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26

27 **Abstract**

28 To classify trees along a spectrum of isohydric to anisohydric behavior is a promising new
29 framework for identifying tree species' sensitivities to drought stress, directly related to the
30 vulnerability of carbon uptake of terrestrial ecosystems with increased hydroclimate variability. Trees
31 with isohydric strategies regulate stomatal conductance to maintain stationary leaf water potential,
32 while trees with anisohydric strategies allow leaf water potential to fall, which in the absence of
33 significant hydraulic cavitation, will facilitate greater rates of carbon uptake. Despite the recognition
34 of the gas exchange consequences of isohydric and anisohydric strategies for individual tree species,
35 there have been few studies regarding whether isohydric trees produces distinct spectral signatures
36 under drought stress that can be remotely sensed. Here, we examined the capability of four vegetation
37 indices (PRI, NDVI, NDVI₇₀₅, and EVI) to capture the differences in spectral responses between
38 isohydric and anisohydric trees within a deciduous forest in central Indiana, USA. Both leaf-level
39 spectral measurements and canopy-scale satellite observations were used to compare peak growing-
40 season spectral signatures between a drought and a non-drought year. At the leaf scale, two vegetation
41 indices (NDVI and NDVI₇₀₅) failed to capture the drought signal or the divergent
42 isohydric/anisohydric behavior. EVI successfully captured the drought signal at both leaf and canopy
43 scales, but failed to capture the divergent behavior between isohydric and anisohydric tree species
44 during the drought. PRI captured both drought signals and divergent isohydric/anisohydric behavior at
45 both leaf and canopy scales once normalized between sunlit (backward direction images) and shaded
46 (forward direction images) portions of canopy, which indicates drought stress and subsequent
47 photosynthetic downregulation are greater in the sunlit portion of canopy. This study presents a
48 significant step forward in our ability to directly mapping emergent isohydricity at different scales
49 based on divergent spectral signatures between sunlit and shaded canopies.

50

51 **Keywords: PRI, isohydrlicity, drought, LUE, imaging spectroscopy, MODIS**

52

53 **Introduction**

54 Recently, many forested regions in the United States have experienced increasing drought stress
55 under ongoing climate change (Dai 2013). Warmer temperatures have increased hydroclimate
56 variability (Seager *et al.*, 2009; O'Gorman and Schneider, 2009) with more frequent droughts,
57 contributing to acute reductions in net ecosystem productivity (Roman *et al.* 2015), prolonged water
58 stress (Brzostek *et al.* 2014; Xu and Baldocchi 2003), and widespread tree mortality (Adams *et al.*
59 2009; Allen and Breshears 2007; McDowell and Allen 2015) in forested ecosystems. Differences
60 between tree species in their response to imposed hydrologic stress can be evaluated by characterizing
61 species along a continuum of isohydric to anisohydric behavior (Choat *et al.* 2012; Klein *et al.* 2014;
62 Martinez-Vilalta *et al.* 2014). Despite the risk of xylem cavitation associated with highly negative leaf
63 water potentials, trees with anisohydric strategies frequently sustain high stomatal conductance during
64 drought (Matheny *et al.* 2016; Meinzer *et al.* 2017; Roman *et al.* 2015; Yi *et al.* 2017). In the absence
65 of significant xylem cavitation, this strategy can also support relatively high levels of photosynthesis
66 during periods of drought stress. In contrast, trees with isohydric strategies regulate stomatal
67 conductance to maintain a relatively stationary leaf water potential as soil moisture declines. This
68 strategy reduces embolism risk, but at a cost of reduced carbon assimilation (McDowell *et al.* 2008).
69 The ratio of drought-driven declines in carbon uptake as compared to stomatal conductance depends
70 on the plasticity of the plant's intrinsic water use efficiency, which can increase during periods of
71 hydraulic stress, but not typically to an extent to completely mitigate losses in carbon uptake linked to
72 stomatal closure (Novick *et al.* 2015). In relatively mesic forests like Eastern US, where drought
73 events can have large effects on carbon and water cycling but rarely cause extensive hydraulic failure,
74 the isohydric/anisohydric continuum is well suited for predicting the extent to which trees regulate
75 their stomates during drought periods (Novick *et al.* 2015; Roman *et al.* 2015; Yi *et al.* 2017), as well
76 as the resulting carbon assimilation strategies. The sensitivity of gross primary production (GPP) to
77 drought stress has been shown to depend strongly on the relative proportion of isohydric vs.

78 anisohydric tree species in forest stands (Roman et al. 2015). Recent works also suggest that isohydric
79 stands are associated with greater interannual variability of evapotranspiration and reduced
80 productivity compared to anisohydric stands, as related to strong isohydric behavior in their stomatal
81 control and cavitation vulnerability (Ford et al. 2011; Novick et al. 2016; Stoy et al. 2006). As a result,
82 the degree to which forest stands respond isohydrically or anisohydrically to drought stress directly
83 influences ecosystem productivity and has the potential to cause cascading effects on regional carbon
84 storage.

85 While diagnosing isohydric vs. anisohydric stomatal responses is relatively straightforward, albeit
86 labor-intensive, at the tree level (Meinzer et al. 2016), we have little understanding of how this
87 behavior can be detected at large scales using remote sensing data. Specifically, there has been limited
88 research regarding whether isohydric and anisohydric trees possess distinct spectral responses under
89 drought stress (Sims et al. 2014) that can be remotely sensed and mapped to parameterize large-scale
90 GPP models (Konings and Gentine 2016; Sperry et al. 2016). Classically, light use efficiency (LUE)
91 based GPP models have been widely used to assess the impact of droughts on carbon sequestration at
92 a global scale (e.g. Song et al. 2013). In these models, GPP is usually proportional to the multiplicative
93 term between LUE and absorbed photosynthetically active radiation (APAR). However, remote
94 sensing metrics of canopy greenness often fail to capture small changes in leaf area or its function (i.e.
95 stomatal closure) that accompany drought (Asner et al. 2004; Hwang et al. 2008; Sims et al. 2014).
96 For this reason, most remote-sensing based terrestrial productivity models use water stress functions
97 that downregulate LUE values under conditions of low water availability, based on either vapor
98 pressure deficit (VPD) or soil moisture metrics (Song et al. 2013). These models are often applied to a
99 wide range of vegetation composition and formulate the stomatal responses with a single set of
100 ecophysiological model parameters based on the major plant functional types (e.g. Biome Parameter
101 Look-Up Table for MODIS GPP; Zhao et al. 2005). Consequently, current model structures cannot
102 reproduce observed differences in drought sensitivity among isohydric and anisohydric species within

103 temperate deciduous forests. This mismatch between our theoretical understanding of plant drought
104 response and its representation in models highlights a critical knowledge gap in understanding the
105 vulnerability of carbon sequestration of terrestrial ecosystems under increased hydroclimate variability,
106 especially in terms of species-specific variability with different water use and carbon assimilation
107 strategies under drought conditions.

108 Recently, there have been advances in directly mapping LUE using imaging spectroscopy from
109 different platforms. LUE mapping using remote sensing data is mostly based on the detection of
110 excess energy not used in photosynthesis especially under drought conditions (Coops et al. 2010;
111 Demmig-Adams and Adams III 1996), using the photochemical reflectance index (PRI) (Gamon et al.
112 1992; Gamon et al. 1997; Penuelas et al. 1995). PRI has been successfully applied across various
113 remote sensing platforms ranging from narrow bandwidth spectro-radiometers to the Moderate
114 Resolution Imaging Spectroradiometer (MODIS) (Drolet et al. 2008; Drolet et al. 2005; Goerner et al.
115 2009). Although PRI values are closely related to photosynthetic LUE values across different scales
116 (see Garbulsky et al. 2011), PRI-LUE relationships often show strong sensitivity to shadow fractions
117 following sun-target-sensor geometry (Drolet et al. 2005; Hall et al. 2008; Hilker et al. 2008), as well
118 as canopy structures and soil background reflectance (Barton and North 2001; Suárez et al. 2008).
119 Across forest ecosystems, the correlation between LUE and PRI for sunlit canopy is higher than
120 shaded canopy during water-stress conditions (Hall et al. 2008; Zhang et al. 2016). Expanding on this
121 framework, known correlations between view angle and canopy LUE suggest that the variability in
122 PRI values with view angle is directly related to the degree of down-regulation between sunlit and
123 shaded portions of the canopy (Coops et al. 2010; Drolet et al. 2005). Using multi-angle satellite
124 systems, such as backward and forward direction images, to capture the sunlit and shaded portion of
125 canopy, respectively (Hall et al. 2012; Hall et al. 2008), it may be possible to derive canopy responses
126 to drought (i.e. isohydry vs. anisohydry) by quantifying the extent to which sunlit and shaded canopies
127 differ in PRI.

