



## Assessing the effectiveness of RapidEye multispectral imagery for vegetation mapping in Madeira Island (Portugal)

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### Abstract

Madeira Island is a biodiversity hotspot due to its high number of endemic/native plant species. In this work we developed and assessed a methodological framework to produce a RapidEye-based vegetation map. Reasonable accuracies were achieved for a 26 categories classification scheme in two different seasons. We tested pixel and object based approaches and the inclusion of a vegetation index band on top of the pre-processed RapidEye bands stack. Object based generally showed to outperform pixel based classification approaches except for linear or highly scattered classes. The addition of a vegetation index to the workflow increased the separability of the Jeffrey-Matusita least separable class pairs, but not necessarily the overall accuracy. The Pontius accuracy assessment highlighted class specific accuracy tradeoffs related to different combinations of the inputs and methods. The approach to be used, in conclusion, should be carefully considered on the basis of the desired result.

**Keywords:** Land cover mapping, biodiversity assessment, land use assessment, oceanic island.

### Introduction

Madeira Island (Archipelago of Madeira, Portugal) is the main island of the Autonomous Region of Madeira, which is one of the European Union Outermost Regions. The majority of European Outermost Regions - except the French Guiana - are small islands or archipelagos [Gil et al., 2012]. Madeira Island is located in the Macaronesia biogeographic region, which is constituted by five oceanic archipelagos located on the eastern part of the Atlantic Ocean,

facing the coasts of the African continent and the Iberian Peninsula. The archipelagos share a common geologic origin and several fauna and flora elements [Fernández-Palacios et al., 2011]. Madeira Island is also considered part of the Mediterranean basin biodiversity hotspot [Medail and Quezel, 1997] mainly due to the high degree of endemism and the presence of Laurisilva, a laurel forest whose most representative species are nowadays occurring only in Macaronesia. Madeira's Laurisilva, representing the world widest continuous and better conserved patch [Fernández-Palacios et al., 2011], was named in 2009 World UNESCO Heritage [World Heritage Committee, 2009] and is protected under the Natural Park of Madeira framework. Despite the conservation efforts, human influence led to a conspicuous recession of these native/endemic species, with the substitution by exotic, invasive and sin-anthropogenic species [Capelo et al., 2004]. Human development is not the only cause for relevant changes in land cover. Madeira is subject to violent catastrophic events such as landslides with debris streams after winter season flash-floods [Lira et al., 2011, 2013] and large wildfires, when the annual grass and the seasonal shrubs curing levels turn them in an extremely flammable fuel [Nasa Visible Earth, 2010; Fontinha et al., 2014]. Since the year 1803 flood (which caused 800-1000 casualties) at least 30 main events of this kind followed, with the most disastrous of our century being dated on 20 February 2010; the flood caused 45 casualties and temporary streams with a range of 200-600 m<sup>3</sup> s<sup>-1</sup> of debris transportation (maximum peak of 663 m<sup>3</sup> s<sup>-1</sup> in Ribeira Brava basin) [Fragoso et al., 2012]. Wildfires can sensibly affect main land cover changes, favoring the spread of certain exotic species which are on one hand, in critic periods, a more flammable fuel than endemic species and on the other, a more aggressive colonizer during the recovery [Fontinha et al., 2014]. During the eighth Convention on Biological Diversity in Brazil it was in fact established that among the main threats in small islands to sustainable development, nature conservation and biodiversity maintainability, there are the climate variability and changes, the proliferation of invasive exotic species, the increasing growth of tourist activity, the natural catastrophes and the overexploitation of natural resources [CBD, 2006]. Remote Sensing data is optimal for vegetation mapping at various scales of interest, as well as time saving, cheaper and more timely than traditional field survey methods [Xie et al., 2008]. Due to climate constraints and the consequent general lack of free-of-charge satellite remote sensing data (e.g. Landsat) covering most oceanic islands, only few examples of vegetation mapping in these insular territories have been based on satellite remote sensing data [Gil et al., 2011, 2013, 2014; Gil and Abadi, 2015] and none directly applied to Madeira Island. RapidEye is the first satellite system on its resolutions (spatial, temporal and radiometric) that provides a red-edge channel [Jung-Rothenhäusler et al., 2007]. The red-edge band has demonstrated to significantly improve the vegetation species recognition [Schuster et al., 2012; Adam et al., 2014], especially when used for the retrieval of RapidEye-specific vegetation indices [Eitel et al., 2007; Förster et al., 2011; Fritsch et al., 2012] and the monitoring of vegetation sanitary status [Adelabu et al., 2014]. RapidEye multispectral imagery also showed good performance in plant communities classification [Förster et al., 2012; Elatawneh et al., 2014], isolated-three genera recognition [Tigges et al., 2013], forest intra-and-inter species biomass prediction [Dube et al., 2014], damage in biodiversity [Cruz-Lopez and Lopez-Saldaña, 2011], species and age specific forest groups [Ivanov et al., 2011], burnt areas detection and retrieval of biophysical products [Vuolo et al., 2010; Jiali et al., 2012; Asam et al., 2013; Cho et al., 2013; Kross et al., 2015]. A few statistically

oriented studies demonstrated that artificial bands derived from combinations of satellite data can be more correlated to certain vegetation characteristics than the reflectance values originally acquired by the sensor. For example Beckschäfer et al. [2014] showed higher correlation of vegetation indices and texture bands retrieved from RapidEye product with the Leaf Area Index than its own bands. Lu [2014] found higher correlation of forest stand structure and aboveground biomass to an artificially calculated texture band than to the reflectance bands used for its calculation. Thus, although the data becomes redundant in the set, it has been demonstrated that the implementation of a calculated band in a multi-spectral dataset improves the final classification accuracy [Schuster et al., 2012; Adelabu et al., 2014; Buck et al., 2015; Godinho et al., 2016]. Due to the high spatial resolution of RapidEye multispectral imagery (6.5 m), and a set of channels strategically positioned for vegetation oriented analysis [Dube et al., 2014], the image processing and classification procedures of this type of remote sensing data can be approached either as pixel oriented or object oriented. A pixel oriented approach enables the possibility not to lose any spatial information, classifying every pixel in the scene; an object oriented approach consists in one additional preliminary step before the classification which clusters the pixels in segments. A segment is a group of spatially adjacent and spectrally similar pixels which, once filtered, will assume a unique spectral value (per band) with a decrease of complexity for the classifier [Blaschke, 2010]. This research paper proposes and describes a methodological framework which aims to produce a high spatial resolution vegetation map with a specific focus on natural vegetation of Madeira Island (Archipelago of Madeira, Portugal), especially its native/endemic and invasive vegetation patches, through the use of pixel-based and segment-based supervised classifications of RapidEye multispectral bands and derived vegetation indices.

