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Mercenaria mercenaria, Eastern Hard Clam, Quahog

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Penaeus aztecus, Brown Shrimp

Brevoortia tyrannus, Atlantic Menhaden

Prepared by C. I. Gibson, F. C. Tone, P. Wilkinson, J. W. Blaylock, R. E. Schirmer

Pacific Northwest Laboratory

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Commission

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ABSTRACT

Bromoform has been identified as the single most abundant halogenated organic compound produced by the chlorination of marine waters. To determine the potential biological effects of its release into marine waters, short-term toxicity bioassays and 28-day uptake/28-day depuration studies were conducted with five marine species: Protothaca staminea, Mercenaria mercenaria, Crassostrea virginica, Penaeus aztecus, and Brevoortia tyrannus. The bioassay studies indicate that 96-hr LC50s ranged from approximately 7 ppm for B. tyrannus to greater than 40 ppm for P. staminea. Behavioral changes were noted in P. aztecus and B. tyrannus exposed to sublethal concentrations of bromoform.

In all species tested, the uptake and depuration of bromoform was rapid. Bromoform was present in all exposed animal tissues within 24 hours and was depurated within 48 hours. In the mollusk species, there was bioaccumulation above water concentrations in the first week of exposure, and then the tissue concentrations fell to levels approximately equal to the water concentrations. The shrimp and menhaden also bioaccumulated bromoform above water concentrations in the first week of exposure, but then the tissue concentrations fell to approximately 0.4 µg/g and remained at this level independent of water concentrations.

SUMMARY

The investigation of the by-products created by the chlorination of sea water showed that bromoform was the major halogenated organic formed. To determine the potential for environmental effects of bromoform and its possible return to man, a study was undertaken to test its toxicity and uptake with five commercial and recreationally important species. The species to be tested, Protothaca staminea (Pacific littleneck clam), Mercenaria mercenaria (Eastern hard clam or quahog), Penaeus aztecus (brown shrimp) and Brevoortia tyrannus (Atlantic menhaden) were selected by NRC. These species have commercial and economic importance and are often found in the vicinity of discharge streams from nuclear fueled steam electric stations.

The research approach was to run a series of 96-hr LC50 screening bioassays to determine the relative toxicity of bromoform and to run a standard 28-day uptake/28-day depuration test. These studies would provide the information needed to determine the necessity of further testing.

The initial problem of the study was the development of a method to introduce bromoform into sea water in a manner that would result in a bromoform/seawater solution. This was to be done in a flow-through system to provide test conditions that were as similar as possible to the conditions an animal may experience in the environment.

The introduction of bromoform into sea water in a soluble form turned out to be very difficult. When liquid-to-liquid mixing was attempted, bromoform tended to pool together in the bottom of the mixing vessel and was slow to go into solution. In addition, a large excess of bromoform was required to bring a given volume of sea water to the saturation level. These features made it impractical to use liquid-to-liquid mixing to provide the needed quantities of seawater/bromoform solutions for the desired toxicity and uptake testing. The sparging of air saturated with bromoform into sea water was found to be the most satisfactory method of producing the desired concentrations of bromoform in sea water under the conditions needed to perform the tests.

Testing of the eastern species was conducted at the Battelle Marine Research facility located in Daytona Beach, Florida, and testing of the west coast species was conducted at the Sequim, Washington, Marine Research laboratory. Analysis of the sea water for bromoform was conducted at the Sequim laboratory. To insure that the transport of the samples to Sequim would not affect the concentration, both storage and shipping tests were conducted with known samples. Under all situations there were no significant changes caused by either storage or shipping.

Preliminary testing of the littleneck clam indicated that a standard 96-hr LC50 would be impractical. At concentrations of bromoform that caused death, littleneck clams would close up and die in the shut position, making it difficult to determine if the clam was dead or when it died. At lower concentrations, pumping by the clams appeared to be sporadic, making the actual exposure time variable depending on the individual clam's activity. Therefore, it was decided to conduct the 28-day uptake/28-day depuration test at 1, 5, 10, and 20 mg bromoform/l concentrations and to record the mortalities. It was felt that this would provide an estimate of the toxicity of bromoform in addition to the desired uptake data.

It was very difficult to maintain the target bromoform concentrations over the 28-day study period. The actual average concentrations were 2, 7, 19, and 27 mg bromoform/l. Mortality occurred only in the 27 mg/l and 19 mg/l test conditions. At 27 mg/l average bromoform concentration, there was mortality on day 7 and day 25; at 19 mg/l average bromoform concentration, mortality occurred only on day 25.

At Daytona, all species were subjected to standard 96-hr tests using paired tanks. As at Sequim, there was difficulty producing mortalities in the mollusks because of their apparent tolerance to bromoform and their ability to close up and reduce their exposure time. Because of this, a standard LC50 value could not be calculated. However, based on latent mortality it was estimated that for these species the 96-hr LC50 would be in the range of 40 to 140 mg/l.

Testing of the menhaden and brown shrimp was successful. They had calculated 96-hr LC50 values of 12 mg/l and 26 mg/l, respectively. Both of these species exhibited behavioral changes at sublethal concentrations.

The 28-day uptake/28-day depuration tests were conducted at concentrations below those found to be lethal in the toxicity tests. However, even at these low concentrations some mortalities occurred in the menhaden and shrimp tests. These mortalities were probably due to the fact that it is difficult to hold these organisms for long periods of time in the laboratory without experiencing some mortality, even in the controls.

All the mollusk species (littleneck clams, quahogs, and oysters) had tissue concentrations that were reflective of the ambient water concentrations. There was some indication of increased concentrations (that is, tissue concentrations above water concentrations) during the first week of exposure, but the general tendency was for these to decrease with time and be equal to water concentrations. The bromoform was depurated from the mollusks within 24 to 48 hours after exposure stopped.

The menhaden and shrimp had a different pattern of uptake. In menhaden, at an average water concentration of 0.21 mg/ℓ, the tissue concentrations were high during the first week (up to 7.61 μg/g) but then fell to near the water concentrations. At the lower test concentration (0.04 mg/ℓ), the tissue concentrations remained around .40 μg/g, a factor of about 10 above water concentrations. The same phenomenon was noted in the shrimp.

TABLE OF CONTENTS

Abstract	iii
Summary	v
List of Figures	xi
List of Tables	xiii
Preface	xv
Acknowledgments	xvii
Introduction	1
Discussion	6
Conclusions and Recommendations	9
References	11

LIST OF FIGURES

1.	Toxicant delivery system for seawater/bromoform bioassays conducted at Sequim	19
2.	A section of the bromoform exposure system used at the Daytona Beach Laboratory	19
3.	Measured concentrations in <u>Protothaca staminea</u> exposure with target bromoform concentration of 20 mg/l	20
4.	Measured concentrations in <u>Protothaca staminea</u> exposure tank with target bromoform concentration of 10 mg/l	20
5.	Measured concentrations in <u>Protothaca staminea</u> exposure tank with target bromoform concentration of 5 mg/l	21
6.	Measured concentrations in <u>Protothaca staminea</u> exposure tank with target bromoform concentration of 1 mg/l	21
7.	<u>Penaeus aztecus</u> mortality bromoform concentration plot for calculation of 96-hr LC50 by the method of Litchfield and Wilcoxon (1949)	22
8.	<u>Brevoortia tyrannus</u> mortality bromoform concentration plot for 96-hr LC50 calculation by the method of Litchfield and Wilcoxon (1949)	22
9.	<u>Crassostrea virginica</u> latent mortality bromoform concentration plot used to determine concentration above which the 96-hr LC50 should fall	23
10.	<u>Mercenaria mercenaria</u> latent mortality bromoform concentration plot used to determine concentration above which the 96-hr LC50 should fall	23
11.	Water and tissue concentrations of bromoform (Eastern oyster, <u>Crassostrea virginica</u>) 28-day uptake/28-day depuration studies. Target water bromoform was 1.0 mg/l	24
12.	Water and tissue concentrations of bromoform (Eastern oyster, <u>Crassostrea virginica</u>) 28-day uptake/28-day depuration studies. target water bromoform was 0.1 mg/l	25
13.	Water and tissue concentrations of bromoform (Eastern oyster, <u>Crassostrea virginica</u>) 28-day uptake/28-day depuration studies. Control	26

14.	Water and tissue concentrations of bromoform (Quahaug, <u>Mercenaria mercenaria</u>) 28-day uptake/28-day depuration studies. Target water bromoform concentration was 1.0 mg/l . . .	27
15.	Water and tissue concentrations of bromoform (Quahaug, <u>Mercenaria mercenaria</u>) 28-day uptake/28-day depuration studies. Target water bromoform concentration was 0.1 mg/l . . .	28
16.	Water and tissue concentrations of bromoform (Quahaug, <u>Mercenaria mercenaria</u>) 28-day uptake/28-day depuration studies. Control	29
17.	Water and tissue concentrations of bromoform (Littleneck clam, <u>Protothaca staminea</u>) 28-day uptake/28-day depuration studies. Target water bromoform concentration was 10 mg/l . . .	30
18.	Water and tissue concentrations of bromoform (Littleneck clam, <u>Protothaca staminea</u>) 28-day uptake/28-day depuration studies. Target water bromoform concentration was 1.0 mg/l . . .	31
19.	Water and tissue concentrations of bromoform (Littleneck clam, <u>Protothaca staminea</u>) 28-day uptake/28-day depuration studies. Control	32
20.	Water and tissue concentrations of bromoform (Menhaden, <u>Brevoortia tyrannus</u>) 28-day uptake/28-day depuration studies. Target water bromoform concentrations was 1.0 mg/l	33
21.	Water and tissue concentrations of bromoform (Menhaden, <u>Brevoortia tyrannus</u>) 28-day uptake/28-day depuration studies. Target water bromoform concentration was 0.1 mg/l	34
22.	Water and tissue concentrations of bromoform (Menhaden, <u>Brevoortia tyrannus</u>) 28-day uptake/28-day depuration studies. Control	35
23.	Water and tissue concentrations of bromoform (Shrimp, <u>Penaeus azectus</u>) 28-day uptake/28-day depuration studies. Target water bromoform concentration was 1.0 mg/l	36
24.	Water and tissue concentration of bromoform (Shrimp, <u>Penaeus azectus</u>) 28-day uptake/28-day depuration studies. Target water bromoform concentration was 0.1 mg/l	37
25.	Water and tissue concentrations of bromoform (Shrimp, <u>Penaeus aztecus</u>) 28-day uptake/28-day depuration studies. Control	38

LIST OF TABLES

1. Concentration of bromoform (mg/ℓ) in tank at time of noted clam responses	12
2. Measured bromoform concentrations (mg/ℓ) in <u>Protothaca staminea</u> in exposure tanks	13
3. Bromoform concentration in each test organism for Eastern oyster (<u>Crassostrea virginica</u>) (μg/g tissue-wet weight)	14
4. Bromoform concentration in each test organisms for Quahaug (<u>Mercenaria mercenaria</u>) (μg/g tissue-wet weight)	15
5. Average (\bar{x}) Body Burden and Standard Deviation (S.D.) for littleneck clam (<u>Protothaca staminea</u>) (μg/g tissue-wet weight)	16
6. Bromoform concentration in each test organism for menhaden (<u>Brevoortia tyrannus</u>) (μg/g tissue-wet weight)	17
7. Bromoform concentration in each test organism for shrimp (<u>Penaeus aztecus</u>) (μg/g tissue-wet weight)	18

PREFACE

This report includes data and analysis for the Marine Biology Task of the program on Biocide By-Products in Aquatic Environments.

Reports prepared for the entire program are:

<u>Title</u>	<u>Author</u>
• Investigation of Halogenated Components Formed from Chlorination of Natural Waters: Preliminary Studies, NUREG/CR-1299	Roger M. Bean Robert G. Riley
• Acute Toxicity and Bioaccumulation of Chloroform to Four Species of Fresh Water Fish <u>Salmo gairdneri</u> , Rainbow Trout <u>Lepomis macrochirus</u> , Bluegill <u>Micropterus salmoides</u> , Largemouth Bass <u>Ictalurus punctatus</u> , Channel Catfish, NUREG/CR-0893	David R. Anderson E. William Lusty
• Chronic Effects of Chlorination By-Products on Rainbow Trout, <u>Salmo gairdneri</u> , NUREG/CR-0892	David R. Anderson Roger M. Bean Roger E. Schirmer
• Toxicity, Bioaccumulation and Depuration of Bromoform in Five Marine Species <u>Protothaca staminea</u> , Littleneck Clam <u>Mercenaria mercenaria</u> , Eastern Hard Clam, Quahog <u>Crassostrea virginica</u> , Eastern oyster <u>Penaeus aztecus</u> , Brown Shrimp <u>Brevoortia tyrannus</u> , Atlantic Menhaden, NUREG/CR-1297	Charles I. Gibson Fredrick C. Tone Peter Wilkinson J. W. Blaylock Roger E. Schirmer
• Growth and Histological Effects to <u>Protothaca staminea</u> , (Littleneck Clam) of Long-Term Exposure to Chlorinated Sea Water, NUREG/CR-1298	Charles I. Gibson Robert E. Hillman Peter Wilkinson Dana L. Woodruff
• Analysis of Organohalogen Products from Chlorination of Natural Waters Under Simulated Biofouling Control Conditions, NUREG/CR-1301	Roger M. Bean Dale C. Mann Robert G. Riley
• Biocide By-Products in Aquatic Environments, Final Report Covering Period September 10, 1976 through September 30, 1979, NUREG/CR-1300	Roger M. Bean Charles I. Gibson David R. Anderson

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INTRODUCTION

Bromoform was identified by Bean et al., (1980) as the major halogenated organic produced by the chlorination of Sequim Bay, Washington sea water. As part of the NRC program to determine the toxic and bioaccumulation potential for chlorination by-products, a series of tests were undertaken to determine the acute toxicity and bioconcentration potential for bromoform. The test organisms were selected, in conjunction with NRC, as species that would have a high potential for exposure to the cooling water discharge streams of nuclear-fueled steam electric stations and were of commercial and/or recreational importance. These species, Protothaca staminea (the Pacific littleneck clam), Crassostrea virginica (the Eastern oyster), Mercenaria mercenaria (the Eastern hard clam, or quahog), Penaeus aztecus (the brown shrimp), and Brevoortia tyrannus (the Atlantic menhaden) were to be used in 96-hour LC50 tests and 28-day uptake/28-day depuration tests. To avoid animal transportation problems, the studies with the Pacific littleneck clam were conducted at the Battelle Marine Research Laboratory, Sequim Bay, Washington, and the studies with the other four species were conducted at the Daytona Beach Marine Laboratory, Daytona Beach, Florida.

