3. Saithe in division Va. Mean weight at age (g) in the catches and in the spawning stock, with predictions in gray.

|  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 1428 | 1983 | 2667 | 3689 | 5409 | 6321 | 7213 | 8565 | 9147 | 9617 | 10066 | 11041 |
| 1981 | 1585 | 2037 | 2696 | 3525 | 4541 | 6247 | 6991 | 8202 | 9537 | 9089 | 9351 | 10225 |
| 1982 | 1547 | 2194 | 3015 | 3183 | 5114 | 6202 | 7256 | 7922 | 8924 | 10134 | 9447 | 10535 |
| 1983 | 1530 | 2221 | 3171 | 4270 | 4107 | 5984 | 7565 | 8673 | 8801 | 9039 | 11138 | 9818 |
| 1984 | 1653 | 2432 | 3330 | 4681 | 5466 | 4973 | 7407 | 8179 | 8770 | 8831 | 11010 | 11127 |
| 1985 | 1609 | 2172 | 3169 | 3922 | 4697 | 6411 | 6492 | 8346 | 9401 | 10335 | 11027 | 10644 |
| 1986 | 1450 | 2190 | 2959 | 4402 | 5488 | 6406 | 7570 | 6487 | 9616 | 10462 | 11747 | 11902 |
| 1987 | 1516 | 1715 | 2670 | 3839 | 5081 | 6185 | 7330 | 8025 | 7974 | 9615 | 12246 | 11656 |
| 1988 | 1261 | 2017 | 2513 | 3476 | 471 | 5932 | 7523 | 8439 | 8748 | 9559 | 10824 | 14099 |
| 1989 | 1403 | 2021 | 2194 | 3047 | 4505 | 5889 | 7172 | 8852 | 10170 | 10392 | 12522 | 11923 |
| 1990 | 1647 | 1983 | 2566 | 3021 | 4077 | 5744 | 7038 | 7564 | 8854 | 10645 | 11674 | 11431 |
| 1991 | 1224 | 1939 | 2432 | 3160 | 3634 | 4967 | 6629 | 7704 | 9061 | 9117 | 10922 | 11342 |
| 1992 | 1269 | 1909 | 2578 | 3288 | 4150 | 4865 | 6168 | 7926 | 8349 | 9029 | 11574 | 9466 |
| 1993 | 1381 | 2143 | 2742 | 3636 | 4398 | 5421 | 5319 | 7006 | 8070 | 10048 | 9106 | 11591 |
| 1994 | 1444 | 1836 | 2649 | 3512 | 4906 | 5539 | 6818 | 6374 | 8341 | 9770 | 10528 | 11257 |
| 1995 | 1370 | 1977 | 2769 | 3722 | 4621 | 5854 | 6416 | 7356 | 6815 | 8312 | 9119 | 11910 |
| 1996 | 122 | 1755 | 2670 | 3802 | 4902 | 5681 | 7182 | 7734 | 9256 | 8322 | 10501 | 11894 |
| 19 | 132 | 193 | 240 | 3906 | 5032 | 6171 | 7202 | 7883 | 8856 | 9649 | 9621 | 10877 |
| 1998 | 1347 | 197 | 2943 | 3419 | 4850 | 5962 | 6933 | 7781 | 8695 | 9564 | 10164 | 10379 |
| 199 | 1279 | 2106 | 275 | 349 | 3831 | 5819 | 7072 | 8078 | 8865 | 10550 | 10823 | 11300 |
| 2000 | 1367 | 1929 | 2751 | 3274 | 4171 | 4447 | 6790 | 8216 | 9369 | 9817 | 10932 | 12204 |
| 2001 | 1280 | 1882 | 2599 | 3697 | 4420 | 5538 | 5639 | 7985 | 9059 | 9942 | 10632 | 10988 |
| 2002 | 1308 | 1946 | 2569 | 3266 | 4872 | 5365 | 6830 | 7067 | 9240 | 9659 | 10088 | 11632 |
| 2003 | 1310 | 1908 | 2545 | 3336 | 4069 | 5792 | 7156 | 8131 | 8051 | 10186 | 10948 | 11780 |
| 2004 | 1467 | 1847 | 2181 | 2918 | 4017 | 5135 | 7125 | 7732 | 8420 | 8927 | 10420 | 10622 |
| 2005 | 1287 | 1888 | 2307 | 2619 | 3516 | 5080 | 6060 | 8052 | 8292 | 8342 | 8567 | 10256 |
| 2006 | 1164 | 1722 | 2369 | 2808 | 3235 | 4361 | 6007 | 7166 | 8459 | 9324 | 9902 | 9636 |
| 2007 | 1140 | 1578 | 2122 | 2719 | 3495 | 4114 | 5402 | 6995 | 7792 | 9331 | 9970 | 10738 |
| 2008 | 1306 | 1805 | 2295 | 2749 | 3515 | 4530 | 5132 | 6394 | 7694 | 9170 | 9594 | 11258 |
| 2009 | 1412 | 1862 | 2561 | 3023 | 3676 | 4596 | 5651 | 6074 | 7356 | 8608 | 9812 | 10639 |
| 2010 | 1287 | 1787 | 2579 | 3469 | 4135 | 4850 | 5558 | 6289 | 6750 | 7997 | 9429 | 10481 |
| 2011 | 1175 | 1801 | 2526 | 3680 | 4613 | 5367 | 5685 | 6466 | 6851 | 7039 | 8268 | 8958 |
| 2012 | 1160 | 1668 | 2369 | 3347 | 4430 | 5486 | 6161 | 6448 | 7220 | 8054 | 8147 | 8901 |
| 2013 | 1056 | 1675 | 2219 | 3244 | 4529 | 5628 | 6397 | 7055 | 7378 | 7955 | 8400 | 8870 |
| 2014 | 1211 | 1575 | 2229 | 2983 | 4378 | 5598 | 6773 | 8023 | 7875 | 8646 | 9179 | 9749 |
| 2015 | 1142 | 1726 | 2217 | 3071 | 4030 | 5532 | 6846 | 7175 | 7491 | 8218 | 8575 | 9173 |
| 2016 | 1142 | 1726 | 2217 | 3071 | 4030 | 5532 | 6846 | 7175 | 7491 | 8218 | 8575 | 9173 |
| 2017 | 1142 | 1726 | 2217 | 3071 | 4030 | 5532 | 6846 | 7175 | 7491 | 8218 | 8575 | 9173 |

Table 8.4. Saithe in division Va. Maturity at age used for calculating the SSB.

|  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 0 | 0.089 | 0.197 | 0.380 | 0.604 | 0.792 | 0.905 | 1 | 1 | 1 | 1 | 1 |
| 1986 | 0 | 0.080 | 0.178 | 0.351 | 0.575 | 0.772 | 0.894 | 1 | 1 | 1 | 1 | 1 |
| 1987 | 0 | 0.072 | 0.162 | 0.325 | 0.547 | 0.751 | 0.883 | 1 | 1 | 1 | 1 | 1 |
| 1988 | 0 | 0.065 | 0.148 | 0.303 | 0.521 | 0.731 | 0.871 | 1 | 1 | 1 | 1 | 1 |
| 1989 | 0 | 0.060 | 0.138 | 0.285 | 0.499 | 0.714 | 0.862 | 1 | 1 | 1 | 1 | 1 |
| 1990 | 0 | 0.057 | 0.131 | 0.273 | 0.484 | 0.701 | 0.854 | 1 | 1 | 1 | 1 | 1 |
| 1991 | 0 | 0.055 | 0.127 | 0.266 | 0.475 | 0.694 | 0.850 | 1 | 1 | 1 | 1 | 1 |
| 1992 | 0 | 0.055 | 0.127 | 0.266 | 0.476 | 0.694 | 0.850 | 1 | 1 | 1 | 1 | 1 |
| 1993 | 0 | 0.057 | 0.131 | 0.274 | 0.485 | 0.702 | 0.855 | 1 | 1 | 1 | 1 | 1 |
| 1994 | 0 | 0.062 | 0.141 | 0.290 | 0.505 | 0.718 | 0.864 | 1 | 1 | 1 | 1 | 1 |
| 1995 | 0 | 0.069 | 0.157 | 0.317 | 0.537 | 0.743 | 0.879 | 1 | 1 | 1 | 1 | 1 |
| 1996 | 0 | 0.081 | 0.181 | 0.355 | 0.579 | 0.775 | 0.896 | 1 | 1 | 1 | 1 | 1 |
| 1997 | 0 | 0.097 | 0.212 | 0.402 | 0.627 | 0.807 | 0.913 | 1 | 1 | 1 | 1 | 1 |
| 1998 | 0 | 0.117 | 0.248 | 0.451 | 0.673 | 0.837 | 0.928 | 1 | 1 | 1 | 1 | 1 |
| 1999 | 0 | 0.137 | 0.284 | 0.497 | 0.712 | 0.860 | 0.939 | 1 | 1 | 1 | 1 | 1 |
| 2000 | 0 | 0.154 | 0.313 | 0.532 | 0.740 | 0.877 | 0.947 | 1 | 1 | 1 | 1 | 1 |
| 2001 | 0 | 0.165 | 0.331 | 0.552 | 0.755 | 0.885 | 0.951 | 1 | 1 | 1 | 1 | 1 |
| 2002 | 0 | 0.169 | 0.337 | 0.560 | 0.760 | 0.888 | 0.952 | 1 | 1 | 1 | 1 | 1 |
| 2003 | 0 | 0.168 | 0.335 | 0.557 | 0.759 | 0.887 | 0.952 | 1 | 1 | 1 | 1 | 1 |
| 2004 | 0 | 0.163 | 0.328 | 0.549 | 0.753 | 0.884 | 0.950 | 1 | 1 | 1 | 1 | 1 |
| 2005 | 0 | 0.157 | 0.318 | 0.538 | 0.744 | 0.879 | 0.948 | 1 | 1 | 1 | 1 | 1 |
| 2006 | 0 | 0.152 | 0.309 | 0.527 | 0.736 | 0.874 | 0.946 | 1 | 1 | 1 | 1 | 1 |
| 2007 | 0 | 0.146 | 0.300 | 0.517 | 0.728 | 0.870 | 0.943 | 1 | 1 | 1 | 1 | 1 |
| 2008 | 0 | 0.141 | 0.291 | 0.506 | 0.719 | 0.865 | 0.941 | 1 | 1 | 1 | 1 | 1 |
| 2009 | 0 | 0.136 | 0.282 | 0.495 | 0.710 | 0.859 | 0.939 | 1 | 1 | 1 | 1 | 1 |
| 2010 | 0 | 0.130 | 0.272 | 0.483 | 0.700 | 0.853 | 0.936 | 1 | 1 | 1 | 1 | 1 |
| 2011 | 0 | 0.124 | 0.261 | 0.469 | 0.688 | 0.847 | 0.932 | 1 | 1 | 1 | 1 | 1 |
| 2012 | 0 | 0.118 | 0.250 | 0.455 | 0.676 | 0.839 | 0.929 | 1 | 1 | 1 | 1 | 1 |
| 2013 | 0 | 0.112 | 0.239 | 0.440 | 0.662 | 0.830 | 0.924 | 1 | 1 | 1 | 1 | 1 |
| 2014 | 0 | 0.106 | 0.228 | 0.424 | 0.648 | 0.821 | 0.920 | 1 | 1 | 1 | 1 | 1 |
| 2015 | 0 | 0.100 | 0.217 | 0.409 | 0.633 | 0.812 | 0.915 | 1 | 1 | 1 | 1 | 1 |
| 2016 | 0 | 0.100 | 0.217 | 0.409 | 0.633 | 0.812 | 0.915 | 1 | 1 | 1 | 1 | 1 |
| 2017 | 0 | 0.100 | 0.217 | 0.409 | 0.633 | 0.812 | 0.915 | 1 | 1 | 1 | 1 | 1 |

Table 8.5. Saithe in division Va. Survey catch at age.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 198 | 0.0 | 0.6 | 0.58 | 2.99 | 5.1 | 1.74 | 1.0 | 0.5 | 1.3 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 |
| 5 | 5 | 1 |  |  | 1 |  | 6 | 0 | 7 | 6 | 8 | 8 | 7 | 7 |
| 198 | 0.0 | 2.3 | 2.40 | 2.06 | 2.0 | 1.42 | 0.6 | 0.2 | 0.1 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 |
| 6 | 2 | 3 |  |  | 9 |  | 2 | 8 | 9 | 2 | 9 | 7 | 3 | 0 |
| 198 | 0.1 | 0.3 | 11.5 | 12.9 | 6.4 | 3.95 | 3.0 | 0.7 | 0.3 | 0.2 | 0.3 | 0.0 | 0.0 | 0.0 |
| 7 | 0 | 9 | 2 | 3 | 2 |  | 7 | 9 | 6 | 6 | 3 | 5 | 1 | 3 |
| 198 | 0.6 | 0.3 | 0.49 | 2.72 | 2.8 | 1.71 | 0.9 | 0.4 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 |
| 8 | 9 | 1 |  |  | 1 |  | 5 | 0 | 7 | 8 | 0 | 5 | 1 | 0 |
| 198 | 0.2 | 1.4 | 3.96 | 5.05 | 6.5 | 2.49 | 1.7 | 0.9 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 9 | 0 | 3 |  |  | 7 |  | 7 | 1 | 0 | 0 | 2 | 0 | 3 | 0 |
| 199 | 0.0 | 0.3 | 1.69 | 4.86 | 6.3 | 12.3 | 3.3 | 1.2 | 0.6 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 |
| 0 | 1 | 5 |  |  | 7 | 3 | 0 | 1 | 4 | 2 | 6 | 2 | 1 | 3 |
| 199 | 0.0 | 0.2 | 1.40 | 1.72 | 2.2 | 1.13 | 2.5 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1 | 1 | 2 |  |  | 2 |  | 0 | 0 | 2 | 3 | 0 | 1 | 0 | 1 |
| 199 | 0.0 | 0.1 | 0.91 | 5.73 | 5.5 | 2.79 | 2.6 | 1.9 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2 | 1 | 5 |  |  | 2 |  | 8 | 1 | 8 | 6 | 6 | 2 | 0 | 0 |
| 199 | 0.0 | 1.2 | 11.0 | 2.00 | 6.8 | 2.41 | 2.2 | 1.0 | 4.0 | 0.6 | 0.0 | 0.0 | 0.0 | 0.0 |
| 3 | 0 | 7 | 4 |  | 0 |  | 5 | 2 | 2 | 4 | 5 | 0 | 2 | 0 |
| 199 | 0.0 | 0.8 | 0.73 | 1.89 | 1.7 | 1.95 | 0.5 | 0.8 | 1.0 | 3.6 | 0.4 | 0.1 | 0.0 | 0.0 |
| 4 | 4 | 2 |  |  | 4 |  | 3 | 4 | 0 | 2 | 1 | 8 | 0 | 4 |
| 199 | 0.0 | 0.4 | 1.98 | 1.12 | 0.5 | 0.28 | 0.3 | 0.1 | 0.1 | 0.1 | 0.3 | 0.0 | 0.0 | 0.0 |
| 5 | 6 | 8 |  |  | 1 |  | 4 | 0 | 5 | 5 | 3 | 2 | 0 | 0 |
| 199 | 0.0 | 0.1 | 0.51 | 3.76 | 1.1 | 0.99 | 0.5 | 1.0 | 0.0 | 0.0 | 0.1 | 0.2 | 0.0 | 0.0 |
| 6 | 3 | 3 |  |  | 2 |  | 8 | 0 | 5 | 9 | 0 | 5 | 3 | 0 |
| 199 | 0.1 | 0.3 | 0.90 | 4.72 | 3.9 | 0.94 | 0.4 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 7 | 6 | 2 |  |  | 6 |  | 0 | 6 | 0 | 5 | 2 | 2 | 2 | 0 |
| 199 | 0.0 | 0.1 | 1.64 | 2.33 | 2.5 | 1.23 | 0.7 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 8 | 1 | 1 |  |  | 3 |  | 1 | 1 | 8 | 7 | 4 | 3 | 5 | 3 |
| 199 | 0.5 | 0.7 | 3.71 | 0.93 | 1.2 | 1.64 | 0.5 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 9 | 7 | 5 |  |  | 5 |  | 7 | 7 | 2 | 2 | 2 | 0 | 0 | 2 |
| 200 | 0.0 | 0.3 | 2.02 | 2.54 | 0.6 | 0.84 | 0.5 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 0 | 0 | 8 |  |  | 1 |  | 3 | 7 | 7 | 3 | 1 | 0 | 1 | 1 |
| 200 | 0.0 | 0.8 | 1.90 | 2.64 | 1.6 | 0.20 | 0.2 | 0.4 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1 | 0 | 9 |  |  | 0 |  | 3 | 0 | 3 | 7 | 4 | 1 | 0 | 0 |
| 200 | 0.0 | 1.0 | 2.23 | 2.97 | 3.0 | 2.15 | 0.4 | 0.4 | 0.3 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2 | 2 | 5 |  |  | 8 |  | 2 | 9 | 2 | 2 | 2 | 3 | 0 | 0 |
| 200 | 0.0 | 0.0 | 9.62 | 5.06 | 2.9 | 1.34 | 0.7 | 0.2 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 |
| 3 | 1 | 5 |  |  | 4 |  | 7 | 1 | 5 | 0 | 2 | 3 | 0 | 0 |
| 200 | 0.0 | 0.9 | 1.38 | 9.39 | 6.0 | 4.35 | 1.4 | 0.8 | 0.1 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 |
| 4 | 1 | 1 |  |  | 4 |  | 8 | 1 | 7 | 6 | 2 | 6 | 2 | 0 |
| 200 | 0.0 | 0.2 | 4.32 | 2.39 | 7.4 | 4.66 | 2.3 | 0.8 | 0.4 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 |
| 5 | 0 | 6 |  |  | 2 |  | 1 | 6 | 4 | 2 | 5 | 8 | 3 | 0 |
| 200 | 0.0 | 0.0 | 2.18 | 6.69 | 1.9 | 8.91 | 3.5 | 1.2 | 0.2 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 |
| 6 | 1 | 0 |  |  | 8 |  | 2 | 1 | 9 | 5 | 3 | 4 | 4 | 0 |
| 200 | 0.0 | 0.0 | 0.31 | 1.73 | 3.2 | 0.81 | 1.6 | 0.7 | 0.2 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 |
| 7 | 0 | 6 |  |  | 2 |  | 2 | 0 | 9 | 6 | 1 | 8 | 2 | 0 |
| 200 | 0.0 | 0.0 | 2.25 | 1.79 | 2.8 | 4.01 | 0.6 | 0.7 | 0.3 | 0.1 | 0.0 | 0.1 | 0.0 | 0.0 |
| 8 | 1 | 8 |  |  | 5 |  | 1 | 8 | 4 | 5 | 9 | 3 | 4 | 2 |


| 200 | 0.0 | 0.2 | 2.43 | 1.80 | 0.6 | 0.91 | 0.8 | 0.1 | 0.2 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 9 | 1 | 1 |  |  | 8 |  | 4 | 2 | 6 | 5 | 3 | 4 | 0 | 2 |
| 201 | 0.0 | 0.0 | 1.23 | 4.99 | 2.4 | 0.63 | 0.6 | 0.4 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 |
| 0 | 0 | 7 |  |  | 9 |  | 0 | 8 | 7 | 3 | 7 | 7 | 7 | 2 |
| 201 | 0.0 | 0.1 | 3.83 | 4.20 | 3.0 | 1.15 | 0.4 | 0.3 | 0.4 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 |
| 1 | 0 | 5 |  |  | 6 |  | 1 | 9 | 4 | 7 | 0 | 9 | 6 | 5 |
| 201 | 0.0 | 0.0 | 1.75 | 12.0 | 6.8 | 2.75 | 0.6 | 0.1 | 0.3 | 0.5 | 0.1 | 0.1 | 0.0 | 0.0 |
| 2 | 2 | 2 |  | 4 | 6 |  | 2 | 7 | 8 | 0 | 3 | 2 | 6 | 8 |
| 201 | 0.0 | 0.1 | 4.27 | 7.43 | 6.7 | 4.65 | 2.5 | 1.1 | 0.3 | 0.4 | 0.3 | 0.2 | 0.1 | 0.0 |
| 3 | 1 | 2 |  |  | 8 |  | 7 | 2 | 0 | 4 | 6 | 6 | 3 | 1 |
| 201 | 0.0 | 0.0 | 0.39 | 3.84 | 3.7 | 2.04 | 0.8 | 0.4 | 0.1 | 0.1 | 0.1 | 0.1 | 0.0 | 0.0 |
| 4 | 1 | 3 |  |  | 8 |  | 6 | 2 | 5 | 1 | 8 | 8 | 7 | 9 |
| 201 | 0.0 | 0.0 | 1.07 | 1.90 | 3.1 | 1.72 | 0.8 | 0.7 | 0.6 | 0.4 | 0.2 | 0.2 | 0.2 | 0.1 |
| 5 | 6 | 4 |  |  | 6 |  | 1 | 2 | 8 | 5 | 6 | 3 | 1 | 5 |

Table 8.6. Saithe in division Va. Commercial catch-at-age residuals $\log (\mathrm{obs} / \mathrm{fit})$.

|  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | -0.79 | -0.60 | 0.25 | 0.16 | -0.08 | 0.22 | -0.05 | 0.22 | -0.30 | -0.41 | -0.70 | -0.04 |
| 1981 | -0.51 | -0.26 | -0.56 | 0.36 | -0.16 | 0.06 | 0.12 | 0.28 | -0.85 | 0.93 | 1.19 | 1.88 |
| 1982 | 0.84 | -0.25 | 0.13 | -0.37 | 0.24 | -0.03 | 0.37 | -0.36 | -1.20 | -1.15 | -0.53 | -0.10 |
| 1983 | -2.48 | 0.94 | -0.67 | 0.11 | 0.46 | 0.20 | -0.01 | -0.80 | -2.43 | -2.79 | -5.17 | -3.83 |
| 1984 | -4.23 | -1.61 | -1.29 | 0.22 | 0.47 | 0.79 | -0.45 | 0.43 | 0.96 | 1.03 | 3.47 | 4.90 |
| 1985 | -0.28 | 1.23 | 0.58 | 0.08 | 0.27 | -0.19 | -1.16 | -0.75 | -1.41 | -3.07 | 0.65 | 2.46 |
| 1986 | 2.28 | -0.72 | 0.24 | -0.32 | -0.09 | 0.10 | -0.41 | 0.93 | -1.09 | -1.39 | -1.83 | -1.75 |
| 1987 | -0.99 | 0.16 | 0.32 | 0.22 | 0.09 | 0.07 | 0.06 | -0.18 | -0.08 | -2.91 | -1.91 | -0.24 |
| 1988 | 0.92 | -0.31 | 0.06 | 0.04 | -0.14 | 0.20 | 0.79 | -0.66 | 0.48 | -1.69 | -3.27 | -6.88 |
| 1989 | -0.84 | 0.61 | 0.01 | 0.23 | -0.58 | 0.04 | 0.27 | -0.18 | 0.71 | 0.37 | -1.13 | -3.73 |
| 1990 | -1.78 | -0.54 | 0.07 | 0.01 | 0.34 | 0.06 | 0.11 | -0.37 | 0.07 | -0.78 | 0.16 | -1.61 |
| 1991 | -1.96 | -1.11 | 0.06 | 0.33 | 0.03 | -0.08 | 0.01 | 0.72 | 0.23 | -1.38 | -3.92 | -3.95 |
| 1992 | -0.22 | 0.58 | 1.05 | 0.43 | 0.06 | -0.83 | -0.25 | -0.40 | -0.28 | -1.17 | 0.48 | -0.88 |
| 1993 | 0.98 | -0.16 | -0.33 | -0.14 | -0.20 | -0.08 | 0.40 | 0.04 | 0.33 | 0.75 | 0.65 | -1.28 |
| 1994 | 1.09 | 0.98 | -0.12 | -0.69 | -0.43 | -0.46 | 0.20 | 0.41 | 0.51 | 0.80 | 1.87 | 1.79 |
| 1995 | 1.59 | 0.28 | 0.10 | -0.02 | 0.04 | -0.08 | -0.20 | -0.19 | -0.26 | -0.87 | -0.68 | -1.83 |
| 1996 | 1.45 | 0.17 | -0.12 | -0.51 | -0.34 | 0.19 | 0.27 | 0.03 | 0.39 | -0.15 | 1.33 | 2.41 |
| 1997 | 0.25 | 0.33 | -0.30 | 0.08 | -0.03 | 0.29 | -0.10 | -0.40 | -0.30 | 0.37 | -1.12 | 0.20 |
| 1998 | -0.36 | -0.07 | -0.47 | -0.76 | 0.18 | 0.00 | 0.84 | 0.70 | 1.24 | 1.15 | 1.57 | 0.84 |
| 1999 | 0.38 | 0.03 | 0.00 | 0.08 | 0.00 | -0.20 | -0.35 | 0.33 | -0.65 | -0.57 | 0.31 | 0.17 |
| 2000 | -0.07 | -0.20 | 0.11 | 0.07 | -0.16 | 0.48 | -0.49 | -0.27 | -0.22 | -0.93 | -0.09 | -1.12 |
| 2001 | -0.09 | 0.23 | -0.27 | -0.14 | -0.03 | -0.18 | 0.39 | 0.06 | 0.13 | 0.04 | 0.37 | 1.00 |
| 2002 | -0.62 | -0.08 | 0.18 | 0.36 | -0.17 | 0.12 | -0.26 | -0.35 | -0.09 | -1.20 | -0.10 | -0.36 |
| 2003 | 0.40 | -0.28 | 0.42 | -0.02 | -0.02 | -0.61 | 0.04 | -0.20 | 0.04 | -1.24 | 0.24 | 1.21 |
| 2004 | -0.16 | -0.39 | -0.11 | 0.27 | -0.10 | 0.28 | 0.61 | 0.41 | -0.05 | -0.47 | 0.75 | -0.26 |
| 2005 | -0.39 | -0.32 | -0.31 | 0.44 | 0.34 | -0.23 | -0.13 | -0.01 | 0.31 | -0.07 | -0.31 | -0.43 |
| 2006 | -0.69 | -0.16 | -0.35 | -0.03 | 0.56 | 0.07 | -0.33 | -0.10 | 0.76 | 0.95 | 1.05 | 0.12 |
| 2007 | 0.79 | 0.22 | 0.14 | 0.25 | -0.13 | -0.04 | -0.40 | -0.48 | -0.82 | 0.30 | 0.25 | -0.26 |
| 2008 | 0.11 | 0.36 | 0.18 | 0.23 | -0.11 | -0.23 | -0.24 | -0.30 | -0.60 | 0.10 | 2.71 | 1.70 |
| 2009 | 0.75 | 0.49 | -0.11 | -0.27 | -0.16 | 0.29 | -0.34 | -0.12 | 0.10 | 0.42 | 1.08 | 2.54 |
| 2010 | 0.46 | 0.28 | 0.11 | -0.46 | 0.03 | -0.12 | 0.47 | -0.43 | 0.07 | -0.06 | 0.84 | 1.18 |
| 2011 | 0.14 | -0.06 | -0.05 | -0.30 | -0.04 | 0.29 | 0.14 | 0.36 | -0.24 | -0.01 | 0.48 | 1.41 |
| 2012 | -0.57 | -0.37 | -0.27 | -0.20 | -0.24 | 0.06 | 0.66 | 0.55 | 1.38 | 0.12 | 0.55 | 0.91 |
| 2013 | -0.50 | -0.04 | 0.33 | -0.02 | -0.19 | -0.11 | 0.28 | 0.12 | -0.36 | 0.35 | -0.53 | 0.01 |
| 2014 | 0.16 | -0.20 | 0.50 | -0.04 | -0.22 | -0.37 | -0.24 | 0.45 | 0.28 | -0.38 | 1.99 | 0.16 |

Table 8.7. Saithe in division Va. Survey catch-at-age residuals $\log (o b s / f i t)$.

|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | -0.43 | -1.53 | -0.46 | 0.55 | 0.20 | 0.36 | -0.17 | 0.83 | -1.01 |
| 1986 | 0.78 | -0.61 | -0.69 | -0.80 | -0.50 | -0.35 | -0.47 | -0.68 | -0.39 |
| 1987 | -0.62 | 0.87 | 0.73 | 0.74 | 0.44 | 1.13 | 0.74 | 0.53 | 0.25 |
| 1988 | -0.35 | -2.15 | -1.48 | -0.96 | -0.29 | -0.46 | -0.39 | -1.31 | -0.59 |
| 1989 | 1.95 | 0.86 | -0.05 | -0.34 | -0.60 | 0.50 | 0.33 | 0.35 | -5.74 |
| 1990 | -0.12 | 0.36 | 0.45 | 0.33 | 0.91 | 0.48 | 0.88 | 0.65 | -0.47 |
| 1991 | 0.15 | -0.28 | -0.27 | -0.36 | -1.18 | -0.62 | -1.45 | -3.10 | -2.26 |
| 1992 | -0.64 | 0.03 | 0.74 | 1.23 | 0.46 | 0.66 | -0.02 | -0.70 | -1.15 |
| 1993 | 2.00 | 2.63 | $0.31$ | 1.07 | 0.80 | 1.02 | 0.44 | 1.69 | 0.97 |
| 1994 | 0.88 | -0.44 | -0.09 | 0.29 | 0.18 | -0.12 | 0.84 | 1.31 | 2.32 |
| 1995 | 0.43 | 0.12 | -0.56 | -1.48 | -1.24 | -0.96 | -1.03 | -0.21 | -0.09 |
| 1996 | -0.62 | -1.29 | 0.24 | -0.41 | -0.09 | 0.52 | 1.34 | -0.87 | 0.00 |
| 1997 | 1.22 | -0.13 | 0.70 | 0.44 | -0.07 | -0.32 | -0.04 | -0.47 | -0.15 |
| 1998 | -1.50 | 1.36 | 0.36 | 0.11 | -0.41 | 0.34 | 0.23 | -0.06 | -0.18 |
| 1999 | 0.72 | 0.85 | 0.06 | -0.24 | 0.09 | -0.64 | -0.57 | -2.24 | -1.04 |
| 2000 | -0.72 | 0.10 | -0.23 | -0.30 | -0.21 | -0.55 | -0.07 | -0.87 | -1.14 |
| 2001 | 0.10 | -0.62 | -0.22 | -0.64 | -1.11 | -1.04 | -0.08 | -0.84 | -0.22 |
| 2002 | 0.13 | -0.61 | -0.72 | 0.08 | 0.18 | 0.42 | 0.60 | 0.34 | 0.36 |
| 2003 | -2.22 | 0.95 | -0.28 | -0.61 | -0.41 | -0.34 | 0.38 | -1.36 | -0.40 |
| 2004 | -0.05 | -0.12 | 0.31 | 0.07 | 0.34 | 0.37 | 0.45 | 0.80 | 0.59 |
| 2005 | -0.87 | 0.00 | -0.06 | 0.26 | 0.33 | 0.29 | 0.40 | 0.25 | 0.83 |
| 2006 | -6.50 | -0.16 | -0.07 | -0.03 | 1.07 | 0.72 | 0.23 | -0.33 | 0.06 |
| 2007 | -2.08 | -1.51 | -1.01 | -0.67 | -0.48 | -0.20 | -0.47 | -0.90 | -0.54 |
| 2008 | -2.23 | 0.39 | -0.03 | -0.17 | 0.19 | -0.09 | -0.36 | -0.77 | -1.19 |
| 2009 | -1.10 | 0.00 | -0.45 | -0.90 | -0.89 | -0.90 | -1.27 | -1.08 | -1.22 |
| 2010 | -2.62 | -0.81 | 0.25 | 0.18 | -0.39 | -0.66 | -0.85 | -1.35 | -1.41 |
| 2011 | -1.60 | 0.30 | 0.04 | -0.10 | -0.17 | -0.21 | -0.50 | -0.45 | 0.12 |
| 2012 | -3.77 | -0.50 | 1.04 | 0.82 | 0.31 | -0.24 | -0.60 | -0.05 | 0.11 |
| 2013 | -0.95 | 0.71 | 0.60 | 0.56 | 0.88 | 0.87 | 1.09 | 0.54 | 0.52 |
| 2014 | -2.84 | -1.35 | -0.03 | 0.00 | -0.32 | -0.45 | -0.62 | -0.80 | -0.24 |
| 2015 | -2.74 | -0.45 | -0.10 | -0.09 | -0.43 | -0.80 | -0.06 | 0.38 | 0.85 |

Table 8.8. Saithe in division Va. Main population estimates. The recruitment column is aligned so that the 2000 cohort is shown in the year 2000, but that cohort size is the estimated $N$ at age 3 in 2003.

|  | B4+ | SSB | Cohort | Y | F4-9 | HR |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 312 | 122 | 32 | 58 | 0.29 | 19\% |
| 1981 | 304 | 130 | 42 | 59 | 0.26 | 19\% |
| 1982 | 294 | 148 | 35 | 69 | 0.30 | 23\% |
| 1983 | 270 | 147 | 67 | 58 | 0.24 | 22\% |
| 1984 | 287 | 149 | 91 | 63 | 0.23 | 22\% |
| 1985 | 299 | 139 | 50 | 57 | 0.25 | 19\% |
| 1986 | 318 | 137 | 32 | 65 | 0.28 | 20\% |
| 1987 | 335 | 128 | 21 | 81 | 0.35 | 24\% |
| 1988 | 415 | 125 | 29 | 77 | 0.32 | 19\% |
| 1989 | 397 | 127 | 15 | 82 | 0.31 | 21\% |
| 1990 | 377 | 134 | 20 | 98 | 0.35 | 26\% |
| 1991 | 336 | 143 | 18 | 102 | 0.37 | 30\% |
| 1992 | 288 | 135 | 30 | 80 | 0.37 | 28\% |
| 1993 | 230 | 112 | 25 | 72 | 0.40 | 31\% |
| 1994 | 187 | 93 | 17 | 64 | 0.45 | 34\% |
| 1995 | 152 | 70 | 9 | 49 | 0.46 | 32\% |
| 1996 | 148 | 61 | 30 | 40 | 0.41 | 27\% |
| 1997 | 155 | 62 | 31 | 37 | 0.37 | 24\% |
| 1998 | 153 | 68 | 53 | 32 | 0.30 | 21\% |
| 1999 | 131 | 72 | 63 | 31 | 0.31 | 24\% |
| 2000 | 141 | 74 | 72 | 33 | 0.33 | 23\% |
| 2001 | 161 | 80 | 26 | 32 | 0.28 | 20\% |
| 2002 | 217 | 96 | 72 | 42 | 0.31 | 19\% |
| 2003 | 276 | 118 | 42 | 52 | 0.30 | 19\% |
| 2004 | 316 | 137 | 19 | 65 | 0.26 | 20\% |
| 2005 | 282 | 147 | 27 | 69 | 0.29 | 25\% |
| 2006 | 307 | 156 | 41 | 76 | 0.31 | 25\% |
| 2007 | 278 | 152 | 41 | 64 | 0.28 | 23\% |
| 2008 | 248 | 149 | 50 | 70 | 0.32 | 28\% |
| 2009 | 224 | 137 | 45 | 61 | 0.30 | 27\% |
| 2010 | 227 | 127 | 39 | 54 | 0.27 | 24\% |
| 2011 | 239 | 122 | 21 | 51 | 0.24 | 21\% |
| 2012 | 253 | 122 | 26 | 52 | 0.23 | 20\% |
| 2013 | 268 | 128 | 32 | 58 | 0.25 | 22\% |
| 2014 | 265 | 132 | 33 | 46 | 0.19 | 18\% |
| 2015 | 255 | 139 | 33 | 57 | 0.26 | 22\% |

Table 8.9. Saithe in division Va. Stock in numbers.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 32.2 | 24.6 | 28.2 | 46.8 | 30.9 | 10.3 | 8.1 | 3.7 | 1.3 | 0.7 | 0.7 | 0.5 | 0.3 | 0.1 |
| 1981 | 47.9 | 26.4 | 20.2 | 22.7 | 35.2 | 21.2 | 6.3 | 4.6 | 2.0 | 0.7 | 0.4 | 0.4 | 0.3 | 0.2 |
| 1982 | 62.4 | 39.3 | 21.6 | 16.3 | 17.2 | 24.6 | 13.3 | 3.7 | 2.6 | 1.1 | 0.4 | 0.2 | 0.2 | 0.2 |
| 1983 | 52.8 | 51.1 | 32.1 | 17.4 | 12.2 | 11.8 | 14.8 | 7.5 | 1.9 | 1.4 | 0.6 | 0.2 | 0.1 | 0.1 |
| 1984 | 99.7 | 43.2 | 41.8 | 26.0 | 13.3 | 8.6 | 7.5 | 9.0 | 4.3 | 1.1 | 0.8 | 0.4 | 0.1 | 0.1 |
| 1985 | 136.4 | 81.6 | 35.4 | 33.8 | 19.9 | 9.4 | 5.6 | 4.6 | 5.2 | 2.5 | 0.7 | 0.5 | 0.2 | 0.1 |
| 1986 | 75.3 | 111.7 | 66.8 | 28.6 | 25.8 | 14.0 | 6.0 | 3.4 | 2.6 | 3.0 | 1.4 | 0.4 | 0.3 | 0.1 |
| 1987 | 47.6 | 61.7 | 91.4 | 53.9 | 21.6 | 17.8 | 8.7 | 3.5 | 1.8 | 1.5 | 1.6 | 0.8 | 0.2 | 0.2 |
| 1988 | 31.0 | 39.0 | 50.5 | 73.4 | 39.8 | 14.3 | 10.2 | 4.6 | 1.7 | 0.9 | 0.7 | 0.9 | 0.4 | 0.1 |
| 1989 | 44.0 | 25.4 | 31.9 | 40.6 | 54.7 | 26.8 | 8.5 | 5.6 | 2.3 | 0.9 | 0.5 | 0.4 | 0.5 | 0.2 |
| 1990 | 22.1 | 36.0 | 20.8 | 25.7 | 30.4 | 37.2 | 16.2 | 4.7 | 2.9 | 1.2 | 0.5 | 0.3 | 0.2 | 0.3 |
| 1991 | 29.5 | 18.1 | 29.5 | 16.7 | 19.0 | 20.2 | 31.4 | 8.6 | 2.3 | 1.5 | 0.6 | 0.2 | 0.1 | 0.1 |
| 1992 | 26.3 | 24.2 | 14.8 | 23.6 | 12.3 | 12.4 | 11.3 | 16.2 | 4.0 | 1.1 | 0.7 | 0.3 | 0.1 | 0.1 |
| 1993 | 44.3 | 21.5 | 19.8 | 11.9 | 17.4 | 8.0 | 7.0 | 5.9 | 7.7 | 2.0 | 0.5 | 0.4 | 0.2 | 0.1 |
| 1994 | 38.0 | 36.3 | 17.6 | 15.9 | 8.7 | 11.2 | 4.4 | 3.5 | 2.7 | 3.6 | 0.9 | 0.3 | 0.2 | 0.1 |
| 1995 | 25.0 | 31.1 | 29.7 | 14.1 | 11.4 | 5.4 | 5.8 | 2.1 | 1.5 | 1.2 | 1.5 | 0.4 | 0.1 | 0.1 |
| 1996 | 12.8 | 20.5 | 25.5 | 23.7 | 10.1 | 7.0 | 2.8 | 2.7 | 0.8 | 0.6 | 0.5 | 0.7 | 0.2 | 0.1 |
| 1997 | 44.9 | 10.5 | 16.8 | 20.4 | 17.3 | 6.5 | 3.8 | 1.4 | 1.2 | 0.4 | 0.3 | 0.2 | 0.4 | 0.1 |
| 1998 | 46.2 | 36.8 | 8.6 | 13.3 | 14.5 | 11.2 | 3.9 | 2.1 | 0.7 | 0.6 | 0.2 | 0.1 | 0.1 | 0.2 |
| 1999 | 79.7 | 37.8 | 30.1 | 6.8 | 9.6 | 9.8 | 7.1 | 2.3 | 1.1 | 0.3 | 0.3 | 0.1 | 0.1 | 0.1 |
| 2000 | 93.4 | 65.2 | 31.0 | 23.9 | 5.0 | 6.5 | 6.1 | 4.1 | 1.2 | 0.6 | 0.2 | 0.1 | 0.0 | 0.0 |
| 2001 | 106.8 | 76.5 | 53.4 | 24.6 | 17.2 | 3.3 | 4.0 | 3.5 | 2.1 | 0.6 | 0.3 | 0.1 | 0.1 | 0.0 |
| 2002 | 38.1 | 87.5 | 62.6 | 42.6 | 18.0 | 11.8 | 2.1 | 2.4 | 1.9 | 1.1 | 0.3 | 0.1 | 0.0 | 0.0 |
| 2003 | 107.7 | 31.2 | 71.6 | 49.8 | 30.9 | 12.2 | 7.4 | 1.2 | 1.3 | 1.0 | 0.6 | 0.2 | 0.1 | 0.0 |
| 2004 | 62.2 | 88.2 | 25.5 | 57.0 | 36.2 | 20.9 | 7.7 | 4.3 | 0.7 | 0.6 | 0.5 | 0.3 | 0.1 | 0.0 |
| 2005 | 28.1 | 50.9 | 72.2 | 19.9 | 38.0 | 22.8 | 13.0 | 4.8 | 2.7 | 0.4 | 0.4 | 0.3 | 0.2 | 0.0 |
| 2006 | 40.4 | 23.0 | 41.7 | 56.2 | 13.0 | 23.3 | 13.8 | 7.9 | 3.0 | 1.6 | 0.2 | 0.2 | 0.1 | 0.1 |
| 2007 | 61.0 | 33.1 | 18.8 | 32.4 | 36.2 | 7.9 | 13.8 | 8.2 | 4.8 | 1.7 | 0.9 | 0.1 | 0.1 | 0.1 |
| 2008 | 61.1 | 50.0 | 27.1 | 14.7 | 21.2 | 22.3 | 4.8 | 8.5 | 5.1 | 2.8 | 0.9 | 0.4 | 0.1 | 0.0 |
| 2009 | 74.2 | 50.0 | 40.9 | 21.0 | 9.3 | 12.6 | 13.0 | 2.8 | 5.0 | 2.9 | 1.5 | 0.5 | 0.2 | 0.0 |
| 2010 | 66.6 | 60.8 | 41.0 | 31.8 | 13.6 | 5.6 | 7.5 | 7.8 | 1.7 | 2.9 | 1.6 | 0.7 | 0.2 | 0.1 |
| 2011 | 58.4 | 54.5 | 49.7 | 32.0 | 21.1 | 8.5 | 3.5 | 4.7 | 4.9 | 1.0 | 1.6 | 0.8 | 0.4 | 0.1 |
| 2012 | 30.6 | 47.8 | 44.6 | 39.0 | 21.7 | 13.6 | 5.4 | 2.2 | 3.0 | 3.0 | 0.6 | 0.9 | 0.5 | 0.2 |
| 2013 | 39.3 | 25.1 | 39.2 | 35.1 | 26.7 | 14.1 | 8.7 | 3.5 | 1.4 | 1.9 | 1.8 | 0.3 | 0.5 | 0.3 |
| 2014 | 47.6 | 32.1 | 20.5 | 30.7 | 23.6 | 17.0 | 8.9 | 5.5 | 2.2 | 0.9 | 1.1 | 1.0 | 0.2 | 0.3 |
| 2015 | 48.8 | 39.0 | 26.3 | 16.2 | 21.6 | 16.0 | 11.4 | 6.0 | 3.7 | 1.5 | 0.6 | 0.7 | 0.6 | 0.1 |

Table 8.10. Saithe in division Va. Fishing mortality rate.

|  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0.02 | 0.09 | 0.18 | 0.30 | 0.36 | 0.44 | 0.41 | 0.44 | 0.36 | 0.36 | 0.36 | 0.36 |
| 1981 | 0.01 | 0.08 | 0.16 | 0.26 | 0.32 | 0.39 | 0.36 | 0.39 | 0.32 | 0.32 | 0.32 | 0.32 |
| 1982 | 0.02 | 0.09 | 0.18 | 0.30 | 0.37 | 0.45 | 0.42 | 0.45 | 0.37 | 0.37 | 0.37 | 0.37 |
| 1983 | 0.01 | 0.07 | 0.15 | 0.24 | 0.30 | 0.36 | 0.34 | 0.36 | 0.30 | 0.30 | 0.30 | 0.30 |
| 1984 | 0.01 | 0.07 | 0.14 | 0.23 | 0.29 | 0.34 | 0.32 | 0.34 | 0.28 | 0.28 | 0.28 | 0.28 |
| 1985 | 0.01 | 0.07 | 0.15 | 0.25 | 0.30 | 0.37 | 0.34 | 0.37 | 0.30 | 0.30 | 0.30 | 0.30 |
| 1986 | 0.02 | 0.08 | 0.17 | 0.28 | 0.35 | 0.42 | 0.39 | 0.42 | 0.34 | 0.34 | 0.34 | 0.34 |
| 1987 | 0.02 | 0.10 | 0.21 | 0.35 | 0.43 | 0.52 | 0.49 | 0.52 | 0.43 | 0.43 | 0.43 | 0.43 |
| 1988 | 0.02 | 0.09 | 0.19 | 0.32 | 0.40 | 0.48 | 0.45 | 0.48 | 0.39 | 0.39 | 0.39 | 0.39 |
| 1989 | 0.02 | 0.09 | 0.19 | 0.31 | 0.38 | 0.46 | 0.43 | 0.46 | 0.37 | 0.37 | 0.37 | 0.37 |
| 1990 | 0.02 | 0.10 | 0.21 | 0.35 | 0.43 | 0.52 | 0.48 | 0.52 | 0.43 | 0.43 | 0.43 | 0.43 |
| 1991 | 0.02 | 0.11 | 0.23 | 0.38 | 0.46 | 0.56 | 0.52 | 0.56 | 0.46 | 0.46 | 0.46 | 0.46 |
| 1992 | 0.02 | 0.11 | 0.22 | 0.37 | 0.45 | 0.55 | 0.51 | 0.55 | 0.45 | 0.45 | 0.45 | 0.45 |
| 1993 | 0.02 | 0.12 | 0.24 | 0.40 | 0.49 | 0.59 | 0.55 | 0.59 | 0.49 | 0.49 | 0.49 | 0.49 |
| 1994 | 0.03 | 0.13 | 0.27 | 0.45 | 0.56 | 0.67 | 0.63 | 0.67 | 0.55 | 0.55 | 0.55 | 0.55 |
| 1995 | 0.03 | 0.13 | 0.28 | 0.46 | 0.57 | 0.69 | 0.64 | 0.69 | 0.56 | 0.56 | 0.56 | 0.56 |
| 1996 | 0.02 | 0.12 | 0.25 | 0.41 | 0.50 | 0.60 | 0.56 | 0.60 | 0.49 | 0.49 | 0.49 | 0.49 |
| 1997 | 0.04 | 0.14 | 0.23 | 0.31 | 0.42 | 0.53 | 0.57 | 0.55 | 0.56 | 0.56 | 0.56 | 0.56 |
| 1998 | 0.03 | 0.12 | 0.19 | 0.26 | 0.34 | 0.43 | 0.46 | 0.45 | 0.46 | 0.46 | 0.46 | 0.46 |
| 1999 | 0.03 | 0.12 | 0.20 | 0.27 | 0.36 | 0.45 | 0.49 | 0.47 | 0.48 | 0.48 | 0.48 | 0.48 |
| 2000 | 0.03 | 0.13 | 0.21 | 0.28 | 0.38 | 0.47 | 0.51 | 0.50 | 0.51 | 0.51 | 0.51 | 0.51 |
| 2001 | 0.03 | 0.11 | 0.18 | 0.24 | 0.32 | 0.40 | 0.43 | 0.42 | 0.43 | 0.43 | 0.43 | 0.43 |
| 2002 | 0.03 | 0.12 | 0.19 | 0.26 | 0.35 | 0.44 | 0.47 | 0.46 | 0.47 | 0.47 | 0.47 | 0.47 |
| 2003 | 0.03 | 0.12 | 0.19 | 0.26 | 0.34 | 0.43 | 0.47 | 0.45 | 0.46 | 0.46 | 0.46 | 0.46 |
| 2004 | 0.05 | 0.21 | 0.26 | 0.28 | 0.27 | 0.26 | 0.30 | 0.37 | 0.43 | 0.43 | 0.43 | 0.43 |
| 2005 | 0.05 | 0.22 | 0.29 | 0.31 | 0.29 | 0.29 | 0.32 | 0.40 | 0.47 | 0.47 | 0.47 | 0.47 |
| 2006 | 0.05 | 0.24 | 0.31 | 0.33 | 0.31 | 0.31 | 0.35 | 0.43 | 0.51 | 0.51 | 0.51 | 0.51 |
| 2007 | 0.05 | 0.22 | 0.28 | 0.30 | 0.29 | 0.28 | 0.32 | 0.39 | 0.46 | 0.46 | 0.46 | 0.46 |
| 2008 | 0.06 | 0.25 | 0.32 | 0.34 | 0.33 | 0.33 | 0.37 | 0.45 | 0.53 | 0.53 | 0.53 | 0.53 |
| 2009 | 0.05 | 0.24 | 0.30 | 0.32 | 0.31 | 0.30 | 0.34 | 0.42 | 0.50 | 0.50 | 0.50 | 0.50 |
| 2010 | 0.05 | 0.21 | 0.27 | 0.28 | 0.27 | 0.27 | 0.30 | 0.37 | 0.44 | 0.44 | 0.44 | 0.44 |
| 2011 | 0.04 | 0.19 | 0.24 | 0.26 | 0.25 | 0.24 | 0.27 | 0.34 | 0.40 | 0.40 | 0.40 | 0.40 |
| 2012 | 0.04 | 0.18 | 0.23 | 0.25 | 0.23 | 0.23 | 0.26 | 0.32 | 0.38 | 0.38 | 0.38 | 0.38 |
| 2013 | 0.04 | 0.20 | 0.25 | 0.27 | 0.25 | 0.25 | 0.28 | 0.35 | 0.41 | 0.41 | 0.41 | 0.41 |
| 2014 | 0.03 | 0.15 | 0.19 | 0.20 | 0.20 | 0.19 | 0.22 | 0.27 | 0.32 | 0.32 | 0.32 | 0.32 |
| 2015 | 0.05 | 0.20 | 0.26 | 0.28 | 0.26 | 0.26 | 0.29 | 0.36 | 0.43 | 0.43 | 0.43 | 0.43 |

Table 8.11. Saithe in division Va. Input values for short-term projections. Same weights are used for catch weights and stock weights.

| 201 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| N | 26.3 | 16.2 | 21.6 | 16.0 | 11.4 | 6.0 | 3.7 | 1.5 | 0.6 | 0.7 | 0.6 | 0.1 |
| M | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| mat | 0.00 | 0.10 | 0.21 | 0.40 | 0.63 | 0.81 | 0.91 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
|  | 0 | 0 | 7 | 9 | 3 | 2 | 5 | 0 | 0 | 0 | 0 | 0 |
| w | 1.14 | 1.72 | 2.21 | 3.07 | 4.03 | 5.53 | 6.84 | 7.17 | 7.49 | 8.21 |  | 9.17 |
|  | 2 | 6 | 7 | 1 | 0 | 2 | 6 | 5 | 1 | 8 | 5 | 3 |
| sel | 0.10 | 0.47 | 0.60 | 0.64 | 0.62 | 0.61 | 0.68 | 0.84 | 1.00 | 1.00 | 1.00 | 1.00 |
|  | 6 | 6 | 8 | 8 | 1 | 1 | 7 | 6 | 0 | 0 | 0 | 0 |
| pF | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| pM | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6 |  |  |  |  |  |  |  |  |  |  |  |  |
| N | 31.9 |  |  |  |  |  |  |  |  |  |  |  |
| M | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| mat | 0.00 | 0.10 | 0.21 | 0.40 | 0.63 | 0.81 | 0.91 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
|  | 0 | 0 | 7 | 9 | 3 | 2 | 5 | 0 | 0 | 0 | 0 | 0 |
| w | 1.14 | 1.72 | 2.21 | 3.07 | 4.03 | 5.53 | 6.84 | 7.17 | 7.49 | 8.21 | 8.57 | 9.17 |
|  | 2 | 6 | 7 | 1 | 0 | 2 | 6 | 5 | 1 | 8 | 5 | 3 |
| sel | 0.10 | 0.47 | 0.60 | 0.64 | 0.62 | 0.61 | 0.68 | 0.84 | 1.00 | 1.00 | 1.00 | 1.00 |
|  | 6 | 6 | 8 | 8 | 1 | 1 | 7 | 6 | 0 | 0 | 0 | 0 |
| pF | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| pM | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| $7$ |  |  |  |  |  |  |  |  |  |  |  |  |
| N | 32.7 |  |  |  |  |  |  |  |  |  |  |  |
| M | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| mat | 0.00 | 0.10 | 0.21 | 0.40 | 0.63 | 0.81 | 0.91 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
|  | 0 | 0 | 7 | 9 | 3 | 2 | 5 | 0 | 0 | 0 | 0 | 0 |
| w | 1.14 | 1.72 | 2.21 | 3.07 | 4.03 | 5.53 | 6.84 | 7.17 | 7.49 | 8.21 | 8.57 | 9.17 |
|  | 2 | 6 | 7 | 1 | 0 | 2 | 6 | 5 | 1 | 8 | 5 | 3 |
| sel | 0.10 | 0.47 | 0.60 | 0.64 | 0.62 | 0.61 | 0.68 | 0.84 | 1.00 | 1.00 | 1.00 | 1.00 |
|  | 6 | 6 | 8 | 8 | 1 | 1 | 7 | 6 | 0 | 0 | 0 | 0 |
| pF | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| pM | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 8.12. Saithe in division Va. Output from short-term projections.

| 2015 |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| B4+ | SSB | Fbar | Landings |  |  |  |
| 255 | 139 | 0.26 | 57 |  |  |  |
|  |  |  |  | 2017 |  |  |
| 2016 |  |  |  | B4+ | SSB | Rationale |
| B4+ | SSB | Fbar | Landings |  |  |  |


| 238 | 138 | 0.26 | 54 | 227 | 130 | $20 \%$ HCR |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |

$20 \% \mathrm{HCR}=$ average between $0.2 \mathrm{~B} 4+$ (current year) and last year's TAC


Figure 8.1 Saithe in Division Va. Landings by gear.

Saithe - Apr 122015
Based on landings data from the Directorate of Fisheries


Figure 8.2 Saithe in division Va. Cumulative landings in the current fishing year (left) and calendar year (right). The vertical (green line) in the left figure shows the quota for the current fishing year.


Figure 8.3 Saithe in division Va. Weight at age in the catches, as relative deviations from the mean The current year's deviation is a preliminary prediction.


Figure 8.4 Saithe in division Va. Maturity at age used for calculating the SSB.


Figure 8.5 Saithe in division Va. Spring survey biomass index and model fit.


Figure 8.6. Estimated selectivity patterns for the 3 periods.

Std catch residuals


Figure 8.7. Saithe in division Va. Commercial and survey catch-at-age residuals from the fitted model. Filled circles are positive log residuals and hollow circles are negative log residuals.


Figure 8.8. Saithe in division Va. Results from the fitted model and short-term forecast. The red line indicates the time of the current assessment.


Figure 8.9. Saithe in division Va. Retrospective pattern for the assessment model.


Figure 8.10. Saithe in division Va. Comparison between the default separable model (ADSEP) and alternative assessment models.

## 9 Icelandic Cod

### 9.1 Summary

The spawning stock ( $\mathrm{SSB}_{2015}$ ) is estimated to be 547 kt and is higher than has been observed over the last five decades. The reference biomass ( $\mathrm{B}_{4+2015}$ ) is estimated to be 1302 kt , the highest observed since the late 1970's. Fishing mortality, being 0.3 in 2014, has declined significantly in recent years and is presently the lowest observed in last 6 decades. Year classes since the mid-1980s are estimated to be relatively stable but with the mean around the lower values observed in the period 1955 to 1985.

According to the adopted management harvest rule the TAC will be 239 kt in the next fishing season. ICES has evaluated the plan and concludes that it is in accordance with the precautionary approach and the ICES MSY framework.
Mean weight at age in the stock and the catches that were record low in 2006-8 have been increasing in recent years and are now around the long term mean.
The input in the analytical assessments are catch at age 1955-2014 and spring groundfish survey (SMB) indices at age from 1985-2015 and fall survey groundfish survey (SMH) indices at age from 1996-2014. The results from the AD-Model builder statistical Catch at Age Model (ADCAM) as was used as the final run. This framework has been the basis for the advice since 2002.

The reference stock ( $\mathrm{B}_{4}$ ) in 2014 is now estimated to be 1181 kt compared to 1106 kt last year. The SSB in 2014 is now estimated to be 425 kt compared to 427 kt estimated last year. Fishing mortality in 2013 is now estimated 0.29 compared to 0.3 estimated last year. Year classes 2011-2013 were estimated to be 181, 160 and 109 million in last years assessment and are now estimated to be 181, 161 and 115 millions.

### 9.2 Stock description and management units

The Icelandic cod stock is distributed all around Iceland and in the assessment cod within Icelandic EEZ waters it is assumed to be a single homogeneous unit in the assessment. Spawning takes place in late winter mainly off the south west coast but smaller, variable regional spawning components have also been observed all around Iceland. A slight but significant genetic difference has been observed between the cod spawning in the northern waters vs cod spawning in the southern waters (Pampoulie et al. 2007). There are indications that different behavioral type (shallow vs. deep migration) may be found within cod spawning in the same areas (Pampoulie et al. 2008). Both these information indicate that management measurements operating on a finer scale may be warranted, although appropriate non-ambiguous management measure in addition to maintaining low fishing mortality have not yet been identified.
The pelagic eggs and larvae from the main spawning grounds off the south west coast drift clockwise around the island to the main nursery grounds off the north coast. A larval drift to Greenland waters has been recorded in some years and substantial immigrations of mature cod from Greenland which are considered to be of Icelandic origin have been observed in some periods. This pattern was quite prevalent prior to 1970, while condition in Greenlandic waters are thought to have been favorable for cod productivity. Periodic immigrations have been estimated in the assessment from anomalies in the catch at age matrix with timing and age of such events being based on expert judgement using external information. The most recent of such migration was from the 1984 year class in 1990, the number estimated around 30 millions. Recent
tagging experiments as well as abnormal decline in survey indices in West Greenlandic waters indicate that part of the 2003 and to some extent the 2002 year classes may have migrated from Greenland to Icelandic waters. In the current assessment the immigration at age 6 in 2009 is estimated around 9.7 millions corresponding an additional biomass of around 31 kt in 2009. The influence of this immigration on the current biomass estimate is minimal.

Extensive tagging experiments spanning with some hiatuses over the last 100 year indicate that significant emigration from Iceland to other areas may be rare. In recent years it has been observed that cod tagged in Iceland has been recaptured inside Faroese waters close to the EEZ line separating Iceland and the Faroes islands. Anecdotal information from the fishing industry indicates that may be some exchange of cod across the Denmark Strait. These migrations may be of different nature than the hypothesized net "life history" immigration of cod described above.

### 9.3 Data

The data used for assessing Icelandic cod landings, catch-at-age composition and indices from standardized bottom trawl surveys. The sampling programs i.e log books, surveys, sampling from landings etc. have been described in previous reports.

### 9.4 Landings

Landings of Icelandic cod in 2014 are estimated to have been 221.343 kt of which 219.682 kt were taken by Icelandic fleet. The landings by gear and month were as follows:

|  | 1 |  |  |  | 3 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 11 | 12 | Total | Proportion |  |  |  |  |  |  |  |  |  |  |
| Longline | 8544 | 6053 | 6553 | 5695 | 6666 | 4065 | 2591 | 3338 | 8364 | 10047 | 11468 | 5424 | 78809 |
| Gill net | 2526 | 4837 | 6522 | 2825 | 790 | 76 | 62 | 142 | 294 | 237 | 482 | 250 | 19043 |
| Hooks | 28 | 161 | 1178 | 1448 | 2872 | 3665 | 3603 | 1874 | 709 | 189 | 100 | 4 | 15831 |
| Danis seine | 574 | 1754 | 2440 | 1206 | 952 | 363 | 229 | 286 | 1032 | 768 | 573 | 176 | 10353 |
| Trawl | 8489 | 6883 | 8514 | 7052 | 9094 | 6168 | 4041 | 4577 | 9881 | 11913 | 11601 | 9094 | 97307 |
|  | 20161 | 19688 | 25207 | 18226 | 20374 | 14337 | 10526 | 10217 | 20280 | 23154 | 24224 | 14948 | 221343 |

Historically the landings of bottom trawlers constituted a larger portion of the total catches than today, in some years prior to 1990 reaching $60 \%$ of the total landings. In the 1990's the landings from bottom trawlers declined significantly within a period of 5 years, and have been just above $40 \%$ of the total landings in the last decade. (Figure 9.1). The share of long line has tripled over the last 20 years and is now on par with bottom trawl. The share of gill net has over the same time period declined and is now only half of what it was in the 1980's.

The trawl fishery is high aggregated with large proportion of the catches taken along the northeast and southeast continental shelf slope boundary (Figure 9.2) while the long line fishery is much more dispersed (Figure 9.3).

The trend in landings in last two decades is largely a reflection of the TAC that is set for the fishing year (starting 1. September and ending 31 august).

According to the HCR the TAC for the fishing year 2013/14 was supposed to be capped to 215 kt . Landings of the Icelandic fleet was however 225.081. Including additional landings from the foreign fleet this amounts to an overshoot of some $4.7 \%$.

The estimate of landings for the current fishing year is 220 thousand tones, this being based on the following:

| Category | Amount |
| :--- | ---: |
| ITQ holders | 204.552 |
| Shell and shrimp | 1.385 |
| Village support | 5.094 |
| Long line allowance | 3.375 |
| Beach fishing | 7.500 |
| Sport fishing | 0.200 |
| Village support II | 2.404 |
| Recalled | -8.200 |
| Estimated foreign | 2.000 |
| Miscallaneous | 2.000 |
| Total | 220.310 |

The catches in the first four months of the current fishing year (September - December 2014) were 83 kt . The remainder of the estimated catch in the fishing year ( $220 \mathrm{kt} \mathrm{)} \mathrm{is}$ 137 kt . Assuming that the same proportion of the allowable catch for the next fishing year is taken in the first four months (September-December 2014) as last year landings (some 0.38), the catch in 2015 is estimated to be 234 kt .
Mean annual discard of cod over the period 2001-2012 is around $1 \%$ of landings (Ólafur Pálsson et al. 2013). The method used for deriving these estimates assumes that discarding only occurs as high grading.

### 9.5 Catch in numbers and weight by age

### 9.5.1 Catch in numbers by age

Sampling protocol for estimating the age composition of the cod has been in effect since 1991 and has been described previous reports. The method for deriving the catch at age are described in the stock annex. The catch at age matrix is reasonably consistent (Table 9.1), with CV estimated to be approximately 0.2 for age groups 4-10 based on a Shepherd-Nicholson model.

### 9.5.2 Mean weight at age in the landings

The mean weight age in the landings (Table 9.2 and Figure 9.4) declined from 2001 to 2007, reaching then a historical low in many age groups. The weight at age have been increasing in recent years and are in 2014 around the average weights observed over the period from 1985 and close to the long term mean (1955-2014). The variation in the pattern of weight at age in the catches is in part a reflection of the variation in the weight in the stock as seen in the measurements from the spring survey (Table 9.5 and Figure 9.5). The latest spring survey weight measurements (in 2015) are below average in younger ages but above average in older ages.

The reference biomass ( $\mathrm{B}_{4+}$ ) upon which the TAC in the fishing year is set (based on the $H C R)$ is derived from population numbers and catch weights in the beginning of the assessment year. In recent years the estimates of mean weights in the landings of age groups 3-9 in the assessment years have been based on a prediction from the spring survey weight measurements in that year using the slope and the intercept from a linear relationship between survey and catch weights in preceding year. The same approach was used this year for predicting weight at age in the catches for 2015. I.e. the alpha and beta were estimated from :
$c W_{a, 2014}=a l p h a+$ beta $\quad s W_{a, 2014}$
and the catch weights for 2015 then from:
$c W_{a, 2015}=a l p h a+$ beta $\quad s W_{a, 2015}$
Based on this the mean weights at age in the catches in 2015 are predicted to be at or above average (Figure [fig:cW01]).

### 9.6 Surveys

### 9.6.1 Length based indices

The total biomass indices from the spring (SMB) and the fall (SMH) surveys (Figure 9.6) indicate that the stock biomass has been increasing substantially in recent years and is in the last 4 years among highest since the start of the spring survey in 1985. The increase in biomass is most pronounced in larger fish.

### 9.6.2 Age based indices

Abundance indices by age from the spring and the fall surveys (Tables 9.6 and 9.8). Indices of older fish are all relatively high in recent years despite the indices of these year classes when younger are low or moderate in size (Figure 9.7).

The variance of age groups 5-9 was abnormally high in the spring 2012 survey but the value for the last three years being normal (Table 9.9). This high cv is in part attributed to one haul having extremely high cod catches. In previous NWWG report it was shown that the influence of the large haul did not have a significant effect on the assessment, this being attributed to survey residuals in a given year being modeled by a multivariate normal distribution (see stock annex).

### 9.7 Commercial cpue and effort

Unstandardised CPUE and effort indices, based on log book records were not considered during this meeting. In previous reports it has been concluded that changes in these parameters, although to some extent a reflection of the dynamics in the stock they are confounded by other factors.

### 9.8 Assessment

Last year, the results from a statistical catch at age model (sometimes refer to as ADCAM) tuned with the spring and the fall survey was used as the final point estimator upon which advice was based (referred to as the SPALY model in the text that follows). In this framework the catch at age are modelled and the fishing mortality changes gradually over time, constrained by a random walk (further explanation of the model setup are provided in the stock annex). In addition to the above model, the data have also been extensively explored in the TSA framework, using a Time Series Analysis developed and run by Guðmundur Guðmundsson (1994, 2004, details of model description are given in WD 29, NWWG 2013). Models where the catch/fishing mortality is not modelled (ADAPT) and where the fishing pattern is not considered to change each year (SEPARABLE) are also routinely run for comparative purpose.

The SPALY framework from last year, i.e. tuning with both the spring and the fall survey using ADCAM show similar diagnostics as that observed last years (see Tables 9.19, 9.11 and 9.12 and Figure 9.8 for the residuals). A negative residual block for spring survey indices age groups 2 to 5 in recent years may indicate that there may have been
some change in catchability. The detailed result from the SPALY ADCAM run are provided in Tables 9.13, 9.14 and the stock summary in Table 9.15 and Figure 9.9. The reference biomass is estimated to be 1302 kt in 2015 and the fishing mortality 0.28 in 2014.

Assessment based on ADCAM tuning with the spring and the fall survey separately have in recent years shown that the fall survey gives a higher estimate than the spring survey (Figure 9.11). Tuning with spring survey only this year resulted in a reference biomass of 1186 kt in 2015 and a fishing mortality of 0.32 in 2014. An assessment based on the fall survey only gave reference biomass of 1378 kt in 2015 and fishing mortality of 0.32 in 2014. There is hence some conflict with respect to the extent of the increase in the biomass and reduction in fishing mortality in recent years between the two survey input sources. In addition there are conflict in the signals between the surveys on one hand and the catch at age matrix, the year classes declining at somewhat faster rate in the fisheries than in the surveys.

Although there are indication that there may be violation in the SPALY ADCAM setup it was considered premature to base the advice this year on one of the alternative models setup and assumptions. If the true reference biomass in 2015 is around 1000 kt and the TAC is set at 239 kt it is equivalent to the decision being based on a 0.24 harvest rate. If this becomes the realized harvest in 2017 it is still among the lowest observed historical rates.

### 9.9 Reference points and stock classification

In 2010 ACOM set the $B_{\lim }$ as 125 kt based on recommendation of the NWWG. The basis for $\mathrm{Blim}_{\text {lim }}$ is $\mathrm{B}_{\text {loss }}$ and/or the SSBbreak in a segmented regression based on recruitment from year classes 1952-1984 on one hand and recruitment from year classes 1985 onwards on the other hand (Figure 9.11). The splitting of the recruitment time series is based on the hypothesis that recruitment productivity as a function of spawning stock biomass, as it is presently measured, is lower in latter period compared with the former period.
An harvest rate limit point derived deterministically from $B_{\lim }$ according to the methodology outlined in SGMAS 2006 indicates that it is in the vicinity of 0.35.
$B_{p a}$ and $F_{p a}$ have never been set for this stock. Based on the ICES default methodology for the derivation of $\mathrm{B}_{\mathrm{pa}}$ and $\mathrm{F}_{\mathrm{pa}}$ from $\mathrm{B}_{\lim }$ and $\mathrm{F}_{\text {lim }}$ these reference points would be somewhere in the vicinity of: $\mathrm{B}_{\mathrm{pa}}=1.4 \mathrm{Blim}_{\lim }=175 \mathrm{kt} \mathrm{HR} \mathrm{pa}_{\mathrm{pa}}=\mathrm{HR} \lim / 1.4=0.25$
The $B_{\text {trigger }}$ and the HR $\quad$ HCR in the HCR are thus respectively above and below the default candidate PA-reference points. Given the current ICES MSY framework, upon which the HCR for iCod has been evaluated, definition of PA-reference points may be deemed as redundant. The NWWG does not suggest a formal establishment of PA-points for iCod at this point.
$\mathrm{F}_{\text {msy }}$ or $\mathrm{HR}_{\mathrm{msy}}$ point estimate, to be used in stock status classification in the advisory text has not been defined for this stock. The harvest rate in the management control rule upon which the TAC is based (landings being based on $20 \%$ of the reference biomass ( $\mathrm{B}_{4+}$ ) and the TAC in the current fishing year) has been deemed by ICES to be in conformity with the ICES PA and MSY approach. The distribution of the realized harvest rate when the HCR is followed show that the $90 \%$ expected range are within a harvest rate of $0.15-0.27$ (Figure 9.13). The recent realized harvest rates are within that range.

The harvest rate (no $\mathrm{B}_{\text {trigger }}$ ) upon which the a catch advise is based that results in maximum sustainable yield has been estimated to be 0.22 , with a $90 \%$ realized range of
$0.16-0.31$. Note that the method used in the derivation of these statistics are not that same as that used to define $\mathrm{F}_{\text {msy }}$ ranges in the WKMSYREF3.

### 9.10 State of the stock

The spawning stock reached a historical low in 1993 (120 kt point estimates) but has since then increased and estimated to be 547 kt at present 9.9, Table 9.15). A spawning stock biomass above the current estimates has not been observed since the early the 1960's. This increase in biomass of older fish occurs despite productivity in terms of recruitment of the year classes now contributing to the spawning stock (2001-2010) having been relatively low (geometric mean $=129$ ). The driving factor is hence attributed to a significant decline in fishing mortality/exploitation rate in recent years, being at present within the same order as observed in the beginning of the time series. The average size of the most recent year classes that are estimated (2011-2014) is 158 millions, this being close to the long term mean (164 million).

### 9.11 Short term deterministic forecast

Input: The stock in numbers in 2015 (Table 9.16) for year classes 2014 and older are obtained from the current assessment (Table 9.14). Given the current harvest control rule, where the TAC 2015/2016 is determined from the $B_{4+, 2015 \text {, the only additional pre- }}$ diction needed is the estimates of weights in 2015. These were described in section 9.3.2. Hence there is no need to carry through a short term prediction so what follows is just to keep up with the ICES convention.

Additional assumptions used in the deterministic predictions are as usual: Weights and proportion mature in the spawning stock from 2015 onwards were kept constant. The fishing pattern used is the average of the years 2011-2014. The estimated landings for the calendar year 2015 is 227 kt as discussed in section 9.3.1. Details of the inputs values are provided in Table 9.16.

Output: The estimated reference biomass in 2015 is 1302 kt . The TAC in the current fishing year is 218 kt . According to the management harvest control rule, given that the current SSB estimates are above the SSB $_{\text {trigger }}(220 \mathrm{kt})$, the TAC in the next fishing year is:

$$
T A C_{2015 / 2016}=0.5 \quad 218+0.10 \quad 1300=239 k t
$$

### 9.12 Stochastic forecast

Medium term forecasts up to year 2019 was run on the three ADCAM runs, both surveys in the tuning (SPALY), spring survey only and fall survey only. The platform used is the same as used in the assessment. Harvest rate of 0.20 (for the SPALY assessment), 0.22 (for the SMB tuned assessment) and 0.18 (for the SMH tuned assessment) were used in the future, the latter two to accounting for potential under or overestimation of the stock.

The analysis indicate there is high probability that the spawning stock size is and will remain above $B_{\text {trigger }}\left(220 \mathrm{kt}\right.$ ) and $\mathrm{B}_{\lim }(125 \mathrm{kt})$ (Figure 9.10).

### 9.13 Uncertainties in assessment and forecast

Alternative model assumptions indicate that the reference biomass may be around range from $1000-1300 \mathrm{kt}$ in 2014, compared with the 1100 kt estimated from the SPALY model. The lower alternative state of nature implies that the reference biomass
upon which the TAC is set may be $20 \%$ lower than used and that the realized harvest rate could materialize to be $24 \%$ given a TAC of 218 kt . According to the HCR evaluation (ICES 2009), this is close to the upper bounds of expected harvest rates.

### 9.14 Comparison with previous assessment, forecast and advise

The reference stock ( $\mathrm{B}_{4}$ ) in 2014 is now estimated to be 1181 kt compared to 1106 kt last year. The SSB in 2014 is now estimated to be 425 kt compared to 427 kt estimated last year. Fishing mortality in 2013 is now estimated 0.29 compared to 0.3 estimated last year. Year classes 2011-2013 were estimated to be 181, 160 and 109 million in last years assessment and are now estimated to be 181, 161 and 115 millions.

A standard ICES retrospective plots (Figure 9.12) show estimates of key metrics in recent years compared with current estimates.

The basis for the advice this year is the same as last year: the management plan/MSY/precautionary approach.

### 9.15 Management considerations

Prior to allocating quota to the Icelandic fleet that is under the ITQ control, the managers should ensure subtracting all estimated catches from other sources, including any landings arising from new regulations.

### 9.16 Regulations and their effects

Exploitation rate and fishing mortality have been reduced significantly after the implementation of the catch rule in 1995 compared with the past. I.e. management measures by restricting landings based on the HCR are manifested in lower fishing mortality and higher stock biomass for the iCod.

A quick closure system has been in force since 1976, aimed at protecting juvenile fish. Fishing is prohibited, for at least two weeks, in areas where the number of small cod (< 55 cm ) in the catches has been observed by inspectors to exceed $25 \%$. A preliminary evaluation of the effectiveness of the system indicates that the relatively small areas closed for a short time do most likely not contribute much to the protection of juveniles. On the other hand, several consecutive quick closures often lead to closures of larger areas for a longer time and force the fleet to operate in other areas. The effect of these longer closures has not been evaluated analytically.

Since 1995, spawning areas have been closed for 2-3 weeks during the spawning season for all fisheries. The intent of this measure was to protect spawning activity of fish. In 2005, the maximum allowed mesh size in gill nets was decreased to 8 inches in order to protect the largest spawners.

The mesh size in the cod-end in the trawling fishery was increased from 120 mm to 155 mm in 1977. Since 1998 the minimum cod-end mesh size allowed is 135 mm , provided that a so-called Polish cover is not used. Numerous areas are closed temporarily or permanently for all fisheries or specific gears for protecting juveniles and habitat, or for socio-political reasons. The effects of these measures have not been evaluated.

### 9.17 Changes in fishing technology and fishing patterns

Changes in the importance of the various gears used to catch cod are described in section 9.3. The decline in the gill net fishery is likely to have resulted in overall shift in the fishing pattern away from the largest fish. The increase in the long line fishery in
the north was partly the reason for the decline in the observed mean weight at age of oldest fish in the catches.

### 9.18 Environmental influence on the stock

Environmental influence on the stock are partly integrated in the annual input data for the analytical assessment, both in terms of weight and stock indices. The causation is however poorly understood.

An increased inflow of Atlantic water has been observed in Icelandic waters since 1997, resulting in higher temperature and higher salinity. A northward shift in distribution of immature capelin may be linked to these hydrographical changes, resulting in lower availability of capelin for cod. In the past low weights-at-age of cod have been related to a low biomass of capelin. The increase in mean weight-at-age in cod in recent years may, however, have more to do with reduction in fishing mortality than with changes in availability of capelin.

Table 9.1: Icelandic cod in Division Va. Estimateded catch in numbers by year and age in millions of fish in 1955-2013.

| year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1955 | 4.790 | 25.164 | 46.566 | 28.287 | 10.541 | 5.224 | 2.467 | 25.182 | 2.101 | 1.202 | 1.668 | 0.665 |
| 1956 | 6.709 | 17.265 | 31.030 | 27.793 | 14.389 | 4.261 | 3.429 | 2.128 | 16.820 | 1.552 | 1.522 | 1.545 |
| 1957 | 13.240 | 21.278 | 17.515 | 24.569 | 17.634 | 12.296 | 3.568 | 2.169 | 1.171 | 6.822 | 0.512 | 1.089 |
| 1958 | 25.237 | 30.742 | 14.298 | 10.859 | 15.997 | 15.822 | 12.021 | 2.003 | 2.125 | 0.771 | 3.508 | 0.723 |
| 1959 | 18.394 | 37.650 | 23.901 | 7.682 | 5.883 | 8.791 | 13.003 | 7.683 | 0.914 | 0.990 | 0.218 | 1.287 |
| 1960 | 14.830 | 28.642 | 27.968 | 14.120 | 8.387 | 6.089 | 6.393 | 11.600 | 3.526 | 0.692 | 0.183 | 0.510 |
| 1961 | 16.507 | 21.808 | 19.488 | 15.034 | 7.900 | 6.925 | 3.969 | 3.211 | 6.756 | 1.202 | 0.089 | 0.425 |
| 1962 | 13.514 | 28.526 | 18.924 | 14.650 | 12.045 | 4.276 | 8.809 | 2.664 | 1.883 | 2.988 | 0.405 | 0.324 |
| 1963 | 18.507 | 28.466 | 19.664 | 11.314 | 15.682 | 7.704 | 2.724 | 6.508 | 1.657 | 1.030 | 1.372 | 0.246 |
| 1964 | 19.287 | 28.845 | 18.712 | 11.620 | 7.936 | 18.032 | 5.040 | 1.437 | 2.670 | 0.655 | 0.370 | 1.025 |
| 1965 | 21.658 | 29.586 | 24.783 | 11.706 | 9.334 | 6.394 | 11.122 | 1.477 | 0.823 | 0.489 | 0.118 | 0.489 |
| 1966 | 17.910 | 30.649 | 20.006 | 13.872 | 5.942 | 7.586 | 2.320 | 5.583 | 0.407 | 0.363 | 0.299 | 0.311 |
| 1967 | 25.945 | 27.941 | 24.322 | 11.320 | 8.751 | 2.595 | 5.490 | 1.392 | 1.998 | 0.109 | 0.030 | 0.106 |
| 1968 | 11.933 | 47.311 | 22.344 | 16.277 | 15.590 | 7.059 | 1.571 | 2.506 | 0.512 | 0.659 | 0.047 | 0.098 |
| 1969 | 11.149 | 23.925 | 45.445 | 17.397 | 12.559 | 14.811 | 1.590 | 0.475 | 0.340 | 0.064 | 0.024 | 0.021 |
| 1970 | 9.876 | 47.210 | 23.607 | 25.451 | 15.196 | 12.261 | 14.469 | 0.567 | 0.207 | 0.147 | 0.035 | 0.050 |
| 1971 | 13.060 | 35.856 | 45.577 | 21.135 | 17.340 | 10.924 | 6.001 | 4.210 | 0.237 | 0.069 | 0.038 | 0.020 |
| 1972 | 8.973 | 29.574 | 30.918 | 22.855 | 11.097 | 9.784 | 10.538 | 3.938 | 1.242 | 0.119 | 0.031 | 0.001 |
| 1973 | 36.538 | 25.542 | 27.391 | 17.045 | 12.721 | 3.685 | 4.718 | 5.809 | 1.134 | 0.282 | 0.007 | 0.001 |
| 1974 | 14.846 | 61.826 | 21.824 | 14.413 | 8.974 | 6.216 | 1.647 | 2.530 | 1.765 | 0.334 | 0.062 | 0.028 |
| 1975 | 29.301 | 29.489 | 44.138 | 12.088 | 9.628 | 3.691 | 2.051 | 0.752 | 0.891 | 0.416 | 0.060 | 0.046 |
| 1976 | 23.578 | 39.790 | 21.092 | 24.395 | 5.803 | 5.343 | 1.297 | 0.633 | 0.205 | 0.155 | 0.065 | 0.029 |
| 1977 | 2.614 | 42.659 | 32.465 | 12.162 | 13.017 | 2.809 | 1.773 | 0.421 | 0.086 | 0.024 | 0.006 | 0.002 |
| 1978 | 5.999 | 16.287 | 43.931 | 17.626 | 8.729 | 4.119 | 0.978 | 0.348 | 0.119 | 0.048 | 0.015 | 0.027 |
| 1979 | 7.186 | 28.427 | 13.772 | 34.443 | 14.130 | 4.426 | 1.432 | 0.350 | 0.168 | 0.043 | 0.024 | 0.004 |
| 1980 | 4.348 | 28.530 | 32.500 | 15.119 | 27.090 | 7.847 | 2.228 | 0.646 | 0.246 | 0.099 | 0.025 | 0.004 |
| 1981 | 2.118 | 13.297 | 39.195 | 23.247 | 12.710 | 26.455 | 4.804 | 1.677 | 0.582 | 0.228 | 0.053 | 0.068 |
| 1982 | 3.285 | 20.812 | 24.462 | 28.351 | 14.012 | 7.666 | 11.517 | 1.912 | 0.327 | 0.094 | 0.043 | 0.011 |
| 1983 | 3.554 | 10.910 | 24.305 | 18.944 | 17.382 | 8.381 | 2.054 | 2.733 | 0.514 | 0.215 | 0.064 | 0.037 |
| 1984 | 6.750 | 31.553 | 19.420 | 15.326 | 8.082 | 7.336 | 2.680 | 0.512 | 0.538 | 0.195 | 0.090 | 0.036 |
| 1985 | 6.457 | 24.552 | 35.392 | 18.267 | 8.711 | 4.201 | 2.264 | 1.063 | 0.217 | 0.233 | 0.102 | 0.038 |
| 1986 | 20.642 | 20.330 | 26.644 | 30.839 | 11.413 | 4.441 | 1.771 | 0.805 | 0.392 | 0.103 | 0.076 | 0.044 |
| 1987 | 11.002 | 62.130 | 27.192 | 15.127 | 15.695 | 4.159 | 1.463 | 0.592 | 0.253 | 0.142 | 0.046 | 0.058 |
| 1988 | 6.713 | 39.323 | 55.895 | 18.663 | 6.399 | 5.877 | 1.345 | 0.455 | 0.305 | 0.157 | 0.114 | 0.025 |
| 1989 | 2.605 | 27.983 | 50.059 | 31.455 | 6.010 | 1.915 | 0.881 | 0.225 | 0.107 | 0.086 | 0.038 | 0.005 |
| 1990 | 5.785 | 12.313 | 27.179 | 44.534 | 17.037 | 2.573 | 0.609 | 0.322 | 0.118 | 0.050 | 0.015 | 0.020 |
| 1991 | 8.554 | 25.131 | 15.491 | 21.514 | 25.038 | 6.364 | 0.903 | 0.243 | 0.125 | 0.063 | 0.011 | 0.012 |
| 1992 | 12.217 | 21.708 | 26.524 | 11.413 | 10.073 | 8.304 | 2.006 | 0.257 | 0.046 | 0.032 | 0.009 | 0.008 |
| 1993 | 20.500 | 33.078 | 15.195 | 13.281 | 3.583 | 2.785 | 2.707 | 1.181 | 0.180 | 0.034 | 0.011 | 0.013 |
| 1994 | 6.160 | 24.142 | 19.666 | 6.968 | 4.393 | 1.257 | 0.599 | 0.508 | 0.283 | 0.049 | 0.018 | 0.006 |
| 1995 | 10.770 | 9.103 | 16.829 | 13.066 | 4.115 | 1.596 | 0.313 | 0.184 | 0.156 | 0.141 | 0.029 | 0.008 |
| 1996 | 5.356 | 14.886 | 7.372 | 12.307 | 9.429 | 2.157 | 0.837 | 0.208 | 0.076 | 0.065 | 0.055 | 0.005 |
| 1997 | 1.722 | 16.442 | 17.298 | 6.711 | 7.379 | 5.958 | 1.147 | 0.493 | 0.126 | 0.028 | 0.037 | 0.021 |
| 1998 | 3.458 | 7.707 | 25.394 | 20.167 | 5.893 | 3.856 | 2.951 | 0.500 | 0.196 | 0.055 | 0.033 | 0.013 |
| 1999 | 2.525 | 19.554 | 15.226 | 24.622 | 12.966 | 2.795 | 1.489 | 0.748 | 0.140 | 0.046 | 0.010 | 0.005 |
| 2000 | 10.493 | 6.581 | 29.080 | 11.227 | 11.390 | 5.714 | 1.104 | 0.567 | 0.314 | 0.074 | 0.022 | 0.006 |
| 2001 | 11.338 | 25.040 | 9.311 | 19.471 | 5.620 | 3.929 | 2.017 | 0.452 | 0.202 | 0.118 | 0.013 | 0.009 |
| 2002 | 5.934 | 18.482 | 24.297 | 6.874 | 8.943 | 2.227 | 1.353 | 0.689 | 0.123 | 0.040 | 0.041 | 0.002 |
| 2003 | 3.950 | 16.160 | 21.874 | 18.145 | 5.063 | 4.419 | 1.124 | 0.401 | 0.172 | 0.034 | 0.020 | 0.015 |
| 2004 | 1.778 | 19.184 | 25.003 | 17.384 | 9.926 | 2.734 | 2.023 | 0.481 | 0.126 | 0.062 | 0.014 | 0.005 |
| 2005 | 5.102 | 5.125 | 26.749 | 16.980 | 8.339 | 4.682 | 1.292 | 0.913 | 0.203 | 0.089 | 0.025 | 0.002 |
| 2006 | 3.258 | 12.884 | 8.438 | 22.041 | 10.418 | 4.523 | 2.194 | 0.497 | 0.336 | 0.067 | 0.027 | 0.002 |
| 2007 | 2.074 | 11.961 | 15.948 | 8.280 | 9.593 | 5.428 | 2.205 | 1.229 | 0.366 | 0.198 | 0.053 | 0.010 |
| 2008 | 2.616 | 4.850 | 12.585 | 11.973 | 5.238 | 4.582 | 2.040 | 0.831 | 0.308 | 0.053 | 0.037 | 0.004 |
| 2009 | 3.660 | 8.150 | 9.480 | 17.330 | 10.060 | 3.910 | 2.290 | 0.770 | 0.310 | 0.090 | 0.020 | 0.010 |
| 2010 | 3.174 | 7.219 | 9.385 | 8.692 | 10.695 | 5.588 | 1.599 | 1.095 | 0.337 | 0.197 | 0.071 | 0.016 |
| 2011 | 4.780 | 7.257 | 9.284 | 10.735 | 6.032 | 6.152 | 2.361 | 0.666 | 0.459 | 0.151 | 0.041 | 0.010 |
| 2012 | 3.839 | 10.010 | 10.400 | 9.435 | 8.866 | 4.834 | 3.206 | 1.269 | 0.369 | 0.218 | 0.101 | 0.030 |
| 2013 | 5.141 | 12.299 | 14.846 | 11.212 | 7.358 | 5.643 | 2.688 | 1.930 | 0.675 | 0.289 | 0.156 | 0.052 |
| 2014 | 5.263 | 7.371 | 13.304 | 12.984 | 8.831 | 4.829 | 3.112 | 1.570 | 1.027 | 0.360 | 0.100 | 0.089 |

Table 9.2: Icelandic cod in Division Va. Estimated mean weight at age in the landings ( kg ) in period the 1955-2014. The weights for age groups 3 to 9 in 2015 are based on predictions from the 2015 spring survey measurements. The weights in the catches are used to calculate the reference biomass ( $B_{4+}$ ).

| year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1955 | 0.827 | 1.307 | 2.157 | 3.617 | 4.638 | 5.657 | 6.635 | 6.168 | 8.746 | 8.829 | 10.086 | 14.584 |
| 1956 | 1.080 | 1.600 | 2.190 | 3.280 | 4.650 | 5.630 | 6.180 | 6.970 | 6.830 | 9.290 | 10.965 | 12.954 |
| 1957 | 1.140 | 1.710 | 2.520 | 3.200 | 4.560 | 5.960 | 7.170 | 7.260 | 8.300 | 8.290 | 10.350 | 13.174 |
| 1958 | 1.210 | 1.810 | 3.120 | 4.510 | 5.000 | 5.940 | 6.640 | 8.290 | 8.510 | 8.840 | 9.360 | 13.097 |
| 1959 | 1.110 | 1.950 | 2.930 | 4.520 | 5.520 | 6.170 | 6.610 | 7.130 | 8.510 | 8.670 | 9.980 | 11.276 |
| 1960 | 1.060 | 1.720 | 2.920 | 4.640 | 5.660 | 6.550 | 6.910 | 7.140 | 7.970 | 10.240 | 10.100 | 12.871 |
| 1961 | 1.020 | 1.670 | 2.700 | 4.330 | 5.530 | 6.310 | 6.930 | 7.310 | 7.500 | 8.510 | 9.840 | 14.550 |
| 1962 | 0.990 | 1.610 | 2.610 | 3.900 | 5.720 | 6.660 | 6.750 | 7.060 | 7.540 | 8.280 | 10.900 | 12.826 |
| 1963 | 1.250 | 1.650 | 2.640 | 3.800 | 5.110 | 6.920 | 7.840 | 7.610 | 8.230 | 9.100 | 9.920 | 11.553 |
| 1964 | 1.210 | 1.750 | 2.640 | 4.020 | 5.450 | 6.460 | 8.000 | 9.940 | 9.210 | 10.940 | 12.670 | 15.900 |
| 1965 | 1.020 | 1.530 | 2.570 | 4.090 | 5.410 | 6.400 | 7.120 | 8.600 | 12.310 | 10.460 | 10.190 | 17.220 |
| 1966 | 1.170 | 1.680 | 2.590 | 4.180 | 5.730 | 6.900 | 7.830 | 8.580 | 9.090 | 14.230 | 14.090 | 17.924 |
| 1967 | 1.120 | 1.820 | 2.660 | 4.067 | 5.560 | 7.790 | 7.840 | 8.430 | 9.090 | 10.090 | 14.240 | 16.412 |
| 1968 | 1.170 | 1.590 | 2.680 | 3.930 | 5.040 | 5.910 | 7.510 | 8.480 | 10.750 | 11.580 | 14.640 | 16.011 |
| 1969 | 1.100 | 1.810 | 2.480 | 3.770 | 5.040 | 5.860 | 7.000 | 8.350 | 8.720 | 10.080 | 11.430 | 13.144 |
| 1970 | 0.990 | 1.450 | 2.440 | 3.770 | 4.860 | 5.590 | 6.260 | 8.370 | 10.490 | 12.310 | 14.590 | 21.777 |
| 1971 | 1.090 | 1.570 | 2.310 | 2.980 | 4.930 | 5.150 | 5.580 | 6.300 | 8.530 | 11.240 | 14.740 | 17.130 |
| 1972 | 0.980 | 1.460 | 2.210 | 3.250 | 4.330 | 5.610 | 6.040 | 6.100 | 6.870 | 8.950 | 11.720 | 16.000 |
| 1973 | 1.030 | 1.420 | 2.470 | 3.600 | 4.900 | 6.110 | 6.670 | 6.750 | 7.430 | 7.950 | 10.170 | 17.000 |
| 1974 | 1.050 | 1.710 | 2.430 | 3.820 | 5.240 | 6.660 | 7.150 | 7.760 | 8.190 | 9.780 | 12.380 | 14.700 |
| 1975 | 1.100 | 1.770 | 2.780 | 3.760 | 5.450 | 6.690 | 7.570 | 8.580 | 8.810 | 9.780 | 10.090 | 11.000 |
| 1976 | 1.350 | 1.780 | 2.650 | 4.100 | 5.070 | 6.730 | 8.250 | 9.610 | 11.540 | 11.430 | 14.060 | 16.180 |
| 1977 | 1.259 | 1.911 | 2.856 | 4.069 | 5.777 | 6.636 | 7.685 | 9.730 | 11.703 | 14.394 | 17.456 | 24.116 |
| 1978 | 1.289 | 1.833 | 2.929 | 3.955 | 5.726 | 6.806 | 9.041 | 10.865 | 13.068 | 11.982 | 19.062 | 21.284 |
| 1979 | 1.408 | 1.956 | 2.642 | 3.999 | 5.548 | 6.754 | 8.299 | 9.312 | 13.130 | 13.418 | 13.540 | 20.072 |
| 1980 | 1.392 | 1.862 | 2.733 | 3.768 | 5.259 | 6.981 | 8.037 | 10.731 | 12.301 | 17.281 | 14.893 | 19.069 |
| 1981 | 1.180 | 1.651 | 2.260 | 3.293 | 4.483 | 5.821 | 7.739 | 9.422 | 11.374 | 12.784 | 12.514 | 19.069 |
| 1982 | 1.006 | 1.550 | 2.246 | 3.104 | 4.258 | 5.386 | 6.682 | 9.141 | 11.963 | 14.226 | 17.287 | 16.590 |
| 1983 | 1.095 | 1.599 | 2.275 | 3.021 | 4.096 | 5.481 | 7.049 | 8.128 | 11.009 | 13.972 | 15.882 | 18.498 |
| 1984 | 1.288 | 1.725 | 2.596 | 3.581 | 4.371 | 5.798 | 7.456 | 9.851 | 11.052 | 14.338 | 15.273 | 16.660 |
| 1985 | 1.407 | 1.971 | 2.576 | 3.650 | 4.976 | 6.372 | 8.207 | 10.320 | 12.197 | 14.683 | 16.175 | 19.050 |
| 1986 | 1.459 | 1.961 | 2.844 | 3.593 | 4.635 | 6.155 | 7.503 | 9.084 | 10.356 | 15.283 | 14.540 | 15.017 |
| 1987 | 1.316 | 1.956 | 2.686 | 3.894 | 4.716 | 6.257 | 7.368 | 9.243 | 10.697 | 10.622 | 15.894 | 12.592 |
| 1988 | 1.438 | 1.805 | 2.576 | 3.519 | 4.930 | 6.001 | 7.144 | 8.822 | 9.977 | 11.732 | 14.156 | 13.042 |
| 1989 | 1.186 | 1.813 | 2.590 | 3.915 | 5.210 | 6.892 | 8.035 | 9.831 | 11.986 | 10.003 | 12.611 | 16.045 |
| 1990 | 1.290 | 1.704 | 2.383 | 3.034 | 4.624 | 6.521 | 8.888 | 10.592 | 10.993 | 14.570 | 15.732 | 17.290 |
| 1991 | 1.309 | 1.899 | 2.475 | 3.159 | 3.792 | 5.680 | 7.242 | 9.804 | 9.754 | 14.344 | 14.172 | 20.200 |
| 1992 | 1.289 | 1.768 | 2.469 | 3.292 | 4.394 | 5.582 | 6.830 | 8.127 | 12.679 | 13.410 | 15.715 | 11.267 |
| 1993 | 1.392 | 1.887 | 2.772 | 3.762 | 4.930 | 6.054 | 7.450 | 8.641 | 10.901 | 12.517 | 14.742 | 16.874 |
| 1994 | 1.443 | 2.063 | 2.562 | 3.659 | 5.117 | 6.262 | 7.719 | 8.896 | 10.847 | 12.874 | 14.742 | 17.470 |
| 1995 | 1.348 | 1.959 | 2.920 | 3.625 | 5.176 | 6.416 | 7.916 | 10.273 | 11.022 | 11.407 | 13.098 | 15.182 |
| 1996 | 1.457 | 1.930 | 3.132 | 4.141 | 4.922 | 6.009 | 7.406 | 9.772 | 10.539 | 13.503 | 13.689 | 16.194 |
| 1997 | 1.484 | 1.877 | 2.878 | 4.028 | 5.402 | 6.386 | 7.344 | 8.537 | 10.797 | 11.533 | 10.428 | 12.788 |
| 1998 | 1.230 | 1.750 | 2.458 | 3.559 | 5.213 | 7.737 | 7.837 | 9.304 | 10.759 | 14.903 | 16.651 | 18.666 |
| 1999 | 1.241 | 1.716 | 2.426 | 3.443 | 4.720 | 6.352 | 8.730 | 9.946 | 11.088 | 12.535 | 14.995 | 15.151 |
| 2000 | 1.308 | 1.782 | 2.330 | 3.252 | 4.690 | 5.894 | 7.809 | 9.203 | 10.240 | 11.172 | 13.172 | 17.442 |
| 2001 | 1.499 | 2.050 | 2.649 | 3.413 | 4.766 | 6.508 | 7.520 | 9.055 | 8.769 | 9.526 | 11.210 | 13.874 |
| 2002 | 1.294 | 1.926 | 2.656 | 3.680 | 4.720 | 6.369 | 7.808 | 9.002 | 10.422 | 13.402 | 9.008 | 16.893 |
| 2003 | 1.265 | 1.790 | 2.424 | 3.505 | 4.455 | 5.037 | 5.980 | 7.819 | 8.802 | 10.712 | 12.152 | 13.797 |
| 2004 | 1.257 | 1.771 | 2.323 | 3.312 | 4.269 | 5.394 | 5.872 | 7.397 | 10.808 | 11.569 | 13.767 | 12.955 |
| 2005 | 1.194 | 1.712 | 2.374 | 3.435 | 4.392 | 5.201 | 6.200 | 5.495 | 7.211 | 9.909 | 12.944 | 18.151 |
| 2006 | 1.070 | 1.614 | 2.185 | 3.052 | 4.347 | 5.177 | 5.382 | 5.769 | 6.258 | 5.688 | 7.301 | 15.412 |
| 2007 | 1.083 | 1.556 | 2.144 | 2.754 | 3.920 | 5.255 | 6.272 | 6.481 | 7.142 | 6.530 | 9.724 | 10.143 |
| 2008 | 1.162 | 1.627 | 2.318 | 3.120 | 3.846 | 5.367 | 6.771 | 7.648 | 8.282 | 11.181 | 14.266 | 17.320 |
| 2009 | 1.109 | 1.680 | 2.204 | 3.206 | 4.098 | 4.884 | 6.744 | 8.505 | 10.126 | 12.108 | 12.471 | 15.264 |
| 2010 | 1.131 | 1.769 | 2.334 | 3.161 | 4.422 | 5.498 | 6.552 | 7.945 | 8.913 | 10.090 | 10.417 | 13.489 |
| 2011 | 1.163 | 1.795 | 2.615 | 3.471 | 4.469 | 5.850 | 6.742 | 7.850 | 8.810 | 9.797 | 13.534 | 13.033 |
| 2012 | 1.256 | 1.667 | 2.448 | 3.728 | 4.713 | 5.894 | 7.616 | 8.358 | 9.543 | 10.916 | 10.884 | 11.758 |
| 2013 | 1.245 | 1.721 | 2.477 | 3.557 | 4.930 | 6.161 | 7.517 | 8.412 | 9.332 | 9.923 | 11.194 | 12.687 |
| 2014 | 1.222 | 1.790 | 2.535 | 3.431 | 4.565 | 6.043 | 7.544 | 9.178 | 9.713 | 10.513 | 11.437 | 12.979 |
| 2015 | 1.421 | 1.814 | 2.757 | 3.542 | 4.673 | 6.039 | 8.145 | 9.187 | 9.722 | 10.523 | 11.448 | 12.992 |

Table 9.3: Icelandic cod in Division Va. Estimated weight at age in the spawning stock (kg. These weights are used to calculate the spawning stock biomass (SSB).

| year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1955 | 0.645 | 1.019 | 1.833 | 3.183 | 4.128 | 5.657 | 6.635 | 6.168 | 8.746 | 8.829 | 10.086 | 14.584 |
| 1956 | 0.645 | 1.248 | 1.862 | 2.886 | 4.138 | 5.630 | 6.180 | 6.970 | 6.830 | 9.290 | 10.965 | 12.954 |
| 1957 | 0.645 | 1.334 | 2.142 | 2.816 | 4.058 | 5.960 | 7.170 | 7.260 | 8.300 | 8.290 | 10.350 | 13.174 |
| 1958 | 0.645 | 1.412 | 2.652 | 3.969 | 4.450 | 5.940 | 6.640 | 8.290 | 8.510 | 8.840 | 9.360 | 13.097 |
| 1959 | 0.645 | 1.521 | 2.490 | 3.978 | 4.913 | 6.170 | 6.610 | 7.130 | 8.510 | 8.670 | 9.980 | 11.276 |
| 1960 | 0.645 | 1.342 | 2.482 | 4.083 | 5.037 | 6.550 | 6.910 | 7.140 | 7.970 | 10.240 | 10.100 | 12.871 |
| 1961 | 0.645 | 1.303 | 2.295 | 3.810 | 4.922 | 6.310 | 6.930 | 7.310 | 0.750 | 8.510 | 9.840 | 14.550 |
| 1962 | 0.645 | 1.256 | 2.218 | 3.432 | 5.091 | 6.660 | 6.750 | 7.060 | 7.540 | 8.280 | 10.900 | 12.826 |
| 1963 | 0.645 | 1.287 | 2.244 | 3.344 | 4.548 | 6.920 | 7.840 | 7.610 | 8.230 | 9.100 | 9.920 | 11.553 |
| 1964 | 0.645 | 1.365 | 2.244 | 3.538 | 4.850 | 6.460 | 8.000 | 9.940 | 9.210 | 10.940 | 12.670 | 15.900 |
| 1965 | 0.645 | 1.193 | 2.184 | 3.599 | 4.815 | 6.400 | 7.120 | 8.600 | 12.310 | 10.460 | 10.190 | 17.220 |
| 1966 | 0.645 | 1.310 | 2.202 | 3.678 | 5.100 | 6.900 | 7.830 | 8.580 | 9.090 | 14.230 | 14.090 | 17.924 |
| 1967 | 0.645 | 1.420 | 2.261 | 3.579 | 4.948 | 7.790 | 7.840 | 8.430 | 9.090 | 10.090 | 14.240 | 16.412 |
| 1968 | 0.645 | 1.240 | 2.278 | 3.458 | 4.486 | 5.910 | 7.510 | 8.480 | 10.750 | 11.580 | 14.640 | 16.011 |
| 1969 | 0.645 | 1.412 | 2.108 | 3.318 | 4.486 | 5.860 | 7.000 | 8.350 | 8.720 | 10.080 | 11.430 | 13.144 |
| 1970 | 0.645 | 1.131 | 2.074 | 3.318 | 4.325 | 5.590 | 6.260 | 8.370 | 10.490 | 12.310 | 14.590 | 21.777 |
| 1971 | 0.645 | 1.225 | 1.964 | 2.622 | 4.388 | 5.150 | 5.580 | 6.300 | 8.530 | 11.240 | 14.740 | 17.130 |
| 1972 | 0.645 | 1.139 | 1.878 | 2.860 | 3.854 | 5.610 | 6.040 | 6.100 | 6.870 | 8.950 | 11.720 | 16.000 |
| 1973 | 0.645 | 1.108 | 2.100 | 3.168 | 4.361 | 6.110 | 6.670 | 6.750 | 7.430 | 7.950 | 10.170 | 17.000 |
| 1974 | 0.645 | 1.334 | 2.066 | 3.362 | 4.664 | 6.660 | 7.150 | 7.760 | 8.190 | 9.780 | 12.380 | 14.700 |
| 1975 | 0.645 | 1.381 | 2.363 | 3.309 | 4.850 | 6.690 | 7.570 | 8.580 | 8.810 | 9.780 | 10.090 | 11.000 |
| 1976 | 0.645 | 1.388 | 2.252 | 3.608 | 4.512 | 6.730 | 8.250 | 9.610 | 11.540 | 11.430 | 14.060 | 16.180 |
| 1977 | 0.645 | 1.491 | 2.428 | 3.581 | 5.142 | 6.636 | 7.685 | 9.730 | 11.703 | 14.394 | 17.456 | 24.116 |
| 1978 | 0.645 | 1.430 | 2.490 | 3.480 | 5.096 | 6.806 | 9.041 | 10.865 | 13.068 | 11.982 | 19.062 | 21.284 |
| 1979 | 0.645 | 1.526 | 2.246 | 3.519 | 4.938 | 6.754 | 8.299 | 9.312 | 13.130 | 13.418 | 13.540 | 20.072 |
| 1980 | 0.645 | 1.452 | 2.323 | 3.316 | 4.681 | 6.981 | 8.037 | 10.731 | 12.301 | 17.281 | 14.893 | 19.069 |
| 1981 | 0.645 | 1.288 | 1.921 | 2.898 | 3.990 | 5.821 | 7.739 | 9.422 | 11.374 | 12.784 | 12.514 | 19.069 |
| 1982 | 0.645 | 1.209 | 1.909 | 2.732 | 3.790 | 5.386 | 6.682 | 9.141 | 11.963 | 14.226 | 17.287 | 16.590 |
| 1983 | 0.645 | 1.247 | 1.934 | 2.658 | 3.645 | 5.481 | 7.049 | 8.128 | 11.009 | 13.972 | 15.882 | 18.498 |
| 1984 | 0.645 | 1.346 | 2.207 | 3.151 | 3.890 | 5.798 | 7.456 | 9.851 | 11.052 | 14.338 | 15.273 | 16.660 |
| 1985 | 1.306 | 1.382 | 1.752 | 2.710 | 3.443 | 4.675 | 7.220 | 10.320 | 12.197 | 14.683 | 16.175 | 19.050 |
| 1986 | 1.306 | 1.604 | 2.892 | 3.234 | 4.572 | 5.805 | 7.247 | 9.084 | 10.356 | 15.283 | 14.540 | 15.017 |
| 1987 | 1.706 | 1.589 | 2.426 | 3.516 | 4.879 | 6.459 | 7.656 | 9.243 | 10.697 | 10.622 | 15.894 | 12.592 |
| 1988 | 0.929 | 1.480 | 2.263 | 3.273 | 4.387 | 4.566 | 8.275 | 8.822 | 9.977 | 11.732 | 14.156 | 13.042 |
| 1989 | 0.822 | 1.501 | 2.346 | 3.428 | 4.676 | 7.388 | 8.506 | 9.831 | 11.986 | 10.003 | 12.611 | 16.045 |
| 1990 | 0.725 | 1.043 | 2.179 | 2.809 | 4.421 | 6.359 | 9.230 | 10.592 | 10.993 | 14.570 | 15.732 | 17.290 |
| 1991 | 0.114 | 1.286 | 2.042 | 2.752 | 3.404 | 6.091 | 9.152 | 9.804 | 9.754 | 14.344 | 14.172 | 20.200 |
| 1992 | 0.448 | 1.344 | 2.096 | 3.029 | 3.755 | 5.143 | 7.562 | 8.127 | 12.679 | 13.410 | 15.715 | 11.267 |
| 1993 | 0.773 | 1.363 | 2.309 | 3.236 | 4.111 | 5.710 | 6.352 | 8.641 | 10.901 | 12.517 | 14.742 | 16.874 |
| 1994 | 1.611 | 1.728 | 2.253 | 3.341 | 4.515 | 6.535 | 10.039 | 8.896 | 10.847 | 12.874 | 14.742 | 17.470 |
| 1995 | 0.514 | 1.636 | 2.346 | 3.186 | 4.488 | 5.528 | 8.620 | 10.273 | 11.022 | 11.407 | 13.098 | 15.182 |
| 1996 | 0.543 | 1.754 | 2.491 | 3.534 | 4.254 | 5.634 | 8.300 | 9.772 | 10.539 | 13.503 | 13.689 | 16.193 |
| 1997 | 1.112 | 1.347 | 2.267 | 3.746 | 5.426 | 5.972 | 6.958 | 8.537 | 10.797 | 11.533 | 10.428 | 12.788 |
| 1998 | 1.112 | 1.821 | 2.261 | 3.263 | 4.468 | 5.784 | 6.812 | 9.304 | 10.759 | 14.903 | 16.651 | 18.666 |
| 1999 | 1.307 | 1.467 | 1.933 | 2.997 | 3.961 | 5.120 | 6.494 | 9.946 | 11.088 | 12.535 | 14.995 | 15.151 |
| 2000 | 0.496 | 1.355 | 1.916 | 2.881 | 4.318 | 5.580 | 8.497 | 9.203 | 10.240 | 11.172 | 13.172 | 17.442 |
| 2001 | 0.816 | 1.583 | 2.108 | 2.700 | 4.086 | 6.202 | 6.907 | 9.055 | 8.769 | 9.526 | 11.210 | 13.874 |
| 2002 | 0.780 | 1.590 | 2.259 | 3.120 | 3.985 | 5.958 | 9.234 | 9.002 | 10.422 | 13.402 | 9.008 | 16.893 |
| 2003 | 1.149 | 1.324 | 2.239 | 3.052 | 4.231 | 5.057 | 6.838 | 7.819 | 8.802 | 10.712 | 12.152 | 13.797 |
| 2004 | 1.149 | 1.430 | 2.099 | 3.049 | 3.743 | 5.319 | 5.682 | 7.397 | 10.808 | 11.569 | 13.767 | 12.955 |
| 2005 | 0.649 | 1.120 | 1.898 | 2.962 | 3.875 | 4.806 | 7.281 | 5.495 | 7.211 | 9.909 | 12.944 | 18.151 |
| 2006 | 0.907 | 1.384 | 1.999 | 2.907 | 4.384 | 5.122 | 6.536 | 5.769 | 6.258 | 5.688 | 7.301 | 15.412 |
| 2007 | 1.403 | 1.264 | 2.022 | 2.582 | 4.081 | 5.725 | 6.736 | 6.481 | 7.142 | 6.530 | 9.724 | 10.143 |
| 2008 | 0.912 | 1.842 | 2.232 | 2.925 | 3.915 | 5.462 | 7.075 | 7.648 | 8.282 | 11.181 | 14.266 | 17.320 |
| 2009 | 0.644 | 1.441 | 2.028 | 2.873 | 3.913 | 4.919 | 7.046 | 8.505 | 10.126 | 12.108 | 12.471 | 15.264 |
| 2010 | 0.644 | 1.588 | 2.153 | 3.131 | 4.173 | 5.197 | 6.356 | 7.945 | 8.913 | 10.090 | 10.417 | 13.489 |
| 2011 | 0.794 | 2.377 | 2.651 | 3.203 | 4.517 | 6.000 | 6.866 | 7.850 | 8.810 | 9.797 | 13.534 | 13.033 |
| 2012 | 1.403 | 1.698 | 2.594 | 3.683 | 4.483 | 5.921 | 7.988 | 8.358 | 9.543 | 10.916 | 10.884 | 11.758 |
| 2013 | 0.944 | 2.282 | 2.983 | 3.827 | 5.206 | 6.543 | 8.298 | 8.415 | 9.336 | 9.926 | 11.195 | 12.691 |
| 2014 | 0.944 | 1.333 | 2.539 | 3.307 | 4.460 | 6.424 | 8.225 | 8.413 | 9.713 | 10.513 | 11.437 | 12.979 |
| 2015 | 0.710 | 1.047 | 3.311 | 3.833 | 4.902 | 6.240 | 8.728 | 9.705 | 9.722 | 10.523 | 11.448 | 12.992 |

Table 9.4: Icelandic cod in Division Va. Estimated maturity at age.

| year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1955 | 0.019 | 0.022 | 0.033 | 0.181 | 0.577 | 0.782 | 0.834 | 0.960 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1956 | 0.019 | 0.025 | 0.033 | 0.111 | 0.577 | 0.782 | 0.818 | 0.980 | 0.980 | 1.000 | 1.000 | 1.000 |
| 1957 | 0.019 | 0.026 | 0.043 | 0.100 | 0.549 | 0.801 | 0.842 | 0.990 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1958 | 0.019 | 0.028 | 0.086 | 0.520 | 0.682 | 0.801 | 0.834 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1959 | 0.019 | 0.029 | 0.070 | 0.535 | 0.772 | 0.818 | 0.834 | 0.990 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1960 | 0.019 | 0.026 | 0.066 | 0.577 | 0.782 | 0.826 | 0.834 | 0.990 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1961 | 0.019 | 0.025 | 0.053 | 0.450 | 0.772 | 0.818 | 0.834 | 0.990 | 0.990 | 1.000 | 1.000 | 1.000 |
| 1962 | 0.019 | 0.025 | 0.048 | 0.281 | 0.791 | 0.834 | 0.834 | 0.990 | 0.990 | 1.000 | 1.000 | 1.000 |
| 1963 | 0.019 | 0.025 | 0.048 | 0.237 | 0.706 | 0.834 | 0.849 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1964 | 0.019 | 0.026 | 0.048 | 0.329 | 0.762 | 0.826 | 0.849 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1965 | 0.019 | 0.025 | 0.045 | 0.354 | 0.751 | 0.826 | 0.842 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1966 | 0.019 | 0.026 | 0.045 | 0.394 | 0.791 | 0.849 | 0.849 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1967 | 0.019 | 0.028 | 0.051 | 0.341 | 0.772 | 0.842 | 0.849 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1968 | 0.019 | 0.025 | 0.051 | 0.292 | 0.682 | 0.801 | 0.842 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1969 | 0.019 | 0.028 | 0.043 | 0.227 | 0.682 | 0.801 | 0.842 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1970 | 0.019 | 0.023 | 0.041 | 0.227 | 0.644 | 0.772 | 0.818 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1971 | 0.019 | 0.025 | 0.037 | 0.074 | 0.657 | 0.706 | 0.772 | 0.979 | 0.994 | 0.982 | 0.993 | 1.000 |
| 1972 | 0.019 | 0.023 | 0.035 | 0.106 | 0.450 | 0.772 | 0.809 | 0.979 | 0.994 | 0.982 | 0.993 | 1.000 |
| 1973 | 0.022 | 0.028 | 0.163 | 0.382 | 0.697 | 0.80 | 0.834 | 0.996 | 0.996 | 1.000 | 1.000 | 1.000 |
| 1974 | 0.020 | 0.031 | 0.085 | 0.346 | 0.636 | 0.790 | 0.818 | 0.989 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1975 | 0.020 | 0.035 | 0.118 | 0.287 | 0.715 | 0.809 | 0.839 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1976 | 0.025 | 0.026 | 0.086 | 0.253 | 0.406 | 0.797 | 0.841 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1977 | 0.019 | 0.024 | 0.060 | 0.382 | 0.742 | 0.817 | 0.842 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1978 | 0.025 | 0.025 | 0.052 | 0.192 | 0.737 | 0.820 | 0.836 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1979 | 0.019 | 0.021 | 0.053 | 0.282 | 0.635 | 0.790 | 0.836 | 0.919 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1980 | 0.026 | 0.021 | 0.047 | 0.225 | 0.653 | 0.777 | 0.834 | 0.977 | 1.000 | 0.964 | 1.000 | 1.000 |
| 1981 | 0.019 | 0.022 | 0.030 | 0.090 | 0.448 | 0.751 | 0.811 | 0.962 | 0.988 | 1.000 | 1.000 | 1.000 |
| 1982 | 0.021 | 0.025 | 0.038 | 0.065 | 0.297 | 0.705 | 0.815 | 0.967 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1983 | 0.019 | 0.030 | 0.047 | 0.116 | 0.264 | 0.530 | 0.715 | 0.979 | 0.985 | 1.000 | 1.000 | 1.000 |
| 1984 | 0.019 | 0.024 | 0.053 | 0.169 | 0.444 | 0.620 | 0.716 | 0.949 | 0.969 | 0.948 | 1.000 | 1.000 |
| 1985 |  | 0.021 | 0.185 | 0.412 | 0.495 | 0.735 | 0.572 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1986 | 0.001 | 0.023 | 0.149 | 0.395 | 0.682 | 0.734 | 0.941 | 0.962 | 0.988 | 1.000 | 1.000 | 1.000 |
| 1987 | 0.002 | 0.033 | 0.093 | 0.360 | 0.490 | 0.885 | 0.782 | 1.000 | 0.979 | 1.000 | 1.000 | 1.000 |
| 1988 | 0.006 | 0.029 | 0.225 | 0.511 | 0.448 | 0.683 | 0.937 | 0.946 | 0.974 | 0.821 | 1.000 | 1.000 |
| 1989 | 0.008 | 0.025 | 0.142 | 0.372 | 0.645 | 0.652 | 0.634 | 0.991 | 1.000 | 0.903 | 0.859 | 1.000 |
| 1990 | 0.006 | 0.012 | 0.155 | 0.437 | 0.581 | 0.796 | 0.814 | 0.986 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1991 |  | 0.055 | 0.149 | 0.369 | 0.637 | 0.790 | 0.682 | 0.842 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1992 | 0.002 | 0.062 | 0.265 | 0.402 | 0.813 | 0.917 | 0.894 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1993 | 0.006 | 0.085 | 0.267 | 0.464 | 0.693 | 0.801 | 0.843 | 0.968 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1994 | 0.008 | 0.110 | 0.339 | 0.591 | 0.702 | 0.917 | 0.698 | 0.852 | 0.985 | 1.000 | 1.000 | 1.000 |
| 1995 | 0.005 | 0.109 | 0.384 | 0.528 | 0.752 | 0.787 | 0.859 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1996 | 0.002 | 0.031 | 0.186 | 0.499 | 0.650 | 0.733 | 0.812 | 1.000 | 1.000 | 0.986 | 0.971 | 1.000 |
| 1997 | 0.006 | 0.037 | 0.246 | 0.424 | 0.685 | 0.787 | 0.804 | 0.932 | 1.000 | 0.913 | 1.000 | 1.000 |
| 1998 |  | 0.061 | 0.209 | 0.491 | 0.782 | 0.814 | 0.810 | 0.925 | 0.998 | 1.000 | 1.000 | 1.000 |
| 1999 | 0.012 | 0.044 | 0.239 | 0.516 | 0.649 | 0.835 | 0.687 | 0.988 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2000 | 0.001 | 0.065 | 0.248 | 0.512 | 0.611 | 0.867 | 0.998 | 0.980 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2001 | 0.004 | 0.043 | 0.261 | 0.589 | 0.750 | 0.742 | 0.862 | 0.987 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2002 | 0.008 | 0.086 | 0.322 | 0.656 | 0.759 | 0.920 | 0.550 | 0.979 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2003 | 0.005 | 0.046 | 0.218 | 0.524 | 0.870 | 0.798 | 0.860 | 0.998 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2004 |  | 0.038 | 0.246 | 0.549 | 0.626 | 0.843 | 0.816 | 0.990 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2005 | 0.003 | 0.109 | 0.281 | 0.493 | 0.792 | 0.805 | 0.951 | 0.908 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2006 | 0.002 | 0.023 | 0.294 | 0.448 | 0.752 | 0.871 | 0.743 | 0.747 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2007 | 0.012 | 0.032 | 0.159 | 0.501 | 0.693 | 0.785 | 0.836 | 0.924 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2008 | 0.001 | 0.041 | 0.276 | 0.549 | 0.727 | 0.827 | 0.846 | 0.954 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2009 | 0.002 | 0.015 | 0.132 | 0.456 | 0.688 | 0.883 | 0.741 | 0.631 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2010 |  | 0.016 | 0.058 | 0.377 | 0.822 | 0.869 | 0.923 | 0.802 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2011 | 0.002 | 0.012 | 0.135 | 0.431 | 0.734 | 0.926 | 0.940 | 0.958 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2012 | 0.004 | 0.029 | 0.126 | 0.411 | 0.728 | 0.882 | 0.961 | 0.830 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2013 | 0.003 | 0.008 | 0.061 | 0.343 | 0.738 | 0.923 | 0.957 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2014 |  | 0.026 | 0.068 | 0.236 | 0.614 | 0.893 | 0.967 | 0.957 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2015 | 0.004 | 0.007 | 0.109 | 0.353 | 0.638 | 0.908 | 0.979 | 0.988 | 1.000 | 1.000 | 1.000 | 1.000 |

