

1 **Environmental DNA as a complementary tool for biodiversity**  
2 **monitoring: A multi-technique and multi-trophic approach to**  
3 **investigate cetacean distribution and feeding ecology**

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## 24 Abstract

25 The use of environmental DNA (eDNA) to assess the presence of biological communities has  
26 emerged as a promising monitoring tool in the marine conservation landscape. Moreover,  
27 advances in Next-Generation Sequencing techniques, such as DNA metabarcoding, enable  
28 multi-species detection in mixed samples, allowing the study of complex ecosystems such as  
29 oceanic ones. We aimed at using these molecular-based techniques to characterise cetacean  
30 communities, as well as potential prey in the northern coast of Mainland Portugal. During  
31 seasonal campaigns, we collected seawater samples, along with visual records of cetacean  
32 occurrence. The eDNA extracted from 64 environmental samples was sequenced in an  
33 Illumina platform, with universal primers targeting marine vertebrates. Five cetacean species  
34 were identified by molecular detection: common dolphin (*Delphinus delphis*), bottlenose  
35 dolphin (*Tursiops truncatus*), Risso's dolphin (*Grampus griseus*), harbour porpoise (*Phocoena*  
36 *phocoena*) and fin whale (*Balaenoptera physalus*). Overall, except for the fin whale (not  
37 sighted during the campaigns), this cetacean community composition was similar to that  
38 obtained through visual monitoring, and the complementary results suggest their presence in  
39 the region all year round. In addition, the positive molecular detections of *B. physalus* are of  
40 special relevance since there are no visual records reported in the area. The detection of  
41 multiple known preys of the identified dolphins indicates they use these coastal areas for  
42 feeding purposes. While this methodological approach remains in a development stage, the  
43 present work highlights the benefits of using eDNA to study marine communities, with specific  
44 applications for research on cetacean distribution and feeding ecology, ultimately serving as  
45 the baseline of a methodological approach for biodiversity monitoring and marine  
46 conservation.

47

48 **Keywords:** eDNA; DNA Metabarcoding; Continental Portugal; Marine Vertebrates; Next-  
49 Generation Sequencing

50

## 51 **Introduction**

52 Environmental DNA (eDNA) is an emerging tool for biodiversity monitoring that has been  
53 gaining prominence in scientific research during the second half of the 21st century, with  
54 increasing numbers of scientific outputs being published every year [1]. For marine  
55 ecosystems, the application of eDNA detection methodologies is still in its infancy, although it  
56 has steadily evolved in the last decades given the interest in its potential [2]. Among the  
57 multiple applications, these allow verifying environmental health by studying microbial  
58 communities, characterising and quantifying stocks of marine vertebrates, such as teleost fish  
59 which represents vital knowledge for good management of the fisheries sector, and  
60 determining the presence and abundance of elusive species, such as cetaceans, allowing a  
61 greater effectiveness of the monitoring work [3-5]. Additionally, technological advances in  
62 Next-Generation Sequencing (NGS) techniques, via DNA metabarcoding, allow for  
63 simultaneous multi-species detections in environmental samples, permitting the study of  
64 multiple trophic levels within the same samples [6-8].

65 Cetaceans are widely dispersed mammals that inhabit most marine environments, from  
66 coastal habitats to neritic waters [9], playing a key ecological role in maintaining the balance  
67 of these ecosystems [10]. As keystone species, the conservation of cetaceans is often a top  
68 priority in international agreements, especially considering the anthropogenic threats they are  
69 currently facing [11]. Addressing the impacting issues in cetacean ecology is crucial, however,  
70 obtaining data that provides a detailed understanding of these animals is rather complex.  
71 Cetaceans are elusive individuals, spending the vast majority of their time underwater. Also,  
72 their distribution range is often very extensive, including areas where access for sampling is  
73 difficult due to logistical and financial limitations inherent to the marine wildlife monitoring work  
74 or even legal constraints [10]. Therefore, the development of new non-invasive methodologies,  
75 such as the metabarcoding analysis of eDNA samples, especially in complement to visual  
76 monitoring is a promising advance in the optimisation of monitoring effectiveness towards the  
77 better understanding of these highly complex species [10].

78 Especially regarding the use of eDNA samples for cetacean monitoring, few published studies  
79 were focused on marine mammals (see review on eDNA application to cetacean monitoring  
80 under [12]). Nevertheless, there are already successful case studies where it has been  
81 possible to identify a variety of cetacean species through environmental samples using both  
82 species-specific [5;13-16] and universal primers [4;17-23]. The possibility of detecting multiple  
83 species within the same environmental samples, enabled through metabarcoding, allows for  
84 a multi-trophic analysis that widens the utility of the samples for the monitoring of several taxa  
85 and application to various fields of research, including the study of trophic chains and species  
86 feeding ecology [23,24].

87 In the present study, a molecular detection methodology was developed, using eDNA samples  
88 as a tool for biodiversity monitoring, especially applied to cetacean species in a coastal region  
89 of the North of Portugal, located in the Eastern North Atlantic (ENA). The ENA region is an  
90 area of great interest regarding the diversity and abundance of cetaceans, with several  
91 different species of dolphins and whales being recurrently reported over the years [25-34].  
92 Specifically, on the northwest Iberia, there is a high richness of cetacean species that feed on  
93 living marine resources [26;35;36]. Here, we sought to obtain additional and concrete data on  
94 the occurrence of cetaceans in this area and infer the ecological reasoning behind it by  
95 recurring to a universal approach to perform a multi-trophic analysis. Furthermore, we  
96 compared the eDNA results on cetacean species detection with the data obtained by  
97 traditional visual monitoring techniques, in order to assess the true potential of eDNA as a  
98 complementary tool across the panoply of methods employed for cetacean monitoring.

99

## 100 **Methods**

### 101 **Study Area**

102 Surveys to collect eDNA samples were performed in the north coast of Continental Portugal.  
103 This subregion, located in the northwest of the Iberia Peninsula, is of particular ecological

104 interest due to the upwelling phenomenon strongly present along the coastline, thus  
105 enhancing primary production and providing great conditions for the development of complex  
106 and rich trophic chains [37-39]. In addition, the area is part of a particularly dynamic coastal  
107 region with several estuaries of rivers that flow into it [40]. Topographically, the study area is  
108 entirely placed on the continental shelf, with a relevant structure in the vicinity offshore, the  
109 Porto canyon [41] (Fig. 1).

110 In order to survey the area, four seasonal monitoring campaigns were carried out between the  
111 summer of 2021 and the winter of 2022/2023. Each campaign consisted of a survey transect  
112 with eight equidistant parallels, perpendicular to the coastline, spaced by approximately 6 nm,  
113 and covering distances of about 10 nm (approximately between 2 to 12 nm from the coastline),  
114 as shown in Fig. 1. For the collection of visual monitoring data, a previously established  
115 protocol [32] was followed to sample the occurrence of cetacean species sighted along the  
116 established transect (Fig. 1). The transect was designed so as to have an equitable range of  
117 observation capacity at all its points, guaranteeing a correct and complete visual monitoring  
118 of the study area. The collected visual monitoring data was imported into ArcGIS Pro for spatial  
119 analysis.

