

1 **Preparation and evaluation of micro and meso porous silica monoliths with embedded**
2 **carbon nanoparticles for the extraction of non-polar compounds from waters**

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7 **Abstract**

8 A novel hybrid micro and meso porous silica monolith with embedded carbon nanoparticles
9 (Si-CNPs monolith) was prepared inside a fused silica capillary (3 cm in length) and used
10 as a sorbent for solid-phase microextraction. The hybrid monolithic capillary was
11 synthesized by hydrolysis and polycondensation of a mixture of tetraethoxysilane (TEOS),
12 ethanol, and three different carbon nanoparticles such as carboxylated single-walled carbon
13 nanotubes (c-SWCNTs), carboxylated multi-walled carbon nanotubes (c-MWCNTs), and
14 oxidized single-walled carbon nanohorns (o-SWNHs) via a two-step catalytic sol-gel
15 process. Compared with silica monolith without carbon nanoparticles, the developed
16 monolithic capillary column exhibited a higher extraction efficiency towards the analytes
17 which can be ascribed to the presence of the carbon nanoparticles. In this regard, the best
18 performance was achieved for silica monolith with embedded c-MWCNTs. The resulted
19 monolithic capillaries were also characterized by scanning electron microscopy (SEM),
20 elemental analysis and nitrogen intrusion porosimetry. Variables affecting to the
21 preparation of the sorbent phase including three different carbon nanoparticles and
22 extraction parameters were studied in depth using polycyclic aromatic hydrocarbons

23 (PAHs) as target analytes. Gas chromatography-mass spectrometry was selected as
24 instrumental technique. Detection limits range from 0.1 to 0.3 $\mu\text{g}\cdot\text{L}^{-1}$, and the inter-
25 extraction units precision (expressed as relative standard deviation) is between 5.9 and 14.4
26 %.

27 **Keywords**

28 Silica monolith, solid-phase microextraction, polycyclic aromatic hydrocarbons, carbon
29 nanoparticles, monolithic hybrid solid.

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42 **1. Introduction**

43 Polycyclic aromatic hydrocarbons (PAHs) are a group of non-polar organic compounds
44 consisting of two or more benzene rings which are byproducts of petroleum processing or
45 combustion of hydrocarbons. PAHs are of environmental concern due to their significant
46 toxicity and potential carcinogenic properties at relatively low levels [1]. In this regard, the
47 concentration of benzo[a]pyrene has been used as indicator of total contamination by PAHs
48 being $0.2 \mu\text{g}\cdot\text{L}^{-1}$ the maximum level established by Agency for toxic substances & Disease
49 Registry (ATSDR) in waters [2].

50 Analytical methods including gas chromatography (GC) [3], gas chromatography coupled
51 with mass spectrometry (GC-MS) [4], high-performance liquid chromatography with
52 fluorescence detection (HPLC-FLD) [5], thin-layer chromatography (TLC) [6], and
53 spectrofluorimetry [7] have traditionally been used to determine PAHs. However,
54 preconcentration and clean-up steps are needed to achieve the required sensitivity and
55 selectivity. In this context, new microextraction techniques have been recently developed
56 based on the simplification and the miniaturization of classical separation techniques in
57 both solid and liquid phase formats. The incorporation of nanostructured materials as well
58 as hybrid sorbents is behind the consolidation of these new approaches particularly due to
59 their high superficial area and variety of interactions [8].

60 Monolithic solids have gained prominence as new separation material because of their
61 unique properties including fast dynamic transport, simplicity of their preparation, frit-free
62 construction, good loading capacity and low backpressure. To date, both organic polymer
63 and silica-based monoliths have been used in several formats in the extraction context such

64 as micropipette-tips, spin columns, microfluidics chips, and capillary columns [9, 10].
65 Carbon nanoparticles, especially carbon nanotubes, have received special research attention
66 since their discovery thanks to their unique and outstanding properties. Their excellent
67 properties in terms of extraction efficiency have resulted in their use in different
68 microextraction formats [11]. Hybrid monoliths prepared via sol-gel from silica precursors
69 and organic frameworks including carbon nanoparticles combine the advantages of both
70 [12-14]. The main challenge is to achieve a good and stable dispersion of the carbon
71 nanoparticles in the polymerization mixture in order to obtain a homogeneous solid.
72 In this article, micro and meso porous silica monoliths with embedded carbon nanoparticles
73 have been successfully synthesized inside a fused silica capillary and evaluated for the
74 determination of non-polar compounds such as PAHs in tap and river water samples. Gas
75 chromatography-mass spectrometry was selected as instrumental technique. All the
76 variables related to the preparation of the monolith as well as those affecting to the
77 microextraction process were evaluated in depth. In addition, the monolithic solid with and
78 without carbon nanoparticles was characterized by elemental analysis to confirm their
79 presence into the silica network and a porosimetry study was also carried out in order to
80 know the size distribution of the pores.

81

82 **2. Experimental section**

83 *2.1 Reagents, materials and samples*

84 The reagents used were of analytical grade or better. Polycyclic aromatic hydrocarbons
85 (naphthalene, phenanthrene, fluoranthene, pyrene, benz[a]anthracene and benzo[a]pyrene)
86 were purchased from Sigma–Aldrich (Madrid, Spain. <http://www.sigmaaldrich.com>). Stock
87 standard solutions of each analyte were prepared in methanol (Sigma-Aldrich) at a

88 concentration of $1 \text{ g}\cdot\text{L}^{-1}$ and stored at $4 \text{ }^\circ\text{C}$. Working solutions were prepared by dilution of
89 the stocks in ultrapure Milli-Q water (Millipore Corp.; Madrid, Spain). In the extraction
90 procedure, methanol from Sigma-Aldrich was employed as eluent.

91 Uncoated fused-silica capillaries ($320 \text{ }\mu\text{m}$ i.d., Sigma Aldrich) were used for the
92 preparation of the monolithic extraction unit. Ferrules 1/16" ID, PEEK tubing 1/16" and
93 internal union zero volume 1/16" to 1/16" (Sigma-Aldrich) were also employed.