Toxicity

Methods and Materials (Sequim):

The first problem that needed to be solved before the toxicity and bioaccumulation testing could be conducted was the delivery of bromoform in solution to the test organisms. Initial tests to produce bromoform/seawater solutions at concentrations desired for the toxicity testing by liquid-to-liquid contact failed. When bromoform was mixed directly with sea water, a large portion would remain in droplet form and eventually settle to the bottom of the tank. The mixing of bromoform and sea water for long periods (24 to 48 hours) did produce bromoform/seawater solutions with desired bromoform concentrations, but the procedure was too inefficient and time consuming to be useful in providing sufficient toxicant stock solutions for flow-through systems. As a confounding factor, the availability of bromoform from United States suppliers dried up, creating a situation where the quantity of bromoform available for use in the tests was limited.

The system that was finally developed and selected for use in the toxicity and bioaccumulation tests was the air sparge method diagrammed in Figures 1 and 2. Figure 1 is the system used at Sequim, and Figure 2 is the system used at Daytona. The basic concept is to flow compressed air into a flask containing bromoform. The exhaust from this flask is carried by silicone tubing to either the test tank as in the case in Figure 1, or a mixing/splitter tank as is the case in Figure 2. The concentration of bromoform in the sea water is regulated by the airflow rate and seawater flow rate.

In preparing to conduct the toxicity test with Protothaca staminea, it was found that at high concentrations (greater than 600 mg bromoform/l) the clams would close up and die in the closed position. Clams were then tested for their sensitivity to bromoform. In this test, we observed five littleneck clams which were placed in a tank receiving clean, running, filtered Sequim Bay sea water. When all five clams had extended their siphons, a bromoform/seawater solution was introduced into the flowing sea water entering the tank. A water sample was collected for bromoform analysis when the first response was noted (siphons slightly retracted), again when all siphons were retracted except one which was again extended, and when all siphons were retracted (Table 1). The flow of bromoform/seawater was continued for 4 hours and then stopped. After one hour, the siphons again extended. The clams were held for 96 hours in clean, running sea water with no mortality.

These preliminary tests, the short supply of bromoform, and the data which indicated that in the real-world situation bromoform would be present at 30 to 80 µg/l concentrations, resulted in the decision to conduct the 28-day exposure studies at bromoform concentrations of 1, 5, 10, and 20 mg/l and observe mortality. The mortality data could then be used to estimate the concentration range of the LC50 and, if warranted, toxicity testing to further refine the LC50 point could be undertaken. Organisms from the lower concentrations (1 and 10 mg/l), would be used for the bioaccumulation studies.

Collection and Exposure (Sequim)

Specimens of P. staminea were collected from Sequim Bay, Washington, and held in unfiltered, ambient running sea water for four days prior to testing. There was less than 1% mortality during the holding period. The exposure of P. staminea was conducted by bubbling air saturated with bromoform (Eastman Spectrograde®) directly into the exposure tanks (Figure 1). The tanks were 30-liter glass aquaria layered with approximately 5 cm of coarse sand. Seventy-five clams were randomly selected and placed in each tank. Concentration of bromoform and mortality was monitored five days a week (Monday through Friday). Bromoform/air flows were adjusted to maintain nominal concentrations of 0, 1, 5, 10, and 20 mg/l.

Water quality was not monitored. Previous studies have shown the water quality to be stable with only slight seasonal variation. Salinity ranges seasonally from 29 to 31 ppt, and temperature from 7° to 13°C. Oxygen is normally at 100% saturation (9.4 mg/l) except in late August and September when concentrations down to 80% saturation have been observed. Water samples for bromoform analysis were collected and stored under refrigeration.

Collection and Exposure (Daytona)

Bromoform exposures with four marine species (Crassostrea virginica, Mercenaria mercenaria, Brevoortia tyrannus, and Penaeus aztecus) were conducted at Battelle's Florida Marine Research Facility in Daytona Beach, Florida. With the exception of shrimp which were purchased from a local bait dealer, clams, oysters, and juvenile menhaden were collected by Battelle staff members in the Halifax River within one mile of the Florida Marine Research Facility. Shrimp and menhaden were held in 11,350- ℓ outdoor circular holding tanks with a continuously flowing supply of Halifax River sea water, filtered through sand and activated carbon. Purina Trout Chow was fed at a daily rate of 5% body wt. Clams and oysters, held in 265- ℓ fiberglass water tables, were supplied with unfiltered water as a food source. Clams were placed in 5 cm of fine beach sand. All organisms were held for at least one week prior to exposure with less than 1% mortality. Average organism individual wet weights were as follows: shrimp - 3.3 g, oysters - 75 g, clams - 141 g, and menhaden - 3.5 g.

During all exposures, salinity, temperature, dissolved oxygen, pH, behavior, mortality and bromoform water concentrations were monitored daily.

In June, 1978 two 96-hr exposures with 5 concentrations plus a control for each were made with shrimp. Each exposure consisted of 12 paired tanks with two tanks for each concentration plus two controls. Ten shrimp were placed in each tank. The first exposure was initiated on June 20, and the second on June 29.

A series of 12 paired tanks were used for simultaneous exposures of menhaden and oysters. This consisted of two controls and two tanks for each of 5 exposure concentrations; one containing 10 juvenile menhaden and the other containing 10 oysters. The experiment was initiated on July 24, 1978. Due to the rapid onset of mortalities of menhaden at the three highest concentrations, mortality data was taken after 2 and 5 hours of exposure and daily thereafter. Daily observations of oysters included noting filtering activity in each tank. Oysters were observed for four days after exposure for latent mortality.

Six tanks were used for clam exposures; 5 bromoform concentrations plus one control. The experiment was initiated on October 2, 1978. Filtering activity was noted daily, and clams were also monitored for four days following exposure for latent mortality.

Uptake and Depuration

For the P. staminea tests, the target bromoform concentrations were 0 (control), 1 and 10 mg/ ℓ . However, because these concentrations were

impossible to maintain, the average measured concentrations are given in the tables listing the results. Target concentrations of 0 (control), 0.1 mg/l and 1.0 mg/l were used in the tests with C. virginica, M. mercenaria, B. tyrannus and P. aztecus. All four species were exposed simultaneously for this 28-day uptake/28-day depuration experiment. Sixty organisms (30 placed in each of a tank pair) of each species were exposed to each target concentration, thereby requiring 12 pairs of 24 exposure tanks. Subsamples of five P. staminea and three each of the other four species were harvested on days 0, 1, 2, 4, 7, 14, 21, 28 of the uptake phase and day 1, 2, 4, 7, 21 and 28 of the depuration phase. P. staminea specimens were frozen in glass jars for later analysis. Specimens of the other species were frozen in collapsed plastic zip-lock bags which were then sealed in 1 l plastic wide mouth jars. In some cases, mortality reduced the number available so the smaller sample groups were collected during the latter phase of the test.

Bromoform Analysis (Water)

Water samples collected from the bioassay tanks were stored in tightly capped and completely filled 60-ml bottles at 4°C prior to analysis. Subsamples (5 to 10 ml) were removed from the bottles and transferred to 25 ml screw-cap vials containing 10 ml hexane. The vials were hand-shaken for 90 seconds and allowed to stand until phase separation occurred. One ml of the hexane phase was transferred to a 1 ml Hewlett Packard[®] autosampler vial with septum cap. An internal standard was then added to the vial by syringe (3 µl of 152 µg/ml, 1, 3-dibromopropane). The samples were then analyzed by electron capture gas chromatography, utilizing a Hewlett Packard model 5840[®] with autosampler.

The analysis conditions were as follows: Column - 30 meter SP2100 glass capillary with 15 to 1 split ratio; carrier gas - helium; oven temperature - 85°C, detector - ⁶³Ni electron capture. Calculation of sample concentration was conducted by the internal standard calibration method.

Bromoform Analysis (Tissue)

Tissues were analyzed for bromoform by homogenizing the tissue in water at 0°C and diluting with enough water to obtain a concentration of approximately 1 g tissue/10 ml tissue suspension. Aliquots (10-20 ml) of the aqueous tissue suspension were extracted with two 5-ml portions of hexane containing 1, 3-dibromopropane as an internal standard. The microliter samples of the hexane solution were injected into a gas chromatograph fitted with an 18-in. Porapak Q[®] column and a ⁶³Ni electron-capture detector. The column was operated at 185°C. The limit of detection of this procedure was 0.0005 µg/g, and the coefficient of variation ranged from 1% at the 1- to 8-µg/g level to 3% at levels below 0.1 µg/g. The coefficient of variation was calculated from 16 replicate analyses of each of 9 tissue samples.

LC50 Calculations

For the menhaden and shrimp, 96-hr LC50 concentrations were calculated by the methods of Litchfield and Wilcoxon (1949). The LC50 concentrations were estimated by the same method from the latent mortality in the case of the quahog and Eastern oyster. No LC50 concentration value was calculated for the Pacific littleneck clam.

TOXICITY RESULTS AND DISCUSSION

Protothaca staminea

Considerable difficulty was experienced in maintaining the desired bromoform concentrations. At the highest level (nominal 20 mg/l), the average of 20 values for the 28-day exposure was 27 mg/l with a range from 9 mg/l to 76 mg/l (Table 2). A plot of the concentrations measured are given in Figure 3. The other exposure tanks had similar variation (Figures 4, 5, and 6) and averaged 2 mg/l, 7 mg/l, and 19/l for the target concentrations of 1, 5, and 10 mg/l, respectively.

Mortality was observed only in the two higher concentrations. At the highest concentrations (average 27 mg/l), 21 of 60 clams were found dead on day 7 of the exposure, and then 10 of 24 clams were found dead on day 25. In the 10 mg/l exposure (average concentration of 19 mg/l), 9 of 45 clams were found dead on day 25. During the depuration cycle of the test, two more mortalities occurred at the highest concentrations, one recorded on day 4, and one recorded on day 7. The exact day of death for these individuals could not be determined. Of interest is the fact that no mortalities occurred at the lower two test levels through the 28-day uptake or 28-day depuration period.

Penaeus aztecus

The calculated 96-hr LC50 for *P. aztecus* was 26 mg/l with a 95% confidence interval between 33 mg/l and 20 mg/l (Figure 7). Of interest was the behavior exhibited by the shrimp at two levels of concentration. At bromoform concentrations of 19 mg/l and above, an avoidance response to the bromoform source occurred within 60 seconds of exposure. At concentrations of 31 mg/l and above, a narcotic-like effect, where the shrimp were observed lying on their sides on the bottom of the tank with their abdominal appendages undulating, occurred within 120 minutes and continued throughout the experiment or until death.

Brevoortia tyrannus

The calculated 96-hr LC50 for *B. tyrannus* was 12 mg/l with a 95% confidence interval between 15 mg/l and 9 mg/l (Figure 8). As the menhaden approached death they began to lose equilibrium and lay on their sides at the bottom of the tank. Opercular movement gradually decreased until all movement stopped.

Crassostrea virginica and Mercenaria mercenaria

A 96-hr exposure period appears to be inadequate to generate meaningful LC50 data on clams and oysters. At concentrations above 10 mg/l, filtering ceases and the bivalves close and remain closed for much of the exposure period. At the end of 96 hours there were no mortalities with either M. mercenaria or C. virginica. However, mortalities occurred during the 3-day period immediately following the 96 hours of exposure to bromoform. Based on this latent mortality data, the 50% mortality concentration of C. virginica and M. mercenaria was estimated to be in the range of 40 mg/l and 140 mg/l, respectively (Figures 9 & 10).

DISCUSSION

The results of the bioassays indicate that bromoform does not cause acute effects to the species tested at concentrations below 1 mg/l. Menhaden were the most sensitive, with a 96-hr LC50 of 12 mg/l. Shrimp were next in sensitivity with a 96-hr LC50 of 26 mg/l. The bivalves tested had 96-hr LC50's that were apparently above 40 mg/l. These bromoform concentrations are well above those one would expect in a power plant discharge, based on the findings of Carpenter and Smith (1978) and Bean et al. (1978). They reported bromoform concentrations of 30 to 350 ppb in sea water that had been chlorinated at a rate of 1 to 4 ppm. This is a conversion rate of about 0.02 to 0.08 parts bromoform for each part chlorine added; for this conversion rate, chlorine would have to be added at a rate of 500 mg/l to form sufficient bromoform to cause acute effects. At this rate of chlorination (unless there is an extremely heavy chlorine demand), the residual oxidant will have a much more pronounced effect than bromoform. The literature reports that total residual oxidant causes acute effects to the tested species in the 1.5 mg/l to 0.005 mg/l range (Roberts et al., 1975; Thatcher, 1978; Scott et al., 1978).

The mortalities noted in the P. staminea 27 mg/l uptake/depuration exposure are curious in that they appear to have occurred at two single points in time. Both occurrences were four days after peak exposure concentrations were experienced (Figure 3). The first mortality occurred after a peak of 56 mg/l bromoform, and the second occurred after a peak of 76 mg/l bromoform. Thus, in this exposure tank it appears that there may be threshold concentration above which mortality begins.

The mortality that occurred in the 19 mg/l exposure did not follow the pattern found at the higher level. The concentration in this system was not as variable as in the 27 mg/l exposure, and the mortality did not occur until ten days after a peak concentration occurred.

The delayed mortalities noted in the oyster tests, and the above clam mortalities indicate that the action of the bromoform at high concentrations can cause severe enough damage to prevent recovery. This action can result from short-term exposure to high concentrations (probably greater than 50 mg/l) or longer term exposure to lower concentrations in the 20 to 30 mg/l range. In regard to concern about the release of bromoform from steam electric stations, the exact concentrations required for either short-term or long-term mortality is academic since these levels are approximately 1000 times those expected to be found.

At sublethal concentrations, the menhaden and shrimp exhibited some qualitative behavioral changes. After 48 hours, juvenile (under 7 cm T.L.) menhaden exposed to 6 mg/l and 9 mg/l bromoform exhibited extreme excitation to external stimuli such as loud noises, quick movements or sudden light changes. These stimuli would cause the fish to swim rapidly in random directions and frequently collide with the tank walls. In control tanks and at the higher concentrations, this response did not occur. This excitability continued for up to 20 days after the exposure to bromoform had been terminated.

The shrimp responded similarly at bromoform concentrations between 0.4 mg/l and 6 mg/l. However, at concentrations below 3 mg/l the response was no longer evident within one hour after bromoform addition was stopped. At concentrations between 3 mg/l and 6 mg/l, the response continued for at least one day.