120

121 Fig. 1 – Study area with surveyed transect of at-sea monitoring campaigns, conducted between the  
122 summer of 2021 and the winter of 2022/2023, in the north coast of Continental Portugal. Isobaths with  
123 bathymetry in meters.

124

## 125 **Water Collection and Filtration**

126 The environmental samples were sampled at pre-defined stations at the vertices of the  
127 campaign transect (Fig. 1, Table S1), using a bucket and a rope to collect 5 litres of water.  
128 Before water sampling, the samples were poured into 5L-containers. All the materials used for  
129 sample collection were previously cleaned with 10% bleach, rinsed with MilliQ water, and  
130 washed with local seawater (i.e., seawater at the sampling station) just before sample

131 collection. After collection, volumes ranging from 1 to 3 litres were filtered immediately on  
132 board through Sterivex units (0.22  $\mu\text{m}$ ) using a peristaltic pump. After filtration, the samples  
133 were kept in liquid nitrogen, and stored at  $-80^{\circ}\text{C}$  upon arrival at the laboratory. In total, 64  
134 samples were collected, 4 at each station (two summers and two winters), with few exceptions:  
135 (1) two additional collections were made at S09 and S10 during the first summer campaign  
136 (2021); (2) during the first winter campaign, it was not possible to collect samples at stations  
137 S7 and S8, due to adverse weather conditions (Table S2).

138

## 139 **DNA Extraction**

140 Total eDNA was isolated using the DNeasy® PowerWater® Sterivex™ Kit (QIAGEN),  
141 following the manufacturer's instructions, with some adjustments to increase the DNA yield,  
142 namely: increased vortex times (10 minutes at all vortex steps) and 5 minutes rest time before  
143 the last centrifugation step. After extraction, DNA concentration for all samples was quantified using  
144 Qubit™ dsDNA High Sensitivity (HS) assay kit (Invitrogen™). Environmental DNA extraction was  
145 performed in a specifically dedicated laboratory for the extraction of genetic material, with the  
146 bench being always cleaned with ethanol prior to its use. All materials used were also sterilised  
147 in UV light before the start of the extraction process.

148

## 149 **Library Preparation, Sequencing and Bioinformatic**

### 150 **Analysis**

151 All samples were sequenced in high throughput sequencing in an Illumina MiSeq300 platform,  
152 using MarVer3(A) primers [8] and the Superfi II Polymerase (Invitrogen™). For marine  
153 vertebrate library preparation, a fragment of the vertebrate mitochondrial 16S rRNA gene was  
154 amplified and reamplified. In the first amplification step, PCRs were carried out in triplicate in  
155 a final volume of 10  $\mu\text{L}$ , containing 2  $\mu\text{L}$  of template DNA, 0,5  $\mu\text{M}$  of the primers, 1X Platinum  
156 SuperFi II DNA Polymerase (Invitrogen), 0,8 mM dNTPs, 1X SuperFi II Buffer, 1X CES [42],

157 and ultrapure water up to 10  $\mu$ L. The PCR protocol was the following: an initial denaturation  
158 step at 98°C for 30 seconds, followed by 35 cycles of 98°C for 10 seconds, 49 °C for 10  
159 seconds, 72°C for 30 seconds, and a final extension step at 72 °C for 5 minutes. Triplicate  
160 PCR products were pooled together. The oligonucleotide indices that are required for  
161 multiplexing different libraries in the same sequencing pool were attached in a second  
162 amplification step with identical conditions but only 5 cycles and with an annealing temperature  
163 of 60°C. Libraries were then purified using the Mag-Bind RXNPure Plus magnetic beads  
164 (Omega Bio-tek), following the instructions provided by the manufacturer. Lastly, the pool was  
165 sequenced in a fraction (3/8) of a MiSeq PE300 flow cell (Illumina). DNA metabarcoding library  
166 preparation and sequencing were carried out by AllGenetics & Biology SL  
167 ([www.allgenetics.eu](http://www.allgenetics.eu)).

168 Following that, sequencing adapters were removed using Cutadapt v3.5 (Martin, 2011) and  
169 the originated reads went through a DADA2 [43] tool pipeline to remove other non-biological  
170 DNA sequences (e.g., primers), filter the reads according to their quality, denoise, dereplicate,  
171 cluster the resulting sequences into Amplicon Sequence Variants (ASVs), merge  
172 corresponding forward and reverse reads, and remove chimeric sequences. The taxonomic  
173 assignment of each ASV (with a minimum of 5 reads) was performed by querying its  
174 representative sequence against a local instance of the NCBI's Nucleotide database (last  
175 updated on 18/09/2023), using the algorithm BLASTn v2.13.0+ [44] with the following  
176 parameters: percent identity of 99%, e-value of 1e-05, and a minimum hit coverage of 80%.  
177 If, based on these parameters, a matched ASV represented multiple species, the one with  
178 100% identity was considered. For this part of the work, the ASVs corresponding to all  
179 cetacean species and Actinopteri superclass species with known occurrence in Continental  
180 Portugal (as of the Ocean Biodiversity Information System [OBIS] up to 05/09/2023) were  
181 selected. For the construction of the databases, all available nucleotide sequences (containing  
182 the 16S rRNA gene) from selected taxa were retrieved from the NCBI nucleotide database.  
183 Then, using the Geneious® software (v.7.0.6), primer sequences (forward and reverse) were  
184 annotated (using the 'Add Primers to the Sequence' tool), sequences were cut (using the

185 'Extract PCR product' tool) so they only contain the fragments spanning (and including) the  
186 PCR primers, and all nucleotide sequences not containing the entire fragment of interest were  
187 excluded from the database. Positively matched ASVs of each sample were then grouped into  
188 the respective sampling stations and seasonal campaigns for clear visual representation of  
189 the results. Representative stacked bar plots of identified taxa relative abundance and  
190 Actinopteri ASVs heatmaps for winter and summer monitoring campaigns were produced  
191 using the *ggplot2* (v.3.4.3) package [45], in R.

192

## 193 **Sampling Statistical Tests**

194 In order to better understand whether there are significant variations in the concentrations obtained  
195 from samples taken in the different seasonal campaigns, Kruskal-Wallis chi-squared tests were carried  
196 out, and if significant differences were found ( $p$ -value  $< 0,05$ ), we followed with a pairwise Wilcoxon  
197 test. The Wilcoxon test was also applied to compare the concentrations obtained from stations located  
198 at different distances from the coast (2 nm versus 12 nm). For this comparative analysis of DNA  
199 quantification results, samples that had DNA concentration below the Qubit's HS kit detection  
200 limit (0,005 ng/ $\mu$ L) were assigned a value of 0,001 ng/ $\mu$ L. A sample log of all collected and  
201 extracted samples can be found in Table S2.

202 Statistical tests (Kruskall-Wallis test and pairwise Wilcoxon tests, with the level of significance  
203 set to 0,05) were carried out to understand the variation in the total number of ASVs obtained  
204 in the samples at the different seasonal campaigns and at the different distance from the coast  
205 (2 nm versus 12 nm).

