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## Pacific Northwest National Laboratory

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# Proposed Occupational Exposure Limits for Non-Carcinogenic Hanford Waste Tank Vapor Chemicals

TS Poet  
C Timchalk

March 2006



Prepared for the U.S. Department of Energy  
under Contract DE-AC05-76RL01830

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Waste Tank Vapor Chemicals**

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Pacific Northwest National Laboratory  
Richland, Washington 99352

## Summary

A large number of volatile chemicals have been identified in the headspaces of tanks used to store mixed chemical and radioactive waste at the U.S. Department of Energy (DOE) Hanford Site, and there is concern that vapor releases from the tanks may be hazardous to workers. The tank farm operations contractor, CH2M Hill Hanford Group, Inc., is contractually obligated to meet occupational exposure limits (OELs) established by the Occupational Safety and Health Administration (OSHA) and American Conference of Governmental Industrial Hygienists (ACGIH). Some of the vapors detected in the tank headspaces have OSHA and/or ACGIH OELs, but not all. To address the need for worker exposure guidelines for those chemicals that lack OSHA or ACGIH OELs, a procedure for assigning Acceptable Occupational Exposure Limits (AOELs) for Hanford Site tank farm workers has been developed and applied to a selected group of 57 headspace chemicals.

The 57 chemicals considered in this report were identified by a toxicological screening process (Poet et al. 2006) and reviews of the analytical data on which the existence of the chemicals in the tanks were based (e.g., Sklarew and Mitroshkov 2006). AOELs developed by this report are considered “proposed” until reviewed and formally adopted by the Environmental Exposure Assessment Review Group.

The proposed AOELs were preferentially based on OELs established by U.S. or foreign agencies and organizations other than OSHA and ACGIH. When other such OELs were available, they were evaluated for use as AOELs. Four chemicals were assigned AOELs on this basis (Table S.1). For the remaining 53 chemicals without established OELs, epidemiological and toxicological information was reviewed to identify potential hazards, select critical effects, and estimate dose-response to determine exposure levels expected to be safe. When information on the chemical itself was insufficient to develop an AOEL, the proposed AOEL was based on one or more structurally related chemicals (surrogates) for which sufficient data were available. Surrogates with close structural and pharmacological activity with established OELs (ACGIH or OSHA) were preferentially selected. Of these 53 chemicals, 47 were assigned proposed AOELs (Table S.2), and six were determined to not warrant complete AOEL development (Table S.3).

**Table S.1.** OELs for Chemicals with Limits from NIOSH, AIHA or MAK

Chemical	CAS <sup>(a)</sup> Number	Proposed AOEL (ppm)	OEL Source
Propanenitrile	107-12-0	6	REL <sup>(b)</sup>
2-Ethyl-1-hexanol	104-76-7	20	MAK <sup>(c)</sup>
Butanal	123-72-8	25	WEEL <sup>(d)</sup>
Butanenitrile	109-74-0	8	REL

(a) CAS = Chemical Abstracts Service.  
 (b) REL - Recommended Exposure Limits, National Institute for Occupational Safety and Health (NIOSH)  
 (c) MAK- Maximum Arbeitsplatz Konzentration, German Commission for the Investigation of Health Hazards of Chemical Compounds in the Work Area (MAK Commission)  
 (d) WEEL - Workplace Environmental Exposure Limits, American Industrial Hygiene Association (AIHA)

**Table S.2.** Summary of Proposed AOELs

Chemical	CAS <sup>(a)</sup> Number	Proposed AOEL (ppm)
<b>HALOCARBONS</b>		
2-Fluoropropene	1184-60-7	0.1
<b>ALCOHOLS</b>		
Cyclopentanol	96-41-3	5
<b>ETHERS</b>		
1-Methoxybutane	628-28-4	17
<b>ALDEHYDES</b>		
2-Methylbut-2-enal	1115-11-3	0.03
2-Ethylhex-2-enal	645-62-5	0.1
<b>KETONES</b>		
4-Methyl-2-hexanone	105-42-0	0.5
3-Hexanone	589-38-8	67
6-Methyl-2-heptanone	928-68-7	8
2-Nonanone	821-55-6	17
3-Dodecanone	1534-27-6	17
2-Tridecanone	593-08-8	17
3-Tridecanone	1534-26-5	17
5-Methyl-2-(1-methylethenyl)cyclohexanone	89-82-7	2.5
4,7,7-Trimethylbicyclo[4.1.0]heptan-3-one	4176-04-9	0.7
3-Methyl-3-buten-2-one	814-78-8	0.02
<b>ESTERS</b>		
Dibutyl butylphosphonate	78-46-6	0.02
<b>NITRILES</b>		
2-Methylene butanenitrile	1647-11-6	0.3
Pentanenitrile	110-59-8	6
Hexanenitrile	628-73-9	6
Heptanenitrile	629-08-3	6
Octanenitrile	124-12-9	6
Nonanenitrile	2243-27-8	6
Decanenitrile	1975-78-6	6
3-Butenenitrile	109-75-1	1
2,4-Pentadienenitrile	1615-70-9	0.3
<b>AMINES &amp; AMIDES</b>		
N-Methylaziridine	1072-44-2	0.05
<b>NITRITES &amp; NITRATES</b>		
Methyl nitrite	624-91-9	0.1
Ethyl nitrite	109955	0.1
Butyl nitrite	544-16-1	0.1
Methyl nitrate	598-58-3	8
Ethyl nitrate	625-58-1	8
Butyl nitrate	928-45-0	8
1,3-Propanediol, dinitrate	3457-90-7	0.05
1,4-Butanediol, dinitrate	3457-91-8	0.05
1,5-Pentanediol, dinitrate	3457-92-9	0.05
1,2,3-Propanetriol, 1,3-dinitrate	623-87-0	0.05

**Table S.2. (contd)**

Chemical	CAS <sup>(a)</sup> Number	Proposed AOEL (ppm)
1,2,3-Propanetriol, 1-nitrate	624-43-1	8
<b>NITRO COMPOUNDS</b>		
2-Nitro-1-propanol	2902-96-7	0.8
2-Nitro-2-methylpropane	594-70-7	0.3
1-Nitrobutane	627-05-4	2.5
<b>HETEROCYCLICS</b>		
Furan	110-00-9	0.001
2-Methylfuran	534-22-5	0.001
2-Propylfuran	4229-91-8	0.001
2-Ethyl-5-methylfuran	1703-52-2	0.001
2,5-Dihydrofuran	1708-29-8	0.001
alpha-Propylfuranacetaldehyde	31681-26-2	0.001
2,4-Dimethylpyridine	108-47-4	0.5
(a) CAS = Chemical Abstracts Service.		

**Table S.3. Chemicals For Which No AOEL is Recommended**

Chemical	CAS <sup>(a)</sup> Number
1-Hexadecanol	36653-82-4
1-Octadecanol	112-92-5
1-Methylethyl tetradecanoate	110-27-0
Butyl tetradecanoate	110-36-1
1-Methylethyl hexadecanoate	142-91-6
Acetamide	60-35-5
(a) CAS = Chemical Abstracts Service.	

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## Acronyms and Abbreviations

AEGL	Acute Exposure Guideline Levels (EPA)
ACGIH	American Conference of Governmental Industrial Hygienists
AIHA	American Industrial Hygienist Association
AOEL	Acceptable Occupational Exposure Limit
BMD	Benchmark dose
BMDL	BMD limit
CAS	Chemical Abstracts Service
CSF	cancer slope factor
DOE	U.S. Department of Energy
EPA	U.S. Environmental Protection Agency (references to particular U.S. state or foreign country EPAs are so noted)
HSDB <sup>®</sup>	Hazardous Substances Data Bank
IARC	International Agency for Research on Cancer
IDLH	Immediately Dangerous to Life or Health
LC <sub>50</sub>	lethal concentration for 50% of test population
LD <sub>50</sub>	lethal dose for 50% of test population
LClo	lethal Concentration low; the lowest concentration at which death occurred
MAK	1) German Commission for the Investigation of Health Hazards of Chemical Compounds in the Work Area (MAK Commission), 2) the maximum permissible concentration of a chemical compound present in the air within a working area, which according to current knowledge, does not impair the health of the employee or cause undue annoyance.
MF	modifying factors
NIEHS	National Institute of Environmental Health Sciences
NIH	National Institutes of Health
NIOSH	National Institute for Occupational Safety and Health
NLM	National Library of Medicine
NOELs	No-Observed-Effect Levels

NTP	National Toxicology Program
OEL(s)	Occupational Exposure Limit(s)
OSHA	Occupational Safety and Health Administration
PCBs	polychlorinated biphenyls
PEL	Permissible Exposure Limit
POD	point of departure
ppm	parts per million
REL	Recommended Exposure Limits
RfC	reference concentration
RfD	reference dose
RTECS <sup>®</sup>	Registry of Toxic Effects of Chemical Substances
SCAPA	DOE Subcommittee on Consequence Assessment and Protective Actions
STEL	short-term exposure limit
TClo	lowest published toxic concentration
TEEL	Temporary Exposure Limit
TLV <sup>®</sup>	threshold limit value (ACGIH-specific)
TWA	time-weighted average
TOXNET <sup>®</sup>	The Toxicology Data Network
TWINS	Tank Waste Information Network System
UF	Uncertainty factor
WEEL	Workplace Environmental Exposure Limit
WHO	World Health Organization

## 1.0 Introduction

A large number of volatile compounds have been identified in the headspaces of single- and double-shelled tanks used to store mixed chemicals and radioactive waste at the U.S. Department of Energy (DOE) Hanford Site. Concern about potential exposure of tank farm workers to vapors has prompted efforts to evaluate the potential health risk associated with exposure to these chemicals. However, occupational exposure limits (OELs) have not been established for many of these chemicals.

This report describes the process that was used to assign Acceptable Occupational Exposure Limits (AOELs) to non-carcinogenic chemicals and families of chemicals that had no previously established OELs but were identified in the tank waste headspace at or above preliminary screening values. An AOEL is defined as the concentration in air expected to lead to no adverse health effects that is acceptable to management, industrial hygiene professionals, and workers at the Hanford Site. These chemicals are listed and discussed in Section 5.0.

Permissible Exposure Limits (PELs) are established by the Occupational Health and Safety Administration (OSHA) of the U.S. Department of Labor and are legal obligations for defined industries. The National Institute for Occupational Safety and Health (NIOSH) is an arm of the Centers for Disease Control (CDC) and makes recommendations to OSHA regarding OEL values. The American Conference of Governmental Industrial Hygienists (ACGIH) is a private organization that recommends Threshold Limit Values (TLVs) to industry for voluntary application. Other human exposure standards have been published, e.g., Emergency Response Planning Guidelines and Workplace Environmental Exposure Levels (WEELs) have been established by the American Industrial Hygiene Association (AIHA), and Acute Exposure Guideline Levels have been developed by the U.S. Environmental Protection Agency (EPA). However, their applicability to occupational exposures must, in each case, be evaluated. The DOE mandates the need to comply with the OSHA and ACGIH standards in its contract with CH2M HILL Hanford Group, Inc.

The 57 chemicals considered in this report were identified by a toxicological screening process (Poet et al. 2006) and reviews of the analytical data on which the existence of the chemicals in the tanks were based (e.g., Sklarew and Mitroshkov 2006). AOELs developed by this report are considered “proposed” until reviewed and formally adopted by the Environmental Exposure Assessment Review Group.

## 2.0 Overview of Approach

The National Academy of Sciences (NAS 1983, 1994; GAO 2001) has provided overall guidance for chemical risk assessment as further developed and applied by U.S. regulatory agencies. The process for establishing an AOEL for tank waste chemical exposure was modeled upon the generalized scientific approaches used by OSHA and ACGIH to establish OELs for worker exposure. In addition, where appropriate, the scientific approach used by other regulatory agencies, such as the EPA, to establish reference concentrations (RfC) for airborne pollutants were considered.

The key steps in the assessment process are shown in Figure 2.1. The approach first required identifying specific chemical agents or classes of chemical agents that were detected in the tank headspace at or above previously established screening values. Chemical agents that exceed the screening values underwent a more extensive evaluation to establish an AOEL for tank farms operations. First, relevant OELs were reviewed in detail. Without appropriate exposure guidelines, available epidemiology and toxicology information on a given chemical or chemical class were reviewed to identify potential hazards, select critical effects, and estimate a dose-response to determine suitable exposure levels

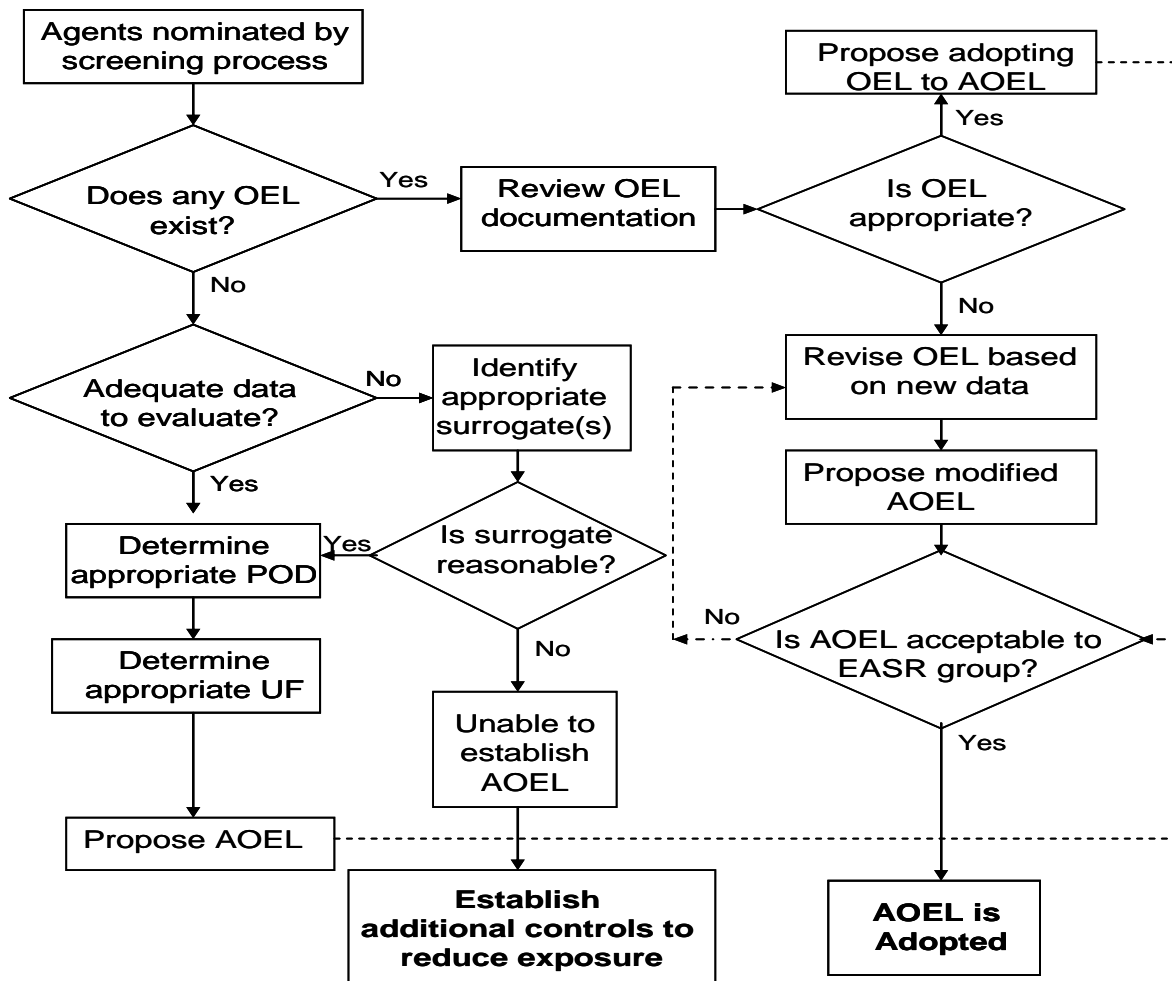


Figure 2.1. Flow Diagram Outlining Processes for Establishing an AOEL

(Haber and Maier 2002). Consistent with the technical approaches ACGIH uses for establishing a Threshold Limit Value (TLV), the AOEL was based on the best scientific information available and included a critical evaluation of all supporting information (ACGIH 2005a). The AOEL review delineated the most important adverse effects. In evaluating the health effects, human data such as epidemiology studies are of prime importance; however, many of the chemicals of concern had little, if any, available human data. The second tier evaluation focused on the most relevant and sensitive animal toxicity data for the chemical and/or chemical class. Here, the basis for the AOEL is the dose-response relationship for the toxic effects of greatest concern, and the point of departure (POD) for calculating the AOEL is the associated no-observed-effect-level/lowest-observed-effect-levels (NOEL/LOEL). The rationale for the selection of a given toxic effect for the AOEL is documented.

For setting an AOEL, an appropriate risk assessment approach was used that was selected based on the critical toxicological effect, the observed dose-response, and the quality of the data used for the assessment (ACGIH 2005a). This included the use of quantitative dose-response models such as the benchmark dose (BMD) and/or the application of added factors to a NOEL/LOEL to address uncertainty.

Before establishing and implementing a formal AOEL for worker protection in the tank farms, the proposed exposure level for a given chemical and/or chemical class was submitted to the Exposure Assessment Strategy Review Group for peer review. The peer reviewers included a range of health and science professionals, including representatives from occupational medicine, toxicology, risk assessment, industrial hygiene, health physics, and worker groups. The review critically assessed the rationale for establishing the AOEL, provided the broad base of stakeholders an understanding of the rationale for the AOEL, and provided stakeholders an opportunity to submit appropriate input in the assessment.

## **2.1 Data Base and Literature Evaluation**

### **2.1.1 Data Base Searches**

It is imperative to consider relevant human exposure, epidemiology, and toxicological information in establishing AOELs. A methodical analysis of the available literature as it relates to hazard identification and quantitative dose-response toxicity evaluation is central to the AOEL assessment process. Internet data bases such as The Toxicology Data Network (TOXNET<sup>®</sup>), TOMES<sup>®</sup>, PUBMED<sup>®</sup>, the International Agency for Research on Cancer (IARC), and/or STN<sup>®</sup>, were utilized as primary sources for initiating searches, which include both the name and Chemical Abstracts Service registry number of the compound of interest (See Table 2.1). These data bases contain information applicable to toxicological assessment of chemicals, and the information they provide may overlap. For example, both TOXNET and TOMES contain the Registry of Toxic Effects of Chemical Substances (RTECS<sup>®</sup>) and Hazardous Substances Data Bank (HSDB<sup>®</sup>) files. When no RTECS and/or HSDB files existed for a compound or were considered insufficient, the TOXNET literature and/or PUBMED databases were searched and the original literature evaluated for any relevant information. If relevant information was not located in this manner, PUBMED was searched directly and/or the chemical abstracts database was searched through STN. If information was found in one or more of the above sources but was incomplete, conflicting or considered insufficient, other sources such as National Institute of Environmental Health Sciences (NIEHS) National Toxicology Program (NTP), IARC, the U.S. EPA and the Agency for Toxic Substances and Disease Registry

**Table 2.1.** Internet Data Bases

TOXNET <sup>®</sup>	The Toxicology Data Network, a set of data bases covering toxicology, hazardous chemicals, and related areas; it is maintained by the National Library of Medicine (NLM). ( <a href="http://toxnet.nlm.nih.gov/">http://toxnet.nlm.nih.gov/</a> )
HSDB <sup>®</sup>	Hazardous Substances Data Bank. Accessible through TOXNET. Provided by the NLM.
PUBMED <sup>®</sup>	PubMed, provided by the NLM, contains citations for biomedical articles back to the 1950s; sources include MEDLINE and additional life science journals. ( <a href="http://www.ncbi.nlm.nih.gov/entrez/query.fcgi?db=PubMed">http://www.ncbi.nlm.nih.gov/entrez/query.fcgi?db=PubMed</a> ).
RTECS <sup>®</sup>	Registry of Toxic Effects of Chemical Substances, provided by Thomson Micromedex, Inc. Accessed through TOMES <sup>®</sup> , which has the same provider.
STN <sup>®</sup>	The Scientific & Technical Information Network, provided by FIZ Karlsruhe, the American Chemical Society, and the Japan Science and Technology Corporation. It links to published research in the world's journal and patent literature back to the beginning of the 20th century.
TOMES <sup>®</sup>	Registry of Toxic Effects of Chemical Substances, provided by Thomson-Micromedex, Inc.

Monographs were also searched. Frequently, more in-depth information of the chemical of interest or its surrogate was located in this way and was useful in assigning an AOEL.

### **2.1.2 Literature Evaluation of Published Exposure Guidelines**

Initial efforts focused on a comprehensive evaluation of the available literature to identify any published exposure guidelines for the chemical of interest or for a reasonable surrogate. The summary documentation used to support the establishment of the OEL was reviewed to ascertain its relevance, and any AOEL adopted based on these published OELs included appropriate documentation justifying the scientific rationale for its use.

### **2.1.3 Use of Surrogate OELs**

If there were no appropriate exposure guidelines for the chemical of interest, it was deemed reasonable to identify and assign an OEL based on a structurally related chemical (surrogate). In addition to being structurally related, ideal surrogates also had a similar toxicological profile (i.e., similar target organs and response), although they may display greater or lesser potency than the chemical of concern.



## 3.0 Toxicology Review

### 3.1 Toxicity Review Criteria

In evaluating the toxicity profile for a chemical or chemical class, the assessment was prioritized based on identification of toxic effects that are particularly relevant for occupational exposure. This would include:

- Identifying chemical agents that have a high acute toxicity potential such as Category 1 chemicals as defined by the Organization for Economic Co-operation and Development (OECD 2004). Chemicals in Category 1, considered by OECD to have high acute toxicity potential, have LD<sub>50</sub> and LC<sub>50</sub> values of ≤ 5 mg/kg oral or ≤ 100 ppm inhalation, respectively. Headspace chemicals such as these are of particular concern if the LD<sub>50</sub>/LC<sub>50</sub> values do not have large safety margins (<100) between observed toxicity and potential exposure levels.
- Giving priority to the evaluation of studies that utilize the most relevant occupational exposure routes (inhalation > oral > dermal > others [intravenous/ intraperitoneal]).
- Giving priority to well-characterized dose-response toxicity studies, particularly those that included a more comprehensive evaluation of the toxicity, which includes quantitative in vivo (acute → sub-chronic → chronic) experiments.

The toxicity testing paradigm suggests that with increasing length of chemical exposures (i.e., acute → chronic) the effective dose levels generally decrease such that the lowest NOEL/LOEL will be determined from long-term studies. In addition, long-term chronic studies can identify chemicals that have the potential to produce a broad range of chronic health effects. An important strength of this testing paradigm is that it provided a fairly comprehensive in vivo toxicological evaluation that could be used to link dose-response results across a very broad range of exposure routes (oral vs. inhalation), doses (low vs. high), durations (sub-chronic vs. chronic), and species (rat vs. dog). The use of this type of testing data provided greater confidence (i.e., less uncertainty) in the AOEL that was established.