Table 9.5. Icelandic cod in Division Va. Estimated spring survey weight at age.

| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1985 | 14 | 137 | 388 | 1117 | 1733 | 2578 | 3222 | 4671 | 5866 | 7035 |
| 1986 | 15 | 159 | 616 | 1219 | 2246 | 2962 | 4328 | 5588 | 7226 | 8314 |
| 1987 | 14 | 117 | 467 | 1198 | 1751 | 2981 | 4197 | 6336 | 6943 | 10048 |
| 1988 | 11 | 122 | 495 | 1076 | 1964 | 3096 | 3554 | 4365 | 8133 | 9439 |
| 1989 | 22 | 150 | 548 | 1140 | 1932 | 3049 | 4387 | 6264 | 7014 | 12540 |
| 1990 | 19 | 135 | 459 | 1039 | 1814 | 2595 | 3873 | 6045 | 8158 | 9583 |
| 1991 | 18 | 147 | 553 | 1166 | 1842 | 2586 | 3267 | 5735 | 7612 | 14448 |
| 1992 | 24 | 134 | 500 | 1012 | 1845 | 2567 | 3652 | 5049 | 7443 | 13540 |
| 1993 | 12 | 171 | 576 | 1166 | 1944 | 2991 | 3961 | 5378 | 5984 | 9335 |
| 1994 | 13 | 174 | 686 | 1412 | 2044 | 3181 | 4133 | 6273 | 8310 | 9890 |
| 1995 | 10 | 134 | 605 | 1377 | 2284 | 2989 | 4449 | 5323 | 8068 | 9251 |
| 1996 | 11 | 155 | 551 | 1350 | 2082 | 3321 | 4044 | 5263 | 7478 | 9965 |
| 1997 | 18 | 140 | 546 | 1194 | 2168 | 3220 | 4864 | 5508 | 6458 | 6901 |
| 1998 | 15 | 158 | 485 | 1208 | 2041 | 3017 | 4253 | 5437 | 6347 | 8384 |
| 1999 | 14 | 140 | 578 | 1070 | 1847 | 2867 | 3820 | 4981 | 5626 | 8193 |
| 2000 | 16 | 124 | 486 | 1195 | 1817 | 2771 | 4066 | 5349 | 8503 | 8401 |
| 2001 | 17 | 152 | 531 | 1186 | 1852 | 2641 | 3760 | 5452 | 6442 | 8174 |
| 2002 | 11 | 132 | 510 | 1206 | 1998 | 2920 | 3780 | 5759 | 6264 | 6285 |
| 2003 | 16 | 131 | 466 | 1179 | 1918 | 2788 | 4139 | 4678 | 6260 | 9595 |
| 2004 | 20 | 147 | 481 | 1062 | 1873 | 2803 | 3458 | 4989 | 5314 | 7794 |
| 2005 | 11 | 118 | 451 | 1029 | 1760 | 2644 | 3646 | 4362 | 7248 | 6672 |
| 2006 | 13 | 105 | 417 | 982 | 1689 | 2600 | 4050 | 4750 | 5623 | 8380 |
| 2007 | 14 | 101 | 410 | 969 | 1663 | 2342 | 3635 | 5017 | 6122 | 7747 |
| 2008 | 11 | 121 | 376 | 937 | 1805 | 2612 | 3592 | 4933 | 6395 | 8407 |
| 2009 | 12 | 113 | 413 | 845 | 1602 | 2633 | 3659 | 4683 | 5769 | 6289 |
| 2010 | 13 | 98 | 391 | 1008 | 1697 | 2570 | 4021 | 4912 | 6101 | 7753 |
| 2011 | 12 | 102 | 395 | 1126 | 2114 | 2986 | 4225 | 5876 | 6645 | 7904 |
| 2012 | 12 | 142 | 477 | 1143 | 1929 | 3180 | 4249 | 5718 | 7826 | 7609 |
| 2013 | 13 | 113 | 495 | 1054 | 1785 | 3022 | 4772 | 6381 | 8053 | 9537 |
| 2014 | 11 | 114 | 359 | 1078 | 1710 | 2632 | 3987 | 6168 | 8068 | 10117 |
| 2015 | 13 | 150 | 418 | 898 | 2054 | 3016 | 4401 | 6074 | 8653 | 9620 |

Table 9.6. Icelandic cod in Division Va. Survey indices of the spring bottom trawl survey (SMB).

| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1985.00 | 16.54 | 110.43 | 35.40 | 48.20 | 64.15 | 22.57 | 14.85 | 4.85 | 3.21 | 1.76 |
| 1986.00 | 15.05 | 60.24 | 95.89 | 22.42 | 21.21 | 26.34 | 6.63 | 2.48 | 0.83 | 0.73 |
| 1987.00 | 3.65 | 28.21 | 103.74 | 81.99 | 21.08 | 12.21 | 12.00 | 2.55 | 0.89 | 0.38 |
| 1988.00 | 3.44 | 6.96 | 72.09 | 101.50 | 66.58 | 7.82 | 5.91 | 6.29 | 0.58 | 0.24 |
| 1989.00 | 4.04 | 16.38 | 21.97 | 77.79 | 67.59 | 34.20 | 4.20 | 1.45 | 1.14 | 0.24 |
| 1990.00 | 5.56 | 11.78 | 26.08 | 14.07 | 27.05 | 32.38 | 14.21 | 1.50 | 0.52 | 0.41 |
| 1991.00 | 3.95 | 16.00 | 18.20 | 30.17 | 15.24 | 18.09 | 20.93 | 4.24 | 0.79 | 0.29 |
| 1992.00 | 0.71 | 16.80 | 33.54 | 18.89 | 16.34 | 6.54 | 5.70 | 5.12 | 1.29 | 0.22 |
| 1993.00 | 3.57 | 4.75 | 30.78 | 36.48 | 13.22 | 9.90 | 2.13 | 1.75 | 1.17 | 0.36 |
| 1994.00 | 14.38 | 14.94 | 9.01 | 26.66 | 21.90 | 5.77 | 3.63 | 0.70 | 0.48 | 0.47 |
| 1995.00 | 1.08 | 29.13 | 24.75 | 8.98 | 23.88 | 17.69 | 3.78 | 1.80 | 0.35 | 0.17 |
| 1996.00 | 3.72 | 5.42 | 42.58 | 29.44 | 12.89 | 14.62 | 14.02 | 3.80 | 1.04 | 0.18 |
| 1997.00 | 1.18 | 22.18 | 13.55 | 56.31 | 29.10 | 9.50 | 8.78 | 6.61 | 0.56 | 0.21 |
| 1998.00 | 8.06 | 5.36 | 29.92 | 16.04 | 61.73 | 28.58 | 6.50 | 5.24 | 3.03 | 0.66 |
| 1999.00 | 7.39 | 32.98 | 7.01 | 42.25 | 13.00 | 23.66 | 11.12 | 2.35 | 1.32 | 0.70 |
| 2000.00 | 18.85 | 27.60 | 54.99 | 6.94 | 30.00 | 8.28 | 8.18 | 4.14 | 0.51 | 0.30 |
| 2001.00 | 12.13 | 21.74 | 36.38 | 38.04 | 4.95 | 15.11 | 3.30 | 1.96 | 0.81 | 0.29 |
| 2002.00 | 0.91 | 37.85 | 41.22 | 40.13 | 36.25 | 7.09 | 8.32 | 1.49 | 0.72 | 0.30 |
| 2003.00 | 11.17 | 4.17 | 46.35 | 36.58 | 28.42 | 16.89 | 3.82 | 4.34 | 1.03 | 0.20 |
| 2004.00 | 6.57 | 24.43 | 7.87 | 61.79 | 35.00 | 24.83 | 14.44 | 2.82 | 2.88 | 0.47 |
| 2005.00 | 2.56 | 14.54 | 38.70 | 9.68 | 43.57 | 22.97 | 10.84 | 5.77 | 0.93 | 0.92 |
| 2006.00 | 8.79 | 6.39 | 22.67 | 38.44 | 10.83 | 27.74 | 10.05 | 3.55 | 1.38 | 0.25 |
| 2007.00 | 5.61 | 18.21 | 8.58 | 21.09 | 27.60 | 9.06 | 9.75 | 5.08 | 2.11 | 0.75 |
| 2008.00 | 6.40 | 11.77 | 22.08 | 9.31 | 20.43 | 20.40 | 8.10 | 6.63 | 2.47 | 0.60 |
| 2009.00 | 21.27 | 11.62 | 15.80 | 21.82 | 14.59 | 23.45 | 14.59 | 4.18 | 2.73 | 1.02 |
| 2010.00 | 18.29 | 20.00 | 18.00 | 17.73 | 23.75 | 13.27 | 16.60 | 8.93 | 2.71 | 1.70 |
| 2011.00 | 3.57 | 21.49 | 26.63 | 19.90 | 22.48 | 25.32 | 13.51 | 12.31 | 4.55 | 0.91 |
| 2012.00 | 19.94 | 9.75 | 37.59 | 56.57 | 41.59 | 30.22 | 26.99 | 9.96 | 6.30 | 2.76 |
| 2013.00 | 10.80 | 31.40 | 17.68 | 43.76 | 46.47 | 25.24 | 16.50 | 13.81 | 6.94 | 3.33 |
| 2014.00 | 3.31 | 23.97 | 38.00 | 23.48 | 47.17 | 37.60 | 17.31 | 8.18 | 4.26 | 2.22 |
| 2015.00 | 20.84 | 10.66 | 27.42 | 41.87 | 20.93 | 40.85 | 28.14 | 16.41 | 4.99 | 3.13 |
|  |  |  |  |  |  |  |  |  |  |  |

Table 9.7. Icelandic cod in Division Va. Survey cv of the spring bottom trawl survey (SMB).

| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1985.00 | 0.08 | 0.44 | 0.19 | 0.11 | 0.11 | 0.10 | 0.10 | 0.10 | 0.14 | 0.16 |
| 1986.00 | 0.09 | 0.10 | 0.10 | 0.10 | 0.09 | 0.08 | 0.09 | 0.07 | 0.08 | 0.07 |
| 1987.00 | 0.13 | 0.11 | 0.09 | 0.10 | 0.10 | 0.10 | 0.09 | 0.11 | 0.10 | 0.14 |
| 1988.00 | 0.19 | 0.18 | 0.10 | 0.10 | 0.11 | 0.10 | 0.09 | 0.09 | 0.12 | 0.12 |
| 1989.00 | 0.12 | 0.10 | 0.15 | 0.21 | 0.16 | 0.12 | 0.10 | 0.10 | 0.12 | 0.13 |
| 1990.00 | 0.14 | 0.09 | 0.13 | 0.13 | 0.10 | 0.09 | 0.09 | 0.11 | 0.13 | 0.17 |
| 1991.00 | 0.12 | 0.10 | 0.07 | 0.12 | 0.12 | 0.10 | 0.10 | 0.11 | 0.16 | 0.31 |
| 1992.00 | 0.11 | 0.08 | 0.07 | 0.09 | 0.10 | 0.09 | 0.08 | 0.09 | 0.12 | 0.23 |
| 1993.00 | 0.20 | 0.10 | 0.09 | 0.11 | 0.11 | 0.10 | 0.10 | 0.10 | 0.09 | 0.10 |
| 1994.00 | 0.26 | 0.12 | 0.09 | 0.12 | 0.14 | 0.14 | 0.13 | 0.11 | 0.14 | 0.16 |
| 1995.00 | 0.17 | 0.08 | 0.09 | 0.10 | 0.10 | 0.10 | 0.10 | 0.12 | 0.15 | 0.16 |
| 1996.00 | 0.12 | 0.10 | 0.11 | 0.15 | 0.14 | 0.11 | 0.10 | 0.10 | 0.15 | 0.22 |
| 1997.00 | 0.14 | 0.08 | 0.08 | 0.10 | 0.10 | 0.09 | 0.10 | 0.11 | 0.15 | 0.25 |
| 1998.00 | 0.12 | 0.15 | 0.09 | 0.12 | 0.17 | 0.15 | 0.11 | 0.11 | 0.13 | 0.17 |
| 1999.00 | 0.11 | 0.08 | 0.07 | 0.10 | 0.10 | 0.09 | 0.09 | 0.10 | 0.08 | 0.10 |
| 2000.00 | 0.07 | 0.07 | 0.08 | 0.08 | 0.09 | 0.09 | 0.08 | 0.10 | 0.10 | 0.08 |
| 2001.00 | 0.09 | 0.10 | 0.10 | 0.15 | 0.17 | 0.20 | 0.15 | 0.12 | 0.10 | 0.16 |
| 2002.00 | 0.18 | 0.09 | 0.13 | 0.16 | 0.18 | 0.15 | 0.15 | 0.10 | 0.14 | 0.11 |
| 2003.00 | 0.10 | 0.11 | 0.07 | 0.12 | 0.11 | 0.10 | 0.10 | 0.15 | 0.19 | 0.19 |
| 2004.00 | 0.10 | 0.08 | 0.10 | 0.16 | 0.15 | 0.16 | 0.15 | 0.13 | 0.17 | 0.21 |
| 2005.00 | 0.12 | 0.12 | 0.07 | 0.09 | 0.12 | 0.12 | 0.12 | 0.12 | 0.15 | 0.19 |
| 2006.00 | 0.09 | 0.11 | 0.08 | 0.10 | 0.10 | 0.11 | 0.12 | 0.15 | 0.13 | 0.20 |
| 2007.00 | 0.09 | 0.12 | 0.10 | 0.10 | 0.11 | 0.09 | 0.08 | 0.09 | 0.11 | 0.18 |
| 2008.00 | 0.11 | 0.09 | 0.07 | 0.09 | 0.10 | 0.10 | 0.10 | 0.09 | 0.10 | 0.09 |
| 2009.00 | 0.10 | 0.10 | 0.09 | 0.10 | 0.13 | 0.13 | 0.12 | 0.11 | 0.09 | 0.10 |
| 2010.00 | 0.08 | 0.10 | 0.12 | 0.10 | 0.11 | 0.10 | 0.10 | 0.09 | 0.08 | 0.09 |
| 2011.00 | 0.11 | 0.12 | 0.10 | 0.12 | 0.14 | 0.14 | 0.12 | 0.14 | 0.11 | 0.09 |
| 2012.00 | 0.09 | 0.14 | 0.08 | 0.32 | 0.41 | 0.34 | 0.24 | 0.17 | 0.11 | 0.14 |
| 2013.00 | 0.06 | 0.13 | 0.08 | 0.11 | 0.13 | 0.12 | 0.10 | 0.11 | 0.14 | 0.16 |
| 2014.00 | 0.13 | 0.35 | 0.11 | 0.13 | 0.13 | 0.12 | 0.11 | 0.13 | 0.19 | 0.30 |
| 2015.00 | 0.08 | 0.15 | 0.08 | 0.09 | 0.09 | 0.10 | 0.14 | 0.18 | 0.17 | 0.18 |
|  |  |  |  |  |  |  |  |  |  |  |

Table 9.8. Icelandic cod in Division Va. Survey indices of the fall bottom trawl survey (SMH).

| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1996.00 | 6.69 | 3.57 | 20.00 | 13.98 | 5.40 | 7.44 | 6.26 | 1.60 | 0.31 | 0.09 |
| 1997.00 | 0.66 | 16.89 | 6.71 | 28.82 | 15.86 | 5.19 | 3.60 | 2.07 | 0.30 | 0.11 |
| 1998.00 | 5.92 | 2.63 | 15.62 | 7.36 | 16.01 | 16.03 | 5.20 | 2.24 | 1.27 | 0.20 |
| 1999.00 | 8.61 | 14.54 | 5.68 | 23.38 | 7.42 | 9.94 | 4.05 | 0.59 | 0.34 | 0.36 |
| 2000.00 | 4.60 | 13.17 | 15.25 | 3.71 | 11.15 | 3.49 | 2.61 | 1.11 | 0.34 | 0.28 |
| 2001.00 | 7.11 | 11.51 | 19.53 | 21.13 | 3.30 | 6.73 | 1.60 | 0.76 | 0.17 | 0.03 |
| 2002.00 | 0.92 | 13.72 | 16.11 | 23.39 | 15.94 | 5.41 | 4.77 | 1.11 | 0.61 | 0.08 |
| 2003.00 | 5.16 | 2.68 | 25.66 | 16.98 | 13.22 | 8.99 | 1.89 | 2.55 | 0.38 | 0.10 |
| 2004.00 | 3.67 | 16.27 | 6.91 | 29.82 | 18.84 | 11.73 | 7.38 | 1.88 | 1.65 | 0.23 |
| 2005.00 | 2.15 | 9.03 | 20.37 | 6.82 | 25.62 | 10.88 | 3.86 | 1.91 | 0.29 | 0.31 |
| 2006.00 | 4.51 | 4.52 | 16.28 | 23.04 | 7.67 | 13.93 | 6.12 | 2.05 | 1.02 | 0.16 |
| 2007.00 | 3.73 | 9.82 | 4.93 | 11.73 | 15.68 | 6.34 | 5.91 | 3.14 | 0.76 | 0.50 |
| 2008.00 | 5.30 | 11.88 | 15.19 | 7.66 | 17.57 | 18.51 | 5.67 | 5.61 | 1.50 | 0.79 |
| 2009.00 | 7.04 | 8.30 | 13.14 | 18.11 | 12.39 | 16.46 | 10.22 | 3.15 | 2.75 | 0.84 |
| 2010.00 | 10.78 | 18.82 | 16.18 | 15.52 | 17.96 | 9.81 | 11.21 | 6.81 | 2.29 | 1.20 |
| 2012.00 | 7.43 | 9.43 | 23.38 | 20.66 | 12.72 | 10.82 | 9.53 | 5.31 | 3.33 | 1.55 |
| 2013.00 | 6.25 | 19.28 | 13.41 | 27.13 | 21.99 | 12.60 | 7.72 | 5.94 | 2.93 | 1.87 |
| 2014.00 | 3.57 | 16.01 | 23.57 | 13.85 | 23.73 | 19.83 | 8.54 | 5.91 | 4.02 | 2.50 |

Table 9.9. Icelandic cod in Division Va. Survey cv of the fall bottom trawl survey (SMH).

| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1996.00 | 0.35 | 0.18 | 0.11 | 0.14 | 0.13 | 0.13 | 0.17 | 0.23 | 0.27 | 0.33 |
| 1997.00 | 0.34 | 0.54 | 0.22 | 0.26 | 0.21 | 0.15 | 0.11 | 0.12 | 0.12 | 0.15 |
| 1998.00 | 0.16 | 0.12 | 0.12 | 0.11 | 0.13 | 0.19 | 0.32 | 0.35 | 0.38 | 0.34 |
| 1999.00 | 0.32 | 0.14 | 0.24 | 0.30 | 0.32 | 0.23 | 0.20 | 0.19 | 0.19 | 0.21 |
| 2000.00 | 0.18 | 0.26 | 0.14 | 0.14 | 0.15 | 0.18 | 0.16 | 0.18 | 0.33 | 0.31 |
| 2001.00 | 0.17 | 0.14 | 0.14 | 0.11 | 0.11 | 0.11 | 0.17 | 0.33 | 0.41 | 0.79 |
| 2002.00 | 0.16 | 0.12 | 0.12 | 0.13 | 0.12 | 0.11 | 0.11 | 0.12 | 0.15 | 0.50 |
| 2003.00 | 0.13 | 0.14 | 0.12 | 0.11 | 0.11 | 0.09 | 0.10 | 0.14 | 0.19 | 0.32 |
| 2004.00 | 0.14 | 0.17 | 0.13 | 0.14 | 0.11 | 0.10 | 0.09 | 0.08 | 0.08 | 0.09 |
| 2005.00 | 0.27 | 0.10 | 0.11 | 0.10 | 0.12 | 0.11 | 0.10 | 0.08 | 0.09 | 0.10 |
| 2006.00 | 0.15 | 0.14 | 0.13 | 0.13 | 0.11 | 0.11 | 0.11 | 0.10 | 0.09 | 0.16 |
| 2007.00 | 0.21 | 0.14 | 0.11 | 0.14 | 0.14 | 0.14 | 0.13 | 0.11 | 0.11 | 0.12 |
| 2008.00 | 0.17 | 0.11 | 0.10 | 0.10 | 0.11 | 0.11 | 0.15 | 0.20 | 0.24 | 0.22 |
| 2009.00 | 0.17 | 0.11 | 0.13 | 0.14 | 0.13 | 0.12 | 0.11 | 0.11 | 0.11 | 0.14 |
| 2010.00 | 0.17 | 0.16 | 0.11 | 0.13 | 0.13 | 0.11 | 0.15 | 0.17 | 0.19 | 0.20 |
| 2012.00 | 0.15 | 0.11 | 0.12 | 0.13 | 0.14 | 0.14 | 0.12 | 0.12 | 0.14 | 0.15 |
| 2013.00 | 0.16 | 0.14 | 0.14 | 0.14 | 0.12 | 0.11 | 0.11 | 0.12 | 0.13 | 0.14 |
| 2014.00 | 0.24 | 0.13 | 0.12 | 0.14 | 0.13 | 0.11 | 0.10 | 0.13 | 0.20 | 0.28 |

Table 9.10: Icelandic cod in Division Va. Catch at age residuals from the ADCAM model tuned with the spring (SMB) and the fall (SMH) surveys.

| year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1955 | -0.122 | -0.207 | 0.078 | 0.115 | 0.208 | -0.115 | -0.163 | 0.132 | -0.099 | -0.451 | -0.206 | -0.001 |
| 1956 | -0.026 | -0.047 | 0.027 | -0.005 | -0.134 | -0.199 | -0.006 | 0.008 | 0.174 | 0.094 | 0.229 | . 219 |
| 1957 | 0.092 | 0.018 | -0.015 | 0.168 | -0.133 | 0.092 | 0.063 | -0.147 | -0.098 | -0.115 | -0.381 | 0.520 |
| 1958 | 0.153 | 0.176 | -0.265 | -0.072 | 0.060 | 0.079 | 0.133 | -0.231 | 0.232 | 0.001 | -0.228 | 0.392 |
| 1959 | -0.214 | 0.211 | 0.260 | -0.242 | -0.218 | -0.061 | -0.069 | 0.280 | -0.264 | 0.380 | -0.229 | -0.405 |
| 1960 | 0.100 | -0.356 | 0.140 | 0.188 | 0.063 | 0.074 | -0.026 | -0.113 | -0.041 | 0.033 | -0.640 | 0.903 |
| 1961 | 0.051 | 0.040 | -0.403 | 0.118 | -0.016 | 0.272 | 0.202 | -0.142 | 0.087 | -0.191 | -0.974 | 0.828 |
| 1962 | 0.089 | -0.008 | 0.125 | -0.243 | 0.116 | -0.295 | 0.091 | 0.260 | -0.064 | 0.034 | -0.401 | 0.697 |
| 1963 | -0.058 | 0.295 | -0.175 | 0.012 | -0.030 | -0.070 | -0.376 | 0.209 | 0.349 | 0.063 | 0.073 | -0.617 |
| 1964 | -0.127 | -0.016 | 0.126 | -0.252 | -0.118 | 0.378 | -0.102 | -0.455 | -0.012 | 0.267 | -0.158 | 0.006 |
| 1965 | -0.031 | -0.114 | 0.085 | 0.163 | -0.128 | 0.049 | 0.474 | -0.481 | -0.055 | -0.507 | -0.361 | 0.635 |
| 1966 | -0.042 | -0.043 | -0.179 | 0.096 | -0.070 | 0.124 | -0.346 | 0.592 | -0.828 | 0.279 | 0.009 | 1.057 |
| 1967 | 0.191 | -0.129 | 0.023 | -0.199 | 0.025 | -0.373 | 0.490 | 0.046 | 0.672 | -0.725 | -0.836 | -0.184 |
| 1968 | 0.035 | -0.021 | -0.273 | -0.120 | 0.232 | 0.157 | -0.418 | 0.367 | -0.124 | 0.600 | -0.658 | 0.654 |
| 1969 | -0.090 | -0.027 | 0.153 | -0.011 | 0.052 | -0.149 | -0.326 | -0.246 | -0.041 | -0.258 | -0.810 | -0.144 |
| 1970 | -0.096 | 0.136 | -0.053 | -0.136 | 0.054 | -0.162 | 0.478 | -0.581 | -0.119 | 0.246 | 0.294 | 0.451 |
| 1971 | -0.103 | 0.071 | 0.091 | 0.177 | -0.184 | 0.283 | -0.170 | 0.055 | -0.452 | -0.021 | 0.122 | 0.360 |
| 1972 | -0.167 | -0.126 | 0.069 | -0.033 | 0.117 | -0.052 | -0.104 | 0.293 | -0.070 | 0.170 | 0.523 | -2.764 |
| 1973 | 0.275 | -0.022 | -0.098 | 0.028 | -0.004 | -0.241 | 0.087 | 0.172 | 0.157 | -0.195 | -1.254 | -2.096 |
| 1974 | -0.159 | 0.210 | -0.021 | -0.177 | -0.006 | -0.003 | -0.222 | 0.289 | 0.011 | 0.186 | -0.436 | 0.805 |
| 1975 | 0.188 | -0.073 | 0.041 | -0.054 | 0.030 | -0.152 | -0.208 | -0.005 | 0.407 | -0.016 | -0.123 | 0.092 |
| 1976 | 0.098 | 0.002 | -0.169 | 0.077 | -0.092 | 0.252 | -0.157 | -0.155 | 0.056 | 0.272 | -0.234 | 0.237 |
| 1977 | -0.400 | -0.062 | 0.046 | -0.093 | 0.126 | 0.052 | 0.307 | 0.029 | -0.702 | -0.478 | -1.224 | -2.495 |
| 1978 | 0.079 | -0.014 | 0.038 | -0.096 | 0.043 | -0.206 | 0.120 | -0.188 | 0.015 | -0.051 | 0.527 | 1.201 |
| 1979 | 0.157 | 0.094 | -0.217 | 0.103 | -0.046 | 0.031 | -0.312 | -0.078 | 0.045 | -0.146 | 0.406 | -0.200 |
| 1980 | 0.211 | 0.010 | 0.078 | 0.061 | -0.010 | -0.090 | 0.125 | -0.486 | 0.295 | 0.097 | 0.153 | -1.085 |
| 1981 | -0.301 | -0.206 | 0.082 | -0.136 | 0.071 | 0.090 | 0.021 | 0.326 | -0.076 | 0.599 | -0.020 | 1.169 |
| 1982 | 0.009 | 0.152 | 0.072 | -0.055 | -0.221 | 0.192 | 0.177 | 0.136 | -0.230 | -0.868 | 0.046 | -0.860 |
| 1983 | -0.321 | -0.357 | 0.111 | 0.142 | 0.043 | 0.009 | -0.038 | -0.028 | 0.004 | 0.373 | -0.197 | 0.586 |
| 1984 | 0.347 | 0.025 | -0.059 | -0.045 | -0.098 | -0.005 | 0.054 | -0.138 | -0.353 | 0.165 | 0.709 | 0.102 |
| 1985 | 0.039 | 0.182 | -0.103 | 0.122 | -0.098 | -0.021 | -0.139 | 0.133 | 0.027 | -0.346 | 0.470 | 0.469 |
| 1986 | 0.147 | -0.118 | 0.014 | -0.015 | 0.179 | -0.046 | 0.117 | -0.212 | 0.077 | 0.052 | -0.598 | 0.182 |
| 1987 | -0.147 | 0.122 | 0.013 | -0.165 | 0.064 | 0.038 | -0.027 | 0.112 | -0.380 | -0.115 | 0.116 | -0.303 |
| 1988 | -0.085 | -0.059 | -0.054 | 0.136 | -0.086 | 0.069 | 0.157 | 0.028 | 0.477 | 0.016 | 0.535 | 0.109 |
| 1989 | -0.212 | 0.044 | 0.147 | -0.071 | -0.003 | -0.152 | -0.325 | -0.093 | -0.024 | 0.516 | -0.031 | -1.426 |
| 1990 | -0.002 | -0.138 | -0.108 | 0.003 | 0.038 | 0.093 | -0.085 | -0.231 | 0.288 | 0.112 | -0.223 | 0.068 |
| 1991 | 0.070 | 0.041 | -0.132 | -0.067 | 0.094 | -0.074 | 0.116 | -0.075 | -0.315 | 0.403 | -0.573 | 0.113 |
| 1992 | -0.227 | 0.080 | 0.043 | 0.028 | 0.103 | -0.003 | -0.043 | -0.067 | -0.746 | -0.768 | -0.573 | -0.158 |
| 1993 | 0.255 | 0.046 | -0.203 | -0.056 | -0.073 | -0.123 | 0.068 | 0.486 | 0.500 | -0.209 | -0.988 | 0.423 |
| 1994 | 0.031 | 0.246 | -0.135 | -0.194 | -0.040 | 0.067 | -0.193 | -0.136 | 0.427 | 0.522 | 0.514 | -0.397 |
| 1995 | 0.275 | -0.033 | 0.083 | -0.035 | -0.040 | -0.117 | -0.128 | -0.293 | -0.214 | 0.733 | 1.110 | 0.611 |
| 1996 | 0.003 | -0.053 | -0.177 | 0.077 | 0.041 | . 016 | 0.125 | 0.172 | -0.383 | -0.401 | 0.601 | -0.063 |
| 1997 | -0.156 | 0.025 | -0.029 | -0.125 | -0.095 | 0.207 | 0.173 | 0.257 | 0.411 | -0.728 | -0.233 | 0.169 |
| 1998 | -0.181 | -0.169 | 0.064 | 0.073 | 0.019 | -0.166 | 0.242 | 0.047 | 0.091 | 0.284 | 0.154 | -0.728 |
| 1999 | -0.101 | 0.033 | 0.034 | 0.026 | 0.088 | -0.044 | -0.244 | -0.185 | -0.259 | -0.394 | -0.476 | -0.901 |
| 2000 | 0.173 | -0.239 | 0.106 | -0.040 | 0.013 | 0.108 | 0.035 | -0.113 | -0.001 | 0.155 | -0.130 | -0.068 |
| 2001 | 0.189 | 0.195 | -0.162 | -0.007 | 0.026 | -0.182 | 0.097 | 0.280 | -0.037 | 0.153 | -0.517 | -0.004 |
| 2002 | -0.020 | 0.085 | 0.032 | -0.078 | -0.025 | -0.008 | -0.153 | 0.290 | 0.277 | -0.308 | 0.387 | -1.110 |
| 2003 | -0.230 | 0.030 | -0.011 | -0.034 | 0.175 | 0.007 | 0.224 | -0.311 | 0.068 | 0.167 | 0.157 | 0.496 |
| 2004 | -0.221 | 0.109 | 0.099 | -0.087 | -0.059 | 0.235 | 0.027 | 0.233 | -0.486 | 0.010 | 0.247 | -0.332 |
| 2005 | 0.196 | -0.293 | 0.144 | -0.058 | -0.119 | -0.087 | 0.321 | 0.098 | 0.335 | 0.104 | 0.052 | -0.817 |
| 2006 | -0.065 | 0.030 | -0.138 | 0.064 | 0.052 | -0.084 | -0.082 | 0.177 | -0.003 | 0.117 | -0.187 | -1.639 |
| 2007 | -0.108 | 0.181 | -0.038 | -0.012 | -0.149 | 0.055 | -0.030 | 0.178 | 0.764 | 0.367 | 0.775 | -0.346 |
| 2008 | 0.023 | -0.195 | 0.077 | -0.113 | 0.080 | -0.179 | 0.013 | 0.062 | -0.000 | 0.080 | -0.001 | -0.539 |
| 2009 | 0.138 | -0.061 | 0.075 | 0.138 | -0.052 | 0.259 | -0.202 | -0.248 | -0.061 | -0.420 | 0.010 | -0.545 |
| 2010 | 0.005 | 0.015 | -0.132 | 0.075 | 0.026 | -0.061 | 0.175 | -0.126 | -0.127 | 0.318 | 0.293 | 0.566 |
| 2011 | 0.125 | -0.043 | 0.020 | 0.026 | -0.033 | 0.007 | -0.128 | 0.040 | -0.055 | -0.133 | -0.259 | -0.888 |
| 2012 | -0.128 | -0.015 | 0.027 | -0.010 | -0.012 | 0.170 | 0.001 | -0.219 | 0.177 | -0.270 | 0.180 | -0.135 |
| 2013 | 0.133 | 0.048 | 0.033 | 0.035 | -0.099 | -0.055 | 0.184 | -0.034 | -0.198 | 0.376 | 0.005 | -0.156 |
| 2014 | -0.041 | -0.002 | -0.016 | -0.015 | 0.037 | -0.061 | 0.004 | 0.187 | 0.097 | -0.239 | 0.082 | -0.065 |

Table 9.11: Icelandic cod in Division Va. Spring survey (SMB) at age residuals from the ADCAM model, assessment tuned with both the spring and the fall survey.

| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1985 | -0.481 | 0.034 | 0.226 | 0.450 | 0.136 | 0.263 | 0.405 | 0.184 | 0.317 | 0.663 |
| 1986 | 0.434 | -0.061 | -0.383 | -0.224 | -0.072 | 0.002 | -0.159 | -0.273 | -0.248 | -0.046 |
| 1987 | 0.648 | 0.007 | 0.137 | -0.449 | -0.020 | -0.071 | 0.045 | -0.087 | -0.093 | -0.008 |
| 1988 | -0.191 | 0.032 | 0.502 | 0.163 | -0.110 | -0.339 | 0.090 | 0.486 | -0.115 | -0.102 |
| 1989 | 0.380 | 0.069 | 0.529 | 0.563 | 0.250 | 0.190 | -0.118 | -0.105 | 0.216 | 0.106 |
| 1990 | -0.474 | 0.128 | 0.074 | 0.061 | -0.143 | -0.148 | 0.081 | -0.150 | -0.039 | 0.157 |
| 1991 | -0.163 | -0.447 | 0.103 | 0.164 | 0.262 | 0.041 | 0.136 | -0.151 | 0.226 | 0.195 |
| 1992 | -0.233 | 0.032 | -0.185 | 0.123 | -0.078 | -0.128 | -0.137 | -0.141 | -0.104 | -0.006 |
| 1993 | -0.505 | -0.023 | 0.190 | -0.040 | 0.062 | -0.036 | -0.209 | -0.155 | -0.219 | -0.213 |
| 1994 | 0.534 | -0.245 | 0.028 | 0.120 | -0.183 | -0.316 | -0.159 | -0.217 | -0.179 | -0.048 |
| 1995 | -0.220 | 0.142 | -0.217 | -0.041 | 0.178 | -0.011 | -0.215 | -0.086 | -0.060 | -0.201 |
| 1996 | -0.631 | -0.100 | 0.104 | -0.114 | 0.218 | -0.038 | 0.260 | 0.398 | 0.210 | 0.058 |
| 1997 | 0.193 | -0.046 | 0.138 | 0.286 | -0.021 | -0.039 | -0.030 | 0.253 | -0.343 | -0.288 |
| 1998 | -0.096 | 0.135 | -0.176 | 0.134 | 0.518 | 0.299 | 0.097 | 0.207 | 0.438 | 0.501 |
| 1999 | -0.025 | 0.184 | -0.033 | 0.058 | -0.036 | 0.086 | 0.033 | -0.017 | -0.012 | 0.141 |
| 2000 | 0.895 | 0.139 | 0.287 | -0.159 | -0.076 | -0.203 | -0.189 | -0.003 | -0.239 | -0.225 |
| 2001 | 0.202 | 0.029 | 0.017 | -0.086 | -0.442 | -0.217 | -0.367 | -0.552 | -0.322 | 0.209 |
| 2002 | -0.153 | 0.251 | 0.153 | 0.073 | 0.058 | -0.142 | -0.171 | -0.273 | -0.397 | -0.131 |
| 2003 | 0.013 | -0.117 | 0.053 | -0.033 | -0.107 | -0.203 | -0.190 | -0.061 | 0.177 | -0.516 |
| 2004 | -0.091 | 0.177 | -0.108 | 0.275 | 0.117 | 0.231 | 0.201 | 0.145 | 0.428 | 0.286 |
| 2005 | -0.143 | 0.076 | 0.203 | -0.109 | 0.098 | 0.109 | 0.009 | 0.046 | 0.034 | 0.244 |
| 2006 | 0.173 | -0.047 | -0.027 | 0.070 | -0.083 | 0.162 | -0.094 | -0.314 | -0.334 | -0.220 |
| 2007 | 0.017 | 0.145 | -0.312 | -0.233 | -0.160 | -0.175 | -0.297 | -0.049 | 0.052 | -0.073 |
| 2008 | -0.034 | 0.000 | -0.087 | -0.409 | -0.274 | -0.105 | 0.126 | -0.042 | 0.114 | -0.170 |
| 2009 | 0.382 | -0.123 | -0.171 | -0.236 | -0.165 | -0.078 | -0.070 | 0.031 | -0.188 | -0.102 |
| 2010 | 0.157 | -0.187 | -0.207 | -0.239 | -0.209 | -0.192 | -0.073 | -0.048 | 0.349 | 0.027 |
| 2011 | -0.408 | -0.182 | -0.365 | -0.293 | -0.110 | 0.048 | 0.110 | 0.094 | -0.035 | -0.111 |
| 2012 | 0.186 | -0.103 | -0.093 | 0.188 | 0.317 | 0.291 | 0.394 | 0.261 | 0.114 | 0.081 |
| 2013 | -0.109 | 0.104 | -0.111 | -0.085 | 0.037 | 0.047 | 0.028 | 0.205 | 0.512 | -0.019 |
| 2014 | -0.363 | 0.033 | -0.140 | -0.009 | 0.018 | 0.158 | -0.019 | -0.195 | -0.306 | -0.050 |
| 2015 | 0.141 | -0.037 | -0.227 | -0.159 | -0.184 | 0.211 | 0.179 | 0.364 | -0.069 | -0.079 |

Table 9.12: Icelandic cod in Division Va. Fall survey (SMH) at age residuals from the ADCAM model, assessment tuned with both the spring and the fall survey.

| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1996 | 0.035 | -0.078 | -0.010 | -0.184 | -0.012 | -0.062 | 0.179 | 0.190 | -0.166 | -0.047 |
| 1997 | -0.167 | 0.112 | -0.018 | 0.250 | 0.057 | -0.152 | -0.125 | -0.036 | -0.323 | -0.057 |
| 1998 | -0.204 | -0.011 | -0.188 | 0.033 | -0.031 | 0.368 | 0.519 | 0.113 | 0.276 | 0.044 |
| 1999 | 0.266 | -0.097 | 0.113 | 0.111 | 0.079 | 0.008 | -0.092 | -0.290 | -0.329 | 0.104 |
| 2000 | -0.262 | -0.075 | -0.261 | -0.081 | -0.221 | -0.210 | -0.363 | -0.310 | 0.029 | 0.206 |
| 2001 | -0.121 | -0.148 | 0.041 | -0.011 | -0.217 | -0.236 | -0.223 | -0.490 | -0.530 | -0.329 |
| 2002 | -0.177 | -0.201 | -0.123 | 0.155 | 0.005 | 0.128 | 0.010 | 0.019 | 0.011 | -0.380 |
| 2003 | -0.102 | -0.104 | 0.092 | -0.146 | -0.108 | -0.138 | -0.122 | 0.062 | -0.047 | -0.437 |
| 2004 | -0.121 | 0.152 | 0.101 | 0.131 | 0.173 | 0.116 | 0.232 | 0.319 | 0.451 | 0.171 |
| 2005 | 0.073 | -0.078 | 0.100 | 0.077 | 0.255 | 0.011 | -0.256 | -0.295 | -0.234 | -0.155 |
| 2006 | 0.074 | -0.068 | 0.096 | 0.100 | 0.074 | 0.060 | 0.046 | -0.214 | -0.078 | -0.098 |
| 2007 | 0.137 | -0.008 | -0.331 | -0.265 | -0.101 | -0.016 | -0.176 | 0.024 | -0.261 | 0.048 |
| 2008 | 0.269 | 0.275 | 0.045 | -0.129 | 0.090 | 0.232 | 0.280 | 0.236 | 0.048 | 0.320 |
| 2009 | -0.057 | -0.081 | 0.092 | 0.069 | 0.142 | 0.059 | 0.135 | 0.216 | 0.253 | 0.108 |
| 2010 | 0.279 | 0.097 | 0.140 | 0.117 | 0.079 | -0.007 | 0.107 | 0.187 | 0.513 | 0.038 |
| 2011 |  |  |  |  |  |  |  |  |  |  |
| 2012 | -0.128 | 0.161 | 0.014 | -0.227 | -0.223 | -0.161 | 0.028 | 0.182 | -0.118 | -0.110 |
| 2013 | -0.016 | 0.039 | 0.104 | -0.010 | -0.038 | -0.072 | -0.087 | -0.058 | 0.159 | -0.154 |
| 2014 | 0.166 | 0.078 | 0.000 | 0.029 | -0.001 | 0.072 | -0.078 | 0.030 | 0.117 | 0.483 |

Table 9.13: Icelandic cod in Division Va. Estimates of fishing mortality 1955-2014 based on ACAM using catch at age and spring and fall bottom survey indices.

| year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1955 | 0.040 | 0.170 | 0.253 | 0.274 | 0.302 | 0.304 | 0.283 | 0.324 | 0.324 | 0.308 | 0.324 | 0.324 |
| 1956 | 0.051 | 0.182 | 0.249 | 0.259 | 0.290 | 0.303 | 0.295 | 0.342 | 0.355 | 0.335 | 0.332 | 0.332 |
| 1957 | 0.081 | 0.215 | 0.274 | 0.272 | 0.301 | 0.328 | 0.327 | 0.363 | 0.364 | 0.331 | 0.298 | 0.298 |
| 1958 | 0.114 | 0.248 | 0.302 | 0.291 | 0.324 | 0.371 | 0.397 | 0.439 | 0.443 | 0.385 | 0.323 | 0.323 |
| 1959 | 0.091 | 0.233 | 0.282 | 0.257 | 0.299 | 0.341 | 0.351 | 0.399 | 0.382 | 0.320 | 0.229 | 0.229 |
| 1960 | 0.101 | 0.233 | 0.295 | 0.292 | 0.338 | 0.398 | 0.428 | 0.476 | 0.475 | 0.385 | 0.270 | 0.270 |
| 1961 | 0.094 | 0.225 | 0.259 | 0.262 | 0.334 | 0.399 | 0.418 | 0.459 | 0.440 | 0.349 | 0.227 | 0.227 |
| 1962 | 0.112 | 0.248 | 0.282 | 0.264 | 0.347 | 0.424 | 0.467 | 0.513 | 0.488 | 0.378 | 0.239 | 0.239 |
| 1963 | 0.130 | 0.283 | 0.328 | 0.309 | 0.383 | 0.492 | 0.587 | 0.646 | 0.625 | 0.462 | 0.285 | 0.285 |
| 1964 | 0.126 | 0.290 | 0.372 | 0.361 | 0.435 | 0.570 | 0.740 | 0.810 | 0.835 | 0.609 | 0.388 | 0.388 |
| 1965 | 0.121 | 0.284 | 0.385 | 0.403 | 0.471 | 0.602 | 0.744 | 0.848 | 0.879 | 0.654 | 0.425 | 0.425 |
| 1966 | 0.094 | 0.253 | 0.341 | 0.382 | 0.491 | 0.622 | 0.780 | 0.915 | 1.007 | 0.784 | 0.532 | 0.532 |
| 1967 | 0.077 | 0.229 | 0.303 | 0.338 | 0.484 | 0.610 | 0.749 | 0.878 | 0.928 | 0.723 | 0.460 | 0.460 |
| 1968 | 0.077 | 0.247 | 0.342 | 0.405 | 0.576 | 0.765 | 1.036 | 1.200 | 1.360 | 1.081 | 0.739 | 0.739 |
| 1969 | 0.056 | 0.232 | 0.322 | 0.354 | 0.505 | 0.608 | 0.719 | 0.837 | 0.871 | 0.714 | 0.444 | 0.444 |
| 1970 | 0.069 | 0.270 | 0.389 | 0.426 | 0.551 | 0.650 | 0.760 | 0.891 | 0.950 | 0.801 | 0.515 | 0.515 |
| 1971 | 0.088 | 0.309 | 0.478 | 0.532 | 0.620 | 0.717 | 0.800 | 0.957 | 1.034 | 0.881 | 0.580 | 0.580 |
| 1972 | 0.088 | 0.302 | 0.480 | 0.553 | 0.649 | 0.730 | 0.791 | 0.959 | 1.060 | 0.912 | 0.602 | 0.602 |
| 1973 | 0.119 | 0.321 | 0.488 | 0.564 | 0.668 | 0.754 | 0.799 | 0.953 | 1.042 | 0.902 | 0.591 | 0.591 |
| 1974 | 0.113 | 0.325 | 0.499 | 0.575 | 0.699 | 0.832 | 0.921 | 1.056 | 1.180 | 1.027 | 0.697 | 0.697 |
| 1975 | 0.108 | 0.310 | 0.502 | 0.601 | 0.722 | 0.884 | 1.022 | 1.127 | 1.253 | 1.101 | 0.772 | 0.772 |
| 1976 | 0.066 | 0.258 | 0.428 | 0.552 | 0.695 | 0.852 | 0.948 | 1.009 | 1.063 | 0.943 | 0.654 | 0.654 |
| 1977 | 0.030 | 0.195 | 0.330 | 0.428 | 0.610 | 0.721 | 0.728 | 0.739 | 0.698 | 0.626 | 0.407 | 0.407 |
| 1978 | 0.027 | 0.174 | 0.281 | 0.354 | 0.525 | 0.603 | 0.547 | 0.549 | 0.484 | 0.447 | 0.281 | 0.281 |
| 1979 | 0.028 | 0.171 | 0.274 | 0.344 | 0.503 | 0.567 | 0.496 | 0.491 | 0.419 | 0.391 | 0.246 | 0.246 |
| 1980 | 0.028 | 0.175 | 0.306 | 0.386 | 0.538 | 0.620 | 0.557 | 0.546 | 0.469 | 0.439 | 0.291 | 0.291 |
| 1981 | 0.023 | 0.176 | 0.353 | 0.488 | 0.648 | 0.819 | 0.850 | 0.818 | 0.751 | 0.691 | 0.519 | 0.519 |
| 1982 | 0.028 | 0.192 | 0.395 | 0.558 | 0.699 | 0.898 | 0.958 | 0.869 | 0.747 | 0.671 | 0.510 | 0.510 |
| 1983 | 0.023 | 0.179 | 0.377 | 0.555 | 0.706 | 0.881 | 0.914 | 0.852 | 0.734 | 0.671 | 0.523 | 0.523 |
| 1984 | 0.039 | 0.200 | 0.378 | 0.531 | 0.675 | 0.805 | 0.752 | 0.702 | 0.596 | 0.559 | 0.432 | 0.432 |
| 1985 | 0.050 | 0.230 | 0.422 | 0.578 | 0.714 | 0.832 | 0.764 | 0.699 | 0.594 | 0.559 | 0.437 | 0.437 |
| 1986 | 0.061 | 0.261 | 0.517 | 0.713 | 0.824 | 0.953 | 0.872 | 0.767 | 0.658 | 0.613 | 0.486 | 0.486 |
| 1987 | 0.055 | 0.272 | 0.555 | 0.816 | 0.905 | 1.058 | 0.991 | 0.847 | 0.742 | 0.691 | 0.568 | 0.568 |
| 1988 | 0.047 | 0.258 | 0.523 | 0.794 | 0.921 | 1.102 | 1.077 | 0.939 | 0.871 | 0.822 | 0.713 | 0.713 |
| 1989 | 0.041 | 0.242 | 0.464 | 0.653 | 0.794 | 0.894 | 0.796 | 0.716 | 0.642 | 0.620 | 0.506 | 0.506 |
| 1990 | 0.050 | 0.250 | 0.472 | 0.662 | 0.789 | 0.856 | 0.745 | 0.683 | 0.614 | 0.595 | 0.484 | 0.484 |
| 1991 | 0.086 | 0.302 | 0.566 | 0.812 | 0.884 | 0.945 | 0.838 | 0.764 | 0.704 | 0.677 | 0.569 | 0.569 |
| 1992 | 0.102 | 0.320 | 0.600 | 0.871 | 0.924 | 1.002 | 0.885 | 0.795 | 0.732 | 0.699 | 0.596 | 0.596 |
| 1993 | 0.138 | 0.313 | 0.556 | 0.805 | 0.889 | 1.030 | 1.017 | 0.923 | 0.887 | 0.841 | 0.753 | 0.753 |
| 1994 | 0.088 | 0.242 | 0.384 | 0.532 | 0.678 | 0.765 | 0.713 | 0.691 | 0.642 | 0.627 | 0.538 | 0.538 |
| 1995 | 0.061 | 0.196 | 0.319 | 0.422 | 0.569 | 0.625 | 0.556 | 0.566 | 0.518 | 0.518 | 0.433 | 0.433 |
| 1996 | 0.036 | 0.161 | 0.283 | 0.412 | 0.558 | 0.624 | 0.576 | 0.591 | 0.545 | 0.538 | 0.460 | 0.460 |
| 1997 | 0.025 | 0.145 | 0.276 | 0.422 | 0.584 | 0.668 | 0.654 | 0.673 | 0.634 | 0.616 | 0.543 | 0.543 |
| 1998 | 0.029 | 0.154 | 0.332 | 0.522 | 0.666 | 0.780 | 0.808 | 0.814 | 0.796 | 0.765 | 0.707 | 0.707 |
| 1999 | 0.044 | 0.177 | 0.395 | 0.656 | 0.751 | 0.871 | 0.918 | 0.891 | 0.874 | 0.834 | 0.789 | 0.789 |
| 2000 | 0.058 | 0.181 | 0.393 | 0.630 | 0.754 | 0.891 | 0.962 | 0.950 | 0.947 | 0.905 | 0.877 | 0.877 |
| 2001 | 0.066 | 0.188 | 0.381 | 0.580 | 0.698 | 0.856 | 0.983 | 1.001 | 1.018 | 0.970 | 0.962 | 0.962 |
| 2002 | 0.043 | 0.164 | 0.338 | 0.485 | 0.595 | 0.704 | 0.807 | 0.859 | 0.856 | 0.826 | 0.802 | 0.802 |
| 2003 | 0.031 | 0.149 | 0.332 | 0.496 | 0.571 | 0.645 | 0.693 | 0.747 | 0.731 | 0.716 | 0.681 | 0.681 |
| 2004 | 0.031 | 0.144 | 0.332 | 0.528 | 0.579 | 0.650 | 0.685 | 0.729 | 0.711 | 0.699 | 0.666 | 0.666 |
| 2005 | 0.030 | 0.126 | 0.292 | 0.480 | 0.547 | 0.624 | 0.662 | 0.703 | 0.693 | 0.684 | 0.651 | 0.651 |
| 2006 | 0.029 | 0.119 | 0.264 | 0.461 | 0.534 | 0.625 | 0.678 | 0.712 | 0.708 | 0.695 | 0.666 | 0.666 |
| 2007 | 0.027 | 0.108 | 0.229 | 0.383 | 0.486 | 0.593 | 0.672 | 0.717 | 0.732 | 0.717 | 0.697 | 0.697 |
| 2008 | 0.021 | 0.087 | 0.177 | 0.292 | 0.398 | 0.472 | 0.491 | 0.516 | 0.486 | 0.477 | 0.427 | 0.427 |
| 2009 | 0.030 | 0.093 | 0.182 | 0.300 | 0.397 | 0.466 | 0.473 | 0.477 | 0.429 | 0.416 | 0.361 | 0.361 |
| 2010 | 0.027 | 0.086 | 0.159 | 0.253 | 0.350 | 0.405 | 0.390 | 0.397 | 0.343 | 0.336 | 0.279 | 0.279 |
| 2011 | 0.027 | 0.084 | 0.151 | 0.230 | 0.316 | 0.356 | 0.323 | 0.326 | 0.265 | 0.256 | 0.200 | 0.200 |
| 2012 | 0.028 | 0.085 | 0.154 | 0.234 | 0.316 | 0.353 | 0.320 | 0.321 | 0.258 | 0.247 | 0.192 | 0.192 |
| 2013 | 0.042 | 0.096 | 0.166 | 0.244 | 0.321 | 0.359 | 0.333 | 0.339 | 0.275 | 0.262 | 0.205 | 0.205 |
| 2014 | 0.034 | 0.090 | 0.153 | 0.226 | 0.309 | 0.345 | 0.321 | 0.330 | 0.262 | 0.242 | 0.187 | 0.187 |

Table 9.14: Icelandic cod in Division Va. Estimates of numbers at age in the stock 1955-2015 based on ACAM using catch at age and spring and fall bottom survey indices.

| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1955 | 170.63 | 152.80 | 151.99 | 217.61 | 211.93 | 115.44 | 36.04 | 24.57 | 12.95 | 87.57 | 9.20 | 7.81 | 8.13 | 2.64 |
| 1956 | 220.71 | 170.63 | 152.80 | 119.55 | 150.28 | 134.76 | 71.84 | 21.81 | 14.84 | 7.99 | 51.87 | 5.45 | 4.70 | 4.82 |
| 1957 | 289.05 | 220.71 | 170.63 | 118.88 | 81.58 | 95.88 | 85.20 | 44.02 | 13.18 | 9.05 | 4.65 | 29.78 | 3.19 | 2.76 |
| 1958 | 154.36 | 289.05 | 220.71 | 128.81 | 78.52 | 50.80 | 59.83 | 51.65 | 35.18 | 7.78 | 5.16 | 2.64 | 17.50 | 1.94 |
| 1959 | 192.90 | 154.36 | 289.05 | 161.18 | 82.26 | 47.54 | 31.10 | 35.44 | 51.54 | 19.36 | 4.11 | 2.71 | 1.47 | 10.37 |
| 1960 | 128.94 | 192.90 | 154.36 | 216.11 | 104.52 | 50.78 | 30.11 | 18.89 | 20.63 | 37.50 | 10.63 | 2.30 | 1.61 | 0.96 |
| 1961 | 177.51 | 128.94 | 192.90 | 114.27 | 140.14 | 63.72 | 31.06 | 17.59 | 10.39 | 11.00 | 19.06 | 5.42 | 1.28 | 1.01 |
| 1962 | 204.04 | 177.51 | 128.94 | 143.78 | 74.70 | 88.53 | 40.16 | 18.22 | 23.60 | 5.60 | 5.69 | 10.05 | 3.13 | 0.83 |
| 1963 | 216.37 | 204.04 | 177.51 | 94.42 | 91.85 | 46.14 | 55.68 | 23.25 | 9.76 | 12.11 | 2.74 | 2.86 | 5.64 | 2.02 |
| 1964 | 229.14 | 216.37 | 204.04 | 127.66 | 58.25 | 54.17 | 27.74 | 31.08 | 11.64 | 4.44 | 5.20 | 1.20 | 1.48 | 3.47 |
| 1965 | 320.13 | 229.14 | 216.37 | 147.31 | 78.18 | 32.87 | 30.93 | 14.70 | 14.38 | 4.54 | 1.62 | 1.85 | 0.54 | 0.82 |
| 1966 | 171.90 | 320.13 | 229.14 | 157.00 | 90.79 | 43.57 | 17.99 | 15.81 | 6.59 | 5.60 | 1.59 | 0.55 | 0.79 | 0.29 |
| 1967 | 247.44 | 171.90 | 320.12 | 170.76 | 99.77 | 52.85 | 24.36 | 9.01 | 6.95 | 2.47 | 1.83 | 0.48 | 0.21 | 0.38 |
| 1968 | 180.46 | 247.44 | 171.90 | 242.76 | 111.19 | 60.32 | 30.86 | 12.29 | 4.01 | 2.69 | 0.84 | 0.59 | 0.19 | 0.11 |
| 1969 | 188.59 | 180.46 | 247.44 | 130.34 | 155.31 | 64.68 | 32.92 | 41.21 | 4.68 | 1.17 | 0.66 | 0.18 | 0.16 | 0.07 |
| 1970 | 139.26 | 188.59 | 180.46 | 191.59 | 84.59 | 92.10 | 37.16 | 32.91 | 18.36 | 1.87 | 0.41 | 0.23 | 0.07 | 0.09 |
| 1971 | 273.05 | 139.26 | 188.59 | 137.94 | 119.79 | 46.92 | 49.26 | 17.53 | 14.06 | 7.03 | 0.63 | 0.13 | 0.08 | 0.03 |
| 1972 | 178.96 | 273.05 | 139.25 | 141.35 | 82.92 | 60.78 | 22.56 | 21.69 | 23.28 | 5.18 | 2.21 | 0.18 | 0.04 | 0.04 |
| 1973 | 260.74 | 178.96 | 273.05 | 104.45 | 85.58 | 42.03 | 28.62 | 9.65 | 8.56 | 8.64 | 1.62 | 0.63 | 0.06 | 0.02 |
| 1974 | 367.20 | 260.74 | 178.96 | 198.53 | 62.06 | 42.99 | 19.57 | 12.02 | 3.71 | 3.15 | 2.73 | 0.47 | 0.21 | 0.03 |
| 1975 | 143.27 | 367.20 | 260.74 | 130.82 | 117.50 | 30.85 | 19.81 | 7.96 | 4.28 | 1.21 | 0.90 | 0.69 | 0.14 | 0.09 |
| 1976 | 227.53 | 143.27 | 367.20 | 191.59 | 78.59 | 58.23 | 13.85 | 7.88 | 2.69 | 1.26 | 0.32 | 0.21 | 0.19 | 0.05 |
| 1977 | 243.31 | 227.53 | 143.27 | 281.34 | 121.16 | 41.95 | 27.46 | 5.66 | 2.75 | 0.85 | 0.38 | 0.09 | 0.07 | 0.08 |
| 1978 | 139.94 | 243.31 | 227.53 | 113.78 | 189.47 | 71.34 | 22.38 | 12.22 | 2.25 | 1.09 | 0.33 | 0.15 | 0.04 | 0.04 |
| 1979 | 140.35 | 139.94 | 243.31 | 181.28 | 78.28 | 117.08 | 40.97 | 10.84 | 5.48 | 1.07 | 0.51 | 0.17 | 0.08 | 0.02 |
| 1980 | 131.63 | 140.35 | 139.94 | 193.66 | 125.12 | 48.71 | 71.85 | 20.29 | 5.03 | 2.73 | 0.54 | 0.28 | 0.09 | 0.05 |
| 1981 | 232.93 | 131.63 | 140.35 | 111.39 | 133.11 | 75.43 | 27.11 | 47.08 | 8.93 | 2.36 | 1.29 | 0.27 | 0.15 | 0.06 |
| 1982 | 139.03 | 232.93 | 131.63 | 112.32 | 76.48 | 76.57 | 37.90 | 11.61 | 16.99 | 3.13 | 0.85 | 0.50 | 0.11 | 0.07 |
| 1983 | 140.29 | 139.03 | 232.93 | 104.83 | 75.87 | 42.18 | 35.89 | 15.42 | 3.87 | 5.33 | 1.07 | 0.33 | 0.21 | 0.06 |
| 1984 | 329.65 | 140.29 | 139.03 | 186.28 | 71.79 | 42.60 | 19.83 | 14.51 | 5.23 | 1.27 | 1.86 | 0.42 | 0.14 | 0.10 |
| 1985 | 260.38 | 329.65 | 140.29 | 109.52 | 124.82 | 40.28 | 20.51 | 8.27 | 5.31 | 2.02 | 0.52 | 0.84 | 0.20 | 0.07 |
| 1986 | 175.50 | 260.38 | 329.65 | 109.26 | 71.25 | 66.99 | 18.51 | 8.22 | 2.95 | 2.02 | 0.82 | 0.23 | 0.39 | 0.10 |
| 1987 | 89.19 | 175.50 | 260.38 | 253.82 | 68.87 | 34.80 | 26.88 | 6.65 | 2.60 | 1.01 | 0.77 | 0.35 | 0.10 | 0.20 |
| 1988 | 130.60 | 89.19 | 175.50 | 201.67 | 158.38 | 32.37 | 12.59 | 8.90 | 1.89 | 0.79 | 0.35 | 0.30 | 0.14 | 0.05 |
| 1989 | 106.87 | 130.60 | 89.19 | 137.09 | 127.58 | 76.85 | 11.98 | 4.10 | 2.42 | 0.53 | 0.25 | 0.12 | 0.11 | 0.06 |
| 1990 | 174.57 | 106.87 | 130.60 | 70.11 | 88.14 | 100.05 | 32.73 | 4.43 | 1.37 | 0.89 | 0.21 | 0.11 | 0.05 | 0.05 |
| 1991 | 135.55 | 174.57 | 106.87 | 101.69 | 44.69 | 45.02 | 42.25 | 12.18 | 1.54 | 0.53 | 0.37 | 0.09 | 0.05 | 0.03 |
| 1992 | 77.73 | 135.55 | 174.57 | 80.31 | 61.57 | 20.77 | 16.37 | 14.29 | 3.88 | 0.55 | 0.20 | 0.15 | 0.04 | 0.02 |
| 1993 | 151.18 | 77.73 | 135.55 | 129.10 | 47.75 | 27.67 | 7.12 | 5.32 | 4.30 | 1.31 | 0.20 | 0.08 | 0.06 | 0.02 |
| 1994 | 165.62 | 151.18 | 77.73 | 96.66 | 77.30 | 22.43 | 10.13 | 2.40 | 1.55 | 1.27 | 0.43 | 0.07 | 0.03 | 0.02 |
| 1995 | 88.26 | 165.62 | 151.18 | 58.25 | 62.15 | 43.09 | 10.79 | 4.21 | 0.91 | 0.62 | 0.52 | 0.18 | 0.03 | 0.01 |
| 1996 | 161.54 | 88.26 | 165.62 | 116.40 | 39.21 | 36.98 | 23.14 | 5.00 | 1.85 | 0.43 | 0.29 | 0.25 | 0.09 | 0.02 |
| 1997 | 70.80 | 161.54 | 88.26 | 130.78 | 81.16 | 24.19 | 20.05 | 10.84 | 2.19 | 0.85 | 0.19 | 0.14 | 0.12 | 0.05 |
| 1998 | 171.61 | 70.80 | 161.54 | 70.44 | 92.62 | 50.43 | 12.99 | 9.16 | 4.55 | 0.93 | 0.36 | 0.08 | 0.06 | 0.06 |
| 1999 | 161.43 | 171.61 | 70.80 | 128.52 | 49.45 | 54.43 | 24.49 | 5.47 | 3.44 | 1.66 | 0.34 | 0.13 | 0.03 | 0.02 |
| 2000 | 158.89 | 161.43 | 171.61 | 55.44 | 88.17 | 27.28 | 23.13 | 9.46 | 1.87 | 1.12 | 0.56 | 0.12 | 0.05 | 0.01 |
| 2001 | 178.28 | 158.89 | 161.43 | 132.54 | 37.86 | 48.71 | 11.89 | 8.91 | 3.18 | 0.59 | 0.36 | 0.18 | 0.04 | 0.02 |
| 2002 | 80.20 | 178.28 | 158.89 | 123.70 | 89.96 | 21.18 | 22.33 | 4.84 | 3.10 | 0.97 | 0.18 | 0.11 | 0.05 | 0.01 |
| 2003 | 154.41 | 80.20 | 178.28 | 124.62 | 85.99 | 52.52 | 10.68 | 10.08 | 1.96 | 1.13 | 0.34 | 0.06 | 0.04 | 0.02 |
| 2004 | 135.22 | 154.41 | 80.20 | 141.47 | 87.90 | 50.53 | 26.18 | 4.94 | 4.33 | 0.80 | 0.44 | 0.13 | 0.02 | 0.02 |
| 2005 | 97.21 | 135.22 | 154.41 | 63.66 | 100.32 | 51.62 | 24.39 | 12.02 | 2.11 | 1.79 | 0.32 | 0.18 | 0.05 | 0.01 |
| 2006 | 133.86 | 97.21 | 135.22 | 122.63 | 45.93 | 61.31 | 26.14 | 11.56 | 5.27 | 0.89 | 0.72 | 0.13 | 0.07 | 0.02 |
| 2007 | 119.41 | 133.86 | 97.21 | 107.57 | 89.13 | 28.89 | 31.66 | 12.55 | 5.07 | 2.19 | 0.36 | 0.29 | 0.05 | 0.03 |
| 2008 | 130.26 | 119.41 | 133.86 | 77.50 | 79.07 | 58.06 | 16.13 | 15.94 | 5.68 | 2.12 | 0.88 | 0.14 | 0.12 | 0.02 |
| 2009 | 172.28 | 130.26 | 119.41 | 107.28 | 58.14 | 63.88 | 35.48 | 8.87 | 8.14 | 2.85 | 1.03 | 0.44 | 0.07 | 0.06 |
| 2010 | 176.98 | 172.28 | 130.26 | 94.88 | 80.02 | 39.68 | 38.74 | 19.54 | 4.56 | 4.16 | 1.45 | 0.55 | 0.24 | 0.04 |
| 2011 | 120.90 | 176.98 | 172.28 | 103.79 | 71.27 | 55.87 | 25.23 | 22.36 | 10.67 | 2.53 | 2.29 | 0.84 | 0.32 | 0.15 |
| 2012 | 180.79 | 120.90 | 176.98 | 137.24 | 78.14 | 50.15 | 36.32 | 15.06 | 12.82 | 6.32 | 1.49 | 1.44 | 0.53 | 0.22 |
| 2013 | 160.72 | 180.79 | 120.90 | 140.96 | 103.19 | 54.85 | 32.48 | 21.68 | 8.66 | 7.62 | 3.76 | 0.94 | 0.92 | 0.36 |
| 2014 | 115.38 | 160.73 | 180.79 | 94.92 | 104.83 | 71.55 | 35.18 | 19.30 | 12.39 | 5.08 | 4.44 | 2.34 | 0.59 | 0.61 |
| 2015 | 186.27 | 115.38 | 160.72 | 143.06 | 71.05 | 73.64 | 46.72 | 21.15 | 11.19 | 7.36 | 2.99 | 2.80 | 1.50 | 0.40 |

Table 9.15: Icelandic cod in Division Va. Landings (thousand tonnes, average fishing mortality of age groups 5 to 10, recruitment to the fisheries at age 3 (millions), reference fishing biomass (B4+, thousand tonnes), spawning stock biomass (thousand tonnes) at spawning time and harvest ratio.

| Year | Yield | F5-10 | SSB | Reference biomass | Recruits | Harvest rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1955 | 545.250 | 0.290 | 941.981 | 2361.190 | 151.988 | 0.222 |
| 1956 | 486.909 | 0.290 | 795.988 | 2085.480 | 152.797 | 0.228 |
| 1957 | 455.182 | 0.311 | 775.735 | 1881.630 | 170.633 | 0.240 |
| 1958 | 517.359 | 0.354 | 875.551 | 1867.800 | 220.714 | 0.271 |
| 1959 | 459.081 | 0.322 | 853.637 | 1829.410 | 289.052 | 0.249 |
| 1960 | 470.121 | 0.371 | 708.947 | 1753.890 | 154.360 | 0.269 |
| 1961 | 377.291 | 0.355 | 467.292 | 1496.600 | 192.903 | 0.253 |
| 1962 | 388.985 | 0.383 | 568.921 | 1492.510 | 128.941 | 0.262 |
| 1963 | 408.800 | 0.458 | 507.766 | 1315.590 | 177.514 | 0.313 |
| 1964 | 437.012 | 0.548 | 451.023 | 1219.020 | 204.038 | 0.352 |
| 1965 | 387.106 | 0.575 | 317.608 | 1022.480 | 216.371 | 0.356 |
| 1966 | 353.357 | 0.589 | 277.232 | 1031.370 | 229.142 | 0.333 |
| 1967 | 335.721 | 0.560 | 256.458 | 1102.690 | 320.125 | 0.300 |
| 1968 | 381.770 | 0.721 | 221.565 | 1222.540 | 171.900 | 0.314 |
| 1969 | 403.205 | 0.558 | 313.603 | 1325.290 | 247.444 | 0.307 |
| 1970 | 475.077 | 0.611 | 331.007 | 1336.640 | 180.463 | 0.345 |
| 1971 | 444.248 | 0.684 | 242.420 | 1097.650 | 188.594 | 0.393 |
| 1972 | 395.166 | 0.694 | 221.685 | 996.656 | 139.255 | 0.401 |
| 1973 | 369.205 | 0.705 | 245.326 | 843.598 | 273.048 | 0.431 |
| 1974 | 368.133 | 0.764 | 186.955 | 918.054 | 178.957 | 0.393 |
| 1975 | 364.754 | 0.810 | 168.249 | 895.165 | 260.742 | 0.407 |
| 1976 | 346.253 | 0.747 | 138.466 | 955.181 | 367.198 | 0.359 |
| 1977 | 340.086 | 0.593 | 198.573 | 1288.760 | 143.272 | 0.261 |
| 1978 | 329.602 | 0.477 | 212.192 | 1296.860 | 227.531 | 0.259 |
| 1979 | 366.462 | 0.446 | 303.839 | 1396.080 | 243.309 | 0.260 |
| 1980 | 432.237 | 0.492 | 356.551 | 1489.030 | 139.936 | 0.288 |
| 1981 | 465.032 | 0.663 | 263.850 | 1241.260 | 140.346 | 0.363 |
| 1982 | 380.068 | 0.730 | 167.079 | 969.996 | 131.631 | 0.384 |
| 1983 | 298.049 | 0.714 | 129.989 | 790.639 | 232.927 | 0.367 |
| 1984 | 282.022 | 0.641 | 140.961 | 913.010 | 139.028 | 0.316 |
| 1985 | 323.428 | 0.668 | 163.443 | 926.811 | 140.289 | 0.350 |
| 1986 | 364.797 | 0.774 | 195.872 | 853.810 | 329.650 | 0.421 |
| 1987 | 389.915 | 0.862 | 150.765 | 1029.860 | 260.379 | 0.372 |
| 1988 | 377.554 | 0.893 | 167.899 | 1031.580 | 175.504 | 0.368 |
| 1989 | 363.125 | 0.719 | 172.633 | 1001.710 | 89.187 | 0.358 |
| 1990 | 335.316 | 0.701 | 213.825 | 840.680 | 130.600 | 0.407 |
| 1991 | 307.759 | 0.802 | 164.307 | 697.921 | 106.873 | 0.443 |
| 1992 | 264.834 | 0.846 | 150.513 | 550.430 | 174.568 | 0.472 |
| 1993 | 250.704 | 0.870 | 121.063 | 595.107 | 135.552 | 0.423 |
| 1994 | 178.138 | 0.627 | 156.469 | 576.020 | 77.726 | 0.313 |
| 1995 | 168.592 | 0.510 | 176.977 | 556.752 | 151.183 | 0.298 |
| 1996 | 180.701 | 0.507 | 158.996 | 670.328 | 165.618 | 0.269 |
| 1997 | 203.112 | 0.546 | 189.069 | 782.972 | 88.259 | 0.258 |
| 1998 | 243.987 | 0.653 | 200.247 | 720.515 | 161.541 | 0.330 |
| 1999 | 260.147 | 0.747 | 178.018 | 731.021 | 70.798 | 0.352 |
| 2000 | 235.092 | 0.763 | 165.971 | 590.023 | 171.607 | 0.385 |
| 2001 | 234.229 | 0.750 | 159.861 | 687.596 | 161.433 | 0.333 |
| 2002 | 208.487 | 0.631 | 197.402 | 728.295 | 158.887 | 0.284 |
| 2003 | 207.543 | 0.581 | 189.739 | 738.905 | 178.277 | 0.279 |
| 2004 | 226.762 | 0.584 | 200.710 | 798.720 | 80.200 | 0.279 |
| 2005 | 213.403 | 0.551 | 228.099 | 721.935 | 154.405 | 0.296 |
| 2006 | 196.077 | 0.545 | 222.803 | 698.554 | 135.220 | 0.279 |
| 2007 | 170.300 | 0.513 | 207.394 | 679.381 | 97.208 | 0.251 |
| 2008 | 146.104 | 0.391 | 269.852 | 703.665 | 133.857 | 0.217 |
| 2009 | 181.151 | 0.382 | 254.908 | 798.690 | 119.407 | 0.221 |
| 2010 | 168.880 | 0.326 | 293.057 | 843.149 | 130.256 | 0.201 |
| 2011 | 170.425 | 0.284 | 367.435 | 936.570 | 172.277 | 0.184 |
| 2012 | 194.795 | 0.283 | 410.695 | 1055.700 | 176.983 | 0.184 |
| 2013 | 223.277 | 0.294 | 452.691 | 1175.500 | 120.903 | 0.189 |
| 2014 | 221.340 | 0.281 | 424.930 | 1180.980 | 180.786 | 0.187 |
| 2015 |  | 0.248 | 546.658 | 1302.040 | 160.725 |  |

Table 9.16: Icelandic cod in Division Va. Inputs in the deterministic predictions.

| Age | Parameter | 2015 | 2016 | 2017 |
| ---: | :--- | ---: | ---: | ---: |
| 3 | Catch weights | 1.421 | 1.421 | 1.421 |
| 4 | Catch weights | 1.814 | 1.814 | 1.814 |
| 5 | Catch weights | 2.757 | 2.757 | 2.757 |
| 6 | Catch weights | 3.542 | 3.542 | 3.542 |
| 7 | Catch weights | 4.674 | 4.674 | 4.674 |
| 8 | Catch weights | 6.039 | 6.039 | 6.039 |
| 9 | Catch weights | 8.145 | 8.145 | 8.145 |
| 10 | Catch weights | 9.187 | 9.187 | 9.187 |
| 11 | Catch weights | 9.722 | 9.722 | 9.722 |
| 12 | Catch weights | 10.523 | 10.523 | 10.523 |
| 13 | Catch weights | 11.448 | 11.448 | 11.448 |
| 14 | Catch weights | 12.992 | 12.992 | 12.992 |
| 3 | SSB weights | 0.710 | 0.709 | 0.709 |
| 4 | SSB weights | 1.047 | 1.047 | 1.046 |
| 5 | SSB weights | 3.311 | 3.310 | 3.309 |
| 6 | SSB weights | 3.833 | 3.832 | 3.830 |
| 7 | SSB weights | 4.902 | 4.900 | 4.898 |
| 8 | SSB weights | 6.240 | 6.238 | 6.235 |
| 9 | SSB weights | 8.728 | 8.726 | 8.722 |
| 10 | SSB weights | 9.705 | 9.703 | 9.698 |
| 11 | SSB weights | 9.722 | 9.720 | 9.715 |
| 12 | SSB weights | 10.523 | 10.520 | 10.515 |
| 13 | SSB weights | 11.448 | 11.445 | 11.440 |
| 14 | SSB weights | 12.992 | 12.988 | 12.982 |
| 3 | Maturity | 0.004 | 0.004 | 0.004 |
| 4 | Maturity | 0.007 | 0.007 | 0.007 |
| 5 | Maturity | 0.109 | 0.109 | 0.109 |
| 6 | Maturity | 0.353 | 0.353 | 0.353 |
| 7 | Maturity | 0.638 | 0.638 | 0.638 |
| 8 | Maturity | 0.908 | 0.908 | 0.908 |
| 9 | Maturity | 0.979 | 0.979 | 0.979 |
| 10 | Maturity | 0.988 | 0.988 | 0.988 |
| 11 | Maturity | 1.000 | 1.000 | 1.000 |
| 12 | Maturity | 1.000 | 1.000 | 1.000 |
| 13 | Maturity | 1.000 | 1.000 | 1.000 |
| 14 | Maturity | 1.000 | 1.000 | 1.000 |
| 3 | Selection | 0.121 | 0.121 | 0.121 |
| 4 | Selection | 0.316 | 0.316 | 0.316 |
| 5 | Selection | 0.552 | 0.552 | 0.552 |
| 6 | Selection | 0.822 | 0.822 | 0.822 |
| 7 | Selection | 1.103 | 1.103 | 1.103 |
| 8 | Selection | 1.232 | 1.232 | 1.232 |
| 9 | Selection | 1.136 | 1.136 | 1.136 |
| 10 | Selection | 1.154 | 1.154 | 1.154 |
| 11 | Selection | 0.791 | 0.791 | 0.791 |
| 12 | Selection | 0.791 | 0.791 | 0.791 |
| 13 | Selection | 0.791 | 0.791 | 0.791 |
| 14 | Selection | 0.791 | 0.791 | 0.791 |
| 3 | Stock numbers | 160.726 | 115.383 | 186.271 |
| 4 | Stock numbers | 143.066 |  |  |
| 5 | Stock numbers | 71.055 |  |  |
| 6 | Stock numbers | 73.645 |  |  |
| 7 | Stock numbers | 46.722 |  |  |
| 8 | Stock numbers | 21.149 |  |  |
| 10 | Stock numbers | 11.189 |  |  |
| 12 | Stock numbers | 7.362 |  |  |
| 14 | Stock numbers numbers | 2.992 |  |  |
|  | Stock numbers numbers | 1.799 |  |  |
|  | 0.404 |  |  |  |

Table 9.17: Icelandic cod in Division Va. Output of the deterministic predictions.

| Year | B4. | Fmult | Fbar | SSB | Landings | 2017.B4. | 2017.SSB | SSB.change | TAC.change |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 1302 |  | 0.26 | 544 | 228 |  |  |  |  |
| 2016 | 1371 | 0.00 | 0.00 | 660 | 0 | 1622 | 896 | 36\% | -100\% |
|  |  | 0.23 | 0.06 | 643 | 60 | 1553 | 823 | 28\% | -74\% |
|  |  | 0.27 | 0.07 | 640 | 70 | 1542 | 812 | 27\% | -69\% |
|  |  | 0.31 | 0.08 | 637 | 79 | 1531 | 801 | 26\% | -65\% |
|  |  | 0.34 | 0.09 | 634 | 89 | 1520 | 789 | 24\% | -61\% |
|  |  | 0.38 | 0.10 | 632 | 98 | 1509 | 778 | 23\% | -57\% |
|  |  | 0.42 | 0.11 | 629 | 108 | 1499 | 768 | 22\% | -53\% |
|  |  | 0.46 | 0.12 | 626 | 117 | 1488 | 757 | 21\% | -49\% |
|  |  | 0.50 | 0.13 | 623 | 126 | 1478 | 746 | 20\% | -45\% |
|  |  | 0.54 | 0.14 | 621 | 135 | 1467 | 736 | 19\% | -41\% |
|  |  | 0.57 | 0.15 | 618 | 144 | 1457 | 726 | 17\% | -37\% |
|  |  | 0.61 | 0.16 | 615 | 153 | 1447 | 716 | 16\% | -33\% |
|  |  | 0.65 | 0.17 | 613 | 162 | 1437 | 706 | 15\% | -29\% |
|  |  | 0.69 | 0.18 | 610 | 171 | 1427 | 696 | 14\% | -25\% |
|  |  | 0.73 | 0.19 | 607 | 180 | 1417 | 686 | 13\% | -21\% |
|  |  | 0.76 | 0.20 | 605 | 188 | 1407 | 677 | 12\% | -17\% |
|  |  | 0.80 | 0.21 | 602 | 197 | 1397 | 668 | 11\% | -14\% |
|  |  | 0.84 | 0.22 | 599 | 205 | 1388 | 658 | 10\% | -10\% |
|  |  | 0.88 | 0.23 | 597 | 214 | 1378 | 649 | 9\% | -6\% |
|  |  | 0.92 | 0.24 | 594 | 222 | 1369 | 640 | 8\% | -3\% |
|  |  | 0.96 | 0.25 | 592 | 230 | 1359 | 632 | 7\% | 1\% |
|  |  | 0.99 | 0.26 | 589 | 239 | 1350 | 623 | 6\% | 5\% |
|  |  | 1.03 | 0.27 | 586 | 247 | 1341 | 614 | 5\% | 8\% |
|  |  | 1.07 | 0.28 | 584 | 255 | 1332 | 606 | 4\% | 12\% |
|  |  | 1.11 | 0.29 | 581 | 263 | 1322 | 598 | 3\% | 15\% |
|  |  | 1.15 | 0.30 | 579 | 271 | 1313 | 590 | 2\% | 19\% |
|  |  | 1.19 | 0.31 | 576 | 279 | 1305 | 582 | 1\% | 22\% |
|  |  | 1.22 | 0.32 | 574 | 286 | 1296 | 574 | -0\% | 26\% |
|  |  | 1.26 | 0.33 | 571 | 294 | 1287 | 566 | -1\% | 29\% |
|  |  | 1.30 | 0.34 | 569 | 302 | 1278 | 558 | -2\% | $32 \%$ |
|  |  | 1.34 | 0.35 | 566 | 309 | 1270 | 551 | -3\% | 36\% |
|  |  | 1.38 | 0.36 | 564 | 317 | 1261 | 543 | -4\% | 39\% |
|  |  | 1.41 | 0.37 | 561 | 324 | 1253 | 536 | -5\% | 42\% |
|  |  | 1.45 | 0.38 | 559 | 332 | 1244 | 529 | -5\% | 45\% |
|  |  | 1.49 | 0.39 | 557 | 339 | 1236 | 521 | -6\% | 49\% |
|  |  | 1.53 | 0.40 | 554 | 346 | 1228 | 514 | -7\% | 52\% |
|  |  | 1.57 | 0.41 | 552 | 353 | 1220 | 508 | -8\% | 55\% |
|  |  | 1.61 | 0.42 | 549 | 361 | 1212 | 501 | -9\% | 58\% |
|  |  | 1.64 | 0.43 | 547 | 368 | 1204 | 494 | -10\% | 61\% |
|  |  | 1.68 | 0.44 | 545 | 375 | 1196 | 487 | -11\% | 64\% |
|  |  | 1.72 | 0.45 | 542 | 382 | 1188 | 481 | -11\% | 67\% |
|  |  | 1.76 | 0.46 | 540 | 389 | 1180 | 474 | -12\% | 70\% |



Figure 9.1: Icelandic cod division Va. Total landings from 1905 to 2014 and landings by principal gear from 1955 to 2014. The proportion of landings by each gear is shown by the red line.


Figure 9.2: Icelandic cod division Va. Distribution of bottom trawl catches (tonnes per square km) in 2014. Shaded areas represent permanent areas closed to bottom trawl fishing.


Figure 9.3: Icelandic cod division Va. Distribution of long line catches (tonnes per square $\mathbf{k m}$ ) in 2014. Shaded areas represent permanent areas closed to long line fishing.


Figure 9.4: Icelandic cod division Va. Estimated weight at age (numbers in panels indicate age classes) in the catches 1985-2015 expressed as deviation from the mean. Weights at age in 2015 are predicted from 2015 spring survey weights. Note that values that are equal to the mean are not visible in this type of a plot.


Figure 9.5: Icelandic cod division Va. Estimated weight at age (numbers in panel indicate age classes) in the spring survey 1985-2014 (SMB) and fall survey 1996-2013 (SMH) expressed as proportional deviations from the mean. No fall survey was conducted in 2011. Note that values that are equal to the mean are not visible in this type of a plot.


Figure 9.6: Icelandic cod division Va. Abundance indices of cod in the groundfish survey in spring 1985-2014 (SMB red, longer time series) and fall 1996-2013 (SMH blue, shorter time series). Bottom left) Biomass index of 55 cm and larger, bottom right) Biomass index 80 cm and larger, top right) Abundance index of $<55 \mathrm{~cm}$, top left) Abundance index of $<18 \mathrm{~cm}$ fish. The shaded area and the vertical bar show 1 standard error of the estimate.


Figure 9.7: Icelandic cod division Va. Age based abundance indices of cod in the groundfish survey in spring 1985-2015 (SMB) and fall 1996-2014 (SMH). The indices are standardized within each age group and within each survey.


Figure 9.8: Catch residuals (left), spring survey residuals (SMB, middle) and fall survey residuals (SMH, right) by year and age from the spaly ADCAM run. Note that values that are equal to the mean are not visible in this type of a plot and that no survey was carried out in the fall 2011.


Figure 9.9: Icelandic cod in division Va. Assessment summary based ADCAM tuned with the spring and the fall survey.


Figure 9.10: Icelandic cod in division Va. Comparison of different stock trajectories using alternative model frameworks, input and assumptions. The medium term simulations are based on the following harvest rates: 0.22 (smb), $0.20(\mathrm{smx})$ and 0.18 ( smh ).


Figure 9.11: Icelandic cod in division Va. Spawning stock biomass and corresponding recruitment at age 3. The numerical values refer to year class with the horizontal lines referring to geometric mean recruitment for year classes 1954-1984 (red line) and 1985-2014 (green line). Vertical lines refer to Blim (Bloss, red) and Btrigger (green).


Figure 9.12: Icelandic cod in division Va. Empirical retrospective patterns from the 2004 to 2015 (this years assessment, marked in red) assessments as summarized in ICES annual advisory sheet.


Figure 9.13: Icelandic cod in division Va. Distribution of realized harvest rate when TAC is based on the current catch rule ( $20 \%$ harvest rate and the catch stabilizer). The upper and lower $5 \%$ of the realized harvest rate is 0.15 and 0.27 . The distribution is based on the last 5 years in the simulations, the number of iterations being 1225 .

## 10 Icelandic haddock

The 2014 year class is estimated to be large, after 6 consecutive small year class from 2008-2013. The Current assessment shows some upward revision of the stock compared to last years assessment, mostly caused by more growth than predicted. The main features are though the same that recruitment is poor and the stock is predicted to decrease somewhat in next two years before the 2014 year-class recruits to the stock.

Growth in 2014 was above average since 1985 and more than predicted and the mean weight of young fish is above average while old fish are close to average. The assessment procedure was the same as last year (SPALY), an Adapt type model tuned with both the surveys.

There are differences in the perception of the state of stock in assessment based on either the spring or autumn survey with autumn survey indicating a larger stock. It has been like that since 2009. Different models using the same tuning data show similar results.

Advice is given according to the adopted Harvest Control Rule, and the advice for the fishing year 2014/2015 (September $1^{\text {st }} 2015$ - August 31st 2016) is 36400 tonnes. The advice for the following fishing year is predicted to be approximately 31000 tonnes but increasing after that when the 2014 year class comes in.

No environmental drivers or ecosystem effects are known that can help in prediction of the development of the haddock stock. Some effect of the environment on the stock can though not be excluded.

### 10.1 Data

Landings of Icelandic haddock in 2014 are estimated to have been 33900 tonnes, see Figure 10.1.1 and Table 10.1.1., the lowest since World War II. Of the landings, 33000 tonnes are caught by Iceland and 900 tonnes by the Faeroese. The landings have decreased from 100 thous. tonnes between 2005-2008. The proportion of haddock caught by longliners was $48 \%$ in 2014, the highest ever (Figure 10.1.2). On longer time scale the share of longlines increasing in last 15 years, while the proportion of haddock caught in gillnets is now very small. Spatial distribution of the landings does not change very much from year to year but catches from the area north of Iceland have increased gradually over the last 10-15 years. (Figure 10.1.3 and 10.1.8).

Catch in numbers at age is shown in Table 10.1.2 and Figure 10.1.4. Age 7 accounted for $35 \%$ of the landings and age 9 and older for $17.5 \%$ while the average contribution of age 9 and older is $4.5 \%$. The results for are close to expectation (Figure 10.1.5).

The index of total biomass from the groundfish surveys in March and October is shown in Figure 10.1.8. Both surveys show much increase between 2002 and 2005 but considerable decrease from 2007-2010. The difference in perception of the stock between the surveys is that the autumn survey shows less contrast between periods of large and small stock. In recent years the assessment has predicted reduced biomass while the reality has been unchanged biomass in the March survey and some increase in the autumn survey, causing upwards revision of the stock in each assessment compared to earlier assessments.

Age disaggregated indices from the March survey are given in Table 10.1.3 and indices from the autumn survey in Table 10.1.4. Abundance of age groups 2-5 and in the 2015 March survey is low while age 8 is among the highest indices observed (Figure 10.1.9). The index of age 12 ( 2003 cohort) is much higher than seen before (large part of $11+$ ), but that cohort will though not contribute much to the landings. Year classes 2008 and 2009 (age 7 and 6) are now close to average, mostly due to reduced fishing mortality in recent years but those year classes were originally small.
The survey results indicate that in recent decade higher and higher proportion of the haddock stock has gradually been inhabiting the waters north of Iceland (Figures 10.1.7 and 10.1.8.).
Mean weight at age in the catch is shown in Table 10.1.6 and Figure 10.1.10. Mean weight at age in the stock is given in Table 10.1.5 and Figure 10.1.9. Those data are obtained from the groundfish survey in March and are also used as mean weight at age in the spawning stock.

Both stock and catch weights have been increasing in recent years, after being very low when the stock was large between 2005-2009. Higher mean weight at age is most apparent for the younger haddock from the small cohorts (2008 and later), but mean weight of the old fish is now near average.

Mean weight of the 2007 year class that will be the most important year class in the fisheries in next years ( $\approx 25 \%$ by weight 2015 and $15 \%$ in 2016) is close average since 1985. Mean weight of the youngest, small year classes is now above the average since 1985. Mean weight at age in the March survey 2015 was on the average higher than predicted last year, but growth in 2014 was approximately $10 \%$ better than in 2013 (Figures 10.2.8, and 10.4.2). In recent year's growth has shown interannual variability, indicating that short term prediction should rather been based on average growth of last two years instead of only last year's growth.
Maturity at age data are given in Table 10.1.7 and Figure 10.1.11. Those data are obtained from the groundfish survey in March. Maturity by age of the youngest age groups has been decreasing in recent years while mean weight at age has been increasing so maturity by size has been decreasing. The most likely explanation is high proportion of those age groups north of Iceland where proportion mature has always been low.

Catch per unit effort data (figure 10.1.12) give somewhat different picture of the development of the stock from the surveys and assessment, much less increase after 2000 but much less decrease in recent years (figures 10.1.8 vs 10.1.15). The interesting thing for the current assessment is the relatively high CPUE, in recent years, confirming fishermen's view that catching haddock in now very easy. The discrepancy observed between CPUE and stock size has not been explained, but a number of plausible reasons mentioned.

- Area inhabited by the stock increased so the density in the traditional fishing area did not increase in relation to the stock size.
- When the stock was large slower growth lead to higher proportion of the stock below "fishable size" 45 cm limiting the areas where large haddock could be caught without too much bycatch of small haddock.
- The opposite is happening in recent years, faster growth and poor recruitment lead to the fisheries not limited by small haddock.
- Bycatch issues, but haddock is often caught as bycatch or one of the species in mixed fisheries where the goal is certain mixture of species.


### 10.2 Assessment.

From 2007 to 2014 the final assessment was based on an Adapt type model calibrated with indices from both the groundfish surveys in March and October. Before that statistical catch at age model calibrated with indices from the March survey was used.

Assessment in recent years has shown some difference between different models, but more difference between different data sources i.e the March and the October surveys. From 2004 to 2008 models calibrated with the October survey indicated smaller stock. In the last five years things have changed and models calibrated with the October survey indicate a better state of the stock, the difference increased with addition of the most recent data points i.e. October 2014 and March 2015. This behaviour is in line with what is seen in the surveys where the contrast in biomass is higher in the March survey (Figure 10.1.8).

The stock was benchmarked in February 2013, (WKROUND 2013) and the assessment procedure used since 2007 was recommended for few more years, if major problems do not show up (see stock annex). This year the method described by WKROUND was changed by basing prediction of growth on the average of last 2 years instead of only the last year. The effect on TAC in next fishing year is reducing it by 2000 tonnes from 38.400 tons to 36.400 tons.

The results of the assessment indicate that the stock decreased from 2008-2011 when large year classes disappeared from the stock and were replaced by smaller yea classes. (Figure 10.2.1) Since 2011 the rate of reduction has slowed down as fishing mortality has been low. The spawning stock has though decreased more than the reference biomass as proportion mature by age/size has been decreasing. Fishing mortality is now estimated to be low and should continue to be so if the adopted HCR will be followed and the stock size not overestimated. Still the stock is predicted to decrease somewhat in next two years before the 2014 year class starts having effects.

The main features of the current assessment are the same as in the assessments 2011 to 2014. The current indicates larger stock than the 2014 assessment (Figures 10.2.7 and 10.2.8). Most of the difference is explained by higher than predicted mean weight at age (Figure 10.2.8). The tendency has been to underestimate recruitment and stock size in recent years.

Residuals from the assessment model are positive for the most recent October survey but close to zero for the most recent March survey. (Figures 10.2.2 and 10.2.3). The March surveys 2011-2014 are on the other hand below predictions. Similar thing seem to be happening in the fishery in 2012-2013 (Fig. 10.1.15) so there are indication that the stock might be underestimated or availability of haddock is unusually high.

Analysis by the TSA model (WD 38 2014) indicate increased natural mortality after age 6. This increase will not have large effect on the harvest rate in the HCR which is mostly selected based on precautionary criteria, i.e low probability of $\mathrm{SSB}<\mathrm{B}_{\text {loss }}$ in periods of poor recruitment. Higher M of older age groups will reduce the possibility of storing old spawners through periods of poor recruitment. To get maximum yield per recruit increased $M$ would call for higher harvest ratio. Increased $M$ of old fish will be investigated further in coming years, but reduced availability with age leads to similar observations.

Standard errors in estimates of SSB in 2015 from the Adapt model are 9 thous. tons for the March survey and 16 thous. for the autumn survey. The difference between the stock biomass is 65 thous. tonnes ( 125 vs 60 thous. tonnes) that does not fit within the confidence intervals (less than $1 \%$ probability of 65 thous tonnes or more difference between autumn survey and March survey results). This is an indication that the estimated confidence intervals are too narrow. The same observation was made last 3 years. The spawning stock according to the model tuned with both the surveys is 79 thous. tonnes.

Plot of observed vs. predicted biomass from the surveys (figure 10.2.3) indicates that historically the autumn survey biomass has been closer to prediction than corresponding values from the March survey where the contrast in observed biomass is more than predicted from the assessment. When the stock was small in 2000 and 2001, the March survey indicated considerably smaller stock while the autumn survey values were reasonably correct and from 2003-2007 the March survey overestimated the stock.

Figure 10.2 .5 shows the estimated "catchability" and CV as a function of age for the surveys, showing that estimated CV is lower in the autumn survey for ages 2 to 6 . Therefore, the autumn survey gets more weight for those age groups. The figure also indicates that estimated CV and "catchability" have not changed much for the March survey since 2008, but catchability of the autumn survey increased as has CV of the oldest age groups. This observation does partly have to do with the length of the series in.

To summarize there are indications from the autumn survey that the stock might be larger than predicted but from the March survey that it is smaller. CPUE data, not used directly in the assessment support that the stock might be larger.

### 10.3 Reference points

In March 2013, ICES evaluated a proposed Harvest Control Rule for Icelandic haddock (Björnsson 2013) and the Icelandic government adopted it in April 2013. The Harvest control rule is

```
The annual total allowable catch (TAC) will be set by applying the following harvest control rule (HCR):
1. When spawing stock biomass in the year following the assessment year (SSB}\mp@subsup{y}{y+1}{})\mathrm{ is equal to or greater than
SSB trigger:
    TACCyy+l}=\alpha \mp@subsup{B}{45+,y+l}{
2. When SSB}\mp@subsup{}{y+1}{}\mathrm{ is below SSB trigger:
    TAC }\mp@subsup{C}{yy+l}{}=\alpha SSB\mp@subsup{B}{y+1}{}/SS\mp@subsup{B}{\mathrm{ trigger }}{}\quad\mp@subsup{B}{45+,y+l}{
Where:
y the assessment year,
y/y+1 the fishing year starting 1 September in year y and ending 31 August in year y+1
y-1/y the fishing year starting 1 September in year y-1 and ending 31 August in year y
B45+,y+1 the reference biomass of 45cm and larger haddock in the year following the assessment year
and were }\alpha=0.40\mathrm{ and SSB 
```

$B_{45+}$, is on the average close to the spawning stock, but is not affected by changes in proportion mature by size/age. Large variability in size at age (Figure 10.1.12) is the reason for basing reference biomass on size rather than age. Proportion of a cohort above $45 \mathrm{~cm}\left(B_{45+}\right)$ is calculated from stock weights by the green curve in Figure 10.4.3.

Blim for Icelandic haddock was defined by ICES in 2011 as 45000 tonnes or Bloss. From $H_{m s y}$ the harvest ratio giving
maximum yield was estimated as 0.52 and $H_{P A}$ harvest ratio giving 5\% probability of $\mathrm{SSB}<\boldsymbol{B}_{\mathrm{lim}}$ as 0.46 , compared to the target harvest rate of 0.4 . These numbers do though not have any meaning when the HCR has been adopted.

### 10.4 Short term forecast

Prediction of weight at age in the stock, weight at age in the catches, maturity at age and selection has been similar since 2006 (WD \#19 in 2006). The procedure is described in the advice part of the report of ADGISHA (Björnsson 2013) and also in the stock annex.

To summarize, TAC for the fishing year 2015/2016 is a function of the biomass of 45 cm and larger haddock and the spawning stock in the beginning of 2016. To be able to predict the stock size in 2016, catch 2015, mean weight at age in the catch 2015 , selection at age in the catch 2015, stock weights in 2016 and maturity at age in 2016 must be predicted. The prediction of these values is described in Björnsson (2013) and the stock annex, but to summarize, catch in the assessment year (2015) is the TAC left in the current fishing year in the beginning of the assessment year plus $1 / 3$ of the predicted TAC next fishing year. The TAC for the fishing year 2014/2015 was 30400 tonnes. The landings in September - December 2014 were 11400 tonnes or $37 \%$ of the TAC. The average contribution of the first 4 month of of the Fishing year is on the other hand around $33 \%$. Landings for the fishing year 2013/2014 are now estimated to be 39600 tonnes while the TAC issued was 38 thous. tonnes. Looking at the rate of landings (Figure 10.4.1) they indicate that the TAC for the current fishing year will not be exceeded. .

In the Icelandic fishery management system certain relatively small transfer is allowed between species, to increase flexibility in mixed fisheries. Currently net transfer is towards haddock, probably because haddock is easy to catch, as demonstrated by high CPUE in 2014. The haddock quota does also seem to be limiting in some mixed fisheries. Looking over longer period quota transfer towards/from haddock has on the average been close to zero. In predictions for current fishing year 1000 transfer towards haddock is assumed.

On January $1^{\text {st }} 2015$, 19 thous. tonnes of quota were left To this are added $1 / 3^{\text {rd }}$ of next years TAC (12 100) and 1000 ton transfer. This leads to 32100 tonnes catch in the calendar year 2015 .

In current fishing year $45 \%$ of the quota is caught in September-December leading to 4 500 tonnes extra catch in 2013 compared to if $1 / 3^{\text {rd }}$ of the quota was caught in that period. It can be argued that when in the fishing year the TAC is caught is not crucial for development of the stock as long as the total catch is according to the TAC. Therefore the predictions are based on catching $1 / 3^{\text {rd }}$ of the TAC in September - December.

Mean weight and maturity at age in 2015 are available and are used to predict catch weights and selection at age (Figure 10.4.2). Growth in 2015 is predicted by the equation

$$
\log \frac{W_{a+1, t+1}}{W_{a, t}}=\alpha+\beta \log W_{a, t}+\delta_{y e a r}
$$

Where according to the stockannex the factor $\delta_{\text {year }}$ for the year 2014 (figure 10.4.2) is used for 2015 and onwards. Looking at the development of the annual growth in recent
years considerable interannual variability can be seen. Therefore, it is proposed to predict growth in the assessment year based on the average of the growth in the 2 preceeding years. This change will not change the average harvest ratio but will reduce variability in advice compared to the method described in the stock annex.

Maturity, selection, catch weights at age and proportion of the biomass above 45 cm are then predicted from stock weights 2016. When those values have been estimated the prediction is done by the same model as used in the assessment.

The model works iteratively as the estimated TAC for the fishing year 2015/2016 has some effect of the biomass in the beginning of 2016, which the TAC is based on. Advice for the following fishing year is predicted to be approximately 31000 tonnes but increasing after that when the 2014 year class comes in

Results of the short term prediction are shown in figure 10.2.1 assuming that the harvest control rule is followed. TAC for the fishing year 2015/2016 will be 36400 tons. Short term prognosis based on the traditional ICES approach are shown in table 10.4.1

### 10.5 References

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Table 10.1.1 Haddock in Division Va Landings by nation.

| Country | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Belgium | 1010 | 1144 | 673 | 377 | 268 | 359 | 391 | 257 |
| Faroe Islands | 2161 | 2029 | 1839 | 1982 | 1783 | 707 | 987 | 1289 |
| Iceland | 52152 | 47916 | 61033 | 67038 | 63889 | 47216 | 49553 | 47317 |
| Norway | 11 | 23 | 15 | 28 | 3 | 3 | + |  |
| $€ U K$ |  |  |  |  |  |  |  |  |
| Total | 55334 | 51112 | 63560 | 69425 | 65943 | 48285 | 50933 | 48863 |

HADDOCK Va

| Country | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Belgium | 238 | 352 | 483 | 595 | 485 | 361 | 458 | 248 |
| Faroe Islands | 1043 | 797 | 606 | 603 | 773 | 757 | 754 | 911 |
| Iceland | 39479 | 53085 | 61792 | 66004 | 53516 | 46098 | 46932 | 58408 |
| Norway | 1 | + |  |  |  |  |  | 1 |
| Total | 40761 | 54234 | 62881 | 67202 | 53774 | 47216 | 48144 | 59567 |

HADDOCK Va

| Country | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Belgium |  |  |  |  |  |  |  |  |
| Faroe Islands | 758 | 664 | 340 | 639 | 624 | 968 | 609 | 878 |
| Iceland | 60061 | 56223 | 43245 | 40795 | 44557 | 41199 | 39038 | 49591 |
| Norway | + | 4 |  |  |  |  |  |  |
| Total | 60819 | 56891 | 43585 | 41434 | 45481 | 42167 | 39647 | 50469 |


| Country | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Belgium |  |  |  |  |  |  |  |  |  |
| Faroe <br> Islands | 833 | 1035 | 1372 | 1499 | 1780 | 828 | 625 | 311 | 207 |
| Iceland | 59970 | 83791 | 95859 | 96115 | 108175 | 101651 | 81418 | 63868 | 49231 |
| Norway | 30 | 9 |  |  | 11 | 11 |  |  |  |
| Total | 60884 | 84835 | 97231 | 97614 | 109966 | 102490 | 82043 | 64179 | 49437 |


| Country | 2012 | 2013 | 2014 |
| :--- | :--- | :--- | :--- |


| Belgium |  |  |  |
| :--- | :--- | :--- | :--- |
| Faroe <br> Islands | 303 | 600 | 800 |
| Iceland | 45888 | 43500 | 33100 |
| Norway |  |  |  |
| Total | 46191 | 44100 | 33900 |

Table 10.1.2 Haddock in division Va. Catch in number by year and age.

| Year/ <br> Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 149 | 1908 | 3762 | 6057 | 9022 | 1743 | 438 | 56 | 112 |
| 1980 | 595 | 1385 | 11481 | 4298 | 3798 | 3732 | 544 | 91 | 37 |
| 1981 | 10 | 514 | 4911 | 16900 | 5999 | 2825 | 1803 | 168 | 57 |
| 1982 | 107 | 245 | 3149 | 10851 | 14049 | 2068 | 1000 | 725 | 201 |
| 1983 | 34 | 1010 | 1589 | 4596 | 9850 | 8839 | 766 | 207 | 280 |
| 1984 | 241 | 1069 | 4946 | 1341 | 4772 | 3742 | 4076 | 238 | 80 |
| 1985 | 1320 | 1728 | 4562 | 6796 | 855 | 1682 | 1914 | 1903 | 296 |
| 1986 | 1012 | 4223 | 4068 | 4686 | 5139 | 494 | 796 | 897 | 400 |
| 1987 | 1939 | 8308 | 6965 | 2728 | 2042 | 1094 | 132 | 165 | 339 |
| 1988 | 237 | 9831 | 15164 | 5824 | 1304 | 1084 | 609 | 66 | 213 |
| 1989 | 188 | 2474 | 22560 | 9571 | 3196 | 513 | 556 | 144 | 141 |
| 1990 | 1857 | 2415 | 8628 | 23611 | 6331 | 816 | 150 | 67 | 74 |
| 1991 | 8617 | 2145 | 5397 | 7342 | 14103 | 2648 | 338 | 40 | 27 |
| 1992 | 5405 | 10693 | 5721 | 4610 | 3691 | 5209 | 999 | 120 | 16 |
| 1993 | 769 | 12333 | 12815 | 2968 | 1722 | 1425 | 2239 | 343 | 38 |
| 1994 | 3198 | 3343 | 28258 | 10682 | 1469 | 726 | 358 | 647 | 108 |
| 1995 | 4015 | 7323 | 5744 | 23927 | 5769 | 615 | 290 | 187 | 331 |
| 1996 | 3090 | 10552 | 7639 | 4468 | 12896 | 2346 | 208 | 79 | 125 |
| 1997 | 1364 | 3939 | 10915 | 4895 | 2610 | 5035 | 719 | 64 | 69 |
| 1998 | 279 | 8257 | 5667 | 7856 | 2418 | 1422 | 1897 | 261 | 45 |
| 1999 | 1434 | 1550 | 17243 | 4516 | 4837 | 915 | 620 | 481 | 64 |
| 2000 | 2659 | 6317 | 2352 | 13615 | 1945 | 1706 | 324 | 222 | 192 |
| 2001 | 2515 | 11098 | 6954 | 1446 | 6262 | 675 | 478 | 105 | 94 |
| 2002 | 1082 | 10434 | 15998 | 5099 | 1131 | 3149 | 262 | 169 | 100 |
| 2003 | 401 | 6352 | 16265 | 12548 | 2968 | 748 | 1236 | 91 | 70 |
| 2004 | 1597 | 4063 | 17652 | 19358 | 8871 | 1940 | 471 | 489 | 155 |
| 2005 | 2405 | 9450 | 6929 | 25421 | 13778 | 4584 | 809 | 251 | 237 |
| 2006 | 241 | 10038 | 21246 | 6646 | 18840 | 7600 | 2180 | 323 | 202 |
| 2007 | 782 | 3884 | 42224 | 22239 | 3354 | 9952 | 2740 | 519 | 181 |
| 2008 | 2316 | 4508 | 9706 | 53022 | 11014 | 1717 | 3033 | 815 | 192 |
| 2009 | 1066 | 3185 | 4886 | 8892 | 35011 | 5733 | 726 | 1381 | 509 |
| 2010 | 121 | 6032 | 7061 | 4806 | 6766 | 17503 | 1874 | 354 | 528 |
| 2011 | 253 | 1584 | 11797 | 5080 | 2853 | 3983 | 6220 | 494 | 183 |
| 2012 | 196 | 1322 | 3421 | 13107 | 2223 | 1231 | 2480 | 2662 | 370 |
| 2013 | 250 | 1042 | 2865 | 4008 | 9222 | 1206 | 668 | 1248 | 1599 |
| 2014 | 238 | 1478 | 1751 | 2725 | 2737 | 4742 | 447 | 387 | 1403 |

Table 10.1.3 Icelandic haddock. Age disaggregated survey indices from the groundfish survey in March .

| YEAR/ |  |  |  |  |  |  |  | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| AGE | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{9}$ | 10 | $11+$ |  |  |  |  |
| 1985 | 28.14 | 32.68 | 18.33 | 23.58 | 26.39 | 3.7 | 10.86 | 4.8 | 5.54 | 0.49 | 0.19 |
| 1986 | 123.87 | 108.48 | 58.97 | 12.79 | 16.31 | 13.12 | 0.97 | 2.71 | 1.22 | 2.25 | 0.19 |
| 1987 | 21.82 | 338.29 | 147.5 | 44.15 | 7.68 | 7.47 | 4.72 | 0.39 | 0.61 | 0.44 | 0.86 |
| 1988 | 15.77 | 40.73 | 184.79 | 88.87 | 22.86 | 1.34 | 2.18 | 1.76 | 0.16 | 0.22 | 0.31 |
| 1989 | 10.58 | 23.33 | 41.16 | 146.61 | 45.09 | 12.88 | 0.79 | 0.81 | 0.41 | 0.28 | 0.23 |
| 1990 | 70.48 | 31.8 | 26.73 | 38.84 | 92.82 | 30.89 | 3.44 | 0.88 | 0.23 | 0 | 0.02 |
| 1991 | 89.73 | 145.95 | 41.43 | 17.73 | 20.19 | 32.85 | 7.63 | 0.3 | 0.1 | 0.08 | 0.08 |
| 1992 | 18.15 | 211.43 | 137.77 | 35.38 | 16.91 | 13.77 | 16.32 | 2.22 | 0.18 | 0.07 | 0 |
| 1993 | 29.99 | 37.8 | 244.96 | 87.19 | 11.23 | 3.85 | 1.66 | 4.46 | 0.88 | 0 | 0 |
| 1994 | 58.54 | 61.34 | 39.83 | 142.35 | 42.18 | 6.9 | 2.87 | 1.42 | 4.44 | 0.17 | 0 |
| 1995 | 35.89 | 82.47 | 47.03 | 19.75 | 69.52 | 7.66 | 1.31 | 0.11 | 0.34 | 0 | 0 |
| 1996 | 95.25 | 66.21 | 119.86 | 36.78 | 19.58 | 40.63 | 5.78 | 0.59 | 0.13 | 0.12 | 0.15 |
| 1997 | 8.6 | 119.35 | 50.81 | 53.33 | 10.88 | 7.37 | 10.9 | 1.35 | 0.07 | 0.03 | 0.13 |
| 1998 | 23.08 | 18 | 107.93 | 28.23 | 23.49 | 4.9 | 3.54 | 4.56 | 0.33 | 0 | 0 |
| 1999 | 80.73 | 85.46 | 25.53 | 98.73 | 12.99 | 9.85 | 1.42 | 1.77 | 1.03 | 0.09 | 0 |
| 2000 | 60.58 | 90.07 | 44.63 | 8.45 | 25.22 | 3.14 | 1.59 | 0.4 | 0.15 | 0.52 | 0.04 |
| 2001 | 81.27 | 147.71 | 115.4 | 22.15 | 4.09 | 10.63 | 0.93 | 0.57 | 0 | 0.1 | 0 |
| 2002 | 20.75 | 298.67 | 200.74 | 112.49 | 23.24 | 3.51 | 7.49 | 0.31 | 0.3 | 0.08 | 0.15 |
| 2003 | 111.59 | 97.54 | 282.28 | 244.81 | 113.45 | 18 | 2.55 | 4.48 | 0.48 | 0.82 | 0.15 |
| 2004 | 325.9 | 291.65 | 70.75 | 208.74 | 109.33 | 33.96 | 6.79 | 1.24 | 0.82 | 0 | 0.31 |
| 2005 | 57.96 | 698.48 | 289.43 | 44.58 | 157.2 | 57.52 | 15.72 | 3.35 | 0.32 | 0.25 | 0.02 |
| 2006 | 39.29 | 88.69 | 575.93 | 179.11 | 19.13 | 62.94 | 16.43 | 6.74 | 0.7 | 0.29 | 0 |
| 2007 | 34 | 65.6 | 88.63 | 436.41 | 85.68 | 7.9 | 21.6 | 4.74 | 2.15 | 0.07 | 0 |
| 2008 | 88.53 | 68.05 | 71.7 | 75.57 | 222.79 | 29.99 | 3.53 | 7.47 | 1.64 | 0.27 | 0.03 |
| 2009 | 10.46 | 111.21 | 53.82 | 41.48 | 41.91 | 105.64 | 12.94 | 2.23 | 3.11 | 0.44 | 0.23 |
| 2010 | 15.15 | 27.71 | 138.2 | 29.95 | 18.28 | 20.59 | 31.59 | 2.92 | 0.46 | 0.69 | 0.2 |
| 2011 | 8.79 | 27.65 | 24.75 | 77.43 | 14.03 | 5.9 | 9.4 | 14.89 | 1.22 | 0.31 | 0.3 |
| 2012 | 12.47 | 14.9 | 31.27 | 27.22 | 58.3 | 5.23 | 2.92 | 5.3 | 6.87 | 0.8 | 0.49 |
| 2013 | 13.91 | 23.32 | 19.72 | 22.9 | 22.51 | 41.93 | 4.78 | 2.52 | 3.83 | 4.52 | 1.02 |
| 2014 | 14.01 | 24.78 | 30.27 | 17.74 | 16.44 | 14.79 | 16.44 | 1.33 | 1.05 | 1.68 | 1.63 |
| 2015 | 62.58 | 19.59 | 26.56 | 34.23 | 12.58 | 11.18 | 9.63 | 9.96 | 1.14 | 0.56 | 2.29 |
|  |  |  |  |  |  |  |  |  |  |  |  |

Table 10.1.4 Icelandic haddock. Age disaggregated survey indices from the groundfish survey in October

| YEAR $/$ |  |  |  |  |  |  | $\mathbf{2}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| AGE | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ |  |  |  |  |  |  |
| 1996 | 16.1 | 461.3 | 109.4 | 85.6 | 18.5 | 7.8 | 18.3 | 1.6 | 0 | 0 | 0 |
| 1997 | 52.9 | 32.4 | 212.9 | 54.5 | 38.7 | 7 | 5.7 | 6.1 | 0.3 | 0 | 0 |
| 1998 | 209.1 | 81.1 | 32.5 | 133.4 | 19.8 | 15.7 | 5.3 | 5.4 | 1.9 | 0 | 0.1 |
| 1999 | 178.6 | 397.4 | 66.9 | 28.6 | 97.1 | 11.9 | 10.4 | 0.5 | 2.1 | 0.3 | 0 |
| 2000 | 56.2 | 161.9 | 260.1 | 46.3 | 8.2 | 28.7 | 2 | 3.2 | 0.1 | 0.3 | 0.6 |
| 2001 | 47 | 387.5 | 281.6 | 170.2 | 35.7 | 4.1 | 13.9 | 0.7 | 1 | 0 | 0.2 |
| 2002 | 150.6 | 85.2 | 237.8 | 197.5 | 98.5 | 19.3 | 3 | 2.3 | 1 | 0.1 | 0 |
| 2003 | 316.5 | 345.5 | 146.9 | 251.9 | 169.1 | 56.6 | 9.5 | 2.4 | 0.7 | 0 | 0 |
| 2004 | 189.4 | 714.2 | 347.3 | 51.2 | 160.3 | 70.6 | 17 | 4 | 0.8 | 0.5 | 0 |
| 2005 | 91.1 | 74.2 | 560.4 | 182.1 | 27.3 | 96.5 | 26.7 | 10.4 | 1.9 | 0 | 0.1 |
| 2006 | 85.9 | 124.1 | 117.6 | 510.4 | 108.5 | 13.8 | 40.4 | 9.8 | 3.9 | 1.5 | 0 |
| 2007 | 203.4 | 93 | 78.4 | 92.8 | 341.4 | 58.6 | 8.5 | 12.3 | 3.8 | 0.6 | 0.3 |
| 2008 | 95.3 | 201.8 | 93.9 | 68.4 | 87.9 | 198.9 | 16.8 | 2.9 | 3.5 | 0.2 | 0.1 |
| 2009 | 52.8 | 47.5 | 269.5 | 68.1 | 31 | 48.5 | 96.5 | 9.5 | 1.5 | 2.2 | 0.4 |
| 2010 | 37.2 | 43.3 | 56.6 | 143.4 | 30.5 | 14.4 | 23.7 | 37.2 | 4.8 | 0.9 | 1.1 |
| 2012 | 26.8 | 53.8 | 29.1 | 34.3 | 37.7 | 70.3 | 9.3 | 3.6 | 9.8 | 10.3 | 1.7 |
| 2013 | 27.1 | 91.9 | 131.4 | 37.3 | 38.6 | 39.3 | 44.8 | 6.2 | 2.3 | 5.8 | 4.9 |
| 2014 | 250.2 | 35.1 | 41.3 | 67.3 | 24.1 | 27.2 | 24.4 | 26.3 | 2.3 | 1.5 | 5 |

Table 10.1.5 Haddock in division Va Weight at age in the stock. Predicted values are shaded

| Year/age | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | 10 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1979 | 37 | 185 | 481 | 910 | 1409 | 1968 | 2496 | 3077 | 3300 | 4000 |
| 1980 | 37 | 185 | 481 | 910 | 1409 | 1968 | 2496 | 3077 | 3300 | 4615 |
| 1981 | 37 | 185 | 481 | 910 | 1409 | 1968 | 2496 | 3077 | 3300 | 4898 |
| 1982 | 37 | 185 | 481 | 910 | 1409 | 1968 | 2496 | 3077 | 3300 | 3952 |
| 1983 | 37 | 185 | 481 | 910 | 1409 | 1968 | 2496 | 3077 | 3300 | 4463 |
| 1984 | 37 | 185 | 481 | 910 | 1409 | 1968 | 2496 | 3077 | 3300 | 3941 |
| 1985 | 36 | 244 | 568 | 1187 | 1673 | 2371 | 2766 | 3197 | 3331 | 4564 |
| 1986 | 35 | 239 | 671 | 1134 | 1943 | 2399 | 3190 | 3293 | 3728 | 4436 |
| 1987 | 31 | 162 | 550 | 1216 | 1825 | 2605 | 3030 | 3642 | 3837 | 3653 |
| 1988 | 37 | 176 | 457 | 974 | 1830 | 2695 | 3102 | 3481 | 3318 | 4169 |
| 1989 | 26 | 182 | 441 | 887 | 1510 | 2380 | 3009 | 3499 | 3195 | 5039 |
| 1990 | 29 | 184 | 457 | 840 | 1234 | 1965 | 2675 | 3052 | 3267 | 4115 |
| 1991 | 31 | 176 | 501 | 1003 | 1406 | 1884 | 2496 | 3755 | 3653 | 5243 |
| 1992 | 28 | 157 | 503 | 894 | 1365 | 1891 | 2325 | 2936 | 3682 | 4674 |
| 1993 | 41 | 168 | 384 | 878 | 1492 | 1785 | 2562 | 2573 | 3266 | 4047 |
| 1994 | 33 | 181 | 392 | 680 | 1235 | 1766 | 1717 | 2977 | 2131 | 3154 |
| 1995 | 37 | 167 | 440 | 755 | 1065 | 1857 | 2689 | 5377 | 1306 | 3119 |
| 1996 | 41 | 174 | 453 | 813 | 1076 | 1477 | 2171 | 2426 | 4847 | 3686 |
| 1997 | 50 | 174 | 424 | 817 | 1221 | 1425 | 1915 | 2390 | 3692 | 3508 |


| 1998 | 41 | 203 | 415 | 753 | 1241 | 1747 | 1996 | 2342 | 3076 | 3275 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1999 | 33 | 206 | 480 | 715 | 1189 | 1956 | 2366 | 2782 | 2922 | 3534 |
| 2000 | 29 | 179 | 552 | 889 | 1159 | 1767 | 2612 | 2917 | 3132 | 3734 |
| 2001 | 36 | 190 | 490 | 1056 | 1437 | 1509 | 2169 | 2765 | 3300 | 4715 |
| 2002 | 67 | 172 | 475 | 889 | 1460 | 1949 | 2137 | 1990 | 3709 | 4078 |
| 2003 | 40 | 230 | 412 | 801 | 1268 | 1873 | 3139 | 2343 | 3301 | 3289 |
| 2004 | 34 | 176 | 556 | 807 | 1282 | 1690 | 2454 | 3236 | 2942 | 3957 |
| 2005 | 40 | 153 | 448 | 920 | 1188 | 1564 | 2128 | 2808 | 2550 | 2755 |
| 2006 | 33 | 127 | 333 | 736 | 1145 | 1512 | 1944 | 2232 | 3272 | 3617 |
| 2007 | 48 | 170 | 350 | 615 | 1053 | 1514 | 1786 | 2073 | 2198 | 2408 |
| 2008 | 27 | 179 | 382 | 595 | 868 | 1295 | 1828 | 2201 | 2340 | 2568 |
| 2009 | 29 | 139 | 442 | 687 | 882 | 1141 | 1495 | 1920 | 2574 | 3070 |
| 2010 | 32 | 150 | 392 | 773 | 942 | 1190 | 1468 | 1829 | 2086 | 2730 |
| 2011 | 35 | 175 | 442 | 757 | 1129 | 1304 | 1583 | 1865 | 2107 | 3094 |
| 2012 | 28 | 202 | 482 | 801 | 1145 | 1480 | 1909 | 2072 | 2353 | 2350 |
| 2013 | 33 | 201 | 589 | 967 | 1312 | 1710 | 1999 | 2265 | 2764 | 2709 |
| 2014 | 36 | 222 | 570 | 1005 | 1372 | 1751 | 2141 | 2298 | 2653 | 3104 |
| 2015 | 32 | 255 | 614 | 1073 | 1637 | 1926 | 2452 | 2774 | 3170 | 3173 |
| 2016 | 32 | 184 | 609 | 1117 | 1643 | 2200 | 2461 | 2908 | 3167 | 3472 |
| 2017 | 32 | 182 | 486 | 1111 | 1689 | 2205 | 2698 | 2915 | 3272 | 3470 |

