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Abstract:

Running Head: Porcupines, pines, and beetles INDIRECT INTERACTIONS AMONG DENDROPHAGES: PORCUPINES PREDISPOSE PINYON PINES TO BARK BEETLE ATTACK LINDA M. ILSE and ERIC C. HELLGREN^{1,2} Department of Zoology, Oklahoma State University, Stillwater, Oklahoma, 74078, USA ¹Corresponding author: Eric C. Hellgren Cooperative Wildlife Research Laboratory Department of Zoology Southern Illinois University Carbondale, IL 62901-6504 e-mail: hellgren@siu.edu ²Present address: Cooperative Wildlife Research Laboratory, Department of Zoology, Southern Illinois University, Carbondale, Il 62901-6504, USA

Abstract

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2 Discerning linkages among trophic levels and untangling indirect interactions is essential to understanding structuring of communities and ecosystems. Indeed, indirect 3 4 interactions among disparate taxa are often essential to the functional role of these species. The goal of this research was to test the hypothesis that the relationship between 5 2 dendrophagous taxa, the North American porcupine (*Erethizon dorsatum*) and the pine 6 7 engraver beetle (*Ips hoppingi*), is an asymmetric indirect interaction mediated by a common host. We proposed that the porcupine predisposes the papershell pinyon pine 8 9 (*Pinus remota*) to colonization by pine engraver beetles. We examined porcupine tree selection, pinyon pine physiology and physiognomy, and beetle-pine associations on a 10 study area in the southwestern Edwards Plateau of Texas from June 1997 to August 1999. 11 Although attacks by beetles were evident on both damaged and undamaged trees, 12 successful colonization was greater ($\chi^2 = 75.3$, df = 3, P < 0.01) on pines damaged by 13 14 porcupines. Intensity of porcupine attack, indexed by number of feeding scars and area of bark removed, also was associated with subsequent colonization by beetles. 15 Porcupines selected pinyon pines over more abundant species (P < 0.001) and were 16 17 selective at the level of morphology, whereas pine engraver beetles were selective of tree 18 morphology and physiology. Trees colonized by beetles had phloem with higher 19 concentrations of fructose and glucose and lower percent composition of limonene, 20 sabinene, and terpinolene than uncolonized trees. Our findings supported our hypothesis of an indirect interaction between these dendrophages. We rejected alternative 21 explanations (e.g., that these dendrophages preferred similar trees or that beetles 22

facilitated porcupine damage) for this relationship based on the biology of *Ips* and their

- selection of host trees. We propose that release of volatile terpenes as a result of
- 2 porcupine feeding and reallocation of carbon resources as a response to stress explains
- the facilitation of beetle colonization in porcupine-damaged trees. Our findings parallel
- 4 those observed in other systems involving indirect effects and fit within the framework of
- 5 integrated theories explaining host plant-herbivore interactions.
- 6 Key Words: dendrophages, Erethizon dorsatum, indirect effects, Ips hoppingi, Pinus
- 7 remota, plant-herbivore interactions

1. Introduction

The role of direct interspecific interactions in structuring communities is evident in extensive examinations of competition (Connell, 1983; Schoener, 1983) and predator-prey relations (Sih et al., 1985; Martin, 1988). The strength of indirect effects occurring among different trophic levels has received much less attention, which is likely a result of the inherent difficulty in detecting these types of relationships (Wooton, 1994). The role of a requisite third species in these events compromises the efficacy and timely detection of indirect interactions (Davidson et al., 1984).

Indirect ecological interactions among species in disparate taxa are essential to the functional roles of these individual species (Christensen and Whitham, 1993; Elkinton et al., 1996; Martinsen et al., 1998). Discounting or ignoring the role of indirect effects can lead to erroneous conclusions regarding community dynamics of a system (Davidson et al., 1984; Wooton, 1992). Insects and vertebrates can interact to influence species composition, organic decomposition, and soil properties within an ecosystem (Sharpe et al., 1995; Elkinton et al., 1996). Several studies have shown that attack by one consumer can increase susceptibility to attack by others. Browsing by mammalian herbivores can lead to increased densities of leaf-eating (Martinsen et al., 1998) and galling insects