94 The reagents needed for the synthesis of the monolithic phase, tetraethyl orthosilicate
95 (TEOS), ethanol, Milli-Q water, 3-(trimethoxysilyl)propyl methacrylate, methanol, sodium
96 hydroxide (NaOH), hydrochloric acid (HCl), acetone and acetic acid were purchased from
97 Sigma-Aldrich.

98 Carboxylated multi-walled carbon nanotubes (c-MWCNTs, $< 8 \text{ nm}$ o.d., $10\text{-}30 \text{ }\mu\text{m}$ length,
99 $> 95 \text{ wt}\%$ purity, $3.86 \text{ wt}\%$ functional content) and carboxylated single-walled carbon
100 nanotubes (c-SWCNTs, $1\text{-}2 \text{ nm}$ o.d., $5\text{-}30 \text{ }\mu\text{m}$ length, $>90 \text{ wt}\%$ purity, $2.73 \text{ wt}\%$ functional
101 content) were obtained from Sigma-Aldrich. Single-walled nanohorns were purchased from
102 Carbonium S.r.l (Padua, Italy. <http://www.carbonium.it/public/site/index.php>). SWNHs
103 form stable dahlia-shaped aggregates with an average diameter of $60\text{-}80 \text{ nm}$. Individually,
104 the lengths of these SWNHs are in the range of 40 to 50 nm , and the diameter in the
105 cylindrical structure varies between $4\text{-}5 \text{ nm}$.

106 The dispersions of the carboxylated carbon nanotubes (c-SWCNTs and c-MWCNTs) were
107 made in ethanol. In brief, 5 mg of c-MWCNTs were weighed, added to a glass vial and
108 ultrasonic-assisted dispersed in 50 mL of ethanol for 30 min . On the other hand, carbon
109 nanohorns (SWNHs) were weighed (5 mg) and added to a glass vial, which was further
110 introduced into a microwave oven, being the solid irradiated at 800 W for 10 min . After

111 cooling at room temperature, the oxidized carbon nanohorns (o-SWNHs) were dispersed in
112 50 mL of ethanol and sonicated for 30 min.

113 Tap and river water samples were selected for the determination of the target compounds
114 using the monolithic microextraction unit. Water samples from the Guadalquivir river were
115 collected in amber glass bottles without headspace. The spiked samples were prepared by
116 adding the analytes at a concentration of $20 \mu\text{g}\cdot\text{L}^{-1}$, and then they were left to stand for 24 h
117 until the analysis.

118 *2.2 Chromatographic analysis*

119 Gas chromatographic/mass spectrometric analyses were carried out on a gas chromatograph
120 (Varian CP-3800)-mass spectrometer (Varian 1200 MS/MS) working under single
121 quadrupole mode and with an electron multiplier detector. The gas chromatograph was
122 equipped with a fused silica capillary column VF-5 ms (30 m x 0.25 mm i.d.) coated with 5
123 % phenyl-95 % dimethylpolysiloxane (film thickness 0.25 μm) (Sigma-Aldrich) to separate
124 the six analytes. System control and data acquisition was achieved with an HP1701CA MS
125 ChemStation software.

126 The column temperature program was as follows: the initial temperature, 80 °C, was kept
127 for 2.5 min, raised up to 200 °C at $25 \text{ }^\circ\text{C}\cdot\text{min}^{-1}$ (maintained for 1 min) and further to 250 °C
128 at $10 \text{ }^\circ\text{C}\cdot\text{min}^{-1}$. Then it was immediately ramped up to 285 °C at $5 \text{ }^\circ\text{C}\cdot\text{min}^{-1}$. The final
129 temperature, 300 °C, was reached with a ramp of $30 \text{ }^\circ\text{C}\cdot\text{min}^{-1}$ and maintained for 1 min.
130 The quadrupole mass spectrometer detector was operated in selected ion monitoring mode,
131 recording the following fragment-ions characteristic of each analyte: 128 (from 3.0 to
132 6.4 min) for naphthalene, 178 (from 6.4 to 10.58 min) for phenanthrene, 202 (from 10.58 to

133 14.0 min) for fluoranthene and pyrene, 228(from 14.0 to 17.0 min) for benz[a]anthracene,
134 252 (from 17.0 to 21.0 min) for benzo[a]pyrene, all of them at 1 scan/s. Electron impact
135 ionization (70 eV) was used for analytes fragmentation. The injector temperature was 270
136 °C and the splitless mode was selected. The injection volume, 2 μL of methanol, was
137 measured with a 5 μL microsyringe (Hamilton Co., Nevada, USA). The carrier gas used
138 was helium (6.0 grade, Air Liquide, Seville, Spain) at a flow rate of $1.0 \text{ mL}\cdot\text{min}^{-1}$, and it
139 was regulated by digital pressure controller. The transfer line and ionization source were
140 maintained at 280 °C and 250 °C, respectively.

141 *2.3 Preparation of Si-CNPs monolithic solid*

142 It is well known that silica monoliths tend to shrink during the post-heat treatment, even in
143 small devices such as capillaries. However, this shrinkage can be minimized or prevented
144 by a covalent attachment of the monolith to the capillary inner walls. Therefore, the hybrid
145 monolith was synthesized inside a fused silica capillary (3 cm in length) which was
146 previously modified with 7% solution of 3-(trimethoxysilyl)propyl methacrylate in ethanol
147 [15].