## Materials and methods

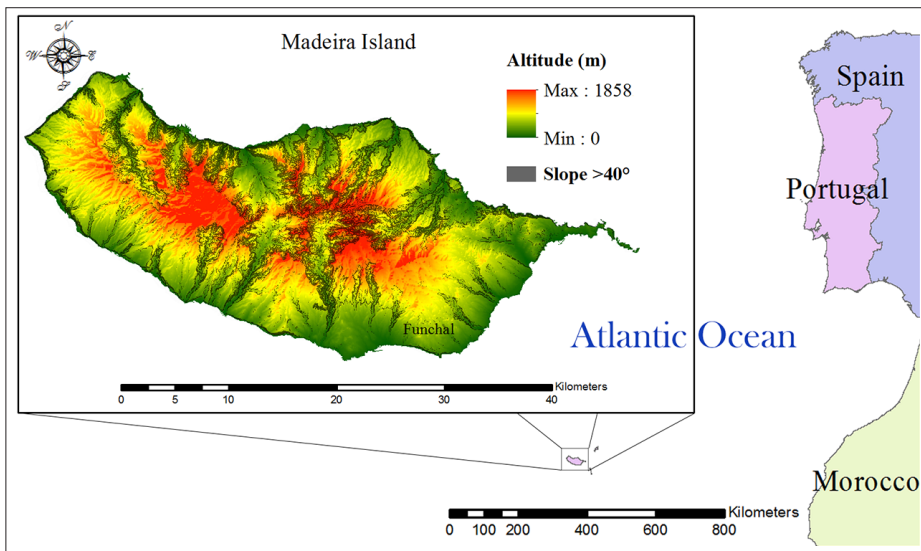
### *Study Area*

The Archipelago of Madeira is located in the biogeographic region of Macaronesia. It includes the islands of Madeira and Porto Santo, as well as the Selvagens islets. Madeira Island (32°38'49.96"N 16°54'29.59"W / 32.6472111°N 16.9082194°W) has an extension of 742 km<sup>2</sup>. From a geological perspective, the archipelago formation is considered due to Miocene volcanism [Zbyszewski, 1971; Emery et al., 2012]. In this context, Madeira has reached the geological evolution stage of "erosion and dismantling" [Fernández-Palacios et al., 2011] which is characterized by a considerable topographic complexity with highly various and often sheer relieves, despite the loss of the summit peaks, turned into plateaus, due to erosive phenomena (Fig. 1). Madeira is positioned on the interface of Temperate and Mediterranean macro-bio-climates [Rivas-Martinez, 2008] due to an East to West mountain range orientation. This configuration, nearly perpendicular to the trade winds, that blow from North-East for almost all the year, does characterize a relevant difference in temperature and rainfall in the North and South side of the Island [Prada et al., 2009]. Hence the southern side is characterized by at least two months of dry summer not compensated by the underground water reservoir, while in the north-side a cool climate with a distributed precipitation is dominant [Mesquita et al., 2004]. Seasonal relevant variations of temperatures and rainfall with altitude and island side contributed to the development of a complex mosaic of vegetation across the Island. According to the

most recent models [Aguiar et al., 2004; Costa et al., 2012] the natural potential vegetation of Madeira corresponds to the following climatophyllous series:

- 1) *Mayteno umbellatae-Oleo maderensis* sigmetum (series of madeiran oleaster tree), inframediterranean, thermophyllous on rocky biotopes, surviving in scarce mosaics in steep ravines (south face up to 200 m a.s.l.);
- 2) *Helichryso melaleuci-Sideroxylo marmulanae* sigmetum (series of marmulano tree), microforest in inframediterranean (north face 0 to 80 m a.s.l., south face, 200-300 m a.s.l.), scarcely represented on current vegetation;
- 3) *Semele androgynae-Apollonio barbujanae* sigmetum (mediterranean laurel/ barbusano-tree forest series), thermophillous, infra- and thermomediterranean forest, nowadays mostly destroyed (South 300-800 m a.s.l., North 50-450 m a.s.l.);
- 4) *Clethro arboreae-Ocoteo foetentis* sigmetum (stink-laurel temperate forest series) Infra to mesotemperate, forest series, still covering large areas of the island, mainly on the North face (800-1450 m a.s.l. South face; 300-1400 m a.s.l. North face);
- 5) *Polysticho falcinelli-Erico arboreae* sigmetum (high altitude tree-heath series), Meso-supra temperate series of tree-heath forests, mostly destroyed cutting and grazing (*Erica arborea*) (1400 to 1650 m a.s.l.).

In more general terms, Madeira’s biodiversity extent reaches 7571 terrestrial species [Borges et al., 2008] with 138 taxa of Madeira’s endemic vascular plant, 67 taxa of Macaronesian endemic vascular plant and 448 taxa of Madeira’s native vascular plant [Jardim and Sequeira, 2008]. The endemism density level (species per surface unit) in this island is the highest in Macaronesia (6.8%), whilst over 400 taxa of vascular plants are considered as introduced [Jardim et al., 2008], among which 91 are considered invasive [Kueffer et al., 2010]. In light of what stated above and all the cited literature, Madeira should be considered a crucial hotspot for biodiversity conservation.



**Figure 1 - Geographic position and orographic map of Madeira Island.**

### ***Satellite imagery***

Two RapidEye scenes processed at level 1B (L1B), fully covering Madeira Island and acquired respectively on December 13th, 2009 (winter) and August 10th, 2011 (summer) were used in this study. RapidEye is constituted by a constellation of five Earth Observation satellites carrying identical multispectral push-broom sensors. Each sensor records in five bands of the spectrum including three visible radiation bands (blue, 440-510 nm; green, 520-590 nm; red, 630-685 nm), a red-edge band (690-730 nm) and a near infrared band (760-850 nm)

L1B RapidEye products are delivered with a spatial resolution of 6.5 m at nadir, a radiometric resolution of 16 bit and with only minimum level of processing. This imagery comes radiometrically inter-calibrated, to uniform responses from different sensors in different geometric conditions and with the bands co-registered to each other, to guarantee the overlap of the bands within a scene. No terrain model is used for the processing of L1B imagery thus an ortho-rectification step must be included to increase the accuracy of the product. Excluding off-nadir and terrain effects, the root mean square error (RMSE) is expected to be of 10m [RapidEye, 2012]. The main characteristics of the scenes used are reported in Table 1.

**Table 1 - RapidEye scenes dates and acquisition geometry details.**

<b>Parameter</b>	<b>Scene 1</b>	<b>Scene 2</b>
<b>Acquisition date</b>	2009-12-13	2011-08-10
<b>Hour (UTC)</b>	12:52	12:52
<b>Spacecraft off-nadir view angle (deg)</b>	-16.58597	+3.499160
<b>Scan line direction azimuth referred to true North (deg)</b>	100.6	279.22
<b>Solar azimuth (deg)</b>	177.5662	163.7326
<b>Solar elevation (deg)</b>	34.2387	72.3995

### ***Other datasets used***

A dataset constituted by 387 geo-referenced (GPS) points indicating the occurrence of endemic/native and alien invasive plants patches (92 classes at sub community or stand level) across whole Madeira Island was used in this study. This dataset comprehends the result of several field surveys performed from 2009 to 2011 by the Madeira Botanical Group specialists (based at the Department of Life Sciences, University of Madeira, Portugal) - However in all these surveys no specific and well-defined statistical sampling design was used. Additional datasets used in the work are indicated in Table 2.

**Table 2 - Additional datasets used in this study.**

<b>Dataset</b>	<b>Data Producer</b>	<b>Format</b>	<b>Geographical Scale / Spatial Resolution</b>	<b>Use in the study</b>
LULC of Madeira 2007	Regional Government of Madeira Autonomous Region (Portugal)	Polygon vector	1:10000; minimum cartographic unit: 0.25 ha	Ground truth reference
Madeira Island Altimetry layer (contour lines) 2010	Regional Government of Madeira Autonomous Region (Portugal)	Polyline vector	Contour interval: 5m	Ortho-rectification
Madeira Island Forest Inventory 2008 [IFRAM, 2008]	Regional Government of Madeira Autonomous Region (Portugal)	Polygon vector (Thiessen polygons)	Polygon area: 62500 m <sup>2</sup>	Ground truth reference
Color Ortophoto-maps 2007	Regional Government of Madeira Autonomous Region (Portugal)	Raster RGB	1:5000	Ground truth reference
Google Earth	Google Inc.	Online DigitalGlobe Quickbird pan-sharpened imagery	0.65m	Ground truth reference
Google street view	Google Inc.	Photos	-	Ground truth reference

Note: Original or re-projected Geographic References: EPSG 3061; projection UTM; Ellipsoid International 1924; Datum: Porto Santo 1995; UTM Coordinate System - Zone 28N.

Figure 2 schematically represents the methodological workflow of this study which can be divided in three different steps, namely: (1) RapidEye data preprocessing; (2) classification scheme set up, training sites collection and spectral separability assessment; and (3) concurrent classification process and accuracy assessment. Every step is described below.

