

1 **Title**

2 LCA for territorial metabolism analysis: an application to organic waste management planning

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22 **Abstract**

23 The management of biodegradable waste in landfills is associated to a range environmental impacts
24 and refers to a linear model regarded as unsustainable. At the same time, many agricultural and urban
25 soils present low organic matter content. Composting is emerging as a good practice for converting
26 organic waste into a new resource able to improve soil properties, thus providing regulation
27 Ecosystem Services. Considering a territorial metabolism perspective, this work discusses the
28 potential contribution to regional planning of a transferable methodology for quantifying
29 environmental impacts and benefits associated to waste management, based on a combined use of
30 Life Cycle Assessment and Geographic Information Systems, and considering Ecosystem Services
31 deriving from the application of compost in agricultural soils. The approach was tested through a site-
32 dependent analysis based on primary data, referring to year 2019 and focusing on the Veneto Region,
33 Italy. Results indicate that direct benefits associated to the use of compost and the thermal energy to
34 district heating are compensating from 29% to 51% of the impacts associated to compost production,
35 while waste transports represent the largest share of the impacts, covering between 52% and 78% of
36 the total flows considered. The proposed methodology is applied for comparing the reference
37 condition to alternative scenarios, in the perspective of providing support in Strategic Environmental
38 Assessment procedures. In this context, results shown markedly lower impacts associated to compost
39 production, with respect to organic waste treatment, for 5 out of 6 of the considered categories
40 (freshwater eutrophication 100:1, climate change 5:1), with the notable exception of water resource
41 depletion. Scenarios produced are discussed with respect to the choice between centralized and non-
42 centralized plants, and the characterization of potential benefits at the territorial scale associated to
43 compost use for urban green infrastructures. With respect to this latter point, results showed an
44 underdeveloped use of compost-related ES flow, compared to its capacity, suggesting an interest for
45 further research aimed at estimating compost requirements by urban and peri-urban soils.

46

47 **Keywords:** Life cycle assessment, regulation ecosystem services, compost, organic waste, GIS,
48 urban policies

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51 **1. Introduction**

52 In 2015 the United Nations (UN) by publishing the Agenda 2030 and the Sustainable Development
53 Goals (SDGs) have brought high on the political agendas the need for management of the organic
54 fraction of municipal solid waste (OFMSW), especially of food fraction, which represents the most
55 consistent waste flow resulting from household (EEA, 2020; Tonini et al., 2020). The SDG 12
56 “Ensure sustainable consumption and production patterns”, in particular, aims to take urgent effort to
57 enhance resource efficiency and reduce waste (UN, 2015; UNEP, 2017). According to the Europe
58 2020 Strategy (EC, 2010), the bio-economy could provide an important contribution to the
59 achievement of the green targets in Europe in the upcoming decades (Pergola et al., 2018). A basic
60 principle of the bio-economy is promoting a sustainable and efficient (see e.g. Sepehri and
61 Sarrafzadeh, 2018) resources transformation and conversion into bio-energy and/or bio-based
62 products, reducing the dependency on natural resources (EC, 2012a). In this framework, composting
63 is a moderately easy and inexpensive way for stabilizing and reducing biodegradable waste (Crowe
64 et al., 2002). Compost represents an organic and biologically stabilized soil conditioner, obtained
65 from the treatment of the OFMSW. During the composting process, the microbial community
66 transforms the degradable organic matter into more stable forms, such as water (H₂O), carbon dioxide
67 (CO₂) and ammonia (NH₃), releasing heat as a metabolic waste product (Ciavatta et al., 1993).
68 Composting thus represents a solution to divert OFMSW from landfills, obtaining a new product
69 suitable, in particular, for agricultural purposes (Al-Rumaihi et al., 2020). In addition, integrated
70 composting and anaerobic digestion plants can provide useful resources to other supply chains, such
71 as energy and water recovery (Hannibal et al., 2018). At the European level, it has been estimated
72 that almost 50% of the entire amount of compost produced is used in agriculture (Saveyn and Eder,
73 2013). Other uses include gardening (13%), horticulture (11%), landscaping (10%), mixture (6%),
74 land restoration and landfill coverage (5%), while the remaining 4% for other uses including export
75 and wholesale (Corden et al., 2019).

76 Composting can be performed at three different scales: centralized composting acts on vast
77 geographic areas; decentralized composting acts at the community scale; finally, composting in
78 individual housing units acts at home/courtyard scale. Centralized composting, which is the most
79 widespread, involves both positive and negative aspects. On the one hand, it is a system that treats a
80 more significant quantity of waste and guarantees greater control of the compost production
81 processes; on the other, it involves high environmental costs for the waste collection and transport
82 over long distances (Bruni et al., 2020). In this context, an interesting approach supporting the
83 evaluation of costs and benefits associated to centralized composting actions can be identified in Life
84 Cycle Assessment (LCA). European Commission promoted LCA as a useful quantitative tool for
85 assessing the potential environmental impacts of products and processes, capable of supporting
86 producers and stakeholders (COM 302, 2003). LCA looks at the product's entire life cycle (extraction
87 of resources, raw material processing, production, usage, recycling, disposal of remaining waste) and
88 has been frequently used for waste and agricultural assessments. Literature review showed that there
89 is a strong connection between compost and regulation Ecosystem Services (ES) (Martínez-Blanco
90 et al., 2013), including carbon, macronutrients and water cycle regulation, pest and disease control
91 for crops, soil formation, pollination, and mitigation of natural hazards. Compost effectiveness in
92 terms of providing regulation services results from the interaction among multiple variables,
93 including OFMSW quality, compost maturity/stability, crop management, soil texture, soil organic
94 matter concentration (SOM), and soil temperature/humidity (Martínez-Blanco et al., 2013; Adhikari
95 et al., 2016). Although this relationship is widely recognized (e.g. Cortés et al., 2020), benefits from
96 compost application to soil have not been included in many LCA studies of compost, reducing its
97 environmental value (Department of Environment and Conservation, 2006).

98 The present study focuses on the usability of LCA to support environmental assessment of compost
99 production and utilization at the territorial scale. This topic is of interest within planning processes,
100 and is framed within a territorial metabolism approach, quantifying of negative and positive

101 environmental effects associated to OFMSW recovery, its treatment, and the use of produced
102 compost. This general purpose is pursued through two specific goals: a methodological one, aimed at
103 performing an LCA based on primary data collected at a centralized composting plant, and referring
104 to the year 2019; a second one focused on scenario building, of interest for comparing different
105 management alternatives. In terms of negative environmental effects, the focus is on emissions related
106 to waste collection and transport, and water and energy consumption for the operation of the plant.
107 In terms of benefits, the focus is primarily on regulation ES related to the use of compost, on energy
108 recovery, and on biomethane production (co-product). Data collection and analyses were carried out
109 in the framework of a project financed by the Veneto region (European Social Fund, Call 2020).

110

111 **2. Methodology and research phases: the LCA approach**

112 The selected case study considered an integrated composting and anaerobic digestion plant located in
113 the municipality of Este, in the district of Padova (Veneto region, Italy). The plant is managed by
114 S.E.S.A. S.p.A., a joint-stock company responsible for the collection, transport, selection, treatment,
115 recovery and disposal of municipal solid waste, the production of compost and energy recovery. This
116 plant represents a large-scale centralized operation and serves vast geographic areas and sectors. The
117 company has provided data related to the vehicle fleet and flows of organic waste, compost, energy
118 and material guaranteeing the operation of the plant.