148 A schematic representation of the synthesis procedure is shown in the Fig. 1. The hybrid
149 monolith was prepared by hydrolysis and polycondensation of precursors via a two-step
150 catalytic sol-gel process according to a previously described procedure with some
151 modifications [16]. It should be noticed that the rate of hydrolysis and condensation of the
152 precursors is the most important factor to control the final monolithic structure. The
153 kinetics of each step of the sol-gel process is pH-dependent. Low pH values produce a fast
154 hydrolysis of alkoxysilanes generating hydroxyl groups, while high pH values accelerate
155 the condensation reactions [16]. This pH variation was regulated by adding hydrochloric

156 acid (0.5 M) or ammonium hydroxide (0.5 M) as required In the acid-catalyzed hydrolysis
157 step, 160 μL of the TEOS was mixed with a solution of 200 μL of ethanol containing one
158 of the three carbon nanoparticles studied ($0.1 \text{ g}\cdot\text{L}^{-1}$), 20 μL of Milli-Q water, and 10 μL of
159 HCl (0.5 M) in a 1.5 mL eppendorf vial. The pre-polymerization mixture solution was
160 sonicated at 20 °C for 4 h, and then 25 μL of ammonium (0.5 M) were added into the
161 solution. After shaking the mixture, the modified capillaries were filled with the
162 homogenous polymerization mixture up to a length of 3 cm, and sealed at both ends with
163 two pieces of rubbers. Next, the capillaries were introduced into an oven at 40 °C for 12 h.
164 After polymerization and using a micro-HPLC pump, the resulting columns were rinsed
165 with ethanol to remove any possible unreacted monomer and the soluble hydrolysis
166 products. Then, the monolithic capillary was flushed with a 5 % (v/v) solution of acetic
167 acid in methanol for 10 min ($0.1 \text{ mL}\cdot\text{min}^{-1}$), followed by Milli-Q water ($0.1 \text{ mL}\cdot\text{min}^{-1}$, 10
168 min) in order to conditioning the surface monolith prior to use.

169

170 *2.4 Characterization*

171 A JEOL JSM 6300 scanning electron microscopy (Isaza, Alcobendas, Spain) was also used
172 to obtain the micrographs of the monolithic solid with and without carbon nanoparticles to
173 evaluate the morphological and structural characteristics of the monolithic sorbent. The
174 monolith was fixed on the stub by a double-sided sticky tape and then coated with gold.
175 Transmission electron microscopy (TEM) images were recorded by use of a JEOL JEM
176 1400 microscope (Isaza, Alcobendas, Spain) operating at an accelerating voltage of 120
177 kV. The TEM micrographs were used to characterize the nanoparticles size and
178 morphology.

179 Elemental analysis was carried out on an elemental analyzer CHNS Eurovector EA 3000.
180 Nitrogen adsorption/desorption experiments were carried out at -196 °C using a
181 Quantachrome® ASiQwin™-Automated Gas Sorption Data. The specific surface area
182 values were calculated according to the BET (Brunauer–Emmett–Teller) equation. T-plot
183 method was used to determine the micropore surface areas, and the average pore volumes
184 were evaluated from the desorption branches of isotherms based on the BJH (Barrett–
185 Joyner–Halenda) model.

186 *2.5 Analytical procedure*

187 The microextraction procedure followed for the isolation and preconcentration of PAHs
188 from waters comprised the following steps. First, aliquots of 2 mL of aqueous standards or
189 water samples, containing the target analytes, were passed through the polymer monolithic
190 capillary at a flow rate of 0.1 mL·min⁻¹ by means of the micro-HPLC pump. The monolith
191 was washed with 500 µL of Milli-Q water (0.1 mL·min⁻¹), and then the aqueous phase
192 remaining in the capillary was eliminated by means of a nitrogen stream (10 min). After
193 this step, the analytes were eluted pumping methanol (500 µL, 0.1 mL·min⁻¹), and the
194 eluent was collected into a vial, evaporated and the residue was redissolved in 10 µL of
195 methanol. Finally, 2 µL of the organic phase with the extracted analytes were analyzed by
196 GC/MS. The chromatographic peak areas were used as analytical signals. Between
197 extractions, the monolith was sequentially conditioned with a 5 % (v/v) solution of acetic
198 acid in methanol and Milli-Q water (10 min in each case) at a flow rate of 0.1 mL·min⁻¹.
199 No degradation effects or deformation of the monolithic structure was observed in terms of
200 extraction efficiency with repetitive microextraction cycles. The extraction unit can be used
201 repeatedly owing to its compatibility with organic solvents as well as water [17].

202 **3. Results and discussion**

203 *3.1 Variables affecting to the preparation of the monolith*

204 The preparation of the bare monolith was carried out as described in section 2.3. The
205 incorporation of carbon nanoparticles (c-SWCNTs, c-MWCNTs and o-SWNHs) in the
206 monolith was deeply studied. As it was previously mentioned, the homogenous and stable
207 dispersion of the nanoparticles is crucial for this aim. Therefore, dispersions of each carbon
208 nanoparticle were prepared at a concentration of $0.1 \text{ g}\cdot\text{L}^{-1}$ and several organic solvents
209 (methanol, ethanol and 2-propanol) were studied to efficiently disperse the carbon
210 nanoparticles. Fig. 2 shows the behavior for each carbon nanoparticle in the different
211 organic media. As it can be seen, ethanol resulted to be the best choice to disperse the
212 carboxylated carbon nanotubes (Fig. 2A and 2B), while good dispersion of o-SWNHs was
213 obtained in the three solvents (Fig. 2C). Therefore, ethanol was selected as optimum
214 solvent for the efficient dispersion of the carbon nanoparticles used in this article. Fig. 2D
215 also presents the TEM micrographs obtained for the dispersions of c-SWCNTs, c-
216 MWCNTs, and o-SWNHs.