Table 10.1.6 Haddock in division Va Weight at age in the catches. Predicted values are shaded.

| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 620 | 960 | 1410 | 2030 | 2910 | 3800 | 4560 | 4720 | 4000 |
| 1980 | 837 | 831 | 1306 | 2207 | 2738 | 3188 | 3843 | 4506 | 4615 |
| 1981 | 584 | 693 | 1081 | 1656 | 2283 | 3214 | 3409 | 4046 | 4898 |
| 1982 | 289 | 959 | 1455 | 1674 | 2351 | 3031 | 3481 | 3874 | 3952 |
| 1983 | 320 | 1006 | 1496 | 1921 | 2371 | 2873 | 3678 | 4265 | 4463 |
| 1984 | 691 | 1007 | 1544 | 2120 | 2514 | 3027 | 2940 | 3906 | 3941 |
| 1985 | 652 | 1125 | 1811 | 2260 | 2924 | 3547 | 3733 | 4039 | 4564 |
| 1986 | 336 | 1227 | 1780 | 2431 | 2771 | 3689 | 3820 | 4258 | 4436 |
| 1987 | 452 | 1064 | 1692 | 2408 | 3000 | 3565 | 4215 | 4502 | 3653 |
| 1988 | 362 | 780 | 1474 | 2217 | 2931 | 3529 | 3781 | 4467 | 4169 |
| 1989 | 323 | 857 | 1185 | 1996 | 2893 | 4066 | 3866 | 4734 | 5039 |
| 1990 | 269 | 700 | 1054 | 1562 | 2364 | 3414 | 4134 | 4946 | 4115 |
| 1991 | 288 | 699 | 979 | 1412 | 1887 | 2674 | 3135 | 4341 | 5243 |
| 1992 | 313 | 806 | 1167 | 1524 | 1950 | 2357 | 3075 | 4053 | 4674 |
| 1993 | 303 | 705 | 1333 | 1875 | 2386 | 2996 | 3059 | 3363 | 4047 |
| 1994 | 337 | 668 | 1019 | 1717 | 2391 | 2717 | 3280 | 3156 | 3154 |
| 1995 | 351 | 746 | 1096 | 1318 | 2044 | 2893 | 3049 | 3675 | 3119 |
| 1996 | 311 | 787 | 1187 | 1560 | 1849 | 2670 | 3510 | 3567 | 3686 |
| 1997 | 379 | 764 | 1163 | 1649 | 1943 | 2342 | 3020 | 3337 | 3508 |
| 1998 | 445 | 724 | 1147 | 1683 | 2250 | 2475 | 2834 | 3333 | 3275 |
| 1999 | 555 | 908 | 1101 | 1658 | 2216 | 2659 | 2928 | 3209 | 3534 |
| 2000 | 495 | 978 | 1333 | 1481 | 2119 | 2696 | 3307 | 3597 | 3734 |
| 2001 | 541 | 945 | 1456 | 1731 | 1832 | 2243 | 3020 | 3328 | 4715 |
| 2002 | 564 | 928 | 1253 | 1737 | 2219 | 2230 | 2911 | 3365 | 4078 |
| 2003 | 498 | 922 | 1283 | 1704 | 2274 | 2744 | 2635 | 2819 | 3289 |
| 2004 | 559 | 1006 | 1258 | 1579 | 2044 | 2809 | 3123 | 2945 | 3957 |
| 2005 | 339 | 886 | 1265 | 1506 | 1916 | 2323 | 3028 | 3211 | 2755 |
| 2006 | 402 | 749 | 1093 | 1495 | 1758 | 2163 | 2555 | 3054 | 3617 |
| 2007 | 510 | 748 | 988 | 1346 | 1840 | 2062 | 2350 | 2525 | 2408 |
| 2008 | 383 | 636 | 857 | 1125 | 1575 | 2149 | 2417 | 2802 | 2568 |
| 2009 | 452 | 841 | 960 | 1131 | 1352 | 1757 | 2364 | 2497 | 3070 |
| 2010 | 447 | 756 | 1092 | 1294 | 1448 | 1685 | 2188 | 2366 | 2657 |
| 2011 | 588 | 905 | 1122 | 1455 | 1688 | 1914 | 2094 | 2455 | 2919 |
| 2012 | 668 | 978 | 1222 | 1492 | 1903 | 2164 | 2366 | 2704 | 2765 |
| 2013 | 678 | 1084 | 1358 | 1675 | 2036 | 2400 | 2554 | 3097 | 3111 |
| 2014 | 536 | 1080 | 1433 | 1793 | 2121 | 2504 | 2624 | 3178 | 3272 |
| 2015 | 559 | 1018 | 1491 | 1989 | 2223 | 2621 | 2852 | 3124 | 3126 |
| 2016 | 447 | 1013 | 1533 | 1994 | 2434 | 2628 | 2945 | 3122 | 3324 |

Table 10.1.7 Haddock in division Va Sexual maturity at age in the stock. (from the March survey). Predicted values are shaded. The numbers for age 10 only apply to the spawning stock.

| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 0.08 | 0.301 | 0.539 | 0.722 | 0.821 | 0.868 | 0.904 | 0.963 | 1 |
| 1980 | 0.08 | 0.301 | 0.539 | 0.722 | 0.821 | 0.868 | 0.904 | 0.963 | 1 |
| 1981 | 0.08 | 0.301 | 0.539 | 0.722 | 0.821 | 0.868 | 0.904 | 0.963 | 1 |
| 1982 | 0.08 | 0.301 | 0.539 | 0.722 | 0.821 | 0.868 | 0.904 | 0.963 | 1 |
| 1983 | 0.08 | 0.301 | 0.539 | 0.722 | 0.821 | 0.868 | 0.904 | 0.963 | 1 |
| 1984 | 0.08 | 0.301 | 0.539 | 0.722 | 0.821 | 0.868 | 0.904 | 0.963 | 1 |
| 1985 | 0.016 | 0.144 | 0.536 | 0.577 | 0.765 | 0.766 | 0.961 | 0.934 | 1 |
| 1986 | 0.021 | 0.205 | 0.413 | 0.673 | 0.845 | 0.884 | 0.952 | 0.986 | 1 |
| 1987 | 0.022 | 0.137 | 0.426 | 0.535 | 0.778 | 0.776 | 1 | 0.969 | 1 |
| 1988 | 0.013 | 0.221 | 0.394 | 0.767 | 0.793 | 0.928 | 0.914 | 1 | 1 |
| 1989 | 0.041 | 0.202 | 0.532 | 0.727 | 0.818 | 0.998 | 1 | 1 | 1 |
| 1990 | 0.114 | 0.334 | 0.634 | 0.814 | 0.843 | 0.918 | 0.882 | 1 | 1 |
| 1991 | 0.063 | 0.224 | 0.592 | 0.739 | 0.817 | 0.894 | 0.495 | 1 | 1 |
| 1992 | 0.05 | 0.227 | 0.419 | 0.799 | 0.901 | 0.901 | 0.858 | 1 | 1 |
| 1993 | 0.124 | 0.362 | 0.481 | 0.67 | 0.904 | 0.977 | 0.908 | 0.867 | 1 |
| 1994 | 0.248 | 0.312 | 0.573 | 0.762 | 0.846 | 1 | 0.907 | 1 | 1 |
| 1995 | 0.124 | 0.479 | 0.382 | 0.75 | 0.753 | 0.606 | 0.985 | 1 | 1 |
| 1996 | 0.191 | 0.362 | 0.59 | 0.648 | 0.787 | 0.739 | 0.949 | 0.908 | 1 |
| 1997 | 0.093 | 0.436 | 0.587 | 0.683 | 0.75 | 0.783 | 0.88 | 1 | 1 |
| 1998 | 0.026 | 0.454 | 0.668 | 0.77 | 0.733 | 0.849 | 0.899 | 1 | 1 |
| 1999 | 0.05 | 0.397 | 0.683 | 0.724 | 0.749 | 0.892 | 0.761 | 0.92 | 1 |
| 2000 | 0.107 | 0.261 | 0.632 | 0.808 | 0.868 | 0.873 | 1 | 0.78 | 1 |
| 2001 | 0.091 | 0.377 | 0.522 | 0.753 | 0.895 | 0.916 | 0.918 | 1 | 1 |
| 2002 | 0.047 | 0.286 | 0.633 | 0.8 | 0.934 | 0.928 | 1 | 1 | 1 |
| 2003 | 0.062 | 0.347 | 0.685 | 0.867 | 0.922 | 0.946 | 1 | 1 | 1 |
| 2004 | 0.037 | 0.361 | 0.57 | 0.831 | 0.91 | 1 | 1 | 1 | 1 |
| 2005 | 0.024 | 0.23 | 0.562 | 0.753 | 0.927 | 0.936 | 0.968 | 1 | 1 |
| 2006 | 0.027 | 0.117 | 0.462 | 0.621 | 0.739 | 0.918 | 1 | 1 | 1 |
| 2007 | 0.078 | 0.208 | 0.418 | 0.68 | 0.77 | 0.875 | 0.959 | 1 | 1 |
| 2008 | 0.027 | 0.263 | 0.418 | 0.621 | 0.828 | 0.87 | 0.904 | 0.975 | 1 |
| 2009 | 0.017 | 0.301 | 0.47 | 0.576 | 0.847 | 0.891 | 1 | 0.968 | 1 |
| 2010 | 0.029 | 0.187 | 0.618 | 0.778 | 0.787 | 0.887 | 0.934 | 1 | 0.958 |
| 2011 | 0.045 | 0.176 | 0.426 | 0.823 | 0.816 | 0.838 | 0.899 | 0.974 | 1 |
| 2012 | 0.106 | 0.167 | 0.445 | 0.627 | 0.819 | 0.903 | 0.852 | 0.911 | 1 |
| 2013 | 0.046 | 0.223 | 0.381 | 0.714 | 0.793 | 0.92 | 0.986 | 0.974 | 0.992 |
| 2014 | 0.107 | 0.192 | 0.391 | 0.567 | 0.675 | 0.735 | 0.925 | 0.906 | 0.883 |
| 2015 | 0.138 | 0.283 | 0.445 | 0.667 | 0.795 | 0.772 | 0.892 | 1 | 0.889 |
| 2016 | 0.067 | 0.417 | 0.696 | 0.827 | 0.894 | 0.912 | 0.935 | 1 | 1 |
| 2017 | 0.066 | 0.317 | 0.694 | 0.835 | 0.894 | 0.925 | 0.935 | 1 | 1 |

Table 10.2.1 Haddock in division Va. Summary table from the SPALY run using the surveys in March and October for tuning.

| Year | Recruitment thousand at age 2 | $\begin{aligned} & \text { Biomass } \\ & 3+\text { tons } \\ & \hline \end{aligned}$ | SSB tons | Landings tons | Yield/SSB | F4-7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 80923 | 162177 | 96072 | 55330 | 0.576 | 0.521 |
| 1980 | 37390 | 192244 | 116521 | 51110 | 0.439 | 0.398 |
| 1981 | 10426 | 206988 | 141628 | 63558 | 0.449 | 0.542 |
| 1982 | 42788 | 180380 | 136817 | 69428 | 0.507 | 0.444 |
| 1983 | 29306 | 148112 | 112589 | 65942 | 0.586 | 0.508 |
| 1984 | 20574 | 112797 | 82961 | 48282 | 0.582 | 0.515 |
| 1985 | 42788 | 102394 | 66652 | 51102 | 0.767 | 0.537 |
| 1986 | 86501 | 96480 | 59837 | 48859 | 0.817 | 0.739 |
| 1987 | 164036 | 105395 | 46298 | 40760 | 0.88 | 0.584 |
| 1988 | 48742 | 153708 | 69391 | 54204 | 0.781 | 0.675 |
| 1989 | 29778 | 168184 | 99537 | 62885 | 0.632 | 0.676 |
| 1990 | 27094 | 145507 | 110745 | 67198 | 0.607 | 0.611 |
| 1991 | 92280 | 122708 | 89825 | 54692 | 0.609 | 0.664 |
| 1992 | 175094 | 106310 | 66379 | 47121 | 0.71 | 0.728 |
| 1993 | 38437 | 130461 | 71000 | 48123 | 0.678 | 0.669 |
| 1994 | 46842 | 127836 | 83295 | 59502 | 0.714 | 0.641 |
| 1995 | 72857 | 124042 | 85054 | 60884 | 0.716 | 0.661 |
| 1996 | 36341 | 108036 | 70008 | 56890 | 0.813 | 0.675 |
| 1997 | 102509 | 87152 | 58993 | 43764 | 0.742 | 0.624 |
| 1998 | 17976 | 97121 | 64203 | 41192 | 0.642 | 0.627 |
| 1999 | 50160 | 91024 | 64439 | 45411 | 0.705 | 0.685 |
| 2000 | 117423 | 90674 | 63509 | 42105 | 0.663 | 0.636 |
| 2001 | 156535 | 115046 | 70366 | 39654 | 0.564 | 0.462 |
| 2002 | 187267 | 168427 | 99344 | 50498 | 0.508 | 0.461 |
| 2003 | 50154 | 219757 | 147519 | 60883 | 0.413 | 0.404 |
| 2004 | 151983 | 252717 | 181270 | 84828 | 0.468 | 0.491 |
| 2005 | 385734 | 259074 | 176986 | 97225 | 0.549 | 0.522 |
| 2006 | 90259 | 299329 | 143539 | 97614 | 0.68 | 0.578 |
| 2007 | 42302 | 297983 | 162862 | 109966 | 0.675 | 0.556 |
| 2008 | 44042 | 250161 | 158820 | 102872 | 0.648 | 0.476 |
| 2009 | 119817 | 193067 | 142937 | 82045 | 0.574 | 0.495 |
| 2010 | 38892 | 168462 | 114292 | 64168 | 0.561 | 0.471 |
| 2011 | 28567 | 152264 | 97645 | 49433 | 0.506 | 0.411 |
| 2012 | 19641 | 141018 | 93636 | 46208 | 0.493 | 0.345 |
| 2013 | 34400 | 131953 | 96665 | 44097 | 0.456 | 0.347 |
| 2014 | 23914 | 114966 | 72366 | 33900 | 0.468 | 0.307 |
| 2015 | 20253 | 112462 | 78319 |  |  |  |
| $\begin{aligned} & \text { Mean79- } \\ & 2014 \end{aligned}$ | 74703 | 155038 | 99792 | 58753 | 0.61 | 0.541 |

Table 10.2.2 Haddock in division Va. Number in stock from the SPALY run using both the surveys. Shaded cells are input to prediction. . Predictions shown are based on HCR.

| Year/Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 45.7 | 80.9 | 117.3 | 27.7 | 19.6 | 20.44 | 3.41 | 0.77 | 0.15 | 0.05 |
| 1980 | 12.7 | 37.4 | 66.1 | 94.3 | 19.3 | 10.54 | 8.57 | 1.21 | 0.23 | 0.07 |
| 1981 | 52.3 | 10.4 | 30.1 | 52.9 | 66.8 | 11.91 | 5.19 | 3.64 | 0.5 | 0.11 |
| 1982 | 35.8 | 42.8 | 8.5 | 24.2 | 38.9 | 39.42 | 4.33 | 1.69 | 1.35 | 0.26 |
| 1983 | 25.1 | 29.3 | 34.9 | 6.8 | 16.9 | 21.99 | 19.56 | 1.67 | 0.48 | 0.45 |
| 1984 | 52.3 | 20.6 | 24 | 27.7 | 4.1 | 9.7 | 9.09 | 8.02 | 0.68 | 0.21 |
| 1985 | 105.7 | 42.8 | 16.6 | 18.6 | 18.2 | 2.14 | 3.63 | 4.06 | 2.88 | 0.34 |
| 1986 | 200.3 | 86.5 | 33.8 | 12.1 | 11.1 | 8.75 | 0.98 | 1.45 | 1.59 | 0.63 |
| 1987 | 59.5 | 164 | 69.9 | 23.9 | 6.2 | 4.88 | 2.51 | 0.35 | 0.46 | 0.49 |
| 1988 | 36.4 | 48.7 | 132.6 | 49.7 | 13.2 | 2.59 | 2.15 | 1.07 | 0.17 | 0.23 |
| 1989 | 33.1 | 29.8 | 39.7 | 99.6 | 27 | 5.58 | 0.94 | 0.78 | 0.32 | 0.08 |
| 1990 | 112.7 | 27.1 | 24.2 | 30.3 | 61.1 | 13.43 | 1.68 | 0.31 | 0.14 | 0.13 |
| 1991 | 213.9 | 92.3 | 20.5 | 17.6 | 17 | 28.7 | 5.27 | 0.63 | 0.12 | 0.05 |
| 1992 | 47 | 175.1 | 67.8 | 14.8 | 9.6 | 7.25 | 10.74 | 1.92 | 0.21 | 0.06 |
| 1993 | 57.2 | 38.4 | 138.5 | 45.8 | 7 | 3.65 | 2.59 | 4.08 | 0.67 | 0.07 |
| 1994 | 89 | 46.8 | 30.8 | 102.2 | 25.9 | 3.03 | 1.43 | 0.83 | 1.31 | 0.23 |
| 1995 | 44.4 | 72.9 | 35.5 | 22.2 | 58.1 | 11.54 | 1.15 | 0.52 | 0.36 | 0.49 |
| 1996 | 125.2 | 36.3 | 56 | 22.4 | 12.9 | 25.93 | 4.23 | 0.38 | 0.16 | 0.13 |
| 1997 | 22 | 102.5 | 27 | 36.3 | 11.4 | 6.56 | 9.56 | 1.34 | 0.13 | 0.06 |
| 1998 | 61.3 | 18 | 82.7 | 18.5 | 19.9 | 4.93 | 3.01 | 3.27 | 0.45 | 0.05 |
| 1999 | 143.4 | 50.2 | 14.5 | 60.2 | 10 | 9.15 | 1.85 | 1.18 | 0.96 | 0.13 |
| 2000 | 191.2 | 117.4 | 39.8 | 10.4 | 33.7 | 4.12 | 3.11 | 0.69 | 0.4 | 0.35 |
| 2001 | 228.7 | 156.5 | 93.7 | 26.9 | 6.4 | 15.28 | 1.61 | 1.01 | 0.27 | 0.13 |
| 2002 | 61.3 | 187.3 | 125.9 | 66.7 | 15.7 | 3.95 | 6.85 | 0.71 | 0.39 | 0.12 |
| 2003 | 185.6 | 50.1 | 152.3 | 93.6 | 40.1 | 8.23 | 2.21 | 2.76 | 0.34 | 0.17 |
| 2004 | 471.1 | 152 | 40.7 | 119 | 61.9 | 21.5 | 4.05 | 1.13 | 1.14 | 0.2 |
| 2005 | 110.2 | 385.7 | 123 | 29.6 | 81.4 | 33.19 | 9.58 | 1.56 | 0.5 | 0.49 |
| 2006 | 51.7 | 90.3 | 313.6 | 92.1 | 18 | 43.68 | 14.71 | 3.7 | 0.55 | 0.18 |
| 2007 | 53.8 | 42.3 | 73.7 | 247.7 | 56.2 | 8.73 | 18.71 | 5.17 | 1.05 | 0.16 |
| 2008 | 146.3 | 44 | 33.9 | 56.8 | 164.6 | 25.9 | 4.11 | 6.31 | 1.75 | 0.39 |
| 2009 | 47.5 | 119.8 | 34 | 23.7 | 37.7 | 86.78 | 11.24 | 1.81 | 2.43 | 0.7 |
| 2010 | 34.9 | 38.9 | 97.1 | 24.9 | 15 | 22.84 | 39.37 | 4.02 | 0.83 | 0.74 |
| 2011 | 24 | 28.6 | 31.7 | 74.1 | 14 | 7.92 | 12.58 | 16.4 | 1.59 | 0.36 |
| 2012 | 42 | 19.6 | 23.2 | 24.6 | 50 | 6.88 | 3.9 | 6.7 | 7.8 | 0.86 |
| 2013 | 29.2 | 34.4 | 15.9 | 17.8 | 17 | 29.05 | 3.62 | 2.08 | 3.24 | 3.98 |
| 2014 | 24.7 | 23.9 | 27.9 | 12.1 | 11.9 | 10.29 | 15.44 | 1.87 | 1.1 | 1.52 |
| 2015 | 130.8 | 20.2 | 19.4 | 21.5 | 8.3 | 7.32 | 5.95 | 8.35 | 1.13 | 0.55 |
| 2016 | 66.8 | 107.1 | 16.3 | 14.3 | 14.1 | 4.91 | 4.19 | 3.25 | 4.53 | 0.61 |
| 2017 | 66.8 | 54.7 | 87.5 | 11.7 | 8.8 | 7.72 | 2.51 | 2.08 | 1.6 | 2.24 |

Table 10.2.3 Haddock in division Va. Fishing mortality from the SPALY run using the March and October surveys for tuning. Predictions based on F4-7 $=0.3$ are highlighted.

| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 0.002 | 0.018 | 0.162 | 0.419 | 0.669 | 0.833 | 0.99 | 0.553 | 0 |
| 1980 | 0.018 | 0.023 | 0.144 | 0.282 | 0.508 | 0.657 | 0.685 | 0.561 | 0.724 |
| 1981 | 0.001 | 0.019 | 0.108 | 0.328 | 0.813 | 0.92 | 0.793 | 0.463 | 0.569 |
| 1982 | 0.003 | 0.032 | 0.156 | 0.369 | 0.501 | 0.751 | 1.056 | 0.903 | 1.288 |
| 1983 | 0.001 | 0.032 | 0.301 | 0.357 | 0.683 | 0.692 | 0.706 | 0.643 | 1.051 |
| 1984 | 0.013 | 0.051 | 0.22 | 0.449 | 0.784 | 0.607 | 0.825 | 0.493 | 0.369 |
| 1985 | 0.035 | 0.122 | 0.315 | 0.532 | 0.582 | 0.719 | 0.737 | 1.314 | 1.184 |
| 1986 | 0.013 | 0.148 | 0.467 | 0.625 | 1.048 | 0.816 | 0.937 | 0.976 | 0.918 |
| 1987 | 0.013 | 0.141 | 0.389 | 0.669 | 0.62 | 0.657 | 0.53 | 0.5 | 0.685 |
| 1988 | 0.005 | 0.086 | 0.411 | 0.665 | 0.811 | 0.815 | 0.998 | 0.557 | 0.557 |
| 1989 | 0.007 | 0.071 | 0.288 | 0.498 | 1.003 | 0.917 | 1.552 | 0.682 | 0.632 |
| 1990 | 0.079 | 0.117 | 0.379 | 0.556 | 0.736 | 0.772 | 0.769 | 0.794 | 0.467 |
| 1991 | 0.109 | 0.123 | 0.413 | 0.651 | 0.783 | 0.811 | 0.89 | 0.473 | 0.25 |
| 1992 | 0.035 | 0.192 | 0.555 | 0.762 | 0.827 | 0.768 | 0.858 | 0.973 | 0.204 |
| 1993 | 0.022 | 0.104 | 0.37 | 0.635 | 0.736 | 0.934 | 0.933 | 0.842 | 0.383 |
| 1994 | 0.078 | 0.128 | 0.365 | 0.608 | 0.769 | 0.821 | 0.643 | 0.786 | 0.575 |
| 1995 | 0.063 | 0.259 | 0.337 | 0.607 | 0.804 | 0.895 | 0.971 | 0.856 | 0.926 |
| 1996 | 0.099 | 0.233 | 0.473 | 0.48 | 0.798 | 0.95 | 0.912 | 0.79 | 0.756 |
| 1997 | 0.015 | 0.176 | 0.404 | 0.641 | 0.579 | 0.873 | 0.9 | 0.819 | 0.253 |
| 1998 | 0.017 | 0.117 | 0.413 | 0.575 | 0.781 | 0.738 | 1.025 | 1.041 | 0.53 |
| 1999 | 0.032 | 0.126 | 0.38 | 0.689 | 0.878 | 0.792 | 0.87 | 0.806 | 0.776 |
| 2000 | 0.025 | 0.193 | 0.286 | 0.591 | 0.737 | 0.93 | 0.74 | 0.933 | 0.807 |
| 2001 | 0.018 | 0.14 | 0.337 | 0.286 | 0.603 | 0.62 | 0.745 | 0.568 | 0.44 |
| 2002 | 0.006 | 0.096 | 0.308 | 0.445 | 0.381 | 0.71 | 0.523 | 0.65 | 0.468 |
| 2003 | 0.009 | 0.047 | 0.213 | 0.424 | 0.508 | 0.469 | 0.685 | 0.345 | 0.383 |
| 2004 | 0.012 | 0.117 | 0.179 | 0.424 | 0.609 | 0.753 | 0.616 | 0.645 | 0.71 |
| 2005 | 0.007 | 0.089 | 0.299 | 0.423 | 0.614 | 0.753 | 0.849 | 0.809 | 0.653 |
| 2006 | 0.003 | 0.036 | 0.294 | 0.524 | 0.648 | 0.846 | 1.056 | 1.057 | 0.829 |
| 2007 | 0.021 | 0.06 | 0.209 | 0.575 | 0.553 | 0.886 | 0.882 | 0.787 | 0.58 |
| 2008 | 0.06 | 0.159 | 0.209 | 0.44 | 0.635 | 0.62 | 0.757 | 0.723 | 0.636 |
| 2009 | 0.01 | 0.109 | 0.259 | 0.302 | 0.59 | 0.829 | 0.585 | 0.992 | 0.987 |
| 2010 | 0.003 | 0.071 | 0.376 | 0.438 | 0.396 | 0.676 | 0.725 | 0.642 | 0.963 |
| 2011 | 0.01 | 0.057 | 0.194 | 0.512 | 0.508 | 0.431 | 0.543 | 0.42 | 0.428 |
| 2012 | 0.011 | 0.065 | 0.167 | 0.342 | 0.442 | 0.429 | 0.526 | 0.474 | 0.372 |
| 2013 | 0.008 | 0.075 | 0.196 | 0.302 | 0.432 | 0.459 | 0.439 | 0.555 | 0.478 |
| 2014 | 0.011 | 0.06 | 0.175 | 0.29 | 0.348 | 0.415 | 0.306 | 0.494 | 0.554 |
| 2015 | 0.018 | 0.106 | 0.224 | 0.326 | 0.358 | 0.406 | 0.411 | 0.411 | 0.411 |
| 2016 | 0.003 | 0.128 | 0.288 | 0.402 | 0.471 | 0.5 | 0.505 | 0.505 | 0.505 |
| 2017 | 0.002 | 0.093 | 0.295 | 0.421 | 0.486 | 0.52 | 0.52 | 0.52 | 0.52 |

Table 10.4.1 Output from short term predictions. Numbers here apply to calendar years.

The adopted HCR lead to TAC of 30.4 kt for the fishing year 2014/2015 and landings of $\mathbf{2 9 . 0}$ thous. tonnes in the calendar year 2015.

| 2015 |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
| Bio 3+ | SSB | Fmult | F4-7 | Landings |
| 113 | 78 | 1.072 | 0.329 | 32 |


|  |  | 2016 |  | 2017 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fmult | F4-7 | Bio 3+ | SSB | Landings | Bio 3+ | SSB |
| 0.1 | 0.031 | 101 | 85 | 3 | 148 | 108 |
| 0.2 | 0.061 | 101 | 85 | 6 | 145 | 105 |
| 0.3 | 0.092 | 101 | 85 | 9 | 142 | 102 |
| 0.4 | 0.123 | 101 | 85 | 12 | 139 | 100 |
| 0.5 | 0.153 | 101 | 85 | 14 | 137 | 97 |
| 0.6 | 0.184 | 101 | 85 | 17 | 134 | 95 |
| 0.7 | 0.215 | 101 | 85 | 20 | 131 | 93 |
| 0.8 | 0.245 | 101 | 85 | 22 | 129 | 90 |
| 0.9 | 0.276 | 101 | 85 | 24 | 127 | 88 |
| 1 | 0.307 | 101 | 85 | 27 | 124 | 86 |
| 1.1 | 0.337 | 101 | 85 | 29 | 122 | 84 |
| 1.2 | 0.368 | 101 | 85 | 31 | 120 | 82 |
| 1.3 | 0.399 | 101 | 85 | 33 | 118 | 80 |
| 1.4 | 0.429 | 101 | 85 | 35 | 116 | 78 |
| 1.5 | 0.46 | 101 | 85 | 37 | 114 | 77 |
| 1.6 | 0.491 | 101 | 85 | 39 | 112 | 75 |
| 1.7 | 0.521 | 101 | 85 | 41 | 110 | 73 |
| 1.8 | 0.552 | 101 | 85 | 43 | 108 | 72 |
| 1.9 | 0.582 | 101 | 85 | 45 | 107 | 70 |
| 2 | 0.613 | 101 | 85 | 47 | 105 | 68 |



Figure 10.1.1 Haddock in division Va. Landings 1905-2013


Figure 10.1.2 Haddock Division VA. Landings in tons and percent of total by gear and year.


Figure 10.1.3 Haddock Division VA. Spatial distribution af landings. The legend show tonnes per square mile.


Figure 10.1.4 Haddock in division Va. Age disaggregated catch in tons.


Figure 10.1.5 Haddock in division Va. Percent of catch in tonnes 2014 compared to last years predictions.


Figure 10.1.6 Icelandic haddock. Total biomass indices from the groundfish surveys in March (lines and shading) and the groundfish survey in October vertical segments. The standard error in the estimate of the indices is shown in the figure. Due to a strike the autumn survey was not conducted in October 2011.

*Figure 10.1.7. Spatial distribution of haddock in the groundfish survey in March. The legend show kg per hour towed.


Figure 10.1.8. Proportion of the landings and the biomass of 42 cm and larger haddock that is in the north area. The small figure shows the northern area.


Figure 10.1.9 Haddock in division Va. Mean weight at age in the survey. Predictions are shown as red. The values shown are used as weight at age in the stock and spawning stock.


Figure 10.1.10 Haddock in division Va. Mean weight at age in the catches. Predictions are shown as red.


Figure 10.1.11 Haddock in division Va. Maturity at age in the survey. The blue bar indicates predictions. The values are used to calculate the spawning stock.


Figure 10.1.12. Catch per unit effort in the most important gear types. The bars are based on locations where more than $50 \%$ of the catch is haddock and the lines on all records where haddock is caught. A change occurred in the longline fleet starting September 1999. Earlier only vessels larger than 10 BRT were required to return logbooks but later all vessels were required to return logbooks. Not updated


Figure 10.2.1 Haddock in division Va. Summary from assessment. Red colours in lower figure indicates predicted values.


Figure 10.2.2. Haddock in division Va. Residuals from the fit to survey data . from Adapt run based on the both the surveys. Coloured circles indicate positive residuals (observed > modelled). The largest circle corresponds to a value of 0.87 . Residuals are proportional to the area of the circles. Lage efri harvest ratio


Figure 10.2.3. Haddock in division Va. Observed and predicted biomass from the surveys according to the SPALY run.


Figure 10.2.4. Haddock in division Va . Results from the spaly run. Catchability and CV from the autumn survey (wide lines) and March survey (thinner lines). Estimates from 2008 shown dashed.


Figure 10.2.5. Haddock in division Va. Retrospective pattern of SSB from the SPALY run. Each retro ends one year after the assesemt, but the biomass in the beginning of the year after the assessment year is the basis for advice. Errors in prediction of weight and maturity at age are not included.

10.2.6 Haddock in division Va. Estimate of the reference biomass 45 cm and larger from some different assessment models and tuning data. (SMB refers to March survey, SMH autumn survey and SMX both.

10.2..7 Haddock in division Va. Comparison of some of the results of 2014 assesement based on different tuning data and 2013 assessment tuned with both the surveys. .


Figure 10.2.8. Comparison of 2013 and 2014 assessment


Figure 10.4.1 Haddock in division Va. Development of the landings during the fishing year 2013/2014 (left side) and calendar year (2014) on the left. Fishing year 2012/2013 and calendar year 2013 shown for comparison. Tac (kvóti) for the fishing year shown in the left figure.


Figure 10.4.2 Haddock in division Va. Input data to prediction. Predictions are based on the period since 2000. . Exponential of the yearfactor (growth multiplier) in the equation
$\log \frac{W_{a+1, t+1}}{W_{a, t}}=\alpha+\beta \log W_{a, t}+\delta_{\text {year }}$


Figure 10.4.3 Haddock in division Va. Proportion of the biomass of a yearclass above certain size. The points show data, compiled from the March survey and the lines a curve fitted to the data and used in simulations.

## 11 Icelandic summer spawning herring

## Executive summary

The total reported landings in 2014/15 fishing season were 95.5 kt but the TAC was set at 83 kt -the difference being caused by a transfer of quota between years. The fishable stock (age $4+$ ) in the winter surveys 2014/15 was estimated at 433 kt , compared to 410 kt in the winter 2013/14. The 2013 year class (age 1 in 2014) appears small.

This is an update assessment where the 2014 data have been added to the input data and no revisions of last year's data. The analytical assessment model, NFT-Adapt, indicates that the biomass of age 3+ is 438 kt and SSB is 342 kt in the beginning of 2015. Record small year class from 2011 entering the spawning stock in 2015, causes a decline in SSB but it is still above BPA. Fishing at $\mathbf{F}_{0.1}=0.22$ in the fishing season 2015/16 will give a catch of 71 thousands tons. SSB in 2016 is expected to be 327 kt .

Changes in the predictions approach, where the geometric mean for number at age-3 in the assessment year (2015) was replaced by a projection of number at age 3 from a survey estimate at age- 1 of the year class, has decreased the uncertainty and has minor impacts on the advice ( $3 \%$ lower). This year's results support that additional natural mortality in the stock due to Ichthyophonus infection should only be applied for the first two years of the outburst.

### 11.1 Scientific data

### 11.1.1 Surveys description

The scientific data used for assessment of the Icelandic summer-spawning herring stock are based on annual acoustic surveys (IS-Her-Aco-4Q/1Q)., which have been ongoing since 1974 (Table 11.1.1.1). These surveys have been conducted in October-December or January. The surveyed area each year is decided on basis of available information on the distribution of the stock in previous and the current year, which include information from the fishery. Thus, the survey area varies spatially as the survey is focused on the adult and incoming year classes but is considered to cover the whole stock each year.

The acoustic abundance index for the adult stock in the winter 2014/2015 derives from three dedicated acoustic surveys in the autumn 2014 (Table 11.1.1.2). During 24 November to 6 December, RV Bjarni Sæmundsson covered areas west, south and southeast off Iceland. The acoustic index for Breiðafjörður, which has been the main over wintering areas of the stock in last eight winters, was then derived from two surveys there (Óskarsson and Reynisson 2015).

Like last six winters, but different from subsequent years, the nursery grounds of the stock were covered on RV Dröfn in a survey during November 13-23. The objective was to get an acoustic estimate of juveniles at age 1 and estimate their prevalence of Ichthyophonus infection (see Óskarsson and Pálsson 2015).

The instrument and methods in the surveys were the same as in previous years and described in the stock annex.

### 11.1.2 The surveys results

The fishable part of the herring stock was observed in four main areas, in an offshore area west of Iceland (Kolluáll), in Breiðamerkurdjúp southeast of Iceland, near Vestmannaeyjar south of Iceland and in Kolgrafafjörður in Breiðafjörður west of Iceland (see locations on Fig. 11.1.2.1). During the winters 2006/07 to 2012/2013, most of the stock was measured acoustically in the southern part of the bay Breiðafjörður, however, that changed in last winter and was more extreme this winter. The total amount of the adult stock (age 4+) in Breiðafjörður (Kolgrafafjörður) came to only 10 kt , while in Kolluáll on the shelf outside of Breiðafjörður, 345 kt were measured. Then, like in recent years, around 92 kt were measured in Breiðamerkurdjúp and around 10 kt near Vestmannaeyjar. The total estimate of the adult stock (age 4+) was therefore 433 kt , compare to 410 kt in the autumn 2013. The total biomass was 474 kt in comparison to 473 kt in the autumn 2013.

Figure 11.1.2.2 shows the total estimated biomass of age $3+$ in the acoustic survey since 1973, how the eastern part of the stock generally decreased in size and the western part increased during around 1995-2007, and then the opposite from 2007-2014.

The 2010 year class (age 4 in the autumn 2014) was the most numerous in the survey or $16 \%$ of the total number of herring and was both off the south and west coast of Iceland (Table 11.1.1.1). The 2008 and 2009 year classes were also numerous (12\%), and were mostly off the west coast (Kolluáll).

The number of juvenile herring (i.e. age 1) observed acoustically amounted to 438 million fish. Around $86 \%$ of it derived from unconventional nursery grounds west of Iceland (Jökuldjúp) and the rest in Eyjafjörður off the mid northcoast. Applying the linearregression provided by Gudmundsdottir et al. (2007) implied that the 2013 year class will be 464 millions at age 3 in 2016, or below average year class size ( 575 millions at age 3). This number is used in the forecast in the 2015 assessment below.
The length composition of the adult part of the stock in the acoustic estimation in 2014/15 was based on total 10 samples, 8 taken west of Iceland while only 2 samples were obtained south and southeast of Iceland (Table 11.1.2.1). The age composition was then derived from length-at-age key from the same samples. The total number of aged scales from these samples was 534 . While the number of samples in the west is considered adequate, the number of samples in the east was not. The low sampling number there was due to combination of bad weathers, restrictions of survey time and behaviour of the fish (Óskarsson and Pálsson 2015). This adds uncertainty to the estimated length/age composition in that area.

### 11.1.3 Prevalence of /chthyophonus infection in the stock

In a working document to NWWG 2013, Óskarsson and Pálsson (2013) addressed the development and nature of the massive and long-lasting Ichthyophonus hoferi outbreak in Icelandic summer-spawning herring since the autumn 2008 to 2013. Their main conclusions were that the infection was only causing significant additional mortality in the first two years, despite a high prevalence of infection for five years. It indicated that the infection to be less lethal for herring than had been assumed in previous assessments. This was followed in the 2013 and 2014 assessments (ICES 2013a; 2014), where additional natural mortality because of the infection, and estimated from catch samples (e.g. Óskarsson et al. 2012a; ICES. 2012), was only be applied for the years 2009 and 2010, but not the following years.

The results of this year's investigations are supporting this main conclusion of not significant infection mortality since 2010.

The prevalence of infection in the Icelandic summer-spawning herring in the winter 2014/2015 in Breiðafjörður and Kolluáll was highest for the 2006, 2005, 2004 and 2013 year classes (Figure 11.1.3.1). This is comparable to results in recent years where the prevalence of infection has been in the range of $35-55 \%$ (Figure 11.1.3.2). The prevalence of infection of the younger age groups continues to be low, suggesting a low rate, if any, of new infection in the stock. For the yearclasses 2006 and older, the prevalence of infection seems to be going slowly down.

### 11.2 Information from the fishing industry

The total landings of Icelandic summer-spawning herring in 2014/2015 season were about 95 kt with no discards reported (Table 11.2.1 and in Figure 11.2.1). Note that the total landings include also bycatches in the mackerel fishery in June-August 2014, even if they belong to the official fishing season 2013/2014. This is a traditional method in assessment of the stock. The quality of the herring landing data regarding discards and misreporting is consider to be adequate as implied in a general summary in section 7 and in the Her-Vasu stock annex. The recommended TAC, provided in the spring 2014, was 83 kt and allowable TAC 83 kt . The difference between the landings and TAC is due to transfer of quota from the previous fishing season (7 kt) and the next season ( $\sim 5$ kt ).

The direct fishery started in end of October in Kolluáll, an offshore area west of Iceland. Most of the catches were taken there in October to December in pelagic trawls, or 83\% of the total catch (Fig. 11.2.2). In January around $4 \%$ were taken in the same area, while the remaining catch ( $13 \%$ ) was taken as bycatch in the fishery for the Norwegian spring-spawning herring, NSSH, and Atlantic mackerel during June to October. Because of the location of the main fishery, and the bycatch fishery, $99.6 \%$ of the catch was taken in pelagic trawls. This winter, drift nets were used in this fishery for the fourth time since mid 1980s. It was because of allocation of catch quota to small fishing vessels (<200 bt) that were allowed to catch some limited catches. The total catch in drift nets amounted though to only 50 tons.

Like in some of the previous winters, spring-spawning herring (Icelandic spring spawners or NSSH) was mixed with the Icelandic summer-spawning herring stock in the catches in the winter 2014/2015. Based on maturity stage of the herring in catch samples, $3 \%$ of the herring west of Iceland were spring spawners.

### 11.2.1 Fleets and fishing grounds

The herring fishing season has taken minor changes in the last three decades as detailed in the stock annex. All seasonal restricted landings, catches and recommended TACs since 1984 are given in thousands tonnes (kt) in Table 11.2.1.

Around $99.6 \%$ of the catch in 2014/15 was taken with purse-seines, and the rest with pelagic trawls and by drift nets (Figure 11.2.1.1). During all fishing seasons since 2007/2008, most of the catches ( $\sim 90 \%$ ) have been taken west off Iceland in Breiðafjörður, while prior to that they were mainly taken off the south-, southeast-, and the east coast. In 2013/2014 there was an indication for changes in this pattern, with less proportion in Breiðafjörður, and then in 2014/2015 almost all of the overwintering west of Iceland took place offshore. These changes in distribution explain the dominating pelagic trawl fishery, which is preferred by the fleet over purse seine in offshore areas.

To protect juveniles herring ( 27 cm and smaller) in the fishery, area closures are enforced based on a regulation of the herring fishery set by the Icelandic Ministry of Fisheries (no. 376, 8. Oktober 1992). No closure was enforced in this herring fishery in $2014 / 15$. Normally, the age of first recruitment to the fishery is age-3, which is fish at length around $26-29 \mathrm{~cm}$.

### 11.2.2 Catch in numbers, weight at age and maturity

Catch at age in 2014/2015:
The procedure for the catch at age estimations, as described in the Stock Annex, was followed for the 2014/15 fishing season. It involves calculations from catch data collected at the harbours by the research personnel or at sea by fishermen (Table 11.2.2.1). This year, the calculations were accomplished by dividing the total catch into five cells confined by season and area as detailed in Óskarsson and Pálsson (2015). In the same way, five weight-at-length relationships derived from the length and weight measurements of the catch samples were used and one length-at-age relations. The catches of the Icelandic summer spawners in number-at-age for this fishing season as well as back to 1982 are given in Table 11.2.2.2. The geographical location of the sampling is shown on Figure 11.2.2.1.

The age compositions of the catches during the overwintering and the bycatch of herring in the mackerel and NSS-herring were similar. The main difference was for the composition of age 3 (2011 year class), which was in higher proportion in the bycatch (4\%) than at the small overwintering areas in Kolluáll (1\%). The bycatch was well sampled over a wide area and is considered to give a better representation of the age class strength, in contrast to the small area of the overwintering grounds. Thus, the age 3 herring was to a less degree overwintering on west of Iceland, and more in the south, and as a result not targeted in the main fishery. This might cause some underestimation of the year class in the assessment, even if all measures indicate that it is small.

## Weight at age:

As stated in the stock annex, the mean weight-at-age of the stock is derived from the catch samples (Table 11.2.2.3). The total number of fish weighed from the catch in 2014/15 was 3093 and 2450 of them were aged from their fish scales.

## Proportion mature:

The fixed maturity ogives were used in this year's assessment, as introduced in the stock annex, where proportion mature-at-age 3 is set $20 \%$ and $85 \%$ for fish at age 4 , while all older fish is considered mature.

## Observed versus predictions of catch composition:

The relative contribution of the different year (age) classes was somewhat different from what was predicted in the analytical assessment in 2014 (Figure 11.2.2.2). The biggest difference was for age 3 (2011 year class) where the prediction was based on geometric mean but not survey results. Otherwise the contribution of the age 5 and 6 (the 2009 and 2008 year classes, respectively) was less than predicted while the other way around for other age groups.

### 11.3 Analytical assessment

### 11.3.1 Analysis of input data

Examination of catch curves for the year classes from 1984 to 2011 (Figure 11.3.1.1) indicates, in general, that the total mortality signal $(Z)$ in the fully recruited age groups is around 0.4 . It is under the assumption that the effort has been the same the whole time. In recent years the effort has changed a lot because of the infection and spatial distribution of the stock, and the mass mortality in 2012/2013, which makes any strong deductions from the catch curves for those recent years meaningless.

Catch curves were also plotted using the age disaggregated survey indices for each year class from 1984-2011 (Figure 11.3.1.2). Even if the total mortalities look at bit noisy for some year classes, they seem to be fairly close to 0.4 . There is an indication that the fish is fully assessable to the survey at age 3, but apparently a year later occasionally.

Mortality in the stock because of the Ichthyophonus outbreak can not be detected clearly from the catch curves of the surveys. There is possibly a small change in level of the curve around 2009 for the big 1999 year classes. However, it should be noted that the highest prevalence of infection has been in the 2004, 2005 and 2006 year classes and they were not all fully in the survey prior to the infection outbreak. Further work on this matter is ongoing.

The year class strength was evaluated independently from the analytical assessment, by sum the total catch of each year class (Figure 11.3.1.3). The 1999 year class is apparently the largest in the time series. For the most recent five year classes, only 2008 and 2009 might become close to average size, where around $50 \%$ and $30 \%$, respectively, of their expected total catch have been already caught considering their current total catch in relation to the cumulative fishing of the age 4 and 5 as estimated from the year classes 1978-1996 (Figure 11.3.1.4).

### 11.3.2 Exploration of different assessment models

In order to explore the data this year, only one assessment tool was used, NFT-ADAPT (VPA/ADPAT version 3.3.0 NOAA Fisheries Toolbox). The NFT-Adapt has been used as the basis for the assessments since 2005 and it was considered appropriate as the principal assessment tool for the stock at benchmark assessment in January 2011 (ICES 2011a). The catch data used were from 1987/88-2014/15 (Table 11.2.2.2) and survey data from 1987/88-2014/15 (Table 11.1.1.1). Other input data consisted of: (i) mean weight at age (Table 11.2.2.3); (ii) maturity ogive (Table 11.2.2.4); (iii) natural mortality, M, that was set to 0.1 for all age groups in all years, except for 2009, where it was set 0.49 because of the Ichthyophonus infection, and for 2010 where M was for same reasons age dependent (Table 11.3.2.1; Óskarsson and Pálsson 2013); (iv) proportion of $M$ before spawning was set to 0.5 ; and (v) proportion of F before spawning was set to 0 . Thus, no changes in the input data from last except for one more year of data.

## NFT-Adapt:

The estimated parameters in NFT Adapt are the stock in numbers at age. The parameters are output by the Levenburg-Marquardt Non-Linear Least Squares minimization algorithm (see VPA/ADAPT Version 3.3.0, Reference Manual). The estimated parameters were stock numbers for ages 4 to 12 in the end of year 2014, while the stock numbers at age 3 were set to the geometric mean from 1991-2011. Like in last years' assessments, the input partial recruitment was set to 1 for ages 4 and older and the classic
method was used to calculate the value of fully-recruited fishing mortality in the terminal year.
The catchability at age in the survey, as estimated by the NFT Adapt, and the CV is shown in Figure 11.3.2.1. The age groups 3-10 were used for tuning (Table 11.1.1.1 as decided at the benchmark in ICES (2011).
The output and model settings of the NFT-Adapt run (the adopted final assessment model; see below) are shown in Table 11.3.2.2. Stock numbers and fishing mortalities derived from the run are shown in Table 11.3.2.3 and Table 11.3.2.4, respectively, and summarized in Table 11.3.2.5 and Figure 11.3.2.2.
Residuals of the model fit are shown in Figure 11.3.2.3 and Table 11.3.2.6, and shows both cohort and year affects. The main pattern is the same as presented in recent assessments. Positive residuals, where the model estimates is smaller than seen in the survey, can be seen for 1994 and 1999 year classes for almost all age groups and a negative residuals for the 2001 year class. Year blocks of positive residuals are apparent for the years $\sim 2000$ to 2006 (i.e. referring to January 1st), indicating that the model estimated the age groups smaller than observed in the surveys. During these years, the stock was overwintering in offshore areas off the east and west coast, compare to mainly easterly distribution before and overwintering in inshore areas there after (from ~2006-2012). These positive blocks could therefore reflect changes in catchability of the survey for these years. Positive residuals, even if relatively weaker, were also observed for 2012. A block of negative residuals was however observed for 2009 (survey in the autumn 2008).

Retrospective analysis (Figure 11.3.2.4) indicate a more stability in the most recent years than often before, i.e. adding new data to the model does not change the present perception of the stock size. The same applies correspondingly to the fishing mortality. Furthermore, to sustain the high M in the input data for 2009 and 2010 because of the infection, SSB of the most recent four years lifts in comparison to the preceding years. It required also an increase in recruitment estimates as apparent on the retrospective plots of number-at-age 3. A revision of the number at age 3 of the 2008 and 2009 year classes (in 2011 and 2012) is also apparent retrospectively, which is related to their high survey indices at age 3. Note that the high F in 2012 (Figure 11.3.2.4) is due to the mass mortality, which was added to the catches that year in the assessment as presented earlier (ICES 2014).
Like described before (ICES 2014), the main difference between observed and predicted survey values from the NFT-Adapt model was for the period 1999-2004, where the observed values were well above the predicted (Figure 11.3.2.5), otherwise they fitted relatively well. Like seen in the residual plot (Figure 11.3.2.3), the observed value for the 2009 survey was lower than predicted and the vice versa for the 2012 survey (referring to the beginning of the year; Figure 11.3.2.5). The low survey value in 2009 is likely underestimate due to distribution of the stock that year in Breiðafjörður (Óskarsson et al. 2010), while the reason for the positive block during 2000-2004 is not fully known even if mainly caused by the large 1999 year class and possibly changes in the catchability of the survey as suggested above. However, an exploratory run in NFT-Adapt done in the 2011 assessment (ICES 2011b) where these years were excluded in the tuning, did not change the point estimate of the stock size in the latest year (January 1st 2011), implying that the terminal point estimates in the final run was not driven by this residual block.

## Comparisons of model run from previous year:

As pointed out above, only NFT-Adapt was run in this year's assessment. Thus, the comparisons here are limited to the final NFT-Adapt runs in 2014 and 2015 with respect to recruitment, biomass, and N weighed average $\mathrm{F}_{5-10}$ (Figure 11.3.2.2). The results of the final NFT run in 2015 were in a good agreement the run in 2014. The main difference is related to the number at age 3 in 2014, which was based on geometric mean in the 2014 assessment while estimated in the 2015 assessment. This single difference is apparent when the estimates of number at age in 2014 are compared (Figure 11.3.2.6).

### 11.3.3 Final assessment

This is an update assessment so the results of the NFT-Adapt were adopted as point estimator for the prediction and thus the basis for the advice as in recent years. The model settings and outputs are shown in Table 11.3.2.2 to Table 11.3.2.4 and Figure 11.3.2.2.

The assessment (Table 11.3.2.5 and Figure 11.3.2.2) indicates that the fishing mortality (weighed average for age $5-10$ ) was 0.26 in 2014 or above $\mathrm{F}_{\mathrm{pa}}=0.22$, while only 0.16 in 2013. This reflects partly a transfer of 7 kt of quota not caught in 2013 to 2014. The low F during 2009 to 2011 was related to cautious TAC and apparently overestimation of the mortality caused by the Ichthyophonus outburst.
As mention above, the estimated number of herring that died in Kolgrafafjörður in the two incidents of the mass mortalities there (Óskarsson et al. 2013) were added to the catches in 2012 and is also included in the high $F$ that year (Table 11.3.2.5 and Figure 11.3.2.2). The F related only to landings in 2012 came to 0.22 .

### 11.4 Reference points

## Precautionary reference points:

The Working Group has pointed out that managing this stock at an exploitation rate at or above $\mathrm{F}_{0.1}$ has been successful in the past, despite biased assessments. Thus, as stated in the Stock Annex, the Northern Pelagic and Blue Whiting Fisheries Working Group agreed in 1998 with the SGPAFM on using $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{0.1}=0.22, \mathrm{~B}_{\mathrm{pa}}=\mathrm{Blim}^{*}{ }^{*} \mathrm{e}^{1.645 \sigma}=300000 \mathrm{t}$ where Blim 200000 t . The Study Group on Precautionary Reference Points for Advice on Fishery Management met in February 2003 and concluded that it was not considered relevant to change the Blim from 200000 t . The WG have not dealt with this issue.

The fishing mortality during 1987 to 2008 was on average 0.31 (weighed $\mathrm{F}_{5-10}$ ), or approximately $40 \%$ higher than the intended target of $\mathrm{F}_{0.1}=0.22$. This is despite the fact that the managers have followed the scientific advice and restricted quotas with the aim of fishing at the intended target. During this period the SSB has remained above $B_{\lim }$ and reached a record high level around 2008.

## MSY based reference points:

The MSY based reference points have not been set for Icelandic summer-spawning herring, but exploratory work was present at the NWWG meeting in 2011 in a form as requested by ICES (ICES 2011b). The HCS program Version 10.3 (Skagen, 2012) was used to evaluate possible points based on the MSY framework that could be a basis for a management plan and Harvest Control Rule later.

Number of different runs was made with varying settings. The results implied that the MSY framework was confirmative with the currently used precautionary reference
points. It means that the currently used $\mathrm{F}_{0.1}=0.22$ could be a valid candidate for $\mathrm{F}_{\mathrm{ms}}$. This however, needs to be explored more thoroughly later.

### 11.5 State of the stock

The stock was at high levels until 2008 but since then a substantial reduction took place despite a low fishing mortality. The reduction was caused by mortality induced by Ichthyophonus infection in the stock in 2008 and 2009. However, the observed high prevalence of infection for all the years since then is not considered to be causing further mortality in the stock and the negative trend in the stock size has reversed due to incoming of year classes at near and just below average size. Record small year class from 2011 entering the spawning stock in 2015 causes a decline in SSB even if still above BPA.

### 11.6 Short term forecast

### 11.6.1 The input data

The final adopted model, NFT-Adapt, which gave the number-at-age on January $1^{\text {st }}$, 2015, was used for the prognosis. All input values for the prognosis are given in Table 11.6.1.1. Note that different from previous four assessments, the number at age 3 in the assessment year (2012 year class) was not set equal to the geometric mean. Instead the predicted number from juvenile survey was used as detailed below.

The weights were estimated from the last year catch weights (see Stock Annex) and as in the recent years, the weights are expected to continue to be high (Figure 11.6.1.1). The selection pattern used in the prognosis was based on averages over 2012 to 2014 from the final run (Figure 11.6.1.2) (see Stock Annex). As traditionally, M was set 0.1, proportion M before spawning was set 0.5 and proportion F before spawning was set 0 . The numbers of recruits in the prognosis were determined as follows:

The 2012 year class: An acoustic survey aimed for getting an abundance index for this year class took place in November 2013 (ICES 2014), and using a relation obtained by Gudmundsdóttir et al. (2007) provides estimate of 477 millions at age 3 in 2015.
The 2013 year class: An acoustic survey aimed for getting an abundance index for this year class took place in November-December 2014 (Óskarsson and Reynisson 2015), and using a relation obtained by Gudmundsdóttir et al. (2007) provides estimate of 464 millions at age 3 in 2016.
The 2014 year class: No acoustic estimates are available for the year class yet thus the number-at-age 3 in 2017 was set to the geometrical mean for age- 3 over 1987-2011, which give 602 millions.

### 11.6.2 Prognosis results

SSB and biomass of age 3+ are estimated to be 325 kt and 417 kt , respectively, in the beginning of the fishing season 2015/16 (approximately the same as at spawning in July 2015). The results of the short term prediction from the final NFT-Adapt run (Table 11.6.1.2) indicate that fishing at $0.22\left(=\mathrm{F}_{0.1}\right.$; the stock is managed at $\left.\mathrm{F}=0.22 \sim \mathrm{~F}_{\text {MSY }}\right)$ would correspond to TAC in 2015/2016 of 71 kt and SSB at the spawning season in 2016 would be 327 kt .

The proposed composition of the catch in the season 2015/16 consists mainly of the 2008, 2009 and 2010 year classes, each contributing to $21-23 \%$ in total biomass of the catch (Figure 11.6.2.1). The small 2011 year class is only believed to give $4 \%$ of the catches.

### 11.7 Medium term predictions

Prognosis was made for the stock until the spawning season 2018 (Table 11.6.1.3) and the input data were the same as introduced above in section 11.6.1. The main features are that fishing at target $\mathrm{F}=0.22$ will give relatively constant catches and the SSB will increase slightly throughout the period.

### 11.8 Uncertainties in assessment and forecast

### 11.8.1 Assessment

There are several factors that could lead to uncertainty in the assessment. As done in the recent two assessments, additional natural mortality caused by the Ichthyophonus infection is only set for the first two years instead of all years since 2009. While this approach is considered to reduce the uncertainty in the assessment, quantification of the infection mortality needs to be improved in the future, and is ongoing currently. However, it should be noted that an exercise in last year's assessment (ICES 2014) showed that changing M for 2009 and 2010 changed the historical perception of the stocks size but had insignificant impacts on the assessment of the final year and the resulting advice.

The 2011 year class is small according to available measures. However, considering both the lower proportions of it in the winter catches taken in a small area compare to the bycatch during the summer over a wide area, and inadequate sample number in the survey southeast of Iceland (Breiðamerkurdjúp), the size of the year class might be underestimated somewhat in the assessment.

### 11.8.2 Forecast

The uncertainties mentioned above regarding the assessment apply also for the forecast, both regarding the mortality due to the Ichthyophonus infection and the size of the 2011 year class.
The number-at-age 3 in the beginning of 2015 used in the prognosis ( 477 millions) was predicted from a survey estimate of number at age 1 in 2013 but not as geometric mean from NFT-Adapt ( 602 millions), as done in previous years. This new approach had minor impact on the advice for 2015/16 (2 kt reduction in TAC), is considered to reduce the uncertainty in the forecast and is in accordance with the approach described in the Stock Annex.

### 11.8.3 Assessment quality

In previous years there has been concerns regarding the assessment because of retrospective patterns of the models. No assessment was provided in the 2005 due to data and model problems and in the two next consecutive years, ACFM rejected the assessment due to the retrospective pattern. In the assessments in 2007-2009 there was observed an improvement in the pattern from NFT-Adapt, while in 2010-2011, a retrospective pattern appeared again which was both related to the high M because of the Ichthyophonus infection but also due to new and more optimistic information about incoming year classes to the fishable stock (particularly the 2008 year class) and fishing pattern in recent year. The retrospective pattern in the last two and this year's assessment are less than seen for many years for SSB and F. That could be interpreted as an indication for improvements in the assessment quality in comparison to recent years.

### 11.9 Comparison with previous assessment and forecast

This year's assessment was conducted in the same way as in last year. In the current assessment, SSB in the beginning of the year 2014 is $5 \%$ lower ( 409 kt versus 430 kt ), size of the 2009 year class $6 \%$ lower, size of the 2010 year class 6 higher, and $\mathrm{WF}_{5-10}$ in 2013 is the same ( 0.16 in both cases), compare to the 2014 assessment.

### 11.10 Management plans and evaluations

The practice has been to manage fisheries on this stock at $\mathrm{F}=\mathrm{F}_{0.1}\left(=0.22=\mathrm{F}_{\mathrm{pa}}\right)$ for more than 20 years. However, no formal management strategy has been adopted.

### 11.11 Management consideration

Inspections indicate still a high prevalence of heart lesions related to Ichthyophonus hoferi in the herring stock, as in the last six years. This average prevalence of $15 \%$ for fish at age $3+$ in 2014/2015 (Óskarsson and Pálsson 2014) is believed to remain for some years but decrease as the infected year classes gets older and disappear from the stock. No indications are for significant new infection in the stock since 2010 and mortality due to the infection is considered insignificant during 2011-2015.

### 11.12 Ecosystem considerations

The reason for the outbreak of Ichthyophonus infection in the herring stock that was first observed in the autumn 2008 is not known but is probably the effect of interaction between environmental factors and distribution of the stock (Óskarsson et al. 2009). It includes that outbreak of Ichthyophonus spores in the environment, which infect the herring via oral intake (Jones and Dawe 2002), could be linked to the observed increased temperature off the southwest coast. Further researches on the causes of such an outbreak are needed and how the herring get infected, i.e. through intake of free floating spores or through zooplankton that contain spores. However, with respect of the impacts of the outbreak on the herring stock, significant additional mortality was estimated to have taken place only in the first two years (ICES 2014; Óskarsson and Pálsson 2013), despite a high prevalence of infection for now seven years. Thus, the infection that is still found in the stock (average prevalence of $15 \%$ for fish at age 3+; Óskarsson and Pálsson 2014) will decrease and disappear over some years as the fish gets older.

The WG does not have any information of direct evidence of environmental effects of the stock but emphasize that increased sea temperature is considered to have generally positive effects on the stock (Jakobsson and Stefansson, 1999; Óskarsson and Taggart 2010). It is manifest in observations of higher number of recruits per SSB during warm years and relatively high mean weight-at-age during recent years. Furthermore, the stock occupies colder water around Iceland than other herring stocks in the N -Atlantic and is therefore on edge of the distribution towards cold water, where warming will generally have a positive impacts on the stock development. The increased temperature in Icelandic waters since 1998 (MRI 2012), has therefore probably positive effects on the stock, possibly apart from the Ichthyophonus outbreak.

### 11.13 Regulations and their effects

The fishery of the Icelandic summer-spawning herring is limited to the period 1 September to 1 May each season, according to regulations set by the Icelandic Fishery Ministry (no. 770, 8. September 2006). Several other regulations are enforced by the

Ministry that effect the herring fishery. They involve protections of juveniles herring ( 27 cm and smaller) in the fishery where area closures are enforced if the proportion of juveniles exceeds $25 \%$ in number (no. 376, 8. October 1992). Another regulation deals with the quantity of bycatch allowed. Then there are regulations that prohibit use of pelagic trawls within the 12 nm fishing zone (no. 770, 8. September 2006), which is enforced to limit bycatch of juveniles of other fish species.

### 11.14 Changes in fishing technology and fishing patterns

There are no recent changes in fishing technology which may lead to different catch compositions. The fishing pattern in 2014/2015 was different than in last seven seasons. Instead of fishing near only in a small inshore area off the west coast in purse seine, the whole directed fishery took place in offshore areas west of Iceland by pelagic trawls. These changes are not considered to affect the selectivity of the fishery because the fishery is still targeting dense schools of overwintering herring in large fishing gears, getting huge catches in each haul and is by none means size selective.

The fishing pattern varies annually as noted in section 11.2 and it is related to variation in distribution of the different age classes of the stock. This variation can have consequences for the catch composition but it is impossible to provide a forecast about this variation.

### 11.15 Species interaction effects and ecosystem drivers

The WG have not dealt with this issue in a thoroughly and dedicated manner. However, some work has been done in this field in recent years in one way or another.
Regarding relevant researches on species interaction, the main work relates to the increasing amount of North East Atlantic mackerel (NEAM) feeding in Icelandic waters since 2007 (Astthorsson et al. 2012; ICES. 2013b). Surveys in the summers since 2010 indicate a high overlap in spatial and temporal distribution of NEAM and Icelandic summer-spawning herring (Nøttestad et al. 2014). Moreover, the diet composition of NEAM in Icelandic waters showed a clear overlap with those of the two herring stocks, i.e. Icelandic summer-spawning herring and Norwegian spring-spawning herring (Óskarsson et al. 2012b). Even if Copepoda was important diet group for all the three stocks its relative contribution to the total diet was apparently higher for NEAM than the two herring stocks. Considering former studies of herring diet, this finding was unexpected, and particularly how little the Copepoda contributed to the herring diet. This difference in the stomach content of NEAM and the two herring stocks indicated that there could be some difference in feeding ecology between them in Icelandic waters, where NEAM preferred Copepoda, or feed in the water column where they dominate over other prey groups, while the opposite would be for the herring and the prey Euphausiacea. Recent studies in the Nordic Seas have shown similar results (Langøy et al. 2012; Debes et al. 2012). The indication for difference in feeding ecology of the species is further supported by the fact that the body condition of the two herring stocks showed no clear decreasing trend since the invasion of NEAM started into Icelandic waters. On the contrary the mean weights-at-age of the summer spawners have been high, for example record high in the autumn 2014 (Figure 11.6.1.1), and the mean weight-at-length have also been relatively high in recent years (Óskarsson and Pálsson 2015). It should though be noted that comparison of the diet composition of herring in recent years to earlier studies, mainly on NSS herring, indicate that the herring might have shifted their feeding preference towards Euphausiacea instead of Copepoda. That
is possibly a consequence of increased competition for food with NEAM, where the herring is overwhelmed and shifts towards other preys.
The WG is not aware of documentations of strong signals from ecosystem or environmental variables that impact the herring stock and could possibly be a basis for implementing ecosystem drivers in the analytical basis for its advice. For example, recruitment in the stock has been positively, but weakly, linked to NAO winter index (North Atlantic Oscillation) and sea temperature (Óskarsson and Taggart 2010), while indices representing zooplankton abundance in the spring have not been found to impact the recruitment (Óskarsson and Taggart 2010) or body condition and growth rate of the adult part of the stock (Óskarsson 2008).

### 11.16 Comments on the PA reference points

The WG have not dealt with this issue recently.

### 11.17 Comments on the assessment

The assessment implies that the stock size has been rather unvarying in recent years following a period of depletion related to the Ichthyophonus infection. This is related to average size recruiting year classes entering the fishable stock, no infection mortality, and moderate fishing mortality. The assessment follows fairly well the pattern in the tuning series for recent years (Figure 11.3.2.5). However, small year class from 2011 entering the spawning stock in 2015 causes a decline in SSB even if it still above BPA. The size of the year class is possibly somewhat too pessimistic in this assessment but the downward trend is foreseen anyway.
This year's research on the Ichthyophonus infection in the stock supports the approach taken since 2013 (ICES 2013a; Óskarsson and Pálsson 2013) that additional natural mortality in the stock due to Ichthyophonus infection should only be applied for the first two years of the outburst. Further research is ongoing to quantify the infection mortality for these two years, but it is important to note that changing the mortality has mainly impacts on the historical perspective of the stock size and insignificant impacts on the present stock status.
In conclusion of the review group for NWWG 2011 (ICES 2011b), the suggestion was "to improve the assessment in order to get a better fitting for the years 2000-2005 and to work on the reference points". In this year's assessment, it was not dealt with these aspects specifically, but they still require attention. The years 2000-2005 fit still poorly to the tuning series and no satisfactory explanation exists for this pattern. The models recently used for the stock (NFT-Adapt, TSA and Coleraine (in Benchmark assessment in 2011; Gudmundsdottir 2011)) are not able to follow this trend in the tuning series. It should be noted that this same pattern was observed in the benchmark assessment in 2011 (Gudmundsdottir 2011) where input data were limited to the period before the infection so assumptions related to the natural mortality-infection are probably only responsible for this pattern to small degree if any. As mention above (section 11.3.2), the discrepancy could be related to the fact that during these years, the stock was overwintering in offshore areas off the east and west coast, compare to mainly easterly distribution before and overwintering in inshore areas there after (from ~2006-2013). These positive blocks could therefore reflect changes in catchability of the survey for these years. This must be kept in mind for the years to come since the stock has now started again to overwinter in offshore areas.

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Table 11.1.1.1. Icelandic summer-spawning herring. Acoustic estimates (in millions) in the winters 1973/74-2014/15 (age refers to the autumns). Years without surveys are marked with*.

| Year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | $15+$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1973/74 | 154.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 154 |
| 1974/75 | 5.000 | 137.000 | 19.000 | 21.000 | 2.000 | 2.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 186 |
| 1975/76 | 136.000 | 20.000 | 133.000 | 17.000 | 10.000 | 3.000 | 3.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 322 |
| 1976/77* |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1977/78 | 212.000 | 424.000 | 46.000 | 19.000 | 139.000 | 18.000 | 18.000 | 10.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 886 |
| 1978/79 | 158.000 | 334.000 | 215.000 | 49.000 | 20.000 | 111.000 | 30.000 | 30.000 | 20.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 967 |
| 1979/80 | 19.000 | 177.000 | 360.000 | 253.000 | 51.000 | 41.000 | 93.000 | 10.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 1004 |
| 1980/81 | 361.000 | 462.000 | 85.000 | 170.000 | 182.000 | 33.000 | 29.000 | 58.000 | 10.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 1390 |
| 1981/82 | 17.000 | 75.000 | 159.000 | 42.000 | 123.000 | 162.000 | 24.000 | 8.000 | 46.000 | 10.000 | 0.000 | 0.000 | 0.000 | 0.000 | 666 |
| 1982/83* |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1983/84 | 171.000 | 310.000 | 724.000 | 80.000 | 39.000 | 15.000 | 27.000 | 26.000 | 10.000 | 5.000 | 12.000 | 0.000 | 0.000 | 0.000 | 1419 |
| 1984/85 | 28.000 | 67.000 | 56.000 | 360.000 | 65.000 | 32.000 | 16.000 | 17.000 | 18.000 | 9.000 | 7.000 | 4.000 | 5.000 | 5.000 | 689 |
| 1985/86 | 652.000 | 208.000 | 110.000 | 86.000 | 425.000 | 67.000 | 41.000 | 17.000 | 27.000 | 26.000 | 16.000 | 6.000 | 6.000 | 1.000 | 1688 |
| 1986/87* |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1987/88 | 115.544 | 401.246 | 858.012 | 308.065 | 57.103 | 32.532 | 70.426 | 36.713 | 23.586 | 18.401 | 24.278 | 10.127 | 3.926 | 4.858 | 1965 |
| 1988/89 | 635.675 | 201.284 | 232.808 | 381.417 | 188.456 | 46.448 | 25.798 | 32.819 | 17.439 | 10.373 | 9.081 | 5.419 | 3.128 | 5.007 | 1795 |
| 1989/90 | 138.780 | 655.361 | 179.364 | 278.836 | 592.982 | 179.665 | 22.182 | 21.768 | 13.080 | 9.941 | 1.989 | 0.000 | 0.000 | 0.000 | 2094 |
| 1990/91 | 403.661 | 132.235 | 258.591 | 94.373 | 191.054 | 514.403 | 79.353 | 37.618 | 9.394 | 12.636 | 0.000 | 0.000 | 0.000 | 0.000 | 1733 |
| 1991/92 | 598.157 | 1049.990 | 354.521 | 319.866 | 89.825 | 138.333 | 256.921 | 21.290 | 9.866 | 0.000 | 9.327 | 0.000 | 0.000 | 1.494 | 2850 |
| 1992/93 | 267.862 | 830.608 | 729.556 | 158.778 | 130.781 | 54.156 | 96.330 | 96.649 | 24.542 | 1.130 | 1.130 | 3.390 | 0.000 | 0.000 | 2395 |
| 1993/94 | 302.075 | 505.279 | 882.868 | 496.297 | 66.963 | 58.295 | 106.172 | 48.874 | 36.201 | 0.000 | 4.224 | 18.080 | 0.000 | 0.000 | 2525 |
| 1994/95* |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1995/96 | 216.991 | 133.810 | 761.581 | 277.893 | 385.027 | 176.906 | 98.150 | 48.503 | 16.226 | 29.390 | 47.945 | 4.476 | 0.000 | 0.000 | 2197 |
| 1996/97 | 33.363 | 270.706 | 133.667 | 468.678 | 269.888 | 325.664 | 217.421 | 92.979 | 55.494 | 39.048 | 30.028 | 53.216 | 18.838 | 12.612 | 2022 |
| 1997/98 | 291.884 | 601.783 | 81.055 | 57.366 | 287.046 | 155.998 | 203.382 | 105.730 | 35.469 | 27.373 | 14.234 | 36.500 | 14.235 | 11.570 | 1924 |
| 1998/99 | 100.426 | 255.937 | 1081.504 | 103.344 | 51.786 | 135.246 | 70.514 | 101.626 | 53.935 | 17.414 | 13.636 | 2.642 | 4.209 | 8.775 | 2001 |
| 1999/00 | 516.153 | 839.491 | 239.064 | 605.858 | 88.214 | 43.353 | 165.716 | 89.916 | 121.345 | 77.600 | 21.542 | 3.740 | 11.149 | 0.000 | 2823 |


| Year $\backslash$ age | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| $2000 / 01$ | 190.281 | 966.960 | 1316.413 | 191.001 | 482.418 | 34.377 | 15.727 | 37.940 | 14.320 | 15.413 | 14.668 | 1.705 | 3.259 | 0.000 |
| $2001 / 02$ | 1047.643 | 287.004 | 217.441 | 260.497 | 161.049 | 345.852 | 62.451 | 57.105 | 38.405 | 46.044 | 38.114 | 21.062 | 3.663 | 0.000 |
| $2002 / 03$ | 1731.809 | 1919.368 | 553.149 | 205.656 | 262.362 | 153.037 | 276.199 | 99.206 | 47.621 | 55.126 | 18.798 | 24.419 | 24.112 | 1.377 |
| $2003 / 04$ | 1115.255 | 1434.976 | 2058.222 | 330.800 | 109.146 | 100.785 | 38.693 | 45.582 | 7.039 | 6.362 | 7.509 | 10.894 | 0.000 | 2.289 |
| $2004 / 05$ | 2417.128 | 713.730 | 1022.326 | 1046.657 | 171.326 | 62.429 | 44.313 | 10.947 | 23.942 | 12.669 | 0.000 | 1.948 | 11.088 | 0.000 |
| $2005 / 06$ | 469.532 | 443.877 | 344.983 | 818.738 | 1220.902 | 281.448 | 122.183 | 129.588 | 73.339 | 65.287 | 10.115 | 9.205 | 3.548 |  |
| $2006 / 07$ | 109.959 | 608.205 | 1059.597 | 410.145 | 424.525 | 693.423 | 95.997 | 123.748 | 48.773 | 0.955 | 0.000 | 0.000 | 0.000 | 0.480 |
| $2007 / 08$ | 90.231 | 456.773 | 289.260 | 541.585 | 309.443 | 402.889 | 702.708 | 221.626 | 244.772 | 13.997 | 22.113 | 68.105 | 10.136 | 2.800 |
| $2008 / 09$ | 149.466 | 196.127 | 416.862 | 288.156 | 457.659 | 266.975 | 225.747 | 168.960 | 29.922 | 26.281 | 17.790 | 9.881 | 0.974 | 3.195 |
| $2009 / 10$ | 151.066 | 315.941 | 490.653 | 554.818 | 271.445 | 327.275 | 149.143 | 83.875 | 156.920 | 36.666 | 13.649 | 8.507 | 1.458 | 5.590 |
| $2010 / 11$ | 106.178 | 280.582 | 228.857 | 304.885 | 296.254 | 138.686 | 301.285 | 60.997 | 141.323 | 97.412 | 37.006 | 0.000 | 4.019 | 0.000 |
| $2011 / 12$ | 704.863 | 977.323 | 434.876 | 313.742 | 272.140 | 239.320 | 154.581 | 175.088 | 84.582 | 92.435 | 89.376 | 17.638 | 6.808 | 4,989 |
| $2012 / 13$ | 178.500 | 781.083 | 631.421 | 166.627 | 126.961 | 142.044 | 110.084 | 97.000 | 74.340 | 69.473 | 43.376 | 38.450 | 7.458 | 0.773 |
| $2013 / 14$ | 15.919 | 314.865 | 218.715 | 344.981 | 151.631 | 132.767 | 120.756 | 118.377 | 89.555 | 74.602 | 48.695 | 44.637 | 31.096 | 11.598 |
| $2014 / 15$ | 152.422 | 90,269 | 330.084 | 260.919 | 259.079 | 187.905 | 111.955 | 91.629 | 37.855 | 76.680 | 30.366 | 10.619 | 22.799 | 10.108 |
|  |  |  |  |  |  |  |  | 1667 |  |  |  |  |  |  |

11.1.1.2. Overview of acoustic surveys conducted in the winter 2014/15 that contributed to the abundance estimates of the fishable stock and juveniles (age-1) of Icelandic summer-spawning herring.

|  | SURVEY <br> CODE | PERIOD | AREA | USED IN 2015 <br> ABUNDANCE |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | B7-2014 | 24 November - 6 <br> December 2014 | West, south and <br> southeast of <br> Iceland | The fishable stock <br> and juvenile <br> herring | Yes |
| 2 | D9-2014 | 13 November - 23 <br> November 2014 | Breiðafjörður <br> (adults) and then <br> fjords and bays <br> west, north and <br> east of Iceland <br> (juveniles) | The fishable stock <br> and juvenile <br> herring | Yes, the |

Table 11.1.2.1. Icelandic summers-spawning herring. Number of scales by ages and number of samples taken in the annual acoustic surveys in the seasons 1987/88-2014/15 (age refers to the former year, i.e. autumns). In 2000 seven samples were used from the fishery. No survey was conducted in 1994/95.

| Number of scales |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | Number of samples |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year $\backslash$ age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | Total | Total | West | East |
| 1987/88 | 11 | 59 | 246 | 156 | 37 | 28 | 58 | 33 | 22 | 16 | 23 | 10 | 5 | 8 | 712 | 8 | 1 | 7 |
| 1988/89 | 229 | 78 | 181 | 424 | 178 | 69 | 50 | 77 | 42 | 29 | 23 | 13 | 7 | 12 | 1412 | 18 | 5 | 10 |
| 1989/90 | 38 | 245 | 96 | 132 | 225 | 35 | 2 | 2 | 3 | 3 | 2 | 0 | 0 | 0 | 783 | 8 |  | 8 |
| 1990/91 | 418 | 229 | 303 | 90 | 131 | 257 | 28 | 6 | 3 | 8 | 0 | 0 | 0 | 0 | 1473 | 15 |  | 15 |
| 1991/92 | 414 | 439 | 127 | 127 | 33 | 48 | 84 | 5 | 3 | 0 | 2 | 0 | 0 | 1 | 1283 | 15 |  | 15 |
| 1992/93 | 122 | 513 | 289 | 68 | 73 | 28 | 38 | 34 | 6 | 2 | 2 | 6 | 0 | 0 | 1181 | 12 |  | 12 |
| 1993/94 | 63 | 285 | 343 | 129 | 13 | 15 | 7 | 14 | 11 | 0 | 1 | 3 | 0 | 0 | 884 | 9 |  | 9 |
| 1994/95* |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1995/96 | 183 | 90 | 471 | 162 | 209 | 107 | 38 | 18 | 8 | 14 | 18 | 2 | 0 | 0 | 1320 | 14 | 9 | 5 |
| 1996/97 | 24 | 150 | 88 | 351 | 141 | 137 | 87 | 32 | 15 | 10 | 7 | 14 | 4 | 2 | 1062 | 11 | 4 | 7 |
| 1997/98 | 101 | 249 | 50 | 36 | 159 | 95 | 122 | 62 | 21 | 13 | 8 | 15 | 8 | 5 | 944 | 14 | 7 | 7 |
| 1998/99 | 130 | 216 | 777 | 72 | 31 | 65 | 59 | 86 | 37 | 22 | 17 | 5 | 6 | 11 | 1534 | 17 | 10 | 7 |
| 1999/00 | 116 | 227 | 72 | 144 | 17 | 13 | 26 | 26 | 27 | 10 | 8 | 2 | 1 | 0 | 689 | 7 | 3 | 4 |
| 2000/01 | 116 | 249 | 332 | 87 | 166 | 10 | 7 | 21 | 8 | 14 | 11 | 3 | 1 | 0 | 1025 | 14 | 10 | 4 |
| 2001/02 | 61 | 56 | 130 | 114 | 62 | 136 | 25 | 24 | 17 | 21 | 17 | 10 | 3 | 0 | 676 | 9 | 4 | 5 |
| 2002/03 | 520 | 705 | 258 | 104 | 130 | 74 | 128 | 46 | 26 | 25 | 13 | 15 | 10 | 1 | 2055 | 22 | 12 | 10 |
| 2003/04 | 126 | 301 | 415 | 88 | 35 | 32 | 15 | 17 | 3 | 4 | 4 | 6 | 1 | 1 | 1048 | 13 | 8 | 5 |
| 2004/05 | 304 | 159 | 284 | 326 | 70 | 29 | 17 | 5 | 8 | 4 | 0 | 3 | 3 | 0 | 1212 | 13 | 4 | 9 |
| 2005/06 | 217 | 312 | 190 | 420 | 501 | 110 | 40 | 38 | 26 | 18 | 5 | 5 | 5 | 7 | 1894 | 22 | 14 | 8 |
| 2006/07 | 19 | 77 | 134 | 64 | 71 | 88 | 22 | 4 | 2 | 2 | 0 | 0 | 0 | 1 | 484 | 6 | 4 | 2 |
| 2007/08 | 58 | 288 | 180 | 264 | 85 | 80 | 104 | 19 | 15 | 2 | 2 | 6 | 1 | 3 | 1107 | 17 | 13 | 4 |
| 2008/09 | 274 | 208 | 213 | 136 | 204 | 123 | 125 | 97 | 18 | 13 | 9 | 7 | 4 | 17 | 1448 | 29 | 19 | 10 |
| 2009/10 | 104 | 100 | 105 | 116 | 60 | 74 | 34 | 19 | 36 | 8 | 3 | 4 | 2 | 2 | 667 | 17 | 10 | 7 |
| 2010/11 | 35 | 74 | 102 | 157 | 139 | 61 | 119 | 22 | 52 | 36 | 13 | 0 | 1 | 0 | 811 | 11 | 8 | 3 |
| 2011/12 | 229 | 330 | 134 | 115 | 100 | 106 | 74 | 87 | 45 | 48 | 51 | 10 | 3 | 3 | 1335 | 15 | 9 | 6 |
| 2012/13 $\ddagger$ | 42 | 266 | 554 | 273 | 220 | 252 | 198 | 165 | 126 | 114 | 69 | 61 | 12 | 2 | 2370 | 60 | $55 \ddagger$ | 5 |
| 2013/14 | 26 | 472 | 275 | 414 | 199 | 200 | 199 | 208 | 163 | 138 | 90 | 85 | 60 | 23 | 2552 | 45 | 37才 | 8 |
| 2014/15 | 83 | 50 | 96 | 71 | 72 | 53 | 32 | 26 | 11 | 22 | 8 | 3 | 6 | 4 | 534 | 10 | 8 | 2 |

*No survey
$\ddagger$ Samples in the western part were mainly from the commercial catch as there was impossible to secure a usable research survey samples from Kolgrafafjörður where most of the herring was observed.

Table 11.2.1. Icelandic summer spawners. Landings, catches, recommended TACs, and set National TACs in thousand tonnes.

| Year | Landings | Catches | Recom. TACs | Nat. <br> TACs | Year | Landings | Catches | Recom. <br> TACs | Nat. <br> TACs |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 0.31 | 0.31 |  |  | 2007/2008 | 158.9 | 158.9 | 130 | 150 |
| 1973 | 0.254 | 0.254 |  |  | 2008/2009 | 151.8 | 151.8 | 130 | 150 |
| 1974 | 1.275 | 1.275 |  |  | 2009/2010 | 46.3 | 46.3 | 40 | 47 |
| 1975 | 13.28 | 13.28 |  |  | 2010/2011 | 43.5 | 43.5 | 40 | 40 |
| 1976 | 17.168 | 17.168 |  |  | 2011/2012 $\ddagger$ | 49.4 | 49.4 | 40 | 45 |
| 1977 | 28.925 | 28.925 |  |  | 2012/2013 $\ddagger$ | 72.0 | 72.0 | 67 | 68.5 |
| 1978 | 37.333 | 37.333 |  |  | 2013/2014 $\ddagger$ | 72.0 | 72.0 | 87 | 87 |
| 1979 | 45.072 | 45.072 |  |  | 2014/2015 $\ddagger$ § | 95.0 | 95.0 | 83 | 83 |
| 1980 | 53.268 | 53.268 |  |  |  |  |  |  |  |
| 1981 | 39.544 | 39.544 |  |  |  |  |  |  |  |
| 1982 | 56.528 | 56.528 |  |  |  |  |  |  |  |
| 1983 | 58.867 | 58.867 |  |  |  |  |  |  |  |
| 1984 | 50.304 | 50.304 |  |  |  |  |  |  |  |
| 1985 | 49.368 | 49.368 | 50 | 50 |  |  |  |  |  |
| 1986 | 65.5 | 65.5 | 65 | 65 |  |  |  |  |  |
| 1987 | 75 | 75 | 70 | 73 |  |  |  |  |  |
| 1988 | 92.8 | 92.8 | 90 | 90 |  |  |  |  |  |
| 1989 | 97.3 | 101 | 90 | 90 |  |  |  |  |  |
| 1990/1991 | 101.6 | 105.1 | 80 | 110 |  |  |  |  |  |
| 1991/1992 | 98.5 | 109.5 | 80 | 110 |  |  |  |  |  |
| 1992/1993 | 106.7 | 108.5 | 90 | 110 |  |  |  |  |  |
| 1993/1994 | 101.5 | 102.7 | 90 | 100 |  |  |  |  |  |
| 1994/1995 | 132 | 134 | 120 | 120 |  |  |  |  |  |
| 1995/1996 | 125 | 125.9 | 110 | 110 |  |  |  |  |  |
| 1996/1997 | 95.9 | 95.9 | 100 | 100 |  |  |  |  |  |
| 1997/1998 | 64.7 | 64.7 | 100 | 100 |  |  |  |  |  |
| 1998/1999** | 87 | 87 | 90 | 70 |  |  |  |  |  |
| 1999/2000 | 92.9 | 92.9 | 100 | 100 |  |  |  |  |  |
| 2000/2001 | 100.3 | 100.3 | 110 | 110 |  |  |  |  |  |
| 2001/2002 | 95.7 | 95.7 | 125 | 125 |  |  |  |  |  |
| 2002/2003* | 96.1 | 96.1 | 105 | 105 |  |  |  |  |  |
| 2003/2004* | 130.7 | 130.7 | 110 | 110 |  |  |  |  |  |
| 2004/2005 | 114.2 | 114.2 | 110 | 110 |  |  |  |  |  |
| 2005/2006 | 103 | 103 | 110 | 110 |  |  |  |  |  |
| 2006/2007 | 135 | 135 | 130 | 130 |  |  |  |  |  |

*Summer fishery in 2002 and 2003 included
** TAC was decided 70 thous. tonnes but because of transfers from the previous quota year the national TAC became 90 thous. tonnes.
$\ddagger$ Landings and catches include bycatch of Icelandic summer-spawning herring in the mackerel and NSS herring fishery during the preceding summer (i.e. from the fishing season before in June-August).
${ }^{\text {§ }}$ The landings and catches in 2014/2015 consist of transfer of 7 kt from the year before and 5 kt from the year to come, which explains the discrepancy to the TACs.

Table 11.2.2.1. Overview of number of catch samples and fish measurements of Icelandic summerspawning herring during June 2014 to March 2015.

|  |  |  | October-March on the main <br> fishing grounds west of <br> Iceland (cells 3-5) |  | Total |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 11.2.2.2. Icelandic summer-spawning herring. Catch in numbers (millions) and total catch in weight (thous. tonnes) ( 1981 refers to season 1981/1982 etc).

| Year $\backslash$ age |  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 5+ | atch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1975 | 1.518 | 2.049 | 31.975 | 6.493 | 7.905 | 0.86 | 0.442 | 0.345 | 0.114 | 0.004 | 0.001 | 0.0 | . | 0.001 | 13.280 |
| 1976 | 0.614 | 9.848 | 3.908 | 34.144 | 7.009 | 5.481 | 1.04 | 0.438 | 0.296 | 0.134 | 0.092 | 0.0 | 0.00 | . 001 | 17.168 |
| 197 | 0.705 | 18.853 | 24.152 | 10.404 | 46.357 | 6.73 | 5.421 | 1.3 | 0.524 | 0.362 | 0.027 | 0. | 0.00 | 1 | 5 |
| 1978 | 2.634 | 22.551 | 50.995 | 13.846 | 8.738 | 39.492 | 7.2 | 6.35 | 1.616 | 0.926 | 0.4 | 0. | 0.02 | 1 | 37.333 |
| 1979 | 0.929 | 15.098 | 47.561 | 69.735 | 16.451 | 8.00 | 26. | 3. | 1.869 | 0.494 | 0.439 | 0. | 0.05 | 006 | 45.072 |
| 1980 | 3.147 | 14.347 | 20.761 | 60.727 | 65.328 | 11.5 | 9.2 | 9.4 | 1.796 | 1.464 | 0.698 | 0.0 | 0.1 | . 079 | 53.268 |
| 1981 | 2.283 | 4.629 | 16.771 | 12.126 | 36.871 | 41.91 | 7.2 | 4.86 | 13.416 | 1.032 | 0.884 | 0.7 | 0.1 | 0.062 | 39.544 |
| 1982 | 0.454 | 19.187 | 28.109 | 38.280 | 16.623 | 38.308 | 43.770 | 6.813 | 6.633 | 10.457 | 2.354 | 0.5 | 0.07 | 0.211 | 56.528 |
| 1983 | 1.475 | 22.499 | 151.718 | 30.285 | 21.599 | 8.667 | 14.0 | 13.713 | 3.728 | 2.381 | 3.436 | 0.5 | . 1 | 0.003 | 58.867 |
| 1984 | 0.421 | 18.015 | 32.2 | 141.354 | 17.043 | 7.113 | 3.9 | 4.113 | 4.517 | 1.828 | 0.202 | 0.2 | 0.260 | . 003 | 50.304 |
| 1985 | 0.112 | 12.872 | 24.659 | 21.656 | 85.210 | 11.903 | 5.74 | 2.336 | 4.363 | 4.053 | 2.773 |  | . 48 | 0.581 | 49.368 |
| 19 | 0. | 8 | 33.938 | 23 | 20 | 77 | 18 | 10.986 | 8. | 9.675 | 3 | 3.682 | 2.91 | 8 | 0 |
| 1987 | 0.029 | 3.144 | 44.590 | 60.285 | 20.622 | 19.75 | 46 | 15.232 | 13.963 |  | 3.216 |  | , | 2.280 | 75.439 |
| 1988 | 0.879 | 4.757 | 41.331 | 99.366 | 69.3 | 22.9 | 20 |  |  | . 2 | 7.378 |  | 4.80 | . 957 | 92.828 |
| 19 | 3.9 | 22.628 | 26.649 | 77 | 188.654 | 43.11 | 8.1 | 5.8 | 7.292 | 4.780 | 3.449 |  | 0.8 | 0.348 | 0 |
| 1990 | 12.567 | 14.884 | 56.995 | 35.593 | 79.7 | 157.2 | 30.2 | 8.18 | 4.372 | 3.379 | 1.786 | 0.7 | 0.4 | 0.5 | 7 |
| 1991 | 37.085 | 88.683 | 49.081 | 86.292 | 34.793 | 55.22 | 110. | 10.07 | 4.155 | 2.735 | 2.003 | 0.5 | . 3 | 0.41 | 9 |
| 1992 | 16.144 | 94.86 | 122.626 | 38.381 | 58.605 | 27.92 | 38. | 3.1 | 11.592 | 1.727 | 1.757 | 0.1 | 0.3 | .001 | 08.504 |
| 1993 | 2.467 | 51.153 | 177.78 | 92.68 | 20.79 | 28.56 | 13.3 | 19.617 | 15.266 | 4.254 | 0.797 | 0.2 | 0.00 | . 001 | 41 |
| 1994 | 5.738 | 134.61 | 113.29 | 142.8 | 87.207 | 24.9 | 20 | 16.301 | 5.6 | 14.68 | 2.936 | 1. | 0.2 | 0.1 | 3 |
| 1995 | 4.555 | 20.9 | 137 | 86.864 | 10 | 76.7 | 21 | 15.225 | 8. | 9.617 | 7.034 |  | 0.6 | . 23 | 1 |
| 1996 | 0.7 | 15 | 40 | 86 | 68 | 84. | 39 | 14.746 | 8. | 5.83 | 3.152 | 5. | 1.99 | 4 | 2 |
| 19 | 2.0 | 39 | 30 | 26 | 36 | 33.705 | 31 | 22.277 | 8. | 3.383 | 1.141 | 10.29 | 0.9 | 4 | 2 |
| 1998 | 23.655 | 45 | 175.529 | 22.69 | 8.613 | 40.898 |  | 32.046 | 14.647 | 2.122 | 2.754 |  | 1.0 | 011 | 86.998 |
| 1999 | 5.306 | 56.315 | 54.779 | 140.913 | 16.093 | 13.506 | 31.4 | 19.845 | 2.0 | 12.609 | 2.673 | 2.7 | 1.4 | 2.514 | 92.896 |
| 2000 | 17.286 | 57.28 | 136.278 | 49.289 | 76.614 | 11.54 | 8.29 | 6.367 | 9.874 | 1.332 | 6.744 | 2.97 | . 5 | 1.10 | 00.332 |
| 2001 | 27.486 | 42.304 | 86.422 | 93.597 | 30.336 | 54.49 | 10.37 | 8.762 | 12.244 | 9.907 | 8.259 | 6.08 | 1.4 | 1.259 | 95.675 |
| 2002 | 11.698 | 80.863 | 70.801 | 45.607 | 54.202 | 21.21 | 42.19 | 9.888 | 4.707 | 6.52 | 9.108 | 9.35 | 3.9 | . 697 | 96.128 |
| 2003 | 24.4 | 211 | 286.017 | 58.120 | 27.979 | 25.592 | 14.2 | 10.944 | 2.230 | 3.424 | 4.225 | 2.56 | 1.5 | . 370 | 30.741 |
| 2004 | 23.144 | 63.35 | 139.543 | 182.45 | 40.489 | 13.727 | 9.3 | 5.769 | 7.021 | 3.136 | 1.861 | 3.87 | 0.9 | . 8 | 37 |
| 2005 | 6.088 | 26.091 | 42.116 | 117. | 133.437 | 27.565 | 12.07 | 9.203 | 5.172 | 5.116 | 1.045 | 1.706 | 2. | . 75 | 03.043 |
| 2006 | 52.567 | 118 | 217.672 | 54.800 | 48.312 | 57.241 | 13.603 | 5.994 | 4.299 | 0.898 | 1.626 | 1.21 | 0.8 | . 93 | 35.303 |
| 2007 | 10.817 | 94.250 | 83.631 | 163.294 | 61.207 | 87.541 | 92.12 | 23.238 | 11.728 | 7.319 | 2.593 | 4.961 | 2.302 | . 420 | 158.917 |
| 2008 | 10.427 | 38.830 | 90.932 | 79.745 | 107.644 | 59.656 | 62.19 | 54.345 | 18.130 | 8.240 | 5.157 | 2.680 | 2.630 | . 178 | 51.780 |
| 2009 | 5.431 | 21.856 | 35.221 | 31.914 | 18.826 | 22.725 | 10.425 | 9.213 | 9.549 | 2.238 | 1.033 | 0.76 | 0.40 | 0.298 | 46.332 |
| 2010 | 1.476 | 8.843 | 22.674 | 29.492 | 24.293 | 14.419 | 17.40 | 10.045 | 7.576 | 8.896 | 1.764 | 1.10 | 0.672 | 0.555 | 43.533 |
| 2011 | 0.521 | 9.357 | 24.621 | 20.046 | 22.869 | 23.706 | 13.74 | 16.967 | 10.039 | 7.623 | 7.745 | 1.441 | 0.61 | 0.785 | 49.446 |
| 2012* | 0.403 | 17.827 | 89.432 | 51.257 | 43.079 | 51.224 | 41.846 | 34.653 | 27.21 | 4.946 | 5.473 | 13.575 | 2.595 | 0.253 | 25.369 |
| 2013 | 6.888 | 46.848 | 24.833 | 35.070 | 17.250 | 18.550 | 19.032 | 21.821 | 15.952 | 15.804 | 10.081 | 9.775 | 6.722 | 2.486 | 72.058 |
| 2014 | 0.000 | 3.537 | 53.241 | 50.609 | 70.044 | 34.393 | 22.08 | 22.138 | 13.298 | 17.761 | 7.974 | 4.461 | 2.862 | 1.746 | 94.975 |

[^2]Table 11.2.2.3. Icelandic summer-spawning herring. The mean weight (g) at age from the commercial catch ( 1981 refers to season 1981/1982 etc).

| Year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1975 | 110 | 179 | 241 | 291 | 319 | 339 | 365 | 364 | 407 | 389 | 430 | 416 | 416 | 416 |
| 1976 | 103 | 189 | 243 | 281 | 305 | 335 | 351 | 355 | 395 | 363 | 396 | 396 | 396 | 396 |
| 1977 | 84 | 157 | 217 | 261 | 285 | 313 | 326 | 347 | 364 | 362 | 358 | 355 | 400 | 420 |
| 1978 | 73 | 128 | 196 | 247 | 295 | 314 | 339 | 359 | 360 | 376 | 380 | 425 | 425 | 425 |
| 1979 | 75 | 145 | 182 | 231 | 285 | 316 | 334 | 350 | 367 | 368 | 371 | 350 | 350 | 450 |
| 1980 | 69 | 115 | 202 | 232 | 269 | 317 | 352 | 360 | 380 | 383 | 393 | 390 | 390 | 390 |
| 1981 | 61 | 141 | 190 | 246 | 269 | 298 | 330 | 356 | 368 | 405 | 382 | 400 | 400 | 400 |
| 1982 | 65 | 141 | 186 | 217 | 274 | 293 | 323 | 354 | 385 | 389 | 400 | 394 | 390 | 420 |
| 1983 | 59 | 132 | 180 | 218 | 260 | 309 | 329 | 356 | 370 | 407 | 437 | 459 | 430 | 472 |
| 1984 | 49 | 131 | 189 | 217 | 245 | 277 | 315 | 322 | 351 | 334 | 362 | 446 | 417 | 392 |
| 1985 | 53 | 146 | 219 | 266 | 285 | 315 | 335 | 365 | 388 | 400 | 453 | 469 | 433 | 447 |
| 1986 | 60 | 140 | 200 | 252 | 282 | 298 | 320 | 334 | 373 | 380 | 394 | 408 | 405 | 439 |
| 1987 | 60 | 168 | 200 | 240 | 278 | 304 | 325 | 339 | 356 | 378 | 400 | 404 | 424 | 430 |
| 1988 | 75 | 157 | 221 | 239 | 271 | 298 | 319 | 334 | 354 | 352 | 371 | 390 | 408 | 437 |
| 1989 | 63 | 130 | 206 | 246 | 261 | 290 | 331 | 338 | 352 | 369 | 389 | 380 | 434 | 409 |
| 1990 | 80 | 127 | 197 | 245 | 272 | 285 | 305 | 324 | 336 | 362 | 370 | 382 | 375 | 378 |
| 1991 | 74 | 135 | 188 | 232 | 267 | 289 | 304 | 323 | 340 | 352 | 369 | 402 | 406 | 388 |
| 1992 | 68 | 148 | 190 | 235 | 273 | 312 | 329 | 339 | 355 | 382 | 405 | 377 | 398 | 398 |
| 1993 | 66 | 145 | 211 | 246 | 292 | 324 | 350 | 362 | 376 | 386 | 419 | 389 | 389 | 389 |
| 1994 | 66 | 134 | 201 | 247 | 272 | 303 | 333 | 366 | 378 | 389 | 390 | 412 | 418 | 383 |
| 1995 | 68 | 130 | 183 | 240 | 277 | 298 | 325 | 358 | 378 | 397 | 409 | 431 | 430 | 467 |
| 1996 | 75 | 139 | 168 | 212 | 258 | 289 | 308 | 325 | 353 | 353 | 377 | 404 | 395 | 410 |
| 1997 | 63 | 131 | 191 | 233 | 269 | 300 | 324 | 341 | 355 | 362 | 367 | 393 | 398 | 411 |
| 1998 | 52 | 134 | 185 | 238 | 264 | 288 | 324 | 340 | 348 | 375 | 406 | 391 | 426 | 456 |
| 1999 | 74 | 137 | 204 | 233 | 268 | 294 | 311 | 339 | 353 | 362 | 378 | 385 | 411 | 422 |
| 2000 | 62 | 159 | 217 | 268 | 289 | 325 | 342 | 363 | 378 | 393 | 407 | 425 | 436 | 430 |
| 2001 | 74 | 139 | 214 | 244 | 286 | 296 | 324 | 347 | 354 | 385 | 403 | 421 | 421 | 433 |
| 2002 | 85 | 161 | 211 | 258 | 280 | 319 | 332 | 354 | 405 | 396 | 416 | 433 | 463 | 460 |
| 2003 | 72 | 156 | 189 | 229 | 260 | 283 | 309 | 336 | 336 | 369 | 394 | 378 | 412 | 423 |
| 2004 | 84 | 149 | 213 | 248 | 280 | 315 | 331 | 349 | 355 | 379 | 388 | 412 | 419 | 425 |
| 2005 | 106 | 170 | 224 | 262 | 275 | 298 | 324 | 335 | 335 | 356 | 372 | 394 | 405 | 413 |
| 2006 | 107 | 189 | 234 | 263 | 290 | 304 | 339 | 349 | 369 | 416 | 402 | 413 | 413 | 467 |
| 2007 | 93 | 158 | 221 | 245 | 261 | 277 | 287 | 311 | 339 | 334 | 346 | 356 | 384 | 390 |
| 2008 | 105 | 174 | 232 | 275 | 292 | 307 | 315 | 327 | 345 | 366 | 377 | 372 | 403 | 434 |
| 2009 | 113 | 190 | 237 | 274 | 304 | 318 | 326 | 335 | 342 | 360 | 372 | 394 | 409 | 421 |
| 2010 | 87 | 204 | 243 | 271 | 297 | 315 | 329 | 335 | 341 | 351 | 367 | 366 | 405 | 416 |
| 2011 | 97 | 187 | 245 | 283 | 309 | 328 | 343 | 352 | 356 | 364 | 375 | 386 | 378 | 432 |
| 2012 | 65 | 206 | 244 | 282 | 301 | 320 | 333 | 344 | 350 | 359 | 364 | 367 | 373 | 391 |
| 2013 | 95 | 182 | 238 | 271 | 300 | 322 | 337 | 349 | 360 | 365 | 362 | 375 | 377 | 394 |
| 2014 |  | 202 | 259 | 288 | 306 | 328 | 346 | 354 | 362 | 366 | 367 | 380 | 383 | 403 |

Table 11.2.2.4. Icelandic summer-spawning herring. Proportion mature at age ( 1981 refers to season 1981/1982 etc).

| Year\age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1975 | 0 | 0.27 | 0.97 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1976 | 0 | 0.13 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1977 | 0 | 0.02 | 0.87 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1978 | 0 | 0.04 | 0.78 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1979 | 0 | 0.07 | 0.65 | 0.98 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1980 | 0 | 0.05 | 0.92 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1981 | 0 | 0.03 | 0.65 | 0.99 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1982 | 0.02 | 0.05 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1983 | 0 | 0 | 0.64 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1984 | 0 | 0.01 | 0.82 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1985 | 0 | 0 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1986 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1987 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1988 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1989 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1990 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1991 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1992 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1993 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1994 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1995 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1996 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1997 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1998 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1999 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2000 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2001 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2002 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2003 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2004 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2005 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2006 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2007 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2008 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2009 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2010 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2011 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2012 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2013 | 0 | 0.2 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2014 | 0 | 0.1 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |

Table 11.3.2.1. Icelandic summer-spawning herring. Natural mortality at age where the deviation from the fixed $\mathrm{M}=0.1$ is due to the Ichthyophonus infection (1981 refers to season 1981/1982 etc).

| Year $\backslash$ age | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| $1987-2008$ | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 |
| 2009 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 | 0.49 |
| 2010 | 0.458 | 0.74 | 0.74 | 0.69 | 0.63 | 0.6 | 0.58 | 0.57 | 0.56 | 0.54 | 0.53 | 0.52 | 0.56 | 0.58 |
| $2011-2014$ | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 |

Table 11.3.2.2. Model settings and results of model parameters from the NFT-Adapt run in 2014 for Icelandic summer spawning herring.