128 In this paper, we hypothesize that the degree to which canopies downregulate photosynthesis under
129 drought stress (isohydry vs. anisohydry) is directly correlated with the magnitude of the differences in
130 PRI between sunlit and shaded portions of canopy (Coops et al. 2010). Further, using both leaf and
131 canopy scale measurements, we seek (1) to capture the difference in leaf-scale spectral signatures
132 between isohydric and anisohydric trees species under drought stress, and (2) to examine whether this
133 behavior can be detected at the canopy scale using multi-angle MODIS images. To meet these
134 objectives, we combined site-level spectral observations and remotely sensed data for a biologically
135 diverse temperate deciduous forest that contains a broad spectrum of isohydric and anisohydric
136 species, and that experienced an exceptionally severe drought in 2012 (Roman et al. 2015).

137

138 **Methods and materials**

139 *Study area*

140 The study site is located in the Morgan Monroe State Forest (MMSF; 39.3232°N, 86.4131°W) in
141 central Indiana, USA. MMSF is a deciduous broadleaf forest with average canopy height and rooting
142 depth of 27 m and 0.44 m, respectively (Ehman et al. 2002). Among the common tree species found in
143 the study region, sugar maple (*Acer saccharum*), tulip poplar (*Liriodendron tulipifera*) and sassafras
144 (*Sassafras albidum*) are generally recognized to be among the most isohydric (or conservative) species,
145 whereas white and red oaks (*Quercus alba* and *Q. rubra*) are anisohydric (or non-conservative) (Choat
146 et al. 2012; Novick et al. 2015; Roman et al. 2015). Based on measurements of trees with diameter at
147 breast height ≥ 10 cm made in 54 large plots in March 2011, the study site is dominated by isohydric
148 species such as sugar maple, tulip poplar and sassafras, which represent 21%, 25%, and 10% of total
149 basal area, respectively, while the anisohydric white and red oaks represent 5% and 3%, respectively.
150 The study site experienced a severe drought during the summer (June-August) of 2012. During this
151 time period, the total rainfall was 135 mm, which is less than 50% of average June-August rainfall
152 (302 mm from 1999-2015 except for 2012) (Figure S1; June and July rainfall was just ~10% of the

153 long-term mean). As a result, there was a dramatic early dry down of soil water content in 2012
154 (Figure 1). In field measurements and remote sensing analyses, we also include data from 2011 and
155 2013 to represent normal and wet years, respectively (Figure S1).

156

157 *Ecosystem measurements*

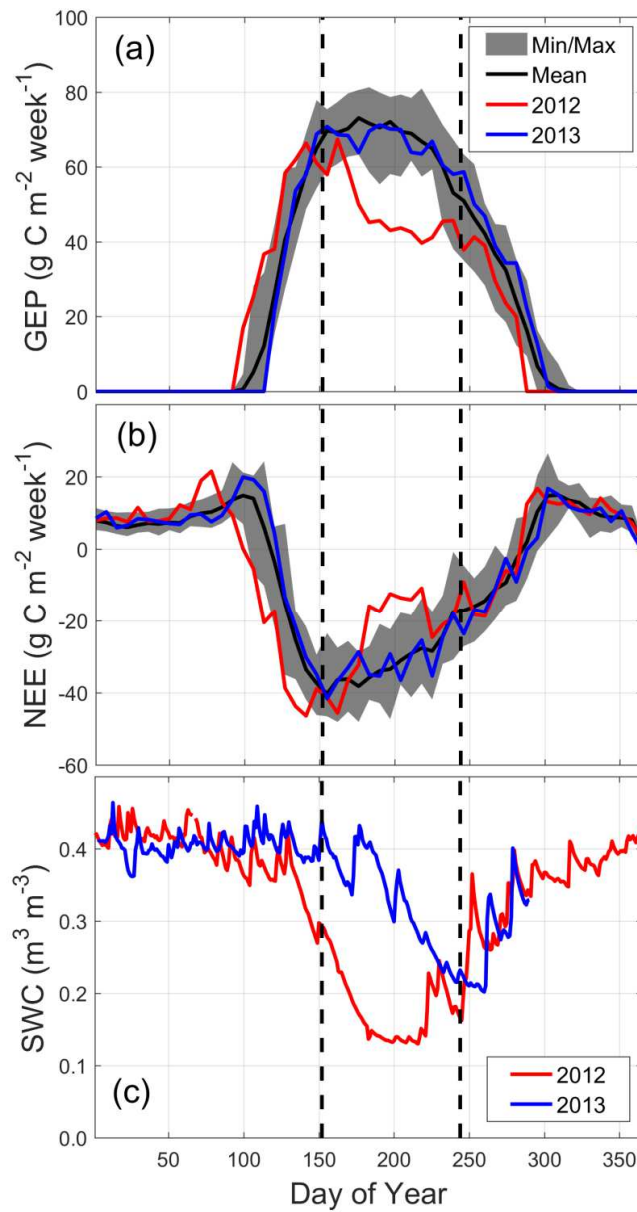
158 The net ecosystem exchange of CO₂ (NEE) has been continuously measured from the MMSF
159 *AmeriFlux* Tower (site code US-MMS) since 1998. Carbon dioxide and water vapor fluxes were
160 measured at the top of the tower (i.e., $z = 46$ m) using a sonic anemometer (CSAT-3; Campbell
161 Scientific Inc.) and a closed-path infrared gas analyzer (LI-7000; LI-COR, Lincoln, NE), which is
162 calibrated weekly. Eddy covariance NEE data were recorded continuously at a frequency of 10 Hz,
163 and filtered for periods of low turbulence (i.e. friction velocity < 0.3 m s⁻¹). Nocturnal and dormant
164 season NEE data were then used parameterize a model for ecosystem respiration (RE), which was
165 subtracted from NEE to derive an estimate of gross ecosystem productivity (GEP), as discussed in
166 detail in Dragoni et al. (2011), Schmid et al. (2000), and Sulman et al. (2016).

167 Absorbed photosynthetically active radiation (APAR) values were estimated from the difference of
168 PAR measurements between top and below the canopy. Total incoming photosynthetically active
169 radiation (IPAR) values were measured at 46-m height using a quantum sensor (BF3, Delta-T Devices
170 Ltd., Cambridge), while below-canopy PAR (PAR_{below}) values measured at 2-m height (Li-190, Li-
171 COR, Lincoln, NE). Hourly albedo (α) values were estimated from shortwave incoming and reflected
172 irradiance measurements from a net radiometer (CNR1, Kipp & Zonen, Netherlands), installed at 46-
173 m height. Hourly absorbed photosynthetically active radiation (APAR) and fraction of absorbed PAR
174 (FPAR) values were calculated as follows:

$$175 \quad APAR = IPAR \cdot (1 - \alpha) - PAR_{below} \quad (1)$$

$$176 \quad FPAR = APAR / IPAR \quad (2)$$

177 Hourly GEP and APAR estimates were used to calculate hourly LUE (= GEP/APAR) during the
178 peak growing season (June through August) in 2011, 2012, and 2013, each representing normal, dry,
179 and wet years, respectively. Weekly leaf area index (LAI) values have been measured along three
180 transects ($n = 30$ total) around the flux tower at dawn (LAI-2200, Li-COR, Lincoln, NE) during the
181 whole growing season (March-November). Volumetric soil water content (SWC) were also
182 continuously measured from the top 30 cm soil layer by reflectometer probes (CS615 and CS616;
183 Campbell Scientific, Logan, UT).
184



185

186 Figure 1: Long-term tower-based measurements of (a) weekly gross ecosystem productivity
 187 (GEP), (b) weekly net ecosystem exchange (NEE), and (c) mean daily volumetric soil water
 188 content (SWC) at the Morgan-Monroe State Forest flux tower in the study site. The gray
 189 regions represent the minimum and maximum ranges from 1999 to 2015, except for 2012
 190 (drought year). The vertical dashed lines show the peak growing season (June-August).
 191 Negative NEE values represent carbon sinks.