These observations are qualitative but in complete opposition to the response noted for those organisms that died. At the higher levels, the bromoform appeared to act as a narcotic. Both shrimp and menhaden gradually slowed down, lost orientation and eventually stopped pleopod or opercular movement. This condition was reversible for the shrimp, which recovered within a few hours if the bromoform input was stopped before pleopod motion ceased.

Based on the 96-hr LC50 studies and mortality data from the uptake and depuration studies, the potential for acute environmental effects (to the studied species) from bromoform created through chlorination of steam electric station cooling waters is minimal. The behavioral responses noted should be considered subjective observations that may or may not be related to bromoform exposures. To determine if the behavioral responses noted are, in fact, real changes and caused by bromoform, further research will be necessary.

RESULTS AND DISCUSSION (Uptake and Depuration)

The results of the tissue analyses are presented in Tables 3 through 7. The average body burdens of bromoform of each harvest date are plotted with the daily water bromoform concentrations in Figures 11 through 25.

It was difficult to hold the water-bromoform concentrations within the desired target concentrations, so it is the average concentrations for the 28-day exposure period that are listed in the tables. However, in discussion about concentration factors, the body burdens are compared to the water concentration on the day of harvest.

In general, the three molluscan species P. staminea, C. virginica and M. mercenaria had tissue concentrations that were about equal to the water concentrations. There were several exceptions to this generalization, the most notable of which was the high body concentration (11.59 mg bromoform/g tissue) found on day 14 in the 1 mg bromoform/l C. virginica test. The 11.59 mg bromoform/g tissue is a concentration factor of approximately 15, three times higher than any other concentration factor observed for the oysters. However, that particular oyster and three others appear to be exceptions to the general trend of body burdens, which are approximately the same as water concentrations (Figure 1). At 0.1 mg bromoform/l the tissue concentrations were close to the water concentrations on the day of harvest.

M. mercenaria body burdens in the 1.0 mg bromoform/l test followed the water concentrations during the first week but remained lower than the water for the rest of the exposure. At 0.1 mg bromoform/l the body burdens were slightly above the water concentrations but, in general, followed the water levels closely except for a single incidence on day 28. P. staminea tissue concentrations were also similar to the water concentrations on day of harvest.

Menhaden and shrimp were different from the molluscs, showing body burdens higher than the water concentrations at the 0.1 mg bromoform exposure condition; at 1.0 mg bromoform/l they were higher than the water during the first week but then fell to approximately the water concentration level for the remaining three weeks. The control organisms also had significant levels of bromoform in their tissues during the period when bromoform was introduced into the other test systems. In fact, even though the exposure systems were completely separate, at times there were measurable concentrations of bromoform in the control systems. However, as soon as use of bromoform in the other systems was stopped, it disappeared from the controls. Apparently, there was sufficient bromoform vapor present in the air to allow some to enter the control exposure systems. The post exposure data, however, indicate that the water and tissue levels of bromoform return to zero and remain there. Therefore, the controls in these tests are actually serving as low-concentration exposure systems. The data for the period after bromoform introduction was stopped show that water concentrations and tissue concentrations returned to zero. This does not provide the classical control condition but serves as a strong indication that if a completely separate system in another laboratory room had been used, the control levels would have stayed at zero.

The average water concentration during the 28-day exposure of menhaden in the control system was 0.03 mg bromoform/l, a concentration not different from the 0.1 mg bromoform/l target test condition average of 0.04 mg bromoform/l. The same was true for the menhaden exposures where the control average bromoform concentrations was 0.04 mg/l and the 0.1 mg/l target was 0.05 mg/l.

These data indicate that bromoform is taken up by these species, but that the degree of concentration depends on the individual, the species and the water concentrations. For molluscs, the general trend is for the tissue concentrations to reflect the water concentrations. Menhaden and shrimp, however, show a tendency to concentrate bromoform by a factor of 3-50 above ambient water concentrations. However, there are also indications that at water concentrations above 0.1 mg/l, the body burdens decrease with time to approximately 0.4 µg/g tissue, a body burden concentration similar to that found in organisms exposed to lower water concentrations.

Testing for bioaccumulation of bromoform is very difficult because of the problem associated with providing a precise concentration of bromoform in water for the 28-day exposure period. In addition, there are problems of vapor contamination of the control systems if the two are operated in the same room. However, the results of the tests presented here indicate that for the molluscs tested, the tissue concentrations will be a reflection of the ambient water concentrations, and that menhaden and shrimp will concentrate bromoform up to a point, after which there appears to be a maximum body burden that is maintained with time. Further testing is needed to determine where within the organism the bromoform is concentrated and the factors that influence its uptake.

CONCLUSIONS AND RECOMMENDATIONS

The toxicity tests conducted indicate that for brown shrimp the calculated 96-hr LC50 is 26 mg/l with a 95% confidence interval between 33 and 20 mg/l.

The 96-hr LC50 concentration for menhaden was calculated to be 12 mg/l with a 95% confidence interval between 15 mg/l and 9 mg/l.

Behavioral changes were observed in both menhaden and shrimp exposed to sublethal concentrations of bromoform.

Standard 96-hr LC50 values were not calculated for littleneck clams, quahogs or Eastern oysters. Based on latent mortalities and mortalities in the 28-day uptake exposures, it is estimated that the 96-hr LC50 value would be greater than 30 to 40 mg/l.

All the tested species rapidly took up and depurated bromoform. The mollusk species had tissue concentrations that were above water concentrations during the first week of exposure, but the tissue concentration decreased during the last three weeks of exposure and were reflective of the ambient water concentrations.

In shrimp and menhaden the tissue concentrations were also highest during the first week of exposure. After that they fell to a concentration of approximately 0.4 $\mu\text{g/g}$ and remained there for the remaining three weeks of exposure. The tissue concentrations of 0.4 $\mu\text{g/g}$ appear to be maintained independent of the water concentration.

The 96-hr LC50 values and estimated LC50 values indicate that relatively large amounts of bromoform would need to be generated in power plant discharges to cause acute toxicities.

The extent and consequences of the behavioral changes noted on the survival of the shrimp and menhaden are not known.

The bromoform uptake data indicates that if bromoform is present in the water, it will be present in the tissue of the organism.

Further investigations would be required to refine the LC50 values and to determine the major factors that determine the extent to which bromoform is bioaccumulated.

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*Available for purchase from the NRC/GPO Sales Program, U.S. Nuclear Regulatory Commission, Washington, DC 20555, and the National Technical Information Service, Springfield, VA 22161.

Table 1. Concentration of bromoform (mg/l) in tank at time of noted clam responses.

Slight retraction of siphon	430 mg/l
Retraction, then extension by one clam	425 mg/l
All siphons retracted	346 mg/l

Table 2. Measured Bromoform Concentrations (mg/l) in Protothaca staminea in Exposure Tanks

DATE	TARGET CONCENTRATION			
	1	5	10	20
5/30/78	1.9	4.8	8.4	11.8
5/31/78	2.3	4.2	14.0	27.1
6/1/78 (Surface)*	4.8	7.5	21.7	45.8
6/1/78 (Bottom)*	4.4	10.2	20.1	48.3
6/2/78	6.5	11.2	30.6	57.2
6/5/78	1.7	6.5	20.8	30.2
6/6/78	1.4	4.9	17.4	29.6
6/7/78	1.1	5.3	24.9	21.7
6/8/78	0.6	3.8	17.0	17.7
6/9/78	0.9	4.9	23.0	9.0
6/12/78	1.0	4.5	20.0	10.8
6/13/78	0.9	11.7	36.3	16.7
6/14/78	0.6	4.0	18.6	15.3
6/15/78	0.9	5.7	15.1	12.3
6/16/78	1.1	14.9	15.0	20.6
6/19/78	0.7	5.1	15.5	76.6
6/20/78	0.8	6.1	16.3	45.9
6/21/78	0.9	6.4	22.8	22.2
6/22/78	0.7	6.7	15.9	20.6
6/23/78	1.0	10.8	17.6	18.2
6/26/78	0.4	8.0	16.4	14.6
6/27/78	0.5	8.2	18.7	23.0

* Duplicate samples collected; surface sample by routine procedure, bottom samples by use of glass siphon with opening drawing sample from 5 cm above bottom.

Table 3. Bromoform Concentration in Each Test Organism for Eastern Oyster
(*Crassostrea virginica*) ($\mu\text{g/g}$ tissue-wet weight)

Average Bromoform Exposure Concentration (mg/l)	Tank #	1	2	4	7	14	21	28
		----- S A M P L E D A Y - U P T A K E -----						
0.03	1	0.24	0.24	0.00	0.00	0.00	0.00	0.00
		0.32	0.00	0.00	0.00	0.00	0.00	0.00
		0.00	0.24	0.00	0.00	0.00	0.00	0.00
0.09	2	0.55	0.20	0.00	0.21	0.00	0.24	0.00
		0.23	0.19	0.00	0.00	0.00	0.24	0.18
		0.25	0.21	0.00	0.00	0.00	0.19	0.00
0.86	3	0.89	1.54	3.33	0.32	0.64	0.85	0.48
		1.21	0.83	1.33	0.61	11.59	1.80	0.42
		1.37	0.78	1.10	0.46	0.35	0.39	0.22

Average Bromoform Exposure Concentration (mg/l)	Tank #	1	2	4	7	14	21	28
		----- S A M P L E D A Y - D E P U R A T I O N -----						
0.00	1	0.00	0.00	0.00	0.00	0.00	0.00	0.00
		0.00	0.00	0.00	0.00	0.00	0.00	0.00
		0.00	0.00	0.00	0.00	0.00	0.00	0.00
0.00	2	0.00	0.00	0.00	0.00	0.00	0.00	0.00
		0.00	0.00	0.00	0.00	0.00	0.00	0.00
		0.00	0.00	0.00	0.00	0.00	0.00	0.00
0.00	3	0.00	0.00	0.00	0.00	0.00	0.00	0.00
		0.00	0.00	0.00	0.00	0.00	0.00	0.00
		0.00	0.00	0.00	0.00	0.00	0.00	0.00

Table 4. Bromoform Concentration in Each Test Organism for Quahaug (Mercenaria mercenaria)
($\mu\text{g/g}$ tissue-wet weight)

Average Bromoform Exposure Concentration (mg/l)	Tank #	1	2	4	7	14	21	28
		----- S A M P L E D A Y - U P T A K E -----						
0.03	1	0.11	0.04	0.12	0.08	0.06	+	0.03
		0.11	0.07	0.00	0.07	0.07		0.00
		0.08	0.05	0.07	0.08	0.00		0.00
0.09	2	0.46	0.09	0.10	0.09	0.00	0.17	1.87
		0.42	0.15	0.09	0.08	0.07	0.15	0.25
		0.50	0.15	0.07	0.08	0.22	0.13	0.23
0.99	3	0.11	0.24	0.72	0.61	0.28	0.48	0.09
		0.06	0.14	1.01	0.33	0.31	0.70	0.21
		0.17	0.15	1.28	0.40	0.19	0.34	0.15

Average Bromoform Exposure Concentration (mg/l)	Tank #	1	2	4	7	14	21	28
		----- S A M P L E D A Y - D E P U R A T I O N -----						
0.00	1	0.00	+	1.83	0.00	0.00	0.00	0.00
		0.00		106.96	0.00	0.00	0.00	0.00
		0.00		0.00	0.00	0.00	0.00	0.00
0.00	2	0.00	0.00	*	0.00	0.00	0.00	0.00
		0.00	0.13		0.00	0.00	0.00	0.00
		0.16	0.40		0.00	0.00	0.00	0.00
0.00	3	0.00	0.00	0.00	0.00	0.00	0.00	+
		0.00	0.00	0.00	0.00	0.00	0.00	
		0.17	0.00	0.00	0.00	0.00	0.00	

+ No data-sample lost
* No organisms remaining due to earlier mortality

Table 5. Average (\bar{x}) Body Burden and Standard Deviation (S.D.) for Littleneck Clam (Protothaca staminea ($\mu\text{g/g}$ tissue-wet weight))

			<u>UPTAKE</u>						
Average Bromoform Exposure Concentration	Tank #		1	2	4	7	14	21	28
			----- S A M P L E D A Y -----						
0 mg/l	1	\bar{x}	.03	.15	.13	0	.50	.01	.04
		S.D.	\pm .03	\pm .17	\pm .08		\pm .07	\pm .02	\pm .03
2 mg/l	2	\bar{x}	9.94	5.59	13.65	4.08	2.22	2.05	1.08
		S.D.	\pm 2.16	\pm 1.38	\pm 4.81	\pm .67	\pm 1.05	\pm .40	\pm .18
19 mg/l	3	\bar{x}	22.14	17.63	38.05	37.29	22.05	24.85	14.25
		S.D.	\pm 4.58	\pm 2.30	\pm 6.35	\pm 7.14	\pm 3.35	\pm 7.90	\pm 7.32

			<u>DEPURATION</u>						
Average Bromoform Exposure Concentration	Tank #		1	2	4	7	14	21	28
			----- S A M P L E D A Y -----						
0 mg/l	1	\bar{x}	0	0	0	0	0	0	0
0 mg/l	2	\bar{x}	.12	.03	0	.02	0	0	0
		S.D.	\pm .26	\pm .03		\pm .02			
0 mg/l	3	\bar{x}	5.28	0.33*	0	0	0	.17	0
		S.D.	\pm 4.81	0.30	0	0	0	\pm .02	

* Range - 2 values only

Table 6. Bromoform Concentration in Each Test Organism for Menhaden (Brevoortia tyrannus)
($\mu\text{g/g}$ tissue-wet weight)

Average Bromoform Exposure Concentration (mg/l)	Tank #	1	2	4	7	14	21	28
		----- S A M P L E D A Y - U P T A K E -----						
0.03	1 (Control)	1.24	0.00	2.10	0.48	0.18	0.43	0.00
		0.78	0.00	0.00	0.26	0.00	*	*
		1.52	0.91	0.00	0.31	0.17	*	*
0.04	2	0.33	0.51	1.16	0.28	0.20	0.38	0.15
		0.64	0.24	1.13	0.35	0.16	*	*
		0.00	1.10	1.10	0.23	0.18	*	*
0.21	3	0.79	7.61	1.09	0.29	0.18	0.30	0.67
		0.59	0.00	0.83	0.34	0.28	*	*
		1.12	0.53	0.78	0.36	0.21	*	*