119 The approach combined Geographical Information System (GIS) analysis with the Life Cycle
120 Assessment (LCA), coherently with the idea of territorial Life Cycle Assessment (Hiloidhari et al.,
121 2017). While GIS is used for spatial data acquisition, processing and visualization, LCA is a
122 standardized method used to quantify the environmental impacts related to a product or a process.
123 For the construction of the LCA analysis, following the ISO 14040 and 14044 standards (International
124 Organization for Standardization, 2006), the methodology was articulated in four steps:

125 - the “goal and scope”, which involves the definition of objectives and scope of the LCA study, the
126 functional unit (FU) and system boundaries;

127 - the “life cycle inventory” (LCI), which includes collecting data for all the processes included in the
128 analysis;

129 - the “life cycle impact assessment” (LCIA), which evaluates the significance of potential
130 environmental impacts based on the LCI flow results, and includes the selection of characterization
131 model and impact categories;

132 - the “interpretation”, in which results of the assessment are analyzed to identify processes of the life
133 stages that have relatively high environmental impacts.

134 GIS was the key tool for the processing of primary data during the LCI construction phase. Based on
135 existing examples (Mutel et al., 2012; Hiloidhari et al., 2017; García-Pérez et al., 2018), GIS was
136 used to set up an area-specific LCA analysis. A spatial analysis of primary data, focusing on OFMSW,
137 compost flows, and soil properties was carried out as a part of the study. Data on waste flows were
138 provided by the partner company, while those concerning the properties of the soils were collected
139 by the Veneto Region environment agency (ARPAV) as part of institutional monitoring activities.
140 The latter data were used for ES quantification. The base year for this analysis is 2019. The following
141 sections will introduce details on the different steps of the LCA methodology (2.1-2.3). Section 2.4
142 will therefore explain the rationale for scenarios comparison.

143

144 **2.1. LCA: Goal, functional unit, and system boundaries definition**

145 The goal of this study is to estimate the life cycle assessment (LCA) of compost produced in a
146 composting and anaerobic digestion plant. The functional unit (FU) chosen is the production of a
147 metric ton (t) of compost derived from organic waste processing, which provided a reference to
148 normalise material and energy fluxes in input and output to the system. The performed LCA study
149 (see Figure 1) considered the collection and transport of municipal solid waste, their processing in
150 the plant and the co-production of energy and bio-methane, the transport of compost, and the
151 regulation ES associated with the use of compost in agriculture, which was considered as the main
152 activity using the compost produced by the plant (data provided by the partner company show that
153 the 84% of the compost produced annually is delivered to the agricultural sector). Instead, the main
154 processes excluded from the LCA are construction of the plant infrastructure, and the use of compost
155 in urban areas. The choice of disregarding the first processes was based on the consideration of the
156 difficulty of evaluating properly the building lifespan and on the consideration that many studies
157 considered a long duration, ~ 50 year, for buildings with characteristics similar to the ones considered

158 (Ji et al., 2021). As for the use of compost in urban areas is concerned, although this accounted for
159 approximately 16% of the total compost produced (on a weight basis), the scarce amount of
160 information available for characterizing the diversity of urban soils prevented from considering it in
161 the baseline LCA, and was therefore considered in the scenario analysis (see par 2.4). Avoided waste
162 disposal in landfill and incineration were considered outside the system boundaries, and included
163 within the subsequent scenario analysis.

164 < **Figure 1** >

165

166 **2.1.1. Analysis and mapping of primary data: the transport system**

167 Based on the primary data provided by the company, an analysis of the transport system, quantities
168 and characterization of waste and compost flows was carried out. All these data were analysed from
169 a spatial point of view, through the software QGIS 3.14, with two main goals: i) obtaining a spatial
170 visualization of the data and mapping all the points of origin (O) of waste and destination (D) of the
171 compost; ii) calculating the mileage related to these flows, using the QGIS “shortest route” tool
172 (Figure 2). As regards the analysis of the transport system, the variables considered are the amount
173 of materials transported (metric tons, here in tons), the distances travelled (km), the type of vehicle
174 differing in terms of capacity (7.5-16 and 16-32 tons) and fuel (diesel or biomethane).

175 < **Figure 2** >

176

177 **2.1.2. Analysis and mapping of primary data: assessing ecosystem services**

178 The characterization of soil parameters allowed to quantify the potential ecosystem services linked to
179 the use of compost in agriculture. This potential is associated with the increase in soil organic carbon

180 (SOC), sequestered CO₂eq, available water capacity (AWC) and macronutrients (N, P, K)
181 concentration in the soil. The analysis focused on the agricultural area of the 34 municipalities in
182 Veneto Region in which the partner company sold the compost for agriculture use in 2019. Areas
183 were defined based on level II of the Corine Land Cover in 2018 (<https://land.copernicus.eu/>).
184 Concerning these municipalities, the main soil parameters have been defined, based on the primary
185 data provided by the Veneto Region Environmental Agency data (ARPAV): the type of soil (sand,
186 silt, clay), SOC in the surface layer of the soil (first 50 cm), expressed both as a percentage (%) and
187 in tonnes per hectare (t ha⁻¹), AWC expressed in millimetres (mm) and subsequently converted in kg
188 (figure 3). As regards the sequestered CO₂eq and AWC, this quantification relies on the increase in
189 soil organic carbon (SOC), assuming that the ideal amount of compost, identified as 30 tons per
190 hectare (ECN, 2010 ; ISWA, 2020; Colombani et al., 2020), is applied to the soil each year over 20
191 years. The prerequisite for this evaluation is that the compost is of quality and therefore does not
192 present physico-chemical contamination due to a mixture of non-organic waste. Estimated SOC
193 increase rates vary depending on two starting soil parameters: the soil texture and the percentage of
194 SOC, differentiated between lower than 2% and higher than 2%, following the European Soil Bureau,
195 which has classified European soils based on their SOC content (Table 1).

196 <Figure 3>

197 <Table 1>

198 Once these percentages of SOC increase for soil texture with different initial organic carbon contents
199 were defined, specific analyses for the 34 municipalities considered were carried out. For each
200 municipality, the agricultural area is defined by the soil texture (sandy, loamy and clayey),
201 distinguishing the areas with low and high SOC (respectively <2% and > 2%), in order to establish
202 the percentage of organic carbon increase, to be applied to the starting SOC value, expressed in
203 tons/hectare (each municipality has multiple parcels of agricultural soil in which these parameters are

204 variable and can be associated with different percentages of SOC increase; therefore, for defining a
205 single percentage of SOC increase per municipality, a weighted average is carried out with respect to
206 the extension of each soil texture). After obtaining an increased value of SOC for each municipality,
207 this was converted into CO₂eq sequestered in the atmosphere (kg of CO₂eq).