206

## 207 **Results**

### 208 **Visual Surveys**



209 A total of 71 sightings of 4 different cetacean species were recorded during monitoring  
210 campaigns (Fig. S1A, Table S3). The majority of these records were from common dolphin  
211 (*Delphinus delphis*) with 45 sightings across the entire surveyed area, representing  
212 approximately 63.4% of all records. As for the other 3 species - bottlenose dolphin (*Tursiops*  
213 *truncatus*), Risso's dolphin (*Grampus griseus*) and harbour porpoise (*Phocoena phocoena*) –  
214 these were sighted on 3 different occasions each. Besides the identified records to the species  
215 level, there were also 13 sightings of unidentified Delphinidae (18,3% of total records) and, on  
216 4 occasions, it was only possible to identify cetacean occurrence at the superorder level  
217 (Cetacea). All sightings recorded with the respective date and geographical coordinates can  
218 be found in Table S3.

219

## 220 **eDNA Sampling**

221 Among the campaigns conducted, the summer of 2021 showed much higher DNA  
222 concentrations compared to the other seasonal campaigns, with an average quantification of  
223  $9.46 \pm 9.05$  ng/ $\mu$ L (Median= 9,22 ng/ $\mu$ L). This was followed by the winter of 2021/22 with  $5.10$   
224  $\pm 6.11$  (Median= 3.02 ng/ $\mu$ L), the winter of 2022/23 with  $2.26 \pm 3.50$  ng/ $\mu$ L (Median= 0.8 ng/ $\mu$ L)  
225 and, finally, the summer of 2022 with  $1.08 \pm 2.38$  ng/ $\mu$ L (Median= 0.67 ng/ $\mu$ L). Statistical tests  
226 showed significant differences between the quantified concentrations of the different  
227 campaigns ( $p$ -value =  $9.63e^{-06}$ ), these being explained by differences between summer 2021  
228 and summer 2022, summer 2021 and winter 2022/23, and also between summer 2022 and  
229 winter 2022/23 (Table S4A). At the seasonal level, the summer shows higher concentrations,  
230 with  $5.52 \pm 8.18$  ng/ $\mu$ L (Median= 3.51 ng/ $\mu$ L) compared to the  $3.59 \pm 5.02$  ng/ $\mu$ L (Median=  
231 1.49 ng/ $\mu$ L) obtained for the winter. However, this difference was not statistically significant  
232 ( $p$ -value = 0.79). Taking distance into account, the sampling stations near the coast (2 nm)  
233 had an average DNA concentration of  $5.48 \pm 8.18$  ng/ $\mu$ L (Median= 3.2 ng/ $\mu$ L), while the DNA  
234 obtained in stations at 12 nm was quantified as  $3.75 \pm 5.21$  ng/ $\mu$ L (Median= 1.26 ng/ $\mu$ L) on  
235 average. These differences were not statistically significant ( $p$ -value = 0.45).

236

## 237 **eDNA Monitoring: Sequencing Output and Taxonomic**

### 238 **Identification**

239 A total of 5 876 226 reads were generated by the Illumina platform with an approximate  
240 average of  $90\,403.48 \pm 42\,311.20$ . After quality-filtering steps, the final output was 3 735 272  
241 total reads with an average of approximately  $57\,465.72 \pm 40\,520.33$  (representing 65.6% of  
242 the input). There were 5 outlier samples that produced less than 50 reads. A total of 173 156  
243 representative ASVs (of previously determined marine vertebrate species: cetaceans and  
244 selected Actinopteri species with known occurrence in Continental Portugal). Among the  
245 different seasonal campaigns, winter 2022/23 obtained the highest number of resulting ASVs  
246 with an average of  $7\,951.5 \pm 23\,991.69$  (Median = 39), followed by winter 2022/23 with 2024  
247  $\pm 2\,980.45$  ASVs (Median = 722.5), the summer 2021 with  $932.33 \pm 1\,716.23$  ASVs (Median  
248 = 201.5), and finally the summer 2022 with an average of  $51.5 \pm 178.56$  ASVs (Median = 0).  
249 The statistical tests carried out indicate that there was a significant difference in the uptake of  
250 target eDNA taxa, both between seasons (p-value = 0.019) and between the different  
251 campaigns (p-value = 0.009). Pairwise tests highlight differences between the two summer  
252 campaigns, and both winter campaigns with the summer of 2022 (Table S4B). With regard to  
253 distance from the shore, samples collected closer (2 nm) resulted in an average number of  
254 ASVs of  $988.13 \pm 2087.06$  (Median = 0), while those further away (12 nm) resulted in  $4,423.31$   
255  $\pm 17,126.03$  ASVs (Median = 52.5). However, differences of obtained ASVs obtained in  
256 relation to the distance to coast were not significant (p-value = 0.64).

257 Of the 64 eDNA samples sequenced through metabarcoding, 33 (51.6%) had positive  
258 detections of representative ASVs, from which 10 of the samples (15.6%) had positive  
259 detections of cetacean ASVs. By grouping the results by sampling station, we noticed that all  
260 sampling stations had positive detections for either Cetacea or Actinopteri. Relative  
261 abundances of all cetacean species identified, and Actinopteri taxa detected in 4 or more  
262 stations, are illustrated in Fig. 2.

263

264 Fig. 2 - Taxonomic assignment from environmental DNA samples collected in 16 sampling stations, in  
265 the North coast of Mainland Portugal, between 2021 and 2023. The plot represents the relative  
266 abundance (%) of sequences matched for each marine vertebrate species, considering pre-determined  
267 Cetacea, and Actinopteri species with positive detection in 4 or more stations (see methods). "Others"  
268 represents Actinopteri superclass species identified in 3 or fewer stations.

269

270 Focusing on cetaceans, 5 different species were detected, with at least one of them identified  
271 in 7 out of the 16 stations (Fig. 2 and Table S4). The common dolphin was the most detected  
272 species, being identified at 6/16 stations, followed by the harbour porpoise and the fin whale  
273 (*Balaenoptera physalus*), occurring at 2/16 sampling stations. The Risso's dolphin and the  
274 bottlenose dolphin were detected on 1/16 stations.

275 Regarding Actinopteri species, 19 species were identified, with at least one of them detected  
276 in all sampling stations (Fig. 2 and 3). Of all these detections, the most notable result was the  
277 highly frequent presence of the sardine (*Sardina pilchardus*) - in 14/16 sampling stations; the  
278 European seabass (*Dicentrarchus labrax*) - in 9/16 stations; and the European anchovy  
279 (*Engraulis encrasicolus*) - in 6/16 of the sampled sites (Fig. 2 and 3). All the Actinopteri species  
280 identified, as well as the respective number of ASVs detected at each station, can be found in  
281 Table S4. Additionally to the frequency of the detections across sampling stations, there is  
282 also a higher abundance of captured ASVs of the species *S. pilchardus* and *D. labrax* (Fig. 3).  
283 In 5/16 sampling stations, 100% of the matched ASVs belonged to either *S. pilchardus* or *D.*  
284 *labrax* (Fig. 2). The highest number of ASVs was obtained for *S. pilchardus* in Station S15 (off  
285 north of Viana do Castelo) (Fig. 5). Station S01 (southwest of Porto, at ~2nm from coast) had  
286 positive detections for a higher number of selected Actinopteri species in both seasons (Fig.  
287 3). The Station S08 (near Póvoa de Varzim) only had one minor detection (9 BLAST hits for  
288 its respective ASV, see Table S5) for *S. pilchardus* across all monitoring campaigns.