Average Bromoform Exposure Concentration (mg/l)	Tank #	1	2	4	7	14	21	28
		----- S A M P L E D A Y - D E P U R A T I O N -----						
0.00	1	0.00	0.00	0.00	0.00	+	+	+
		*	*	*	*			
		*	*	*	*			
0.00	2	0.00	0.00	0.00	0.00	+	+	+
		*	*	*	*			
		*	*	*	*			
0.00	3	0.00	0.00	0.00	+	+	+	+
		*	*	*				
		*	*	*				

+ No data

* No organisms remaining due to earlier mortality

Table 7. Bromoform Concentration in Each Test Organism for Shrimp (*Penaeus aztecus*)
($\mu\text{g/g}$ tissue-wet weight)

Average Bromoform Exposure Concentration (mg/l)	Tank #	----- S A M P L E D A Y - U P T A K E -----							
		1	2	4	7	14	21	28	
0.03	1	0.60	0.64	0.33	0.41	0.33	0.42	0.26	
		0.55	0.79	0.41	0.43	0.23	*	*	
		0.49	0.57	0.31	0.38	0.32	*	*	
0.05	2	0.51	0.67	0.32	0.39	0.33	0.38	0.00	
		0.40	0.50	0.39	0.24	0.30	*	*	
		0.44	0.81	0.40	0.34	0.35	*	*	
0.29	3	+	0.83	@	@	0.31	0.70	0.37	
			0.63			0.35	*	*	
			0.96			0.37	*	*	

Average Bromoform Exposure Concentration (mg/l)	Tank #	----- S A M P L E D A Y - D E P U R A T I O N -----							
		1	2	4	7	14	21	28	
0.00	1	0.00	0.00	0.00	0.00	0.15	0.00	*	
		*	*	*	0.00	0.00	0.00	*	
		*	*	*	*	*	0.00	*	
0.00	2	0.00	0.00	0.00	0.00	0.00	0.00	0.00	
					0.00	0.00	0.00	0.00	
						0.00	0.00		
0.00	3	0.00	0.00	0.00	0.00	0.00	0.00	*	
						0.00	0.00	*	
						*	*	*	

@ Sample not processed
+ Sample Lost
* No organism remaining due to earlier mortality

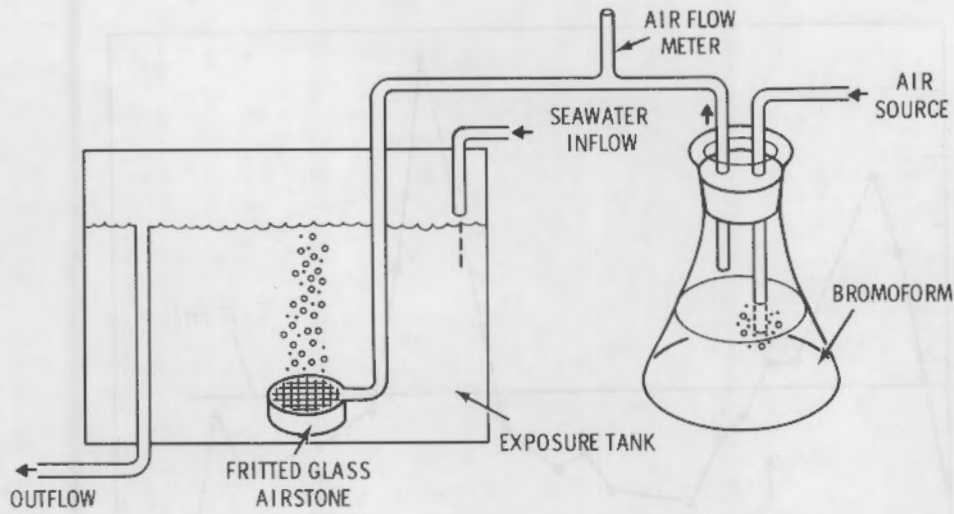


Figure 1. Toxicant delivery system for seawater/bromoform bioassays conducted at Sequim

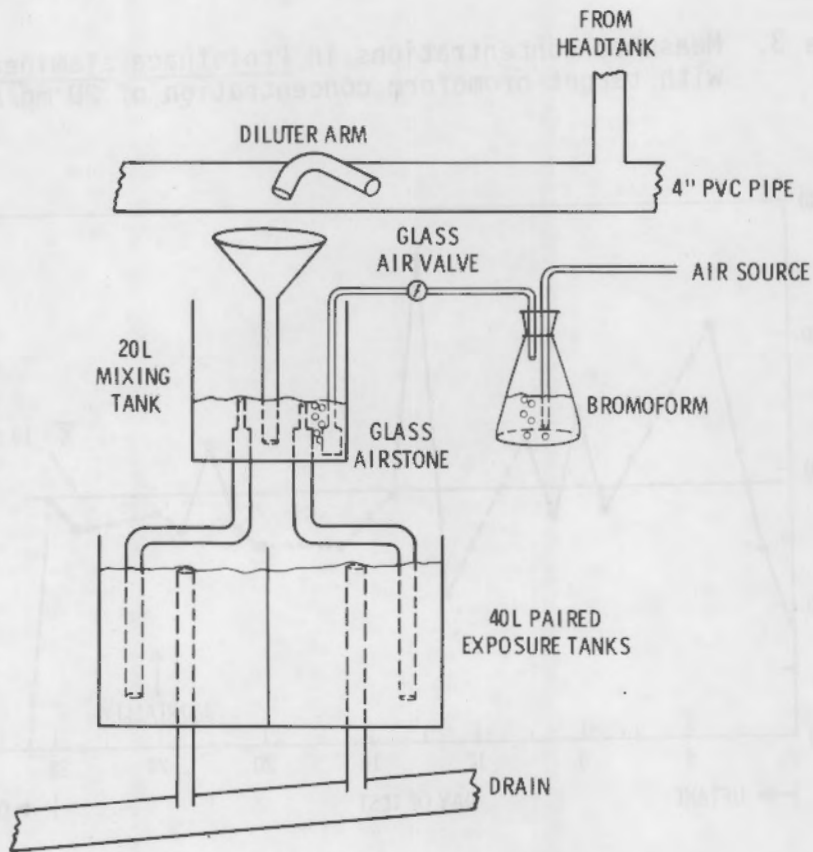


Figure 2. A section of the bromoform exposure system used at the Daytona Beach Laboratory

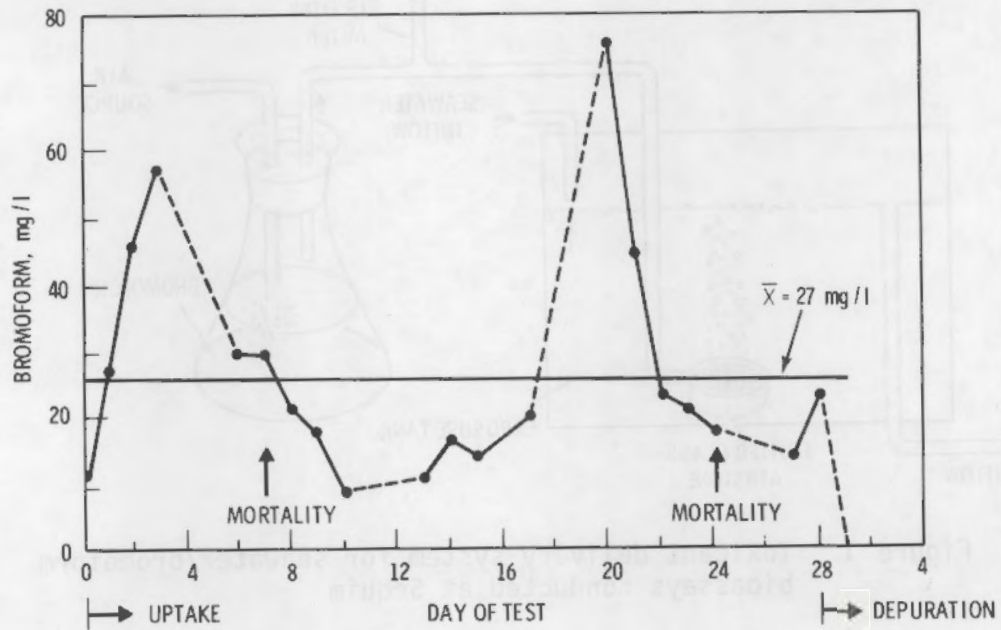


Figure 3. Measured concentrations in *Protothaca staminea* exposure with target bromoform concentration of 20 mg/l

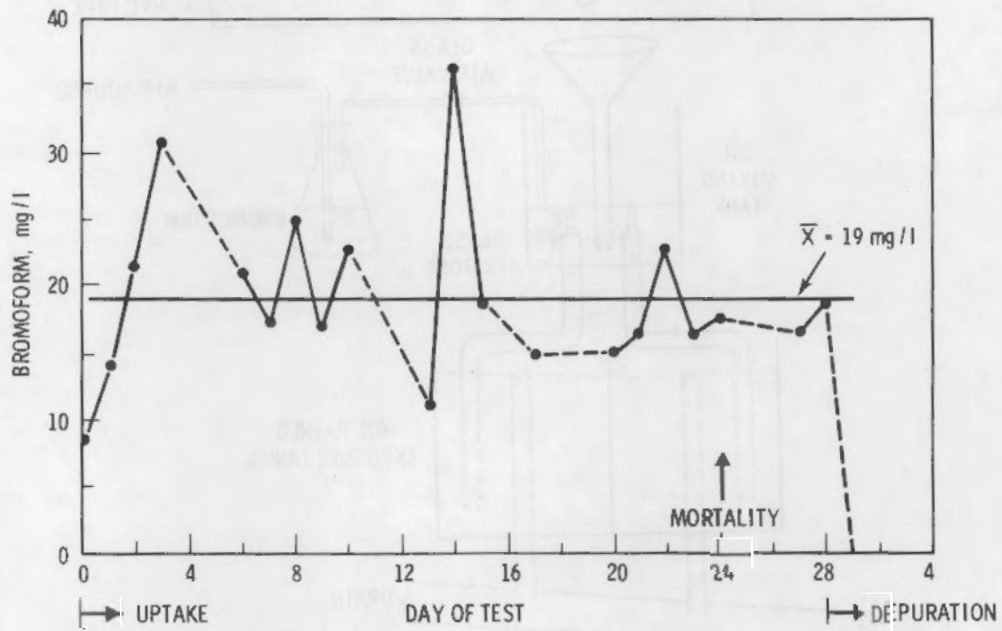


Figure 4. Measured concentrations in *Protothaca staminea* exposure tank with target bromoform concentration of 10 mg/l

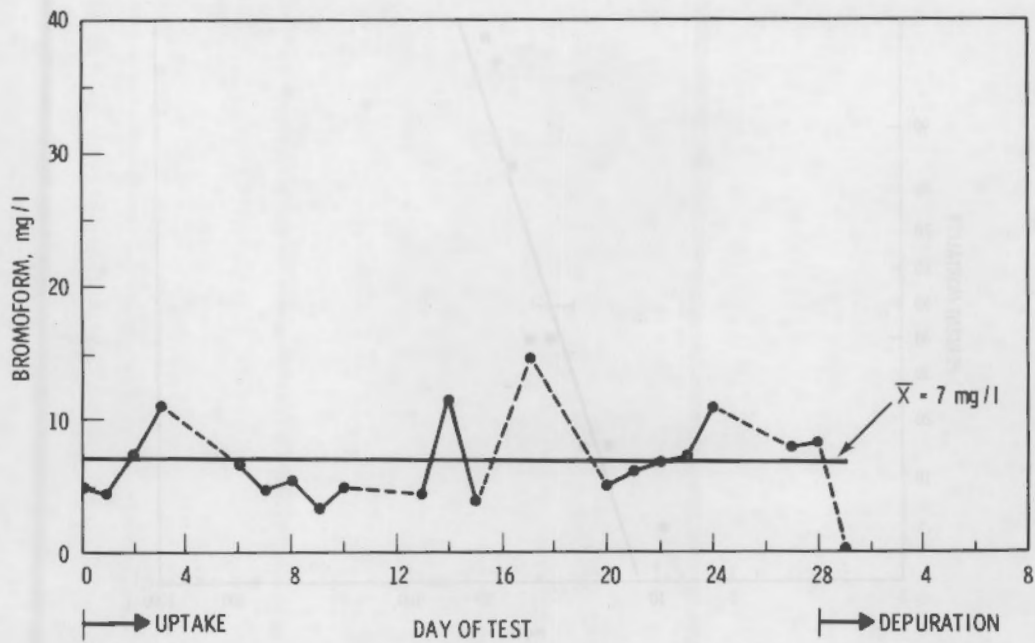


Figure 5. Measured concentrations in *Protothaca staminea* exposure tank with target bromoform concentration of 5 mg/l

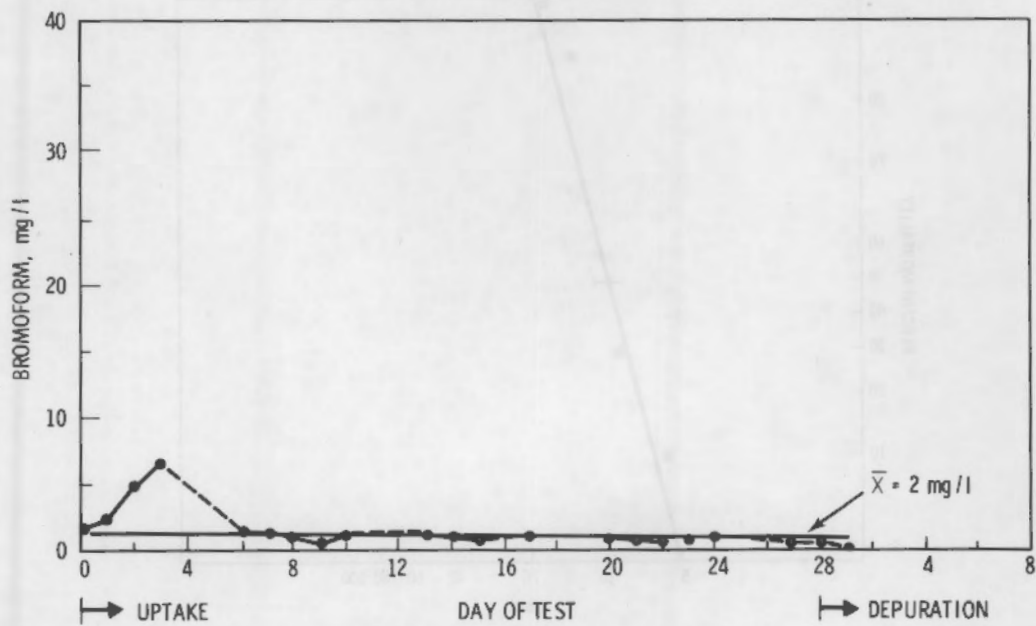


Figure 6. Measured concentrations in *Protothaca staminea* exposure tank with target bromoform concentration of 1 mg/l