208 The increase in SOC defined for each municipality was then used as a basis to calculate the rise in
209 soil water availability (AWC), applying a corresponding increase of 2.1% of soil AWC to every 1%
210 increase in SOC. To define the rise in soil nutrient macronutrients (N, P, K), different increase values
211 were applied (expressed in kg tons⁻¹ of compost) depending on the type of compost applied (Table
212 2), whether mixed (ACM) or green (ACV). Once these ES were quantified for the agricultural area
213 of the 34 municipalities analysed, the values obtained were compared to the real ha in which the
214 compost was applied. Assuming that the optimal amount of compost for 1 ha is 30 tons, for each
215 municipality the amount of compost delivered by S.E.S.A. was divided by 30. Finally, the quantified
216 ES for each municipality were normalized to the considered period (1 year), and scaled to the
217 functional unit, 1 ton of compost produced (Table 2).

218 <Table 2>

219

220 **2.2. Life cycle inventory (LCI)**

221 The processing of primary data related to flows of matter and energy involved in the compost supply
222 chain, allowed to produce the inventory (LCI) reported in Table 3. Inputs, resources, energy and waste
223 associated to each process were quantified based on the Ecoinvent 3 database, allocating waste
224 disposal to the compost production process. All values were scaled to the functional unit of 1 ton of
225 compost. Energy production and consumption of integrated waste treatment systems were the subject
226 of different works (e.g. Colón et al., 2012; Fernández-Rodríguez et al., 2016). The analysis performed

227 here took into consideration biomethane as a compost co-product. Biomethane produced was used in
228 part for transport operation, and therefore introduced in the analysis by constraining the impacts of
229 biomethane vehicles, both small and medium. This was done by setting to 0 the fuel consumption of
230 such vehicles, while leaving all the other emissions associated to transportation (infrastructures,
231 vehicles, combustion). A fraction of biomethane was distributed to the area nearby the plant in the
232 form of district heating. Avoided impacts associated to this practice were quantified by accounting
233 for the impacts associated to methane production. Regulation ES associated to the use of compost
234 were accounted in the model by introducing in the inventory avoided CO₂ emissions in the
235 atmosphere. AWC increase was accounted by considering avoided consumption of groundwater for
236 irrigation, while the N, P and K increase in the soil were accounted as avoided production of N, P, K
237 fertilizers.

238 <Table 3>

239

240 **2.3. Life cycle impact assessment (LCIA): impact categories and methods**

241 LCA was performed in SimaPro v. 84 (PRé Sustainability), allowing the classification and
242 characterization phases defined by the ISO regulation (2006). In the classification stage, each burden
243 is linked to one or more impact categories, while in the characterization stage the contribution of each
244 burden to the environmental impact categories (EICs) is calculated by multiplying the burdens by a
245 characterization factor (Guinée, 2001). In accordance with Martínez-Blanco et al. (2009) and Avadí
246 (2020), we considered for the analysis two evaluation methods: the “ILCD 2011 Midpoint”, including
247 climate change (CC), ozone depletion (OD), acidification (AC), freshwater eutrophication (EU),
248 water resource depletion (WD), and the “Cumulative Energy Demand” (CED) including the non-
249 renewable fossil (NF). Emissions to air, soil, surface and groundwater, and resource consumption are
250 considered with respect to their contribution to these EICs and are associated to the different

251 processes. These EICs have been related to the main processes analyzed in the life cycle of compost
252 and included in Table 3.

253 **2.4. Steps of the analysis: reference and scenarios**

254 The quantification of negative and positive environmental effects associated to OFMSW was carried
255 out in three steps. First the reference condition (REF) was assessed, performing LCIA for the base
256 year 2019. As a second step, an exploratory analysis was carried out, by comparing to REF the
257 following three scenarios, here synthetically presented:

258 1) the substitution of diesel-fueled vehicles with bio-methane vehicles. This comparison was
259 suggested by a recent change in the company fleet, which was partly converted to bio-methane power
260 supply. Being based on company management actions, this scenario analysis was named “company
261 management” (MGM);

262 2) the alternative treatment of OFMSW by landfill. This is regarded as a worst-case alternative,
263 considering an entirely linear approach in treating the organic matter currently converted into
264 compost. For this reason, this scenario was named “worst case” (WRS);

265 3) how current ES flows can increase considering the capacity of urban areas (ES-URB). Literature
266 on ecosystem services addresses the difference between flow and capacity, which distinguishes the
267 current ES supply from the potential supply (Burkhard et al., 2012; Baró et al., 2016). The analysis
268 focused on the potential use of compost for the maintenance of existing green infrastructures and the
269 ES related to the management of public green spaces. In fact, green infrastructures in urban contexts
270 can benefit from the application of compost, being urban soils generally low in organic matter (Sæbø
271 et al., 2006). ES-URB was applied to the city of Padova, the closest one to the plant, with the idea of
272 pursuing the utilization of compost within the area of waste production, in order to promote a higher
273 circularity of the biogeochemical flows. Estimation of ES capacity in the urban context, was

274 performed assuming to apply 30 tons of compost per year for each hectare of urban green
275 infrastructure.

276 3. Results

277 Results of the LCIA for REF (reference condition) are illustrated in Figure 4, showing the assessment
278 obtained for the 6 impact categories selected for the analysis (CC, OD, AC, EU, WD, NF). Positive
279 values indicate impacts on the environment, while negative ones stand for avoided emissions, while
280 different colours indicate the contribution of processes to the overall category. Results show how the
281 direct benefits associated to the use of compost and the thermal energy to district heating are
282 compensating, considering an average among all the impact categories, 29% of the impacts, with
283 peaks of 51% for EU and 31% for AC.

284 <Figure 4>

285 In all the categories, OFMSW transportation represents the largest share of the impacts, with values
286 ranging between 52% and 78% of the total flows considered. Waste landfill represent the second term
287 in the budget, although its contribution is always below 30% (achieved in the case of climate change).
288 As expected, the renewal of the fleet and its transition to biomethane-fueled vehicles, under the
289 company management scenario, MGM (Figure 5), implied a decrease in all the indicators considered,
290 with higher effects on CC and a more restricted on AC.

291 Figure 6 compares the net balance given by the sum of positive and negative contributions for each
292 category calculated in REF, to the WRS scenario (impacts associated to OFMSW waste disposal by
293 landfill), showing a markedly lower impact for all the categories, with the notable exception of WD.
294 These differences are more pronounced for EU (~100:1), and CC (~5:1). The difference in impacts
295 for the various categories can be explained based on the combination of processes included in the
296 REF condition and the WRS scenario, and their relative share. Effects on the different matrices and
297 the associated impact categories are ultimately defined by this share.

298 <Figure 5>

299 <Figure 6>

300

301 ES-URB scenario estimated the potential for ecosystem services associated to the use of compost in
302 the urban context (Table 4). Urban green areas considered (Figure 7), included: uncultivated green
303 areas (23 ha), sports fields (175 ha), green areas associated with viability (100 ha), urban parks and
304 permanent lawn areas with spontaneous grassing (247 ha), for a total of 719 ha. Padova has a silty
305 soil texture and a low SOC percentage (< 2%) homogeneous throughout the territory, which defines
306 an increase of 46% in SOC. This percentage, applied to the current amount of SOC, leads to an
307 estimation of $2.4 \cdot 10^6$ kg of CO₂eq sequestered, $71.9 \cdot 10^6$ kg of AWC, $3.8 \cdot 10^5$ kg of N, $1.4 \cdot 10^5$ kg of
308 P, and $2.1 \cdot 10^5$ kg of K per year.

