



**The BNL ASTD Field Lab – Near-Real-time
Characterization of BNL Stockpiled Soils to Accelerate
Completion of the EM Chemical Holes Project**

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April 2003

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**Environmental Research and Technology Division
Environmental Sciences Department**

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Executive Summary

As of October 2001, approximately 7,000 yd³ of stockpiled soil remained at Brookhaven National Laboratory (BNL) after the remediation of the BNL Chemical/Animal/Glass Pits disposal area. The soils were originally contaminated with radioactive materials and heavy metals, depending on what materials had been interred in the pits, and how the pits were excavated. During the 1997 removal action, the more hazardous/radioactive materials were segregated, along with, chemical liquids and solids, animal carcasses, intact gas cylinders, and a large quantity of metal and glass debris. Nearly all of these materials have been disposed of.

In order to ensure that all debris was removed and to characterize the large quantity of heterogeneous soil, BNL initiated an extended sorting, segregation, and characterization project directed at the remaining soil stockpiles. The project was co-funded by the Department of Energy Environmental Management Office (DOE EM) through the BNL Environmental Restoration program and through the DOE EM Office of Science and Technology Accelerated Site Technology Deployment (ASTD) program. The focus was to remove any non-conforming items, and to assure that mercury and radioactive contaminant levels were within acceptable limits for disposal as low-level radioactive waste. Soils with mercury concentrations above allowable levels would be separated for disposal as mixed waste.

Sorting and segregation were conducted simultaneously. Large stockpiles (ranging from 150 to 1,200 yd³) were subdivided into manageable 20 yd³ units after powered vibratory screening. The ½-inch screen removed almost all non-conforming items (plus some gravel). Non-conforming items were separated for further characterization. Soil that passed through the screen was also visually inspected before being moved to a 20 yd³ “subpile.” Eight samples from each subpile were collected after establishing a grid of four quadrants: north, east, south and west, and two layers: top and bottom. Field personnel collected eight 100-gram samples, plus quality assurance (QA) duplicates for chemical analysis, and a 1-liter jar of material for gamma spectroscopy. After analyses were completed and reviewed, the stockpiles were reconstructed for later disposal as discrete entities within a disposal site profile.

A field lab was set up in a trailer close to the stockpile site, equipped with instrumentation to test for mercury, RCRA metals, and gamma spectroscopy, and a tumbler for carrying out a modified Toxicity Characteristic Leaching Procedure (TCLP) protocol. Chemical analysis included X-ray fluorescence (XRF) to screen for high (>260 ppm) total mercury concentrations, and modified TCLP tests to verify that the soils were not RCRA hazardous. The modified TCLP tests were 1/10th scale, to minimize secondary (leachate) waste and maximize tumbler capacity and sampler throughput. TCLP leachate analysis was accomplished using a Milestone Direct Mercury Analyzer (DMA80). Gamma spectroscopy provided added assurance of previously measured Am-241, Cs-137, and Co-60 contamination levels.

The ASTD field laboratory completed more than 2,500 analyses of total Hg (XRF) and TCLP/DMA analyses over an 18-week period. Reliable statistical verification was accomplished for more than 98% of the stockpile subpiles. For most subpiles, TCLP analyses were completed within two days. One of the most significant aspects of the project success was schedule

acceleration. The original schedule projected activities extending from early April until September 30. Due to efficiency and reliability of the vibratory screening operation and cooperative, dry summer weather, stockpile reconstruction was completed in the third week of August. Reduction of the planned sample collection rate, from three samples per five cubic yards to two, resulted in further schedule acceleration. The resulting sample frequency, however, was still 22 times greater than the baseline frequency (1 per 55 yd³).

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1. Introduction

Approximately 11,500 cubic yards (yd³) of contaminated soil were excavated from Brookhaven National Laboratory's (BNL) former Animal/Chemical and Glass Holes (Chemical Holes) in 1997 as part of activities required under the Comprehensive Environmental Response, Compensation, and Liability Act (CERCLA, or Superfund). The Chemical Holes remedial action was initiated to remove laboratory glassware, chemicals, and related wastes from disposal pits used at BNL from 1958 through 1976. The soils were placed into separate stockpiles and characterized following an approved sampling plan (Procedure for Sampling Soil Stockpiles Chemical Holes Project, BNL, October 1997). As part of the removal action, the materials removed from the pits were segregated according to size and type. Larger items were collected manually, and the remaining materials were separated using a 2-inch screen. The remaining (less than 2-inch) material, consisting mostly of soil and gravel, but also including small bottles and vials potentially containing hazardous material, was collected and stored in 18 stockpiles at the removal project site. The stockpiles range in size from 100 yd³ to 1,800 yd³. In the ensuing years, some of the stockpile soils, which had been characterized as non-radioactive and non-hazardous, were disposed at Subtitle D facilities. During the removal action, the more hazardous/radioactive materials were segregated, along with a large quantity of metal and glass debris. Nearly all of these latter materials have been disposed of. The majority of the stockpiled soil had been identified as low-level radioactive waste (LLW). One stockpile (Stockpile 12, 700 yd³) was to be handled as mixed waste based on the presence of visible mercury reported by workers during the original removal action.

From September 1999 to January 2000, 29 railcars that included materials from Stockpiles 10 and 13 were shipped to Envirocare of Utah for LLW disposal after obtaining the appropriate approvals. Envirocare's routine sampling program (every 10th rail car) indicated that the Stockpile 10 soils exceeded RCRA criteria for allowable mercury levels of 0.2 mg/L in TCLP leachate. Subsequently, some of the Stockpile 10 soils were re-classified as mixed waste, treated (stabilized), and disposed of at Envirocare. This resulted in a non-conformance incident, and significant additional treatment/disposal costs (nearly \$450,000). Additional costs were identified at BNL before all the waste materials were shipped in this period. During loading of some of these wastes for disposal, "non-conforming items" were identified (e.g., vials, bottles, etc. less than 2 inches and potentially containing hazardous materials such as mercury) still entrained in the waste soil, requiring that about 380 yd³ of soil from Stockpile 13 be sorted a second time.

Evaluations of the root causes of the Envirocare non-conformance incident focused on the need for improved segregation of non-conforming items, as well as improved sampling and analysis protocols applied to the stockpiles for characterization. Thus, manual hand raking of non-conforming items from the remaining portion of Stockpile 13 was initiated. This was the baseline sorting technique, but it proved to be an extremely slow, labor-intensive, and expensive process. Following the occurrence investigation (ORPS No. CH-BH-BNL-BNL-2000-0002), the *Bulk Waste Determination Guidance Document* (also referred to here as the Toolbox [1]) was prepared, based on sampling protocols identified in EPA SW-846. This procedure represents the baseline characterization currently in place for the stockpiles prior to disposal. While intended to

provide a standard method for characterizing remaining stockpiles, it requires relatively few samples be taken for large volumes of contaminated soil. For example, application of this methodology to Stockpile 6B, which contained a total of 440 yd³ of soil, resulted in identification of only 8 soil samples for analysis. In the application of the procedure, an interactive spreadsheet (referred to as the Toolbox) is used to input characterization data. The Toolbox then gives an evaluation as to whether additional samples are needed for characterization to a 95% confidence level.

In order to continue disposal efforts and improve performance, BNL tested a powered ½-inch screen sorting method on Stockpile 6B, and conducted characterization sampling following the Toolbox method. Stockpile 6B was shipped and disposed of without incident in July 2001. As part of this effort, ten samples were collected from each of the other stockpiles, and analyzed for TCLP and total inorganic constituents. As of October 2001, approximately 7,000 yd³ of soil in 10 stockpiles still remained, requiring final disposition.

To prevent more non-conformance incidents (and potential regulatory problems) during subsequent soil disposal activities, BNL initiated an expanded sorting, segregation, and characterization project directed at the ten remaining soil stockpiles. The project was co-funded by the BNL Environmental Restoration program and the DOE EM Office of Science and Technology Accelerated Site Technology Deployment (ASTD) program. The focus was to remove non-conforming items and assure that mercury and radioactive contaminant levels were within acceptable levels for disposal as low-level radioactive waste. Extensive sampling was planned to provide a sound statistical basis for confidence in the measured contaminant levels. Soils with mercury concentrations above allowable levels would be separated for disposal as mixed waste.

The project involved the use of a power screen for sorting and segregation, and setting up and operating a field laboratory near the stockpile area that would provide rapid sample analyses for total mercury, other RCRA metals, and TCLP mercury. Samples were to be collected and carried to the field lab for analysis with a planned one-day turnaround. Cost-effective, timely analysis of soil contamination allowed many more samples be taken and analyzed, significantly improving confidence in the data.

The next two sections describe planning and preparation for the project, and the work completed. The final section compares the ASTD Field Lab to off-site and duplicate results, and discusses lessons learned.

2. Project Planning

The main project goals were:

1. The complete removal of all non-conforming items in a safe manner from the soil stockpiles, and
2. The analysis of each stockpile for total mercury and TCLP mercury, to demonstrate in a statistically reliable manner that the soil is non-hazardous on average.

2.1 Stockpile Sorting and Sampling

The first goal involved the use of the power screening method applied to Stockpile 6B, as mentioned earlier. During the screening, soils were to be separated into 20 yd³ subpiles for sampling and subsequent analysis. A field laboratory, set up in a trailer near the stockpile area, was to be the means of achieving the second goal. The field lab would have the capability to test soils directly for total mercury, and would conduct modified TLCP extractions and test the extract for mercury. Finally, an ISOCS unit would collect gamma spectra for radiological characterization.

BNL's EM Directorate completed the plans for the sorting and segregation activities. A health and safety plan [2], a sampling and analysis plan [3], and a technical work document [4] were prepared, following BNL EM standard procedures. The Environmental Research and Technology Division (ERTD) was responsible for the field laboratory and, prepared similar plans for its operation [5, 6].

Based on the earlier work with Stockpile 6B, stockpile sorting rates were estimated at a maximum of 100 yd³ per day, or five 20 yd³ subpiles per day. Sampling and analysis plans were designed around this sorting volume estimate and the assumption that 3 samples were to be collected for every 5 yd³. Three samples per 5 yd³ were considered more than adequate for statistical certainty, particularly in comparison to the 1 sample per 55 yd³ determined to be acceptable for Stockpile 6B with the Toolbox.

Crumbling, et al [7] reported that uncertainty in environmental characterization is often a trade-off involving the number of samples and sampling methods, field screening analysis methods, and certified analytical laboratory methods. While precision in analytical methods has been steadily increasing with improved technology, accuracy in characterization is much more dependent on how well the sample reflects the actual condition of the waste. Uncertainty in the data therefore is much more closely tied to the extent of sampling. In many environmental remediation characterization efforts, sampling uncertainty offsets analytical laboratory reliability. Explicitly: "If representativeness cannot be established, the quality of the chemical analysis is irrelevant." [7] Data quality in support of remediation decisions can be improved considerably with increased sampling and field screening methods to supplement certified laboratory results.