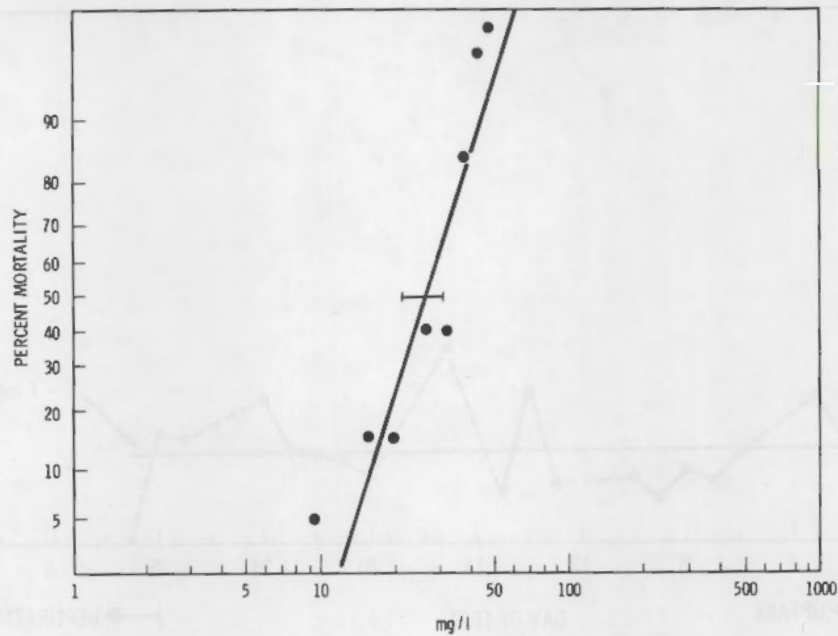


Figure 7. *Penaeus aztecus* mortality bromoform concentration plot for calculation of 96-hr LC50 by the method of Litchfield and Wilcoxon (1949)

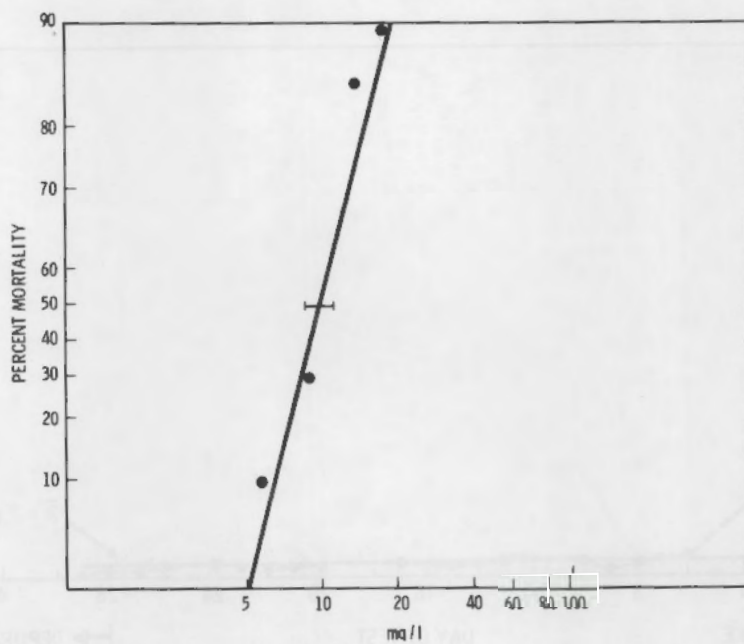


Figure 8. *Brevoortia tyrannus* mortality bromoform concentration plot for 96-hr LC50 calculation by the method of Litchfield and Wilcoxon (1949)

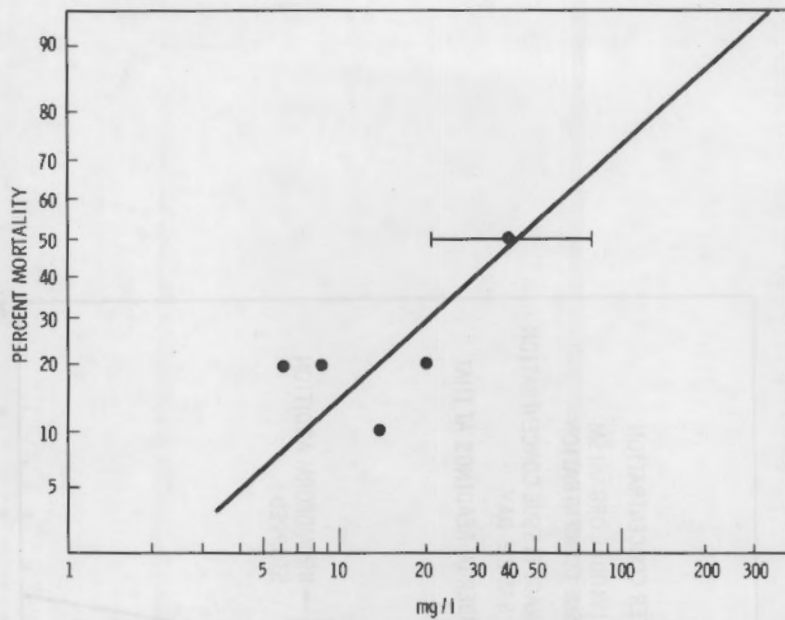


Figure 9. Crassostrea virginica latent mortality bromoform concentration plot used to determine concentration above which the 96-hr LC50 should fall

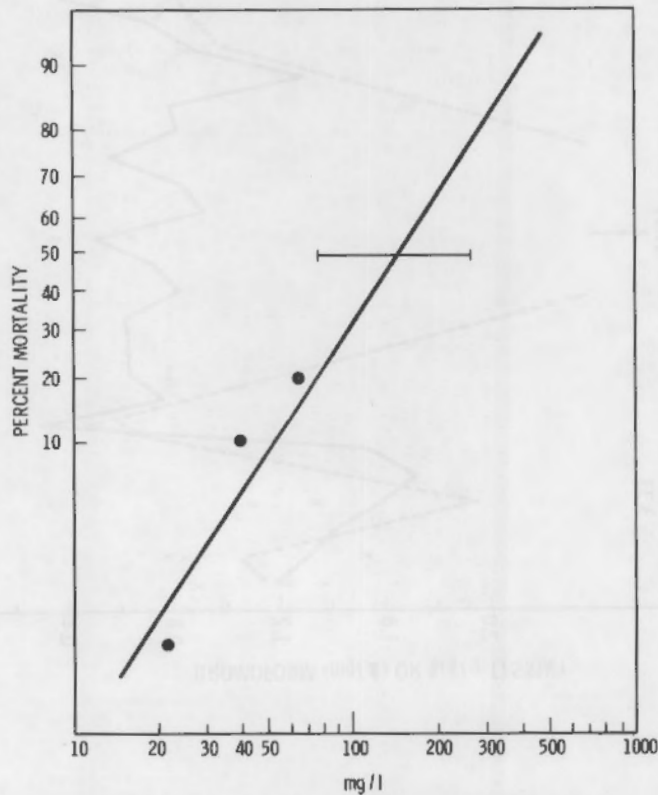


Figure 10. Mercenaria mercenaria latent mortality bromoform concentration plot used to determine concentration above which the 96-hr LC50 should fall

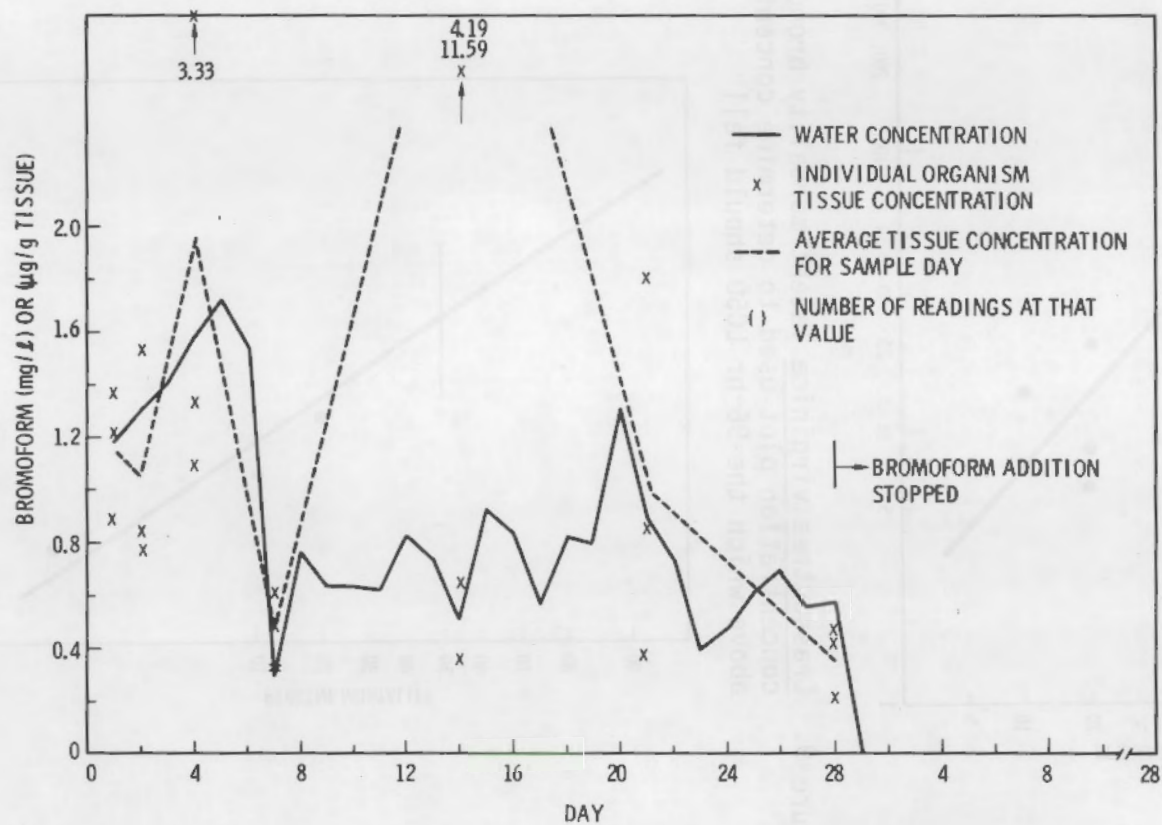


Figure 11. Water and tissue concentrations of bromoform (Eastern oyster, *Crassostrea virginica*) 28-day uptake/28-day depuration studies. Target water bromoform was 1.0 mg/l

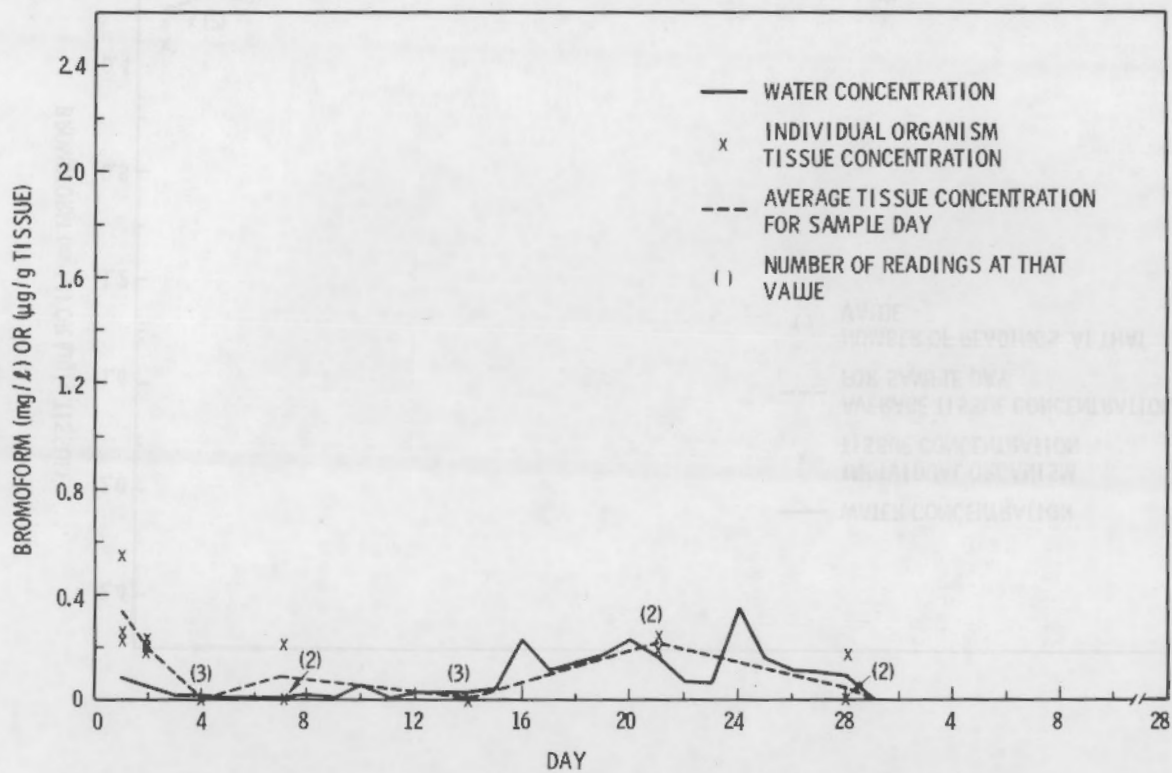


Figure 12. Water and tissue concentrations of bromoform (Eastern oyster, *Crassostrea virginica*) 28-day uptake/28-day depuration studies. target water bromoform was 0.1 mg/l