309

310 <Table 4>

311 <Figure 7>

312

313 4. Discussion

314 The study of biogeochemical fluxes associated to organic matter exchanges in the territory can be
315 framed within the broader context of territorial metabolism (Wolman, 1965), and analysed treating
316 cities as heterotrophic systems (e.g. Odum, 1963; Grimm et al., 2000). In this perspective, compost
317 production and its use pose challenges and opportunities, of interest for territorial planning. CC values
318 obtained in this study for the REF condition, ~ 2000 kg CO₂ eq per ton produced, are higher than
319 previous estimations by Cortes et al. (2020), 472.59 kg CO₂ eq which, nonetheless used as functional
320 unit 1 ton of feedstock mixture fed to the composting facility, thus making difficult the overall

321 comparison. Avadí et al. (2020) in a screening LCA study carried out in France, reported values closer
322 to those found here, ranging between 517 and 3084 kg CO₂ eq for ton of compost produced. Beside
323 the comparison of the absolute values, it is worth nothing that these analyses are referring to territories
324 characterized by different specificities, both in terms of waste production and transport systems (fuel,
325 road infrastructures availability). In order to support territorial planning it seems of relevance, instead,
326 the possibility of remaining site-specific, associating alternative scenarios to the analysis of the
327 reference state. This was attempted in this work with the goal of providing an approach aligned with
328 the requirements of strategic environmental assessment procedures, which in EU are accompanying
329 the development of territorial plans (Directive 2001/42/EC). The REF scenario was first compared
330 with WRS, indicating that, for 5 out of 6 indicators (WD the notable exception), the linear model
331 production-consumption-disposal, is far less efficient than the one based on organic matter re-
332 circulation through compost (differences ~ 100:1 for EU). In this respect, it seems of interest to direct
333 future research efforts at improving the understanding of water metabolism, by performing a water
334 footprint analysis (e.g. Boulay et al., 2018), also in consideration of the expected trends in water
335 scarcity events affecting this specific region (<https://cordex.org/>).

336 Results showed how the conversion to bio-methane have led to a reduction of the environmental
337 footprint, quantifiable with an average decrease of 28% considering all the categories, 10% in the
338 case of CC and 58% for OD. This scenario, MGM, allowed to show the potential of an integrated
339 GIS-LCA analysis for supporting the design of management choices aimed at putting the circular
340 economy concept into practice (Sassanelli et al., 2019). On the other hand, the high impact of the
341 transport sector points the attention to the potential limitations of centralized plants operating at a
342 large scale, due to the high costs of collecting and transporting waste over long distances, as also
343 discussed by other studies (De Feo et al., 2016). In this respect, existing literature (Righi et al., 2013;
344 Bruni et al., 2020) highlight how decentralized composting systems, on a community and/or
345 neighbourhood scale, could minimize negative impacts on the environment. This aspect can be

346 regarded as relevant for fostering the development of innovative decentralized models of organic
347 waste valorisation (e.g. Giwa et al, 2022), although territorial-specific considerations must be
348 performed to consider the real applicability of these solutions in alternative contexts. Indeed,
349 decentralized composting is currently not a common practice in Italy and does not benefit from
350 adequate support from local authorities and citizens, who consider this infrastructure unsuitable and
351 dangerous for urban contexts. Sensitizing the population to the waste issue could be the first step to
352 provide information of what happens to waste once they have been collected and raise collective
353 awareness of the real need to reuse waste and transform them into new resources.

354 Green Public Procurement (EC, 2012b) is part of the strategies identified by the European Union to
355 implement the Integrated Product Policy (COM 68, 2001; COM 302, 2003), which constitutes an
356 approach aimed at reducing the environmental impact of products and services over their entire life
357 cycle, in a broader perspective of sustainable development. Compost, as defined by Italian national
358 legislation, is one of these products defined as recycled goods and should therefore be included in
359 public expenditure. Worldwide, about one-third of the soil is in conditions of degradation due to
360 erosion, decrease in organic matter, waterproofing, salinization, contamination, compaction, decrease
361 in soil biodiversity, floods and landslides (COM 347, 2002). Urban soils in particular can be highly
362 affected by human activities and, therefore, in many cases must receive a pre-treatment, in order to
363 facilitate the supply of ES and meet the needs of urban inhabitants (Morel et al., 2015). Adding
364 compost to urban soil for planting new tree species, leads to improved tree growth, and associated
365 regulation and habitat services, although it may take several years before its benefits could manifest
366 themselves (Oldfield et al. 2015). In some cases, a proper soil restoration intervention is required,
367 before any type of further action can take place – ecosystem restoration actions are receiving
368 increasing attention by policies at the EU-level (COM 304, 2022). Scenario ES-URB, rooted on the
369 distinction between ES flow and capacity, presents a first accounting of the potential regulation
370 services provided within the administrative boundaries of the city of Padova, estimated in $2.41 \cdot 10^6$

371 kg CO₂ eq, representing approximately 2.7 times the overall flow associated to the current REF
372 condition. It is worth remarking that this scenario is not analysed through a further LCA, and thus not
373 considers the additional quantity of organic matter required to produce more compost, its source
374 location, and the associated transport burdens. It also assumed that urban soils have the same
375 requirements in terms of amendment use than agricultural areas.

376 Results showed how the production of biogas, which self-feeds the plant and part of the transport
377 fleet, entails additional benefits through the district heating, quantified in 11.6 kg CO₂ eq, which
378 represents almost 12% of the CO₂ eq associated to compost use. This result highlights that compost
379 use has not yet reached a large market within Veneto Region. The current compost flow is associated
380 with a small portion of the agricultural soil, which could potentially be extended, guaranteeing a
381 higher level of CO₂eq sequestration, improving the quality of soils and thus reducing their
382 degradation. An interesting possibility to extend the flow-capacity scenario carried out within this
383 work, would be estimating the potential provision of regulating ES associated with the use of compost
384 in the entire region. ES quantification performed in this work did not consider the type of crops due
385 to the lack of data. This lack did not allow to include additional ES in the analysis, such as crop yield
386 improvement, crop nutritional quality, pest and disease suppression; moreover, knowing the types of
387 crops, it would have been possible to define more realistic ES values since each crop can react
388 differently to the compost. Another limitation concerns the close relationship between the
389 quantification of ES and the quantity and quality of compost (Ramos, 2017). This study assumes a
390 high compost quality, but its parameters (level of humidity, nutrient content, etc.) were not analysed.
391 Another perspective for expanding the present work would therefore be the use of soil
392 biogeochemical models (e.g. Abrahamsen and Hansen, 2000), to study organic matter degradation in
393 different conditions, thus allowing to relate more specifically compost features to its effects. In this
394 case, an attempt should also be made to bridge local modelling efforts (data intensive) to the wide
395 scale picture, which is provided by operational data collected by regional environmental agencies,

396 and used for territorial planning and management purposes. The final aim would be to improve the
397 picture at the territorial scale, rather than obtaining very detailed local-level descriptions.