192

193 *Leaf-scale spectroscopic and gas exchange measurements*

194 Using a portable field spectroradiometer (UniSpec-SC Spectral Analysis System, PP Systems Inc.,
195 Amesbury, MA) and photosynthetic system (Li-6400Xt, Li-COR, Lincoln, NE), both leaf reflectance
196 and photosynthesis were measured for the intact leaves around midday (10 am-3 pm) weekly from
197 June through August in 2012 and 2013. We selected three trees with similar sizes from each species
198 group: sugar maples (mean diameter at breast height: 44.1, 44.2, and 48.4 cm), tulip poplars (62.5,
199 62.8, and 86.7 cm), sassafras (41.4, 34.0, and 36.1 cm), and oaks (two white oaks: 39.2 and 43.8 cm,
200 one red oak: 47.6 cm). Using a 25-m boom lift, five sunlit and five shaded leaves were selected for
201 each tree from the top and bottom of the canopy, respectively. Both spectroscopic and gas exchange
202 measurements were performed on the same leaves attached to the trees, but we did not measure the
203 same leaves over the entire experiment. Leaf gas exchange measurement was performed over a short
204 period of time (less than 2 min) with chamber conditions (CO₂ concentration, temperature, humidity,
205 and PAR) set to match ambient values (Roman et al. 2015). Previous analysis of the dynamics of leaf
206 water potential (pre-dawn and midday), also collected concurrently with the gas exchange
207 measurements, have been used to show that oaks behaved very anisohydrically during the 2012
208 drought, while other species were more isohydric (Roman et al. 2015). Leaf reflectance values were
209 also measured at 3-nm intervals within the visible and near infrared electromagnetic spectrum (400-
210 1000 nm), later interpolated to 1-nm intervals. All spectral and photosynthetic measurements were
211 averaged separately for sunlit and shaded leaves of each tree. These concurrent measurements of leaf
212 reflectance and photosynthesis provide a unique opportunity to test the ability to capture isohydric vs.
213 anisohydric behavior during the severe drought using remote sensing data.

214

215 *Leaf-scale spectral indices*

216 In this study, we used four spectral indices based on narrow bands of the electromagnetic spectrum:

217 photochemical reflectance index (PRI), normalized difference vegetation index (NDVI), red edge
218 NDVI (NDVI₇₀₅), and enhanced vegetation index (EVI) at both leaf- and canopy-scale spectral
219 measurements (Table 1). PRI directly quantifies the reflectance change associated with xanthophyll
220 absorption (at 531 nm) (Penuelas et al. 1995) and the associated LUE response (Drolet et al. 2008;
221 Drolet et al. 2005; Garbulsky et al. 2013; Rahman et al. 2004; Zhang et al. 2016). A simple
222 transformation was applied to PRI values to avoid negative or near-zero values, known as scaled-PRI
223 (sPRI; Table 1). For NDVI and EVI calculations, we used the reflectance of 645, 858.5, and 469 nm
224 wavelengths, the middle of the MODIS band 1 (red), 2 (near infrared), and 3 (blue) to facilitate the
225 comparison with MODIS vegetation indices. The red edge NDVI (NDVI₇₀₅) represents the normalized
226 difference between the 705 and 750 nm wavelengths, developed based on the chlorophyll index (Sims
227 and Gamon 2002). While PRI captures drought-induced instantaneous physiological changes in the
228 xanthophyll cycle, other leaf-scale indices were sensitive to chlorophyll-related changes during
229 droughts (Marchin et al. 2010; Munne-Bosch et al. 2001). Additionally, the canopy-scale NDVI and
230 EVI values also capture structural changes (e.g. LAI, leaf inclination angle etc.) accompanying
231 droughts. We compared these four spectral indices and photosynthesis measurements during the peak
232 drought period (July to August) in 2012 with the same period in a following wet year (2013). We
233 examined the differences in the spectral indices in regards to the photosynthesis between dry and wet
234 years, and between isohydric and anisohydric species both at the leaf and canopy scales. To capture
235 the divergence in spectral indices between sunlit and shaded portions of canopy, we calculated the
236 sunlit-to-shaded ratios in this study, such as $PRI_{sunlit}/PRI_{shaded}$. These ratio indices also make it possible
237 to normalize seasonal changes of PRI, driven by the changes in leaf color and carotenoids-to-
238 chlorophyll contents (Filella et al. 2009; Gamon et al. 2016; Wong and Gamon 2015).

239

240 Table 1: Four narrow-band spectral indices (PRI, NDVI, NDVI₇₀₅, and EVI) used in this study.

Spectral index	Equation	Reference
PRI or sPRI	$PRI = \frac{\rho_{531} - \rho_{570}}{\rho_{531} + \rho_{570}}; \quad sPRI = \frac{PRI + 1}{2}$	Gamon et al. (1992), Penuelas et al. (1995) Rahman et al. (2004)
NDVI	$NDVI = \frac{\rho_{858.5} - \rho_{645}}{\rho_{858.5} + \rho_{645}}$	Tucker (1979)
NDVI ₇₀₅	$NDVI_{705} = \frac{\rho_{750} - \rho_{705}}{\rho_{750} + \rho_{705}}$	Sims and Gamon (2002)
EVI	$EVI = 2.5 \frac{\rho_{858.5} - \rho_{645}}{\rho_{858.5} + 6\rho_{645} - 7.5\rho_{469} + 1}$	Huete et al. (1997)

241 ρ_n is reflectance at n nm. PRI: photochemical reflectance index, NDVI: normalized difference
 242 vegetation index, NDVI₇₀₅: red edge NDVI, and EVI: enhanced vegetation index

243

244 *MODIS data processing*

245 We extracted the 250-m MODIS NDVI and EVI data (MOD13Q1) around the flux tower from
 246 2001 to 2014 using the MODIS Web Service Tool
 247 (https://modis.ornl.gov/data/modis_webservice.html). We used only the pixel values with a ‘good’
 248 quality, based on the Pixel Reliability parameter (Didan and Huete 2006). The NDVI and EVI results
 249 were presented by averaging a 3×2-km area around the flux tower, which represents relatively
 250 homogenous forest cover. We also calculated the canopy-scale sPRI values using MODIS Terra and
 251 Aqua data (collection 6) at the same spatial extent. Three MODIS products (MOD/MYD021KM,
 252 MOD/MYD03, and MOD/MYD35) were downloaded from the Level-1 and Atmosphere Archive and
 253 Distribution System (LAADS; <https://ladsweb.nascom.nasa.gov/>) for June, July, and August of 2011,
 254 2012, and 2013. The MOD/MYD021KM product includes calibrated and geolocated at-aperture
 255 radiances at 1-km resolution for 36 MODIS bands. We only analyzed the pixels which centroids are
 256 located within the 3×2-km area in the sPRI calculation. Since the reference band used in PRI (i.e., the
 257 band at 570 nm) is not available in MODIS products, we apply a modified approach proposed by
 258 Drolet et al. (2005) to produce the reflectance data required to calculate sPRI.