### 3.2 Minimal Data Sets

As discussed by Haber and Maier (2002), a number of regulatory authorities including the Health Council of the Netherlands (HCN 2000) and the EPA (1994) have established minimum data requirements based on the rationale that any value derived from data that is less than the minimum prescribed would have too much uncertainty. An alternative approach is to identify the types of data that are particularly useful and utilize a weight-of-evidence approach in the evaluation that looks at the sum total of all available information (SCOEL 1999; Haber and Maier 2002). In this regard, a weight-of-evidence review for the development of an AOEL focused on evaluating the types of studies that had greater relevance to occupational exposure and were amendable to establishing a dose-response relationship.

For this AOEL assessment process, a minimum data set was not established, and a weight-of-evidence approach was used. Due to the anticipated lack of robust toxicity data, a minimum data set was hard to achieve for the broad number of chemical agents being evaluated, particularly where there was a need to apply surrogate chemical data in the assessment, which was often the case. The AOEL documentation and the peer reviewers provided the means to assess the degree of confidence that was placed on the

AOEL. It is important to recognize that with the unique chemistry associated with chemical/radioactive tank waste, appropriate toxicity data on the chemicals of interest or their reasonable surrogates was often absent, making the establishment of a defensible AOEL problematic. In these cases, the confidence in the AOEL was assessed and depended upon the availability of additional toxicity data from the literature.

### **3.3 Use of Surrogate Toxicity Data**

When forced to use a surrogate chemical or chemical class for developing an AOEL, it was important that the surrogate had as robust a toxicology database as possible. Using a surrogate chemical with a well characterized toxicity database ensured it had a clear dose-response relationship and a clearly defined NOEL to use as a POD for establishing an AOEL. Substantially more uncertainty was assigned when utilizing surrogates that lacked a robust toxicity database. A written assessment of the overall strengths and weaknesses of the surrogate chemical was included within the AOEL documentation (Section 5).

## 4.0 Procedure for Calculating an AOEL

### 4.1 Point of Departure

The overarching goal in evaluating the toxicity data bases is to determine a POD for developing an AOEL. Haber and Maier (2002) defined the POD as the concentration to which uncertainty factors are applied to derive an AOEL. The POD is most likely the NOEL/LOEL that was determined from the most appropriate toxicity study. In practice, this is usually the highest NOEL or lowest LOEL that was experimentally derived. It is also possible to utilize a BMD approach, as described below, to determine a POD, particularly when the experimental studies did not identify a NOEL (Fillipsson et al. 2003; Haber and Maier 2002; Crump 1984; Dourson et al. 1985).

### 4.2 Approach for Developing Exposure Limits for Non-Cancer Effects

As suggested by Bailey (2002) and others, there are numerous sources of uncertainty in the establishment of an acceptable exposure level. The approach used for identifying an acceptable exposure level for the general population or for occupational-related exposures is to adjust the NOEL or LOEL downward. The magnitude of the downward adjustment reflects the degree of uncertainty concerning the acceptable exposure limits. To address these uncertainties, empirical factors may be used to account for inadequate experimental data, interspecies variability, human variability, or extrapolation for short-term to long-term studies (EPA 2002; Dorne and Renwick 2005). In addition to uncertainty factors, additional modifying factors (MF) have been used occasionally by some regulatory agencies, such as EPA, to reflect uncertainties not addressed by other factors. Equation 1 was used to calculate the AOELs.

$$\text{AOEL} = \frac{\text{NOEL or LOEL}}{\text{UF} \times \text{MF}} \quad (4.1)$$

The application of empirical uncertainty factors to determine the AOEL was based on the methods utilized by EPA for deriving an RfC (EPA 2002). The default uncertainty factor (UF) generally covers a single order of magnitude (i.e., 10), or a value of 3 is used in place of one-half powers (i.e., 100.5). Additional factors could also be considered for inadequate data and for extrapolation from less than lifetime-to-lifetime exposures. As EPA suggest, supporting documentation included the justification used for the individual factors selected. In addition, as EPA (2002) recommends, the total UF applied to any particular chemical was limited to no more than 3000 and deriving exposure guidelines that involved the full 10-fold UF in four or more areas of extrapolation was not done. The following uncertainty and modifying factors were applied when necessary:

- **Extrapolation from animal data to humans (interspecies UF).** This factor is intended to account for the uncertainty in extrapolating animal data to the case of an average healthy human. It assumes that humans are more susceptible to the toxicity than the animal species evaluated.
- **Variability in the human population (intraspecies UF).** This factor is intended to account for the variation in sensitivity among humans.
- **LOEL to NOEL UF.** This factor is intended to address the uncertainty associated with extrapolation from LOELs to NOELs.

- **Sub-chronic to chronic duration UF.** This factor is intended to account for the uncertainty in extrapolating from less than chronic NOELs to chronic NOELs.
- **Inadequate data base UF.** This factor is intended to account for the inability of any single animal study to adequately address all possible adverse outcomes in humans.
- **Modifying factors (MF).** This factor is intended to account for any other scientific uncertainties in the study or data bases that are not explicitly treated by other UFs. The magnitude of the MF principally depends on professional judgment.

### 4.3 Establishing a NOEL Using Benchmark Dose (BMD)

As Dorne and Renwick (2005) note, using a BMD as proposed by Crump (1984) defines a lower statistical confidence for the dose, corresponding to a predefined low-level increase in adverse effects over background. The BMD approach provides a more quantitative alternative to the first step in the dose-response assessment than the NOEL/LOEL process (EPA 2000). This is particularly useful when a NOEL has not been adequately defined from the experimental data.

The approach utilized EPA's BMD software available on the internet (<http://www.epa.gov/ncea/bmds.htm>). The EPA guidance document (EPA 2000) provides a detailed discussion on a number of important considerations including the types of studies that are appropriate for BMD, selection of the benchmark response values, choice of models for computing BMD, and details concerning the computation of confidence limits for the BMD. Based on the EPA criteria, a 10% response is at or near the limit of sensitivity in most cancer and non-cancer bioassays and was used as an appropriate effective dose (ED10) for the BMD. As the EPA guidance (EPA 2000) indicates, the primary goal of the mathematical modeling is to fit a model to dose-response data, particularly at the low end of the observable dose-response range. The recommended criterion for selection of an appropriate model to obtain a Benchmark Dose Limit (BMDL) is the Akaike's Information Criterion. The Akaike's Information Criterion values were computed for each of the models used in the BMD calculation and compared the most appropriate model was selected from the analysis. Once a BMD value was selected, a lower confidence was placed on the BMDL that assured, with high confidence (95%), the benchmark response was not exceeded. The BMDL was then used in the numerator of Equation 1 to calculate an AOEL, as described above.

### 4.4 Format for this AOEL Document

The AOEL documentation for a given chemical or chemical class is included in the following sections. An Executive Summary communicates the overall process and conclusions of the analysis. A methodology section briefly describes the approach used for developing the AOEL, and this is followed by a section describing the Available Guidelines that are particularly relevant in developing the AOEL. The Toxicology Summary section reviews pertinent human epidemiology and animal toxicology results that are directly relevant to the setting of the AOEL. Rather than describing all the available data, this section focuses on the key studies and results pertinent to the evaluation. The final Data Analysis section describes the process used for setting the AOEL.

## 5.0 AOEL Documentation

### 5.1 Halocarbons

#### 5.1.1 Fluoropropene (CAS# 1184-60-7)

An AOEL is proposed for 2-fluoropropene based on a comprehensive review of the available toxicology and guideline literature. 2-Fluoropropene is a fluorinated propene that is similar in structure to halogenated ethylenes such as fluoroethylene (vinyl fluoride), chloroethylene (vinyl chloride), or bromoethylene (vinyl bromide). As a class of compounds, the halogenated ethylenes have been classified as known (A1) or suspected (A2) human carcinogens. Although there are no ACGIH or OSHA guidelines for 2-fluoropropene, the ACGIH has set TLV-Time Weighted Average (TWA) exposure concentrations for fluoroethylene at 1 ppm (1.9 mg/m<sup>3</sup>) largely based on analogy to the TLV-TWAs for chloroethylene (vinyl chloride) and bromoethylene (vinyl bromide) (ACGIH 2001bb). Based on the uncertainty associated with chemical differences between 2-fluoropropene and the halogenated ethylenes, an AOEL of **0.1 ppm** is proposed for 2-fluoropropene.

##### 5.1.1.1 Methodology

The Standard Operating Procedures described in Sections 2.0 - 4.0 were used as a guide for developing this AOEL. To identify and access available toxicity data on 2-fluoropropene, a number of Internet data bases, including PUBMED<sup>®</sup>, TOXNET<sup>®</sup>, and TOMES<sup>®</sup>, were searched and summary toxicity profiles were obtained, including the identification of primary references. In addition, ACGIH documentation was likewise reviewed for 2-fluoropropene and relevant surrogate chemicals. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profile and used for the development of the AOEL.

##### 5.1.1.2 Available Guidelines

No OEL for 2-fluoropropene has been proposed by OSHA or ACGIH. However, ACGIH has assigned a TLV-TWA for the structural surrogate fluoroethylene (ACGIH 2001n). The ACGIH documentation for fluoroethylene is based upon the structural similarity to the known human carcinogen vinyl chloride and the suspected human carcinogen vinyl bromide. However, it has been suggested that the slower relative metabolism of fluoroethylene vs. vinyl chloride to a similar toxic metabolite supports fluoroethylene being less potent than vinyl chloride (Filser and Bolt 1979). Likewise, a direct comparison of rodent bioassay results between fluoroethylene (Bogdanffy et al. 1995) and vinyl bromide (Benya et al. 1982) suggests that vinyl bromide is more than twice as potent as fluoroethylene. Therefore, ACGIH recommended a 1 ppm TLV-TWA for fluoroethylene, which is twice the recommended TLV-TWA (0.5 ppm) for vinyl bromide.

##### 5.1.1.3 Toxicology Summary and Data Analysis

There was no available toxicity information for 2-fluoropropene. A search of the RTECS data base found very limited and dated toxicity data on 2-chloropropene but suggested a low acute toxicity with an LC<sub>50</sub> in mice of 267 g/m<sup>3</sup> and a lowest published concentration (TClo) in rats of 191 g/m<sup>3</sup> following repeated inhalation exposure. For the structural surrogate fluoroethylene, the acute toxicity is very low with lethal

concentrations in rats reported to be >800,000 ppm for an ~12-hour exposure, and no fatalities were noted in rats exposed for 7 hours/day, 5 days/week for 30 days at 100,000 ppm fluoroethylene (Clayton 1967). A 13-week exposure of rats and mice to fluoroethylene concentrations ranging from 200 – 20,000<sup>(a)</sup> ppm reported a concentration-dependent increase in urinary fluoride excretion (Bogdanffy et al. 1990). The oncogenic potential of vinyl fluoride was evaluated in both rats and mice by inhalation at concentrations of 0, 25, 250 and 2500 ppm, and fluoroethylene was found to be carcinogenic at the lowest concentration (25 ppm) in both species (Bogdanffy et al. 1995). The tumorigenic response was reported to be similar to that seen with vinyl chloride and vinyl bromide (ACGIH 2001bb; Maltoni et al. 1981; Benya et al. 1982).

#### 5.1.1.4 Conclusion

The ACGIH has set the TLV-TWA for vinyl fluoride at 1 ppm (1.9 mg/m<sup>3</sup>) based primarily by analogy with vinyl chloride and vinyl bromide. Based on the uncertainty associated with chemical differences between 2-fluoropropene and vinyl fluoride, an additional 10-fold factor is applied to the vinyl fluoride OEL to establish an AOEL of **0.1 ppm** for 2-fluoropropene.

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(a) ACGIH 2001bb incorrectly identifies 2000 ppm as the maximum vapor concentration.

## 5.2 Alcohols

### 5.2.1 1-Hexadecanol (CAS # 36653-82-4), 1-Octadecanol (CAS # 112-92-5)

There are currently no published exposure guidelines for 1-hexadecanol (16-carbon) or 1-octadecanol (C-18). Both these long-chain aliphatic alcohols have relatively low toxicity based on acute toxicity testing. These alcohols are common components of cosmetic compounds. Octadecanol (stearyl alcohol) is found naturally in various mammalian tissues and is readily converted to stearic acid. Octadecanol is an 18-carbon straight chain aliphatic alcohol often used as an emollient to prevent drying and chapping of skin and as a thickener and pearling agent in cosmetics. The common use in cosmetics and relatively low toxicity indicate an AOEL is not needed for these chemicals at this time. As additional new data become available for these compounds, it would be reasonable to reassess the appropriateness of assigning an AOEL.

### 5.2.2 Cyclopentanol (CAS# 96-41-3)

An AOEL is proposed for cyclopentanol based on a comprehensive review of the available toxicology and guideline literature. Cyclopentanol is a cyclic (5-carbon) alcohol. Although there are no ACGIH or OSHA guidelines for cyclopentanol, the ACGIH has set TLV-TWA exposure concentrations for cyclohexanol (ACGIH 2001). The TLV-TWA for the cyclohexanol was set at 50 ppm (206 mg/m<sup>3</sup>) to minimize the potential for eye irritation and possible central nervous system effects including narcosis and incoordination. Based on the uncertainty associated with chemical and toxicological differences between cyclopentanol and cyclohexanol, an AOEL of **5 ppm** is proposed for cyclopentanol.

#### 5.2.2.1 Methodology

The procedures described in Sections 2.0 - 4.0 were used as a guide for developing this AOEL. To identify and access available toxicity data on cyclopentanol, a number of Internet data bases including PUBMED<sup>®</sup>, TOXNET<sup>®</sup>, and TOMES<sup>®</sup> were searched and summary toxicity profiles were obtained and primary references identified. ACGIH documentation was also reviewed for cyclohexanol, which is a relevant surrogate chemical. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profile and develop the AOEL.

#### 5.2.2.2 Available Guidelines

No OEL for cyclopentanol has been proposed by OSHA or ACGIH. However, ACGIH has assigned a TLV-TWA for the surrogate, cyclohexanol. The ACGIH documentation indicates that a TLV-TWA of 50 ppm will protect from cyclohexanol vapor irritation based on both animal and human data.

#### 5.2.2.3 Toxicology Summary and Data Analysis

Limited toxicity evaluation has been conducted with cyclopentanol. Acute oral exposure to doses of 0.6 – 1.3 ml/kg of body weight did result in mortality, and toxic response included unsteady gait and convulsion. Gross pathology indicated effects on the liver and kidneys. Acute inhalation exposure to a saturated atmosphere (~1300 ppm) for 1 hour resulted in no mortality; however, a ~3-hour exposure at this same concentration resulted in 100% mortalities.

The surrogate cyclohexanol has a low acute oral toxicity in rats with LD<sub>50</sub>s at ~2 g/kg of body weight. Inhalation exposure to ~1000 ppm for 6 hours/day, 5 days/week, for 5 to 11 weeks resulted in intoxication and death in 50% of rabbits. Rabbits exposed to concentrations of 145 ppm showed only slight degeneration in the liver and kidneys (Treon et al. 1943). In humans exposed to cyclohexanol (3-5 minutes), an 8-hour acceptable air concentration was calculated as < 100 ppm (Nelson 1943).

#### 5.2.2.4 Conclusion

The ACGIH has set the TLV-TWA for the cyclohexanol at 50 ppm (206 mg/m<sup>3</sup>) to minimize the potential for eye irritation and possible central nervous system effects including narcosis and incoordination. Based on the uncertainty associated with chemical and toxicological differences between cyclopentanol and cyclohexanol, an additional 10-fold factor is applied to the cyclohexanol OEL to establish an AOEL of **5 ppm** for cyclopentanol.

#### 5.2.3 2-Ethyl-1-hexanol (CAS# 104-76-7)

An AOEL is proposed for 2-ethyl-1-hexanol based on a comprehensive review of the available toxicology and guideline literature. 2-Ethyl-1-hexanol is an aliphatic branched (8-carbon) alcohol. Although there are no ACGIH or OSHA guidelines for 2-ethyl-1-hexanol, the German Commission for the Investigation of Health Hazards of Chemical Compounds in the Work Area (MAK Commission) set a MAK of 20 ppm (MAK set in 2005). In addition, the ACGIH has set a TLV-TWA exposure concentration for isooctyl alcohol, which has the same molecular formula but a slightly different structural formula than 2-ethyl-1-hexanol. The TLV-TWA for isooctyl alcohol was set at 50 ppm (266 mg/m<sup>3</sup>)<sup>(a)</sup> to minimize the potential for upper respiratory tract irritation (ACGIH 2001p). The German MAK of **20 ppm** will be adopted for 2-ethyl-1-hexanol.

##### 5.2.3.1 Methodology

The procedures described in Sections 2.0 - 4.0 were used as a guide for developing this AOEL. To identify and access available toxicity data on 2-ethyl-1-hexanol, a number of Internet data bases including PUBMED, TOXNET, and TOMES were searched and summary toxicity profiles were obtained and primary references identified. In addition, ACGIH documentation was likewise reviewed for isooctyl alcohol as a relevant surrogate chemical. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profile and develop the AOEL.

##### 5.2.3.2 Available Guidelines

No OEL for 2-ethyl-1-hexanol has been proposed by OSHA or ACGIH, and the German MAK commission has recently updated their OEL for 2-ethyl-1-hexanol from 50 ppm to 20 ppm. There is no available documentation explaining the reasons for the change (MAK set in 2005). In addition, ACGIH has assigned a TLV-TWA for a structural surrogate isooctyl alcohol (ACGIH 2001p). The ACGIH documentation indicates that a TLV-TWA of 50 ppm will minimize the potential for upper respiratory tract irritation.

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(a) ACGIH incorrectly applied units of mg/kg to this value.



### 5.2.3.3 Toxicology Summary and Data Analysis

There appears to be substantial toxicity data on 2-ethyl-1-hexanol, however no U.S. guidelines have yet been established for occupational exposure. In mice, rats, and guinea pigs, the LC<sub>50</sub> for 2-ethyl-1-hexanol was >227 ppm (6-8 hour exposure) (Scala and Burtis 1973; Smyth et al. 1969; Treon 1963). The oral LD<sub>50</sub> in rats reported in several studies was ~3 g/kg of body weight (Scala and Burtis 1973; Albro 1975; Treon 1963; NIOSH 1976). In these studies, the animals were observed for 24 hours prior to necropsy, and no deaths occurred during exposure or observation. All animals exhibited central nervous system depression, labored breathing, and one guinea pig had a clonic convulsion. The results of long-term oral carcinogenicity studies in both rats and mice indicate that 2-ethyl-1-hexanol is not carcinogenic (BASF 1991; 1992). Although the German MAK has established an OEL for 2-ethyl-1-hexanol (20 ppm), for comparison isooctyl alcohol was also chosen as a surrogate because it has the same molecular formula as 2-ethyl-1-hexanol, and an ACGIH TLV-TWA has been established and documented (ACGIH 2001p). An acute oral LD<sub>50</sub> for isooctyl alcohol in rats of 1.5 g/kg of body weight has been reported, and the principal signs of effect were central nervous system depression and labored respiration. In addition, mice, rats, and guinea pigs that inhaled 200 ppm isooctyl alcohol for 6 hours showed only moderate local irritation of the upper respiratory tract, but no signs of systemic intoxication were observed (Scala and Burtis 1973). These toxicological responses seen with isooctyl alcohol following either oral or inhalation exposure are reasonably comparable to the response seen with 2-ethyl-1-hexanol. Hence, isooctyl alcohol is a reasonable surrogate based on similar molecular structure and toxicological profile.

### 5.2.3.4 Conclusion

The German MAK commission has set an OEL TWA for 2-ethyl-hexanol at 20 ppm; however, we lack documentation explaining the rationale for the selection of the OEL. The ACGIH has set the TLV-TWA for isooctyl alcohol at 50 ppm (266 mg/m<sup>3</sup>) to minimize the potential for upper respiratory tract irritation. The German MAK of **20 ppm** will be adopted for 2-ethyl-hexanol. For perspective, this AOEL is 2.5-fold lower than the ACGIH TLV-TWA for the surrogate isooctyl alcohol.

## 5.3 Ethers

### 5.3.1 1-Methoxybutane (CAS# 628-28-4)

An AOEL is proposed for 1-methoxybutane (n-butyl methyl ether) based on a comprehensive review of the available toxicology and guideline literature. 1-methoxybutane has the same molecular formula ( $C_5H_{12}O$ ) and is similar in structure to methyl tert-butyl ether (MTBE). Although there are no ACGIH or OSHA guidelines for 1-methoxybutane, the ACGIH has set TLV-TWA exposure concentrations for MTBE at 50 ppm ( $180 \text{ mg/m}^3$ ) based upon extensive animal toxicity testing and human exposure data (ACGIH 2002a). Based on the uncertainty associated with slight chemical differences between 1-methoxybutane and MTBE, an AOEL of **17 ppm** is proposed for 1-methoxybutane.

#### 5.3.1.1 Methodology

The procedures described in Sections 2.0 - 4.0 were used as a guide for developing this AOEL. To identify and access available toxicity data on 1-methoxybutane, a number of Internet data bases including PUBMED, TOXNET, and TOMES were searched, and summary toxicity profiles were obtained and primary references identified. In addition, ACGIH documentation was likewise reviewed for 1-methoxybutane and relevant surrogate chemicals. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profile and develop the AOEL.

#### 5.3.1.2 Available Guidelines

No OEL for 1-methoxybutane has been proposed by OSHA or ACGIH. However, ACGIH has assigned a TLV-TWA of 50 ppm ( $180 \text{ mg/m}^3$ ) for the structural surrogate MTBE (ACGIH 2002a). The TLV-TWA for MTBE is based upon the observance of no adverse symptoms in humans (10 subjects) exposed to up to 50 ppm MTBE (Johanson et al. 1995) and a no-observed-adverse-effect level (NOAEL) for repeated inhalation exposure in rats at 800 ppm (Daughtrey et al. 1997; Lington et al. 1997) and a NOEL of 400 ppm in rats for a two-generation study (Bevan et al. 1997).

#### 5.3.1.3 Toxicology Summary and Data Analysis

A search of the RTECS data base found very limited toxicity data on 1-methoxybutane but suggested a low acute toxicity with an  $LC_{50}$  in mice of  $176 \text{ g/m}^3$  ( $48,800 \text{ ppm}$ ). The primary observed effect is general anesthesia. For the structural surrogate MTBE, acute toxicity is very low with lethal concentrations reported to be  $>35,000 \text{ ppm}$  in both mice and rats (Marsh and Leake 1950; Conaway et al. 1985). Since MTBE is used extensively as an octane-enhancing agent in gasoline, there is substantial exposure both occupationally and to the general public (ACGIH 2002a). Rats were repeatedly exposed to MTBE vapors 6 hours/day, 5 days/week for 13 weeks at 0, 800, 4000 and 8000 ppm (Lington et al. 1997). The authors reported no treatment-related mortalities, and the NOEL for repeated exposure was 800 ppm. In addition, MTBE has not been found to induce adverse effects in fetal development (rats, mice, and rabbits) or on reproduction (ACGIH 2002a; Caprino and Togna 1998). In chronic vapor inhalation studies, rats and mice were exposed to 0, 400, 3000 and 8000 ppm MTBE for  $>18$  months (Burleigh-Flayer et al. 1992; Chun et al. 1992). MTBE has been associated with an increase in hepatocellular carcinoma in female mice and increased renal cancer in male rats ( $\alpha_2$ -microglobulin nephropathy). Based upon the observed tumorigenic response in rodent studies, MTBE has been classified as an A3

Confirmed Animal Carcinogen with Unknown Relevance to Humans. A number of human in vitro and in vivo metabolism studies have been conducted with MTBE. Johanson et al. (1995) conducted a controlled human exposure study in which 10 adult male volunteers were exposed to 5, 25, and 50 ppm MTBE for 2 hours while conducting light exercise. This pharmacokinetic study reported that the kinetics were linear at all concentrations with no observed adverse symptoms in humans at concentrations up to 50 ppm.