398

400 **5. Conclusions**

401 This work allowed to identify and test a novel and transferable methodology for quantifying
402 environmental impacts and benefits associated to compost production and its utilization at the
403 territorial scale, based on the combined use of LCA and GIS, and considering ES deriving from the
404 application of compost in agricultural soils. The proposed comparison of alternative scenarios provide
405 a means for integrating this type of evaluation in Strategic Environmental Assessment procedures,
406 thus supporting regional planning efforts. Results pointed out a lower impact associated to compost
407 production with respect to organic waste treatment, for most of the impact categories considered, with
408 the notable exception of water resource depletion. A major contribution of transports to the overall
409 impact was detected, covering up to 78% of the total flows. Scenario comparison indicate an
410 underdeveloped use of compost-related ES flow, compared to its capacity. Further research in this
411 area should target a more robust estimation of ES capacity of urban soils at the regional scale, and
412 extending the capacity evaluation to agricultural soils, also including considerations on site-specific
413 biogeochemical conditions.

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419 activities of S.E.S.A. Spa to support the construction of the Inventory. The authors would like to
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421

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565 **Figure captions**

566

567 Figure 1. LCA Analysis. The square represents the system boundaries, i.e. processes, inputs and
568 outputs taken into account in the LCA analysis (what is outside the square is excluded from the
569 present analysis). Processes considered relevant in the present case study are: transport, avoided
570 disposal, waste processing and compost use.

571

572 Figure 2. Mapping of analysed organic matter flows within the Veneto Region (north-east of Italy).
573 A) wet waste flows origin and quantity per municipality; B) green waste flows origin and quantity
574 per municipality; C) compost flows, aggregated based on their different use.

575

576 Figure 3. Properties of agricultural soils in the municipalities of the Veneto Region where the compost
577 was used in 2019 (Elaboration based on data provided by CLC – 2018 and ARPAV – 2015, 2016).
578 These represent the main soil properties that influence the benefits given by compost (in this case,
579 compared to the quantity of 30 t/ha).

580

581 Figure 4. Results of the Life Cycle Impact Assessment for the Reference condition (REF).
582 Contribution analysis showing the weight of each process on the different EICs. Impact categories
583 considered: climate change (CC), ozone depletion (OD), acidification (AC), freshwater
584 eutrophication (EU), water resource depletion (WD), and non-renewable fossil (NF).

585

586 Figure 5. MGM scenario: transport processes for the REF condition. The current fleet partly
587 composed by biomethane-fueled vehicles (BT) is compared with a former fleet, fully composed by
588 diesel-fueled traditional vehicles (TV).

589

590 Figure 6. Results obtained for REF (reference condition, net balance) compared to impacts associated
591 to WRS (worst case, OFMSW waste disposal by landfill). Impact categories considered: climate
592 change (CC), ozone depletion (OD), acidification (AC), freshwater eutrophication (EU), water
593 resource depletion (WD), and non-renewable fossil (NF).

594

595 Figure 7. ES-URB scenario. Mapping of the green infrastructures of the Municipality of Padua,
596 categorized by 5 main types.

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606 **Tables**

607 Table 1. Annually increase of Soil Organic Carbon (SOC) and soil macro-elements (N, P, K) with
608 the use of 30 t/ha of mixed and green compost, for different soil textures with different initial SOC
609 contents. Data based on ISWA (2020).

Soil texture	Increase in SOC percentage in soil with SOC < 2% with 30 t/ha of compost	Increase in SOC percentage in soil with SOC > 2% with 30 t/ha of compost	Increase in soil macro-elements (kg/t compost)	Mixed compost	Green compost
Sand	40 %	20 %	Nitrogen (N)	17.9 kg	13.7 kg
Silt	46 %	23 %	Phosphorus (P)	6.3 kg	5.1 kg
Clay	55 %	27 %	Potassium (K)	9.8 kg	9.9 kg

610

611 Table 2. Quantification of regulation ES related to the use of compost, including: CO₂eq
 612 sequestration; Increase of available water capacity (AWC); Increase of soil nitrogen (N); soil
 613 phosphorus (P) and soil potassium (K).

Ecosystem services	Benefits in one year (kg)	Benefits for one ton of compost (kg)
CO ₂ eq sequestration	0.89 10 ⁶	124
Increase of available water capacity (AWC)	26.55 10 ⁶	3.7 10 ³
Increase of soil nitrogen (N)	0.13 10 ⁶	17.8
Increase of soil phosphorus (P)	0.05 10 ⁶	6.3
Increase of soil potassium (K)	0.07 10 ⁶	9.8

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615

616 Table 3. Life Cycle Inventory compiled within this work. Processes were aggregated in Transport,
 617 Waste to final disposal, Electricity, Thermal energy, Water, Avoided waste to final disposal, and
 618 Ecosystem Services from compost use; the LCI reports, for each process, the value corresponding to
 619 one ton of compost with a specific unit of measurement.

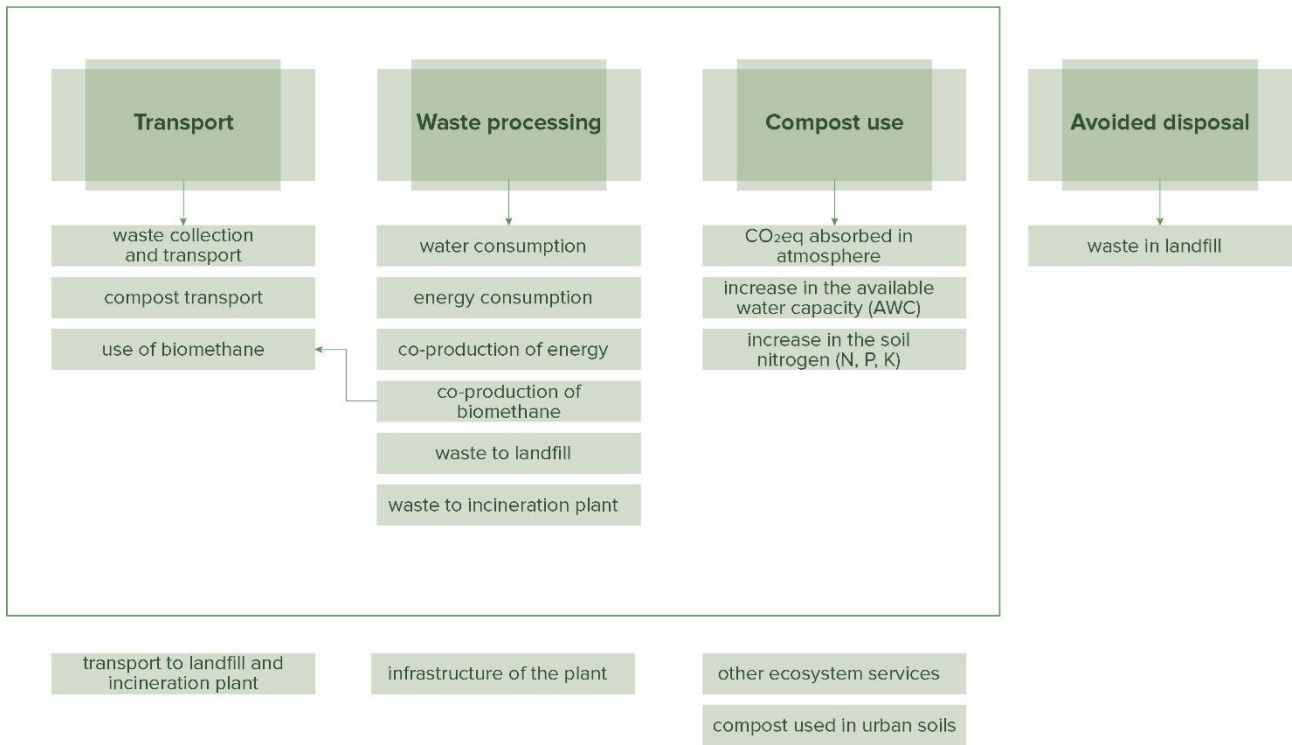
Category	Process	Value	Units
Transport	transport with small biomethane vehicles	2247	tkm
	transport with medium biomethane vehicles	1181	tkm
	transport with small diesel vehicles	627	tkm
	transport with medium diesel vehicles	308	tkm
Waste disposal	landfill	1.77	t
	incineration	0.04	t