217 Next, the influence of these carbon nanoparticles incorporated to the monolith was
218 evaluated in terms of extraction efficiency using aqueous standards containing
219 benzo[a]pyrene as model compound at a concentration of $1 \text{ }\mu\text{g}\cdot\text{mL}^{-1}$. As it can be seen in
220 Fig. 3, the introduction of the c-MWCNTs, c-SWCNTs, and o-SWNHs into the network of
221 silica monoliths improved the extraction of benzo[a]pyrene in comparison with the bare
222 monolith. In fact, the best performance was achieved using c-MWCNTs due to their large
223 specific surface area and number of sheets. On the other hand, o-SWNHs present an
224 extensive surface area and horn interstices which provide them with an exceptional
225 adsorption capacity. Besides, the oxidation conditions can produce nanowindows,

226 generating additional interaction sites on the conical and curved ends. All this, combined
227 with the formation of stable aggregates improve their extraction capacity explaining the
228 better performance in comparison with c-SWCNTs. Based on the results obtained, this, c-
229 MWNTs were selected for further experiments. Next, the amount of c-MWCNTs was
230 evaluated at concentrations of 0.01, 0.05 and 0.1 g·L⁻¹. Monolithic solid formed at
231 concentrations of nanoparticles higher than 0.1 g·L⁻¹ showed non homogenous distribution
232 of the c-MWCNTs due to the aggregation of the carbon nanoparticles decreasing the
233 extraction efficiency. As it is shown in Fig. 4, the signals increased with the concentration,
234 and therefore a concentration of 0.1 g·L⁻¹ of the c-MWCNTs in ethanol was selected.

235

236 *3.2 Characterization*

237 Once optimized, the synthesized monoliths were characterized by SEM, elemental analysis,
238 and nitrogen adsorption/desorption measurements.

239 Micrographs of the monolithic sorbents were obtained for the bare silica, Si-SWNHs, Si-
240 SWCNTs, and Si-MWCNTs monolith. As it can be seen in the Fig. 5, the silica monolithic
241 solids exhibit homogeneous and porous networks that consisted of interconnected
242 microglobules. It is worth to mention that the graph of the Si-SWNHs monolith (Fig. 5B)
243 shows that SWNHs form spherical aggregates with dahlia-like structures, while the
244 monolithic solids modified with carbon nanotubes (Fig. 5C and 5D) maintain practically
245 the same microscopic morphology as it showed for the bare silica monolith (Fig. 5A).

246 The C% of the bare silica, Si-SWNHs, Si-SWNTs, and Si-MWCNTs monolith was
247 determined by elemental analysis and resulted to be 0.72%, 4.24%, 1.94%, and 2.48%,
248 respectively. The increase in the proportion of the carbon in the synthesized material
249 demonstrated the successful incorporation of the carbon nanoparticles into the silica

250 network. As expected, the highest percentage was found for the single walled carbon
251 nanohorns due to its size and weight.

252 Isotherms of nitrogen adsorption/desorption for each type of synthesized monolithic solid
253 are described in Fig 6. As it can be seen, the isotherms for silica, Si-SWNHs, and Si-
254 SWCNTs monoliths were type I in the BDDT classification [18], while the nitrogen
255 isotherms obtained for the Si-MWCNTs monolith was the type IV, exhibiting H1 hysteresis
256 loops which indicates a mesoporous solid. The BET surface and pore volume values are
257 compiled in the Table 1 together with the micropore surface areas.

258

259 *3.3 Evaluation of the variables affecting to the microextraction process*

260 As it is a novel extraction procedure, an univariate approach was selected to study the effect
261 of each single variable in the extraction process, using aqueous standards containing
262 benzo[a]pyrene as model compound at a concentration of $1 \mu\text{g}\cdot\text{mL}^{-1}$. The initial
263 experimental conditions for the microextraction process were: 3 mL of aqueous standard
264 and an elution volume of 200 μL of methanol. Flow rates used for preconcentration and
265 elution step were fixed at $0.1 \text{ mL}\cdot\text{min}^{-1}$.

266 The sample and eluent volumes are critical parameters for the method sensitivity, since they
267 will determine the preconcentration factor. The sample volume was studied within the
268 interval 1-4 mL. The maximum extraction efficiency for the target triazines was achieved
269 for a volume of 2 mL, decreasing over this value due to a breakthrough effect. The volume
270 of methanol required for analytes elution was studied between 100 and 500 μL . The
271 methanolic extract was collected in a 1.5 mL glass vial, and a volume of 500 μL was
272 necessary to quantitatively elute the target analytes which was fixed as optimum. An
273 evaporation–redissolution step was included in order to reduce the final volume to 10 μL ,

274 thus increasing the method sensitivity. Finally, 2 μL of the organic phase with the extracted
275 analytes were injected into the gas chromatograph/mass spectrometer for their separation
276 and detection. The chromatographic peak areas were used as analytical signals.

277 Between samples, the monolithic capillary was conditioned with a 5 % (v/v) solution of
278 acetic acid in methanol for 10 min. The pH of the solvent is a critical parameter in the
279 conditioning in order to achieve a reproducible elution as underivatized silica exhibit
280 secondary interactions between analytes and negative charged silanol groups at $\text{pH} > 4$ [19].
281 Milli-Q water was sequentially passed through the hybrid monolithic capillary at a flow
282 rate of $0.1 \text{ mL} \cdot \text{min}^{-1}$ for 10 min.

283

284 *3.4 Analytical figures of merit and analysis of water samples* The analytical figures of merit
285 of the proposed method are summarized in Table 2. The calibration curves for PAHs was
286 constructed by using working aqueous standards prepared in duplicate at controlled
287 concentrations which were subjected to the optimized extraction procedure, and 2 μL of the
288 organic extract was injected into the GC/MS for analysis. The corresponding equations
289 were obtained by plotting the peak areas of the characteristic m/z fragment ions against the
290 concentration for each target analyte. The method was characterized on the basis of its
291 linearity, sensitivity, and precision.