289

290 Fig. 3 - Number of Amplicon Sequence Variants (ASVs) detected for each species of the Actinopteri  
291 superclass molecularly identified in environmental DNA samples, collected at the sampling stations  
292 surveyed in the summer (top) and winter (bottom) in the North coast of Mainland Portugal, between  
293 2021 and 2023. The number of ASVs was logarithmised for scaling purposes.

294

## 295 **Cross-checking eDNA with Visual Monitoring Data**

296 Overall, by cross-referencing the data obtained by the two sampling methodologies applied in  
297 this study, we can verify the occurrence of the following four species in both datasets (visual  
298 records, eDNA molecular detection): *D. delphis*, *P. phocoena*, *T. truncatus* and *G. griseus*. In  
299 addition to these, positive molecular detections of *B. physalus* were obtained through eDNA  
300 analysis. Regarding the spatial distribution of these records, it is clear that *D. delphis* was  
301 sighted not only in greater abundance but also with a dispersed distribution throughout the  
302 region studied, which is also in line with the data obtained by molecular detection. For the  
303 other identified species, their DNA abundance was lower and there were fewer visual records  
304 obtained as well.

305 Overall, no pattern of spatial or temporal correlation between molecular detections and visual  
306 records is clear. An inter-seasonal comparison of the spatial distribution of the records  
307 obtained by the visual surveys with the eDNA detections reveals contrasting results (Fig. 4).  
308 In terms of composition of the cetacean community, in the summer campaigns, all the 4  
309 species mentioned above were sighted; while in the winter campaigns, sightings of *D. delphis*  
310 accounted for all records, excepting one sighting of *T. truncatus* (considering sightings  
311 identified at the species level) (Fig. 4A). Through eDNA analysis though, all the species were  
312 molecularly detected in the winter campaigns (including also *B. physalus*), while in the summer  
313 season the only species identification through molecular detection of cetacean species was of  
314 *D. delphis* (Fig. 4B). Disregarding seasonality and accounting for spatial distribution only, the  
315 location of molecular detections of each species at the sampling stations seems to align better  
316 with the visual records at the latitudinal level than in relation to distance to the coastline (i.e.,  
317 longitudinally). In summary, it is worth highlighting that: i) in both datasets, *D. delphis* was

318 found to be distributed across the entire area with higher frequency north of Póvoa de Varzim;  
319 ii) *T. truncatus* was distributed in the north of the study area, visually recorded at different  
320 distances from the coast but only detected with molecular methods in the most coastal  
321 sampling station (S13); iii) *P. phocoena* was detected in both datasets south of Póvoa de  
322 Varzim, and mostly in stations near coast (with only one sighting further from the coast); iv)  
323 the molecular detections of *B. physalus* occurred in the stations closer to the coast, one in the  
324 north and another in the south of the study area (Fig. 4B).

325

326 Fig. 4 – Seasonality of cetacean records obtained by visual monitoring (A) and molecular analysis of  
327 environmental DNA (B), for summer (left) and winter (right) campaigns carried out in the North coast of  
328 Mainland Portugal, between 2021 and 2023.

329

## 330 Multi-trophic Analysis

331 Across all campaigns, multiple Actinopteri detections coincided with the molecular presence  
332 of cetacean species (Table 1). *D. delphis* was detected by eDNA simultaneously to *S.*  
333 *pilchardus* (on two occasions), *B. belone*, *E. encrasicolus*, *D. labrax*, *S. saurus* and *L. bergylta*.  
334 *P. phocoena* detections were coincident with a higher richness of Actinopteri species,  
335 including *S. pilchardus*, *D. labrax*, *S. scombrus*, *L. bergylta*, *Pleuronectidae spp.*, *H.*  
336 *lanceolatus* and *C. conger* (Table 1). *G. griseus* and *B. physalus* were also detected at the  
337 same time as *S. pilchardus* (Table 1), while *T. truncatus* was solely detected without  
338 associated potential Actinopteri prey.

339

340 Table 1 - Coincident Cetacea and Actinopteri species detections by environmental DNA at  
341 sampling stations surveyed across seasonal monitoring campaigns carried in the North coast  
342 of Mainland Portugal, between 2021 and 2023.

Monitoring campaign	Sampling station	Cetacean	Actinopteri
Summer 2021	S11	<i>Delphinus delphis</i>	<i>Belone belone</i>

	S13	<i>Delphinus delphis</i>	<i>Sardina pilchardus</i> <i>Engraulis encrasicolus</i> <i>Dicentrarchus labrax</i>
	S15	<i>Delphinus delphis</i>	<i>Scorpaenopsis scorpaenoides</i>
Winter 2021/22	S01	<i>Phocoena phocoena</i>	<i>Sardina pilchardus</i> <i>Dicentrarchus labrax</i> <i>Labrus bergylta</i> <i>Alosa alosa</i> <i>Pleuronectidae spp.</i> <i>Dicoglossa cuneata</i> <i>Hyperoplus lanceolatus</i> <i>Conger conger</i>
	S04	<i>Phocoena phocoena</i>	<i>Sardina pilchardus</i> <i>Dicentrarchus labrax</i> <i>Alosa alosa</i> <i>Dicoglossa cuneata</i> <i>Scorpaenopsis scorpaenoides</i> <i>Sarpa salpa</i>
Winter 2022/23	S01	<i>Delphinus delphis</i>	<i>Labrus bergylta</i>
	S02	<i>Grampus griseus</i>	<i>Sardina pilchardus</i>
	S04	<i>Delphinus delphis</i> <i>Balaenoptera physalus</i>	<i>Sardina pilchardus</i>

## 344 **Discussion**

### 345 **Environmental DNA Sampling and Uptake**

346 The sampling techniques here employed were chosen due to their on-field practicability  
347 (collection of seawater and filtration) and lower risk of contamination (no direct handling of  
348 filters during DNA extraction). To date, there are no generalised standard protocols for the  
349 detection and identification of cetacean species through eDNA, with a wide variation of  
350 methods currently being implemented [12]. The lack of a standardised, reliable, and  
351 repeatable protocol for sample collection (and subsequent processing and analysis) still  
352 constitutes a major obstacle to the use of this monitoring tool for marine biodiversity monitoring  
353 [46].

354 On the northwest coast of the Iberian Peninsula, upwelling phenomena tends to occur in the  
355 summer, inducing an increase in primary productivity and subsequent development of the  
356 trophic chain [39]. Therefore, it was expected DNA concentrations to be uniformly higher in  
357 the samples collected during summer campaigns, which was not the case. However, while the  
358 2021 summer campaign presented the highest values of DNA concentration, which could be  
359 explained by the mentioned oceanographic phenomena, the exact opposite was observed in  
360 2022. This significant difference, although not ruling out possible human error in the different  
361 sampling steps, may be due to some unknown oceanographic event happening during  
362 sampling days or to other environmental variables with known impact on DNA preservation,  
363 such as temperature [47;48], pH [49;50] and/or UV radiation exposure [51]. The lack of inter-  
364 seasonal significant differences in the DNA concentration obtained from eDNA samples, in  
365 contrast to the significant inter-campaign differences observed, may evidence that the  
366 seasonal (i.e., inter-annual cycles such as seasonal upwelling) influence is less relevant than  
367 the impact of less cyclic and more ephemeral phenomena on the DNA preservation and  
368 retrieval. Regarding the number of target ASVs, significant differences found between both  
369 seasons and monitoring campaigns, suggest a higher target DNA uptake in winter months.