Energy	electricity consumption	1540	kWh
	co-production of electricity	3214	kWh
	Thermal energy to district heating	49	m ³
Water	water consumption	985	kg
	Water recovery	3638	kg
Avoided waste to final disposal	avoided waste in landfill	20.03	t
Regulation services	CO ₂ eq storage	124	kg
	increase in the available water capacity (AWC)	3698	kg
	increase in the soil nitrogen (N)	17.87	kg
	increase in the soil phosphorus (P)	6.30	kg
	increase in the soil potassium (K)	9.80	kg

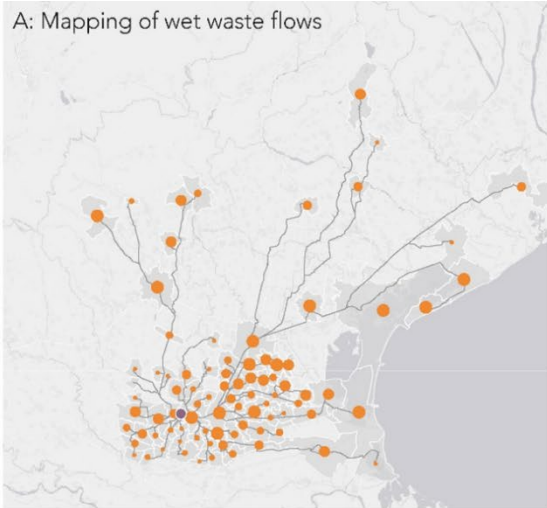
620

621 Table 4. REF and ES-URB scenarios: flow and capacity for the 4 regulation ES considered in this
622 work. Capacity estimation was based on the green infrastructures in which compost could be used
623 within the municipality of Padova.

Regulation ES	REF Flow of ecosystem services for one year (kg)	ES-URB Capacity of ecosystem services for one year (kg)
CO ₂ eq storage	0.89 10 ⁶	2.41 10 ⁶
Increase of available water capacity	26.55 10 ⁶	71.89 10 ⁶
Increase of soil nitrogen (N)	0.13 10 ⁶	0.38 10 ⁶
Increase of soil phosphorus (P)	0.05 10 ⁶	0.14 10 ⁶
Increase of soil potassium (K)	0.07 10 ⁶	0.21 10 ⁶