259 While Drolet et al. (2005) use band 13 as the reference band, previous research at MMSF has
 260 shown that band 13 is likely saturated for terrestrial observations (Goerner et al. 2009). As a result, our
 261 calculation of PRI uses band 1 as the reference and band 11 as the detection band. Recently, this index
 262 was suggested to be renamed as a Chlorophyll/Carotenoid Index (CCI) to monitor evergreen
 263 photosynthetic activity (Gamon et al. 2016). Additionally, previous studies have shown that
 264 atmospheric correction of MODIS data does not improve the PRI results and it may even degrade the
 265 correlation between PRI and LUE (Goerner et al. 2011). So, instead of using atmospherically corrected
 266 reflectance values, we used the top-of-atmosphere (TOA) reflectance to calculate PRI (He et al. 2016).
 267 The MOD/MYD35 products were also used for cloud masking and removing cloud-covered pixels.
 268 We used only the pixels with the ‘confident clear’ quality flag in the MOD/MYD35 products. To
 269 remove pixels with large footprints, we eliminated all pixels with sensor zenith angle larger than 40°.

270 From the MODIS Terra and Aqua data, we calculated the PRI values for backward (sun is behind
 271 the sensor) and forward (sun is in front of the sensor) direction images which effectively capture the
 272 spectral signatures of sunlit and shaded portion of canopy, respectively. The solar/sensor azimuth and
 273 zenith angles and coordinates of all pixels from the MOD/MYD03 product were used to determine the
 274 backward and forward direction images. Using the solar and sensor azimuth angles obtained from
 275 MOD/MYD03 products, the relative azimuth angle Δ can be defined as follows:

$$276 \quad \Delta = \begin{cases} |\theta - \theta_0| & \text{if } 0^\circ \leq |\theta - \theta_0| \leq 180^\circ \\ 360^\circ - |\theta - \theta_0| & \text{if } |\theta - \theta_0| > 180^\circ, \end{cases} \quad (3)$$

277 where θ and θ_0 are the sensor and solar azimuth angles, respectively. Backward direction images
 278 have a relative azimuth angle less than or equal to 60°, while forward direction images have a relative
 279 azimuth angle greater than 60° (Drolet et al. 2005). Note that the sunlit portion of canopy dominates
 280 the field of view of the sensor with smaller Δ angles (backward direction) with increasing the shaded
 281 portion with larger Δ angles (forward direction) (Cheng et al. 2012; Hall et al. 2008). In a manner
 282 similar to the leaf-scale sunlit-to-shaded sPRI ratios, the sPRI ratios were also calculated between

283 backward and forward direction images. However, the MODIS PRI values are not essentially paired
284 simply because cloud-free backward and forward images were not collected on the same days,
285 different with the tree-level measurements. To avoid no PRI values in either backward or forward
286 direction images at a short time interval, we calculated the PRI ratios over three-week intervals during
287 the peak growing season (June-August) from 2011 to 2013.

288

289 *Statistical analyses*

290 We performed two-sample Welch's (unequal variances) *t*-tests with a null hypothesis that the
291 group mean of 2012 (dry year) is greater than that of 2013 (wet year) during the peak drought period
292 (July to August). We applied this test to all four spectral indices (PRI, NDVI, NDVI₇₀₅, and EVI;
293 Table 1) and their ratios at the tree level, as well as to all pixel-level MODIS PRI values at the canopy
294 scale. For three groups among normal (2011), dry, (2012), and wet (2013) years, we applied the
295 analysis of variance (ANOVA) tests to evaluate whether the means of three groups are equal.

296

297 **Results**

298 *Flux and gas exchange measurements*

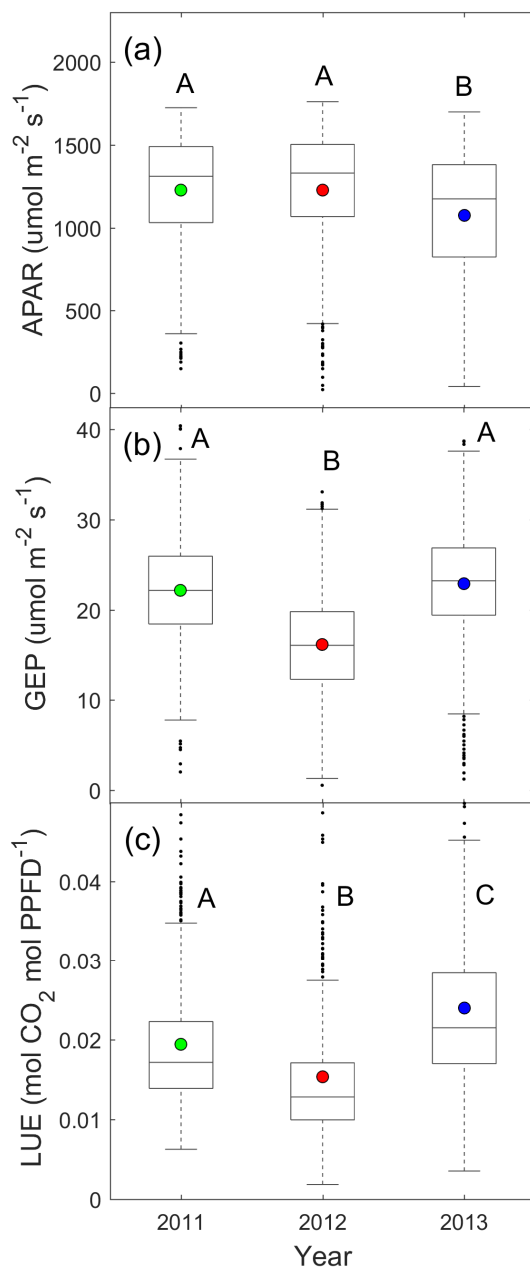
299 At the canopy scale, both GEP and NEE from the flux tower were reduced to an unprecedented
300 level during the 2012 drought event (Figure 1). During the peak drought period (July-August), the
301 magnitudes of weekly GEP and NEE were over 30 g C m⁻² (less than 40% of mean value) and 20 g C
302 m⁻² lower (less than 55% of mean value) respectively, compared to the baseline mean GEP and NEE
303 (1999-2015 except for 2012). Significant differences in midday (10 am-3 pm) hourly APAR and GEP
304 between 2012 and 2013 were also observed (Figure 2). Mean APAR values were significantly higher
305 ($p < 0.01$) in 2012 (dry) than in 2013 (wet), which were largely driven by significantly higher
306 incoming PAR due to less frequent clouds with little changes in fraction of absorbed PAR (FPAR)

307 (Figure S2). As a result, mean hourly LUE (=GEP/APAR) values were significantly lower in 2012
308 than in 2013 ($p < 0.01$). It is worthwhile to note that LUE values in the normal year (2011) were
309 between the values in wet and dry years.

310 Leaf gas exchange measurements showed similar significant decreases in mean net photosynthesis
311 (PS) during the drought ($p < 0.001$), as well as divergent behavior between isohydric and anisohydric
312 species (Figure 3). In 2012, mean leaf PS values were reduced by about 30-50% compared to those in
313 2013 for the three isohydric species (sugar maple, tulip poplar, and sassafras). In the sunlit leaves,
314 tulip poplar showed the largest percent decrease (41.6 %) in the mean values, followed by sassafras
315 (37.7 %) and sugar maple (31.7 %). In the shaded leaves, sassafras showed the largest percent
316 decrease (46.2 %), followed by sugar maple (40.9 %) and tulip poplar (33.8 %). Sunlit leaves usually
317 had higher levels of PS than the shaded ones, and they showed larger decreases in the absolute
318 magnitudes of PS during the 2012 drought. However, oaks (anisohydric) still maintained relatively
319 high levels of PS both in sunlit and shaded leaves. While stomatal conductance of oak species was
320 unaffected by the drought (Roman et al. 2015), there was a small (9.4 %) but significant difference (p
321 < 0.05) in mean PS of sunlit oak leaves between years (Figure 3d). This could reflect non-stomatal (i.e.
322 biochemical) limitations to the photosynthetic machinery imposed by the particularly low leaf water
323 potential observed for these species during the drought (Roman et al. 2015). It is worthwhile to note
324 that the decrease in PS of sunlit leaves was somewhat compensated by a significant increase in PS of
325 shaded PS ($p < 0.05$; 21.2 %) for oak species (Figure 3d).