Soil stockpiles were located in two areas, shown as green rectangles on the map in Figure 1. For eight of the stockpiles, the power-screen was set up at the former Glass Holes Area between Stockpiles 12 and 8. (The power-screen is shown in Figure 2.) As sorting progressed and soil from each stockpile was placed as 20 yd³ subpiles with a front-loader, samples were collected from the subpiles. The subpiles were placed in the area to the southeast of the power-screen, south of Stockpile 15, west of and adjacent to the woods. This served as the staging area for stockpile reconstruction in the Animal/Chemical Pits Area (also shown in Figure 1 as red-outlined rectangles). Subpiles were placed on 3-mil polyethylene sheets, and, after sampling, covered with polyethylene as well, until reconstruction was approved. Once the eight stockpiles in the Glass Holes area were sorted and reconstructed, the power-screen was moved to the Animal Chemical pits area for sorting the last two stockpiles, 6C and 6R.

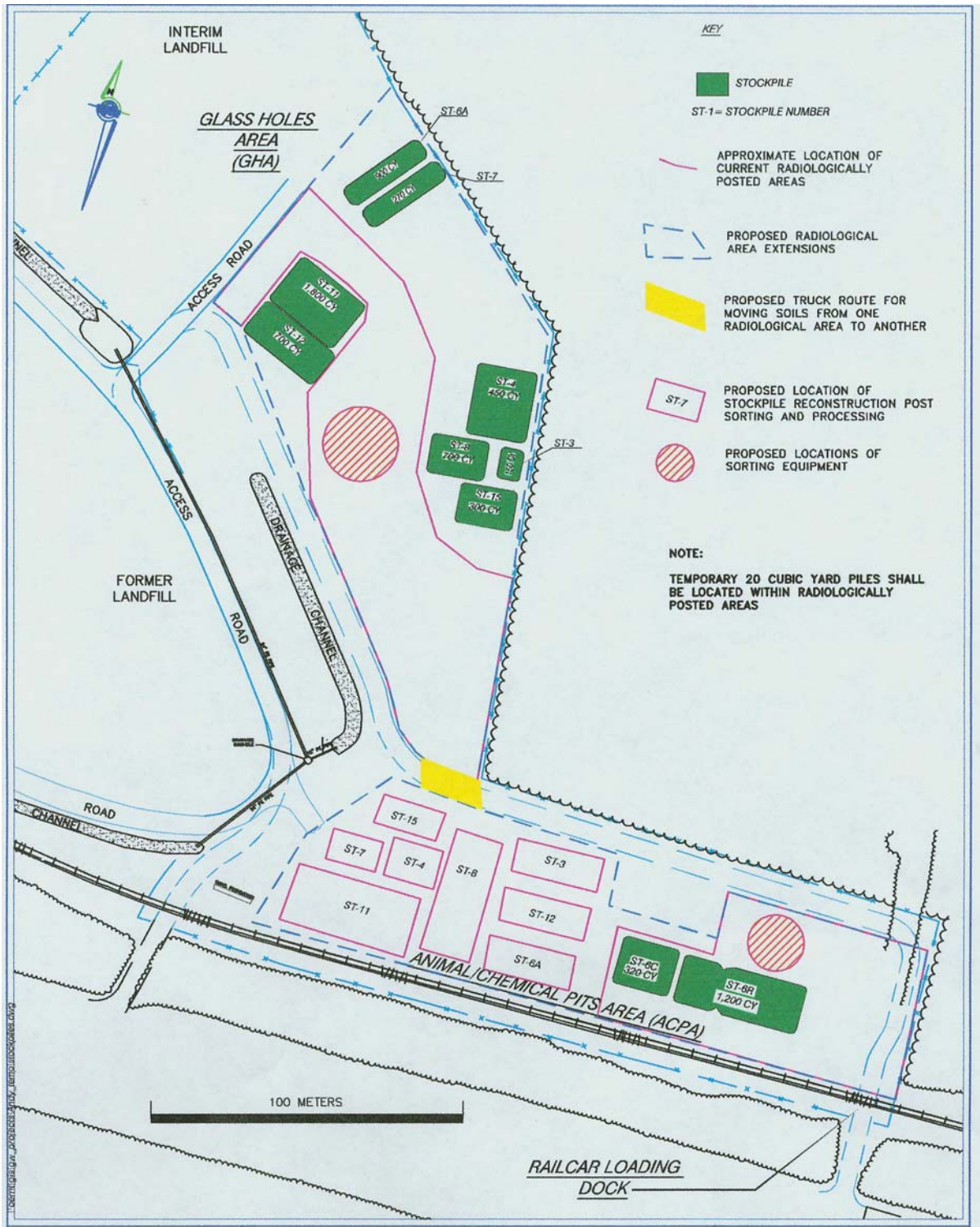


Figure 1. Stockpiles, sorting areas, and stockpile reconstruction areas at the former chemical/animal/glass disposal pits



Figure 2. The half-inch vibratory screen unit (Powerscreen Chieftain 600)

2.2 ASTD Field Laboratory

Samples from the subpiles were transported to the ASTD Field Laboratory for analysis each day they were collected. As noted earlier, 60 samples per day were expected, based on Stockpile 6B sorting experience. Sample analysis procedures were based on this throughput estimate, and on testing each sample for total mercury and most (but not all) samples using a modified EPA Toxicity Characteristic Leaching Procedure (TCLP). Originally, the testing scheme was to follow that shown in Figure 3. X-ray Fluorescence (XRF) was to be used as a screening tool to measure total mercury in the soil sample, since this mercury detection limit was identified by the manufacturer as achievable in solids with an analysis time of between 5 and 10 minutes. If mercury concentration for a sample was above 10 ppm, total mercury measurements were to be verified using the PDV 5000, and the sample was to be tested using the modified TCLP. If the XRF indicated that the sample was below 10 ppm mercury, the total mercury was to be verified with the Milestone Direct Mercury Analyzer (DMA) or the PDV 5000. From the latter test, if total mercury was less than 4 ppm, no further testing was deemed necessary, since a concentration of < 4 ppm total mercury could not result in TCLP concentrations in excess of the TCLP allowable limit of 0.2 ppm, assuming all of the mercury was leached. If more than 4ppm was detected, TCLP was still required.

TCLP failure means that the results for a particular subpile would have to be reviewed and averaged, to determine if the subpile or a portion of it had to be segregated for ultimate hazardous or mixed waste disposal. The Toolbox [Ref. 1] provides algorithms for determining the acceptability of the averages and confidence levels for subpiles and whole stockpiles.

In addition to the chemical analyses, the ASTD Field Lab also collected radiological content information using the *In Situ* Object Counting System (ISOCS) to determine the presence of gamma-emitting radionuclides.

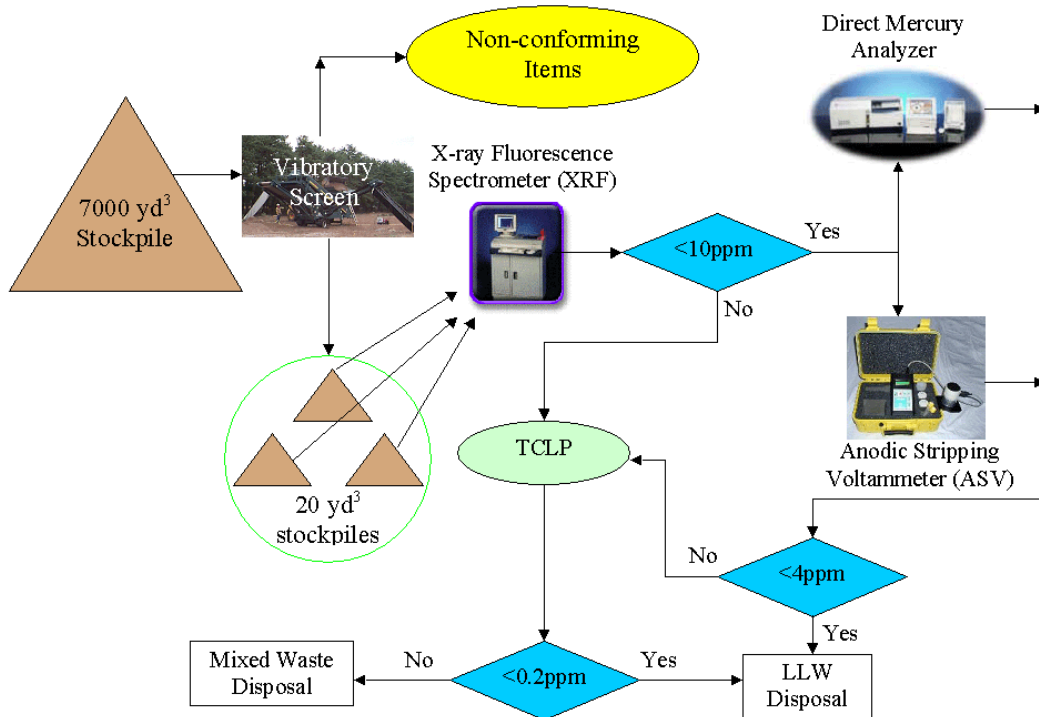


Figure 3. Planned process diagram for innovative waste segregation and near-real-time characterization of stockpiled soils

2.3 Field Laboratory Analytical Methods

2.3.1 X-ray Fluorescence

X-ray Fluorescence (XRF) is a mature technology that has been used for decades for elemental analysis in research laboratories and industrial process monitoring. Detection limits to between 10 and 100 ppm are easily achieved for most elements. With higher strength X-ray sources and secondary targets, sensitivities as low as 1.0 ppm or less are also possible. The Model EX-6600A field-deployable XRF unit was purchased for this project from Jordan Valley; the reported detection limit for mercury was 1.5 ppm.

An important advantage of the XRF method is that sample preparation is minimal, and, for mercury, the RCRA metal of concern in this project, detection can be achieved in air or a helium-flushed system, rather than vacuum. XRF is a non-destructive method that can be applied to solid, powdered, or liquid samples. Generally, no secondary wastes are generated as a result of sample preparation. Sample analysis time, including preparation, was expected to be approximately 10 minutes. A 10-position automatic sample-changer was included in the purchased unit.

2.3.2 Anodic Stripping Voltammetry

The PDV 5000 was chosen for the project because it promised to be an inexpensive, simple, rapid technique for analyzing mercury and other toxic metals. It is a commercially available version of a method that in the past had been generally relegated to research laboratories. The method involves anodic stripping voltammetry (ASV), in which metals in solution are first electroplated onto an electrode. After this, current is reversed to strip the metals from the electrode back into solution. Each metal becomes stripped at a characteristic voltage, and the current density at that voltage is proportional to the total quantity of metal.

Because ASV involves the analysis of liquids, applications have usually focused on process liquid monitoring. Environmental applications have included river water testing for mercury and other metals. For soil testing, metals can be detected and quantified electrochemically after acid digestion with an acidic electrolyte. With the metal constituents in solution, electroplating followed by anodic stripping becomes possible. The detection limit for mercury in field applications was stated as 0.01 ppm in the liquid, which, according to the published procedure, corresponds to 0.02 ppm in the soil matrix.

2.3.3 Direct Mercury Analyzer

The Milestone DMA-80 measures low concentrations of mercury in environmental samples, in accordance with EPA Method 7473, "Mercury In Solids And Solutions By Thermal Decomposition, Amalgamation, and Atomic Absorption Spectrophotometry." Its reproducibility, low detection limits, rapid throughput, and the fact that it does not generate any secondary waste, makes it ideally suited for environmental applications.