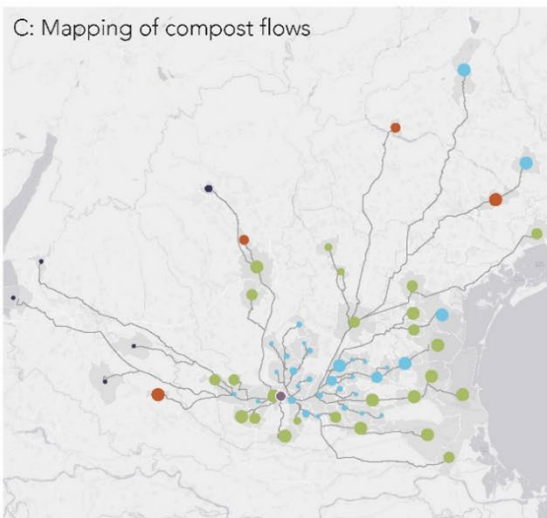
A: Mapping of wet waste flows

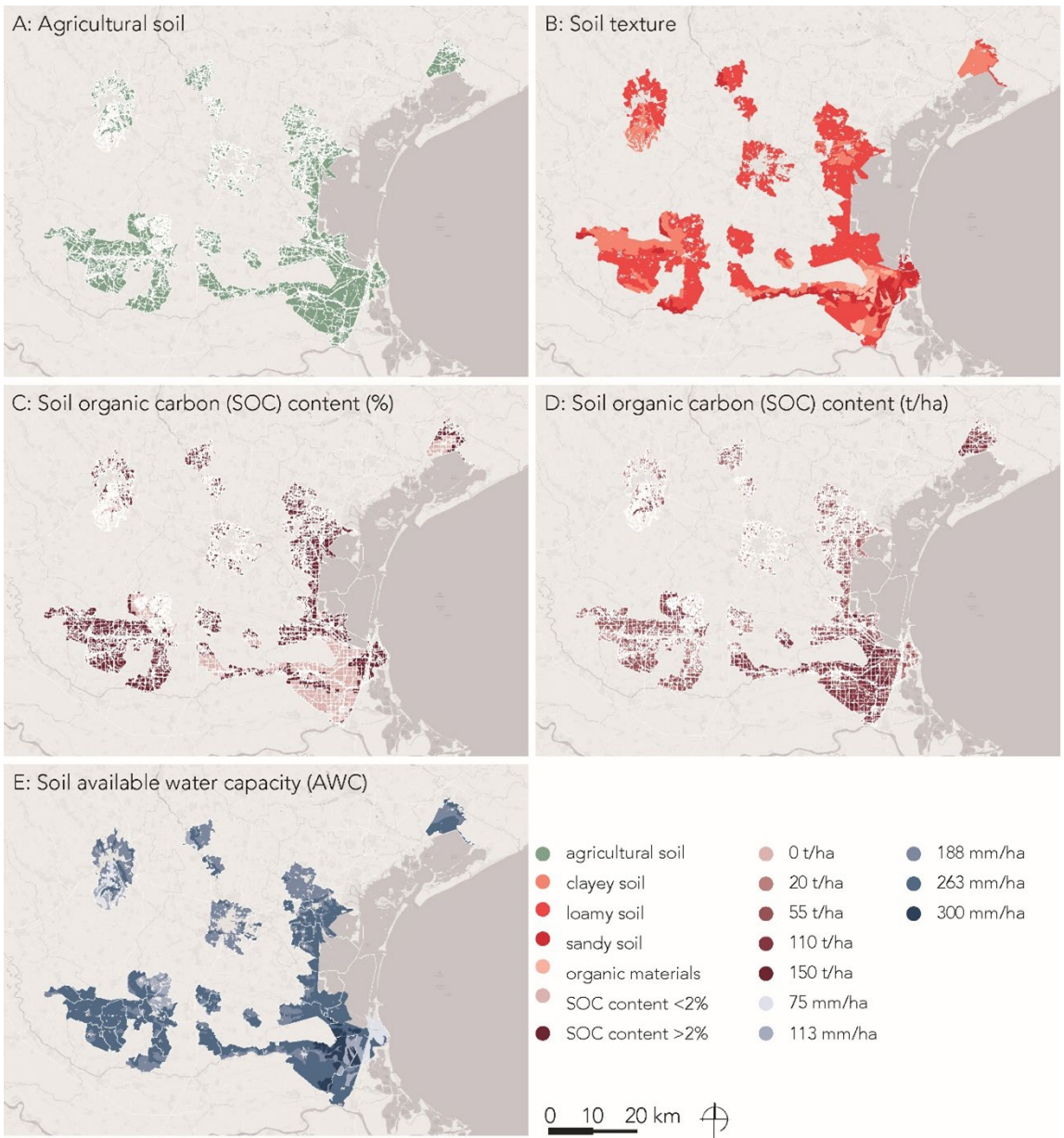


B: Mapping of green waste flows

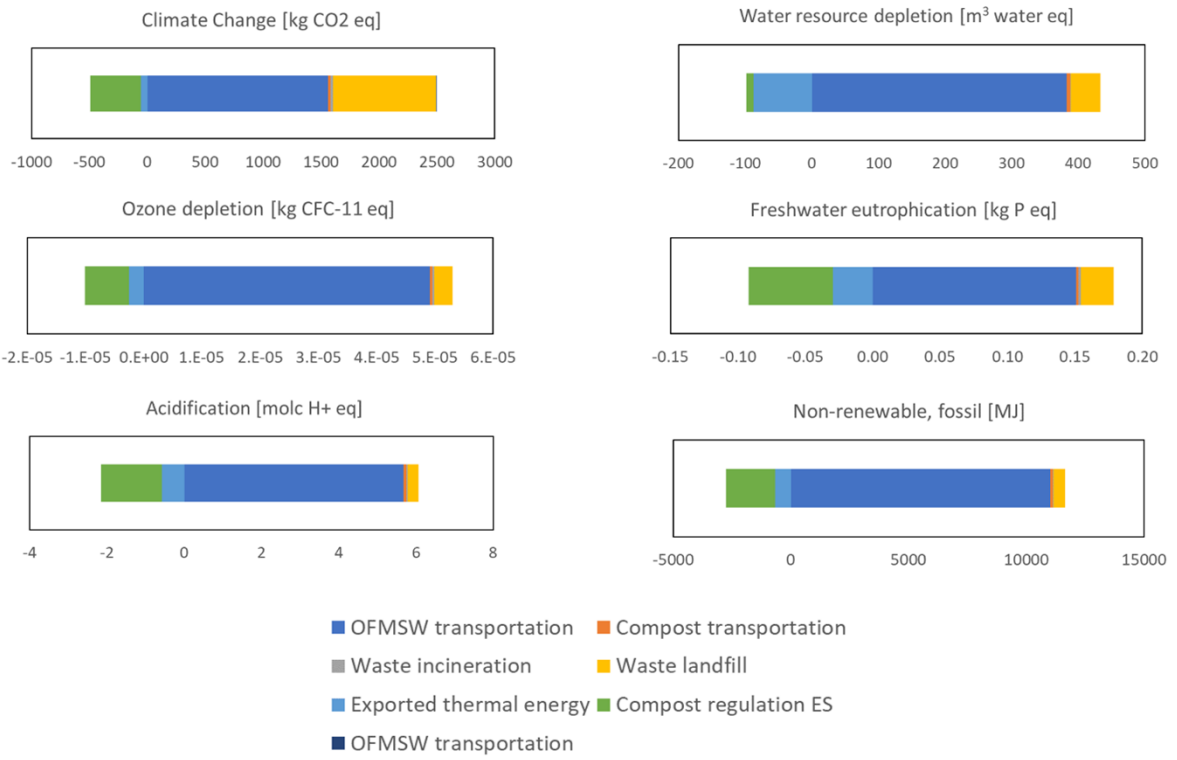


C: Mapping of compost flows

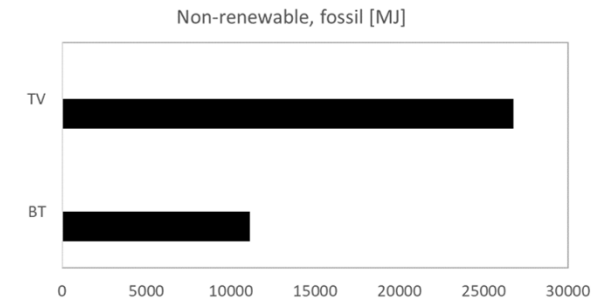
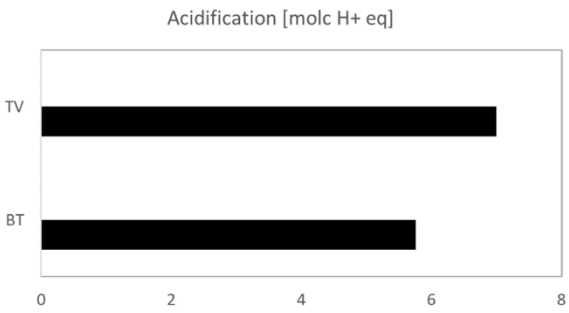
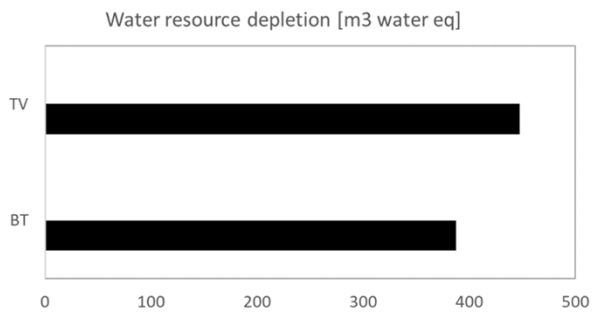
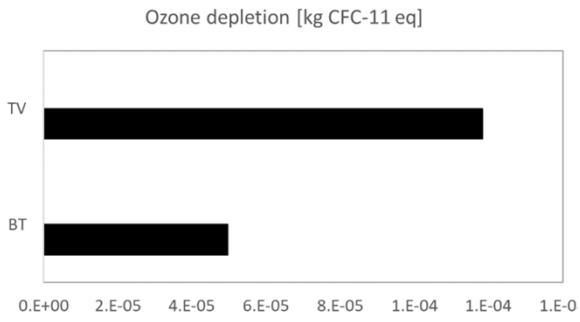
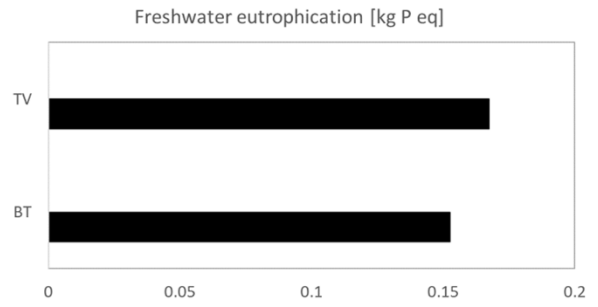
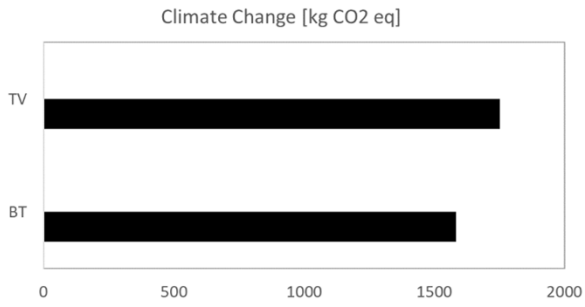