(Roininen et al., 1997; Olofsson and Strengbom, 2000). Herbivory may induce 1 qualitative changes in defensive chemistry (Tallamy and Raupp, 1991; Loreto et al., 2 2000), resulting in increased occurrence of related and non-related herbivorous taxa 3 (Martinsen et al., 1998; Redman and Scriber, 2000; Tomlin et al., 2000). Although the 4 impact of herbivory on dendrophagic insects has received less attention, Conner and 5 Rudolph (1995) suggested that pecking at resin wells by red-cockaded woodpeckers 6 (*Picoides borealis*) may increase susceptibility of cavity trees to infestation by bark 7 beetles (*Dendroctonus frontalis*.) Similarly, defoliation of jack pine (*Pinus banksiana*) by 8 the jack pine budworm (*Choristoneura pinus pinus*) led to increased colonization by 9 subcortical insects (Wallin and Raffa, 2001). 10 11 We examine a novel interaction involving 2 dendrophagic species and their tree 12 hosts to advance our understanding of plant-herbivore interactions. Previous research effort in this area has focused on leaf-eaters and defoliation. The cost to defend, repair, 13 14 and replace bark damaged by dendrophages likely differs from damage produced by defoliators, warranting examination of a model system involving indirect interactions 15 among a host plant and 2 dendrophages. Our study animals, the porcupine (Erethizon 16 dorsatum) and bark beetles (Coleoptera: Scolytidae), are taxonomically distinct 17 dendrophagous taxa occurring sympatrically in many wooded ecosystems. Their 18 19 phloem-feeding activities impact nutrient flow within the host tree, as opposed to nutrient production (by defoliators) or nutrient storage and acquisition (by root pathogens). 20 Phloem feeding causes wounding or girdling, thereby altering translocation of 21 22 carbohydrates and resulting in increased activity of bark beetles (Dunn and Lorio 1992). Mechanisms of host attack exhibited by porcupines and bark beetles differ. The 23 24 porcupine exploits a variety of habitat types (Roze and Ilse, 2003) and prefer to attack healthy, vigorous trees (Sharpe et al., 1995). A strict herbivore, it consumes deciduous 25 leaves and herbaceous matter during spring and summer, but feeds primarily on inner 26 bark and coniferous foliage during the winter and fall (Roze and Ilse, 2003). Although 27

1 trees may be girdled in the process, damage frequently is restricted to rectangular or ovate patches positioned within grasping distance of branches occupied by the porcupine 2 (Spencer, 1964). Conversely, scolytid species of the *Ips* genus, emerge in spring and 3 4 aggregation and development of multiple generations occurs through summer (Stark. 1982). Furthermore, they attack, feed, and oviposit in stressed or otherwise compromised 5 conifers (Raffa et al., 1993) and exhibit greater host specificity at the species level 6 (Wood, 1963; Cane et al., 1990). Events known or presumed to precipitate pine engraver 7 and other bark beetle infestations include fire, severe drought, mechanical injury, 8 lightning, and even cavity-nesting by the red-cockaded woodpecker (Blanche et al., 1985; 9 Nebeker and Hodges, 1985; Conner and Rudolph, 1995). As with porcupine damage, 10 11 host mortality following infestation is contingent upon attack intensity and tree vigor. 12 The constitutive oleoresin system of healthy conifers acts as a primary defense response against bark beetle attack by flushing out invaders. A secondary, or induced, 13 response contains the infestation when the primary response is insufficient to repel attack 14 (Berryman, 1972; Cates and Alexander, 1982; Raffa, 1991). Chemical and nutritional 15 imbalances resulting from a variety of stressors diminish a host's ability to mount a 16 defensive response and increase the potential for pathogenic infection and increased 17 susceptibility to bark beetle invasion (Hodges et al., 1979; Lorio 1993; Paine and Baker, 18 19 1993). Porcupines and pine engravers feed upon the papershell pinyon pine (*Pinus* 20 remota) in the pinyon-juniper woodlands of the Edwards Plateau region of Texas (Ilse, 21 22 2001; Ilse and Hellgren 2001). Close observation of vigorous, healthy pines in this region often indicated extensive porcupine feeding scars. However, examination of dead 23 24 and dying trees indicated additional presence of pine engraver beetles (*I. hoppingi*). These observations led to development of our hypothesis that porcupine-feeding activity 25 acts similar to other mechanical stressors (e.g., fire, lightning, cavity-building), 26 predisposing these pinyon pines to subsequent colonization by bark beetles and 27