The method involves weighing the sample and placing it directly in a small "boat," for drying and thermal treatment in a stream of heated oxygen gas. The gas stream is passed through a gold amalgamation trap that captures all mercury in the vapor phase. The gold amalgam is subsequently heated and mercury vapor detected with an atomic absorption spectrophotometer tuned to the absorption wavelength for mercury, 254 nm. The Milestone DMA detection limit is 0.11 nanograms mercury. For the listed maximum sample size of 0.5 grams, therefore, the theoretical detection limit in terms of concentration is 0.00022 ppm (0.22 ppb).

2.3.4 Modified TCLP

The modified Toxicity Characteristic Leaching Procedure (TCLP) used was essentially a 1/10th scale version of the test recommended by the U.S. Environmental Protection Agency (EPA) [Ref. 8]. Extraction fluid #1 was used for all tests, as determined by the related EPA procedure [Ref. 9]. Soil samples of 10 gram size (rather than 100 g) were weighed out in 250 mL plastic bottles, and 200 mL (rather than 2L) of fluid #1 was added to the soil. Five leach samples in small bottles were placed inside a plastic 1-gallon jar to serve as secondary containment, which was then placed into a compartment of the tumbler apparatus. A TCLP tumbler purchased from Miller Analytical, with 12 compartments, allowed for 60 samples to be tested simultaneously. The samples were tumbled for 18 hours, per the TCLP procedure. The modified small-scale

procedure meant that more samples could be tumbled concurrently, and less waste was produced. Results were reported as parts per million (ppm) or parts per billion (ppb) TCLP mercury, representing mercury concentration in the leach solution, not the solid.

2.3.5 *In Situ* Object Counting System

The Canberra *In Situ* Object Counting System (ISOCS) consists of a portable germanium detector controlled by proprietary software for detector calibration and evaluation of specific activity. Standard sample configurations with shielding may be used, or large areas or equipment may be surveyed with the detector, provided models are available for geometric data interpretation. ISOCS has been used at BNL in two earlier ASTD projects. For the Chemical Holes Field Lab, a standard sample configuration was used for verifying gamma-emitting radionuclide contamination in the stockpiles. One sample per 20 yd³ subpile was counted.

2.3.6 Integrated Operations and Quality Assurance

Since the stockpiles had been classified as low-level waste, the field lab was set up as a radioactive materials area for sample storage. Sample handling, namely opening bottles, weighing out soils, and preparing TCLP bottles for tumbling, was carried out in the west half of the trailer. Because of the potential for loose soil releases in this half of the lab, it was designated as a radioactive dispersibles area when transfer operations were being conducted. The XRF, PDV5000, and DMA-80 were also located in this section. The east end of the lab trailer was used for receiving samples, ISOCS counting, and running the TCLP tumbling apparatus. A schematic layout of the Lab is shown in Figure 4.

Sample receipt involved signing Chain of Custody (COC) forms after verifying that all bottles were labeled clearly and that the COC information was correct. ISOCS samples were to be stored and counted in the east end of the trailer. The remainder of the samples was to be transferred for processing and analysis by the chemical techniques described above.

After a targeted one-day turnaround, analytical results were compiled and transferred to the Project Manager for review. Unused soils and liquid extracts were returned to the stockpiles for final disposal. When all data for a stockpile had been reviewed, the Project Manager approved stockpile reconstruction. The criteria for reconstruction were that the soils contained less than 260 ppm total mercury and less than 0.2 ppm TCLP mercury.

Quality assurance focused on maintaining proper chain of custody protocols and taking a subset of field duplicates: one for analysis in the ASTD Field Lab, and one for analysis at an independent off-site laboratory. In addition, the analytical instruments were to be calibrated according to manufacturer's recommended guidelines.

ASTD Field Laboratory

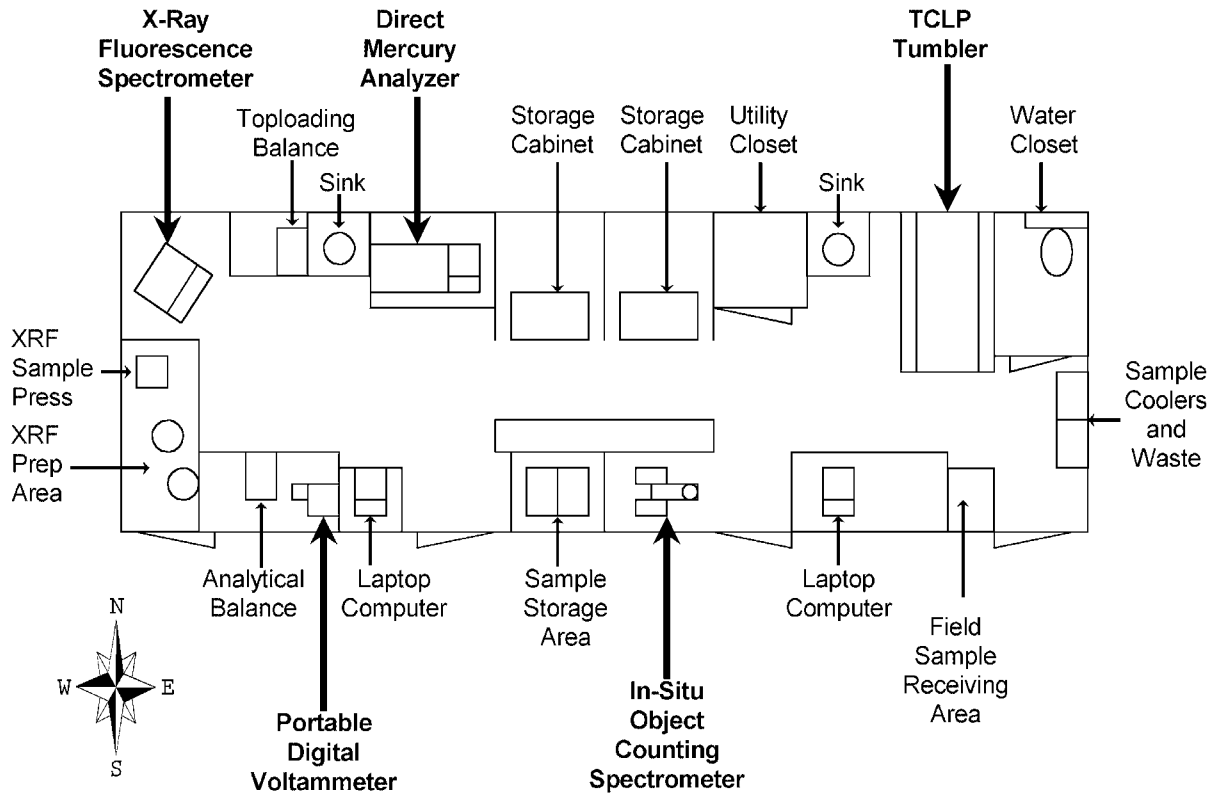


Figure 4. ASTD Field Lab - Schematic layout

3. Work Completed

3.1 Sample Production Rates and Chemical Analysis Capacity

Sorting began the first week of May, 2002, on Stockpile 8. After about 2 weeks, the soil sorting team had developed a standard routine for their activities, and they were processing more than 100 yd³ per day. In fact, on their best days they processed as much as 230 yd³, or more than twice the originally projected capacity. This resulted in a significantly higher production of analytical samples than was anticipated (60 samples per day).

In addition to receiving more samples than planned for, ASTD Field Lab operations required modification due to poor performance of the PDV 5000 (primarily unsatisfactory reproducibility) which eventually resulted in eliminating this method from the suite of analyses (see discussion below). Consequently a revised analytical strategy was implemented to improve productivity and provide the critical data needed for decision-making on the status of individual soil stockpiles. These three essential criteria were:

- confirmation that total mercury levels were < 260 ppm
- evaluation of TCLP mercury levels to ensure they did not exceed 0.2 ppm
- radiological analysis

Specifically, the XRF was used to screen for >260 ppm total mercury and the DMA-80 was assigned to measure mercury exclusively in the TCLP leachates.

The DMA-80 was assigned for TCLP results because it could be used for liquid as well as solid samples and gave highly reproducible results, as evidenced by blank and calibration standards tested after every five samples. With the DMA-80, preparation of the TCLP leach liquids for analysis was minimal, namely filtering the solution (0.45 μ) and pipetting 0.4 mL. Because of its reproducibility, the DMA-80 was deemed most important for application to the regulatory compliance test.

The XRF was assigned to total mercury analysis, but the procedure was modified to shorten preparation and analysis time. The Jordan Valley EX-6600 had been chosen because the company demonstrated detection limits of less than 10 ppm total mercury on samples provided by BNL. However, this was accomplished after milling/grinding the samples for ten minutes, and pressing the fine powder into pellets with a hydraulic press. With these steps, sample preparation time was about twenty minutes per sample, far too long for the throughput needed for the field lab application. In addition, the counting time on the x-ray unit was about fifteen minutes per sample. A shorter sample preparation time was developed which required sieving the soil to less than 2 mm and compacting it in a disposable plastic cup with a transparent Mylar film bottom. This and a 5-minute analysis time meant that the detection limit for the samples was approximately 20 ppm total mercury. The higher detection limit meant that the XRF unit was best used as a screening tool for higher levels of mercury, i.e. more than 260 ppm. Soils with this high a concentration require treatment before disposal under EPA land disposal restrictions. After Stockpile 15 was completed, it was determined that a shorter analysis time of 3 minutes, with a detection limit of 50 ppm was necessary to analyze 60 samples per day.

The PDV 5000 was dropped from use because reproducibility of calibration standards with it was often less than 20%. Discussions with the instrument manufacturer failed to resolve the issue completely, primarily because there was insufficient man-power to work on the recommended method modifications. Anodic stripping voltammetry depends on obtaining reproducibly dissolved mercury analyte in an interference-free electrolyte solution. It was not clear from the limited tests conducted whether:

1. Interfering elements were present in the soils and were being dissolved along with mercury,
2. The acid solution used was not dissolving mercury reproducibly from the soils, or
3. The active carbon electrode on the instrument was becoming fouled.

The revised laboratory procedure process flow diagram is shown in Figure 5. Several weeks were required to make all adjustments, and several important decisions about the data requirements were needed to make data production match sample production. The most important decision was to reduce the sample collection rate from three for every five cubic yards

to two. The resulting sample frequency (1 per 2.5 yd³) was still 22 times greater than the baseline frequency (1 per 55 yd³).