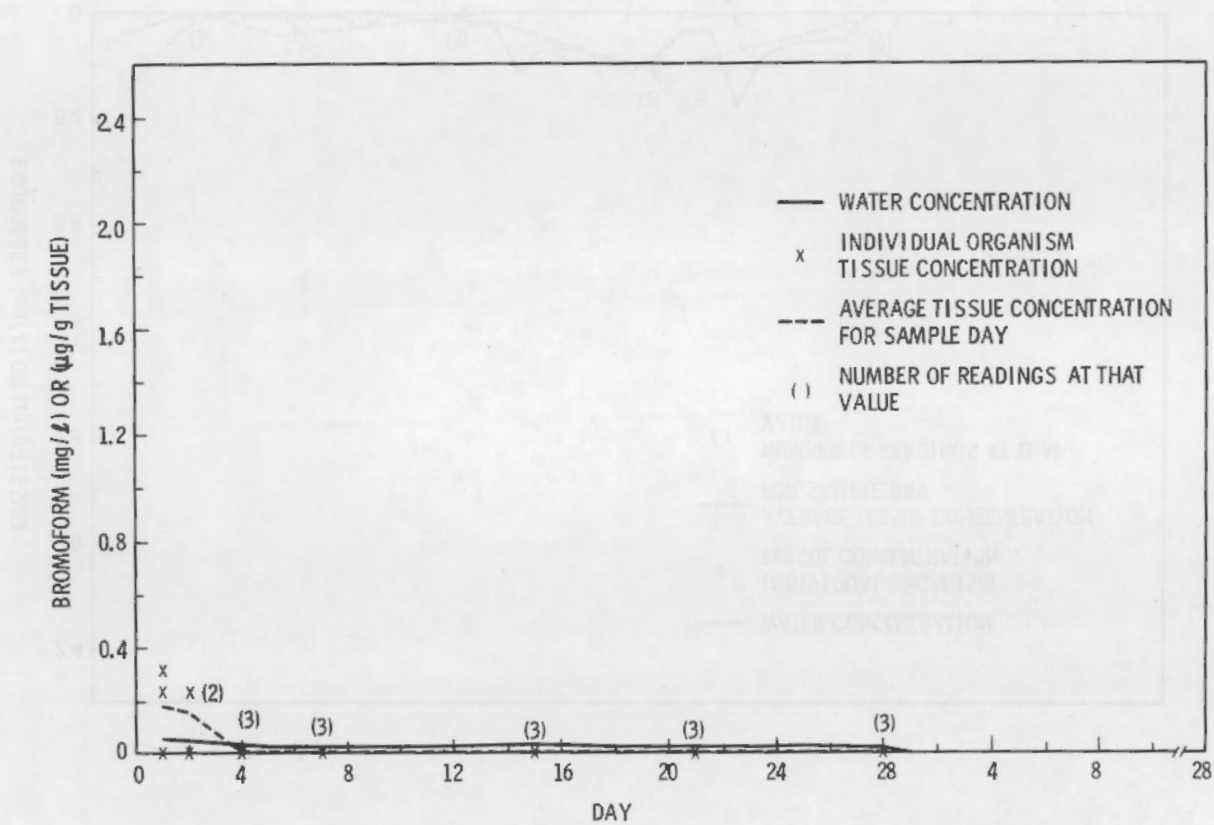


Figure 13. Water and tissue concentrations of bromoform (Eastern oyster, *Crassostrea virginica*) 28-day uptake/28-day depuration studies. Control

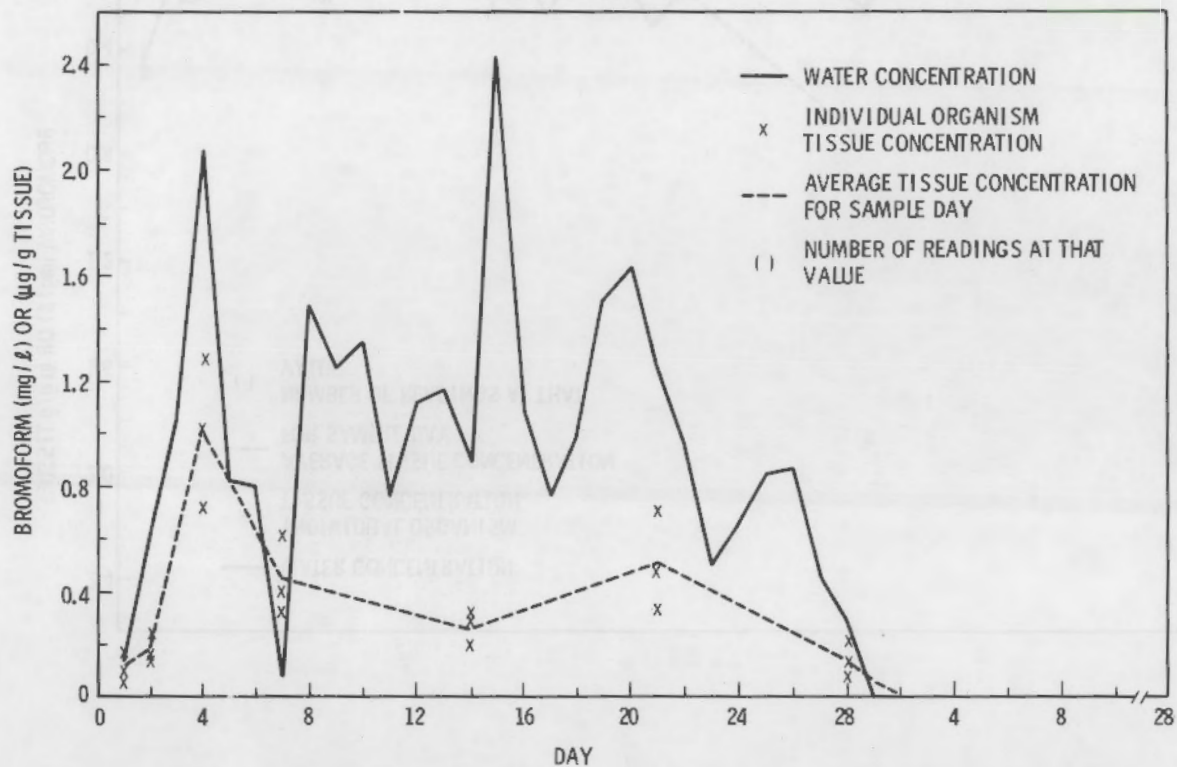


Figure 14. Water and tissue concentrations of bromoform (Quahaug, *Mercenaria mercenaria*) 28-day uptake/28-day depuration studies. Target water bromoform concentration was 1.0 mg/l

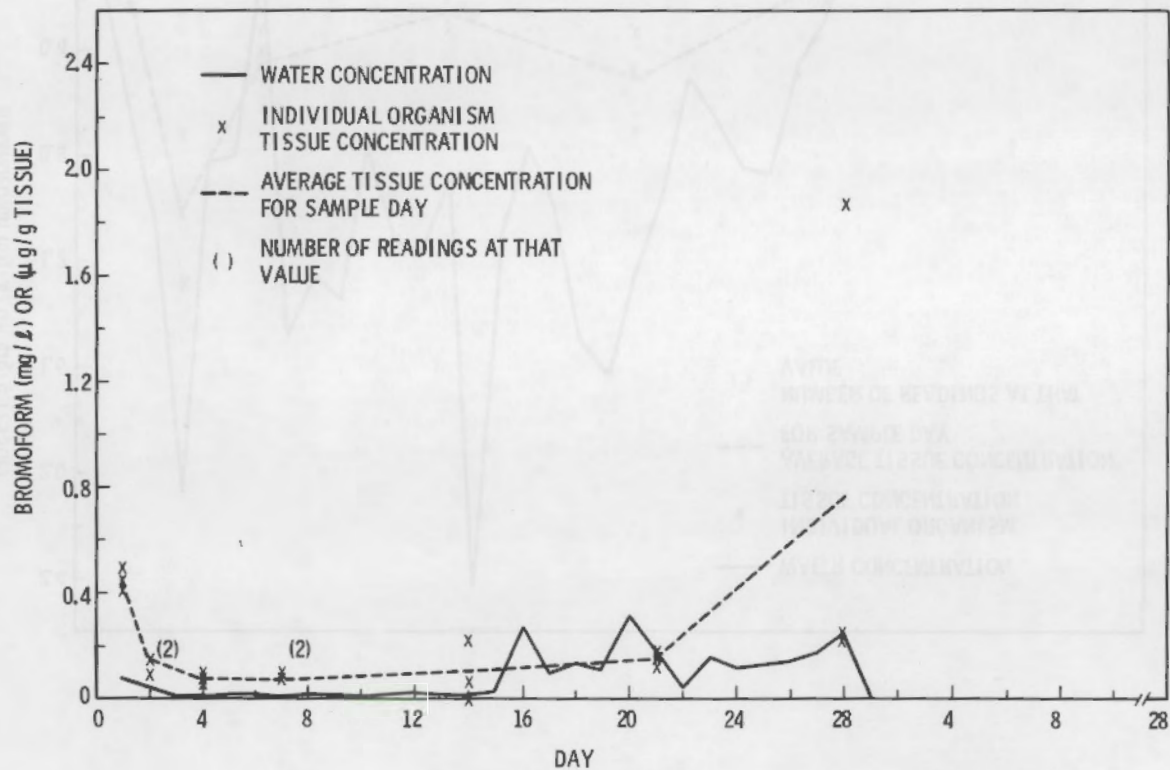


Figure 15. Water and tissue concentrations of bromoform (*Quahaug, Mercenaria mercenaria*) 28-day uptake/28-day depuration studies. Target water bromoform concentration was 0.1 mg/l

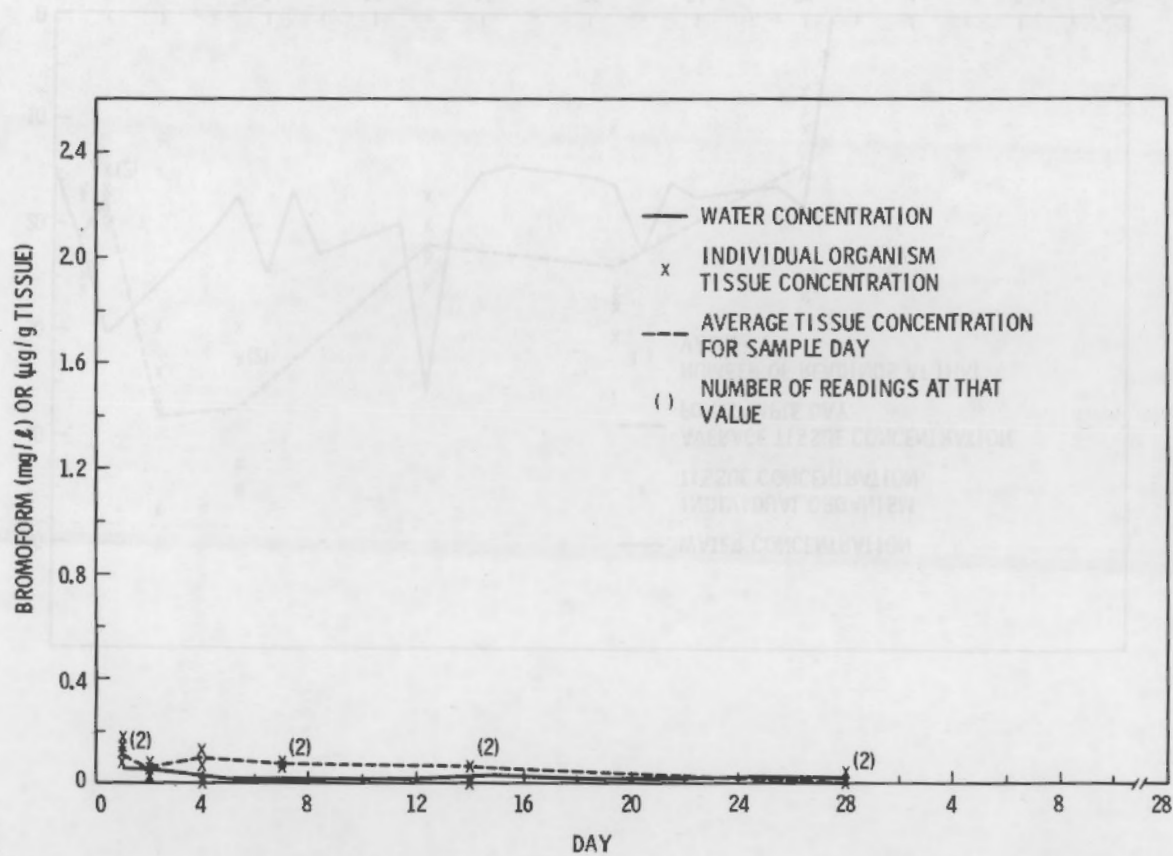


Figure 16. Water and tissue concentrations of bromoform (Quahaug, *Mercenaria mercenaria*) 28-day uptake/28-day depuration studies. Control

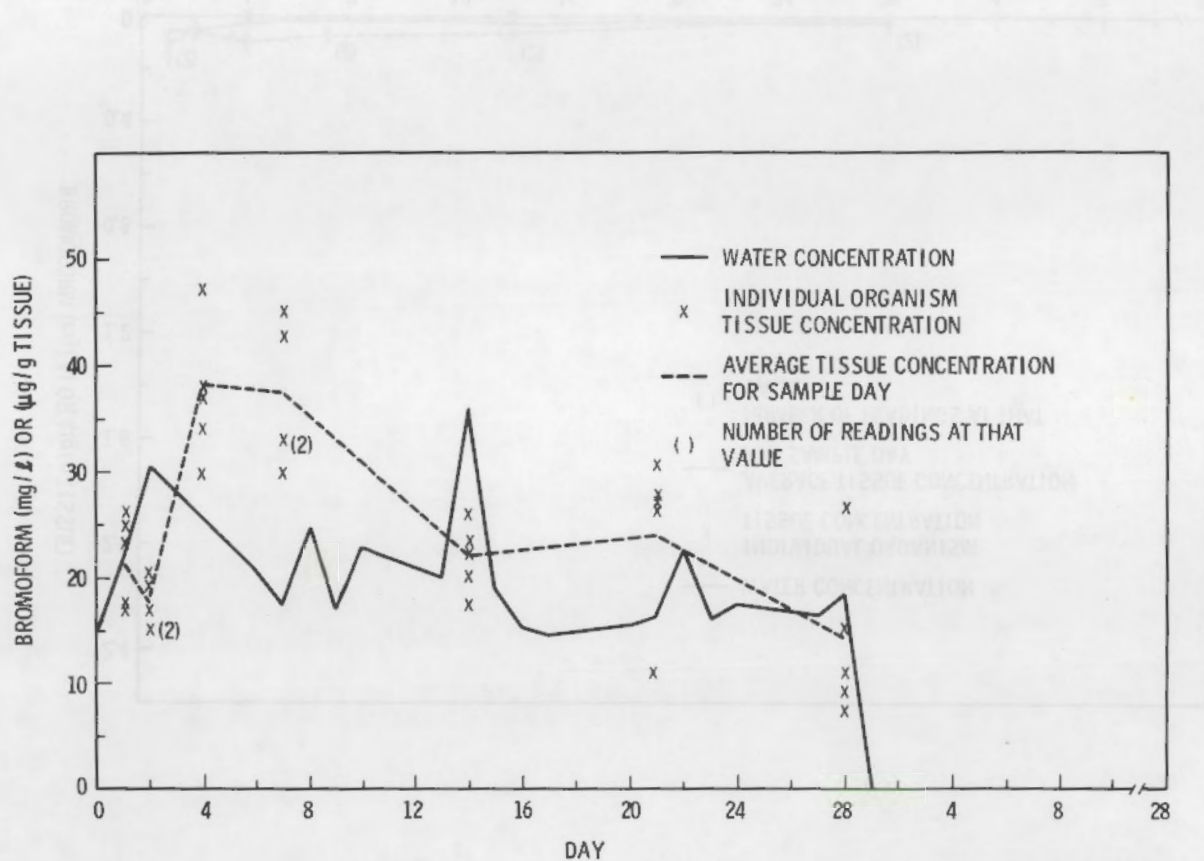


Figure 17. Water and tissue concentrations of bromoform (Littleneck clam, *Protothaca staminea*) 28-day uptake/28-day depuration studies. Target water bromoform concentration was 10 mg/l