producing an asymmetric, indirect interaction between these 2 dendrophages. We 1 compared morphological and physiological characteristics, as well as colonization 2 success of *I. hoppingi*, of trees that were (target) or were not (nontarget) attacked by 3 4 porcupines. We predicted that if a facultative association occurred among these taxa, we would observe greater colonization of pine engraver beetles on those trees that had been 5 previously damaged by porcupines. We also predicted that we would discern differences 6 in morphology and physiology of target and nontarget trees attributed to taxon-specific 7 host selection. 8 9 2. Study area and methods 10 11 12 2.1 Study area 13 14 Research was conducted on the 2577-ha Kickapoo Caverns State Park (KCSP; Fig. 1) located about 35 km north of Brackettville, Texas. The site (formerly recognized 15 as Kickapoo Caverns State Natural Area) straddles Kinney and Edwards counties in the 16 southwestern region of the Edwards Plateau. Topography was predominantly steep 17 limestone hills and deep canyons with elevations of 482–610 m. Average annual rainfall 18 19 is about 45 cm (National Oceanic and Atmospheric Administration, 1999). No standing 20 water or active springs were present on the site. Shallow clay soils of east- and north-facing slopes supported pinyon-juniper-oak 21 22 plant communities. In addition to papershell pinyon pine, dominant tree species included 23 Ashe juniper (Juniperus ashei), plateau live oak (Ouercus fusiformis), Texas persimmon 24 (Diospyros texana), and vasey oak (O. pungens var. vasevana). Woody shrubs included 25 evergreen sumac (*Rhus virens*), guajillo (*Acacia berlandieri*) prickly pear (*Opuntia* spp.), and Roemer acacia (Acacia roemeriana). Ground cover was represented by cedar 26

panicgrass (Dichanthelium pedicellatum) and cedar sedge (Carex planostachys) in

shaded areas, and sideoats grama (Bouteloua curtipenula) and hairy tridens (Erioneuron 1 pilosum) in more open areas. Shallow soils of the south- and west-facing slopes were 2 dominated by guajillo plant communities, with pinyon pines restricted to lower slopes. 3 4 Shrub species included guajillo, coyotillo (Karwinskia humboldtiana) and leatherstem (Jatropha dioica). Grasses included threeawn (Aristida spp.) and red grama (Bouteloua 5 trifida). Mottes of plateau live oak mixed with vasey oak and Ashe juniper were common 6 in canyons and drainages where moisture was more abundant and soil was deeper. 7 Pinyon pines and netleaf hackberry (*Celtis reticulata*) also occurred in these areas. 8 Dominant grasses include threeawn and annual dropseed (*Sporobolus* spp.). 9 10 11 2.2 Animal capture and handling 12 Porcupines were captured in cage-type live traps (Tomahawk Live Trap Co., 13 Tomahawk, WI) using apples and salt as bait (Hale and Fuller 1996) or by immobilizing 14 15 animals in trees or dens by use of an adjustable pole-mounted syringe. We immobilized porcupines with a mixture of tiletamine hydrochloride and zolazepam hydrochloride 16 (Telazol®, A.H. Robbins, Richmond, VA) at a rate of 7 mg/kg body weight (Hale et al., 17 1994). All animals were marked with self-piercing ear tags (National Band and Tag Co., 18 Newport, KY), and individuals > 1.5 kg body mass were outfitted with radio-transmitters 19 20 (L&L Electronics, Manomet, Illinois) secured by nylon mesh collars. 21 22 2.3 Selection of trees by porcupines 23 24 We used radiotelemetry to locate and obtain visual observations on all animals at least twice weekly during 1997–1999. Triangulation was used only when an animal left 25 the study site and lack of appropriate authorization or hunting seasons precluded our safe 26 and/or lawful access. We recorded location and activity of each porcupine and the tree 27

species for every animal located within a tree. Three hundred, 0.04–ha fixed-radius plots

were randomly established across the study site and sampled to assess relative availability 1 of tree species. We tallied all trees > 1.5 m in height because porcupines were rarely 2 observed using trees shorter than this threshold. 3 4 2.4 Morphology of target vs. nontarget trees 5 6 7 We randomly established 20, 500-m transects (Fig.1) in woodland habitat used by porcupines across the study area. Transects were not placed in open grassland or guajillo 8 9 habitats where porcupine activity was limited. At 25-m intervals on each transect, we 10 tagged the nearest porcupine-damaged (target) tree and a neighboring undamaged (nontarget) tree for a maximum of 20 pairs of trees per transect. Most transects had 11 fewer than 20 pairs of trees due to low availability of nontarget trees. Diameter at root 12 collar (Cognac, 1996), height, crown diameter, crown density and bark thickness were 13 recorded for each tree. Basal area was determined using a 10-factor prism at each tagged 14 15 tree. Limb structure of each tree was characterized as vertical if lateral limbs generally 16 extended upward and horizontal if lateral limbs extended outward. Total number of 17 porcupine feeding scars was recorded and area of bark removed (cm²) was estimated for 18 each target tree. Activity of bark beetles at each tree was characterized as: no beetle 19 20 activity (N); attack only (A), evident by the presence of resin tubules but no successful colonization; or colonization (C), evident by the presence of > 1 dead stems or branches 21 resulting from beetle engravings. Morphological and beetle activity data on all transects 22 were collected during June-August in 1997 and 1998. 23 24 25 2.5 Physiology of transect trees 26 Physiological characteristics of pinyon pines were evaluated by measuring plant 27