#### **5.3.1.4 Conclusion**

The ACGIH has set the TLV-TWA for MTBE at 50 ppm (180 mg/m<sup>3</sup>). Based on the uncertainty associated with chemical differences between 1-methoxybutane and MTBE, an additional 3-fold factor is applied to the MTBE OEL to account for the limited 1-methoxybutane data to establish an AOEL of **17 ppm** for 1-methoxybutane.

## 5.4 Aldehydes

### 5.4.1 2-Methylbut-2-enal (CAS# 1115-11-3) and 2-ethylhex-2-enal (CAS# 645-62-5)

2-Methylbut-2-enal and 2-ethylhex-2-enal are acrolein derivatives, which contain a branched aliphatic sidechain with a double bond and oxygen on the C-1 carbon. The principal toxic effect that stems from exposures to aldehydes and aliphatic hydrocarbons is irritation to mucous membranes. Irritation caused by similar compounds with both a double bond and a double-bonded oxygen can be severe, and the ACGIH has recommended TLV-ceilings of 0.1 and 0.3 ppm for acrolein and crotonaldehyde, respectively. Very little information was found regarding 2-methylbut-2-enal and 2-ethylhex-2-enal, and proposed OELs were based on crotonaldehyde as a surrogate with an UF of 30 (2-methylbut-2-enal) or 100 (2-ethylhex-2-enal), depending on the availability of toxicology data. **The recommended OELs for 2-methylbut-2-enal and 2-ethylhex-2-enal are 0.03 and 0.1, respectively.**

#### 5.4.1.1 Methodology

The procedures described in Sections 2.0 - 4.0 were used as a guide for developing this AOEL. To identify and access available toxicity data on 2-methylbut-2-enal and 2-ethylhex-2-enal, a number of Internet data bases including PUBMED, TOXNET, and TOMES were searched and summary toxicity profiles were obtained and primary references identified. In addition, ACGIH documentation was likewise reviewed for crotonaldehyde, acrolein, and propionaldehyde as potentially relevant surrogate chemicals. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profile and develop the AOEL.

#### 5.4.1.2 Available Guidelines

No OELs have been proposed by OSHA or ACGIH for either 2-methylbut-2-enal or 2-ethylhex-2-enal. However, ACGIH has assigned TLV ceilings for the structural surrogates, crotonaldehyde and acrolein, and a TLV-TWA for propionaldehyde. Propionaldehyde lacks the double bond, and the ACGIH TLV-TWA is 20 ppm (ACGIH 2002b). The ACGIH TLV ceilings for crotonaldehyde and acrolein are 0.3 and 0.1 ppm, respectively (ACGIH 2001e; 2001a). The toxicological endpoint of concern for all of these aldehydes is irritation of the eyes and upper respiratory tract. The irritation caused by exposures to crotonaldehyde and acrolein are rapid and can be lethal due to changes in pulmonary function. Crotonaldehyde is an animal carcinogen (hepatocellular carcinomas).

#### 5.4.1.3 Toxicology Summary

No toxicity data were found for 2-methylbut-2-enal. 2-Methylbut-2-enal is a pheromone released in rabbit milk apparently needed by newborn rabbit pups to find the nipple (Schaal et al. 2003). The structural surrogate of 2-methylbut-2-enal lacking the double-bonded oxygen, 2-methyl-2-butene has a high LC<sub>50</sub> (low toxicity), and an inhalation NOEL in a 26-week study was 580 ppm (Olefins Panel 2004).

There are limited toxicity data available for 2-ethylhex-2-enal. The European Commission of the European Chemicals Bureau has collected toxicology data since 2-ethylhex-2-enal is a high-production volume chemical in Europe. All animals exposed acutely to 2-ethylhex-2-enal via inhalation at levels of 8406 ppm died. No animals died or were grossly affected by exposures to 2425 ppm. The oral LD<sub>50</sub> for

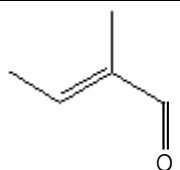
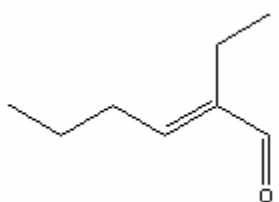
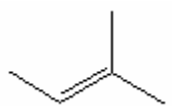


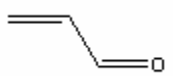
2-ethylhex-2-enal is 3-4.7 g/kg (Table 5.1). 2-Ethylhex-2-enal was negative in Ames Salmonella tests for genotoxicity.

Aldehydes such as these are generally irritating to mucous membranes, especially of the eyes or upper respiratory tract. The unsaturated aldehydes such as acrolein and crotonaldehyde are much more acutely irritating than the similarly sized aldehydes (compare propionaldehyde to acrolein).

#### 5.4.1.4 Conclusion

In reviewing the available toxicity data bases on acrolein derivatives with similar structures to 2-ethylhex-2-enal and 2-methylbut-2-enal, the most robust data were found for crotonaldehyde and acrolein, which differ in that they lack branching and have smaller carbon chains. The ACGIH-recommended TWA ceilings for crotonaldehyde and acrolein are 0.3 and 0.1 ppm, respectively. An UF of 3 is recommended to extrapolate a recommended OEL for 2-ethylhex-2-enal from crotonaldehyde due to the structural differences inherent with branching of the side-chain, which may result in metabolic or toxicological differences. Thus, the recommended **OEL for 2-ethylhex-2-enal is 0.1 ppm**. An additional UF of 3 is recommended to extrapolate a recommended OEL for 2-methylbut-2-enal from crotonaldehyde due to the limited toxicological data for 2-methylbut-2-enal. Thus, the recommended **OEL for 2-methylbut-2-enal is 0.03 ppm** (Table 5.2).

**Table 5.1.** Comparative Toxicity Parameters

Chemical	NOEL1	LC <sub>50</sub>	LD <sub>50</sub>	Structure
2-Methylbut-2-enal				
2-ethylhex-2-enal			3-4.7 g/kg	
Surrogates				
2-Methylbut-2-ene	580 ppm 26 weeks			
Crotonaldehyde	10 ppm/3.3 hr LOEL	600- 1500/30mi n 100/4 hours		
Propionaldehyde	150 ppm 6 hours/day, 7 days/week, 52 weeks (reversible nasal lesions)	26000 ppm/30 min	~0.8 - 1.7 g/kg	
Acrolein		~4 ppm 6 hours/day 62 days	46 mg/kg	

**Table 5.2.** OEL Summary

	OEL	Comment
2-Methylbut-2-enal	0.03	UF = 3 for ceiling adjustment, and additional 3 for limited data
2-ethylhex-2-enal	0.1	UF = 3 for ceiling adjustment
Surrogates		
2-Methylbut-2-ene	NA	
Crotonaldehyde	0.3 ppm	ACGIH TLV-Ceiling
Propionaldehyde	20 ppm	ACGIH TLV-TWA
Acrolein	0.1 ppm	ACGIH TLV-Ceiling

## 5.4.2 Butanal (CAS# 123-72-8)

An AOEL is proposed for butanal based on a comprehensive review of the available toxicology and guideline literature. Butanal (butyraldehyde) is an aliphatic aldehyde. Aliphatic aldehydes are generally volatile organic compounds that irritate skin, eyes, and the upper respiratory tract. Although there are no ACGIH or OSHA guidelines for butanal, the AIHA has proposed a TWA Workplace Environmental Exposure Limit (WEEL) of 25 ppm. The AIHA procedures for developing WEELs are similar to procedures described in Sections 2.0-4.0 for developing proposed AOELs for the current tank waste headspace analysis and include an assessment of the chemicals by qualified personnel, a review of the toxicological literature, and a committee assessment ([http://www.aiha.org/committees/documents/weel\\_procedures.doc](http://www.aiha.org/committees/documents/weel_procedures.doc)). The proposed AOEL for butanal is **25 ppm** and is equivalent to the WEEL developed by the AIHA.

### 5.4.2.1 Methodology

To identify and access available toxicity data on butanal, a number of Internet data bases including PUBMED, TOXNET, TOMES, and IARC, were searched and summary toxicity profiles were obtained and primary references identified. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profile and develop the AOEL. Primary references for the most relevant toxicology endpoints were obtained for review.

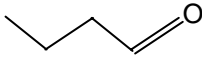
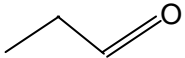
### 5.4.2.2 Available Guidelines

No OEL for butanal has been proposed by OSHA or ACGIH. However, the AIHA has assigned a WEEL for butanal at 25 ppm. Several aliphatic aldehydes have had TLVs and WEELS set, and values assigned by these two groups are generally the same. For comparison, basic toxicity/exposure guidelines for butanal and propanal are given in Table 5.3. The Spacecraft Maximum Allowable Concentration for butanal is 41 ppm (CLS 2000), and the Danish EPA (<http://www.mst.dk/chemi/01084008.htm>) has set a TLV of an average 8-hour mean limit of 25 ppm based on the WEEL.

### 5.4.2.3 Toxicology Summary and Data Analysis

Butanal is used in the production of resins, solvents, and plasticizers. It is metabolized to butyric acid, primarily in the liver. Acids from aliphatic aldehydes can serve as substrates for fatty acid oxidation within the Krebs cycle. The primary toxicological endpoint from butanal exposures is irritation to the skin and eyes. Acute toxicity is low (See Table 5.3). The NOEL following a 1-month inhalation exposure (6 hours/day, 5 days/week) in rats was 320 ppm (EPA 1989). No evidence has been presented

**Table 5.3.** Butanal and Propanal

Chemical	Structure	LC <sub>50</sub> (rats, ppm) <sup>(a)</sup>	WEEL	TLV
Butanal		59000	25	n.a.
Propanal		26000	20	20

(a) Skog 1950, as reported in CLS 2000.  
(<http://books.nap.edu/books/0309067952/html/1.html>)

that butanal exposures cause cancer. In humans, exposures to 200 ppm for 30 minutes were non-irritating (CLS 2000). The odor threshold for butanal is approximately 3 ppm (Cometto-Muniz et al. 1998).

No guidelines have been established for butanal by either ACGIH or OSHA. The 8-hour TWA WEEL of 25 ppm proposed by the AIHA is within the same range as available ACGIH TLVs given for several related aliphatic aldehydes (ACGIH 2005b), including the closely related propanal (ACGIH 2002b; Table 5.3).

#### **5.4.2.4 Conclusion**

Based on the high  $LC_{50}$  in rats, the lack of irritation observed in human exposures to 200 ppm, and the 320 ppm NOEL for subchronic inhalation exposures in rats, the 25 ppm WEEL is appropriate. In addition, the WEELs assigned by AIHA that are available for other aliphatic aldehydes generally match the TLVs assigned by ACGIH. Therefore, an AOEL of **25 ppm** is proposed for butanal.



## 5.5 Ketones

### 5.5.1 4-Methyl-2-hexanone (CAS# 105-42-0)

An AOEL is proposed for 4-methyl-2-hexanone based on a comprehensive review of the available toxicology and guideline literature. 4-Methyl-2-hexanone is an aliphatic branched (7-carbon) ketone that is structurally similar to 2-hexanone. Aliphatic ketones are generally volatile organic compounds that can be irritating to the skin, eyes, and upper respiratory tract. Although there are no ACGIH or OSHA guidelines for 4-methyl-2-hexanone, the ACGIH has set TLV-TWA exposure concentrations for 2-hexanone which has extensive animal toxicity and human data available. 2-Hexanone is a known animal and human neurotoxicant in which the mechanism of neurotoxicity has been well studied. The TLV-TWA for 2-hexanone was set at 5 ppm (20 mg/m<sup>3</sup>) to minimize the potential for distal peripheral neurotoxicity, primarily nerve fiber conduction, with weakness in the hands and feet and loss of coordination. Based on the uncertainty associated with chemical and toxicological differences between 4-methyl-2-hexanone and 2-hexanone, an AOEL of **0.5 ppm** is proposed for 4-methyl-2-hexanone.

#### 5.5.1.1 Methodology

The procedures described in Sections 2.0 - 4.0 were used as a guide for developing this AOEL. To identify and access available toxicity data on 4-methyl-2-hexanone, a number of Internet data bases, including PUBMED, TOXNET, and TOMES, were searched and summary toxicity profiles were obtained and primary references identified. In addition, ACGIH documentation was likewise reviewed for 4-methyl-2-hexanone and relevant surrogate chemicals. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profile and develop the AOEL.

#### 5.5.1.2 Available Guidelines

No OEL for 4-methyl-2-hexanone has been proposed by OSHA or ACGIH. However, ACGIH has assigned a TLV-TWA for the structural surrogate 2-hexanone. The ACGIH documentation indicates that a TLV-TWA of 5 ppm was reasonable based on the decrements in rodent and nonhuman primate sciatic-tibial nerve conduction following 9 months of inhalation exposure to 2-hexanone at concentrations equivalent to an 8-hour TWA of 75 ppm (Johnson et al. 1978). The ACGIH indicated that the TLV-TWA of 5 ppm is anticipated to include an adequate margin of safety to minimize the potential for adverse neurotoxicity.

#### 5.5.1.3 Toxicology Summary and Data Analysis

The toxicity data base for the surrogate 2-hexanone was utilized for the assessment because no substantial information was available on 4-methyl-2-hexanone. In the case of 2-hexanone, there is very robust toxicity information particularly as it relates to neurotoxicity. Subchronic inhalation exposure to 2-hexanone at concentrations ranging from 200 ppm to 1600 ppm is associated with the observance of peripheral neuropathy in chickens, rats, cats, and monkeys (Abou-Donia et al. 1985; Mendell et al. 1974; Spencer et al. 1975; Duckett et al. 1974). Johnson et al. (1979) reported that rats and monkeys that inhaled 100 ppm of 2-hexanone for 9 months developed abnormal neurophysiologic indicators, and this exposure was equivalent to a daily 8-hour TWA of 75 ppm. In addition to the observed neurotoxicity, rats that inhaled 700 ppm 2-hexanone for 11 weeks developed testicular atrophy (Katz et al. 1980);

however, observed testicular toxicity occurs at doses that are greater than those leading to axonopathy. Extensive pharmacokinetic and metabolism investigations have identified the ultimate neurotoxic metabolite as 2,5-hexanedione, which is a metabolite of both 2-hexanone and hexane. Anthony et al. (1983) evaluated the neurotoxicity of the 3,4-dimethyl-2,5-hexanedione, and it was reported as 20-30 times more potent on a molar basis than 2,5-hexanedione. It was noted that alkyl substitution causes branched-chain compounds to cyclize more rapidly than unbranched analogs, implicating pyrrole formation in the pathogenesis of the axonopathy. Although no data were available on 4-methyl-2-hexanone, it is feasible that the alkyl substitution could likewise enhance pyrrole formation.

The ACGIH documentation provides numerous examples of occupational exposure to 2-hexanone that have led to peripheral neuropathy, and the ACGIH has established a Biological Exposure Index of 0.4 mg/L for 2,5-hexanedione in urine taken at the end of a work shift (ACGIH 2003).

#### **5.5.1.4 Conclusion**

The ACGIH has set the TLV-TWA for 2-hexanone at 5 ppm (20 mg/m<sup>3</sup>) to minimize the potential for neurotoxic effects. Based on uncertainty associated with chemical and toxicological differences between 2-hexanone and 4-methyl-2-hexanone, an additional 10-fold factor is applied to the 2-hexanone OEL to establish an AOEL of **0.5 ppm** for 4-methyl-2-hexanone.

### **5.5.2 3-Hexanone (CAS# 589-38-8)**

An AOEL is proposed for 3-hexanone based on a comprehensive review of the available toxicology and guideline literature. 3-Hexanone is an aliphatic (6-carbon) ketone. Aliphatic ketones are generally volatile organic compounds that can be irritating to the skin, eyes, and the upper respiratory tract. Although there are no ACGIH or OSHA guidelines for 3-hexanone, the ACGIH has set TLV-TWA exposure concentrations for 2-pentanone and 3-pentanone (ACGIH 2001h; 2001r) with 2-pentanone having the more robust data set. The TLV-TWA for the pentanones was set at 200 ppm (705 mg/m<sup>3</sup>) to minimize the potential objectionable narcotic effects and significant irritation. Based on the uncertainty associated with chemical differences between 3-hexanone and 2- or 3-pentanone, an AOEL of **67 ppm** is proposed for 3-hexanone.

#### **5.5.2.1 Methodology**

The procedures described in Sections 2.0 - 4.0 were used as a guide for developing this AOEL. To identify and access available toxicity data on 3-hexanone, a number of Internet data bases, including PUBMED, TOXNET, and TOMES were searched, and summary toxicity profiles were obtained and primary references identified. ACGIH documentation was likewise reviewed for 3-hexanone and relevant surrogate chemicals. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profile and develop the AOEL.

#### **5.5.2.2 Available Guidelines**

No OEL for 3-hexanone has been proposed by OSHA or ACGIH. However, ACGIH has assigned a TLV-TWA for the structural surrogates 2- and 3-pentanone. The ACGIH documentation indicates that occupational exposure to 1500 ppm 2-pentanone has been associated with complaints of ocular and mucous membrane irritation in humans, and high-dose exposure to 2-pentanone was associated with

narcosis in animals. Therefore, a 200 ppm TLV-TWA was recommended by ACGIH and is believed to be low enough to protect from narcotic and irritation effects.

### 5.5.2.3 Toxicology Summary and Data Analysis

The lowest published lethal concentration for 3-hexanone was reported as a 4-hour inhalation exposure to 4000 ppm in rats. No other relevant information was obtained for 3-hexanone. However, for the surrogate 2-pentanone, occupational exposure to 1500 ppm was associated with a strong odor and ocular and upper respiratory tract irritation in humans. Inhalation exposure of guinea pigs to 5000 ppm for 8 hours resulted in narcosis, but obvious ocular and upper respiratory tract irritation was likewise noted. The 4-hour LC<sub>50</sub> in rats was >2000 ppm.

### 5.5.2.4 Conclusion

The ACGIH has set the TLV-TWA for the pentanones at 200 ppm (705 mg/m<sup>3</sup>) to minimize the potential objectionable narcotic effects and significant irritation. Based on the uncertainty associated with chemical differences between 3-hexanone and 2- or 3-pentanone, an additional 3-fold factor is applied to the pentanone OEL to establish an AOEL of **67 ppm** for 3-hexanone.

## 5.5.3 6-Methyl-2-heptanone (CAS# 928-68-7)

An AOEL is proposed for 6-methyl-2-heptanone based on a comprehensive review of the available toxicology and guideline literature. 6-methyl-2-heptanone is an aliphatic branched (8-carbon) ketone. Aliphatic ketones are generally volatile organic compounds that can be irritating to the skin, eyes, and the upper respiratory tract. Although there are no ACGIH or OSHA guidelines for 6-methyl-2-heptanone, the ACGIH has set TLV-TWA exposure concentrations for 5-methyl-3-heptanone and 2-heptanone (ACGIH 2001i; 2001q) with 5-methyl-3-heptanone having available human data. The TLV-TWA for 5-methyl-3-heptanone was set at 25 ppm (131 mg/m<sup>3</sup>) as a comfort level in unconditioned workers. Based on the uncertainty associated with chemical differences between 6-methyl-2-heptanone, 5-methyl-3-heptanone, and 2-heptanone an AOEL of **8 ppm** is proposed for 6-methyl-2-heptanone.

### 5.5.3.1 Methodology

The procedures described in Sections 2.0 - 4.0 were used as a guide for developing this AOEL. To identify and access available toxicity data on 3-hexanone, a number of Internet data bases including PUBMED, TOXNET, and TOMES were searched, and summary toxicity profiles were obtained and primary references identified. ACGIH documentation was likewise reviewed for 6-methyl-2-heptanone and relevant surrogate chemicals. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profile and develop the AOEL.

### 5.5.3.2 Available Guidelines

No OEL for 6-methyl-2-heptanone has been proposed by OSHA or ACGIH. However, ACGIH has assigned a TLV-TWA for the structural surrogates 5-methyl-3-heptanone and 2-heptanone. The ACGIH documentation indicates that humans are sensitive to 5-methyl-3-heptanone vapor exposures at concentrations of >25 ppm, which results in both eye and respiratory irritation. Therefore, a 25 ppm

TLV-TWA was recommended by ACGIH as a comfort level for unconditioned workers. Data for 2-heptanone was limited to animal studies, and a TLV-TWA of 50 ppm was recommended to minimize the potential for irritant effects.

### 5.5.3.3 Toxicology Summary and Data Analysis

No substantial information was available for 6-methyl-2-heptanone. The ACGIH documentation indicates that unconditioned human occupational exposure to 50 ppm 5-methyl-3-heptanone was the threshold for eye and nose irritation in 50% of the exposed subjects, while 6 ppm was an odor threshold. A 100-ppm exposure was reported to be irritating to the mucous membranes and produced headache and nausea that were too intense to tolerate for more than a few minutes. In rats and mice, 4-hour exposure to a saturated atmosphere (3000 ppm) rapidly resulted in signs of eye and respiratory system irritation. Oral LD<sub>50</sub>s ranged from 2.5- to 3.8 g/kg in rats, mice, and guinea pigs. There were no reported human data with 2-heptanone, although NIOSH reported that ~67,000 workers are potentially exposed. Exposure to 1500 ppm 2-heptanone produced irritation, while 2000 ppm was strongly narcotic, and 4800 ppm caused narcosis and death in 4-8 hours in rats. The oral LD<sub>50</sub> is ~1.7g/kg in rats. A subchronic study conducted in both rats and monkeys found no clinical evidence of toxicity or neurological impairment following a 9-month exposure to either 131 or 1025 ppm 2-heptanone (Johnson et al. 1978). The TLV-TWA of 50 ppm was recommended by ACGIH to minimize any potential for irritant effects.

### 5.5.3.4 Conclusion

The ACGIH has set the TLV-TWA for 5-methyl-3-heptanone and 2-heptanone at 25 ppm (131 mg/m<sup>3</sup>) and 50 ppm (233 mg/m<sup>3</sup>), respectively, to minimize the potential irritation effects. The results observed in humans with 5-methyl-3-heptanone are the most relevant and will be the primary surrogate for establishing an AOEL for 5-methyl-2-heptanone. Based on minimal uncertainty associated with chemical differences between 5-methyl-3-heptanone and 6-methyl-2-heptanone, an additional 3-fold factor is applied to the 5-methyl-3-heptanone OEL to establish an AOEL of **8 ppm** for 6-methyl-2-heptanone.