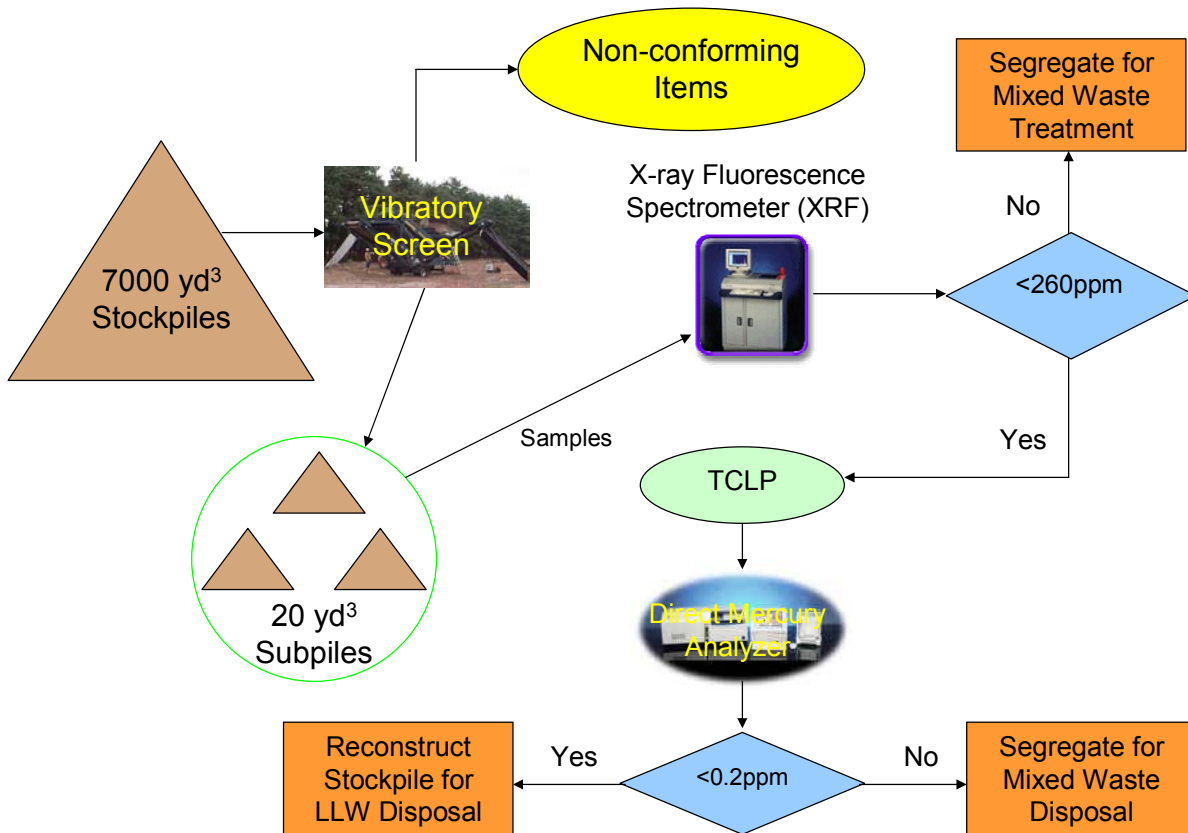


Figure 5. Final sorting and analysis scheme for ASTD Field Lab

3.2 Field Laboratory Results

Table 1 compares original estimated stockpile volumes to those determined during sorting operations. Stockpile sorting included moving sorted soils from the power-screen to the subpile staging area with a 1-yd³ front-loader. Thus, the soil stockpile volumes were measured more accurately than when they were accumulated initially, because the subpiles were built to a specified number of front-loader loads. Overall, the total screened soil volume, 5,660 yd³, was smaller than the original estimate of 6,790 yd³. Final disposal volume may also be slightly smaller, depending on how much compaction occurs during loading and transport in railcars. Some of the volume reduction may be associated with the separation of old plastic cover material, which had been covered with new layers, as weather-induced tears required the replacement of the first covers. All cover materials were segregated for disposal as debris.

Soil sorting activities produced 283 subpiles, which were then re-assembled into ten new stockpiles, corresponding to the original ten. Radioactive measurements with the ISOCS unit and chemical test results are discussed separately.

3.2.1 ISOCS Results

The number of samples collected for ISOCS analysis totaled 283, or 1 for each subpile. Results are listed in full in Appendix 1, and summarized by stockpile in Table 2. The results include average values, maximum values, and the total number of non-detected measurements (NDs) for each stockpile. These quantities show that the stockpiles are slightly contaminated, overall.

Previous characterization data indicated that radioactive contamination was low and restricted to a few radionuclides, primarily cesium-137 (Cs-137) and americium-241 (Am-241). These general results are verified with the additional ISOCS data. Other isotopes detected include cobalt-60 (Co-60), radium-226 (Ra-226), thorium-232 (Th-232), and uranium-235 (U-235). Of all these isotopes, Ra-226 and Th-232 are potentially naturally occurring at low levels because of Long Island's geology [10]. All are associated with some aspect of nuclear power or research, e.g., fuel (U-235, Th-232), fuel by-products and waste (Co-60, Cs-137) or isotope application studies (Co-60, Cs-137, Am-241, Ra-226, Th-232).

Am-241 had the highest levels of contamination, with 4 readings greater than 10 pCi/g. in 4 separate stockpiles. It is worth noting that in 3 of these stockpiles (St-6A, St-11, and St-12) more than half of the subpiles had Am-241 non-detects, and in the fourth stockpile, 11 out of 34 subpiles were non-detected for Am-241. It is also significant that the average Am-241 values for the stockpiles were all less than 5 pCi/g, and that Am-241 was the only radionuclide that exhibited a concentration in excess of 1.0 pCi/g.

Cs-137 has been identified as a radionuclide of concern at BNL, and cleanup levels for it are set at 23 pCi/g. The maximum concentration of Cs-137 found was 2.3 pCi/g in a subpile from Stockpile 11. The highest average concentration for a stockpile was 0.8 pCi/g, for St-7.

3.2.2 Total Mercury and TCLP Mercury Results

The total number of chemical analyses performed in this project was in excess of 2,264 (8 times 283). During sorting and sampling of the first stockpile, St-8, it was immediately obvious that stockpile soils were being screened and sampled at a rate that far exceeded the field lab's projected daily capacity. To keep up with sample production, the center level sample from the three-level sampling grid was ignored. Thus only four samples from the top and four from the bottom level quadrants were actually analyzed. Further, because it was being used as a higher-concentration screening tool, XRF analysis was performed on 132 fewer samples (for Stockpiles 3, 4, and 8). (These can be identified in Appendix 2, which lists all total mercury and TCLP mercury results side-by-side.) After Stockpile 8 was completed, the sampling plan was modified so that 8 rather than 12 samples were collected for every 20 yd³ (or 2 for every 5 cubic yards). Additionally, a sample for offsite analysis and a field duplicate (FD) were collected for every 20th sample. For Stockpile 8, 28 of 132 "center level" samples collected were analyzed for

TCLP (subpiles 1 to 3 and 20 to 23, inclusive). Thus, TCLP analyses in the ASTD Field Lab totaled 2,406 (2,264 + 28 + 114 FDs). There were a total of 2,246 XRF analyses.

Table 3 lists average total and TCLP mercury concentrations for each stockpile. Stockpile 7 had the highest total mercury concentration of 65.1 (\pm 2.8) mg/kg, well below the EPA action level of 260 mg/kg for which treatment is required. Stockpile 6C showed the highest statistical deviation of about 20% (or 57.5 \pm 12.0 mg/kg), and Stockpile 3 was uniformly at the detection limit of 18.8 mg/kg total mercury. (Recall that later measurements with a shorter run time meant that the minimum detection level for stockpiles tested after St-15 was 50.5 mg/kg). The highest subpile average for total mercury was 86.6 mg/kg in St-6C. Stockpile 6C also had the highest single sample total mercury value at 174.0 mg/kg. In all, there were 16 single samples above 100 ppm, and these were distributed over 5 stockpiles. Stockpiles 8, 3, 4, 6A, and 7 had no samples with total mercury concentration greater than 100 ppm.

More importantly, in terms of classifying the soils as hazardous or mixed wastes, stockpile averages for TCLP mercury are all well below 200 μ g/L, the hazardous waste definition for EPA’s Toxicity Characteristic. As can be seen in Table 3, the stockpile TCLP mercury averages are well below the BNL administrative action level of 160 μ g/L. Stockpile 12 had the highest levels of TCLP mercury, at 73.0 (\pm 29.6) μ g/L. This is consistent with Stockpile 12 having the highest subpile average TCLP of 153.7 μ g/L, and 5 subpiles with average TCLP mercury above 100 μ g/L (see Table 4). As can be seen in Table 4, only two other subpiles, one in Stockpile 11 and one in Stockpile 6R, had an average TCLP mercury level above 100 μ g/L.

There were 14 individual samples with TCLP mercury above the administrative action level of 160 μ g/L, as listed below, from Stockpiles 11, 12, and 6R. When TCLP leachate samples above the 160 μ g/L level were found, the original leachates were re-analyzed in the DMA to verify the levels. Those shown below were the higher of the two test results. Stockpiles 3, 4, 6A, 6C, 7, 8, and 15 had no single samples above the BNL administrative limit.

Sample	TCLP Hg	Sample	TCLP Hg	Sample	TCLP Hg
6R29SWT	338.7	1204SET	336.0	1127SET	210.9
6R33SWT	661.7	1217NEB	238.5	131NWT	206.4
6R47FD	267.9	1217NWT	612.8	1143SWB	595.8
6R48NWT	203.5	1220SWT	285.6	1150NEB	161.5
6R51SWT	893.1	1229SET	183.4		

4. Discussion and Conclusions

The ASTD Field Laboratory results presented above indicate that the soil stockpiles on average can be classified as non-hazardous, because total mercury is less than 260 ppm, and TCLP mercury is less than 200 μ g/L. However, waste disposal facilities require that a certified laboratory must provide characterization data for waste disposal purposes. Thus, the QA program for the ASTD Field Laboratory included sending data to a certified off-site laboratory.

In this way, the ASTD Field Lab data serves as broad statistical support for higher confidence levels in limited sampling and analysis at an off-site lab.

4.1 Comparison with Off-site Results

The BNL EM Directorate Quality Assurance plan generally requires that one field duplicate and/or blind duplicate sample be sent to a contract lab for every twenty samples, or five percent. For the Chemical Holes Sorting And Analysis Project, a field duplicate was collected every 20 samples for analysis in the ASTD Field Lab, and a duplicate sent to an off-site laboratory for independent total mercury and TCLP mercury analysis. A complete listing of all off-site results tabulated side-by-side with the field lab results is contained in Appendix 3.

There are several approaches for inter-laboratory data comparisons. For the purposes of this discussion, two are used: one compares individual sample results with the off-site duplicates to determine if there is a correlation between the two sets of results. The second involves comparisons of subpile and stockpile averages obtained from the two data sets, to see if similar characteristics apply (i.e. do the data sets agree that the soils are not hazardous).

Figure 6 displays an X-Y plot of the ASTD Field Lab results *versus* results from the off-site contract laboratory. The slope of a best-fit line calculated by least squares regression is a quantitative means of demonstrating how well the two data sets' values match each other. Slope values close to unity (1.00) indicate that the data values match closely. Correlation coefficients can also be calculated mathematically; values close to one are indicative of a close correlation (lower scatter) between data sets. Both calculations are straightforward and accessible as functions through Microsoft Excel. The values obtained are shown below.

	Slope	Correlation Coefficient
TCLP Hg	0.76	0.58
Total Hg	0.82	0.46

The scatter shown by the data in Figure 6 is not unexpected, and a visual inspection of the plot along the slope = 1.0 line suggests that the data are distributed roughly evenly about the slope. However, the slope for the TCLP mercury is only 0.76, and the correlation coefficient is low at 0.58. A slightly "better" slope of 0.82 is found for total mercury, but scatter is worse, with a correlation coefficient of 0.46. In both instances, a slope of less than 1.0 indicates that the ASTD Field lab measurements were higher for any given sample. This in turn suggests that, if one assumes that the "true" value is that reported by the off-site contract laboratory, the ASTD field lab erred on the conservative side by providing mercury measurements slightly higher than the "true" value.