moisture stress, 24-hr resin flow, monoterpene content of resin, and carbohydrate content

of phloem in August-September 1998. Plant moisture stress was evaluated using the 1 pressure-bomb technique (Waring and Cleary, 1967; Ritchie and Hinckley, 1975). Stems 2 representing current year's growth were excised and sampled during pre-dawn hours to 3 4 ensure trees were at equilibrium with regard to water potential. Exudate from an arch punch wound (1.25 cm) was collected in plastic vials to 5 determine 24-hr resin flow (Hodges et al., 1979) for all tagged trees. All trees were 6 tapped between 0700 and 0900 h to alleviate photoperiod effects. Tissue removed from 7 the arch punch wound from 3 randomly selected pairs of trees from each transect was 8 placed in plastic bags and frozen before processing for carbohydrate analyses. Phloem 9 was separated from the outer bark and dried at 45° C to a constant weight. Samples were 10 ground using a mortar and pestle. Sugar extraction was performed using a modification 11 12 of the method described by McCready et al. (1950) and modified by Wood and McMeans (1981) for woody tissues. Three extractions were completed using 80% ethanol and then 13 brought to volume using 80% ethanol for a 1:400 dilution. Glucose, fructose, and 14 sucrose were identified using high-pressure liquid chromatography (HPLC) and 15 expressed in mole fractions (umol/mL; Russo et al., 1998). 16 Resin for monoterpene analyses was collected from the same 3 randomly selected 17 pairs of trees used in carbohydrate analyses. Holes (12 mm) were drilled at an upward 18 19 angle into the tree and 1-dram glass vials were screwed directly into the holes to alleviate evaporative loss of hydrocarbons. Vials were removed after 24 h and the resin they 20 contained was processed for subsequent gas chromatograph analysis of 7 monoterpenes 21 22 $(\alpha$ -pinene, camphene, sabinene, β -pinene, myrcene, limonene, and terpinolene). Equal volumes of chromatographic grade pentane were added to aliquots of each sample 23 24 (Snyder, 1992) to facilitate injection into a gas chromatograph (Smith, 1977). 25 2.6 Statistical analyses 26

1	Selection of tree species by porcupines was determined by comparing use with
2	availability using compositional analysis (Aebischer et al. 1993) and individual
3	porcupines as the experimental unit. Analyses were restricted to porcupines with >20
4	locations in trees ($n = 19$) and to 4 groups of tree species: oaks, junipers, pinyon pines,
5	and other. Tree availability was based on plot sampling.
6	Morphological and physiological data for all trees were averaged across each
7	transect ($n = 20$); hence, transects were the experimental units. To explore porcupine
8	selection, average values for target and nontarget trees on each transect were compared
9	using paired <i>t</i> -tests and significance was set at $P < 0.05$. The relationship between
10	porcupine damage (target, nontarget) and beetle activity (none, attacked, colonized) on
11	individual trees was examined with a χ^2 test of independence. In addition,
12	morphological and physiological characteristics of target and nontarget trees was
13	compared using 2-way ANOVA, with level of beetle activity and tree classification
14	(target, nontarget) as main effects and the activity-classification interaction. Resin
15	chemistry data were entered as percent composition of 8 dominant monoterpenes and
16	subjected to angular transformation before analysis. Correlation analyses were
17	performed to determine if percent composition of individual monoterpenes was
18	associated with area of exposed xylem resulting from porcupine herbivory. Alpha levels
19	were Bonferroni-adjusted for correlation analyses of monoterpene concentrations to $P =$
20	0.0024.
21	
22 23	3. Results
24 25	3.1 Selection of trees by porcupines
26	Thirty-seven porcupines (24F;13M) were equipped with radio collars and were
27	tracked > 1 month during the 3-year study period, yielding 1,496 total locations. Visual
28	observations comprised 1,401 of the locations. Eighty percent of those locations ($n =$