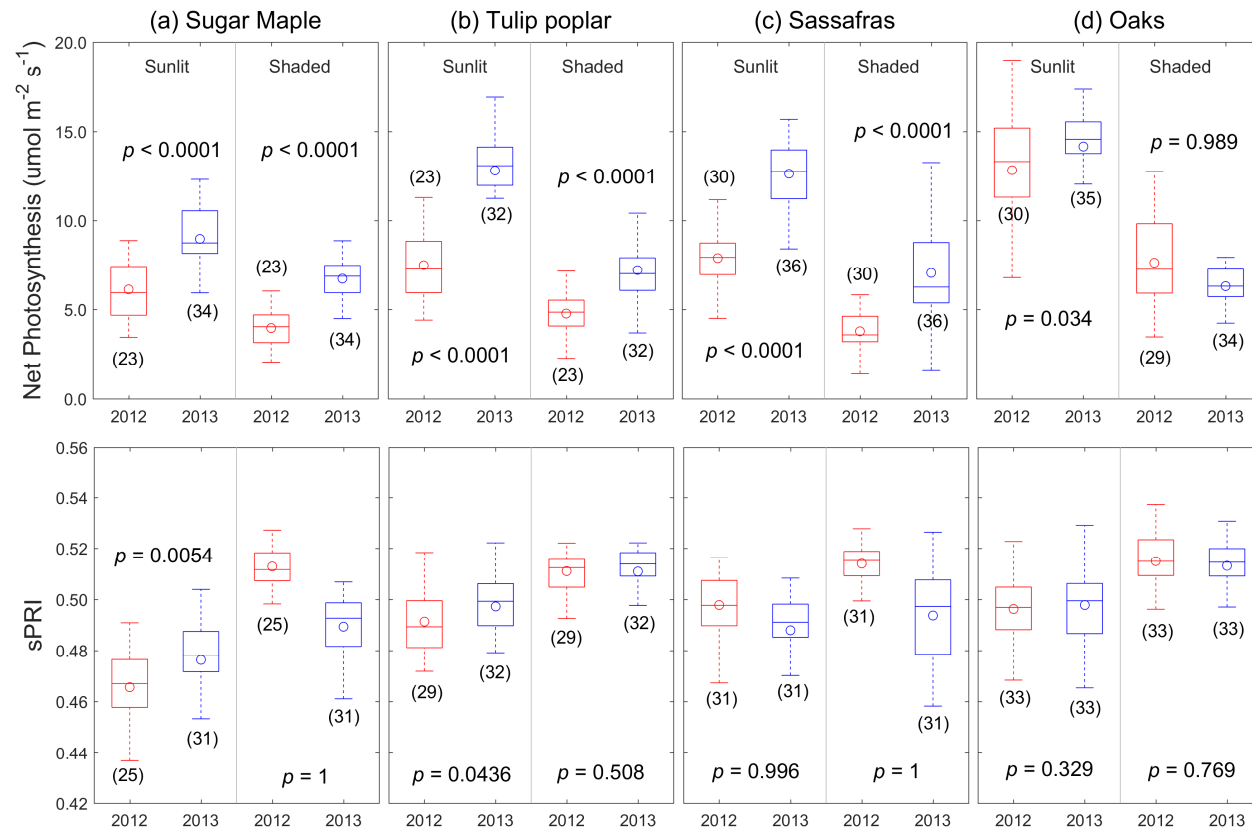
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329 Figure 2: Hourly midday (10 am-3 pm) (a) absorbed photosynthetically active radiation
 330 (APAR), (b) gross ecosystem productivity (GEP), and (c) light use efficiency (LUE) values
 331 from June through August in normal (2011 – green), drought (2012 – red), and wet y(2013 –
 332 blue) years from the Morgan Monroe State Forest flux tower. Circles are the mean values,
 333 while black dots are outliers. Different letters (A-C) denote significant differences in the group
 334 means using an analysis of variance (ANOVA) test ($p < 0.01$).



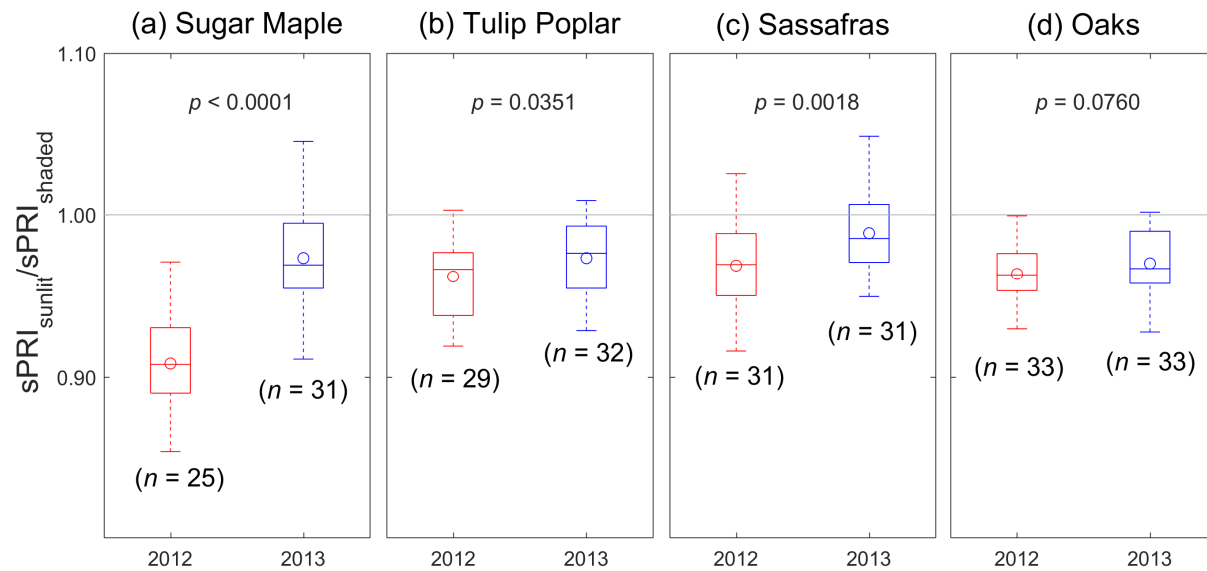
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336 Figure 3: Boxplots for the measured leaf net photosynthesis ($\mu\text{mol m}^{-2} \text{s}^{-1}$) and sPRI values (dimensionless) from July to August in
 337 2012 (drought year - red) and 2013 (wet year - blue) for three isohydric species: (a) sugar maple, (b) tulip poplar, (c) sassafras, and
 338 (d) two anisohydric oaks (two white oaks and one red oak). Leaf reflectance and photosynthesis were measured weekly around
 339 midday (10 am-3 pm) for sunlit and shaded leaves separately (Figure S3 and S4). All measurements represent the tree-level
 340 observations from three trees in each group ($n = 12$ total), which are actually the averaged values from five leaf samples from each
 341 tree. The numbers in the parentheses are the total number of tree-level observations, and circles are the mean values. The p -
 342 values were from the two-sample t -test with a null hypothesis that the group mean of 2012 is greater than that of 2013.