292 The corresponding calibration graphs were constructed by extracting in duplicate nine
293 working aqueous standards containing the analytes at concentrations in the range from 0.3
294 to $1000 \mu\text{g} \cdot \text{L}^{-1}$. The sensitivity of the method was evaluated according to the limit of
295 detection (LOD) and quantification (LOQ). LODs, calculated by using a signal-to-noise
296 ratio (S/N) of 3, were $0.1 \mu\text{g} \cdot \text{L}^{-1}$ and the LOQs, calculated by using a S/N of 10, were 0.3
297 $\mu\text{g} \cdot \text{L}^{-1}$ for all analytes. The linearity was maintained at least in the interval $0.3\text{-}1000 \mu\text{g} \cdot \text{L}^{-1}$

298 for all analytes. Figure 7 shows a SIM chromatograph with the different m/z fragments ions
299 obtained after monolithic microextraction procedure of a blank water and a standard sample
300 with the target analytes at $0.1 \mu\text{g}\cdot\text{L}^{-1}$. The precision of the method was studied under
301 repeatability and reproducibility experimental conditions. Repeatability (intra-extraction
302 units conditions), expressed as relative standard deviation (RSD) was calculated from five
303 individual standards prepared at a concentration of $1 \mu\text{g}\cdot\text{mL}^{-1}$ and it was lower than 14.1 %
304 for all the analytes. The reproducibility between extraction units (inter-extraction units) was
305 also evaluated and the results, expressed as RSD (n=5) ranged from 5.9 to 14.4 %.

306 Once optimized and analytically characterized, the proposed method was applied to the
307 determination of the target PAHs in environmental water samples (tap and river). The
308 samples were analyzed in order to find any potential presence of the analytes. Since the
309 analytes were not detected, validation samples were prepared using blank waters samples
310 (tap and river waters) fortified with the six target analytes (naphthalene, phenanthrene,
311 fluoranthene, pyrene, benz[a]anthracene and benzo[a]pyrene) at a concentration of $20 \mu\text{g}\cdot\text{L}^{-1}$.
312 Samples were maintained in amber glass bottles without headspace and in the dark for 24
313 h until analysis. Then, the fortified samples were analyzed using the extraction method, and
314 the concentration for each PAH was calculated by interpolating the peak area obtained in
315 the corresponding calibration graph. The recovery values were calculated dividing the
316 concentration found by the concentration added, and expressed in percentage. The recovery
317 values (average of three replicates) obtained for each of the fortified samples analyzed are
318 shown in Table 3, and they ranged from 72 to 124 %.

319

320 **4. Conclusions**

321 In this study, a novel silica monolith with embedded carbon nanoparticles was synthesized
322 inside a fused silica capillary for the extraction of non-polar compounds from water
323 samples. The hybrid monolithic solid sorbent was prepared by two-step catalytic sol-gel
324 process and the NPs were directly added into the polymerization mixture. The most critical
325 point of the synthesis was to achieve stable dispersions of the NPs due to their tendency to
326 form aggregates over time. A comparison between the silica monolith without and with
327 carbon nanoparticles was carried out. Hence, the three selected NPs were evaluated as
328 sorbents improving the extraction capacity in any case. Among them, the best results were
329 obtained for the c-MWCNTs being selected as optimum. Variables affecting to the
330 preparation of the monolithic solid as well as the extraction conditions were evaluated by
331 optimizing experimental parameters.

332 This work, demonstrates that the addition of the carbon NPs into silica solid monolithic
333 significantly increases the extraction efficiency for the target analytes due to π - π stacking
334 interaction and hydrophobic effect. In addition, the proposed microextraction method
335 shows favorable analytical features in comparison to the previously reported analytical
336 methods for the determination of PAHs in water samples. By way of example, Table 4
337 summarizes the main analytical information of the selected references related to the
338 proposed microextraction method, including LODs and LOQs as well as extraction format
339 used. As it can be derived from the data shown in the table, the proposed method presents
340 similar limits of quantification compared with the majority of the other approaches. The
341 lowest LOQs were obtained for [21], where an on-line thermal desorption step of the
342 retained analytes was implemented. This fact dramatically improves the method sensitivity
343 as it avoids the dilution inherent to the chemical elution. Also, the inclusion of the
344 fluorimetric detection resulted in a better sensitivity thanks to the native fluorescence of

345 PAHs. Nevertheless, it should be highlighted that present hybrid sorbent has been
346 optimized for achieve the concentration level established by ATSDR, and therefore, a better
347 sensitivity is not required in water samples.

348

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354 the service provided to obtain the micrographs and elemental analysis measurements.

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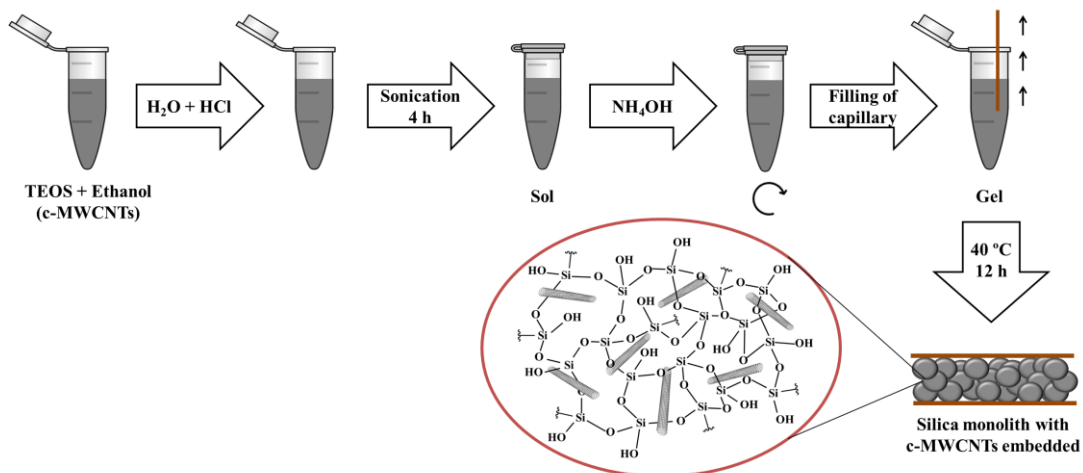
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365 **Figures**