### 5.5.4 2-Nonanone (CAS# 821-55-6), 3-Dodecanone (CAS#1534-27-6), 2-Tridecanone (CAS#593-08-8), 3-Tridecanone (CAS#1534-26-5)

An AOEL is proposed for 2-nonanone, 3-dodecanone, 2-tridecanone, and 3-tridecanone based on a comprehensive review of the available toxicology and guideline literature. These long-chained aliphatic ketones range from 9-carbon to 13-carbon in length with the ketone present on either the 2nd or 3rd carbon. Aliphatic ketones can be irritating to the skin, eyes and the upper respiratory tract. Although there are no ACGIH or OSHA guidelines for these long chained aliphatic ketones, the ACGIH has set TLV-TWA exposure concentrations for the structural isomers of heptanone (2-heptanone, 3-heptanone and 4-heptanone [molecular formula: C<sub>7</sub>H<sub>14</sub>O]) at 50 ppm (233 mg/m<sup>3</sup>) to minimize the potential for eye and skin irritation (ACGIH 2001i; 2001b; 2001q). Based on the uncertainty associated with chemical differences between 2-nonanone, 3-dodecanone, 2-tridecanone, and 3-tridecanone, and the isomers of heptanone, an AOEL of **17 ppm** is proposed for these long (C-9 to C-13) chained aliphatic ketones.

#### 5.5.4.1 Methodology

The procedures described in Sections 2.0 - 4.0 were used as a guide for developing this AOEL. To identify and access available toxicity data on 2-nonanone, 3-dodecanone, 2-tridecanone, and

3-tridecanone a number of Internet data bases including PUBMED, TOXNET, and TOMES were searched, and summary toxicity profiles were obtained and primary references identified. In addition, ACGIH documentation was reviewed for 2-heptanone, 3-heptanone, 4-heptanone, and relevant surrogate chemicals. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profiles and develop the AOEL.

#### **5.5.4.2 Available Guidelines**

No OELs for 2-nonanone, 3-dodecanone, 2-tridecanone, and 3-tridecanone have been proposed by OSHA or ACGIH. However, ACGIH has assigned a TLV-TWA for the structural surrogate isomers 2-heptanone, 3-heptanone, and 4-heptanone. The ACGIH documentation recommends a 50-ppm TLV-TWA to minimize the potential for ocular, dermal, and mucus membrane irritation.

#### **5.5.4.3 Toxicology Summary and Data Analysis**

These higher molecular weight ketones generally have low volatility; therefore, toxicity information is more available from oral and dermal exposures. Acute oral LD<sub>50</sub>s are high (greater than 5 g/kg), indicating low toxicity. For 2-nonanone, the LD<sub>50</sub> is greater than 3 and 7 g/kg in rats and mice, respectively (Tanii 1986). The ACGIH documentation for 2-heptanone, 3-heptanone, and 4-heptanone ketone isomers was based upon the similar toxicological response. Exposure to 1500 ppm 2-heptanone produced irritation, while 2000 ppm was strongly narcotic, and 4800 ppm caused narcosis and death in 4-8 hours in rats. The oral LD<sub>50</sub> in rats is ~1.7g/kg (ACGIH 2001i). A report suggested that irritation potency might increase with increasing ketone carbon chain length (Specht et al. 1940). A subchronic study conducted in both rats and monkeys found no clinical evidence of toxicity or neurological impairment following a 9-month exposure to either 131 or 1025 ppm 2-heptanone (Johnson et al. 1978). Likewise, limited inhalation toxicity studies (4-hour acute exposure) with 3-heptanone resulted in no fatalities at 2000 ppm, but all animals died at 4000 ppm (Smyth et al. 1949). Rats exposed to 700 ppm for 24-weeks exhibited no signs of neurotoxic effects (Katz et al. 1980), even though the neurotoxic 2,5-heptanedione metabolite was detected. Katz et al (1980) concluded that even though the neurotoxic metabolite was detectable, the systemic concentration achieved was insufficient to elicit a neurotoxic response. Limited toxicity studies have also been conducted with the 4-heptanone isomer. Following a 6-hour inhalation exposure, the LC<sub>50</sub> for 4-heptanone was 2690 ppm, with narcosis occurring at 1600 ppm, and three out of four rats killed at 3200 ppm (Krasavage et al. 1982). A 6-hours/day exposure to 1200 ppm of 4-heptanone resulted in a slightly decreased response to stimulation during exposure, and marginal liver enlargement but no changes in clinical chemistry or pathology (Krasavage et al. 1982). Hence, for these structurally related heptanone compounds, the ACGIH recommended a TLV-TWA of 50 ppm primarily to minimize any potential for irritant effects.

#### **5.5.4.4 Conclusion**

The ACGIH has set the TLV-TWA for 2-heptanone, 3-heptanone, and 4-heptanone ketone isomers at 50 ppm (233 mg/m<sup>3</sup>). Based on the uncertainty associated with chemical differences between 2-nonanone, 3-dodecanone, 2-tridecanone, 3-tridecanone, and the isomers of heptanone, an additional 3-fold factor is applied to the OEL to establish an AOEL of 17 ppm for these long (C-9 to C-13) chained aliphatic ketones.

#### **5.5.4.5 5-Methyl-2-(1-methyethenyl)cyclohexanone (CAS# 89-82-7)**

An AOEL is proposed for 5-methyl-2-(1-methyethenyl) cyclohexanone, which is also known as Pulegone, based on a comprehensive review of the available toxicology and guideline literature. This methyl and isopropyl substituted cyclohexanone is structurally similar to cyclohexanone. Although there are no ACGIH or OSHA guidelines for pulegone, the ACGIH has set TLV-TWA exposure concentrations for cyclohexanone at 25 ppm (100 mg/m<sup>3</sup>) to minimize the potential for eye, skin, and throat irritation (ACGIH 2001j). Based on the uncertainty associated with chemical differences between cyclohexanone and 5-methyl-2-(1-methyethenyl) cyclohexanone, an AOEL of **2.5 ppm** is proposed.

#### **5.5.4.6 Methodology**

The procedures described in Sections 2.0 - 4.0 were used as a guide for developing this AOEL. To identify and access available toxicity data on 5-methyl-2-(1-methyethenyl) cyclohexanone, a number of Internet data bases including PUBMED, TOXNET, and TOMES were searched, and summary toxicity profiles were obtained and primary references identified. In addition, ACGIH documentation was reviewed for 5-methyl-2-(1-methyethenyl) cyclohexanone and relevant surrogate chemicals. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profiles and develop the AOEL.

#### **5.5.4.7 Available Guidelines**

No OELs for 5-methyl-2-(1-methyethenyl) cyclohexanone have been proposed by OSHA or ACGIH. However, ACGIH has assigned a TLV-TWA for the structural surrogate cyclohexanone at 25 ppm to minimize the potential for eye, nose, and throat irritation.

#### **5.5.4.8 Toxicology Summary and Data Analysis**

5-Methyl-2-(1-methyethenyl) cyclohexanone, commonly known as pulegone, is a major component of oil of pennyroyal, which has been used for flavoring foods and drinks. Pennyroyal oil has also been used as a fragrance agent and as an herbal medicine. Tice (1998) prepared a detailed review of the toxicology literature on pulegone and one of its metabolites, menthofuran, in support of pulegone and its metabolite being nominated by the NTP for additional chronic testing. The rationale for further testing was based on the potential for human exposure and the absence of any carcinogenicity data. The NTP bioassay on pulegone is currently ongoing, and updates on project results can be obtained from the NTP website (<http://ntp-server.niehs.nih.gov/>). Human exposure to pulegone is primarily through ingestion of food products, and human acute toxicity has been associated with the ingestion of high doses of pennyroyal oil. There are no available animal inhalation studies with pulegone; however, the LD<sub>50</sub> values for subcutaneous or intraperitoneal routes ranged from 150 – 1700 mg/kg in mice and rats. In a 28-day oral gavage study in rats, pulegone was administered at doses of 0, 20, 60 and 160 mg/kg/day. The NOEL for pulegone was 20 mg/kg/day (Thorup et al. 1983). In experimental animals, pulegone is metabolically transformed (CYP450 involvement) to menthofuran and other metabolites. The primary toxic response at high doses is associated with hepatic effects. There are currently no published data found on chronic exposure, teratogenicity, embryotoxicity, or carcinogenicity of pulegone.

Cyclohexanone (CAS 108-94-1) is a structural surrogate for pulegone, and ACGIH (2001j) has established a TLV-TWA of 25 ppm (100 mg/m<sup>3</sup>) for occupational exposure, which is intended to

minimize the potential for eye, nose, and throat irritation. The acute toxicity of cyclohexanone is low; the intraperitoneal and subcutaneous LD<sub>50</sub> values ranged from 1130 – 1535 mg/kg (Smyth et al. 1969; Gupta et al. 1979; Deichman and LeBlanc 1943). A 6-hour exposure to 4000 ppm in guinea pigs resulted in narcosis, hypothermia, and decreased respiration (Specht et al. 1940). Rabbits were exposed to 190 to 1414 ppm cyclohexanone for 6 hours/day for 50 days. The 190-ppm exposure showed only minor degenerative changes in liver and kidney (Treon et al. 1943). A 2-year carcinogenicity study was conducted in both rats and mice, and the neoplasms in the treated groups did not differ significantly from controls (Lijinsky and Kovatch 1986). The major metabolite of cyclohexanone is a reduction to cyclohexanol and subsequent conjugation with glucuronic acid. In humans, exposure to 25 ppm cyclohexanone was not uncomfortable for most subjects; 50 ppm was irritating to the throat, while exposure to 75 ppm for 3-5 minutes resulted in more pronounced irritation of the eyes, nose, and throat (Nelson et al. 1943).

#### **5.5.4.9 Conclusion**

The ACGIH has set the TLV-TWA for cyclohexanone at 25 ppm (100 mg/m<sup>3</sup>). Based on the uncertainty associated with chemical, metabolic, and toxicological differences between cyclohexanone and 5-methyl-2-(1-methylethenyl) cyclohexanone, an additional 10-fold factor is applied to the cyclohexanone OEL to establish an AOEL of **2.5 ppm** for 5-methyl-2-(1-methylethenyl) cyclohexanone.

#### **5.5.5 4,7,7-Trimethyl-bicyclo[4.1.0]heptan-3-one (CAS# 4176-04-9)**

An AOEL is proposed for 4,7,7-trimethyl-bicyclo[4.1.0]heptane-3-one, which is also known as caranone, based on a comprehensive review of the available toxicology and guideline literature. This cyclo ketone is structurally similar to camphor (1,7,7-trimethyl-bicyclo[2.2.1]heptan-2-one). Although there are no ACGIH or OSHA guidelines for caranone, the ACGIH has set TLV-TWA exposure concentrations for the structurally similar camphor at 2 ppm (12 mg/m<sup>3</sup>) to minimize the potential for eye and nose irritation and loss of sense of smell (ACGIH 2001b). Based on the uncertainty associated with chemical differences between camphor and caranone, an AOEL of **0.7 ppm** is proposed.

##### **5.5.5.1 Methodology**

The procedures described in Sections 2.0 - 4.0 were used as a guide for developing this AOEL. To identify and access available toxicity data on 4,7,7-trimethyl-bicyclo[4.1.0]heptane-3-one (caranone), a number of Internet data bases, including PUBMED, TOXNET, and TOMES were searched, and summary toxicity profiles were obtained and primary references identified. In addition, ACGIH documentation was reviewed for camphor (1,7,7-trimethyl-bicyclo[2.2.1]heptan-2-one), which is a relevant surrogate chemical. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profiles and develop the AOEL.

##### **5.5.5.2 Available Guidelines**

No OELs for 4,7,7-trimethyl-bicyclo[4.1.0]heptane-3-one have been proposed by OSHA or ACGIH. However, ACGIH has assigned a TLV-TWA for the structural surrogate camphor (1,7,7-trimethyl-bicyclo[2.2.1]heptan-2-one) at 2 ppm to minimize the potential for eye and nose irritation and loss of sense of smell.

### 5.5.5.3 Toxicology Summary and Data Analysis

4,7,7-Trimethyl-bicyclo[4.1.0]heptane-3-one is commonly known as caranone. There are no available animal inhalation studies with caranone. Camphor (CAS 76-22-2) is a structural surrogate for caranone, and ACGIH (2001b) has established a TLV-TWA of 2 ppm (12 mg/m<sup>3</sup>) for occupational exposure, which is intended to minimize the potential for eye and nose irritation and loss of sense of smell. The acute inhalation toxicity of camphor was evaluated; the LC<sub>50</sub> in mice was reported as 75 ppm, and the TCLo in mice was 18 ppm for a repeated (7-week) exposure. No tumorigenic response was reported in mice following subcutaneous or topical treatment with camphor. In humans, exposure at or below 2 ppm over a 5-day work week showed that irritation to the eyes and nose and loss of sense of smell did not occur (Gronka et al. 1969). In addition, physical examination of exposed workers failed to reveal any significant findings apart from nose and throat inflammation, with the longest exposures lasting ~10 months. High-level exposures (concentration not specified) are reported to cause nausea, anxiety, dizziness, confusion, headache, and twitching of facial muscles (Gronka et al. 1969; Proctor et al. 1988; Gosselin et al. 1984; Arnow 1976).

### 5.5.5.4 Conclusion

The ACGIH has set the TLV-TWA for camphor at 2 ppm (12 mg/m<sup>3</sup>). Based on the uncertainty associated with chemical and toxicological differences between caranone and camphor, an additional 3-fold factor is applied to the camphor OEL to establish an AOEL of **0.7 ppm** for 4,7,7-trimethyl-bicyclo[4.1.0]heptane-3-one.

### 5.5.6 3-Methyl-3-buten-2-one (CAS# 814-78-8)

An AOEL is proposed for 3-methyl-3-buten-2-one, based on a comprehensive review of the available toxicology and guideline literature. This ketone is structurally similar to 3-butene-2-one. Although there are no ACGIH or OSHA guidelines for 3-methyl-3-buten-2-one, the American Conference of Governmental Industrial Hygienists (ACGIH<sup>®</sup>) has set Threshold Limit Value-Ceiling (TLV<sup>®</sup>-Ceiling) exposure concentrations for 3-butene-2-one at 0.2 ppm (0.6 mg/m<sup>3</sup>) to minimize the potential for irritation of the skin, eyes and respiratory tract and the risk of dermal sensitization or allergic type reaction (ACGIH 2001s). Based on the uncertainty associated with chemical and toxicological differences between 3-buten-2-one and 3-methyl-3-buten-2-one an AOEL of 0.02 ppm is proposed.

#### 5.5.6.1 Methodology

The procedures described in Sections 2.0 - 4.0 were used as a guide for developing this AOEL. To identify and access available toxicity data on 3-methyl-3-buten-2-one, a number of Internet data bases, including PUBMED, TOXNET, and TOMES were searched, and summary toxicity profiles were obtained and primary references identified. In addition, ACGIH documentation was reviewed for 3-buten-2-one, which is a relevant surrogate chemical. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profiles and develop the AOEL.



### 5.5.6.2 Available Guidelines

No OELs for 3-methyl-3-buten-2-one have been proposed by OSHA or ACGIH. However, ACGIH has assigned a TLV-ceiling for the structural surrogate 3-buten-2-one at 0.2 ppm to minimize the potential for irritation of the skin, eyes, and respiratory tract and the risk of dermal sensitization or allergic reaction. The German MAK commission has reported 3-methyl-3-buten-2-one as a skin sensitizing substance based on animal tests (MAK 2000).

### 5.5.6.3 Toxicology Summary and Data Analysis

For 3-methyl-3-buten-2-one, the lowest published lethal concentration in rats and guinea pigs ranged from 125 – 250 ppm for 4 hours. A 30 ppm, 7 hours/day exposure to rats, guinea pigs, and rabbits for a total of 20 exposures in 28 days resulted in ocular and nasal irritation in all species, death and weight loss in rats, and decreased growth in guinea pigs. In addition, 100 7 hours/day exposures of 15 ppm over 140 days resulted in ocular and nasal irritation in all animals (Clayton and Clayton 1982). In humans, 3-methyl-3-buten-2-one is reported to produce skin irritation upon contact. Vapors have also been reported to be highly irritating to the respiratory tract (Clayton and Clayton 1982).

The surrogate 3-buten-2-one has an odor threshold of ~0.2 ppm, which corresponds to the current ACGIH TLV-ceiling. The 4-hour LC<sub>50</sub> values for 3-buten-2-one ranged from 2.4 to 2.8 ppm in rats and mice, and a 50% respiratory depression of 5.3 ppm was reported for mice (Schaper 1993). These acute results with 3-buten-2-one suggest that this surrogate may be acutely more toxic than 3-methyl-3-buten-2-one. In humans, 3-buten-2-one is a severe skin, eye, and respiratory tract irritant.

### 5.5.6.4 Conclusion

The ACGIH has set the TLV-ceiling for 3-buten-2-one at 0.2 ppm (0.6 mg/m<sup>3</sup>). Based on the uncertainty associated with chemical and toxicological differences between 3-methyl-3-buten-2-one and 3-buten-2-one, an additional 3-fold factor is applied; in addition, a 3-fold factor is applied for extrapolation from the TLV-ceiling to a TWA. Hence, a net 9-fold factor is applied to the 3-buten-2-one OEL to establish an AOEL of 0.02 ppm for 3-methyl-3-buten-2-one.

## 5.6 Esters

### 5.6.1 1-Methylethyl tetradecanoate (Cas # 110-27-0), Butyl tetradecanoate (Cas # 110-36-1), and 1-Methylethyl hexadecanoate (CAS # 142-91-6)

The screening value for these decanoic acids was obtained from a Swedish TWA of 5 mg/m<sup>3</sup> for the related butyl stearate (butyl octadecanoate, CAS # 123-95-5). The ACGIH also has established a TLV-TWA for stearates of 10 mg/m<sup>3</sup> (~ 0.72 ppm), and OSHA has set a nuisance dust level for stearates of 5 mg/m<sup>3</sup>. These values are for total particulates or nuisance dusts and do not apply to vapors. These decanoic acids are common components of cosmetic compounds. They break down into innocuous metabolic products and are non-toxic, similar to the related fatty acids. The common use in cosmetics and the known and expected metabolic breakdown to non-toxic products indicates an AOEL is not needed for these chemicals at this time. As additional new data become available for these compounds, it would be reasonable to reassess the appropriateness of assigning an AOEL.

### 5.6.2 Dibutyl butylphosphonate (CAS# 78-46-6)

An AOEL is proposed for dibutyl butylphosphonate based on a comprehensive review of the available toxicology and guideline literature. Dibutyl butylphosphonate is a phosphoric acid similar in structure to dibutyl phosphate (DBP, CAS# 107-66-4) and tributyl phosphate (TBP, CAS# 126-73-8). Although there are no ACGIH or OSHA guidelines for dibutyl butylphosphonate, the ACGIH has set TLV-TWA exposure concentrations for TBP at 0.2 ppm (2.2 mg/m<sup>3</sup>) and for DBP at 1 ppm (8.6 mg/m<sup>3</sup>) (ACGIH 2001aa; 2001g). Based on the uncertainty associated with chemical and toxicological differences between dibutyl butylphosphonate and DBP or TBP, an AOEL of **0.02 ppm** is proposed for dibutyl butylphosphonate by adding additional uncertainty factors to the TBP TLV-TWA.

#### 5.6.2.1 Methodology

The procedures described in Sections 2.0 - 4.0 were used as a guide for developing this AOEL. To identify and access available toxicity data on dibutyl butylphosphonate, a number of Internet data bases including PUBMED, TOXNET, and TOMES were searched, and summary toxicity profiles were obtained and primary references identified. In addition, ACGIH documentation was reviewed for dibutyl phosphate and relevant surrogate chemicals. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profile and develop the AOEL.

#### 5.6.2.2 Available Guidelines

No OEL for dibutyl butylphosphonate has been proposed by OSHA or ACGIH. However, ACGIH has assigned a TLV-TWA for the structural surrogates DBP and TBP (ACGIH 2001aa; 2001g). The ACGIH documentation for DBP and TBP are based upon very limited toxicity information (primarily sensory irritation), and in the case of TBP, by structural analogy with triphenyl phosphate. The current TLV-TWA for TBP and DBP are 0.2 ppm (2.2 mg/m<sup>3</sup>) and 1 ppm (8.6 mg/m<sup>3</sup>), respectively.

#### 5.6.2.3 Toxicology Summary and Data Analysis

A search of the RTECS data base found very limited toxicity data on dibutyl butylphosphonate. The LD<sub>50</sub> following intraperitoneal and intravenous dosing in the mouse were 125 mg/kg and 56 mg/kg,

respectively. For the structural surrogate DBP there is very limited information supporting the TLV-TWA, with some indication of respiratory tract irritation and headache in some workers exposed to DBP (unknown concentration/time) (ACGIH 2001g). In contrast, there are slightly more toxicity data available for TBP (ACGIH 2001aa). A 6-hour inhalation exposure of rats to TBP resulted in an LC<sub>50</sub> of 123 ppm, and a 4- to 5-hour inhalation LC<sub>50</sub> of 227 ppm (as determined by the World Health Organization [WHO] in 1991). Workers exposed to 1.4 ppm TBP have complained of nausea and headaches (ACGIH 2001aa); hence, the ACGIH recommended a TLV-TWA of 0.2 ppm for TBP because it was less than the concentration reported to be associated with worker complaints.

#### 5.6.2.4 Conclusion

The ACGIH has set the TLV-TWA for tributyl phosphate at 0.2 ppm (2.29 mg/m<sup>3</sup>) based primarily by analogy with triphenyl phosphate and reports of worker complaints at 1.4 ppm exposure. Based on the uncertainty associated with chemical and toxicological differences between dibutyl butylphosphonate and TBP, a 9-fold factor is applied to the TBP TLV-TWA to establish an AOEL of **0.02 ppm** for dibutyl butylphosphonate. A 3-fold factor was used to account for the lack of a robust toxicity data set, and an additional 3-fold factor used to account for structural differences between dibutyl butylphosphonate and TBP.

## 5.7 Alkyl Nitriles and Alkene Nitriles

### 5.7.1 Propanenitrile (CAS #107-12-0), Butanenitrile (CAS # 109-74-0), 2-Methylene butanenitrile (CAS#1647-11-6), Pentanenitrile (CAS# 110-59-8), Hexanenitrile (CAS # 628-73-9), Heptanenitrile (CAS# 629-08-3), Octanenitrile (CAS # 124-12-9), Nonanenitrile (CAS #2243-27-8), Decanenitrile (CAS # 1975-78-6) and 3-Butanenitrile (CAS# 109-75-1), and 2,-4-Pentadienenitrile (CAS# 1615-70-9)

AOELs are proposed for a family of saturated alkyl nitriles with two to nine carbon alkyl side-chains and unsaturated alkenes with three to four carbon side-chains, based on a comprehensive review of the available toxicology literature. The principal toxic effects of these nitriles have been proposed to be mediated by metabolically released cyanide. The chemical in this family with the most robust toxicological data set is propanenitrile. The proposed **AOELs for these compounds range from 0.3 to 8 ppm** and are based on toxicity information and pre-existing exposure guidelines from acceptable private and government agencies for the nitriles of interest and structurally similar compounds. An additional margin of safety was added, where appropriate, for the compounds in this class without pre-existing guidelines.