370 Nevertheless, this result should be interpreted with caution, since these differences are mainly  
371 justified by the low numbers evidenced in the summer 2022 campaign (for which a very low  
372 DNA concentration was obtained). As far as the proximity to the coast, although with some  
373 variation, there were no significant differences in terms of DNA concentrations or the number  
374 of ASVs obtained. Researchers have reported a faster DNA degradation in inshore areas  
375 associated with intense anthropogenic pressures [52]. Nevertheless, coastal areas are also  
376 likely to have higher biomass, in comparison with offshore areas, due to the coastal upwelling  
377 phenomena [39] and the input of the rivers [53;54]. The coastal waters of the north coast of  
378 Continental Portugal are heavily influenced by the input of various fluvial water masses such  
379 as the Douro, Cávado, and Minho rivers [55;56], which could also be interfering with sampling  
380 results. The river plume of the Douro river is known to vary substantially inter annually, with  
381 increasing extension into the ocean in winter months [57]. Overall, we consider that further  
382 studies are required in the future to monitor the influence of environmental conditions on eDNA  
383 uptake, and therefore evaluate sampling success in dynamic marine coastal ecosystems.

384

## 385 **Bioinformatic Analysis Criteria**

386 To compile the BLASTn databases, we established a strict criterion, comprising the amplified  
387 fragment of interest and the sequences of the forward and reverse primers for the  
388 aforementioned species of interest. However, this type of approach for sequence selection  
389 has its pros and cons. On the positive side, it allows for greater certainty that a given sequence  
390 belongs to a respective species since smaller fragments can result in greater similarity  
391 between species (conservative approach). However, this method can lead to a loss of  
392 information by possibly removing sequences of interest. Moreover, and particularly for  
393 assembling the database of sequences corresponding to the Actinopteri species, the approach  
394 was only focused on the species already recorded along the Portuguese coast, substantially  
395 reducing the size of the database. Such a decision was made since, in this context, the



396 purpose for the detection of Actinopteri species was tied to the ecological role they play in the  
397 diet of cetaceans.

398 Given the parameters applied in the BLASTn algorithm used for the taxonomic assignment  
399 (identity higher than 99%), there were only two cases in which multiple species matched to a  
400 given ASV: one within Cetacea taxa, and another within Actinopteri. A specific example of  
401 multi-species assignment to an ASV was a case where the ASV was assigned to *D. delphis*  
402 with 100% identity, but other taxa were assigned with a percentage identity of 99.14% (the  
403 genera *Tursiops* spp. and *Stenella* spp.). This was to be expected given the genetic similarity  
404 of these species pertaining to the oceanic dolphin family (Delphinidae), as described by  
405 McGowen et al. [58]. Especially between the common dolphin and the bottlenose dolphin,  
406 there are only two Single Nucleotide Polymorphisms (SNPs) on the amplicon obtained with  
407 the MarVer3 primers. Besides this case, there was a double identification with 100% percent  
408 identity for the same ASV - 7 sequences for the Atlantic horse mackerel (*Trachurus trachurus*)  
409 and 1 for the megrim (*Lepidorhombus whiffiagonis*) in the NCBI database. In this case, the  
410 species *T. trachurus* was selected as the correct identification. Such decision was because  
411 the matched sequence of *L. whiffiagonis* referred to samples collected in the Mediterranean  
412 Sea, in which case the authors of the work demonstrated a great variation in the mitochondrial  
413 16S rRNA region between the Mediterranean and the Atlantic Ocean populations [59].

414

## 415 **Cetacean Detections: eDNA versus Visual Monitoring**

416 As described in the results, the species detected through visual monitoring were also detected  
417 by eDNA, with the additional record of the species *B. physalus*, which was only recorded  
418 through molecular methods in the environmental samples. This is a species with non-existent  
419 sightings in this region of Continental Portugal, even though it has been recorded in other  
420 parts of the country's coastline [60]. Additionally, in the Galician coastal waters, adjacent to  
421 the study area, the species occurs rather frequently [30;61;62]. Offshore Galicia, there are  
422 relevant topographical structures, seamounts, that may act as hotspots for pelagic biodiversity

423 [63], such as the Galicia Bank [64]. The obtained detection of *B. physalus* DNA may be  
424 evidence of an occasional occurrence of the species in waters off northern Portugal.  
425 As shown by the results, the common dolphin represented the vast majority of the sightings in  
426 the study area. The wide distribution and frequent presence of *D. delphis* along the entire  
427 northern Portuguese coast was corroborated by molecular detection on eDNA samples. The  
428 harbour porpoise was detected via eDNA at two stations located closer to the coastline. That  
429 result is in line with the frequent occurrence of the species at the mouth of the Douro River  
430 (located in proximity to the sampling sites), as described by Gil et al. [65] with occurrence  
431 records obtained through visual monitoring. As for *T. truncatus* and *G. griseus*, both species  
432 were only sighted on very few occasions. By eDNA, each of these was detected at one station,  
433 emphasising the less frequent occurrence in the north of Continental Portugal, in comparison  
434 with *D. delphis* and *P. phocoena*. *T. truncatus* was detected via eDNA in northern latitudes,  
435 where the sightings of the species were also registered. As for *G. griseus*, the DNA was  
436 detected in the southern quadrant of the study area, where the species was also recorded  
437 through visual monitoring, where they were also recorded by molecular detection. On the other  
438 hand, The bottlenose dolphin is a frequent species in the northwest of the Iberian Peninsula  
439 [66], therefore detection through visual and molecular methods was anticipated, although  
440 more records were to be expected. Overall, the combination of the results from visual  
441 monitoring and eDNA suggests the year-round presence of the dolphin and porpoise species  
442 on the north coast of Continental Portugal. Given the difficult weather conditions for visual  
443 monitoring in the ENA for a large part of the year, mostly in winter months, and consequently  
444 low monitoring effort in the region, complementary results from eDNA prove to be very useful  
445 to increase baseline knowledge of these coastal populations.  
446 Besides multi-species detection on eDNA samples, the application of the methodology for  
447 single-species detection may be of particular importance. This is the case given the need to  
448 monitor resident populations in decline, as shown by Ma et al. [14] where eDNA tools were  
449 applied for conservation efforts. In the north coast of Continental Portugal, we believe that this  
450 approach is extremely pertinent for species of conservation interest, such as the harbour

451 porpoise, whose population in the mainland Portuguese coastline is predicted to be extinct in  
452 20 years (Critically Endangered, [67]).