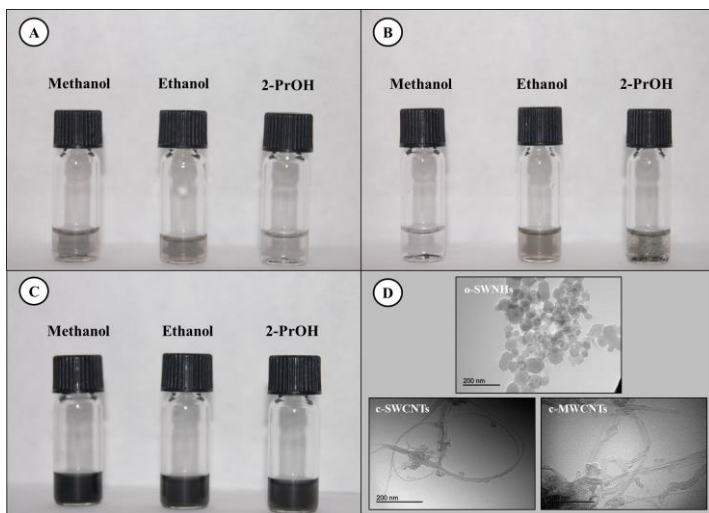
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368 Fig. 1 Scheme of the sol-gel process and the filling of the fused silica capillary.

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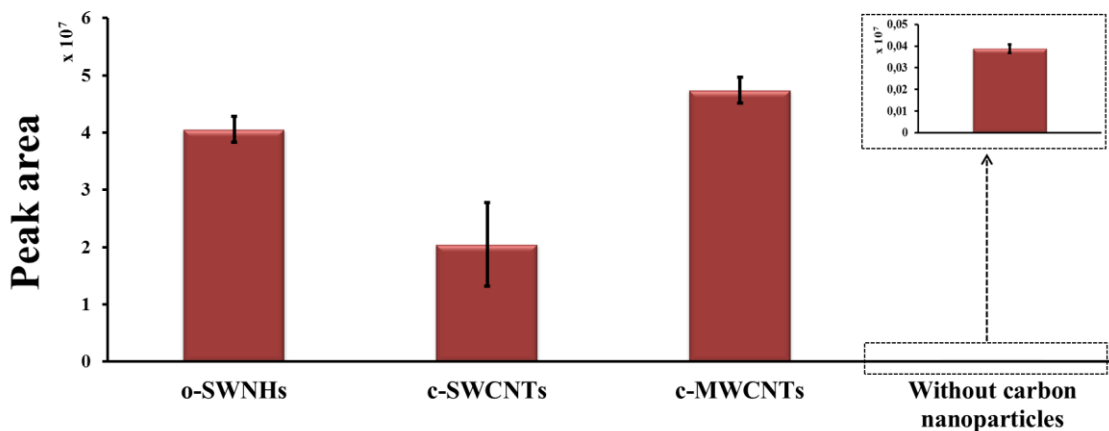
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371 Fig. 2 Photographs of the dispersions of c-SWCNTs (A), c-MWCNTs (B), and o-SWNHs

372 (C) in methanol, ethanol, and 2-PrOH; and TEM images (D) of the three carbon

373 nanostructures.

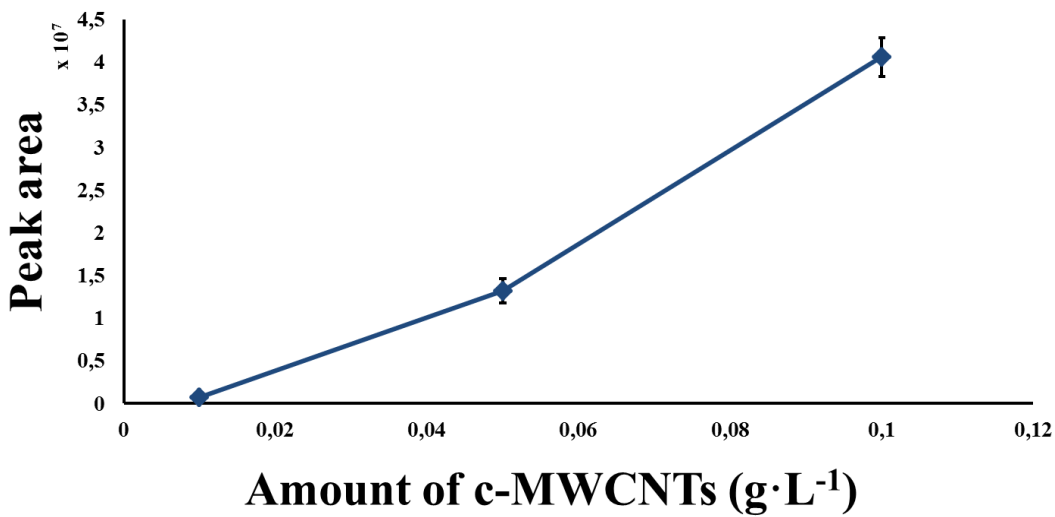
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376 Fig. 3 Comparison of the analytical performance of silica monolith without nanoparticles
 377 and the monolith containing embedded o-SWNHs, c-SWCNTs and c-MWCNTs.

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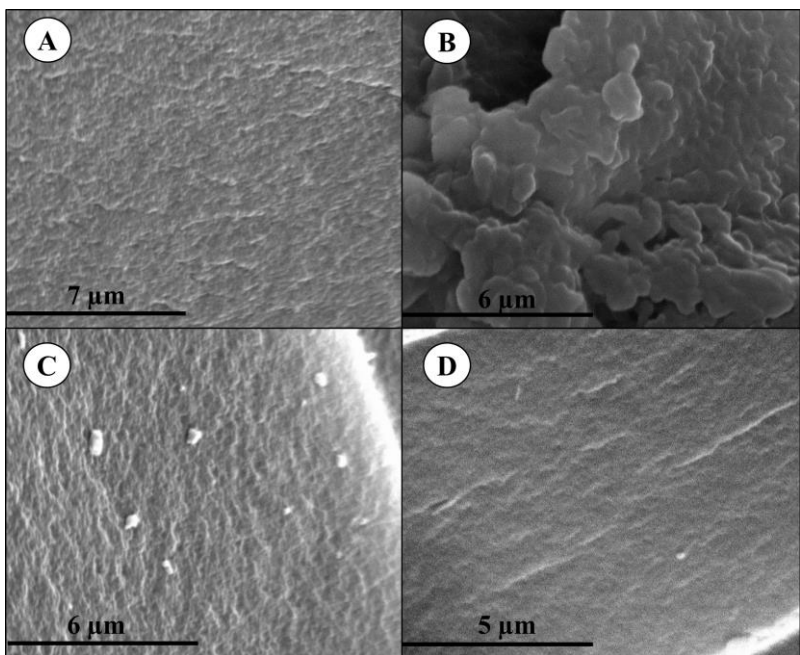


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380 Fig. 4 Influence of the concentration of the c-MWCNTs dispersion on the PAHs retention.