343 *Leaf-scale spectral indices*

344 Contrary to expectation, scaled PRI (*sPRI*) values of sunlit and shaded leaves did not show any
345 consistent patterns between dry and wet years (Figure 3), and the correlation analyses with the leaf PS
346 measurements did not show any significant results at tree and species levels. The *sPRI* values of the
347 sunlit leaves were generally lower compared to those of the shaded leaves in both years, which may
348 effectively represent higher LUE in shaded portion of canopy. The mean *sPRI* values of the sunlit
349 leaves (except for sassafras) were generally lower in 2012 (dry) than in 2013 (wet), while those of the
350 shaded leaves were generally higher in 2012 (except for tulip poplar) (Figure 3). The *sPRI* values
351 showed more interannual variability for the three isohydric species, while they were not significantly
352 different for two oak species (Figure 3d). These contrasting patterns between isohydric and
353 anisohydric species led to the difference in the sunlit-to-shaded PRI ratios between dry and wet years
354 (Figure 4). For all three isohydric species, the mean *sPRI* ratios ($sPRI_{sunlit}/sPRI_{shaded}$) were significantly
355 lower in 2012 ($p < 0.05$), while not for the two oak species ($p > 0.05$). This demonstrates that the
356 decreases in *sPRI* were much greater for sunlit than for shaded leaves during the drought, consistent
357 with the leaf-level PS measurements (Figure 3).

358 For the other spectral indices (NDVI, NDVI₇₀₅, and EVI), only EVI showed consistently and
359 statistically lower mean values in 2012 for both sunlit and shaded leaves, compared to 2013 (except
360 for the sunlit leaves of sassafras; Figure S5). NDVI₇₀₅ showed slightly better performance to capture
361 the drought effect and isohydric/anisohydric behavior than NDVI, but these patterns were still not
362 consistent. All three spectral indices did not show any distinct inter-annual differences between the
363 oak species and others, including their sunlit-to-shaded ratios (Figure S6). Thus all three of the
364 greenness spectral indices failed to capture isohydric vs. anisohydric behavior under the drought
365 condition, supporting that there is no clear evidence that drought-induced changes in leaf PS has any
366 direct implications on these three spectral indices.



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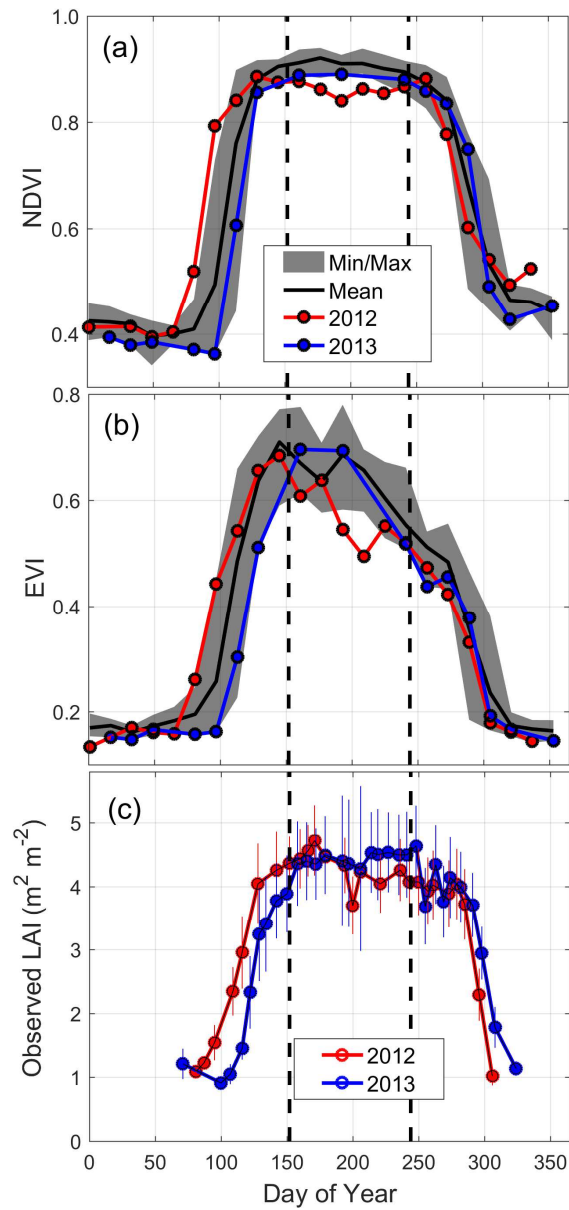
368 Figure 4: Boxplots for the sunlit-to-shaded sPRI ratios (dimensionless) from July to August in 2012 (drought year - red) and 2013
 369 (wet year - blue) for three isohydric species: (a) sugar maple, (b) tulip poplar, (c) sassafras, and (d) anisohydric oak species (two
 370 white oaks and one red oak). All measurements represent the tree-level observations from three trees in each group, calculated
 371 from five sunlit and five shaded leaves from each tree. The numbers in the parentheses are the total number of tree-level
 372 observations, and circles show the mean values. The p -values were from the two-sample t -test with a null hypothesis that the group
 373 mean of 2012 is greater than that of 2013.

374 *Canopy-scale spectral indices*

375 The canopy-scale NDVI and EVI values showed clear differences between 2012 (dry) and 2013
376 (wet) years during the peak drought season (July-August) (Figure 5). Although both vegetation indices
377 show little interannual variability in their maximum values for the period from 2001 to 2014, the mean
378 NDVI and EVI values (July-August) were 5% and 11% less, respectively, in 2012 compared to the
379 mean for 2001-2014 (except 2012). Since leaf NDVI was largely unaffected by drought (Figure S5)
380 and the decrease in canopy NDVI roughly corresponded with the observed decrease in LAI values
381 (about 5-10 %) (Figure 5c), the changes in canopy structure appear to account for much of the change
382 in NDVI during the drought. On the other hand, canopy EVI declined slightly more during the drought
383 than did NDVI, indicating that both canopy structural and leaf spectral changes were factors in the
384 change in canopy EVI. However, note that decreases in both NDVI and EVI values were substantially
385 smaller than the changes in GEP and NEE values during the drought period (Figure 1).

386 Similar to the leaf-scale measurements, the pixel-level MODIS sPRI values during the peak
387 drought period (July-August) did not show consistent or statistically different patterns between dry
388 and wet years for both Terra and Aqua images (Figure 6). The MODIS sPRI values were generally
389 higher than the field observations, possibly due to the difference in the reference band. However, the
390 sPRI values from backward direction images were significantly lower than forward ones in 2012 for
391 both Terra and Aqua images ($p < 0.1$) (Figure 6). This demonstrates that the decreases in sPRI were
392 much greater in the sunlit portion of the canopy than in shaded portions during the peak drought,
393 consistent with the leaf-level spectral and PS measurements (Figure 3 and 4). As a result, the
394 backward-to-forward sPRI ratios were consistently lower during the peak drought period (July-August)
395 in 2012 than those in the normal and wet years (Figure 7). This was similar to the lower sunlit-to-
396 shade sPRI ratios at the leaf scale in 2012 (Figure 4), although the absolute ratio values are slightly
397 higher at the canopy scale. Given the <10% basal area composition of the anisohydric species in the
398 study site, it is possible that the leaf-level PRI signal of the isohydric species should be visible at the