VPA Version 3.3.0

| Model ID: Run2. Final last yeat + one year more data. |
| :--- |
| Input File: C:IUSERS\ASTAINFT\VPA\2015\RUN2\RUN2.DAT |
| Date of Run: 10-APR-2015 |

Levenburg-Marquardt Algorithm Completed 5 Iterations
Residual Sum of Squares $=50.2737$
Number of Residuals $=216$
Number of Parameters $=9$
Degrees of Freedom $=207$
Mean Squared Residual $=0.242868$
Standard Deviation $=0.492817$

Number of Years $=28$
Number of Ages = 11
First Year $=1987$
Youngest Age $=3$
Oldest True Age $=12$

Number of Survey Indices Available $=10$
Number of Survey Indices Used in Estimate $=8$

VPA Classic Method - Auto Estimated Q's

Stock Numbers Predicted in Terminal Year Plus One (2015)
Age Stock Predicted Std. Error CV
$4 \quad 88460.210 \quad 0.444726 \mathrm{E}+05 \quad 0.502741 \mathrm{E}+00$
$5 \quad 248124.780 \quad 0.960257 \mathrm{E}+05 \quad 0.387006 \mathrm{E}+00$
$6 \quad 229312.373 \quad 0.765804 \mathrm{E}+05 \quad 0.333957 \mathrm{E}+00$
$7 \quad 249087.051 \quad 0.793252 \mathrm{E}+05 \quad 0.318464 \mathrm{E}+00$
$8 \quad 98087.946 \quad 0.310708 \mathrm{E}+05 \quad 0.316764 \mathrm{E}+00$
$9 \quad 57259.569 \quad 0.186145 \mathrm{E}+05 \quad 0.325090 \mathrm{E}+00$
$10 \quad 40071.703 \quad 0.141892 \mathrm{E}+05 \quad 0.354095 \mathrm{E}+00$
$11 \quad 26023.962 \quad 0.956065 \mathrm{E}+04 \quad 0.367379 \mathrm{E}+00$
$12 \quad 11004.627 \quad 0.930594 \mathrm{E}+04 \quad 0.845639 \mathrm{E}+00$

Catchability Values for Each Survey Used in Estimate
INDEX Catchability Std. Error CV
$1 \quad 0.102045 \mathrm{E}+01 \quad 0.104443 \mathrm{E}+00 \quad 0.102350 \mathrm{E}+00$

| 2 | $0.125781 \mathrm{E}+01$ | $0.120301 \mathrm{E}+00$ | $0.956428 \mathrm{E}-01$ |
| :--- | :--- | :--- | :--- |
| 3 | $0.125774 \mathrm{E}+01$ | $0.875790 \mathrm{E}-01$ | $0.696322 \mathrm{E}-01$ |
| 4 | $0.134448 \mathrm{E}+01$ | $0.951709 \mathrm{E}-01$ | $0.707864 \mathrm{E}-01$ |
| 5 | $0.147853 \mathrm{E}+01$ | $0.117278 \mathrm{E}+00$ | $0.793206 \mathrm{E}-01$ |
| 6 | $0.169227 \mathrm{E}+01$ | $0.160301 \mathrm{E}+00$ | $0.947256 \mathrm{E}-01$ |
| 7 | $0.180443 \mathrm{E}+01$ | $0.212041 \mathrm{E}+00$ | $0.117511 \mathrm{E}+00$ |
| 8 | $0.171735 \mathrm{E}+01$ | $0.211353 \mathrm{E}+00$ | $0.123069 \mathrm{E}+00$ |

-- Non-Linear Least Squares Fit --

Maximum Marquadt Iterations $=100$
Scaled Gradient Tolerance $=6.055454 \mathrm{E}-05$
Scaled Step Tolerance $=1.000000 \mathrm{E}-18$
Relative Function Tolerance $=1.000000 \mathrm{E}-18$
Absolute Function Tolerance $=4.930381 \mathrm{E}-32$
Reported Machine Precision $=2.220446 \mathrm{E}-16$

VPA Method Options

- Catchability Values Estimated as an Analytic Function of N
- Catch Equation Used in Cohort Solution
- Plus Group Forward Calculation Method Used
- Arithmetic Average Used in F-Oldest Calculation
- F-Oldest Calculation in Years Prior to Terminal Year

Uses Fishing Mortality in Ages 8 to 11

- Calculation of Population of Age 3 In Year 2015
$=$ Geometric Mean of First Age Populations Year Range Applied = 1991 to 2011
- Survey Weight Factors Were Used

Stock Estimates:
Age 4
Age 5
Age 6
Age 7
Age 8
Age 9
Age 10
Age 11
Age 12

Full F in Terminal Year $\quad=0.3953$
F in Oldest True Age in Terminal Year $=0.5133$
Full F Calculated Using Classic Method

F in Oldest True Age in Terminal Year has been
Calculated in Same Manner as in All Other Years

|  | Input Partial | Calc P | ial Fis |  | Used In |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | ecruitment | Recruit | t Mo |  | Il F Comments |
| 3 | 0.500 | 0.040 | 0.0373 | NO | Stock Estimate in $\mathrm{T}+1$ |
| 4 | 0.800 | 0.200 | 0.1854 | NO | Stock Estimate in $\mathrm{T}+1$ |
| 5 | 1.000 | 0.206 | 0.1902 | YES | Stock Estimate in $\mathrm{T}+1$ |
| 6 | 1.000 | 0.256 | 0.2365 | YES | Stock Estimate in $\mathrm{T}+1$ |
| 7 | 1.000 | 0.310 | 0.2871 | YES | Stock Estimate in $\mathrm{T}+1$ |
| 8 | 1.000 | 0.337 | 0.3117 | YES | Stock Estimate in $\mathrm{T}+1$ |
| 9 | 1.000 | 0.455 | 0.4210 | YES | Stock Estimate in $\mathrm{T}+1$ |
| 10 | 1.000 | 0.427 | 0.3949 | YES | Stock Estimate in $\mathrm{T}+1$ |
| 11 | 1.000 | 1.000 | 0.9255 | YES | Stock Estimate in $\mathrm{T}+1$ |
| 12 | 1.000 | 0.555 | 0.5133 |  | F-Oldest |

Table 11.3.2.3. Icelandic summer spawners stock estimates (from NFT-Adapt in 2014) in numbers (thousands) by age (years) at January $1^{\text {st }}$ during 1987-2015.

| Age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | $13+$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year |  |  |  |  |  |  |  |  |  |  |  |  |
| 1987 | 529937 | 989084 | 300694 | 84605 | 69140 | 107466 | 42635 | 38034 | 26408 | 34264 | 34292 | 2256559 |
| 1988 | 271093 | 476517 | 852580 | 214871 | 56994 | 43836 | 53489 | 24150 | 21192 | 14258 | 36997 | 2065977 |
| 1989 | 447691 | 240772 | 391903 | 677068 | 128726 | 29843 | 20627 | 18028 | 10184 | 9485 | 26105 | 2000432 |
| 1990 | 301186 | 383582 | 192547 | 280755 | 433772 | 75629 | 19307 | 13073 | 9410 | 4695 | 26465 | 1740421 |
| 1991 | 842220 | 258379 | 292963 | 140442 | 178424 | 243589 | 39800 | 9723 | 7687 | 5314 | 24864 | 2043403 |
| 1992 | 1035186 | 677832 | 187209 | 183286 | 94077 | 109103 | 116243 | 26453 | 4866 | 4365 | 24199 | 2462818 |
| 1993 | 637975 | 846555 | 496934 | 132973 | 110308 | 58658 | 62328 | 54949 | 12969 | 2768 | 23679 | 2440096 |
| 1994 | 694264 | 528661 | 597305 | 361681 | 100580 | 72727 | 40446 | 37806 | 35246 | 7705 | 22930 | 2499351 |
| 1995 | 204385 | 500441 | 370860 | 404938 | 244544 | 67379 | 46557 | 21168 | 19355 | 17999 | 23159 | 1920786 |
| 1996 | 182845 | 164995 | 322700 | 253168 | 262915 | 148509 | 40725 | 27700 | 11069 | 8424 | 27594 | 1450645 |
| 1997 | 778 | 150 | 111059 | 210265 | 163721 | 15 | 96765 | 22883 | 17084 | 4503 | 22258 | 1734894 |
| 19 | 32 | 66 | 10737 | 7553 | 155382 | 11 | 113228 | 66 | 12627 | 12248 | 10153 | 65 |
| 1999 | 56 | 25 | 437111 | 7562 | 601 | 101 | 804 | 72071 | 46206 | 9411 | 13652 | 1711458 |
| 2000 | 40706 | 456945 | 174950 | 261987 | 53156 | 4162 | 62301 | 54008 | 44332 | 29854 | 12024 | 1598249 |
| 2001 | 495361 | 313934 | 284286 | 111572 | 164429 | 37143 | 29796 | 40852 | 39496 | 29366 | 26182 | 1572417 |
| 2002 | 1566444 | 408028 | 202118 | 168549 | 72190 | 97153 | 23772 | 18655 | 25359 | 26342 | 34059 | 2642668 |
| 2003 | 1180030 | 1340526 | 301992 | 139617 | 101148 | 45214 | 47984 | 12152 | 12415 | 16762 | 28045 | 3225884 |
| 2004 | 802061 | 866987 | 941576 | 218095 | 99779 | 67251 | 27451 | 33035 | 8879 | 7987 | 31310 | 3104411 |
| 2005 | 1159326 | 665538 | 652005 | 678821 | 158911 | 77249 | 51980 | 19365 | 23229 | 5064 | 27416 | 3518904 |
| 2006 | 863161 | 1024198 | 562181 | 478041 | 487590 | 117623 | 58434 | 38298 | 12618 | 16165 | 24057 | 3682366 |
| 2007 | 912519 | 668468 | 720197 | 456624 | 386656 | 386823 | 93509 | 47180 | 30570 | 10564 | 32005 | 3745114 |
| 2008 | 877954 | 736676 | 526781 | 499891 | 352810 | 264864 | 261386 | 62397 | 31536 | 20686 | 27825 | 3662807 |
| 2009 | 802971 | 757499 | 580211 | 400936 | 350186 | 262605 | 180662 | 184944 | 39272 | 20721 | 32849 | 3612856 |
| 2010 | 646270 | 475008 | 436837 | 330794 | 231071 | 196984 | 152817 | 103558 | 105922 | 22330 | 30882 | 2732474 |
| 2011 | 629267 | 302381 | 211388 | 198742 | 158833 | 116336 | 97522 | 79001 | 53528 | 55051 | 28057 | 1930107 |
| 2012 | 421941 | 560489 | 250212 | 172228 | 158108 | 121209 | 92207 | 72136 | 61949 | 41195 | 65143 | 2016818 |
| 2013 | 413974 | 364844 | 422246 | 177763 | 114981 | 94524 | 70034 | 50619 | 39502 | 32441 | 65992 | 1846920 |
| 2014 | 101479 | 330081 | 306527 | 348744 | 144459 | 86428 | 67468 | 42689 | 30685 | 20783 | 61516 | 1540860 |
| 2015* | 476547 | 88460 | 248125 | 229312 | 249087 | 98088 | 57260 | 40072 | 26024 | 11005 | 58307 | 1582287 |

[^3]Table 11.3.2.4. Estimated fishing mortality at age of Icelandic summer-spawning herring (from NFT-Adapt in 2015) by age (years) during 1987-2014 and weighed average F by numbers for age 510.

| Year\age | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3 +}$ | WF 5-10 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| $\mathbf{1 9 8 7}$ | 0.006 | 0.049 | 0.236 | 0.295 | 0.356 | 0.598 | 0.468 | 0.485 | 0.516 | 0.517 | 0.517 | 0.347 |
| 1988 | 0.019 | 0.096 | 0.131 | 0.412 | 0.547 | 0.654 | 0.988 | 0.764 | 0.704 | 0.777 | 0.506 | 0.266 |
| 1989 | 0.055 | 0.124 | 0.234 | 0.345 | 0.432 | 0.336 | 0.356 | 0.550 | 0.674 | 0.479 | 0.111 | 0.322 |
| 1990 | 0.053 | 0.170 | 0.216 | 0.353 | 0.477 | 0.542 | 0.586 | 0.431 | 0.471 | 0.508 | 0.071 | 0.400 |
| 1991 | 0.117 | 0.222 | 0.369 | 0.301 | 0.392 | 0.640 | 0.309 | 0.592 | 0.466 | 0.502 | 0.055 | 0.436 |
| 1992 | 0.101 | 0.210 | 0.242 | 0.408 | 0.372 | 0.460 | 0.649 | 0.613 | 0.464 | 0.547 | 0.023 | 0.415 |
| 1993 | 0.088 | 0.249 | 0.218 | 0.179 | 0.317 | 0.272 | 0.400 | 0.344 | 0.421 | 0.359 | 0.011 | 0.247 |
| 1994 | 0.227 | 0.255 | 0.289 | 0.291 | 0.301 | 0.346 | 0.548 | 0.570 | 0.572 | 0.509 | 0.090 | 0.311 |
| 1995 | 0.114 | 0.339 | 0.282 | 0.332 | 0.399 | 0.404 | 0.419 | 0.548 | 0.732 | 0.526 | 0.154 | 0.342 |
| 1996 | 0.096 | 0.296 | 0.328 | 0.336 | 0.411 | 0.328 | 0.476 | 0.383 | 0.799 | 0.497 | 0.349 | 0.358 |
| 1997 | 0.054 | 0.236 | 0.286 | 0.203 | 0.243 | 0.231 | 0.276 | 0.495 | 0.233 | 0.309 | 1.034 | 0.247 |
| 1998 | 0.159 | 0.323 | 0.251 | 0.128 | 0.323 | 0.267 | 0.352 | 0.263 | 0.194 | 0.269 | 0.572 | 0.276 |
| 1999 | 0.111 | 0.260 | 0.412 | 0.253 | 0.268 | 0.391 | 0.299 | 0.386 | 0.337 | 0.353 | 0.715 | 0.371 |
| 2000 | 0.160 | 0.375 | 0.350 | 0.366 | 0.259 | 0.234 | 0.322 | 0.213 | 0.312 | 0.270 | 0.670 | 0.327 |
| 2001 | 0.094 | 0.340 | 0.423 | 0.335 | 0.426 | 0.346 | 0.368 | 0.377 | 0.305 | 0.349 | 0.436 | 0.400 |
| 2002 | 0.056 | 0.201 | 0.270 | 0.411 | 0.368 | 0.605 | 0.571 | 0.307 | 0.314 | 0.449 | 0.876 | 0.392 |
| 2003 | 0.208 | 0.253 | 0.226 | 0.236 | 0.308 | 0.399 | 0.273 | 0.214 | 0.341 | 0.307 | 0.231 | 0.256 |
| 2004 | 0.087 | 0.185 | 0.227 | 0.217 | 0.156 | 0.158 | 0.249 | 0.252 | 0.462 | 0.280 | 0.255 | 0.218 |
| 2005 | 0.024 | 0.069 | 0.210 | 0.231 | 0.201 | 0.179 | 0.206 | 0.328 | 0.263 | 0.244 | 0.192 | 0.218 |
| 2006 | 0.156 | 0.252 | 0.108 | 0.112 | 0.132 | 0.129 | 0.114 | 0.125 | 0.078 | 0.112 | 0.140 | 0.118 |
| 2007 | 0.114 | 0.138 | 0.265 | 0.158 | 0.278 | 0.292 | 0.305 | 0.303 | 0.291 | 0.298 | 0.335 | 0.252 |
| 2008 | 0.048 | 0.139 | 0.173 | 0.256 | 0.195 | 0.283 | 0.246 | 0.363 | 0.320 | 0.303 | 0.280 | 0.229 |
| 2009 | 0.035 | 0.061 | 0.072 | 0.061 | 0.085 | 0.051 | 0.067 | 0.067 | 0.075 | 0.065 | 0.058 | 0.068 |
| 2010 | 0.020 | 0.070 | 0.098 | 0.104 | 0.086 | 0.123 | 0.090 | 0.100 | 0.114 | 0.107 | 0.102 | 0.100 |
| 2011 | 0.016 | 0.089 | 0.105 | 0.129 | 0.170 | 0.133 | 0.202 | 0.143 | 0.162 | 0.160 | 0.113 | 0.141 |
| 2012 | 0.045 | 0.183 | 0.242 | 0.304 | 0.414 | 0.449 | 0.500 | 0.502 | 0.547 | 0.499 | 0.307 | 0.364 |
| 2013 | 0.127 | 0.074 | 0.091 | 0.108 | 0.186 | 0.237 | 0.395 | 0.401 | 0.542 | 0.394 | 0.359 | 0.161 |
| 2014 | 0.037 | 0.185 | 0.190 | 0.237 | 0.287 | 0.312 | 0.421 | 0.395 | 0.926 | 0.513 | 0.168 | 0.255 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 11.3.2.5. Summary table from NFT-Adapt run in 2015 for Icelandic summer spawning herring.

| Year | Recruits, age 3 (millions) | Biomass age 3+(kt) | SSB (kt) | Landings age 3+ (kt) | Yield/SSB | WFage 5-10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 | 530 | 504 | 384 | 75 | 0.20 | 0.35 |
| 1988 | 271 | 495 | 423 | 93 | 0.22 | 0.27 |
| 1989 | 448 | 459 | 386 | 101 | 0.26 | 0.32 |
| 1990 | 301 | 410 | 350 | 104 | 0.30 | 0.40 |
| 1991 | 842 | 424 | 310 | 107 | 0.34 | 0.44 |
| 1992 | 1035 | 503 | 344 | 107 | 0.31 | 0.41 |
| 1993 | 638 | 547 | 425 | 103 | 0.24 | 0.25 |
| 1994 | 694 | 555 | 442 | 134 | 0.30 | 0.31 |
| 1995 | 204 | 464 | 408 | 125 | 0.31 | 0.34 |
| 1996 | 183 | 350 | 309 | 96 | 0.31 | 0.36 |
| 1997 | 778 | 371 | 271 | 65 | 0.24 | 0.25 |
| 1998 | 325 | 370 | 302 | 86 | 0.29 | 0.28 |
| 1999 | 564 | 378 | 294 | 93 | 0.31 | 0.37 |
| 2000 | 407 | 396 | 313 | 100 | 0.32 | 0.33 |
| 2001 | 495 | 361 | 281 | 94 | 0.33 | 0.40 |
| 2002 | 1566 | 545 | 314 | 96 | 0.31 | 0.39 |
| 2003 | 1180 | 628 | 421 | 129 | 0.31 | 0.26 |
| 2004 | 802 | 693 | 542 | 112 | 0.21 | 0.22 |
| 2005 | 1159 | 821 | 610 | 102 | 0.17 | 0.22 |
| 2006 | 863 | 934 | 730 | 130 | 0.18 | 0.12 |
| 2007 | 913 | 875 | 702 | 158 | 0.23 | 0.25 |
| 2008 | 878 | 944 | 757 | 151 | 0.20 | 0.23 |
| 2009 | 803 | 969 | 642 | 46 | 0.07 | 0.07 |
| 2010 | 646 | 745 | 453 | 43 | 0.10 | 0.10 |
| 2011 | 629 | 519 | 393 | 49 | 0.13 | 0.14 |
| 2012* | 422 | 555 | 442 | 125 | 0.28 | 0.38 |
| 2013 | 414 | 493 | 399 | 71 | 0.17 | 0.16 |
| 2014 | 101 | 460 | 410 | 95 | 0.23 | 0.26 |
| 2015§ | 477 | 438 | 342 |  |  |  |
| Mean | 640 | 559 | 428 | 100 | 0.25 | 0.28 |

* The mass mortality of 52 thousands tons in Kolgrafafjörður in the winter 2012/13 is included in the landings, yield/SSB, and WF.
${ }^{\S}$ Number at age 3 in 2015 is predicted from an survey index of number at age 1 in 2012 (see section 11.6.1)

Table 11.3.2.6. The residuals from survey observations and NFT-Adapt 2015 results for Icelandic summer spawning herring (no surveys in 1987 and 1995) on $1^{\text {st }}$ January.

| Year $\backslash$ Age | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 |  |  |  |  |  |  |  |  |
| 1988 | -0.192 | -0.223 | 0.131 | -0.294 | -0.689 | -0.251 | -0.171 | -0.434 |
| 1989 | -0.199 | -0.750 | -0.803 | 0.085 | 0.051 | -0.003 | 0.000 | 0.000 |
| 1990 | 0.515 | -0.300 | -0.236 | 0.017 | 0.474 | -0.387 | -0.001 | -0.002 |
| 1991 | -0.690 | -0.354 | -0.627 | -0.228 | 0.356 | 0.164 | 0.008 | -0.003 |
| 1992 | 0.417 | 0.409 | 0.328 | -0.342 | -0.154 | 0.267 | -0.807 | 0.002 |
| 1993 | -0.039 | 0.155 | -0.052 | -0.126 | -0.471 | -0.091 | -0.026 | 0.097 |
| 1994 | -0.065 | 0.161 | 0.087 | -0.703 | -0.612 | 0.439 | -0.333 | -0.514 |
| 1995 |  |  |  |  |  |  |  |  |
| 1996 | -0.230 | 0.629 | -0.136 | 0.085 | -0.216 | 0.354 | -0.030 | -0.158 |
| 1997 | 0.568 | -0.044 | 0.572 | 0.204 | 0.334 | 0.283 | 0.812 | 0.637 |
| 1998 | -0.123 | -0.511 | -0.504 | 0.318 | -0.096 | 0.060 | -0.125 | 0.492 |
| 1999 | 0.000 | 0.677 | 0.083 | -0.446 | -0.107 | -0.658 | -0.247 | -0.386 |
| 2000 | 0.588 | 0.083 | 0.609 | 0.211 | -0.350 | 0.452 | -0.080 | 0.466 |
| 2001 | 1.105 | 1.303 | 0.308 | 0.780 | -0.468 | -1.165 | -0.664 | -1.555 |
| 2002 | -0.372 | -0.156 | 0.206 | 0.506 | 0.879 | 0.440 | 0.529 | -0.126 |
| 2003 | 0.339 | 0.376 | 0.158 | 0.657 | 0.828 | 1.224 | 1.509 | 0.804 |
| 2004 | 0.484 | 0.553 | 0.187 | -0.206 | 0.014 | -0.183 | -0.268 | -0.008 |
| 2005 | 0.050 | 0.220 | 0.204 | -0.221 | -0.604 | -0.686 | -1.161 | -0.511 |
| 2006 | -0.856 | -0.718 | 0.309 | 0.622 | 0.481 | 0.212 | 0.629 | 1.219 |
| 2007 | -0.115 | 0.157 | -0.337 | -0.203 | 0.193 | -0.500 | 0.374 | -0.074 |
| 2008 | -0.498 | -0.829 | -0.149 | -0.427 | 0.028 | 0.463 | 0.677 | 1.508 |
| 2009 | -1.372 | -0.560 | -0.560 | -0.028 | -0.375 | -0.303 | -0.681 | -0.813 |
| 2010 | -0.428 | -0.113 | 0.288 | -0.135 | 0.117 | -0.550 | -0.801 | -0.148 |
| 2011 | -0.095 | -0.150 | 0.199 | 0.327 | -0.215 | 0.602 | -0.849 | 0.430 |
| 2012 | 0.536 | 0.323 | 0.370 | 0.247 | 0.289 | -0.009 | 0.296 | -0.229 |
| 2013 | 0.741 | 0.173 | -0.294 | -0.197 | 0.016 | -0.074 | 0.060 | 0.091 |
| 2014 | -0.067 | -0.567 | -0.240 | -0.248 | 0.038 | 0.056 | 0.430 | 0.530 |
| 2015 | 0.000 | 0.056 | -0.100 | -0.257 | 0.259 | 0.144 | 0.237 | -0.166 |

Table 11.6.1.1. The input data used for prognosis of the Icelandic summer-spawning herring in the 2015 assessment: the predicted weights, the selection pattern, $M$, proportion of $M$ before spawning, and the number-at-age derived from NFT-Adapt run.

| Age (year <br> class) | Mean <br> weights <br> (kg) | $\mathbf{M}$ | Maturity <br> ogive | Selection <br> pattern | Mortality prop. <br> before spawning | Number at <br> age |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  | F | M | Jan. 1st 2015 |
| $3(2012)$ | 0.164 | 0.10 | 0.200 | 0.355 | 0.000 | 0.500 | 476.5 |
| $4(2011)$ | 0.247 | 0.10 | 0.850 | 0.599 | 0.000 | 0.500 | 88.5 |
| $5(2010)$ | 0.292 | 0.10 | 1.000 | 1.000 | 0.000 | 0.500 | 248.1 |
| $6(2009)$ | 0.314 | 0.10 | 1.000 | 1.000 | 0.000 | 0.500 | 229.3 |
| $7(2008)$ | 0.328 | 0.10 | 1.000 | 1.000 | 0.000 | 0.500 | 249.1 |
| $8(2007)$ | 0.345 | 0.10 | 1.000 | 1.000 | 0.000 | 0.500 | 98.1 |
| $9(2006)$ | 0.359 | 0.10 | 1.000 | 1.000 | 0.000 | 0.500 | 57.3 |
| $10(2005)$ | 0.365 | 0.10 | 1.000 | 1.000 | 0.000 | 0.500 | 40.1 |
| $11(2004)$ | 0.372 | 0.10 | 1.000 | 1.000 | 0.000 | 0.500 | 26.0 |
| $12(2003)$ | 0.375 | 0.10 | 1.000 | 1.000 | 0.000 | 0.500 | 11.0 |
| $13+(2002+)$ | 0.375 | 0.10 | 1.000 | 1.000 | 0.000 | 0.500 | 58.3 |

Table 11.6.1.2. Icelandic summer-spawning herring. Short term prediction where the basis is: SSB(2015): 342 kt ; Biomass age 3+ (2015): 417 kt (at spawning time); Catch(2014/15): 95 kt ; WF 5 ${ }_{10}(2014)=0.255$. The fishery has been managed on basis of F $_{0.1}=0.22$ for over 20 years. SSB is in the spawning seasons, which is approximately the beginning of the subsequent fishing season. Catches and SSB are in thousands tons.

| Rationale | Landings | Basis | F | SSB | \%SSB <br> change ${ }^{1)}$ | \% TAC change ${ }^{2)}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | (2015/16) |  | (2015/2016) | 2016 |  |  |
| MSY <br> approach | 71 | Fmsy | 0.22 | 327 | -5 | -34 |
| F0.1 | 71 | F0.1=Fpa=0.22 | 0.22 | 327 | -5 | -34 |
| Zero catch | 0 | $\mathrm{F}=0$ | 0.00 | 496 | 31 |  |
| Status quo | 53 | F (2013) | 0.16 | 343 | 0 | -79 |
| Fmult | 9 | $0.1 \times$ (F0.1) | 0.02 | 383 | 11 | -1018 |
|  | 21 | $0.25 \times$ (F0.1) | 0.05 | 372 | 8 | -352 |
|  | 37 | $0.5 \times(\mathrm{F} 0.1)$ | 0.11 | 358 | 4 | -157 |
|  | 57 | $0.75 \times$ (F0.1) | 0.17 | 339 | -1 | -67 |
|  | 65 | $0.9 \times$ (F0.1) | 0.20 | 332 | -3 | -46 |
|  | 76 | $1.1 \times$ (F0.1) | 0.24 | 322 | -6 | -25 |
|  | 87 | $1.25 \times$ (F0.1) | 0.27 | 312 | -9 | -9 |
|  | 101 | $1.5 \times$ (F0.1) | 0.33 | 300 | -14 | 6 |

1) SSB 2016 relative to SSB 2015.
${ }^{2)}$ TAC 2015/16 relative to landings 2014/15.

Table 11.6.1.3. Icelandic summer-spawning herring. Medium term prediction where the basis is : $\operatorname{SSB}(2015): 342 \mathrm{kt}$; Catch(2014/15): $95 \mathrm{kt} ; \mathrm{WF}_{5-10}(2014)=0.255$. The prognosis of the Icelandic summer spawning herring for the next fishing season $(2015 / 2016)$ and the two subsequent seasons under five different options ( $F 0.1=0.22$, constant TAC of $60 \mathrm{kt}, 70 \mathrm{kt}, 80 \mathrm{kt}$ and 90 kt ) from the final NFTAdapt run in 2015. SSBs are in the spawning seasons, which is approximately the beginning of the subsequent fishing season.

| 2015/2016 | Spawning 2016 | $2016 / 2017$ |  | Spawning 2017 | $2017 / 2018$ | Spawning 2018 |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| TAC | F | Biomass 3+ | SSB (kt) | TAC | F | Biomass 3+ | SSB (kt) | TAC | F | Biomass 3+ | SSB (kt) |
| $(\mathrm{kt})$ | $(5-10)$ | $(\mathrm{kt})$ |  | $(\mathrm{kt})$ | $(5-10)$ | $(\mathrm{kt})$ |  | $(\mathrm{kt})$ | $(5-10)$ | $(\mathrm{kt})$ |  |
| 71 | 0.22 | 417 | 327 | 64 | 0.22 | 454 | 345 | 69 | 0.22 | 491 | 376 |
| 60 | 0.18 | 428 | 337 | 60 | 0.20 | 468 | 358 | 60 | 0.18 | 513 | 398 |
| 70 | 0.22 | 418 | 328 | 70 | 0.24 | 449 | 340 | 70 | 0.23 | 485 | 371 |
| 80 | 0.25 | 409 | 319 | 80 | 0.29 | 429 | 322 | 80 | 0.28 | 456 | 343 |
| 90 | 0.29 | 399 | 310 | 90 | 0.35 | 410 | 304 | 90 | 0.35 | 427 | 316 |



Figure 11.1.2.1. The locations of the areas that are referred to in the text.


Figure 11.1.2.2 Total biomass index for Icelandic summer-spawning herring from the acoustic surveys for ages $3+$ in the areas east and west of $18^{\circ} \mathrm{W}$ (except in 2011 and 2012 where fish outside of Breiðafjörður was set to the eastern part), combined over all areas and age 3-10 which are used in tuning of the analytical assessment. The years in the plot (1973-2014) refer to the autumn of the fishing seasons.


Figure 11.1.3.1. The prevalence of Ichthyophonus infection for the different age groups of Icelandic summer-spawning herring in Kolluáll in the winter 2014/2015.


Figure 11.1.3.2. The prevalence of Ichthyophonus infection for the different year classes of Icelandic summer-spawning herring in Breiðafjörður and Kolluáll as estimated in the autumns 2008 to 2014.


Figure 11.2.1. Icelandic summer spawning herring. Seasonal total landings (in thousand tonnes) during 1947-2014, referring to the autumns.


Figure 11.2.2. The distribution of the fishery (in tonnes) of Icelandic summer spawning herring during the fishing season 2014/15, including the bycatch in the mackerel fishery in June-September 2014.


Figure 11.2.1.1. Icelandic summer spawning herring. Proportion of the total catches of the Icelandic summer-spawning herring in 1975/76-2014/15 taken by different gears.
a)

b)


Figure 11.2.2.1. The distribution of samples, and number of fish taken for length measurements (to left) and age determination (to right) as indicated on graphs form the fishery of Icelandic summerspawning herring in June 2014 to March 2015.


Figure 11.2.2.2. Proportion of the different age groups of Icelandic summer-spawning herring to the total catches (biomass) as observed in 2014/2015 fishing season (June 2014-March 2015), predicted in the 2014 assessment (ICES 2014) for the 2014/2015 fishing season, and the summer catches in June-September 2014 in comparison to the age composition in the stock according to the acoustic measurements in the winter 2014/2015.


Figure 11.3.1.1. Icelandic summer-spawning herring. Catch curves by year classes 1984-2011. Grey lines correspond to $\mathrm{Z}=0.4$. Note that the mass mortality in Kolgrafafjörður is added to the catches in 2012.


Figure 11.3.1.2. Icelandic summer spawning herring. Catch curves from survey data by year classes 1984-2011. Grey lines correspond to $\mathrm{Z}=0.4$.
a)


Figure 11.3.1.3. The sum of total catch of each year class of Icelandic summer-spawning herring from 1971 to 2009 based on catch data from 1975-2014. The provided summary statistic is based on year classes from 1973 to 2002.


Figure 11.3.1.4. The cumulative total biomass in the catch (in proportion) of Icelandic summerspawning herring for different age group for the year classes 1978 to 1996.


Figure 11.3.2.1. Icelandic summer-spawning herring. The catchability ( $\pm 2 \mathrm{SE}$ ) and its CV for the acoustic surveys used in the final Adapt run in 2015 (1987-2014) compare to the assessment in 2014.


Figure 11.3.2.2. Icelandic summer-spawning herring. Comparisons of NFT-Adapt runs in 2015 and in 2014 concerning (a) biomass of age 3-12, (b) biomass of age 4-12, (c) number at age 3, and Nweighed $F$ for age $5-10$. Note that the mass mortality in Kolgrafafjörður in the winter 2012/13 is included in weighed $F$ for that year ( $\mathrm{WF}_{5-10}$ without the mass mortality was $\sim 0.22$ ) and the number at age 3 in 2014 from Adapt 2014 was geometric mean while number at age 3 in 2015 was predicted from juvenile index (see section 11.6.1).


Figure 11.3.2.3. Icelandic summer spawning herring. Residuals of NFT-Adapt run in 2015 from survey observations (moved to $1^{\text {st }}$ January). Filled bubbles are positive and open negative. Max bubble $=1.55$.


Figure 11.3.2.4. Icelandic summer spawning herring. Retrospective pattern from NFT-Adapt in 2015 in spawning stock biomass (the top panel), $\mathbf{N}$ weighted $\mathrm{F}_{5-10}$ (middle panel) and recruitment as number at age 3 (lowest panel).


Figure 11.3.2.5. Icelandic summer-spawning herring. Observed versus. predicted survey values from NFT-Adapt run in 2015 for ages 4-11 with respect to numbers (upper) and biomass (lower).


Figure 11.3.2.6. Icelandic summer-spawning herring. Comparison of number-at-age on Jan. 1st. 2014 from the final NFT model runs in 2014 and 2015 assessments.


Figure 11.6.1.1. Icelandic summer spawning herring. The mean weight at age for age groups 3 to 12 (+ group) as mean weight across 1987-2006, 2009-2013, observed in the winter 2014/2015, and finally predicted weights for the autumn 2015 from the weights in 2014, which was used in the stock prognosis.


Figure 11.6.1.2. Icelandic summer spawning herring. The selection pattern for age groups 3 to 12 (+ group) for the years 2012 to 2014, the average selection across these three years, and the selection used in the prognosis (three years average for age 3 and 4 , but fixed at 1.0 for age $4+$ ).

## Predicted herring landings in weights in 2015/2016



Figure 11.6.2.1. Icelandic summer spawning herring. The predicted biomass contribution of the different year classes to the catches in the fishing season 2015/2016.

## 12 Capelin in the Iceland-East Greenland-Jan Mayen area

## Summary

In May 2014 ICES advised on the basis of precautionary considerations that the initial quota be set at the $50 \%$ of the predicted quota, implying an initial quota of 225000 t . The quota was revised to $260000 t$ after an acoustic survey in October. The final TAC of $580000 t$ for the fishing season 2014/2015 was set after an acoustic survey in January 2015.

The total landings in the fishing season 2014/2015 amounted 517 thousand $t$ (preliminary data). Around 40000 t were taken in July 2014, 5000 t in November 2014 and the rest during the winter months January-March 2015.

The stock was benchmarked in January 2015. New methods for setting a final TAC and an initial quota were established. They will both apply for the next fishing season. Blim was defined during the meeting.

The acoustic index of 1 year old capelin from the acoustic survey is of an average size. On the basis of the new prediction model the initial quota in 2015/16 is 54 thousand t .

As the capelin increases its weight rapidly over the summer it is recommended that the fishery doesn't start until late autumn.

### 12.1 Stock description and management units

See stock-annex.

### 12.2 Scientific data

The capelin stock in Iceland-East Greenland-Jan Mayen area has been assessed by acoustics annually since 1978. The surveys have taken place in autumn (SeptemberDecember) and in winter (January-February). An overview is given in the stock annex.

### 12.2.1 Surveys in autumn 2014

In 2014 the survey took place in the period 16 September- 10 October.
The survey area extended along the shelf edge off East Greenland from about $73^{\circ} 30^{\prime}$ N to about $65^{\circ} 30^{\prime} \mathrm{N}$, as well as Denmark Strait and the continental slope west and north off Iceland (Figure 12.2.1). Weather conditions during the survey were adverse for the first few days and the survey had to be discontinued several times because of storms but no drift ice was encountered during the survey.
Capelin was westerly distributed (as in recent years) and similar to last year capelin was found almost up to $73^{\circ} \mathrm{N}$ (Figure 12.2.1). In 2010 the northern limit was $71^{\circ} \mathrm{N}$, in $201270^{\circ} 30^{\prime} \mathrm{N}$, but there was no survey in 2011 (due to strike). Immature capelin dominated south of $68^{\circ} 15^{\prime} \mathrm{N}$, but was now also found north of Iceland. Further north, along the Greenland shelf, older, maturing capelin predominated.
The combined index of young capelin (immature at age 1 and 2) in 2014 is at the average of years 1980-2013 (Tables 12.2.1-12.2.2 and Figure 12.2.2). The index of young capelin in the autumn surveys has been the basis for the starting quota for many years, see further chapter 12.7.

In this survey around 660 thousand $t$ of mature capelin was measured (Tables 12.2.1 12.2.3). On the basis of this estimate of the mature stock the Marine Research Institute
recommended a TAC of 260 thousand $t$ for the fishing season 2014/2015. This recommendation was in accordance with existing HCR and management plan between Iceland, Norway and Greenland.

### 12.2.2 Surveys in winter 2015

One survey was conducted in $5^{\text {th }}-30^{\text {th }}$ January 2014. The survey tracks are shown in Figure 12.2.3. Survey conditions were for the most part rather difficult. The survey had to be discontinued after 5 days in good recordings because of bad weather and could not be resumed again until 5 days later. The survey had to start again in the Denmark Strait because of uncertainty of migration speed and distance in the meantime. This part of the survey was done 17-29 January. The migration was not concentrated along the slope as usually but was spread over greater distances resulting in long survey transects and therefore rather slow progress of the survey. On the other hand the migration speed of the capelin appeared rather slow.

The main part of the mature capelin was situated from Denmark Strait in the west (from approx. $26^{\circ} 30^{\prime} \mathrm{W}$ ) outside the shelf to the Kolbeinsey ridge $\left(18^{\circ} \mathrm{W}\right)$. Further, there were some capelin found in a relatively small area east of Langanes in about $66^{\circ} 20^{\prime} \mathrm{N}$ and $12^{\circ} 30^{\prime} \mathrm{W}$. In the most western part of the distribution area mixing of mature and immature capelin was common but was less prominent further east. Age, length and weight disaggregated abundance is shown in table 12.2.4. The total biomass was estimated 1071 thousand tonnes where about 970 thousand tonnes were maturing to spawn.

On the basis of this estimate of the mature stock the Marine Research Institute recommended a TAC of 580 thousand $t$ for the fishing season 2014/2015. This recommendation was in accordance with existing HCR and management plan between Iceland, Norway and Greenland.

Since in year 2010 the autumn surveys have started in September, a month earlier than in previous years. The winter survey 2015 makes the third successful pair of autumn/winter surveys since 2010, as in 2011 there was no autumn survey in September due to a strike and in winter 2014 the survey did not give successful coverage of the capelin spawning stock. The 2014/15 pair gave $47 \%$ increase in SSB biomass following the winter survey estimate which is an extreme compared to the $14 \%$ and $9 \%$ differences of the 2010/11 and 2012/13 pairs respectively.

### 12.2.3 Fishery dependent data

A preliminary catch quota of 225 thousand $t$ was recommended for the 2014/2015 fishing season. A total of 517 thousand $t$ were caught from the final quota of 580 thousand $t$ in the 2014/2015 fishing season.

In July 2014, around 100 nmi southwest off Scoresby, the Norwegian and Danish vessels caught 30 thousand $t$ and 10 thousand $t$ of capelin respectively. Further, in autumn 2014, the Greenlandic fleet caught 5 thousand $t$ in Denmark strait, close to the midline between Iceland and Greenland.

The distribution of the winter catches, based on logbooks for the Icelandic fleet, is shown in Figure 12.3.1. The beginning of the 2015 winter fishery had a slow start where in weeks 2-4 the fleet was fishing north of the north-eastern Icelandic continental shelf from rather scattered observations of capelin, meanwhile a considerable abundance of capelin was observed further west by a research survey. During the following 3 weeks the fisheries took mainly place much closer to the shore off north Iceland with few casts
east of the Eastfjords. This is unusual development of the fishery as traditionally, at this time, the fleet would be fishing east and south of Iceland from the spawning migration of capelin. However, it is not until in week 8 that the fleet starts fishing from dense schools of capelin close to shore south-east off Iceland. This migration moved fast westward and was followed by the fishery south of the coast of west-Iceland in following two weeks and ended west of Breidafjordur in weeks 11-13. The landings during winter amounted 417 thousand t (preliminary information).
The total annual catch of capelin in the Iceland-East Greenland-Jan Mayen area since 1964 is given by weight, season, and fleet in Table 12.3.1 and Figure 12.3.2. The catches of the 2014/2015 fishing season amounted 517 thousand $t$, which is close to the average of catches since the beginning of the fishery.

Sampling from commercial catches is not considered to be adequate. From the fishery in winter 24 samples ( 2331 length measured and age read) were taken by the Icelandic fleet, 14 samples were taken by the Norwegian fleet, but no samples were taken by either the Faroes or the Greenlandic fleet. The Norwegian Fleet took 14 samples during the fishery in July and further 7 samples were taken in Iceland from Danish capelin fishery landed to Icelandic factories in summer 2014.

The total catches in numbers by age during the summer/autumn since 1985 are given in Table 12.3.2 and for the winter since 1986 in Table 12.3.3. Similar age distribution was observed in the catches 2015 as in the survey in January 2015.

Preliminary and final TAC as well as landings for the fishing seasons since 1992/93 are given in Table 12.3.4.

### 12.3 Growth

Seasonal growth pattern, with considerably increased growth rate during summer and autumn has been observed in this capelin stock in a study of the period 1979-1992. Where immature fish had slower growth during winter, the maturing fish had faster summer growth that continued throughout the winter until spawning in in March/April, followed by almost $100 \%$ spawning mortality (Vilhjalmsson, 1994). Further, examination of the growth of immature capelin at age 1 in autumn to mature at age 2 in autumn the year after in the period 1979-2013 showed on average almost 4 fold weight increase during one year (Gudmundsdottir and Thorsteinsson, WD in 2014). This considerable weight increase and seasonal pattern in growth the year before spawning should be taken into account when deciding the timing of the capelin fisheries.

### 12.4 Methods

Since early 1980s the stock has been managed according to an escapement strategy, leaving 400 thousand t to spawning. To predict the TAC for the next fishing season a model was developed early 1990s. These models were not endorsed by the benchmark working group WKSHORT 2009.

New methods for setting the final TAC and setting an initial quota were endorsed by the benchmark working group WKICE in 2015. The working frame is now a stochastic one, similar to the one for the Barents Sea capelin. The total TAC will be given so, that $\mathrm{p}(\mathrm{SSB}<\mathrm{Blim}=150000 \mathrm{t})<0.05$. This method will be applied for the first time in the next fishing season (2015/16).

The rationale behind the model of setting an initial quota is to set it low so that there is a minimum risk of it being higher than the final TAC. This method is used for the first
time during this working group, for setting the initial quota for the fishing season 2015/16.

See WKICE (ICES, 2015) and the Stock Annex for the capelin in the Iceland-East Green-land-Jan Mayen area.

### 12.5 Reference points

During WKICE (ICES, 2015) Blim of 150000 t was defined. No other reference points are defined for this stock.

### 12.6 State of the stock

The objective of the existing HCR for the stock is to leave 400 thousand $t$ for spawning (escapement strategy). It is estimated that 460 thousand t were left for spawning in spring 2015 (Table 12.7.1 and 12.7.2).

The acoustic indices of recruits are considered to be close to an average size.

### 12.7 Short term forecast

Figure 12.8.1 (taken from WKICE 2015, stock-annex) shows the relation between the index of immature capelin from an autumn survey and the estimated final advice for the fishing season starting a year later (see stock-annex). An initial quota is set on the basis of precautionary considerations (blue line). The risk of setting an initial quota exceeding the final TAC is considered very low.

By using the acoustic indices on immature capelin at age 1 and immature at age 2 from autumn 2014 (Table 12.2.2 and figure 12.2.2) the predicted TAC for the fishing season 2015/2016 is estimated as 313600 t and the initial quota as 53600 t (table 12.8.1).

## 12.8 (Medium term forecasts)

### 12.9 Uncertainties in assessment and forecast

The uncertainty of the acoustic estimates of the stock depends largely on the uncertainty of the echo abundance. The CV for the mature biomass was estimated by bootstrapping the SSB biomass values within the rectangles and strata used in the acoustic assessment. In the autumn 2014 survey the CV was estimated as 0.19 for the mature stock and 0.12 in the winter 2015 survey.

This fishing season there is a discrepancy between the estimate of the mature stock according to the acoustic surveys in autumn and winter, but these surveys are used to revise/set the TAC. The winter survey gives $47 \%$ higher mature stock, than the autumn survey, but both surveys are considered to have covered the mature stock. In the fishing seasons 2010/11 and 2012/13 the ratio between the corresponding pairs were $14 \%$ and $9 \%$ respectively.

The uncertainty when calculating the stock size by a deterministic method used so far is the value of M . A fixed value of $\mathrm{M}=0.035 /$ month has been used, but it may be too low according to WKSHORT 2009.

The acoustic survey in September-October 2014 had a good coverage of the spatial distribution of the capelin stock. The uncertainty of the immature capelin estimate is considered low with a CV of 0.18 . The numbers of immatures are used as input in the prediction model.

### 12.10Comparison with previous assessment and forecast

For the fishing season 2014/2015 an initial quota of 225 thousand $t$ were set. The final TAC was 580 thousand $t$. The landings were 517 thousand $t$.
According to the HCR 400 thousand $t$ shall be left for spawning. It is assumed that around 460 thousand t spawned in spring 2015 as not all quota was taken.

### 12.11 Management plans and evaluations

In June 1989 Greenland, Iceland and Norway signed a management plan. It has been revised several times since then, most recently in 2003.

The fishery is managed according to a two-step management plan which requires a minimum spawning-stock biomass of 400 thousand $t$ by the end of the fishing season. The first step in this plan is to set a preliminary TAC based on the results of an acoustic survey carried out to evaluate the abundance of immature (age 1 and age 2) part of the capelin stock about a year before it enters the fishable stock. The preliminary TAC is set at $2 / 3$ of the predicted TAC, calculated on the condition that 400 thousand $t$ of the SSB should be left for spawning. The second step is based on the results of another survey conducted during the fishing season for the same year classes. This result is used to revise the TAC and set the final TAC, still based on the condition that 400 thousand $t$ should be left for spawning.

ICES has not evaluated the management plan with respect to its conformity to the precautionary approach.

### 12.1 2 Management considerations

The fishing season for capelin has since 1975 started in the period from late June to July/August (when surveys on the juvenile part of the stock the year before has resulted in the setting of a preliminary catch quota). At that time the availability of plankton is at its highest and the fishable stock of capelin is feeding very actively over large areas north of Iceland between Greenland and Jan Mayen, increasing rapidly in size, weight and fatness. By late September/beginning of October this period of rapid growth is over. The growth is fasted the first two years, but the weight increase is most in the year before spawning.

Taken into account the large weight increase in the summer before spawning (section 12.4 ) it is clear that more catches are gained by the same effort if the fishery starts late autumn instead of summer. This is also supported by information for the Barents Sea capelin, but is has been shown for that stock that fishing during autumn would maximize the yield, but from the ecosystem point of view a winter fishery were preferable (Gjøsæter et.al., 2002). As the biology of these two capelin stocks is similar and their effect on the ecosystem too, this is considered to be valid for the Icelandic capelin too.

Seasonal variation of fat content is also observed. During the summer period, the fat content rises from approximately $5 \%$ to $20 \%$ in late autumn before spawning (Figure 12.1.1, Engilbertsson et. al. 2012). In the following fall and winter the fat content slowly declines, until the spawning migration begins in early January where the fat content drops drastically from about $15 \%$ to $5 \%$ in mid-April. Immature capelin has much lower fat content, usually less than 3-4\%.

During the summer and autumn, survey results show often overlap between juveniles and adult capelin. It has been reported by fishermen that while fishing with pelagic trawl in such areas, the catches are often poorer than expected from echo signals than
when fishing in areas where there is only adult capelin. That might indicate greater escapement of juveniles through meshes. The effect of such escapement on the fish is unknown.

### 12.13 Ecosystem considerations

Capelin is an important forage fish and its dynamics are expected to have implications on the productivity of their predators (see further in section 7.3). Capelin is the main single item in the diet of Icelandic cod and a key prey to several species of marine mammals and seabirds and also important as food for several other commercial fish species (see e.g. Vilhjálmsson, 2002).

### 12.14Regulations and their effects

Over the years the fishery has been closed during April - late June and the season has started in July/August or later, depending on the state of the stock.

Areas with high abundances of juvenile age 1 and 2 capelin (on the shelf region off $\mathrm{NW}-\mathrm{N}-\mathrm{and}$ NE-Iceland) have usually been closed to the summer and autumn fishery.

It is permissible to transfer catches from the purse-seine of one vessel to another vessel, in order to avoid slippage. However, if the catches are beyond the carrying capacity of the vessel and no other vessel is nearby, slippage is allowed. In recent years, reporting of such slippage has not been frequent. Industrial trawlers do not have the permission to slip capelin in order to harmonize catches to the processing.

In Icelandic waters, fishing with pelagic trawl is only allowed in limited area off the NE-coast (fishing in January) to protect capelin juveniles and to reduce the risk of affecting the spawning migration route.

A regulation calling for immediate, temporary area closures when high abundance of juveniles are measured in the catch (more than $20 \%$ of the catch composed of fish less than 14 cm ) is enforced in Icelandic waters, using on-board observers.

### 12.15Changes in fishing technology and fishing patterns

Variable amount of the catches have been taken with pelagic trawl through the fishing seasons. Total landings in 2015/16 amounted 517 kt (preliminary numbers) ( $88 \%$ purseseine, $12 \%$ pelagic trawl). Discards are considered negligible.

### 12.16Changes in the environment

Icelandic waters are characterized by highly variable hydrographical conditions, with temperatures and salinities depending on the strength of Atlantic inflow through the Denmark Strait and the variable flow of polar water from the north. Since 1996 the quarterly monitoring of environmental conditions of Icelandic waters shows a rise in sea temperatures north and east of Iceland, which probably also reaches farther north and northwest, as well as on the spawning grounds at South- and Southwest Iceland. It has been put forward in the 2000s that this temperature increase, may have led to displacements of the juvenile part and the spawning areas of the capelin stock (Vilhjálmsson, 2007). The acoustic surveys in autumn 2010, 2012-2014 partly confirmed this, but major part of the spawning still takes place on the usual grounds.

More detailed environmental description is in section 7.3.

### 12.17 Recommendations

A considerable difference was evident between SSB estimates from autumn and winter acoustic surveys during this fishing season, meanwhile both surveys are believed to have a good coverage of the maturing stock component. Hence, it is recommended that it should be evaluated if and then how the SSB estimates from autumn and winter surveys should be weighted in the future when final TAC is defined.

Studies of optimal harvesting of capelin should be conducted and presented to the NWWG. These estimates should take account of growth, mortality and gear selection in relation to the timing of the fishery (as recommended by WKICE 2015).

The NWWG should initiate a review of the role of capelin in the Icelandic Sea ecosystem and in particular whether the population size and growth of capelin predators shows a response to changes in capelin abundance (as recommended by WKICE 2015).

### 12.18Special request

ICES got the following request from the Coastal States of the capelin in the IcelandEast Greenland-Jan Mayen area:
"Capelin in East Greenland, Iceland and Jan Mayen area was benchmarked by ICES in January this year and consequently the basis for the TAC advice was changed to a new PA approach. I assume therefore that ICES in May will provide advice in accordance to the new approach. As you are aware, the coastal states of this stock, namely Greenland, Iceland and Norway, manage the stock according to an agreement where the main element are an escapement target and an initial TAC of $2 / 3$ of the final TAC. The coastal states appreciate the work conducted at the benchmark and aim to follow this new approach, however, the present management agreement is likely not revised according to the benchmark before later this year. On behalf of the coastal states, Greenland, Iceland and Norway, I therefore kindly request ICES in its May 2015 advice to provide a catch option according to the present management agreement."

## Short term by the old rule

WKSHORT (ICES, 2009) rejected the model that had been used since 1991 for predicting a quota for the next fishing season. The main reasoning was the value of $\mathrm{M}=0.035$ per month, which is considered too low. Another issue identified by the WKSHORT was the intercept. The model predicts a quota for a fishing season starting in June in year Y , based on indices from autumn survey in year $\mathrm{Y}-1$ and from winter survey in year Y , with the restriction that 400000 t have to spawn 1 April in year $\mathrm{Y}+1$ (Gudmundsdottir and Vilhjálmsson, 2002). This means that the spawning stock biomass has to be estimated around 18 months later than the estimates of the juveniles are obtained. An initial quota for the fishing season was then set as $2 / 3$ rd of the predicted quota.

Since 2011 additionally two other methods have been used to explore the SSB one and a half year after the survey takes place in autumn.

Based on the results from the models an initial quota could only be set twice; for the 2011/12 and 2014/15 fishing seasons. Then more caution was also applied, namely $1 / 2$ instead of $2 / 3$ of the predicted quota was set as an initial quota.

Like in the last years the three methods will be explored.

## Projection model.

Based on the results from the acoustic surveys in autumn 2014 and from winter 2015 the new data was added to the time-series. Input data is given in table 12.19.1 and
12.19.2. Data from the acoustic surveys in 2004, 2005, 2007-2009 and 2011 were not used as they are not considered to be valid due to different reasons, but the main reason being lack of coverage due to drift ice in E-Greenland waters.

The updated regressions are shown in Figure 12.19.1. By applying this method a zero index for age 1 gives 19.44 million at age 2 (intercept of 19.44), resulting in a SSB of 280 thousand tonnes at spawning time one and a half year after the survey is conducted and using long term average mean weight at age.

The residuals from the regression indicate that the regression performs better at medium and high values than at low values, as then the observed values lie all below the fitted line, which makes it questionable if the regression can be used at low values. This year the index of 1 year old is at the lower range of the zone where the regression is considered to behave better.

In the last two years the predictor in regression 2 was in the lower range, contributing to the SSB only with 30 thousand tonnes in last year and nothing the year before. This time the predictor is higher than in the two preceding years, however well below the average.

This method estimates the SSB to be almost 920 thousand $t$ in spring 2016 if no fishery takes place (Table 12.19.3). The contribution of the younger year class is 784 thousand $t$ and the older year class 132 thousand t . Based on this model the predicted TAC for the fishing season 2015/2016 is 519 thousand $t$. According to the management plan $2 / 3 \mathrm{rd}$ of it is allocated as an initial quota, that is 346 thousand t . If the same ratio is used as in the last two times this method was applied, then the initial quota would be set as $1 / 2$ of the predicted quota, resulting in an initial quota of 260 thousand $t$.

The predictor in regression 2 describes the total number of two years old in the stock in autumn $Y$ (2014) (the number of mature and immature at age 2). The immatures are taken as the measured number of immature capelin at age 2 in the acoustic autumn survey. The mature part is however derived by projecting the number mature at age 3 in the winter survey backwards in time to autumn and adding the catches of the same year class taken before the survey. There is then a relation between the total number of capelin at age 2 in autumn to the mature number at age 3, a year later. By using this relation then higher number at age 2 gives higher number mature at age 3 (even though they are already dead). With the negative intercept, low values provide "negatively" to the SSB!

This year the biomass derived from the winter survey was $47 \%$ higher than the autumn survey. Theoretically the autumn survey could have been used for setting a final the TAC. Therefore it was decided to explore how the model would estimate the SSB in 2016 if only values from the autumn survey were applied. Applying these values the SSB in 2016 is estimated to be 800 thousand $t$, whereas the contribution of the younger year class is 776 thousand $t$ and the older year class 23 thousand $t$. ( 0 index gives $\sim 260$ thousand t . Based on this model the predicted TAC for the fishing season 2015/2016 is 401 thousand $t$ implying an initial quota of 267 thousand $t$ ( $2 / 3$ rd of predicted catch) or 200 thousand t ( $1 / 2$ of predicted catch).

So depending on whether the results from the acoustic surveys in autumn or in winter are used, the estimate of the SSB differ of 116 thousand $t$, thereof stem 109 thousand $t$ from the older year class. The use of the regression for the older age group must be considered questionable. Using only the contributions from the younger age group the SSB in 2016 would be around 780 thousand t .

## Zero intercept regressions.

An alternative procedure to the one above, is to make the regressions go through the origin (figure 12.18.3). For low indices the residuals have similar pattern as in the 'projection model', but in the opposite direction. Bigger changes are observed in the residuals for higher indices. This 'zero intercept regression' causes less problems at low index values. For the older age group the regression gives slightly higher values that in the 'projection model'. This method estimates the SSB in spring 2016 to be 863 thousand $t$ if no fishery takes place ( 675 thousand $t$ for age 2 and 188 thousand $t$ for age 3). (Table 12.19.3). The high number at age 3 (already dead) in the survey in winter 2015 mean high number at age 2 last autumn, which again mean higher number of mature fish at age 3 next autumn!

## Simple forward projection.

By using a standard ICES procedure (a simple forward projection) it is assumed that the indices are absolute. A natural mortality of $M=0.035 /$ month is used, but this assumption of M was, among other things, the reasoning for not endorsing the projection model by WKSHORT 2009 as it was considered too low. This method doesn't involve specific issues when index is low. By assuming no fishery on juveniles in 2014/15 and 2015/16 this method gives a SSB of 626 thousand t in spring 2016 (Table 12.19.3).

## Summary.

The three methods above estimate the SSB in 2016 to be in the range from 626 - 920 thousand t if no fishery takes place (Table 12.19.3). According to the management plan 400 thousand $t$ have to be left for spawning. This means that approximately 226-520 thousand tonnes are left for fishing. An initial quota is then calculated on that basis. Due to uncertainties in the projection model and the approach for calculating the TAC, it has been suggested that the fishery should not be opened until after an acoustic survey (in autumn/winter) if the predicted TAC $<500$ thousand t (Gudmundsdottir and Vilhjálmsson, 2002).

### 12.19References

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Table 12.2.1 Capelin. Acoustic assessment of capelin in the Iceland/Greenland/Jan Mayen area, by r/v Arni Fridriksson 16/9-10/10 2014 (Numbers in millions, biomass in tonnes).

| N Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{array}{\|l\|} \hline \text { Length } \\ 7.5 \\ \hline \end{array}$ | W | 1 | 2 | 3 | 4 | Numbers | Biomass | Total |
|  | 1.3 | 478.1 | 0.0 | 0.0 | 0.0 | 478.1 | 643.9 |  |
| 8 | 1.5 | 3050.3 | 0.0 | 0.0 | 0.0 | 3050.3 | 4706.8 |  |
| 8.5 | 2.0 | 4301.5 | 0.0 | 0.0 | 0.0 | 4301.5 | 8458.4 |  |
| 9 | 2.4 | 8007.4 | 0.0 | 0.0 | 0.0 | 8007.4 | 19208.3 |  |
| 9.5 | 2.9 | 7183.0 | 0.0 | 0.0 | 0.0 | 7183.0 | 20993.4 |  |
| 10 | 3.5 | 7879.0 | 0.0 | 0.0 | 0.0 | 7879.0 | 27671.7 |  |
| 10.5 | 4.3 | 6334.5 | 0.0 | 0.0 | 0.0 | 6334.5 | 26997.1 |  |
| 11 | 5.3 | 4567.7 | 0.0 | 0.0 | 0.0 | 4567.7 | 24010.5 |  |
| 11.5 | 6.1 | 5106.0 | 0.0 | 0.0 | 0.0 | 5106.0 | 30957.0 |  |
| 12 | 7.0 | 5028.4 | 0.0 | 0.0 | 0.0 | 5028.4 | 35211.9 |  |
| 12.5 | 8.0 | 3721.8 | 116.1 | 0.0 | 0.0 | 3837.9 | 30773.8 |  |
| 13 | 9.4 | 1930.6 | 617.4 | 0.0 | 0.0 | 2548.0 | 23915.3 |  |
| 13.5 | 11.3 | 432.9 | 1806.2 | 107.4 | 0.0 | 2346.5 | 26450.2 |  |
| 14 | 12.9 | 0.0 | 3128.8 | 294.3 | 0.0 | 3423.1 | 44099.3 |  |
| 14.5 | 14.7 | 22.2 | 4830.1 | 722.5 | 0.0 | 5574.7 | 81888.6 |  |
| 15 | 16.6 | 0.0 | 5492.6 | 938.0 | 9.7 | 6440.2 | 107001.5 |  |
| 15.5 | 18.6 | 0.0 | 5512.1 | 1453.0 | 0.0 | 6965.1 | 129495.1 |  |
| 16 | 21.5 | 0.0 | 3767.6 | 1556.6 | 11.7 | 5335.9 | 114518.3 |  |
| 16.5 | 23.5 | 0.0 | 2785.9 | 1004.4 | 29.6 | 3819.9 | 89714.4 |  |
| 17 | 26.3 | 0.0 | 1165.5 | 996.8 | 19.9 | 2182.3 | 57449.0 |  |
| 17.5 | 29.6 | 0.0 | 558.6 | 435.4 | 8.3 | 1002.2 | 29650.7 |  |
| 18 | 33.4 | 0.0 | 75.0 | 190.3 | 11.7 | 276.9 | 9258.8 |  |
| 18.5 | 35.8 | 0.0 | 11.7 | 97.9 | 0.0 | 109.6 | 3926.7 |  |
| 19 | 39.2 | 0.0 | 11.7 | 8.3 | 0.0 | 19.9 | 781.4 |  |
| 19.5 | 42.6 | 0.0 | 0.0 | 8.3 | 0.0 | 8.3 | 351.6 |  |
|  | Total ${ }^{*}$ 10-6 | 58043.3 | 29879.2 | 7813.1 | 90.9 | 95826.4 |  |  |
|  | Total B (tt) | 252027.3 | 528961.6 | 164877.4 | 2267.3 |  | 948133.7 |  |
|  | Average W (gr) | 4.3 | 17.7 | 21.1 | 25.0 |  |  | 9.9 |
|  | Average L (cm) | 10.3 | 14.8 | 14.2 | 12.8 |  |  | 12.0 |
|  | \%N | 60.6 | 31.2 | 8.2 | 0.1 |  |  | 100.0 |

Immature

| Age | 1 | 2 | 3 |  |
| :--- | ---: | ---: | ---: | ---: |
| Total |  |  |  |  |
| Total $\mathrm{N}^{*} 10-6$ | 57018.4 | 3345.7 | 209.6 | 60573.7 |
| Total B (tt) | 241896.8 | 42454.8 | 3479.9 | 287831.5 |
| Average W (gr) | 4.2 | 12.7 | 16.6 | 4.8 |
| Average L (cm) | 10.2 | 14.1 | 15.0 | 10.5 |
| $\% \mathrm{~N}$ | 94.1 | 5.5 | 0.3 | 100.0 |

Mature

| Age | 1 | 2 | 3 | 4 | Total |
| :--- | ---: | ---: | ---: | ---: | ---: |
|  | Total $\mathrm{N}^{*} 10-6$ | 1024.9 | 26533.5 | 7603.5 | 90.9 |
| Total B (tt) | 10130.5 | 486506.8 | 161397.1 | 2267.3 | 660301.7 |
| Average W (gr) | 9.9 | 18.3 | 21.2 | 25.0 | 18.7 |
| Average L (cm) | 12.9 | 15.3 | 15.9 | 16.7 | 15.4 |
| $\% \mathrm{~N}$ | 2.9 | 75.3 | 21.6 | 0.3 | 100.0 |

Table 12.2.2. Icelandic Capelin. Abundance of age-classes in numbers (109) measured in acoustic surveys in autumn.

| Year | Mon | Day | age 1 | age 1 | age2 | age2mat | age3 | age3mat | age4 | age5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | imm | mat | imm | mat | imm | Mat | mat | mat |
| 1978 | 10 | 16 |  |  |  | 60.0 |  | 13.9 | 0.4 |  |
| 1979 | 10 | 14 | 10.0 |  |  | 49.7 |  | 9.1 | 0.4 |  |
| 1980 | 10 | 11 | 23.5 |  |  | 19.5 |  | 4.8 |  |  |
| 1981 | 11 | 26 | 21.0 |  | 1.1 | 11.9 |  | 0.6 |  |  |
| 1982 | 10 | 2 | 68.0 |  | 1.7 | 15.0 |  | 1.6 |  |  |
| 1983 | 10 | 3 | 44.1 |  | 8.2 | 58.6 |  | 5.6 | 0.1 |  |
| 1984 | 11 | 1 | 73.8 |  | 4.6 | 31.9 |  | 10.3 | 0.3 |  |
| 1985 | 10 | 8 | 33.8 |  | 12.6 | 43.7 |  | 14.4 | 0.4 | 0.1 |
| 1986 | 10 | 4 | 58.6 |  | 1.4 | 19.9 |  | 29.8 | 0.3 |  |
| 1987 | 11 | 18 | 21.3 |  | 2.5 | 52.0 |  | 13.5 |  |  |
| 1988 | 10 | 6 | 43.9 |  | 6.7 | 53.0 |  | 17.0 | 0.4 |  |
| 1989 | 10 | 26 | 29.2 |  | 1.8 | 2.9 |  | 0.6 |  |  |
| 1990 | 11 | 8 | 24.9 |  | 1.3 | 16.4 |  | 2.7 | 0.1 |  |
| 1991 | 11 | 15 | 60.0 |  | 5.3 | 44.7 |  | 4.2 |  |  |
| 1992 | 10 | 13 | 104.6 |  | 2.3 | 54.5 |  | 4.3 | 0.1 |  |
| 1993 | 11 | 18 | 100.4 |  | 9.8 | 55.1 |  | 4.9 |  |  |
| 1994 | 11 | 25 | 119.0 |  | 6.9 | 29.2 |  | 4.4 |  |  |
| 1995 | 11 | 30 | 165.0 |  | 30.1 | 84.6 |  | 7.0 |  |  |
| 1996 | 11 | 27 | 111.9 |  | 16.4 | 70.0 |  | 15.9 |  |  |
| 1997 | 11 | 1 | 66.8 |  | 30.8 | 52.5 |  | 8.5 |  |  |
| 1998 | 11 | 13 | 121.0 |  | 5.9 | 20.5 |  | 3.3 |  |  |
| 1999 | 11 | 15 | 89.8 |  | 4.4 | 18.1 |  | 0.9 |  |  |
| 2000 | 11 | 10 | 103.7 |  | 10.9 | 11.6 | 0.1 | 0.6 |  |  |
| 2001 | 11 | 12 | 101.8 |  | 2.4 | 22.1 | 0.0 | 0.7 |  |  |
| 2002 | 11 | 12 | 1.0 |  | 0.5 |  |  |  |  |  |
| 2003 | 11 | 6 | 4.9 |  | 3.1 | 1.7 | 0.1 | 0.2 |  |  |
| 2004 | 11 | 22 | 7.9 |  | 0.1 | 7.3 |  | 0.8 | 0.0 |  |
| 20053) | 11 |  |  |  |  |  |  |  |  |  |
| 2006 | 11 | 6 | 44.7 |  | 0.3 | 5.2 |  | 0.4 |  |  |
| 2007 | 11 | 7 | 5.7 |  | 0.1 | 1.3 |  | 0.0 |  |  |
| 2008 | 11 | 17 | 7.5 | 5.1 | 0.4 | 12.1 |  | 1.8 |  |  |
| 2009 | 11 | 24 | 13.0 | 2.4 |  | 5.0 |  | 0.7 |  |  |
| 2010 | 10 | 1 | 91.6 | 9.6 | 6.3 | 25.8 | 0.1 | 0.8 | 0.02 |  |
| 20114) | 11 | 29 | 9.0 | 0.6 | 3.6 | 19.9 | 0.05 | 2.1 |  |  |
| 2012 | 10 | 3 | 18.5 | 0.9 | 2.0 | 21.2 | 0.07 | 11.4 | 0.1 |  |
| 2013 | 9 | 17 | 60.1 | 0.6 | 6.9 | 25.0 | 1.3 | 6.9 | 0.1 |  |
| 2014 | 9 | 16 | 57.0 | 1.0 | 3.3 | 26.5 | 0.2 | 7.6 | 0.1 |  |

1987 - The number at age 1 was from survey earlier in autumn.
2005 - Scouting vessels searched for capelin. r/s ÁF measured. No samples taken for age determination. Estimated to be < 50 thous. tonnes.
2011-Only limited coverage of the traditional capelin distribution area.

Table 12.2.3. Icelandic Capelin. Mean weight (g) of age-classes measured in acoustic surveys in autumn. (imm=immature, mat=mature). See footnotes in table 12.2.2.

| Year | Mon. | age 1 | Age 1 | age2 | age2 | age3 | age3 | age4 | age5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Imm. | Mat. | Imm. | Mat. | Imm. | Mat. | mat | mat |
| 1978 | 10 |  |  |  | 19.8 |  | 25.4 | 26.3 |  |
| 1979 | 10 | 6.2 |  |  | 15.7 |  | 23.0 | 20.8 |  |
| 1980 | 10 | 7.3 |  |  | 19.4 |  | 26.7 |  |  |
| 1981 | 11 | 3.6 |  | 12.3 | 19.4 |  | 22.5 |  |  |
| 1982 | 10 | 3.8 |  | 8.5 | 16.5 |  | 24.1 |  |  |
| 1983 | 10 | 5.1 |  | 9.5 | 16.8 |  | 22.5 | 23.0 |  |
| 1984 | 11 | 2.9 |  | 8.3 | 15.8 |  | 25.7 | 23.2 |  |
| 1985 | 10 | 3.8 |  | 8.5 | 15.5 |  | 23.8 | 29.5 | 31.0 |
| 1986 | 10 | 4.0 |  | 6.1 | 18.1 |  | 24.1 | 28.8 |  |
| 1987 | 11 | 2.8 |  | 8.7 | 17.9 |  | 25.8 |  |  |
| 1988 | 10 | 3.0 |  | 8.0 | 15.4 |  | 23.4 | 20.9 |  |
| 1989 | 10 | 3.5 |  | 8.0 | 12.9 |  | 24.0 |  |  |
| 1990 | 11 | 3.9 |  | 8.4 | 18.0 |  | 25.5 | 36.0 |  |
| 1991 | 11 | 4.7 |  | 7.9 | 16.3 |  | 25.4 |  |  |
| 1992 | 10 | 3.7 |  | 8.6 | 16.5 |  | 22.6 | 22.0 |  |
| 1993 | 11 | 3.6 |  | 8.9 | 16.2 |  | 23.3 |  |  |
| 1994 | 11 | 3.3 |  | 7.9 | 15.9 |  | 23.6 |  |  |
| 1995 | 11 | 3.7 |  | 7.0 | 14.0 |  | 20.8 |  |  |
| 1996 | 11 | 3.1 |  | 7.4 | 15.8 |  | 20.6 |  |  |
| 1997 | 11 | 3.3 |  | 8.5 | 14.3 |  | 20.1 |  |  |
| 1998 | 11 | 3.5 |  | 9.9 | 13.7 |  | 18.8 |  |  |
| 1999 | 11 | 3.6 |  | 8.0 | 15.4 |  | 19.5 |  |  |
| 2000 | 11 | 3.9 |  | 8.5 | 13.4 | 13.0 | 20.8 |  |  |
| 2001 | 11 | 3.8 |  | 8.8 | 16.3 | 15.7 | 23.9 |  |  |
| 2002 | 11 |  |  |  |  |  |  |  |  |
| 2003 | 11 | 7.2 |  | 14.9 | 17.0 | 22.6 | 23.7 |  |  |
| 2004 | 11 | 7.4 |  | 7.6 | 16.0 |  | 18.0 | 14.5 |  |
| 2005 |  |  |  |  |  |  |  |  |  |
| 2006 | 11 | 3.7 |  | 7.9 | 15.0 |  | 16.7 |  |  |
| 2007 | 11 | 5.5 |  | 8.6 | 14.9 |  | 15.8 |  |  |
| 2008 | 11 | 6.2 | 11.0 | 6.9 | 18.6 |  | 22.4 |  |  |
| 2009 | 11 | 5.1 | 9.8 |  | 20.0 |  | 23.8 |  |  |
| 2010 | 10 | 5.8 | 12.9 | 12.2 | 19.0 | 12.9 | 24.0 | 21.2 |  |
| 2011 | 11 | 6.8 | 11.4 | 11.1 | 18.7 | 15.8 | 24.4 |  |  |
| 2012 | 10 | 6.5 | 16.0 | 15.3 | 22.0 | 22.4 | 28.0 | 26.6 |  |
| 2013 | 9 | 5.8 | 12.6 | 10.9 | 18.0 | 11.2 | 20.9 | 23.6 |  |
| 2014 | 9 | 4.2 | 9.9 | 12.7 | 18.3 | 16.6 | 21.2 | 25.0 |  |

Table 12.2.4. Icelandic Capelin. Assessment of mature capelin in the Iceland/EastGreenland/Jan Mayen area, by r/v Arni Fridriksson in January 2015 (Numbers in millions, biomass in tonnes).

| Length |  |  | Age |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |

Table 12．3．1 Capelin．The international catch since 1964 （thousand tonnes）．

|  | Winter season |  |  |  |  | Summer and autumn season |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\underset{\text { だ }}{\text { た }}$ | $\begin{aligned} & \text { تِ } \\ & \text { تِ } \\ & \text { 区्U } \end{aligned}$ | $\begin{aligned} & \text { 心. } \\ & \text { 30 } \\ & \text { Z } \end{aligned}$ |  |  |  | $\begin{aligned} & \stackrel{\rightharpoonup}{\tilde{W}} \\ & \stackrel{\rightharpoonup}{巳} \end{aligned}$ |  |  |  | 吻 |  | $\begin{aligned} & \stackrel{\rightharpoonup}{0} \\ & \stackrel{\ominus}{0} \end{aligned}$ |
| 1964 | 8.6 | － | － |  | 8.6 | － | － | － |  | － | － | 8.6 |
| 1965 | 49.7 | － | － |  | 49.7 | － | － | － |  | － | － | 49.7 |
| 1966 | 124.5 | － | － |  | 124.5 | － | － | － |  | － | － | 124.5 |
| 1967 | 97.2 | － | － |  | 97.2 | － | － | － |  | － | － | 97.2 |
| 1968 | 78.1 | － | － |  | 78.1 | － | － | － |  | － | － | 78.1 |
| 1969 | 170.6 | － | － |  | 170.6 | － | － | － |  | － | － | 170.6 |
| 1970 | 190.8 | － | － |  | 190.8 | － | － | － |  | － | － | 190.8 |
| 1971 | 182.9 | － | － |  | 182.9 | － | － | － |  | － | － | 182.9 |
| 1972 | 276.5 | － | － |  | 276.5 |  | － | － |  | － | － | 276.5 |
| 1973 | 440.9 | － | － |  | 440.9 | － | － | － |  | － | － | 440.9 |
| 1974 | 461.9 | － | － |  | 461.9 | － | － | － |  | － | － | 461.9 |
| 1975 | 457.1 | － | － |  | 457.1 | 3.1 | － | － |  | － | 3.1 | 460.2 |
| 1976 | 338.7 | － | － |  | 338.7 | 114.4 | － | － |  | － | 114.4 | 453.1 |
| 1977 | 549.2 | － | 24.3 |  | 573.5 | 259.7 | － | － |  | － | 259.7 | 833.2 |
| 1978 | 468.4 | － | 36.2 |  | 504.6 | 497.5 | 154.1 | 3.4 |  | － | 655.0 | 1，159．6 |
| 1979 | 521.7 | － | 18.2 |  | 539.9 | 442.0 | 124.0 | 22.0 |  | － | 588.0 | 1，127．9 |
| 1980 | 392.1 | － | － |  | 392.1 | 367.4 | 118.7 | 24.2 |  | 17.3 | 527.6 | 919.7 |
| 1981 | 156.0 | － | － |  | 156.0 | 484.6 | 91.4 | 16.2 |  | 20.8 | 613.0 | 769.0 |
| 1982 | 13.2 | － | － |  | 13.2 | － | － | － |  | － | － | 13.2 |
| 1983 | － | － | － |  | － | 133.4 | － | － |  | － | 133.4 | 133.4 |
| 1984 | 439.6 | － | － |  | 439.6 | 425.2 | 104.6 | 10.2 |  | 8.5 | 548.5 | 988.1 |
| 1985 | 348.5 | － | － |  | 348.5 | 644.8 | 193.0 | 65.9 |  | 16.0 | 919.7 | 1，268．2 |
| 1986 | 341.8 | 50.0 | － |  | 391.8 | 552.5 | 149.7 | 65.4 |  | 5.3 | 772.9 | 1，164．7 |
| 1987 | 500.6 | 59.9 | － |  | 560.5 | 311.3 | 82.1 | 65.2 |  | － | 458.6 | 1，019．1 |
| 1988 | 600.6 | 56.6 | － |  | 657.2 | 311.4 | 11.5 | 48.5 |  | － | 371.4 | 1，028．6 |
| 1989 | 609.1 | 56.0 | － |  | 665.1 | 53.9 | 52.7 | 14.4 |  | － | 121.0 | 786，1 |
| 1990 | 612.0 | 62.5 | 12.3 |  | 686.8 | 83.7 | 21.9 | 5.6 |  | － | 111.2 | 798.0 |
| 1991 | 202.4 | － | － |  | 202.4 | 56.0 | － | － |  | － | 56.0 | 258.4 |
| 1992 | 573.5 | 47.6 | － |  | 621.1 | 213.4 | 65.3 | 18.9 | 0.5 | － | 298.1 | 919.2 |
| 1993 | 489.1 | － | － | 0.5 | 489.6 | 450.0 | 127.5 | 23.9 | 10.2 | － | 611.6 | 1，101．2 |
| 1994 | 550.3 | 15.0 | － | 1.8 | 567.1 | 210.7 | 99.0 | 12.3 | 2.1 | － | 324.1 | 891.2 |
| 1995 | 539.4 | － | － | 0.4 | 539.8 | 175.5 | 28.0 | － | 2.2 | － | 205.7 | 745.5 |
| 1996 | 707.9 | － | 10.0 | 5.7 | 723.6 | 474.3 | 206.0 | 17.6 | 15.0 | 60.9 | 773.8 | 1，497．4 |
| 1997 | 774.9 | － | 16.1 | 6.1 | 797.1 | 536.0 | 153.6 | 20.5 | 6.5 | 47.1 | 763.6 | 1，561．5 |
| 1998 | 457.0 | － | 14.7 | 9.6 | 481.3 | 290.8 | 72.9 | 26.9 | 8.0 | 41.9 | 440.5 | 921.8 |
| 1999 | 607.8 | 14.8 | 13.8 | 22.5 | 658.9 | 83.0 | 11.4 | 6.0 | 2.0 | － | 102.4 | 761.3 |
| 2000 | 761.4 | 14.9 | 32.0 | 22.0 | 830.3 | 126.5 | 80.1 | 30.0 | 7.5 | 21.0 | 265.1 | 1，095．4 |
| 2001 | 767.2 | － | 10.0 | 29.0 | 806.2 | 150.0 | 106.0 | 12.0 | 9.0 | 17.0 | 294.0 | 1，061．2 |
| 2002 | 901.0 | － | 28.0 | 26.0 | 955.0 | 180.0 | 118.7 | － | 13.0 | 28.0 | 339.7 | 1，294．7 |
| 2003 | 585.0 | － | 40.0 | 23.0 | 648.0 | 96.5 | 78.0 | 3.5 | 2.5 | 18.0 | 198.5 | 846.5 |


|  | Winter season |  |  |  |  | Summer and autumn season |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| む゙ | $\begin{aligned} & \text { ت} \\ & \text { Z } \\ & \text { UUU } \end{aligned}$ | $\begin{aligned} & \text { ते } \\ & 3 \\ & \vdots \\ & \text { Z } \end{aligned}$ | $\begin{gathered} \mathscr{U} \\ \stackrel{0}{0} \\ \text { تَ } \end{gathered}$ |  |  | $\begin{aligned} & \text { ت} \\ & \text { ت} \\ & \text { UU } \end{aligned}$ | $\begin{aligned} & \text { ते } \\ & \text { 3 } \\ & \text { Z } \end{aligned}$ | $\begin{gathered} \mathscr{U} \\ \stackrel{\tilde{U}}{\substack{1}} \end{gathered}$ |  | Pr |  | $\begin{gathered} \text { 프 } \\ 0 \end{gathered}$ |
| 2004 | 478.8 | 15.8 | 30.8 | 17.5 | 542.9 | 46.0 | 34.0 | - | 12.0 |  | 92.0 | 634.9 |
| 2005 | 594.1 | 69.0 | 19.0 | 10.0 | 692.0 | 9.0 | - | - | - | - | 9.0 | 701.1 |
| 2006 | 193.0 | 8.0 | 30.0 | 7.0 | 238.0 | - | - | - | - |  | - | 238.0 |
| 2007 | 307.0 | 38.0 | 19.0 | 12.8 | 376.8 | - | - | - | - | - | - | 376.8 |
| 2008 | 149.0 | 37.6 | 10.1 | 6.7 | 203.4 | - | - | - | - | - | - | 203.4 |
| 2009 | 15.1 | - | - | - | 15.1 | - | - | - | - | - | - | 15.1 |
| 2010 | 110.6 | 28.3 | 7.7 | 4.7 | 150.7 | 5.4 | - | - | - | - | 5.4 | 156.1 |
| 2011 | 321.8 | 30.8 | 19.5 | 13.1 | 385.2 | 8.4 | 58.5 | - | 5.2 | - | 72.1 | 457.3 |
| 2012 | 576.2 | 46.2 | 29.7 | 22.3 | 674.4 | 9 | - | - | 1 | - | 10.0 | 684.4 |
| 2013 | 454.0 | 40.0 | 30.0 | 17.0 | 541.0 | - | - | - | - | - | - | 541.0 |
| 2014* | 111.4 | 6.2 | 8.0 | 16.1 | 141.7 | - | 30.5 | - | 5.3 | 9.7 | 45.5 | 187.2 |
| 2015* | 353.6 | 50.6 | 29.9 | 37.9 | 471.9 |  |  |  |  |  |  |  |

*preliminary, provided by working group members.

Table 12.3.2 Icelandic capelin. The total international catch of capelin in the Iceland-East Green-land-Jan Mayen area by age group in numbers (billions) and the total catch by numbers and weight (thousand tonnes) in the autumn season (August-December) since 1985.
\(\left.\begin{array}{llllllll}\hline \& \& \& \& \& \& Total \& Total <br>

Year \& age 1 \& age 2 \& age 3 \& age 4 \& Age \& number\end{array}\right]\)| weight |
| :--- |
| 1985 |
| 0.8 |

Table 12.3.3 Icelandic capelin. The total international catch of capelin in the Iceland-East Green-land-Jan Mayen area by age group in numbers (billions) and the total catch by numbers and weight (thousand tonnes) in the winter season (January-March) since 1986.

| Year | age 1 | age 2 | age 3 | age 4 | age 5 | Total number | Total weight |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 |  | 0.1 | 9.8 | 6.9 | 0.2 | 17.0 | 391.8 |
| 1987 |  | + | 6.9 | 15.5 | - | 22.4 | 560.5 |
| 1988 |  | + | 23.4 | 7.2 | 0.3 | 30.9 | 657.2 |
| 1989 |  | 0.1 | 22.9 | 7.8 | + | 30.8 | 665.1 |
| 1990 |  | 1.4 | 24.8 | 9.6 | 0.1 | 35.9 | 686.8 |
| 1991 |  | 0.5 | 7.4 | 1.5 | + | 9.4 | 202.4 |
| 1992 |  | 2.7 | 29.4 | 2.8 | + | 34.9 | 621.1 |
| 1993 |  | 0.2 | 20.1 | 2.5 | + | 22.8 | 489.6 |
| 1994 |  | 0.6 | 22.7 | 3.9 | + | 27.2 | 567.1 |
| 1995 |  | 1.3 | 17.6 | 5.9 | + | 24.8 | 539.8 |
| 1996 |  | 0.6 | 27.4 | 7.7 | + | 35.7 | 723.6 |
| 1997 |  | 0.9 | 29.1 | 11 | + | 41.0 | 797.6 |
| 1998 |  | 0.3 | 20.4 | 5.4 | + | 26.1 | 481.3 |
| 1999 |  | 0.5 | 31.2 | 7.5 | + | 39.2 | 658.9 |
| 2000 |  | 0.3 | 36.3 | 5.4 | + | 42.0 | 830.3 |
| 2001 |  | 0.4 | 27.9 | 6.7 | + | 35.0 | 787.2 |
| 2002 |  | 0.1 | 33.1 | 4.2 | + | 37.4 | 955.0 |
| 2003 |  | 0.1 | 32.2 | 1.9 | + | 34.4 | 648.0 |
| 2004 |  | 0.6 | 24.6 | 3 | + | 28.3 | 542.9 |
| 2005 |  | 0.1 | 31.5 | 3.1 | - | 34.7 | 692.0 |
| 2006 |  | 0.1 | 10.4 | 0.3 | - | 10.8 | 230.0 |
| 2007 |  | 0.3 | 19.5 | 0.5 | - | 20.3 | 376.8 |
| 2008 |  | 0.5 | 10.6 | 0.4 | - | 11.5 | 202.4 |
| 2009 |  | 0.1 | 0.6 | 0.1 | - | 0.7 | 15.1 |
| 2010 |  | 0.7 | 5.3 | 0.9 | 0.01 | 6.9 | 150.7 |
| 2011 |  | 0.1 | 16.2 | 0.6 | - | 17.0 | 385.2 |
| 2012 | 0.02 | 0.6 | 25.0 | 6.1 | 0.02 | 31.8 | 674.4 |
| 2013 | - | 0.3 | 12.1 | 9.7 | 0.2 | 22.3 | 541.0 |
| 2014 | - | 0.1 | 4.8 | 1.3 | + | 6.1 | 141.8 |
| 2015 | - | 0.3 | 17.5 | 4.7 | 0.1 | 22.7 | 471.9 |

Table 12.3.4. Initial quota and final TAC by seasons.

| Fishing season | Initial quota | Final TAC | Landings |
| :---: | :---: | :---: | :---: |
| 1992/93 ${ }^{1}$ | 500 | 900 | 788 |
| 1993/94 ${ }^{1}$ | 900 | 1250 | 1179 |
| 1994/95 | 950 | 850 | 842 |
| 1995/96 ${ }^{1}$ | 800 | 1390 | 930 |
| 1996/97 ${ }^{1}$ | 1100 | 1600 | 1571 |
| 1997/98 | 850 | 1265 | 1245 |
| 1998/99 | 950 | 1200 | 1100 |
| 1999/00 | 866 | 1000 | 934 |
| 2000/01 | 650 | 1090 | 1065 |
| 2001/02 | 700 | 1300 | 1249 |
| 2002/03 | 690 | 1000 | 988 |
| 2003/04 ${ }^{2}$ | 555 | 900 | 741 |
| 2004/05 ${ }^{3}$ | 335 | 985 | 783 |
| 2005/06 | No fishery | 235 | 238 |
| 2006/07 | No fishery | 385 | 377 |
| 2007/08 | 207 | 207 | 202 |
| 2008/09 ${ }^{4}$ | No fishery |  | 15 |
| 2009/10 | No fishery | 150 | 151 |
| 2010/11 | No fishery | 390 | 391 |
| 2011/12 | 366 | 765 | 747 |
| 2012/13 | No fishery | 570 | 551 |
| 2013/14 ${ }^{1}$ | No fishery | 160 | 142 |
| 2014/15 ${ }^{5}$ | 225 | 580 | 517 |

1) The final TAC was set on basis of autumn surveys in the season.
2) Indices from April 2003 were projected back to October 2002.
3) The initial quota was set on a basis of an acoustic survey in June/July 2004
4) No fishery was allowed, 15000 t was assigned to scouting vessels.
5) Preliminary landings.

Table 12.7.1 Icelandic capelin. The estimated number (billions) of capelin on 1 January since 1979 by age and maturity groups. The total number (billions) and weight (thousand tonnes) of the immature and maturing (fishable) stock components and the remaining spawning stock by number and weight ( March) are also given.

|  | Age <br> 2 | Age 3 | Age <br> 2 | Age 3 | Age <br> 4 | Age 5 | Numb |  | weig |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Juv. | Imm. | Mat. | Mat. | Mat. | Mat. | Imm. | Mat. | Imm. | Mat. | SSN | SSB |
| 1979 | 137.6 | 12.8 |  | 51.8 | 14.8 | 0.3 | 150.4 | 66.9 | 1028 | 1358 | 29 | 600 |
| 1980 | 50.6 | 13.8 |  | 53.4 | 3.6 | 0.2 | 64.4 | 57.2 | 502 | 980 | 17.5 | 300 |
| 1981 | 55.3 | 3.5 |  | 16.3 | 4.9 | + | 58.8 | 21.2 | 527 | 471 | 7.7 | 170 |
| 1982 | 41.2 | 3 |  | 8 | 0.5 | + | 44.2 | 8.5 | 292 | 171 | 6.8 | 140 |
| 1983 | 123.7 | 12.6 |  | 14.3 | 2 | + | 136.3 | 16.3 | 685 | 315 | 13.5 | 260 |
| 1984 | 105 | 35.7 |  | 39.8 | 7.6 | 0.1 | 140.7 | 47.5 | 984 | 966 | 21.6 | 440 |
| 1985 | 211.6 | 34.3 |  | 25.2 | 15.6 | 0.3 | 245.9 | 41.1 | 1467 | 913 | 20.7 | 460 |
| 1986 | 83.2 | 83.9 |  | 34.5 | 10.5 | 0.2 | 167.1 | 45.2 | 1414 | 1059 | 19.6 | 460 |
| 1987 | 131.9 | 25.6 |  | 22.1 | 37 | 0.2 | 157.5 | 59.1 | 1003 | 1355 | 18.3 | 420 |
| 1988 | 120.5 | 31.2 |  | 34.1 | 11.7 | + | 151.3 | 45.8 | 1083 | 993 | 18.5 | 400 |
| 1989 | 67.8 | 20.1 |  | 48.8 | 16 | 0.3 | 87.9 | 64.8 | 434 | 1298 | 22 | 440 |
| 1990 | 53.9 | 8.6 |  | 31.2 | 12.1 | + | 62.5 | 43.3 | 291 | 904 | 5.5 | 115 |
| 1991 | 98.9 | 8.6 |  | 22.3 | 4.5 | $+$ | 107.5 | 26.8 | 501 | 544 | 16.3 | 330 |
| 1992 | 111.6 | 8.1 |  | 54.8 | 5.3 | + | 119.7 | 60.1 | 487 | 1106 | 25.8 | 475 |
| 1993 | 124.6 | 13.9 |  | 46.5 | 3.5 | + | 138.5 | 50 | 622 | 1017 | 23.6 | 499 |
| 1994 | 121.3 | 16.9 |  | 50.5 | 4.6 | + | 138.2 | 55.1 | 573 | 1063 | 24.8 | 460 |
| 1995 | 188.1 | 29.5 |  | 35.1 | 8.7 | + | 217.6 | 43.8 | 696 | 914 | 19.2 | 420 |
| 1996 | 165.2 | 37.9 |  | 75.5 | 20.1 | + | 203.1 | 95.6 | 800 | 1820 | 42.8 | 830 |
| 1997 | 160 | 24.1 |  | 72.4 | 24.8 | + | 184.1 | 97.2 | 672 | 1881 | 21.8 | 430 |
| 1998 | 138.8 | 29.5 |  | 50.1 | 7.9 | $+$ | 168.3 | 58 | 621 | 1106 | 27.6 | 492 |
| 1999 | 140.9 | 16.1 |  | 53.2 | 16 | + | 157 | 69.3 | 585 | 1171 | 29.5 | 500 |
| 2000 | 115.8 | 20.5 |  | 68.2 | 10 | + | 136.3 | 78.2 | 535 | 1485 | 34.2 | 650 |
| 2001 | 122.2 | 21 |  | 46.3 | 10.5 | + | 161.2 | 56.8 | 655 | 1197 | 21.3 | 450 |
| 2002 | 117.3 | 7.6 |  | 59.3 | 10.5 | + | 126.6 | 69.8 | 510 | 1445 | 22.9 | 475 |
| 2003 | 109.4 | 9.4 |  | 58.4 | 2.9 |  | 105.1 | 61.3 | 487 | 1214 | 20.7 | 410 |
| 2004 | 134.6 | 11.4 |  | 54.2 | 6.2 | + | 143.5 | 60.4 | 597 | 1204 | 28.2 | 535 |
| 2005 | 48.0 | 2.9 |  | 86.6 | 7.5 | + | 50.9 | 72.5 | 570 | 1450 | 36.3 | 602 |
| 2006 | 81.7 | 2.1 |  | 29.4 | 1.9 |  | 83.8 | 31.3 | 761 | 639 | 18.8 | 400 |
| 2007 | 55.8 | 1.1 |  | 52.5 | 1.4 |  | 56.9 | 53.9 | 515 | 997 | 19.1 | 410 |
| 2008 | 32.4 | 4.0 |  | 32.5 | 0.7 |  | 36.3 | 33.2 | 339 | 619 | 22.2 | 406 |
| 2009 | 37.3 | 6.4 |  | 14.5 | 2.6 | + | 43.7 | 17.1 | 413 | 343 | 17.3 | 328 |
| 2010 | 77.0 | 2.9 |  | 21.5 | 4.2 |  | 79.9 | 25.2 | 728 | 548 | 21.5 | 410 |
| 2011 | 117.7 | 13.6 |  | 36.2 | 1.9 | - | 131.3 | 38.1 | 1235 | 765 | 22.3 | 411 |
| 2012 | 49.1 | 28.8 |  | 46.4 | 7.9 | + | 77.9 | 54.4 | 678 | 1112 | 20.7 | 418 |
| 2013 | 60.8 | 9.6 | 2.2 | 22.0 | 18.8 | 0.4 | 70.4 | 42.1 | 574 | 983 | 17.9 | 417 |
| 2014 | 69.6* | 17.2* | 0.6 | 22.5 | 6.3 | 0.1 | 86.8* | 29.4 | 591* | 545 | 21.1 | 424 |
| 2015 | 51.1* | 3.0* | 0.6 | 40.9 | 10.9 | 0.1 | 54.1* | 52.5 | 254* | 1013 | 19.5 | 460 |

* preliminary

Table 12.7.2 Icelandic capelin in the Iceland-East Greenland-Jan Mayen area since the fishing season 1978/79. (A fishing season e.g. 1978/79 starts in summer 1978 and ends in March 1979). Recruitment of 1 year old fish (unit 109) are given for 1 August in the beginning of the season. Spawning stock biomass (' 000 t ) is given at the time of spawning at the end of the fishing season. Landings (' 000 t ) are the sum of the total landings in the season

| Season (Summer/winter) | Recruitment | Landings | Spawning stock biomass |
| :---: | :---: | :---: | :---: |
| 1978/79 | 164 | 1195 | 600 |
| 1979/80 | 60 | 980 | 300 |
| 1980/81 | 66 | 684 | 170 |
| 1981/82 | 49 | 626 | 140 |
| 1982/83 | 146 | 0 | 260 |
| 1983/84 | 124 | 573 | 440 |
| 1984/85 | 251 | 897 | 460 |
| 1985/86 | 99 | 1312 | 460 |
| 1986/87 | 156 | 1333 | 420 |
| 1987/88 | 144 | 1116 | 400 |
| 1988/89 | 81 | 1037 | 440 |
| 1989/90 | 64 | 808 | 115 |
| 1990/91 | 118 | 314 | 330 |
| 1991/92 | 133 | 677 | 475 |
| 1992/93 | 148 | 788 | 499 |
| 1993/94 | 144 | 1179 | 460 |
| 1994/95 | 224 | 864 | 420 |
| 1995/96 | 197 | 929 | 830 |
| 1996/97 | 191 | 1571 | 430 |
| 1997/98 | 165 | 1245 | 492 |
| 1998/99 | 168 | 1100 | 500 |
| 1999/00 | 138 | 933 | 650 |
| 2000/01 | 146 | 1071 | 450 |
| 2001/02 | 140 | 1249 | 475 |
| 2002/03 | 130 | 988 | 410 |
| 2003/04 | 160 | 741 | 535 |
| 2004/05 | 57 | 783 | 602 |
| 2005/06 | 97 | 238 | 400 |
| 2006/07 | 67 | 377 | 410 |
| 2007/08 | 39 | 202 | 406 |
| 2008/09 | 44 | 15 | 328 |
| 2009/10 | 92 | 151 | 410 |
| 2010/11 | 140 | 391 | 411 |
| 2011/12 | $58$ | 747 | 418 |
| 2012/13 | $72$ | 551 | 417 |
| 2013/14 | $83^{*}$ | 142 | 424 |
| 2014/15 | $61^{*}$ | 517* | 460* |

[^4]Table 12.8.1.

| Rationale | Catches <br> $(2015 / 2016)$ | Basis |
| :--- | :--- | :--- |
| Predicted final TAC | $313600 t$ | Black regression line, see figure 2.3.5.2 |
| Initial quota, precautionary <br> considerations | $53600 t$ | Blue line, see figure 2.3.5.2. |

Table 12.19.1. Icelandic capelin. Input data for "old" short term predictions. Abundance at age in numbers (billions). Age1.ac and Age2.imm.ac are measured in acoustic surveys in autumn, Sep-tember-October. *back* are numbers 1 August, derived from stock projections from winter surveys in January-February.

| Year | Agel.ac | Age2.imm.ac | Age2.back.mat | Age2.back.tot | Age3.back.mat |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 23.5 | -1 | -1 | -1 | -1 |
| 1981 | 23.7 | 1.1 | 39.7 | 43.2 | -1 |
| 1982 | 68 | 1.7 | 17.1 | 32 | 2.3 |
| 1983 | 44.1 | 8.2 | 53.7 | 96.2 | 9.8 |
| 1984 | 73.8 | 4.6 | 40.7 | 81.8 | 27.9 |
| 1985 | 33.8 | 12.6 | 64.6 | 164.7 | 27 |
| 1986 | 58.6 | 1.4 | 35.6 | 66.2 | 65.8 |
| 1987 | 70.2 | 2.5 | 65.4 | 102.7 | 20.1 |
| 1988 | 43.9 | 6.7 | 70.3 | 94.3 | 24.5 |
| 1989 | 29.2 | 1.8 | 42.8 | 53.1 | 15.8 |
| 1990 | 39.2 | 1.3 | 31.9 | 42.1 | 6.8 |
| 1991 | 60 | 5.3 | 67.7 | 77.4 | 6.7 |
| 1992 | 104.6 | 2.3 | 70.7 | 87.3 | 6.4 |
| 1993 | 100.4 | 9.8 | 86.9 | 107 | 10.9 |
| 1994 | 119 | 6.9 | 59.8 | 94.8 | 13.2 |
| 1995 | 165 | 30.1 | 102.2 | 147.3 | 23 |
| 1996 | 111.9 | 16.4 | 100.7 | 129.6 | 29.6 |
| 1997 | 128.5 | 30.8 | 90.3 | 125.6 | 19 |
| 1998 | 121 | 5.9 | 89.5 | 108.7 | 23.2 |
| 1999 | 89.8 | 4.4 | 85.9 | 110.3 | 12.6 |
| 2000 | 103.7 | 10.9 | 65.7 | 91.4 | 16 |
| 2001 | 101.8 | 2.4 | 86.7 | 95.7 | 16.9 |
| 2002 | 74.4 | 0.5 | 68 | 91.9 | 5.9 |
| 2003 | 86.4 | 3.1 | 82.1 | 93.5 | 15.7 |
| 2004 | 7.9 | 0.1 | 86.6 | 90.1 | 7.5 |
| 2005 | -1 | -1 | 37.2 | 38.9 | 2.3 |
| 2006 | 44.7 | 0.3 | 62.5 | 63.7 | 1.1 |
| 2007 | 5.7 | 0.1 | 38.7 | 43.4 | 0.8 |
| 2008 | 7.5 | 0.4 | 17.2 | 23.3 | 3.1 |
| 2009 | 13 | -1 | 20.8 | 24.1 | 4 |
| 2010 | 91.6 | 6.3 | 41.5 | 58.4 | 2.2 |
| 2011 | 9 | 3.6 | 62.1 | 96 | 11.1 |
| 2012 | 18.5 | 2 | 26.9 | 38 | 21.7 |
| 2013 | 60.1 | 6.9 | 23.7 | 43.3 | 6.6 |

Table 12.19.2. Icelandic capelin. Mean weight at age in autumn, used in the short-term projections.

| Year | age2 | age3 |
| :---: | :---: | :---: |
| 1980 | 19.4 |  |
| 1981 | 19.4 | 22.5 |
| 1982 | 16.5 | 24.1 |
| 1983 | 16.8 | 22.5 |
| 1984 | 15.8 | 25.7 |
| 1985 | 15.5 | 23.8 |
| 1986 | 18.1 | 24.1 |
| 1987 | 17.6 | 25.8 |
| 1988 | 15.4 | 23.4 |
| 1989 | 17.8 | 26 |
| 1990 | 18.1 | 25.5 |
| 1991 | 16.3 | 25.4 |
| 1992 | 16.5 | 22.6 |
| 1993 | 16.2 | 23.3 |
| 1994 | 16 | 23.6 |
| 1995 | 14 | 20.8 |
| 1996 | 15.8 | 20.6 |
| 1997 | 14.3 | 20.1 |
| 1998 | 13.7 | 18.8 |
| 1999 | 15.4 | 19.5 |
| 2000 | 13.4 | 20.8 |
| 2001 | 16.3 | 23.9 |
| 2002 |  |  |
| 2003 | 17 | 23.7 |
| 2004 | 16 | 18 |
| 2005 |  |  |
| 2006 | 15 | 16.7 |
| 2007 | 14.9 | 15.8 |
| 2008 | 18.6 | 22.4 |
| 2009 | 20 | 23.8 |
| 2010 | 19 | 24.1 |
| 2011 | 18.7 | 24.3 |
| 2012 | 22 | 28 |
| 2013 | 18 | 20.9 |
| 2014 | 18.3 | 21.2 |

Table 12.19.3 Icelandic capelin. Outlook for 2015/2016 based on "old" prediction model.

|  |  | Landings |  |  |
| :--- | :--- | :--- | :--- | :--- |
| Method | Rationale | $\mathbf{2 0 1 5 / 2 0 1 6}$ | Basis | SSB $\mathbf{2 0 1 6}$ |
| Projection model | Zero catch | 0 | No fishing | 920 |
| Projection model - only younger age group | Zero catch | 0 | No fishing | 784 |
| Projection model - all data from autumn survey | Zero catch | 0 | No fishing | 800 |
| Zero Intercept Regression | Zero catch | 0 | No fishing | 863 |
| ICES Short Term Forecast | Zero catch | 0 | No fishing | 626 |
| Projection model | Management plan | 519 | Bescapement | 400 |



Figure 12.2.1. Icelandic capelin. Cruise tracks, relative density and distribution of capelin during an acoustic survey by r/v Arni Fridriksson during 16 September - 10 October 2014

## Acoustic index (immatures at age 1 \& 2)



Figure 12.2.2. Icelandic capelin. Indices of immature 1 and immature 2 years old capelin from acoustic surveys in autumn since 1980.


Figure 12.2.3. Icelandic capelin. Survey tracks of r/s Arni Fridriksson during 17 - 29 January 2015.


Figure 12.3.1. Icelandic capelin. Distribution of the catches in the fishing season 2014/15 based on data from logbooks of the Icelandic fleet.


Figure 12.3.2. Icelandic capelin. The total catch (in thousand tonnes) of the Icelandic capelin since 1963/64 by season.


Figure 12.8.1. Estimated final advice according to the proposed stochastic HCR against the measured number of immature capelin $\sim 15$ months earlier. The lines indicate the final TAC (unbroken) and the preliminary TAC (broken) when it is set using a $U_{\text {trigger }}$ (red vertical line) of 50 billion immature fish and a cap on the initial TAC of 400 kt . (The figure is taken from the stock-annex, WKICE 2015).

|  |  |
| :---: | :---: |
| Regression 1 (black line) and prediction interval showing relation between acoustic index at age 1 and mature capelin at age 2 one year later. Blue line is the zero intercept regression. | Regression 2 (black line) and prediction interval showing relation between total number of capelin at age 2 and mature capelin at age 3 one year later. Blue line is the zero intercept regression. |
|  |  |
| Relation between total number at age 2 and 3 in the stock in year $Y$ and mean weight for capelin at age 2 a year later. | Relation between total number at age 2 and 3 in the stock in year $Y$ and mean weight for capelin at age 3 a year later. |

Figure 12.18.1 Regressions used in "old" short-term model.

## 13 Overview on ecosystem, fisheries and their management in Greenland waters.

### 13.1 Ecosystem considerations

The marine ecosystem around Greenland is located from arctic to subarctic regions. The water masses in East Greenland are composed of the polar East Greenland Current and the warm and saline Irminger Current of Atlantic origin. As the currents round Cape Farewell at Southernmost Greenland the saline, warm Irminger water subducts the colder polar water and forms the relatively warm West Greenland Current. This flows along the West Greenland coast mixing extensively as it flows north. This current is of importance in the transport of larval and juvenile fish along the coast for important species such as cod and Greenland halibut. Additionally, cod from Icelandic waters spawning south and west of Iceland occasionally enters Greenland waters via the Irminger current and is distributed along both the Greenland East and West coast (Figure 1).


Figure 1. Spawning areas, egg and larval transport of Atlantic cod (Gadus morhus) in Greenlandic and Icelandic waters.

Depending of the relative strength of the two East Greenland currents, the Polar Current and the Irminger Current, the marine environment experience extensive variability with respect to the hydrographical properties of the West Greenland Current. The general effects of such changes have been increased production during warm periods as compared to cold ones, and resulted in extensive distribution and productivity changes of many commercial stocks. Historically, cod is the most prominent example of such a change (Holger \& Wieland 2008).
In recent year's temperature have increased significantly in Greenland waters. In West Greenland the sea temperature have increased particularly compared to the years in 1970'ies to mid 1990'ies and historical highs was registered in 2005 for the time series 1880-2012 (Figure 2).


Figure 2. Mean temperature on top of Fylla Bank (located outside Nuuk Fjord, 0-40 m depth) in the middle of June for the period 1950-2013. The curves are 3 year running mean values. The magenta/purple line is extended back to 1876 using Smed-data for area A1. From Ribergaard (2014).

Temperature in the centre of the Irminger Sea, in the depth interval 200-400m, shows no such clear long-term trend (ICES 2013c). However Rudels el al. (2012) finds that between 1998 and 2010, the salinity and temperature of the deep water in the Greenland Sea increased. Furthermore increasing temperatures in salinity the Atlantic Water entering the Arctic in the Fram Strait has increased throughout the period 1996-2012, though with the highest observation in 2006 (ICES 2013c). Such environmental changes might well propagate to different trophic levels. Accordingly, shrimp biomass fluctuations in Greenland waters as a result of environmental changes could affect fish predators such as cod (Hvingel \& Kingsley 2006) and the other way around.
The primary production period in Greenland is timely displaced along the coast due to increasing sea ice cover and a shorter summer period moving north (Blicher et al. 2007) but the main primary production takes place in May-June (Figure 3). The large latitudinal gradient spanned by Greenland, the ecosystem structure shifts moving north. For instance, the secondary producer assembly (e.g. mainly copepods) shifts from being dominated by smaller Atlantic species (Calanus finmarchicus and Calanus glacialis) to being increasingly dominated by the (sub)arctic species Calanus hyperboreus.


Figure 3. Annual variation in algal biomass and productivity at the inlet of Nuuk Fjord. a: chlorophyll ( $\mu \mathrm{g} \mathrm{l}^{-1}$ ), b: fluorescence, c : primary production ( $\mathrm{mg} \mathrm{C} \mathrm{m}{ }^{-2} \mathrm{~d}^{-1}$ ). Dots represent sampling points. From Mikkelsen et al. (2008).

Recently, the distribution of commercial species such as cod and shrimp has shifted considerably north. Such shifts have previously been associated with temperature, and
may very well be linked to the observed increase in temperature. Additionally, changes in growth of fishes may also increase as a result of temperature changes as seen for both Greenland halibut (Sünksen et al. 2008) and cod (Hovgård and Wieland, 2008).
In recent years more southerly distributed species not normally seen in Greenland water such as pearlside (Maurolicus muelleri), Whiting (Merlangius merlangus), blackbelly rosefish (Helicolenus dactylopterus), angler (Lophius piscatorius) and snake pipefish (Entelurus aequoreus) have been observed in surveys in offshore West and East Greenland and inshore West Greenland and their presence is possibly linked to increases in termperature (Møller et al. 2010).

In 2011 a mackerel (Scomber scombrus) fishery was initiated in East Greenland waters. Previous to this, no catches had ever been reported for this area and in 2013 mackerel was for the first time documented along the West Greenland coast. The reasons) for the increased abundance of mackerel in Greenlandic waters has not been clarified, however factors such as changes in the regime for their usual food resources, a density dependent effect and increased temperatures have been proposed (ICES 2013a). The effects of increased pelagic fishes abundance and their distributional shifts on demersal fishes are unknown.

### 13.1.1 Atmospheric conditions

Cod and possibly other species recruitment in Greenland waters is significantly influenced by environmental factors such as sea surface temperatures in the important Dohrn Bank region during spawning and hence by air temperatures together with the meridional wind in the region between Iceland and Greenland (Stein and Borovkov 2004). The effect of the meridional wind component in the region off South Greenland on the first winter of the offspring appears to play a vital role for the cod recruitment process. For instance, during 2003, when the strong 2003 YC was born, negative anomales were more than $-2.0 \mathrm{~m} / \mathrm{sec}$, and that particular YC was large in East Greenland waters. In general, it seems that during anomalous east wind conditions during summer months, anomalous numbers of 0-group cod are also found in Greenland waters.


Figure 4. NAO Index (Dec-Feb) 1950-2012.

## The NAO index

The NAO index, as given for 1950-2012 (Figure 4), shows negative values for winter (December-February) 2008/2009, 2009/2010 and 2010/2011. The 2009/2010 index is the strongest negative index (-1.64), encountered since 1950.
During the second half of the last century the 1960s were generally "low-index" years while the 1990s were "high-index" years. A major exception to this pattern occurred between the winter preceding 1995 and 1996, when the index flipped from being one
of its most positive (1.36) values to a negative value (-0.62). The direct influence of NAO on Nuuk winter mean air temperatures is as follows: A "low-index" year corresponds with warmer-than-normal years. Colder-than-normal temperature conditions at Nuuk are linked to "high-index" years and hence indicate a negative correlation of Nuuk winter air temperatures with the NAO. Correlation between both time series is significant ( $\mathrm{r}=-0.73, \mathrm{p} \ll 0.001$; Stein 2004). This is seen for instance in 2009, 2010 and 2011 where air temperature anomalies at Nuuk (1.0K, 4.8 K and 2.9 K ) where associated with low NAO values (Fig. 5). The 2010 air temperature anomaly (4.0K) was the highest recorded, and was associated with the largest negative NAO anomaly (see Fig. 6).


Figure 5. Time series of annual mean winter (DEC-FEB) air temperature anomalies (K) at Nuuk (1876-2012, rel. 1961-1990)


Figure 6. Time series of annual mean air temperature anomalies (K) at Nuuk (1876-2011, rel. 19611990), and 13 year running mean.

## Zonal wind components

A negative anomaly of zonal wind components for the Northwest Atlantic is associated with atmospheric conditions in the Iceland-Greenland region enclosing strong easterly winds (Figure 7, top left panel in). These winds favour surface water transports from Iceland to East Greenland and was particularly strong in 2009, while it was completely different during the same months in 2010 (Figure 7). During May-August in 2011, the cells of negative anomalies were seen to the east of Newfoundland (anomalies < 3.0 $\mathrm{m} / \mathrm{sec}$ ), and to the east of Iceland.


Figure 7. Zonal wind components for the North Atlantic (May-Aug), anomalies from 1981-2010. top left: 2009; top right: 2010; bottom left: 2011.

## Meridional wind components

As discussed in Stein and Borovkov (2004), the meridional wind component (Dec-Jan) from the Southwest Greenland region correlated positively with the trend in Greenland cod recruitment time series (first winter of age-0 cod). During winter 2009/2010, positive meridional wind anomalies were observed Southwest Greenland (Figure 8, top left panel). During winter 2010/2011, the center of positive meridional wind anomalies had moved to the Davis Strait region (Figure 5, top right panel), and during winter 2011/2012, positive meridional wind anomalies had moved to the Northeast off Newfoundland (bottom left panel in Fig. 8).


Figure 8. Meridional wind component (Dec-Jan), anomalies from 1981-2010. top left: 2009/2010; top right: 2010/2011;bottom left: 2011/2012;

### 13.1.2 Description of the fisheries

Fisheries targeting marine resources off Greenland can be divided into inshore and offshore fleets. The majority of the Greenland fleet has been built up through the 60s and is today comprised of approx. 450 larger vessels and a big fleet of small boats. It is estimated that around 1700 small boats are dissipating in some sort of artisanal fishery mainly for private use or in the pound net fishery.
Active fishing fleet reported to Greenland statistic by GRT in 1996 - no later number is available:

All fleet $(\mathrm{N})<5 \mathrm{GRT}$ 6-10GRT 11-20GRT 21-80GRT >80GRT
$441 \quad 31 \% \quad 34 \% \quad 2 \% \quad 9 \% \quad 6 \%$

There is a large difference between the fleet in the northern and southern part of Greenland. In south, were the cod fishery has historically been important the average vessel age is 22 years, in north only 9 years as it is mostly comprised of smaller boats targeting Greenland halibut using longlines.

### 13.1.3 Inshore fleets

The fleet is constituted by a variety of different platforms from dog sledges used for ice fishing, to small multipurpose boats engaged in whaling or deploying passive gears such as gill nets, pound nets, traps, dredges and longlines.

In the northern areas from Disko Bay at $72^{\circ} \mathrm{N}$ and north to Upernavik at $74^{\circ} 30 \mathrm{~N}$, dog sledge are the platforms in winter and small open vessels the units in summer, both fishing with longlines to target Greenland halibut in the ice fjords. The main by-catch from this fishery is redfish, Greenland shark, roughhead grenadier and in recent years cod in Disko Bay.

The coastal shrimp fisheries are distributed along most of the West coast from $61-72^{\circ} \mathrm{N}$. The main by-catch with the inshore shrimp trawlers is juvenile redfish, cod and Greenland halibut. An inshore shrimp fishery is conducted mainly in Disko Bay. Sorting grid is mandatory for the shrimp fishery; however several small inshore shrimp trawlers have dispensation for using sorting grid.

Cod is targeted all year, but with a peak in effort in June - July as cod in this period as accessible in shallow waters facilitating the use of the main gear types, pound and gill nets. By-catches are limited and are mainly Greenland cod (Gadus ogac) and wolffish.

In the recent years there has been an increasing exploitation rate for lumpfish. The fishing season is short, with the majority of the catch being caught in May-June. Lumpfish is caught along most of the West coast and is caught using gill nets. In small areas there is a substantial by catch of birds, especially common eiders (Somateria mollissima)

The scallop fishery is conducted with dredges at the West coast from $64-72{ }^{\circ} \mathrm{N}$, with the main landings at $66^{\circ} \mathrm{N}$. By-catch in this fishery is considered insignificant.

Snow crabs are caught in traps in areas $62-70^{\circ}$ N. Problems with by-catch are at present unknown, but are believed to be insignificant.

Salmon are caught in August-October with drifting nets and gillnets. The fishery is a mix of salmon of European and North American origin.

The coastal fleets fishing for Atlantic cod, snow crab, scallops and shrimp are regulated by licenses, TAC and closed areas. Fishery for salmon and lumpfish are unregulated.

### 13.1.4 Offshore fleets

Apart from the Greenland fleet, the marine resources in Greenland waters are exploited by several nations, mainly EU, Iceland and Norway using bottom and pelagic trawls as well as longlines.

The demersal offshore fishery is comprised of vessels primarily fishing Greenland halibut, shrimp, redfish and cod. Greenland halibut and redfish have been targeted since 1985 using demersal otter board trawls with a minimum mesh size of 140 mm . A cod fishery has previously been conducted since 1920s in West Greenland offshore waters but was absent from 1992 to the 00ies. In 2010 the cod fishery was closed off West Greenland and catches has been insignificant since. The Greenland offshore shrimp fleet consist of 15 freezer trawlers. They exclusively target shrimp stocks off West and East Greenland with landings slightly below 100000 t . The shrimp fleet is close to or above 80 BT and $75 \%$ of the fleet process the shrimp onboard. Shrimp trawls are used with a minimum mesh size of 44 mm and a mandatory sorting grid ( 22 mm ) to avoid by-catch of juvenile fish. The three most economically important fish species in Greenland: Greenland halibut, redfish and cod are found in relatively small proportions in
the by-catch. However, when juvenile fish are caught, even small biomasses can correspond to relatively high numbers.

Longliners are operating on both the East and West coast with Greenland halibut and cod as targeted species. By-catches include roundnose grenadier, roughhead grenadier, tusk, Atlantic halibut and Greenland shark (Gordon et al. 2003).
The pelagic fishery in Greenland waters is conducted in East Greenland and currently targeted species are mackerel and pelagic redfish. A relatively small fishery after herring is carried out in the border area between Greenland, Iceland and Jan Mayen. A capelin fishery has previously been done but as the Greenland share of the TAC is taken in other waters. Generally, the pelagic fishery in Greenland is very clean, with small amounts of by-catch seen.
The demersal and pelagic offshore fishing, together with longlines are managed by TAC, minimum landing sizes, gear specifications and irregularly closed areas.

### 13.2 Overview of resources

In the last century the main target species of the various fisheries in Greenland waters have changed. A large international fleet in the 50s and 60s landed large catches of cod reaching historic high in 1962 with about 450 000t. The offshore stock collapsed in the late 60 s-early 70 s due to heavy exploitation and possible due to environmental conditions. Since then the stock has been low, with occasional larger YC being transported from Iceland (i.e. 1984 and 2003). Since 2010 the cod biomass has been concentrated in the spawning grounds off East Greenland. Following the cod collapse, the offshore shrimp fishery started in 1969 and has been increasing up to 2003 reaching a catch level close to 150000 t . The stock decreased thereafter and is now at the low 1990 level with an advised TAC for 2015 of 60000 t .

### 13.2.1 Shrimp

The shrimp (Pandalus borealis) stock in Greenland waters has been declining since 2003. The stock in East Greenland is at a low level based on available information. The 2003 West Greenland shrimp biomass was at the highest in the time series but it has since decreased.

### 13.2.2 Snow crab

The biomass of snow crab (Chionoecetes opilio) in West Greenland waters has decreased substantially since 2001 . Snow crab has been exploited inshore since the mid 90s and offshore since 1999. Total landings have since 2010 been reported at around 2 000 t a decrease from a high level in 2001 at 15000 t . After several years of decreasing CPUE it now appears to have stabilized at low levels in the majority of areas.

### 13.2.3 Scallops

The status of scallops in Greenland is unknown. From the mid 80s to the start 90s landings were between 4-600 t yearly, increased to around 2000 t in late 1990ies. Catches decreased again and is below 600 tons in 2014. The fishery is based on license and is exclusively at the west coast between $20-60 \mathrm{~m}$. The growth rate is considered very low reaching the minimum landing size on 65 mm in 10 years.

### 13.2.4 Squids

The status of squids in Greenland waters are unknown.

### 13.2.5 Cod

Since 2015 assessment and advice for cod in Greenland water take into account that 3 different stocks, based on spawning areas and genetics, are the basis for the cod fishery and the following management is therefore recommended for different 3 areas: a) inshore in Western Greenland, b) offshore Western Greenland (NAFO 1A-1E) and offshore Eastern Greenland (ICES subarea XIVb and NAFO subarea 1F). Current landings for inshore cod are 18300 t , and have steadily increased since 2009 where landings were 7000 t . Landing from offshore Western Greenland was minor (less than 500 tons since 2006) and from offshore Eastern Greenland area 2014 landing was 7900 t , an increase from the 2011-2013 level at 5000 t .

Catches are high compared to the last three decades, however they are only a fraction of the landings caught in the 1950's and 1960's. Recruitment has been negligible since the 1984 and 1985 year class, though it has improved in the last decade, especially inshore, where the 2009 YC is the best seen since the 1984 YC. In 2007 and 2009 dense concentrations of unusual large cod were documented to be actively spawning off East Greenland, and management actions have been taken to protect these spawning aggregations. The inshore fishery has been regulated since 2009 and the offshore fishery is managed with license and minimum size $(40 \mathrm{~cm})$. As a response to the favourable environmental conditions (large shrimp stock, high temperatures) there is a possibility that the offshore cod will rebuild to historical levels if managed with this objective. A management plan with the objective of achieving this goal has been implemented for the fishing seasons 2014-2016. Several YC are present in the inshore fishery, and with the stable recruitment in recent years and widespread fishery there are several indications that the stock is experiencing favourable conditions and that recruitment is not impaired in spite of an increased fishing effort in later years.

### 13.2.6 Redfish

Redfish (Sebastes mentella and Sebastes norvegicus) are primarily caught of East Greenland. Catches have been small since 1994, but recently large year classes have given rise to a significant fishery with 2010-14 catches being around 8000 t . This includes both redfish species, but the majority (e.g. $\sim 80 \%$ ) is most likely S. mentella. Recent East Greenland survey estimates indicate i decline i S. mentella while S. norvegicus is increasing.

### 13.2.7 Greenland halibut

Greenland halibut in the Greenland area consist of at least two stocks and several components; the status of the inshore component is not known but it has sustained catches of $15-20000 \mathrm{t}$ annually, taken primarily in the northern area (north of $68^{\circ} \mathrm{N}$ ). The offshore stock component in West Greenland (NAFO SA $0+1$ ) is a part of a shared stock between Greenland and Canada. The stock has remained stable in the last decade, sustaining a fishery of about 30000 t annually ( 15000 t in Greenland water). The East Greenland stock is a part of a stock complex extending from Greenland to the Barents Sea. The stock size is currently estimated as being at a historical low. Catches exceeds advice in most years with catches in Greenland waters being around 8000 t .

### 13.2.8 Lumpfish

The status of the lumpfish is unknown. The landing of lumpfish has increased dramatically in the last decades with catches being close to 13000 t in 2013. Catches are highest in the southern-mid section of the Greenland west coast. There are no indications of
the impact on the stock. A management plan was implemented in 2014 regulating the fishery with TAC and number of fishing days.

### 13.2.9 Capelin

On the Greenland Eastcoast an offshore pelagic fleet have been conducting a fishery on capelin ( 45000 t landed in 2014 by Greenland, EU, Norway and Iceland). The capelin has shifted distribution more west and north in recent years, and are believed to spent a substantial amount of time in Greenland waters. The west Greenland capelin stock is not fished and its size is unknown.

### 13.2.10 Mackerel

A mackerel fishery in Greenland waters initiated in 2011 with catches of 162 t and increased to more than 78000 t in 2014. Mackerel is known to feed on various species, including fish larvae, and it competes with others pelagic species, such as herring, for resources (Langøy et al. 2012). Thus it might/can have a key role on the ecosystem of many commercial important species in Greenland.

### 13.2.11 Herring

A fishery for Norwegian spring spawning herring in Greenland water has increased in recent years and in 2014 catches increased to 9000 t . The herring has shifted distribution more west in recent years

### 13.3 Advice on demersal fisheries

ICES recommends that the offshore cod stock is protected to allow for rebuilding. Inshore cod advice is based on the DLS approach. For the offshore cod, a recovery plan is recommended to ensure a sustainable increase in SSB and recruitment. Such initiatives must include appropriate measures to avoid any cod by-catch in other fisheries deploying mobile gears capable of catching cod. Observers must monitor functionalism of measures.

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## 14 Cod (Gadus morhua) in NAFO Subdivisions 1A-1E (Offshore West Greenland)

## Executive summary

From 2015 the advice for cod in Greenland offshore waters has been split in two stock components (advice year 2016). The West Greenland offshore stock component is now comprised of the NAFO subdivisions 1A-E in West Greenland. The East Greenland stock component is comprised of the area NAFO subdivision 1F in South Greenland and ICES subarea XIV in East Greenland.

Some mixing occurs between the two stocks in West Greenland which at present is considered to act as a nursing area for juveniles of the East Greenland stock component. The offshore fishery in West Greenland was closed in accordance with an implemented management plan in 2014. However, a dispensation was given to a small trawler that fished 116 tons and the 2009 YC dominated the catches.

Survey indices show that the biomass and abundance has increased due to the 2009 YC which is present in considerable numbers. This YC is distributed further south in 2012-2014 than in 2011.

The spatial distribution of the 2009 YC is different than previous year classes that usually migrate out of the area at age 4 , but a large part of the 2009 YC still remains in the southern area (NAFO 1E) at age 5 in 2014.

No formal assessment is conducted and there are no biological reference points for the stock. Information from survey indices (German Groundfish survey and Greenland Shrimp and Fish survey) are used as basis for advice.

No significant spawning has been observed in the area, and fish older than 6 yrs are lacking in the area.

### 14.1 Stock definition

The cod found in Greenland is derived from four separate "stocks" that each is labelled by their spawning areas: I) offshore West Greenland waters; II) West Greenland fiords; III) offshore East Greenland and Icelandic waters and IV) inshore Icelandic waters (Therkildsen et al. 2013), (Fig. 14.1).

From 2012 the inshore component (West Greenland, NAFO Subarea 1) was assessed separately from all offshore components. From 2015 the offshore West Greenland (NAFO subdivisions 1A-E) and East Greenland (NAFO subdivision 1F and ICES subarea XIV) components was assessed separately. The Stock Annex provides more details on the stock identities including the references to the primary literature.

### 14.2 Fishery

### 14.2.1 The emergence and collapse of the Greenland offshore cod fisheries

The Greenland commercial cod fishery in West Greenland started in the 1920s. The fishery gradually developed culminating with catch levels at 400000 tonnes annually in the 1960s. Due to overfishing and deteriorating environmental conditions the stock size declined and the fishery completely collapsed in the early 1990's (Tab 14.2.1, Fig 14.2.1). No fishery has developed since. More details on the historical development in the fisheries are provided in the stock annex.

### 14.2.2 The fishery in 2014

In 2014 a management plan for the offshore fisheries for cod was implemented with the overall objective of rebuilding the stock in West Greenland by closing the area for fisheries. However a dispensation was given to a small trawler (<75BRT/120BRT) that fished offshore on the inshore quota.
Catches in the fishery in 2014 amounted to a total of 116 t with 30 t being caught on Tovqussaq Bank (NAFO div 1C, $64^{\circ} 25 N$ ) and $86 t$ being caught primarily on Dana Bank (NAFO div 1D and 1E, between $62^{\circ} 00-63^{\circ} 00 \mathrm{~N}$ ), and a smaller amount on Fiskenæs Bank (NAFO div. 1D, $63^{\circ} 30 \mathrm{~N}$, figure 14.2.2.1 and 14.2.2.2).

The fishery occurred in spring Mar, Apr and May with $50 \%$ being caught in May (table 14.2.2.1).

### 14.2.3 Length, weight and age distributions in the fishery.

Length measurements were taken by Greenland Institute of Natural Resources (GINR) personal on the landings in the city of Paamiut from catches taken on Tovqussaq Bank in NAFO div. 1C $\left(64^{\circ} 25 \mathrm{~N}\right)$, whereas a crew member took length measurements directly on the ship from catches taken on Dana Bank in NAFO div 1D and 1E $\left(62-63^{\circ} \mathrm{N}\right)$. In total 1070 cod were measured.

Overall mean length in the fishery was 54 cm and 5 year olds (YC 2009) dominated the catches (figure 14.2.3.1).

Mean length differed between areas with smallest fish caught furthest to the north in NAFO 1C (mean length $=48 \mathrm{~cm}$ ) and largest fish caught on Dana Bank (mean length $=$ 58 cm , figure 14.2.3.2). The 2009 YC dominated the catches in both areas, but there were larger and older fish ( $6-8 \mathrm{yrs}$ ) present on Dana Bank but not on Tovqussaq bank where younger fish ages $4-5$ yrs were present.

### 14.3 Surveys

At present, two offshore trawl surveys (Greenland and German) provide the core information relevant for stock assessment purposes. For details of survey design see stock annexes.

The German survey targets cod and has since 1982 covered the main cod grounds off West Greenland up to $67^{\circ} \mathrm{N}$ at depths down to 400 m , thus including periods of both high and low cod abundance. The Greenland survey targets shrimp and cod off West Greenland up to $72^{\circ} \mathrm{N}$ and from 0 to 600 m from 1992, hereby extending into northern areas where large cod concentrations are not expected. Although most of the effort has previously been allocated towards shrimp the recent addition of additional fish stations implies a fair coverage of the West Greenland cod habitat in this survey.

### 14.3.1 Results of the Greenland Shrimp and Fish survey

The numbers of valid trawl hauls in 2014 was 194 (table 14.3.1.1).
The 2014 survey abundance index of Atlantic cod in West Greenland was estimated at 110 million individuals and the survey biomass at 84900 tons. Survey abundance and biomass increased with $58 \%$ and $147 \%$ respectively compared to 2013 (table 14.3.1.2 \& 14.3.1.3). Abundance and biomass was primarily in NAFO Div. 1C and E (figure 14.3.1.1 and 14.3.1.2).

The survey catches were dominated by the 2009 YC in 2011, 2012 and 2013 accounting for $84 \%, 64 \%$ and $52 \%$ of the total abundance respectively (table 14.3.1.4, figure 14.3.1.3). In 2014 the 2010 YC dominated the abundance with $51 \%$ of the total abundance followed by the 2009 YC accounting for $33 \%$ of the total abundance. The 2010 YC was primarily found in one large haul (in NAFO 1C) that accounted for $66 \%$ of the abundance estimate, whereas the 2009 YC was also found in other hauls (NAFO 1E) (figure 14.3.1.4 and 14.3.1.5).

The 2009 was mainly found in the northern part of the survey (NAFO 1B) at age 2 in 2011 (figure 14.3.1.5). In 2012 and 2013 this YC was however mainly found in the southern part of the survey (NAFO Div. 1D and 1E) (table 14.3.1.5, figure 14.3.1.4 and 14.3.1.5). The 2010 YC show the same distribution pattern of being in the northern part of the survey area (NAFO 1A and 1B) at ages 1 and 2, and moving further to the south at ages 3 and 4 . Younger year classes (2011- and 2012 YC ) were primarily found in the northern part of the survey area (NAFO Div. 1A and 1B).

The majority of cod found offshore in West Greenland are younger than 7 years, and the 2014 survey confirmed that older and larger cod barely exist offshore in West Greenland (figure 14.3.1.5 and 14.3.1.6).

The B4+ biomass has increased in recent years, but prior to 2013 the B4+ was much lower than the total biomass indicating that until 2012 cod younger than 4 yrs were dominating the area (table 14.3.1.6).

The offshore cod start to spawn at age 5-6 yrs and the spawning stock biomass in the survey show an increasing trend in recent years with spawning stock being concentrated in the southern area (NAFO 1E, figure 14.3.1.7 and 14.3.1.8). The spawning stock estimate for NAFO 1C is based on one large haul.

The 2014 survey shows a very high increase in both abundance and biomass compared to previous years. However, one station constituted $66 \%$ of the abundance estimate and $61 \%$ of the biomass estimate, which is also reflected in a high index uncertainty ( $\mathrm{CV}=67$ ). This haul was located in NAFO division 1C resulting in higher abundance index in this area compared to other areas (figure 14.3.1.4) and high spawning stock biomass in this area compared to other areas in the survey (figure 14.3.1.7). The large haul was taken on Tovqussaq Bank where a small trawl fishery also took place in May. The length of fish caught in the commercial fishery was similar to the length found in the survey (figure 14.2.3.2).