#### 5.7.1.1 Methodology

To identify and access available toxicity data on alkyl and alkene nitriles, a number of Internet data bases, including PUBMED, TOXNET, TOMES, EPA-IRIS, and the IARC were searched, and summary toxicity profiles were obtained and primary references identified. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profile and for the development of the AOEL. Primary references for the most relevant toxicology endpoints were obtained for review.

#### 5.7.1.2 Available Guidelines

Guidelines have been proposed for propanenitrile and butanenitrile (see Table 5.4). Guidelines have also been proposed for related alkyl nitriles and alkene nitriles (see Tables 5.4 and 5.5). For many of these nitrile compounds, skin notations, but no classifications as carcinogens, have also been recommended.

#### 5.7.1.3 Toxicology Summary

The LC<sub>50s</sub> for butanenitrile and propanenitrile are similar (see Table 5.6). The endpoint of concern for developing an AOEL based on animal toxicity studies is likely to be developmental effects. At high exposure concentrations, embryoletality was observed; at lower concentrations, fetotoxicity was observed (Saillenfait et al. 1993). The NOELs for developmental effects are very similar between butanenitrile propanenitrile and pentanenitrile (see Table 5.6).

**Table 5.4.** Exposure Limit Values Alkyl Nitriles

Nitrile	Agency	Description	Limit (ppm)
Acetonitrile	ACGIH <sup>®</sup>	TLV <sup>®</sup> -TWA	20
	EPA	AEGL3 (8-hr)	100
Butanenitrile	NIOSH	REL	8
Propanenitrile	NIOSH	REL	6
	EPA	AEGL3 (8-hr)	1.4

REL = Recommended exposure limit  
 TLV = Threshold limit value  
 TWA = Time-weighted average  
 AEGL<sup>(a)</sup> = Acute Exposure Guideline Levels, once-in-a-lifetime, or rare, exposure to airborne chemicals. The AEGLs for acetonitrile and propanenitrile are under review and do not have technical support documentation available at this time.

(a) AEGL Levels 1-3. AEGL Level 1 is an exposure at which an individual may have some discomfort. Level 3 is the level at which an individual could experience life-threatening effects. Level 2 is the action level, the level at which an individual may experience long-lasting effects.

**Table 5.5.** Acute or Subchronic LD<sub>50</sub>s and OELs for Alkene Nitriles

Nitrile	C #	Species	LD <sub>50</sub> <sup>(a)</sup> (mmol/kg)	Exposure Limit: Agency
Acrylonitrile	2	Mouse	0.73	2 ppm: OSHA PEL 1 ppm: NIOSH REL
Methacrylonitrile	3	Mouse	0.26	1 ppm: NIOSH REL
3-Butenenitrile	3	Mouse	1.0	--

OSHA = Occupational Health and Safety Administration  
 PEL = Permissible exposure limit  
 (a) All LD<sub>50</sub> values are from Tanii and Hashimoto (1984a).

Acetonitrile, pentanenitrile, and propanenitrile were all negative in genotoxicity assays conducted by the NTP of the NIEHS.<sup>(a)</sup> Lifetime inhalation exposures to the structurally related acetonitrile did not result in increased incidences of cancers in rats (NTP 1996).

Toxicity of these saturated alkyl nitriles and unsaturated alkene nitriles is most likely mediated through metabolically released cyanide (Willhite and Smith 1981; Tanii and Hashimoto 1984b). All of these nitriles contain alkyl or alkene chains with nine carbons or less and a nitrile functional group (R-C≡N). The EPA Office of Pollution Prevention and Toxics has accepted a proposal from Eastman Kodak Co. and Solutia, Inc. to combine alkyl nitriles with up to three carbons (C-3) for the purposes of addressing data gaps.<sup>(a)</sup> For the purpose of developing AOELs, nitriles with longer side-chains are also considered within this family.

- (a) Assay information can be found online:  
<http://ntp.niehs.nih.gov/ntpweb/index.cfm?objectid=123DED4A-F1F6-975E-7CB19C212F28F788>.
- (a) See EPA's comments to the proposal online: <http://www.epa.gov/chemrtk/alkyntrl/c14860ct.htm>.

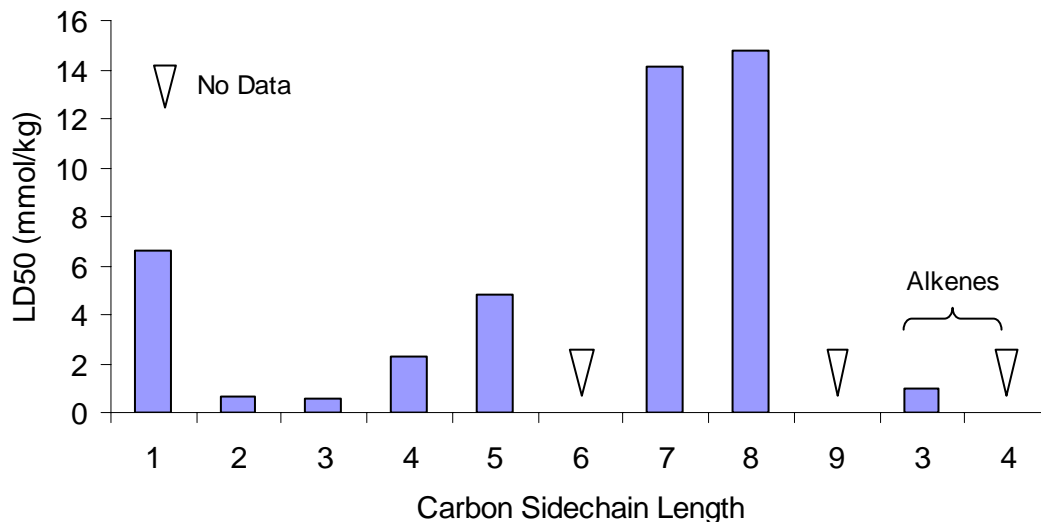
**Table 5.6.** Acute or Subchronic NOELs and LC<sub>50</sub>s for Alkyl Nitriles

Nitrile	C # <sup>(a)</sup>	Species	LC <sub>50</sub> (ppm)	LD <sub>50</sub> <sup>(b)</sup> (mg/kg)	NOEL	Reference
Acetonitrile	1	Rat/Developmental	--	--	190 mg/kg	Johannsen et al. 1986
		Mouse	2673	6.6	--	Willhite et al. 1981
Propanenitrile	2	Rat/Developmental	--	--	50 ppm	Saillenfait et al. 1993
		Rat/Developmental	--	--	20 mg/kg	Johannsen et al. 1986
		Mouse	163	0.65	--	Willhite et al. 1981
Butanenitrile	3	Mouse	249	0.56	--	Willhite et al. 1981
		Rat/Developmental	--	--	50 ppm	Saillenfait et al. 1993
Pentanenitrile	4	Rat/Developmental	--	--	30 mg/kg	Munley et al. 2001
		Mouse		2.3		
Hexanenitrile	5	Mouse	--	4.8	--	
Heptanenitrile	6	NA	--	--	--	
Octanenitrile	7	Mouse	--	14.1	--	
Nonanenitrile	8	Mouse	--	14.8	--	
Decanenitrile	9	NA	--	--	--	
-- = Not applicable						
(a) Sidechain carbon number						
(b) All LD <sub>50</sub> values are from Tanii and Hashimoto (1984a).						

Toxicity data for most of these chemicals are limited. The decision to include the alkyl and alkene nitriles as a family was driven by the comparable available data, the similarity of the compounds, and the common active group (C≡N). The research reported by Tanii and Hashimoto (1984a 1984b) found cyanide metabolism in all the tested members of this family. The LD<sub>50</sub> values determined by Tanii and Hashimoto (1984a) formed the basis for the proposed AOELs for chemicals in this family that currently lack AOELs (see Tables 5.5 and 5.6).

#### 5.7.1.4 Data Analysis

NIOSH has set exposure guidelines for two of the four short-chain nitriles of interest (i.e., butanenitrile and pentanenitrile) and ACGIH has set exposure guidelines for the structurally similar acetonitrile (see Table 5.4). The TLV for acetonitrile is about three-fold higher than the NIOSH-recommended exposure limits (RELs) for the other nitriles. Likewise, the NOELs for developmental effects are about eight-fold higher for acetonitrile than propanenitrile and pentanenitrile (see Table 5.5). The LD<sub>50</sub> values follow a similar trend, indicating toxicity of acetonitrile is less than alkyl and alkene nitriles with C-2 to C-5 chains. Alkyl nitriles with chains longer than seven carbons have increased LD<sub>50</sub> values (decreased toxicity) compared to the shorter-chain nitriles (see Tables 5.5 and 5.6; Figure 5.1).



**Figure 5.1.** Comparison of Mice LD<sub>50</sub> Values (Tanii and Hashimoto 1984a) Based on Carbon Length

To determine reasonable AOELs, the guidelines for propanenitrile (6 ppm: NIOSH REL) and butanenitrile (8 ppm: NIOSH REL) have been compared to the toxicity data available for the family of compounds in the literature. For alkyl nitriles with toxicity data below those of propanenitrile, an AOEL of 6 ppm is proposed. NIOSH has recommended RELs for the structurally similar acrylonitrile and methacrylonitrile. The toxicity data indicate similar LD<sub>50</sub>s for the alkene nitriles (see Table 5.5) and an AOEL of 1 ppm is proposed for 3-butenenitrile, for which toxicity data are available, and 0.3 ppm for 2,4-pentadienenitrile and for 2-methylene butanenitrile, which lack toxicity data. Table 5.7 summarizes the AOELs for each nitrile in this family.

It is unclear why the EPA Acute Exposure Guideline Level 3<sup>(a)</sup> (AEGL3) for propanenitrile is lower than the NIOSH REL (see Table 5.4). At this time, no documentation is available to assess the derivation of the AEGL3. As additional new data become available for nitriles, it would be reasonable to reassess the appropriateness of this proposed AOEL.

### 5.7.1.5 Conclusion

The nitriles are structurally similar and consist of an alkyl or alkene side-chain and a nitrile functional group. Toxicity is most likely mediated through the cyanide metabolite. The known NOELs for developmental effects and LD<sub>50</sub>s are similar for this class of chemicals. The proposed AOELs for this family of nitriles are based on the NIOSH RELs for nitriles, with an additional safety factor when toxicity data are not available. The proposed **AOELs of 0.3 to 8 ppm** are comparable to the NIOSH RELs for propanenitrile and butanenitrile.

(a) The EPA proprionitrile results are available online: <http://www.epa.gov/oppt/aegl/results27.htm>.

**Table 5.7.** Proposed AOELS

Nitrile	Carbon #	AOEL (ppm)	Basis
Propanenitrile (CAS #107-12-0)	2	6	NIOSH
Butanenitrile (CAS # 109-74-0)	3	8	NIOSH
3-Butenenitrile (CAS# 109-75-1)	3	1	Methacrylonitrile Surrogate: NIOSH REL, ACGIH TLV
2-Methylene butanenitrile (CAS#1647-11-6)	3	0.3	Methacrylonitrile Surrogate: NIOSH REL, ACGIH TLV, with UF of 3 due to lack of toxicity data
Pentanenitrile (CAS# 110-59-8)	4	6	Propanenitrile
Hexanenitrile (CAS # 628-73-9)	5	6	Propanenitrile
Heptanenitrile (CAS# 629-08-3)	6	6	Propanenitrile
Octanenitrile (CAS # 124-12-9)	7	6	Propanenitrile
Nonanenitrile (CAS #2243-27-8)	8	6	Propanenitrile
Decanenitrile (CAS # 1975-78-6)	9	6	Propanenitrile
2,-4-Pentadienenitrile (CAS# 1615-70-9)	4	0.3	Methacrylonitrile Surrogate: NIOSH REL, ACGIH TLV, with UF of 3 due to lack of toxicity data



## 5.8 Amines and Amides

### 5.8.1 Acetamide (CAS #60-35-5)

In the interest of protecting the health of tank farm workers at DOE's Hanford Site, chemicals found within tank headspaces that the IARC lists as "possibly carcinogenic to humans" (2B) are being evaluated. The evaluation process will determine if there is a need to establish an AOEL for worker exposure based on a review of all available exposure and toxicological data as described in Sections 2.0 - 4.0. IARC assigned acetamide a 2B carcinogen classification based on a paper by Fleischman et al. (1980) in which a diet of 2.36% acetamide resulted in hepatocarcinomas in male and female rats, and diets of 1.18 and 2.36% resulted in malignant lymphomas in male mice (Table 5.8). Only IARC and California EPA have assigned a carcinogen classification for acetamide; neither the EPA nor OSHA has made such a designation. The NTP investigated the mutagenic potential of acetamide in drosophila and salmonella and the results were negative. The bioassay results from the NTP were "inconclusive and not reportable" (<http://ntp.niehs.nih.gov/index.cfm?objectid=07058954-ED59-D6FE-5E91CDDBC78C556E>).

There are more data available for the related chemicals N,N-dimethylacetamide and formamide (Figure 5.2). N,N-Dimethylacetamide is metabolized to acetamide in the liver. Acetamide and formamide are very similar, but acetamide is metabolized more slowly. The ACGIH, OSHA, and NIOSH have all set exposure limits for N,N-dimethylacetamide at 10 ppm TWA. The ACGIH recommended AOEL for formamide is also 10 ppm. Neither formamide nor dimethylacetamide have been assigned possible human carcinogen status.

Acetamide, dimethylacetamide, formamide, and dimethylformamide all cause progressive hepatotoxicity (Kennedy 1986). The hepatocellular carcinomas observed by Fleischman et al. (1980) may have been due to hepatocellular cytotoxicity and concomitant compensatory cellular regeneration, which would have a threshold effect. Also, the ACGIH TLV of 10 ppm for demethylacetamide should provide a good estimate for an acetamide TLV, and even after application of a 10x safety factor, the margin of exposure for acetamide (and related compounds, see Table 5.9) should be sufficient (estimated OEL of 1 ppm compared to tank headspace concentration of 0.0032 ppm). Therefore, no OEL is proposed for acetamide at this time. As additional new data become available for acetamide, it would be reasonable to reassess the appropriateness of not assigning an OEL based on the potential carcinogenicity.

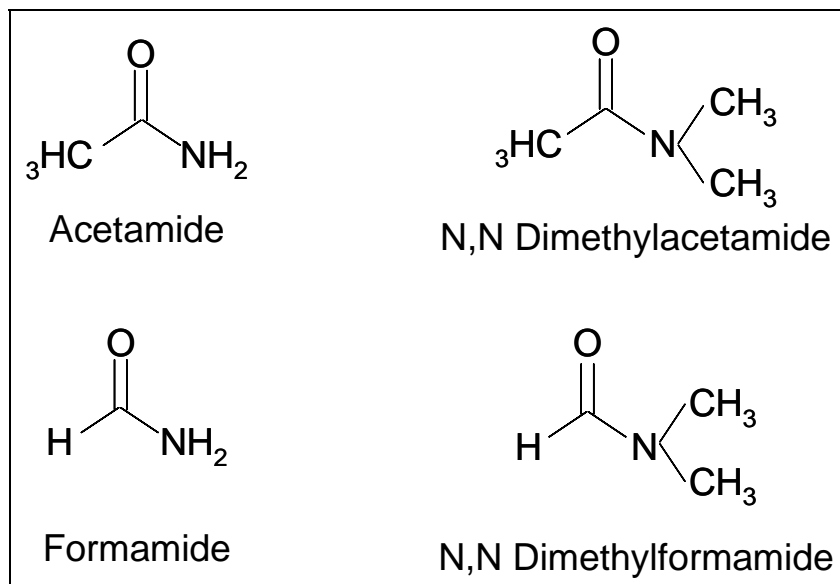
**Table 5.8.** Incidence of Cancers in Rats and Mice Fed Diets Containing Acetamide

Species	Acetamide (%)	Endpoint	Incidence	Control
Rat (male) <sup>(a)</sup>	2.36	Hepatocellular Carcinomas	41/47	0/50
Mouse (male)	1.18	Malignant lymphoma	7/50	0/79
Mouse (male)	2.36	Malignant lymphoma	7/46	0/79

(a) Male and female rats were dosed, with similar outcomes, but male rats had slightly higher incidence and quicker time to tumor development Fleischman et al. (1980).

**Table 5.9.** Acetamide and Related Chemicals

CAS or TWINS Number	Chemical	Max. Conc. (ppm)	Surrogate	Surrogate OEL Source	Safety Factor	Screening Value (ppm)
60-35-5	Acetamide	0.0032	N,N-Dimethylacetamide	TLV	1000	0.01
79-16-3	N-Methylacetamide	0.00037	N,N-Dimethylacetamide	TLV	100	0.1
541-35-5	Butanamide	0.020	N,N-Dimethylacetamide	TLV	100	0.1
10264-17-2	N-Hexylbutanamide	0.000052	N,N-Dimethylacetamide	TLV	100	0.1



**Figure 5.2.** Acetamide and Related Compounds

### 5.8.2 N-Methylaziridine (CAS# 1072-44-2)

An AOEL is proposed for N-methylaziridine. Aziridines are 3 membered heterocyclic compounds with an amine group. N-Methylaziridine has a methyl group bound to the ring nitrogen. Aziridines have high-angle strains due to the tight nature of the 3-member ring compared to larger ring structures. No toxicological data were found for N-methylaziridine, but it is similar to ethylenimine and propyleneimine, which have ACGIH-recommended TLVs of 0.5 and 2 ppm, respectively. These ACGIH TLVs are intended to minimize respiratory tract irritation. Ethylenimine and propyleneimine are confirmed animal carcinogens. **The proposed AOEL for N-methylaziridine is 0.05 ppm** based on the ACGIH TLVs for the closely relate compound ethylenimine with an UF of 10.

#### 5.8.2.1 Methodology

The procedures described in Sections 2.0 - 4.0 were used as a guide for developing this AOEL. To identify and access available toxicity data on N-methylaziridine, a number of Internet data bases including PUBMED, TOXNET, and TOMES were searched, and summary toxicity profiles were obtained and primary references identified. In addition, ACGIH documentation was reviewed for relevant surrogate chemicals. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profile and develop the AOEL.

### 5.8.2.2 Available Guidelines

No OEL for N-methylaziridine has been proposed by the OSHA or ACGIH. The ACGIH has proposed TLVs of 0.5 and 2 ppm for ethylenimine and propylenimine, structurally similar compounds. In addition, OSHA has assigned a Permissible Exposure Limit (PEL) of 2 ppm, and NIOSH has assigned an Immediately Dangerous to Life or Health (IDLH) of 100 ppm for propylenimine. Skin notations have been recommended for these aziridines.

### 5.8.2.3 Toxicology Summary

Aziridines are reactive due to the ring strain, and they interact readily with nucleophiles. The substituted polyethylenimine is not mutagenic in salmonella strains. Aziridine and aziridines with smaller substitutions (i.e., propylenimine) are mutagenic and carcinogenic in laboratory animals. Aziridines are irritating to mucous membranes, skin, eyes, and lungs. Acute exposures of laboratory animals have resulted in congestion of the respiratory tract, edema, hemorrhage, and kidney damage. The 10-minute LC<sub>50</sub> for aziridine is 2200 ppm, but exposures of 4 or more hours to as low as 25 ppm has resulted in death in rats and guinea pigs. Embryotoxicity (decreased body weights, hematomas, decreased live births) have been observed with inhalation exposure concentrations as low as 0.6 ppm (NTP). Although animal studies to low concentrations result in serious effects, exposures in humans have resulted in less notable outcomes. Vapor exposures in human studies have resulted in no ill-effects for up to 3 hours. After 3 hours, vomiting and eye and nose irritation was reported. Acute dermal or inhalation exposures have resulted in central nervous system effects, fluid in the lungs, and damage to liver and kidneys (ACGIH 20011).

Propylenimine has less potent toxicity than aziridine. The physiological action of propylenimine, however, resembles that of aziridine (ACGIH 2001z). The LCL0 for rats is 500 ppm for a 4-hour exposure. N-Aziridineethanol is an N-substituted aziridine with an ethanol group. Like the other aziridines, N-aziridineethanol is an eye and skin irritant that also has limited data indicating it is an animal carcinogen (National Institutes of Health [NIH]).

### 5.8.2.4 Data Analysis

The most data were available for the unsubstituted aziridine. The ACGIH has recommended TLV-TWAs of 0.5 and 2 ppm for aziridine and propylenimine, respectively. The toxicity and carcinogenicity of aziridines with substitutions smaller than three carbons is similar. Due to the ring strain, these compounds are reactive. The primary endpoint of concern is irritation of skin, eyes, and lungs, and, for some aziridines, kidney damage. The ACGIH TLV for aziridine is four-times lower than for propylenimine due to the decreased toxicity for that substituted aziridine over the unsubstituted ring. The N-substitute aziridine is not likely to result in significant decrease in ring strain.

### 5.8.2.5 Conclusion

Based on the lack of toxicological data for N-methylaziridine, it is proposed that the ACGIH TLV of 0.5 ppm be used with an UF of 10 to propose an OEL for **N-methylaziridine of 0.05 ppm**. As additional new data become available for the N-methylaziridine, it would be reasonable to reassess the appropriateness of basing the AOEL upon the WEEL for substituted pyridines.

## 5.9 Nitrites and Nitrates

### 5.9.1 Short-Chain Alkyl Nitrites: n-Butyl nitrite (CAS# 544-16-9), Ethyl nitrite (CAS# 109-95-5), and Methyl nitrite (CAS# 624-91-1)

The principal toxic effect of exposures to volatile organic nitrites results from methemoglobinemia. Evidence indicates that the methemoglobinemia is mediated by the nitrite ion after the organic nitrite is oxidized; therefore, it should be possible to set an AOEL for the family of organic nitrites of interest. The ACGIH has suggested a TLV of 1 ppm for the related organic volatile nitrite, isobutyl nitrite. Using the TLV for isobutyl nitrite and applying a safety margin for chemical differences, the proposed AOEL for the nitrite family is 0.1 ppm (100 ppb).

#### 5.9.1.1 Toxicology Summary

Nitrate ( $\text{NO}_2^-$ ) is a naturally occurring inorganic ion. Nitrites react with the iron ( $\text{Fe}^{2+}$ ) of deoxyhemoglobin and form methemoglobin, resulting in iron incapable of transporting oxygen ( $\text{Fe}^{3+}$ ). The principal toxicology of nitrite exposures stems from methemoglobinemia and includes cyanosis, cardiac dysrhythmias, hemolytic anemia, circulatory failure, and progressive central nervous system effects. Children are especially at risk from oral exposures.