1,118) were in trees, 14% (n = 197) were on the ground, and only 6% (n = 86) were 1 located in dens. 2 3 We tallied 1,046 trees representing 10 distinct genera in the plot sampling. Ashe juniper, Texas persimmon, oaks, and pinyon pines accounted for 92% of all available tree 4 species, and 97% of all porcupine observations occurred in these species. Trees were 5 used nonrandomly ($\chi^2_3 = 35.6$, P < 0.001) by the 19 porcupines included in the analysis. 6 Pinyon pines ranked highest in selection, and were preferred (P < 0.05) in all pairwise 7 comparisons with other tree groups. Oaks ranked second, and were preferred (P < 0.05)8 relative to junipers and other trees. 9 10 3.2 Morphology of transect trees 11 12 Data were collected on 366 trees (183 pairs). Four morphological characteristics 13 differed between target and nontarget trees (Table 2). Porcupines used trees that were 14 greater in girth, taller, greater in crown diameter, and thicker-barked. *Post hoc* analyses 15 of those 5 characteristics revealed that diameter at root collar was significantly correlated 16 with height (r = 0.79, P < 0.0001), bark thickness (r = 0.83, P < 0.0001), crown diameter 17 (r = 0.89, P < 0.0001), and resin flow (r = 0.50, P < 0.001). Additionally, 72 % (n = 131)18 of all trees damaged by porcupines (n = 183) exhibited horizontal instead of vertical 19 lateral limb structure, indicating disproportionate use of this structural characteristic (χ^2 = 20 35.2, df = 1, P < 0.0001). Similarly, beetle colonization was associated with horizontal 21 limb structure ($\chi^2 = 13.6$, df = 2, P < 0.001). Of beetle-colonized trees, 83% (83 of 100) 22 23 had horizontal limb structure vs. 62.5% (15 of 24) attacked trees and 56.9% (33 of 58) of trees with no beetle activity. 24 25 Trees damaged by porcupines were more likely to be colonized by bark beetles, and undamaged trees were more likely to free from beetle attack ($\chi^2 = 75.3$, df = 3, P <26

0.01, Fig. 2). However, tree characteristics varying by level of beetle activity were not

- entirely consistent with those associated with porcupine use of trees (Table 2). Larger
- trees were more likely to be fed upon by porcupines and colonized by beetles; however,
- 3 resin flow did not vary across levels of beetle activity. Pinyon pine basal area varied by
- 4 the interacting effects of beetle activity and tree classification ($F_{2,92} = 3.94$, P = 0.02).
- Basal area of pinyon pines (Table 2) was greater (P < 0.055) on colonized, nontarget
- trees ($\overline{x} + SE = 3.1 + 0.8 \text{ m}^2/\text{ha}$) than any other combination of beetle activity and tree
- 7 class (pooled $\overline{x} \pm SE = 1.4 \pm 0.3 \text{ m}^2/\text{ha}$, all cell \overline{x} between 1.1 and 1.8 m²/ha).
- Number of porcupine scars was correlated positively with area of bark that had
- been removed (r = 0.90, P < 0.0001). Area of bark removed was less (P < 0.001) on
- trees that exhibited no beetle activity or attack only (pooled $\bar{x} \pm SE = 303 \text{ cm}^2 \pm 75$) than
- on trees that had been colonized ($\bar{x} \pm SE = 989 \text{ cm}^2 \pm 152$).

12

3.3 Physiology of transect trees

- Target trees had greater resin flow than nontarget trees (Table 2). Plant moisture
- stress did not differ between target and non-target trees (P > 0.05) and measurable
- amounts of sucrose were found in only 2 samples collected from transects. Glucose and
- fructose dominated all samples, but did not differ between target and nontarget trees
- 19 (glucose: $\bar{x} \pm SE = 0.410 \,\mu\text{mol/mL} \pm 0.020$; fructose: 0.417 $\mu\text{mol/mL} \pm 0.023$). Trees
- that were colonized by beetles had higher levels of both these sugars (glucose: 0.451
- μ mol/mL + 0.021, P < 0.01; fructose: 0.455 μ mol/mL + 0.022, P < 0.03) than trees that
- had no beetle activity or had been attacked only (glucose: 0.389 μmol/mL + 0.019;
- fructose: $0.395 \,\mu\text{mol/mL} + 0.022$). An interaction in fructose concentration across levels
- of beetle activity and by tree classification approached significance (P < 0.08).
- 25 Undamaged trees that were colonized by beetles tended to have higher levels of fructose
- $(0.544 \mu mol/mL + 0.04)$ than target trees $(0.422 \mu mol/mL + 0.03)$ and undamaged trees
- with no beetle activity or attack only (0.386 μ mol/mL \pm 0.02). Fructose concentrations