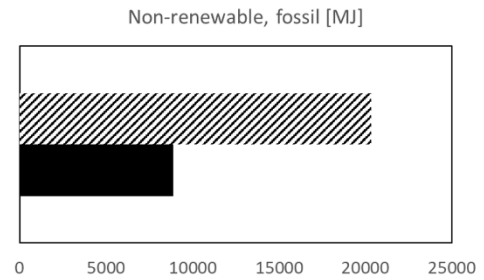
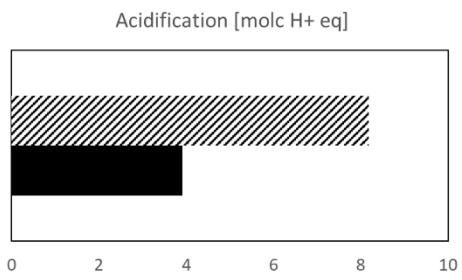
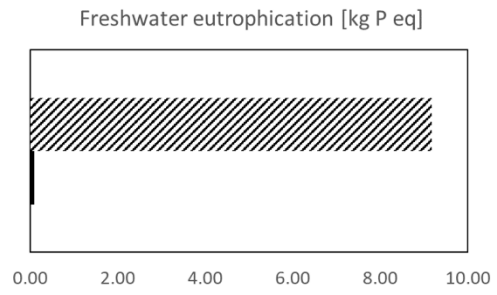
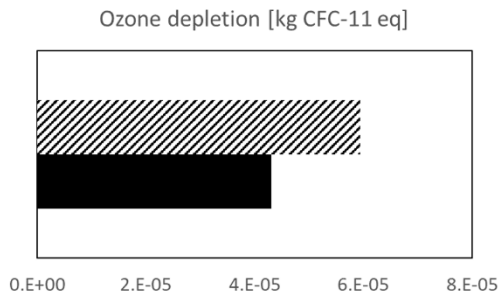
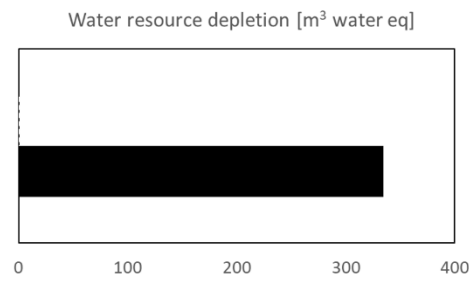
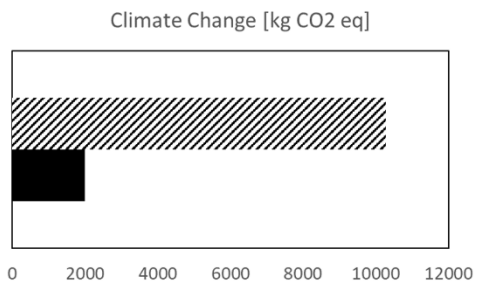
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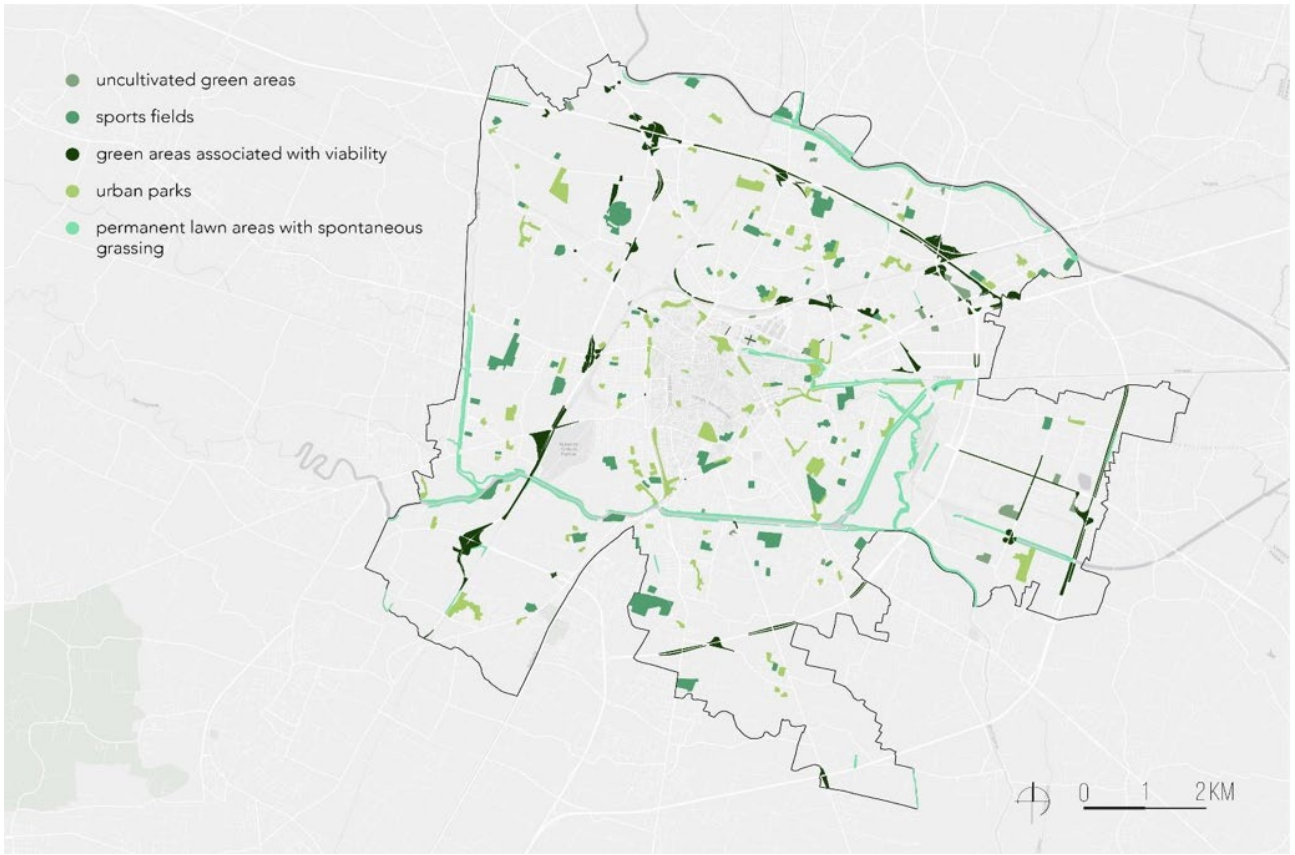
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▨ WRS ■ REF



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