453 Despite the aforementioned promising results, some challenges remain to be addressed for  
454 the future consistent application of this novel monitoring tool. In previous eDNA metabarcoding  
455 studies, it was usually not possible to distinguish between species of the same genus, as  
456 observed in Valsecchi et al. [22] for the *Tursiops* and *Stenella* genera, and even within the  
457 Delphinidae family [4;20]. In the present work, this was also a challenge during the assignment  
458 of ASVs, even though a conservative selection criterion was established. Therefore, this  
459 limitation represents a significant obstacle to meet monitoring objectives. In addition, although  
460 the use of universal primers may provide us with relevant information at the multi-trophic level,  
461 there is a major drawback when the technique is used to detect a certain taxon (e.g.,  
462 monitoring cetacean occurrence). Besides the likelihood that primers bind to non-target DNA  
463 in greater abundance at the time of amplification, which can result in false negative detections  
464 [68;69], it is impossible to pre-select the samples with the target DNA before the sequencing  
465 step as the detections in the post-PCR electrophoresis may allude to different taxonomic  
466 groups. Thus, in the context of the applicability of eDNA for more efficient and accessible (i.e.,  
467 cost-effective) taxon-specific monitoring, it is advisable to use taxon-specific primers. In the  
468 specific case of cetaceans, where, in addition, relative abundance may often be rather low (in  
469 comparison to other marine taxa), the development of cetacean-specific primers is one of the  
470 next key steps to address in the continued emergence of the application of eDNA tools for  
471 cetacean monitoring.

472

## 473 **Monitoring Multiple Trophic Levels**

474 The introduction of universal primers in the study of marine communities allows the collection  
475 of data from multiple trophic levels, with posterior inference of possible ecological interactions  
476 and prey/predator relationships. Universal primers also provide a deeper understanding of  
477 unusual distribution patterns, as was the case in Zhang et al. [24] where the authors used data

478 obtained by eDNA metabarcoding to assess the food resources of Eden's whale  
479 (*Balaenoptera edeni edeni*), and in Visser et al. [70] where the cephalopod community  
480 composition was studied to better understand Risso's dolphins (*G.griseus*) and Cuvier's  
481 beaked whales (*Ziphius cavirostris*) foraging behaviours.

482 In this study, it was possible to detect numerous species from the Actinopteri superclass. In  
483 addition to the sardine, highly abundant and dispersed in the study region, and reported as  
484 main target for the diet of *D. delphis* on the Iberian coast [71;72], other species detected have  
485 also been found in its stomach contents in the portuguese continental coast, such as the  
486 European anchovy (*Engraulis encrasicolus*), the sole (*Soleidae spp.*), the needlefish (*B.*  
487 *belone*), the scads (*Trachurus spp.*) and the Atlantic mackerel (*Scomber spp.*) [72]. Another  
488 species with abundant and dispersed detections in this study, the European seabass, was  
489 also found in the stomach contents of *D. delphis* in the English Channel [73]. This latter coastal  
490 fish species, of high economic interest, is of particular relevance from a cetacean conservation  
491 perspective since its feeding preferences overlap with those of the dolphins, with reported  
492 implications for their bycatch by promoting interactions with the fisheries sector [74]. The  
493 aforementioned sardine is also extremely relevant to this topic due to its commercial  
494 importance in Portugal, being a main target of purse seine fisheries in the country [75]. In  
495 Continental Portugal, interactions between various dolphin species with this fishing method  
496 are frequent and can result in the bycatch of these animals [76;77]. In the case of *P. phocoena*,  
497 many of the species identified simultaneously, such as *Scomber scombrus* and the ballan  
498 wrasse (*Labrus bergylta*), and others occurring in other stations, such as *T. trachurus*, are also  
499 known prey for these animals in the ENA [78;79], which suggests that the population sighted  
500 regularly near the mouth of the Douro River [65] uses this coastal area for feeding purposes.  
501 Although not detected in coincidence, a wide variety of fish taxa identified in the studied area  
502 were found in *T. truncatus* stomach contents in Iberia, namely *S. pilchardus*, European conger  
503 (*Conger conger*), *E. encrasicolus*, *Trachurus spp.* and *Soleidae spp.* [80].

504 Therefore, data obtained from metabarcoding techniques evidenced the ecological  
505 importance of the region as a feeding area for coastal cetaceans, also highlighting the prey

506 species available to their populations, which may have relevant implications for management  
507 and conservation strategies. Environmental DNA is thus proving to be a promising tool for  
508 multitrophic assessments, not only to study biodiversity occurrence, but also to infer about  
509 ecological processes and investigate trophic relationships.

510

## 511 **Conclusion**

512 In summary, we demonstrated the potential of metabarcoding methods applied to eDNA  
513 samples, for biodiversity assessments, with special relevance as a cetacean monitoring  
514 technique for the study of cetacean distribution and feeding ecology. However, there are still  
515 obstacles and difficulties to overcome. More studies are necessary to better understand this  
516 novel sampling method, as there are still knowledge gaps in the application of the method,  
517 from environmental sampling to the analysis of the sequencing results.

518 Positive detections of cetacean species in this work constitute important data, as it was  
519 possible to characterise the northern coast of Continental Portugal in terms of cetacean  
520 occurrence, not only reproducing similar results but also complementing the data obtained  
521 through visual monitoring. Here, eDNA monitoring allowed us to conclude that dolphin and  
522 porpoise species that are less often sighted during visual surveys, even with known  
523 occurrence in the study area, have a probable all-year-round presence in this region of the  
524 ENA. Additionally, this study also shows the potential of this molecular-based technique to  
525 collect data when the monitoring effort is hampered by the inherent conditions of the sampled  
526 region (e.g., weather conditions), especially for less frequent species (possibly, without  
527 existing visual records). Furthermore, by expanding the analysis to lower trophic levels, the  
528 detection of multiple species known as cetacean prey species evidenced their use of these  
529 coastal areas for feeding purposes.

530 In conclusion, this work has demonstrated the potential of an innovative monitoring  
531 methodology for studying complex marine biological communities, such as cetaceans, even  
532 allowing for a multi-trophic approach, essential for conservation efforts. Therefore, although

533 the effectiveness of using eDNA as a tool in cetacean monitoring programmes remains under  
534 development, this work represents a step forward towards that goal.