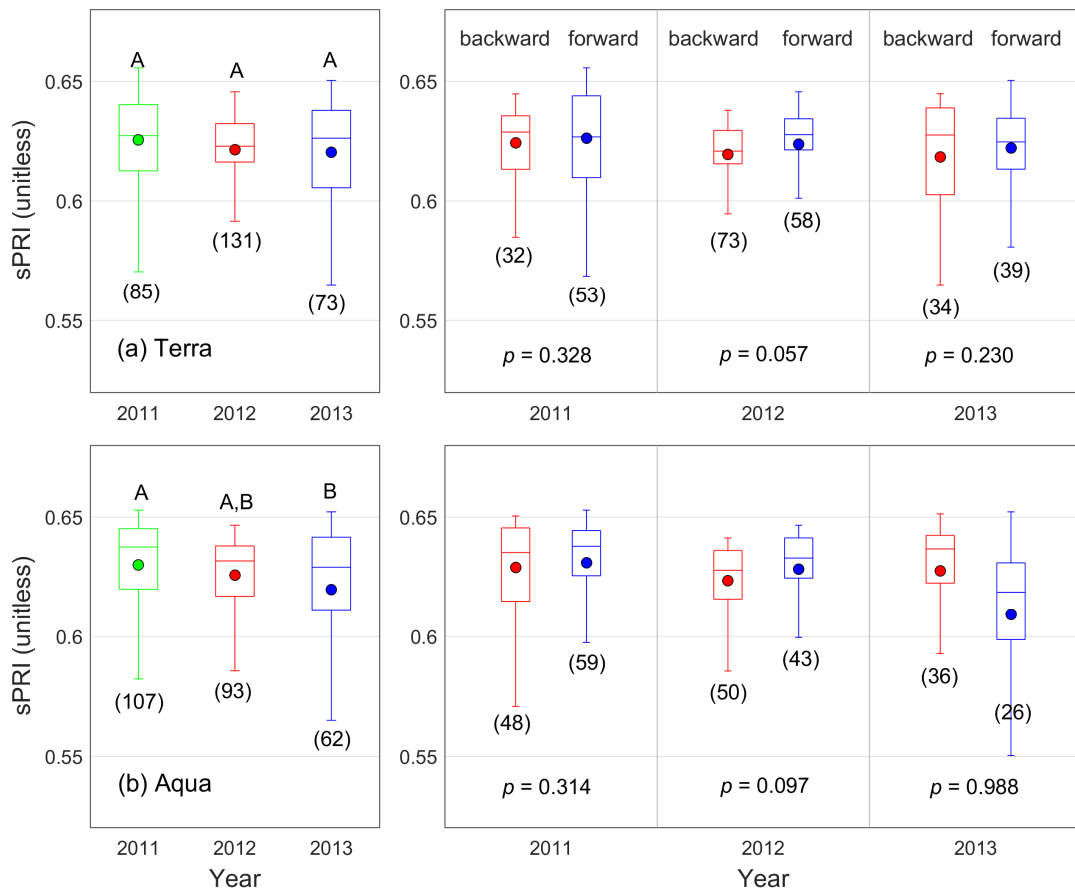
399 MODIS pixel scale. However, note that the seasonal drought patterns were more obvious in the Aqua
400 data (local overpass time of approximately 1:30 PM; Figure 7b) than in Terra data (10:30 AM; Figure
401 7a), which more resemble GEP and NEE patterns in 2012 and 2013 (Figure 1). Additionally, these
402 ratios generally remained higher in 2013 (wet), compared to those at the same time in the other two
403 years.
404



405

406 Figure 5: Seasonal patterns of MODIS (a) NDVI and (b) EVI (MOD13Q1), (c) observed LAI
 407 (leaf are index) in 2012 (dry) and 2013 (wet) years. The gray regions represent the minimum
 408 and maximum ranges of NDVI and EVI values from 2001 to 2015, except for 2012. LAI
 409 values were measured weekly along the three transects around the flux tower ($n = 30$ total),
 410 and the vertical lines represent the standard deviations. The vertical dashed lines show the
 411 peak growing season (June-August) of the year.

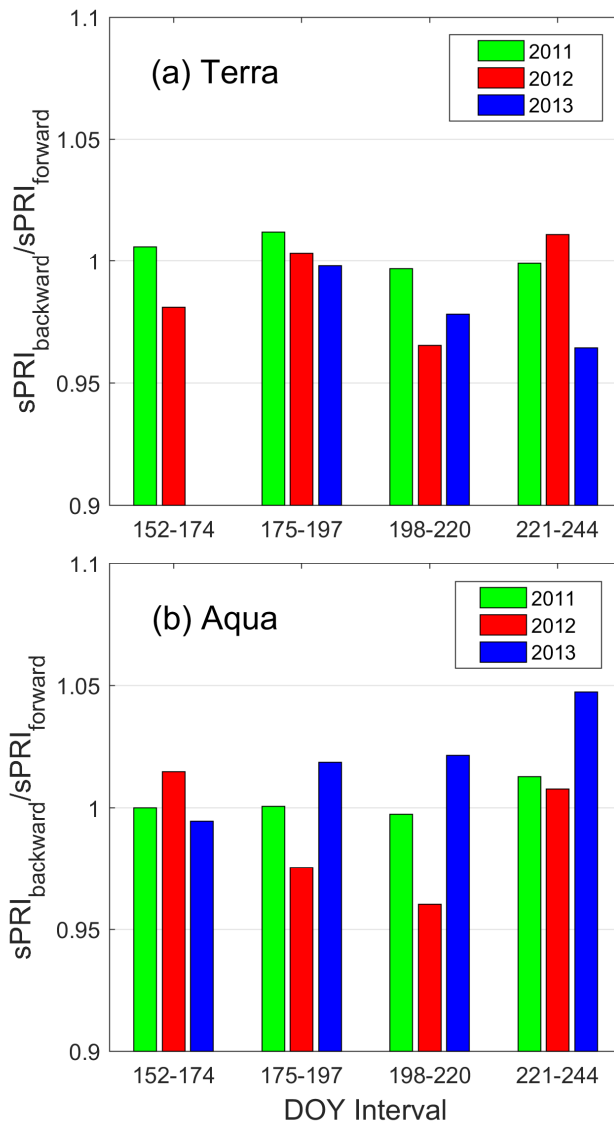
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414 Figure 6: All sPRI values obtained from MODIS (a) Terra and (b) Aqua data from July
 415 through August in normal (2011 – green), dry (2012 – red), and wet (2013 – blue) years. Bar
 416 graphs in the right column show the pixel-level PRI values from the backward (red) and
 417 forward (blue) direction MODIS images each year. Backward direction images include more
 418 sunlit portion of the canopy, while forward direction images include more shaded portion. The
 419 numbers in the parentheses are the total number of pixel-level MODIS PRI values used in
 420 the study (Figure S7), and circles show the mean values. Different letters (A-B) denote
 421 significant differences in the group means using an analysis of variance (ANOVA) test ($p <$
 422 0.05). The p -values were from the two-sample t -test with a null hypothesis that the group
 423 mean of 2012 is greater than that of 2013.

424



425

426 Figure 7: The sPRI ratios of backward to forward direction images for MODIS (a) Terra and
 427 (b) Aqua data from June through August in normal (2011 – green), dry (2012 – red), and wet
 428 (2013 – blue) years in the study site. The definition of backward and forward direction
 429 MODIS images are provided in Eq. 3. The sPRI ratios were calculated roughly at three-week
 430 intervals from the group means of the pixel-level MODIS PRI values at each day of year
 431 (DOY) interval (Figure S7).

432

433 **Discussion and conclusions**

434 In this study, we examined the capability of four spectral indices to capture peak drought signals
435 and isohydric/anisohydric behavior at both leaf and canopy scales in the closed deciduous broadleaf
436 forest. NDVI and NDVI₇₀₅ failed to capture the drought signal or the divergent isohydric/anisohydric
437 behavior at the leaf scale, while NDVI did show the canopy-scale structural changes (e.g. LAI) caused
438 by the drought. EVI successfully captured the drought signals at both leaf and canopy scales, but failed
439 to capture the isohydric/anisohydric behavior. Finally, PRI captured both drought signals and
440 divergent isohydric/anisohydric behavior at both leaf and canopy scales once normalized between
441 sunlit (backward direction) and shaded (forward direction) portions of canopy. Consistent with
442 previous literature (e.g. Cheng et al. 2012; Hall et al. 2008), our results showed the directional
443 responses of PRI were larger when drought stress and subsequent photosynthetic downregulation were
444 greater in the sunlit portion of canopy. It is worthwhile to note that the seasonal drought signals in
445 backward/forward PRI ratios from the MODIS Aqua data were more similar with GEP, NEE, and soil
446 moisture patterns than those from Terra data.