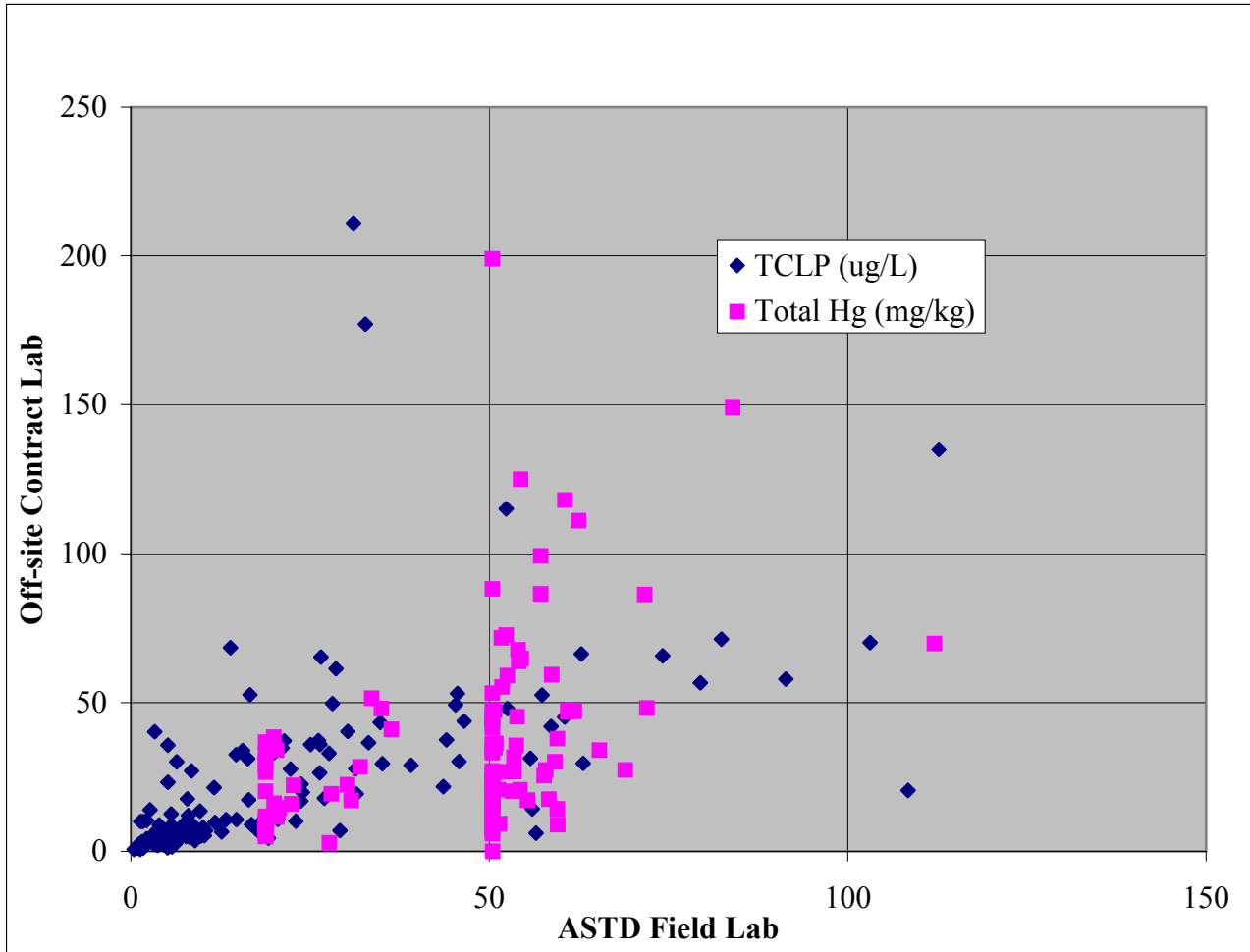


Figure 6. Field lab and contract lab data comparison

In spite of the observed scatter and the deviations of the slopes from 1.0, the agreement between the data sets is reasonable because of the type of contamination involved and the deposition method. As noted in the beginning of this report, the mercury contamination is believed to be primarily in the elemental, liquid form. It was originally placed in the disposal pits in whole containers (glass or plastic). The containers may have been broken at this stage or during the excavation phase in 1997. As a result of breakage, liquid mercury would physically disperse through the soil pores, driven by gravity, until the droplets became small enough the capillary forces would hold them up. Further dispersion would be expected during excavation or subsequent sorting and handling. In either event, the distribution of the mercury would not be uniform unless some sort of homogenization was conducted with the soils. A significant degree of heterogeneity can therefore be expected, as evidenced in the scatter seen in Figure 6, and can be expected from the proposed deposition model. The real question is – what valid average values can be used as data to describe the materials and be acceptable to disposal sites and regulators? Toward this end, independent sets of measurements can provide assurance that average values (as opposed to individual value comparisons displayed in Figure 6) and the error associated with those values are “true.”

Data comparisons for TCLP mercury are listed in Tables 5 and 6, and shown as a bar graph in Figure 7. Table 5 compares stockpile averages calculated using the off-site samples and the corresponding ASTD Field lab samples. These common data point averages are less representative than those shown for the Field Lab in Table 6 because the number of points for the overall average is lower, e.g. for St 3, 6 points rather than 42, or for St 11, 25 rather than 545 points. St 8 is an anomaly, because 103 off-site data points represents the analysis of many additional samples collected under the initial sample rate of 12 per 20 yd³. When the ASTD Field Lab was unable to complete all analyses at the pace that samples were being produced, many were sent off-site. Thus for St 8 there were only 41 common data points, as shown in Table 5.

St 7 was also a special case. The Contract Lab reported one sample (ST-7-06-NWB) as having TCLP mercury at 367 µg/L. A repeat analysis of the remaining sample at the contract lab (soil extraction then analysis of the leachate) was reported as being 5.7 µg/L. The Field Lab had found 3.5 µg/L TCLP mercury in leachate from the same soil sample. The overall average and standard deviation for St 7 (79.3 ± 161.0 µg/L) based on off-site data and including the 367 µg/L were the highest found to that point in the project. Because the average plus the uncertainty suggested the stockpile might be at or slightly above the BNL 160 ppb action level, the Project Manager requested an off-site repeat of all Stockpile 7, and all samples stored in the Field Lab were transferred to the Contract Lab for extraction and analysis. The repeat run found the same average TCLP mercury, but one sample (ST-7-04-NET) was reported as having 4,520 µg/L!! When this sample was repeated, a value less than 50 was reported. Thus the second run had a much greater uncertainty (standard deviation), with TCLP mercury equal to 79.1 ± 546.9 µg/L, if the value above 4,000 µg/L is used. (These latter data are listed in Appendix 3.) Even though the high levels in individual samples could not be repeated, Stockpile 7 was assigned for disposal as mixed waste. This was supported in part by total mercury values, which were highest for Stockpile 7, at 65 mg/kg, as determined in the ASTD Field Lab, and 61 mg/kg, as determined by the off-site contract Lab (Table 8).