535

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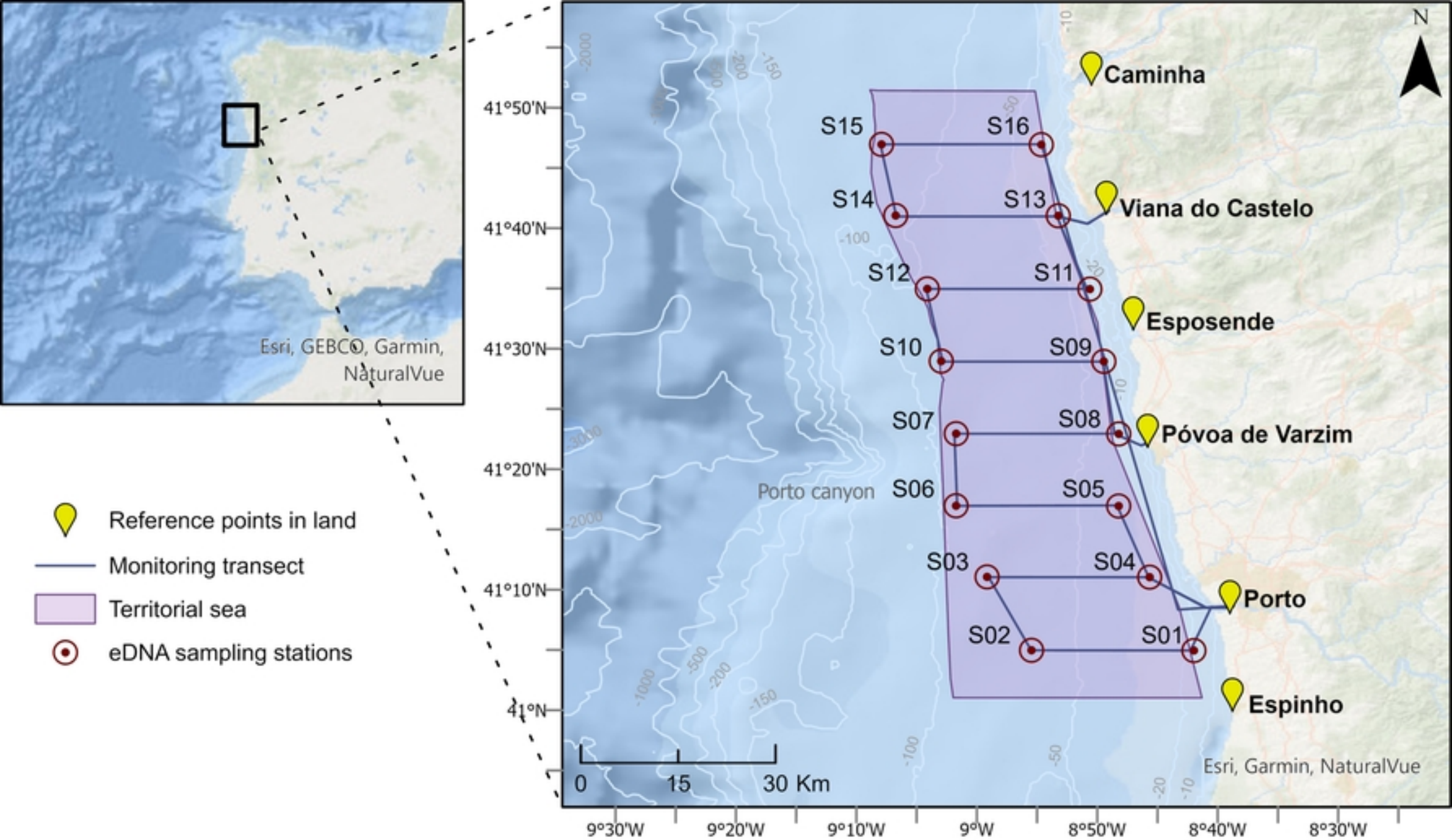


Fig. 1

Species

<span style="color: red;">■</span> <i>Phocoena phocoena</i>	<span style="color: lightgreen;">■</span> <i>Grampus griseus</i>	<span style="color: gold;">■</span> <i>Tursiops truncatus</i>	<span style="color: purple;">■</span> <i>Engraulis encrasicolus</i>	<span style="color: gray;">■</span> Others
<span style="color: cyan;">■</span> <i>Balaenoptera physalus</i>	<span style="color: blue;">■</span> <i>Delphinus delphis</i>	<span style="color: darkslategray;">■</span> <i>Sardina pilchardus</i>	<span style="color: orange;">■</span> <i>Dicentrarchus labrax</i>	