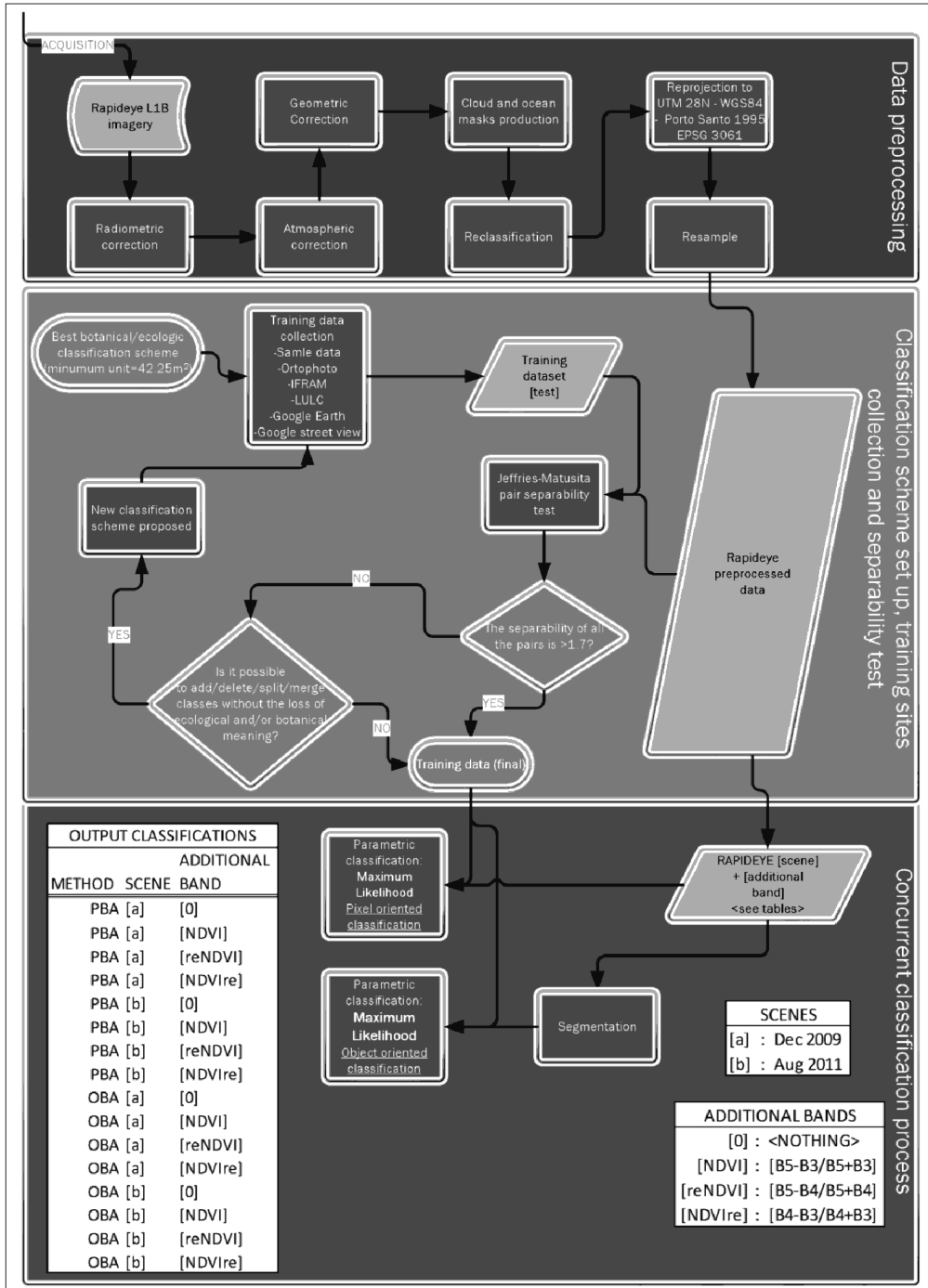


Figure 2 - Schematic representation of the methodological workflow followed in this study.

### ***Preprocessing***

The two RapidEye images were converted to top-of-atmosphere reflectance (radiometric correction), and subsequently atmospherically corrected with the Fast Line-of-sight Atmospheric Analysis of Hypercubes (FLAASH) algorithm [Adler-Golden et al., 1998]. FLAASH method is based on the Moderate Atmospheric Transmittance code (MODTRAN) [Abreu and Anderson, 1996; Matthew et al., 2000] and adapted to the observations made by Kaufman et al. [1997]. Both images were ortho-rectified using the Rational Polynomial Coefficients (RPC) provided within the imagery and the digital elevation model layer. Since the result was not accurate as expected, the scenes were further warped with 30 ground control points each, to fit a polynomial second order equation [Richards and Jia, 1999; Jensen, 2005] and further resampled to co-register the two scenes to ensure an effective overlay (geometric correction) - The digital elevation model was obtained by interpolation of the contour data with ANUDEM technique [Hutchinson, 1988]. Cloud masks were manually digitized through on-screen photo-interpretation. The pixels within the ocean and those covered by cloud masks have been reclassified as no data. Shadows originated by clouds and by both relief/sun and relief/sun/sensor geometries have been left to be posteriorly recognized and mapped by the classifier. The imagery was re-projected in the local reference system UTM 28N - Porto Santo 1995 (EPSG 3061)

### ***Classification scheme set up, training sites collection and spectral separability assessment***

An “ideal classification scheme” for vegetation mapping in Madeira Island was initially set up to be consistent in the representation of dominant plant species within their ecological communities for the Island according to the works of Capelo et al. [2004] and Costa et al. [2004]. With every point representing a pixel in the RapidEye scene, hence a given land-cover’s spectral signature, an iterate process was initialized to obtain a training data set formed by at least 45 point elements per class with each class’ spectral response being statistically separable from the others. For every classification scheme since the “ideal”, the iteration thus included two steps: the collection of new point sets for the classification scheme under assessment and the spectral separability assessment of the points for the RapidEye scene. The available surveys developed from 2009 to 2011 by the Madeira Botanical Group were used as training basis for the photointerpretation process. The point elements were collected through visual likelihood of the field survey points within the ortophoto-maps. In addition Google Earth and Google Street View online resources were used as supplementary sources of information where the ortophoto quality was low or the land cover uncertain. The points however were finally accepted only if the result was meaningful with respect to the classes of the Land use Land cover map (LULC) of Madeira and the Madeira Island Forest Inventory [IFRAM, 2008]. The Jeffries-Matusita (*JM*) spectral separability measure was computed for every pair of classes in order to assess the viability of the classification scheme. The Totally Separable class Pairs index (TSP) was calculated as initially proposed by Michelson and Seaquist [1995] and modified with the higher threshold value of 1.7 for considering a pair separable:



$$TSP_{1.7} = \frac{JM_n}{N} \quad [1]$$

where  $JM_n$  are the class pairs with a  $JM$  distance  $\geq 1.7$  (1.4 in the original version of  $TSP$ ) and  $N$  is the total number of class pairs. The higher threshold of 1.7 was introduced to better split problematic from easily separable classes. In Table 2 are reported the main changes operated to the “ideal classification scheme” towards an operational and spectrally viable classification scheme. The resulting sample was finally randomly stratified into 30 points per class for the training data and 15 points per class for ground truth data.

### Data segmentation and classification

The classification step can be divided in three different levels of analysis:

- Winter (December 2009) vs. Summer (August 2011) scenes;
- Pixel-based approach (PBA) vs. object-based approach (OBA);
- Pertinence of using RapidEye-derived vegetation indices as ancillary bands in the classification procedure.

Pixel and object-based Maximum Likelihood supervised classifications were applied to eight initial datasets including solely RapidEye multispectral bands (from December 2009 and August 2011, respectively) and also combinations of these same multispectral bands with three different vegetation indices - the classic NDVI, the already tested reNDVI and a newly tested NDVIre, respectively - in order to test the full potential of the RapidEye red-edge band for vegetation studies, as presented in Table 3. Sixteen different input combinations and, subsequently, classification maps were thus obtained (see output table printed in Fig. 2), assessed and compared.

**Table 3 - Vegetation indices used in this study.**

Index	Acronym	Formula	Reference
<b>Normalized Difference Vegetation Index - Classic version: NIR-red bands</b>	NDVI	$(\text{NIR-Red})/(\text{NIR+Red})$	[Rouse et al., 1974]
<b>Normalized Difference Vegetation Index - var 1 Near infrared to red edge difference</b>	reNDVI	$(\text{NIR-RE})/(\text{NIR+RE})$	[Tucker, 1979; Gitelson and Merzlyak, 1994; Mutanga et al., 2012]
<b>Normalized Difference Vegetation Index - var 2 Red edge to red difference</b>	NDVIre	$(\text{RE-Red})/(\text{RE+Red})$	<b>proposed</b>

NIR: near infrared band, wavelength 760-850 nm, RapidEye band 5; RE: red-edge band, wavelength 690-730 nm, RapidEye band 4; Red: red band, wavelength: 630-685 nm, RapidEye band 3.