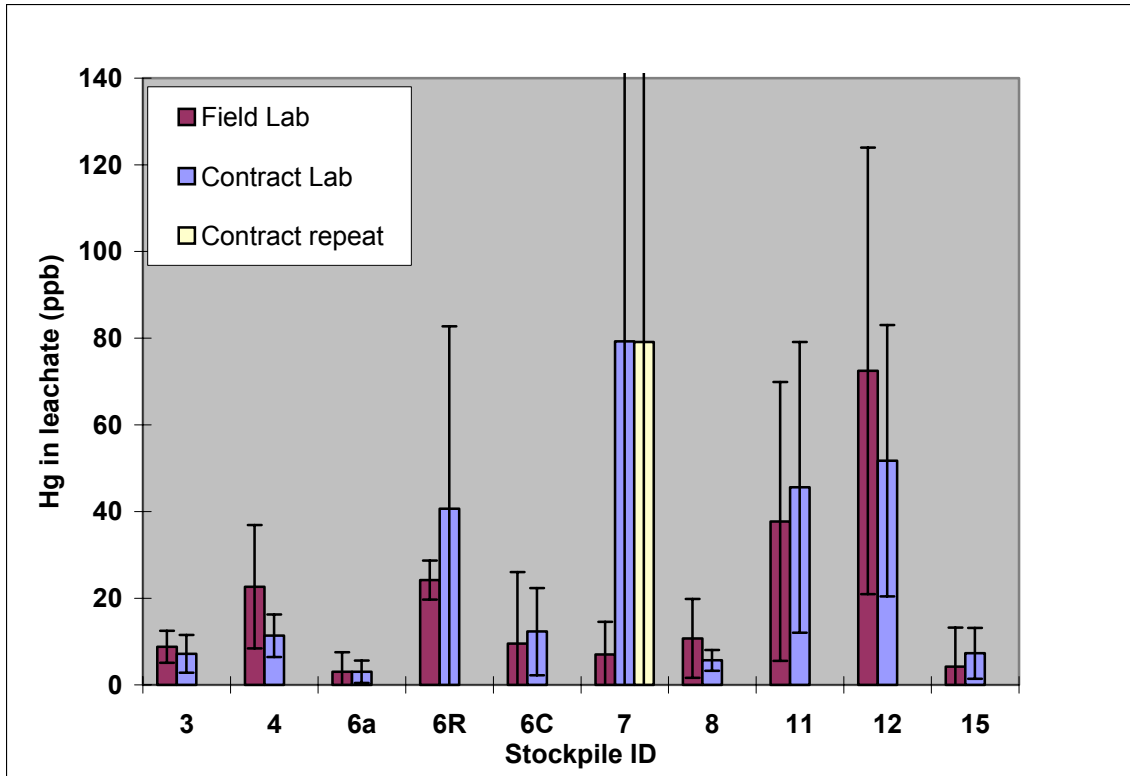


Figure 7. Comparison of TCLP mercury results for soil stockpiles

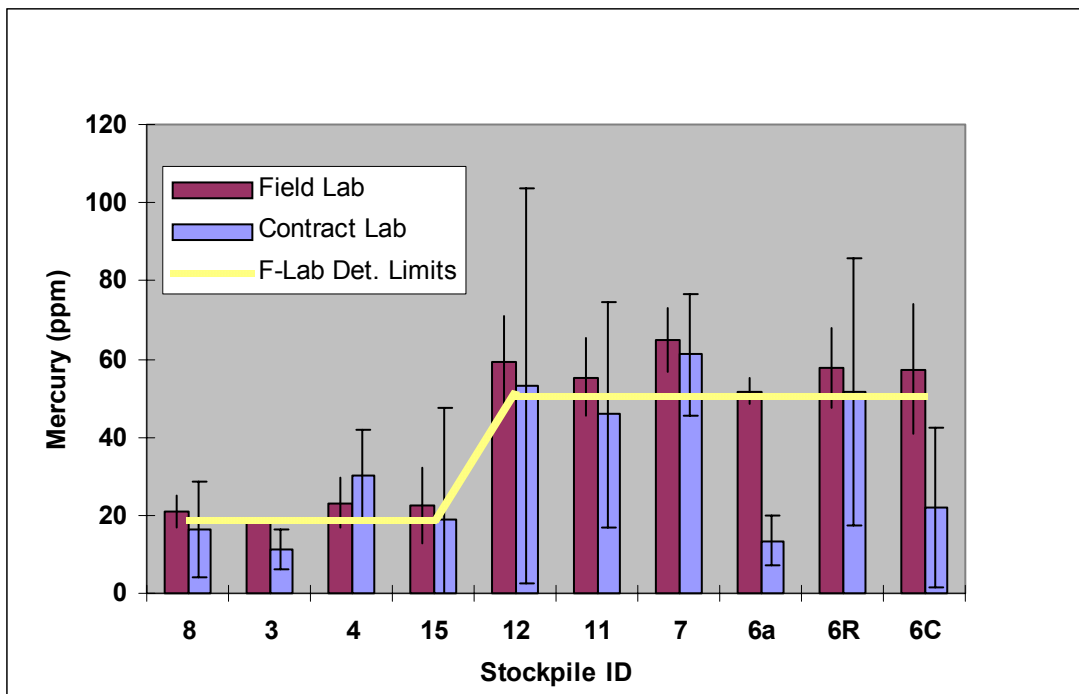


Figure 8. Comparison of total mercury results for stockpiled soils

4.2 Comparison with 2001 Sampling Campaign Results

A significant point of comparison is a 2001 sampling and characterization campaign, undertaken with the Toolbox to address some of the uncertainties with the earlier characterization work. In this campaign, 10 samples were collected for analysis from each stockpile in random fashion [Ref. 11]. From that campaign results, eight of the ten Stockpiles could be classified as non-hazardous from all TCLP metals and total mercury. However, the Toolbox called for further characterization, with 293 and 647 samples, respectively, taken for Stockpiles 6R and 7. A comparison of the 10-sample 2001 data for TCLP mercury and total mercury with results from the ASTD Field Lab is shown in Figure 9. For Stockpiles 6R, 7, and 12, the 2001 data standard deviations (shown as error bars) for TCLP mercury are off-scale. The ASTD Field Lab data in general reduced the uncertainty (standard deviations) for all total mercury data. The general conclusion from the ASTD Field Lab TCLP data was that all Stockpiles were within acceptable levels for classification as non-hazardous.

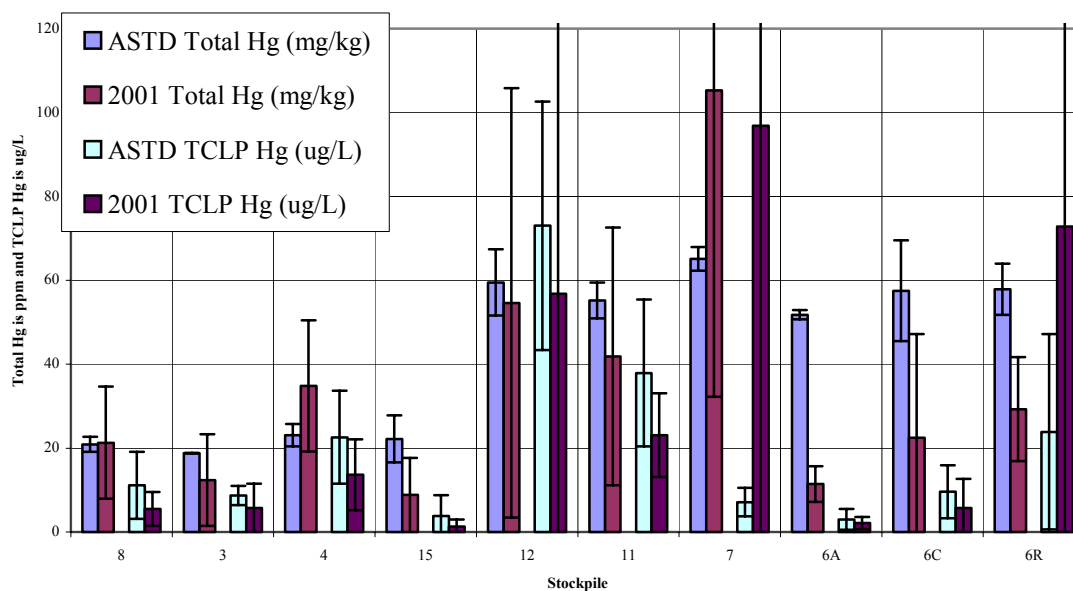


Figure 9. Comparison of field lab results with 2001 sampling campaign

4.3 Conclusions

The ASTD field laboratory completed more than 2,200 analyses for total mercury (XRF) and more than 2,400 TCLP mercury (DMA) analyses, over the project. Reliable statistical verification of the original characterization of the stockpiles as low-level wastes was accomplished. The Classification for Stockpile 12 Stockpile 12 was For most of the subpiles, TCLP analyses were completed within two days. One of the most significant aspects of the

project success was schedule acceleration. The original schedule projected activities extending from early April until September 30. Stockpile reconstruction was completed in the third week of August.

5. References

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3. BNL EM Directorate, "Sampling And Analysis Plan For Waste Characterization Of Stockpiled Soil," Brookhaven National Laboratory, January 2002.
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7. Crumbling, D.M., et al "Managing Uncertainty in Environmental Decisions," Environmental Science and Technology, American Chemical Society, pp. 405-409, October 1, 2001.
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10. W. de Laguna, "Geology of Brookhaven National Laboratory and Vicinity, Suffolk County, New York," Geological Survey Bulletin 1156-A, U.S. Geological Service (1963)
11. BNL EM Directorate, "Animal/Chemical Pits And Glass Holes Project Stockpile Characterization Summary Report," Brookhaven National Laboratory, March 28, 2002.

Table 1
Stockpiles Volumes and Samples Collected

Stockpile Number	Original Estimated Volume (yd³)	Number of Subpiles	Actual Volume* (yd³)	Samples**	Sorting Completed
3	150	5	100	42	May
4	450	34	680	286	May
6A	900	30	600	243	July
6C	320	10	200	84	August
6R	1,200	53	1,060	446	August
7	270	8	160	68	July
8	700	33	660	278	May
11	1,800	65	1,300	546	June
12	700	31	620	261	June
15	300	14	280	118	June
<i>Total</i>	<i>6,790</i>	<i>283</i>	<i>5,660</i>	<i>2,372</i>	

* Based on the number of 20 yd³ Subpiles

** For total Hg/TCLP Hg analyses. Includes field duplicates but not ISOCS or Off-site samples.

Total number of subpiles = 283 = number of ISOCS samples

Total Regular samples = 283x8 = 2,264

Total FD = 2372-2264 = 108

Total off-site is higher because all of ST 7 was re-analyzed off-site

**Table 2
Stockpiles Radionuclide Contamination Summary**

		Estimated Activity Concentration											
Stockpile	Total	Co-60				Cs-137				Ra-226			
		Maximum Value (pCi/g)	Subpile w/Max Value	Average (pCi/g)	Total ND's	Maximum Value (pCi/g)	Subpile w/Max Value	Average (pCi/g)	Total ND's	Maximum Value (pCi/g)	Subpile w/Max Value	Average (pCi/g)	Total ND's
St-8	33	0.7	21	0.5	21	0.6	19	0.3	31	0.5	3,19	0.5	30
St-3	5	0.5	1	0.5	4	ND		ND	5	ND		ND	5
St-4	34	0.6	15,21,24,27,29,31,34	0.5	17	0.2	5,6,10,15,16,25	0.2	18	0.4	30	0.4	33
St-15	14	0.7	22	0.5	2	0.3	8	0.2	12	0.5	11	0.5	13
St 12	31	ND		ND	31	1.1	24	0.4	10	ND		ND	31
St-11	65	0.8	49	0.5	36	2.3	45	0.3	19	0.5	49	0.5	64
St-7	8	ND		ND	8	1.0	1	0.8	0	0.3	6	0.3	7
St-6A	30	0.8	23	0.5	19	0.3	4,5	0.2	23	ND		ND	30
St 6C	10	0.6	5,7	0.5	3	ND		ND	10	ND		ND	10
St-6R	53	0.5	58,56,55,45,35	0.4	44	0.3	2,41,48,50	0.2	39	0.5	24	0.6	52
		Estimated Activity Concentration											
Stockpile	Total	Th-232				Am-241				U-235			
		Maximum Value (pCi/g)	Subpile w/Max Value	Average (pCi/g)	Total ND's	Maximum Value (pCi/g)	Subpile w/Max Value	Average (pCi/g)	Total ND's	Maximum Value (pCi/g)	Subpile w/Max Value	Average (pCi/g)	Total ND's
St-8	33	1.2	5	0.7	27	1.8	22	0.9	17	ND		0	65
St-3	5	0.5	3	0.5	4	ND		ND	5	ND		0	33
St-4	34	0.5	6,20,25	0.5	31	10.4	16	1.6	11	ND		0	31
St-15	14	0.5	11	0.5	12	0.8	8	0.8	13	ND		0	14
St 12	31	ND		0.0	31	11.4	11	3.1	17	ND		0	5
St-11	65	0.5	10	0.5	64	16.4	60	3.5	49	ND		0	34
St-7	8	0.7	3,4	0.7	6	1.1	5	1.0	3	ND		0	8
St-6A	30	0.5	13,15	0.5	35	20.1	29	2.7	20	ND		0	37
St 6C	10	ND		0.0	10	1.6	8	1.6	9	ND		0	10
St-6R	53	0.6	2	0.6	52	0.8	6	0.5	46	ND		0	53

Table 3
Stockpile Summary Data – Total and TCLP Mercury Results

Stockpile	Stockpile Average Results				Maximum Values	
	Total Hg	Std.Dev.	TCLP Hg	Std.Dev.	Total Hg	TCLP Hg
	(mg/kg)	(mg/kg)	(µg/L)	(µg/L)	(mg/kg)	(µg/L)
8	20.9	1.8	11.1	8.0	25.2	38.2
3	18.8	0.0	8.7	2.3	18.8	11.9
4	23.1	2.7	22.6	11.1	33.1	47.0
15	22.2	5.6	3.8	5.0	36.2	20.3
12	59.5	7.9	73.0	29.6	82.7	153.7
11	55.2	4.3	37.9	17.5	76.1	113.4
7	65.1	2.8	7.1	3.4	68.3	13.6
6A	51.8	1.1	3.0	2.5	54.5	10.7
6C	57.5	12.0	9.6	6.3	86.6	19.3
6R	57.9	6.1	23.9	23.3	83.2	120.7