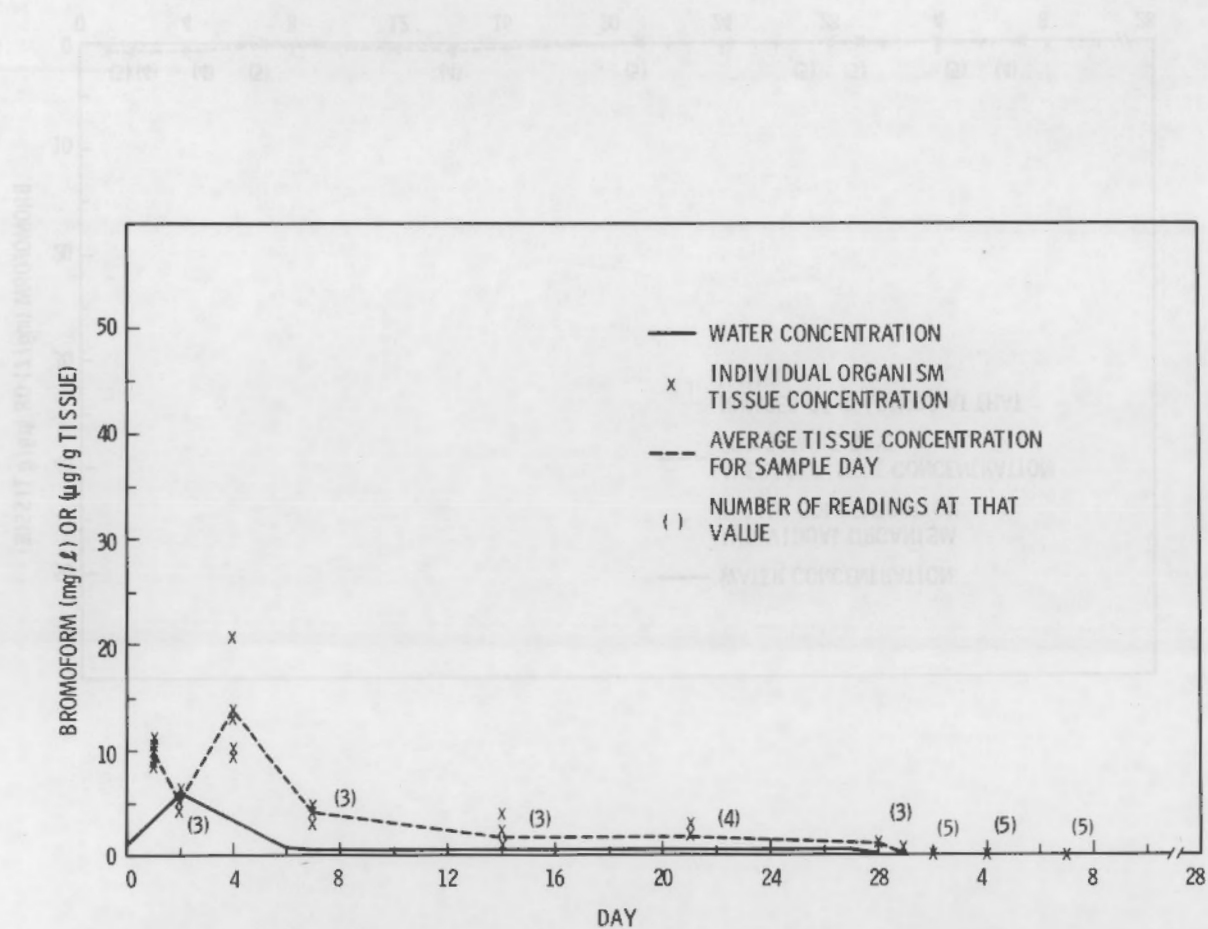


Figure 18. Water and tissue concentrations of bromoform (Littleneck clam, *Protothaca staminea*) 28-day uptake/28-day depuration studies. Target water bromoform concentration was 1.0 mg/l

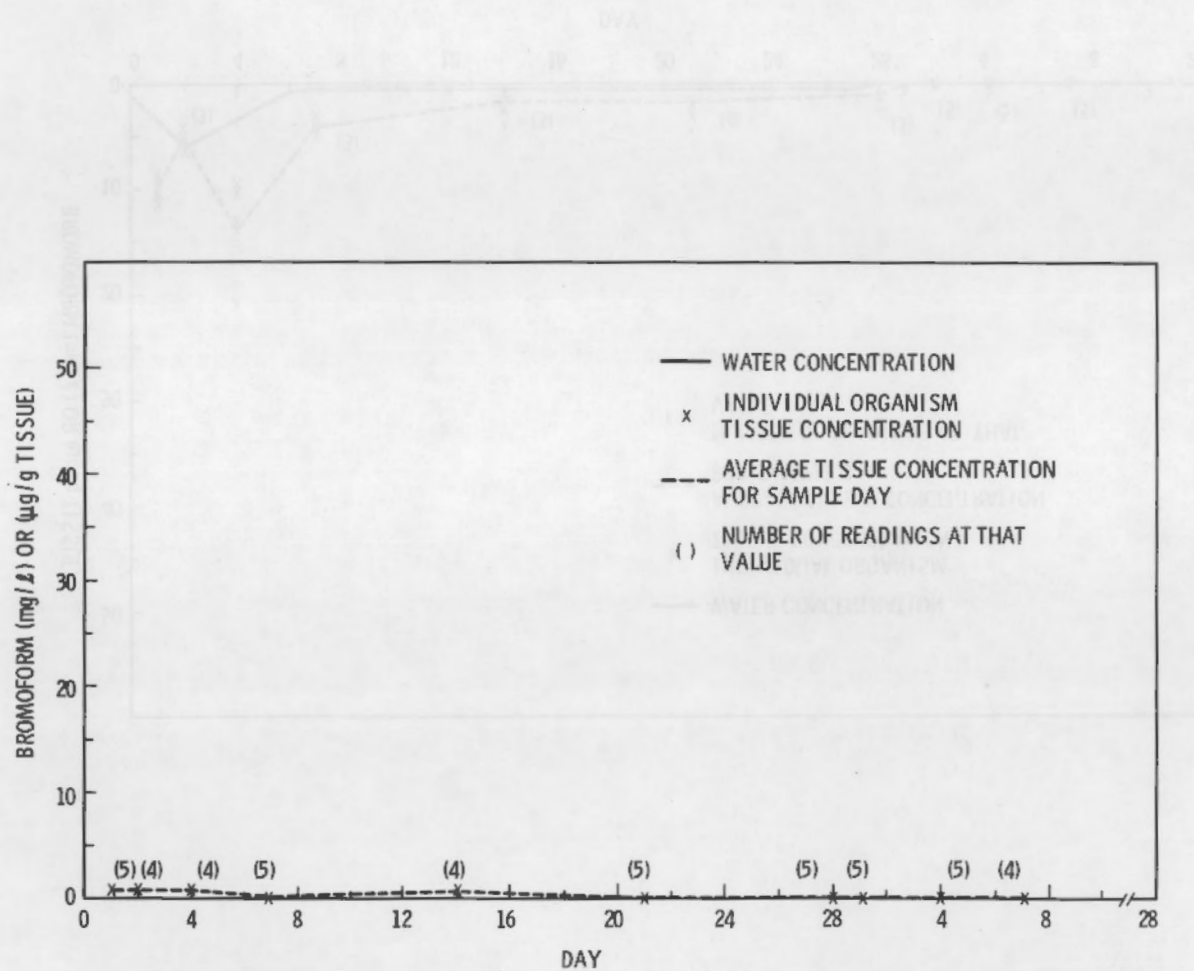


Figure 19. Water and tissue concentrations of bromoform (Littleneck clam, *Protothaca staminea*) 28-day uptake/28-day depuration studies. Control

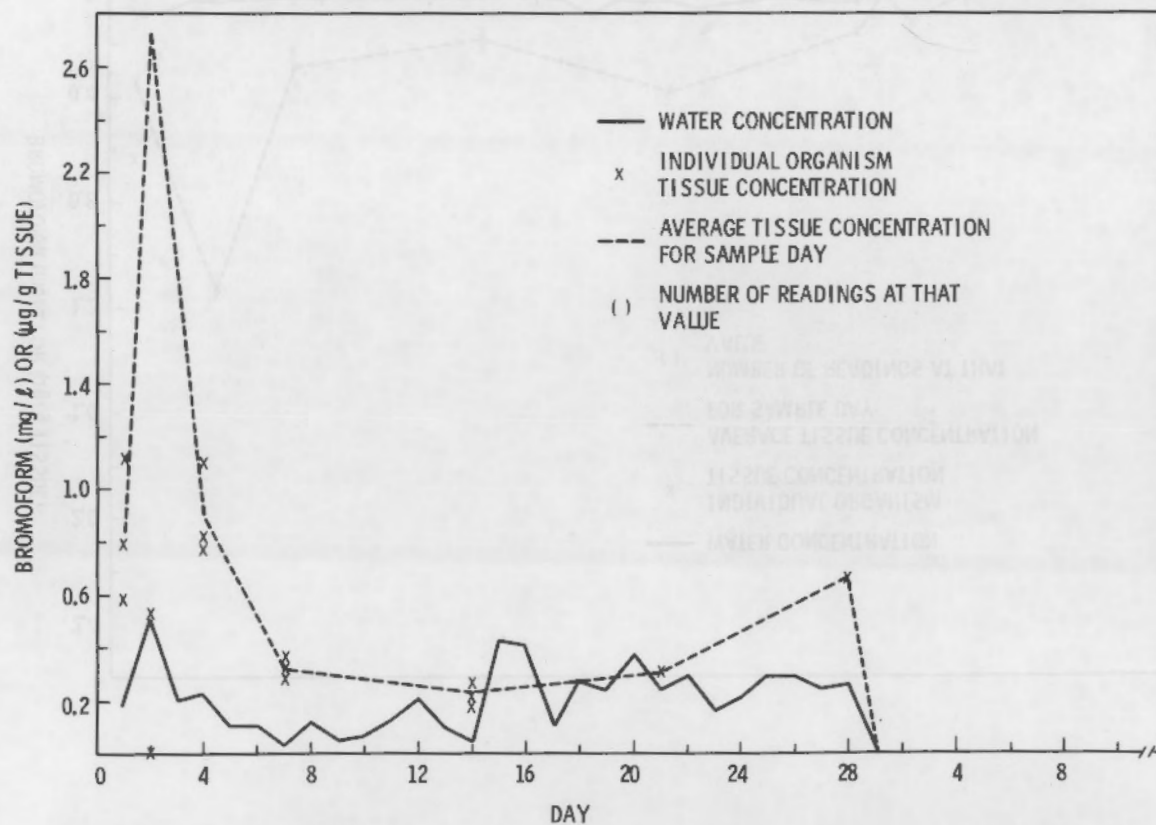


Figure 20. Water and tissue concentrations of bromoform (*Menhaden, Brevoortia tyrannus*) 28-day uptake/28-day depuration studies. Target water bromoform concentrations was 1.0 mg/l

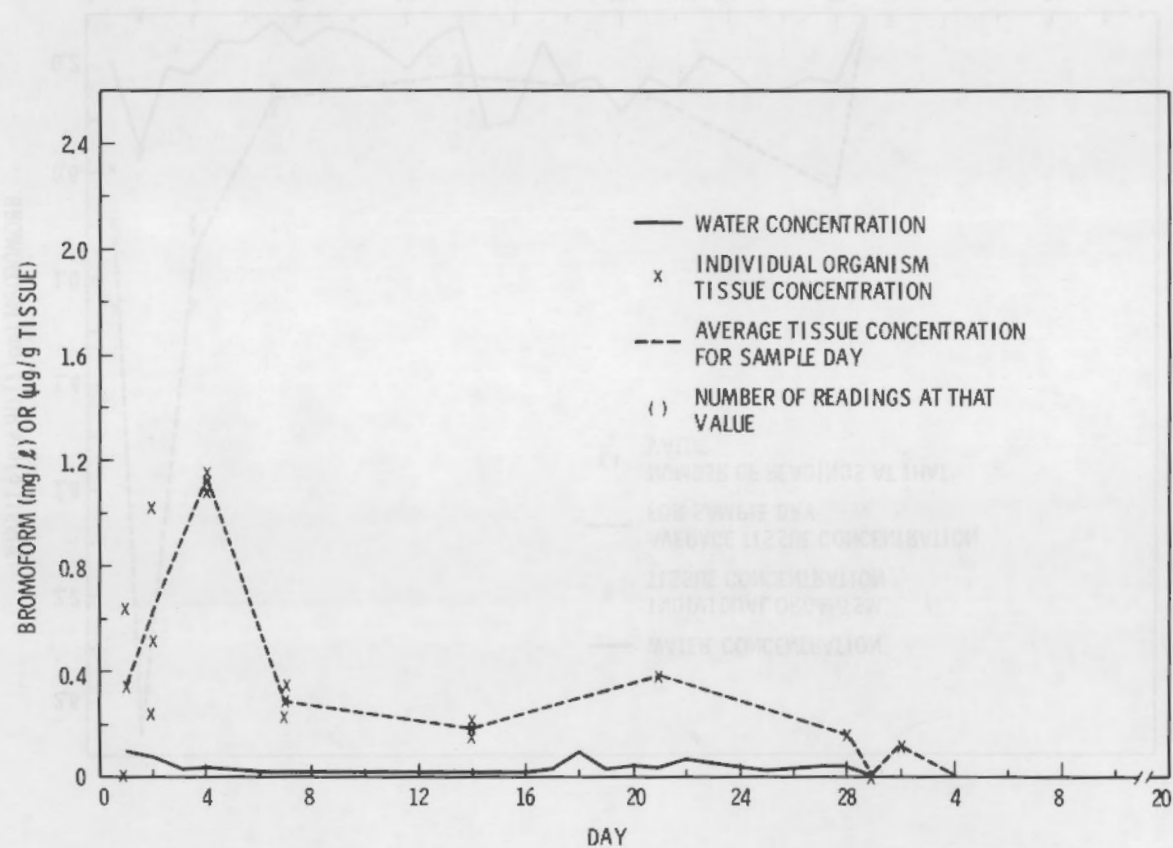


Figure 21. Water and tissue concentrations of bromoform (*Menhaden*, *Brevoortia tyrannus*) 28-day uptake/28-day depuration studies. Target water bromoform concentration was 0.1 mg/l