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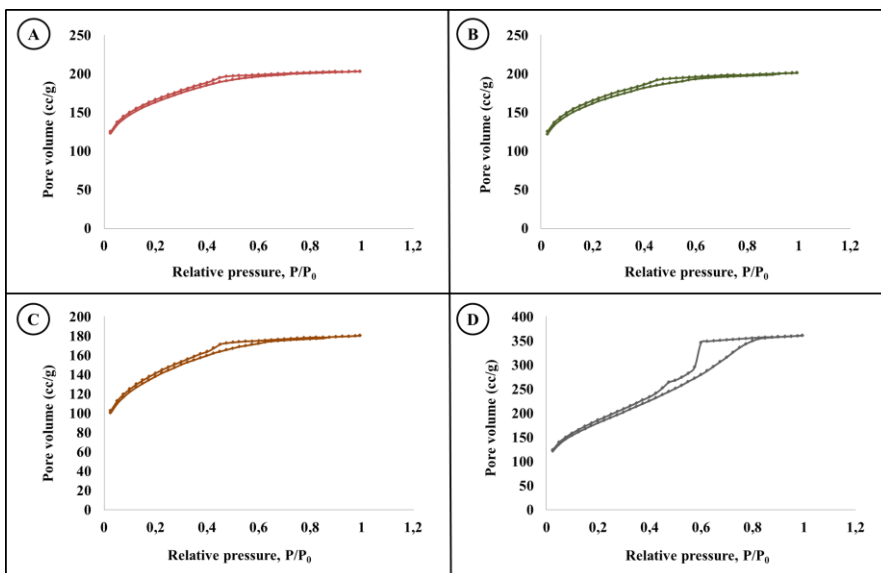
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384 Fig. 5 Scanning electron microscopy of silica monolith (A), Si-SWNHs (B), Si-SWCNTs
 385 (C), and Si-MWCNTs (D).

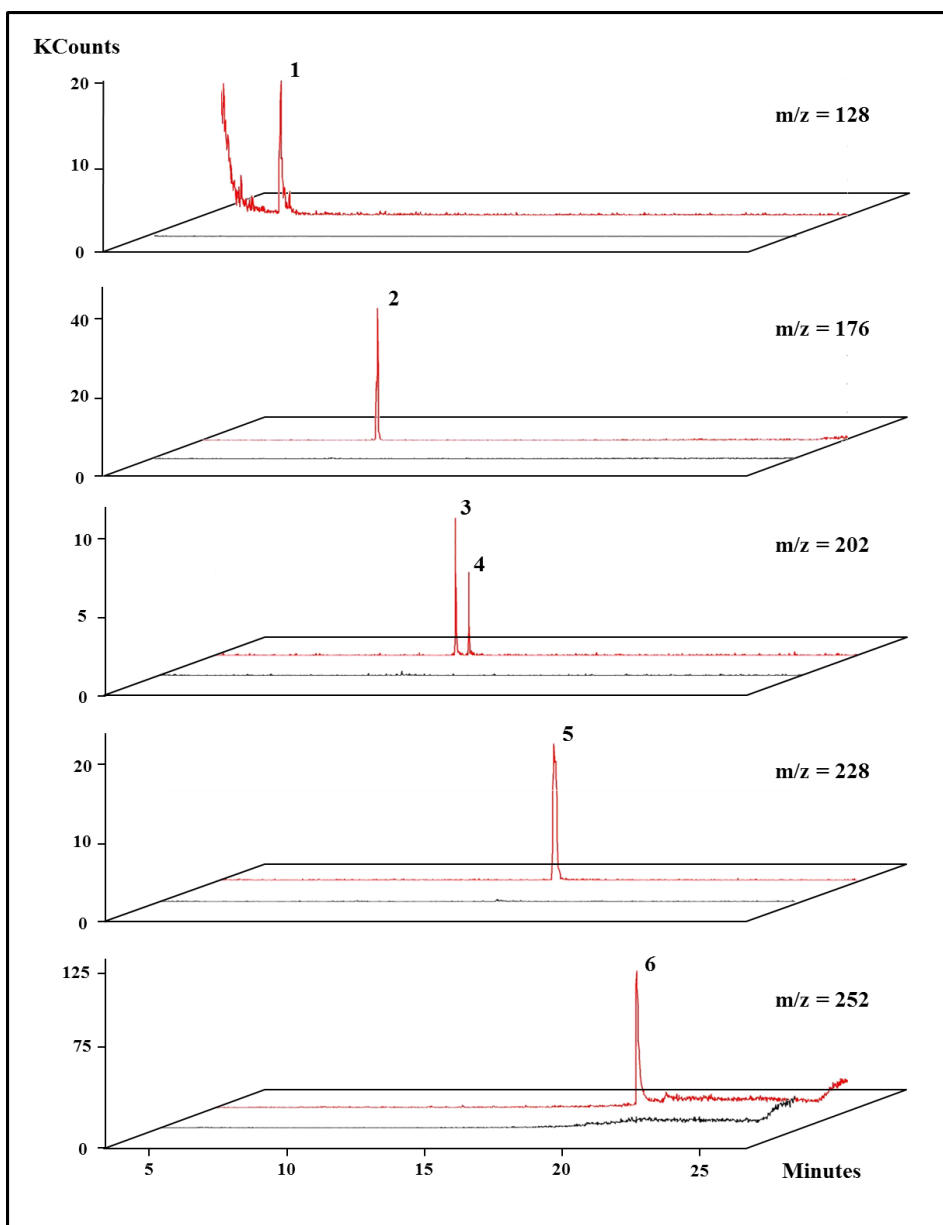
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388 Fig. 6 N₂ adsorption and desorption isotherms of silica monolith (A), Si-SWNHs (B), Si-
 389 SWCNTs (C), and Si-MWCNTs (D). P: sample pressure; P₀: saturation pressure.