Volatile organic nitrites (e.g., butyl and amyl nitrite) are in medications and drugs of abuse. Production of nitric oxide from exposures to several nitrites has been observed (Cederqvist et al. 1994), and is likely to represent a common mechanism between chemicals of this class. Cyanosis is the major effect seen in animals following inhalation exposures to butyl nitrite. In rats, the  $\text{LC}_{50\text{s}}$  for the primary three nitrites of interest were 176, 160, and 420 ppm/4 hours for methyl, ethyl, and butyl nitrite, respectively (Klonne et al. 1987). No reports on chronic effects from nitrite exposures were found.

The potential of nitrites to cause cancer in humans is unknown. Isobutyl nitrite vapor was mutagenic in the Ames assay (Mirvish et al. 1993) and in salmonella tests (NTP 1996). Several nitrites have been shown to be mutagenic in vitro, including n-butyl, isobutyl, iso-amyl, sec-butyl, and n-propyl in mice lymphoma cells (Dunkel et al. 1989), and methyl nitrite (with and without an S9 activating system) in Salmonella (Tornqvist et al. 1983).

Although no in vivo animal carcinogenicity studies were identified for n-butyl, ethyl, or methyl nitrites, a two-year bioassay with the structurally similar isobutyl nitrite was reported (NTP 1996). In the two-year inhalation studies, there was “clear evidence” of increased incidences of alveolar and bronchiolar adenomas and carcinomas in rats, and some evidence of carcinogenic activity in B6C3F1 mice (NTP 1996). Based on the results of this study, ACGIH has recommended that isobutyl nitrite be given the notation of Confirmed Animal Carcinogen with Unknown Relevance to Humans (A3).

#### 5.9.1.2 Conclusion

The most robust set of toxicological data were for butyl nitrite. A single study did show that the  $\text{LC}_{50}$  for nitrites was lowest, with the smaller side chains (methyl  $\approx$  ethyl < butyl). In reviewing the available exposure guidelines for organic nitrites, only an ACGIH guideline for isobutyl nitrite was found. An additional safety factor of 10 is proposed to account for chemical differences between isobutyl nitrite and methyl- or ethyl- nitrite. Since the TLV for isobutyl nitrite has been set by ACGIH at 1 ppm, a proposed

AOEL for the family of short-chain alkyl nitrites (such as methyl, ethyl, and butyl) is 0.1 ppm. As additional and appropriate new data become available for nitrites, it would be reasonable to reassess the appropriateness of basing the alkyl nitrite's AOEL upon isobutyl nitrite.

### **5.9.2 Short-Chain Alkyl Nitrates Methyl Nitrate (CAS #598-58-3), Ethyl Nitrate (CAS # 625-58-1), Butyl Nitrate (CAS # 120-92-3), and Propyl Nitrate (CAS # 627-13-4)**

An AOEL is proposed for the short-chain alkyl nitrates (methyl, ethyl, and butyl) based on a comprehensive review of the available toxicology literature. These four nitrates are structurally similar and consist of a one- to four-carbon alkyl side-chain and a nitrite functional group. The principal toxic effects of exposures to volatile short-chain alkyl nitrates result from methemoglobinemia. Nitrates are converted to nitrites, and the toxicity is mediated by the nitrite ion after the organic nitrite is oxidized; therefore, it should be possible to set an AOEL for the family of short-chain alkyl nitrates of interest. Although the organic nitrate identified with the most available data is propyl nitrate, the nitrates of interest are methyl and ethyl nitrate. A TLV of 25 ppm has been assigned by ACGIH for n-propyl nitrate. Based on the uncertainty associated with chemical differences between propyl nitrate and methyl and ethyl nitrate, an AOEL of 8 ppm is proposed for these nitrates.

#### **5.9.2.1 Methodology**

To identify and access available toxicity data on organic nitrates, a number of Internet data bases including PUBMED, TOXNET, TOMES, the EPA-IRIS, and IARC were searched, and summary toxicity profiles were obtained and primary references identified. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profile to develop the AOEL. Primary references for the most relevant toxicology endpoints were obtained for review.

#### **5.9.2.2 Available Guidelines**

No OEL for ethyl, methyl, or butyl nitrate has been proposed by OSHA or ACGIH. However, NIOSH, OSHA, and ACGIH have all set OELs for propyl nitrate (see Table 5.10). From 1971 to 2001, the TLVs have not changed and are 25 or 40 ppm, depending on the length of exposure, for all agencies.

#### **5.9.2.3 Toxicology Summary**

Organic nitrates are used in the synthesis of other compounds as a room deodorizer propellant and in rocket propellants. Butyl nitrate and the related amyl nitrate are in the inhalant drugs of abuse known as Poppers and, therefore, exposures to these nitrates can be high. Butyl nitrate is a degradation product of tributyl phosphate (Tashiro et al. 2000). Environmental contamination with nitrates often stems from nitrogen-containing fertilizers. In the environment, organic nitrogen is converted to nitrite and nitrate by microbes, and nitrite is oxidized to nitrate so nitrate is found in high levels in groundwater. Worldwide, nitrate-contaminated groundwater is common, especially around farming communities where nitrogen-containing fertilizers are used. Epidemiological studies indicate that nitrate contamination of water supplies may be associated with risk of Type I diabetes, but only at concentrations greater than 10 to 25 mg/L (Van Maanen et al. 2000).

**Table 5.10.** Exposure Limit Values for Propyl Nitrate

Agency	Description	Value (ppm)	Date <sup>(a)</sup>
ACGIH	TLV-STEL	40	2001
ACGIH	TLV-TWA	25	2001
CDC <sup>2</sup>	IDLH	500	1994
NIOSH	STEL	40	1994
NIOSH	REL	25	1994
OSHA	PEL-STEL	40	1989
OSHA	PEL-TWA	25	1994

(a) The date listed is the most recent found; in some cases an earlier OEL had been proposed.  
(b) See NTIS (1995).  
IDLH = Immediately dangerous to life or health.  
PEL = Permissible exposure limit.  
REL = Recommended exposure limit.  
STEL = Short-term Exposure Limit.  
TWA = Time-weighted Average.

Nitrate ( $\text{NO}_3^-$ ) and nitrite ( $\text{NO}_2^-$ ) are naturally occurring inorganic ions. Nitrates are converted to nitrites by intestinal bacteria. Nitrites, in turn, react with the iron ( $\text{Fe}^{2+}$ ) of deoxyhemoglobin and form methemoglobin, resulting in iron incapable of transporting oxygen ( $\text{Fe}^{3+}$ ). The principal toxicity of nitrate exposures stems from methemoglobinemia and includes cyanosis, cardiac dysrhythmias, hemolytic anemia, circulatory failure, and progressive central nervous system effects. Children are especially at risk from oral exposures. The EPA has derived a reference dose (RfD) of 1.6 mg/kg/day for inorganic nitrate (IRIS 2002); a RfC has not been derived.

Orally, nitrates are rapidly absorbed from the small intestine. Regardless of exposure route, nitrates are delivered to the intestine via blood circulation and converted, primarily by bacterial oxidation, to the more reactive nitrites; therefore, nitrites are expected to have lower  $\text{LD}_{50}$  and  $\text{LC}_{50}$  values than nitrates (Bruning-Fann and Kaneene 1993). The  $\text{LC}_{50}$  value for methyl nitrate in rats is 1275 ppm/4 hr; the  $\text{LC}_{50}$  for the corresponding nitrite is seven times lower (176 ppm/4 hr).<sup>(a)</sup> Organic nitrites are rapidly converted in the liver to inorganic nitrite. The toxic mechanism of action leading to methemoglobinemia proceeds through either direct oxidation of the heme iron by the nitrite, or through production of NO from the nitrite (Cederqvist et al. 1994). Production of nitric oxide from exposures to several nitrites and nitrates has been observed (Cederqvist et al. 1994) and is likely to represent a common mechanism among chemicals of this class. The risk of methemoglobinemia from exposure of nitrate depends on the dose of nitrate and on the number and type of enteric bacteria.

The potential of nitrates to cause cancer in humans is unknown. Nitrates can react with amino acids to form nitrosamines, which can cause cancer in animals. An epidemiological study in workers exposed to nitrate fertilizers has shown a slight increase in stomach cancer incidence, but no increase in total cancer (Zandjani et al. 1994). Other studies have shown no relationship between occupation exposures to nitrate containing dust and any cancers (Fraser et al. 1982, 1989).

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(a) Refer to the Hazardous Substances Data Bank online: <http://csi.micromedex.com/assm.asp?HS7197>.

#### 5.9.2.4 Data Analysis

No guidelines have been established for methyl, ethyl, or butyl nitrate (C-1, C-2, and C-4 side-chains, respectively). Exposure-level guidelines have been established by several agencies for propyl nitrate (see Table 5.10). The ACGIH TLVs were recommended based on acute and subchronic animal studies.

Propyl nitrate causes sensory irritation at exposure levels much lower than needed to produce measurable methemoglobinemia (Rinehart et al. 1958). Therefore, there were no relevant human studies identified by the ACGIH. The TLV was based on the acute inhalation exposure LC<sub>50</sub> in dogs (2000 to 2500 ppm) with a “reasonable uncertainty” factor. The LD<sub>50</sub> for methyl nitrate in rats is about two times lower than the LC<sub>50</sub> value reported for propyl nitrate in dogs. Due to the potential increase in toxicity of the shorter-chain nitrates and the uncertainty associated with chemical differences, an additional uncertainty factor of 10 is recommended for the proposed AOEL for propyl nitrate, which results in an **AOEL of 2.5 ppm**. The ACGIH also concluded that there is insufficient data to warrant skin or carcinogenicity notations for propyl nitrate.

#### 5.9.2.5 Conclusion

The most robust set of toxicological data were for propyl nitrate. The LC<sub>50</sub> for acute inhalation of methyl nitrate in rats is 1275 ppm/4 hours, the acute LC<sub>50</sub> for propyl nitrate in dogs is 2000 to 2500 ppm/4 hours. Several agencies have established OELs for propyl nitrate, and these values were used to propose an AOEL for methyl and ethyl nitrate. An additional safety factor of 3 is proposed to account for chemical differences between propyl nitrate and the shorter-chain nitrates. Since the TLV for propyl nitrate has been set by ACGIH, OSHA, and NIOSH at 25 ppm (TWA), a proposed AOEL for the family of short-chain alkyl nitrates (methyl and ethyl) is **8 ppm**. As additional new data become available for nitrates, it would be reasonable to reassess the appropriateness of basing the alkyl nitrate’s AOEL upon propyl nitrate.

### 5.9.3 Dinitrates: 1,3-Propanediol, dinitrate (CAS # 3457-90-7), 1,4-Butanediol, dinitrate (CAS # 3457-91-8), 1,5-Pentanediol, dinitrate (CAS # 3457-92-9), and 1,2,3-Propanetriol, 1,3-dinitrate (CAS # 623-87-0)

An AOEL is proposed for the dinitrates, based on a comprehensive review of the available toxicology literature. The principal toxic effects of exposures to nitrates result from methemoglobinemia. Dinitrates are also vasodilators, but the mechanism of action is incompletely understood. The development of proposed AOELs for the dinitrates are based on short-chain aliphatic nitrites, propylene glycol dinitrate, and nitroglycerine. Nitrates are converted to nitrites, and the methemoglobinemia is mediated by the nitrite ion after the organic nitrite is oxidized. The ACGIH has assigned TLVs of 0.05 ppm for ethylene glycol dinitrate, propylene glycol dinitrate, and nitroglycerine. The ACGIH TLV has already considered the incomplete understanding of the mechanism of action and the use of surrogate chemicals, therefore, an **AOEL of 0.05 ppm** is proposed for these dinitrates.

#### 5.9.3.1 Methodology

To identify and access available toxicity data on dinitrates, a number of Internet data bases, including PUBMED, TOXNET, TOMES, the EPA-IRIS, and IARC were searched, and summary toxicity profiles were obtained and primary references identified. These toxicology summaries were then reviewed to

identify the most critical data sets that could be used to establish the toxicity profile and develop the AOEL. Primary references for the most relevant toxicology endpoints were obtained for review.

### 5.9.3.2 Available Guidelines

No OELs are available for the dinitrates found in the tank waste headspace. The ACGIH has set TLVs of 0.05 ppm for the related dinitrates propylene glycol dinitrate, (ACGIH 2001y), ethylene glycol dinitrate (ACGIH 2001k) and the trinitrate, nitroglycerine (ACGIH 2001t). The TLVs for all of these compounds are based on assessment of propylene glycol dinitrate (ACGIH 2001y). OSHA and NIOSH have also set REL and PELs for propylene glycol at a TWA of 0.05 ppm, respectively.

### 5.9.3.3 Toxicology Summary

Very little data were found concerning the toxicity of these dinitrates. By analogy with propylene glycol dinitrate, and other nitrates, a probable toxicity profile can be estimated. When ingested orally, nitrates are rapidly absorbed from the small intestine. Regardless of exposure route, nitrates are delivered to the intestine via blood circulation and converted, primarily by bacterial oxidation, to more reactive nitrites. Organic nitrites are rapidly converted in the liver to inorganic nitrite. The toxic mechanism of action leading to methemoglobinemia proceeds through either direct oxidation of the heme iron by the nitrite or through production of NO from the nitrite (Cederqvist et al. 1994). Production of nitric oxide from exposures to several nitrites and nitrates has been observed (Cederqvist et al. 1994) and is likely to represent a common mechanism among chemicals of this class. The risk of methemoglobinemia from exposure of nitrate depends on the dose of nitrate and on the number and type of enteric bacteria.

The related compound with the most available toxicological information is propylene glycol dinitrate (CAS# 6423-43-4). The primary toxicological endpoints of concern for propylene glycol dinitrate exposures are its effects on the cardiovascular and central nervous systems. Human exposures have led to symptoms of dizziness, eye irritation, nasal congestion, heart palpitations, and chest pain. Dizziness has been reported for inhalation exposures as low as 1.5 ppm, and the threshold for mild headache was 0.5 ppm (reviewed in: National Research Council 2001).

### 5.9.3.4 Data Analysis

No guidelines have been established for the dinitrates within the tank waste headspace. The ACGIH has set TWA-TLV of 0.05 ppm for related dinitrates based primarily on the data available for propylene glycol dinitrate. This TLV is 500-fold lower than the ACGIH TWA for n-propyl nitrate (see Section 5.9.2). Due to the vasodilation effects from exposures to dinitrates and potential increase in toxicity of the dinitrates over the nitrates, it is proposed that the **AOEL of 0.05 ppm** set by ACGIH for propyl glycol dinitrate, ethyl glycol dinitrate, and nitroglycerine be applied to the dinitrates within the tank waste headspace.

### 5.9.3.5 Conclusion

The most robust set of toxicological data were for propylene glycol dinitrate. The ACGIH has used this dataset to base AOELs for other di- and tri-nitrates. By analogy with the short-chain aliphatic mono-nitrates, this AOEL of 0.05 ppm is conservative for these dinitrates. As additional new data become available for dinitrates, it would be reasonable to reassess the proposed AOEL.



#### **5.9.4 1-Nitrate 1,2,3-propanetriol (CAS # 624-43-1)**

An AOEL is proposed for 1-nitrate 1,2,3-propanetriol based on a comprehensive review of the available toxicology literature. 1-Nitrate 1,2,3-propanetriol consists of a nitrate functional group and a glycerol sidechain. The principal toxic effects of exposures to nitrates result from methemoglobinemia. Nitrates are converted to nitrites, and the methemoglobinemia is mediated by the nitrite ion after the organic nitrite is oxidized. Very high exposures to glycerol mist may result in hemolysis, hemoglobinuria, and renal failure. The ACGIH has assigned TLVs of 2.65 and 25 ppm for glycerin mist and n-propyl nitrate vapor, respectively. Following metabolism in the intestine and liver, the likely metabolites will be inorganic nitrite and propylene glycol. Since propylene glycol is relatively non-toxic, the proposed AOEL for 1-nitrate 1,2,3-propanetriol is based on the 25 ppm OEL for n-propyl nitrate with a safety factor of 3 for uncertainties between chemical differences. The proposed AOEL for 1-nitrate 1,2,3-propanetriol is **8 ppm**.

##### **5.9.4.1 Methodology**

To identify and access available toxicity data on 1-nitrate 1,2,3-propanetriol, a number of Internet data bases, including PUBMED, TOXNET, TOMES, and IARC were searched, and summary toxicity profiles were obtained and primary references identified. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profile and develop the AOEL. Primary references for the most relevant toxicology endpoints were obtained for review.

##### **5.9.4.2 Available Guidelines**

No OEL for 1-nitrate 1,2,3-propanetriol has been proposed by either OSHA or ACGIH. NIOSH, OSHA, and ACGIH have all set 25 ppm TWA OELs for propyl nitrate. Likewise, OSHA and ACGIH have both set an OEL of 10 mg/m<sup>3</sup> (2.65 ppm) for glycerin particulate exposures.

##### **5.9.4.3 Toxicology Summary**

Very little data were found concerning the toxicity of 1-nitrate 1,2,3-propanetriol. By analogies with propyl nitrate, nitroglycerine, glycerin, and propylene glycol, a probable toxicity profile can be estimated. When ingested orally, nitrates are rapidly absorbed from the small intestine. Regardless of exposure route, nitrates are delivered to the intestine via blood circulation and converted, primarily by bacterial oxidation, to more reactive nitrites. Organic nitrites are rapidly converted in the liver to inorganic nitrite. The toxic mechanism of action leading to methemoglobinemia proceeds through either direct oxidation of the heme iron by the nitrite, or through production of NO from the nitrite (Cederqvist et al. 1994). Production of nitric oxide from exposures to several nitrites and nitrates has been observed (Cederqvist et al. 1994) and is likely to represent a common mechanism among chemicals of this class. The risk of methemoglobinemia from exposure of nitrate depends on the dose of nitrate and on the number and type of enteric bacteria.

##### **5.9.4.4 Data Analysis**

No guidelines have been established for 1-nitrate 1,2,3-propanetriol. Exposure level guidelines have been established by several agencies for propyl nitrate (see alkyl nitrates AOEL). The ACGIH TLVs were recommended based on acute and subchronic animal studies. Propyl nitrate causes sensory irritation at

exposure levels much lower than needed to produce measurable methemoglobinemia (Rinehart et al. 1958). Propylene glycol is relatively non-toxic, and human ingestion of up to 100 ml primarily leads to gastrointestinal upset. Due to the uncertainty associated with chemical differences, an additional uncertainty factor of 3 is recommended for the proposed AOEL for 1-nitrate 1,2,3-propanetriol, which results in an AOEL of 8 ppm. The proposed AOEL for glycerol is based on mist exposures, rather than vapor, and is therefore not applicable for vapors in the tank waste headspace.

#### **5.9.4.5 Conclusion**

The most likely active moiety of 1-nitrate 1,2,3-propanetriol will be the nitrite following metabolism in the gastrointestinal tract and liver and leaving inorganic nitrite and propylene glycol. Alternative metabolism may lead to the production of glycerin. Inorganic nitrite and glycerin have more toxicity associated with exposures than does propylene glycol. Based on the supposition that the metabolism will follow that typical for alkyl nitrates, an **AOEL of 8 ppm**, based on n-propyl nitrate is proposed. As additional new data become available for 1-nitrate 1,2,3-propanetriol, it would be reasonable to reassess the proposed AOEL.

## 5.10 Nitro Compounds

### 5.10.1 2-Nitro-1-propanol (CAS# 2902-96-7), 2-nitro-methylpropane (CAS# 594-70-7), and 1-nitrobutane (CAS# 627-05-4)

The principal toxic effects that stem from exposures to nitro compounds are irritation of eyes and respiratory tract and liver damage. Because no OELs are available for the nitro compounds of concern, 2-nitropropane (branched) and 1-nitropropane (unbranched) were chosen as surrogates with the most data and available OELs. The branched surrogate nitro compounds are more potent hepatotoxicants than the non-branched. Likewise, 2-nitropropane is an animal hepatocarcinogen, probably due to cellular damage and compensatory proliferation. The ACGIH has recommended TLVs of 10 and 25 ppm for 2-nitropropane (branched) and 1-nitropropane, respectively. 1-Nitrobutane is an unbranched nitro compound with a longer side chain than 1-nitropropane, the recommended **OEL for 1-nitrobutane is 2.5 ppm** and was based on the OEL for 1-nitropropane with an uncertainty factor of 10 for minor structural differences and limited data. 2-Nitro-methylpropane is a branched nitro compound, with 1 more methyl group than 2-nitropropane, the recommended **OEL for 2-nitro-methylpropane is 0.3 ppm** and was based on the OEL for 2-nitropropane with an uncertainty factor of 10 for structural differences and 3 for limited toxicity data. 2-Nitro-1-propanol has an OH group that the other two chemicals of concern do not have. The recommended OEL for **2-nitro-1-propanol is 0.8 ppm** based on the OEL for 1-nitropropane with an uncertainty factor of 10 for structural differences and an additional UF of 3 for limited data.

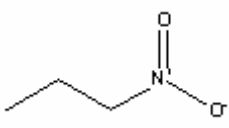
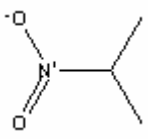
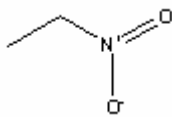
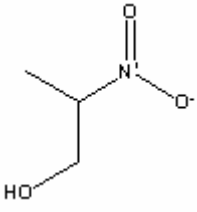
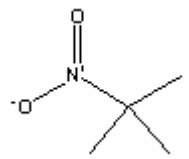
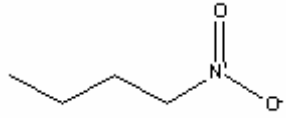
#### 5.10.1.1 Methodology

The procedures described in Sections 2.0 - 4.0 for were used as a guide for developing this AOEL. To identify and access available toxicity data on 2-nitro-1-propanol, 2-nitro-methylpropane, and 1-nitrobutane, a number of Internet data bases including PUBMED, TOXNET, and TOMES were searched, and summary toxicity profiles were obtained and primary references identified. In addition, ACGIH documentation was reviewed for nitroethane, 2-nitropropane, and 1-nitropropane and relevant surrogate chemicals. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profile and develop the AOEL.

#### 5.10.1.2 Available Guidelines

No OELs have been proposed by OSHA or ACGIH for 2-nitro-1-propanol, 2-nitro-methylpropane, or 1-nitrobutane. However, ACGIH has assigned a TLV-TWA for the structural surrogates, nitroethane, 2-nitropropane, and 1-nitropropane. 1-Nitropropane is structurally very similar to 1-nitrobutane, differing only in one carbon. Likewise, 2-nitropropane is comparable to 2-nitro-methylpropane, differing in the side-chains; the former is secondary and the latter is a tertiary structure (Table 5.11). The ACGIH documentation for 1-nitropropane indicates that human exposures to 100 ppm or more caused irritation to the eyes and recommended a TLV of 25 ppm to minimize the potential of irritation to eyes and respiratory tract (ACGIH 2001u). The branched 2-nitropropane causes liver damage in laboratory animals, and hepatocellular carcinomas were noted in rats necropsied after 6 months of exposures to 207 ppm. For shorter exposure durations, progressive liver damage was observed, suggesting that the hepatocellular carcinomas were a consequence of cellular regeneration and proliferation secondary to damage. ACGIH recommended a TLV of 10 ppm for 2-nitropropane (ACGIH 2001v).