### 14.3.2 Results of the German groundfish survey

In 2014, 45 valid trawl stations were sampled during autumn in the German Greenland offshore groundfish survey in West Greenland NAFO 1C-1E (Table 14.3.2.1).

Overall, abundance increased by $24 \%$ from 2013 to 2014 (Table 14.3.2.2) and biomass increased by $57 \%$ (Table14.3.2.3). The main reason for the increase in abundance and biomass was one very large haul, located in NAFO 1D (figure 14.3.2.2) which contributed with $69 \%$ of the biomass estimate and resulted in high SD (table 14.3.2.3). The hauls resulted in increased numbers of the 2009 and 2010 YC in 2014 compared to 2013 (Table 14.3.2.4). Since 2012 the 2009 YC has dominated the catches and the 2010 has been the second most abundant YC in the survey. These two year classes are mainly observed in NAFO div. 1C and 1D in 2014 (figure 14.3.2.3) which is further to the north for the 2009 YC than in the Greenland survey.

The survey time series shows three abundance peaks: one in 1987-1989 caused by the 1984 and 1985 YC, one in 2006 caused by the 2003 YC and one in 2012 caused by the

2009 YC (figure 14.3.2.4). Biomass indices show the same peaks, although an increase in biomass in the period 2012-2014 compared to the previous periods (figure 14.3.2.5).

Overall findings are the same in the Greenland and the German survey: the 2009 YC dominates catches in recent years, followed by the 2010 YC.

### 14.4 Information on spawning

No spawning of significance has been documented on the banks in West Greenland, and few large cod are found in the survey. Spawning is therefore assumed to be limited.

### 14.5 Tagging experiments

A total of 16030 cod have been tagged in different regions of Greenland in the period of 2003-2014 (table 14.5.1). 3884 cod have been tagged in the offshore area in West Greenland NAFO 1D+1E (primarily on Dana Bank) in 2007, 2012 and 2013.

Offshore recaptures are found both in West- and East Greenland and Iceland (table 14.5.2). Tagged fish in the offshore area in West Greenland are more often caught in the same area (21 individuals), but some also migrate eastward (5 individuals recaptured in East Greenland, and 13 in Iceland, table 14.5.2). Limited fishing in several areas influences the signal from the recaptures, and more analysis needs to be performed taking the fishing effort into account in order to investigate magnitude of the eastward migration rate.

### 14.6 State of the stock

The West Greenland offshore stock component has been severely depleted since the 1970ies and collapsed in the 1990ies. The surveys show only a minor increase in biomass in recent years. Abundance however has fluctuated since 2005, indicating that small fish enter the survey but are not caught at older ages. This is caused by an eastward migration out of the area, and the area is presently considered to act mainly as a nursing area for the East Greenland and Icelandic stock components.
Recently the 2009 YC has been caught in considerable numbers and is believed to be of East Greenland and/or Icelandic origin and will probably migrate out of West Greenland when reaching maturity. However, at age 5 a part of this YC still remains in the southern part of the area (NAFO 1E), which has not been the typical pattern observed with the recent larger than average YC's from 2003, 2005 and 2007.
The stock is considered to be at a very low level compared to historic.

### 14.7 Implemented management measures for 2015

According to a management plan implemented in 2014 no offshore fishery is to take place in NAFO subdivision 1A-1E in 2015.

### 14.8 Management plan

In 2014 a management plan was implemented for the offshore cod fishery in Greenland (2014-2016). The management plan is built on the distinction between the inshore and two offshore stocks components.


Management area West Greenland covers NAFO Subdivisions 1A-E and management area SouthEast Greenland covers ICES Subarea XIVb (survey area Q1-6) + NAFO Subdivision 1 F corresponding to the ICES distinction.

According to the management plan, management area West TAC should be $0 t$ for the period 2014-2016 in order to protect the West offshore stock component. The TAC in management area South East is $10000 \mathrm{t} /$ year between 2014 and 2016.

The management plan has not been evaluated by ICES.

### 14.9 Management considerations.

The fishery in West Greenland should be considered a mixed stock fishery, containing fish from both Greenland and Iceland stocks. There is currently no standardized procedure to determine the proportional contribution of each stock to the landings. However, given the current state of the stock, catches taken in West Greenland waters will primarily consist of fish from other cod stocks.

The traditional spawning grounds in West Greenland are well described and if any fishing is allowed such areas should be protected. This will both protect any present spawning stock and minimize the proportion of the West Greenland stock in the catches.

### 14.10References

Retzel, A. 2015. Greenland Shrimp and Fish survey results for Atlantic cod in NAFO subareas 1A-1E (West Greenland) in 2014. ICES North Western Working Group (NWWG) April 28May 5, 2015, WD 19.

Retzel, A. 2015. Greenland commercial data for Atlantic cod in West Greenland offshore waters for 2014. ICES North Western Working Group (NWWG) April 28- May 5, 2015, WD 20.

Therkildsen, N.O.,Hemmer-Hansen, J., Hedeholm, R.B., Wisz, M.S., Pampoulie, C., Meldrup, D., Bonanomi, S., Retzel, A., Olsen, S.M., Nielsen, E.E. 2013. Spatiotemporal SNP analysis reveal
pronounced biocomplexity at the northern renge margin of Atlantic cod Gadus morhua. Evoltutionary Applications. DOI 10.1111/eva. 12055

Table 14.2.1. Offshore catches ( $\mathbf{t}$ ) divided into NAFO divisions in West Greenland. 1924-1991: Horsted 2000, 2004-present: Greenland Fisheries License Control.

| Year | NAFO $1 \mathrm{~A}$ | NAFO $1 \text { B }$ | NAFO 1 C | NAFO 1D | NAFO 1E | NAFO IF | Unknown NAFO div. | NAFO $1 A-1 E$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1924 |  |  |  |  |  |  | 200 |  |
| 1925 |  |  |  |  |  |  | 1871 |  |
| 1926 |  |  |  |  |  |  | 4452 |  |
| 1927 |  |  |  |  |  |  | 4427 |  |
| 1928 |  |  |  |  |  |  | 5871 |  |
| 1929 |  |  |  |  |  |  | 22304 |  |
| 1930 |  |  |  |  |  |  | 94722 |  |
| 1931 |  |  |  |  |  |  | 120858 |  |
| 1932 |  |  |  |  |  |  | 87273 |  |
| 1933 |  |  |  |  |  |  | 54351 |  |
| 1934 |  |  |  |  |  |  | 88422 |  |
| 1935 |  |  |  |  |  |  | 65796 |  |
| 1936 |  |  |  |  |  |  | 125972 |  |
| 1937 |  |  |  |  |  |  | 90296 |  |
| 1938 |  |  |  |  |  |  | 90042 |  |
| 1939 |  |  |  |  |  |  | 62807 |  |
| 1940 |  |  |  |  |  |  | 43122 |  |
| 1941 |  |  |  |  |  |  | 35000 |  |
| 1942 |  |  |  |  |  |  | 40814 |  |
| 1943 |  |  |  |  |  |  | 47400 |  |
| 1944 |  |  |  |  |  |  | 51627 |  |
| 1945 |  |  |  |  |  |  | 45800 |  |
| 1946 |  |  |  |  |  |  | 44395 |  |
| 1947 |  |  |  |  |  |  | 63458 |  |
| 1948 |  |  |  |  |  |  | 109058 |  |
| 1949 |  |  |  |  |  |  | 156015 |  |
| 1950 |  |  |  |  |  |  | 179398 |  |
| 1951 |  |  |  |  |  |  | 222340 |  |
| 1952 | 0 | 261 | 2996 | 18188 | 707 | 37905 | 257488 | 117126* |
| 1953 | 4546 | 46546 | 10611 | 38915 | 932 | 25242 | 98225 | 180220* |
| 1954 | 2811 | 97306 | 18192 | 91555 | 727 | 15350 | 60179 | 266682 * |
| 1955 | 773 | 50106 | 32829 | 87327 | 3753 | 4655 | 68488 | 241499** |
| 1956 | 15 | 56011 | 38428 | 128255 | 8721 | 4922 | 66265 | 296315* |
| 1957 | 0 | 58575 | 32594 | 62106 | 29093 | 16317 | 47357 | 225836* |
| 1958 | 168 | 55626 | 41074 | 73067 | 21624 | 26765 | 75795 | 258062 * |
| 1959 | 986 | 74304 | 10954 | 30254 | 12560 | 11009 | 67598 | 191343* |
| 1960 | 35 | 58648 | 18493 | 35939 | 16396 | 9885 | 76431 | 200522 * |
| 1961 | 503 | 78018 | 43351 | 70881 | 16031 | 14618 | 90224 | 293104* |
| 1962 | 1017 | 122388 | 75380 | 57972 | 25336 | 17289 | 125896 | 400719* |
| 1963 | 66 | 70236 | 73142 | 76579 | 46370 | 16440 | 122653 | 381917* |
| 1964 | 96 | 49049 | 49102 | 82936 | 33287 | 13844 | 99438 | 307878* |
| 1965 | 385 | 80931 | 66817 | 71036 | 15594 | 15002 | 92630 | 321829 * |
| 1966 | 12 | 99495 | 43557 | 62594 | 19579 | 18769 | 95124 | 313044 * |
| 1967 | 361 | 58612 | 78270 | 122518 | 34096 | 12187 | 95911 | 385949 * |
| 1968 | 881 | 12333 | 89636 | 94820 | 61591 | 16362 | 97390 | 350870 * |
| 1969 | 490 | 7652 | 31140 | 65115 | 41648 | 11507 | 35611 | 179055* |
| 1970 | 278 | 3719 | 13244 | 23496 | 23215 | 15519 | 18420 | 78775* |
| 1971 | 39 | 1621 | 28839 | 21188 | 9088 | 20515 | 26384 | 80501* |


| Year | NAFO $1 \mathrm{~A}$ | NAFO $1 \text { B }$ | NAFO $1 \mathrm{C}$ | NAFO 1D | NAFO 1 E | NAFO $1 F$ | Unknown NAFO div. | NAFO $1 A-1 E$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 0 | 3033 | 42736 | 18699 | 7022 | 4396 | 20083 | 90410* |
| 1973 | 0 | 2341 | 17735 | 18587 | 10581 | 2908 | 1168 | 50347 * |
| 1974 | 36 | 1430 | 12452 | 14747 | 8701 | 1374 | 656 | 37999 * |
| 1975 | 0 | 49 | 18258 | 12494 | 6880 | 3124 | 549 | 38188* |
| 1976 | 0 | 442 | 5418 | 10704 | 8446 | 2873 | 229 | 25215* |
| 1977 | 127 | 301 | 4472 | 7943 | 8506 | 2175 | $35477{ }^{1}$ | 53546 * |
| 1978 | 0 | 0 | 11856 | 2638 | 3715 | 549 | $34563{ }^{1}$ | 51760 * |
| 1979 | 0 | 16 | 6561 | 4042 | 1115 | 537 | $51139{ }^{1}$ | 60635* |
| 1980 | 0 | 1800 | 2200 | 2117 | 1687 | 384 | $7241{ }^{1}$ | 14705* |
| 1981 | 0 | 0 | 4289 | 4701 | 4508 | 255 | 0 | 13498 |
| 1982 | 0 | 133 | 6143 | 10977 | 11222 | 692 | 1174 | 29621* |
| 1983 | 0 | 0 | 717 | 6223 | 16518 | 4628 | 293 | 23703* |
| 1984 | 0 | 0 | 0 | 4921 | 5453 | 3083 | 0 | 10374 |
| 1985 | 0 | 0 | 0 | 145 | 1961 | 1927 | 2402 | 3360 * |
| 1986 | 0 | 0 | 0 | 2 | 72 | 24 | 1203 | 982 * |
| 1987 | 0 | 0 | 5 | 815 | 67 | 43 | 3041 | 3787* |
| 1988 | 0 | 0 | 919 | 17463 | 10913 | 6466 | 8101 | 35931* |
| 1989 | 0 | 0 | 0 | 11071 | 48092 | 14248 | 2 | 59165 |
| 1990 | 0 | 0 | 2 | 563 | 21513 | 10580 | 7503 | 27151* |
| 1991 | 0 | 0 | 0 | 0 | 104 | 1942 | 0 | 104 |
| 1992 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 5 | 3 | 1 | 0 | 8 |
| 2005 | 0 | 0 | 1 | 0 | 0 | 71 | 0 | 1 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 414 | 0 | 0 |
| 2007 | 0 | 0 | 0 | 31 | 435 | $2011^{2}$ | 0 | 466 |
| 2008 | 0 | 0 | 0 | 23 | 526 | $11370{ }^{2}$ | 0 | 549 |
| 2009 | 0 | 0 | 0 | 0 | 6 | $3323{ }^{2}$ | 0 | 6 |
| 2010 | 0 | 0 | 0 | 0 | 2 | 281 | 0 | 2 |
| 2011 | 0 | 0 | 0 | 0 | 8 | 542 | 0 | 8 |
| 2012 | 0 | 0 | 1 | 95 | 236 | 1470 | 0 | 332 |
| 2013 | 0 | 0 | 0 | 209 | 270 | 1405 | 0 | 479 |
| 2014 | 0 | 0 | 30 | 68 | 18 | 1833 | 0 | 116 |

1) Estimates for assessment include estimates of unreported catches. The total estimated value for West Greenland (inshore + offshore) was 73000 t in 1977 and 1978, 1979: 99000 t , 1980: 54000 t . The value given in the table are these values minus the inshore catches minus known offshore NAFO division catches.
2) Include catches taken with small vessels and landed to a factory in South Greenland (Qaqortoq), 2007:597 t, 2008: 2262 t , 2009: 136 t .
*) Unknown NAFO division catches added accordingly to the proportion of known catch in NAFO divisions 1A-1E to known total catch in all NAFO divisions.

Table 14.2.2.1: 2014 cod catches ( $t$ ) divided into month and NAFO areas, caught by the offshore fisheries.

| NAFO | Jan | Feb | Mar | Apr | May Jun | Jul | Aug | Sep | Oct | Nov | Dec | Total $\%$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 C |  |  |  | 28 | 2 |  |  |  |  |  |  |  |
| 1 D | 2 | 15 | 4 | 47 |  |  |  |  |  | 30 | $26 \%$ |  |
| 1 E |  | 7 | 0.4 | 11 |  |  | 68 | $58 \%$ |  |  |  |  |
| Total | 2 | 22 | 33 | 59 |  |  | 18 | $16 \%$ |  |  |  |  |
| $\%$ | $2 \%$ | $19 \%$ | $29 \%$ | $51 \%$ |  |  | 116 |  |  |  |  |  |

Table 14.3.1.1. Number of hauls in the Greenland Shrimp and Fish survey in West Greenland by year and NAFO subdivisions.

| West Greenland |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year/NAFO 0A | 1A | 1B | 1C | 1D | 1E | Total |
| 1992 | 92 | 44 | 18 | 18 | 11 | 183 |
| 1993 | 69 | 49 | 21 | 15 | 12 | 166 |
| 1994 | 76 | 58 | 23 | 8 | 9 | 174 |
| 1995 | 83 | 61 | 29 | 13 | 14 | 200 |
| 1996 | 71 | 57 | 29 | 12 | 9 | 178 |
| 1997 | 84 | 56 | 32 | 12 | 12 | 196 |
| 1998 | 77 | 80 | 27 | 19 | 14 | 217 |
| 1999 | 84 | 81 | 33 | 16 | 14 | 228 |
| 2000 | 56 | 62 | 37 | 23 | 14 | 192 |
| 2001 | 60 | 75 | 36 | 24 | 15 | 210 |
| 2002 | 50 | 80 | 32 | 18 | 20 | 200 |
| 2003 | 51 | 63 | 30 | 18 | 15 | 177 |
| 2004 | 54 | 55 | 24 | 22 | 20 | 175 |
| New Survey Gear Introduced |  |  |  |  |  |  |
| 20056 | 65 | 56 | 26 | 19 | 23 | 195 |
| 20065 | 86 | 60 | 26 | 20 | 21 | 218 |
| 2007 8 | 73 | 58 | 26 | 27 | 31 | 223 |
| 2008 6 | 69 | 61 | 28 | 23 | 25 | 212 |
| 20098 | 74 | 75 | 28 | 22 | 24 | 231 |
| 201010 | 95 | 76 | 30 | 23 | 25 | 259 |
| 20110 | 73 | 64 | 24 | 18 | 12 | 191 |
| 20120 | 73 | 64 | 21 | 18 | 18 | 194 |
| 2013 4 | 73 | 52 | 20 | 13 | 21 | 183 |
| 2014 0 | 78 | 57 | 19 | 17 | 23 | 194 |

Table 14.3.1.2 Cod abundance indices ('000) from the Greenland Shrimp and Fish survey in West Greenland by year and NAFO subdivisions.

| West Greenland |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0A | 1A | 1B | 1C | 1D | 1E | Total | CV |
| 1992 |  | 4 | 53 | 243 | 345 | 0 | 645 |  |
| 1993 |  | 2 | 16 | 54 | 135 | 286 | 493 |  |
| 1994 |  | 10 | 41 | 87 | 0 | 6 | 144 |  |
| 1995 |  | 0 | 51 | 380 | 44 | 62 | 537 |  |
| 1996 |  | 0 | 0 | 46 | 68 | 87 | 201 |  |
| 1997 |  | 0 | 7 | 31 | 0 | 0 | 38 |  |
| 1998 |  | 0 | 4 | 0 | 26 | 26 | 56 |  |
| 1999 |  | 32 | 136 | 16 | 23 | 6 | 213 |  |
| 2000 |  | 585 | 437 | 71 | 58 | 9 | 1160 |  |
| 2001 |  | 26 | 305 | 110 | 448 | 305 | 1194 |  |
| 2002 |  | 13 | 203 | 78 | 3294 | 114 | 3702 |  |
| 2003 |  | 492 | 1395 | 351 | 727 | 214 | 3179 |  |
| 2004 |  | 197 | 152 | 379 | 2630 | 1538 | 4896 |  |
| New Survey Gear Introduced |  |  |  |  |  |  |  |  |
| 2005 | 143 | 198 | 871 | 1845 | 4796 | 6683 | 14537 | 25 |
| 2006 | 453 | 371 | 4454 | 2564 | 15703 | 3359 | 26905 | 45 |
| 2007 | 737 | 1318 | 3302 | 7353 | 3624 | 3296 | 19628 | 31 |
| 2008 | 1209 | 897 | 4185 | 4068 | 9008 | 11553 | 30913 | 27 |
| 2009 | 881 | 889 | 4195 | 3272 | 2788 | 1252 | 13277 | 12 |
| 2010 | 338 | 720 | 2837 | 2712 | 8295 | 2745 | 17647 | 23 |
| 2011 |  | 8756 | 47092 | 2179 | 26510 | 1013 | 85549 | 14 |
| 2012 |  | 7661 | 10228 | 3017 | 1270 | 27081 | 49258 | 54 |
| 2013 | 4613 | 8951 | 12864 | 5673 | 7887 | 29924 | 69911 | 43 |
| 2014 |  | 6911 | 5670 | 78854 | 2456 | 16254 | 110145 | 67 |

Table 14.3.1.3. Cod biomass indices (tons) from the Greenland Shrimp and Fish survey in West Greenland by year and NAFO subdivisions.

| West Greenland |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0A | 1A | 1B | 1 C | 1D | 1E | Total | CV |
| 1992 |  | 23 | 54 | 75 | 118 | 0 | 270 |  |
| 1993 |  | 2 | 5 | 25 | 39 | 124 | 195 |  |
| 1994 |  | 3 | 9 | 38 | 0 | 1 | 51 |  |
| 1995 |  | 5 | 6 | 120 | 23 | 3 | 157 |  |
| 1996 |  | 0 | 0 | 15 | 23 | 27 | 65 |  |
| 1997 |  | 0 | 2 | 53 | 0 | 0 | 55 |  |
| 1998 |  | 1 | 1 | 0 | 47 | 50 | 99 |  |
| 1999 |  | 29 | 28 | 1 | 17 | 1 | 76 |  |
| 2000 |  | 226 | 130 | 21 | 9 | 2 | 388 |  |
| 2001 |  | 140 | 155 | 56 | 178 | 98 | 627 |  |
| 2002 |  | 67 | 128 | 41 | 1489 | 42 | 1767 |  |
| 2003 |  | 444 | 323 | 264 | 453 | 118 | 1602 |  |
| 2004 |  | 542 | 53 | 176 | 680 | 685 | 2136 |  |
| New Survey Gear Introduced |  |  |  |  |  |  |  |  |
| 2005 | 38 | 69 | 364 | 458 | 1084 | 1141 | 3155 | 26 |
| 2006 | 114 | 62 | 677 | 537 | 5131 | 525 | 7046 | 64 |
| 2007 | 247 | 387 | 872 | 1562 | 628 | 659 | 4355 | 31 |
| 2008 | 413 | 377 | 2046 | 929 | 1633 | 3227 | 8625 | 28 |
| 2009 | 208 | 230 | 1251 | 711 | 439 | 253 | 3092 | 14 |
| 2010 | 180 | 263 | 999 | 543 | 2426 | 908 | 5319 | 22 |
| 2011 |  | 1569 | 9654 | 408 | 5316 | 191 | 17140 | 14 |
| 2012 |  | 1932 | 2938 | 1125 | 464 | 14103 | 20562 | 69 |
| 2013 | 2395 | 2692 | 3960 | 1732 | 4551 | 19017 | 34345 | 53 |
| 2014 |  | 2639 | 2305 | 56061 | 2511 | 21381 | 84897 | 64 |

Table 14.3.1.4: Abundance indices ('000) by year-class/age from the Greenland Shrimp and Fish survey in West Greenland (NAFO 1A-1E).

## West Greenland

| Year/age 0 |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | $10+$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2005 | 134 | 815 | 10247 | 1604 | 1514 | 186 | 35 | 2 | 0 | 0 | 0 |
| 2006 | 249 | 6543 | 3577 | 12677 | 3395 | 401 | 47 | 16 | 0 | 0 | 0 |
| 2007 | 152 | 270 | 13792 | 3439 | 1934 | 37 | 4 | 0 | 0 | 0 | 0 |
| 2008 | 31 | 3472 | 2692 | 18780 | 4904 | 868 | 121 | 44 | 0 | 0 | 0 |
| 2009 | 0 | 124 | 9442 | 1666 | 1717 | 326 | 3 | 0 | 0 | 0 | 0 |
| 2010 | 209 | 2703 | 2094 | 10566 | 1252 | 775 | 42 | 7 | 0 | 0 | 0 |
| 2011 | 19 | 4940 | 71837 | 4453 | 3735 | 391 | 175 | 0 | 0 | 0 | 0 |
| 2012 | 0 | 204 | 11264 | 31593 | 3648 | 2427 | 116 | 7 | 0 | 0 | 0 |
| 2013 | 0 | 2904 | 8912 | 15168 | 36226 | 5665 | 848 | 142 | 22 | 25 | 0 |
| 2014 | 0 | 471 | 4792 | 8088 | 56469 | 35839 | 2597 | 1718 | 125 | 35 | 11 |

Table 14.3.1.5 Abundance indices ('000) by age from the Greenland Shrimp and Fish survey in West Greenland by NAFO divisions, 2014.

| West Greenland | 2014 | 2013 | 2012 | 2011 | 2010 | 2009 | 2008 | 2007 | 2006 | $2005<2005$ |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year-class | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | $10+$ |
| Age |  |  |  |  |  |  |  |  |  |  |  |
| Div. 0A | 0 | 135 | 2328 | 25321323 | 352 | 148 | 47 | 34 | 0 | 11 |  |
| Div. 1A | 0 | 336 | 1308 | 20321451 | 494 | 48 | 0 | 0 | 0 | 0 |  |
| Div. 1B | 0 | 0 | 101528264929724238 | 1245 | 233 | 0 | 0 | 0 |  |  |  |
| Div. 1C | 0 | 0 | 140 | 306 | 994 | 843 | 41 | 119 | 13 | 0 | 0 |
| Div. 1D | 0 | 0 | 0 | 391 | 3403 | 9912 | 1115 | 131978 | 35 | 0 |  |
| Div. 1E |  |  |  |  |  |  |  |  |  |  |  |

Table 14.3.1.6 Biomass of ages $4+$ of cod from the Greenland Shrimp and Fish survey in West Greenland (NAFO 1A-1E).

| Year | B4+ |
| :--- | :--- |
| 2005 | 1197 |
| 2006 | 2057 |
| 2007 | 1233 |
| 2008 | 3588 |
| 2009 | 1512 |
| 2010 | 1882 |
| 2011 | 3152 |
| 2012 | 5879 |
| 2013 | 28671 |
| 2014 | 83027 |

Table 14.3.2.1 German survey. Numbers of valid hauls by stratum in West Greenland (NAFO 1C-E).

|  | NAFO 1 C |  | NAFO 1D |  | NAFO 1E |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | Str 1.1 | Str 1.2 | Str 2.1 | Str 2.2 | Str 3.1 | Str 3.2 | Sum |
| 1981 | 1 | 1 | 13 | 2 | 3 | 1 | 21 |
| 1982 | 20 | 11 | 16 | 7 | 9 | 6 | 69 |
| 1983 | 26 | 11 | 25 | 11 | 17 | 5 | 95 |
| 1984 | 25 | 13 | 26 | 8 | 19 | 6 | 97 |
| 1985 | 10 | 8 | 26 | 10 | 17 | 5 | 76 |
| 1986 | 27 | 9 | 21 | 9 | 16 | 7 | 89 |
| 1987 | 25 | 19 | 21 | 4 | 18 | 4 | 91 |
| 1988 | 34 | 21 | 28 | 5 | 18 | 5 | 111 |
| 1989 | 25 | 14 | 30 | 9 | 8 | 3 | 89 |
| 1990 | 19 | 7 | 23 | 8 | 16 | 3 | 76 |
| 1991 | 19 | 11 | 23 | 7 | 13 | 6 | 79 |
| 1992 | 6 | 6 | 6 | 5 | 6 | 6 | 35 |
| 1993 | 9 | 7 | 9 | 6 | 10 | 8 | 49 |
| 1994 | 16 | 13 | 13 | 8 | 10 | 6 | 66 |
| 1995 | . | . | 3 | . | 10 | 7 | 20 |
| 1996 | 5 | 5 | 8 | 5 | 12 | 5 | 40 |
| 1997 | 5 | 6 | 5 | 5 | 6 | 5 | 32 |
| 1998 | 9 | 5 | 10 | 7 | 11 | 6 | 48 |
| 1999 | 8 | 7 | 14 | 8 | 13 | 6 | 56 |
| 2000 | 13 | 6 | 15 | 6 | 14 | 5 | 59 |
| 2001 | $\cdot$ | - | 15 | 7 | 15 | 5 | 42 |
| 2002 | - | - | 7 | 2 | 5 | 6 | 20 |
| 2003 | . | . | 7 | 6 | 7 | 7 | 27 |
| 2004 | 8 | 8 | 11 | 9 | 9 | 5 | 50 |
| 2005 | . | . | 9 | 7 | 8 | 6 | 30 |
| 2006 | 6 | 5 | 7 | 5 | 7 | 7 | 37 |
| 2007 | 5 | 5 | 7 | 5 | 6 | 5 | 33 |
| 2008 | 5 | - | 7 | 7 | 7 | 9 | 35 |
| 2009 | 2 | - | 5 | 5 | 6 | 6 | 24 |
| 2010 | 5 | 5 | 10 | 5 | 7 | 9 | 41 |
| 2011 | - | . | 5 | 5 | 5 | 5 | 20 |
| 2012 | 5 | 5 | 10 | 8 | 9 | 7 | 44 |
| 2013 | 6 | 6 | 8 | 6 | 10 | 7 | 43 |
| 2014 | 5 | 5 | 10 | 8 | 10 | 7 | 45 |

Table 14.3.2.2 German survey. Cod abundance indices ('000) from the German survey in West Greenland (NAFO 1C-1E) by year and stratum.

|  | NAFO 1 C |  | NAFO 1D |  | NAFO 1 E |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | str1_1 | str1_2 | str2_1 | str2_2 | str3_1 | str3_2 | Sum | SD |
| 1982 | 2364 | 408 | 27594 | 920 | 7401 | 1801 | 40488 | 18605 |
| 1983 | 177 | 196 | 7079 | 2230 | 8678 | 1230 | 19590 | 7266 |
| 1984 | 189 | 90 | 2524 | 98 | 2666 | 364 | 5931 | 3629 |
| 1985 | 8094 | 1107 | 7237 | 2348 | 4984 | 840 | 24610 | 10809 |
| 1986 | 14716 | 630 | 22985 | 108 | 16570 | 609 | 55618 | 29631 |
| 1987 | 173517 | 482 | 115172 | 3790 | 72349 | 186 | 365496 | 331763 |
| 1988 | 46027 | 1106 | 186523 | 43090 | 21037 | 51 | 297834 | 216925 |
| 1989 | 1362 | 483 | 16280 | 325 | 129005 | 678 | 148133 | 65933 |
| 1990 | 619 | 299 | 2279 | 235 | 3827 | 61 | 7320 | 5462 |
| 1991 | 142 | 116 | 88 | 92 | 474 | 387 | 1299 | 412 |
| 1992 | 274 | 334 | 72 | 127 | 57 | 38 | 902 | 314 |
| 1993 | 327 | 243 | 105 | 109 | 53 | 21 | 858 | 195 |
| 1994 | 95 | 53 | 16 | 17 | 34 | 11 | 226 | 79 |
| 1995 | . | . | 27 | . | 72 | 34 | 133 | 60 |
| 1996 | 82 | 70 | 42 | 20 | 65 | 0 | 279 | 80 |
| 1997 | 0 | 24 | 17 | 0 | 57 | 3 | 101 | 45 |
| 1998 | 793 | 0 | 23 | 28 | 7 | 0 | 851 | 573 |
| 1999 | 103 | 33 | 33 | 11 | 197 | 7 | 384 | 171 |
| 2000 | 205 | 250 | 50 | 174 | 288 | 9 | 976 | 383 |
| 2001 | . | . | 584 | 36 | 3020 | 9 | 3649 | 3481 |
| 2002 | . | . | 238 | 21 | 342 | 23 | 624 | 257 |
| 2003 | . | . | 625 | 99 | 1625 | 73 | 2422 | 945 |
| 2004 | 503 | 213 | 1522 | 123 | 2709 | 638 | 5708 | 1592 |
| 2005 | . | . | 1586 | 264 | 5666 | 419 | 7935 | 3115 |
| 2006 | 495 | 485 | 87439 | 858 | 4481 | 1323 | 95081 | 99523 |
| 2007 | 1430 | 3261 | 3417 | 687 | 9861 | 71 | 18727 | 8645 |
| 2008 | 2666 | . | 916 | 911 | 23527 | 616 | 28636 | 26712 |
| 2009 | 72 | . | 1370 | 850 | 1068 | 378 | 3738 | 879 |
| 2010 | 2644 | 464 | 4451 | 631 | 5148 | 274 | 13612 | 6231 |
| 2011 | . | . | 716 | 375 | 1242 | 337 | 2670 | 782 |
| 2012 | 99609 | 1253 | 6007 | 442 | 8455 | 1251 | 117017 | 68441 |
| 2013 | 4457 | 1585 | 20122 | 221 | 7138 | 252 | 33775 | 22438 |
| 2014 | 9952 | 2008 | 28102 | 413 | 1261 | 86 | 41822 | 38616 |

Table 14.3.2.3 German survey, Cod biomass indices (tons) from the German survey in West Greenland (NAFO 1C-1E) by year and stratum.

|  | NAFO 1C |  | NAFO 1D |  | NAFO 1 E |  |  | SD |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | str1_1 | str1_2 | str2_1 | str2_2 | str3_1 | str3_2 | Sum |  |
| 1982 | 1113 | 163 | 37404 | 1280 | 9970 | 4483 | 54413 | 26014 |
| 1983 | 144 | 87 | 9052 | 3381 | 12953 | 5015 | 30632 | 10295 |
| 1984 | 406 | 104 | 3998 | 137 | 3643 | 551 | 8839 | 5507 |
| 1985 | 1046 | 112 | 6543 | 1181 | 4700 | 506 | 14088 | 18209 |
| 1986 | 4858 | 254 | 11787 | 36 | 12381 | 651 | 29967 | 13885 |
| 1987 | 148896 | 156 | 93292 | 2446 | 54178 | 107 | 299075 | 299459 |
| 1988 | 47085 | 579 | 190073 | 39548 | 19663 | 54 | 297002 | 227428 |
| 1989 | 384 | 124 | 15061 | 211 | 113614 | 710 | 130104 | 55334 |
| 1990 | 130 | 66 | 1948 | 123 | 3652 | 56 | 5975 | 4986 |
| 1991 | 45 | 38 | 36 | 28 | 549 | 374 | 1070 | 529 |
| 1992 | 65 | 104 | 15 | 33 | 10 | 7 | 234 | 97 |
| 1993 | 77 | 45 | 27 | 27 | 30 | 6 | 212 | 53 |
| 1994 | 13 | 17 | 3 | 12 | 11 | 5 | 61 | 17 |
| 1995 | . | . | 14 | . | 13 | 7 | 34 | 12 |
| 1996 | 13 | 35 | 12 | 11 | 28 | 0 | 99 | 29 |
| 1997 | 0 | 21 | 11 | 0 | 50 | 3 | 85 | 43 |
| 1998 | 38 | 0 | 1 | 7 | 1 | 0 | 47 | 25 |
| 1999 | 16 | 11 | 6 | 3 | 63 | 5 | 104 | 57 |
| 2000 | 54 | 71 | 11 | 83 | 73 | 5 | 297 | 117 |
| 2001 | - | . | 163 | 17 | 1024 | 5 | 1209 | 1212 |
| 2002 | . | . | 89 | 16 | 136 | 7 | 248 | 108 |
| 2003 | . | . | 98 | 44 | 736 | 32 | 910 | 461 |
| 2004 | 172 | 83 | 274 | 45 | 547 | 186 | 1307 | 342 |
| 2005 | - | - | 605 | 124 | 1796 | 146 | 2671 | 1057 |
| 2006 | 102 | 138 | 45616 | 250 | 2046 | 614 | 48766 | 52298 |
| 2007 | 319 | 885 | 1579 | 244 | 7804 | 43 | 10874 | 7524 |
| 2008 | 872 | . | 193 | 206 | 11479 | 175 | 12925 | 13686 |
| 2009 | 19 | . | 309 | 293 | 372 | 153 | 1146 | 255 |
| 2010 | 1012 | 244 | 2234 | 312 | 2703 | 173 | 6678 | 3057 |
| 2011 | . | - | 189 | 128 | 1040 | 194 | 1551 | 602 |
| 2012 | 52497 | 588 | 4185 | 240 | 8203 | 848 | 66561 | 35693 |
| 2013 | 2703 | 1670 | 17316 | 142 | 11251 | 544 | 33626 | 18801 |
| 2014 | 10597 | 2154 | 35741 | 422 | 3561 | 397 | 52872 | 47451 |

Table 14.3.2.4 German survey, West Greenland (NAFO 1C-1D). Age disaggregated abundance indices ('1000).

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | $11+$ | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 |  | 77 | 505 | 14266 | 5195 | 14798 | 4144 | 908 | 178 | 344 | 35 | 34 | 40484 |
| 1983*) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1984 | 80 | 3 | 13 | 709 | 604 | 3495 | 289 | 628 | 32 | 61 | 13 | 0 | 5927 |
| 1985 | 202 | 16823 | 623 | 330 | 2271 | 1100 | 2982 | 112 | 164 | 2 | 3 | 0 | 24612 |
| 1986 |  | 3600 | 45772 | 1686 | 321 | 2386 | 652 | 1098 | 22 | 74 | 3 | 1 | 55615 |
| 1987 |  | 147 | 22578 | 318948 | 13977 | 2930 | 4603 | 649 | 1506 |  | 131 | 13 | 365482 |
| 1988 |  | 124 | 1357 | 44364 | 247618 | 2660 | 311 | 521 | 318 | 529 | 12 | 15 | 297829 |
| 1989 | 0 | 163 | 1293 | 3821 | 79642 | 62126 | 1008 |  | 47 | 7 | 24 | 0 | 148131 |
| 1990 | 11 | 17 | 595 | 1242 | 368 | 4089 | 990 | 6 | 0 | 0 |  | 1 | 7319 |
| 1991 |  | 86 | 94 | 193 | 350 | 36 | 461 | 57 | 2 |  |  | 0 | 1279 |
| 1992 |  | 88 | 672 | 100 | 17 | 25 |  | 0 |  |  |  | 0 | 902 |
| 1993 |  | 8 | 499 | 318 | 12 | 21 |  |  |  |  |  | 0 | 858 |
| 1994 |  | 98 | 18 | 90 | 14 | 3 |  | 2 |  |  |  | 0 | 225 |
| 1995 |  |  | 111 | 6 | 16 |  |  |  |  |  |  | 0 | 133 |
| 1996 |  | 76 | 6 | 193 | 5 |  | 0 |  |  |  |  | 0 | 280 |
| 1997 |  | 6 | 13 | 7 | 76 |  |  |  |  |  |  | 0 | 102 |
| 1998 | 0 | 845 |  | 3 | 3 | 0 |  |  |  |  |  | 0 | 851 |
| 1999 | 8 | 165 | 166 | 36 | 3 |  | 3 |  |  |  |  | 0 | 381 |
| 2000 |  | 60 | 524 | 328 | 62 |  |  |  |  |  |  | 0 | 974 |
| 2001 |  | 266 | 2753 | 527 | 65 | 20 |  |  |  |  |  | 0 | 3631 |
| 2002 | 0 | 6 | 309 | 290 | 17 |  |  |  |  |  |  | 0 | 622 |
| 2003 |  | 1368 | 205 | 511 | 284 | 36 | 9 |  |  |  |  | 0 | 2413 |
| 2004 | 132 | 3078 | 2008 | 307 | 108 | 55 | 15 | 0 |  |  |  | 0 | 5703 |
| 2005 | 91 | 156 | 6893 | 653 | 40 | 16 | 14 | 0 | 0 |  |  | 0 | 7863 |
| 2006 | 157 | 1949 | 6961 | 83106 | 2708 | 45 | 51 | 67 | 0 |  |  | 0 | 95044 |
| 2007 | 139 | 229 | 9402 | 1655 | 6989 | 227 | 35 | 38 | 12 |  |  | 0 | 18726 |
| 2008 | 8 | 1224 | 2317 | 20080 | 3747 | 1235 | 20 | 3 | 2 | 0 | 0 | 0 | 28636 |
| 2009 | 36 | 326 | 2513 | 363 | 406 | 37 | 40 | 14 |  |  |  | 0 | 3735 |
| 2010 | 208 | 1531 | 1726 | 9201 | 577 | 259 | 51 | 48 | 3 | 3 |  | 5 | 13612 |
| 2011 |  | 195 | 1572 | 385 | 368 | 68 | 33 | 26 | 24 | 0 | 0 | 0 | 2671 |
| 2012 | 142 | 1191 | 37872 | 66947 | 7682 | 2847 | 227 | 76 | 8 | 18 |  | 0 | 117010 |
| 2013 |  | 152 | 1562 | 12824 | 15859 | 1783 | 1135 | 234 | 86 | 23 | 18 | 4 | 33680 |
| 2014 |  |  | 880 | 4629 | 17021 | 17863 | 1080 | 277 | 32 | 0 | 4 | 0 | 41786 |

*) calculated proportionally using age compositions reported by the ICES Working Group on Cod Stocks off East Greenland (ICES, 1984).

Table 14.5.1. Number of tagged cod in the period of 2003 to 2014 in different regions. Bank (West) = NAFO division 1D+1E. East Greenland = NAFO division 1F + ICES division XIVb.

| Tagged |  |  |  |
| :--- | :--- | :--- | :--- |
| Year | Fjord | Bank (West) | East Greenland |
| 2003 | 599 |  |  |
| 2004 | 658 |  |  |
| 2005 | 565 |  | 1387 |
| 2006 | 41 | 721 | 1296 |
| 2007 | 1140 |  | 525 |
| 2008 | 231 |  |  |
| 2009 | 633 |  | 403 |
| 2010 | 88 | 2359 |  |
| 2011 | 28 |  | 1203 |
| 2012 | 86 |  |  |
| 2013 | 183 |  |  |
| 2014 |  |  |  |

Table 14.5.2: Number of recaptured cod in the period of 2003 to 2014 in different regions. Bank $($ West $)=$ NAFO division 1D $+1 E$. East Greenland = NAFO division 1F + ICES division XIVb.

| Recaptures |  |  |  |
| :--- | :--- | :--- | :--- |
|  | Fjord (West) | Bank (West) | East Greenland |
| Fjord (West) | 436 | 8 | 3 |
| Bank (West) |  | 21 | 2 |
| East Greenland |  | 5 | 89 |
| Fjord (East) |  | 13 | 1 |
| Iceland | 3 |  | 98 |



Figure. 14.1. Sampling location of spawning cod in Greenland and Iceland in the genetic project. The colours of the dots represent the blends of sample mean of the different spawning population: West offshore, Nuuk (inshore), East (Greenland and offshore Iceland) and Iceland inshore as signal intensities of green and red respectively. After Therkildsen et al. 2013.


Figure 14.2.1. Annual catch of cod in offshore West Greenland (NAFO subdivisions 1A-1E) used by the Working Group.


Figure 14.2.2.1: Annual distribution of total catches of Atlantic cod in West and East Greenland. Q1-Q6 illustrates survey areas (strata) in the East Greenland shrimp and fish survey.


Figure 14.2.2.1: Continued. Annual distribution of total catches of Atlantic cod in West and East Greenland. Q1-Q6 illustrates survey areas (strata) in the East Greenland shrimp and fish survey.


Figure 14.2.2.2: Distribution of Longline and Trawl catches of Atlantic cod in West and East Greenland 2014. Q1-Q6 illustrates survey areas (strata) in the East Greenland shrimp and fish survey.


Figure 14.2.3.1: Total length and age distributions of commercial cod catches in the West Greenland (NAFO 1A-1E) offshore fishery in 2014.


Figure 14.2.3.2: Length and age distributions of commercial cod catches in different NAFO divisions in West Greenland in 2014. NAFO 1C is furthest to the north.


Figure14.3.1.1. Greenland shrimp and fish survey 2008-2014. Abundance per $\mathrm{Km}^{2}$


Figure 14.3.1.2. Greenland shrimp and fish survey 2008-2014. Catch weight kg per $\mathrm{Km}^{2}$

## Survey CAA 1A-1E



Figure 14.3.1.3: Abundance index by age in NAFO 1A-1E combined. Size of circles represents index value.


Figure 14.3.1.4: West Greenland Shrimp and fish survey, 2014. Abundance index by length (cm) and area. Areas from north (top) to south (bottom ) are: NAFO div. 1A; 1B+0A; 1C, 1D, 1E.

































 6.



Figure 14.3.1.5. Abundace ( $\mathrm{no} / \mathrm{km}^{2}$ ) pr. station of ages 1-10 in the years 2008-2014.


Figure 14.3.1.6: Total abundance indices by length in West Greenland shrimp and fish survey (NAFO 1A-1E).


Figure 14.3.1.7: Estimated SSB (tons) by NAFO subdivisions from the West Greenland Shrimp and Fish survey, 2014. Maturity taken from proportion mature by length as recorded on observer trips off East Greenland in 2007.


Figure 14.3.1.8: Estimated SSB (tons) by year from the West Greenland Shrimp and Fish survey (NAFO 1A-1E).

Abundance distribution Atlantic cod


## German survey 2014

Figure 14.3.2.1 German survey, 2014. Abundance (num per km2) pr haul.

Biomass distribution Atlantic cod



German survey 2014
Figure 14.3.2.2 German survey, 2014. Biomass (kg per km2) pr haul.


Figure 14.3.2.3 German survey, Cod off Greenland. Abundance per age group and stratum. Strata 1 -4 is West Greenland from north to south; strata 5-9 is East Greenland from south to north.


Figure 14.3.2.4 German survey, Cod off Greenland. Abundance indices for West Greenland (NAFO subdivisions 1C-1E).


Figure 14.3.2.5 German survey, Cod off Greenland. Biomass indices for West Greenland (NAFO subdivisions 1C-1E).

## 15 Cod in inshore waters of NAFO Subarea 1 (Greenland cod)

## Executive Summary

Total catches from the inshore fishery were 18331 t in 2014 which is the highest since early 1990'ies. Several year-classes were caught in the fishery but catches were dominated by the 2009 YC .
he mean length in the fisheries has increases from 44 cm in 2006 to 58 cm in 2014 . Survey recruitment indices from the inshore area show that incoming year classes (2011 and 2012) are below average.

The stock was benchmarked in 2015 and a new procedure for making catch advice was adopted. The procedure is based on a linear regression on pairs of survey values (ages $3-8)$ and catches in the following year. The advice is based on the average of the 2013 and 2014 survey values for ages 3-8 multiplied by a scaling factor.

### 15.1 The fishery

Details on the historical development in the fisheries are provided in the stock annex.

### 15.1.1 The present fishery

The original TAC for the coastal fishery in 2014 was set at 15000 tons. During the season first 1500 tons was added in October and another 2000 tons were added in November resulting in a total TAC of 18500 t . Further a dispensation was given to three small vessels to fish offshore on the inshore quota. Only one vessel used its license and caught 166 t .
The coastal fishery took 18400 tons in 2014, which is an increase of $39 \%$ compared to 2013 (table 15.1.1.1, figure 15.1.1.1). The majority of the catches ( $79 \%$ ) were taken in Mid Greenland in NAFO Div. 1B, 1C and 1D (table 15.1.1.1, figure 15.1.1.2). The most important fishery is the pound net fishery that takes place during summer followed by the fishery with jigs that takes place in autumn (table 15.1.1.2 and 15.1.1.3). In 2014 half of the total catch was taken by pound net which is a decrease compared to 2013 where $2 / 3$ of the total catch where taken by pound net and 2012 where $3 / 4$ of the total catch was taken by pound net (Fig. 15.1.1.1). Since 2012 jigs have become more dominant from $7 \%$ of the total catch in 2012 to $25 \%$ of the total catch in 2014. Gillnets and longlines constitutes the rest of the total catch.
The commercial fishery for the inshore cod is carried out along the entire coastline of West Greenland from Disko bay to Cap Farewell (Figure 15.1.1.3). Gillnets and Jigs are used more often in mid Greenland (NAFO 1B and1C) compared to other areas where they comprised $29 \%$ of the total catches (table 15.1.1.3). The pound-net fishery is the dominating gear (especially in the Nuuk area, NAFO 1D), although to a lesser extent in Disko bay (NAFO 1A, figure 15.1.1.3) due to the fishing industry being concentrated on Greenland Halibut, therefore cod is mostly caught as bycatch in the longline and gillnet settings for Greenland Halibut. Cod catches in the Disko Bay area, especially in the southern part of Disko Bay, have however increased in recent years and $17 \%$ of the total inshore catches was taken in Disko Bay in 2014. The cod fishery is now larger in Disko Bay than in South Greenland (NAFO 1E and 1F). The catches in South Greenland were the lowest recorded since 2004 (table 15.1.1.1) and comprised only $0.4 \%$ of the total inshore catches in 2014 (table 15.1.1.2).

### 15.1.2 Length, weight and age distributions

In 2014 the Greenland inshore length frequencies were measured from 49 inshore samples (6 446 cod measured).
everal year-classes were caught in the inshore fishery in 2014 and ages 4-6 (YC 20082010) comprised the catches in 2014, with the 2009 YC dominating the catches (figure 15.1.2.1, table 15.1.2.1). Mean length in catches have increased from 53 cm in 2010-2013 to 58 cm in 2014. This increase is caused by the 2009 YC being the dominating YC at age 4 in 2013 and age 5 in 2014. The 2010 YC does not seem to be as abundant as the 2009 YC and therefore the mean length has increased (figure 15.1.2.2).

### 15.1.3 Information on spawning

In 2011 a survey was conducted in spring in order to investigate the extent of spawning in fjords not traditionally surveyed. The results show that spawning occurs in most fjords and is especially pronounced between Sisimiut (NAFO 1B) and Paamiut (NAFO 1E).

### 15.1.4 Results of the West Greenland gillnet survey

The numbers of valid net settings in 2014 was 41 in NAFO 1B and 60 in NAFO 1D (Table 15.1.4.1). Area and site specific catch rates can be seen in Fig. 15.1.4.1.
In 1B age 2 and 3 fish ( 2011 and $2012 \mathrm{YC}^{\prime}$ s), which the survey mainly targets, appear to be small cohorts, and are smaller than the time series mean (Table 15.1.4.2, Fig. 15.1.4.2). The 2009 YC that has been large the most resent years in 1B is no longer observed as a particular strong YC at the age of 5 (Table 15.1.4.2, Fig. 15.1.4.3). Overall, the NAFO 1B index (including all age groups) declined from 2013 by $64 \%$.

The 2014 catches in NAFO 1D were dominated by 3 year old cod (2011 YC, Table 15.1.4.2). Catch rates of this YC was the second highest in the time series. The 2009 YC was not an outstanding cohort at age 2 and 3 in 2011 and 2012 in 1D but the index increased for age 4 in 2013 and was one of the highest recorded indices in 2013, and at age 5 in 2014. The overall index for NAFO 1D (including all age groups) is the second highest in the time series, but decreased by $26 \%$ compared to 2013.

Combining the two NAFO divisions in a joint index shows an overall decline of 49 \% in total index for all ages from 2013 and it has not been as low since 2006 (Fig.15.1.4.4). This overall trend is driven by the development in NAFO 1B, where the catch rates and index values are normally higher than 1D (Table 15.1.4.2, Fig. 15.1.4.3). However this has changed in 2014 with the total index being higher in 1D. This is caused by the index being higher in 1D for especially ages 3 and 4 (2011 and 2010 YC). The combined index for 1B and 1D for age 2 and 3 jointly has decreased by $61 \%$ compared to the average of the preceding four years.

The combined 1B and 1D index for ages 3-8 that are used together with catch data for making catch advice for the inshore management area (ICES 2015) is very similar to the index for all age groups (Fig. 15.1.4.4) and has also decreased compared to the most recent years. But the decrease is more moderate, $41 \%$ compared to 2013 and $30 \%$ compared to the preceding four years.
Recruitment in all three areas (one is no longer surveyed) declined from the start of the survey period until the late 1990's where recruitment was very low in the inshore areas. Around 2000, recruitment started to increase, and has been stable and occasionally very high since (Table 15.1.4.2).

In NAFO 1B, namely the 2009 YC resulted in high catch rates and the combined survey index was the highest observed in the years it was dominant. The catch rates of this YC in 1B was the highest recorded in the time series at ages 1, 2 and 3, but was not an outstanding cohort at age 5 in 2014. In 2014 the YC was "gone" from 1B which can be an effect of the survey design (bad at catching older fish) and some emigration to offshore areas which was seen for the 1984 YC. In 1D the 2009 YC has not been extraordinarily large before it reached the age of 4 in 2013 and in 2014 it still remains large. This might be an effect of immigration from the offshore area to the inshore area as the 2009 YC is also registered as a large year class offshore (Retzel 2015, a) and a southwards migration from the northern inshore regions (1B).

As the larger cohorts have disappeared from the survey, especially in NAFO 1B, and no new cohorts are entering, the index in 1B has decreased to a very low level compared to the increasing trend which has been the case in recent years. Historically, the combined index has mainly been driven by dynamics in 1B. The 2 and 3 year old fish (2011 and 2012 YC ) in 1D are well above average recruitment. However, in 1B it was substantially lower, leading to a reduction of the combined index compared to 20102013.

### 15.1.5 State of the stock

There have been several years of steady and relatively high recruitment and the biomass estimate is increasing and has been doing so for more than ten years. Several year classes are in the catches, and the large 2009 YC has now entered the fishery. However after the 2009 YC has entered the fishery no new incoming yearclasses of the same size has been observed. Spawning has been documented in most fjords on the west coast, with key areas in NAFO 1B and 1D. Hence the overall state of the stock is considered good, but the lack of incoming large yearclasses is cause for concern.

### 15.1.6 Implemented management measures for 2015

Until 2009 the inshore fishery was unregulated by a TAC. The TAC in 2009-2014 can be seen in figure 15.1.1.2. The TAC for 2014 is set at 25000 t . No other management measures have been taken.

### 15.1.7 Management plan

No management plan currently exists for the inshore cod stock.

### 15.1.8 Management considerations

When managing this stock, it should be taken into consideration that the inshore cod tend to form very dense spawning aggregations in limited areas. It could be considered to limit the fishery in certain areas or certain periods, especially if the stock shows a declining trend. These areas include specifically certain areas in the Nuuk and Sisimiut fjord systems.

Genetic and tagging results indicate limited migration between fjords and management should therefore ensure that not all catches are taken in a limited area. This is especially important in areas that are considered to have maintained the stocks in periods of overall stock decline in Greenland (i.e. Nuuk and Sisimiut fjords).

### 15.1.9 Basis for advice

The survey index in a given year was related to the catch in the next year (Figure D.1.1). The advice is then based on the survey index multiplied by a factor. The validity of this
approach rests on a number of assumptions. Among others, the fishery has been at a stable sustainable level (ideally the same across years). Based on model outputs and catch curves (Hedeholm and Post, 2015) this seems to be a reasonable assumption, at least during the last 15 years. Some years in the 1980s did not follow the overall trend, and were most likely subjected to a very high fishing intensity and a very high offshore input to the fishery, and these years are therefore excluded from the regression analyses. The fish enter the fishery at age 4 . Accordingly the survey index of ages 3-8 was used to generate advice.


Figure D.1.1. Survey index of 3-8 year olds vs. the catch the following year. $\mathrm{r}^{2}=0.76$. Based on data from 1991-2014. Points are labelled by survey year.

Given that this approach is based on variable data a precautionary approach should be taken. So rather than having the regression pass through the origin, the intercept with the $x$-axis is set at a survey index value of 50 and slope is 37.9 . The survey tend to vary considerably between years, and to avoid having the advice fluctuate accordingly the average of the last two years survey index values were used when calculating the catch advice. Consequently, the advice is generated as follows:
$\mathrm{C}_{\mathrm{y}+1}=37.9$ * ( $\left.\mathrm{U}_{3-8 y}-\mathrm{U}_{\text {trigger }}\right)(1)$
where $U_{3-8 y}$ is the combined survey value for ages 3-8 and $U_{\text {trigger }}$ is 50 .

### 15.1.10 References

Retzel, A. 2015.a Greenland Shrimp and Fish survey results for Atlantic cod in NAFO subareas 1A-1E (West Greenland) in 2014. ICES North Western Working Group (NWWG) April 28May 5, 2015, WD 19.

Retzel, A. 2015.b Greenland commercial data for Atlantic cod in Greenland inshore waters for 2014. ICES North Western Working Group (NWWG) April 28- May 5, 2015, WD 18.

Retzel, A., Post, S. L. 2015. Greenland Shrimp and Fish survey results for Atlantic cod in ICES subarea XIVb (East Greenland) and NAFO subarea 1F (SouthWest Greenland) in 2014. ICES North Western Working Group (NWWG) April 28- May 5, 2015, WD 21.

Table 15.1.1.1. Cod catches ( t ) divided into NAFO divisions, caught in the inshore fishery (19111993: Horsted 2000, 1994-2006: ICES 2007, Statistic Greenland, 2007-present: Greenland Fisheries License Control). ICES XIVb=inshore East Greenland.

| NAFO divisions |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1A | 1B | 1 C | 1D | 1E | 1F | Unknown NAFO div | Total <br> WestGreenland | $\begin{aligned} & \text { ICES } \\ & \text { XIVb } \end{aligned}$ |
| 1911 |  |  |  | 19 |  |  |  | 19 |  |
| 1912 |  |  |  | 5 |  |  |  | 5 |  |
| 1913 |  |  |  | 66 |  |  |  | 66 |  |
| 1914 |  |  |  | 60 |  |  |  | 60 |  |
| 1915 |  | 47 | 6 | 45 |  |  |  | 98 |  |
| 1916 |  | 66 | 24 | 103 |  |  |  | 193 |  |
| 1917 |  | 67 | 28 | 59 |  |  |  | 154 |  |
| 1918 |  | 106 | 26 | 140 |  | 169 |  | 441 |  |
| 1919 |  | 39 | 37 | 140 | 148 | 137 |  | 501 |  |
| 1920 |  | 117 | 32 | 187 | 23 | 95 |  | 454 |  |
| 1921 |  | 116 | 92 | 97 | 7 | 196 |  | 508 |  |
| 1922 |  | 82 | 178 | 144 | 40 | 158 |  | 602 |  |
| 1923 |  | 120 | 116 | 147 | 0 | 307 |  | 690 |  |
| 1924 |  | 131 | 223 | 221 | 1 | 267 |  | 843 |  |
| 1925 |  | 122 | 371 | 318 | 45 | 168 |  | 1024 |  |
| 1926 |  | 97 | 785 | 673 | 170 | 499 |  | 2224 |  |
| 1927 |  | 282 | 974 | 982 | 305 | 1027 |  | 3570 |  |
| 1928 |  | 426 | 888 | 1153 | 497 | 1199 |  | 4163 |  |
| 1929 |  | 1479 | 1572 | 1335 | 642 | 2052 |  | 7080 |  |
| 1930 | 137 | 2208 | 2326 | 1681 | 994 | 2312 |  | 9658 |  |
| 1931 | 315 | 1905 | 2026 | 1520 | 835 | 2453 |  | 9054 |  |
| 1932 | 358 | 1713 | 2130 | 1042 | 731 | 3258 |  | 9232 |  |
| 1933 | 304 | 1799 | 1743 | 1148 | 948 | 2296 |  | 8238 |  |
| 1934 | 451 | 2080 | 1473 | 652 | 921 | 3591 |  | 9168 |  |
| 1935 | 524 | 1870 | 1277 | 769 | 670 | 2466 |  | 7576 |  |
| 1936 | 329 | 2039 | 1199 | 705 | 717 | 2185 |  | 7174 |  |
| 1937 | 135 | 1982 | 1433 | 854 | 496 | 2061 |  | 6961 |  |
| 1938 | 258 | 1743 | 1406 | 703 | 347 | 1035 |  | 5492 |  |
| 1939 | 416 | 2256 | 1732 | 896 | 431 | 1430 |  | 7161 |  |
| 1940 | 482 | 2478 | 1600 | 1061 | 646 | 1759 |  | 8026 |  |
| 1941 | 636 | 3229 | 1473 | 823 | 593 | 1868 |  | 8622 |  |
| 1942 | 879 | 3831 | 2249 | 1332 | 1003 | 2733 |  | 12027 |  |
| 1943 | 1507 | 5056 | 2016 | 1240 | 1134 | 2073 |  | 13026 |  |
| 1944 | 1795 | 4322 | 2355 | 1547 | 1198 | 2168 |  | 13385 |  |
| 1945 | 1585 | 4987 | 2844 | 1207 | 1474 | 2192 |  | 14289 |  |
| 1946 | 1889 | 5210 | 2871 | 1438 | 1139 | 2715 |  | 15262 |  |
| 1947 | 1573 | 5261 | 3323 | 2096 | 1658 | 4118 |  | 18029 |  |
| 1948 | 1130 | 5660 | 3756 | 1657 | 1652 | 4820 |  | 18675 |  |
| 1949 | 1403 | 4580 | 3666 | 2110 | 2151 | 3140 |  | 17050 |  |
| 1950 | 1657 | 6358 | 4140 | 2357 | 2278 | 4383 |  | 21173 |  |
| 1951 | 1277 | 5322 | 3324 | 2571 | 2101 | 3605 |  | 18200 |  |
| 1952 | 646 | 4443 | 2906 | 2437 | 2216 | 4078 |  | 16726 |  |
| 1953 | 1092 | 5030 | 3662 | 5513 | 3093 | 4261 |  | 22651 |  |

Table 15.1.1.1. continued

| Year | NAFO divisions |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1A | 1B | 1C | 1D | 1 E | 1F | Unknown NAFO div | Total <br> WestGreenland | $\begin{aligned} & \text { ICES } \\ & \text { XIVb } \end{aligned}$ |
| 1954 | 950 | 6164 | 3118 | 3275 | 1773 | 3418 |  | 18698 |  |
| 1955 | 591 | 5523 | 3225 | 4061 | 2773 | 3614 |  | 19787 |  |
| 1956 | 475 | 5373 | 3175 | 5127 | 3292 | 3586 |  | 21028 |  |
| 1957 | 277 | 6146 | 3282 | 5257 | 4380 | 5251 |  | 24593 |  |
| 1958 | 19 | 6178 | 3724 | 5456 | 3975 | 6450 |  | 25802 |  |
| 1959 | 237 | 6404 | 5590 | 5009 | 3767 | 6570 |  | 27577 |  |
| 1960 | 188 | 6741 | 6230 | 3614 | 3626 | 6610 |  | 27009 |  |
| 1961 | 601 | 6569 | 6726 | 4178 | 6182 | 9709 |  | 33965 |  |
| 1962 | 315 | 7809 | 6269 | 3824 | 5638 | 11525 |  | 35380 |  |
| 1963 | 295 | 4877 | 3178 | 2804 | 3078 | 9037 |  | 23269 |  |
| 1964 | 275 | 3311 | 2447 | 8766 | 2206 | 4981 |  | 21986 |  |
| 1965 | 325 | 5209 | 4818 | 6046 | 2477 | 5447 |  | 24322 |  |
| 1966 | 483 | 8738 | 5669 | 7022 | 2335 | 4799 |  | 29046 |  |
| 1967 | 310 | 5658 | 6248 | 6747 | 2429 | 6132 |  | 27524 |  |
| 1968 | 142 | 1669 | 2738 | 6123 | 2837 | 7207 |  | 20716 |  |
| 1969 | 57 | 1767 | 4287 | 7540 | 2017 | 5568 |  | 21236 |  |
| 1970 | 136 | 1469 | 2219 | 3661 | 2424 | 5654 |  | 15563 |  |
| 1971 | 255 | 1807 | 2011 | 3802 | 1698 | 3933 |  | 13506 |  |
| 1972 | 263 | 1855 | 3328 | 3973 | 1533 | 3696 |  | 14648 |  |
| 1973 | 158 | 1362 | 1225 | 3682 | 1614 | 1581 |  | 9622 |  |
| 1974 | 454 | 926 | 1449 | 2588 | 1628 | 1593 |  | 8638 |  |
| 1975 | 216 | 1038 | 1930 | 1269 | 964 | 1140 |  | 6557 |  |
| 1976 | 204 | 644 | 1224 | 904 | 1367 | 831 |  | 5174 |  |
| 1977 | 216 | 580 | 2505 | 2946 | 3521 | 4231 |  | 13999 |  |
| 1978 | 348 | 1587 | 3244 | 2614 | 4642 | 7244 |  | 19679 |  |
| 1979 | 433 | 1768 | 2201 | 6378 | 9609 | 15201 |  | 35590 |  |
| 1980 | 719 | 2303 | 2269 | 7781 | 10647 | 14852 |  | 38571 |  |
| 1981 | 281 | 2810 | 3599 | 6119 | 7711 | 11505 | 7678 | 39703 |  |
| 1982 | 206 | 2448 | 3176 | 7186 | 4536 | 3621 | 5491 | 26664 |  |
| 1983 | 148 | 2803 | 3640 | 7430 | 5016 | 2500 | 7205 | 28742 |  |
| 1984 | 175 | 3908 | 1889 | 5414 | 1149 | 1333 | 6090 | 19958 |  |
| 1985 | 149 | 2936 | 957 | 1976 | 1178 | 1245 |  | 8441 |  |
| 1986 | 76 | 1038 | 255 | 1209 | 1456 | 1268 |  | 5302 |  |
| 1987 | 77 | 2366 | 423 | 6407 | 3602 | 1326 | 403 | 14604 |  |
| 1988 | 333 | 6294 | 1342 | 2992 | 3346 | 4484 |  | 18791 |  |
| 1989 | 634 | 8491 | 5671 | 8212 | 10845 | 4676 |  | 38529 |  |
| 1990 | 476 | 9857 | 1482 | 9826 | 1917 | 5241 |  | 28799 |  |
| 1991 | 876 | 8641 | 917 | 2782 | 1089 | 4007 |  | 18312 |  |
| 1992 | 695 | 2710 | 563 | 1070 | 239 | 450 |  | 5727 |  |
| 1993 | 333 | 327 | 168 | 970 | 19 | 109 |  | 1926 |  |
| 1994 | 209 | 332 | 589 | 914 | 11 | 62 |  | 2117 |  |
| 1995 | 53 | 521 | 710 | 332 | 4 | 81 |  | 1701 |  |
| 1996 | 41 | 211 | 471 | 164 | 11 | 46 |  | 944 |  |
| 1997 | 18 | 446 | 198 | 99 | 13 | 130 | 282 | 1186 |  |
| 1998 | 9 | 118 | 79 | 78 | 0 | 38 |  | 322 |  |

Table 15.1.1.1. continued

| Year | NAFO divisions |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1A | 1B | 1C | 1D | 1E | 1F | Unknown NAFO div | Total WestGreenland | $\begin{aligned} & \text { ICES } \\ & \text { XIVb } \end{aligned}$ |
| 1999 | 68 | 142 | 55 | 336 | 8 | 4 |  | 613 |  |
| 2000 | 154 | 266 | 0 | 332 | 0 | 12 |  | 764 |  |
| 2001 | 117 | 1183 | 245 | 54 | 0 | 81 |  | 1680 |  |
| 2002 | 263 | 1803 | 505 | 214 | 24 | 813 |  | 3622 |  |
| 2003 | 1109 | 1522 | 334 | 274 | 3 | 479 | 1494 | 5215 |  |
| 2004 | 535 | 1316 | 242 | 116 | 47 | 84 | 2608 | 4948 |  |
| 2005 | 650 | 2351 | 1137 | 1162 | 278 | 382 | 83 | 6043 |  |
| 2006 | 922 | 1682 | 577 | 943 | 630 | 1461 | 1173 | 7388 |  |
| 2007 | 416 | 2547 | 1195 | 1842 | 659 | 4391 |  | 11050 | 42 |
| 2008 | 870 | 3066 | 1539 | 3172 | 225 | 1133 |  | 10005 | 6 |
| 2009 | 325 | 1288 | 1189 | 2009 | 1142 | 1581 |  | 7534 | 2 |
| 2010 | 559 | 2990 | 1607 | 1795 | 1458 | 859 |  | 9268 | 2 |
| 2011 | 567 | 2364 | 2850 | 2905 | 1274 | 1047 |  | 11007 | 0 |
| 2012 | 546 | 1376 | 2061 | 4375 | 1989 | 325 |  | 10672 | 0.02 |
| 2013 | 1506 | 2552 | 2784 | 4711 | 1450 | 198 |  | 13202 | 35 |
| 2014 | 3084 | 6142 | 3710 | 4629 | 684 | 82 |  | 18331 | 38 |

Table 15.1.1.2: Catches $(t)$ divided into month and NAFO Divisions, caught by the coastal fisheries.

| NAFO | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec | Total | $\%$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1A | 96 | 108 | 91 | 78 | 65 | 172 | 406 | 419 | 400 | 601 | 425 | 223 | 3084 | $17 \%$ |
| 1B | 36 | 43 | 195 | 193 | 400 | 1051 | 1255 | 548 | 561 | 833 | 566 | 460 | 6142 | $33 \%$ |
| 1C | 104 | 51 | 33 | 57 | 372 | 813 | 460 | 312 | 475 | 386 | 382 | 264 | 3710 | $20 \%$ |
| 1D | 85 | 53 | 73 | 38 | 343 | 693 | 1245 | 356 | 878 | 507 | 135 | 223 | 4629 | $25 \%$ |
| 1E | 0.01 | 2 | 6 | 0.1 | 4 | 17 | 256 | 224 | 126 | 27 | 20 | 4 | 684 | $4 \%$ |
| 1F | 0.3 | 0.03 | 0.4 | 2 | 3 | 8 | 1 | 3 | 3 | 30 | 28 | 4 | 82 | $0.4 \%$ |
| ICES | 0.2 | 0.2 |  |  |  |  |  | 17 | 21 |  |  |  | 38 |  |
| XIVb |  |  |  |  |  |  |  |  |  |  |  |  |  | $0.2 \%$ |
| Total | 322 | 257 | 397 | 369 | 1186 | 2753 | 3624 | 1879 | 2463 | 2384 | 1556 | 1177 | 18369 |  |
| $\%$ | $2 \%$ | $1 \%$ | $2 \%$ | $2 \%$ | $6 \%$ | $15 \%$ | $20 \%$ | $10 \%$ | $13 \%$ | $13 \%$ | $8 \%$ | $6 \%$ |  |  |

Table 15.1.1.3: Landings (\%) divided into month and gear and NAFO Divisions and gear.

| Gear/Month | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec | Total |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Poundnet | 0.001 |  | 0.002 | 0.3 | 6 | 13 | 15 | 4 | 5 | 2 | 1 | 1 | 47 |
| Gillnet | 1 | 1 | 1 | 1 | 0.4 | 0.4 | 0.3 | 0.2 | 1 | 4 | 3 | 3 | 18 |
| Jig | 0.1 | 0.1 | 0.4 | 0.1 | 0.2 | 1 | 4 | 6 | 6 | 5 | 2 | 0.4 | 25 |
| Longline | 1 | 0.3 | 0.3 | 0.2 | 0.3 | 0.5 | 1 | 0.2 | 1 | 2 | 2 | 2 | 10 |
| Total | 2 | 1 | 2 | 2 | 6 | 15 | 20 | 10 | 13 | 13 | 8 | 6 | 100 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Gear/NAFO | $\mathbf{1 A}$ | $\mathbf{1 B}$ | $1 \mathbf{C}$ | $\mathbf{1 D}$ | $\mathbf{1 E}$ | 1 F | ICES |  |  |  |  |  | Total |
| Poundnet | 3 | 13 | 9 | 19 | 3 | 0.1 |  |  |  |  |  |  | 47 |
| Gillnet | 5 | 9 | 2 | 1 | 0.03 | 0.2 |  |  |  |  |  |  | 18 |
| Jig | 5 | 11 | 7 | 2 | 0.3 | 0.1 | 0.2 |  |  |  |  |  | 25 |
| Longline | 4 | 1 | 3 | 3 | 0.1 | 0.1 |  |  |  |  |  |  | 10 |
| Total | 17 | 33 | 20 | 25 | 4 | 0.4 | 0.2 |  |  |  |  |  | 100 |

Table 15.1.2.1. Estimated catches in numbers ('000) at age, and total catch by year ( $t$ ).

| Year/age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | Catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1976 | 2508 | 924 | 556 | 287 | 38 | 31 | 11 | 7 | 5174 |
| 1977 | 467 | 5437 | 1100 | 883 | 179 | 7 | 142 | 46 | 13999 |
| 1978 | 97 | 1262 | 9904 | 132 | 68 | 7 | 3 |  | 19679 |
| 1979 | 323 | 2297 | 2380 | 8281 | 170 | 96 | 4 | 14 | 35590 |
| 1980 | 4343 | 4334 | 1646 | 806 | 6492 | 106 | 29 | 37 | 38571 |
| 1981 | 87 | 15793 | 5225 | 725 | 499 | 2906 | 61 | 17 | 39703 |
| 1982 | 3013 | 1587 | 6309 | 1545 | 798 | 152 | 610 | 154 | 26664 |
| 1983 | 229 | 16877 | 1381 | 4352 | 368 | 139 | 65 | 75 | 28742 |
| 1984 | 520 | 4451 | 9269 | 346 | 634 | 18 | 42 | 12 | 19958 |
| 1985 | 5 | 2400 | 1028 | 2229 | 196 | 363 | 14 | 78 | 8441 |
| 1986 | 286 | 178 | 896 | 460 | 721 | 16 | 102 | 38 | 5302 |
| 1987 | 5503 | 1334 | 228 | 710 | 340 | 1084 | 46 | 265 | 14604 |
| 1988 | 419 | 15588 | 150 | 51 | 39 | 90 | 161 | 12 | 18791 |
| 1989 | 15 | 5962 | 23956 | 271 | 46 | 2 | 93 | 176 | 38529 |
| 1990 | 212 | 2997 | 15403 | 6732 | 33 | 11 | 7 | 16 | 28799 |
| 1991 | 124 | 6022 | 4910 | 5695 | 330 | 0 |  |  | 18312 |
| 1992 | 8 | 2408 | 2344 | 452 | 139 | 46 | 13 | 5 | 5727 |
| 1993 | 28 | 661 | 575 | 206 | 34 | 41 | 10 | 7 | 1926 |
| 1994 | 22 | 1468 | 342 | 62 | 45 | 8 | 11 | 1 | 2117 |
| 1995 | 1 | 834 | 773 | 37 | 5 | 0 | 0 |  | 1701 |
| 1996 | 2 | 165 | 362 | 130 | 25 | 3 | 1 | 0 | 944 |
| 1997 | 1 | 397 | 311 | 179 | 31 | 0 |  |  | 1186 |
| 1998* |  |  |  |  |  |  |  |  | 322 |
| 1999 | 87 | 465 | 105 | 1 | 0 | 0 |  |  | 613 |
| 2000 | 4 | 228 | 336 | 7 | 0 | 0 |  |  | 764 |
| 2001* |  |  |  |  |  |  |  |  | 1680 |
| 2002 | 532 | 2243 | 657 | 29 | 9 | 1 | 0 | 0 | 3622 |
| 2003 | 152 | 581 | 1547 | 258 | 51 | 16 | 15 | 11 | 5215 |
| 2004 | 530 | 1669 | 1095 | 228 | 37 | 3 |  |  | 4948 |
| 2005 | 1387 | 2400 | 941 | 185 | 36 | 10 | 4 |  | 6043 |
| 2006 | 4256 | 3363 | 680 | 22 | 0 | 0 | 0 |  | 7388 |
| 2007 | 1945 | 7913 | 1010 | 116 | 38 | 13 | 8 | 4 | 11050 |
| 2008 | 1177 | 5015 | 2794 | 319 | 36 | 6 | 2 |  | 10005 |
| 2009 | 487 | 3540 | 2372 | 194 | 13 | 3 | 0 | 4 | 7534 |
| 2010 | 301 | 1091 | 2475 | 1524 | 141 | 32 | 21 | 27 | 9268 |
| 2011 | 129 | 2929 | 2567 | 1480 | 255 | 90 | 12 | 7 | 11007 |
| 2012 | 735 | 1725 | 2681 | 850 | 182 | 21 | 13 | 13 | 10672 |
| 2013 | 143 | 3806 | 2477 | 1083 | 361 | 115 | 67 | 9 | 13202 |
| 2014 | 40 | 1394 | 4033 | 2296 | 330 | 169 | 103 | 52 | 18369 |

Table 15.1.2.2. West Greenland inshore cod. Estimated weight at age (kg).

| y | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1976 | 0.811 | 1.114 | 1.662 | 2.738 | 3.226 | 4.062 | 5.831 | 12.747 |
| 1977 | 0.674 | 1.382 | 2.201 | 2.649 | 3.322 | 6.363 | 3.92 | 4.616 |
| 1978 | 0.668 | 0.965 | 1.801 | 2.472 | 2.845 | 3.649 | 4.733 |  |
| 1979 | 0.8 | 1.309 | 2.111 | 3.153 | 3.696 | 4.371 | 6.861 | 8.007 |
| 1980 | 0.753 | 1.017 | 1.884 | 2.58 | 3.823 | 4.107 | 5.715 | 7.902 |
| 1981 | 0.308 | 1.045 | 1.576 | 2.19 | 2.59 | 4.029 | 3.529 | 7.831 |
| 1982 | 0.844 | 1.118 | 1.604 | 2.605 | 3.875 | 5.495 | 5.425 | 6.278 |
| 1983 | 0.552 | 0.937 | 1.337 | 2.039 | 2.795 | 3.378 | 4.218 | 4.109 |
| 1984 | 0.624 | 0.967 | 1.385 | 1.869 | 2.469 | 3.286 | 3.985 | 4.433 |
| 1985 | 0.42 | 0.754 | 1.134 | 1.662 | 2.065 | 2.669 | 3.486 | 4.337 |
| 1986 | 0.582 | 1.248 | 1.414 | 2.043 | 2.689 | 3.188 | 3.893 | 8.401 |
| 1987 | 0.872 | 1.187 | 2.043 | 2.302 | 2.963 | 3.294 | 4.114 | 5.107 |
| 1988 | 0.659 | 1.106 | 1.251 | 1.691 | 2.677 | 3.046 | 3.478 | 5.111 |
| 1989 | 0.558 | 0.855 | 1.308 | 1.821 | 3.161 | 4.252 | 4.397 | 5.862 |
| 1990 | 0.649 | 0.889 | 1.031 | 1.452 | 2.614 | 3.765 | 5.846 | 10.868 |
| 1991 | 0.802 | 0.966 | 1.088 | 1.146 | 1.595 | 3.964 |  |  |
| 1992 | 0.567 | 0.869 | 1.028 | 1.697 | 1.849 | 2.845 | 3.253 | 4.402 |
| 1993 | 0.585 | 0.82 | 1.239 | 1.83 | 1.802 | 2.873 | 3.976 | 8.777 |
| 1994 | 0.43 | 0.883 | 1.359 | 1.706 | 3.103 | 3.9 | 4.976 | 16.271 |
| 1995 | 0.768 | 0.93 | 1.093 | 1.799 | 2.493 | 4.13 | 6.49 |  |
| 1996 | 0.501 | 0.814 | 1.201 | 2.176 | 2.955 | 4.151 | 5.507 | 6.577 |
| 1997 | 0.560 | 0.956 | 1.397 | 1.767 | 1.830 | 3.239 |  |  |
| 1998* |  |  |  |  |  |  |  |  |
| 1999 | 0.739 | 0.895 | 1.24 | 2.254 | 3.387 | 4.556 |  |  |
| 2000 | 0.642 | 1.121 | 1.453 | 2.378 | 2.621 | 2.409 |  |  |
| 2001* |  |  |  |  |  |  |  |  |
| 2002 | 0.708 | 0.999 | 1.397 | 2.318 | 1.884 | 2.853 | 3.560 | 3.356 |
| 2003 | 1.046 | 1.391 | 2.069 | 2.565 | 3.3 | 3.988 | 5.095 | 6.958 |
| 2004 | 0.988 | 1.236 | 1.584 | 2.158 | 3.149 | 6.132 |  |  |
| 2005 | 0.811 | 1.106 | 1.728 | 2.415 | 2.81 | 6.955 |  |  |
| 2006 | 0.724 | 0.944 | 1.560 | 3.102 | 4.522 | 9.931 | 9.931 |  |
| 2007 | 0.703 | 0.95 | 1.543 | 2.574 | 4.003 | 5.136 | 6.541 | 10.25 |
| 2008 | 0.615 | 0.884 | 1.406 | 2.332 | 3.709 | 5.463 | 7.263 |  |
| 2009 | 0.641 | 0.898 | 1.461 | 2.348 | 4.055 | 5.132 | 5.869 | 14.181 |
| 2010 | 0.659 | 0.976 | 1.517 | 2.12 | 3.204 | 4.872 | 6.929 | 9.796 |
| 2011 | 0.657 | 0.918 | 1.466 | 2.013 | 3.305 | 5.396 | 7.527 | 10.366 |
| 2012 | 0.764 | 1.109 | 1.81 | 2.7 | 3.554 | 5.964 | 6.91 | 14.345 |
| 2013 | 0.766 | 1.258 | 1.623 | 2.235 | 3.059 | 3.636 | 4.114 | 7.43 |
| 2014 | 0.691 | 1.226 | 1.934 | 2.534 | 3.407 | 5.326 | 5.745 | 7.772 |

Table 15.1.4.1: Survey effort in the Greenland Inshore Gill-net survey (nos. of valid net settings).

| Division | 1 B | 1D | 1F | Total |
| :---: | :---: | :---: | :---: | :---: |
| 1985 | 3 | 38 | 27 | 68 |
| 1986 | 26 | 22 | 23 | 71 |
| 1987 | 24 | 27 | 26 | 77 |
| 1988 | 21 | 24 | 24 | 69 |
| 1989 | 28 | 19 | 32 | 79 |
| 1990 | 18 | 21 | 18 | 57 |
| 1991 | 23 | 24 | 20 | 67 |
| 1992 | 27 | 29 | 23 | 79 |
| 1993 | 23 | 25 | 19 | 67 |
| 1994 | 20 | 29 | 17 | 66 |
| 1995 | 24 | 21 | 20 | 65 |
| 1996 | 26 | 25 | - | 51 |
| 1997 | 20 | 23 | - | 43 |
| 1998 | 24 | 26 | 22 | 72 |
| 1999 | - | 24 | - | 24 |
| 2000 | - | 27 | 20 | 47 |
| 2001 | - | - | - | - |
| 2002 | 21 | 20 | - | 41 |
| 2003 | 33 | 27 | - | 60 |
| 2004 | 27 | 31 | - | 58 |
| 2005 | 25 | 28 | - | 53 |
| 2006 | 45 | 51 | - | 96 |
| 2007 | 52 | - | 39 | 91 |
| 2008 | - | 58 | 60 | 118 |
| 2009 | - | 58 | 18 | 76 |
| 2010 | 66 | 52 | - | 118 |
| 2011 | 57 | 44 | - | 101 |
| 2012 | 54 | 52 | - | 106 |
| 2013 | 58 | 52 | - | 110 |
| 2014 | 41 | 60 | - | 101 |

Table 15.1.4.2 : NAFO Div. 1B. Cod abundance indices (numbers of cod caught per 100 hours net settings) by age in the West Greenland inshore gill-net survey. na= data not available.

| Year/Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ | All |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 26 | 23 | 0 | 6 | 0 | 0 | 0 | 0 | 54 |
| 1986 | 4 | 245 | 16 | 8 | 2 | 2 | 0 | 0 | 278 |
| 1987 | 0 | 122 | 233 | 25 | 1 | 0 | 0 | 0 | 381 |
| 1988 | 0 | 33 | 130 | 111 | 2 | 0 | 0 | 0 | 276 |
| 1989 | 1 | 110 | 83 | 57 | 32 | 1 | 0 | 0 | 283 |
| 1990 | 0 | 109 | 108 | 62 | 53 | 12 | 0 | 0 | 344 |
| 1991 | 0 | 3 | 131 | 53 | 11 | 3 | 0 | 0 | 202 |
| 1992 | 0 | 43 | 10 | 18 | 3 | 0 | 0 | 0 | 74 |
| 1993 | 0 | 22 | 22 | 2 | 1 | 0 | 0 | 0 | 47 |
| 1994 | 4 | 8 | 19 | 12 | 0 | 0 | 0 | 0 | 43 |
| 1995 | 2 | 115 | 19 | 7 | 1 | 0 | 0 | 0 | 143 |
| 1996 | 0 | 28 | 40 | 7 | 1 | 0 | 0 | 0 | 77 |
| 1997 | 0 | 14 | 8 | 3 | 1 | 0 | 0 | 0 | 26 |
| 1998 | 2 | 7 | 4 | 6 | 3 | 0 | 0 | 0 | 23 |
| 1999 | na | na | na | na | na | na | na | na | na |
| 2000 | na | na | na | na | na | na | na | na | na |
| 2001 | na | na | na | na | na | na | na | na | na |
| 2002 | 31 | 207 | 72 | 21 | 9 | 1 | 0 | 0 | 340 |
| 2003 | 1 | 68 | 69 | 21 | 3 | 0 | 0 | 0 | 163 |
| 2004 | 32 | 28 | 29 | 9 | 5 | 0 |  | 0 | 102 |
| 2005 | 47 | 123 | 35 | 7 | 5 | 1 | 3 | 0 | 221 |
| 2006 | 32 | 148 | 60 | 24 | 1 | 1 | 0 | 0 | 170 |
| 2007 | 7 | 170 | 82 | 15 | 1 | 0 | 0 | 0 | 275 |
| 2008 | na | na | na | na | na | na | na | na | na |
| 2009 | na | na | na | na | na | na | na | na | na |
| 2010 | 138 | 155 | 120 | 58 | 12 | 1 | 0 | 0 | 484 |
| 2011 | 20 | 526 | 106 | 44 | 19 | 1 | 0 | 0 | 717 |
| 2012 | 7 | 184 | 304 | 30 | 8 | 3 | 0 | 0 | 536 |
| 2013 | 4 | 158 | 105 | 104 | 27 | 8 | 1 | 1 | 408 |
| 2014 | 7 | 46 | 45 | 25 | 19 | 4 | 0 | 1 | 146 |

Table 15.1.4.2, continued : NAFO Div. 1D. Cod abundance indices (numbers of cod caught per 100 hours net settings) by age in the West Greenland inshore gill-net survey.

| Year/Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ | All |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 68 | 77 | 0 | 3 | 3 | 3 | 0 | 1 | 155 |
| 1986 | 0 | 96 | 15 | 0 | 0 | 1 | 2 | 0 | 114 |
| 1987 | 1 | 16 | 68 | 5 | 0 | 0 | 0 | 0 | 90 |
| 1988 | 0 | 20 | 48 | 30 | 1 | 0 | 0 | 0 | 99 |
| 1989 | 0 | 78 | 47 | 13 | 13 | 0 | 0 | 0 | 152 |
| 1990 | 0 | 14 | 35 | 4 | 4 | 3 | 0 | 0 | 60 |
| 1991 | 124 | 3 | 17 | 6 | 2 | 1 | 0 | 0 | 154 |
| 1992 | 0 | 61 | 22 | 10 | 7 | 1 | 0 | 0 | 100 |
| 1993 | 0 | 4 | 57 | 20 | 2 | 0 | 0 | 0 | 83 |
| 1994 | 0 | 0 | 6 | 5 | 1 | 0 | 0 | 0 | 12 |
| 1995 | 0 | 3 | 2 | 4 | 4 | 0 | 0 | 0 | 12 |
| 1996 | 0 | 1 | 1 | 0 | 2 | 0 | 0 | 0 | 4 |
| 1997 | 3 | 3 | 1 | 0 | 0 | 1 | 0 | 0 | 8 |
| 1998 | 0 | 10 | 17 | 1 | 0 | 0 | 0 | 0 | 28 |
| 1999 | 0 | 0 | 1 | 3 | 0 | 0 | 0 | 0 | 5 |
| 2000 | 0 | 2 | 2 | 1 | 1 | 0 | 0 | 0 | 6 |
| 2001 | na | na | na | na | na | na | na | na | na |
| 2002 | 0 | 7 | 4 | 3 | 0 | 0 | 0 | 0 | 14 |
| 2003 | 0 | 6 | 4 | 2 | 1 | 0 | 0 | 0 | 13 |
| 2004 | 3 | 43 | 6 | 3 | 1 | 1 | 0 | 0 | 57 |
| 2005 | 9 | 27 | 7 | 2 | 0 | 0 | 0 | 0 | 45 |
| 2006 | 2 | 114 | 37 | 13 | 4 | 0 | 0 | 0 | 170 |
| 2007 | na | na | na | na | na | na | na | na | na |
| 2008 | 4 | 4 | 47 | 63 | 7 | 0 | 0 | 0 | 124 |
| 2009 | 4 | 52 | 14 | 72 | 23 | 1 | 0 | 0 | 166 |
| 2010 | 1 | 33 | 107 | 18 | 27 | 3 | 0 | 0 | 189 |
| 2011 | 10 | 45 | 3 | 18 | 6 | 4 | 1 | 0 | 88 |
| 2012 | 2 | 52 | 46 | 21 | 28 | 2 | 0 | 1 | 151 |
| 2013 | 0 | 91 | 61 | 77 | 25 | 8 | 3 | 2 | 267 |
| 2014 | 0 | 41 | 74 | 46 | 27 | 6 | 1 | 0 | 196 |

Table 15.1.4.2, continued : NAFO Div. 1F. Cod abundance indices (numbers of cod caught per 100 hours net settings) by age in the West Greenland inshore gill-net survey.

| Year/ |  | 2 | 3 | 4 | 5 | 6 | 7 | 8+ | All |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 204 | 8 | 1 | 1 | 1 | 1 | 1 | 0 | 217 |
| 1986 | 17 | 112 | 5 | 0 | 2 | 0 | 0 | 0 | 136 |
| 1987 | 0 | 143 | 147 | 1 | 0 | 0 | 0 | 0 | 291 |
| 1988 | 0 | 1 | 83 | 6 | 0 | 0 | 0 | 0 | 89 |
| 1989 | 0 | 5 | 2 | 19 | 2 | 0 | 0 | 0 | 29 |
| 1990 | 0 | 0 | 3 | 2 | 13 | 1 | 0 | 0 | 18 |
| 1991 | 2 | 2 | 0 | 2 | 0 | 1 | 0 | 0 | 7 |
| 1992 | 0 | 3 | 1 | 0 | 1 | 0 | 1 | 0 | 6 |
| 1993 | 0 | 5 | 2 | 1 | 0 | 0 | 0 | 0 | 8 |
| 1994 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 3 |
| 1995 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1996 | na | na | na | na | na | na | na | na | na |
| 1997 | na | na | na | na | na | na | na | na | na |
| 1998 | 0 | 4 | 12 | 0 | 0 | 0 | 0 | 0 | 17 |
| 1999 | na | na | na | na | na | na | na | na | na |
| 2000 | 0 | 14 | 8 | 0 | 2 | 0 | 1 | 0 | 24 |
| 2001 | na | na | na | na | na | na | na | na | na |
| 2002 | na | na | na | na | na | na | na | na | na |
| 2003 | na | na | na | na | na | na | na | na | na |
| 2004 | na | na | na | na | na | na | na | na | na |
| 2005 | na | na | na | na | na | na | na | na | na |
| 2006 | na | na | na | na | na | na | na | na | na |
| 2007 | 6 | 90 | 9 | 21 | 1 | 0 | 0 | 0 | 108 |
| 2008 | 8 | 17 | 30 | 4 | 2 | 0 | 0 | 0 | 62 |
| 2009 | 3 | 39 | 14 | 15 | 0 | 0 | 0 | 0 | 71 |
| 2010-2014 na |  | na | na | na | na | na | Na | na | na |



Figure 15.1.1.1 Inshore landings from West Greenland (Horsted 1994, 2000).


Figure 15.1.1.2. Total catches and TAC in the inshore fishery by NAFO Divisions from 2000-2014.


Figure 15.1.1.3. Distribution of the inshore commercial fishery by gear.


Figure 15.1.2.1. Total length and age distributions of inshore cod catches


Figure 15.1.2.2. Length distribution in the inshore fishery in the period 2006-2014.


Figure 15.1.4.1 The inshore gill net survey area on the Greenland West coast. Top picture is the Sisimiut fjord system in NAFO 1B and bottom picture is the Nuuk fjord system in NAFO 1D. Survey estimates of catch rates are indicated on both maps as \#caught/100h.


Figure 15.1.4.2. : Recruitment indices (numbers caught/100 hr.) for ages 2 and 3 in 1B (top), 1D (middle) and 1B and 1D combined (lower) in West Greenland. Simultaneous surveys were not carried out 1999-2001 and 2007-2009.


Figure 15.1.4.3 Recruitment indices (numbers caught/100 hr.) for ages 1-5 in 1B (left), 1D (middle) and 1B and 1D combined (right) in West Greenland from 1985-2014. Size of circles represents the size of the index values and the values are standardized within each area and are not comparable among each other.


Figure 15.1.4.4 Recruitment indices (numbers caught/100 hr. netsetting) for all age groups (left) and ages 3-8 (right - used in the current catch advice) in 1B and 1D combined. Simultaneous surveys were not carried out 1999-2001 and 2007-2009.

## 16 Cod in offshore waters of ICES Subarea XIV and NAFO subarea 1

## Executive summary

From 2014 the management for cod in Greenland offshore waters has been split in two stock components according to areas: NAFO subdivisions 1A-E in West Greenland and NAFO subdivision 1F in South Greenland combined with ICES subarea XIV in East Greenland. The ICES advice for 2016 has for the first time been given according to these two areas.

The offshore fishery in East and South Greenland in 2014 was conducted as an experimental fishery with a TAC of 10000 tons. Total catches were 7893 tons. The year-class dominating the catches was the 2007 YC, which it has done since 2012. The largest cod (mean length of 83 cm ) were caught by trawlers on Dohrn Bank close to the Iceland EEZ.

Available survey biomass indices from the Greenland and German surveys show that the biomass has increased due to the growth of the 2009 YC and in part the 2007 YC. Abundance has however decreased as fewer young fish are observed.

The 2009 YC followed by the 2007 YC has dominated the survey since 2012. The 2009 YC is primarily distributed in South Greenland, whereas the 2007 YC is distributed more to the north in East Greenland. Spawning offshore cod are only found in East Greenland in local high densities.

The procedure suggested as basis for advice at the Benchmark in 2015 was not implemented by NWWG due to shortcomings. Instead, advice was based on an Fproxy multiplier generated from the relationship between the catches and survey index in a period with a considered sustainable fishery, multiplied by the latest year's smoothed survey index (Greenland Shrimp and Fish survey).

### 16.1 Stock definition

The cod found in Greenland is derived from four separate "stocks" that each is labelled by their spawning areas: I) offshore West Greenland waters; II) West Greenland fiords; III) offshore East Greenland and Icelandic waters and IV) inshore Icelandic waters (Therkildsen et al. 2013), (Fig. 16.1).

From 2012 the inshore component (West Greenland, NAFO Subarea 1) was assessed separately from all offshore components. From 2016 the offshore West Greenland (NAFO subdivisions 1A-E) and East Greenland (NAFO subdivision 1F and ICES subarea XIV) components was assessed separately. The Stock Annex provides more details on the stock identities including the references to primary works.

### 16.2 Fishery

### 16.2.1 The emergence and collapse of the Greenland offshore cod fisheries

The Greenland commercial cod fishery in East Greenland started in 1954, but started earlier in Southwest Greenland (NAFO subdiv. 1F, table 16.2.2.1). The fishery gradually developed culminating with catch levels above 40,000 tons annually in the 1960s. Due to overfishing and deteriorating environmental conditions the stock size declined and the fishery completely collapsed in the early 1990's (Fig 16.2.1). In the 2000s catches have gradually increased with maximum catches in 2008 of 14500 tons. Between 2008 and 2010 offshore areal closures were implemented in order to protect the spawning
stock in offshore areas. More details on the historical development in the fisheries are provided in the stock annex.

### 16.2.2 The offshore fishery in 2014

In 2014 a management plan was implemented for the offshore cod fishery in South and East Greenland. According to the management plan the TAC is 10000 tons/year between 2014 and 2016, The TAC between 2014 and 2016 is to be taken in equal amounts in four areas: Survey area Q1+Q2, Survey area Q3+Q4, Survey area Q5+Q6 and NAFO subdivision 1F.

In 2014 the offshore TAC of 10000 tons was divided with 6600 tons to Greenland, 2200 tons to EU and 1200 tons to Norway. Further 355 tons was allocated to the Faroe Islands as part of a mix quota (OTH). EU, Norway and the Faroe Islands fished their quota whereas Greenland fished 4300 tons.

Sampling of length frequencies and information on length, weight and age were collected by the crew on the ships who length measured 50 randomly selected cod each day and took individual measurements (length, weight, gutted weight and otoliths) from 20 randomly selected cod each day.

Offshore catches in South and East Greenland in 2014 amounted to a total of 7893 tons with 1833 tons caught in Southwest Greenland (NAFO 1F) and 6060 tons caught in East Greenland (table 16.2.2.1).
$77 \%$ of the total catches were taken in East Greenland where the fishery peaked in June. Catches in Southwest Greenland peaked in March and November (table 16.2.2.2). The fishery where distributed from Julianehåb Bight in Southwest Greenland $\left(60^{\circ} \mathrm{N}\right)$ to Dohrn Bank $\left(66^{\circ} \mathrm{N}\right)$ in East Greenland (figure 16.2.2.1).

A dispensation was given to a Greenlandic longliner to fish within the 3 nm from the baseline in East Greenland where vessels larger than 75BRT/120BT are not allowed. In total the vessel caught 230 tons, concentrated around the Tasiilaq area $\left(65-66^{\circ} \mathrm{N}\right.$, figure 16.2.2.2).
$67 \%$ of the total catch was taken by trawlers and $33 \%$ by longliners primarily in June (table 16.2.2.3). About half ( $48 \%$ ) of the total trawl catches were taken in a small area between $65-66^{\circ} \mathrm{N}$ and $29-31^{\circ} \mathrm{W}$ on the edge of the continental shelf on Dohrn Bank (management area Q1Q2, figure 16.2.2.2). $50 \%$ of the total longline catches were taken along the continental shelf and on Kleine Bank $\left(64^{\circ} \mathrm{N}\right)$ in management area Q3Q4 in East Greenland.

The offshore fishery peaked in summer and fall, and the majority of the catches where taken in management areas Q1Q2 and Q3Q4 in East Greenland. The trawlers conducted most of their fishing in a small area on the edge of Dohrn bank close to the Icelandic EEZ (management area Q1Q2), whereas the longliners conducted most of their fishing further to the south along the continental shelf and on Kleine Bank (management area Q3Q4). The reason for this separation is likely because the largest cod are found in the Dohrn Bank area and therefore attractive for the trawlers.

Larger and older fish (7+ yr old) are located furthest to the north on Dohrn Bank, whereas younger fish dominate in the South ( $5-7 \mathrm{yr}$ old) which also corresponds to the composition of age classes in the different areas observed in the survey (Retzel \& Post 2015).

The 2007 YC was the dominating YC in the catches at age 5 and 6 in 2012 and 2013. In 2014 the 2007 YC is still the dominating YC at age 7 and it is therefore believed that this

YC is primarily of Greenlandic origin and not Icelandic. The YC is now contributing to the spawning stock in East Greenland in considerable numbers, and it is still located in South Greenland indicating that spawning could also be going on here.

### 16.2.3 Length, weight and age distributions in the offshore fishery 2014

There is limited landing sample information from the 1990's where the cod fishery was very low in East Greenland. For that period length frequency information is generally lacking for the offshore fisheries where cod was only taken as a by-catch. Sampling intensities have increased considerably in the later years, and in 2014 the offshore fisheries was very well covered.
Catch-at-age and weight-at-age has been compiled for the offshore area since 2005 (table 16.2.3.1).
Length measurement amounted to 10259 in Southwest Greenland and 3657 in East Greenland.
The overall mean length in the catches was 74 cm , and the YC 2007 (7 yr old fish) dominated the catches (figure 16.2.3.1). The mean length differed between areas with the largest cod (mean length $=83 \mathrm{~cm}$ ) and oldest ( $7-10 \mathrm{yr}$ old) being caught on Dohrn Bank furthest to the north in the fishing area in East Greenland (management area Q1Q2), and the smallest (mean length $=66 \mathrm{~cm}$ ) and youngest ( $5-7 \mathrm{yr}$ old) being caught in Southwest Greenland (NAFO 1F, figure 16.2.3.2).

In 2012 and 2013 the 2007 YC dominated the total catches (Table 16.2.3.1). This YC was especially abundant in the catches in West Greenland and East Greenland south of Dohrn Bank in 2013. In 2014 this YC is abundant in all areas, especially in management area Q3Q4 in East Greenland. In Southwest Greenland (NAFO 1F) the 2008 and 2009 YC is more abundant than in the other management areas in East Greenland (figure 16.2.3.2).

### 16.2.4 CPUE index

Log books on a haul by haul basis from the cod fishery since 1975 where compiled in 2014. But due very low catches and few hauls in the 90'ies and closed areas in 20082010, the logbook data are not used in the assessment process. Nevertheless, CPUE results generated by a GLM model are presented here.

As EU and Greenland vessels have participated in the fisheries in the entire period, data from these were used in the GLM model. Hauls made in the closed area in the period of 2008-2010 were excluded from the analysis, as they were considered being by-catches.

The CPUE index was relatively stable in the first part of the time series (1975-1992, mean 0.675 tons $/ \mathrm{hr}$ ), except 1989 where it increased to 1.676 tons/hr (figure 16.2.4.1). This increase was likely caused by the large 1984 YC entering the fishery. The CPUE then declined from 1993-2005 ( 0.149 tons/hr), but sampling of the fishery was low in this period due to very low catches of about 200-300 tons, and catches where taken primarily as by-catch in the redfish fishery. In 2006-2008 the CPUE increased (mean 1.623 tons $/ \mathrm{hr}$ ) as catches started to increase. In 2009 the CPUE decreased to 0.436 tons/hr, which was most likely caused by the east ward migration of the 2003 year class out of the allowed fishing areas (table 16.2.4.1).
In 2010, where almost all of the offshore area was closed, except of a small area in South East Greenland, the index increased to 0.670 tons $/ \mathrm{hr}$, but catches were taken by very
few vessels. In 2011 all closed areas where reopened and fishery started again, especially in East Greenland north of $63^{\circ} \mathrm{N}$, resulting in an increase in CPUE to 1.164 tons/hr. This trend follows the development in survey index from 2008-2011 (Retzel \& Post 2015), with several YC's being present and a steady increase in biomass since 2006. Since 2011 CPUE has declined (2012-2014, mean 0.937 tons/hr). In contrast the survey index has increased which is properly caused by a large 2009 YC observed in the survey, but not yet in the fishery as the 2009 YC is observed primarily in South Greenland in the survey and the majority of the fishery is concentrated further north in East Greenland.

Overall the CPUE index indicates a relatively stable biomass in recent years.

### 16.3 Surveys

At present, two offshore trawl surveys (Greenlandic and German) provide the core information relevant for stock assessment purposes. For details of survey design see stock annex.

The German survey targets mainly cod and has since 1982 covered the main cod grounds off both East and West Greenland at depths down to 400 m . The Greenland survey in West Greenland targets shrimp and cod down to 600 m . The Greenland survey is believed to provide a better coverage of the cod distribution in especially East Greenland as the survey has twice as many stations covering both shelf edge and top, whereas the stations in the German survey are usually concentrated at the shelf edge. The Greenland survey time series is however limited to the 2008-2014 period as the survey in East Greenland first started in 2008.

### 16.3.1 Results of the Greenland Shrimp and Fish survey in South and EastGreenland

A total number of 113 valid hauls were made in 2014 (Error! Reference source not found. 16.3.1.1).

For Atlantic cod the abundance index was estimated at 58 million individuals and the survey biomass at 184500 tons. Survey abundance and biomass decreased with $44 \%$ and $11 \%$ respectively compared to 2013 (Error! Reference source not found. 16.3.1.2 and Error! Reference source not found. 16.3.1.3). SSB biomass increased with $10 \%$.

The majority of the abundance and biomass was primarily located in Southwest Greenland (NAFO 1F) (figure 16.3.1.1 and figure 16.3.1.2). However, opposite to previous years a large amount was also found in mid-East Greenland (Q4). In this stratum (Q4, $1-200 \mathrm{~m}$ ) a high index for both abundance and biomass was calculated as a result of one single big trawl haul containing $>4$ t Atlantic cod. This haul accounted for $30 \%$ and $44 \%$ of the total abundance and biomass index respectively which also resulted in large overall survey CV of 0.46 .

The dominating cohort is the 2009 YC accounting for $38 \%$ in abundance and the second largest cohort was the 2007 YC (21\%) (Error! Reference source not found.16.4.1.4,

Figure 16.3.1.3). The 2009 YC is dominating in South Greenland (NAFO 1F and Q6) where $90 \%$ of the total 2009 YC abundance is found (Error! Reference source not found.16.3.1.5, Error! Reference source not found.16.3.1.4). The 2007 YC is dominating in the northern part of the survey area in East Greenland (Dohrn Bank (Q1) Skjoldungen Bank (Q4)) where $88 \%$ of the total 2007 YC abundance is found. However $63 \%$ of the total abundance of the 2007 YC is found in Q4 where the one large haul was
taken. The size of fish caught in this one haul is comprised of the 2004-2008 YC (figure 16.3.1.5).

In general younger cod (3-6 yrs) are predominantly found in South Greenland (NAFO $1 \mathrm{~F}+\mathrm{Q} 6$ ), whereas older cod (> 7 yrs ) are found in the northern survey area in East Greenland (Error! Reference source not found. 16.3.1.5, Error! Reference source not found.16.3.1.5). Estimated mean length has increased throughout the time series and now catches mainly comprises of fish from 45-100cm (Error! Reference source not found.16.3.1.6). SSB was estimated to 150000 tons, which is by far the highest observed in the time series and 10\% larger than the second highest, observed in 2013 (Error! Reference source not found.16.3.1.7).

Out of the 113 stations one contributed with $30 \%$ and $44 \%$ of the total abundance and biomass index estimates in East Greenland. This haul was located in Q4 with few stations and contained primarily older fish (>6 yrs), resulting in higher abundance index (Error! Reference source not found.16.3.1.4) and high spawning stock biomass compared to other areas in the survey (Error! Reference source not found.16.3.1.8). NAFO 1 F show the same high abundance index, but this is not based on a single haul in this area, but on several hauls with above average catches. The fish caught in NAFO 1F where smaller than in Q4 and was mainly comprised of the 2009YC which is in agreement with the general perception of the migration pattern of the stock in this area.

The survey show a decline in both biomass and especially in abundance compared to 2013, but biomass is $62 \%$ above the average biomass for the period 2008-2014. Abundance is on the same level as the average abundance for the period 2008-2014. The decline in abundance is compensated by the growth of the existing cod stock, but indicates that the new cohorts entering the stock are smaller. West Greenland north of NAFO 1F is believed to be a nursing area for the East Greenland cod stock. The survey in this area (Retzel 2015a) indicates that new year classes (age 1-3 yrs) are present, but none are of the same size as the 2009 YC.

The last YC that gave rise to a substantial fishery (15 000 tons, (Retzel 2015,b)) was the 2003 YC in 2008. Comparing the size of the 2009 YC with 2003 YC at age 5 in all of Greenland offshore waters (West+East Greenland) indicates the 2009 YC is double the size of the 2003 YC. The distribution pattern of the two year classes seems different (Error! Reference source not found.16.3.1.5). The 2003 YC at age 5 was almost exclusively found in South Greenland, whereas the 2009 YC at age 5 is found more spread out in West Greenland from NAFO 1D - NAFO 1F. At younger ages ( $1-2 \mathrm{yrs}$ ) the 2009 YC was mainly distributed in Northwest Greenland in 2010 and 2011, but moved further south in 2012-2014.

The abundance of the 2003 YC at age 6 in 2009 declined by almost $3 / 4$ compared to the abundance at age 5 in 2008. Cod start to spawn at age 6 and the sharp decline of this YC in Greenlandic waters compared with an increase in the abundance of this YC in Icelandic waters (ICES 2014) indicate that this YC has originated from Iceland and has returned for spawning. The 2007 YC, that is dominating the fishery (Retzel 2015,b) and the survey in the northern part of East Greenland, has not declined in abundance from age 5 to age 6, and is therefore believed to be of East Greenlandic origin, and is now contributing substantially to the spawning stock in East Greenland

### 16.3.2 Results of the German groundfish survey off West and East Greenland

In 2014, 75 valid trawl stations were sampled during autumn in the German Greenland offshore groundfish survey (Table 16.3.2.1, Figure 16.3.2.1).

Abundance increased by $78 \%$ from 2013 to 2014 (table 16.3.2.2), and biomass doubled (94\%, Table16.3.2.3). One haul located in NAFO 1F (Southwest Greenland, figure 16.3.2.2) accounted for $25 \%$ of the biomass index. In East Greenland several large hauls were taken from north to south, but none contributed with more than $10 \%$ to the biomass index. The large increase in biomass can therefore not be explained by one large haul. The main reason for the increase in abundance and biomass was increased numbers of the 2009 and 2010 YC in 2014 compared to 2013 (Table 16.3.2.4). The 2009 YC first appeared in considerable numbers in the survey in 2012, and has dominated the survey in 2013 and 2014. The 2010 YC first appeared in considerable numbers in the survey in 2014. The large increase in biomass is probably caused by a southward migration of especially the 2009 YC, which at a younger age was observed more in West Greenland than in South Greenland. In 2014 the 2009 YC and 2010 YC was dominating in South Greenland, whereas older year classes dominated in the northern areas in East Greenland (figure 16.3.2.3).

The survey time series (figure 16.3.2.4) shows three abundance peaks in 1987-1989 caused by the 1984 and 1985 YC, in 2005-2007 caused by the 2003 YC and in 20132014 caused by the 2009 YC. Biomass indices show the same peaks, although a large increase in biomass in 2014 compared to the previous periods (figure 16.3.2.5).

Overall findings where the same between the Greenland and the German survey: a 2009 YC dominating the catches in recent years in South Greenland. However the German survey observed increased numbers of the 2010 YC that was not observed in the Greenland survey. This YC was however observed in the Greenland survey in West Greenland north of NAFO subdivision 1F. The German survey is conducted in October, whereas the Greenland survey in West Greenland is conducted in June/July. The findings might results from a southward migration of the 2010 YC during summer/fall.
Both surveys show that older and larger cod are found furthest to the north in East Greenland, especially in the Dohrn Bank region.

### 16.3.3 Smoothed surveys

The East Greenland area is highly dynamic due to migrations to and from adjacent areas. Inflow of eggs and larvae from Iceland is a common and sometimes large event and some year classes found in East Greenland are primarily from this area (e.g. 2003 YC). West Greenland functions to a very large extent as nursing grounds for East Greenland juveniles and the return migration often produce very sudden and large biomass increases. Jointly, this dynamic can cause large between year variations in survey indices that may appear unrealistic. Furthermore, survey indices are associated with large uncertainties in this area, particularly because of single very large hauls. As the surveys form the basis for the advice, such uncertainty is unwanted and a random effects survey smoother was applied to the estimates of biomass. The underlying survey biomass is modelled with a random walk with process errors, and the observations of survey biomass estimates are estimated with observation errors:
$\mathrm{Zt}_{\mathrm{t}}=\mathrm{Z} \mathrm{t}-1+\mathrm{at}_{\mathrm{t}}$
$\mathrm{yt}_{\mathrm{t}}=\mathrm{Zt}-1+\mathrm{e}_{\mathrm{t}}$
where $\mathrm{zt}_{\mathrm{t}}$ is the natural $\log$ of true survey biomass at time $\mathrm{t}, \mathrm{y}_{\mathrm{t}}$ is the natural $\log$ of estimated survey biomass, and at and et are process and observation errors, respectively, modelled with normal distributions. For a more throughout description see ICES (2015). The smoothed biomass estimates are very close to the observed mean estimates in 2014 (Error! Reference source not found.16.3.1.3, Figure 16.3.3.116.3.3.1), but the CV
was also very high. The observed CV was 0.45 , while the smoothed CV estimate was 0.31 . The process SD was 0.37 .

### 16.3.4 Alternative basis for advice

This stock was benchmarked in 2015 and the benchmark concluded that catch advice for this stock should be based on the 3.2 DLS approach (ICES 2015). The NWWG however concluded that this approach had some major drawbacks for this stock. At NWWG the applicability of the DLS approach was explored and several shortcomings in relation to this stock were found:

- Using the DLS is a slow responding approach not suited to a species with a very dynamic stock development. Applying it for this stock would not allow managers to react to sudden increases or decreases in biomass due to the $20 \%$ cap/change limit from catches. To adjust for this, additional exploratory analyses looked at the consequence of having a cap $10 \%, 20 \%, 40 \%$ and without (Figure 16.3.4.1). The $40 \%$ and no cap however entailed that the advice rose very quickly in response to increasing survey values. On the other hand, the lower caps were not able to react in periods with low biomasses.
- The level of advised catch depends to a very large extent on the offset (Figure 16.3.4.2). For instance, if the approach is implemented at a time of low catches ( $<500 \mathrm{t}$ ) it will take a long time for the advice to adjust to increasing biomass. The other case with a starting point with high catches could also be chosen and that would result in an advice that starts high and stays at that level for a long time.

These issues raised above are particularly important for this stock were large year-toyear variations are a natural occurrence. Therefore, the NWWG concluded that the DLS category 3.2 method was not the best available options, and instead developed a new method to be used. As a period of relatively stable catches is co-occurring with rising survey indices (2009-2014), a derived $\mathrm{F}_{\text {proxy }}$ would be a better basis for advice and more precautionary. The fishing mortality in this period was explored by log catch ratios (Figure 16.3.4.3) and NWWG concluded that as F appeared to be very low in this period no precautionary buffer should be applied. Also, as the stock status is well described through two surveys no uncertainty cap should be applied. Hence, the catch advice should be based on an $\mathrm{F}_{\text {proxy }}$ multiplier on the Greenland survey (smoothed) which has the best coverage of the stock. The catch was divided by the survey from 2009-2014 and the average of this (0.051) was multiplied with the smoothed 2014 Greenland survey index (175 160) to give the 2016 catch advice.

### 16.4 Information on spawning

Adequate maturity information has been lacking for the offshore cod stock as the Greeland and German surveys are conducted well outside the spawning period. The offshore fishery has however shown dense concentrations of large spawning cod off East Greenland at least since 2004. The fishery showed that spawning is concentrated on banks north of $62^{\circ} \mathrm{N}$ in East Greenland. For further information on spawning see stock annex.

### 16.5 Tagging experiments

A total of 16030 cod have been tagged in different regions of Greenland in the period of 2003-2014 (table 16.5.1). Cod in the offshore area in West Greenland have been
tagged in 2007, 2012 and 2013 on Dana Bak (NAFO 1DE). Cod offshore in East Greenland have been tagged in 2007-2009, 2011, 2012 and 2014 from Julianhåbs Bight (NAFO 1F) in SouthWest Greenland to Dohrn Bank in East Greenland.

Inshore recaptures are almost exclusively recaptured in the same place as tagged (table 16.5.2). No tags from the inshore area have been recaptured offshore except 3 that were recaptured in Iceland. These three cod were tagged in the inshore area in South Greenland.

Offshore recaptures are found both in West-, East Greenland and Iceland (table 16.5.2). Most recaptured tags in both West and East Greenland are recaptured in the same place as they were tagged. Recaptured tags from Iceland are mostly tagged in East Greenland, but also in West Greenland typically in South Greenland. The majority of the recaptured tags in Iceland are caught in the northeast area close to Dohrn Bank (figure 16.5.1). More analysis needs to be performed on the tagging data in order to investigate the interaction between Iceland and East Greenland.

### 16.6 State of the stock

The offshore component has been severely depleted since 1990. However, the surveys indicate an improvement in recruitment with all year classes since 2002, and estimated at sizes above the very small year classes seen in the 1990s. These YC's has lead to a stock increase during the 00 s and an increase in catches.

The overall trend in the two surveys is the same: the 2009 YC is distributed in South Greenland, whereas older yearclasses are distributed further north in East Greenland.

The German survey shows a doubling in biomass which cannot be explained by one large haul. The increase was caused by increasing numbers of especially the 2009 YC, but also the 2010 YC in South Greenland. The same increase was not observed in the Greenland survey in South and East Greenland, but instead in West Greenland. The Greenland survey takes place during summer whereas the German survey takes place a couple of month later in autumn. A southward migration of the 2009 YC and 2010 YC during fall might explain the difference between the two surveys.
The fishery confirmed the distribution found in the surveys with younger yearclasses ( $<7 \mathrm{yrs}$ ) dominating the catches in South Greenland, and older yearclasses dominating the catches further north in East Greenland, especially in the Dohrn Bank area.

Indicators show that fishing pressure has been low the last $5-6$ yrs and the stock is considered to be improving. The stock size is however still low in comparison to the 1950's and 1960', where catches exceeded 30000 tons for a number of years.

### 16.7 Implemented management measures for 2015

The offshore quota for the total international fishery is set at 10000 t according to a management plan that was implemented in 2014. The conditions of the fishery are as followed:

1) To spread the fishery so a maximum of 2.500 t are allowed to be taken in 4 management areas (see figure below).

To protect the spawning stock no fishing is allowed from April $1^{\text {st }}$ to May $31^{\text {st }}$ in all areas.

To obtain biological information of the cod stock, hence the vessels must register the catch composition as prescribed in the logbook regulation and conduct sampling of length (length measurements), weight and age (Otolith).

### 16.8 Management plan

In 2014 a management plan was implemented for the offshore cod fishery in Greenland (2014-2016) but it has not been evaluated by ICES. The management plan is built on the distinction between the inshore and two offshore stocks components.


Management area West covers NAFO Subarea 1A-E and management area Southeast covers ICES Subarea XIVb (survey area Q1-6) + NAFO 1F.

According to the management plan the TAC in management area Southeast is 10000 t/year between 2014 and 2016 and no fishery should be done north of 1F in West Greenland.

The TAC in management area South and East Greenland is divided equally between four areas: Survey area Q1+Q2, Q3+Q4, Q5+Q6 and NAFO area 1F.

The management plan has not been evaluated by ICES.

### 16.9 Management considerations.

Larger and older fish (7+ yr old) are located furthest to the north on Dohrn Bank, whereas younger fish dominate in the South ( $5-7 \mathrm{yr}$ old). This reflects the eastward migration behaviour towards the spawning grounds in East Greenland. Further, the genetic studies combined with tagging results suggest that the spawning stock component in East Greenland is associated with the offshore spawning population in Iceland, but the extent and exact dynamics of this association is unknown.

### 16.10References

ICES (2015). Report of the Benchmark Workshop on Icelandic Stocks (WKICE), 26-30 January 2015, Copenhagen, Denmark. ICES CM 2015/ACOM:31. 133 pp.