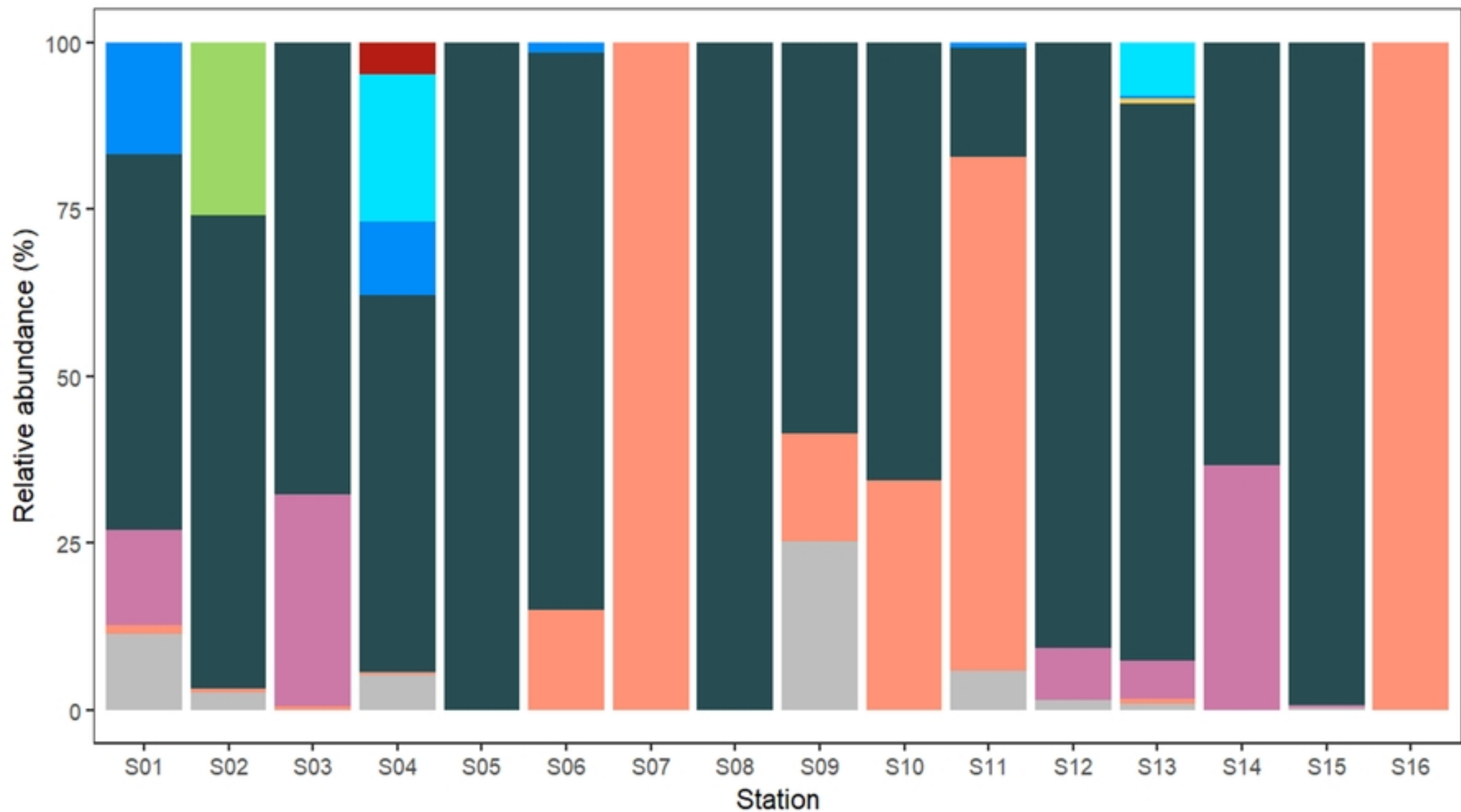


Fig. 2

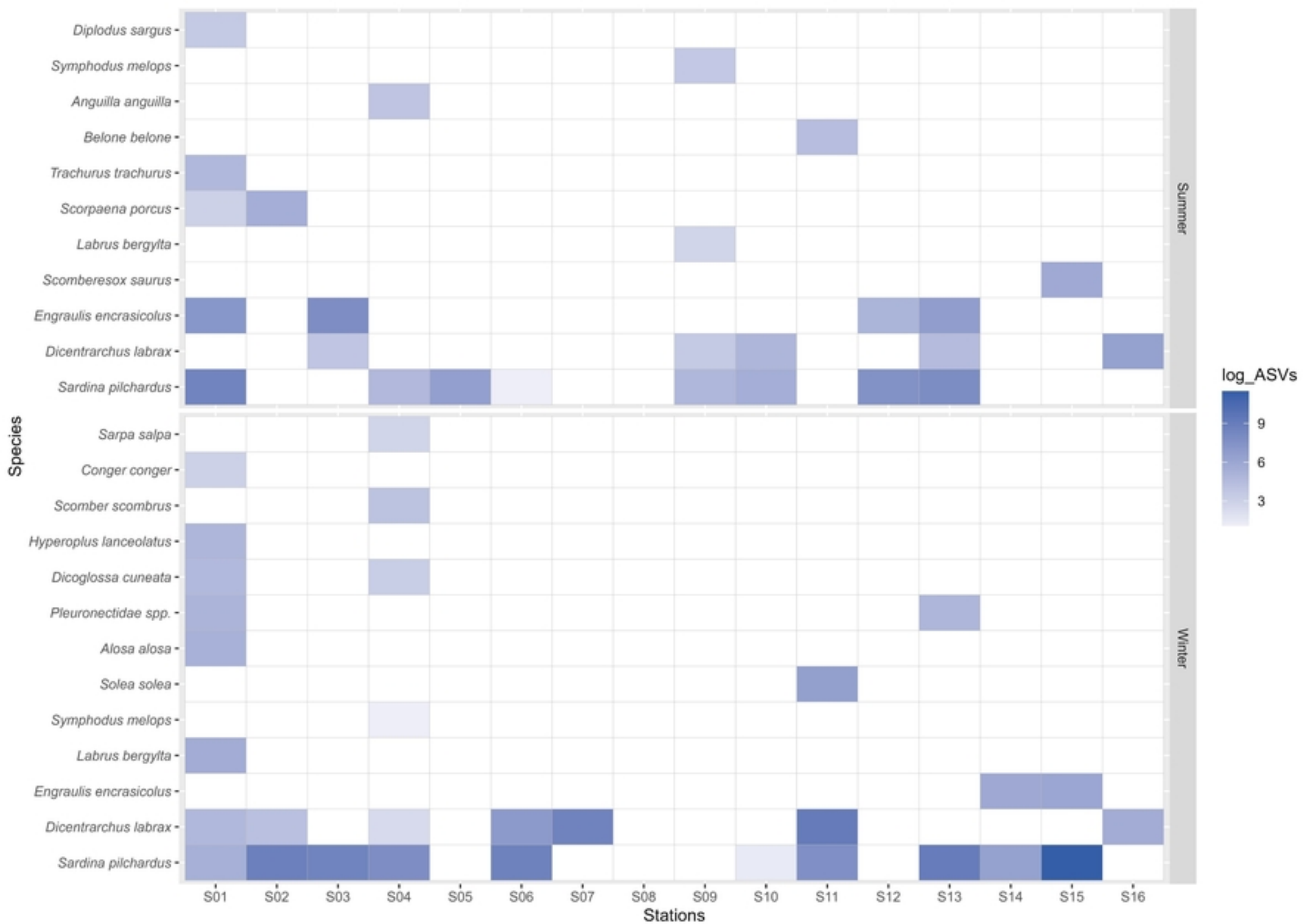


Fig. 3

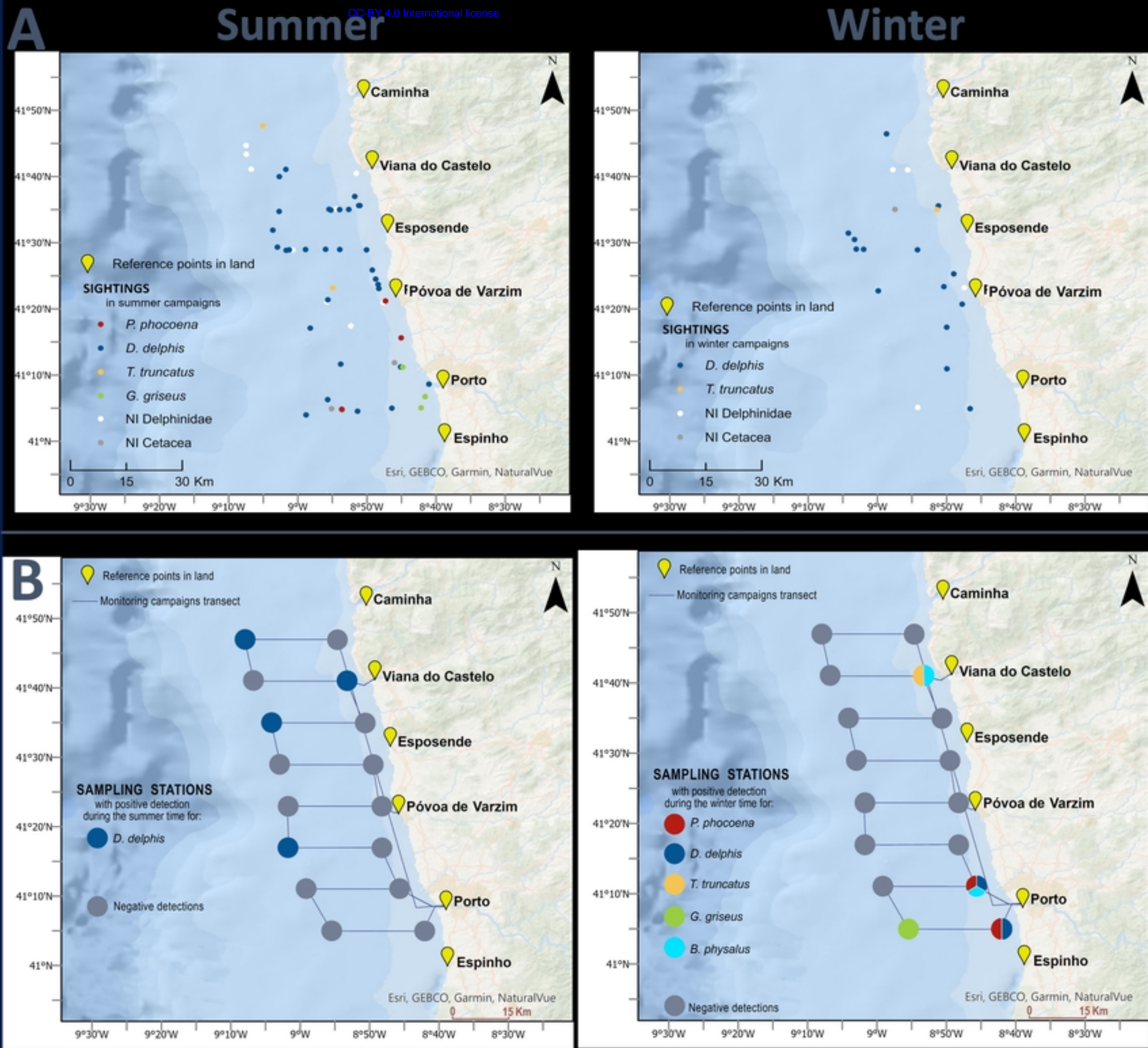


Fig. 4