Table 4
List of Subpiles Averages – Total and TCLP Mercury Concentrations

COC	Subpile	Subpile Average Results			
		Total Hg (mg/kg)	Std.Dev. (mg/kg)	TCLP (ug/L)	Std.Dev. (ug/L)
	Stockpile 8				
13442	st801	21.6	3.4	2.0	0.5
13444	st802	20.8	1.0	4.8	1.3
13445	st803	20.7	0.9	4.8	3.5
13446	st804	25.2	5.9	6.8	1.9
13447	st805	24.2	6.4	12.0	6.7
13448	st806	22.3	3.1	9.1	2.2
13449	st807	20.4	0.0	7.0	1.2
13450	st808	22.1	3.1	11.3	7.1
13458	st809	21.7	1.7	9.0	1.4
13459	st810	21.8	2.5	8.1	1.3
13461	st811	20.3	0.0	9.7	4.0
13462	st812	20.3	0.0	4.8	0.5
13464	st813	23.3	4.3	11.7	4.3
13465	st814	21.0	1.9	10.9	0.9
13467	st815	20.3	0.0	5.0	0.8
13505	st816	20.6	0.6	4.2	0.7
13506	st817	20.5	0.4	5.2	1.1
13507	st818	24.7	11.6	8.2	1.4
13508	st819	23.0	7.9	10.8	3.6
13509	st820	18.8	0.0	7.0	3.5
13510	st821	18.9	0.4	15.3	17.9
13511	st822	21.3	4.6	16.4	6.8
13512	st823	18.8	0.0	13.8	4.3
13513	st824	22.1	9.2	38.3	9.8
13514	st825	19.2	0.7	38.2	8.0
13533	st826	19.7	1.9	17.4	8.3
13534	st827	18.8	0.0	16.0	8.5
13535	st828	19.7	1.6	6.8	1.7
13536	st829	20.1	2.6	14.3	1.7
13537	st830	18.8	0.0	13.4	2.7
13538	st831	20.9	3.1	8.9	1.3
13539	st832	18.8	0.0	9.8	0.9
13540	st833	18.8	0.0	4.8	3.1
	St 8 Overall average	20.9	1.8	11.1	8.0
	Stockpile 3				
13541	ST301	18.8	0.0	7.6	2.7
13542	ST302	18.8	0.0	5.9	3.1
13560	ST303	18.8	0.0	8.3	3.6

COC	Subpile	Subpile Average Results			
		Total Hg (mg/kg)	Std.Dev. (mg/kg)	TCLP (ug/L)	Std.Dev. (ug/L)
13561	ST304	18.8	0.0	11.9	3.6
13562	ST305	18.8	0.0	9.9	3.2
	St 3 Overall average	18.8	0.0	8.7	2.3
	Stockpile 4				
13563	st401	24.6	12.7	8.5	1.2
13564	st402	33.1	15.9	11.9	3.9
13565	st403	24.4	7.6	13.0	3.0
13566	st404	24.2	7.1	14.1	2.7
13567	st405	22.8	3.3	20.8	9.1
13568	st406	26.4	8.9	21.4	2.5
13569	st407	26.2	5.5	22.7	3.3
13570	st408	23.6	4.6	17.3	13.6
13571	st409	18.8	0.0	12.2	7.2
13572	st410	20.9	3.8	13.3	7.4
13575	st411	23.9	5.4	13.5	5.0
13576	st412	20.2	2.6	17.3	5.3
13577	st413	24.4	4.7	18.8	3.5
13578	st414	24.9	7.7	18.1	4.2
13579	st415	19.6	1.7	12.2	3.9
13599	st416	19.9	2.7	10.5	6.2
13600	st417	19.4	1.1	8.8	2.7
13601	st418	22.2	5.2	11.0	2.2
13602	st419	20.5	2.8	24.6	6.3
13603	st420	25.2	6.0	24.9	9.7
13609	st421	22.1	3.7	35.5	5.8
13604	st422	22.0	4.0	32.0	2.9
13605	st423	19.5	1.3	35.3	7.2
13606	st424	22.8	4.4	28.4	6.5
13607	st425	23.5	4.5	36.4	25.7
13608	st426	24.3	6.1	39.0	6.8
13610	st427	25.0	7.0	47.0	12.3
13611	st428	23.2	4.1	44.5	28.0
13612	st429	23.6	6.1	35.1	20.6
13613	st430	20.4	2.8	29.6	12.9
13614	st431	22.4	4.9	31.3	4.2
13615	st432	23.3	4.3	30.5	8.4
13616	st433	24.2	5.8	21.2	2.2
13617	st434	22.1	6.4	6.3	3.3
	ST 4 - Overall Average	23.1	2.7	22.6	11.1
	Stockpile 15				
13618	st151	19.4	1.9	2.0	0.7
13619	sta52	20.0	2.4	5.5	1.6
16620	st153	18.8	0.0	4.1	1.3
13641	st154	18.8	0.0	1.5	1.4
13642	st155	18.8	0.0	1.0	0.6

COC	Subpile	Subpile Average Results			
		Total Hg (mg/kg)	Std.Dev. (mg/kg)	TCLP (ug/L)	Std.Dev. (ug/L)
13643	st156	18.8	0.0	3.2	1.3
13644	st157	18.8	0.0	2.4	2.6
13645	st158	18.8	0.0	3.6	4.5
13646	st159	18.8	0.0	0.6	0.3
13647	st1510	36.2	31.3	20.3	22.7
13648	st1511	18.8	0.0	1.7	0.3
13649	st1512	27.7	0.0	1.5	0.4
13650	st1513	27.7	0.0	1.5	0.5
13651	st1514	29.1	3.0	4.1	6.6
	St 15 Overall Average	22.2	5.6	3.8	5.0
	Stockpile 12				
13652	st1201	54.6	4.1	32.3	13.4
13653	st1202	53.9	3.9	39.2	13.8
13654	st1203	54.3	7.2	53.4	23.5
13655	st1204	56.4	8.3	102.4	96.7
13656	st1205	61.3	17.6	84.2	28.6
13657	st1206	54.1	8.8	34.2	14.6
13658	st1207	51.3	1.8	25.0	9.2
13739	st1208	51.2	1.8	43.7	7.0
13740	st1209	51.1	1.8	94.7	57.1
13741	st1210	53.3	4.0	57.4	14.4
13742	st1211	52.2	3.7	80.5	13.0
13743	st1212	64.4	20.6	80.0	11.8
13744	st1213	59.7	10.0	84.4	17.0
13745	st1214	60.8	10.9	60.8	9.2
13746	st1215	61.1	4.7	68.4	13.6
13749	st1216	59.0	12.4	70.5	24.3
13750	st1217	56.7	5.5	153.7	196.0
13751	st1218	60.7	9.0	85.2	19.0
13752	st1219	56.4	5.4	88.0	25.9
13753	st1220	53.9	4.0	89.2	83.0
13754	st1221	51.3	2.3	62.8	12.2
13755	st1222	54.4	6.2	56.0	14.0
13756	st1223	56.2	5.8	52.2	18.5
13757	st1224	57.2	4.7	52.7	8.5
13758	st1225	60.0	8.7	67.5	27.6
14315	st1226	68.7	6.9	98.4	36.3
14316	st1227	69.5	12.1	103.0	16.8
14317	st1228	72.6	11.1	112.8	14.2
14318	st1229	76.0	14.9	122.5	27.3
14319	st1230	82.7	20.2	83.0	18.4
14320	st1231	70.1	15.5	26.3	15.5
	St 12 Overall Average	59.5	7.9	73.0	29.6
	Stockpile 11				

COC	Subpile	Subpile Average Results			
		Total Hg	Std.Dev.	TCLP	Std.Dev.
		(mg/kg)	(mg/kg)	(ug/L)	(ug/L)
14322	st1101	50.5	0.0	20.7	4.6
14323	st1102	50.9	1.2	26.2	3.1
14324	st1103	55.7	8.8	28.6	7.5
14325	st1104	52.1	2.3	29.8	7.9
14326	st1105	59.7	13.9	43.8	9.8
14327	st1106	57.1	6.1	46.6	18.2
14328	st1107	55.7	6.7	40.5	12.3
14329	st1108	54.9	4.3	23.9	3.7
14330	st1109	51.7	3.2	39.0	12.4
14331	st1110	56.9	4.9	33.3	23.1
14332	st1111	54.5	5.7	30.7	5.9
14333	st1112	51.8	3.3	30.9	6.3
14334	st1113	51.2	1.8	30.7	9.5
14341	st1114	53.9	6.9	27.7	3.8
14342	st1115	51.7	2.3	49.7	23.7
14343	st1116	52.5	5.2	30.3	8.1
14344	st1117	52.5	3.6	30.3	6.0
14345	st1118	53.5	5.2	29.9	6.0
14346	st1119	59.1	16.4	24.8	12.7
14347	st1120	53.0	2.6	48.0	23.7
14348	st1121	52.3	3.1	26.9	11.2
14349	st1122	56.5	10.0	22.1	15.8
14587	st1123	51.7	3.2	21.2	5.9
14588	st1124	58.1	15.6	16.8	5.7
14589	st1125	51.7	2.2	20.8	9.6
14590	st1126	53.5	2.9	26.7	6.2
14591	st1127	52.1	3.3	66.2	67.9
14592	st1128	51.2	1.4	39.7	11.0
14593	st1129	54.5	6.2	32.6	17.9
14594	st1130	51.9	4.2	32.3	8.6
14595	st1131	59.5	9.2	61.5	58.8
14596	st1132	53.5	4.5	40.6	7.8
14597	st1133	56.6	9.4	42.0	6.8
14599	st1134	53.3	3.0	17.1	3.2
14601	st1135	59.6	9.5	75.8	6.2
14602	st1136	54.5	5.5	50.4	10.6
14603	st1137	56.9	5.2	49.2	6.5
14604	st1138	55.4	5.5	31.4	5.5
14605	st1139	57.1	10.1	76.3	16.9
14606	st1140	62.7	11.7	54.6	8.9
14607	st1141	57.3	9.0	63.4	12.3
14608	st1142	59.7	22.3	28.9	5.7
14609	st1143	59.8	10.6	113.4	120.6
14610	st1144	56.7	10.9	63.2	14.5
14611	st1145	55.6	5.6	47.4	19.8

COC	Subpile	Subpile Average Results			
		Total Hg (mg/kg)	Std.Dev. (mg/kg)	TCLP (ug/L)	Std.Dev. (ug/L)
14612	st1146	59.2	3.9	43.3	6.7
14613	st1147	58.0	9.8	29.9	4.2
14614	st1148	55.0	5.2	38.8	12.9
14615	st1149	76.1	18.4	72.3	14.8
14616	st1150	55.3	5.8	47.4	28.5
14617	st1151	55.4	6.1	38.7	14.2
14644	st1152	64.1	20.1	29.9	3.9
14645	st1153	53.0	4.1	22.8	3.1
14646	st1154	52.0	2.4	26.4	4.8
14647	st1155	51.6	2.4	34.0	4.8
14648	st1156	58.4	10.6	37.0	8.1
14649	st1157	64.2	19.0	45.3	12.0
14650	st1158	53.1	3.4	36.6	2.6
14651	st1159	53.4	4.0	37.8	5.5
14652	st1160	50.5	0.0	35.3	5.0
14653	st1161	56.3	10.0	23.2	5.7
14654	st1162	53.1	4.7	33.6	4.0
14655	st1163	50.5	0.0	18.4	6.7
14656	st1164	50.5	0.0	8.8	1.5
14657	st1165	50.7	0.8	18.0	3.8
	St 11 Overall average	55.2	4.3	37.9	17.5
	Stockpile 7				
14659	st701	66.9	9.2	6.0	3.2
14660	st702	63.8	10.5	13.6	17.7
14661	st703	68.1	5.9	3.5	5.1
14662	st704	64.2	7.4	4.9	4.5
14663	st705	60.6	8.6	6.1	3.6
14677	st706	66.6	9.4	5.3	2.7
14678	st707	68.3	4.5	11.2	4.2
14679	st708	62.6	8.7	5.4	4.8
	St 7 Overall average	65.1	2.8	7.0	3.5
	Stockpile 6A				
14680	st6A01	50.5	0.0	1.3	0.5
14681	st6A02	50.5	0.0	4.3	4.1
14682	st6A03	51.3	2.5	10.7	17.3
14683	st6A04	52.4	3.9	4.5	4.1
14684	st6A05	52.6	5.7	4.1	1.2
14685	st6A06	52.4	4.5	4.4	3.1
14686	st6A07	53.2	3.9	3.5	0.9
14687	st6A08	53.0	2.8	5.6	2.6
14688	st6A09	50.7	0.4	5.1	1.6
14689	st6A10	54.5	9.2	7.3	9.3
14690	st6A11	53.3	4.5	6.2	1.7
14691	st6A12	52.3	3.4	6.1	2.0
14692	st6A13	52.5	2.9	5.0	3.0