447 Although PRI captures instantaneous energy dissipations in the xanthophyll cycle, it can be also
448 influenced by seasonal and interannual changes including carotenoid-to-chlorophyll pigment ratios
449 (Filella et al. 2009; Sims et al. 2006; Wong and Gamon 2015) and sun-target-sensor geometry (Gamon
450 and Bond 2013). It is also well-known that trees retranslocate their leaf nitrogen under severe drought
451 conditions (Heckathorn and DeLucia 1994), which might also affect the interannual LUE and PRI
452 values between wet and dry years (Garbulsky et al. 2011). Therefore, seasonal and interannual
453 variability of PRIs can be readily confounded by multiple environmental factors, which makes it
454 difficult to detect and standardize drought stress using PRIs (Soudani et al. 2014; Zarco-Tejada et al.
455 2013; Zhang et al. 2016). Although there were the significant interannual changes in carbon uptakes
456 both at the leaf and canopy scales, we also did not see any consistent shifts or apparent trends in the
457 PRI values between wet and dry years. Likewise, we could not find any significant correlations

458 between the PRI values and the leaf PS measurements both at tree and species levels. However, once
459 normalized using sunlit-to-shade (leaf scale) and backward-to-forward image (canopy scale) ratios
460 under the assumption of increasing the degree of down-regulation during droughts, we could
461 successfully capture the drought effect on divergent behaviors between wet and dry years, between
462 isohydric and anisohydric tree species, and possibly between morning (Terra) and afternoon (Aqua)
463 MODIS images.

464 Our analysis also agrees well with a recent study by Vicca et al. (2016), who reported that only
465 EVI is capable of detecting drought signals at Hesse Forest in the northeastern France (also a
466 deciduous forest), while PRI and NDVI values did not capture the interannual changes in GPP caused
467 by drought. EVI was originally designed to reduce the spectral variances driven by atmospheric
468 aerosols by including the blue band (Huete et al. 2002). However, it is also sensitive to carotenoid
469 contents, which have high absorption between 400 and 500 nm (Sims and Gamon 2002) and usually
470 increases during droughts (Liu et al. 2011). However, EVI failed to capture divergent isohydric vs.
471 anisohydric behavior upon drought stress, clearly manifested in the leaf-level PS measurements.
472 Therefore, it is likely that the reduction of EVI during the drought period was driven not by emergent
473 differences in photosynthetic activities, but by the accompanying leaf physiological changes as well as
474 the structural changes at the canopy level, as suggested by Vicca et al. (2016).

475 Recently, two normalization methods for PRI were proposed by Zarco-Tejada et al. (2013) and
476 Soudani et al. (2014) at different time scales. Zarco-Tejada et al. (2013) used both renormalized
477 difference vegetation and red edge ratio indices to normalize PRI with the drought-induced changes in
478 chlorophyll content and vegetation structure at diurnal scale. Soudani et al. (2014) proposed the
479 correction of PRI values using intercepts of the linear regressions between APAR and PRI (called PRI₀)
480 at seasonal scale. Our normalization method intended to detect species-level responses to severe
481 droughts (isohydricity vs. anisohydricity) for closed deciduous forests, where traditional spectral
482 indices often have limited capability to capture drought-induced structural and/or chlorophyll-related

483 changes due to saturation problems (e.g. Daughtry et al. 2000). Additionally, we reported the
484 divergence of PRI values between sunlit and shaded canopy portions consistently across leaf, canopy,
485 and species levels, which would provide a unique advantage in upscaling drought effect on ecosystem
486 carbon sequestration using spectral remote sensing information. However, the suggested method using
487 backward and forward direction images may not be suitable for sparse forests. It is simply because this
488 method explicitly assumes mutual shadowing between trees (viewing and illumination shadows on
489 vegetative surfaces) are dominant factors controlling canopy bidirectional reflectance distribution
490 function (BRDF) (Li and Strahler 1992; Li et al. 1995).

491 Although the divergence in MODIS PRI values between backward (sunlit) and forward (shaded)
492 direction images were found in both Terra and Aqua data, the drought pattern in the sPRI ratios were
493 more obvious in the Aqua data, compared to the Terra. There are two possible explanations for this.
494 First, it might be mainly due to calibration issues of Terra by the problem in the solar diffuser door
495 (e.g. Franz et al. 2008), which affects more the shorter wavelength bands. This issue was expected to
496 be improved in the MODIS Collection 6, but it is quite possible that it did not completely remove the
497 problems identified in the Collection 5 (Wu et al. 2013). Note that Gamon et al. (2016) also only used
498 Aqua data to calculate CCI to monitor evergreen photosynthetic activity. Second, the drought stress
499 and subsequent photosynthetic downregulation are likely to be greater in the afternoon rather than in
500 the morning (Sims et al. 2005). The hourly LUE values from the flux tower also usually showed
501 strong diurnal patterns especially in the non-cloudy days (Figure S8). These LUE values and their
502 variances are often higher in the morning and lower in the afternoon, which might make drought
503 signals unclear for the Terra data.

504 Many coupled modeling efforts have shown that separate modeling between sunlit and shaded
505 portions of canopy was the most efficient way to simulate daily photosynthesis without multilayer
506 simulations (Chen et al. 1999; de Pury and Farquhar 1997; Song et al. 2009; Wang and Leuning 1998).
507 The dynamic separation between sunlit and shaded leaves is also justified in that the upper sunlit

508 canopy is usually light saturated (lower LUE), and therefore more vulnerable to drought stress, while
509 the lower shaded canopy responds linearly to irradiance with higher LUE (de Pury and Farquhar 1997;
510 Wang and Leuning 1998). This vertical transition of rate determining factors for photosynthesis
511 typically led to the distributions of leaf biochemical and structural traits along canopy depth profile,
512 such as carbon-to-nitrogen ratios and specific leaf area (Field, 1983; de Pury & Farquhar, 1997).
513 Recently, Coops et al. (2017) also showed that incorporating vertical gradients of PRI-derived LUE is
514 critical in upscaling leaf-scale physiological behavior into canopy-level GPP in the closed deciduous
515 canopy. However, to date few remote sensing based global GPP models incorporate two-leaf modeling
516 structures because of computational efficiency (Zhou et al. 2015). In this study, we suggest that it
517 would be critical to incorporate the two-leaf model structure to properly simulate the drought effect on
518 ecosystem carbon sequestrations under climate change considering emergent divergence between
519 sunlit and shade portions of canopy.

520 Parameterizing water use strategies into terrestrial ecosystem models will essentially require
521 mapping the distribution of anisohydric and isohydric trees across the different spatial scales. However,
522 mapping isohydricity from the parameter space along the isohydric/anisohydric continuum to
523 landscape space requires very detailed vegetation information with intensive field observations during
524 drought (Meinzer et al. 2016). To date, we are aware of only one attempt to develop a map of the
525 distribution of isohydricity across the landscape (Konings and Gentine 2016). That work relied on
526 remotely-sensed observations of vegetation optical depth using Advanced Microwave Scanning
527 Radiometer-E (AMSR-E) data, which has been previously shown to be correlated with leaf water
528 potential. Our approach suggests another path forward for using emergent spectral signatures during
529 droughts to map the distribution of isohydric and anisohydric species. This study presents a theoretical
530 framework for large-scale isohydricity mapping based on emergent spectral signatures under drought
531 stress using multi-angle MODIS images. This direct mapping of ecosystem sensitivity to drought
532 stress may represent a significant step forward from the simplistic drought scalars currently used in

533 remote-sensing based productivity models.

534

535 *Conclusions*

536 In this study, we examined the capability of four spectral indices to capture peak drought signals
537 and isohydric/anisohydric behavior in a deciduous broadleaf forest using in-situ spectral
538 measurements and multi-angle remote sensing images. After normalizing between sunlit (backward)
539 and shaded (forward) portions of canopy, PRI successfully captured both drought signals and species-
540 level drought responses at both leaf and canopy scales. Drought stress and subsequent photosynthetic
541 downregulation were greater in the sunlit portion of canopy, which provides us a key to capture
542 species- and canopy-level drought responses in the study site. Importantly, this work may represent a
543 significant step forward in our ability to dynamically predict the impact of water stress on forest
544 carbon gain by providing a theoretical framework for mapping large-scale isohydricity based on
545 emergent spectral responses under drought stress.

546

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555

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