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392 Fig. 7 SIM chromatogram obtained after monolithic microextraction procedure of a blank

393 water (green) and a standard sample with the target analytes at $0.1 \mu\text{g}\cdot\text{L}^{-1}$ (red). Peaks: (1)

394 Naphthalene, (2) Phenantrene, (3) Fluoranthene, (4) Pyrene, (5) Benz[a]anthracene, (6)

395 Benzo[a]pyrene.

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398 *Table 1. Textural properties of the synthesized monolithic solids.*

	Silica monolith	Si-SWNHs monolith	Si-SWCNTs monolith	Si-MWCNTs monolith
Specific surface area, BET (m ² /g)	594.07	588.25	497.37	646.82
Micropore surface area (m ² /g)	345.41	352.80	255.52	108.89
Average pore volume (cc/g)	0.314	0.311	0.279	0.557

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405 *Table 2. Analytical figures of merit of Si-MWCNTs monolithic microextraction unit to the*
 406 *determination of the target polycyclic aromatic hydrocarbons.*

Analyte	m/z	LOD ($\mu\text{g}\cdot\text{L}^{-1}$)	LOQ ($\mu\text{g}\cdot\text{L}^{-1}$)	Precision	
				Intra-extraction units RSD (% n=5)	Inter-extraction units RSD (% n=5)
Naphthalene	128	0.1	0.3	14.1	12.9
Phenanthrene	178	0.1	0.3	10.2	14.4
Fluoranthene	202	0.1	0.3	8.8	10.9
Pyrene	202	0.1	0.3	8.4	7.9
Benz[a]anthracene	228	0.1	0.3	10.8	10.7
Benzo[a]pyrene	252	0.1	0.3	13.3	13.9

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412 *Table 3. Recovery study for naphthalene, phenanthrene, fluoranthene, pyrene,*
 413 *benz[a]anthracene, and benzo[a]pyrene spiked to water samples analyzed following Si-*
 414 *MWCNTs monolithic extraction unit.*

Analyte	Tap water (% , n=3)	River water (% , n=3)
Naphthalene	113 ± 7	105 ± 5
Phenanthrene	85 ± 11	82 ± 14
Fluoranthene	99 ± 9	76 ± 9
Pyrene	101 ± 8	72 ± 8
Benz[a]anthracene	124 ± 5	121 ± 10
Benzo[a]pyrene	114 ± 14	109 ± 14

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425 *Table 4. Comparison of the developed Si-CNPs monolith with other monolithic solids*
 426 *reported for the determination of PAHs.*

Monolithic sorbent	Selected PAHs	Packing	Sample	LODs	LOQs	Detection	Reference
Si-CNPs monolith	Naphthalene, phenanthrene, fluoranthene, pyrene, benz[a]anthracene, and benzo[a]pyrene	Monolithic capillary column	Water samples	0.1 µg·L ⁻¹	0.3 µg·L ⁻¹	GC-MS	This work
Molecularly imprinted sol-gel polymer	Naphthalene, acenaphthylene, acenaphthene, fluorene, phenanthrene, anthracene, fluoranthene, pyrene, benz[a]anthracene, chrysene, benzo[b]fluoranthene, benzo[k]fluoranthene, benzo[a]pyrene, indeno[1,2,3-cd]pyrene, dibenz[a,h]anthracene, and benzo[ghi]perylene	Packed SPE	Water samples	5.2-12.6 ng·L ⁻¹	0.10-1.0 µg·L ⁻¹	GC-MS	[20]
Poly(butyl methacrylate-co-ethylene dimethacrylate) monolith	Naphthalene, acenaphthene, fluorene, anthracene, fluoranthene, pyrene, benz[a]anthracene, benzo[k], fluoranthene, benzo[a]pyrene, and indeno[1,2,3-cd]pyrene	Monolithic capillary column	Water samples	2.8-11.5 ng·L ⁻¹	0.012-0.05 µg·L ⁻¹	GC-MS	[21]
Graphene monolith	Naphthalene, acenaphthene, fluorene, phenanthrene, fluoranthene, pyrene, and benz[a]anthracene	Monolithic fiber	Water samples	4.0-50.0 ng·L ⁻¹	0.05-0.5 µg·L ⁻¹	GC	[22]
Hybrid organic-inorganic	Biphenyl, fluorene, phenanthrene, and fluoranthene	Monolithic capillary column	Water samples	2.4-8.1 µg·L ⁻¹	8.0-27.8 µg·L ⁻¹	µ-HPLC-UV	[23]
Poly(vinylpyrrolidone-co-divinylbenzene)	Naphthalene, fluorene, and fluoranthene	Stir bar	Water samples	0.045-0.093 µg·L ⁻¹	-	GC-MS	[24]
PolyHIPE monolith	Naphthalene, acenaphthylene, acenaphthene, fluorene, phenanthrene, anthracene, fluoranthene, pyrene, benz[a]anthracene, chrysene, benzo[b]fluoranthene, benzo[k]fluoranthene, benzo[a]pyrene, dibenz[a,h]anthracene, and benzo[ghi]perylene	Pipette tips	Water samples	0.004-0.228 ng·L ⁻¹	0.1-1.0 µg·L ⁻¹	HPLC-FLD	[25]

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