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Table 16.2.2.1. Offshore catches ( t ) divided into NAFO divisions in West Greenland and East Greenland (ICES XIVb). 1924-1995: Horsted 2000, 1995-2000: ICES Catch Statistics, 2001-present: Greenland Fisheries License Control.

| Year | $\begin{aligned} & \text { NAFO } \\ & \text { 1A } \end{aligned}$ | $\begin{aligned} & \text { NAFO } \\ & 1 \text { B } \end{aligned}$ | $\begin{aligned} & \text { NAFO } \\ & 1 \mathrm{C} \end{aligned}$ | $\begin{aligned} & \text { NAFO } \\ & 1 \mathrm{D} \end{aligned}$ | $\begin{aligned} & \text { NAFO } \\ & 1 \mathrm{E} \end{aligned}$ | NAFO $1 \mathrm{~F}$ | Unknown NAFO div. | ICES <br> XIVb | NAFO <br> 1F + <br> ICES <br> XIVb |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1924 |  |  |  |  |  |  | 200 |  |  |
| 1925 |  |  |  |  |  |  | 1871 |  |  |
| 1926 |  |  |  |  |  |  | 4452 |  |  |
| 1927 |  |  |  |  |  |  | 4427 |  |  |
| 1928 |  |  |  |  |  |  | 5871 |  |  |
| 1929 |  |  |  |  |  |  | 22304 |  |  |
| 1930 |  |  |  |  |  |  | 94722 |  |  |
| 1931 |  |  |  |  |  |  | 120858 |  |  |
| 1932 |  |  |  |  |  |  | 87273 |  |  |
| 1933 |  |  |  |  |  |  | 54351 |  |  |
| 1934 |  |  |  |  |  |  | 88422 |  |  |
| 1935 |  |  |  |  |  |  | 65796 |  |  |
| 1936 |  |  |  |  |  |  | 125972 |  |  |
| 1937 |  |  |  |  |  |  | 90296 |  |  |
| 1938 |  |  |  |  |  |  | 90042 |  |  |
| 1939 |  |  |  |  |  |  | 62807 |  |  |
| 1940 |  |  |  |  |  |  | 43122 |  |  |
| 1941 |  |  |  |  |  |  | 35000 |  |  |
| 1942 |  |  |  |  |  |  | 40814 |  |  |
| 1943 |  |  |  |  |  |  | 47400 |  |  |
| 1944 |  |  |  |  |  |  | 51627 |  |  |
| 1945 |  |  |  |  |  |  | 45800 |  |  |
| 1946 |  |  |  |  |  |  | 44395 |  |  |
| 1947 |  |  |  |  |  |  | 63458 |  |  |
| 1948 |  |  |  |  |  |  | 109058 |  |  |
| 1949 |  |  |  |  |  |  | 156015 |  |  |
| 1950 |  |  |  |  |  |  | 179398 |  |  |
| 1951 |  |  |  |  |  |  | 222340 |  |  |
| 1952 | 0 | 261 | 2996 | 18188 | 707 | 37905 | 257488 |  |  |
| 1953 | 4546 | 46546 | 10611 | 38915 | 932 | 25242 | 98225 |  |  |
| 1954 | 2811 | 97306 | 18192 | 91555 | 727 | 15350 | 60179 | 4321 | 23759* |
| 1955 | 773 | 50106 | 32829 | 87327 | 3753 | 4655 | 68488 | 5135 | 11567* |
| 1956 | 15 | 56011 | 38428 | 128255 | 8721 | 4922 | 66265 | 12887 | 19189* |
| 1957 | 0 | 58575 | 32594 | 62106 | 29093 | 16317 | 47357 | 10453 | 30659* |
| 1958 | 168 | 55626 | 41074 | 73067 | 21624 | 26765 | 75795 | 10915 | 46972* |
| 1959 | 986 | 74304 | 10954 | 30254 | 12560 | 11009 | 67598 | 19178 | 35500* |
| 1960 | 35 | 58648 | 18493 | 35939 | 16396 | 9885 | 76431 | 23914 | 39219* |
| 1961 | 503 | 78018 | 43351 | 70881 | 16031 | 14618 | 90224 | 19690 | 40212* |
| 1962 | 1017 | 122388 | 75380 | 57972 | 25336 | 17289 | 125896 | 17315 | 41874* |
| 1963 | 66 | 70236 | 73142 | 76579 | 46370 | 16440 | 122653 | 23057 | 46626* |
| 1964 | 96 | 49049 | 49102 | 82936 | 33287 | 13844 | 99438 | 35577 | 55451 * |
| 1965 | 385 | 80931 | 66817 | 71036 | 15594 | 15002 | 92630 | 17497 | 38063* |
| 1966 | 12 | 99495 | 43557 | 62594 | 19579 | 18769 | 95124 | 12870 | 38956* |
| 1967 | 361 | 58612 | 78270 | 122518 | 34096 | 12187 | 95911 | 24732 | 40738* |
| 1968 | 881 | 12333 | 89636 | 94820 | 61591 | 16362 | 97390 | 15701 | 37844* |
| 1969 | 490 | 7652 | 31140 | 65115 | 41648 | 11507 | 35611 | 17771 | 31879* |



1) Estimates for assessment include estimates of unreported catches. The total estimated value for West Greenland (inshore + offshore) was 73000 t in 1977 and 1978, 1979: 99000 t , 1980: 54000 t . The value given in the table are these values minus the inshore catches minus known offshore NAFO division catches.
2) Estimates for assessment include estimates of unreported catches in East Greenland.
3) Include catches taken with small vessels and landed to a factory in South Greenland (Qaqortoq), 2007:597 t, 2008: 2262 t , 2009: 136 t .
*) Unknown NAFO division catches added accordingly to the proportion of known catch in NAFO division 1F to known total catch in all NAFO divisions.

Table 16.2.2.2: 2014 cod catches ( $t$ ) by area and month. East Greenland (XIVb) divided into three management areas.

| ICES/NAFO | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec | Total | $\%$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| XIVb |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (Q1Q2) | 1 | 0 | 30 | 55 | 37 | 40 | 626 | 635 | 92 | 482 | 602 | 5 | 2606 | $33 \%$ |
| XIVb |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (Q3Q4) | 4 | 2 | 32 | 10 | 73 | 1030 | 329 | 412 | 197 | 198 | 193 |  | 2479 | $31 \%$ |
| XIVb |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (Q5Q6) |  | 3 |  | 0.3 | 135 | 497 | 37 | 35 | 48 | 19 | 12 | 187 | 975 | $12 \%$ |
| 1F |  | 77 | 463 | 45 | 3 | 71 | 14 | 40 | 274 | 229 | 412 | 203 | 1833 | $23 \%$ |
| Total | 4 | 83 | 525 | 110 | 248 | 1639 | 1007 | 1123 | 611 | 929 | 1220 | 395 | 7893 |  |
| $\%$ | $0.1 \%$ | $1 \%$ | $7 \%$ | $1 \%$ | $3 \%$ | $21 \%$ | $13 \%$ | $14 \%$ | $8 \%$ | $12 \%$ | $15 \%$ | $5 \%$ |  |  |

Table 16.2.2.3: 2014 cod catches ( $t$ ) by gear, area and month. East Greenland (XIVb) divided into three management areas.

| Gear | ICES/NAFO | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Longline | $\begin{aligned} & \text { XIVb } \\ & \text { (Q1Q2) } \end{aligned}$ |  |  |  |  |  | 0.3 | 4 | 28 | 4 | 6 |  |  | 42 |
|  | $\begin{aligned} & \text { XIVb } \\ & \text { (Q3Q4) } \end{aligned}$ |  |  | 32 |  | 3 | 1004 | 24 | 8 | 40 | 172 | 28 |  | 1310 |
|  | $\begin{aligned} & \text { XIVb } \\ & \text { (Q5Q6) } \end{aligned}$ |  |  |  |  | 4 | 453 |  |  | 0 | 16 |  |  | 472 |
|  | 1F |  |  | 242 | 42 | 0.2 |  |  | 4 | 68 | 131 | 171 | 115 | 773 |
|  | Total |  |  | 274 | 42 | 7 | 1457 | 28 | 41 | 112 | 324 | 199 | 115 | 2598 |
| Trawl | $\begin{aligned} & \text { XIVb } \\ & \text { (Q1Q2) } \end{aligned}$ | 1 | 0.2 | 30 | 55 | 37 | 40 | 622 | 607 | 88 | 477 | 602 | 5 | 2564 |
|  | $\begin{aligned} & \text { XIVb } \\ & \text { (Q3Q4) } \end{aligned}$ | 4 | 2 |  | 10 | 70 | 27 | 304 | 404 | 157 | 26 | 165 |  | 1168 |
|  | $\begin{aligned} & \text { XIVb } \\ & \text { (Q5Q6) } \end{aligned}$ |  | 3 |  | 0.3 | 132 | 44 | 37 | 35 | 48 | 4 | 12 | 187 | 503 |
|  | 1F |  | 77 | 222 | 3 | 3 | 71 | 14 | 36 | 206 | 99 | 242 | 87 | 1060 |
|  | Total | 4 | 83 | 251 | 69 | 241 | 181 | 979 | 1082 | 499 | 605 | 1021 | 280 | 5295 |

Table 16.2.3.1. Cod in Greenland. Catch at age ('000) and Weight at age ( $\mathbf{k g}$ ) for offshore fleets in East Greenland (ICES XIVb + NAFO 1F). Yellow highlights dominating yearclasses in the catches.

| Catch at age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year/age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| 2005 | 5 | 33 | 57 | 103 | 94 | 57 | 16 | 7 |
| 2006 | 232 | 376 | 135 | 175 | 115 | 14 | 1 | 0 |
| 2007 | 49 | 1529 | 668 | 158 | 124 | 120 | 18 | 15 |
| 2008 | 77 | 586 | 6015 | 2417 | 592 | 44 | 26 | 12 |
| 2009 | 307 | 1287 | 1231 | 434 | 119 | 28 | 16 | 2 |
| 2010 | 10 | 87 | 331 | 193 | 334 | 58 | 8 | 5 |
| 2011 | 3 | 70 | 137 | 425 | 355 | 371 | 96 | 31 |
| 2012 | 13 | 109 | 471 | 281 | 258 | 253 | 148 | 59 |
| 2013 | 0 | 36 | 127 | 615 | 237 | 226 | 153 | 104 |
| 2014 | 1 | 4 | 279 | 434 | 658 | 335 | 173 | 131 |
| Weight at age |  |  |  |  |  |  |  |  |
| 2005 | 0.354 | 0.717 | 1.073 | 1.963 | 2.737 | 3.699 | 5.271 | 7.366 |
| 2006 | 1.323 | 1.602 | 2.349 | 3.608 | 4.420 | 5.440 | 7.191 | 8.127 |
| 2007 | 0.387 | 0.917 | 1.597 | 3.294 | 6.092 | 8.524 | 11.114 | 14.435 |
| 2008 | 0.359 | 0.644 | 1.266 | 1.799 | 3.025 | 4.936 | 5.840 | 8.290 |
| 2009 | 0.489 | 0.776 | 1.396 | 2.797 | 4.634 | 6.453 | 7.804 | 9.993 |
| 2010 | 0.699 | 1.125 | 1.636 | 2.494 | 3.354 | 5.334 | 8.063 | 10.475 |
| 2011 | 0.553 | 1.026 | 1.541 | 2.297 | 3.377 | 4.685 | 6.285 | 10.022 |
| 2012 | 0.502 | 0.892 | 1.440 | 2.380 | 3.570 | 5.142 | 7.172 | 11.417 |
| 2013 | 0.480 | 0.998 | 1.698 | 2.272 | 3.408 | 4.745 | 6.827 | 9.024 |
| 2014 | 0.564 | 1.163 | 1.853 | 2.603 | 3.636 | 4.732 | 6.400 | 8.841 |

Table 16.2.4.1: Data used in the Atlantic cod CPUE. N are number of hauls from vessels from EU and Greenland used in the analysis.

| year | N | In CPUE (ton/hr) | SE |
| :---: | :---: | :---: | :---: |
| 1975 | 82 | -1.12595665 | 0.1602324 |
| 1976 | 5 | -0.95595168 | 0.6044689 |
| 1977 | 304 | 0.06369107 | 0.1044638 |
| 1978 | 232 | -0.24907508 | 0.1131963 |
| 1979 | 313 | -0.21623841 | 0.1128258 |
| 1980 | 106 | -0.8545272 | 0.1464227 |
| 1981 | 10 | -1.46262129 | 0.431476 |
| 1982 | 15 | -1.25247236 | 0.3515014 |
| 1983 | 52 | -0.67873822 | 0.2232734 |
| 1984 | 211 | -0.52057286 | 0.135287 |
| 1985 | 41 | -0.33559598 | 0.2228306 |
| 1986 | 0 | na | na |
| 1987 | 0 | na | na |
| 1988 | 368 | -0.03086541 | 0.0867818 |
| 1989 | 1637 | 0.51658905 | 0.0649446 |
| 1990 | 4374 | -0.03188192 | 0.0427326 |
| 1991 | 3007 | -0.65147626 | 0.0455893 |
| 1992 | 2392 | -0.48332143 | 0.0492012 |
| 1993 | 244 | -2.12548215 | 0.0978487 |
| 1994 | 124 | -3.71358685 | 0.1276957 |
| 1995 | 6 | -3.65761952 | 0.548426 |
| 1996 | 123 | -2.17860376 | 0.1686357 |
| 1997 | 16 | -0.73948754 | 0.3393005 |
| 1998 | 40 | -2.35003708 | 0.2294059 |
| 1999 | 177 | -2.42845091 | 0.1444615 |
| 2000 | 22 | -2.08389238 | 0.288056 |
| 2001 | 94 | -1.90463936 | 0.1449361 |
| 2002 | 140 | -2.85274685 | 0.1629792 |
| 2003 | 144 | -1.66609054 | 0.1322349 |
| 2004 | 89 | -2.14857017 | 0.1856508 |
| 2005 | 55 | -1.07553493 | 0.3548567 |
| 2006 | 261 | 0.49576899 | 0.1070256 |
| 2007 | 358 | 0.66449746 | 0.0819418 |
| 2008 | 1530 | 0.24979789 | 0.0578003 |
| 2009 | 710 | -0.8291972 | 0.0717022 |
| 2010 | 255 | -0.37128087 | 0.1019519 |
| 2011 | 500 | 0.427879 | 0.0818517 |
| 2012 | 493 | 0.15186911 | 0.0808314 |
| 2013 | 435 | -0.32866291 | 0.1015757 |
| 2014 | 947 | -0.07650472 | 0.0873358 |
| Total | 19912 |  |  |

Table 16.3.1.1. Number of hauls in the Greenland Shrimp and Fish survey in ICES XIVb and NAFO 1F.

|  | ICES XIVb |  |  | NAFO |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year/Strata | Q1 | Q2 | Q3 | Q4 | Q5 | Q6 | 1F | Total |
| 1992 |  |  |  |  |  |  | 15 |  |
| 1993 |  |  |  |  |  |  | 13 |  |
| 1994 |  |  |  |  |  |  | 9 |  |
| 1995 |  |  |  |  |  |  | 11 |  |
| 1996 |  |  |  |  |  |  | 11 |  |
| 1997 |  |  |  |  |  |  | 19 |  |
| 1998 |  |  |  |  |  |  | 14 |  |
| 1999 |  |  |  |  |  |  | 17 |  |
| 2000 |  |  |  |  |  |  | 29 |  |
| 2001 |  |  |  |  |  |  | 26 |  |
| 2002 |  |  |  |  |  |  | 27 |  |
| 2003 |  |  |  |  |  |  | 22 |  |
| 2004 |  |  |  |  |  |  | 34 |  |
| 2005 |  |  |  |  |  |  | 23 |  |
| 2006 |  |  |  |  |  |  | 31 |  |
| 2007 |  |  |  |  |  |  | 39 |  |
| 2008 | 8 | 6 | 12 | 7 | 7 | 11 | 47 | 98 |
| 2009 | 22 | 11 | 25 | 20 | 6 | 13 | 48 | 145 |
| 2010 | 19 | 14 | 24 | 9 | 6 | 10 | 40 | 122 |
| 2011 | 20 | 11 | 21 | 12 | 7 | 14 | 25 | 110 |
| 2012 | 20 | 16 | 28 | 13 | 7 | 15 | 26 | 125 |
| 2013 | 25 | 12 | 22 | 14 | 5 | 14 | 28 | 120 |
| 2014 | 22 | 14 | 12 | 9 | 8 | 16 | 32 | 113 |

Table 16.3.1.2 Cod abundance indices ('000) from the Greenland Shrimp and Fish survey by year and strata divisions in ICES XIVb and NAFO 1F. Q1 being the northern strata in East Greenland.

| ICES XIVb |  |  |  |  | NAFO |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Q1 | Q2 | Q3 | Q4 | Q5 | Q6 | 1F | Total | CV |
| 1992 |  |  |  |  |  |  | 8 |  |  |
| 1993 |  |  |  |  |  |  | 18 |  |  |
| 1994 |  |  |  |  |  |  | 0 |  |  |
| 1995 |  |  |  |  |  |  | 39 |  |  |
| 1996 |  |  |  |  |  |  | 107 |  |  |
| 1997 |  |  |  |  |  |  | 0 |  |  |
| 1998 |  |  |  |  |  |  | 3 |  |  |
| 1999 |  |  |  |  |  |  | 0 |  |  |
| 2000 |  |  |  |  |  |  | 189 |  |  |
| 2001 |  |  |  |  |  |  | 313 |  |  |
| 2002 |  |  |  |  |  |  | 457 |  |  |
| 2003 |  |  |  |  |  |  | 211 |  |  |
| 2004 |  |  |  |  |  |  | 1610 |  |  |
| New survey Gear Introduced |  |  |  |  |  |  |  |  |  |
| 2005 |  |  |  |  |  |  | 86410 |  |  |
| 2006 |  |  |  |  |  |  | 39475 |  |  |
| 2007 |  |  |  |  |  |  | 32575 |  |  |
| 2008 | 5456 | 1361 | 13043 | 1975 | 1635 | 7958 | 22887 | 54314 | 22 |
| 2009 | 14304 | 2191 | 28539 | 4374 | 548 | 4753 | 1776 | 56486 | 15 |
| 2010 | 5844 | 732 | 30042 | 3975 | 115 | 4633 | 6557 | 51897 | 45 |
| 2011 | 7843 | 1357 | 5178 | 7733 | 1470 | 19072 | 6330 | 48983 | 22 |
| 2012 | 5475 | 2164 | 3658 | 2453 | 352 | 8635 | 21238 | 43975 | 20 |
| 2013 | 11102 | 1420 | 5667 | 17360 | 537 | 27145 | 49874 | 113104 | 32 |
| 2014 | 4168 | 3445 | 2622 | 19267 | 493 | 5412 | 22702 | 58106 | 36 |

Table 16.3.1.3. Cod biomass indices (tons) from the Greenland Shrimp and Fish survey by year and strata divisions in ICES XIVb (Q1-Q6) and NAFO 1F. Smoothed index is a random effects survey smoother applied to the total index.

| ICES XIVb |  |  | NAFO |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Q1 | Q2 | Q3 | Q4 | Q5 | Q6 | 1F | Total | CV | Smoothed index |
| 1992 |  |  |  |  |  |  | 2 |  |  |  |
| 1993 |  |  |  |  |  |  | 5 |  |  |  |
| 1994 |  |  |  |  |  |  | 0 |  |  |  |
| 1995 |  |  |  |  |  |  | 4 |  |  |  |
| 1996 |  |  |  |  |  |  | 49 |  |  |  |
| 1997 |  |  |  |  |  |  | 0 |  |  |  |
| 1998 |  |  |  |  |  |  | 3 |  |  |  |
| 1999 |  |  |  |  |  |  | 0 |  |  |  |
| 2000 |  |  |  |  |  |  | 46 |  |  |  |
| 2001 |  |  |  |  |  |  | 100 |  |  |  |
| 2002 |  |  |  |  |  |  | 150 |  |  |  |
| 2003 |  |  |  |  |  |  | 46 |  |  |  |
| 2004 |  |  |  |  |  |  | 305 |  |  |  |
| New survey Gear Introduced |  |  |  |  |  |  |  |  |  |  |
| 2005 |  |  |  |  |  |  | 56163 |  |  |  |
| 2006 |  |  |  |  |  |  | 16828 |  |  |  |
| 2007 |  |  |  |  |  |  | 23346 |  |  |  |
| 2008 | 8692 | 2430 | 24101 | 1482 | 2173 | 8838 | 21232 | 68948 | 23 | 68370 |
| 2009 | 10844 | 8874 | 27251 | 7827 | 252 | 3094 | 502 | 58644 | 28 | 66876 |
| 2010 | 16014 | 3151 | 81064 | 6202 | 23 | 4203 | 3142 | 113798 | 51 | 82766 |
| 2011 | 27064 | 8128 | 5561 | 12486 | 5235 | 22665 | 3279 | 84417 | 19 | 85030 |
| 2012 | 24732 | 10058 | 9347 | 5802 | 160 | 14322 | 16212 | 80639 | 16 | 89797 |
| 2013 | 45018 | 9639 | 15017 | 48518 | 977 | 40319 | 47818 | 207306 | 22 | 168570 |
| 2014 | 17182 | 20637 | 15574 | 90795 | 734 | 8884 | 30751 | 184558 | 45 | 175160 |

Table 16.3.1.4: Abundance indices (' ${ }^{\prime} 000$ ) by age from the Greenland Shrimp and Fish survey by year in ICES XIVb + NAFO 1F.

## East Greenland

| Year/age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | $10+$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2008 | 4355 | 372 | 1113 | 7968 | 6582 | 23794 | 5412 | 2235 | 736 | 1006 | 739 |
| 2009 | 14970 | 7642 | 8019 | 4504 | 5378 | 5664 | 6610 | 2537 | 225 | 554 | 385 |
| 2010 | 150 | 2436 | 3959 | 5759 | 3253 | 12785 | 7969 | 11264 | 2958 | 450 | 914 |
| 2011 | 315 | 162 | 5682 | 8288 | 16346 | 5409 | 4707 | 2226 | 3382 | 1834 | 634 |
| 2012 | 0 | 258 | 1208 | 12748 | 7154 | 12041 | 4155 | 2428 | 1345 | 1849 | 790 |
| 2013 | 0 | 157 | 1432 | 1954 | 44843 | 25373 | 26654 | 5209 | 3440 | 1852 | 2190 |
| 2014 | 692 | 15 | 207 | 1849 | 1558 | 21863 | 8805 | 12411 | 2875 | 3790 | 4041 |

Table 16.3.1.5 The abundance indices ('000) by year class/age from the Greenland Shrimp and Fish survey subareas in ICES XIVb and NAFO 1F, 2014.

| Year class | 2014 | 2013 | 2012 | 2011 | 2010 | 2009 | 2008 | 2007 | 2006 | $2005<2005$ |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | $10+$ |
| ICES Q1 | 454 | 0 | 0 | 0 | 76 | 644 | 591 | 1291 | 515 | 238 | 358 |
| ICES Q2 | 0 | 0 | 4 | 0 | 0 | 107 | 519 | 1099 | 495 | 655 | 566 |
| ICES Q3 | 98 | 0 | 0 | 309 | 0 | 154 | 451 | 752 | 309 | 0 | 549 |
| ICES Q4 | 0 | 0 | 38 | 33 | 27 | 1190 | 3575 | 7759 | 1309 | 2799 | 2536 |
| ICES Q5 | 97 | 15 | 0 | 0 | 8 | 186 | 160 | 25 | 1 | 0 | 0 |
| ICES Q6 | 43 | 0 | 88 | 1142 | 86 | 2486 | 663 | 608 | 199 | 67 | 31 |
| NAFO 1F | 0 | 0 | 76 | 366 | 1361 | 17096 | 2846 | 877 | 47 | 32 | 0 |

Table 16.3.2.1 German survey. Numbers of valid hauls by stratum in South and East Greenland, stratum 9 furthest to the north.


Table 16.3.2.2 German survey. Cod abundance indices ('000) from the German survey in South and East Greenland by year and stratum.

| NAFO 1 F |  |  | ICES XIVb |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | str4_1 | str4_2 | str5_1 | str5_2 | str7_1 | str7_2 | str8_2 | str9_2 | Sum | SD |
| 1982 | 8540 | 1245 | . | 366 | 297 | 1493 | 664 | 385 | 12990 | 4973 |
| 1983 | 5267 | 2870 | 209 | 715 | 149 | 564 | 529 | 726 | 11029 | 3796 |
| 1984 | 3296 | 42 | 1268 | 413 | 138 | 750 | 173 | 333 | 6413 | 3845 |
| 1985 | 3492 | 1164 | 920 | 166 | 560 | 1554 | 401 | 310 | 8567 | 1978 |
| 1986 | 8967 | 492 | 3509 | 359 | 776 | 2641 | 1207 | 337 | 18288 | 5097 |
| 1987 | 23219 | 306 | 5655 | 4145 | 399 | 6298 | 1293 | 234 | 41549 | 14816 |
| 1988 | 28259 | 17 | 2590 | 2073 | 302 | 1175 | 738 | 601 | 35755 | 16719 |
| 1989 | 31810 | 31442 | 9979 | . | 880 | . | 2128 | 639 | 76878 | 42682 |
| 1990 | 7052 | 6306 | 2808 | 1155 | 861 | 4295 | 2799 | 468 | 25744 | 7720 |
| 1991 | 1367 | 233 | 790 | 937 | 122 | 368 | 652 | 510 | 4979 | 1548 |
| 1992 | 113 | 134 | . | . | . | . | 228 | 367 | 842 | 192 |
| 1993 | 0 | . | 613 | 62 | 127 | 317 | 114 | 148 | 1381 | 521 |
| 1994 | 44 | 12 | . | . | . | . | - | 234 | 290 | 135 |
| 1995 | 27 | 8 | 89 | 25 | 450 | 3082 | 77 | 91 | 3849 | 1314 |
| 1996 | 156 | 0 | 109 | 0 | 37 | 279 | 29 | 160 | 770 | 173 |
| 1997 | 49 | 0 | 25 | 17 | 200 | 54 | 145 | 1107 | 1597 | 479 |
| 1998 | 40 | 8 | 97 | 0 | 57 | 57 | 24 | 266 | 549 | 142 |
| 1999 | 155 | 0 | 198 | 8 | 165 | 1267 | 116 | 105 | 2014 | 582 |
| 2000 | 76 | 13 | 348 | 15 | 431 | 180 | 25 | 143 | 1231 | 251 |
| 2001 | 343 | 3 | 319 | 27 | 309 | 299 | 204 | 1071 | 2575 | 544 |
| 2002 | 1739 | 0 | 116 | 273 | 769 | 459 | 186 | 875 | 4417 | 1352 |
| 2003 | 840 | 8 | 199 | 183 | 1250 | 1399 | 1100 | 1438 | 6417 | 1004 |
| 2004 | 10902 | 107 | 1684 | 133 | 285 | 1817 | 1401 | 1073 | 17402 | 8499 |
| 2005 | 24438 | 1399 | 16577 | 3078 | 718 | 7157 | 1580 | 2070 | 57017 | 11411 |
| 2006 | 28894 | 486 | 14733 | 3686 | 6044 | 7378 | 2779 | 2700 | 66700 | 15653 |
| 2007 | 67049 | 772 | 2283 | 3256 | 758 | 5363 | 2080 | 2093 | 83654 | 56843 |
| 2008 | 18730 | 292 | 2036 | 4898 | 2203 | 9460 | 1285 | 2678 | 41582 | 10268 |
| 2009 | 1286 | 283 | 1017 | 567 | 3129 | 8755 | 1566 | 3275 | 19878 | 3581 |
| 2010 | 2372 | 141 | 532 | 1703 | 1101 | 8875 | 933 | 1748 | 17405 | 2958 |
| 2011 | 7547 | 162 | 3027 | 1326 | 868 | 1971 | 1243 | 2816 | 18960 | 3196 |
| 2012 | 23964 | 132 | 5689 | 167 | 901 | 2117 | 1114 | 3982 | 38066 | 22168 |
| 2013 | 41722 | 1947 | 2193 | 818 | 874 | 3121 | 1157 | 1342 | 53174 | 43105 |
| 2014 | 73612 | 111 | 8612 | 4013 | 228 | 1089 | 1436 | 5461 | 94562 | 77704 |

Table 16.3.2.3 German survey. Cod biomass indices (tons) from the German survey in South and East Greenland by year and stratum.

| NAFO 1 F |  |  | ICES XIVb |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | str4_1 | str4_2 | str5_1 | str5_2 | str7_1 | str7_2 | str8_2 | str9_2 | Sum | SD |
| 1982 | 14607 | 3690 | . | 1201 | 1036 | 3342 | 2576 | 1900 | 28352 | 8415 |
| 1983 | 9797 | 6219 | 653 | 2209 | 402 | 2294 | 2605 | 4442 | 28621 | 8201 |
| 1984 | 5326 | 82 | 3115 | 1444 | 346 | 1782 | 540 | 2553 | 15188 | 6650 |
| 1985 | 2942 | 1976 | 1812 | 803 | 1393 | 3875 | 1187 | 1605 | 15593 | 3099 |
| 1986 | 8005 | 943 | 1044 | 873 | 2537 | 3921 | 2301 | 709 | 20333 | 6054 |
| 1987 | 17186 | 276 | 2889 | 3735 | 504 | 10243 | 4558 | 1414 | 40805 | 16521 |
| 1988 | 26349 | 17 | 2812 | 4605 | 964 | 2297 | 3475 | 2012 | 42531 | 18651 |
| 1989 | 36912 | 35281 | 23605 | . | 2518 | . | 6889 | 2174 | 107379 | 61579 |
| 1990 | 9212 | 5897 | 5361 | 3215 | 2517 | 10386 | 6551 | 1620 | 44759 | 10905 |
| 1991 | 2088 | 200 | 1465 | 2759 | 196 | 1008 | 2610 | 2100 | 12426 | 4657 |
| 1992 | 79 | 50 | . | . | . | . | 171 | 734 | 1034 | 286 |
| 1993 | 0 | . | 431 | 73 | 247 | 532 | 254 | 547 | 2084 | 588 |
| 1994 | 2 | 7 | . | . | - | - | - | 779 | 788 | 514 |
| 1995 | 6 | 4 | 32 | 62 | 166 | 11744 | 250 | 123 | 12387 | 5550 |
| 1996 | 101 | 0 | 63 | 0 | 109 | 708 | 99 | 511 | 1591 | 333 |
| 1997 | 53 | 0 | 18 | 20 | 358 | 70 | 337 | 4017 | 4873 | 1800 |
| 1998 | 12 | 11 | 29 | 0 | 87 | 122 | 123 | 986 | 1370 | 554 |
| 1999 | 39 | 0 | 24 | 1 | 162 | 2229 | 492 | 201 | 3148 | 1184 |
| 2000 | 13 | 9 | 132 | 17 | 206 | 616 | 75 | 540 | 1608 | 366 |
| 2001 | 88 | 5 | 130 | 19 | 345 | 382 | 387 | 3005 | 4361 | 1593 |
| 2002 | 976 | 0 | 38 | 224 | 1547 | 531 | 541 | 2214 | 6071 | 1306 |
| 2003 | 361 | 17 | 121 | 266 | 3787 | 2440 | 1716 | 4169 | 12877 | 2817 |
| 2004 | 1945 | 177 | 359 | 55 | 957 | 2319 | 3264 | 3240 | 12316 | 3070 |
| 2005 | 9055 | 1870 | 8135 | 2537 | 3155 | 17882 | 3590 | 6806 | 53030 | 7772 |
| 2006 | 31616 | 681 | 8616 | 4130 | 3557 | 10291 | 6084 | 11567 | 76542 | 24680 |
| 2007 | 74671 | 1045 | 3749 | 5042 | 1363 | 14456 | 5374 | 8540 | 114240 | 58452 |
| 2008 | 18543 | 344 | 3630 | 9790 | 5075 | 26506 | 3772 | 11908 | 79568 | 12433 |
| 2009 | 583 | 277 | 1361 | 1726 | 10145 | 28613 | 6351 | 15520 | 64576 | 13358 |
| 2010 | 3629 | 273 | 741 | 5085 | 5244 | 31745 | 4282 | 10932 | 61931 | 11626 |
| 2011 | 12398 | 385 | 5839 | 4364 | 1658 | 8051 | 5735 | 17487 | 55917 | 10240 |
| 2012 | 33871 | 370 | 15679 | 579 | 2596 | 6245 | 5445 | 26885 | 91670 | 30054 |
| 2013 | 74193 | 6525 | 6672 | 2737 | 2577 | 9752 | 4853 | 7575 | 114884 | 75148 |
| 2014 | 132706 | 428 | 31885 | 15935 | 1060 | 4322 | 6480 | 29358 | 222174 | 132209 |

Table 16.3.2.4 German survey, South and East Greenland (NAFO 1F and ICES XIV). Age disaggregate abundance indices ('1000 ).

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | $11+$ | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 |  | 23 | 214 | 2500 | 1760 | 4451 | 1952 | 793 | 223 | 927 | 57 | 74 | 12974 |
| 1983 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1984 | 23 | 8 | 54 | 1134 | 507 | 2434 | 582 | 1242 | 229 | 125 | 17 | 49 | 6404 |
| 1985 | 279 | 2521 | 242 | 160 | 1658 | 947 | 1439 | 344 | 831 | 96 | 27 | 27 | 8571 |
| 1986 |  | 3367 | 9255 | 1128 | 273 | 1631 | 603 | 1300 | 165 | 473 | 31 | 58 | 18284 |
| 1987 |  | 4 | 10193 | 24656 | 2689 | 720 | 1368 | 296 | 966 | 80 | 487 | 49 | 41508 |
| 1988 | 6 | 18 | 335 | 9769 | 23391 | 876 | 200 | 559 | 83 | 337 | 31 | 146 | 35751 |
| 1989 | 12 | 2 | 111 | 732 | 23945 | 49864 | 1007 | 44 | 756 | 70 | 282 | 76 | 76901 |
| 1990 | 58 | 36 | 58 | 715 | 706 | 11679 | 12101 | 139 | 15 | 74 |  | 148 | 25729 |
| 1991 |  | 73 | 150 | 171 | 539 | 102 | 2128 | 1762 | 31 | 11 | 3 | 9 | 4979 |
| 1992 | 214 | 10 | 196 | 103 | 61 | 53 | 67 | 67 | 51 |  |  | 21 | 822 |
| 1993 |  | 4 | 15 | 869 | 152 | 95 | 97 | 31 | 83 | 34 |  | 2 | 1382 |
| 1994 |  | 71 | 5 | 16 | 84 | 39 | 22 | 38 |  | 8 |  | 0 | 283 |
| 1995 |  | 1 | 621 | 347 | 260 | 1399 | 372 | 120 | 403 | 32 | 192 | 102 | 3849 |
| 1996 |  | 0 | 0 | 353 | 130 | 131 | 110 | 23 | 25 |  |  | 0 | 772 |
| 1997 |  | 0 | 12 | 17 | 687 | 557 | 191 | 78 | 48 |  |  | 5 | 1595 |
| 1998 | 51 | 73 | 39 | 4 | 11 | 173 | 138 | 48 | 10 |  |  | 0 | 547 |
| 1999 | 105 | 426 | 389 | 346 | 118 | 257 | 174 | 156 |  | 29 | 16 | 0 | 2016 |
| 2000 |  | 202 | 243 | 323 | 208 | 40 | 72 | 20 | 46 | 61 | 15 | 0 | 1230 |
| 2001 |  | 166 | 568 | 493 | 631 | 362 | 190 | 60 | 50 | 18 | 10 | 2 | 2550 |
| 2002 | 40 | 1 | 395 | 2119 | 601 | 477 | 454 | 217 | 61 | 21 | 11 | 7 | 4404 |
| 2003 | 579 | 629 | 53 | 553 | 1761 | 1026 | 1015 | 541 | 220 | 37 | - | 4 | 6418 |
| 2004 | 386 | 10687 | 1770 | 448 | 617 | 1667 | 921 | 620 | 228 | 39 | 10 | 8 | 17401 |
| 2005 | 80 | 1603 | 39549 | 8091 | 1250 | 2819 | 2549 | 727 | 189 | 40 |  | 0 | 56897 |
| 2006 | 80 | 439 | 3375 | 48140 | 9269 | 1328 | 2404 | 1309 | 193 | 30 | 9 | 0 | 66576 |
| 2007 | 128 | 154 | 2007 | 5149 | 65974 | 8166 | 713 | 658 | 634 | 70 |  | 0 | 83653 |
| 2008 | 14 | 265 | 513 | 8213 | 4401 | 22939 | 4201 | 516 | 220 | 199 | 44 | 29 | 41554 |
| 2009 | 98 | 322 | 1057 | 391 | 1620 | 2863 | 11241 | 1964 | 111 | 134 | 64 | 17 | 19882 |
| 2010 | 22 | 700 | 1425 | 1388 | 845 | 2887 | 2518 | 5707 | 1362 | 236 | 163 | 139 | 17392 |
| 2011 |  | 120 | 1246 | 3475 | 4874 | 2402 | 2949 | 1179 | 2324 | 310 | 23 | 49 | 18951 |
| 2012 | 6 | 50 | 1624 | 10093 | 10233 | 9846 | 2827 | 1778 | 1166 | 379 | 35 | 5 | 38042 |
| 2013 |  | 17 | 35 | 4312 | 27014 | 11146 | 7455 | 1314 | 517 | 291 | 126 | 68 | 52295 |
| 2014 |  | 7 | 55 | 602 | 20847 | 58174 | 9275 | 3284 | 1316 | 494 | 441 | 52 | 94547 |

Table 16.5.1. Number of tagged cod in the period of 2003 to 2014 in different regions. Bank (West) = NAFO division 1D+1E. East Greenland = NAFO subdivision 1F + ICES subarea XIV.

| Year | Fjord | Bank (West) | East Greenland |
| :--- | :--- | :--- | :--- |
| 2003 | 599 |  |  |
| 2004 | 658 |  |  |
| 2005 | 565 | 721 | 1387 |
| 2006 | 41 |  | 1296 |
| 2007 | 1140 |  | 525 |
| 2008 | 231 |  | 403 |
| 2009 | 633 | 1563 | 2359 |
| 2010 | 28 | 2321 | 1203 |
| 2011 | 183 |  |  |
| 2012 |  |  |  |
| 2013 |  |  |  |
| 2014 |  |  |  |

Table 16.5.2: Number of recaptured cod in the period of 2003 to 2014 in different regions. Bank (West) = NAFO division 1D+1E. East Greenland = NAFO subdivision 1F + ICES subarea XIV.

|  | Fjord (West) | Bank (West) | East Greenland |
| :--- | :--- | :--- | :--- |
| Fjord (West) | 436 | 8 | 3 |
| Bank (West) |  | 21 | 2 |
| East Greenland |  | 5 | 89 |
| Fjord (East) |  | 13 | 1 |
| Iceland | 3 |  | 98 |



Figure. 16.1. Sampling location of spawning cod in Greenland and Iceland in the genetic project. The colours of the dots represent the blends of sample mean of the different spawning population: West offshore, Nuuk (inshore), East (Greenland and offshore Iceland) and Iceland inshore as signal intensities of green and red respectively. After Therkildsen et al. 2013.


Figure 16.2.1. Annual total catch in South and East Greenland (NAFO subarea 1F and ICES subarea XIVb). From 2001 divided into gear. TAC until 2013 is for all the offshore area including West Greenland (NAFO subarea 1A-1E).


Figure 16.2.2.1: Annual distribution of total catches of Atlantic cod in West and East Greenland. Q1-Q6 illustrates survey areas (strata) in the East Greenland shrimp and fish survey.


Figure 16.2.2.1: Continued. Annual distribution of total catches of Atlantic cod in West and East Greenland. Q1-Q6 illustrates survey areas (strata) in the East Greenland shrimp and fish survey.


Figure 16.2.2.2: Distribution of Longline and Trawl catches of Atlantic cod in West and East Greenland 2014. Q1-Q6 illustrates survey areas (strata) in the East Greenland shrimp and fish survey.
i)


Figure 16.2.3.1: Combined length and age distributions of commercial cod catches in the South and East Greenland offshore fishery in 2014.


Figure 16.2.3.2: Length and age distributions of commercial cod catches in the four management areas of SouthWest (NAFO 1F) and East Greenland (Q1Q2 furthest north) in 2014.


Figure 16.2.4.1: Ln CPUE (ton/hr) for Atlantic Cod caught in the fishery in East (ICES XIVb) and SouthWest (NAFO 1F) Greenland. Based on model: lncpue = year + management area (Q1Q2, Q3Q4, Q5Q6 and 1F) + ship. Dashed lines are 2*SE.


Figure16.3.1.1. Greenland shrimp and fish survey 2008-2014. Abundance per $\mathrm{Km}^{2}$.


Figure 16.3.1.2. Greenland shrimp and fish survey 2008-2014. Catch weight $\mathbf{~ k g ~ p e r ~} \mathbf{K m}^{2}$


Figure 16.3.1.3: Abundance index pr. age in ICES XIVb and NAFO 1F combined. Size of circles represents size of index.








Figure 16.3.1.4: Abundance index by length (cm) and area in 2014. Areas from north (top) to south (bottom ) is: Q1, Q2, Q3, Q4, Q5,Q6 (ICES XIVb) and NAFO 1F.


Figure 16.3.1.5. Abundace ( $\mathrm{no} / \mathrm{km}^{2}$ ) pr. station of ages 1-9 in the years 2008-2014.


Figure 16.3.1.6: Total abundance indices by length in East Greenland (ICES XIVb + NAFO 1F) shrimp and fish survey.


Figure 16.3.1.7: Estimated SSB (tons) by year from the East Greenland (ICES XIVb + NAFO 1F) Shrimp and Fish survey.


Figure 16.3.1.8: Estimated SSB (tons) by survey areas from the East Greenland (ICES XIVb + NAFO 1F) Shrimp and Fish survey, 2014. NAFO Div 1F (SouthWest Greenland) to the left, "Q" areas (East Greenland) to the right. Cape Farewell is between 1F and Q6.


Figure 16.3.2.1 German survey, 2014. Abundance (num per km2) pr haul.



Figure 16.3.2.2 German survey, 2014. Biomass (kg per km2) pr haul.


Figure 16.3.2.3 German survey, Cod off Greenland. Abundance per age group and stratum. Strata 1 -4 is West Greenland from north to south; strata 5-9 is East Greenland from south to north.


Figure 16.3.2.4 German survey, Cod off Greenland. Abundance indices for South and East Greenland.


Figure 16.3.2.5 German survey, Cod off Greenland. Biomass indices for South and East Greenland.


Figure 16.3.3.1: Biomass index for NAFO 1F and ICES Subarea XIVb. Red squares are the estimated mean value from the survey and the vertical connected lines are upper and lower $95 \%$ confidence intervals. The smoothed estimates are displayed as the blue line and the $95 \%$ confidence intervals of the smoothed values are shown as dashed lines.



Figure 16.3.4.1: Catch, catch advice with different cap's on the 3.2 DLS approach and German survey index (without scale). The two figures show the same, but with two different scales on the Y -axis.


Figure 16.3.4.2: Catch (black), catch advice with setoff in year 1994 (blue), catch advice with setoff in year 1997 (green). Both $20 \%$ cap and 0.8 buffer were used for the two catch-advice calculations.


Figure 16.3.3.4: Log catch ratios in the period 2005-2014 using commercial data.


Figure 16.5.1: Tag and recapture of Greenlandic tagged cod in the period 2003-2014.

## 17 Greenland Halibut in Subareas V, VI, XII, and XIV

Greenland halibut in ICES Subareas V, VI, XII and XIV are assessed as one stock unit although precise stock associations are not known.

The stock was benchmarked in November 2013 (ICES 2013, WKBUT) and changes for the stock assessment are provided in the Stock Annex. The NWWG suggested further changes from the benchmark, however, the advice drafting group referred to the benchmark decision in their minutes as follows:
"The ADG used the stock production model as a basis for the advice.
A number of issues were raised regarding the model. The early part of the CPUE series (pre1995) may not be consistent with the later part. The model fitted with this series had a substantial process error component particularly in the early part of the series. This results in high autocorrelation in the error structure which increases the uncertainties in the predictions. The latter half of the assessment fits well to both CPUE and survey tuning series. It was considered that the main issue with the assessment was the relative magnitude of the stock and exploitation rate in the early part of the series. Although the error structure was a considerable concern the advantages of using the full catch time series to obtain plausible population growth parameters appear to out-weigh the problems that were apparent using a truncated series. The catch advice derived from this model was similar to the catch advice from DLS 3.3 analysis based on catch and the survey tuning series. The NWWG discussed this at length and although the majority considered the DLS approach was preferable it was not possible to reach consensus. The ADG commends the NWWG for their effort in providing a number of WGDOCs on this topic; however, the ADG would have been grateful for a synthesis of the issues. It is recognised that one of the reasons why this might not have been done was the lack of consensus in the NWWG. Given the outcome of the benchmark conducted in November 2013, and the considerations given above the ADG decided it was preferable to base the advice on the surplus production model, than to reject the benchmark and use the DLS approach. This should not be taken as an endorsement of the current model, but as interim solution for the advice this year. The NWWG should organise inter-sessional evaluation of CPUE series and modelling environment (stock production models, Gadget). If progress can be made and a report produced before the end of December 2014, ICES would put in place an inter-benchmark process to review any new proposed assessment."

### 17.1 Executive summary

Input data to the assessment: current surveys have continued and sampling intensity and coverage remains also unchanged. Logbooks from the fishery are available as haul by haul data. Since 2001 no age readings of otoliths were available from the main fishing areas which impede age based assessment.
A logistic production model in a Bayesian framework has been used to assess stock status and for making predictions. The model includes an extended catch series going back to the assumed virgin status of the stock at the beginning of the fishery in 1961. Estimated stock biomass showed an overall decline along with the high catches in the late 1980s and early 1990s. Since 2004/2005 the stock has increased slowly and is now at $70 \% \mathrm{~B}_{\text {msy }}$ and fishing mortality has decreased to $\mathrm{F}_{\text {msy }}$. Although the indices that are used for input to the assessment model (combined survey index at Greenland and Iceland) and logbook information from Iceland trawler fishery all show a slight decrease
in 2014, the remaining available indices from East Greenland (logbook from trawl fishery) and from Faroese waters (logbooks from trawl fishery and a survey) all suggest high and stable biomass in recent years.

### 17.2 Catches, Fisheries, Fleet and Stock Perception

### 17.2.1 Catches

Total annual catches in Divisions Va, Vb, and Subareas VI, XII and XIV are presented for the years 1981-2014 in Tables 17.2.1-17.2.6 and since 1961 in Figure 17.2.1. Catches decreased in 2014 by $22 \%$ to 21.069 t mainly due to TAC restrictions.
Landings in Icelandic waters (usually allocated to Division Va) have historically predominated the total landings in areas V+XIV, but since the mid 1990s also fisheries in XIV and Vb have developed. Landings have since 1997 been between 20 and 31 kt .

### 17.2.2 Fisheries and fleets

In 2014 quotas in Greenland EEZ and Iceland EEZ were almost fully utilized as in the preceding fishing years. In the Faroe EEZ the fishery is regulated by a fixed numbers of licenses and technical measures like by-catch regulations for the trawlers and depth and gear restrictions for the gillnetters.
Most of the fishery for Greenland halibut in Divisions $\mathrm{Va}, \mathrm{Vb}$ and XIVb is a directed trawl fishery, and also an insignificant gillnet and longline fishery takes place. Only minor catches in Va and XIVb are taken as by-catches in a redfish fishery (see section 21 on Greenland slope redfish). No or insignificant discarding has been observed in this fishery.

Spatial distribution of 2014 fishery and historic effort and catch in the trawl fishery in XIV and V is provided in Figures 17.2.2-5. Fishery in the entire area did in the past occur in a more or less continuous belt on the continental slope from the slope of the Faroe plateau to southeast of Iceland extending north and west of Iceland and further south to southeast Greenland. Fishing depth ranges from $350-500 \mathrm{~m}$ southeast, east and north of Iceland to about 1500 m at East Greenland. In 2014 the distribution of the fishery covered all areas but was discontinuous in its distribution.

In 2001-2008 a directed and a by-catch fishery by Spain, France, Lithuania, UK and Norway developed in the Hatton Bank area of Division VIb. However, most of these fisheries ceased after 2008 and is presently insignificant. Landings in Divisions XII and VIb in Tables 17.2.5-17.2.6 derive from the Hatton Bank area.

### 17.2.3 By-catch and discard

The Greenland halibut trawl fishery is commonly a clean fishery with respect to bycatches. Eventual by-catches are mainly redfish and cod. Southeast of Iceland the cod fishery and a minor Greenland halibut fishery are coinciding spatially. In East Greenland where fishery is on the steep slope, fishing grounds for cod and redfish are close to the Greenland halibut fishing grounds, but nevertheless the catches from single hauls are clean.

The mandatory use of sorting grids in the shrimp fishery in Icelandic and Greenland waters since 2002 are observed to have reduced by-catches considerably. Based on sampling in 2006-2007, scientific staff observed by-catches of Greenland halibut to be less
than $1 \%$ compared to about $50 \%$ by weight observed before the implementation of sorting grids (Sünksen 2007, WD \# 18). No information has since been available but the fishery in XIVb report discard less than $1 \%$ by weight.

### 17.3 Trends in Effort and CPUE

### 17.3.1 Division Va

Indices of CPUE for the Icelandic trawl fleet directed at Greenland halibut for the period 1985-2014 is provided in Table 17.3.1 and Figures 17.3.1-3. At the benchmark (WKBUT 2013) the CPUE series from this fishery was questioned due to a marked change in season and area, and also because the regulations might have caused a changed behaviour in the fishing fleets. The important fishing grounds west of Iceland, where approximately $70-80 \%$ of the landings historically came from, are the areas where the season shift mainly has affected the CPUE. A simple standardization procedure was not considered sufficient to account for these changes (Figure 17.3.2.). Therefore a rough estimate on stock biomass distribution in Iceland in each of the fuor areas is that each area account equal. The overall CPUE index for the Icelandic fishery was therefore compiled as the average of the standardised indices from the four areas (Fig 17.3.1-2.).

Catch rates of Icelandic bottom trawlers decreased for all fishing grounds during 1990-1996 (Figure 17.3.1) but have since peaked in 2001 and have in recent years been stable and slowly increasing. The overall tendency is the same for all fishing grounds in Va (Figure 15.3.2) although the less important fishing grounds in north, east and southeast more variable in trend. Both observed and derived effort are about historic average in 2014 (Figure 17.3.3).

### 17.3.2 Division Vb

Information from logbooks from the Faroese otterboard trawl fleet ( $>1000 \mathrm{hp}$ ) was available for the years 1991-2014 (Table 17.3.1, Figure 17.3.4.). The bulk of the fishery has historically been on the south-east slope of the Faroe Plateau. CPUE decreased drastically in the early period by more than $50 \%$ coinciding with a significant increase in effort. In 2011 CPUE increased sharply by more than $60 \%$ and remained at that level in 2012 and 2013.

### 17.3.3 Division XIVb

CPUE and effort from logbooks in XIV are provided in Table 17.3.1 and Figure 17.3.67. In 2005-2008 catch rates were high and above the average, but decreased by nearly $20 \%$ in 2009-2011 along with a massive increase in effort ( $85 \%$ ). CPUE in 2012-14 has increased being at a 2009-2011 level. The CPUE series from Divisions Va, Vb and XIVb have different trends over the time series indicating that the populations/areas are affected by different dynamics.

### 17.3.4 Divisions VI and XIIb

Since 2001 a fishery developed in divisions VIb and XIIb in the Hatton Bank area, but in both divisions the recent catches are relatively small. Limited fleet information is available (ICES WGDEEP).

### 17.4 Catch composition

Length compositions of catches from the commercial trawl fishery in Div. Va are rather stable from year to year. In Figure 17.4.1 length distributions are shown since 1996 from the western area of Iceland, comprising the most important fishing grounds. Distributions are rather stable over the entire period and lack recruitment to the fishery in recent years is obvious. Little or no information is available of the catch composition in XIV and Vb.

### 17.5 Survey information

The total surveyed area in 2014 for Greenland halibut in Divisions Va and XIVb is provided in Figure 17.5.1. Most of the areas where commercial fishing takes place (Figure 15.2.2.) are covered by the annual surveys. The two surveys in Va and XIVb are compiled to one index and used as input in the assessment model.

### 17.5.1 Division Va

The fall survey for Greenland halibut was resumed in 2012 after no survey was conducted in 2011. Since 2008 the fishable biomass of Greenland halibut (fish of length equal to or greater than 50 cm ) increased significantly in Icelandic waters (Figures 17.5.2) and the abundance of fish less 70 cm was historic high in 2013. (Figures 17.5.3. -17.5.4.). Abundance and to a lesser degree biomass decreased markedly in 2014.

### 17.5.2 Division Vb

The catch rates from the available time series of the Faroese survey (1995-2014) declines until 2007 but since then increase to record high levels the last three years (Figure 17.5.5).

### 17.5.3 Division XIVb

Total biomass in the Greenlandic survey (Figure 17.5.6) in 2014 was estimated at 9858 tons (S.E. 1394) which is a $33 \%$ increase from the lowest value in the time series in 2013. A GLM analysis performed on the survey catch rates, taking into account the scattered coverage of area and depth between years did support this development (Figure 17.5.7.). The text table below provides information on the coverage and numbers of stations in 2014 along with the Va survey.

| Survey <br> /Division | No. hauls in 2014 <br> (planned hauls) | Depth range (m) | Coverage (km2) |
| :--- | :--- | :--- | :--- |
| Va | $203(219)$ | $32-1309 ?$ | -130000 |
| XIVb | $76(70)$ | $400-1500$ | 29000 |

The stock annex provides more extensive descriptions of the surveys.

### 17.6 Stock Assessment

### 17.6.1 Benchmark decisions for the stock assessment

The assessment model remains the stock production model using the new combined survey index and the Icelandic cpue index. Reference points as derived from this model are $30 \%$ Bmsy $_{\text {as }} \mathrm{B}_{\text {lim, }} 1.7 \times \mathrm{F}_{\text {msy }}$ as $\mathrm{F}_{\text {lim }}$ and an MSYB trigger defined as $50 \% \mathrm{Bmsy}_{\text {ms }} \mathrm{B}_{\text {msy }}$ and $\mathrm{F}_{\mathrm{msy}}$ are inherited references in the model approach.

The combined Icelandic and Greenland survey index was the input index assuming to cover most of the distributional area for the stock.

### 17.6.2 Summary of the various observation data

A number of indices from surveys and from the commercial fishery are available as indicators for the biomass development.

The surveys in V and XIV are considered to cover the adult stock distribution in the three divisions adequately. A detailed description of the survey/fishery design is provided in the stock annex.

The main fishing grounds are covered well by the available logbook data in Va and XIV, while in Vb the logbook information does not include the second principal fleet, gill netters, that covers other areas within Vb . The fleet behaviour in the entire area is likely influenced by a number of factors other than management restrictions, such as weather conditions and sea ice especially in the north-western areas. Over the years also technological development of the fishing gear has probably caused improved catchability. Therefore CPUE series is considered less qualified as biomass indicators than surveys.
Div. Va: Fishery and survey indices from Va show similar trends although of varying magnitude. The fall groundfish survey in Va (since 1996) indicate a full recovery from a low level in 2004-2006 for all sizes of fish and in all surveyed areas. Icelandic trawl CPUE show signs of a recovery from a low in 2004 although recovery is slow.
Div. Vb: Both standardised survey/exploratory fishery and commercial cpues show a dramatic increase since 2011.
Div. XIVb: The Greenland survey in XIV have remained low since 2008 and reached record low in 2013. The survey index though increased in 2014 due to higher catch rates in the entire survey area. In contradiction to this trend CPUE's from the various trawl fleets in XIVb have been high since 2012.

Subarea VI and XII: No biomass indices are available for these areas. However, the areas are considered negligible with respect to stock size distribution.

### 17.6.3 A model based assessment

The assessment uses a stochastic version of the logistic production model and Bayesian inference according to the Stock Annex in which a more detailed formulation of the model and its performance is found.

### 17.6.3.1 Input data

The model synthesized information from input priors and two independent series of Greenland halibut biomass indices and one series of catches by the fishery (Table 17.6.1). The two series of biomass indices were: a revised and standardised series of annual commercial-vessel catch rates for 1985-2014, CPUE $E_{\mathrm{t}}$; and a combined trawlsurvey biomass index for 1996-2014, Isurt,.

Total reported catch or WGs best estimates in ICES Subareas V, VI, XII and XIV 19612014 was used as yield data (Table 17.6.1, Figure. 17.2.1). Since the fishery has no major discarding problems or misreporting, the reported catches were entered into the model as error-free.

Three additional biomass series were available. However, the Greenland CPUE series showed trends conflicting with those of the other biomass indices - even if restricted
to data just opposite the midline next to the Icelandic fishery and were therefore not included. The Faroese indices of stock biomass (survey and logbooks from trawl fishery) were neither used in the assessment The omission of these indices from the analytical assessment only reflects that the model due to conflicting signals is not able to accommodate them but the indices are though considered to provide true populations trends and should therefore be treated as auxiliary information to the stock assessment.

### 17.6.3.2 Model performance

Inference was made from samples from the converged part of the MCMC samples as identified by appropriate statistics (Boje et al. 2015 WD 27). The model was able to produce a reasonable simulation of the observed data (Figure 17.6.2). The probabilities of getting more extreme observations than the realised ones given in the data series on stock size were in the range of 0.05 to 0.95 i.e. the observations did not lay in the extreme tails of their posterior distributions (Table 17.6.4). Exceptions are observed for the survey in 1997 ( $\mathrm{p}=0.96$ ). The CPUE series was generally better estimated than the survey series (Figure. 17.6.2).
The data could not be expected to carry much information on the parameter $P_{1960}$ - the stock size 25 years prior to when the series of stock biomass series start - and the posterior resembled the prior (Figure 17.6.1). The prior for K was somewhat updated to slightly higher values. However, the posterior still had a wide distribution. The retrospective runs suggest high consistency (Figure. 17.6.3).

The priors for MSY was significantly updated (Figure. 17.6.1). As mentioned above MSY was relatively insensitive to changes in prior distributions. The posterior $K$ had an inter-quartile range of 719-1059 ktons (Table 17.6.3).

### 17.6.3.3 Assessment results

The time series of estimated median biomass-ratios starts in 1960 as a virgin stock at K (Figure. 17.6.4-5). The fishery starts in 1961. Under continuously increasing fishing mortality the stock declined sharply in the mid 1990s to levels below the optimum, $B_{m s y}$. Some rebuilding towards $B_{m s y}$ was then seen in the late 1990s. Since then the stock started to increase from its lowest level in 2004-5 of approx. $45 \%$ of Bмš. In 2014 biomass was at $71 \%$ of $B_{\text {msy }}$. The risk of the biomass being below $B_{m s y}$ in 2014 is $100 \%$ and $0 \%$ of being below $\mathrm{B}_{\text {LIM }}$ (Table 17.6.5). The median fishing mortality ratio ( $\mathrm{F} / \mathrm{F}_{\mathrm{msy}}$ ) has exceeded $F_{m s y}$ since the 1990s and estimed at $1.3 F_{m s y}$ in 2014. (Figure. 17.6.4 and 17.6.5). This parameter can only be estimated with relatively large uncertainty and the posteriors therefore also include values below $F_{m s y}$. However, the probability that the $F$ has exceeded $F_{m s y}$ is high for most of the series.

The posterior for MSY was positively skewed with upper and lower quartiles at 26 ktons and 39 ktons (Table 17.6.3). As mentioned above MSY was relatively insensitive to changes in prior distributions.

Within a one-year perspective the sensitivity of the stock biomass to alternative catch options seems rather low. This is due to the inertia of the model used (see annex) and the low growth rate of the population. Risk associated with five optional catch levels for 2016 are given in Table 17.6.5.

The risk trajectory associated with ten-year projections of stock development assuming a maintained annual catch in the entire period ranging from 0 to 30 ktons were investigated (Figure. 17.6.6. -7 ). The calculated risk is a result of the projected development of the stock and the increase in uncertainty as projections are carried forward. It must
be noted that a catch scenario of a maintained constant catch over a decade without considering arrival of new biological information and advice is highly unrealistic.
Catches around 20 ktons are likely to lead to an increase in stock size and annual catches of 15 kt or less will result in a $50 \%$ probability of reaching BMSY within 10 years (Figure 17.6.6).
Scenarios of fixed levels of fishing mortality ratios within the range of 0.3 to 1.7 were conducted and are shown in Figure 17.6.8. Present biomass is above the MSY Btrigger ( $50 \%$ of $B_{M S Y}$ ) and a fishery at F MSY is then advised according the ICES MSY approach. Fishing at Fmsy will result in catches of 22 kt in 2016 (Figure 15.6 .8 panel D) and a stock size of $71 \%$ of BMSY in 2016 (Table 17.6.5).

### 17.6.3.4Conclusions

Stock status 2014-2015

- Stock size:
- Stock biomass $0.70 B_{m s y}$ (median)
- $100 \%$ probability of being below $B_{m s y}$
- $0 \%$ risk of being below Blim
- Stock production:
- MSY $=26-39$ ktons (inter-quartile range)
$-\quad$ Actual $\approx 0.9 \mathrm{MSY}$ (median)
- Exploitation:
- 21 ktons
- $\sim F_{\text {msy }}$ (median)
- $10 \%$ risk of exceeding Flim


## Predictions

- Risk of exceeding MSY $B_{\text {trigger }}$
- As the stock has improved since 2004-5 and is now further away from Bum the projected risk of exceeding this reference point is low ( between 0 and $1 \%$ ) at any catch at or below 30 kt .
- Catch option of 22 ktons/yr (FMSY level)
- Stock biomass is projected to be maintained or increase (0.71 of BMSY).
- Moratorium
- In the order of 5 years to rebuild to BMSY


### 17.6.4 Reference points

Reference points were defined at the benchmark in 2013 (WKBUT): BLim as $30 \% \mathrm{~B}_{\text {msy }}$ corresponding to production is reduced to $50 \%$ of its maximum. This is equivalent to the SSB-level (spawning stock biomass) at $50 \% \mathrm{R}_{\max }$ (maximum recruitment). Greenland halibut is believed to be a slow growing species i.e. with relative low $r$ (intrinsic rate of increase). This means that even without fishery it would take some 10 years to
rebuild the stock from $30 \% \mathrm{~B}_{\text {msy }}$ to $\mathrm{B}_{\text {msy }}$ (calculated by setting $r=0.21$, the $75^{\text {th }}$ percentile) - but likely longer.

MSY $B_{\text {trigger, }}$ the biomass level that triggers a deviation from Fmsy advice, was defined as $50 \%$ Bmsy. Flim was defined as 1.7 Fmsy.

### 17.7 Exploratory assessment: Gadget

An exploratory assessment on Greenland halibut using Gadget was presented at the meeting. This year only one run was presented based on growth data supplied by Norway. The resulting growth curves from Norway are identical to those estimated from mark-recapture data from Iceland that were presented at WKBUT-2013. The model setup was similar to the one presented at NWWG-2014.

### 17.7.1 Input data

The data used in the model were sex and length dis-aggregated indices from the Icelandic and Greenland surveys (combined in one index). Length distributions from the Icelandic and Greenland trawler fleet. Data on sex-ratio from the Icelandic trawler fleet and length at age data from Norway that was used to estimate growth inside the model. Catches from Iceland, Greenland and the Faroe Islands are included in the model.

### 17.7.2 Model settings

The model time was 1980 to 2015, with recruitment estimated annually. Two stocks are defined in the model, females and males that have different growth rates. Recruitment to the two stocks is equal that is 50:50.

In the model three fleets are defined, Icelandic, Greenland and Faroe Islands trawlers. The Icelanidc fleet has two selection patterns, one for females and another for males modeled as different L50 in the selection curves but with the same slope. The Greenland fleet has its own selection pattern but is not divided by sex. As there are no length distributions from the Faroe Islands it is assumed that the selection pattern in the same as in Iceland (for females).

In the model, natural mortality is set at 0.1 for all ages. The age range in the model is from age 3 to 25 with 25 being a plus-group.

### 17.7.3 Likelihood components

The likelihood components in the assessment are listed in table xxx. In all the model has 15 likelihood components but two of those are mainly for constraining the minimization routines (understocking and bounds).

Text Table: Components in Gadget model.

| Component | Type | Weight | Notes |
| :---: | :---: | :---: | :---: |
| SexRatioIceTrawl.lik | stockdistribution | 10 | Sex ratio from Icelandic trawlers |
| FemaleSmhLD.lik | catchdistribution | 12715 | Female length distribution from the combined survey |
| MaleSmhLD.lik | catchdistribution | 5847 | Male length distribution from the combined survey |
| IceTrawlLD.lik | catchdistribution | 1415 | Length distributions from the Icelandic trawl fleet |
| GreTrawlLD.lik | catchdistribution | 415 | Length distributions from the Greenland trawl fleet |
| FemSmh2045.si | surveyindices | 12 | Female abundance survey index for length 20 to 45 cm |
| FemSmh4665.si | surveyindices | 20 | Female abundance survey index for length 46 to 65 cm |
| FemSmh4665.si | surveyindices | 28 | Female abundance survey index for length larget than 65 cm |
| MaleSmh2045.si | surveyindices | 36 | Male abundance survey index for length 20 to 45 cm |
| MaleSmh4665.si | surveyindices | 28 | Male abundance survey index for length 46 to 65 cm |
| MaleSmh4665.si | surveyindices | 15 | Male abundance survey index for length larget than 65 cm |
| FemSmhML.lik | catchstatistics | 2 | Mean length at age for females based on Norwegian data, allocated to the survey. |
| MaleSmhML.lik | catchstatistics | 2 | Mean length at age for males based on Norwegian data, allocated to the survey. |
| understocking | understocking | 1 | To constrain minimization so that the stock will always be larger than the catches |
| bounds | bounds | 10 | To constrain minimization to respect the bounds of the parameters. |

### 17.7.4 Fit to data

In general the model captures the changes in the length distributions from both commercial catches and from the survey. However the model has problems with estimating the proportions of males and females from commercial catches. The fit to the survey indices is mostly reasonable, the model following the main trends in them. Due to time constraints the re-iterative weighing procedure was not used to assign weights to the likelihood components

### 17.7.5 Model estimates

The three estimated selection curves from commercial catches (Iceland females/males and Greenland) are virtually identical and might in future runs be merged to one selection curve. The estimated selection curve for the combined survey index has much lower L50 than the trawler fleets.

According to the model, recruitment is highly variable but that is most likely the result of very little information in the data on recruitment. Age-based data would help to
stabilize this. SSB is estimated to have increased from 1980 to the early nineties when it started to decrease gradually and has been more or less stable since 2008-2010. Fishing mortality was relatively stable at around $0.12-0.15$ from 1990.

When the model is run excluding the last years (analytical retrospective analysis) the runs ending in 2011 to 2013 show roughly the same pattern and even though there is some scaling issue the terminal estimates are quite consistent. However the run ending in 2015 is quite different from the other three where SSB is estimated considerably higher in the last years and the fishing mortality considerably lower.

### 17.7.6 Future work

The model will need more work to be usable as basis for assessment. Main issues to explore at present is the length aggregation of the survey indices, for example it may be more prudent to have only one index for males and one for females rather than the 6 currently used for tuning the model. The retrospective pattern also has to be addressed. This year the re-weighting of likelihood components procedure was not employed in assigning weights to the different data sets. Data from the Faroe Islands is missing and should be included. Work on aging is planned in Iceland based on the Norwegian method for age determination. When this gets started and a few years of ageing data become available it is expected that the model will be more stable as was the case for tusk and ling in Va (assessed in WGDEEP). Estimates of reference points were not presented based on the current model at the meeting due to computational difficulties.

### 17.8 Management Considerations

Available biological information and information on distribution of the fisheries suggest that Greenland halibut in XIV and V belong to the same entity and do mix. Historic information on tag-recapture experiments in Iceland have shown that Greenland halibut migrate around Iceland. Similar information from Greenland suggests some mix, both between West Greenland, East Greenland and Iceland. Therefore, management of the stock needs to be in accordance with the present three distinct management areas, XIV, Va and Vb. Recent information of tagging experiments in the Barents Sea suggests high mixing between the Barents Sea and Iceland. This connectivity is not accommodated for in the present assessment.
In 2012 the coastal states have initiated work on a common management plan for Greenland halibut. The plan is aimed to have two steps, a graduals lowering of the total catches until biological reference points have been defined by ICES, and thereafter implementation of a harvest control rule in accordance with ICES MSY approach. The plan will include continuous monitoring of the resource and requirements on information from the fishery. As a first step Greenland and Iceland decided on a TAC for 2013 at 26000 t and further agreed to reduce catches in 2014 by $15 \%$ corresponding to catches of 22100 t . The aim is to have a management plan to be implemented in 2015.

### 17.9 Data consideration and Assessment quality

The Icelandic CPUE series has for many years been used as a biomass indicator in the assessment of the stock. The CPUE of the Greenlandic trawlers and the biomass indices from the Faroese waters have not been used in the assessment, mainly because the stock production model is not able to accommodate contrasting indices (Icelandic CPUE and Greenlandic/Icelandic autumn surveys).

A number of issues on data and assessment quality were addressed in last year's report. The aim was that these issues should have been solved intersessionally and reviewed by an interbenchmark. However, no work has been done by the parties and it is realised that a meeting is necessary to be established to complete the work (see section 17.9).

### 17.10 Proposals and recommendations

Stock structure and connectivity between the main fishing areas remains partly unknown. Basic biological information on spawning and nursery grounds for the juveniles also remains poorly known. Biomass indices over the entire assessment area are not similar with respect to trend over time and may suggest different dynamics between areas. Further, recent tagging experiments in the Barents Sea suggest a high connectivity with Iceland waters. Therefore a compilation of present knowledge of stock identification for Greenland halibut in the East Greenland, Iceland, Faroese and Norwegian waters should be made in order to review whether present stock areas are appropriate for assessment purposes. Such a compilation should be evaluated outside NWWG, eg. by WGSIM.

A number of issues on the quality of the input biomass indices to the present assessment model were raised last year. The Icelandic CPUE series that is based on the principal trawler fleet is assumed to have undergone marked changes with respect to management regulations and spatial distribution. The possibility to estimate these effects by standardization of catch rates should be explored. Similar analyses should be conducted on the remaining CPUE series, in order to evaluate them as indicative of biomass development.
The present assessment model, a stock production model in Bayesian framework, is criticized for its behavior in relation to the biomass indices. The models use of process error and sensitivity to various priors should be further scrutinized. A generic review of the model's performance could potentially be by WGMG.
At the benchmark in 2013 (WKBUT) an alternative assessment model, Gadget, was presented. The group encouraged this model to be fully developed in order to replace the stock production model. Presently the Gadget model is not fully developed and several issues need further exploration (see section 17.7).

Ageing of Greenland halibut ceased for many of the marine institutes in Greenland, Iceland, Faroe Island and Norway around 2000 due to reading difficulties and lack of calibration. However, IMR in Norway have now developed a promising method to age Greenland halibut. With the aim to revert to an age based assessment, it is suggested that cooperation between institutes is initiated and an inter calibration protocol is established. This task is a major task since a number of sampled otoliths back in time have to be read, and the time horizon for this project is therefore expected to exceed the near future. NWWG recommends an ageing workshop to be conducted in collaboration with AFWG.

The above issues on input data and model approaches are suggested to be solved at a meeting in the autumn with a minimum attendance of experts from Greenland, Iceland and Faroe Islands. Such a meeting is expected to be coordinated by the relevant institutes. Eventual review of stock id and methods from WGSIM and WGMG should be evaluated externally by ICES along with any recommendations from the expert meeting in autumn prior to next year's NWWG.

Table 17.2.1 GREENLAND HALIBUT. Nominal landings (tonnes) by countries,
in Sub-areas V, VI, XII and XIV 1981-2013, as officially reported to ICES and estimated by WG

| Country | 1981 | 1982 | 1983 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Denmark | - | - | - | - | - | 6 | + | - |
| Faroe Islands | 767 | 1,532 | 1,146 | 1,052 | 853 | 1,096 | 1,378 | 2,319 |
| France | 8 | 27 | 236 | 845 | 52 | 19 | 25 | - |
| Germany | 3,007 | 2,581 | 1,142 | 863 | 858 | 565 | 637 | 493 |
| Greenland | + | 1 | 5 | 81 | 177 | 154 | 37 | 11 |
| Iceland | 15,457 | 28,300 | 28,360 | 29,231 | 31,044 | 44,780 | 49,040 | 58,330 |
| Norway | - | - | 2 | 3 | + | 2 | 1 | 3 |
| Russia | - | - | - | - | - | - | - | - |
| UK (Engl. and Wales) | - | - | - | - | - | - | - | - |
| UK (Scotland) | - | - | - | - | - | - | - |  |
| United Kingdom | - | - | - | - | - | - | - |  |
| Total | - | - | - | - | - | - | - |  |
| Working Group estimate | $-2,239$ | 32,441 | 30,891 | 32,075 | 32,984 | 46,622 | 51,118 | 61,156 |


| Country | 1990 | 1991 | 1992 | 1994 | 1995 | 1996 | 1997 | 1998 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | - | - | - | - | - | 1 | - |  |
| Faroe Is lands | 1,803 | 1,566 | 2,128 | 6,241 | 3,763 | 6,148 | 4,971 | 3,817 |
| France | - | - | 3 | - | - | 29 | 11 | 8 |
| Germany | 336 | 303 | 382 | 648 | 811 | 3,368 | 3,342 | 3,056 |
| Greenland | 40 | 66 | 437 | 867 | 533 | 1,162 | 1,129 | 747 |
| Iceland | 36,557 | 34,883 | 31,955 | 27,778 | 27,383 | 22,055 | 18,569 | 10,728 |
| Norway | 50 | 34 | 221 | 1,173 ${ }^{1}$ | 1,810 | 2,164 | 1,939 | 1,367 |
| Russia | - | - | 5 | - | 10 | 424 | 37 | 52 |
| Spain |  |  |  |  |  |  |  | 89 |
| UK (Engl. and Wales) | 27 | 38 | 109 | 513 | 1,436 | 386 | 218 | 190 |
| UK (Scotland) | - | - | 19 | 84 | 232 | 25 | 26 | 43 |
| United Kingdom |  |  |  |  |  |  |  |  |
| Total | 38,813 | 36,890 | 35,259 | 37,305 | 36,006 | 35,762 | 30,242 | 20,360 |
| Working Group estimate | 39,326 | 37,950 | 35,423 | 36,958 | 36,300 | 35,825 | 30,309 | 20,382 |


| Country | 1999 | 2000 | 2001 | $2003{ }^{1}$ | $2004{ }^{1}$ | $2005{ }^{1}$ | $2006{ }^{1}$ | $2007{ }^{1}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  | - | - | - | - | - | - | - |
| Estonia |  | - | - | - | - | 5 | 3 | - |
| Faroe Islands | 3,884 | - | 121 | 458 | 338 | 1,150 | 855 | 1,141 |
| France | - | 2 | 32 | 177 | 157 | - | 62 | 17 |
| Germany | 3,082 | 3,265 | 2,800 | 2,948 | 5,169 | 5,150 | 4,299 | 4,930 |
| Greenland | 200 | 1,740 | 1,553 | 1,459 | - | - | - | - |
| Iceland | 11,180 | 14,537 | 16,590 | 20,366 | 15,478 | 13,023 | 11,798 | - |
| Ireland |  | - | 56 | - | - | - | - | - |
| Lithuania |  | - | - | 2 | 1 | - | 2 | 3 |
| Norway | 1,187 | 1,750 | 2,243 | 1,074 | 1,233 | 1,124 | 1,097 | 692 |
| Poland |  | - | 2 | 93 | 207 | - | - | - |
| Portugal |  | - | 6 | - | - | - | 1,094 | - |
| Russia | 138 | 183 | 187 | - | 262 | - | 552 | 501 |
| Spain |  | 779 | 1,698 | 3,075 | 4,721 | 506 | 33 | - |
| UK (Engl. and Wales) | 261 | 370 | 227 | 40 | 49 | 10 | 1 | - |
| UK (Scotland) | 69 | 121 | 130 | 367 | 367 | 391 | 1 | - |
| United Kingdom | - | 166 | 252 | 841 | 1,304 | 220 | 93 | 17 |
| Total | 20,001 | 22,913 | 25,897 | 30,900 | 29,286 | 21,579 | 19,890 | 7,301 |
| Working Group estimate | 20,371 | 26,644 | 27,291 | 30,891 | 27,102 | 24,978 | 21,466 | 21,873 |


| Country | $2008^{1}$ | $2009^{1}$ | $2010^{1}$ | $2011^{1}$ | $2012^{1}$ | $2013^{1}$ | $2014^{1}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | - | - | - | - | - | - | - |
| Estonia | - | - | - | - | - | - | 429 |
| Faroe Islands | - | 270 | 1,408 | 1,705 | 2,811 | 2,788 | 3,393 |
| France | 114 | - | - | 9 | 67 | 133 | - |
| Germany | 4,846 | 427 | 5,287 | 5,782 | 4,620 | 3,814 | 3,701 |
| Greenland | - | - | -819 | - | 3,415 | 5,239 | 3,251 |
| Iceland | - | - | 13,293 | 13,192 | 13,749 | 14,859 | 9,861 |
| Ireland | - | - | - | - | - | - |  |
| Lithuania | 566 |  | - | - | 97 | - | - |
| Norway | 639 | 124 | 233 | 176 | 856 | 614 | 764 |
| Poland | 1,354 | 988 | 960 | - | 786 | - | - |
| Portugal | - | - | - | - | - | - | - |
| Russia | 799 | 762 | 1,070 | 1,095 | 1,168 | 1,369 | 587 |
| Spain | - | - | - | - | - | 67 |  |
| United Kingdom | 422 | 581 | 577 | 323 | 12 | 95 | 44 |
| Total | 9,744 | 5,974 | 22,901 | 25,618 | 29,405 | 26,923 | 20,743 |
| Working Group estimate | 24,481 | 28,197 | 25,995 | 26,347 |  |  | 21,069 |

1) Provisional data

Table 17.2.2 GREENLAND HALIBUT. Nominal landings (tonnes) by countries, in Division Va 1981-2011, as officially reported to ICES and estimated by WG.


[^5]2) Includes $223 t$ catch by Norway.

Table 17.2.3 GREENLAND HALIBUT. Nominal landings (tonnes) by countries, in Division Vb 1981-2009 as officially reported to ICES and estimated by WG.


1) Provisional data
2) WGestimate includes additional catches as described in Working Group reports for each year and in the report from 2001.

Table 17.2.4 GREENLAND HALIBUT. Nominal landings (tonnes) by countries, in Sub-area XIV 1981-2009, as officially reported to ICES and estimated by WG.

| Country | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | - | - | - | - | - | 78 | 74 | 98 | 87 |
| Germany | 2,893 | 2,439 | 1,054 | 818 | 636 | 745 | 456 | 595 | 420 |
| Greenland | + | 1 | 5 | 15 | 81 | 177 | 154 | 37 | 11 |
| Iceland | - | - | 1 | 2 | 36 | 17 | 136 | 40 | + |
| Norway | - | - | - | + | - | - | - | - | - |
| Russia | - | - | - | - | - | - | - | - | + |
| UK (Engl. and Wales) | - | - | - | - | - | - | - | - | - |
| UK (Scotland) | - | - | - | - | - | - | - | - | - |
| United Kingdom | - | - | - | - | - | - | - | - | - |
| Total | 2,893 | 2,440 | 1,060 | 835 | 753 | 1,017 | 820 | 770 | 518 |
| Working Group estimate | - | - | - | - | - | - | - | - | - |
| Country | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| Denmark | - | - | - | - | - | - | 1 | + | + |
| Faroe Islands | - | - | - | 181 | 168 | 147 | 130 | 148 | 151 |
| Germany | 293 | 279 | 311 | 391 | 639 | 808 | 3,343 | 3,301 | 3,399 |
| Greenland | 40 | 66 | 437 | 288 | 866 | 533 | 1,162 | 1,129 | $747{ }^{1,7}$ |
| Iceland | - | - | - | 19 | 82 | 7 | - | 1,803 | 148 |
| Norway | 8 | 18 | 196 | 511 | 1,120 | 1,668 | 1,881 | 1,897 ${ }^{1}$ | 1,253 ${ }^{1}$ |
| Russia | - | - | 5 | - | - | 10 | 424 | 37 | 52 |
| UK (Engl. and Wales) | 27 | 38 | 108 | 796 | 513 | 1405 | 264 | 218 | 190 |
| UK (Scotland) | - | - | 18 | 26 | 84 | 205 | 13 |  |  |
| United Kingdom | - | - | - | - | - | - | - |  |  |
| Total | 368 | 401 | 1,075 | 2,212 | 3,472 | 4,783 | 7,218 | 8,533 | 5,940 |
| Working Group estimate | $736{ }^{2}$ | $875{ }^{3}$ | 1,176 ${ }^{4}$ | 2,249 ${ }^{5}$ | $3,125{ }^{6}$ | 5,077 ${ }^{7}$ | 7,283 | 8,558 |  |
| Country | 1999 | 2000 | $2001{ }^{1}$ | $2002{ }^{1}$ | $2003{ }^{1}$ | $2004{ }^{1}$ | $2005{ }^{1}$ | $2006{ }^{1}$ | $2007{ }^{\text {1 }}$ |
| Denmark |  |  |  |  |  |  |  |  |  |
| Faroe Islands | 2 |  |  | 274 | 366 | 274 | 186 | 22 |  |
| Germany | 3,047 | 3,243 | 2,750 | 2,019 | 2,925 | 5,159 | 5,144 | 4,298 | 4,702 |
| Greenland | $200{ }^{1,4}$ | 1,740 | 1,553 | 1,887 | 1,459 |  |  |  |  |
| Iceland | 93 | 30 | 14,280 | 16,947 | 6 |  |  |  |  |
| Ireland |  |  | 7 |  |  |  |  |  |  |
| Norway | 1,100 | 1,161 | 1,424 | 1,660 | 846 | 1,114 | 1,023 | 1,094 |  |
| Poland |  |  |  |  |  | 205 |  |  |  |
| Portugal |  |  | 6 | 130 |  |  |  | 1,094 |  |
| Russia | 138 | 183 | 186 | 44 |  | 261 |  | 505 | 500 |
| Spain |  | 8 | 10 |  | 2,131 | 3,406 | 2 |  |  |
| UK (Engl. and Wales) | 226 | 262 | 100 |  |  |  |  |  |  |
| UK (Scotland) |  |  |  | 24 | 188 | 278 | 160 |  |  |
| United Kingdom |  |  |  | 178 | 799 | 1,294 |  |  |  |
| Total | 4,806 | 6,627 | 20,316 | 22,889 ${ }^{\circ}$ | 8,720 | 11,991 | 6,515 | 7,013 | 5,202 |
| Working Group estimate | 5376 | 6958 | 6,588 ${ }^{6}$ | 6,750 ${ }^{6}$ | 8,017 | 9,854 | 10,185 | 8,589 | 10,261 |


| Country | $2008{ }^{\text {I }}$ | $2009{ }^{\text {1 }}$ | $2010{ }^{\text {1 }}$ | $2011{ }^{1}$ | $2012{ }^{1}$ | $2013{ }^{1}$ | 2014 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Estonia |  |  |  |  |  |  | 429 |
| Faroe Islands |  | 270 | 333 |  | 77 | 125 | 409 |
| Germany | 4,842 | 4 | 4,490 | 5,206 | 4,351 | 3,428 | 3,114 |
| Greenland |  | 2,819 |  | 3,258 | 5,239 | 3,159 | 1,897 |
| Iceland |  |  |  |  | 7,290 |  | 3 |
| Ireland |  |  |  |  |  |  |  |
| Norway | 637 | 29 | 226 | 164 | 853 | 613 | 761 |
| Poland | 1,354 | 718 | 960 |  | 786 |  |  |
| Portugal |  |  |  |  |  |  |  |
| Russia | 763 |  | 1,070 | 1,095 | 1,168 | 1,369 | 587 |
| Spain |  |  |  |  |  |  |  |
| United Kingdom | 131 | 452 | 229 | 309 | 1 | 1 |  |
| Total | 7,727 | 4,292 | 7,308 | 10,032 | 19,765 | 8,694 | 7,200 |
| Working Group estimate | 9,102 | 9,805 | 10,402 | 10,761 |  |  | 7,526 |

1) Provisional data
2)WG estimate includes additional catches as described in working Group reports for each year and in the report from 2001.
2) Includes $125 t$ by Faroe Islands and $206 t$ by Greenland.
3) Excluding 4732 t reported as area unknown.
4) Includes 1523 t by Norway, 102 t by Faroe Islands, 3343 t by Germany, 1910 t by Greenland, 180 t by Russia, as reported to Greenland authorities.
5) Does not include most of the Icelandic catch as those are included in WG estimate of Va.
6) Excluding 138 t reported as area unknown.

Table 17.2.5 GREENLAND HALIBUT. Nominal landings (tonnes) by countries in Sub-area XII, as officially reported to the ICES and estimated by WG

| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | $2003^{1}$ | $2004^{1}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Faroe Islands |  | 47 |  |  |  |  | 40 |  |  |
| France |  |  |  |  | 1 |  |  | 4 | 30 |
| Ireland |  |  |  |  |  |  |  |  |  |
| Lithuania |  |  |  |  |  |  |  |  |  |


| Country | $2005{ }^{1}$ | $2006{ }^{1}$ | $2007{ }^{1}$ | $2008{ }^{1}$ | $2009{ }^{1}$ | $2010{ }^{1}$ | $2011{ }^{1}$ | $2012{ }^{1}$ | $2013{ }^{1}$ | $2014{ }^{1}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands |  |  |  |  |  |  | 106 |  |  |  |
| France |  |  |  |  |  |  |  |  |  |  |
| Ireland |  |  |  |  |  |  |  |  |  |  |
| Lithuania |  | 2 | 3 | 566 |  |  |  | 97 |  |  |
| Poland |  |  |  |  |  |  |  |  |  |  |
| Spain ${ }^{2}$ | 501 |  |  |  |  |  |  |  |  | 67 |
| UK | 3 |  |  |  |  |  |  |  |  |  |
| Russia |  | 46 | 1 |  | 762 |  |  |  |  |  |
| Norway |  |  |  |  | 94 |  |  |  |  |  |
| Estonia |  | 2 |  |  |  |  |  |  |  |  |
| Total | 504 | 50 | 4 | 566 | 856 | 0 | 106 | 97 | 0 | 67 |
| WGestimate | 504 | 50 | 4 | 566 | 856 | 0 | 106 | 97 | 0 | 67 |

[^6]Table 17.2.6 GREENLAND HALIBUT. Nominal landings (tonnes) by countries
in Sub-area VI, as officially reported to the ICES and estimated by WG.

| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | $2003{ }^{1}$ | $2004{ }^{1}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Estonia |  |  |  |  |  |  | 8 |  |  |  |
| Faroe Islands |  |  |  |  |  |  |  |  |  |  |
| France |  |  |  |  |  |  | 286 | 165 | 110 |  |
| Poland |  |  |  |  |  |  | 16 | 91 | 1 |  |
| Spain ${ }^{2}$ |  |  | 22 | 88 | 20 | 350 | 1367 | 214 | 170 |  |
| UK |  |  |  |  | 159 | 247 | 77 | 42 | 10 |  |
| Russia |  |  |  |  |  | 1 |  |  | 1 |  |
| Norway |  |  |  |  | 35 | 317 | 21 | 26 |  |  |
| Total | 0 | 0 | 22 | $88^{\prime}$ | 214 | 915 | 1775 | 538 | 292 |  |
| WGestimate |  |  |  |  |  |  |  |  |  |  |
| Country | $2005{ }^{1}$ | $2006{ }^{1}$ | $2007{ }^{1}$ | $2008{ }^{1}$ | $2009{ }^{1}$ | $2010{ }^{1}$ | $2011{ }^{1}$ | $2012{ }^{1}$ | $2013{ }^{1}$ | $2014{ }^{1}$ |
| Estonia | 5 | 1 |  |  |  |  |  |  |  |  |
| Faroe Islands |  |  |  |  |  | 1 |  |  | 0 | 1 |
| France |  | 22 | 8 | 114 |  | 38 | 8 | 54 | 113 |  |
| Poland |  |  |  |  |  |  |  |  |  |  |
| Spain ${ }^{2}$ | 3 | 33 |  |  |  |  |  |  |  |  |
| UK | 217 | 74 | 15 | 80 | 12 | 11 | 3 | 11 | 93 | 42 |
| Russia |  | 1 |  | 32 |  |  |  |  |  |  |
| Norway |  | 3 |  | 1 | 3 | 2 | 7 | 3 | 1 | 0 |
| Lithuania |  |  |  | 968 |  |  |  | 2 |  |  |
| Total | 225 | 134 | 23 | 1195 | 15 | 52 | 18 | 70 | 207 | 43 |
| WGestimate | 225 | 134 | 23 | 1195 | 15 | 52 | 18 | 70 | 207 | 43 |

${ }^{1}$ Provisional data
${ }^{2}$ Based on estimates by observers onboard vessels

Table 17.3.1. CPUE indices of trawl fleets in Div $\mathrm{Va}, \mathrm{Vb}$ and XIVb as derived from GLM multiplicative models.

| area | year | cpue | \% change <br> in CPUE <br> between years | landings | relative <br> derived effort | \% change in effort between years |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Iceland Va | 1985 | 1.00 |  | 29,197 | 100 |  |
|  | 1986 | 0.99 | -1 | 31,027 | 108 | 8 |
|  | 1987 | 0.96 | -3 | 44,659 | 148 | 38 |
|  | 1988 | 0.91 | -5 | 49,379 | 117 | -21 |
|  | 1989 | 1.05 | 16 | 59,272 | 103 | -12 |
|  | 1990 | 0.75 | -29 | 37,308 | 88 | -15 |
|  | 1991 | 0.73 | -2 | 35,413 | 97 | 10 |
|  | 1992 | 0.67 | -9 | 31,978 | 100 | 2 |
|  | 1993 | 0.54 | -20 | 34,134 | 133 | 33 |
|  | 1994 | 0.44 | -18 | 28,608 | 102 | -23 |
|  | 1995 | 0.36 | -19 | 27,391 | 118 | 15 |
|  | 1996 | 0.30 | -15 | 22,073 | 95 | -20 |
|  | 1997 | 0.32 | 5 | 16,792 | 72 | -24 |
|  | 1998 | 0.50 | 57 | 10,595 | 40 | -44 |
|  | 1999 | 0.54 | 9 | 11,138 | 97 | 140 |
|  | 2000 | 0.58 | 7 | 14,607 | 122 | 27 |
|  | 2001 | 0.59 | 2 | 16,752 | 113 | -8 |
|  | 2002 | 0.48 | -19 | 19,714 | 146 | 29 |
|  | 2003 | 0.35 | -26 | 20,415 | 139 | -4 |
|  | 2004 | 0.30 | -16 | 15,477 | 90 | -36 |
|  | 2005 | 0.27 | -9 | 13,172 | 93 | 4 |
|  | 2006 | 0.37 | 33 | 11,817 | 67 | -28 |
|  | 2007 | 0.46 | 25 | 10,525 | 71 | 6 |
|  | 2008 | 0.39 | -14 | 9,580 | 106 | 49 |
|  | 2009 | 0.41 | 5 | 15,782 | 157 | 48 |
|  | 2010 | 0.41 | -1 | 13,565 | 87 | -45 |
|  | 2011 | 0.43 | 4 | 14,048 | 99 | 15 |
|  | 2012 | 0.44 | 3 | 7,312 | 51 | -49 |
|  | 2013 | 0.45 | 2 | 15,439 | 207 | 310 |
|  | 2014 | 0.41 | -8 | 10,475 | 74 | -64 |
| Greenland, XIVb | 1991 | 1.00 |  | 875 | 100 | 0 |
|  | 1992 | 0.91 | -9 | 1,176 | 148 | 48 |
|  | 1993 | 2.48 | 172 | 2,249 | 70 | -52 |
|  | 1994 | 3.19 | 29 | 3,125 | 108 | 54 |
|  | 1995 | 3.31 | 4 | 5,077 | 157 | 45 |
|  | 1996 | 3.34 | 1 | 7,283 | 142 | -9 |
|  | 1997 | 3.45 | 3 | 8,558 | 114 | -20 |
|  | 1998 | 3.38 | -2 | 5,940 | 71 | -38 |
|  | 1999 | 2.54 | -25 | 5,376 | 120 | 70 |
|  | 2000 | 2.15 | -15 | 6,958 | 153 | 27 |


|  | 2001 | 2.22 | 3 | 7,216 | 101 | -34 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2002 | 2.40 | 8 | 6,621 | 85 | -16 |
|  | 2003 | 2.36 | -2 | 8,017 | 123 | 45 |
|  | 2004 | 2.30 | -2 | 9,854 | 126 | 2 |
|  | 2005 | 3.18 | 38 | 10,185 | 75 | -41 |
|  | 2006 | 3.27 | 3 | 8590 | 82 | 10 |
|  | 2007 | 3.11 | -5 | 10261 | 126 | 54 |
|  | 2008 | 3.15 | 1 | 8,952 | 86 | -32 |
|  | 2009 | 2.60 | -17 | 10,567 | 143 | 66 |
|  | 2010 | 2.73 | 5 | 10,402 | 94 | -35 |
|  | 2011 | 2.68 | -2 | 10,761 | 105 | 13 |
|  | 2012 | 3.17 | 18 | 12,475 | 98 | -7 |
|  | 2013 | 2.95 | -7 | 12,476 | 107 | 9 |
|  | 2014 | 3.14 | 6 | 7,526 | 57 | -47 |
| Faroe Islands, Vb | 1991 | 1.00 |  | 1,662 | 100 | 34 |
|  | 1992 | 0.34 | -21 | 2,269 | 397 | 297 |
|  | 1993 | 0.24 | -11 | 4,434 | 282 | -29 |
|  | 1994 | 0.23 | -2 | 5,225 | 121 | -57 |
|  | 1995 | 0.16 | -28 | 3,832 | 103 | -15 |
|  | 1996 | 0.17 | 4 | 6,469 | 160 | 55 |
|  | 1997 | 0.19 | 12 | 4,870 | 67 | -58 |
|  | 1998 | 0.14 | -34 | 3,825 | 112 | 67 |
|  | 1999 | 0.16 | 12 | 4,265 | 96 | -15 |
|  | 2000 | 0.17 | 11 | 5,079 | 109 | 14 |
|  | 2001 | 0.20 | 19 | 3,245 | 55 | -50 |
|  | 2002 | 0.16 | -24 | 2,694 | 104 | 91 |
|  | 2003 | 0.10 | -29 | 2,426 | 141 | 35 |
|  | 2004 | 0.08 | -12 | 1,771 | 89 | -37 |
|  | 2005 | 0.09 | 4 | 892 | 48 | -46 |
|  | 2006 | 0.10 | 19 | 873 | 83 | 72 |
|  | 2007 | 0.12 | 16 | 1,060 | 107 | 28 |
|  | 2008 | 0.18 | 60 | 1735 | 100 | -6 |
|  | 2009 | 0.21 | 26 | 1760 | 107 | 7 |
|  | 2010 | 0.17 | -21 | 1,413 | 87 | -19 |
|  | 2011 | 0.31 | 65 | 1,489 | 98 | 13 |
|  | 2012 | 0.30 | -4 | 2163 | 59 | -40 |
|  | 2013 |  |  |  | 148 | 153 |
|  | 2014 |  |  | 2,958 |  |  |

Table 17.6.1. Model input data series: Catch by the fishery; three indices of stock biomass - a standardized catch rate index based on fishery data (CPUE) from the Iceland EEZ, a Icelandic (Ice) and a Greenlandic (Green) research survey index.

| Year | Catch (ktons) | CPUE (index) | $\qquad$ |
| :---: | :---: | :---: | :---: |
| 1960 | 0 | - | - |
| 1961 | 0.029 | - | - |
| 1962 | 3.071 | - | - |
| 1963 | 4.275 | - | - |
| 1964 | 4.748 | - | - |
| 1965 | 7.421 | - | - |
| 1966 | 8.030 | - | - |
| 1967 | 9.597 | - | - |
| 1968 | 8.337 | - | - |
| 1969 | 26.200 | - | - |
| 1970 | 33.823 | - | - |
| 1971 | 28.973 | - | - |
| 1972 | 26.473 | - | - |
| 1973 | 20.463 | - | - |
| 1974 | 36.280 | - | - |
| 1975 | 23.494 | - | - |
| 1976 | 6.045 | - | - |
| 1977 | 16.578 | - | - |
| 1978 | 14.349 | - | - |
| 1979 | 23.622 | - | - |
| 1980 | 31.157 | - | - |
| 1981 | 19.239 | - | - |
| 1982 | 32.441 | - | - |
| 1983 | 30.891 | - | - |
| 1984 | 34.024 | - | - |
| 1985 | 32.075 | 1.76 | - |
| 1986 | 32.984 | 1.74 | - |
| 1987 | 46.622 | 1.69 | - |
| 1988 | 51.118 | 1.60 | - |
| 1989 | 61.396 | 1.86 | - |
| 1990 | 39.326 | 1.32 | - |
| 1991 | 37.950 | 1.29 | - |
| 1992 | 35.487 | 1.17 | - |
| 1993 | 41.247 | 0.94 | - |
| 1994 | 37.190 | 0.77 | - |
| 1995 | 36.288 | 0.63 | - |
| 1996 | 35.932 | 0.53 | 65.72 |
| 1997 | 30.309 | 0.56 | 90.24 |
| 1998 | 20.382 | 0.88 | 90.96 |
| 1999 | 20.371 | 0.96 | 89.75 |
| 2000 | 26.644 | 1.03 | 100.56 |
| 2001 | 27.291 | 1.04 | 109.96 |
| 2002 | 29.158 | 0.84 | 83.53 |
| 2003 | 30.891 | 0.62 | 52.10 |
| 2004 | 27.102 | 0.53 | 36.64 |
| 2005 | 24.249 | 0.48 | 55.92 |
| 2006 | 21.432 | 0.64 | 38.78 |
| 2007 | 20.957 | 0.81 | 49.76 |
| 2008 | 22.169 | 0.69 | 57.57 |
| 2009 | 27.349 | 0.73 | 79.77 |
| 2010 | 25.995 | 0.72 | 59.03 |
| 2011 | 26.424 | 0.75 | 70.97 |
| 2012 | 29.309 | 0.77 | 81.89 |
| 2013 | 27.045 | 0.79 | 84.29 |
| 2014 | 21.069 | 0.72 | 75.99 |
| 2015* | 21.000 |  |  |

Table 17.6.2. Priors used in the assessment model. ~ means "distributed as..", dunif = uniform-, dlnorm = lognormal-, dnorm= normal- and dgamma = gammadistributed. Symbols as in text.

| Parameter |  | Prior |  |
| :---: | :---: | :---: | :---: |
| Name | Symbol | Type | Distribution |
| Maximal Suatainable Yield | MSY | reference | dunif( 1,300 ) |
| Carrying capacity | K | low informative | dnorm( 750,300 ) |
| Catchability Iceland survey | $q_{l c e}$ | reference | $\ln \left(\mathrm{q}_{\text {cee }}\right) \sim \operatorname{dunif}(-3,1)$ |
| Catchability Greenland survey | $q_{\text {Green }}$ | reference | $\ln \left(\mathrm{q}_{\text {Green }}\right) \sim$ dunif( $-3,1$ ) |
| Catchability Iceland CPUE | $q_{\text {cpue }}$ | reference | $\ln \left(\mathrm{q}_{\text {cpue }}\right) \sim$ dunif( $-10,1$ ) |
| Initial biomass ratio | $P_{1}$ | informative | dnorm( $2,0.071$ ) |
| Precision Iceland survey | $1 / \sigma_{\text {lce }}{ }^{2}$ | low informative | dgamma(2.5,0.03) |
| Precision Greenland survey | $1 / \sigma_{\text {Green }}{ }^{2}$ | low informative | dgamma(2.5,0.03) |
| Precision Iceland CPUE | $1 / \sigma_{\text {cpue }}{ }^{2}$ | low informative | dgamma(2.5,0.03) |
| Precision model | $1 / \sigma_{P}{ }^{2}$ | reference | dgamma(0.01,0.01) |

Table 17.6.3. Summary of parameter estimates: mean, standard deviation (sd) and 25, 50, and 75 percentiles of the posterior distribution of selected parameters (symbols as in the text).

|  | Mean | sd | $25 \%$ | Median | $75 \%$ |
| :--- | :---: | :---: | :---: | :---: | :---: |
| $M S Y$ (ktons) | 33.00 | 10.99 | 26.22 | 32.39 | 38.86 |
| $K$ (ktons) | 896 | 250 | 719 | 885 | 1059 |
| $r$ | 0.16 | 0.07 | 0.11 | 0.15 | 0.20 |
| $q_{\text {cpue }}$ | 0.003 | 0.001 | 0.002 | 0.003 | 0.003 |
| $q_{\text {Survey }}$ | 0.26 | 0.10 | 0.19 | 0.24 | 0.31 |
| $P_{1985}$ | 1.56 | 0.12 | 1.48 | 1.56 | 1.65 |
| $P_{\text {2014 }}$ | 0.72 | 0.10 | 0.65 | 0.71 | 0.78 |
| $\sigma_{\text {cpue }}$ | 0.09 | 0.02 | 0.08 | 0.09 | 0.11 |
| $\sigma_{\text {Survey }}$ | 0.19 | 0.04 | 0.16 | 0.18 | 0.21 |
| $\sigma_{P}$ | 0.16 | 0.03 | 0.14 | 0.16 | 0.17 |

Table 17.6.4. Model diagnostics: residuals (\% of observed value), probability of getting a more extreme observation (p.extreame; see text for explanation).

|  | CPUE |  |  | Survey |  |
| :---: | ---: | ---: | ---: | ---: | ---: |
| Year | resid (\%) | Pr | resid (\%) | Pr |  |
| 1985 | -2.02 | 0.56 | - | - |  |
| 1986 | -0.93 | 0.53 | - | - |  |
| 1987 | -0.60 | 0.52 | - | - |  |
| 1988 | -1.37 | 0.54 | - | - |  |
| 1989 | 1.93 | 0.44 | - | - |  |
| 1990 | -0.52 | 0.52 | - | - |  |
| 1991 | -2.50 | 0.58 | - | - |  |
| 1992 | -3.25 | 0.60 | - | - |  |
| 1993 | 0.20 | 0.49 | - | - |  |
| 1994 | 0.78 | 0.48 | - | - |  |
| 1995 | 3.85 | 0.38 | - | - |  |
| 1996 | 12.09 | 0.16 | -14.44 | 0.76 |  |
| 1997 | 15.76 | 0.10 | -36.21 | 0.96 |  |
| 1998 | -3.56 | 0.62 | -11.43 | 0.72 |  |
| 1999 | -1.51 | 0.55 | 0.31 | 0.49 |  |
| 2000 | -1.39 | 0.55 | -3.90 | 0.58 |  |
| 2001 | -3.17 | 0.61 | -13.32 | 0.75 |  |
| 2002 | -1.48 | 0.55 | -6.12 | 0.62 |  |
| 2003 | 0.52 | 0.48 | 13.49 | 0.25 |  |
| 2004 | -1.59 | 0.55 | 29.83 | 0.07 |  |
| 2005 | 7.84 | 0.25 | -12.10 | 0.73 |  |
| 2006 | -8.69 | 0.76 | 36.31 | 0.04 |  |
| 2007 | -14.98 | 0.89 | 28.77 | 0.08 |  |
| 2008 | -0.28 | 0.50 | 12.53 | 0.27 |  |
| 2009 | 0.48 | 0.48 | -13.21 | 0.75 |  |
| 2010 | -1.12 | 0.54 | 14.22 | 0.24 |  |
| 2011 | -0.20 | 0.51 | 0.69 | 0.49 |  |
| 2012 | 1.67 | 0.45 | -9.19 | 0.68 |  |
| 2013 | 0.22 | 0.50 | -10.47 | 0.70 |  |
| 2014 | 4.00 | 0.37 | -5.97 | 0.62 |  |
|  |  |  |  |  |  |

Table 17.6.5. Upper: stock status for 2014 and predicted to the end of 2015. Lower: predictions for 2016 with catch options from 0 to 30 ktons and the catch option corresponding to Fmsy.

| Status | $2014^{* *}$ | $2015^{*}$ |
| :--- | ---: | ---: |
| Risk of falling below Blim |  |  |
| (0.3BMSY) | $0 \%$ | $0 \%$ |
| Risk of falling below BMSY | $100 \%$ | $95 \%$ |
| Risk of exceeding FMSY | $47 \%$ | $44 \%$ |
| Risk of exceeding Flim (1.7FMSY) | $10 \%$ | $11 \%$ |
| Stock size (B/Bmsy), median | 0.71 | 0.70 |
| Fishing mortality (F/Fmsy), | 1.30 | 0.94 |
| Productivity (\% of MSY) | $91 \%$ | $91 \%$ |

*Predicted catch $=21 \mathrm{ktons}$
**Catches=21069 t

| Catch option 2016(ktons) | 0 | 5 | 10 | 15 | 20 | 22 | 30 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Risk of falling below Blim $(0.3 B M S Y)$ | 0\% | 0\% | 0\% | 0\% | 0\% | 0\% | 0\% |
| Risk of falling below BMSY | 85\% | 86\% | 87\% | 88\% | 89\% | 90\% | 93\% |
| Risk of exceeding FMSY |  | 1\% | 6\% | 18\% | 38\% | 50\% | 79\% |
| Risk of exceeding Flim (1.7FMSY) | - | 0\% | 1\% | 4\% | 10\% | 15\% | 38\% |
| Stock size (B/Bmsy), median | 0.76 | 0.75 | 0.74 | 0.74 | 0.72 | 0.71 | 0.65 |
| Fishing mortality (F/Fmsy), | 0.00 | 0.21 | 0.42 | 0.64 | 0.87 | 1.00 | 1.46 |
| Productivity (\% of MSY) | 94\% | 94\% | 93\% | 93\% | 92\% | 91\% | 88\% |



Figure 17.2.1. Landings of Greenland halibut in Divisions V, XI and XIV. As the landings within Icelandic waters, since 1976, have not officially been separated and reported according to the defined ICES statistical areas, they are set under area Va by the North Western Working Group. In 2012 Icelandic landings in XIV were recorded in XIV, while for remaining years all landings are recorded in Va.


Figure 17.2.2 Greenland halibut V+XIV. Distribution of fishing effort in 2014. 500 m and 1000 m depth contours are shown.


Figure 17.2.3. Greenland halibut V+XIV. Distribution of catches in the fishery in 2014. 500 m and 1000 m depth contours are shown


Figure 17.2.4. Greenland halibut V+XIV. Distribution of total fishing effort 1991-2012. The 500m and 1000 m depth contours are shown.


Figure 17.2.5. Greenland halibut V+XIV. Distribution of total catches in the fishery 1991-2012 500m and 1000 m depth contours are shown.


Figure 17.3.1. Standardised CPUEs from the Icelandic trawler fleet in Va. Area 1-4 are west, north, east and south-east. The average index of the four areas are used as biomass indicator input to the stock production model.


Figure 17.3.2 Standardised CPUE from the Icelandic trawler fleet in Va by four main fishing areas in Va. 95\% CI indicated.


Figure 17.3.3. Standardised CPUE, observed and derived effort from Icelandic trawl fishery.


Figure 17. 3.4. Standardised CPUE from the Faroese trawler fleet. 95\% CI indicated


Figure 17.3.6. Standardised CPUE from trawler fleets in XIVb. 95\% CI indicated. Points are raw observations.


Figure 17.3.7. Standardised CPUE from trawler fleets in XIVb shown by subdivisions in XIVb in a north-south direction. 95\% CI indicated.


Figure 17.4.1. Length distributions from the commercial trawl fishery in the western fishing grounds of Iceland (Va) in the years 1996-2012. Blue indicate males and red indicates females.


Figure 17.5.1. Stations covered by scientific surveys in XIV+V indicated as station positions in 2013 by the Greenland ( $n=76$ ) and Iceland ( $n=203$ ).


Figure 17.5.2. Distribution of Greenland halibut catches from the Icelandic fall survey since 1996. No survey was conducted in 2011.
Abundance

Biomass


$$
\triangle>50 \mathrm{~cm} \quad \longrightarrow>60 \mathrm{~cm} \quad \longrightarrow>70 \mathrm{~cm} \quad \longrightarrow \rightarrow 80 \mathrm{~cm}
$$

Figure 17.5.3. Greenland halibut in Icelandic fall groundfish survey. No survey was conducted in 2011.


Figure 17.5.4. Abundance indices by length for the Icelandic fall survey 1996-2014. No survey was conducted in 2011.


Figure 17.5.5. Catch rates from a combined survey/fisherman's survey in Vb. Estimates are from a GLM model.


Figure 17.5.6. Distribution of catches of Greenland halibut at East Greenland in 1998-2014 in the Greenland deep-water survey.


Figure 17.5.6 continued. Distribution of catches of Greenland halibut at East Greenland in 1998 2014 in the Greenland deep-water survey.


Figure 17.5.7. Standardised catch rates from the Greenland survey.(95\% CI indicated.)


Figure 17.6.1. Probability density distributions of model parameters: estimated posterior (solid line) and prior (broken line) distributions.


Figure 17.6.2. Observed (red curve) and predicted (dashed lines) series of the two biomass indices input to the model. Dashed lines are inter-quartile range of the posteriors.


Figure 17.6.3. Retrospective plot of median relative biomass ( $B / B_{m s y}$ ). Relative biomass series are estimated by consecutively leaving out from 0 to 9 years of data.


Figure 17.6.4. Stock trajectory. Estimated annual median biomass-ratio (B/BMSY) and fishing mor-tality-ratio (F/FMSY) 1985-2014. Blim indicated by red line.


Figure 17.6.5. Stock summary, upper panel right: fishing mortality (F/Fmsy), left: total biomass (B/Bmsy) and lower panel is landings since start of the fishery.


Figure 17.6.6 Estimated time series of relative biomass ( $B_{t} / B_{m s y}$ ) under different catch option scenarios: $0,5,10,15$ and 20 kt from upper to lower panel. Bold red lines are inter-quartile ranges and the solid black line is the median; the error bars extend to cover the central 90 per cent of the distribution.


Figure 17.6.7. Projections: Medians of estimated posterior biomass- and fishing mortality ratios; estimated risk of exceeding $F_{m s y}$ or going below and $B_{M S Y \text { rtriger }}$ given catches at $\mathbf{0 , 5 , 1 0 , 1 5 , 2 0}$ and 30 ktons.


Figure 17.6.8. Historic landings and projected landings 2015-2025 under various F ratio options from 0.3-1.7 F/Fmsy Solid red line is median, quartiles and $90 \%$ conf limit indicated.


Figure 17.6.9. The logistic production curve in relation to stock biomass (B/Bmsy) (upper) and fishing mortality ( $\mathbf{F} / \mathrm{Fmsy}$ ) (lower). Upper: points of maximum sustainable yield (MSY) and corresponding stock size are shown as well as the slope (red line) of the production curve (blue line); lower: points of MSY and corresponding fishing mortality and Fcrash ( $\mathrm{F} \geq$ Fcrash do not have stable equilibriums and will drive the stock to zero).

## 18 Redfish in Subareas V, VI, XII and XIV

This chapter deals with fisheries directed to Sebastes species in Subareas V, VI, XII and XIV (chapters 18.4 and 18.7), and the abundance and distribution of juveniles (chapter 17.2.1), among other issues.

The "Workshop on Redfish Stock Structure" (WKREDS, 22-23 January 2009, Copenhagen, Denmark; ICES 2009) reviewed the stock structure of Sebastes mentella in the Irminger Sea and adjacent waters. ACOM concluded, based on the outcome of the WKREDS meeting, that there are three biological stocks of S. mentella in the Irminger Sea and adjacent waters:
a 'Deep Pelagic' stock (NAFO 1-2, ICES V, XII, XIV >500 m) - primarily pelagic habitats, and including demersal habitats west of the Faeroe Islands;
a 'Shallow Pelagic' stock (NAFO 1-2, ICES V, XII, XIV <500 m) - extends to ICES I and II, but primarily pelagic habitats, and includes demersal habitats east of the Faeroe Islands;
an 'Icelandic Slope' stock (ICES Va, XIV) - primarily demersal habitats.
This conclusion is primarily based on genetic information, i.e. microsatellite information, and supported by analysis of allozymes, fatty acids and other biological information on stock structure, such as some parasite patterns. The Russian Federation maintains the point of view that there is only one stock of $S$. mentella in the pelagic waters of the Irminger Sea. Accordingly, the Russian Federation presented alternative approaches to stock assessment as well as environmental influence on stock dynamics. Briefly, it is claimed that the current survey based assessment does not adequately reflect stock status and that environmental factors - temperature causes major distributional changes of redfish - affect stock status more than fisheries and the use of the current management areas is rejected (see WD28, WD33 and Annex XX). The other NWWG members did not agree with the Russian Federation's view on stock structure and did not consider the presented assessment approach sufficiently documented.
The adult redfish on the Greenland shelf has traditionally been attributed to several stocks, and there remains the need to investigate the affinity of adult S. mentella in this region. The East-Greenland shelf is most likely a common nursery area for the three biological stocks.

ICES past advice for $S$. mentella fisheries was provided for two distinct management units, i.e. a demersal unit on the continental shelves and slopes and pelagic unit in the Irminger Sea and adjacent waters. However, based on the new stock identification information, ICES recommends three potential management units that are geographic proxies for biological stocks that were partly defined by depth and whose boundaries are based on the spatial distribution pattern of the fishery to minimize mixed stock catches (see Figure 18.1.1):

Management Unit in the northeast Irminger Sea: ICES Areas Va, XII, and XIV.
Management Unit in the southwest Irminger Sea: NAFO Areas 1 and 2, ICES areas Vb, XII and XIV.

Management Unit on the Icelandic slope: ICES Areas Va and XIV, and to the north and east of the boundary proposed in the MU in the northeast Irminger Sea.

The pelagic fishery in the Irminger Sea and adjacent waters shows a clear distinction between two widely separated grounds fished at different seasons and depths. Spatial analysis of the pelagic fishery catch and effort by depth, inside and outside the boundaries proposed for the management units in the northeast Irminger Sea, indicate that the boundaries effectively delineate the pelagic fishery in the northeast Irminger Sea from the pelagic fishery in the southwest Irminger Sea, with a small portion of mixedstock catches. In the last decade the majority (more than $95 \%$ ) of the catches have been taken in the northeast Irminger Sea. The northeastern fisheries on the pelagic S. mentella occur at the start of the fishing season at depths below 500 m and overlap to some extent with demersal fisheries on the continental slopes of Iceland (Sigurdsson et al., 2006).

A schematic illustration of the relationship between the management units and biological stocks is given in Figure 18.1.2.

For the abovementioned reasons, the Group now provides advice for the following Sebastes units:
the S. marinus on the continental shelves of ICES Divisions Va, Vb and Subarea VI and XIV (chapter 19),
the demersal S. mentella on the Icelandic slope (chapter 20),
the shallow and deep pelagic S. mentella units in the Irminger Sea and adjacent waters (chapters 21 and 22, respectively),
the Greenland shelf S. mentella (chapter 23).

### 18.1 Environmental and ecosystem information

Species of the genus Sebastes are common and widely distributed in the North Atlantic. They are found off the coast of Great Britain, along Norway and Spitzbergen, in the Barents Sea, off the Faroe Islands, Iceland, East and West Greenland, and along the east coast of North America from Baffin Island to Cape Cod. All Sebastes species are viviparous. Copulation occurs in autumn-early winter and larvae extrusion takes place in late winter-late spring/early summer. Little is known about the copulation areas.
The increase of water temperature in the Irminger Sea may have an effect on spatial and vertical distribution of S. mentella in the feeding area (Pedchenko, 2005). The abundance and distribution of pelagic S. mentella in relation to oceanographic conditions were analyzed in a special multistage workshop (ICES 2012). Based on 20 years of survey data, the results reveal the average relation of pelagic redfish to their physical habitat in shallow and intermediate waters: The most preferred latitude, longitude, depth, salinity and temperature for $S$. mentella are approximately $58^{\circ} \mathrm{N}, 40^{\circ} \mathrm{W}, 300 \mathrm{~m}, 34.89$ and $4.4^{\circ} \mathrm{C}$, respectively. The spatial distribution of S. mentella in the Irminger Sea mainly in waters $<500 \mathrm{~m}$ (and thus mainly relating to the "shallow" stock) appears strongly influenced by the Irminger Current Water (ICW) temperature changes, linked to the Subpolar Gyre (SPG) circulation and the North Atlantic Oscillation (NAO). The fish avoid waters mainly associated with the ICW $\left(>4.5^{\circ} \mathrm{C}\right.$ and $\left.>34.94\right)$ in the northeastern Irminger Sea, which may cause displacement of the fish towards the southwest, where fresher and colder water occurs.

Results based on international redfish survey data suggest that the interannual distribution of fish above 500 m will shift in a southwest/northeast direction depending on integrated oceanographic conditions (ICES 2012).

### 18.2 Environmental drivers of productivity

### 18.2.1 Abundance and distribution of 0 -group and juvenile redfish

Available data on the distribution of juvenile S. marinus indicate that the nursery grounds are located in Icelandic and Greenland waters. No nursery grounds have been found in Faroese waters. Studies indicate that considerable amounts of juvenile S. marinus off East Greenland are mixed with juvenile S. mentella (Magnússon et al. 1988; 1990, ICES CM 1998/G:3). The 1983 Redfish Study Group report (ICES CM 1983/G:3) and Magnússon and Jóhannesson (1997) describe the distribution of 0-group S. marinus off East Greenland. The nursery areas for S. marinus in Icelandic waters are found all around Iceland, but are mainly located west and north of the island at depths between 50 and 350 m (ICES CM 1983/G:3; Einarsson, 1960; Magnússon and Magnússon 1975; Pálsson et al. 1997). As they grow, the juveniles migrate along the north coast towards the most important fishing areas off the west coast.

Indices for 0-group redfish in the Irminger Sea and at East Greenland areas were available from the Icelandic 0-group surveys from 1970-1995. Thereafter, the survey was discontinued. Above average year class strengths were observed in 1972, 1973-74, 1985-91, and in 1995.

There are very few juvenile demersal S. mentella in Icelandic waters (see chapter 19), and the main nursery area for this species is located off East Greenland (Magnússon et al. 1988, Saborido-Rey et al. 2004). Abundance and biomass indices of redfish smaller than 17 cm from the German annual groundfish survey, conducted on the continental shelf and slope of West and East Greenland down to 400 m , show that juveniles were abundant in 1993 and 1995-1998 (Figure 18.2.1). Since 2008, the survey index has been very low and was in 2013the lowest value recorded since 1982. Juvenile redfish were only classified to the genus Sebastes spp., as identification of small specimens to species level is difficult due to very similar morphological features. The 1999-2013 survey results indicate low abundance and are similar to those observed in the late 1980s. Observations on length distributions of $S$. mentella fished deeper than 400 m indicate that a part of the juvenile $S$. mentella on the East Greenland shelf migrates into deeper shelf areas and into the pelagic zone in the Irminger Sea and adjacent waters (Stransky 2000), with unknown shares.

### 18.3 Ecosystem considerations

Information on the ecosystems around the Faroe Islands, Iceland and Greenland is given in chapters 2, 7 and 13.

Analysis of the oceanographic situation in the Irminger Sea during the 2013 international survey and long-term data including 2003, allows the following conclusions:

Strong positive anomalies of temperature observed in the upper layer of the Irminger Sea with a maximum in 1998 are related to an overall warming of water in the Irminger Sea and adjacent areas in 1994-2013. These changes were also observed in the Irminger Current above the Reykjanes Ridge (Pedchenko, 2000), off Iceland (Malmberget al., 2001) and in the Labrador Sea water (Mortensen and Valdimarsson, 1999). Thus, temperature and salinity in the Irminger Current have increased since 1997 to the highest values seen for decades (ICES, 2001).

The 2003 survey detected high temperature anomalies within the $0-200 \mathrm{~m}$ layer in the Irminger Sea and adjacent waters. At $200-500 \mathrm{~m}$ depth and deeper waters, positive
anomalies were observed in most of the surveyed area. However, increasing temperature as compared to the survey in June-July 2001 was detected only north of $60^{\circ} \mathrm{N}$ in the flow of the Irminger Current above the Reykjanes Ridge and the northwestern part of the Irminger Sea. These changes in oceanographic conditions might have an effect on the seasonal distribution of redfish and its aggregations in the layer shallower than 500 m in the survey area (ICES, 2003).

In June/July 2005 and 2007, water temperature in the shallower layer ( $0-500 \mathrm{~m}$ ) of the Irminger Sea was higher than normal (ICES, 2005). As in the surveys 1999-2003, the redfish were aggregating in the southwestern part of the survey area, partly influenced by these hydrographic conditions. Favorable conditions for aggregation of redfish in an acoustic layer have been marked only in the southwestern part of the survey area with temperatures between $3.6-4.5^{\circ} \mathrm{C}$, as confirmed by the survey results obtained in 2009.

The hydrography in the survey of June/July 2013 shows that temperature in the survey area is above average but it was lower than in 2011 in most of the surveyed area, except for the Irminger Current (ICES, 2013).