To obtain the segmented scenes for the object-based classification, the watershed

segmentation [Roerdink and Meijster, 2000] was performed. Watershed segmentation algorithms require the determination of the best fitting combination of two coefficients, the scale factor and the merge factor within an iterative process aimed to reach a segment size lower than the class object size [Blaschke, 2010]. Several combinations have been assessed until reaching a fitting combination for all the scenes, in order to reduce the disturbance in accuracy due to segmentation parameters. The scale factor was set to 10 and the merge factor to 45.

### ***Accuracy assessment***

The sixteen classifications accuracy results were assessed by overlaying ground truth data to each map to compute confusion matrices, from which statistics have been calculated such as, class-wise quantity disagreement and allocation disagreement [Pontius and Millones, 2011], overall accuracy [Congalton, 1991; Foody, 2002] and Kappa coefficient of agreement [Cohen, 1960]. Quantity disagreement (QD) takes into account the difference in distribution for every class in a set of classes (the classification scheme) between the map in evaluation and the reference given for the accuracy assessment. Allocation disagreement (AD) oppositely takes into account the spatial position mismatch for every class in a set of classes between the map in evaluation and the reference given for the accuracy assessment. Overall accuracy (OA) is the averaged sum of the correct classified pixels proportion per class. Kappa coefficient of agreement ( $\hat{K}$ ) compares the observed proportion of correctly classified pixels to the proportion that would accidentally be classified as correct.

## **Results**

### ***Preprocessing***

The ortho-rectification process-derived RMSE resulted lower than a pixel for both images: 6.21 meters for the scene of December 2009 and 4.84 meters for the scene of August 2011. The unavailable data in RapidEye for 2011 scene amounted to 4 km<sup>2</sup> due to shadow coverage and to 22 km<sup>2</sup> due to cloud coverage, which aggregated correspond to 3.53% of Madeira's land. For 2009 scene, the unavailable data amounted to 98 km<sup>2</sup> due to shadow coverage and to 70 km<sup>2</sup> due to cloud coverage, which aggregated correspond to an unavailable spectral coverage of 22.57% of the island territory. Although the atmospheric correction significantly improved the contrast in the visible wavelengths, an over-reflection of the blue band remained in the December scene due to haze on the eastern part of the Island.

### ***Classification scheme set up, training sites collection and spectral separability assessment***

The final classification scheme was categorized in three levels for a total of 7 first level classes, 12 second level classes and 26 third level classes. The classification scheme is presented in Table 4. A total of 1244 ground truth points have been collected for the December 2009 dataset and a total of 1341 ground truth points have been collected for the August 2011 dataset. The random stratification procedure for the August 2011 datasets left respectively 931 points for classification and 410 points for accuracy assessment. This same procedure for the December 2009 datasets left respectively 851 points for classification and 393 points for accuracy assessment.

**Table 4 (Continued on the next page) - Land-cover classification scheme adopted with brief class description.**

Level #1	Level #2	Level #3	Code	Brief Description
	Water bodies		WaB	Inner artificial and natural water courses, lakes and reservoirs.
Agriculture	Arboriculture	Banana plantation	BaA	Banana crops
		Fruit orchards	FrA	Orchard of citrus ( <i>Citrus</i> spp.), tropical (e.g. <i>Mangifera indica</i> , <i>Carica papaya</i> , <i>Persea Americana</i> , <i>Cyphomandra betacea</i> , <i>Passiflora edulis</i> ) and temperate fruits (e.g. <i>Malus domestica</i> , <i>Prunus</i> spp.)
	Herbaceous crops	Sugar cane	SuA	Sugar cane plantations
		Irrigated crops	IrA	Annual irrigated and flooded crops
	Vineyard	Vineyards	ViA	Vineyards
Forest	Laurisilva forest	<i>Ocotea foetens</i> Laurisilva forest	LstempF	Laurel forest dominated by <i>Ocotea foetens</i> with <i>Laurus novocanariensis</i> and <i>Persea indica</i>
		<i>Apollonias barbujana</i> laurisilva forest	LsmedF	Laurel forest dominated by <i>Apollonias barbujana</i> and <i>Laurus novocanariensis</i>
	Anthropic or synanthropic forest	Chestnut tree forest	ChF	Chestnut tree managed forest
		Eucalyptus forest	EuF	Patches of <i>Eucalyptus</i> spp. dominated woodland to forest
		Pine forest	PiF	Patches of <i>Pinus</i> spp. dominated forest
Woodland	Invasive species woodland	Acacia woodland	AeW	<i>Acacia</i> spp. woodland and isolated trees
	Endemic species woodland	<i>Olea maderensis</i> woodland	OIW	<i>Olea maderensis</i> scattered woodland to isolated trees
		<i>Salix canariensis</i> woodland	SaW	Riparian communities dominated by <i>Salix canariensis</i> trees
		Tree heather woodland	HeW	Woodland of tree heather ( <i>Erica arborea</i> )

**Table 4 (Continued from preceding page) - Land-cover classification scheme adopted with brief class description.**

Level #1	Level #2	Level #3	Code	Brief Description
Shrubland		Shrubland of <i>Arundo donax</i>	ArS	Dense patches of <i>Arundo donax</i> shrubland
		Shrubland of <i>Euphorbia piscatoria</i>	EuS	<i>Euphorbia piscatoria</i> single bushes
		Shrubland of <i>Genista tenera</i>	GeS	Shrub communities dominated by <i>Genista tenera</i> and <i>Teline maderensis</i>
		Shrubland of <i>Ulex latebracteatus</i>	UIS	<i>Ulex europaeus</i> subsp. <i>latebracteatus</i> and <i>Cytisus scoparius</i> subsp. <i>scoparius</i>
		Heath shrubland	HeS	Shrubland dominated by <i>Erica platycodon</i> subsp. <i>maderincola</i>
Grassland	Permanent cultivated pastures	Permanent pastures	PeG	Perennial sown pastures
	Natural grassland	Dry grassland	DrG	Grassland which presents as dry and/or scattered due to the phenological status or temporary unavailability of water (e.g. <i>Hyparrhenia sinaica</i> grassland)
		Wet grassland	WeG	Grassland which presents as wet and generally dense due to the phenological status or temporary availability of water
Non-vegetated areas		Bare soil	Bso	Soil not covered by a vegetation plane
		Bare rocks	Rck	Rocky outcrops and cliffs
		Built-up areas	Bui	Civil, commercial and industrial buildings, roads, infrastructures

In Table 5 are presented the land cover/vegetation classes ordered by frequency of occurrence among the least separable pairs according to Jeffrey-Matusita separability test (*JM* lower than 1.4 and *JM* lower than 1.7, respectively) - This table shows the most problematic classes for the given set of input data.

**Table 5 - Jeffrey-Matusita separability test results for December 2009 and August 2011 imagery: TSP<sub>1,4</sub> and TSP<sub>1,7</sub> resume respectively the percentage of classes with JM separability above the threshold of 1.4 and 1.7, thus the easy-to-separate classes. The lists of the classes with separability below these thresholds are given. Since the JM results are pairwise, every class in a n classes classification scheme has n-1 values of separability. The percentages aside to each class represent the frequency with which each class appears in the group of pairs whose separability is below the thresholds; the highest values represent classes with separability problems towards many classes.**