COC	Subpile	Subpile Average Results			
		Total Hg	Std.Dev.	TCLP	Std.Dev.
		(mg/kg)	(mg/kg)	(ug/L)	(ug/L)
14693	st6A14	51.7	1.4	2.5	0.7
14694	st6A15	52.5	3.1	2.4	1.0
14695	st6A16	50.5	0.0	1.2	1.1
14776	st6A17	52.0	3.0	1.2	1.4
14777	st6A18	50.5	0.0	1.0	0.8
14778	st6A19	50.7	0.5	0.9	0.3
14779	st6A20	53.9	4.6	1.2	0.5
14780	st6A21	51.2	1.3	0.9	0.2
14781	st6A22	52.5	4.2	3.2	1.6
14782	st6A23	52.7	3.7	1.4	1.4
14783	st6A24	50.8	0.9	1.3	0.4
14784	st6A25	50.5	0.2	1.2	0.8
14785	st6A26	51.0	1.5	1.3	0.3
14786	st6A27	51.9	3.1	0.8	0.3
14787	st6A28	50.5	0.0	0.7	0.7
14788	st6A29	50.6	0.2	0.6	0.1
14789	st6A30	51.2	1.2	0.8	0.4
	St 6A Overall Average	51.8	1.1	3.0	2.5
	Stockpile 6C				
14872	st6C01	50.8	0.8	2.3	2.1
14874	st6C02	51.5	2.8	3.2	1.1
14875	st6C03	50.5	0.0	7.3	2.5
14876	st6C04	51.6	2.8	19.3	32.6
14877	st6C05	51.2	1.7	3.3	1.3
14878	st6C06	50.5	0.0	15.9	37.6
14880	st6C07	51.6	2.5	7.2	11.0
14881	st6C08	61.1	15.3	8.9	11.6
14882	st6C09	86.6	36.4	18.2	2.8
14883	st6C10	69.6	11.2	10.3	3.1
	St 6C Overall average	57.5	12.0	9.6	6.3
	Stockpile 6R				
14884	st6R01	55.6	4.6	4.0	2.9
14885	st6R02	53.8	5.1	5.3	3.5
14886	st6R03	53.8	4.7	10.7	15.4
14887	st6R04	59.4	8.8	5.5	1.3
14905	st6R05	52.7	3.0	2.5	2.0
14906	st6R06	52.9	5.7	6.6	2.1
14907	st6R07	55.6	5.5	8.6	4.8
14909	st6R08	55.6	6.8	7.1	4.0
14910	st6R09	83.2	27.6	52.4	45.1
14911	st6R10	69.0	7.1	36.6	14.9
14912	st6R11	59.6	8.6	12.8	2.6
14913	st6R12	67.9	7.6	19.6	19.5
14914	st6R13	62.7	9.7	16.7	9.4
14915	st6R14	59.3	8.1	27.2	13.5

COC	Subpile	Subpile Average Results			
		Total Hg	Std.Dev.	TCLP	Std.Dev.
		(mg/kg)	(mg/kg)	(ug/L)	(ug/L)
14916	st6R15	61.6	12.6	10.3	3.0
14917	st6R16	51.9	2.7	21.5	14.5
14918	st6R17	58.5	11.2	44.2	11.7
14919	st6R18	59.3	11.8	22.8	5.4
14920	st6R19	59.9	5.8	23.9	9.2
14921	st6R20	56.0	6.6	12.5	7.6
14922	st6R21	59.1	10.3	18.5	3.2
14923	st6R22	59.6	6.5	15.3	4.3
14924	st6R23	59.9	10.8	24.4	6.8
14985	st6R24	73.7	8.3	37.3	43.1
14986	st6R25	59.6	6.7	20.2	9.6
14987	st6R26	53.4	2.8	33.5	21.8
14988	st6R27	55.9	4.9	23.6	25.0
14989	st6R28	58.3	13.2	7.2	5.2
14990	st6R29	58.1	4.3	63.7	111.2
14991	st6R30	53.4	5.0	32.7	45.9
14992	st6R31	61.8	11.4	16.2	3.4
15406	st6R32	61.6	8.1	25.8	11.0
15407	st6R33	59.3	5.6	97.1	228.2
15408	st6R34	61.1	18.9	19.8	16.6
15409	st6R35	51.2	2.2	16.9	7.9
15410	st6R36	56.0	6.8	13.8	5.4
15411	st6R37	53.4	4.3	15.0	19.0
15412	st6R38	57.1	6.0	11.0	3.2
15413	st6R39	54.2	4.0	8.0	0.7
15414	st6R40	55.9	4.3	7.1	1.0
15415	st6R41	56.2	5.3	5.8	1.3
15416	st6R42	50.5	0.0	4.2	1.3
15418	st6R43	58.7	6.0	11.6	6.2
15419	st6R44	51.1	1.3	16.3	8.8
15420	st6R45	50.8	0.6	21.0	10.0
15421	st6R46	54.2	5.4	21.4	4.5
15422	st6R47	63.2	22.7	35.9	14.1
15423	st6R48	56.9	11.9	53.2	62.9
15424	st6R49	68.6	10.3	77.9	39.6
15425	st6R50	51.5	2.0	30.9	6.3
15426	st6R51	50.5	0.0	120.7	312.1
15417	st6R52	52.7	3.2	5.8	2.0
15427	st6R53	54.0	4.8	3.4	1.8
	ST 6R Overall Average	57.9	6.1	23.9	23.3

Table 5
TCLP Mercury Results for Common Data Points

Stockpile	# of Data	ASTD Field Lab			Hits		Contract Lab		
		Average	Std.	95%*	>100 ppb	Average	Std.	95%*	
#	Points	µg/L	Dev.	95%*	F-lab	Con	µg/L	Dev.	95%*
3	6	9.1	4.8	18.7	0	0	7.2	4.3	15.9
4	14	19.2	10.2	39.6	0	0	11.3	4.9	21.2
6a	13	3.5	5.0	13.5	0	0	3.0	2.6	8.2
6R	22	20.7	19.8	60.3	0	1	40.7	42.0	124.7
6C	3	3.6	1.5	6.6	0	0	12.3	10.1	32.4
7	4	2.4	1.6	5.6	0	1	97.6	179.7	457.0
8	41	7.4	8.5	24.4	0	0	5.1	2.1	9.3
11	25	31.8	12.3	56.4	0	2	45.6	33.5	112.7
12	13	71.1	26.6	124.3	3	1	51.8	31.3	114.3
15	3	2.7	2.1	6.9	0	0	4.1	5.1	14.3

* Refers to the 95% confidence value, calculated as [Average + (2 x Std.Dev.)]

Table 6
TCLP Mercury Results Comparison for All Data Points

Stockpile	ASTD Field Lab				Hits		Contract Lab			
	Average	Std.	95%*	# of Data	>100 ppb	Average	Std.	95%*	# of Data	
#	µg/L	Dev.	95%*	Points	F-lab	Con	µg/L	Dev.	95%*	Points
3	8.8	3.7	16.2	42	0	0	7.2	4.3	15.9	6
4	22.6	14.2	51.1	286	0	0	11.3	4.9	21.2	14
6a	3.1	4.5	12.0	251	10	0	3.0	2.6	8.2	13
6R	24.2	4.5	33.2	443	10	1	40.7	42.0	124.7	22
6C	9.5	16.5	42.6	83	1	0	12.3	10.1	32.4	3
7	7.0	7.5	22.1	66	0	1	79.3	161.0	401.2	5
8	10.7	9.1	28.9	322	0	0	5.7	2.4	10.5	103
11	37.7	32.2	102.1	545	8	2	45.6	33.5	112.7	25
12	72.5	51.5	175.5	260	42	1	51.8	31.3	114.3	13
15	4.2	9.0	22.3	117	0	1	7.3	5.9	19.0	5
all	27.2	34.8	96.9	2415	71	6	16.7	34.4	85.4	209

* Refers to the 95% confidence value, calculated as [Average + (2 x Std.Dev.)]

**Table 7
Total Mercury Results for Common Data Points**

Stockpile	# of Data	ASTD Field Lab			Contract Lab		
		Average	Std.	95%*	Average	Std.	95%*
#	Points	mg/Kg	Dev.	95%*	mg/Kg	Dev.	95%*
3	3	18.8	0.0	18.8	12.4	6.7	25.8
4	14	27.8	10.4	48.7	30.3	11.6	53.5
6a	13	51.2	2.5	56.3	13.4	6.4	26.3
6R	22	57.7	9.0	75.6	51.4	34.3	120.0
6C	3	51.6	2.0	55.6	21.7	20.4	62.6
7	4	64.2	12.0	88.1	61.1	15.7	92.5
8	17	20.3	2.8	25.8	13.8	8.1	30.1
11	25	54.4	4.6	63.6	45.8	28.8	103.4
12	12	52.2	3.7	59.5	53.1	50.6	154.3
15	5	39.2	40.9	121.1	18.9	28.6	76.0

* Refers to the 95% confidence value, calculated as [Average + (2 x Std.Dev.)]

**Table 8
Total Mercury Results Comparison for All Data Points**

Stockpile	ASTD Field Lab				Contract Lab			
	Average	Std.	95%*	# of Data	Average	Std.	95%*	# of Data
#	mg/Kg	Dev.	95%*	Points	mg/Kg	Dev.	95%*	Points
3	18.8	0.0	18.8	20	11.3	4.9	21.0	6
4	23.1	6.3	35.7	253	30.3	11.6	53.5	14
6a	51.7	3.2	58.1	252	13.4	6.4	26.3	13
6R	57.8	10.1	78.0	445	51.4	34.3	120.0	22
6C	57.3	16.6	90.4	83	21.7	20.4	62.6	3
7	64.8	8.3	81.5	68	61.1	15.7	92.5	4
8	20.3	2.8	25.8	17	16.5	12.3	41.1	17
11	55.4	9.9	75.2	546	45.8	28.8	103.4	25
12	59.3	11.8	83.0	260	53.1	50.6	154.3	12
15	22.3	9.6	41.6	116	18.9	28.6	76.0	5
All	47.2	17.4	82.0	2,060	15.8	10.8	37.4	121

* Refers to the 95% confidence value, calculated as [Average + (2 x Std.Dev.)]

Appendices
(On Compact Disc)

1. *In-Situ* Object Counting System (ISOCS) – Gamma Spectroscopy Data Summaries
2. ASTD Field Lab – Total Mercury and TCLP Mercury Data
3. QA Data – Comparisons with Off-site Lab results