**Table 5.11.** Structure and Proposed OELs for Nitro Compounds

		OEL/proposed OEL	Basis
1-Nitropropane		25	ACGIH
2-Nitropropane		10	ACGIH
Nitroethane		100	ACGIH
2-Nitro-1-propanol		0.8	1-Nitropropane, UF 3 for limited data, UF 10 for structural differences
2-Nitro-methylpropane		0.3	2-Nitropropane, UF 30 for structural differences and limited data
1-Nitrobutane		2.5	1-Nitropropane UF 10 for lack of data

### 5.10.1.3 Toxicology Summary

No toxicity data were found for any of these nitro compounds, necessitating analysis based solely on structurally similar surrogates. For 2-nitro-methylpropane and 1-nitrobutane, the structural similarities to 2-nitropropane and 1-nitropropane, respectively, results in a reasonable level of confidence in using these surrogates. Structural differences for 2-nitro-1-propanol compared to the surrogates are more extensive (discussed in more detail below).

The surrogate for 1-nitrobutane, 1-nitropropane was not mutagenic in Salmonella tester strains. The primary endpoint of concern for 1-nitropropane exposures is irritation of the eyes and respiratory tract (ACGIH 2001u). Acute (1-3 hours) inhalation exposures in rats to 10,000 ppm resulted in conjunctival irritation, lacrimation, and slowed respiration. Rats exposed via inhalation for 7 hours/day, 5 days/week for ~22 months showed no gross or histopathological effects of the liver upon necropsy, and there was also no effect of exposure on organ or body weights or hematology (Table 5.12).

**Table 5.12.** Comparative Toxicity Parameters for Nitro Compound Surrogates

	NOEL <sup>(a)</sup>	LC <sub>50</sub> (ppm)	LD <sub>50</sub> (mg/kg)	Other
Surrogates				
2-methyl-2-nitro-1-propanol			844-1480	
1-nitropropane	101 ppm		250-500	LC100 = 5000 ppm/3 hr LC10 = 714-4622 ppm/4.5 hr
2-nitropropane	27 ppm	400	500-750	
nitroethane	500 ppm			
(a) See text for exposures conditions.				

The endpoint of concern for 2-nitropropane is liver damage. The acute lethality of 2-nitropropane was less than the non-branched 1-nitropropane (See Table 5.12). Rats exposed to 2-nitropropane at 207 ppm for 7 hours/day, 5 days/week, for 6 months developed hepatocellular carcinomas. 2-Nitropropane was also positive for genotoxicity in the Ames test but not in other in vitro genotoxicity studies (i.e., Drosophila recessive assay). Worker exposures to 20-45 ppm resulted in nausea, vomiting, diarrhea, anorexia, and headache.

2-Nitro-1-propanol is a branched nitro compound with a hydroxyl (-OH) group; no surrogate with these structural features and an OEL proposed by ACGIH or OSHA was identified. Some acute toxicity data for a structural analogue, 2-methyl-2-nitro-1-propanol (CAS# 76-39-1) was identified (See structure in Table 5.12). 2-Methyl-2-nitro-1-propanol is produced by the reaction of nitroparaffin with formaldehyde and is used in the production of other alkanolamines. The Dow Chemical Company has assessed the acute toxicity of 2-methyl-2-nitro-1-propanol as a part of its High Production Volume Chemical Challenge Test Plan (<http://www.epa.gov/chemrtk/nitroalc/c14355tp.pdf>) and reports an LD<sub>50</sub> of 845-1480 mg/kg in rats. 2-Methyl-2-nitro-1-propanol was not genotoxic in the Ames assay. The primary toxicological effects seen in laboratory animals following exposures to 2-methyl-2-nitro-1-propanol is eye irritation.

#### 5.10.1.4 Conclusion

In reviewing the available toxicity data bases on nitro compounds with similar structures to 2-nitro-1-propanol, 2-nitro-methylpropane, and 1-nitrobutane the most robust data were found for 2-nitropropane and 1-nitropropane, which differ from 2-nitro-methylpropane and 1-nitrobutane by the addition of an extra carbon group, respectively. The ACGIH-recommended TWA-TLVs for 2-nitropropane and 1-nitropropane are 10 and 25 ppm, respectively. An UF of 10 is recommended to extrapolate a recommended OEL for 2-nitro-methylpropane from 2-nitropropane due to the structural differences inherent with the 3° branching of the sidechain compared to the 2° branching of the surrogate, which may result in metabolic and/or toxicological differences. Thus, the recommended **OEL for 2-nitro-methylpropane is 0.3 ppm**. An UF of 10 is recommended to extrapolate a recommended OEL for

1-nitrobutane from 1-nitropropane due to the lack of data for 1-nitrobutane. The structural differences between the non-branched compounds are less extreme than for the branched nitro compounds. Thus, the recommended **OEL for 1-nitrobutane is 2.5 ppm**.

2-Nitro-1-propanol is a nitro alcohol and therefore structurally different from the surrogates. Several nitro alcohols are used in high quantities in the manufacture of alkanolamines (The Dow Chemical Company 2002). The acute LD<sub>50</sub> for 2-methyl-2-nitro-1-propanal is higher than the acute LD<sub>50</sub> for the 2-nitropropane and 1-nitropropane surrogates. The recommended **OEL for 2-nitro-1-propanol is 0.8 ppm** based on the TLV-TWA for 1-nitropropane with an UF of 10 for structural differences and an additional UF of 3 for limited data. 1-Nitropropane was used as the surrogate due to the listed toxicological endpoint of concern for the surrogate, 2-methyl-2-nitro-1-propanal, of sensory irritation, which was more similar to 1-nitropropane than 2-nitropropane.

## 5.11 Heterocyclics

### 5.11.1 Furan (CAS # 110-00-9)

An AOEL is proposed for furan based on a comprehensive review of the available toxicology literature. The most sensitive and relevant toxicological endpoint for establishing an AOEL for furan is the observed hepatic biliary tract hyperplasia (13-week sub-chronic study) leading to carcinogenicity following oral administration in rats and mice. Based on the available experimental data, the dose-response for both the hyperplasia and tumor response was very steep (i.e., near maximal response at all dose levels) with no experimentally determined NOEL. The results from the 13-week sub-chronic study were used to extrapolate from an observed 70% incidence at 4 mg/kg/day to a dose that would produce a 0.01% incidence ( $1 \times 10^{-4}$ ). The AOEL calculated for furan assumed a human inhalation rate of 10 m<sup>3</sup>/day, and 100% of the inhaled furan was absorbed systemically. The proposed AOEL for furan based on this analysis is ~4 µg/m<sup>3</sup> or 1 ppb.

#### 5.11.1.1 Methodology

To identify and access available toxicity data on furan, a number of Internet data bases including PUBMED, TOXNET, TOMES, EPA-developed IRIS, and one by IARC were searched, and summary toxicity profiles were obtained and primary references identified. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profile and to help identify or establish NOEL that could be used to develop the AOEL. Primary references for the most relevant toxicology endpoints were obtained for review.

#### 5.11.1.2 Available Guidelines

A review of the literature indicates that no documented OELs have been established by ACGIH, OSHA, or state regulatory agencies. In reviewing the Registry of Toxic Effects of Chemical Substances, it was noted that Russia had set a short-term exposure limit (STEL) of 0.5 mg/m<sup>3</sup> with a skin notation in January 1993; however, specific documentation was unavailable to evaluate how this STEL was established. In 1989, EPA established an oral reference dose (RfD) of 0.001 mg/kg/day based on sub-chronic studies in mice using a NOEL of 1.4 mg/kg/day for hepatic lesions (cellular necrosis), with a 1000-fold uncertainty factor applied (10x extrapolation sub-chronic to chronic; 10x interspecies; 10x for added protection sensitive subpopulation). This EPA assessment did not consider the results obtained from NTP chronic bioassay for furan (NTP 1993).

#### 5.11.1.3 Toxicology Summary

Furan is the parent compound for a broad class of structurally related compounds. IARC has classified furan as a possible carcinogen to humans (Group 2B), and it is a known animal carcinogen based on the observed carcinogenic response following oral administration in both mice and rats (NTP 1993). In rats, sub-chronic exposure to furan produced a high incidence of hepatic biliary tract hyperplasia and cholangiofibrosis in both male and females. In rats, chronic furan exposure produced hepatocellular adenomas in both male and female rats, and hepatocellular carcinomas in males. A very high incidence (~100%) of cholangiocarcinomas was also observed in both genders. The incidence of mononuclear-cell leukemia was also increased. Mice, likewise, showed an increased incidence of hepatocellular adenomas

and carcinomas in both genders. The cholangiocarcinomas showed the steepest dose-response. There is some suggestion that furan is bioactivated to a reactive metabolite that can readily alkylate proteins, resulting in significant cytotoxicity.

#### 5.11.1.4 Data Analysis

The most sensitive and relevant toxicological endpoint for establishing an AOEL for furan is the observed carcinogenicity following oral administration in rats and mice. In assessing the carcinogenic dose-response, the observed cholangiocarcinoma response in both male and female rats demonstrated the steepest dose-response and represents the most conservative data set for extrapolation. However, the data for this effect lacked an adequate dose-response to fully evaluate. Therefore, the most sensitive and relevant toxicological endpoint for establishing an AOEL for furan is the observed hepatic billiary tract hyperplasia leading to carcinogenicity following oral administration in rats and mice. The non-neoplastic hepatic hyperplasia response in male and female rats following oral sub-chronic exposure to furan is presented in Table 5.13. In assessing the hepatic non-neoplastic dose-response in rats the billiary tract hyperplasia response in the females is the most sensitive, with a 70% compared to a 40% response in males at the lowest dose (4 mg/kg/day). In assessing the carcinogenic dose-response, the observed cholangiocarcinoma response in both male and female rats demonstrated the steepest dose-response. The observed tumor incidence is presented in Table 5.14. Both genders demonstrated reasonably comparable responses; females had slightly more tumors at the low dose (2 mg/kg/day) than males (49 vs. 43, respectively).

**Table 5.13.** Hepatic Non-Neoplastic Response: Number of Animals with Lesion/Total Number of Animals (NTP 1993)

Dose (mg/kg/day)	Billiary Tract Hyperplasia	
	Male (w/Hyperplasia/Total)	Female (w/ Hyperplasia /Total)
0	0/10	0/10
4	4/10	7/10
8	9/10	10/10
15	10/10	10/10
30	10/10	10/10
60	10/10	10/10

**Table 5.14.** Tumor Response: Number of Animals with Tumors/Total Number of Animals (NTP 1993)

Dose (mg/kg/day)	Cholangiocarcinoma	
	Male (w/Tumors/Total)	Female (w/Tumors/Total)
0	0/50	0/50
2	43/50	49/50
4	48/50	50/50
8	49/50	49/50



The steep dose-response curves for both the pre-neoplastic and carcinogenic responses are very comparable and consistent with a non-genotoxic mode of carcinogenic action (i.e., tumors are secondary to cytotoxicity). The lack of observable tumors in the control groups also indicates that this response in the Fisher 344 (F344) rat is of a low spontaneous frequency; yet, a near-maximum tumor response is observed in both genders following chronic exposure to the low 2 mg/kg/day dose level. The lack of a clearly definable NOEL for either the pre-neoplastic or carcinogenic response does makes it difficult to establish with confidence a point of departure for calculating an AOEL for furan.

#### 5.11.1.5 Conclusion

Initial evaluation of the data were done with the hope of applying a BMD calculation to establish an appropriate NOEL as a POD for calculating a cancer slope factor (CSF) for furan. Benchmark dose software developed by EPA was used (<http://cfpub.epa.gov/ncea/cfm/recordisplay.cfm?deid=20167>). This software was developed as a tool to facilitate the application of BMD methods to EPA hazardous pollutant risk assessments. However, due to the lack of a clear dose-response for the chronic bioassay (Table 5.14) there was some concern about using a BMD approach since there was substantial uncertainty concerning the calculated POD. A second approach attempted to conduct a similar evaluation of the dose-response for the non-neoplastic hepatic response (Table 5.13); however, these results likewise demonstrated a poor dose-response over the range evaluated. Hence, it was decided that since it was not satisfying to apply the BMD model to either the 13-week or chronic experimental data, a simple linear extrapolation approach would be utilized.

This extrapolation was based on the observed dose-response in female rats for billiary tract hyperplasia, where a 70% response (7 out of 10 animals) was observed following 13-weeks of oral exposure to 4 mg/kg/day. Based on this dose-response ratio, it was possible to calculate the dose associated with a 0.01% response as noted in equation 5.1.

$$dose(mg / kg / day) = \frac{(0.01\% * 4mg / kg / day)}{70\%} \quad (5.1)$$

The calculated oral dose for a 0.01% response ( $1 \times 10^{-4}$ ) was  $5.71 \times 10^{-4}$  mg/kg/day or 0.571  $\mu$ g/kg/day. Assuming a human weighed 65 kg, the daily oral equivalent dose would be  $\sim 37$   $\mu$ g/day. To convert the oral dose ( $\mu$ g/kg) to an equivalent inhalation dose ( $\mu$ g/ $m^3$ ), it was assumed that an average human would inhale  $\sim 10$   $m^3$ /day of air, using equation 2.

$$\mu g / m^3 = \frac{\mu g / day}{10m^3 / day} \quad (5.2)$$

These calculations also assumed that 100% of the inhaled furan would be absorbed systemically through the respiratory system. Using these assumptions, the proposed AOEL for furan based on this analysis is  $\sim 4$   $\mu$ g/ $m^3$  or 1 ppb. An AOEL of 1 ppb inhalation (equivalent to  $\sim 0.5$   $\mu$ g/kg/day oral) is reasonably comparable to the 1  $\mu$ g/kg/day proposed as an oral RfD for furan.

### 5.11.2 Furan Family: Substituted Furans 2-methylfuran (CAS# 534-22-5), 2-5-dihydrofuran and 2-propylfuran (CAS# 4229-91-8) and Others

For the broad class of furan-based chemicals, there is strong experimental evidence suggesting that observed toxicity in several target organs, including the lung and liver, involves the formation of a chemically reactive metabolite. The weight of evidence suggests there is a high probability that a broad range of chemicals containing the unsaturated furan nucleus will be metabolized in animals and humans to a highly reactive and toxic furan epoxide metabolite. For these reactive furans, toxicity is observed in organ systems like the liver and lung, which are capable of CYP450 metabolic activation. In the absence of any substantial toxicity data to fully evaluate the range of substituted furans, the most rational and defensible approach is to set an AOEL for all unsubstituted furans based upon the results obtained with furan itself. Therefore, the proposed AOEL for the substituted furans based on this analysis is  $\sim 4 \mu\text{g}/\text{m}^3$  or **1 ppb**.

#### 5.11.2.1 Toxicology Summary

The furans are a class of chemicals that have as their core a common 5-member, oxygen-containing, unsaturated heterocyclic ring. A number of chemicals that contain the furan heterocyclic ring structure have been identified as occurring naturally or being used as industrial intermediates. For the broad class of furan-based chemicals, there is strong experimental evidence suggesting that observed toxicity in several target organs, including the lung and liver, involves the formation of a chemically reactive metabolite. Boyd (1980) has published several reviews on the role of metabolic activation and chemical-induced lung injury and has reviewed a number of furan-based derivatives including furan, 2-methylfuran, 3-methylfuran, and other substituted furans such as 4-ipomeanol. The review provides strong evidence that CYP450 metabolism in microsomes prepared from both the liver and lung of rats and humans appears to mediate the *in vitro* and *in vivo* formation of reactive chemical species that were shown to alkylate microsomal proteins. Through the use of metabolic inhibitors, it has also been shown that in the absence of metabolism, there was insufficient reactivity of the parent furan to alkylate tissue proteins. Additional support for the important role of metabolism was gleaned by showing that alkylation was not observed in non-metabolically active tissues (e.g., blood, heart, skeletal muscle) following exposure to 4-ipomeanol.

Important to the assessment of the potential toxicological significance for the broad range of furan chemistry was the observation that when the furan ring of 4-ipomeanol was replaced by phenyl or methyl, there was no evidence of covalent binding in the rat lung or liver microsomes *in vitro* (Boyd et al. 1978a). In addition, neither of these analogs became covalently bound to the rat lung or liver *in vivo*, and they were relatively non-toxic at doses  $>800 \text{ mg}/\text{kg}$  (Boyd and Burka 1978). As noted by Boyd (1980), data from several other laboratories have suggested that a highly reactive and toxic furan epoxide metabolite can be formed due to the metabolism of the unsaturated furan ring. Evidence for this pathway was shown with the formation of an epoxide of the dihydrofuran moiety of aflatoxin B1 and the metabolic activation of the furan ring of the diuretic, furosemide (Swenson et al. 1974; Mitchell et al. 1976; McMurtry and Mitchell 1977). The importance of the unsaturated furan ring as a prerequisite for toxicity was demonstrated by Wirth et al. (1976), who reported that a furosemide analog that had a fully saturated tetrahydrofuran group could not be metabolized to a reactive epoxide and did not become covalently bound to liver microsomes. There is evidence that both 3-methyl and 2-methylfuran are also converted by CYP450 metabolism to highly reactive metabolites that are covalently bound to protein (Ravindranath et al. 1986; Boyd et al. 1978b). Boyd (1980) noted that a number of other furan derivatives, including

1-ipomeanol, 1,4-ipomeanol, ipomeanine, 2-methylfuran, and 2-furamide, as well as furan have been reported to cause pulmonary and/or hepatic necrosis in mice. In all cases, each of these chemicals contained an unsaturated furan ring. Further, while studies have not been conducted with all these substituted furans to determine whether they are metabolized to reactive metabolites, it is reasonable to speculate that they have the potential for such metabolism.

The weight of evidence suggests that a broad range of chemicals containing the unsaturated furan nucleus have a high probability to be metabolized in animals and humans to a highly reactive and toxic furan epoxide metabolite. For these reactive furans, toxicity is observed in organ systems like the liver and lung, which are capable of CYP450 metabolic activation.

### 5.11.2.2 Conclusion

In reviewing the available toxicity data bases on furan and substituted furans, the most robust data were found for the parent furan and includes a 2-year oral gavage bioassay conducted in rats and mice (NTP 1993). In rats, sub-chronic exposure to furan produced a high incidence of hepatic biliary tract hyperplasia and cholangiofibrosis in both males and females, and chronic furan exposure produced hepatocellular adenomas in both males and females, and hepatocellular carcinomas in males. A very high incidence (~100%) of cholangiocarcinomas was also observed in both genders. There is some suggestion that furan is bioactivated to a reactive metabolite that can readily alkylate proteins, resulting in significant cytotoxicity. (See Section 5.11.1). Hence, there is some suggestion that furan's mode of carcinogenic action is non-genotoxic (i.e., tumors are secondary to cytotoxicity). In the absence of any substantial toxicity data to fully evaluate the range of substituted furans, the most rationale and defensible approach is to set the AOEL for all unsubstituted furans based upon the results obtained with furan itself. When assessing the potential exposures of workers to furan, the ambient furan concentrations should be taken into account. As additional and appropriate new data become available for substituted furans, it would be reasonable to reassess the appropriateness of basing that chemical's AOEL upon furan.

### 5.11.3 2,4-Dimethylpyridine (CAS# 108-47-4)

An AOEL is proposed for 2,4-dimethylpyridine. Toxicological data for 2,4-dimethylpyridine are minimal, but it is structurally similar to pyridine, 3-picoline, and 5-ethyl-2-picoline, for which more data are available. 2,4-Dimethylpyridine is a food additive recognized as having "no safety concern" by the WHO. Repeated exposure to high concentrations of a mixture of pyridine derivatives can cause central nervous system and gastrointestinal disturbances, facial paralysis, ataxia, and anisocoria (unequal pupils). A Temporary Exposure Limit (TEEL) of 0.5 ppm has been recommended by DOE for 2,4-dimethylpyridine, ACGIH has recommended TLVs of 1 ppm for pyridine, and the AIHA has recommended WEELs of 2 ppm for 2-methylpyridine, 3-methylpyridine, and 4-methylpyridine. **The proposed AOEL for 2,4-Dimethylpyridine is 0.5 ppm**, based on the DOE TEEL-1 and consistent with ACGIH TLVs and AIHA WEELs for closely related compounds.

#### 5.11.3.1 Methodology

To identify and access available toxicity data on 2,4-dimethylpyridine, a number of Internet data bases including PUBMED, TOXNET, TOMES, EPA-IRIS, and the IARC were searched, and summary toxicity profiles were obtained, and primary references were identified. These toxicology summaries were then

reviewed to identify the most critical data sets that could be used to establish the toxicity profile and develop the AOEL. Primary references for the most relevant toxicology endpoints were obtained for review.

### 5.11.3.2 Available Guidelines

No OEL for 2,4-dimethylpyridine has been proposed by the OSHA or ACGIH. A TLV of 1 ppm has been proposed by the ACGIH for pyridine, a structurally similar compound. The AIHA has proposed WEELs of 2 ppm for 2-methylpyridine, 3-methylpyridine, and 4-methylpyridine.

Other available guidelines include the DOE Subcommittee on Consequence Assessment and Protective Actions (SCAPA) TEELs. The TEELs are designed to serve as temporary guidance for chemicals needing guidelines for hazard assessment and safety analysis for DOE personnel and general public adjacent to DOE sites. These TEELs are assigned to chemicals on a temporary basis until they can be replaced with AEGLs or emergency response planning guidelines, which require additional documentation and assessment. The protocol used to consider available OEL and toxicological data in the development of TEELs has been published (Craig et al. 2000). Four levels of TEELs (TEEL-0, TEEL-1, TEEL-2, TEEL-3) are calculated as fractions based on set criteria. TEEL-1 is defined as “the maximum airborne concentration below which it is believed that nearly all individuals could be exposed without experiencing other than mild transient adverse health effects or perceiving a clearly defined, objectionable odor.” When a PEL-STEL or TLV-STEL is available, TEEL-1 should be lower (more protective) than the STEL. Although not specifically stated, TEEL-1 is usually based on a 1-hour TWA. The DOE SCAPA-recommended TEEL-1 for 2,4-dimethylpyridine is 0.5 ppm. TEEL-3, defined as “the maximum airborne concentration below which it is believed that nearly all individuals could be exposed without experiencing or developing life-threatening health effects,” is 20 ppm.

### 5.11.3.3 Toxicology Summary

Repeated industrial exposure to mixtures of pyridine derivatives, which include 2,4-dimethylpyridine, have resulted in central nervous system and gastrointestinal disturbances, facial paralysis, ataxia, and anisocoria. Short-term exposures to high levels of pyridine mixes can cause a burning sensation, coughing, dizziness, drowsiness, headache, labored breathing, nausea and sore throat. The LD<sub>50</sub> for 2,4-dimethylpyridine in rats was 200 mg/kg. The WHO has commented that there is no safety concern with 2,4-dimethylpyridine as a food additive (WHO 2005).