### 18.4 Description of fisheries

There are three species of commercially exploited redfish in ICES Subarea V, VI, XII, and XIV: S. norvegicus (in publication both names S. norvegicus and S. marinus can be found, but according to Fernholm and Wheeler (1983) the first name is the correct name), S. mentella and S. viviparus. S. viviparus has only been of a minor commercial value in Icelandic waters and it is exploited in two small areas south of Iceland at depths of $150-250 \mathrm{~m}$. The landings of S. viviparus decreased from 1160 t in 1997 to $2-$ 9 t in 2003-2006 (Table 18.4.1) due to decreased commercial interest in this species. The landings in 2009 amounted to 37 t , more than a twofold increase in comparison with 2008. After a directed fishery developed in 2010, with a total catch of $2600 t$, the MRI advised on a 1500 t TAC for the 2012-2013 fishing year. Annual catches since 2012 are about 530 t .

The Group has in the past included the fraction of $S$. mentella that are caught with pelagic trawls above the western, south-western and southern continental slope of Iceland as part of the landing statistics of the demersal S. mentella. This practice has been in accordance with Icelandic legislation, where captains are obligated to report their $S$. mentella catch as either "pelagic redfish" or as "demersal redfish" depending in which fishing area they fish. According to this legislation, all catch outside the Icelandic EEZ and west of the 'redfish line' (red line shown in Figure 18.1.1, which is drawn approximately over the 1000-m isoclines within the Icelandic EEZ) shall be reported as pelagic S. mentella. All fish caught east of the 'redfish line' shall be reported as demersal $S$. mentella. Most of the catches since 1991 have been taken by bottom trawlers along the shelf west, southwest, and southeast of Iceland at depths between 500 and 800 m . The Group accepts this praxis as pragmatic management measure, but notes that there is no biological information that could support this catch allocation.

As the Review Group in 2005 noted that this issue needed more elaboration, detailed portrayals of the geographical, vertical and seasonal distribution of the demersal $S$. mentella fisheries with different gears are presented here, as done previously (see below). Quantitative information on the fractions of the pelagic catches of demersal $S$. mentella is given in chapter 18 . The proportion of the total demersal S. mentella catches taken by pelagic trawls has ranged since 1991 between $0 \%$ and $44 \%$ (Table 19.3.2), and is on average $15 \%$. With exception of 2007 , no demersal S. mentella has been caught
with pelagic trawls since 2004. The geographic distribution of the Icelandic fishery for S. mentella since 1991 was in general close to the redfish line, off South Iceland, and has expanded into the NAFO Convention Area since 2003 (Figure 18.4.1). The pelagic catches of demersal S. mentella were taken in similar areas and depths as the bottom trawl catches (Figure 18.4.2). The vertical and horizontal distribution of the pelagic catches focused, however, on smaller areas and shallower depth layers than the bottom trawl catches. The seasonal distribution by depth (Figure 18.4.3) shows that the pelagic catches of demersal S. mentella were in general taken in autumn, and overlapped in June with the traditional pelagic fishery only in 2003 and 2007. The bottom trawl catches of the demersal S. mentella were mainly taken in the first quarter of the year and during autumn/winter. The length distributions of the demersal S. mentella catches in Iceland by gear and area are given in Figure 18.4.4. During 1994-1999 and in 2003, the fish taken with pelagic trawls were considerably larger than the fish caught with bottom trawls, but they were of similar length during 2000-2002. The fish caught in the north-eastern area were on average about 5 cm larger than those caught in the south-western area.

### 18.5 Russian pelagic S. mentel/a fishery

Russia's position regarding the structure of redfish stock in the Irminger Sea remains unchanged and it has been expressed in previous reports (ICES, 2009, Annex 4; ICES, 2013; Makhrov et al. 2011; Zelenina et al. 2011). The Russian Federation still maintains its point of view that there is only one stock of beaked redfish S. mentella in the pelagic waters of the Irminger Sea and that is why no split catches information about the fisheries is presented to the NWWG. Russia reiterates its standpoint that studies of the redfish stock structure should be continued (Artamonova et. al 2013) with the aim of developing agreed recommendations using all available scientific and fisheries data as a basis.

In 2014 the fishery was conducted from April to October in ICES Subareas XII and XIV and NAFO Divisions 1F (Tables 21.2.1, 21.2.2, 22.2.1 and 22.2.2) with average CPUE 26.4 t /day and 15.0 t / day in ICES Subareas XII and XIV, respectively; and $18.8 \mathrm{t} /$ day in NAFO.

### 18.6 Biological sampling

Biological samples are taken both in national and international surveys and from the commercial catches. They consist of length measurements, otolith collection, stomach contents, sex and maturity stages. The following samples were taken by several nations during 2013:

|  | Area | No. of <br> samples | No. of fish measured |
| :--- | :--- | :--- | :--- |
| Russia | XIV | 250 | 37086 |
| Russia | NAFO 1F | 50 | 2822 |
| Iceland | XIV (deep) |  |  |
| Greenland | XIVb |  |  |
| Spain | XIVb (deep) | 51 | 4949 |

### 18.7 Demersal S. mentel/a in Vb and VI

### 18.7.1 Demersal S. mentel/a in Vb

### 18.7.1.1 Surveys

The Faroese spring and summer surveys in Division Vb are mainly designed for species inhabiting depths down to 500 m and do not cover the vertical distribution of demersal S. mentella fully. Therefore, the surveys are not used to evaluate the stock status.

### 18.7.1.2 Fisheries

In Division Vb, landings gradually decreased from 15000 t in 1986 to about 5000 t in 2001 (Table 18.6.1). Between 2002 and 2011 annual landings varied between 1100 and 4,000 t. In 2012 landings decreased drastically and in 2014 they remained below 400 t .

Length distributions from the landings in 2001-2014 indicate that the fish caught in Vb in 2014 are 42 cm (Figure 18.6.1).

Non-standardized CPUE indices in Division Vb were obtained from the Faroese otter board ( OB ) trawlers ( $>1000 \mathrm{HP}$ ) towing deeper than 450 m and where demersal S . mentella composed at least $70 \%$ of the total catch in each tow. The OB trawlers have in recent years landed about $50 \%$ of the total demersal S. mentella landings from Vb . CPUE decreased from $500 \mathrm{~kg} / \mathrm{hour}$ in 1991 to $300 \mathrm{~kg} /$ hour in 1993 and remained at that level until 2012 (Figure 18.6.2). In 2014, the CPUE increased slightly to about 200 kg /hour (Figure 18.6.2).

Fishing effort has decreased since the beginning of the time series and remains very low since 2008.

### 18.7.2 Demersal S. mente/la in VI

### 18.7.2.1 Fisheries

In Subarea VI, the annual landings varied between 200 t and 1100 t in 1978-2000 (Table 18.6.1). The landings from VI in 2004 were negligible ( 6 t ), the lowest recorded since 1978. They increased again to 111 t in 2005 and 179 t in 2006. The reported landings in 2008 were 50 t and no catches were taken since 2009.

### 18.8 Regulations (TAC, effort control, area closure, mesh size etc.)

Management of redfish differs between stock units and is described in sections 17.14 for S. norvegicus, section 18.7 for demersal S. mentella, section 19.10 for shallow pelagic S. mentella and section 20.10 for deep pelagic S. mentella.

The allocation of Icelandic S. mentella catches to the pelagic and demersal management unit has been based on the "redfish line" (see section 17.4).

### 18.9 Mixed fisheries, capacity and effort

The official statistics reported to ICES do not divide catch by species/stocks, and since the Review Group in 2005 recommended that "multispecies catch tables are not relevant to management of redfish resources", these data are not given here and the best estimates on the landings by species/stock unit are given in the relevant chapters. Preliminary official landings data were provided by the ICES Secretariat, NEAFC and NAFO, and various national data were reported to the Group. The Group, however, repeatedly faced problems in obtaining catch data, especially with respect to pelagic $S$.
mentella (see chapter 19.11). Detailed descriptions of the fisheries are given in the respective chapters: S. norvegicus in chapter 17.3, demersal S. mentella in chapter 18.3, shallow pelagic $S$. mentella in chapter 19.2, deep pelagic $S$. mentella in chapter 20.2 and Greenland slope redfish in chapter 21.3.

Information from various sources is used to split demersal landings into two redfish species, S. norvegicus and S. mentella (see stock annexes for Icelandic slope S. mentella and S. norvegicus). In Division Va, if no direct information is available on the catches for a given vessel, the landings are allocated based on logbooks and samples from the fishery. According to the proportion of biological samples from each cell (one fourth of ICES statistical square), the unknown catches within that cell are split accordingly and raised to the landings of a given vessel. For other areas, samples from the landings are used as basis for dividing the demersal redfish catches between S. marinus and $S$. mentella.

### 18.10References

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Table 17.4.1. Landings of S. viviparus in Division Va 1996-2014.

| Year | Landings (t) |
| :--- | :--- |
| 1996 | 22 |
| 1997 | 1159 |
| 1998 | 994 |
| 1999 | 498 |
| 2000 | 227 |
| 2001 | 21 |
| 2002 | 20 |
| 2003 | 3 |
| 2004 | 2 |
| 2005 | 4 |
| 2006 | 9 |
| 2007 | 24 |
| 2008 | 15 |
| 2009 | 37 |
| 2010 | 2602 |
| 2011 | 1427 |
| 2012 | 535 |
| 2013 | 532 |
| 2014 | 550 |

Table 17.6.1. Nominal landings (tonnes) of demersal S. mentella 1978-2014 in ICES DivisionsVb and VI.

| Year | Vb | VI |
| :---: | :---: | :---: |
| 1978 | 7767 | 18 |
| 1979 | 7869 | 819 |
| 1980 | 5119 | 1109 |
| 1981 | 4607 | 1008 |
| 1982 | 7631 | 626 |
| 1983 | 5990 | 396 |
| 1984 | 7704 | 609 |
| 1985 | 10560 | 247 |
| 1986 | 15176 | 242 |
| 1987 | 11395 | 478 |
| 1988 | 10488 | 590 |
| 1989 | 10928 | 424 |
| 1990 | 9330 | 348 |
| 1991 | 12897 | 273 |
| 1992 | 12533 | 134 |
| 1993 | 7801 | 346 |
| 1994 | 6899 | 642 |
| 1995 | 5670 | 536 |
| 1996 | 5337 | 1048 |
| 1997 | 4558 | 419 |
| 1998 | 4089 | 298 |
| 1999 | 5294 | 243 |
| 2000 | 4841 | 885 |
| 2001 | 4696 | 36 |
| 2002 | 2552 | 20 |
| 2003 | 2114 | 197 |
| 2004 | 3931 | 6 |
| 2005 | 1593 | 111 |
| 2006 | 3421 | 179 |
| 2007 | 1376 | 1 |
| 2008 | 750 | 50 |
| 2009 | 1077 | 0 |
| 2010 | 1202 | 0 |
| 2011 | 1126 | 0 |
| 2012 | 263 | 0 |
| 2013 | 398 | 0 |
| 20141 | 370 | 0 |

${ }^{1)}$ Provisional


Figure 17.1.1Potential management unit boundaries. The polygon bounded by blue lines, i.e. 1, indicates the region for the 'deep pelagic' management unit in the northwest Irminger Sea, 2 is the "shallow pelagic" management unit in the southwest Irminger Sea, and 3 is the Icelandic slope management unit.


Figure 17.1.2 Schematic representation of biological stocks and potential management units of $S$. mentella in the Irminger Sea and adjacent waters. The management units are shown in Figure 17.1.1. Included is a schematic representation of the geographical catch distribution in recent years. Note that the shallow pelagic stock includes demersal S. mentella east of the Faroe Islands and the deep pelagic stock includes demersal S. mentella west of the Faroe Islands.


Figure 17.2.1 Survey abundance indices of Sebastes spp. ( $<17 \mathrm{~cm}$ ) for East and West Greenland from the German groundfish survey 1982-2014.


Figure 17.4.1Geographical distribution of the Icelandic catches of S. mentella 1991-2002. The color scale indicates catches (tonnes per $\mathrm{NM}^{2}$ ).


Figure 17.4.1 cont. Geographical distribution of the Icelandic catches of S. mentella 2003-2014. The color scale indicates catches (tonnes per $\mathrm{NM}^{2}$ ).


Figure 17.4.2Distance-depth plot for Icelandic S. mentella catches, where distance (in NM) from a fixed position ( $52^{\circ} \mathrm{N} 50^{\circ} \mathrm{W}$ ) is given. The contour lines indicate catches in a given area and distance. The colored contours represent the fishery on pelagic S. mentella, the black contours indicate bottom trawl catches of demersalS. mentella, and the red contours represent catches of demersal $S$. mentella taken with pelagic trawls.


Figure 17.4.3 Depth-time plot for Icelandic S. mentella catches 1991-2014 where the y-axis is depth, the $x$-axis is day of the year and the color indicates the catches. The colored contours represent the fishery on pelagic S. mentella, the black contours indicate bottom trawl catches of demersal S. mentella, and the red contours represent catches of demersal S. mentella taken with pelagic trawls.
























Figure 17.4.4 Length distributions from different Icelandic S. mentella fisheries, 1991-2014. The blue lines represent the fishery on pelagic S. mentella in the northeastern area, the red lines the pelagic fishery in the southwestern area, the black lines indicate bottom trawl catches of demersal S. mentella, and the green lines represent catches of demersal S. mentella taken with pelagic trawls.


Figure 17.6.1 Length distribution of demersal S. mentella from landings of the Faeroese fleet in Division Vb 2001-2014.


Figure 17.6.2 Demersal S. mentella, CPUE (t/hour) and fishing effort (in thousands hours) from the Faeroese CUBA fleet 1991-2014 and where 70\% of the total catch was demersal S. mentella.

## 19 Golden redfish (Sebastes norvegicus) in Subareas V, VI and XIV

## Executive summary

Total landings in 2014 were about 50800 t , which is about 2500 t less than in 2013. About $94 \%$ of the catches were taken in Division Va. A substantial increase in landings from XIVb since 2010, the highest since early 1990s, and is in relation to re-established redfish fishery in 2010. Very little redfish is now taken in Vb .

Catch-at-age data from Va show that the catch was dominated by two strong year classes from 1985 and 1990. From 2008-2011 year-classes 1996-1999 were the most important in the fisheries. Their share has reduced relatively fast and the 2000-2005 year classes are now most important contributing about $60 \%$ of the total catch.

Recruitment seems to be low in all areas, both according to the Icelandic groundfish surveys, and the German survey and the Greenland shrimp and fish survey in EastGreenland. Recruitment is likely to be underestimated as the surveys do not adequately cover nursery areas of the stock.

The stock was benchmarked in 2014 and a management plan evaluated and adopted. The Gadget model was used as basis for advice but the main difference in settings from earlier years was inclusion data from the German survey in East-Greenland and changes in growth rate.

The management plan was based on $\mathrm{F}_{9-19}=0.097$ reducing linearly if the spawning stock is estimated below $220000 t$ ( $\mathrm{B}_{\text {trigger }}$ ). Blim was proposed as 160000 t , lowest SSB in the 2012 run.

According to the management plan the TAC for 2016 will be 51000 t .

### 19.1 Stock description and management units

Golden redfish (Sebastes norvegicus) in ICES Subareas V and XIV have been considered as one management unit.
Catches in ICES Subarea VI have traditionally been included in this report and the Group continues to do so.

### 19.2 Scientific data

This chapter describes results from various surveys conducted annually on the continental shelves and slopes of Subareas V and XIV.

### 19.2.1 Division Va

Two bottom trawl surveys are conducted in Icelandic waters: the Spring Survey in March 1985-2015 and the Autumn Survey in October 1996-2014. The autumn survey was not conducted in 2011. Two survey indices are calculated from these surveys and used in the assessment of golden redfish in ICES Va. Length disaggregated indices from the Spring Survey are used in the Gadget model. Age disaggregated indices from the autumn survey are used in TSA model but the age data as age-length keys in 2 cm length groups in the Gadget model.

The survey stratification and subsequent survey indices for golden redfish were recalculated for the Autumn Survey in 2008 and for the Spring Survey in 2011. The method is described in the Stock Annex for the species. Further changes were made in
the calculation of the survey indices in 2012 by taking into account length dependent diurnal vertical migration of the species. Golden redfish is known for its diurnal vertical migration showing semi-pelagic behaviour. Usually the species is in the pelagic area during the night time and close to the bottom during the day time. However, there is also a size or age difference in this pelagic behaviour where smaller fish shows opposite vertical migration pattern compared to larger fish. The method is described in more details in the Stock Annex.
This scaled diurnal variation by length was used for calculating Cochran index for redfish. The sum of those abundance indices multiplied by mean weight at length or age are the total indices shown in Figure 19.2.1 and Table 19.2.1.
Figure 19.2.1a shows the total biomass index from the Icelandic spring and autumn groundfish surveys with $\pm 1$ standard deviation in the estimate ( $68 \%$ confidence interval). The total biomass of golden redfish as observed in the spring survey decreased from 1988 to a record low in 1995. Between 1996 and 2002 the stock showed signs of improvement but was low compared to the beginning of the series. In 2003 the biomass increased significantly and has since then been high. The 2012 and 2013 estimates were the highest in the time series, but decreased again in 2014 but remained high. The total biomass index in 2015 was similar as in 2014. The index is $25 \%$ higher than in the beginning of the time series. The CV of the measurement error has been considerably higher since 2003 than before that.

The total biomass index from the autumn survey gradually increased from 2000 to 2014 and was in 2014 the highest in the time series (Figure 19.2.1a).

Length distribution from the spring survey shows that the peaks, which can be seen first in 1987 and then in 1991-1992, reached the fishable stock approximately 10 years later (Figure 19.2.2). The increase in the survey index between 1995 and 2005 reflects the recruitment of a relatively strong year classes (1985-year class and then the 1990year class). Abundance of small redfish has since then been much smaller, highest in 1998-2000, but in recent six years very little has been observed of small redfish (Figure 19.2.1d). This has been confirmed by age readings (Figure 19.2.4). In recent years the modes of the length distribution in both surveys has shifted to the right and is narrower. Much less is now observed of golden redfish less than 30 cm compared to other years (Figures 19.2.2 and 19.2.3).
Age disaggregated abundance indices from the autumn survey is shown in Figure 19.2.4 and Table 19.2.2. The sharp increase in the survey indices since 2005 reflects the recruitment of the year-classes from 1996-2005. The year-classes 1996-1999 are gradually disappearing from the stock. The 2000-2005 year-classes are now similar to the indices of the large 1990 year-class at same age and the 2004 year-class the biggest recorded as a 10 year old fish. In 2013 and 2014, the abundance of fish 7 years' old and younger was at the lowest level in the time series for all age groups although slightly improving in 2014 (Table 19.2.2).

### 19.2.2 Division Vb

In Division Vb, CPUE of S. norvegicus were available from the Faeroes spring groundfish survey from 1994-2015 and the summer survey 1996-2014. Both surveys show similar trends in the indices from 1998 onwards with sharp declines between 1998 and 1999 (Figure 19.2.5). After an increase in the mid-1990s, CPUE decreased drastically. CPUE in the spring survey was between 2000 and 2008 stable at low level. In the period 2009-2015 it has been at the lowest level since the beginning of the series. The CPUE
index in the summer survey has gradually decreased and is also at the lowest level recorded.

### 19.2.3 Subarea XIV

Relative abundance and biomass indices from the German groundfish survey from 1982 to 2014 for S. norvegicus (fish $>17 \mathrm{~cm}$ ) are illustrated in Figure 19.2.6. In 2013, the survey was re-stratified, with 4 strata in West Greenland resembling NAFO sub-area structure, and 5 strata in East Greenland. Depth zones considered are $0-200 \mathrm{~m}$ and $200-400 \mathrm{~m}$. The time series was recalculated accordingly. In general, the survey indices are much lower with the new stratification scheme but show similar trend (WD 30 of the 2013 NWWG report).
After a severe depletion of the $S$. norvegicus stock on the traditional fishing grounds around East Greenland in the early 1990's, the survey estimates showed a significant increase in both abundance and biomass with the highest value observed in 2007 (Figure 17.2.7). The survey indices were high although fluctuating until 2013. The survey index increased in 2014 to the highest level in the time series and was almost two times higher than in 2013 (Figure 19.2.6a and Figure 19.2.6b). It should be noted that the CV for the indices are high and the increase is driven by few very large hauls. During the recent period of increase, both the fishable biomass ( $>30 \mathrm{~cm}$ ) and the biomass of prefishery recruits ( $17-30 \mathrm{~cm}$ ) have increased considerably (Figures 19.2.7c and 19.2.8). In 2010-2014 the biomass of $17-30 \mathrm{~cm}$ fish has decreased compared to previous five years whereas the fishable biomass has remained high since 2007.

Abundance indices of redfish smaller than 18 cm from the German annual groundfish survey show that juveniles were abundant in 1993 and 1995-1998 (Figure 18.2.1). Since 2008, the survey index has been very low and was in 2014 the lowest value recorded since 1982. Juvenile redfish were only classified to the genus Sebastes spp., as species identification of small specimens is difficult due to very similar morphological features. The 1999-2014 survey results indicate low abundance and are similar to those observed in the late 1980s. The Greenland shrimp and fish shallow water survey also shows no juvenile redfish ( $<18 \mathrm{~cm}$, not classified to species) were present.

### 19.3 Information from the fishing industry

### 19.3.1 Landings

Total landings gradually decreased by more than $70 \%$ from about 130000 t in 1982 to about $43,000 \mathrm{t}$ in 1994 (Table 19.3.1 and Figure 19.3.1). Since then, the total annual landings have varied between 33,500 and 54,000 $t$. The total landings in 2014 were 50700 t , which is about 2500 t less than in 2013. The majority of the golden redfish catch is taken in ICES Division Va that contributes to about $94-98 \%$ of the total landings.
Landings of golden redfish in Division Va declined from about 98000 t in 1982 to 39 000 t in 1994 (Table 19.3.1). Since then, landings have varied between 32000 and 51000 t , highest in 2013. The landings in 2014 were about 47800 t , about 3500 t less than in 2013. Between $90-95 \%$ of the golden redfish catch is taken by bottom trawlers targeting redfish (both fresh fish and factory trawlers; vessel length $48-65 \mathrm{~m}$ ). The remaining catches are partly caught as by-catch in gillnet, long-line, and lobster fishery. In 2014, as in previous years, most of the catches were taken along the shelf southwest, west and northwest of Iceland (Figure 19.3.2). A notable change is that higher proportion of the catches is now taken along the shelf northwest of Iceland and less south and southwest.

In Division Vb, landings dropped gradually from 1985 to 1999 from 9000 t to 1500 t and varied between 1500 and 2500 t from 1999-2005 (Table 19.3.1). In 2006-2014 annual landings were less than 1000 t which has not been observed before in the time series. The landings in 2014 were 201 t which is 170 less than in 2013 and the lowest landings in the time series. The majority of the golden redfish caught in Division Vb is taken by pair and single trawlers (vessels larger than 1000 HP ).

Annual landings from Subarea XIV have been more variable than in the other areas (Table 19.3.1). After the landings reached a record high of 31000 t in 1982, the golden redfish fishery drastically reduced within the next three years (the landings from XIV were about 2000 t in 1985). During the period 1985-1994, the annual landings from Subarea XIV varied between 600 and 4200 t , but from 1995 to 2009 there was little or no direct fishery for golden redfish and landings were 200 t or less mainly taken as bycatch in the shrimp fishery. In 2010, landings of golden redfish increased considerable and were 1650 t , similar to it was in early 1990s. This increase is mainly due to increased S. mentella fishery in the area. Annual landings 2010-2014 have been between 1000 t and 2700 t , highest in 2014.

Annual landings from Subarea VI increased from 1978 to 1987 followed by a gradual decrease to 1992 (Table 19.3.1). From 1995 to 2004, annual landings have ranged between 400 and 800 t , but decreased to 137 t in 2005. Little or no landings of golden redfish were reported from Subarea VI in 2006-2014 and were 60 t in 2014.

### 19.3.2 Discard

Comparison of sea and port samples from the Icelandic discard sampling program does not indicate significant discarding due to high grading in recent years (Palsson et al 2010), possibly due to area closures of important nursery grounds west off Iceland. Substantial discard of small redfish took place in the deep-water shrimp fishery from 1986 to 1992 when sorting grids became mandatory. Since then the discard has been insignificant both due to the sorting grid and much less abundance of small redfish in the region.

Discard of redfish species in the shrimp fishery in Subarea XIVb is currently considered insignificant (see Chapter 18).

### 19.3.3 Biological data from the commercial fishery

The table below shows the fishery related sampling by gear type and ICES Divisions in 2014. No sampling of the commercial catch from subdivision VI was carried out.

|  |  |  |  | No. <br> length <br> measured | No. Age <br> read |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Area | Nation | Gear | Landings (t) | Samples |  |  |
| Va | Iceland | Bottom trawl | 4,769 | 229 | 40818 | 1492 |
| Vb | Faroe Islands | Bottom trawl | 201 | 10 | 444 |  |
| XIV | Greenland | Bottom trawl | 2706 |  |  |  |

### 19.3.4 Landings by length and age

The length distributions from the Icelandic commercial trawler fleet in 1975-2014 show that the majority of the fish caught is between 30 and 45 cm (Figure 19.3.3). The modes of the length distributions range between 35 and 37 cm . The length distributions in 2012-2014 are narrower than previously, with less than average of both small and large fish caught.

Catch-at-age data from the Icelandic fishery in Division Va show that the 1985-year class dominated the catches from 1995-2002 (Figure 19.3.4 and Table 19.3.2) and in 2002 this year class still contributed to about $25 \%$ of the total catch in weight. The strong 1990-year class dominated the catch in 2003-2007 contributing between 25-30\% of the total catch in weight. This year class contributed about $6 \%$ of the total catch in weight in 2013 and the 1985-year class about 1.5\%, but their share has gradually been decreasing in recent years. In 2007-2010 the 1996-1999 year classes dominated in the catches, but are now gradually decreasing. The 2000-2004 year classes contributed in total about $60 \%$ of the total catch in 2014.

The average total mortality $(Z)$, estimated from the 16-year series of catch-at-age data (Figure 19.3.5) is about 0.20 for age groups 15 and older.

Length distribution from the Faroese commercial catches for 2001-2014 indicates that the fish caught are on average larger than 40 cm with modes between 45 cm and 50 cm (Figure 19.3.6).

No length data from the catches have been available for several years in Subareas XIV and VI.

### 19.3.5 CPUE

CPUE in Va was calculated as non-standardized CPUE and standardised using GLM multiplicative model. Description is given in the stock annex. The outcome of the GLM model run is given in Table 19.3.3 and Figure 19.3.7 and the model residuals in Figure 19.3.8. CPUE derived from logbooks is not considered indicative of stock trends however the information contained in the logbooks on effort, spatial and temporal distribution the fishery is of value.

The CPUE index derived from the GLM model increased considerably in 2001 after being at low level 1993-1999 and was until 2006 high but stable (Figure 19.3.7). In 2006, the CPUE index decreased by $12 \%$ compared to the previous year but has since then increased rapidly. Both the un-standardized CPUE index and the one derived from the GLM model was in 2014 the highest in the time series with sharp increase in recent 8 years. Effort towards golden redfish has since 1986 gradually decreased and is at the lowest level recorded (Figure 19.3.7).

Un-standardized CPUE of the Faroese otter-board (OB) trawlers has been presented in previous reports. They are however considered unreliable and un-representative about the stock in Division Vb. This is because no separation of S. norvegicus/S. mentella is made in the catches.

### 19.4 Methods

### 19.4.1 Changes to the assessment model in January 2014.

The stock was benchmarked in January 2014 and a management plan evaluated and adopted (WKREDMP, ICES 2014). The benchmark group agreed to base the advice for next five years on the Gadget model. The settings are described in the Stock Annex. The following changes were done to the model compared to previous runs:
Abundance indices from the German survey in East Greenland were included in the tuning. The indices were added to the Icelandic spring survey.

Tuning data were limited to $19-54 \mathrm{~cm}$ instead of $25-54 \mathrm{~cm}$ as larger part of the stock area is included. 19 cm is around the length at which redfish in the German survey is
classified to species. Earlier, smaller fish had gradually been removed from the tuning fleet as the nursery area for year classes 1996-2003 seemed to be outside Icelandic waters.

Length at recruitment was estimated separately for year classes 1996-2000 and 2001 and onwards. The reason was higher mean weight at age in landings and autumn survey.
Of the changes mentioned above, the first one has the largest effect on the estimated stock size but the third one does also have considerable effect as when growth increases fishes recruit to the fisheries at younger age if selection is size dependent.
The German survey did get half weight compared to the results in Figure 19.2.6. This was done to avoid extrapolation to areas not surveyed, and hence reduce noise, but the indices are calculated as numbers per square $\mathrm{km}^{2}$ multiplied by an area drawn around the stations (Figure 19.4.1). By using the stratification used to calculate indices shown in Figure 19.2.6, each station in the German survey would get 2.5 times more weight compared to the Icelandic survey. Several things are not comparable between the two surveys, for example different gears are used and the German survey is not conducted during night while the Icelandic survey is conducted both day and night. Therefore the "correct" weight of each survey in the total is difficult to estimate and part of the benchmark work 2014 was to look at the sensitivity to the weight.

The German survey has in recent decade provided increased proportion of the total biomass, but is still only $10 \%$ of the total biomass (Figure 19.4.2). The contribution for each length group (Figure 19.4.3) does though show that large redfish is abundant in East Greenland and large part of the largest redfish $(45+\mathrm{cm})$ is found there. This affects the model results as the relatively large abundance of middle size redfish in the Icelandic spring survey (Figure 19.2.1a) has not lead to subsequent increase in large fish (Figure 19.2.1c). Including the large fish from East Greenland does therefore affect model results and estimated SSB is $20 \%$ higher when the German survey is included, even though the German survey does only account for $10 \%$ of the total biomass as it is weighted. The recruitment signal from the German survey (Figure 19.4.3) is on the other hand not explaining much of the "missing recruitment" from Icelandic waters in recent years.
The weighing of individual data sets in the GADGET model is done using an iterative re-weighing algorithm. The process essentially assigns weights to each input data set on the basis of the inverse variance of the fitted residuals. This is done to reduce the effect of low quality input data. It can also help to identify data discrepancy as shown in Figure 19.4.4 (taken from the WKREDMP report; ICES 2014) which shows that information from the commercial catches indicate status quo state of the stock while the increase is caused by the survey data. In this year assessment the weights were the same as in the benchmark runs in January 2014 and the assessment in May 2014.

### 19.4.2 Gadget model

### 19.4.2.1 Data and model settings

Below is a brief description of the data used in the model and model settings is given. A more detailed description is given in the Stock annex.
Data used in the GADGET model are:
Length disaggregated survey indices $19-54 \mathrm{~cm}$ in 2 cm length increments from the Icelandic groundfish survey in March 1985-2015 and the German survey in East

Greenland 1984-2014. Indices are added together and the German survey gets half the weight compared to what is presented in Figure 19.2.6.
Length distributions from the Icelandic, Faroe Islands and East Greenland commercial catches since 1970.

Landings by 6 month period from Iceland, Faroe Islands and East Greenland.
Age-length keys and mean length at age from the Icelandic groundfish survey in October 1996-2014.
Age-length keys and mean length at age from the Icelandic commercial catch 19952014.

The simulation period is from 1970 to 2019 using data until the first half of 2015 for estimation. Two time steps are used each year. The ages used were 5 to 30 years, where the oldest age is treated as a plus group (fish 30 years and older). Recruitment was set at age 5 .
Estimated parameters are:
Number of fishes when the simulation starts (8 parameters).
Recruitment at age 5 each year ( 43 parameters).
Length at recruitment (3 parameters).
Parameters in the growth equation; (2 parameters).
Parameter $\beta$ of the beta-binomial distribution controlling the spread of the length distribution.

Selection pattern of the three commercial fleets assuming logistic selection (S-shape) ( $3 \times 2$ parameters).
Selection pattern of the survey fleet assuming an Andersen selection curve (bell-shape) (3 parameters).
It needs to be mentioned that the length disaggregated indices are from the spring survey but the age data are from the autumn survey conducted six months later. The surveys could have different catchability but the age data are used as proportions within each 2 cm length group so it should not matter. Growth in between March and October is taken care of by the model.
Projections were run using the Gadget model based fishing mortality of equal to 0.097 for ages 9 to 19 according to agreed management plan.
Assumptions done in the predictions:
Recruitment at age 5 in 2013 and onwards was set as the average of the recruitment in 1970-2012.
Catches in the first time step in 2015 (first 6 months) were set at the same as in the first time step of 2014 for all the fleets. In step 2 in 2015 and onwards the model was run at fixed effort corresponding to $\mathrm{F}_{9-19}=0.097$
The estimated selection pattern from the Icelandic fleet was used for projections.

### 19.4.2.2 Results of the assessment model and predictions

Summary of the assessment is shown in Figure 19.4.5 and Table 19.4.1. The spawning stock has increased in recent years and fishing mortality decreased but annual landings have been relatively stable. The last year class estimated is the 2007 year class but the
following year-classes are assumed to be the average. Compared to last year's assessment the 2006 year-class is estimated smaller than assumed last year but the 2004 and 2005 year-classes are estimated larger. Later year-classes are likely to be smaller than assumed here based on information from the surveys in East-Greenland and Iceland that all indicate low abundance of small redfish. Assumptions about those year-classes will not have much effect on the advice this year but later advice will be affected as well as the development of the spawning stock in short term. The surveys do not seem to cover the nursery areas of the stock.

The results of the assessment presented here are similar to what was presented at WKREDMP (ICES 2014) (Figure 19.4.6). This similarity is expected as only one year of data has been added and the model is a is a low pass filter that does usually not respond rapidly to new data except they are very far from predicted values.

Estimated selection patterns of different fleets are shown in Figure 19.4.7. The Greenlandic and Faeroese fleet catch much larger fish than the Icelandic fleet. This is in line with the results from the German survey in East-Greenland that show most of the large fish in East-Greenland (Figure 19.4.3)

### 19.4.2.3 Fit to data

An aggregated fit to the survey index (converted to biomass) is presented in Figure 19.4.8. It shows a greater level of agreement than most runs based only on the Icelandic data but does mostly show negative residuals for the last 13 years. Residuals by length group show positive residuals in size groups $33-38 \mathrm{~cm}$ in recent years but negative for most other size groups, indicating narrower length distributions in the survey than predicted (Figure 19.5.9).

This lack of fit between observed and predicted survey biomass was one of the main critics of WKRED 2012 (ICES 2012). As can be seen in Figure 19.4.8 the fit is still not good. That lack of fit is caused by too narrow length distribution, with both small and large fish missing but they weight much more in the tuning data than in the total biomass. When looking at the number of years with observed > predicted biomass it must be noted that the assessment converges very slowly and 10 years are in some sense comparable to less than 5 years in other species. Discussions about the problem in WKRED 2014 are still valid.

The correlation between observed and predicted survey indices is good for $33-50 \mathrm{~cm}$ fish (Figures 19.4.10 and 19.4.11). As the model converges slowly, predicted indices could change a number of years back when more data are added. However, it is not the magnitude of the residuals but rather the temporal pattern that is worrying (Figure 19.4.9).

Length distributions from the Icelandic commercial catch does usually show good fit except in the most recent period when the large fish is missing and the length distribution narrower than ever (Figure 19.4.11). One explanation could be that selection in recent years is dome shaped as the large fish is in East-Greenland where the fisheries are less.

The discrepancy between predicted and observed age distributions is not as apparent as for the length distributions (Figures 19.4.12 and 19.4.13). The model uses the data as age-length keys in 2 cm intervals for tuning. Presenting the residuals on that scale is difficult so here the age distributions are shown as aggregates over all length groups. This is not a problem for the catches where the otolith sampling is random, which is
not the case for the survey as there is a maximum limit on the number of otoliths sampled in each tow and therefore lower proportion sampled in hauls with many fish.

### 19.5 Information from catch curves.

The discrepancy in different data sources can be seen by looking at catch curves from age disaggregated catch in numbers and survey indices. The 1995-1999 year-classes have disappeared more rapidly from the fisheries than predicted (Figure 19.4.14) with average $Z$ being $0.24(F=0.19)$ for ages $12-20$. Comparable number for year-classes $1985-1990$ is $\mathrm{Z}=0.15$.

The analyses indicate that fishing mortality was higher than predicted by the assessment models. One explanation is that we are overestimating the stock but there can be a number of alternative explanations.

1. The cohorts grow faster and mature earlier than earlier cohorts. Natural mortality, $M$, might have increased
2. The selection of the fisheries is more dome shaped than before. The fisheries concentrate on the dense schools west of Iceland where the length distribution is narrow.
3. Compared to cohorts 1985-1990 the later cohorts seem to come from other nursery areas.
4. Most of the biomass in the Icelandic surveys in the last decade comes from very dense schools west of Iceland. Catchability in those schools might be different from less dense aggregations.

### 19.6 Alternative assessment

Time Series Analysis (TSA) was run this year as in recent year (ICES 2014). This year, only the age based TSA model was presented. Two model runs were predicted.

1. Catch in numbers by age $1995-2014$, ages $9-19$.
2. Catch in numbers by age 1995-2014, ages 9-19 and age disaggregated indices from the Icelandic autumn survey 1995-2014, ages 7-19.

Figure 19.4.15 shows the biomass of age $9-19$ based on the TSA and Gadget model and Figure 19.4.16 the average fishing mortality. The model tuned with the autumn survey shows similar biomass as the results from the Gadget model. The difference is that data from Greenland is not included in the TSA model. The TSA model based on catch in numbers only has to rely on how the cohorts disappear from the fisheries as it does not have any information on incoming cohorts other than proportions in catch in numbers. Also, the lack of information for older than 19 years old fish is a problem in that model. The difference between the two TSA models does indicate some of the discrepancies seen in the data but the TSA model tuned with the autumn survey does, as the Gadget model, depends on data from the catches and the surveys.

### 19.7 Reference points

Harvest control rule (HCR) was evaluated in January 2014 based on stochastic simulations using the Gadget model. Taking into account conflicting information by different data continuing for many consequent years (sections 19.4-19.6) the simulations were conducted using large assessment error with very high autocorrelation. (CV=0.25,
rho=0.9). It can therefore be argued that the problems described in sections 19.4-19.6 were taken into account by relatively conservative HCR.

Yield-per-recruit analysis show that when average size at age 5 was allowed to change after year class 1996 F9-19,MAX changed from 0.097 to 0.114 . FMAX of fully recruited fish or size based $\mathrm{F}_{\mathrm{MAX}}$ does not change. This is a known phenomenon, for example taken into account in the management of Icelandic haddock and George bank haddock. The proposed fishing mortality of 0.097 is therefore around $85 \%$ of $\mathrm{F}_{\text {max }}$ with current settings. Stochastic simulations indicate that it leads to very low probability of spawning stock going below $B_{\text {trigger }}$ and $B_{l i m}$, even with relatively large auto-correlated assessment error

Yield-per-recruit reference points from the Gadget model (length-based) are not comparable to age based reference points. The proposed harvest ratio, 0.097 , is well above $\mathrm{F}_{0.1}$ and $\mathrm{F}_{5 s 835}$ estimate from the Gadget model. These reference points have previously been proposed for this stock, but these points are also lower than from age based models.

The recruitment pattern observed from year classes 1975-2003 (Figure 19.4.6) does not show long periods of poor recruitment. From a management perspective this is beneficial since overly cautious rules (i.e. low harvest rates) may not be needed to see the stock through long periods of very low recruitment. A spawning stock generated by poor recruitment and low fishing mortality has much broader, and hence more resilient, age distribution than the same size spawning stock generated under higher fishing mortality and a few large recruitment events. Therefore, if poor recruitment lead to the stock declining towards Bloss after adoption of the HCR, 19+ biomass (or another measure of old fish) would still be relatively high, potentially benefitting the stock due the disproportionate reproductive output of older fish.
$B_{\text {trigger }}$ was defined as 220 kt by adding a precautionary buffer to the proposed $\mathrm{B}_{\mathrm{lim}}$ of 160 kt : $160^{*} \exp \left(0.2^{*} 1.645\right)$. The probability of current $\mathrm{SSB}<\mathrm{B}_{\text {trigger }}$ is estimated $2.7 \%$. For simplicity, the action of $B_{\text {trigger }}$ is not included in the simulations since Gadget is not keeping track of "perceived spawning stock". Analysis of the stochastic prediction in R shows that if SSB is below $\mathrm{B}_{\text {trigger }}$ it will only be noted in $<15 \%$ of the cases. The reason is that the spawning stock is only likely to go below $\mathrm{B}_{\text {trigger }}$ in periods of severe overestimation of the stock that occur due to the assumed high autocorrelation in assessment error. This situation differs from that of the stock going below $\mathrm{B}_{\text {trigger }}$ due to poor recruitment (worse than observed in recent decades). In this case the spawning stock should still have a resilient age structure (as discussed above) and this could reduce the need to take further action below $\mathrm{B}_{\text {trigger }}$.

Data on recruitment are still poor and data from other surveys at East-Greenland than the German survey need to be investigated. The Icelandic surveys indicate that recruitment has been very poor for at least the last five years (Figure 19.2.1). The applicability of the Icelandic surveys as measure of recruitment of redfish has been questioned but this is at least a negative signal and in long periods of poor recruitment a low harvest ratio is preferable.

Finally, it must be noted that the $\mathrm{F}_{\text {target }}$ suggested implies a substantial reduction in fishing mortality compared to the last three decades. The stock is not at present considered to be in a very unhealthy state despite this three decade period of relatively high fishing pressure in relation to that proposed for the HCR. Still, the adoption of the HCR should not lead to major changes in the advice from recent years, which has partly been based on similar considerations.

The deliberations above offer some justification that the proposed harvest rate ( $\mathrm{F}_{9-19}=$ 0.097 ) is a sensible target for this stock. This of course depends also on the assumption that assessment is based on natural mortality $M=0.05$.

### 19.8 State of the stock

The results from GADGET indicate that fishing mortality has reduced in recent years and is now close to Fmsy (Figure 19.4.5). Spawning stock and fishable stock have been increasing in recent years and are now the highest since 1986.

In Vb , survey indices are stable at low level and do not indicate an improved situation in the area. In Subarea XIV, the biomass of the fishable stock has been relatively high since 2007. No information is available on exploitation rates in Divisions Vb and XIV.

Results from surveys in Iceland and East Greenland indicate that most recent year classes are poor. The reliability of the surveys as an indicator of recruitment is not known.

### 19.9 Short term forecast

The Gadget model is length based where growth is modelled based on estimated parameters. The only parameters needed for short term forecast are assumptions about size of those cohorts that have not been seen in the surveys. These year classes were assumed to be the average of year classes 1975-2003 (Figure 19.4.5).

The results from the short term simulations based on $\mathrm{F}_{9}-19$ is shown in Figure 19.4.5 and from short term prognosis with varying fishing mortality in 2015 in Table 19.4.2.

### 19.1 0 Medium term forecast

No medium term forecast was carried out.

### 19.1 1 Uncertainties in assessment and forecast

Various factors regarding the uncertainty and modelling challenges are listed in the WKRED-2012 and WKREDMP-2014 reports. The main things relate to the lack of explanation of the GADGET model (or any model for that matter) to account for the increase of abundance in intermediate length groups in the Icelandic March survey. These factors were discussed in sections 19.4-19.7 but a short list is repeated below.
Immigration of intermediate sized redfish in to Va, most likely from Greenland.
Increased aggregation of redfish in areas closed to fishing. These areas on the western part of the Icelandic shelf make up most but not all of the increase in intermediate sized golden redfish in the Icelandic surveys. However eliminating the hauls from these areas in calculation of indices does to some extend reduce this increase.

There are indications that growth of golden redfish has changed over time. This can be seen for example in the 2001 year class which is on average larger than fish of the same age in the earlier year classes (for example, the 1985-1990 year classes). Size at maturity has also decreased that could lead to growth ceasing earlier than before explaining lack of large fish in recent years

### 19.12Comparison with previous assessment and forecast

The current assessment gives similar state of the stock compared to last year's assessment and the assessment presented at the benchmark 2014. Management plans and evaluation

See chapter 19.7

### 19.13 Basis for advice

Harvest control rule accepted at WKREDMP 2014 (ICES 2014).

### 19.14 Management consideration

In 2009 a fishery targeting redfish was initiated in ICES XIV with annual catches of more than 8300 tonnes in 2010-2013 and 7300 in 2014. The fishery does not distinguish between species, but based on survey information, golden redfish is estimated to be between 1000 and 2700, highest in 2014..

Redfish and cod in XIV are found in the same areas and depths and historically these species have been taken in the same fisheries. An increased redfish fishery may therefore affect cod. ICES presently advise that no fishery should take place on offshore cod in Greenland waters. ICES therefore recommend measures that will keep effort on cod low in the redfish fishery.

Greenland opened an offshore cod fishery in 2008. To protect spawning aggregations of cod present management measures in Greenland EEZ prohibits trawl fishery for cod north of $63^{\circ} \mathrm{N}$ latitude. Restrictions on cod bycatch in fisheries directed towards other demersal fish (i.e. redfish and Greenland halibut) provide some protection of cod, but additional measures such as a closure of potential redfish fisheries north of $63^{\circ} \mathrm{N}$ could be considered.

Subarea XIV is an important nursery area for the entire resource. Measures to protect juvenile in Subarea XIV should be continued (sorting grids in the shrimp fishery).

No formal agreement on the management of $S$. norvegicus exists among the three coastal states, Greenland, Iceland and the Faeroe Islands. In Greenland and Iceland the fishery is regulated by a TAC and in the Faeroe Islands by effort limitation. The regulation schemes of those states have previously resulted in catches well in excess of TACs advised by ICES.

### 19.15 Ecosystem consideration

Not evaluated for this stock.

### 19.16 Regulation and their effects

The separation of golden redfish and Icelandic slope S. mentella quota was implemented in the 2010/2011 fishing season.

In the late 1980's, Iceland introduced a sorting grid with a bar spacing of 22 mm in the shrimp fishery to reduce the by-catch of juveniles in the shrimp fishery north of Iceland. This was partly done to avoid redfish juveniles as a by-catch in the fishery, but also juveniles of other species. Since the large year classes of golden redfish disappeared out of the shrimp fishing area, there in the early 1990's, observers report small redfish as being negligible in the Icelandic shrimp fishery. If the sorting grids work where the abundance of redfish is high is a question but not a relevant problem at the moment in Vb as abundance of small redfish is low and shrimp fisheries limited.

There is no minimum landing size of golden redfish in Va. However, if more than 20\% of a catch observed onboard is below 33 cm a small area can be closed temporarily. A large area west and southwest of Iceland is closed for fishing in order to protect young golden redfish.

There is no regulation of the golden redfish in Vb .
Since 2002 it has been mandatory in the shrimp fishery in Subarea XIV to use sorting grids in order to reduce by-catches of juvenile redfish in the shrimp fishery.

### 19.17 Changes in fishing technology and fishing patterns

There have been no changes in the fishing technology and the fishing pattern of golden redfish in Subareas V and XIV.

### 19.18Changes in the environment

No information available.

### 19.19References

ICES 2012. Report of the Benchmark Workshop on Redfish (WKRED 2012). ICES CM 2012/ACOM:48, 291 pp.
ICES 2014. Report of the Workshop on Redfish Management Plan Evaluation (WKREDMP). ICES CM 2014/ACOM:52, 269 pp .

Pálsson, Ó., Björnsson, H., Björnsson, E., Jóhannesson, G. and Ottesen P. 2010. Discards in demersal Icelandic fisheries 2009. Marine Research in Iceland 154.

Table 19.2.1 Survey indices and CV from the spring survey 1985-2015 and the autumn survey 19962014.

|  | Spring Survey |  | Autumn Survey |  |
| :---: | :---: | :---: | :---: | :---: |
| Year | Biomass | CV | Biomass | CV |
| 1985 | 308,145 | 0.095 |  |  |
| 1986 | 328,081 | 0.120 |  |  |
| 1987 | 322,175 | 0.122 |  |  |
| 1988 | 251,597 | 0.095 |  |  |
| 1989 | 281,241 | 0.122 |  |  |
| 1990 | 242,496 | 0.223 |  |  |
| 1991 | 199,179 | 0.114 |  |  |
| 1992 | 160,561 | 0.088 |  |  |
| 1993 | 179,369 | 0.130 |  |  |
| 1994 | 171,167 | 0.097 |  |  |
| 1995 | 146,124 | 0.102 |  |  |
| 1996 | 195,661 | 0.164 | 197,168 | 0.248 |
| 1997 | 212,249 | 0.216 | 117,771 | 0.284 |
| 1998 | 206,544 | 0.136 | 186,466 | 0.348 |
| 1999 | 297,008 | 0.143 | 262,029 | 0.310 |
| 2000 | 221,273 | 0.176 | 137,959 | 0.200 |
| 2001 | 192,739 | 0.176 | 171,417 | 0.158 |
| 2002 | 249,371 | 0.173 | 187,889 | 0.147 |
| 2003 | 334,004 | 0.161 | 196,233 | 0.156 |
| 2004 | 327,174 | 0.236 | 214,472 | 0.239 |
| 2005 | 310,706 | 0.129 | 224,245 | 0.237 |
| 2006 | 257,208 | 0.157 | 272,377 | 0.332 |
| 2007 | 339,967 | 0.224 | 215,214 | 0.249 |
| 2008 | 248,119 | 0.154 | 280,348 | 0.242 |
| 2009 | 302,498 | 0.253 | 285,841 | 0.281 |
| 2010 | 383,771 | 0.245 | 223,862 | 0.172 |
| 2011 | 401,870 | 0.235 |  |  |
| 2012 | 462,078 | 0.204 | 333,915 | 0.224 |
| 2013 | 457,736 | 0.177 | 302,568 | 0.156 |
| 2014 | 403,083 | 0.174 | 416,901 | 0.233 |
| 2015 | 407,050 | 0.280 |  |  |

Table 19.2.2 Golden redfish in Va. Age disaggregated indices (in numbers) from the autumn groundfish survey 1996-2014. The survey was not conducted in 2011.

| Year/ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |
| 1 | 0.3 | 1.0 | 3.7 | 3.3 | 0.8 | 0.4 | 0.1 | 0.0 | 0.0 | 0.1 | 0.2 | 0.1 | 0.0 | 0.1 | 0.0 |  | 0.0 | 0.0 | 0.0 |
| 2 | 2.4 | 0.3 | 1.5 | 3.3 | 1.7 | 1.0 | 1.0 | 0.6 | 0.2 | 0.1 | 0.6 | 1.3 | 0.3 | 0.3 | 0.0 |  | 0.0 | 0.0 | 0.3 |
| 3 | 0.7 | 2.2 | 0.9 | 3.3 | 1.4 | 2.0 | 1.5 | 1.1 | 1.0 | 0.2 | 0.7 | 1.2 | 2.5 | 0.4 | 1.7 |  | 0.1 | 0.0 | 0.3 |
| 4 | 1.6 | 1.6 | 2.3 | 1.5 | 1.6 | 2.4 | 6.1 | 1.1 | 1.9 | 1.0 | 0.5 | 1.1 | 2.7 | 4.6 | 0.3 |  | 1.1 | 0.2 | 0.1 |
| 5 | 8.4 | 2.2 | 0.9 | 4.7 | 1.2 | 5.2 | 5.8 | 12.2 | 3.2 | 4.2 | 5.0 | 2.1 | 4.1 | 12.2 | 4.3 |  | 3.9 | 1.1 | 0.8 |
| 6 | 40.4 | 6.9 | 3.5 | 2.8 | 7.8 | 2.2 | 11.7 | 17.4 | 28.1 | 4.8 | 6.8 | 10.2 | 7.7 | 11.6 | 14.3 |  | 3.1 | 4.0 | 1.7 |
| 7 | 11.4 | 22.4 | 16.7 | 10.5 | 6.6 | 10.6 | 3.2 | 37.5 | 35.9 | 39.0 | 15.2 | 25.6 | 38.3 | 13.7 | 15.0 |  | 23.1 | 3.0 | 12.6 |
| 8 | 19.0 | 14.2 | 58.5 | 47.2 | 6.1 | 10.7 | 25.5 | 9.6 | 63.8 | 43.9 | 79.8 | 35.0 | 73.1 | 72.4 | 23.0 |  | 68.6 | 40.8 | 23.8 |
| 9 | 14.7 | 12.8 | 22.4 | 100.0 | 25.5 | 6.8 | 10.9 | 47.4 | 20.3 | 61.2 | 79.1 | 74.8 | 65.7 | 94.0 | 53.4 |  | 58.9 | 82.3 | 93.8 |
| 10 | 28.6 | 10.8 | 26.0 | 43.4 | 92.8 | 16.6 | 15.9 | 12.2 | 44.2 | 24.1 | 83.2 | 36.3 | 103.3 | 56.9 | 67.8 |  | 61.0 | 54.0 | 146.8 |
| 11 | 103.4 | 17.3 | 18.7 | 20.3 | 11.0 | 109.3 | 30.8 | 16.5 | 18.6 | 43.1 | 25.4 | 35.2 | 61.2 | 98.2 | 31.8 |  | 100.9 | 39.3 | 87.7 |
| 12 | 15.7 | 67.4 | 19.0 | 16.5 | 13.8 | 23.0 | 114.6 | 39.0 | 12.9 | 19.0 | 36.4 | 18.5 | 53.5 | 44.6 | 56.6 |  | 71.8 | 65.2 | 67.3 |
| 13 | 9.7 | 5.9 | 105.4 | 20.6 | 7.6 | 22.7 | 19.5 | 109.6 | 25.9 | 15.0 | 17.5 | 23.2 | 13.1 | 41.7 | 28.2 |  | 42.1 | 45.2 | 65.3 |
| 14 | 16.6 | 5.1 | 10.0 | 148.1 | 7.8 | 7.6 | 11.0 | 12.1 | 101.5 | 26.3 | 14.6 | 7.9 | 17.7 | 9.8 | 19.3 |  | 38.1 | 25.1 | 48.8 |
| 15 | 34.0 | 7.0 | 7.6 | 5.8 | 50.6 | 8.7 | 9.5 | 10.6 | 13.3 | 80.8 | 17.9 | 6.6 | 8.8 | 17.7 | 8.9 |  | 19.1 | 30.1 | 26.1 |
| 16 | 15.9 | 9.8 | 7.7 | 9.6 | 5.1 | 57.4 | 10.3 | 6.0 | 9.4 | 9.3 | 74.0 | 16.6 | 7.6 | 6.7 | 10.8 |  | 16.2 | 17.8 | 25.7 |
| 17 | 1.7 | 6.8 | 14.2 | 10.8 | 2.5 | 4.1 | 45.1 | 7.5 | 5.8 | 6.5 | 8.5 | 48.8 | 12.8 | 6.2 | 4.6 |  | 5.9 | 12.2 | 16.5 |
| 18 | 1.6 | 3.9 | 7.6 | 11.1 | 2.5 | 4.9 | 4.5 | 32.5 | 5.9 | 3.7 | 4.2 | 10.2 | 36.0 | 7.1 | 3.0 |  | 5.8 | 6.8 | 11.9 |
| 19 | 4.2 | 2.0 | 0.5 | 8.4 | 4.5 | 3.5 | 2.7 | 4.5 | 21.2 | 5.0 | 2.7 | 4.4 | 6.0 | 27.7 | 6.6 |  | 3.8 | 4.9 | 5.8 |
| 20 | 6.5 | 1.4 | 3.2 | 3.9 | 6.5 | 4.1 | 3.2 | 1.6 | 3.0 | 21.8 | 3.1 | 1.5 | 5.6 | 4.6 | 22.0 |  | 3.8 | 4.3 | 5.7 |
| 21 | 1.0 | 0.8 | 2.4 | 2.8 | 1.0 | 3.6 | 3.9 | 1.1 | 1.8 | 2.5 | 17.6 | 3.9 | 2.0 | 2.1 | 3.1 |  | 3.4 | 4.6 | 4.7 |
| 22 | 4.9 | 1.5 | 0.8 | 1.0 | 1.6 | 2.2 | 3.1 | 2.7 | 1.7 | 2.0 | 1.9 | 13.6 | 2.3 | 1.3 | 1.2 |  | 17.9 | 2.3 | 3.5 |
| 23 | 3.9 | 2.4 | 2.2 | 2.0 | 0.4 | 0.3 | 0.8 | 1.0 | 2.4 | 2.3 | 1.7 | 1.3 | 10.8 | 1.9 | 1.6 |  | 2.8 | 17.3 | 3.3 |
| 24 | 4.5 | 0.8 | 0.4 | 0.5 | 1.0 | 0.5 | 0.4 | 0.3 | 0.0 | 0.9 | 1.0 | 1.2 | 1.4 | 10.0 | 0.7 |  | 2.0 | 2.4 | 12.3 |
| 25 | 3.8 | 2.7 | 1.4 | 2.8 | 0.7 | 0.2 | 0.5 | 0.3 | 1.2 | 1.2 | 1.7 | 0.2 | 0.8 | 0.7 | 5.7 |  | 1.2 | 1.2 | 1.4 |
| 26 | 0.8 | 1.1 | 0.2 | 1.1 | 0.6 | 0.5 | 0.5 | 0.2 | 0.4 | 0.3 | 0.9 | 0.6 | 0.8 | 0.9 | 0.6 |  | 1.6 | 1.1 | 0.9 |
| 27 | 0.8 | 0.2 | 0.9 | 2.9 | 0.5 | 0.7 | 0.3 | 0.3 | 0.0 | 0.1 | 0.9 | 0.3 | 1.2 | 1.3 | 0.4 |  | 7.4 | 0.8 | 0.8 |
| 28 | 0.8 | 0.4 | 0.5 | 1.5 | 0.6 | 0.5 | 0.2 | 0.0 | 0.2 | 0.2 | 0.2 | 0.0 | 0.5 | 0.2 | 0.7 |  | 0.4 | 8.3 | 0.4 |
| 29 | 0.1 | 0.0 | 0.4 | 1.2 | 0.5 | 0.2 | 0.7 | 0.1 | 0.2 | 0.0 | 0.4 | 0.4 | 0.8 | 1.5 | 0.4 |  | 0.4 | 0.4 | 3.2 |
| $30+$ | 0.8 | 1.3 | 3.1 | 1.1 | 1.3 | 2.1 | 1.4 | 1.5 | 1.5 | 2.1 | 1.0 | 0.9 | 1.4 | 1.6 | 2.0 |  | 2.0 | 3.3 | 2.5 |
| Total | 358.0 | 211.8 | 342.3 | 492.0 | 265.5 | 313.9 | 344.4 | 386.4 | 425.4 | 420.5 | 502.6 | 382.9 | 542.3 | 551.9 | 387.9 |  | 566.4 | 477.9 | 674.0 |

Table 19.3.1 Official landings (in tonnes) of golden redfish, by area, 1978-2013 as officially reported to ICES. Landings statistics for 2014 are provisional.

| Area |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Va | Vb | VI | XIV | Total |
| 1978 | 31300 | 2039 | 313 | 15477 | 49129 |
| 1979 | 56616 | 4805 | 6 | 15787 | 77214 |
| 1980 | 62052 | 4920 | 2 | 22203 | 89177 |
| 1981 | 75828 | 2538 | 3 | 23608 | 101977 |
| 1982 | 97899 | 1810 | 28 | 30692 | 130429 |
| 1983 | 87412 | 3394 | 60 | 15636 | 106502 |
| 1984 | 84766 | 6228 | 86 | 5040 | 96120 |
| 1985 | 67312 | 9194 | 245 | 2117 | 78868 |
| 1986 | 67772 | 6300 | 288 | 2988 | 77348 |
| 1987 | 69212 | 6143 | 576 | 1196 | 77127 |
| 1988 | 80472 | 5020 | 533 | 3964 | 89989 |
| 1989 | 51852 | 4140 | 373 | 685 | 57050 |
| 1990 | 63156 | 2407 | 382 | 687 | 66632 |
| 1991 | 49677 | 2140 | 292 | 4255 | 56364 |
| 1992 | 51464 | 3460 | 40 | 746 | 55710 |
| 1993 | 45890 | 2621 | 101 | 1738 | 50350 |
| 1994 | 38669 | 2274 | 129 | 1443 | 42515 |
| 1995 | 41516 | 2581 | 606 | 62 | 44765 |
| 1996 | 33558 | 2316 | 664 | 59 | 36597 |
| 1997 | 36342 | 2839 | 542 | 37 | 39761 |
| 1998 | 36771 | 2565 | 379 | 109 | 39825 |
| 1999 | 39824 | 1436 | 773 | 7 | 42040 |
| 2000 | 41187 | 1498 | 776 | 89 | 43550 |
| 2001 | 35067 | 1631 | 535 | 93 | 37326 |
| 2002 | 48570 | 1941 | 392 | 189 | 51092 |
| 2003 | 36577 | 1459 | 968 | 215 | 39220 |
| 2004 | 31686 | 1139 | 519 | 107 | 33451 |
| 2005 | 42593 | 2484 | 137 | 115 | 45329 |
| 2006 | 41521 | 656 | 0 | 34 | 42211 |
| 2007 | 38364 | 689 | 0 | 83 | 39134 |
| 2008 | 45538 | 569 | 64 | 80 | 46251 |
| 2009 | 38442 | 462 | 50 | 224 | 39177 |
| 2010 | 36155 | 620 | 220 | 1653 | 38648 |
| 2011 | 43773 | 493 | 83 | 1005 | 45354 |
| 2012 | 43089 | 491 | 41 | 2017 | 45635 |
| 2013 | 51330 | 372 | 92 | 1499 | 53263 |
| 20141) | 47769 | 201 | 60 | 2706 | 50736 |

[^7]Table 19.3.2 Golden redfish in Va. Observed catch in weight (tonnes) by age and years in 1995-2014. It should be noted that the catch-at-age results for 1996 are only based on three samples, which explains that there are no specimens older than 23 years.

| Year/ <br> Age | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7 | 62 | 0 | 33 | 24 | 7 | 40 | 122 | 129 | 201 | 226 | 235 | 187 | 136 | 464 | 109 | 60 | 143 | 71 | 56 | 151 |
| 8 | 374 | 360 | 230 | 284 | 350 | 65 | 138 | 900 | 211 | 845 | 779 | 1063 | 453 | 1279 | 979 | 356 | 559 | 585 | 625 | 467 |
| 9 | 1596 | 825 | 481 | 595 | 1623 | 849 | 394 | 759 | 1366 | 497 | 1917 | 2217 | 1760 | 2244 | 1756 | 2204 | 1561 | 1603 | 2395 | 1747 |
| 10 | 9436 | 3701 | 1039 | 1208 | 1259 | 4290 | 1620 | 833 | 1120 | 2098 | 1519 | 3721 | 2480 | 5173 | 3153 | 2710 | 4519 | 3271 | 3991 | 5620 |
| 11 | 2719 | 9127 | 2701 | 1129 | 1855 | 1888 | 7746 | 3155 | 1194 | 789 | 3120 | 2143 | 3356 | 4053 | 5069 | 2770 | 5453 | 6532 | 6015 | 6014 |
| 12 | 1319 | 2102 | 11572 | 3245 | 2523 | 2268 | 1802 | 10939 | 3945 | 975 | 1908 | 2837 | 1923 | 4721 | 4503 | 4893 | 4869 | 7322 | 9500 | 5786 |
| 13 | 3518 | 1317 | 2822 | 12501 | 2441 | 1686 | 1977 | 3046 | 9749 | 2020 | 1371 | 1640 | 3070 | 2285 | 3426 | 3873 | 6248 | 4034 | 6876 | 5863 |
| 14 | 5671 | 1477 | 1365 | 2077 | 15504 | 2346 | 1246 | 2580 | 2349 | 8594 | 3007 | 1300 | 1048 | 2758 | 1827 | 2727 | 3811 | 4948 | 4003 | 4925 |
| 15 | 5971 | 4347 | 3108 | 2026 | 1238 | 14677 | 835 | 1820 | 1958 | 2131 | 11771 | 2827 | 953 | 1491 | 1974 | 1371 | 2462 | 2896 | 4424 | 3195 |
| 16 | 1730 | 5456 | 3599 | 2392 | 1246 | 1744 | 11486 | 2938 | 1204 | 1675 | 2056 | 10097 | 2150 | 1056 | 1229 | 1192 | 1381 | 1310 | 3010 | 2744 |
| 17 | 852 | 934 | 2981 | 3376 | 1791 | 1167 | 512 | 11695 | 2223 | 804 | 1433 | 2063 | 9261 | 1800 | 664 | 814 | 915 | 781 | 1711 | 2012 |
| 18 | 368 | 379 | 877 | 2025 | 2606 | 1574 | 766 | 2038 | 6330 | 1366 | 1231 | 1154 | 1308 | 8032 | 1482 | 643 | 639 | 696 | 1190 | 1347 |
| 19 | 1134 | 259 | 620 | 1002 | 2183 | 2359 | 1021 | 1119 | 748 | 5129 | 1229 | 666 | 733 | 1464 | 6023 | 1081 | 802 | 389 | 757 | 501 |
| 20 | 1128 | 340 | 910 | 714 | 1236 | 2099 | 1683 | 626 | 402 | 1104 | 6331 | 946 | 713 | 876 | 938 | 4972 | 845 | 899 | 474 | 1247 |
| 21 | 503 | 1157 | 444 | 512 | 452 | 528 | 914 | 1360 | 593 | 331 | 386 | 5433 | 861 | 516 | 635 | 897 | 5156 | 709 | 516 | 531 |
| 22 | 644 | 988 | 511 | 389 | 210 | 435 | 400 | 983 | 773 | 482 | 457 | 597 | 4708 | 802 | 561 | 757 | 1162 | 3557 | 705 | 456 |
| 23 | 1427 | 791 | 651 | 416 | 325 | 266 | 400 | 703 | 737 | 605 | 765 | 221 | 718 | 4062 | 330 | 569 | 754 | 499 | 3171 | 517 |
| 24 | 647 | 0 | 564 | 652 | 214 | 62 | 156 | 357 | 375 | 556 | 598 | 365 | 111 | 363 | 2495 | 661 | 220 | 368 | 204 | 3261 |
| 25 | 745 | 0 | 711 | 510 | 821 | 384 | 119 | 281 | 292 | 250 | 410 | 452 | 595 | 241 | 96 | 2147 | 66 | 257 | 197 | 200 |
| 26 | 365 | 0 | 267 | 391 | 264 | 330 | 109 | 176 | 73 | 102 | 97 | 71 | 323 | 407 | 96 | 264 | 1589 | 217 | 170 | 214 |
| 27 | 350 | 0 | 134 | 420 | 597 | 192 | 264 | 79 | 80 | 178 | 264 | 248 | 341 | 329 | 189 | 383 | 86 | 1408 | 99 | 139 |
| 28 | 725 | 0 | 192 | 352 | 226 | 508 | 182 | 288 | 26 | 136 | 162 | 194 | 195 | 163 | 91 | 131 | 177 | 208 | 803 | 69 |
| 29 | 0 | 0 | 136 | 52 | 104 | 357 | 142 | 479 | 102 | 134 | 28 | 161 | 35 | 163 | 381 | 176 | 47 | 83 | 36 | 721 |
| 30+ | 232 | 0 | 394 | 480 | 747 | 1076 | 1033 | 1287 | 524 | 660 | 1520 | 916 | 1131 | 795 | 438 | 506 | 309 | 447 | 406 | 40 |
| Total | 41516 | 33560 | 36342 | 36772 | 39822 | 41190 | 35067 | 48570 | 36575 | 31687 | 42594 | 41519 | 38362 | 45537 | 38444 | 36157 | 43773 | 43090 | 51334 | 47767 |

Table 19.3.3 Results of the GLM model to calculate standardized CPUE for Icelandic golden redfish fishery in Va. Note that the residuals are shown in Fig. 19.3.9.

Call:
glm(formula $=$ lafli $\sim$ ltogtimi + factor(ar) + as.factor(veman) + factor(skipnr) + factor(reitur), family $=$ gaussian(), data $=t m p$ )

Deviance Residuals:
Min 1Q Median 3Q Max $-6.4139-0.4713 \quad 0.0315 \quad 0.5031 \quad 7.8381$

Coefficients:
Estimate Std. Error t value $\operatorname{Pr}(>|t|)$
(Intercept) $\quad 4.9173330 .876835 \quad 5.6082 .05 \mathrm{e}-08$ ***
ltogtimi $\quad 1.1235160 .003586313 .334<2 \mathrm{e}-16^{* * *}$
factor(ar)1979 $\begin{array}{llllll} & 0.041680 & 0.047178 & 0.883 & 0.376997\end{array}$
factor(ar)1980 0.1430950 .0444943 .2160 .001300 **
factor(ar)1981 0.1943870 .0438194 .436 9.17e-06 ***
factor(ar)1982 $0.139999 \quad 0.0435633 .2140 .001311$ **
$\begin{array}{lllll}\text { factor(ar)1983 } & 0.003101 & 0.042357 & 0.073 & 0.941638\end{array}$
factor(ar)1984 $-0.003777 \quad 0.043701-0.0860 .931123$
factor(ar)1985 $\quad 0.048689 \quad 0.044108 \quad 1.1040 .269653$
factor(ar)1986 $0.012195 \quad 0.043864 \quad 0.2780 .781005$
factor(ar)1987 $0.098549 \quad 0.044939 \quad 2.1930 .028314$ *
$\begin{array}{llllll}\text { factor(ar)1988 } & 0.047076 & 0.045351 & 1.038 & 0.299258\end{array}$
factor(ar)1989 $\begin{array}{lllll}0.069109 & 0.045381 & 1.523 & 0.127803\end{array}$
factor(ar)1990 $\quad 0.060588 \quad 0.045236 \quad 1.3390 .180459$
factor(ar)1991 $\quad 0.057138 \quad 0.040117 \quad 1.4240 .154374$
factor(ar)1992 $-0.136530 \quad 0.040346-3.3840 .000715$ ***
factor(ar)1993 $-0.254184 \quad 0.040047-6.3472 .21 \mathrm{e}-10^{* * *}$
factor(ar)1994 $-0.265108 \quad 0.040882-6.4858 .95 \mathrm{e}-11^{* * *}$
factor(ar)1995 $-0.234088 \quad 0.041133-5.6911 .27 \mathrm{e}-08$ ***
factor(ar)1996 $-0.225875 \quad 0.041519-5.4405 .34 \mathrm{e}-08$ ***
factor(ar)1997 $-0.224488 \quad 0.041658-5.3897 .12 \mathrm{e}-08$ ***
factor(ar)1998 $-0.150249 \quad 0.041890-3.5870 .000335$ ***
factor(ar)1999 $-0.195277 \quad 0.041412-4.7152 .42 \mathrm{e}-06^{* * *}$
factor(ar)2000 $\quad-0.049658 \quad 0.041488-1.197 \quad 0.231349$
factor(ar)2001 $0.1005190 .042527 \quad 2.3640 .018100$ *
factor(ar)2002 0.1268040 .0420643 .0150 .002574 **
$\begin{array}{llllll}\text { factor(ar)2003 } & 0.129320 & 0.043185 & 2.995 & 0.002750 \text { ** }\end{array}$
factor(ar)2004 $0.178750 \quad 0.043846 \quad 4.0774 .57 \mathrm{e}-05^{* * *}$
factor(ar)2005 $0.1132240 .042494 \quad 2.6640 .007713$ **
factor(ar)2006 $-0.032306 \quad 0.041827-0.7720 .439907$
$\begin{array}{llllll}\text { factor(ar)2007 } & 0.016620 & 0.042755 & 0.389 & 0.697474\end{array}$
factor(ar)2008 $\begin{array}{lllll}0.013847 & 0.042118 & 0.329 & 0.742324\end{array}$
$\begin{array}{llllll}\text { factor(ar)2009 } & 0.046115 & 0.042368 & 1.088 & 0.276408\end{array}$
$\begin{array}{llllll}\text { factor(ar)2010 } & 0.077359 & 0.042528 & 1.819 & 0.068917 .\end{array}$
factor(ar)2011 $0.308057 \quad 0.042833 \quad 7.1926 .45 \mathrm{e}-13^{* * *}$
factor(ar)2012 $0.4694630 .04342310 .811<2 \mathrm{e}-16^{* * *}$
factor(ar)2013 $0.5568980 .04333512 .851<2 \mathrm{e}-16^{* * *}$
factor(ar)2014 $0.5231930 .04372911 .965<2 \mathrm{e}-16^{* * *}$
as.factor(veman)2 $0.1445510 .017712 \quad 8.1613 .38 \mathrm{e}-16^{* * *}$
as.factor(veman)3 $0.3454410 .01698820 .335<2 \mathrm{e}-16^{* * *}$
as.factor(veman)4 $0.327880 \quad 0.017528 \quad 18.706<2 \mathrm{e}-16^{* * *}$
as.factor(veman)5 $0.1657370 .0199848 .294<2 \mathrm{e}-16^{* * *}$
as.factor(veman)6 $0.3625720 .01832319 .787<2 \mathrm{e}-16^{* * *}$
as.factor(veman) $70.327200 \quad 0.01767418 .513<2 \mathrm{e}-16^{* * *}$

```
as.factor(veman)8 0.245485 0.017707 13.864<2e-16***
as.factor(veman)9 0.161327 0.017198 9.381<2e-16***
as.factor(veman)10 0.101764 0.017231 5.906 3.52e-09 ***
as.factor(veman)11 0.037417 0.018019 2.077 0.037844*
as.factor(veman)12-0.075603 0.019886 -3.802 0.000144***
```

(Dispersion parameter for gaussian family taken to be 0.7377318 )
Null deviance: 159817 on 62727 degrees of freedom
Residual deviance: 45975 on 62319 degrees of freedom AIC: 159344
Number of Fisher Scoring iterations: 2

Analysis of Deviance Table
Model: gaussian, link: identity
Response: lafli
Terms added sequentially (first to last)
Df Deviance Resid. Df Resid. Dev $\quad$ F $\quad \operatorname{Pr}(>F)$
NULL 62727159817
ltogtimi $1 \begin{array}{llll}96166 & 62726 & 63651 & 130353.525<2.2 e-16 ~ * * * ~\end{array}$
factor(ar) $36 \quad 4115 \quad 62690 \quad 59536 \quad 154.955<2.2 \mathrm{e}-16^{* * *}$ as.factor(veman) $11 \quad 11756267958361 \quad 144.740<2.2 \mathrm{e}-16^{* * *}$
factor(skipnr) $206917462473 \quad 49187 \quad 60.367<2.2 \mathrm{e}-16^{* * *}$
factor(reitur) $1543212 \quad 62319 \quad 45975 \quad 28.275<2.2 \mathrm{e}-16^{* * *}$

Table 19.4.1 Results from the Gadget model of total biomass, spawning stock biomass, recruitment at age 5 , catch and fishing mortality, projections are in italic.

| Year | Biomass | SSB | R(age5) | Catches | F9-19 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 605.4 | 369.8 | 202.9 | 67.9 | 0.100 |
| 1972 | 607.7 | 365.2 | 184.8 | 50.9 | 0.077 |
| 1973 | 650.3 | 374.2 | 443.3 | 43.7 | 0.066 |
| 1974 | 683.4 | 388.4 | 210.5 | 50.6 | 0.073 |
| 1975 | 701.9 | 398.4 | 124.1 | 61.9 | 0.088 |
| 1976 | 707.0 | 396.1 | 207.8 | 94.4 | 0.134 |
| 1977 | 716.7 | 400.1 | 196.6 | 53.8 | 0.079 |
| 1978 | 743.8 | 423.7 | 129.2 | 48.7 | 0.066 |
| 1979 | 761.2 | 440.5 | 161.7 | 77.2 | 0.099 |
| 1980 | 751.2 | 441.8 | 105.4 | 89.1 | 0.114 |
| 1981 | 721.6 | 431.8 | 75.3 | 102.0 | 0.135 |
| 1982 | 664.4 | 402.6 | 63.4 | 130.3 | 0.184 |
| 1983 | 598.9 | 366.0 | 67.6 | 106.0 | 0.162 |
| 1984 | 546.3 | 337.0 | 73.4 | 95.3 | 0.154 |
| 1985 | 509.2 | 313.8 | 131.7 | 78.5 | 0.132 |
| 1986 | 479.2 | 294.1 | 122.1 | 76.9 | 0.140 |
| 1987 | 443.3 | 271.8 | 64.7 | 76.6 | 0.152 |
| 1988 | 395.7 | 241.3 | 41.4 | 89.8 | 0.205 |
| 1989 | 355.1 | 215.1 | 44.9 | 56.6 | 0.145 |
| 1990 | 355.3 | 199.3 | 352.7 | 66.3 | 0.192 |
| 1991 | 334.2 | 182.4 | 59.0 | 56.0 | 0.179 |
| 1992 | 315.4 | 168.9 | 40.1 | 55.8 | 0.196 |
| 1993 | 299.1 | 157.7 | 54.8 | 50.2 | 0.194 |
| 1994 | 288.9 | 151.9 | 64.8 | 42.5 | 0.172 |
| 1995 | 308.1 | 151.6 | 334.8 | 44.3 | 0.182 |
| 1996 | 313.9 | 154.1 | 89.9 | 35.6 | 0.144 |
| 1997 | 313.7 | 156.1 | 41.3 | 39.0 | 0.153 |
| 1998 | 315.5 | 161.0 | 41.7 | 39.7 | 0.153 |
| 1999 | 313.4 | 162.2 | 86.9 | 42.5 | 0.162 |
| 2000 | 308.6 | 164.1 | 50.9 | 42.6 | 0.158 |
| 2001 | 315.4 | 168.3 | 115.5 | 36.7 | 0.131 |
| 2002 | 319.0 | 169.4 | 125.7 | 50.7 | 0.178 |
| 2003 | 335.4 | 173.4 | 193.6 | 38.2 | 0.134 |
| 2004 | 353.3 | 184.7 | 112.6 | 32.8 | 0.111 |
| 2005 | 375.4 | 193.7 | 183.9 | 46.6 | 0.154 |
| 2006 | 402.1 | 204.3 | 197.0 | 42.1 | 0.140 |
| 2007 | 418.2 | 216.3 | 107.1 | 39.2 | 0.125 |
| 2008 | 443.3 | 233.9 | 134.5 | 46.2 | 0.139 |
| 2009 | 476.2 | 251.3 | 218.4 | 39.3 | 0.110 |
| 2010 | 510.0 | 276.5 | 142.7 | 38.5 | 0.098 |
| 2011 | 525.5 | 298.7 | 42.4 | 45.1 | 0.106 |
| 2012 | 531.8 | 312.4 | 79.9 | 45.2 | 0.100 |
| 2013 | 544.7 | 327.8 | 120.0 | 53.1 | 0.111 |
| 2014 | 547.8 | 335.4 | 120.0 | 50.6 | 0.102 |
| 2015 | 557.2 | 346.7 | 120.0 | 49.1 | 0.096 |
| 2016 | 562.8 | 354.2 | 120.0 | 51.0 | 0.098 |
| 2017 | 567.0 | 359.8 | 120.0 | 51.6 | 0.098 |
| 2018 | 570.2 | 363.9 | 120.0 | 51.9 | 0.098 |
| 2019 | 572.9 | 366.9 | 120.0 | 52.1 | 0.098 |

Table 19.4.2 Output from short term prognosis. Multiplier is based on reference to the adopted HCR For-9 $^{9}=0.097$. Biomasses are in the beginning of the year to apply to ICES standard in short term prognosis in other places in the report they are in the middle of the year. NOT UPDATED
$F(2014)=0.101 C(2014)=47.300$ tons.

| $\mathbf{2 0 1 5}$ |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
| Bio 5+ | SSB | Fmult | F9-19 | Landings |
| 507 | 342 | 1.043 | 0.101 | 48.5 |


|  |  | 2015 |  | 2016 |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Fmult | F9-19 | Bio $5+$ | SSB | Landings | Bio $5^{+}$ | SSB |
| 0.0 | 0 | 516 | 352 | 0 | 574 | 404 |
| 0.1 | 0.01 | 516 | 352 | 4.9 | 569 | 399 |
| 0.2 | 0.019 | 516 | 352 | 9.7 | 564 | 395 |
| 0.3 | 0.029 | 516 | 352 | 14.5 | 559 | 391 |
| 0.4 | 0.038 | 516 | 352 | 19.3 | 554 | 387 |
| 0.5 | 0.048 | 516 | 352 | 24 | 549 | 382 |
| 0.6 | 0.058 | 516 | 352 | 28.7 | 544 | 378 |
| 0.7 | 0.067 | 516 | 352 | 33.4 | 539 | 374 |
| 0.8 | 0.077 | 516 | 352 | 38 | 535 | 370 |
| 0.9 | 0.087 | 516 | 352 | 42.7 | 530 | 366 |
| 1.0 | 0.097 | 516 | 352 | 47.3 | 525 | 362 |
| 1.1 | 0.107 | 516 | 352 | 51.8 | 520 | 358 |
| 1.2 | 0.117 | 516 | 352 | 56.4 | 516 | 354 |
| 1.3 | 0.127 | 516 | 352 | 60.9 | 511 | 350 |
| 1.4 | 0.137 | 516 | 352 | 65.4 | 506 | 346 |
| 1.5 | 0.147 | 516 | 352 | 69.8 | 502 | 342 |
| 1.6 | 0.158 | 516 | 352 | 74.3 | 497 | 338 |
| 1.7 | 0.168 | 516 | 352 | 78.7 | 493 | 334 |
| 1.8 | 0.178 | 516 | 352 | 83 | 488 | 330 |
| 1.9 | 0.189 | 516 | 352 | 87.4 | 484 | 326 |
| 2.0 | 0.199 | 516 | 352 | 91.7 | 479 | 322 |



Figure 19.2.1 Indices of golden redfish from the groundfish surveys in March 1985-2015 (line, shaded area) and October 1996-2014 (red and vertical lines). a) Total biomass; b) biomass of fish larger than 32 cm ; c) biomass of fish larger than 40 cm ; d) indices of juvenile golden redfish (4-11) cm in millions. The shaded area and the vertical bar show $\pm 1$ standard error of the estimate.


Figure 19.2.2. Length disaggregated abundance indices of golden redfish from the bottom trawl survey in March 1985-2015 conducted in Icelandic waters. The black line is the mean of total indices 1985-2015.


Figure 19.2.3. Length disaggregated abundance indices of golden redfish from the bottom trawl survey in October 1996-2014 conducted in Icelandic waters. The black line is the mean of total indices 1996-2014. The survey was not conducted in 2011.


Figure 19.2.4 Age disaggregated abundance indices of golden redfish in the bottom trawl survey in October conducted in Icelandic waters 1996-2014. The survey was not conducted in 2011.


Figure 19.2.5 CPUE of golden redfish in the Faeroes spring groundfish survey 1994-2015 and the summer groundfish survey 1996-2014 in ICES Division Vb.


Figure 19.2.6 Golden redfish ( $>17 \mathrm{~cm}$ ). Survey abundance indices for East and West Greenland from the German groundfish survey 1985-2013. a) Total biomass index, b) total abundance index, c) biomass index divided by size classes (17-30 cm and > 30 cm ).


Figure 19.2.7 Golden redfish ( $>17 \mathrm{~cm}$ ). Length frequencies for East and West Greenland 1982-2014.


Figure 19.3.1 Nominal landings of golden redfish in tonnes by ICES Divisions 1978-2014. Landings statistics for 2014 are provisional.


Figure 19.3.2 Geographical distribution of golden redfish bottom trawl catches in Division Va 2001 2014.


Figure 19.3.3 Length distribution (gray shaded area) of golden redfish in the commercial landings of the Icelandic bottom trawl fleet 1975-2014. The blue line is the mean of the years 1975-2014.


Figure 19.3.4 Catch-at-age of golden redfish in numbers in ICES Subdivision Va 1995-2014.


Figure 19.3.5 Catch curve of golden redfish based on the catch-at-age data in ICES Division Va 19952014.


Figure 19.3.6 Length distribution of golden redfish from Faroese catches in 2001-2014.


Figure 19.3.7 CPUE of golden redfish from Icelandic trawlers based on results from the GLM model 1978-2014 where golden redfish catch composed at least $50 \%$ of the total catch in each haul. The figure shows the raw CPUE index (sum(yield)/sum(effort)), standardized CPUE index estimated using a generalized linear model, and effort (blue dotted line).


Figure 19.3.8 Results from the GLM modle (section 8.2.1) for the CPUE series of golden redfish in Va. From left to right, top to bottom: Residuals against fitted values; square root of the absolute value of residuals against predicted values; response against fitted values; normal QQplot of standardized residuals.


Figure 19.4.1 Stations in the German survey in East Greenland with an area used to compile the indices for Gadget shown. This area corresponds to giving a weight of 0.5 to the results in Figure 19.2.7.


Figure 19.4.2 Biomass index from Iceland (blue) and Greenland black, based on weighting the German survey data in Figure 19.2.7 by 0.5.


Figure 19.4.3. Indices from the Icelandic March survey (red) and Icelandic March survey plus German survey in Greenland (blue) by length group.


Figure 19.4.4. Development of SSB from run where certain components of the likelihood function weighted much more than the other components.


Figure 19.4.5. Summary from the assessment. Red values are predictions. Spawning stock is compiled using a fixed maturity ogive with $L 50=33 \mathrm{~cm}$.


Figure 19.4.6. Comparison of the current assessment and the same assessment done in 2014.


Figure 19.4.7. Estimates of selection curves from commercial catches (upper panel) and from the Icelandic March survey. The black line is the estimated selection curve fitted to the length distributional data (Figure 19.4.14) and the red line is the estimated $q$ from the disaggregated tuning indices, scaled to one.


Figure 19.4.8. Comparison of observed and predicted survey biomass from the 2014 (red line) and 2015 (blue) runs.


Figure 19.4.9. Residuals from the fit between model and survey indices. The red circles indicate positive residuals (survey results exceed model prediction). Largest residuals correspond to $\log (\mathrm{obs} / \mathrm{mod})=1$


Figure 19.4.10. Fit to length disaggregated survey indices from Gadget run as XY-scatter. The red line is fitted going through the 0 -point, the green cross goes over the terminal year.


Figure 19.4.11. Fit (red lines) to length disaggregated survey indices (broken lines and points) from Gadget run as time series.


Figure 19.4.12. Fit (red line) to Icelandic commercial length distributions aggregated by 3 years.


Figure 19.4.12. Fit to survey age data (run 1). Bars represent the data and red lines the fit. The likelihood data are used in the model as proportions in each 2 cm length group but presented here as total for each age group something that should only be comparable if catchability was independent of size (age).


Figure 19.4.13. Predicted (red) and observed (blue) age distributions from Icelandic commercial fishery.


Figure 19.4.14 Catch in numbers by age for year-classes 1995-1999 plotted on $\log$ scale. The grey lines correspond to $\mathrm{Z}=0.24$ that is the calculated average Z for age $12+$ for these year-classes.


Figure 19.4.15 Estimated. Biomass of ages 9 - 19 from the Gadget and TSA assessments. All biomasses compiled based on the calculated mean weight at age in the catches.


Figure 19.4.15 Estimated average fishing mortality of age $9-19$ from the Gadget and TSA assessments.

## 20 Icelandic slope Sebastes mentel/a in Va and XIV

## Executive summary

ICES concluded in February 2009 that S. mentella is to be divided to three biological stocks and that the S. mentella on the continental shelf and slope of Iceland should be treated as a separate biological stock and management unit.

Total landings of demersal S. mentella in Icelandic waters in 2014 were about 9500 t , 750 t more than in 2013.

No analytical assessment was conducted and there are no biological reference points for the species. Survey indices from the annual autumn survey since 2000 are used as basis for advice.

Available survey biomass indices show that in Division Va the biomass has gradually decreased from 2006 to 2013, but increased in 2014.

The East-Greenland shelf is most likely a nursery area for the stock. No new recruits $(>18 \mathrm{~cm})$ are seen in the survey catches of the German survey and the Greenland shrimp and fish shallow water survey conducted in the area and no juveniles are present (<18 $\mathrm{cm})$ recent years.

Icelandic slope S. mentella is considered a data limited stock (DLS) and follows the ICES framework for such (Category 3.2). When the precautionary approach is applied, catches in 2016 should be no more than 9954 t . All catch are assumed to be landed.

### 20.1 Stock description and management units

The stock structure of S. mentella in the Irminger Sea and adjacent water is described in Chapter 18 and Stock Annex. The S. mentella on the continental shelf and slope of Iceland is treated as separate biological stock and management unit. Only the fishable stock of Icelandic slope S. mentella is found in Icelandic waters, i.e. mainly fish larger than 30 cm . The East-Greenland shelf is most likely a common nursery area for the three biological stocks described in Stock Annex, including the Icelandic slope one.

### 20.2 Scientific data

Only the fishable stock of Icelandic slope S. mentella is found in Icelandic waters. The Icelandic autumn survey on the continental shelf and slope in Division Va, covering depths down to 1500 m , does, therefore, not cover the whole distribution of the stock. Data for Icelandic slope S. mentella from the Autumn Survey is available from 20002014. No survey was conducted in 2011. A description of the autumn survey is given in Stock Annex for the species.

The total biomass index and the abundance indices from the autumn survey were highest in 2001. After a decrease in 2003 the index increased again in 2006 but gradually decreased until 2013 and to similar level as in 2003 when it was lowest in the time series (Table 20.2.1 and Figure 20.2.1 $a$ and $b$ ). The biomass index increased again in 2014 to similar level as in 2004 (Table 20.2.1 and Figure 20.2.1 $a$ and $b$ ). The biomass index of fish larger than 45 cm was at lowest level in 2007 but increased again in 2009 were it was at similar level until 2013 (Figure 19.2.1c). Sharp increase in biomass index of 45 cm and larger fish in 2014 and was at the highest in the time series.

The abundance index of fish smaller than 30 cm has in 2007-2014 been at lowest level (Figure 20.2.2d). The length of the Icelandic slope $S$. mentella in the autumn survey is
between 25 and more than 50 cm . Since 2000, the mode has shifted to the right, that is, from $36-39 \mathrm{~cm}$ in 2000 to about $42-43 \mathrm{~cm}$ in 2012-2014 (Figure 20.2.2). Very little Icelandic slope S. mentella smaller than 35 cm was observed in the 2014 survey.
Otoliths have been sampled since 2000 and otoliths from the 2000, 2009 and 2010 surveys have been age read. Figure 20.2.3 shows that the 1985 and the 1990 year classes are the most abundant ones in this samples.

### 20.3 Information from the fishing industry

### 20.3.1 Landings

Total annual landings of Icelandic slope S. mentella from ICES Division Va 1978-2014 are presented in Table 20.3.1 and from 1950-2014 in Figure 20.3.1. Annual landings gradually decreased from a record high of 57000 t in 1994 to 17000 t in 2001. Landings in 2001-2010 fluctuated between 17000 t and 20500 except in 2003 and 2008 when annual landings were 28500 t and 24000 respectively. The landings in 2014 were about 9 $500 \mathrm{t}, 700 \mathrm{t}$ more than 2013. The decrease is related to lower TAC for the species.

### 20.3.2 Fisheries and fleets

Most of the fishery for Icelandic slope S. mentella in Va is a directed bottom trawl fishery taken by bottom trawlers along the shelf and slope west, southwest, and southeast of Iceland at depths between 500 and 800 m (Figure 20.3.2). The proportion of Icelandic slope S. mentella catches taken by pelagic trawls 1991-2000 varied between 10 and $44 \%$ of the total landings (Table 20.3.2). In 2001-2014, no pelagic fishery occurred or it was negligible except in 2003 and 2007 (see Stock Annex). In general, the pelagic fishery was mainly in the same areas as the bottom trawl fishery (Figure 20.3.3), but usually in later months of the year (Figure 19.3.4). The bottom trawl catches in the third and fourth quarter of the year decreased considerable in 2001-2007 compared with earlier years but increased again in 2008-2012 (Figure 20.3.4).

A notable change in the catch pattern is that catches taken in the southeast fishing area has been gradually decreasing since 2000 and in recent years very little Icelandic slope S. mentella was taken on these fishing grounds (Figure 20.3.2). This area has historically been an important fishing area for Icelandic slope S. mentella.

### 20.3.3 Sampling from the commercial fishery

The table below shows the 2014 biological sampling from the catch and landings of Icelandic slope S. mentella in ICES Division Va. This is considered to be adequate sampling from the fishery. Otoliths from the commercial catch have been collected, but no systematic age reading is done.

| Year | Nation | Gear | Landings (t) | No. samples | No. length <br> measured |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Va | Iceland | Bottom trawl | 9500 | 84 | 14490 |

### 20.3.4 Length distribution from the commercial catch

Length distributions of Icelandic slope S. mentella in Va from the bottom trawl fishery show an increase in the number of small fish in the catch in 1994 compared to previous years (Figure 20.3.5). The peak of about 32 cm in 1994 can be followed by approximately 1 cm annual growth in 1996-2002. The fish caught in 2004-2014 peaked
around $39-42 \mathrm{~cm}$. The length distribution of Icelandic slope S. mentella from the pelagic fishery, where available, showed that in most years the fish was on average bigger than taken in the bottom trawl fishery (Figure 20.3.5).

### 20.3.5 Catch per unit effort

Trends in both standardized (glm) and raw CPUE and effort are shown in Figure 20.3.6. CPUE gradually decreased from 1978 to a record low in 1994, but has since then slightly increased annually to 2000. The CPUE estimate in 2014 was at similar level as in late 1980s and about $40 \%$ higher than it was in 1994. The CPUE has been stable since 2010. From 1991 to 1994, when CPUE decreased, the fishing effort increased drastically. Since then, effort decreased and is now at similar level as in the early 1980s. Output of the model is given in Table 20.3.3 and the model residuals in Figure 20.3.7.

### 20.3.6 Discard

Although no direct measurements are available on discards, it is believed that there are no significant discards of Icelandic slope S. mentella in the Icelandic redfish fishery.

### 20.4 Methods

No analytical assessment was conducted on this stock.

### 20.5 Reference points

There are no biological reference points for the species. Previous reference points established were based upon commercial CPUE indices, but are now considered to be unreliable indicators of stock size. ICES has withdrawn these reference points.

Icelandic slope beaked redfish in ICES Division Va has previously been assessed based on trends in survey biomass indices from the Icelandic Autumn survey or in ICES "trends based assessment". Supplementary data used in the assessment includes information from the fishery and length distributions from the commercial catch and the Autumn Survey. ICES advised in 2013, based on DLS approach (Method 3.2), that catches are set no higher than 9875 t in 2014. Same advice was applied for the 2015 fishing year. The TAC set by the Icelandic government was 10000 t in 2014 and 2015.

### 20.6 State of the stock

The Group concludes that the state of the stock is on a low level. With the information at hand, current exploitation rates cannot be evaluated for the Icelandic slope S. mentella in Division Va.

The fishable biomass index of Icelandic slope S. mentella from the Icelandic autumn survey shows that the biomass index for 2004-2013 has decreased to similar level as in 2003 when it was at lowest level, but increased again in 2014. The survey was not conducted in 2011. Standardised CPUE indices show a reduction from highs in the late 1980s, but there is an indication that the stock has started a slow recovery since the middle of 1990s, when CPUE was close to $50 \%$ of the maximum. The CPUE index has been stable since 2000 .

In 2000-2008, good recruitment was been observed in the German survey on the East Greenland shelf (growth of about $2 \mathrm{~cm} / \mathrm{yr}$ ) which is assumed to contribute to both the Icelandic slope and pelagic stock at unknown shares. The German survey and the Greenland shrimp and fish shallow water survey both show no new recruits ( $>18 \mathrm{~cm}$ )
and no juveniles are present ( $<18 \mathrm{~cm}$ ). This suggests that the fishery in coming years will be based on the same cohorts.

### 20.7 Management considerations

S. mentella is a slow growing, late maturing deep-sea species and is therefore considered vulnerable to overexploitation and advice has to be conservative.

The CPUE has slightly increased annually since a record low in 1994, especially in recent 3-4 years and is now $40 \%$ higher than in 1994. It is, however, not known to what extent CPUE series reflect change in stock status of Icelandic slope S. mentella. The nature of the redfish fishery is targeting schools of fish using advancing technology. The effect of technological advances is to increase CPUE, but is unlikely to reflect biomass increase.

The advice for 2008-2012 was that a management plan to be developed and implemented which takes into account the uncertainties in science and the properties of the fisheries. ICES suggested that catches of S. mentella are set no higher than 10000 t as a starting point for the adaptive part of the management plan. The advice for 2014 and 2015 were 9875 t based on the DLS approach (Category 3.2).

The Icelandic slope S. mentella fishery southeast of Iceland has gradually ceased since 2000 and very little fishing is conducted in this area. This fishing area was prior to 2000 very important fishing area for Icelandic slope $S$. mentella.

The landings increased in Division Va between 2002 and 2003 by about 10000 t when the fishery of pelagic S. mentella merged with the Icelandic slope fishery at the redfish line. Those two fisheries merged again in 2007.

There are no explicit management for Icelandic slope S. mentella but the species is within the TAC system described in Chapter 7.5. Icelandic authorities gave until the 2010/2011 a joint quota for golden redfish and Icelandic slope S. mentella in Icelandic waters, but now give separate quotas for the species.

### 20.8 Basis for advice

Icelandic slope S. mentella is considered a data limited stock (DLS) and should follow the ICES framework for such (Category 3.2). Below is the description of the formulation of the advice for the 2016 fishing year.

Based on the North Western Working Group recommendation, the stock is treated as a stock with survey data, but no proxies for MSY $B_{\text {trigger }}$ or F values, are known. This means that the catch advice for 2015 is based on the survey adjusted status quo catch equation:

$$
C_{y+1}=C_{y-1}\left(\frac{\sum_{i=y-x}^{y-1} I_{i} / x}{\sum_{i=y-z}^{y-x-1} I_{i} /(z-x)}\right)
$$

Where $I$ is the survey index, $x$ is the number of years in the survey average, $\mathrm{z}=5$ and $\mathrm{C}_{\mathrm{y}}$ 1 is the advice last year. The biomass is estimated to have increased by $0.8 \%$ between 2009-2012 (average of the three years, no survey conducted in 2011) and 2013 and 2014 (average of the two years). This implies an increase of catches of at most $0.8 \%$ in relation to the last year advise $(9875 \mathrm{t}$ ), corresponding to catch of no more than $9,954 \mathrm{t}$. A precautionary buffer of $20 \%$ consistent with the ICES approach is not applied as it was applied the first time DLS was used in 2013.

### 20.9 Regulation and their effects

There are no explicit management for Icelandic slope $S$. mentella. The species is managed under the ITQ system (see Chapter 7.5.1). Icelandic authorities gave until the 2010/2011 fishing year a joint quota for golden redfish (S. marinus) and Icelandic slope S. mentella. The separation of quotas was implemented in the fishing year that started September 1, 2010.

A general description of management and regulation of fish populations in Icelandic waters is given in Chapter 7.5 and in Stock Annex A. 2 with emphasis on Icelandic slope S. mentella where applicable.

Table 20.2.1 Total biomass index ofIcelandic slope S. mentella in the Icelandic Autumn Groundfish survey 2000-2014. No survey was conducted in 2011.

| Year | Iceland | cv |
| :--- | :--- | :--- |
| 2000 | 138924 | 0.145 |
| 2001 | 164030 | 0.172 |
| 2002 | 96923 | 0.137 |
| 2003 | 64621 | 0.127 |
| 2004 | 98373 | 0.164 |
| 2005 | 114953 | 0.249 |
| 2006 | 124509 | 0.172 |
| 2007 | 85469 | 0.183 |
| 2008 | 82703 | 0.139 |
| 2009 | 99767 | 0.183 |
| 2010 | 81963 | 0.149 |
| 2011 |  |  |
| 2012 | 78016 | 0.144 |
| 2013 | 70250 | 0.139 |
| 2014 | 104307 | 0.185 |

Table 20.3.1 Nominal landings (in tonnes) of Icelandic slope S. mentella 1978-2014 ICES Division Va.

| Year | Iceland | Others | Total |
| :--- | :--- | :--- | :--- |
| 1978 | 3693 | 209 | 3902 |
| 1979 | 7448 | 246 | 7694 |
| 1980 | 9849 | 348 | 10197 |
| 1981 | 19242 | 447 | 19689 |
| 1982 | 18279 | 213 | 18492 |
| 1983 | 36585 | 530 | 37115 |
| 1984 | 24271 | 24580 | 18750 |

1) Provisional

Table 20.3.2 Proportion of the landings of Icelandic slope S. mentella taken in ICES Division Va by pelagic and bottom trawls 1991-2014.

| Year | Pelagic trawl | Bottom trawl |
| :--- | :--- | :--- |
| 1991 | $22 \%$ | $78 \%$ |
| 1992 | $27 \%$ | $73 \%$ |
| 1993 | $32 \%$ | $68 \%$ |
| 1994 | $44 \%$ | $56 \%$ |
| 1995 | $36 \%$ | $64 \%$ |
| 1996 | $31 \%$ | $69 \%$ |
| 1997 | $11 \%$ | $89 \%$ |
| 1998 | $37 \%$ | $63 \%$ |
| 1999 | $10 \%$ | $90 \%$ |
| 2000 | $24 \%$ | $76 \%$ |
| 2001 | $3 \%$ | $97 \%$ |
| 2002 | $3 \%$ | $97 \%$ |
| 2003 | $28 \%$ | $72 \%$ |
| 2004 | $0 \%$ | $100 \%$ |
| 2005 | $0 \%$ | $100 \%$ |
| 2006 | $0 \%$ | $100 \%$ |
| 2007 | $17 \%$ | $83 \%$ |
| 2008 | $0 \%$ | $100 \%$ |
| 2009 | $0 \%$ | $100 \%$ |
| 2010 | $0 \%$ | $100 \%$ |
| 2011 | $0 \%$ | $100 \%$ |
| 2012 | $0 \%$ | $100 \%$ |
| 2013 | $0 \%$ | $100 \%$ |
| 2014 | $0 \%$ | $100 \%$ |
|  |  |  |

Table 20.3.3 Results of the GLM model to calculate standardized CPUE for Icelandic slope redfish fishery in Va. Note that the residuals are shown in Figure 18.3.8.

```
Call:glm(formula = lafli ~ ltogtimi + factor(ar) + as.factor(veman) +
factor(skipnr) + factor(reitur), family = gaussian(), data = tmp)
DevianceResiduals:
\begin{tabular}{lllll} 
Min & \(1 Q\) & Median & 3Q & Max \\
-5.0420 & -0.3340 & 0.0145 & 0.3499 & 4.7053
\end{tabular}
```

Coefficients:
EstimateStd. Error $t$ valuePr (>|t|)

| (Intercept) | 7.815942 | 0.635480 | 12.299 | $<2 \mathrm{e}-16$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| ltogtimi | 1.128215 | 0.003627 | 311.058 | < 2e-16 |  |
| factor(ar) 1979 | 0.048608 | 0.075981 | 0.640 | 0.522341 |  |
| factor(ar) 1980 | 0.162845 | 0.070851 | 2.298 | 0.021543 | * |
| factor(ar) 1981 | 0.055250 | 0.071480 | 0.773 | 0.439557 |  |
| factor (ar) 1982 | 0.117208 | 0.067764 | 1.730 | 0.083703 |  |
| factor (ar) 1983 | -0.016378 | 0.065831 | -0.249 | 0.803531 |  |
| factor(ar) 1984 | 0.002665 | 0.066417 | 0.040 | 0.967989 |  |
| factor (ar) 1985 | -0.036672 | 0.066483 | -0.552 | 0.581230 |  |
| factor (ar) 1986 | -0.009475 | 0.066984 | -0.141 | 0.887520 |  |
| factor (ar) 1987 | 0.068589 | 0.067871 | 1.011 | 0.312222 |  |
| factor (ar) 1988 | -0.005756 | 0.067071 | -0.086 | 0.931613 |  |
| factor (ar) 1989 | -0.052207 | 0.066528 | -0.785 | 0.432612 |  |
| factor (ar) 1990 | -0.103960 | 0.064947 | -1.601 | 0.109455 |  |
| factor (ar) 1991 | -0.071487 | 0.062601 | -1.142 | 0.253487 |  |
| factor(ar) 1992 | -0.325045 | 0.062333 | -5.215 | 1.85e-07 |  |
| factor(ar) 1993 | -0.416666 | 0.062303 | -6.688 | $2.30 \mathrm{e}-11$ |  |
| factor(ar) 1994 | -0.532328 | 0.062345 | -8.538 | < 2e-16 |  |
| factor (ar) 1995 | -0.497444 | 0.062501 | -7.959 | $1.79 \mathrm{e}-15$ |  |
| factor (ar) 1996 | -0.480674 | 0.062892 | -7.643 | $2.18 \mathrm{e}-14$ |  |
| factor (ar) 1997 | -0.419092 | 0.062762 | -6.677 | $2.47 \mathrm{e}-11$ |  |
| factor (ar) 1998 | -0.423970 | 0.063917 | -6.633 | $3.34 \mathrm{e}-11$ |  |
| factor (ar) 1999 | -0.370873 | 0.063400 | -5.850 | $4.97 e-09$ |  |
| factor (ar) 2000 | -0.316698 | 0.063806 | -4.963 | $6.96 e-07$ |  |
| factor (ar) 2001 | -0.315265 | 0.064868 | -4.860 | $1.18 \mathrm{e}-06$ |  |
| factor (ar) 2002 | -0.353717 | 0.064062 | -5.521 | 3.39e-08 |  |
| factor (ar) 2003 | -0.278988 | 0.064086 | -4.353 | $1.34 e-05$ |  |
| factor (ar) 2004 | -0.350058 | 0.064470 | -5.430 | 5.68e-08 |  |
| factor(ar) 2005 | -0.353155 | 0.063799 | -5.535 | $3.13 e-08$ |  |
| factor(ar) 2006 | -0.360114 | 0.064295 | -5.601 | $2.15 e-08$ |  |
| factor (ar) 2007 | -0.355708 | 0.065965 | -5.392 | $7.00 \mathrm{e}-08$ |  |
| factor (ar) 2008 | -0.280466 | 0.064961 | -4.317 | 1.58e-05 |  |
| factor (ar) 2009 | -0.324753 | 0.065550 | -4.954 | $7.30 e-07$ |  |
| factor(ar) 2010 | -0.297453 | 0.065913 | -4.513 | 6.42e-06 |  |
| factor (ar) 2011 | -0.188108 | 0.066109 | -2.845 | 0.004438 |  |
| factor (ar) 2012 | -0.292790 | 0.066402 | -4.409 | $1.04 \mathrm{e}-05$ |  |
| factor (ar) 2013 | -0.264235 | 0.067864 | -3.894 | $9.90 \mathrm{e}-05$ |  |
| factor (ar) 2014 | -0.271598 | 0.067844 | -4.003 | $6.26 e-05$ |  |
| as.factor (veman) 2 | 0.119721 | 0.014585 | 8.208 | $2.32 e-16$ |  |
| as.factor (veman) 3 | 0.134197 | 0.015365 | 8.734 | $<2 e-16$ |  |
| as.factor (veman) 4 | 0.114174 | 0.015363 | 7.432 | 1.10e-13 |  |
| as.factor (veman) 5 | 0.024324 | 0.016765 | 1.451 | 0.146817 |  |
| as.factor (veman) 6 | -0.007176 | 0.019157 | -0.375 | 0.707973 |  |
| as.factor (veman) 7 | -0.107690 | 0.019793 | -5.441 | $5.34 \mathrm{e}-08$ |  |
| as.factor (veman) 8 | -0.098225 | 0.019047 | -5.157 | $2.52 e-07$ |  |
| as.factor (veman) 9 | -0.050757 | 0.016390 | -3.097 | 0.001957 |  |
| as.factor (veman) 10 | -0.040110 | 0.015276 | -2.626 | 0.008650 |  |
| as.factor (veman) 11 | -0.078574 | 0.015573 | -5.046 | 4.55e-07 |  |
|  | -0.128913 | 0.016345 | -7.887 |  |  |

```
Analysis of DevianceTable
Model: gaussian, link: identity
Response: lafli
Termsaddedsequentially (first to last)
```

| DfDevianceResid. | Df | Dev | F | $\operatorname{Pr}(>\mathrm{F})$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| NULL |  |  | 33222 | 60961 |  |  |  |
| ltogtimi | 1 | 43574 | 33221 | 17387 | 111790.442 | $<2.2 e-16$ | *** |
| factor(ar) | 36 | 1392 | 33185 | 15995 | 99.231 | $<2.2 e-16$ | * |
| as.factor (veman) | 11 | 270 | 33174 | 15725 | 62.956 | $<2.2 e-16$ | * |
| factor(skipnr) | 157 | 1898 | 33017 | 13826 | 31.022 | $<2.2 e-16$ | *** |
| factor(reitur) | 133 | 1009 | 32884 | 12818 | 19.459 | < 2.2e-16 | * |



Figure 20.2.1 Survey indices of the Icelandic slope S. mentella in the autumn survey in ICES Division Va 2000-2014. No survey was conducted in 2011. a) Total biomass index. b) Total abundance index in millions of fish. c) Biomass index of fish larger than 45 cm . d) Abundance index of fish smaller than 30 cm .


Figure 20.2.2 Length distribution of Icelandic slope S. mentella in the Autumn Groundfish Survey in October 2000-2014 in ICES Division Va. No survey was conducted in 2011. The black line is the mean of 2000-2014.


Figure 20.2.3 Age distribution of Icelandic slope S. mentella from the Autumn Survey in 2000 (n=1 405), $2009(\mathrm{n}=1101)$, and $2010(\mathrm{n}=1$ 206). The age class 50 are the combined age-classes of 50 years and older.


Figure 20.3.1 Nominal landings (in tonnes) of Icelandic slope S. mentella from ICES Divisions Va 1950-2014.




Figure 20.3.2 Geographical location of the Icelandic slope S. mentella catches in Icelandic waters (ICES Division Va and XIV) 1992-2014 as reported in log-books of the Icelandic fleet using bottom trawl. The blue line indicates part of the management unit for the deep-pelagic redfish stock. The dotted line represents the 500 m isobaths.











| $<0.1$ | $\square$ |
| :--- | :--- |
| $0.1-0.2$ |  |
| $0.2-1.0$ |  |
| $1.0-2.0$ |  |
| $2.0-4.0$ |  |
| $4.0-6.0$ |  |
| $6.0-8.0$ |  |
| $>8$ |  |





Figure 20.3.3 Geographical location of the Icelandic slope S. mentella catches in Icelandic waters (ICES Division Va and XIV) 1991-2003 and 2007 as reported in log-books of the Icelandic fleet using pelagic trawl. The blue line indicates part of the proposed management unit for the deep-pelagic redfish stock. The dotted line represents the 500 m isobaths.


Figure 20.3.4 Nominal landings (in tonnes) of Icelandic slope S. mentella in Icelandic waters (ICES Division Va and XIV) of the Icelandic fleet using either bottom trawl (red line) or pelagic trawl (blue line) 1991-2014 divided by month.


Figure 20.3.5 Length distributions of Icelandic slope S. mentella from the Icelandic landings taken with bottom trawl (blue line) and pelagic trawl (red line) in ICES Division Va 1991-2014.


Figure 20.3.6 CPUE relative to 1978 of Icelandic slope S. mentella from the Icelandic bottom trawl fishery in Division Va. CPUE based on a GLM model based on data from log-books and where at least $50 \%$ of the total catch in each tow was Icelandic slope S. mentella. Also shown is fishing effort (hours fished in thousands).


Figure 20.3.7 Residual of the GLM model (section 18.3.5) for the CPUE series of Icelandic slope $S$. mentella.


Figure 20.5.1. Icelandic slope S. mentella. Number aged plotted on log-scale. Grey lines correspond to $Z=0.1$ and $Z=0.3$.

## 21 Shallow Pelagic Sebastes mentella

## Executive summary

ICES concluded in February 2009 that S. mentella is to be divided into three biological stocks and that the shallow pelagic S. mentella in the Irminger Sea and adjacent areas should be treated as separate biological stock and management unit.

Total landings of shallow pelagic S. mentella in 2014 were 6423 t , a significant increase compared to 1527 t in 2013. The catches were almost entirely taken in ICES XII.

No analytical assessment was conducted and there are no biological reference points for the species. Survey indices from the biennial international acoustic redfish survey conducted in the Irminger Sea and adjacent waters since 1991 are used as basis for advice.

The last survey was conducted in June/July 2013. Since 1994, the results of the acoustic survey show a drastic decreasing trend within the deep scattering layer (DSL) from 2.2 million $t$ to 91000 t in 2013. With the trawl method within the DSL ( $350-500 \mathrm{~m}$ ) the biomass was estimated 200000 t , significantly below the 361000 t of 2011. The next international acoustic redfish survey was scheduled to be conducted in June/July 2015.

No signs of recruitment have been observed in the latest German survey on the EastGreenland shelf.

### 21.1 Stock description and management unit

This section addresses the fishery for shallow pelagic S. mentella in the Irminger Sea and adjacent areas (parts of Division Va, Subareas XII and XIV; eastern parts of NAFO Divisions 1F, 2H and 2J) at depths shallower than 500 m .
The following text table summarizes the available information from fishing fleets in the Irminger Sea and adjacent waters fishing for the shallow pelagic redfish in 2014. Only Russia conducted directed fishery on the stock. It should be noted that they also fished the deep pelagic stock:

| Russia | 17 factory trawlers |
| :--- | :--- |

### 21.2 Summary of the development of the fishery

The historic development of the fishery can be found in the Stock Annex. The clear changes in the spatial pattern of the fishery can be seen in Figure 20.2.1, based on logbook data from the Faroe Islands, Greenland, Iceland and Norway. A summary of the catches by ICES Divisions/NAFO regulatory area as estimated by the Working Group is given in Table 20.2.1 and Figure 20.2.2. The estimated catch for 2014 is 6423 t , a significant incecrease from the 1527 t caught in 2013. The catches were almost entirely produced by Russia with 5780 t fromICES XII and NAFO 1F (Tables 20.2.1 and 20.2.2).

There are no new CPUE data for 2014. The standardized CPUE index trend for the period 1994-2006 is shown in Figure 20.2.3. This standardized CPUE series includes data from Faroe Islands, Iceland, Germany, Greenland, and Norway, and it is estimated with a GLM model including the factors year, ship, month and towing time. The model output is shown in Table 20.2.3 and the residuals are in Figure 20.2.4.

### 21.3 Biological information

There are no new data. The length distributions for the period 1989-2006 of biological stocks based on Icelandic data are shown in Figure 20.3.1. The length of the largest proportion of caught fish oscillates around 35 cm for the whole period.

### 21.4 Discards

Redfish form aggregations composed of individuals with a narrow size range, which results in very clean catches. Thus, discards are negligible according to available data from various institutes.

### 21.5 Illegal Unregulated and Unreported Fishing (IUU)

The Group had again difficulties in obtaining catch estimates from several fleets. Furthermore, there are problems with misreported catches from some nations. The Group requests NEAFC and NAFO to provide ICES in time with all the necessary information.

### 21.6 Surveys

The last international trawl-acoustic survey was carried out in 2013 and it is described in detail in ICES WGRS Report 2013 (ICES, 2013). The next survey was scheduled to be carried out in June/July 2015 (ICES, 2013) but it may have to be cancelled after Russia withdrew its participation, unless other countries take part in it.

### 21.6.1 Survey acoustic data

Since 1994, the results of the acoustic survey show a drastic decreasing trend from 2.2 kt to 0.6 kt in 1999 and have fluctuated between $0.7 \mathrm{kt}-0.09 \mathrm{kt}$ in 2001-2013 (Table 20.6.1). The 2003 estimate, however, was considered to be inconsistent with the time series due to a shift in the timing of the survey.
The most recent trawl-acoustic survey on pelagic redfish (S. mentella) in the Irminger Sea and adjacent waters was carried out by Iceland, Germany and Russia in June/July 2013. Approximately $341000 \mathrm{NM}^{2}$ were covered. Figures 20.6 .1 and 20.6 .2 show the biomass estimates for depth shallower than the DSL (Depth Scattering Layer). A total biomass of 91000 t was estimated acoustically in the layer shallower than the DSL (Table 20.6.1 and Figure 20.6.4). The results showed a substantial biomass decline in subarea B compared to 2011 but in other areas the biomass was similar as in 2011 (Table 20.6.2 and Figure 20.6.5 for area definition). Biological samples from the acoustic estimate within the DSL and shallower than 500 m showed a mean length of 36.0 cm (Figure 20.6.6).

### 21.6.2 Survey trawl estimates

In addition to the acoustic measurements, redfish biomass was estimated by correlating catches and acoustic values at depths shallower than 500 m at 200000 t , a $45 \%$ decrease respect the estimation of 360,000 for 2011 (Table 20.6.1 and Figure 20.6.4). Figure 20.6 .3 shows the distribution of the redfish catches within the DSL and shallower than 500 m . It should be noted that the estimate for 2013 was recalculated due to technical error made in 2013 (ICES 2014).

The obtained correlation was used to convert the trawl data at greater depths to acoustic values and from there to abundance. For that purpose, standardized trawl hauls were carried out at depth $350-500 \mathrm{~m}$, evenly distributed over the survey area (Figure 20.6.3). For the time being, the correlation between the catch and acoustic values is
based on few data points only and it is highly variable. It is also assumed that the catchability of the trawl is the same, regardless of the trawling depth, thus the abundance estimate obtained is questionable and must only be considered as a rough attempt to measure the abundance within the DSL. Evaluation on the consistency of the method has to wait until more data points are available.
Biological samples from the trawls taken at depth $<500 \mathrm{~m}$ showed a mean length of 35.5 cm . Figure 20.6 .3 shows the spatial distribution of samples used in the survey and Figure 20.6.6 shows the corresponding length distribution.

### 21.6.3 Methods

The assessment of pelagic redfish in the Irminger Sea and adjacent waters is based on survey indices, catches, CPUE and biological data. See Stock Annex and Section 20.6 for details.

### 21.6.4 Reference points

For pelagic redfish in the Irminger Sea and adjacent waters, no analytical assessment is carried out due to data uncertainties and the lack of reliable age data. Thus, no reference points can be derived.

### 21.7 State of the stock

### 21.7.1 Short term forecast

For pelagic redfish in the Irminger Sea and adjacent waters, no analytical assessment is carried out due to data uncertainties and the lack of reliable age data. Thus, no shortterm forecasts can be derived.

### 21.7.2 Uncertainties in assessment and forecast

### 21.7.2.1 Data considerations

Preliminary official landings data were provided by the ICES Secretariat, NEAFC and NAFO, and various national data were reported to the Group. The Group, however, repeatedly faces problems to obtain reliable catch data due to unreported catches of pelagic redfish and lack of catch data disaggregated by depth from some countries. There are indications that reported effort (and consequently landings) could represent only around $80 \%$ of the real effort in certain years (see Chapter 20.3.3 in the 2008 NWWG report, ICES, 2008). No new data in IUU have been available since 2008.
As in previous years, detailed descriptions on the horizontal, vertical and seasonal distribution of the fisheries were given.
The need and importance of having catch and biological data disaggregated by depth from all nations taking part in the fishery cannot be stressed strongly enough, and the Group urges all nations involved on supplying better data. With this need in mind, ICES sent a data call to all EU countries participating in the redfish fishery, encouraging stockholders to deliver detailed catch data before the WG would meet, but the response was very limited.