RapidEye December 2009				RapidEye August 2011			
TSP <sub>1,4</sub> =	94.00%	TSP <sub>1,7</sub> =	82.90%	TSP <sub>1,4</sub> =	94.60%	TSP <sub>1,7</sub> =	87.80%
JM<1.4		JM<1.7		JM<1.4		JM<1.7	
IrA	11.90%	FrA	11.70%	PeG	10.50%	FrA	8.10%
FrA	11.90%	ArS	7.50%	IrA	10.50%	IrA	7.00%
ArS	9.50%	AcW	6.70%	ViA	7.90%	LsmedF	7.00%
SaW	9.50%	IrA	5.80%	FrA	7.90%	AcW	7.00%
PeG	7.10%	PiF	5.80%	AcW	7.90%	WeG	7.00%
PiF	7.10%	LstempF	5.80%	OIW	7.90%	GeS	5.80%
EuF	7.10%	GeS	5.00%	GeS	5.30%	PeG	4.70%
AcW	4.80%	SaW	5.00%	UIS	5.30%	ViA	4.70%
SuA	4.80%	PeG	4.20%	EuF	5.30%	UIS	4.70%
GeS	4.80%	SuA	4.20%	WeG	5.30%	OIW	4.70%
WeG	4.80%	EuF	4.20%	Rck	5.30%	ArS	4.70%
BaA	2.40%	ChF	4.20%	LsmedF	2.60%	SaW	4.70%
Bso	2.40%	WeG	4.20%	ArS	2.60%	Bui	3.50%
ChF	2.40%	BaA	3.30%	Bui	2.60%	BaA	3.50%
Rck	2.40%	OIW	3.30%	Bso	2.60%	HeS	3.50%
LstempF	2.40%	LsmedF	2.50%	LstempF	2.60%	EuF	2.30%
UIS	2.40%	DrG	2.50%	HeW	2.60%	Bso	2.30%
LsmedF	2.40%	HeS	2.50%	PiF	2.60%	SuA	2.30%
		UIS	2.50%	BaA	2.60%	HeW	2.30%
		Rck	2.50%			PiF	2.30%
		Bui	1.70%			Rck	2.30%
		HeW	1.70%			ChF	1.20%
		Bso	0.80%			EuS	1.20%
		EuS	0.80%			WaB	1.20%
		WaB	0.80%			LstempF	1.20%
		ViA	0.80%			DrG	1.20%

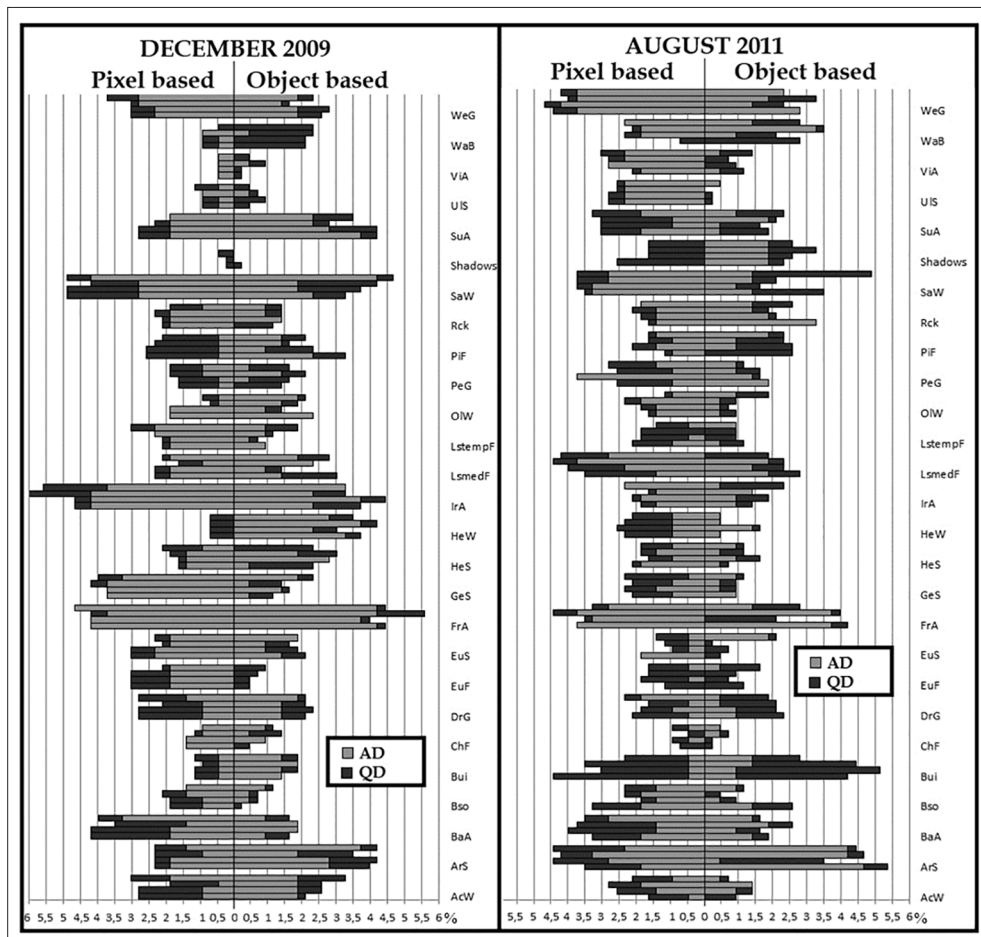
**Data segmentation, classification and accuracy assessment**

In Table 6 are presented the results in terms of OA of the classification and  $\hat{K}$ .

**Table 6 - Overall Accuracy (OA) and Kappa coefficient of agreement ( $\hat{K}$ ) for the processed stacks. OBA: Object Based Approach; PBA: Pixel Based Approach; [a]: RapidEye December scene; [b]: RapidEye August scene. The vegetation index reference is reported in Figure 2.**

Stack	OA	$\hat{K}$
OBA [b]+[NDVI]	74.53%	0.754
OBA [b]+[reNDVI]	74.77%	0.738
OBA [b]+[NDVIre]	73.83%	0.728
OBA [b] +[0]	72.90%	0.718
OBA [a]+[0]	67.52%	0.699
OBA [a]+[NDVI]	67.29%	0.697
OBA [a]+[reNDVI]	66.59%	0.689
PBA [a]+[reNDVI]	65.19%	0.674
OBA [a]+[NDVIre]	65.19%	0.674
PBA [a]+[NDVIre]	64.25%	0.664
PBA [b]+[0]	67.06%	0.657
PBA [a]+[0]	63.32%	0.653
PBA [a]+[NDVI]	63.32%	0.653
PBA [b]+[NDVIre]	66.59%	0.653
PBA [b]+[reNDVI]	65.65%	0.643
PBA [b]+[NDVI]	64.95%	0.636

The best OA and  $\hat{K}$  values have been scored by August scene with an object-based classification approach with the utilization of the additional reNDVI band. In Figure 3 are presented the Allocation and Quantity Disagreement for each class of the scheme for the stacks considered in each scene.



**Figure 3 - Accuracy assessment results per class. December results on the left and August on the right. For each period have been plotted Pixel-based results versus Object-based results. For each class, approach and period, the stack of four horizontal columns represents the Pontius accuracy for - bottom to top - [0], [NDVI], [reNDVI] and [NDVIre]. AD: Allocation Disagreement, QD: Quantity Disagreement.**

In Figure 4 are represented the best land cover/vegetation map produced for December 2009 and August 2011. The map for December was obtained by applying an object-based Maximum Likelihood supervised classification to the dataset including only the RapidEye multispectral bands (OA: 71%) - The map for August was obtained by applying an object-based Maximum Likelihood supervised classification to the dataset including the RapidEye multispectral bands and the reNDVI band (OA: 75%) - The color scheme codifies different saturation levels of the same main hue, for family of covers: greys represent masked pixels and shadows, reds bare and rocky soils and built-up areas, purples agriculture, yellows grasslands, oranges shrublands, light-blues woodlands, greens forests and deep-blue water bodies.