Exposures to the surrogate, pyridine, result in irritation. Repeated exposures can also result in kidney and liver damage. The acute LD<sub>50</sub> for pyridine in rats and mice ranges from 900-1500 mg/kg (ACGIH 2004). The NOEL in rats and mice in 28-day drinking water exposures was 7-15 mg/kg/day. Pyridine is not mutagenic in bacteria or CHO cells. Human exposure assessment in a chemical plant indicated that repeated exposures to 6-12 ppm resulted in headache, temporary vertigo, nervousness, and sleeplessness. The recommended ACGIH TLV for pyridine is 1 ppm based on minimizing the potential for nasal irritation and liver and kidney toxicity.

2,4-Dimethylpyridine is a substituted pyridine, with slightly higher acute toxicity than pyridine; therefore, a substituted pyridine surrogate was sought to assess the potential for higher toxicity in substituted pyridines. 5-Ethyl-2-picoline was the closest surrogate with sufficient toxicity data identified. The NOEL for 5-ethyl-2-picoline was 30 mg/kg/day in a 28-day oral study in mice.

#### 5.11.3.4 Data Analysis

2,4-Dimethylpyridine and mixed pyridine exposures occur in industrial settings. The most data were available for the unsubstituted pyridine. The ACGIH first assigned an OEL of 10 ppm for pyridine in 1956; this value has been changed twice and is 1 ppm as of 2004. A comparison of acute toxicity between pyridine, methylpyridine, and 5-ethyl-2-picoline suggests that the substitutions do not result in increased toxicity. The primary endpoint of concern for 5-ethyl-2-picoline is likely due to its corrosive properties and therefore nasal irritation. The AOEL of 1 ppm for pyridine, and the TEEL-1 for 2,4-dimethylpyridine of 0.5 ppm are similar.

#### 5.11.3.5 Conclusion

Based on the comparison of acute and subchronic effects of 5-ethyl-2-picoline, pyridine, and 2,4-dimethylpyridine in laboratory animals and the effects from human exposures to pyridine mixtures, it is proposed that the **DOE TEEL-1 of 0.5 ppm be adopted as the AOEL for 2,4-dimethylpyridine.** TEEL-1 is similar to the ACGIH TLV of 1 ppm for pyridine and the AIHA WEEL of 2 ppm for 2-methylpyridine, 3-methylpyridine, and 4-methylpyridine. Based on the WHO's recognition of 2,4-dimethylpyridine as having "no safety concern" for oral exposures, this is a very conservative TLV, and TEEL-1 may be based on route-dependent irritation. As additional new data become available for the 2,4-dimethylpyridine, it would be reasonable to reassess the appropriateness of basing the AOEL upon the DOE SCAPA TEEL-1.

#### 5.11.4 1,2,3,6-Tetrahydropyridine (CAS# 694-05-3)

An AOEL is proposed for 1,2,3,6-tetrahydropyridine. No toxicological data were found for 1,2,3,6-tetrahydropyridine but it is structurally similar to both pyridine or piperidine, for which more data are available. Repeated exposures to high concentrations of a mixture of pyridine derivatives can cause central nervous system and gastrointestinal disturbances, facial paralysis, ataxia, and anisocoria. ACGIH has recommended TLVs of 1 ppm for pyridine, and the AIHA has recommended WEELs of 2 ppm for 2-methylpyridine, 3-methylpyridine, and 4-methylpyridine. Piperidines are used as nonsedating antihistamines since they do not readily cross the blood-brain barrier. The AIHA has recommended a WEEL of 1 ppm for piperidine. The proposed AOEL for 1,2,3,6-tetrahydropyridine is 0.1 ppm based on the AIHA WEEL for surrogate piperidine with an UF of 10 for the lack of data and structural differences. **This proposed OEL is also consistent with the ACGIH TLVs of 1 ppm for pyridine and 0.1 for the piperidine derivative, phenylhydrazine.**

#### 5.11.4.1 Methodology

To identify and access available toxicity data on 1,2,3,6-tetrahydropyridine, a number of Internet data bases, including PUBMED, TOXNET, TOMES, EPA-IRIS, and the IARC, were searched and summary toxicity profiles were obtained, including the identification of primary references. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profile and used for the development of the AOEL. Primary references for the most relevant toxicology endpoints were obtained for review.

#### 5.11.4.2 Available Guidelines

No OEL for 1,2,3,6-tetrahydropyridine has been proposed by the OSHA or ACGIH. A TLV of 1 ppm has been proposed by the ACGIH for pyridine and 0.1 ppm for phenylhydrazine, which are structurally similar compounds. The AIHA has proposed WEELs of 2 ppm for 2-methylpyridine, 3-methylpyridine, and 4-methylpyridine. The structurally similar piperidine has an AIHA-recommended WEEL of 1 ppm with a skin notation.

#### 5.11.4.3 Toxicology Summary

Repeated industrial exposure to mixtures of pyridine derivatives have resulted in central nervous system and gastrointestinal disturbances, facial paralysis, ataxia, and anisocoria (unequal pupils). Short-term exposures to high levels of pyridine mixes can cause a burning sensation, coughing, dizziness, drowsiness, headache, labored breathing, nausea and sore throat. Piperidines are nasal and dermal irritants. 1,2,3,6-Tetrahydropyridine is also explosive.

Exposures to the surrogate, pyridine, result in irritation. Repeated exposures can also result in kidney and liver damage. The acute LD<sub>50</sub> for pyridine in rats and mice ranges from 900-1500 mg/kg (ACGIH 2004). The NOEL in rats and mice in 28-day drinking water exposures was 7-15 mg/kg/day. Pyridine is not mutagenic in bacteria or CHO cells. Human exposure assessment in a chemical plant indicated that repeated exposures to 6-12 ppm resulted in headache, temporary vertigo, nervousness, and sleeplessness. The recommended ACGIH TLV for pyridine is 1 ppm based on minimizing the potential for nasal irritation and liver and kidney toxicity.

#### 5.11.4.4 Data Analysis

The most data were available for the unsubstituted pyridine. The ACGIH first assigned an OEL of 10 ppm for pyridine in 1956; this value has been changed twice and is 1 ppm as of 2004. The primary endpoint of concern is likely due to corrosive properties and therefore nasal irritation. The AOEL of 1 ppm for pyridine, and the WEELs for other pyridines of 2 ppm are similar.

#### 5.11.4.5 Conclusion

Based on the lack of toxicological data for 1,2,3,6-tetrahydropyridine, it is proposed that the AIHA WEEL of 1 ppm for piperidine be used with an UF of 10 to propose an OEL for **1,2,3,6-tetrahydropyridine of 0.1 ppm**. As additional new data become available for the 1,2,3,6-tetrahydropyridine, it would be reasonable to reassess the appropriateness of basing the AOEL upon the WEEL for substituted pyridines.

## 5.12 Chlorinated Compounds

### 5.12.1 Polychlorinated Biphenyls (PCBs)

An AOEL is proposed for a family of polychlorinated biphenyls (PCBs). There are 209 different congeners of PCBs that vary in the extent and position of chlorination on the aromatic rings. They are not naturally occurring and primarily come from industrial manufacture of such items as capacitors and transformers. They persist in the environment and fat stores in the body. Exposures to high concentrations can lead to skin lesions (rashes and acne). High chronic concentrations can lead to liver damage, anemia, and potentially to adverse reproductive effects. The toxic potential of PCBs is dependent upon the extent and location of chlorine substitution on the aromatic rings. The PCB congeners found in the tank waste headspace are much more similar to the PCBs with lower percentages of chlorine, and the proposed AOEL for these PCBs is  $0.03 \text{ mg/m}^3$  (~0.003 ppm) based on the ACGIH TLV for 42% chlorinated PCBs with an UF of 3 to account for uncertainty associated with the exact mixture of PCBs in the tank and the potential for chloracne.

#### 5.12.1.1 Methodology

To identify and access available toxicity data on polychlorinated biphenols, a number of Internet data bases, including PUBMED, TOXNET, TOMES, EPA-IRIS, and IARC were searched, and summary toxicity profiles were obtained and primary references identified. These toxicology summaries were then reviewed to identify the most critical data sets that could be used to establish the toxicity profile and develop the AOEL. Primary references for the most relevant toxicology endpoints were obtained for review.

#### 5.12.1.2 Available Guidelines

ACGIH has proposed guidelines for PCBs with 42% (Aroclor 1242<sup>®</sup>) or 54% (Aroclor 1254<sup>®</sup>) chlorine. Aroclor 1242 and 1254 are mixtures of different congeners of chlorobiphenyl. The environmental fate, potential carcinogenicity, and toxicity of PCBs is dependant upon the degree of chlorination (ACGIH 2001c; ACGIH 2001d; Chana et al. 2002). The ACGIH assigned an A3 - Confirmed Animal Carcinogen with Unknown Relevance to Humans notation for the mixture of PCBs containing 54% chlorine, but stated there are insufficient data to assign cancer or skin notations for Aroclor 1242 (42% chlorine).

Other available guidelines include the OSHA PEL, which matches the ACGIH TLV of 1 or  $0.5 \text{ mg/m}^3$  depending on the level of chlorination; the NIOSH REL, which is  $1 \text{ } \mu\text{g/m}^3$ ; and the NIOSH IDLH, which is  $5 \text{ mg/m}^3$ . The IDLH is based on acute inhalation toxicity data of PCBs in humans and animals and is not specific for the chlorine content. The  $1 \text{ } \mu\text{g/m}^3$  NIOSH REL is based on potential carcinogenicity and is not specific for chlorine content. The IARC has given PCBs a 2A Carcinogens status irrespective of the amount of chlorine substitution.

The EPA has calculated CSFs for different mixtures of PCBs based on their associated risks and persistence. For PCBs with the lowest risk and persistence (congeners with more than four chlorines on less than one-half of the total PCBs in the mixture, which includes the PCBs in the tank waste headspace) the upper-bound CSF is 0.07 per mg/kg/day, and the central estimate CSF is 0.04 per mg/kg/day).

### 5.12.1.3 Toxicology Summary

The endpoint of concern for the guideline set by the ACGIH for 42% chlorine PCBs is hepatotoxicity. The toxicity of a PCB is dependent not only upon the number of chlorines present on the biphenyl structures but also the positions of the chlorines. Table 5.15 shows the general rules for relative toxicity of congeners based on chlorine number and position. It has been theorized that the ability or inability of the aromatic rings to become coplanar is linked to the toxicity and or carcinogenicity of the PCB congeners (Chana et al. 2002).

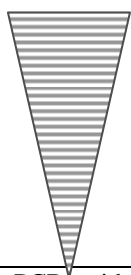
Exposures of laboratory animals (cats, rabbits, guinea pigs, rats, and mice) to 1.9 to 8.6 mg/m<sup>3</sup> 42% chlorine PCB for up to 7 months at 7 hours/day on intermittent days resulted in no obvious adverse affects (ACGIH 2001c). Human exposures to 42% chlorine PCB resulted in acne, respiratory irritation, and liver injury (ACGIH 2001c).

The carcinogenic effects of PCBs have been shown to be dependent on chlorine content in mice and rats fed diets containing 30-60% chlorine PCBs (Ito et al. 1974; Ito et al. 1973, reported in ACGIH 2001c). In mice, no tumors occurred in groups fed less than 54% chlorine PCB.

### 5.12.1.4 Data Analysis

The PCBs in the headspaces of tanks used to store high-level radioactive waste at DOE sites are almost entirely congeners with less than 42% chlorine. There are only three tetrachloro isomers (49% chlorine). The ACGIH TLVs are based on mixtures of PCBs, and the tetrachloro isomers in the headspace within mixtures containing the PCBs with lower percentages of chlorine will still result in mixtures of less than 42% chlorine. In addition, the congeners in the tank waste headspace generally have chlorine substitutions at positions suggesting relatively lower toxicity (Table 5.15). Therefore, the appropriate guideline comparison is to PCB 1242, which has a TLV-TWA of 1 mg/m<sup>3</sup>. However, this TLV may not protect exposed individuals from chloracne.

**Table 5.15.** Relative Toxicities of PCB Congeners in the Tank Waste Head Space

Relative Toxicity <sup>(a)</sup>	PCB Congener Description	Comments <sup>(b)</sup>
	5-6 Chlorines	None in Head Space
	“Dioxin-like” 4, 4' and 3 or 3', or 5 or 5'	None in Head Space
	4 and 4'	2 Congeners in Head Space
	none in ortho (2, or 2' or 6 or 6' position)	3 Congeners in Head Space
	1 ortho (2, or 2' or 6 or 6' position)	10 Congeners in Head Space
	2 ortho (2, or 2' or 6 or 6' position)	6 Congeners in Head Space
<p>(a) PCBs with 5-6 chlorines are the most toxic (Chana et al. 2002), congeners with chlorines in both para positions (4 and 4') and at least 2 chlorines at the meta positions (3, 5, 3', 5') are considered to be 'dioxin like' and are 'particularly toxic' (U.S. EPA, <a href="http://www.epa.gov/toxteam/pcb/pcbtable.htm">http://www.epa.gov/toxteam/pcb/pcbtable.htm</a>). Congeners with ortho-substituted chlorines (2,2',6,6' positions) are less toxic than those without ortho-substituted chlorines</p> <p>(b) There were insufficient data to rank the additional five congeners or isomers present in the headspace on this chart.</p>		



Applying the EPA slope factors for upper and central estimate slope factors for PCBs of the type found in the tank waste headspace, estimated OELs are 26 and 45  $\mu\text{g}/\text{m}^3$ , respectively (for calculation methodology, see Section 5.11.1).

### 5.12.1.5 Conclusion

The use of the ACGIH TLV-TWA of 1  $\text{mg}/\text{m}^3$  based on exposures to 42% chlorine PCB to propose an AOEL for the tank waste PCB headspace content results in a conservative guideline (Table 5.16). The PCB congeners in this population are generally predicted to have relatively low toxicities based on the number and location of chlorine substitutions (see Table 5.16), and the percentage of chlorine in the tank waste headspace will be lower than the 42% in the ACGIH assessment. The proposed AOEL for the PCBs in the tank waste headspace is therefore **0.03  $\text{mg}/\text{m}^3$**  and includes an UF of 30 to account for

**Table 5.16.** PCBs in Tank Waste

Tank Name	Chemical Name	Molecular Weight	Chemical Id	Conc. ( $\text{mg}/\text{m}^3$ )	Conc. (PPM)	TLV/ Max HS <sup>(a)</sup>
241-AX-102	1,1'-Biphenyl, 2-chloro-	188.7	2051-60-7	0.0063	0.0008	158
241-B-103	1,1'-Biphenyl, 3-chloro-	188.7	2051-61-8	0.033	0.0043	<b>30</b>
241-BY-103	1,1'-Biphenyl, 2-chloro-	188.7	2051-60-7	0.0078	0.0010	128
241-BY-104	1,1'-Biphenyl, 2,6-dichloro and siloxane	223.1	MRHUSIO-01	0.0092	0.0010	109
241-BY-112	1,1'-Biphenyl, 2,2'-dichloro-	223.1	13029-08-8	0.0057	0.0006	176
241-S-105	Chloro-1,1'-Biphenyl	188.7	URH000-03	0.023	0.0030	<b>44</b>
241-S-111	1,1'-Biphenyl, 2-chloro-	188.7	2051-60-7	0.0016	0.0002	641
241-TX-105	1,1'-Biphenyl, 2,3,4',6-tetrachloro-	292.0	52663-58-8	0.0023	0.0002	442
Tanks with More than 1 PCB Detected						
241-TX-118	1,1'-Biphenyl, 3,3'-dichloro-	223.1	2050-67-1	0.0069	0.0008	
241-TX-118	1,1'-Biphenyl, 3-chloro-	188.7	2051-61-8	0.0064	0.0008	
			Total	0.013		75
241-TY-101	1,1'-Biphenyl, 2,3,3',5'-tetrachloro-	292.0	41464-49-7	0.0045	0.0004	
241-TY-103	1,1'-Biphenyl, 2-chloro-	188.7	2051-60-7	0.035	0.0045	
			Total	0.039		<b>25</b>
241-TY-104	1,1'-Biphenyl, 2,2'-dichloro-	223.1	13029-08-8	0.0011	0.0001	
241-TY-104	1,1'-Biphenyl, 2,3-dichloro-	223.1	16605-91-7	0.0010	0.0001	
241-TY-104	1,1'-Biphenyl, 2,4',5-trichloro-	257.5	16606-02-3	0.0035	0.0003	
241-TY-104	1,1'-Biphenyl, 4,4'-dichloro-	223.1	2050-68-2	0.0037	0.0004	
241-TY-104	1,1'-Biphenyl, 2-chloro-	188.7	2051-60-7	0.0037	0.0005	
241-TY-104	1,1'-Biphenyl, 4-chloro-	188.7	2051-62-9	0.0026	0.0003	
241-TY-104	1,1'-Biphenyl, trichloro-	257.5	25323-68-6	0.0032	0.0003	
241-TY-104	1,1'-Biphenyl, 2,6-dichloro-	223.1	33146-45-1	0.0032	0.0004	
241-TY-104	1,1'-Biphenyl, 2,5-dichloro-	223.1	34883-39-1	0.0009	0.0001	
241-TY-104	1,1'-Biphenyl, 2,3,3',5'-tetrachloro-	292.0	41464-49-7	0.0026	0.0002	
241-TY-104	1,1'-Biphenyl, 2,3,4',6-tetrachloro-	292.0	52663-58-8	0.0006	0.00005	
(a) The total concentration of PCBs in the Headspace ( $\text{mg}/\text{m}^3$ ) compared to the ACGIH TLV of 1 $\text{mg}/\text{m}^3$ . Only 3 tanks had headspace concentrations with a margin of exposure less than 50 (in bold).						

uncertainty associated with the precise composition of the tank waste PCBs and to protect against chloracne. In addition,  $0.03 \text{ mg/m}^3$  also approximates the OEL achieved using the EPA cancer slope factor calculation. As additional new data become available for these PCBs, it would be reasonable to reassess the appropriateness of basing the AOEL upon Aroclor 1242.

## 6.0 Summary of Proposed AOELs

The AOEL development process described in Sections 2.0 - 4.0 was applied to 57 tank headspace chemicals. Proposed AOELs for four chemicals were based on OELs established by U.S. or foreign government agencies. These are listed in Table 6.1, along with the source of the established OEL. For the remaining 53 chemicals without established OELs, epidemiological and toxicological information were reviewed to identify potential hazards, select critical effects, and estimate dose-response to determine exposure levels expected to present no health risks. When information on the chemical itself was insufficient to develop an AOEL, the proposed AOEL was based on one or more structurally related chemicals (surrogates) for which sufficient data were available. Surrogates with close structural and pharmacological activity with established OELs (ACGIH or OSHA) were preferentially selected. Of these 53 chemicals, 47 were assigned proposed AOELs and six were determined to not warrant complete AOEL development, based on expected low potential for adverse health effects. Table 6.2 lists the 47 proposed AOELs, and Table 6.3 lists the six chemicals for which no AOEL was deemed necessary.

**Table 6.1.** Proposed AOELs for Chemicals with Limits from NIOSH, AIHA or MAK

Chemical	CAS Number	Proposed AOEL (ppm)	OEL Source
Propanenitrile	107-12-0	6	NIOSH REL
2-Ethyl-1-hexanol	104-76-7	20	MAK
Butanal	123-72-8	25	AIHA WEEL
Butanenitrile	109-74-0	8	NIOSH REL

**Table 6.2.** Summary of Proposed AOELs

Chemical	CAS Number	Proposed AOEL (ppm)
<b>HALOCARBONS</b>		
2-Fluoropropene	1184-60-7	0.1
<b>ALCOHOLS</b>		
Cyclopentanol	96-41-3	5
<b>ETHERS</b>		
1-Methoxybutane	628-28-4	17
<b>ALDEHYDES</b>		
2-Methylbut-2-enal	1115-11-3	0.03
2-Ethylhex-2-enal	645-62-5	0.1
<b>KETONES</b>		
4-Methyl-2-hexanone	105-42-0	0.5
3-Hexanone	589-38-8	67
6-Methyl-2-heptanone	928-68-7	8
2-Nonanone	821-55-6	17
3-Dodecanone	1534-27-6	17
2-Tridecanone	593-08-8	17
3-Tridecanone	1534-26-5	17
5-Methyl-2-(1-methylethenyl) cyclohexanone	89-82-7	2.5

**Table 6.2. (contd)**

Chemical	CAS Number	Proposed AOEL (ppm)
4,7,7-Trimethyl bicyclo[4.1.0]heptan-3-one	4176-04-9	0.7
3-Methyl-3-buten-2-one	814-78-8	0.02
<b>ESTERS</b>		
Dibutyl butylphosphonate	78-46-6	0.02
<b>NITRILES</b>		
2-Methylene butanenitrile	1647-11-6	0.3
Pentanenitrile	110-59-8	6
Hexanenitrile	628-73-9	6
Heptanenitrile	629-08-3	6
Octanenitrile	124-12-9	6
Nonanenitrile	2243-27-8	6
Decanenitrile	1975-78-6	6
3-Butenenitrile	109-75-1	1
2,4-Pentadienenitrile	1615-70-9	0.3
<b>AMINES &amp; AMIDES</b>		
N-Methylaziridine	1072-44-2	0.05
<b>NITRITES &amp; NITRATES</b>		
Methyl nitrite	624-91-9	0.1
Ethyl nitrite	109955	0.1
Butyl nitrite	544-16-1	0.1
Methyl nitrate	598-58-3	8
Ethyl nitrate	625-58-1	8
Butyl nitrate	928-45-0	8
1,3-Propanediol, dinitrate	3457-90-7	0.05
1,4-Butanediol, dinitrate	3457-91-8	0.05
1,5-Pentanediol, dinitrate	3457-92-9	0.05
1,2,3-Propanetriol, 1,3-dinitrate	623-87-0	0.05
1,2,3-Propanetriol, 1-nitrate	624-43-1	8
<b>NITRO COMPOUNDS</b>		
2-Nitro-1-propanol	2902-96-7	0.8
2-Nitro-2-methylpropane	594-70-7	0.3
1-Nitrobutane	627-05-4	2.5
<b>HETEROCYCLICS</b>		
Furan	110-00-9	0.001
2-Methylfuran	534-22-5	0.001
2-Propylfuran	4229-91-8	0.001
2-Ethyl-5-methylfuran	1703-52-2	0.001
2,5-Dihydrofuran	1708-29-8	0.001
alpha-Propylfuranacetaldehyde	31681-26-2	0.001
2,4-Dimethylpyridine	108-47-4	0.5

**Table 6.3.** Chemicals For Which No AOEL is Recommended

Chemical	CAS Number
1-Hexadecanol	36653-82-4
1-Octadecanol	112-92-5
1-Methylethyl tetradecanoate	110-27-0
Butyl tetradecanoate	110-36-1
1-Methylethyl hexadecanoate	142-91-6
Acetamide	60-35-5

## 7.0 References

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