### 21.7.2.2 Assessment quality

The results of the international trawl-acoustic survey are given in section 20.6. Given the high variability in the correlation between trawl and acoustic estimates as well as
the assumptions that need to be made about constant catchability across depth and areas, the uncertainty of these estimates is very high.

The reduction in biomass observed in the surveys within the hydroacoustic layer (about 2 million $t$ in the last decade) cannot be explained by the reported removal by the fisheries (about 500,000 $t$ in the entire depth range in 1995-2011) alone. A decreasing trend in the relative biomass indices in the acoustic layer, however, is visible since 1991.

It is not known to what extent CPUE reflects changes in the stock status of pelagic $S$. mentella, since the fishery focuses on aggregations. Therefore, stable or increasing CPUE series might not indicate or reflect actual trends in stock size, although decreasing CPUE indices are likely to reflect a decreasing stock. The new data available to the NWWG were insufficient to estimate the CPUE for 2013.

NEAFC set for 2013 a 0 TAC for Shallow Pelagic S. mentella. However, the Russian Federation decided on an unilateral quota of 27300 t . This quota was taken from both Shallow and Deep pelagic stocks, since the Russian Federation does not agree on the division of the $S$. mentella management units.

### 21.7.3 Comparison with previous assessment and forecast

The data available for evaluating the stock status are similar to last year.

### 21.7.4 Management considerations

The Group needs more and better data and requests that NEAFC and NAFO provide ICES with all information leading to more reliable catch statistics.

The main feature of the fishery since 1998 is a clear distinction between two widely separated fishing grounds with pelagic redfish fished at different seasons and different depths. Since 2000, the southwestern fishing grounds extended also into the NAFO Convention Area. Biological data, however, suggest that the aggregations in the NAFO Convention Area do not constitute a separate stock. The NAFO Scientific Council agreed with this conclusion (NAFO, 2005). The Group concludes that at this time there are not enough scientific bases available to propose an appropriate split of the total TAC among the two fisheries/areas.

### 21.7.5 Ecosystem considerations

The fisheries on pelagic redfish in the Irminger Sea and adjacent waters are generally regarded as having negligible impact on the habitat and other fish or invertebrate species due to very low bycatch and discard rates, characteristic of fisheries using pelagic gear.

### 21.7.6 Changes in the environment

The hydrography in the June/July 2013 survey show that temperature in the survey area is above average but it was lower than in 2011 in most of the surveyed area, except for the Irminger Current (ICES, 2013).

The increase of water temperature in the Irminger Sea may have an effect on spatial and vertical distribution of S. mentella in the feeding area (Pedchenko, 2005). The abundance and distribution of $S$. mentella in relation to oceanographic conditions were analysed in a special multistage workshop (WKREDOCE1-3). Based on 20 years of survey data, the results reveal the average relation of redfish to their physical habitat in shallow and intermediate waters: The most preferred latitude, longitude, depth, salinity
and temperature for S. mentella are approximately $58^{\circ} \mathrm{N}, 40^{\circ} \mathrm{W}, 300 \mathrm{~m}, 34.89$ and $4.4^{\circ} \mathrm{C}$, respectively. The spatial distribution of S. mentella in the Irminger Sea mainly in waters < 500 m (and thus mainly relating to the "shallow" stock) appears strongly influenced by the Irminger Current Water (ICW) temperature changes, linked to the Subpolar Gyre (SPG) circulation and the North Atlantic Oscillation (NAO). The fish avoid waters mainly associated with the ICW $\left(>4.5^{\circ} \mathrm{C}\right.$ and $\left.>34.94\right)$ in the north-eastern Irminger Sea, which may cause displacing towards the southwest, where fresher and colder water occurs (ICES 2012).

Results based on international redfish survey data suggest that the interannual distribution of fish above 500 m will shift in a southwest/northeast direction depending on integrated oceanographic conditions (ICES 2012).

### 21.8 References

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Pedchenko, A. P. 2005. The role of interannual environmental variations in the geographic range of spawning and feeding concentrations of redfish Sebastes mentella in the Irminger Sea. ICES Journal of Marine Science 62: 1501-1510.

Table 21.2.1 Shallow Pelagic S. mentella (stock unit < 500 m ). Catches (in tonnes) by area as used by the Working Group.

| Year | Va | XII | XIV | NAFO 1F | NAFO 2J | NAFO 2H | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 0 | 39783 | 20798 | 0 | 0 | 0 | 60581 |
| 1983 | 0 | 60079 | 155 | 0 | 0 | 0 | 60234 |
| 1984 | 0 | 60643 | 4189 | 0 | 0 | 0 | 64832 |
| 1985 | 0 | 17300 | 54371 | 0 | 0 | 0 | 71671 |
| 1986 | 0 | 24131 | 80976 | 0 | 0 | 0 | 105107 |
| 1987 | 0 | 2948 | 88221 | 0 | 0 | 0 | 91169 |
| 1988 | 0 | 9772 | 81647 | 0 | 0 | 0 | 91419 |
| 1989 | 0 | 17233 | 21551 | 0 | 0 | 0 | 38784 |
| 1990 | 0 | 7039 | 24477 | 385 | 0 | 0 | 31901 |
| 1991 | 0 | 9689 | 17048 | 458 | 0 | 0 | 27195 |
| 1992 | 106 | 22976 | 38709 | 0 | 0 | 0 | 62564 |
| 1993 | 0 | 66458 | 32500 | 0 | 0 | 0 | 100771 |
| 1994 | 665 | 77174 | 18679 | 0 | 0 | 0 | 96869 |
| 1995 | 77 | 78895 | 17895 | 0 | 0 | 0 | 100136 |
| 1996 | 16 | 22474 | 18566 | 0 | 0 | 0 | 41770 |
| 1997 | 321 | 18212 | 8245 | 0 | 0 | 0 | 27746 |
| 1998 | 284 | 21976 | 1598 | 0 | 0 | 0 | 24150 |
| 1999 | 165 | 23659 | 827 | 534 | 0 | 0 | 25512 |
| 2000 | 3375 | 17491 | 687 | 11052 | 0 | 0 | 33216 |
| 2001 | 228 | 32164 | 1151 | 5290 | 8 | 1751 | 41825 |
| 2002 | 10 | 24004 | 222 | 15702 | 0 | 3143 | 43216 |
| 2003 | 49 | 24211 | 134 | 26594 | 325 | 5377 | 56688 |
| 2004 | 10 | 7669 | 1051 | 20336 | 0 | 4778 | 33951 |
| 2005 | 0 | 6784 | 281 | 16260 | 5 | 4899 | 28229 |
| 2006 | 0 | 2094 | 94 | 12692 | 260 | 593 | 15734 |
| 2007 | 71 | 378 | 98 | 2843 | 175 | 2561 | 6126 |
| 2008 | 32 | 25 | 422 | 1580 | 0 | 0 | 2059 |
| 2009 | 0 | 210 | 2170 | 0 | 0 | 0 | 2380 |
| 2010 | 15 | 686 | 423 | 1074 | 0 | 0 | 2198 |
| 2011 | 0 | 0 | 234 | 0 | 0 | 0 | 234 |
| 2012 | 28 | 0 | 0 | 3113 | 32 | 0 | 3173 |
| 2013 | 32 | 13 | 40 | 1443 | 1 | 0 | 1529 |
| 2014 | 153 | 5068 | 489 | 713 | 0 | 0 | 6423 |

1982-1991 All pelagic catches assumed to be of the shallow pelagic stock
1992-1996 Guesstimates based on different sources (see text)
1997-2014 Catches from calculations based on jointed catch database and total landings

Table 21．2．2 Shallow pelagic S．mentella catches（in tonnes）in ICES Div．Va，Subareas XII，XIV and NAFO Div．1F，2H and 2J by countries used by the Working Group．＊Prior to 1991，the figures for Russia included Estonian，Latvian and Lithuanian catches．

|  | $\begin{aligned} & \text { 厄 } \\ & \text { N } \\ & \text { 苞 } \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { ర్ల } \\ & \text { ভ゙ } \\ & \hline \end{aligned}$ |  |  |  | $\begin{aligned} & \text { 入 } \\ & \text { N} \\ & \text { Ey } \\ & \text { U } \\ & \hline \end{aligned}$ |  |  |  |  | $\sum_{\substack{0}}^{0}$ |  |  |  | $\begin{aligned} & \text { 㐅} \\ & \text { 3 } \\ & \text { z} \\ & \text { z } \end{aligned}$ |  |  | $\begin{aligned} & \stackrel{*}{\sim} \\ & \stackrel{\pi}{n} \\ & \check{a} \\ & \hline \underline{\alpha} \\ & \hline \end{aligned}$ | － | $\underset{ }{\square}$ |  | $\stackrel{\text { ®0 }}{\substack{0}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 581 |  | 60000 |  |  |  | 60581 |
| 1983 |  |  |  |  |  | 155 |  |  |  |  |  |  |  |  |  |  |  | 60079 |  |  |  | 60234 |
| 1984 | 2961 |  |  |  |  | 989 |  |  |  |  |  |  |  |  |  | 239 |  | 60643 |  |  |  | 64832 |
| 1985 | 5825 |  |  |  |  | 5438 |  |  |  |  |  |  |  |  |  | 135 |  | 60273 |  |  |  | 71671 |
| 1986 | 11385 |  |  | 5 |  | 8574 |  |  |  |  |  |  |  |  |  | 149 |  | 84994 |  |  |  | 105107 |
| 1987 | 12270 |  |  | 382 |  | 7023 |  |  |  |  |  |  |  |  |  | 25 |  | 71469 |  |  |  | 91169 |
| 1988 | 8455 |  |  | 1090 |  | 16848 |  |  |  |  |  |  |  |  |  |  |  | 65026 |  |  |  | 91419 |
| 1989 | 4546 |  |  | 226 |  | 6797 | 567 |  | 3816 |  |  |  |  |  |  | 112 |  | 22720 |  |  |  | 38784 |
| 1990 | 2690 |  |  |  |  | 7957 |  |  | 4537 |  |  |  |  |  | 7085 |  |  | 9632 |  |  |  | 31901 |
| 1991 |  |  | 2195 | 115 |  | 201 |  |  | 8724 |  |  |  |  |  | 6197 |  |  | 9747 |  |  |  | 27179 |
| 1992 | 628 |  | 1810 | 3765 | 2 | 6447 | 9 |  | 12080 |  | 780 | 6656 |  |  | 14654 |  |  | 15733 |  |  |  | 62564 |
| 1993 | 3216 |  | 6365 | 6812 |  | 16677 | 710 |  | 10167 |  | 6803 | 7899 |  |  | 14112 |  |  | 25229 |  |  | 2782 | 100771 |
| 1994 | 3600 |  | 17875 | 2896 | 606 | 15133 |  |  | 5897 |  | 13205 | 7404 |  |  | 6834 |  | 1510 | 16349 |  |  | 5561 | 96869 |
| 1995 | 2660 | 421 | 11798 | 3667 | 158 | 10714 | 277 |  | 8733 | 841 | 3502 | 16025 | 9 |  | 4288 |  | 2170 | 28314 | 1934 |  | 2230 | 100136 |
| 1996 | 1846 | 343 | 3741 | 2523 |  | 5696 | 1866 |  | 5760 | 219 | 572 | 5618 |  |  | 1681 |  | 476 | 9348 | 1671 | 137 | 273 | 41770 |
| 1997 |  | 102 | 3405 | 3510 |  | 9276 |  |  | 4446 | 28 |  |  |  |  | 330 | 776 | 367 | 3693 | 1812 |  |  | 27746 |
| 1998 |  |  | 3892 | 2990 |  | 9679 | 1161 |  | 1983 | 30 |  | 1734 |  |  | 701 | 12 | 60 | 89 | 1819 |  |  | 24150 |
| 1999 |  |  | 2055 | 1190 |  | 8271 | 998 |  | 3662 |  |  |  |  |  | 2098 | 6 | 62 | 6538 | 447 | 183 |  | 25512 |
| 2000 |  |  | 4218 | 486 |  | 5672 | 956 |  | 3766 |  |  | 430 |  |  | 2124 |  | 37 | 14373 | 1154 |  |  | 33216 |
| 2001 |  |  | 9 | 4364 |  | 4755 | 1083 |  | 14745 |  |  | 8269 |  |  | 947 |  | 256 | 5964 | 1433 |  |  | 41825 |
| 2002 |  |  |  | 719 |  | 5354 | 657 |  | 5229 |  | 1841 | 12052 |  |  | 1094 | 428 | 878 | 13958 | 1005 |  |  | 43216 |


| 㐫 | $\begin{aligned} & \text { 䓘 } \\ & \text { 厄 } \\ & \frac{0}{5} \end{aligned}$ | $\begin{aligned} & \text { ర్ } \\ & \text { た } \\ & \tilde{\pi} \\ & \hline \end{aligned}$ | $\begin{array}{r} \stackrel{\pi}{\bar{y}} \\ \stackrel{\rightharpoonup}{山} \\ \hline \end{array}$ |  | $\begin{aligned} & \text { シ } \\ & \stackrel{\text { 匕 }}{4} \end{aligned}$ |  |  |  |  |  | $\begin{aligned} & \text { r } \\ & \text { 厄 } \\ & \\ & \hline \end{aligned}$ |  |  |  | $\begin{aligned} & \text { त } \\ & \text { 3 } \\ & \text { zo } \\ & \mathbf{z} \end{aligned}$ | $\begin{aligned} & \text { 들 } \\ & \text { त } \end{aligned}$ |  | $\begin{aligned} & * \\ & \stackrel{*}{4} \\ & \hat{n} \\ & \underset{\alpha}{2} \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { ᄃ̄ } \\ & \text { ĩ } \end{aligned}$ | $\stackrel{y}{J}$ | ハّ | － |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 |  |  |  | 1955 |  | 3579 |  | 1047 |  | 4274 |  | 1269 | 21629 |  | 3214 | 917 | 1926 | 15418 | 1461 |  |  | 56688 |
| 2004 |  |  |  | 777 |  | 1126 |  | 750 |  | 5728 |  | 1114 | 3698 |  | 2721 | 1018 | 2133 | 13208 | 1679 |  |  | 33951 |
| 2005 |  |  |  | 210 |  | 1152 |  |  |  | 3086 |  | 919 | 1169 |  | 624 | 1170 | 2780 | 15562 | 1557 |  |  | 28229 |
| 2006 |  |  |  | 334 |  | 994 |  |  |  | 1293 |  | 1803 | 466 |  | 280 | 663 | 1372 | 4953 | 3576 |  |  | 15734 |
| 2007 |  |  | 209 | 98 |  | 0 |  |  |  | 71 |  | 186 | 467 |  |  | 189 | 529 | 4037 | 339 |  |  | 6126 |
| 2008 |  |  |  | 319 |  |  |  |  |  | 63 |  |  | 8 |  |  |  |  | 1597 | 73 |  |  | 2059 |
| 2009 |  |  |  | 87 |  |  |  |  |  | 5 |  |  | 138 |  |  |  |  | 649 | 1438 |  |  | 2380 |
| 2010 |  |  |  | 653 |  |  |  |  |  | 22 |  |  | 551 |  | 12 |  | 377 | 567 | 16 |  |  | 2198 |
| 2011 |  |  |  | 162 |  |  |  |  |  | 72 |  |  |  |  |  |  |  |  |  |  |  | 234 |
| 2012 |  |  |  |  |  |  |  |  |  | 28 |  |  |  |  |  |  |  | 3145 |  |  |  | 3173 |
| 2013 |  |  |  |  |  |  |  |  |  | 72 |  |  |  |  |  |  |  | 1457 |  |  |  | 1529 |
| 2014 |  |  |  |  |  |  |  |  |  | 355 |  |  | 287 |  |  |  |  | 5781 |  |  |  | 6423 |

Table 21.2.3 Output from the GLM model used to standardize CPUE
Call:
glm(formula $=$ lafli $\sim$ ltogtimi + factor(land $)+$ factor(yy) + factor(mm) + factor(skip), family = gaussian(), data $=$ south $)$

Deviance Residuals:
Min 1Q Median 3Q Max
$-2.67560-0.27475 \quad 0.01545 \quad 0.28216 \quad 1.70226$
Coefficients:
Estimate Std. Error $t$ value $\operatorname{Pr}(>|t|)$
(Intercept) $\quad 7.2883306000 .6215319011 .726398294 .487183 \mathrm{e}-27$ ltogtimi $\quad 1.0311890890 .0286543435 .987192171 .185172 \mathrm{e}-120$ factor(land)46 $0.3071080070 .196774191 .560712821 .194800 \mathrm{e}-01$ factor(land) $58-0.6092223840 .59427534-1.02515171 \quad 3.059877 \mathrm{e}-01$ factor(yy)1995 $-0.0145441450 .17246972-0.084328699 .328425 \mathrm{e}-01$ factor(yy)1996 $-0.5399670920 .20301506-2.659739058 .173648 \mathrm{e}-03$ factor(yy)1997 $-0.7810973750 .19187694-4.07082472$ 5.775636e-05 factor(yy)1998 $-0.5982056820 .20022972-2.98759682 \quad 3.006814 \mathrm{e}-03$ factor(yy)1999 -1.032123656 0.19849297-5.19979958 3.371986e-07 factor(yy)2000 $-0.4490670150 .18062595-2.48617105$ 1.337053e-02 factor(yy)2001 $-0.2940957490 .18731402-1.57006796 \quad 1.172876 \mathrm{e}-01$ factor(yy)2002 $-0.5534226980 .20779476-2.663314038 .089018 \mathrm{e}-03$ factor(yy)2003 $-0.4485304620 .20695582-2.16727635 \quad 3.087629 \mathrm{e}-02$ factor(yy)2004 $-0.9404675620 .19921557-4.72085375 \quad 3.382253 \mathrm{e}-06$ factor(yy)2005 $-0.8742280870 .21534893-4.059588746 .047701 \mathrm{e}-05$ factor(yy)2006 $-0.7925136220 .23511568-3.370739078 .318962 \mathrm{e}-04$ factor $(\mathrm{mm}) 3 \quad 0.4035399150 .62653390 \quad 0.64408313 \quad 5.199363 \mathrm{e}-01$ factor $(\mathrm{mm}) 4 \quad 0.0808863360 .599655290 .134888058 .927766 \mathrm{e}-01$ factor $(\mathrm{mm}) 5 \quad 0.6972894820 .597294181 .16741383$ 2.438246e-01 factor $(\mathrm{mm}) 6 \quad 0.1065815040 .595821120 .178881728 .581323 \mathrm{e}-01$ factor $(\mathrm{mm}) 7 \quad 0.1560065390 .599133890 .260386777 .947160 \mathrm{e}-01$ factor $(\mathrm{mm}) 8 \quad 0.2886879020 .60200469 \quad 0.47954427 \quad 6.318459 \mathrm{e}-01$ factor $(\mathrm{mm})^{9} \quad 0.1473727450 .603507550 .24419370 \quad 8.072215 \mathrm{e}-01$ factor(mm)10 $\quad-0.0731373960 .61289180-0.119331669 .050799 \mathrm{e}-01$ factor $(\mathrm{mm}) 11 \quad-0.1114296360 .62872288-0.177231728 .594272 \mathrm{e}-01$ factor $(\mathrm{mm}) 12 \quad-0.6872076540 .84232729-0.815843994 .151349 \mathrm{e}-01$ factor(skip)118 $-0.3091797780 .22143007-1.396286291 .634983 \mathrm{e}-01$ factor(skip)1270 $0.0376031490 .448280910 .083883009 .331966 \mathrm{e}-01$ factor(skip)1273-0.628141253 0.22041607-2.84979787 4.629299e-03 factor(skip)1279-1.173362444 0.44513557-2.63596647 8.756942e-03 factor(skip) 1308 - $0.2669192650 .22303502-1.19675943 \quad 2.321967 \mathrm{e}-01$ factor(skip) 1328 - $0.2716542510 .21750992-1.24892811 \quad 2.125120 \mathrm{e}-01$ factor(skip)1345-0.389432255 0.27300563-1.42646238 1.546113e-01 factor(skip) $1351-0.2109225670 .30230014-0.69772567$ 4.858042e-01 factor(skip)1360 - $0.1603370350 .37131520-0.431808436 .661421 \mathrm{e}-01$ factor(skip)1365-0.037778373 0.28528994-0.13242098 8.947261e-01 factor(skip)1369 $0.0082218780 .23222821 \quad 0.03540430 \quad 9.717772 \mathrm{e}-01$ factor(skip)1376-0.079339629 0.21104413-0.37593857 7.071865e-01 factor(skip)1408-0.360954071 0.46295849-0.77966833 4.361041e-01 factor(skip)1412 -0.186735060 $0.60272438-0.30981833 \quad 7.568804 \mathrm{e}-01$ factor(skip)1459-0.659207386 0.22905256-2.87797434 4.243932e-03 factor(skip)1471-0.067779436 0.39810737-0.17025416 8.649070e-01 factor(skip)1472 -0.243213212 $0.33706786-0.72155563$ 4.710413e-01 factor(skip)1473 -0.831933012 0.45025953-1.84767443 6.547885e-02 factor(skip) 1552 -1.308585894 0.61116338 -2.14113925 3.294138e-02 factor(skip) 1578 -1.486687432 0.38045634-3.90764269 1.115534e-04 factor(skip)1579-0.474709749 0.30501933-1.55632678 1.205189e-01 factor(skip)1585-0.553949127 $0.61783175-0.89660191 \quad 3.705373 \mathrm{e}-01$ factor(skip)1628 0.0488619840 .452916860 .10788290 9.141494e-01 factor(skip)180 $-0.5326137340 .18564922-2.868925304 .364387 \mathrm{e}-03$ factor(skip)1833-0.296067754 0.22785023-1.29939633 1.946488e-01
factor(skip)1868-0.104954736 0.22921245-0.45789282 6.473088e-01 factor(skip)1880 $0.0041530550 .258263610 .016080689 .871790 \mathrm{e}-01$ factor(skip)1902 $0.2040439870 .284172820 .718027824 .732111 \mathrm{e}-01$ factor(skip)1976 -0.380940434 0.61538320-0.61902963 5.362928e-01 factor(skip)1977-0.774106835 0.33815309-2.28922009 2.265145e-02 factor(skip)2165 $0.1050475900 .205808960 .510413116 .100784 \mathrm{e}-01$ factor(skip)2170-0.122213348 $0.20408250-0.598842865 .496585 \mathrm{e}-01$ factor(skip)2182 -0.454140930 $0.23283220-1.950507375 .190006 \mathrm{e}-02$ factor(skip)2184-0.295249414 0.25222782-1.17056639 2.425561e-01 factor(skip)2203-0.136558045 0.20059787-0.68075523 4.964689e-01 factor(skip)2212 $0.1833021430 .304962760 .601064025 .481798 \mathrm{e}-01$ factor(skip)2236 -0.581565095 0.26502996-2.19433717 $2.885678 \mathrm{e}-02$ factor(skip)2265 $0.2397188650 .439515190 .545416575 .858086 \mathrm{e}-01$ factor(skip)2592 -0.282434578 $0.59801605-0.472285956 .370121 \mathrm{e}-01$ factor(skip)3033-0.283499142 $0.72458991-0.391254616 .958431 \mathrm{e}-01$ factor(skip)3135-0.016478186 $0.66105345-0.02492716$ 9.801270e-01 factor(skip) $3156-0.2608053620 .61679034-0.422842816 .726652 \mathrm{e}-01$ factor(skip) $3382-0.4234249190 .62159313-0.681193054 .961922 \mathrm{e}-01$ factor(skip)3523-0.395258535 0.72563919-0.54470396 5.862982e-01 factor(skip)3542 $0.0183557450 .619941890 .029608829 .763956 \mathrm{e}-01$ factor(skip)3709-0.609676578 $0.64465767-0.94573695$ 3.449242e-01 factor(skip) $934-1.0546467130 .17107235-6.16491621$ 1.912436e-09--Signif. codes: $0{ }^{\prime * * * '} 0.001^{\text {'**' }} 0.01^{\prime * \prime} 0.05^{\prime}$.' $0.1^{\prime}$ ' 1
(Dispersion parameter for gaussian family taken to be 0.3450127 )
Null deviance: 989.53 on 458 degrees of freedom
Residual deviance: 131.45 on 381 degrees of freedom AIC: 886.64

```
Number of Fisher Scoring iterations: }
Analysis of Deviance Table
Model: gaussian, link: identity
Response: lafli
Terms added sequentially (first to last)
Df Deviance Resid. DfResid. Dev F Fr(>F)
NULL 428 934.30
ltogtimi 1 682.16 427 252.14 2126.3228<2.2e-16***
factor(land) 2 38.99 425 213.15 60.7682<2.2e-16***
factor(yy) 12 43.18 413 169.96 11.2167<2.2e-16 ***
factor(mm) 10}1017.04 403 152.92 5.3122 2.600e-07****
factor(skip) 47 38.71 356 114.21 2.5673 5.376e-07***
Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1
```

Table 21.6.1 Shallow Pelagic S. mentella. Results for the acoustic survey indices from shallower than the scattering layer, trawl estimates within the deep scattering layer and shallower than 500 $m$, and area coverage of the survey in the Irminger Sea and adjacent waters.

|  |  | Acoustic estimates |  |
| :--- | :--- | :--- | :--- |
| Year | Area covered (1000 NM2) | Trawl estimates $\mathbf{1 0 0 0} \mathbf{t}$ |  |
| 1991 | 105 | 2235 |  |
| 1992 | 190 | 2165 |  |
| 1993 | 121 | 2556 |  |
| 1994 | 190 | 2190 |  |
| 1995 | 168 | 2481 | 565 |
| 1996 | 253 | 1576 | $92^{*}$ |
| 1997 | 158 | 1225 | 392 |
| 1999 | 296 | 614 | 283 |
| 2001 | 420 | 716 | 331 |
| $2003^{*}$ | 405 | $89^{*}$ | 361 |
| 2005 | 386 | 550 | 200 |
| 2007 | 349 | 372 |  |
| 2009 | 360 | 108 |  |
| 2011 | 343 | 123 |  |
| 2013 | 340 | 91 |  |

* The 2003 biomass estimate is considered as inconsistent as the survey was carried out about one month earlier than usual, and a marked seasonal effect was observed.

Table 21.6.2. Results (biomass in '000 t) for the international surveys conducted since 1994, for redfish shallower than the DSL for each subarea (see Figure 21.6.5 for area definition) and total.

|  | Sub-area |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | A | B | C | D | E | F | Total |
| 1994 | 673 | 1228 | - | 63 | 226 |  | 2190 |
| 1996 | 639 | 749 | - | 33 | 155 |  | 1576 |
| 1999 | 72 | 317 | 16 | 42 | 167 |  | 614 |
| 2001 | 88 | 220 | 30 | 267 | 103 | 7 | 716 |
| 2003 | 32 | 46 | 1 | 2 | 10 | 0 | 89 |
| 2005 | 121 | 123 | 0 | 87 | 204 | 17 | 551 |
| 2007 | 80 | 95 | 0 | 53 | 142 | 3 | 372 |
| 2009 | 39 | 48 | 4 | 1 | 15 | 1 | 108 |
| 2011 | 5 | 74 | 0 | 3 | 40 | 1 | 123 |
| 2013 | 9 | 33 | 2 | 5 | 42 | 0 | 91 |



Figure 21.2.1 Fishing areas and total catch of pelagic redfish (S. mentella) in the Irminger Sea and adjacent waters 1989-2012. Data are from the Faroe Islands (1995-2012), Iceland (1989-2012) and Norway (1992-2003). The catches in the legend are given as tonnes per square nautical mile. The blue box represents the management unit for the northern fishing area.


Figure 21.2.1 (Cont.) Fishing areas and total catch of pelagic redfish (S. mentella) in the Irminger Sea and adjacent waters 1989-2012. Data are from the Faroe Islands (1995-2012), Iceland (1989-2012) and Norway (1992-2003). The catches in the legend are given as tonnes per square nautical mile. The blue box represents the management unit for the northern fishing area.


Figure 21.2.2 Landings of shallow pelagic S. mentella (Working Group estimates, see Table 21.2.1).


Figure 21.2.3 Trends in standardised CPUE of the shallow pelagic S. mentella fishery in the Irminger Sea and adjacent waters, based on log-book data from Faroes, Iceland, Norway, and Greenland.


Figure 21.2.4 Residuals from the GLM model used to standardize CPUE, based on log-book data from Faroe Islands, Iceland, Greenland and Norway.


Figure 21.3.1 Length distribution from Icelandic landings of shallow pelagic S. mentella.


Figure 21.6.1 Pelagic S. mentella. Acoustic estimates (average sA values by 5 NM sailed) shallower than the deep-scattering layer (DSL) from the joint trawl-acoustic survey in June/July 2013.



Figure 21.6.2. Redfish acoustic estimates shallower than the DSL. Average $s_{A}$ values within statistical rectangles during the joint international redfish survey in June/July 2013.


Figure 21.6.3 Redfish trawl estimates within the DSL shallower than 500 m (type 2 trawls). sa values calculated by the trawl method (chapter 2.2.3) during the joint international redfish survey in June/July 2013.


Figure 21.6.4. Overview of acoustic survey indices (thousand tonnes) from above the scattering layer (red filled circle), trawl estimates within the scattering layer and shallower than 500 m (black triangle), and aerial coverage (NM2) of the survey (black open circle) in the Irminger Sea and adjacent waters.


Figure 21.6.5 Sub-areas A-F used on international surveys for redfish in the Irminger Sea and adjacent waters, and divisions for biological data (Northeast, Southwest and Southeast; boundaries marked by broken lines).


Figure 21.6.6 Length distribution of redfish in the trawls, by geographical areas and total, from fish caught shallower than 500 m (in 2013).

## Executive summary

ICES concluded in February 2009 that S. mentella is to be divided to three biological stocks and that the deep pelagic S. mentella in the Irminger Sea and adjacent areas should be treated as separate biological stock and management unit.

Total landings of deep pelagic S. mentella s in 2014 were 23755 t , half of the 2013 total catch.

No analytical assessment was conducted and there are no biological reference points for the species. Survey indices from the biennial international trawl-acoustic redfish survey conducted in the Irminger Sea and adjacent waters since 1999 are used as basis for advice.

The survey was conducted in June/July 2013. A total biomass of 280900 t was estimated, a $41 \%$ less than in 2011 (474 000 t).Trawl survey estimates in 2011 and 2013 are lower than the average for 1999-2009 and the estimate for 2013 is the lowest observed. The next international trawl-acoustic redfish survey in the Irminger Sea will be conducted in June/July 2015.

No recruitment has been observed on the East-Greenland shelf during 2013 and 2014, which is a concern because it is assumed to contribute to the three $S$. mentella stocks at unknown shares.

### 22.1 Stock description and management unit

This section addresses the fishery for the biological stock deep pelagic S. mentella in the Irminger Sea and adjacent areas: NAFO 1-2, ICES V, XII, and XIV at depths $>500 \mathrm{~m}$, including demersal habitats west of the Faeroe Islands. This stock corresponds to the management unit in the northeast Irminger Sea (ICES areas Va, XII and XIV).
The following text table summarizes the available information from fishing fleets in the Irminger Sea and adjacent waters in 2014. It should be noted that some these fleets are also fishing the Shallow Pelagic stock:

| Country | Number of trawlers |
| :--- | :---: |
| Faroes | 2 factory trawlers |
| Iceland | 10 factory trawlers |
| Germany | 1 factory trawlers |
| Latvia | 1 factory trawlers |
| Lithuania | 1 factory trawlers |
| Norway | 5 factory trawlers |
| Russia | 17 factory trawlers |
| Spain | 3 factory trawlers |

### 22.2 The fishery

The historic development of the fishery can be found in the Stock Annex. Tables 22.2.1 and 22.2.2 show annual catches, as estimated by the Working Group, disaggregated by ICES and NAFO regulatory areas and by country, respectively.

The changes in the spatial pattern of the fishery for the period 1992-2014 are shown in Figure 22.2.1, and annual catches are presented in Figure 22.2.2. Catches decreased by nearly $50 \%$ from 45594 t in 2013 to 23755 t (Table 22.2.2).

Standardized CPUE series for Faroe Islands, Iceland, Greenland, and Norway 19942014 are estimated with a GLM model including the factors year, ship, month and towing time. The results from the model show that the CPUE oscillates without trend since 1995 (Figure 22.2.3). The model output is shown in Table 21.2.3 and the residuals are in Figure 22.2.4. The CPUE index increased from about 0.3 in 2012 to $>1.0$ in 2013

### 22.3 Biological information

The length distribution from Icelandic landings for the period 1991-2014 is shown in Figure 22.3.1. Peak length between 1994 and 1997 was about 37 cm , but increased to roughly 42 from 1998 to 2005, although in 2002 the distribution showed two peaks, at 37 and 42 cm , and in 2003 the peak declined to 40 cm . Mean length has decreased further over the past years, but an increase was observed with 38.3 cm in 2012 compared with 37.8 cm in 2011. The length distribution in 2014 decreased to 35.8 cm (Figure 22.3.1).

### 22.4 Discards

Discards are not considered to be significant for the time being, according to available data from various institutes.

### 22.5 Illegal, Unregulated and Unreported Fishing (IUU)

The Group had again difficulties in obtaining catch estimates from several fleets. Furthermore, there are problems caused by misreported catches. The Group requests NEAFC and NAFO to provide ICES in time with all the necessary information.

### 22.6 Surveys

The last international trawl-acoustic survey took place in 2013 and it is described in detail in ICES CM WGRS REPORT 2013 (ICES, 2013). The next survey was scheduled to be carried out in June/July 2015 but after the withdrawal of Russia it may be cancelled unless other countries take part. It should be noted that the 2013 estimate was recalculated during the WKREDMP meeting in January 2014 (ICES 2014) as it was wrong because of technical error. The 1999 estimate was also recalculated.

### 22.6.1 Survey trawl estimates

Considering the conclusion of WKREDS (ICES, 2009a) and the recommendation of ICES on stock structure of redfish in the Irminger Sea and adjacent waters, the Group decided in the planning meeting (ICES, 2009b) to sample redfish separately above and below 500 m , i.e. to sample redfish as was done in the 1999, 2001 and 2003 surveys . The deep identification hauls covered the depth layers (headline) $550 \mathrm{~m}, 700 \mathrm{~m}$, and 850 m .

The most recent trawl-acoustic survey on pelagic redfish (S. mentella) in the Irminger Sea and adjacent waters was carried out by Iceland, Germany and Russia in June/ July 2013. Approximately $341000 \mathrm{NM}^{2}$ were covered. A total biomass of 280000 t was estimated, significantly below the 474000 t of 2011 (Table 22.6.2). The results showed large biomass declines in subareas A, B and E (see Figure 22.6.1 for area definition) (Table 21.6.2). Biological samples from the trawls taken at depth $>500 \mathrm{~m}$ showed a mean
length of 38.5 cm , which is 0.5 cm larger than the mean length in 2011. Figure 22.6.2 shows the spatial distribution of samples used in the survey and Figure 22.6.3 shows the corresponding length distribution.

### 22.7 Methods

The assessment of pelagic redfish in the Irminger Sea and adjacent waters is based on survey indices, catches, CPUE and biological data.

### 22.8 Reference points

For pelagic redfish in the Irminger Sea and adjacent waters, no analytical assessment is carried out due to data uncertainties and the lack of reliable age data. Thus, no reference points can be derived.

### 22.9 State of the stock

### 22.9.1 Short term forecast

For pelagic redfish in the Irminger Sea and adjacent waters, no analytical assessment is being carried out due to data uncertainties and the lack of reliable age data. Thus, no short-term forecasts can be derived.

### 22.9.2 Uncertainties in assessment and forecast

### 22.9.2.1 Data considerations

Preliminary official landings data were provided by the ICES Secretariat, NEAFC and NAFO, and various national data were reported to the Group. The Group, however, repeatedly faces problems to obtain reliable catch data due to unreported catches of pelagic redfish and lack of catch data disaggregated by depth from some countries.

As in previous years, detailed descriptions on the horizontal, vertical and seasonal distribution of the fisheries are given.

The need and importance of having catch and biological data disaggregated by depth from all nations taking part in the fishery cannot be stressed strongly enough, and the Group urges all nations involved on supplying better data. With this need in mind, ICES sent a data call to all EU countries participating in the redfish fishery, encouraging stockholders to deliver detailed catch data before the WG would meet, but the response was very limited.

### 22.9.2.2 Assessment quality

The results of the international trawl-acoustic survey are given in section 21.6. Given the high variability in the correlation between trawl and acoustic estimates as well as the assumptions that need to be made about constant catchability across depth and areas, the uncertainty of these estimates is very high.

It is not known to what extent CPUE reflect changes in the stock status of pelagic $S$. mentella, since the fishery focuses on aggregations. Therefore, stable or increasing CPUE series might not indicate or reflect actual trends in stock size, although decreasing CPUE indexes are likely to reflect a decreasing stock.

### 22.9.3 Comparison with previous assessment and forecast

The data available for evaluating the stock status are similar to last year.

### 22.9.4 Management considerations

The Group needs more and better data and requests that NEAFC and NAFO provide ICES with all information leading to more reliable catch statistics.

The main feature of the fishery since 1998 is a clear distinction between two widely separated fishing grounds with pelagic redfish fished at different seasons and different depths. Since 2000, the southwestern fishing grounds extended also into the NAFO Convention Area. Biological data, however, suggest that the aggregations in the NAFO Convention Area do not constitute a separate stock. The NAFO Scientific Council agreed with this conclusion (NAFO, 2005). The Group concludes that at this time there is not enough scientific basis available to propose an appropriate split of the total TAC among the two fisheries/areas.

The 20000 t TAC set by NEAFC for 2014 was overshot by a much smaller percentage than usual, $20 \%$ vs $80 \%$ in 2013. This excess is due to the unilateral decision of the Russian Federation to self-allocate an annual TAC, which was 23700 t for 2014. It was taken from both Shallow and Deep pelagic ( 15475 t) stocks, since the Russian Federation does not agree on the division of the S. mentella management units.

### 22.9.5 Ecosystem considerations

The fisheries on pelagic redfish in the Irminger Sea and adjacent waters are generally regarded as having negligible impact on the habitat and other fish or invertebrate species due to very low bycatch and discard rates, characteristic of fisheries using pelagic gear.

### 22.9.6 Changes in the environment

The hydrography in the survey of June/July 2013 show that temperature in the survey area is above average but it was lower than in 2011 in most of the surveyed area, except for the Irminger Current (ICES, 2013).

The increase of water temperature in the Irminger Sea may have an effect on spatial and vertical distribution of S. mentella in the feeding area (Pedchenko, 2005). The abundance and distribution of $S$. mentella in relation to oceanographic conditions were analysed in a special multistage workshop (WKREDOCE1-3, see ICES 2012b). Based on 20 years of survey data, the results reveal the average relation of redfish to their physical habitat in shallow and intermediate waters: The most preferred latitude, longitude, depth, salinity and temperature for S. mentella are approximately $58^{\circ} \mathrm{N}, 40^{\circ} \mathrm{W}, 300 \mathrm{~m}$, 34.89 and $4.4^{\circ} \mathrm{C}$, respectively. The spatial distribution of $S$. mentella in the Irminger Sea mainly in waters $<500 \mathrm{~m}$ (and thus mainly relating to the "shallow" stock) appears strongly influenced by the Irminger Current Water (ICW) temperature changes, linked to the Subpolar Gyre (SPG) circulation and the North Atlantic Oscillation (NAO). The fish avoid waters mainly associated with the ICW ( $>4.5^{\circ} \mathrm{C}$ and salinity $>34.94$ ) in the north-eastern Irminger Sea, which may cause displacing towards the southwest, where fresher and colder water occurs (ICES 2012b).
Results based on international redfish survey data suggest that the inter-annual distribution of fish above 500 m will shift in a southwest/northeast direction depending on integrated oceanographic conditions (ICES 2012). Whether the results of the study mentioned are applicable to the conditions for the deep pelagic stock needs further investigation.

### 22.10 WKREDMP 2014

At WKREMP 2014 ICES was requested by Faroe Islands, Iceland and Greenland to evaluate proposed harvest control rules for deep pelagic redfish in the Irminger Sea and adjacent waters (ICES 2014). ICES reanalysed the survey time-series, which is the main source of information for the assessment. This changed the perception of stock status and productivity: The stock appears to be at a historical low. ICES also evaluated the proposed harvest control rules, and none of them are expected to lead to an increase in stock size by 2025. Therefore, ICES considers none of these options as being in accordance with the precautionary approach. It is suggested that managers discuss other options with ICES that might be more suitable, including a starting phase to reverse the decline of the stock.

### 22.11References

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Pedchenko, A. P. 2005. The role of interannual environmental variations in the geographic range of spawning and feeding concentrations of redfish Sebastes mentella in the Irminger Sea. ICES Journal of Marine Science 62: 1501-1510.

Table 22.2.1 Deep Pelagic S. mentella (stock unit > $\mathbf{5 0 0} \mathbf{m}$ ). Catches (in tonnes) by area as used by the Working Group.

| Year | Va | XII | XIV | NAFO 1F | Total |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1991 | 0 | 7 | 52 | 0 | 59 |
| 1992 | 1862 | 280 | 1257 | 0 | 3398 |
| 1993 | 2603 | 6068 | 6393 | 0 | 15064 |
| 1994 | 14807 | 16977 | 20036 | 0 | 51820 |
| 1995 | 1466 | 53141 | 21100 | 0 | 75707 |
| 1996 | 4728 | 20060 | 113765 | 0 | 138552 |
| 1997 | 14980 | 1615 | 78485 | 0 | 95079 |
| 1998 | 40328 | 444 | 52046 | 0 | 92818 |
| 1999 | 36359 | 373 | 47421 | 0 | 84153 |
| 2000 | 41302 | 0 | 51811 | 0 | 93113 |
| 2001 | 27920 | 0 | 59073 | 0 | 86993 |
| 2002 | 37269 | 2 | 65858 | 0 | 103128 |
| 2003 | 46627 | 21 | 57648 | 0 | 104296 |
| 2004 | 14446 | 0 | 77508 | 0 | 91954 |
| 2005 | 11726 | 0 | 33759 | 0 | 45485 |
| 2006 | 16452 | 51 | 50531 | 254 | 67288 |
| 2007 | 17769 | 0 | 40748 | 0 | 58516 |
| 2008 | 4602 | 0 | 25443 | 0 | 30045 |
| 2009 | 16828 | 4658 | 32920 | 0 | 54406 |
| 2010 | 8552 | 0 | 50736 | 0 | 59288 |
| 2011 | 0 | 7 | 47326 | 0 | 47333 |
| 2012 | 5530 | 608 | 26668 | 0 | 32806 |
| 2013 | 5274 | 0 | 40778 | 0 | 46052 |
| 2014 | 603 | 0 | 23152 | 0 | 23755 |
|  |  |  |  |  |  |

## Table 22．2．2 Deep pelagic S．mentella catches（in tonnes）in ICES Div．Va，Subareas XII，XIV and NAFO Div．1F，2H and 2J by countries used by the Working Group．

| $\begin{aligned} & \text { だ } \\ & \text { ঠ̀ } \end{aligned}$ |  |  |  | $\begin{aligned} & \tilde{0} \\ & \stackrel{0}{0} \\ & \stackrel{0}{\widetilde{u}} \end{aligned}$ |  |  | $\begin{aligned} & \text { D } \\ & \frac{\tilde{\sigma}}{\bar{J}} \\ & \text { む̀ } \end{aligned}$ | $\begin{aligned} & \underset{\sim}{c} \\ & \underline{\sigma} \\ & \underline{U} \end{aligned}$ | $\begin{aligned} & \text { 厄్ర } \\ & \text { ご } \end{aligned}$ | $\sum_{\pi}^{\pi}$ |  |  | $\begin{aligned} & \text { त } \\ & \text { 3 } \\ & \text { z} \\ & \text { z } \end{aligned}$ | $\begin{aligned} & \text { ס } \\ & \text { 등 } \\ & \text { R } \end{aligned}$ | $\begin{aligned} & \bar{\sigma} \\ & \text { N } \\ & \text { 흥 } \end{aligned}$ | $\begin{aligned} & \underline{\pi} \\ & \tilde{n} \\ & \underset{\sim}{z} \end{aligned}$ | $\begin{aligned} & \text { 드출 } \\ & \text { in } \end{aligned}$ | $\stackrel{\checkmark}{J}$ | 先 | $\begin{aligned} & \overline{\boxed{\circ}} \\ & \stackrel{\rightharpoonup}{\circ} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 |  |  |  |  |  |  |  | 59 |  |  |  |  |  |  |  |  |  |  |  | 59 |
| 1992 |  |  |  |  |  |  |  | 3398 |  |  |  |  |  |  |  |  |  |  |  | 3398 |
| 1993 |  |  |  | 310 |  | 1135 |  | 12741 |  |  |  |  | 878 |  |  |  |  |  |  | 15064 |
| 1994 |  |  |  |  |  | 2019 |  | 47435 |  |  |  |  | 523 |  | 377 | 1465 |  |  |  | 51820 |
| 1995 | 1140 | 181 | 5056 | 1572 | 68 | 8271 | 1579 | 25898 | 396 | 1501 | 6868 | 4 | 3169 |  | 2955 | 15868 | 227 |  | 956 | 75707 |
| 1996 | 1654 | 307 | 3351 | 3748 |  | 15549 | 1671 | 57143 | 196 | 512 | 5031 |  | 5161 |  | 1903 | 36400 | 5558 | 123 | 245 | 138552 |
| 1997 |  | 9 | 315 | 435 |  | 11200 |  | 36830 | 3 |  |  |  | 2849 |  | 3307 | 33237 | 6895 |  |  | 95079 |
| 1998 |  |  | 76 | 4484 |  | 8368 | 302 | 46537 | 1 |  | 34 |  | 438 |  | 4073 | 25748 | 2758 |  |  | 92818 |
| 1999 |  |  | 53 | 3466 |  | 8218 | 3271 | 40261 |  |  |  |  | 3337 |  | 4240 | 11419 | 9885 | 5 |  | 84153 |
| 2000 |  |  | 7733 | 2367 |  | 6827 | 3327 | 41466 |  |  | 0 |  | 3108 |  | 3694 | 14851 | 9740 |  |  | 93113 |
| 2001 |  |  | 878 | 3377 |  | 5914 | 2360 | 27727 |  |  | 7515 |  | 4275 |  | 2488 | 23810 | 8649 |  |  | 86993 |
| 2002 |  |  | 15 | 3664 |  | 7858 | 3442 | 39263 |  |  | 9771 |  | 4197 |  | 2208 | 25309 | 7402 |  |  | 103128 |
| 2003 |  |  |  | 3938 |  | 7028 | 3403 | 44620 |  |  | 0 |  | 5185 |  | 2109 | 28638 | 9374 |  |  | 104296 |
| 2004 |  |  |  | 4670 |  | 2251 | 2419 | 31098 |  |  | 0 |  | 6277 | 1889 | 2286 | 31067 | 9996 |  |  | 91954 |
| 2005 |  |  |  | 1800 |  | 1836 | 1431 | 12919 |  |  | 1027 |  | 3950 | 1240 | 1088 | 16323 | 3871 |  |  | 45485 |
| 2006 |  |  |  | 3498 |  | 1830 | 744 | 20942 |  |  | 1294 |  | 5968 | 1356 | 1313 | 23670 | 6673 |  |  | 67288 |
| 2007 |  |  |  | 2902 |  | 1110 | 1961 | 18097 |  | 575 | 1394 |  | 4628 | 636 | 2067 | 21337 | 3810 |  |  | 58516 |
| 2008 |  |  |  | 2632 |  |  | 1170 | 6723 |  |  | 749 |  | 571 | 219 | 1733 | 15106 | 1142 |  |  | 30045 |
| 2009 |  |  |  | 3206 |  |  | 1519 | 15125 |  | 1355 | 2613 |  |  | 178 | 1596 | 25309 | 2907 |  |  | 54006 |
| 2010 |  |  |  | 3195 |  |  | 1932 | 14772 |  | 1963 | 2228 |  | 2388 | 3 | 2203 | 22803 | 7801 |  |  | 59288 |
| 2011 |  |  |  | 2028 |  | 1787 |  | 11994 |  | 845 | 1348 |  | 1066 |  | 1540 | 22364 | 4361 |  |  | 47333 |
| 2012 |  |  |  | 1438 |  | 1523 |  | 5912 |  | 724 | 558 |  | 3362 |  | 250 | 18377 | 632 |  |  | 32806 |
| 2013 |  |  |  | 1882 |  | 1176 |  | 8545 |  | 1200 | 1163 |  | 2979 |  |  | 26463 | 2644 |  |  | 46052 |
| 20141） |  |  |  | 721 |  | 890 |  | 2081 |  | 867 | 1024 |  | 1965 |  |  | 15475 | 732 |  |  | 23755 |

Provisional．Official Spanish catch data were lower than the data provided by NEAFC and the WG decided to use the highest catch data as a precautionary measure．

Table 22.2.3 Output from the GLM model used to standardize CPUE - NOT UPDATED

## Call:

glm(formula $=$ lafli $\sim$ ltogtimi + factor(land $)+$ factor(yy) + factor(mm) + factor(skip), family $=$ gaussian(), data $=$ north $)$

Deviance Residuals:
Min 1Q Median 3Q Max
$\begin{array}{llllll}-3.5126 & -0.2410 & 0.0168 & 0.2924 & 1.4568\end{array}$
Coefficients: (3 not defined because of singularities)
Estimate Std. Error t value $\operatorname{Pr}(>|t|)$
(Intercept) $\quad 7.8715952790 .3909419420 .134946938$ 2.522077e-78 ltogtimi $\quad 1.0514389930 .01695572 \quad 62.0108867870 .000000 \mathrm{e}+00$ factor(land)6 $\quad-0.1987745460 .35054379-0.5670462565 .707845 \mathrm{e}-01$ factor(land)46 $0.3564163710 .13062029 \quad 2.728644788$ 6.448602e-03 factor(land)58 $0.3859056630 .35297372 \quad 1.093298568$ 2.744732e-01 factor(land)69 $0.1314272160 .21480114 \quad 0.6118552955 .407447 \mathrm{e}-01$ factor(yy)1995 -0.544478224 0.09104861 -5.980082924 2.906186e-09 factor(yy) $1996 \quad-0.5913507580 .08559433-6.9087610037 .764598 \mathrm{e}-12$ factor(yy)1997 $-1.0745797370 .08522551-12.608663463$ 2.120951e-34 factor(yy) $1998 \quad-0.6956380420 .08486144-8.1973398116 .028342 \mathrm{e}-16$ factor(yy)1999 -0.797915967 0.08474619 -9.415361145 2.186144e-20 factor(yy)2000 $-0.4317900780 .08594177-5.0242164755 .788362 \mathrm{e}-07$ factor(yy)2001 $-0.9557691590 .08486379-11.2623911064 .296930 \mathrm{e}-28$ factor(yy)2002 $-0.5755450270 .08596398-6.695188372$ 3.244414e-11 factor(yy)2003 -0.316293204 0.08682628 -3.642828098 2.806914e-04 factor(yy)2004 $-1.0160983160 .08870892-11.454296393$ 5.867807e-29 factor(yy)2005 -1.325075546 0.09344673-14.180009215 1.955001e-42 factor(yy)2006 -0.919905830 0.09688906 -9.494424199 1.079864e-20 factor(yy)2007 $-0.6814049910 .10068657-6.7675859402 .007085 \mathrm{e}-11$ factor(yy)2008 $-1.0392929730 .11665575-8.9090594271 .776186 \mathrm{e}-18$ factor(yy)2009 $-0.5758115150 .10378067-5.548350380$ 3.516366e-08 factor(yy)2010 $-0.3373305720 .10886481-3.098618992$ 1.987589e-03 factor(yy)2011 -0.715714357 0.10794155 -6.630573586 4.961323e-11 factor(yy)2012 -1.309676579 0.11568546-11.321012872 2.345629e-28 factor(yy)2013 $0.0088683250 .15200923 \quad 0.0583407009 .534866 e-01$ factor $(\mathrm{mm}) 3 \quad-0.8126598030 .39174110-2.074481862$ 3.823868e-02 factor $(\mathrm{mm}) 4 \quad-0.3785501090 .37671843-1.0048622013 .151575 \mathrm{e}-01$ factor $(\mathrm{mm}) 5 \quad-0.1804825970 .37822841-0.4771788536 .333181 \mathrm{e}-01$ factor $(\mathrm{mm}) 6 \quad-0.3330639150 .37796344-0.8812066923 .783752 \mathrm{e}-01$ factor $(\mathrm{mm}) 7 \quad-0.5038716480 .37798537-1.3330453631 .827596 \mathrm{e}-01$ factor $(\mathrm{mm}) 8 \quad-0.6088381370 .38197175-1.593934978$ 1.112032e-01 factor(mm) $9 \quad-0.4593266970 .39365610-1.1668222632 .435045 \mathrm{e}-01$ factor $(\mathrm{mm}) 10 \quad-0.7581709300 .43452708-1.7448185828 .126201 \mathrm{e}-02$ factor(mm)11 $\quad-0.7468334340 .50556889-1.4772139941 .398700 \mathrm{e}-01$ factor(skip) $118-0.2673502770 .14387130-1.8582600046 .336686 \mathrm{e}-02$ factor(skip)1265-0.305357861 $0.22335074-1.367167461$ 1.718184e-01 factor(skip) $1268-0.2668523540 .48481836-0.5504171805 .821315 \mathrm{e}-01$ factor(skip) $1270-0.1522207600 .13168835-1.1559166532 .479360 \mathrm{e}-01$ factor(skip) $1273-0.4194008280 .13225616-3.171125008$ 1.555377e-03 factor(skip) $1279-0.4474498210 .22139223-2.0210727974 .348488 \mathrm{e}-02$ factor(skip) $1308-0.0384590050 .12407714-0.3099604507 .566427 \mathrm{e}-01$ factor(skip) $1328-0.1446770280 .13096482-1.1047014502 .695014 \mathrm{e}-01$ factor(skip) $1345-0.4583722160 .12962963-3.5360141154 .210561 \mathrm{e}-04$ factor(skip) $1351-0.3570880240 .13798946-2.5877920709 .771199 \mathrm{e}-03$ factor(skip) $1360-0.1152608780 .13128164-0.877966489$ 3.801305e-01 factor(skip) $1365-0.2959657600 .14818615-1.9972566104 .601360 \mathrm{e}-02$ factor(skip) $1369-0.0113205740 .14158965-0.0799534009 .362871 \mathrm{e}-01$ factor(skip) $1376-0.1694737950 .12691871-1.3352940731 .820230 \mathrm{e}-01$ factor(skip) $1395-0.4099959240 .26206284-1.5644947241 .179543 \mathrm{e}-01$ factor(skip) 1408 -1.101773815 $0.48734615-2.260762308$ 2.394521e-02 factor(skip) $1412-0.1113579480 .29537612-0.377003899$ 7.062346e-01
factor(skip)1459-0.583761981 0.13242885 -4.408117879 1.132118e-05 factor(skip)1471-0.613726934 0.17679359-3.471432056 5.353333e-04 factor(skip)1472-0.511423665 0.16493944-3.100675325 1.973945e-03 factor(skip)1473-0.952361655 0.21265005 -4.478539431 8.201613e-06 factor(skip)1484-1.433135836 0.48533948-2.952852375 3.207303e-03 factor(skip)1497-1.418776803 0.35333866 -4.015345477 6.288939e-05 factor(skip)1530 -0.740040738 $0.48506624-1.5256488311 .273501 \mathrm{e}-01$ factor(skip)1536-1.814090714 0.48651089-3.728777180 2.010160e-04 factor(skip)1552-1.430344940 0.29607192 -4.831072584 1.525962e-06 factor(skip)1553-0.003402939 0.35390074 -0.009615518 9.923296e-01 factor(skip)1578-0.354319033 0.15025452 -2.358125591 1.852067e-02 factor(skip)1579-0.071762545 0.12696644 -0.5652087495.720331e-01 factor(skip)1585 -0.417372263 0.17163124-2.431796521 1.516373e-02 factor(skip)1628-1.097387934 0.29532540 -3.715860332 2.114507e-04 factor(skip) $180-0.5226640610 .11197449-4.6677065743 .373762 \mathrm{e}-06$ factor(skip)1833-0.025339243 0.12377825 -0.204714826 8.378282e-01 factor(skip)1868-0.141108499 0.12362729 -1.141402470 $2.539209 \mathrm{e}-01$ factor(skip) $1880-0.3005382110 .13797726-2.1781720892 .957946 \mathrm{e}-02$ factor(skip)1902-0.122037774 0.13376781 -0.912310478 3.617811e-01 factor(skip)1903-0.503244254 0.29824691-1.687341029 9.178708e-02 factor(skip)1976-0.628200177 0.20114205 -3.123166888 1.830228e-03 factor(skip)1977-0.169902957 0.14830443 -1.145636427 2.521647e-01 factor(skip)2107-0.796278905 0.35043313 -2.272270619 2.323928e-02 factor(skip)2165 $0.0724955880 .13389141 \quad 0.5414506385 .882934 \mathrm{e}-01$ factor(skip)2170 -0.090472494 0.12251178 -0.738479960 4.603614e-01 factor(skip)2182-0.227474427 0.12989539-1.751212416 8.015439e-02 factor(skip)2184-0.083675892 $0.12979356-0.6446844745 .192499 \mathrm{e}-01$ factor(skip)2203-0.195064716 0.12387778 -1.574654639 1.155890e-01 factor(skip)2212 $0.0433651600 .160826960 .2696386257 .874828 \mathrm{e}-01$ factor(skip)2220 $0.1705528460 .486218440 .3507741227 .258169 \mathrm{e}-01$ factor(skip)2236-0.240310005 0.20254007-1.186481294 $2.356576 \mathrm{e}-01$ factor(skip)2248-0.468803577 0.35307606 -1.327769382 1.844965e-01 factor(skip)2265-0.081118608 $0.13404819-0.6051451275 .451923 \mathrm{e}-01$ factor(skip)2410 -0.192166371 0.21987506 -0.873979834 3.822970e-01 factor(skip)2549-0.144974322 $0.14686104-0.9871530313 .237585 \mathrm{e}-01$ factor(skip)2550 $0.0275683340 .48384311 \quad 0.0569778379 .545719 \mathrm{e}-01$ factor(skip)2592 $0.2291146610 .354616050 .6460921985 .183381 \mathrm{e}-01$ factor(skip)3033-1.789786537 $0.44139343-4.0548553825 .326475 \mathrm{e}-05$ factor(skip)3135-0.358279004 0.40653274 -0.881304184 3.783225e-01 factor(skip) $3156-1.2769434310 .36406219-3.5074870844 .683906 \mathrm{e}-04$ factor(skip) $3382-1.1742600490 .37396205-3.140051352$ 1.728735e-03 factor(skip)3523-1.415480845 0.45907228 -3.083350696 2.091640e-03 factor(skip)3542-0.818668375 $0.38925897-2.1031458413 .565233 \mathrm{e}-02$ factor(skip) 3709-1.212729719 $0.37572436-3.2277112211 .280216 \mathrm{e}-03$ factor(skip)934 -0.560484938 $0.10397540-5.3905533168 .388350 \mathrm{e}-08$ factor(skip)A $\quad 0.4788879000 .44197157 \quad 1.083526488$ 2.787836e-01 factor(skip)B $\quad 0.7474847330 .39901053 \quad 1.873345878$ 6.125365e-02 ---
Signif. codes: $0^{\text {'***' }} 0.001^{\text {'**' }} 0.01^{\text {'*' }} 0.05^{\text {'.' }} 0.1^{\text {' ' }} 1$
(Dispersion parameter for gaussian family taken to be 0.2216236 )
Null deviance: 1901.61 on 1335 degrees of freedom Residual deviance: 274.59 on 1239 degrees of freedom AIC: 1873.7

Number of Fisher Scoring iterations: 2
Analysis of Deviance Table

Model: gaussian, link: identity
Response: lafli

Terms added sequentially (first to last)

Df Deviance Resid. DfResid. Dev F $\operatorname{Pr}(>F)$
ltogtimi $11353.661347559 .666153 .6477<2.2 \mathrm{e}-16^{* * *}$
factor(land) $4 \quad 68.12 \quad 1343 \quad 491.55 \quad 77.4139<2.2 \mathrm{e}-16^{* * *}$ factor(yy) $19 \begin{array}{lllll}132.78 & 1324 & 358.76 & 31.7700<2.2 \mathrm{e}-16^{* * *}\end{array}$ factor(mm) $9117.82 \quad 1315 \quad 340.94 \quad 9.00212 .899 \mathrm{e}-13$ *** factor(skip) $64 \quad 65.75 \quad 1251 \quad 275.19 \quad 4.6701<2.2 \mathrm{e}-16^{* * *}$


Table 22.6.1 Deep pelagic $S$. mentella. Survey estimates for depth $>500 \mathrm{~m}$ from trawl samples taken in 2013.

|  | A | B | C | D | E | F | Total |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Area (NM2) | 123531 | 83385 | 4181 | 51185 | 62730 | 15683 | 340695 |
| Mean length | $(\mathrm{cm})$ | 38.8 | 37.5 | 36.1 | 36.3 | 38.2 | 37.7 |
| Mean weight | $(\mathrm{g})$ | 717 | 653 | 615 | 595 | 482 | 654 |
| Biomass $(\mathrm{t})$ | 193000 | 75000 | 0 | 2000 | 10000 | 0 | 280000 |

Table 22.6.2. Results (biomass in ' 000 t ) for the international redfish surveys conducted since 1999 for deep pelagic S. mentella for each subarea (see Figure 22.6.2) and total.

| Sub-area |  |  | B | C | D | E | F |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | A | B | 12 | 27 | 52 | 0 | 935 |
| 1999 | 277 | 568 | 28 | 79 | 64 | 18 | 1001 |
| 2001 | 497 | 316 | 20 | 13 | 27 | 0 | 678 |
| 2003 | 476 | 142 | 0 | 8 | 65 | 3 | 392 |
| 2005 | 221 | 95 | 166 | 1 | 5 | 62 | 11 |
| 2007 | 276 | 291 | 121 | 0 | 8 | 37 | 1 |
| 2009 | 342 | 112 | 0 | 1 | 18 | 0 | 458 |
| 2011 | 193 | 75 | 0 | 2 | 10 | 0 | 474 |
| 2013 |  |  |  |  |  | 280 |  |



Figure 22.2.1 Fishing areas and total catch of deep pelagic redfish (S. mentella) in the Irminger Sea and adjacent waters 1992-2014. Data are from the Faroe Islands (1995-2013), Germany (2011-2014) Greenland (1999-2003 and 2009-2010), Iceland (1995-2014), and Norway (1995-2003 and 2010-2014). The catches in the legend are given as tones per square nautical mile. The blue box represents the proposed management unit.


Figure 22.2.1 (Cont.) Fishing areas and total catch of deep pelagic redfish ( $S$. mentella) in the Irminger Sea and adjacent waters 1992-2014. Data are from the Faroe Islands (1995-2013), Germany (2011-2014) Greenland (1999-2003 and 2009-2010), Iceland (1995-2014), and Norway (1995-2003 and 2010-2014). The catches in the legend are given as tones per square nautical mile. The blue box represents the proposed management unit.


Figure 22.2.2 Landings of deep pelagic S. mentella (Working Group estimates, see Table 21.2.1).


Figure 22.2.3 Trends in standardised CPUE of the deep pelagic S. mentella fishery in the Irminger Sea and adjacent waters, based on log-book data from Faroe Islands, Iceland, Germany, Greenland and Norway.


Figure 22.2.4 Residuals from the GLM model used to standardize CPUE, based on log-book data from Faroe Islands, Iceland, Greenland and Norway.


Figure 22.3.1 Length distribution from Icelandic landings of deep pelagic S. mentella.


Figure 22.6.1 Sub-areas A-F used on international surveys for redfish in the Irminger Sea and adjacent waters, and divisions for biological data (Northeast, Southwest and Southeast; boundaries marked by broken lines).


Figure 22.6.2. Redfish trawl estimates deeper than 500 m (type 3 trawls). sa values calculated by the trawl method (see WGRS Report, 2013) during the joint international redfish survey in June/July 2013.


Figure 22.6.3 Length distribution of redfish in the trawls, by geographical areas (see Fig. 22.6.1) and total, from fish caught deeper than 500 m .

## 23 Greenlandic slope Sebastes mentella in XIVb

## Summary

ICES concluded in February 2009 that demersal S. mentella is to be divided into three biological stocks and that the S. mentella on the continental shelf and slope should be treated as a separate biological stock and management unit. This separation of the stocks did not include the adult S. mentella on the Greenlandic slopes. ICES therefore decided that NWWG will conduct a separate assessment of S. mentella in subarea XIVb until further information is available to assign stock origin. This chapter therefore deals only with the S. mentella on the Greenlandic Slope.

Total landings of demersal S. mentella in East Greenland waters in 2014 were about 4600 tons, which is less than 2010-2012 landings. The lower catches are partly due to a lower presence of mentella in the mixed stock fishery and partly due to a lower total landing of demersal redfish.

In the decade before 2009 S. mentella was mainly a valuable by-catch in the fishery for Greenland halibut. However, since 2009 a fishery directed towards demersal redfish has taken place.

Available survey biomass indices show that in Division XIVb the biomass decreased further in 2014. No new recruits ( $>18 \mathrm{~cm}$ ) are seen in the surveys since 2012, and no juveniles are present $(<18 \mathrm{~cm})$ in both 2013 and 2014 surveys
The advice is based on the DLS approach (3.2) using the Greenland shallow water survey as basis for advice. The ratio is applied to the 2014 advice as catches are well above the current advice. The advice for 2016 is 2240 t .

### 23.1 Stock description and management units

See chapter 18 for description of the stock structure of S. mentella in the Irminger Sea and adjacent waters. ICES has advised separately for $S$. mentella found demersal in ICES XIVb since 2011, and will do so until all available information on stock origin in this area is analysed and a new procedure is agreed upon.

### 23.2 Scientific data

Indices were available from three surveys in XIVb. A German survey directed towards cod in Greenlandic waters ( $0-400$ meters, Fock et al. 2015), the Greenland deep water survey (400-1500 meters) targeting Greenland halibut (Hedeholm et al. 2015a) and the Greenland shrimp and fish survey in shallow water ( $0-600$ meters) which has been conducted since 2008 (Hedeholm et al. 2015b). The German survey on the slope in XIVb has since 1982 been covering the slopes in East Greenland waters. Cod is the target species in this survey and it operates at depths of 400 meters and shallower. The survey was re-stratified in 2009 (see Stock Annex). From 1993-1998 a large number of Sebastes sp. smaller than 17 cm . was found in the survey (Figure 23.2.1). This coincided with a large increase in the amount of 17-30 cm large S. mentella from 1995-1998. From 1998 to 2003 the total biomass increased as a result of many small fish $(<17 \mathrm{~cm})$ in the survey, followed by a few years of high biomass estimates for S. mentella from 2003-2009. This increase occurred in one particular stratum only, i.e. stratum 8.2. From 2009 onward, a declining trend was observed, with the low biomass estimates resembling the conditions before (Figure 23.2.1). In the same period, the amount of small fish ( $17-30 \mathrm{~cm}$ ) has steadily declined causing an increase in the amount of larger fish (Figure 23.2.1c)
until the overall biomass declines in 2010 and 2011. The depletion of the small size group has led to a progressive decline in the juvenile biomass index to a current low level, and no new recruits have been seen in the survey since 2012. This pattern is also reflected in the abundance estimates (Figure 23.2.1). The modal size of the adult fish has increased from 25 cm in 2001 to around 37 cm in 2010, but declined slightly in 2011 and the distribution has becomes flat with clearly defined mode in 2013 (Figure 23.2.2).
The Greenland deep water survey has since 1998, except in 2001, surveyed the slopes of east Greenland from 400 to 1500 meters with the majority of stations deeper than 600 meters targeting Greenland halibut. The biomass indices in the Greenland deep water survey peaked in 2012, but has decreased since then (Figure 23.2.3). The overall length distribution from the entire area in 2013 and 2014 shows a mode around 31 cm which is a $1-2 \mathrm{~cm}$ increase compared to 2011 and 2012 values (Figure 23.2.4).

The Greenland shrimp and fish survey in shallow water in East Greenland started in 2007, and surveys the East Greenland shelf and shelf edge at depths between 0-600 m . However, 2007 was mostly exploratory and is not reported. In general, survey estimates of schooling fish are associated with large uncertainties due to their patchy distribution. This, in conjunction with the relatively short time series, makes overall conclusions regarding stock trends based solely on this survey tentative although it is probably the survey with the best coverage of redfish distribution. The 2014 biomass estimate was the lowest observed in the time series (Figure. 23.2.5). The German survey shows very similar trends both with regards to adult fish and juveniles. The juveniles are at the lowest level in the 30 year time series, and the adult biomass index has declined for the past five years and is at the lowest level since 2005. Both survey length distributions showed no clear mode, but a rather flat distribution (Figure 23.2.6). The German survey and the Greenland shrimp and fish shallow water survey both show overall declines in the S. mentella biomass since 2010 (Figure 23.2.7). Fproxy values based on surveys indicate an increasing fishing mortality in recent years (Figure 23.2.8).

### 23.3 Information from the fishing industry

### 23.3.1 Landings

From the Greenland and German surveys we know that the demersal redfish found on the Greenland slope is a mixture of S. marinus and S. mentella. Based on the surveys and fourteen samples from the commercial fishery the 7314 tonnes of demersal redfish caught in ICES XIVb, was estimated to be 63 \% S. mentella (4608) and $37 \%$ S. norvegicus (2706), which is the lowest proportion of S.mentella observed. Prior to 1974 all catches were reported as S. marinus and the split was determined by working groups on a yearly bases.

Total annual landings of demersal S. mentella from Divisions XIVb since 1974 are presented in Table 23.3.1.1. From 1976 to 1994 annual landings were at a relatively high level with landings ranging between 2000 tons to 20000 tons with a very high peak at nearly 60000 t in 1976. However, this fishery was ended abruptly in 1995 due to large amounts of very small redfish in the catches. From 1998 to 2002 the landings ranged from 1000 to 2000 tons and from 2003 to 2008 landings remained at lower levels ( $<500$ tons). In 2009 an exploratory fishery landed 895 tons of S. mentella. This was a large increase compared to 2008 and for the first time in ten years the fishery was limited by a TAC. In 2010, a quota on 5,000 tons demersal redfish was initially given and of these, $400 t$ were allocated to the Norwegian fleet. After this amount was fished, an extraordinary research quota of 1000 tons was given to a Greenlandic vessel. Since 2010 the catches have been around $8300 t$ (S. mentella and S. marinus combined) and 2013 catches
were 8246 t (Figure 23.3.1.1). The TAC for 2014 is 8500 . In 2010 there was no jurisdiction that clearly delimited the pelagic stocks from the redfish found on the shelf. A few vessels benefitted from this by fishing their pelagic quota on the shelf ( 2179 tons) making catches on the shelf exceed the TAC. This led to the introduction of a "redfish line" that separates the demersal slope stock from the pelagic stocks (see stock annex).

### 23.3.2 CPUE and by-catch CPUE

A redfish by-catch CPUE was introduced at the redfish 2012 benchmark (WKRED). This is based on catches from the Greenland halibut directed fishery (Hedeholm and Boje 2014a) which covers redfish distribution better than data from the redfish directed fishery and covers a longer period (1999-2014). The CPUE has very low values in the initial two years of the time series, but following an increase in 2001, values have remained at the same level until 2006 after which a decline followed. From 2010 to 2012 the CPUE increased, followed by a small decline in 2013-2014 (Figure 23.3.2.1).

The index does not show the decline in biomass index seen in the shallow water surveys (German and Greenland). This could be associated with the nature of the decline, which appears to be confined to the commercial area. The Greenland halibut fishery is not as spatially restricted as the redfish fishery, so it will not be as sensitive to local changes. Based on the CPUE there does not appear to be any large decline in stock size.

The CPUE from the redfish directed fishery showed a drastic decline from 2010 (3.7 $\mathrm{t} / \mathrm{h})$ to 2014 ( $1.3 \mathrm{t} / \mathrm{h}$, Figure 23.3.2.2). This fishery takes place in a geographically limited area between $63.5^{\circ} \mathrm{N}$ and $65^{\circ} \mathrm{N}$, where approximately $90 \%$ of the catches are taken. Accordingly, the CPUE series can only be used as an index on local stock development. Both the Greenland shallow water survey $(0-600 \mathrm{~m})$ and German survey ( $0-400 \mathrm{~m}$ ) show that the main fishing area coincides with the area of highest overall abundance. Hence, the CPUE decline indicates a severe local stock depletion that is also reflected in the overall stock trend.

### 23.3.3 Fisheries and fleets

The fishery for S. mentella on the slopes in XIVb is mainly conducted with bottom trawl. From 1998-2012 only 1\% were caught with longlines. The area where S. mentella is caught is closely related to the area where fishery for Greenland halibut and cod takes place. The majority of the catches are taken at depths from 300 m to 400 m . (Figure 23.3.3.1)

The directed fishery was stopped in 1995, but in 1998 Germany restarted a directed fishery for redfish with annual landings of approximately 1000 t in 1998-2001 increasing to 2100 t in 2002 (Bernreuther et al. 2013). Samples taken from the German fleet indicated that substantial quantities of the redfish caught, especially in 2002, were juveniles, i.e. fish less than 30 cm . There was very little demersal redfish fishery in XIVb in 2003-2004 (less than 500 t ). This continued in 2005-2008 and most S. mentella were caught as by-catch in the Greenland halibut fishery.
After the German fleet stopped fishing in 2002 the majority of the catches have been taken by the British, Faroese, Norwegian and Greenland fleet. The British fishery took place from 2001 to 2005 and since 2006 only Greenland, Faroese Islands, Norway and Germany have had any significant catches (Table 23.3.3.2).

In 2009 three Greenland vessels started a fishery targeting demersal redfish. Each was given an explorative quota of 250 t . This fishery was very successful and led to an increased fishery in 2010 (seven boats), 2011 ( 15 boats) and 2012 ( 21 boats). However, in 2012 95\% of the catch was taken by six vessels and $97 \%$ by five vessels in 2013.

On the steep slopes very little horizontal distance separates the distribution of cod, redfish and Greenland halibut (Figure 23.3.3.2). The part of the fleet with both quotas for redfish and Greenland halibut takes advantage of this by shifting between very short hauls targeting redfish and long hauls directed to Greenland halibut. Thereby avoiding time where the vessel is not fishing due to processing of the catch.

### 23.3.4 By-catch/discard in the shrimp fishery

To minimize by-catch of fish species in the fishery for shrimp the trawls have since 2002 been equipped with grid separators (G.H. 2001). However, the 22 mm spacing between the bars in the separator allows small fish to enter the codend. In a study of the amount of by-catch in the shrimp fishery the mean length of the redfish that entered the cod end was $13-14 \mathrm{~cm}$. The same study also documented that redfish by weight accounted for less than $1 \%$ of the amount of shrimp that were caught (Sünksen 2007). Coincident with the introduction of these separator grids the amount of juvenile redfish caught by the shrimp fishery dropped from annual 100-200 tons to a lower level near 100 tons. Since 2006 not much shrimp fishery has taken place in ICES XIVb and the current level of by-catch must be considered negligible (Table 23.3.4.1). Since 1999 the fishery has started in April-May due to poor winter conditions such as ice and wind that prevents fishing. Only in 2000 and 2002 the fishery started already in February (Table 23.3.4.2). Since 2010, the fishery has been starting already in January. The depth distribution of cod and redfish overlaps (Figure 23.3.3.2) and therefore the fishery for redfish led to a by-catch of cod on 96 tons in 2013. The vessels are allowed a $10 \%$ bycatch of cod.

### 23.3.5 Sampling from the commercial fishery

In 2013 the catch length distribution was estimated from 14 samples ( $\mathrm{N}=1$ 019, Figure 23.3.6.1). Length distribution of 752 redfish analysed by the Greenland Institute of Natural Resources separated into S.mentella $(\mathrm{N}=548)$ and S.norvegicus $(\mathrm{N}=204)$. It showed a clear mode around 34 cm . which is a decrease of 4 cm compared to the last three years. All samples were analyzed by the Greenland Institute of Natural Resources, and it was found that $S$. mentella constituted $76 \%$ of the total sample weight (Figure 23.3.6.2). In both species a mode was seen between $34-36 \mathrm{~cm}$. and for $S$. marinus an additional increase in frequency was seen at $45-50 \mathrm{~cm}$.

### 23.4 Methods

No analytical assessment was conducted.

### 23.5 Reference points (Benchmark, WKRED)

There are no biological reference points defined for this stock. However, part of the benchmark in 2012 (WKRED) was to evaluate the possible use of a stock production model in generating a quantitative advice for this stock. Under certain assumptions and for various intrinsic growth rates (r), current sustainable yields (and MSY) were calculated using the German survey and landings as input data. Across the range of r's, results seemed robust (CV range: 0.03-0.17), and the current sustainable yield was estimated at approximately 3.5 Kt . However, this procedure was criticized at the
benchmark due to lack of coverage of redfish distribution in the survey and questionable landings, and it is stated in the benchmark report that: "The panel does not suggest that the Schaefer model approach used here is to be final; to the contrary it is offered as a first step (from which interim management advice might be formulated)". As there are doubts on stock structure, species determination (and hence catch data accuracy), migration and the quality of the surveys used as basis for the model approach, the applicability of the proposed reference points from WKRED is questionable. Indeed, the use of a stock production model on an aggregation of fish that is not clearly defined as a stock is questionable.

### 23.6 State of the stock

The German survey and the Greenland shrimp and fish shallow water survey both show overall declines in the S. mentella biomass since 2010 (Figure 23.2.7, and both show a complete absence of small fish since $2013(<18 \mathrm{~cm})$. The adult stock decline is caused by a large decline in a small area which coincides with the main fishing area. The directed fishery CPUE for this area has declined from $3.7 \mathrm{t} / \mathrm{h}(2010)$ to $1.3 \mathrm{t} / \mathrm{h}$ suggesting a large local decline. Changes in length distributions in both surveys also suggests that no new cohorts are present on the slope and that the adult biomass decline is caused by the gradual decline of a single/few cohorts. Especially the complete absence of juveniles is cause for concern.

The biomass estimate declines and the concentrated fishery could point to a fishery induced decline. However, the declines are of a magnitude that seems beyond what a limited number of years catches can cause. Hence, surveys may either overestimate the biomass in especially Q3, not survey the entire area of distribution or S. mentella is disappearing due to migration. Survey overestimation may result from the large aggregations of redfish in Q3, which may cause two different survey scenarios, a low-density and high-density situation. If large redfish aggregations changes the catchability, the assumptions of linearity between catch and abundance are rendered invalid - high fish concentration may simply reduce the trawl escape potential. Such a situation would produce disproportionally high catches and subsequently biomass estimates in high density areas such as Q3. Hence, the decline may be a synergetic effect of a reduced biomass caused by the local fishery, and the reduced catchability inferred from the less dense fish aggregations following some years of intense fishing. This is further complicated by the lack of knowledge on the stocks connection to the pelagic (deep and shallow) and Icelandic slope stocks and the degree of migration. Based on this, care must be taken when evaluating stock status, but nevertheless, the consistency in both the German and shallow Greenlandic surveys suggests that the biomass has decreased, especially in area Q3, but the magnitude of the decline is probably not attributable to the fishery alone. Also, the apparent lack of juveniles in all the East Greenland area means that no new fish will grow into the fishable part of the stock for at least 6-8 years, and there is reason for concern. Fproxy values based on surveys indicate an increasing fishing mortality in recent years.

The advice is based on the Data Limited Stock approach (DLS) including biomass indices from the Greenland shallow water survey in the most recent 5 years combined with the recent advice, applying a cap and a precautionary buffer. The ratio is applied to the 2014 advice as catches are well above current advice. The advice for 2016 is 2240 t.

### 23.7 Management considerations

S. mentella is a slow growing, late maturing deep-sea species and is therefore considered vulnerable to overexploitation and advice has to be conservative. The fact that the fishery is targeting a localized aggregation of fish is cause for concern as is the absence of juveniles in the area. Given the biology of the species and the uncertainty in the biomass trend, any advice should consider this a hotspot fishery as it is potentially detrimental to this local and potentially important aggregation of larger fish. The fishery should still be at a low level involving few vessels. This should be maintained until the effect of the fishery can be clarified, especially with the recent declines in biomass estimates (Figures 23.2.1 and 23.2.6) and the fishery should preferably cover a larger area.

### 23.8 References

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Table 23.3.1.1 Nominal landings (tonnes) of demersal S. mentella 1974-2013 ICES division XIVb.

## Demersal S.mentella

| 1974 | 0 |
| :---: | :---: |
| 1975 | 4400 |
| 1976 | 59700 |
| 1977 | 0 |
| 1978 | 5403 |
| 1979 | 5131 |
| 1980 | 10406 |
| 1981 | 19391 |
| 1982 | 12140 |
| 1983 | 15207 |
| 1984 | 9126 |
| 1985 | 9376 |
| 1986 | 12138 |
| 1987 | 6407 |
| 1988 | 6065 |
| 1989 | 2284 |
| 1990 | 6097 |
| 1991 | 7057 |
| 1992 | 7022 |
| 1993 | 14828 |
| 1994 | 19305 |
| 1995 | 819 |
| 1996 | 730 |
| 1997 | 199 |
| 1998 | 1376 |
| 1999 | 853 |
| 2000 | 982 |
| 2001 | 901 |
| 2002 | 2109 |
| 2003 | 446 |
| 2004 | 482 |
| 2005 | 267 |
| 2006 | 202 |
| 2007 | 226 |
| 2008 | 92 |
| 2009 | 895 |
| 2010 | 6613 |
| 2011 | 6705 |
| 2012 | 6572 |
| 2013 | 6597 |
| 2014 | 4608 |

Table 23.3.3.2 Landings (tons) of demersal redfish caught in ICES XIVb by nation. By far the largest proportion were probably $S$. mentella but none of these amounts were converted by the mentella/marinus ratio $(80 \%$ S. mentella) found by the two surveys covering the area.

| Year | DEU | ESP | EU | FRO | GBR | GRL | ISL | NOR | POL | RUS | UNK | Sum |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1999 |  |  |  |  |  |  |  |  |  |  | 853 | 853 |
| 2000 | 884 |  | 11 |  |  | 19 | 65 |  | 3 | 982 |  |  |
| 2001 | 782 |  |  |  | 11 | 9 |  | 99 |  |  |  | 901 |
| 2002 | 1703 |  |  | 48 | 16 | 246 | 29 | 32 |  | 36 |  | 2109 |
| 2003 | 3 | 2 | 2 | 20 | 155 | 232 |  | 32 |  |  | 446 |  |
| 2004 | 5 | 1 | 79 | 12 | 221 | 93 | 68 | 3 |  | 482 |  |  |
| 2005 | 2 |  | 4 | 38 | 96 | 72 | 56 |  |  | 267 |  |  |
| 2006 | 1 |  |  |  |  | 152 | 48 |  |  | 202 |  |  |
| 2007 | 7 |  | 15 | 138 |  | 35 |  | 30 |  |  |  | 226 |
| 2008 | 1 |  | 8 | 50 | 5 | 5 |  | 23 |  |  | 92 |  |
| 2009 |  |  | 203 |  | 822 |  | 93 |  |  | 1118 |  |  |
| 2010 | 10 |  | 12 | 381 |  | 5672 |  | 2190 | 1 |  | 8266 |  |
| 2011 | 1262 |  | 26 | 2 |  | 6757 |  | 334 |  | 1 |  | 8381 |
| 2012 | 1810 |  | 5 | 32 |  | 5964 | 1 | 403 |  | 1 |  | 8216 |
| 2013 | 1957 |  |  | 32 | 30 | 5863 |  | 356 |  | 8 |  | 8246 |
| 2014 | 1973 | 0.2 | 13 |  | 4611 | 98 | 613 |  | 5 |  | 7314 |  |
| Sum | 10400 | 3 | 162 | 969 | 534 | 30552 | 128 | 4442 | 3 | 52 | 856 | 47248 |

Table 23.3.4.1 Discarded by-catch (tons) of Sebastes sp. from the shrimp fishery in ICES XIVb

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec | Sum |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1999 | 6 | 16 | 17 | 5 | 1 | 13 | 2 | 48 | 22 | 30 | 40 | 33 | 234 |
| 2000 | 10 | 3 | 31 | 17 | 15 | 4 | 21 | 78 | 28 | 18 | 9 | 6 | 239 |
| 2001 | 7 | 9 | 10 | 16 | 9 | 11 | 4 | 5 | 3 | 3 | 28 | 6 | 111 |
| 2002 | 3 | 11 | 9 | 6 | 1 | 0 | 0 | 5 | 4 | 8 | 3 | 5 | 55 |
| 2003 | 5 | 6 | 8 | 5 | 5 | 8 | 8 | 15 | 2 | 10 | 12 | 4 | 88 |
| 2004 | 7 | 10 | 17 | 13 | 4 | 2 | 27 | 20 | 7 | 2 | 9 | 0 | 118 |
| 2005 | 7 | 14 | 16 | 8 | 7 | 5 | 6 | 21 | 14 | 4 | 5 | 20 | 126 |
| 2006 | 6 | 2 | 4 | 1 | 3 | 5 | 2 | 4 | 4 | 0 | 0 | 4 | 35 |
| 2007 | 7 | 3 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 14 |
| 2008 | 0 | 2 | 2 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 7 |
| 2009 | 1 | 2 | 11 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 16 |
| 2010 | 1 | 2 | 2 | 1 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 2 | 10 |
| 2011 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 2012 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| 2013 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| 2014 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sum | 60 | 81 | 131 | 75 | 48 | 49 | 71 | 196 | 84 | 75 | 106 | 81 | 1056 |

Table 23.3.4.2 Landings (tons) of demersal redfish caught in ICES XIVb by month. By far the largest proportion were probably $S$. mentella but none of these amounts were converted by the mentella/marinus ratio $(80 \%$ S. mentella) found by the two surveys covering the area

| Year | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep | Oct | Nov | Dec | Sum |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1999 |  | 10 |  | 108 |  | 4 | 42 | 10 | 15 | 34 | 481 | 149 | 853 |
| 2000 | 18 | 238 | 286 | 260 | 10 | 4 | 79 | 72 | 13 | 0 | 3 |  | 982 |
| 2001 |  | 1 |  |  |  | 108 | 2 |  | 184 | 369 | 236 | 901 |  |
| 2002 | 183 | 445 | 354 | 390 | 50 | 472 | 35 | 44 | 59 | 77 |  | 2109 |  |
| 2003 |  |  | 9 | 4 | 26 | 27 | 135 | 195 | 20 | 16 | 12 |  | 446 |
| 2004 |  |  | 35 | 41 | 63 | 75 | 48 | 64 | 96 | 25 | 35 | 482 |  |
| 2005 |  | 1 | 15 | 66 | 24 | 80 | 29 | 13 | 18 | 19 |  | 267 |  |
| 2006 |  | 3 | 7 | 50 | 14 | 39 | 20 | 61 | 2 | 1 | 1 | 2 | 202 |
| 2007 | 6 | 13 | 8 | 8 | 14 | 42 | 4 | 106 | 16 | 7 | 1 | 1 | 226 |
| 2008 | 4 | 3 | 1 | 6 | 12 | 11 | 31 | 12 | 10 | 2 |  |  | 92 |
| 2009 |  |  |  | 1 | 84 | 346 | 148 | 105 | 128 |  | 288 | 17 | 1118 |
| 2010 | 799 | 786 | 708 | 1058 | 2149 | 2100 | 108 | 134 | 88 | 301 | 36 |  | 8266 |
| 2011 | 419 | 1396 | 1661 | 1017 | 268 | 250 | 236 | 598 | 255 | 583 | 1223 | 475 | 8381 |
| 2012 | 899 | 2197 | 628 | 852 | 577 | 699 | 966 | 143 | 44 | 23 | 474 | 712 | 8215 |
| 2013 |  |  | 709 | 1290 | 925 | 1423 | 1218 | 1086 | 723 | 227 | 119 | 527 | 8246 |
| 2014 | 10 | 421 | 206 | 1210 | 1187 | 1709 | 231 | 401 | 376 | 448 | 632 | 479 | 7314 |
| Sum | 2155 | 5250 | 4670 | 6268 | 5763 | 6791 | 3953 | 3037 | 1811 | 1999 | 3760 | 2633 | 48100 |



Figure 23.2.1. Indices from the German East Greenland survey of S. mentella larger than 17 cm . Abundance (a), biomass (b), and biomass split on length (c). On figure (c) the grey bars represent the biomass of $S$. mentella larger than 30 cm and the light bars biomass in fish from 17-30 cm.


Figure 23.2.2. Length distributions from the German East Greenland survey 1985-2014.


Figure 23.2.3. Biomass of $S$.mentella and Sebastes sp derived from the deep Greenland survey. Bars indicate 2SE of the biomass of S. mentella including Sebastes sp.. No survey in 2001. In 2004, 2005 and 2007 a large proportion of the redfish were not determined to species and only reported as "Sebastes sp". It is most likely that the majority of these fish were S. mentella.


Figure 23.2.4. Overall length distribution of Sebastes mentella (number per $\mathbf{k m}^{2}$ ) from the deep Greenland survey.


Figure 23.2.5: Biomass ( Kt ) indices for S. mentella (left) and Sebastes sp . $(<18 \mathrm{~cm})$ off East Greenland in 2008-2014 from the Greenlandic shallow water survey. All surveyed areas (Q1-Q6) are combined.


Figure 23.2.6. Overall length distributions for juvenile redfish S.mentella and S.norvegicus $<17 \mathrm{~cm}$ combined (left) and S. mentella $>17 \mathrm{~cm}$ from the Greenlandic shallow water survey. All surveyed areas combined (Q1-Q6).


Figure 23.2.7: Biomass indices from the German survey in ICES XIVb and the Greenland shallow water survey combined with historical catches.


Figure 23.2.8: Fproxy developed for the German survey excluding years with poor coverage (1990, 1992 and 1994) and the Greenland shallow survey.


Figure 23.3.1.1 Landings of S. mentella in subarea XIVb. Landings of "redfish" have been split based on estimates from survey and commercial catches


Figure 23.3.2.1 Standardized redfish by-catch CPUE in the directed fishery for Greenland halibut in ICES XIVb as a function of year. CPUE was estimated from the GLM model: $\ln C P U E=y e a r+I C E S$ subdivision+depth. Bars represent standard error. Only hauls made below 1000 m were used in the analyses.


Figure 23.3.2.2 Standardized redfish CPUE in the redfish directed fishery ICES XIVb as a function of year. CPUE was estimated from the GLM model: lnCPUE=year+ICES subdivision+depth. Dashed lines represent standard error.


Figure 23.3.3.1 Distribution of catches of demersal redfish in 2014.


Figure 23.3.3.2. Lines represent the share of the total commercial catch caught at a given depth from 1999-2011 in G. morhua, demersal redfish (mixed S. mentella and S. marinus) and R. hippoglossoides.


Figure 23.3.6.1: Length distribution of Sebastes sp. in the commercial catches from 2009-2014. In 2009-2011 the measurements were conducted onboard the trawlers by inspectors that were unable to separate S. mentella from S. marinus.


Figure 23.3.6.2: Length distribution of 752 redfish analysed by the Greenland Institute of Natural Resources separated into S.mentella ( $\mathbf{N}=548$ ) and S.norvegicus $(\mathbf{N}=\mathbf{2 0 4})$.

## Annex 1: List of Participants

North-Western Working Group
28 April - 05 May 2015

| Name | Address | Phone/Fax | Email |
| :---: | :---: | :---: | :---: |
| Rasmus Hedeholm <br> (Chair) | Greenland Institute for Natural Resources $\text { P.O. Box } 570$ GL-3900 Nuuk <br> Greenland | $\begin{aligned} & \text { Phone }+299361200 \\ & \text { Fax }+299361212 \end{aligned}$ | rahe@natur.gl |
| Petur Steingrund | Faroe Marine Research Institute P.O. Box 3051 FO-110 Tórshavn Faroe Islands | $\begin{aligned} & \text { Phone }+298353900 \\ & \text { Fax }+298353901 \end{aligned}$ | peturs@hav.fo |
| Sergey Melinkov | Russian Federal <br> Research Institute of Fisheries \& Oceanography (VNIRO) <br> 17 Verkhne <br> Krasnoselskaya <br> 107140 Moscow <br> Russian Federation | $\begin{aligned} & \text { Phone +7 (495) } \\ & 2644583 \\ & \text { Fax +7 (499) } 26431 \\ & 87 \end{aligned}$ | melnikov@vniro.ru |
| Alexander Bobyrev | Russian Federal <br> Research Institute of <br>  <br> Oceanography <br> (VNIRO) <br> 17 Verkhne <br> Krasnoselskaya <br> 107140 Moscow <br> Russian Federation |  | abobyrev@mail.ru |
| Søren Post | Greenland Institute for Natural Resources P.O. Box 570 GL-3900 Nuuk Greenland | $\begin{aligned} & \text { Phone +299 } 361200 \\ & \text { Fax +299 } 361212 \end{aligned}$ | sopo@natur.gl |
| Rasmus Nygaard | Greenland Institute for Natural Resources P.O. Box 570 GL-3900 Nuuk Greenland | $\begin{aligned} & \text { Phone }+299361200 \\ & \text { Fax }+299361212 \end{aligned}$ | rany@natur.gl |
| Höskuldur <br> Björnsson | Marine Research Institute PO Box 1390 121 Reykjavík Iceland | $\begin{aligned} & \text { Phone }+354575 \\ & 2000 \\ & \text { Fax }+3545752001 \end{aligned}$ | hoski@hafro.is |


| Jesper Boje | The National Institute of Aquatic Resources Section for Fisheries Advice Charlottenlund Slot, Jægersborg Alle 1 DK-2920 Charlottenlund Denmark | $\begin{aligned} & \text { Phone }+45339634 \\ & 64 \\ & \text { Fax }+4533963333 \end{aligned}$ | jbo@aqua.dtu.dk |
| :---: | :---: | :---: | :---: |
| Luis Ridao Cruz | Faroe Marine Research Institute P.O. Box 3051 FO-110 Tórshavn Faroe Islands | $\begin{aligned} & \text { Phone }+298353900 \\ & \text { Fax +298 } 353901 \end{aligned}$ | Luisr@hav.fo |
| Heino Fock | Johann Heinrich von <br> Thünen-Institute, <br> Institute for Sea <br> Fisheries <br> Palmaille 9 <br> D-22767 Hamburg <br> Germany | $\begin{aligned} & \text { Phone }+494038905 \\ & 169 \\ & \text { Fax }+494038905 \\ & 263 \end{aligned}$ | heino.fock@vti.bund.de |
| Einar Hjörleifsson | Marine Research Institute Skúlagata 4 IS-121 Reykjavík Iceland | $\begin{aligned} & \text { Phone +354 } 552 \\ & 0240 \\ & \text { Fax +354 } 5623790 \end{aligned}$ | einarhj@hafro.is |
| Kristjan <br> Kristinsson | Marine Research Institute Skúlagata 4 IS-121 Reykjavík Iceland | $\begin{aligned} & \text { Phone +354 } 575 \\ & 2000 \\ & \text { Fax +354 } 5752091 \end{aligned}$ | krik@hafro.is |
| Arni Magnusson | Marine Research Institute PO Box 1390 IS-121 Reykjavík Iceland | $\begin{aligned} & \text { Phone +354575 } \\ & 2000 \\ & \text { Fax +354 } 5752001 \end{aligned}$ | arnima@hafro.is |
| Gudmundur J. <br> Oskarsson | Marine Research Institute Skúlagata 4 IS-121 Reykjavík Iceland | $\begin{aligned} & \text { Phone }+354575 \\ & 2000 \\ & \text { Fax +354 } 5752001 \end{aligned}$ | gjos@hafro.is |
| Jákup Reinert | Faroe Marine Research Institute P.O. Box 3051 FO-110 Tórshavn Faroe Islands | $\begin{aligned} & \text { Phone }+298353900 \\ & \text { Fax }+298353901 \end{aligned}$ | jakupr@hav.fo |
| Anja Retzel | Greenland Institute for Natural Resources P.O. Box 570 GL-3900 Nuuk Greenland | Phone +299 361200 <br> Fax +299 361212 | AnRe@natur.gl |
| Alexey Rolskiy | Knipovich Polar Research Institute of | $\begin{aligned} & \text { Phone +7 } 815245 \\ & 0568 \end{aligned}$ | rolskiy@pinro.ru |


|  | Marine Fisheries and Oceanography | $\begin{aligned} & \text { Fax }+781524733 \\ & 31 \end{aligned}$ |  |
| :---: | :---: | :---: | :---: |
|  | 6 Knipovitch Street |  |  |
|  |  |  |  |
|  | Russian Federation |  |  |
| Gudmundur <br> Thordarson | Marine Research Institute PO Box 1390 121 Reykjavík Iceland | $\begin{aligned} & \text { Phone }+354 \\ & 5752000 \\ & \text { Fax }+3545752001 \end{aligned}$ | gudthor@hafro.is |
| Elena Guijarro Garcia | Instituto Español de Oceanografía Centro Oceanográfico de Vigo P.O. Box 1552 E-36200 Vigo <br> Spain | $\begin{aligned} & \text { Phone }+34986462 \\ & 201 \\ & \text { Fax +34 } 986498626 \end{aligned}$ | elena.guijarro@vi.ieo.es |
| Agnes Gundersen | Møreforsking Marin, PO Box 5075, 6021 Ålesund, Norway | $\begin{aligned} & +4770111621(00) \\ & /+4792611524 \end{aligned}$ | agnes@mfaa.no |
| Asta <br> Gudmundsdóttir | Marine Research Institute PO Box 1390 IS-121 Reykjavík Iceland | $\begin{aligned} & \text { Phone }+354- \\ & 5752000 \\ & \text { Fax }+354-5752001 \end{aligned}$ | asta@hafro.is |
| Birkir Bardarson | Marine Research Institute PO Box 1390 IS-121 Reykjavík Iceland | $\begin{aligned} & \text { Phone }+354- \\ & 5752000 \\ & \text { Fax }+354-5752001 \end{aligned}$ | birkir@hafro.is |

## Annex 2: Recommendations

1) For the attention of WGMG, the NWWG recommends that: The present model used to assess the Greenland halibut in V,VI, XII and XIV, a stock production model in Bayesian framework, should be reviewed with respect to performance. A number of issues was identified in the NWWG 2014 report, a.o. process error handling and autocorrelation.
2) For the attention of WKARGH and WGBIOP, the NWWG recommends that: Ageing of Greenland halibut has been resumed since 2000 for assessment of the Greenland halibut stock in V,VI,XII and XIV. WKARGH in 2011 made progress that could timely be followed up by a new workshop in 2016 or 2017.

## Annex 3: ToRs for the Next Meeting

## NWWG - North-Western Working Group

The North-Western Working Group (NWWG), chaired by Rasmus Hedeholm, Greenland, will meet at ICES Headquarters, 26 April - 3 May, 2016 to:
a) Address generic ToRs for Regional and Species Working Groups.

The assessments will be carried out on the basis of the stock annex in National Laboratories, prior to the meeting. This will be coordinated as indicated in the table below.

For capelin in Iceland-East Greenland-Jan Mayen area,. NWWG will agree any changes to the WG type report and the Advice sheet no later than X May. An ADG will work by correspondence XX May. The WEBEX will be XX May, and the Advice Release date XX May.

Other material and data relevant for the meeting must be available to the group no later than 14 days prior to the starting date.

NWWG will report by XX May 2015 for the attention of ACOM.

## Annex 4: List of Stock Annexes

A list of the stock annexes (including direct hyperlinks) will be placed here as soon as the work on the annexes is finished.

## Annex 5: Audit reports

## Audit of Cod in Division Va (Icelandic cod; cod-iceg)

Date $4^{\text {st }}$ of May 2015
Reviewer: Anja Retzel

## General

The stock has been assessed in close agreement with the stock annex.

## For single stock summary sheet advice:

1) Assessment type: update
2) Assessment: analytical
3) Forecast: presented
4) Assessment model: statistical catch-at-age (ADCAM) tuned with two (spring and fall) surveys.
5) Data issues: All data is available as described in the Stock Annex.
6) Consistency: SPALY assessment was consistent with last year's assessment.
7) Stock status: Blim is 125 kt , MSY Btrigger is 220 kt and SSB in 2015 is estimated at 547 kt . Reference biomass, (B4+) is estimated at 1302 kt in 2015. PA and MSY reference points have not been set for this stock.
8) Man. Plan.: Because SSB> Btrigger, the $\mathrm{TAC}_{2015 / 2016}$ is set as (TAC $2014 / 2015+$ $0.2^{*} \mathrm{~B}_{\mathrm{B} 4+2015)} / 2$. In accordance with this plan, the proposed TAC for 2015/2016 is 239 kt . According to the advice sheet, ICES has evaluated the plan and concludes that it is in accordance with the precautionary approach and the ICES MSY framework.

## General comments

This was a well documented, well ordered and considered section. It was easy to follow and interpret.

## Technical comments

None

## Conclusions

The assessment has been performed in as close proximity to the Stock annex and the results can be used as basis for advice.

## Audit of Golden redfish (Sebastes norvegicus) in Subareas V, VI and XIV.docx

Date 5 of May 2015
Reviewer: Alexey Rolskiy

## General

A comprehensive review of scientific, fishery and assessment data. Some corrections are needed.

## For single stock summary sheet advice:

1) Assessment type: update
2) Assessment: analytical
3) Forecast: presented
4) Assessment model: the stock was benchmarked in January 2014. Gadget model. Alternative assessment: Time Series Analysis (TSA)
5) Data issues: All data is available as described in the Stock Annex. The assessment indicates conflicting stock trends from different data sources, the catch-at-age data showing less increase in stock size than the survey data.
6) Consistency: Consistent with last year.
7) Stock status: Blim is 160 kt , MSY B $\mathrm{B}_{\text {triger }}$ is 220 kt and SSB in 2015 is estimated at 347 kt . Biomass ( $\mathrm{B}_{5}$ ) is estimated at 557 kt in 2015. PA reference points have not been set for this stock.
8) Man. Plan.: Management plan evaluated and adopted in 2014 (WKREDMP, ICES 2014). According to management plan catches in 2016 should be no more than 51000 tonnes ( $\mathrm{F}_{9-19}=0.097$ ).

## General comments

The results from GADGET indicate that fishing mortality has reduced in recent years and is now close to Fmsy Spawning stock and fishable stock have been increasing in recent years and are now the highest since 1986. Results from surveys in Iceland and East Greenland indicate that most recent year classes are poor.

## Technical comments

None

## Conclusions

The assessment has been performed in as close proximity to the Stock annex and the results can be used as basis for advice. Further work should be invested into the data sources used for the assessment.

## Audit of Faroe Bank Cod (cod-farb)

Date 5 May 2015
Auditor: Guðmundur J. Óskarsson

## For single stock summary sheet advice:

1) Assessment type: Update
2) Assessment: trends
3) Forecast: not presented
4) Assessment model: None
5) Data issues: None, in relation to Stock Annex
6) Consistency: Same advice as in last year, the area closed for fishery, which has been followed
7) Stock status: Poor status of the stock, the biomass indices below average and no signs of improvements
8) Man. Plan.: There is no management plan for this stock

## General comments

The message in the text is clearly put forward and the graphs illustrative. However, the structure and subheadings of the Assessment report is a bit different from other stocks, and should be improved and made more along the lines with others at some point (e.g. during benchmark assessment). For example, the section "Status of the stock" is normally few sentences but fills 1.5 pages out of total 3 pages for this stock. Most of the text there should be under different sections (e.g. "Landings","Information from Surveys", "Exploratory assessment" etc).

## Technical comments

I did not see any errors in the report and the assessment was done according to the Stock Annex.

## Conclusions

The assessment has been performed correctly

## Checklist for review process

## General aspects

- Has the EG answered those TORs relevant to providing advice? -Yes
- Is the assessment according to the stock annex description? -Yes
- Is general ecosystem information provided and is it used in the individual stock sections. -Not provided in the stock report, and therefore not used
- If a management plan has been agreed, has the plan been evaluated? -Not relevant


## For update assessments

- Have the data been used as specified in the stock annex? -Yes
- Has the assessment, recruitment and forecast model been applied as specified in the stock annex? -Not relevant, it is surveys trends based assessment
- Is there any major reason to deviate from the standard procedure for this stock? No
- Does the update assessment give a valid basis for advice? If not, suggested what other basis should be sought for the advice? -Yes, the update assessment gives a valid basis for advice.


## Annex 7: List of Working Documents. (NWWG 2015)

Guðmundur J.Ó. and Páll R.. 2015. Results of acoustic measurements of Icelandic summerspawning herring in the winter 2014/2015. ICES NWWG 2015 Working Document no. 01.
Guðmundur J.Ó. and Jónbjörn P. 2015. Estimation on number-at-age of the catch of Icelandic summer-spawning herring in 2014/2015 fishing season and the development of Ichthyophonus hoferi infection in the stock. ICES NWWG 2015 Working Document no. 02.
Hedeholm R. and Boje J. 2015. Survey for Greenland halibut in ICES Division 14B, August September 2014. ICES NWWG 2015 Working Document no. 03.
Steingrund P. 2015. Greenland halibut CPUE for the research vessel operating on the slope on the Faroe Plateau in May-June 1995-2014. ICES NWWG 2015 Working Document no. 04.

Steingrund P. 2015. Greenland halibut CPUE for commercial trawlers operating on the slope on the Faroe Plateau 1991-2014. ICES NWWG 2015 Working Document no. 05.
Steingrund P. 2015. Survey biomass indices of Greenland halibut on the slopes of the Faroe Plateau 1983-2014. ICES NWWG 2015 Working Document no. 06.

Steingrund P. 2015. A combined biomass index of Greenland halibut on the slopes of the Faroe Plateau 1983-2014. ICES NWWG 2015 Working Document no. 07.

Steingrund P. 2015. Greenland halibut in the Faroese September deepwater survey 2014. ICES NWWG 2015 Working Document no. 08.
Steingrund P. 2015. Maturity status of Greenland halibut in Faroese waters as observed in the Greenland halibut May-June survey 1995-2014. ICES NWWG 2015 Working Document no. 09.

Kristinsson K. 2015. Fishery of Golden Redfish (Sebastes norvegicus) in ICES Divisions Va in 2014. ICES NWWG 2015 Working Document no. 10.
Kristinsson K. 2015.Golden Redfish (Sebastes norvegicus) in ICES Division Va as observed in groundfish surveys. ICES NWWG 2015 Working Document no. 11.

Kristinsson K. 2015. The Fishery of Icelandic Slope Deep-Water Redfish (Sebastes mentella) in ICES Division Va. ICES NWWG 2015 Working Document no. 12.

Kristinsson K. 2015. Icelandic Slope Beaked Redfish (Sebastes mentella) in ICES Division Va as Observed in the Icelandic Autumn Survey 2000-2014. ICES NWWG 2015 Working Document no. 13.

Kristinsson K. 2015. The fishery for shallow and deep pelagic deep-water redfish (Sebastes mentella) in the Irminger Sea and adjacent waters. ICES NWWG 2015 Working Document no. 14.

Kristinsson K. 2015. Golden redfish and slope beaked redfish catch statistics in ICES Divisions Va, Vb and XIV 1906-2014. ICES NWWG 2015 Working Document no. 15.

Post S. L., Retzel A. and Hedeholm R. 2015 West Greenland inshore gillnet survey results for juvenile Atlantic cod in 2014. ICES NWWG 2015 Working Document no. 16.

Boje J. and Hedeholm R. 2015. The fishery for Greenland halibut in ICES Div. XIVb in 2014. ICES NWWG 2015 Working Document no. 17.
Retzel A. 2015. Greenland commercial data for Atlantic cod in Greenland inshore waters for 2014. ICES NWWG 2015 Working Document no. 18.

Retzel A. 2015. Greenland Shrimp and Fish survey results for Atlantic cod in NAFO subareas 1A-1E (West Greenland) in 2014. ICES NWWG 2015 Working Document no. 19.
Retzel A. 2015. Greenland commercial data for Atlantic cod in West Greenland offshore waters for 2014. ICES NWWG 2015 Working Document no. 20.

Retzel A. and Post S. 2015. Greenland Shrimp and Fish survey results for Atlantic cod in ICES subarea XIVb (East Greenland) and NAFO subarea 1F (SouthWest Greenland) in 2014. ICES NWWG 2015 Working Document no. 21.

Retzel A. 2015. Greenland commercial data for Atlantic cod in East Greenland offshore waters for 2014. ICES NWWG 2015 Working Document no. 22.
Popov V. and Rolskiy A. 2015. Preliminary information on the results of Russian fishery and biological samples of pelagic redfish from the ICES subarea XII, XIV and NAFO Div. 1F in 2014. ICES NWWG 2015 Working Document no. 23.

Hedeholm R., Nygaard R. and Boje J. 2015. The fishery for demersal Redfish (S.mentella) in ICES Div. XIVb in 2014. ICES NWWG 2015 Working Document no. 24.

Hedeholm R., Nygaard R. and Boje J. 2015. Greenland Shrimp and Fish Survey Results for Redfish in East Greenland Offshore Waters in 2014. ICES NWWG 2015 Working Document no. 25.

Bobyrev . and Vasilyev. 2015. The assessment of stock status of redfish, Sebastes mentella, in the Irminger Sea. ICES NWWG 2015 Working Document no. 26.

Boje J., Hvingel C. and Hedeholm R. 2015. An assessment of Greenland halibut (Reinhardtius hippoglossoides) off East Greenland, Iceland and the Faroe Islands. ICES NWWG 2015 Working Document no. 27.

Kristinsson K. 2015. Methodology of biomass estimation of deep pelagic beaked redfish (Sebastes mentella) from the international redfish survey in the Irminger Sea and adjacent waters. ICES NWWG 2015 Working Document no. 28.
Fock H. 2015. Update of Groundfish Survey Results for the Atlantic Cod Greenland offshore component After re-stratification of the survey 1982-2014. ICES NWWG 2015 Working Document no. 29.

Fock H. 2015. An age4plus index for the Atlantic Cod Greenland offshore component Based on recommendations from WKICE 2015 1989-2014. ICES NWWG 2015 Working Document no. 30.

Fock H., Stransky C. and Bernreuther. 2015. Abundance for Sebastes norvegicus L., deep sea S. mentella and juvenile redfish (Sebastes spp.) off Greenland based on groundfish surveys 1985-2014. ICES NWWG 2015 Working Document no. 31.

Steingrund P. 2015 A simple reconstruction of the stock size of Faroe Plateau cod back to 1906. ICES NWWG 2015 Working Document no. 32.

## Annex 7: Russian statement

## Russian statements regarding the stock assessment, influence in environmental conditions on pelagic redfish distribution and estimates biomass during the surveys in the Irminger Sea

## Statement regarding Russian assessment of beaked redfish, Sebastes mentel/a, stock status

According to the opinion of Russian experts, survey-based assessment of redfish stock biomass is not efficient enough due to the limited data involved into the analysis and some methodological problems related with surveys design and data processing.

Russian fisheries scientists do not agree with the stock subdivision into different management units and assess stock status using consolidated data. Assessment is based on Russian fisheries statistics and biological data, which appear to be sufficient for applying analytical methods of stock assessment such as surplus-production models and cohort models.

The results of stock assessment using a surplus-production model were presented at NWWG meetings in 2013 (WD 24) and 2014 (WD 28). The conclusion was made that redfish stock is in rather safe condition and relatively stable. The management strategy, based on the MSY concept and the precautionary approach, was implemented through a nonlinear (sigmoid) harvest control rule. With this HCR, the relevant advised TAC for 2015 was estimated as 100 thou. tones (with MSY $=185$ thou. tones and BMSY about 1700 thou. tones).
At the NWWG meeting in 2015 (WD 26), the results of the cohort model TISVPA application for redfish stock assessment have been presented. In the analysis, the data on total annual redfish catches by all countries were used with the division by age-groups following from Russian data on age composition over the period 1990-2013. In the calculations, the age interval starting from 8-years class was used; the natural mortality coefficient was assumed to be equal to 0.1 year $^{-1}$ for all age classes. The data on standardized catch per unit of effort (cpue) were involved as an index of stock biomass. In Figure 1, the reconstructed dynamics of stock biomass (left panel) and fishing mortality (right panel) are shown.

The estimates obtained suggest that current level of redfish stock biomass exceeds 1000 thou. tones, with fishing mortality being at the minimum level. It should be noted that the lack of reliable data on redfish maturation rate impedes estimation of spawning stock biomass (SSB). Due to the same reason, it is not possible, as yet, to optimize the HCR using the results from the TISVPA model. To this end, the estimation of fisheries perspectives is still based on the data from the surplus-production model, and recommended TAC remains at the last year level, i.e., 100 thou. tones.


Figure 1. Estimates of redfish stock biomass (left) and average fishing mortality (right) resulting from the TISVPA model

## Statement regarding the impact of variations in environmental conditions on estimates S. mentel/a biomass during the surveys in the Irminger Sea by results of analysis of SST satellite monitoring data

According opinion of the Russian experts, survey-based assessment of redfish alone does not reflect actual state of the stock due to some methodical difficulties associated with surveys i.e. limited number of trawl stations, changes in survey methodology e.t.c.
In this connection for the first time used the satellite data from meteorological satellites during the studies of environmental long-term fluctuations' influence on distribution and evaluation of redfish biomass.

There have been analytical calculations of the average values of sea surface temperature (SST) for the reference zone, integral acoustic values (SA) for the reference zone, and the average values of the SA only for those places of the reference zone where the aggregations of redfish were found in the layer 0-500 m.

Obtained results demonstrates strong correlation between SST and average values of redfish density in upper layer of the Irminger Sea. (Figure 1, 2). Revealed: the lower the value of SST is in on the lower integral and average values on density of redfish. In other words, the cooling of the sea surface temperature leads to a decrease in the density of redfish aggregations above 500 m , and accordingly, to a decrease in the biomass estimated during international survey in 2001-2013.

Revealed strong correlation between SST and average values of redfish density in upper layer of the Irminger Sea. The cooling of the sea surface temperature leads to a decrease in the density of redfish aggregations above 500 m . In contrast, analysis of oceanographic observations within the different depth layers suggests reduction of redfish density in upper layer with increasing of water temperature in the intermediate layers above 500 m depth. Both cases (variations in SST or changes in water temperature on the depth of redfish distribution) provide the evidence of the influence of oceanographic conditions on vertical (Melnikov et al., 2009) and spatial distribution of redfish aggregations (Riboni et al., 2013). It could be the one of the main reason of stock underestimation during the surveys.
The obtained results once again confirm the decisive role of fluctuations in environmental conditions on estimates of redfish biomass during the survey, rather than the impact of fishing on the stock.

## References

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Riboni I.N., Kristinsson K., Bernreuther M., Hendrik M., Stransky C., Cisewski B., Rolskiy A. 2013. Impact of interannual changes of large-scale circulation and hydrography on the spatial distribution of beaked redfish (Sebastes mentella) in the Irminger Sea, Deep Sea Research Part I: Oceanographic Research Papers, Vol. 82, P. 80-94


$$
Y=61,54 X-398,18 \quad R^{2}=0,54 \quad R=0,73
$$

Figure 1. Comparison quasi-synchronous average values of SST and integral values of SA depended on the distribution and abundance of redfish in reference zone of the Irminger Sea


Figure 2. Comparison quasi-synchronous average values of SST and average density of aggregations redfish (in SA m${ }^{2} /$ mile $^{2}$ ) in reference zone of the Irminger Sea for $1^{\circ} \times 1^{\circ}$ (only for places its discovery)

## Statement about influence of the oceanographic condition on pelagic redfish distribution in the Irminger Sea

According to the international survey results, in 1999-2007 strong positive anomalies of water temperature in the upper 0-200 m layer of the Irminger Sea and adjacent waters were observed. This led to redistribution of part redfish aggregation (above 500 m ) to southwestern part of the survey area (ICES, 2007). Surveys observations are also consistent with fishery development, which considerably spread since 1999 in the southwestern direction, covering Divisions 1F, 2GHJ of the NAFO area (Figure 1).

Later influence of oceanographic condition on abundance and distribution of pelagic redfish was studied during ICES Workshop on Redfish and Oceanographic Conditions (WKREDOCE) with the primary objective to compile and evaluate available hydrographical, hydroacoustic, and trawl data from the Irminger Sea and adjacent waters. A study examining changes in the distribution of redfish over 20 years revealed that at the interannual time scale, the spatial distribution of S. mentella in the Irminger Sea mainly in waters < 500 m appears strongly influenced by the Irminger Current Water (ICW) temperature changes, linked to the Subpolar Gyre (SPG) circulation and the North Atlantic Oscillation (NAO). Acceleration of the SPG due to increase of the NAO index leads to displacement of part redfish aggreagitons in the southwest. An SPG weakening has the opposite effect. A decrease in the NAO index strength since 2008 and the present deceleration of the SPG suggest a subsequent northeast displacement of part of the redfish aggreagitons in upper layer northward in the coming years (ICES, 2012, Riboni et. al, 2013).

In our opinion, these year-to-year northeast/southwest displacement of the redfish (above 500 m depth) influenced by the interannual hydrographic changes in the Irminger Sea - the one of the reason of periodic transformation of the fishery pattern. These fluctuations make impractical using of current management unit boundaries, which are based on spatial patterns of the fishery only (ICES, 2009).

## References

ICES. 2007. Report of the Study Group on Redfish Stocks (SGRS), 31 July - 2 August 2007, Hamburg, Germany. ICES CM 2007/RMC:12. 54 pp.

ICES. 2009. Report of the Workshoip on Redfish Stock Structure (WKREDS), 22-23 January 2009, ICES Headquarters, Copenhagen. Diane. 71 pp.

ICES. 2013. Report of the Working Group on Redfish Surveys (WGRS), 6-8 August 2013, Hamburg, Germany. ICES CM 2013/SSGESST:14. 56 pp.
ICES. 2012. Report of the Second Workshop on Redfish and Oceanographic Conditions, 16-17 August 2012, Johann Heinrich von Thunen Institute, Hamburg, Germany. ICES CM 2012/ACOM:25. 70 pp.

Riboni I.N., Kristinsson K., Bernreuther M., Hendrik M., Stransky C., Cisewski B., Rolskiy A. 2013. Impact of interannual changes of large-scale circulation and hydrography on the spatial distribution of beaked redfish (Sebastes mentella) in the Irminger Sea, Deep Sea Research Part I: Oceanographic Research Papers, Vol. 82, P. 80-94.


Figure 1. Fishing areas and total catch of pelagic redfish (S. mentella) in 1999-2014, derived from catch statistics provided by Russia. The colour scale indicates catches (tonnes per NM ${ }^{2}$ )


[^0]:    * Asterisk indicates missing value(s).

[^1]:    * Asterisk indicates missing value(s).

[^2]:    * Includes both the catches and the herring that died in the mass mortality in the winter 2012/13 in Kolgrafafjörður ${ }^{{ }^{\prime}}$

[^3]:    * Number at age 3 in 2015 is predicted from an survey index of number at age 1 in 2012 (see section 11.6.1)

[^4]:    * preliminary

[^5]:    1) Provisional data
[^6]:    ${ }^{1}$ Provisional data
    ${ }^{2}$ Based on estimates by observers onboard vessels

[^7]:    1) Provisional