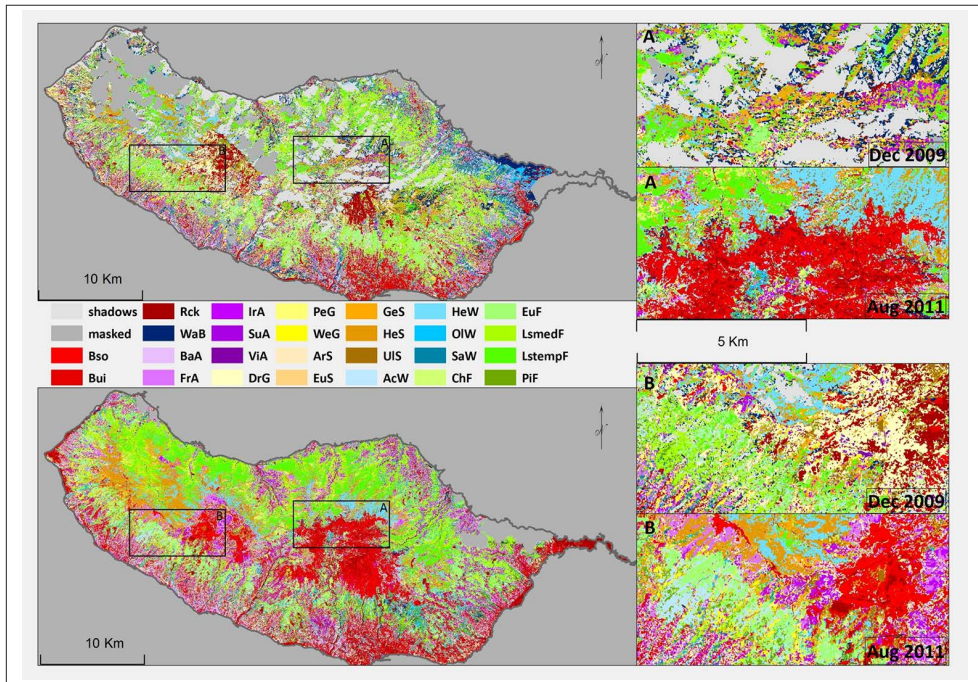


Figure 4 - Best classification maps produced for December 2009 (top) and August 2011 (bottom).

## Discussion

### Preprocessing

The RapidEye imagery 1B processing level is delivered without geometric and atmospheric corrections. The FLAASH model for atmospheric correction is a robust method widely used and accepted by the literature and its usage is relatively simple, requiring only information about the sensor and the atmospheric conditions [some examples: Matthew et al., 2002; Mutanga et al., 2012; Dube et al., 2014; Gerstmann et al., 2016]. The geometric correction instead, for rugged terrains as Madeira Island, requires a substantial user input in the form of ground control points for a polynomial warping procedure [Richards et al., 1999; Jensen, 2005]. Since the provided RPC did not determine a satisfactory result it was necessary to carry it out manually. The Unusable Data Masks (UDM) provided within the imagery data package and delivered as cloud masks, were incomplete, hence both scenes needed an effective cloud masking procedure. Furthermore, some clouds reflectances showed to be practically inseparable from certain artificial materials' saturated spectral responses, as seen for example in Elatawneh et. al. [2014]. The authors thus opted for a manual on-screen digitization for masking the cloud bodies. Regarding the complex shadow coverage across the whole study area (especially for the December 2009 scene, which was acquired with a relevant spacecraft off-nadir view angle, resulting in a huge amount of shaded areas), it was chosen to assume it as a regular class within the classification scheme to be recognized by Maximum Likelihood algorithm. In fact, the variability in size and extension and the fragmentation of the shaded areas across the whole Island made their manual on-screen digitizing too complex and difficult.



### ***Training sample collection and ancillary data issues***

A point sampling-based approach was selected for the training sites collection, in order to fully explore the potential of the available GPS-based botanical surveys developed from 2009 to 2011 by the Madeira Botanical Group. However some vegetation classes have been very difficult to identify and collect, either because mixed, sub-pixel sized or spectrally similar to other classes. Through the *JM* separability test it was possible to assess spectral separability for every pair of land cover/vegetation class signatures, in order to progressively rearrange the classification scheme to guarantee higher spectral homogeneity and coherence within each final class. However, a spectral-based classification scheme definition, due to some similarities between classes (see Tab. 5) may have led to ecologically senseless land cover/vegetation classes. This lower spectral separability between some classes may be mainly derived from the synergistic combination between the medium spectral resolution of RapidEye imagery (despite the existence of the red edge band) and the extensive and ambitious classification scheme (frequently at the plant species/community level) that was adopted in this study. Therefore the final classification scheme alteration was stopped when an ecologically meaningful and an inter-class spectrally separable classification scheme was obtained (i.e. the loop-stopping condition represented in the second part of Fig. 2) - The December 2009 scene was affected by relevant lack of spectral data due to the cloudy and shadowy covers which forced to concentrate the training data in the reduced ground truth areas. For the August 2011 scene, instead, the training data has been collected in a wider range of areas having more ground truth examples to base on, resulting in a more scattered set of points. A more various training dataset gives more flexibility to the classification algorithm, making it able to properly classify pixels whose spectral signature presents slight to relevant differences from the average of the training class. A too various training dataset on the other hand can confuse the algorithm leading to bad classification results. This gives an introduction to the different behaviors of December and August datasets after the increase in variability of the spectral signature through the addition of an ancillary vegetation index band, and after a decrease in variability through a segmentation process.

### ***Seasonal RapidEye data acquisition for improving vegetation mapping accuracy in Madeira Island: winter vs. summer***

General accuracy assessment presented in Table 6 depicts a relevant superiority of the August scene, particularly in the OBA (a OA gap of 5.37% between OBA [b]+[0] and OBA [a]+[0]) - As shown in Figure 3, forest (ChF, LStempF and PiF), woodland (SaW, AcW and OIW) and shrubland (EuS, GeS, HeS and UIS) classes obtained a lower Quantity and Allocation Disagreement in the August period mainly due to the summer phenological stages, which made the vegetation more reflectant, and to better weather conditions. Instead the winter supremacy chiefly regards non-vegetated areas and grasslands. Even though unfavorable phenological stages may have negatively affected the accuracy of some classes in December, we consider its inferiority in accuracy mostly due to the influence of unclear/misty atmosphere that remained in the processed data, even if the 23% of the island had already been excluded from the analysis.

### ***Most suitable supervised classification approach for improving vegetation mapping accuracy in Madeira Island: pixel versus object based classification***

General accuracy assessment presented in Table 6 shows a decisive superiority of the OBA over the PBA (5.84% of OA gap between OBA [b] + [0] and PBA [b] + [0] and 4.2% of OA gap between OBA [a] + [0] and PBA [a] + [0]) - We conclude that these results are due to the enhanced intra-class coherence given by the segmentation process. According to Figure 3, it can be resumed that classes with wide and extended objects score better results in the OBA (AcW, SaW, LStempF, GeS, EuS, EuF, Bso, BaA, IrA and WeG and less noticeably PeG, DrG and ChF) and classes constituted by linear, pointy and very scattered objects perform better in PBA (ArS, Bui and WaB) - The segmentation configuration procedure however may represent a main issue jointly with the training data collection, as both processes are human-driven and directly dependent on the previous and accurate knowledge of the case-study area.

### ***Pertinence of using RapidEye-derived vegetation indices as ancillary bands in the classification procedure for improving vegetation mapping accuracy in Madeira Island***

A vegetation index is a combination of spectral bands that amplifies spectral information of high significance for a distinction of different cover classes [Gerstmann et al., 2016]; the data carried on a vegetation index band may be more correlated to specific vegetation features than the bands that originated it. The vegetation index can be considered a layer carrying information such as biomass [Vuolo et al., 2010; Dube et al., 2014], LAI [Beckschäfer et al., 2014], leaf pigment absorption [Godinho et al., 2016] and can be used as information layer to improve the classification results [Buck et al., 2015]. The 3 vegetation indices tested in this study demonstrated to have generally positive effects in the classification results for an image of good quality . However some controversial results emerged when the vegetation index layer was used to improve the accuracy of an image affected by high atmospheric disturbance. Observing Figure 3, some class accuracy behaviors due to the addition of a vegetation index can be pointed out. A gain in accuracy was determined by a better discrimination between classes of the same genus and characterized by different habitus, at this resolution meaning mainly a difference in canopy density, as observed in HeS and HeW (see the decrease of AD+QD of HeS and HeW after the addition of [NDVIre], above the other vegetation index, in PBA [b] and [NDVI] or [NDVIre] in OBA [a]). Another source of accuracy increase was inspected as the better discrimination among PeG and WeG grassland classes, whose spectral and textural differences may be slight (see the decrease of AD+QD of PeG and WeG after the addition of [reNDVI] in PBA [b], [NDVI] in OBA [b] and [reNDVI] in OBA [a]). Moreover it is observable a possible better detection of vegetated/non-vegetated areas transitions, if taken in consideration for example DrG, a very low density and low reflectance vegetation class (or EuS, who loses the leaves in summer) and Bso (see the decrease of AD+QD of DrG, EuS and Bso after the addition of [NDVI], [reNDVI] or, less markedly, [NDVIre] for PBA [b], [reNDVI], [NDVIre] or [NDVI] for OBA [b] and [reNDVI] or [NDVIre] for PBA [a]). Furthermore the addition of a vegetation index band showed to be useful into rising up the accuracies of the least separable classes for both periods: the classes with at least 5% of appearance in the least separable classes (with  $JM < 1.4$ ) in Table 5 show an increased accuracy in Figure 3. [NDVI] showed to increase accuracy more in the OBA (EuF, FrA, OIW, PeG, Rck, ViA and WeG in [b] and FrA and PiF