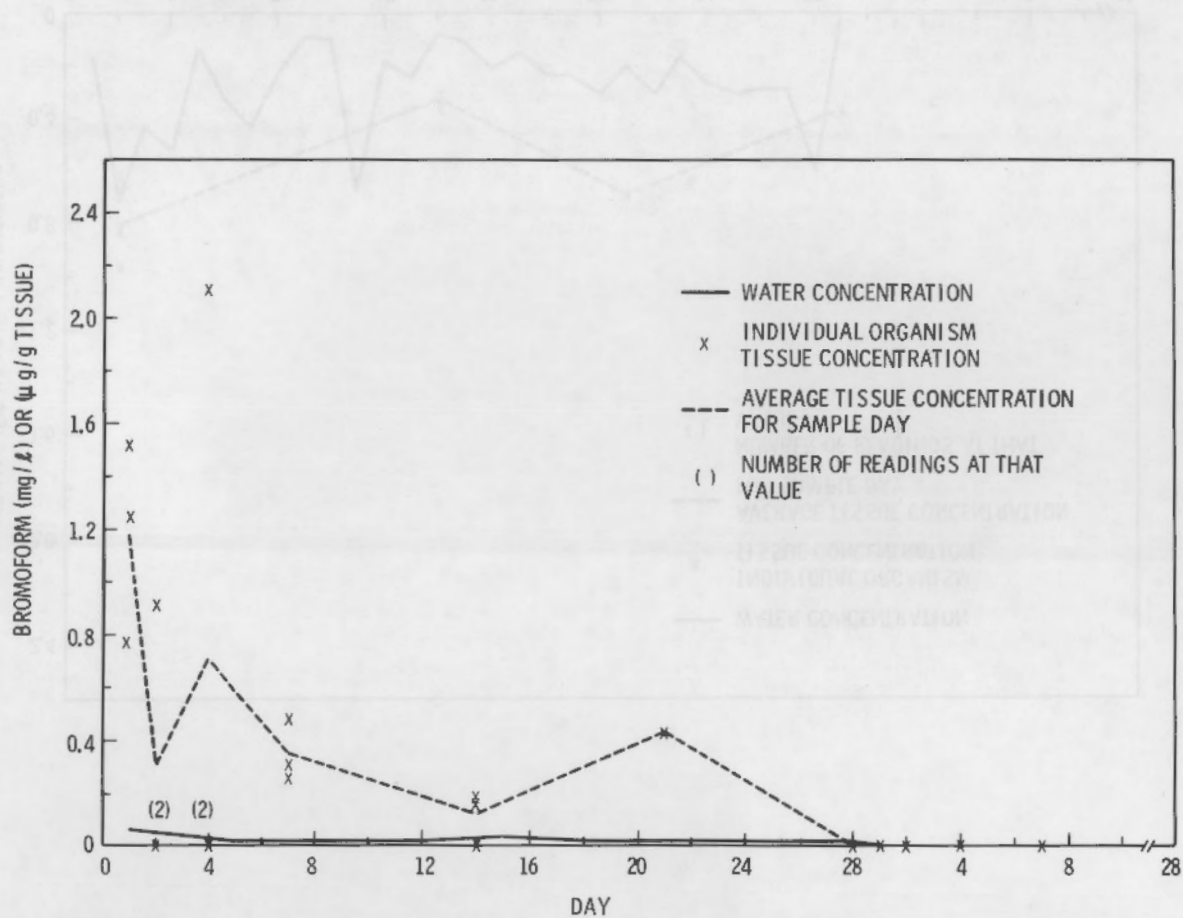


Figure 22. Water and tissue concentrations of bromoform (*Menhaden, Brevoortia tyrannus*) 28-day uptake/28-day depuration studies. Control

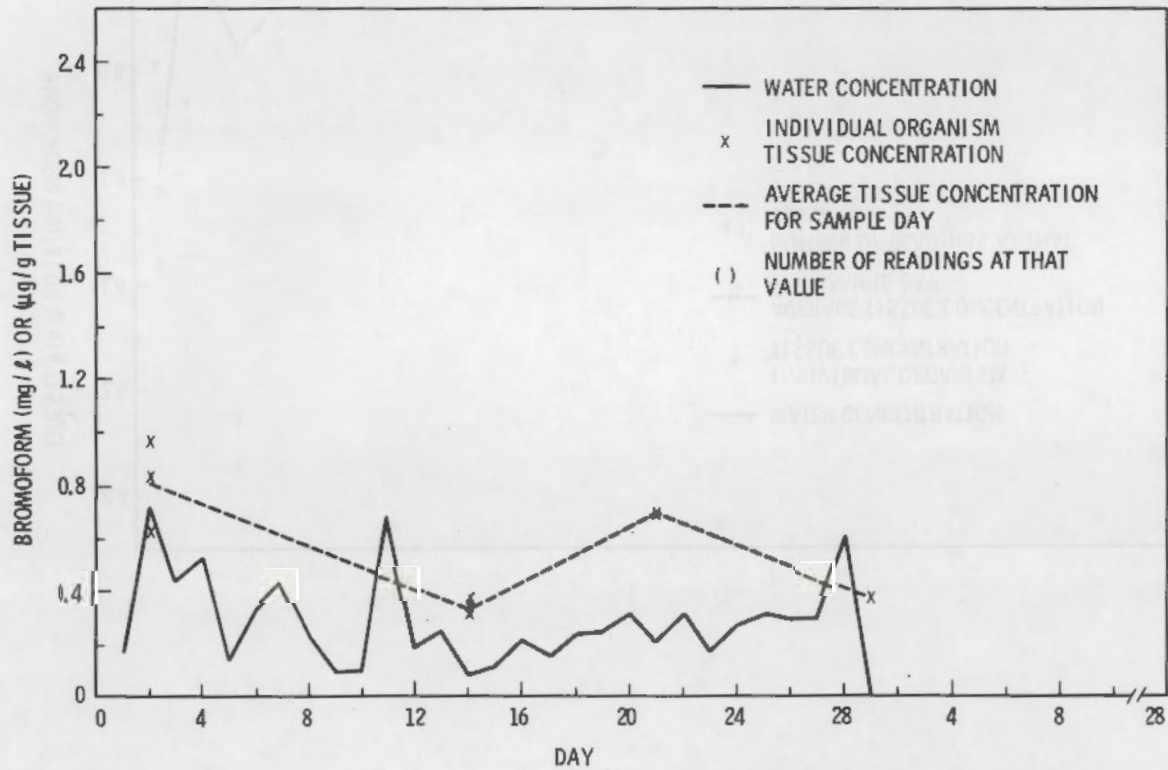


Figure 23. Water and tissue concentrations of bromoform (Shrimp, *Penaeus azectus*) 28-day uptake/28-day depuration studies. Target water bromoform concentration was 1.0 mg/l

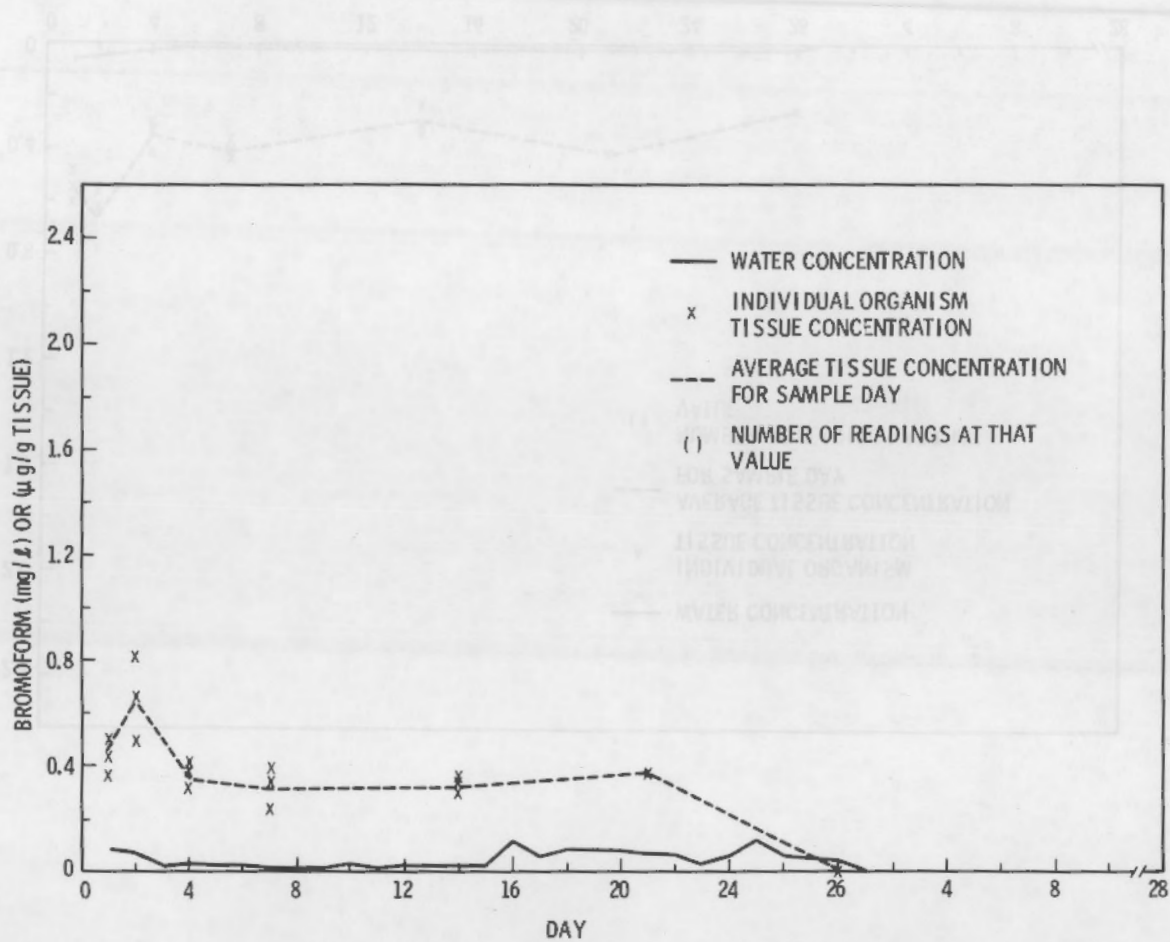


Figure 24. Water and tissue concentration of bromoform (Shrimp, *Penaeus aztecus*) 28-day uptake/28-day depuration studies. Target water bromoform concentration was 0.1 mg/l

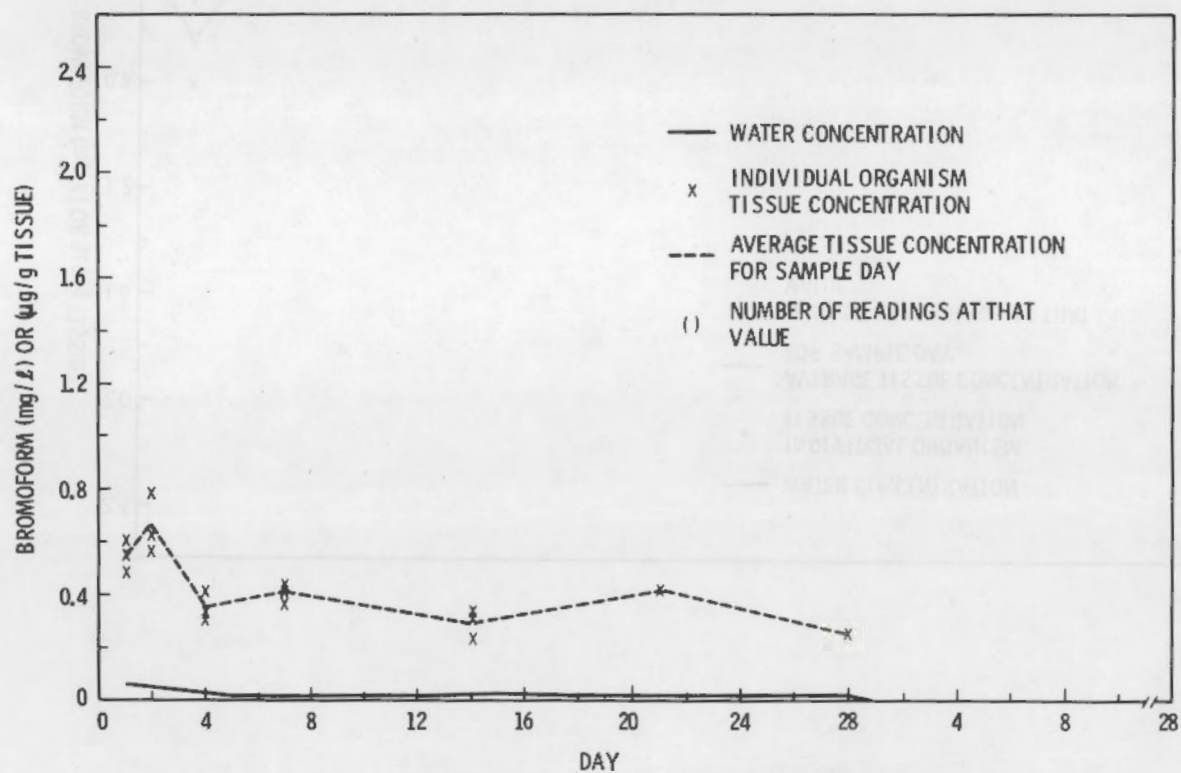


Figure 25. Water and tissue concentrations of bromoform (Shrimp, *Penaeus aztecus*) 28-day uptake/28-day depuration studies. Control

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