in [a]), than in the PBA (only FrA in [b]). [reNDVI] showed to increase the accuracy of the least separable classes slightly more in the OBA (EuF, FrA, PeG, Rck, UIS and ViA in [b] and ArS, IrA and PiF in [a]) than in the PBA (IrA, UIS and WeG in [b] and PiF and SaW in [a]). [NDVIre] showed to increase the accuracy of the least separable classes slightly more in the OBA (AcW, FrA, PeG, Rck and WeG in [b] and IrA and PiF in [a]) than in the PBA (FrA, OIW, UIS and WeG in [b] and EuF and PiF in [a]). The addition of a vegetation index guaranteed an increase in general accuracy (see Table 6) in OBA [b] (all vegetation index sets better than [0]) of 1.63% of OA for the best set, [NDVI] and in PBA [a] ([reNDVI] and [NDVIre] better than [0] and [NDVI] worse than [0]) of 1.87% of OA for the best set, [reNDVI]. We assume that the negative results resulting from the addition of a vegetation index band to the OBA [a] and PBA [b] sets are due to different reasons: the OBA [a] set may have suffered for a too high concentration of the training data, resulting in very little variable class' spectral signature (because collected in the segmented objects with an average reflectance value) combined with an aberrant vegetation index value (this at pixel level) due to atmospheric disturbance; instead PBA [b] set with the vegetation index additional band may have confused the algorithm with too much variance of the spectral signature. However, the decrease in accuracy in PBA [b] with vegetation index additional band can be considered extremely slight (0.47% of OA for the best set, [NDVIre]) as that in OBA [a] (all vegetation index sets worse than [0]) of 0.23% of OA for the best set, [NDVIre].

### ***Main remarks on the output maps***

The main objective of this work was to develop and assess a methodological framework to produce a RapidEye multispectral imagery-based vegetation map. The OA achieved by the best classifications per period (74.8% for OBA [b] + [reNDVI] and 67.52% for OBA [a] + [0]) are high enough to be compared to other supervised classifications of RapidEye scene results found in literature such as in Adam et al. [2014] was scored an OA of 93.07% for an 11 classes scheme and in Polychronaki et al. [2015] were scored OAs of 81% to 86% for a 6 classes scheme. The visual analysis of the best December 2009 classification map in Figure 4 recalls immediately the high disturbance given by the clouds and the shadows covers. In this map an uninterrupted cover is classified only for the South side and the North-East coast of the island. The best August 2011 classification map (Fig. 4), with far less masked areas, was nevertheless less accurate in detecting shaded areas. However, it constitutes a most reliable output for mapping and analyzing vegetation patterns in Madeira Island. This output shows clearly the laurel forest patterns on the North side, interrupted on the lower altitudes. Particularly interesting in terms of landscape evolution patterns recognition, is the detection of the Cural das Freiras fire scar, represented magnified in frame "A" of Figure 4. The fire, visible also in a MODIS scene at an informative NASA site [Nasa Visible Earth, 2010] developed in 2010 and burnt endemic shrubland, woodlands and forests [Fontinha et al., 2014]. Although the masked pixels due to shadow covered land and the misclassified deep blue pixels (shadows recognized as water bodies) present both in the 2009 and in the 2011 maps, the transition between vegetated to bare areas due to the wildfire is evident in frame "A" in a pattern West/East and allows some speculation about the main vegetation groups wiped out by the fire event. Even if in 2009 map are only visible the tallest relieves successively affected by the fire (compare Fig. 4 and Fig. 1), they are recognizable, from West to East, a large patch of laurisilva forests, heath shrublands, eucalypt forests and mixed laurisilva forest with heath

shrubland and agricultural land, that may be misclassified humid grassland. In frame “B” it is well shown the seasonal (it is assumed to be non-permanent changes, only related to the season) land cover transition on the western plateau of Paul da Serra. In December a large part of the plateau is covered in grassland while in August it disappears for bare soil. The heat is dominant in the top part of frame B and while in December is mainly detected as woodland, in August it is detected as shrubland; this is probably due to different vegetative status in the two periods. The green forest strip top left to bottom seems to evolve from mixed endemic laurisilva to managed eucalypt. Such change behavior is not uncommon observing the two classifications, and it would hold a certain relevance if confirmed by future studies for an assessment of the forest management policies. The authors nevertheless prefer to consider the data with prudence due to difference in vegetative status possible misclassification. Furthermore, as expected, many bare areas (in red) during the winter, then turn in to agriculture lands during the summer (in purple). Some grasslands and pastures, inversely disappear with the warm season (an example is represented by the dry grassland around Ponta do Pargo, the westernmost cape of the island). Generally the natural vegetation follows an altitude gradient rule, especially on the southern coast, with three main strips: agriculture, pine forest and eucalyptus forest. The northern side of the island is dominated by the laurel forest in the wild steep valleys and diversely covered in the flat areas (where it is visible a seasonal change from bare to agriculture areas). At the higher altitudes, plateaus and mountains are covered by herbaceous, bushy and woody native/endemic species and by the invasive *Acacia* woodlands. According to the Pontius class accuracy in Figure 3, FrA are largely erroneously classified; this phenomenon is probably due to the high variety of fruit tree kinds and the related differences in phenology, plant density and vegetation spectral reflectance. AcW class has been found mainly mixed thus extremely difficult to sample. SaW low accuracy in the best December 2009 classification map may be explained by the riparian nature of the class, which affected the sampling process, frequently occurring covered by slope-projected shadows in this RapidEye scene. Wbo are largely misclassified because of their linear nature with widths that often approximate the pixel size. Jointly, the high reflective surrounds of rocky beds, which frequently results in water-rock mixed pixels, negatively affect the classifier response, both in PBA and OBA. Furthermore some shaded areas are commonly misclassified as water bodies. SuA classification quality suffers mainly from a spectral likelihood to ArS. This is possibly due to two different reasons: (1) they have the same name in Portuguese (“cana”) which may have made these classes subject to on-screen digitizing misinterpretations in previous works; (2) the difficulty of collecting ArS training data because it is usually used as hedgerow along the border of agricultural land, therefore with very narrow (sub-pixel) width. LSmedF achieved low accuracy values because it was not possible to find large patches of this class and therefore it was difficult to collect related training data.

## Conclusions

The geographic and atmospheric conditions of Madeira Island turn particularly challenging any attempt of remote sensing-based vegetation mapping of this insular territory, namely by using high spatial resolution RapidEye imagery. Some examples are high cloud coverage, with shadow projection caused by relieves, high clouds and off-nadir sensor position. Moreover the topographic complexity of the island consists in a major challenge for the geometric correction. The Madeira’s unique interface climate pushed for an analysis both

of summer and winter imagery. The phenological differences of the vegetation showed not to consist in a multi-temporal series classification issue, if taken in consideration also that most of the natural/protected vegetation grows on the North side of the island, where the humidity is more constant. Nevertheless a main problem remains the availability of quality picture in the winter season considering the oceanic aerosol as further disturbance factor in addition to the clouds and the mist. Although the high spatial resolution would suggest the possibility of a singular tree-by-tree detection (whether the crown is twice the pixel size), we demonstrate that for a relatively important area as Madeira Island, with many spectral classes, considering the high radiometric depth of the RapidEye data, the best results are given by applying an object-based approach to reduce the classifier uncertainty. The five spectral bands, with the fourth placed on the wavelength of the vegetation red-edge spectral response, demonstrate to give accurate details for vegetation mapping. The use of a vegetation index stacked with the RapidEye dataset (pixel-sized and segmented) for improving the classification accuracy returned interesting effects of improvement of the accuracy of the statistically least separable classes, of different classes of plants from the same genus, of very low-reflectance and spread classes (as the dry grassland in summer) on the bare soil and classes whose differences may be slight but economically and practically necessary (as annual grassland and pastures). The addition of a vegetation index showed to be nevertheless a delicate phase whose results depend mainly on the quality of the imagery and training data. The maps produced overcome the previous existing cartography offering a huge variety of uses. We find relevant to stress the correlation between training step, which means certain knowledge of the land cover and accuracy of the classification; as observed in this research we conclude that the prior knowledge (and its resolution) of the cover in the training areas is the first driver of the map accuracy. Regarding the RapidEye imagery we recommend a careful selection of the scenes, with a particular attention to the off-nadir position of the sensor, because can lead to a massive data loss especially in oceanic islands whereas the topography is rugged.

### **Final remarks**

The main objective of this work was to develop and assess a methodological framework to produce a RapidEye multispectral imagery-based vegetation map. RapidEye data demonstrated to provide a high value product allowing the classification of a high resolution scheme with the use of only one scene, achieving reasonable accuracy levels. The addition to the processing of a vegetation index showed potential for the segmentation and classification; however the results disagree and it is needed further study to confirm this point. Two different seasons have been tested for a classification; the comparison of the results showed different seasonal behavior of some classes, particularly grassland and agriculture land. The authors recommend therefore, depending on the purpose of the map to be produced, caution to choose an image acquisition date in which those classes are fully represented. Segmented inputs overcame pixel based approaches demonstrating a necessity of homogenizing spectral signatures from this sensor for the maximum likelihood classifier. The authors additionally recommend taking in consideration the team pre-processing capabilities before acquiring a non-geometrically corrected dataset because the amount of labor needed for these preliminary adjustments can easily over-pass the price difference between a L1B and L3A product.

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## Appendices

**Appendix 1 (Continued on the next page) - Main actions operated to the original classification scheme through the JM separability assessment and training data adjustment iterative process to reach the final classification scheme.**

Ideal classification scheme	Action	Final classification scheme
Natural water streams	Merged <sup>(1)</sup>	Water bodies
Coastal ponds		
Natural inner lakes and ponds		
Artificial inner lakes and ponds		
Fruit orchards	Splitted <sup>(2)</sup>	Fruit orchards
		Banana plantation
Irrigated crops	Splitted <sup>(3)</sup>	Irrigated crops
		Sugar cane
non-irrigated crops	Removed <sup>(4)</sup>	-
Vineyards	none <sup>(5)</sup>	Vineyards
Apollonias barbujana laurisilva forest	Merged <sup>(6)</sup>	Apollonias barbujana laurisilva forest
Open <i>Apollonias barbujana</i> laurisilva forest		
<i>Ocotea foetens</i> Laurisilva forest	Merged <sup>(7)</sup>	<i>Ocotea foetens</i> Laurisilva forest
Open <i>Ocotea foetens</i> laurisilva forest		
<i>Persea indica</i> laurisilva forest	Removed <sup>(8)</sup>	-
Chestnut tree forest	Merged <sup>(9)</sup>	Chestnut tree forest
Chestnut tree forest with broad-leaf trees		
Chestnut tree forest with resinous trees		
Open chestnut tree forest		
<i>Eucalyptus</i> spp. forest	Merged <sup>(10)</sup>	<i>Eucalyptus</i> spp. forest
<i>Eucalyptus</i> spp. forest with broad-leaf trees		
<i>Eucalyptus</i> spp. forest with resinous trees		
Open <i>Eucalyptus</i> spp. forest		
Open <i>Eucalyptus</i> forest with resinous trees		

**Appendix 1 (Continued from preceding page and on the next page) - Main actions operated to the original classification scheme through the *JM* separability assessment and training data adjustment iterative process to reach the final classification scheme.**

Ideal classification scheme	Action	Final classification scheme
Pine forest	Merged <sup>(11)</sup>	Pine forest
Pine forest with resinous trees		
Open pine forest		
Open pine forest with broad-leaf trees		
Forest of other broad-leaf trees	Removed <sup>(12)</sup>	-
Forest of other resinous trees	Removed <sup>(13)</sup>	-
Open forest of other broad-leaf trees	Removed <sup>(14)</sup>	-
Open forest of other resinous trees	Removed <sup>(15)</sup>	-
Community of <i>Acacia dealbata</i>	Merged <sup>(16)</sup>	Acacia woodland
Community of <i>Acacia mearnsii</i>		
Community of <i>Acacia melanoxylon</i>		
Open community of <i>Acacia mearnsii</i>		
Open community of <i>Acacia melanoxylon</i>		
<i>Olea maderensis</i> woodland	Merged <sup>(17)</sup>	<i>Olea maderensis</i> woodland
Open <i>Olea maderensis</i> woodland		
<i>Salix canariensis</i> woodland	none <sup>(18)</sup>	<i>Salix canariensis</i> woodland
Tree heather woodland	Merged <sup>(19)</sup>	Tree heather woodland
Open tree heather woodland		
Community of <i>Pittosporum undulatum</i>	Removed <sup>(20)</sup>	-
Open community of <i>Pittosporum undulatum</i>	Removed <sup>(21)</sup>	-
Community of <i>Arundo donax</i>	none <sup>(22)</sup>	Community of <i>Arundo donax</i>
Community of <i>Euphorbia piscatoria</i>	Merged <sup>(23)</sup>	Community of <i>Euphorbia piscatoria</i>
Open community of <i>Euphorbia piscatoria</i>		
Community of <i>Genista tenera</i>	none <sup>(24)</sup>	Community of <i>Genista tenera</i>
Community of <i>Ulex latebracteatus</i>	none <sup>(25)</sup>	Community of <i>Ulex latebracteatus</i>
Heath shrubland with <i>Myrtus</i> spp.	Merged <sup>(26)</sup>	Heath shrubland
Heath shrubland with <i>Vaccinium</i> spp.		
Open heath shrubland with <i>Myrtus</i> spp.		
Open heath shrubland with <i>Vaccinium</i> spp.		
<i>Myria faya</i> Shrubland	Removed <sup>(27)</sup>	-
<i>Sideroxylum mirmolano</i> community	Removed <sup>(28)</sup>	-
Permanent pastures	none <sup>(29)</sup>	Permanent pastures

**Appendix 1 (Continued from preceding page) - Main actions operated to the original classification scheme through the JM separability assessment and training data adjustment iterative process to reach the final classification scheme.**

Ideal classification scheme	Action	Final classification scheme
High altitude herbaceous community	Rearranged <sup>(30)</sup>	Dry grassland
Low altitude herbaceous community		
Invasive herbaceous community		Wet grassland
Built-up areas	Merged <sup>(31)</sup>	Built-up areas
Road system and associated areas		
Reservoirs and cisterns		
Beach, dunes and coastal areas	Removed <sup>(32)</sup>	-
Burnt areas	Merged <sup>(33)</sup>	Bare rocks
Quarries		
Bare rock		
-	Added <sup>(34)</sup>	Bare soil

(1)(20)(27)(28) scarce availability of data for comparison due to low spatial frequency of the object and/or high pixel mixture due to very scattered class combined with homogeneous patch size smaller than one pixel; (2)(3)(34) existence of stand-alone spectral natural aggregations in the previous class(es) which matched to the interpreter's assessment; (4)(8)(12)(13)(14)(15)(21) unavailability of photo-interpreted data for comparison and impossibility to distinguish the class from others; (5)(18)(22)(24)(25)(29) no action required; (6)(7)(9)(10)(11)(16)(17)(19)(23) (26) dense, open and/or mixed classes have non relevant spectral difference whether the single object roughly approximates the pixel size; (30) unavailability of photo-interpreted data for comparison and impossibility to distinguish the class from others but existence of stand-alone spectral natural aggregations in the previous classes which matched to the interpreter's assessment; (31) merged into anthropic impervious areas; (32)(33) non-relevant class differences; spectral inseparability of the class due to lack of information in the medium infrared; (34) class mainly consequent to seasonal changes between vegetated and non-vegetated areas. Spectrally separable from the others non-vegetated areas.

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