also were higher in undamaged colonized trees than all target trees (0.422 µmol/mL + 1 0.03). 2 The most abundant monoterpenes in target and nontarget trees (n = 120) were α -3 4 pinene, β-pinene, and limonene (Table 3). Alpha-pinene concentration was correlated negatively with sabinene (r = -0.36, P < 0.0017), β -pinene (r = -0.89, P < 0.0001), 5 myrcene (r = -0.73, P < 0.0001), limonene (r < -0.88, P < 0.001), and terpinolene (r = -0.73, P < 0.0001) 6 0.35, P < 0.002). Positive correlations were evident between sabinene and myrcene (r =7 0.40, P < 0.0005), β -pinene and myrcene (r = 0.72, P < 0.0001), β -pinene and limonene 8 (r = 0.67, P < 0.0001), myrcene and limonene (r = 0.48, P < 0.001), myrcene and 9 terpinolene (r = 0.45, P < 0.0001) and terpinolene and sabinene (r = 0.97, P < 0.0001). 10 11 Of the terpenes we examined, only myrcene concentration differed between target and nontarget trees (P = 0.008; Table 3). Negative associations were detected 12 between area of bark removed and levels of sabinene (r = -0.48, P = 0.03), terpinolene (13 r = -0.47, P = 0.03), and myrcene (r = -0.16, P = 0.08). Sabinene, limonene, and 14 terpinolene occurred in lower proportions in trees that had been colonized or attacked 15 than in trees with no beetle activity (Table 3). The only monoterpene affected by the 16 interaction of beetle activity and porcupine damage was camphene ($F_{2.63} = 5.16$, P =17 0.008). Undamaged, colonized trees had higher proportions of this monoterpene (\bar{x} + 18 SE = 2.4% + 1.9) than trees with other combinations of beetle activity and porcupine 19 damage (pooled $\bar{x} = 0.4\% + 0.1$). 20 21 4. Discussion 22 23 24 Our findings are compatible with our predictions that porcupine bark-feeding activity predisposes pinyon pines to subsequent bark beetle activity. Our results do not 25 unequivocally define the mechanism that explains our observations. However, we 26 propose and elaborate 2 non-exclusive and likely interacting mechanisms responsible for 27

the facilitative association observed in our study. These mechanisms are, first, that 1 porcupine damage is a stressor needed for successful colonization by bark beetles, and 2 second, that porcupine damage causes release of volatile terpenes, which in turn cue pine 3 4 engraver beetles to the status of potential colonization sites. In addition, we will reject alternative explanations regarding the direction of the indirect interaction. 5 The first mechanism, that porcupine damage represents an additional stressor 6 requisite to infestation by these bark beetles, is supported by our result that trees damaged 7 by porcupines were more likely to be colonized by bark beetles than undamaged trees. In 8 addition, trees colonized by bark beetles had, on average, 3 times as much bark removed 9 by porcupines as non-colonized trees. Association of beetles with trees in areas of higher 10 pinyon pine basal area in this study further emphasizes the role of stress in insect 11 12 outbreaks (Hodges and Lorio, 1975; Mattson and Haack, 1987; Paine and Baker, 1993). Under these conditions, balance of nutrients necessary for growth and defense responses 13 of the tree is compromised because of lower levels of photosynthates and increased 14 competition for nutrients among within-plant processes (growth-differentiation 15 hypothesis; Lorio, 1986). Consistent with our results, Lombardero et al. (2000) found 16 that resin flow induced by wounding in *Pinus taeda* was lowest in trees with smaller 17 crowns and in areas of high basal density. Successful colonization by beetles was likely 18 19 facilitated because of the diminished resistance by the host tree. We postulate that injury to trees by porcupines elicits a similar response, causing the tree to displace nutrients 20 used for growth to the wound site for defense (Christiansen et al., 1987). Disruption of 21 22 carbon allocation by reallocation of photosynthates to terpenes and resins surrounding feeding scars, and therefore away from the remainder of the tree, puts the tree at 23 24 increased vulnerability to beetle infestation. 25 Porcupines may select pinyon pines over more abundant species to optimize nutrient acquisition and thermoregulation. Coniferous species provide thermal 26 advantages over deciduous species (Clarke and Brander, 1973) and frequently constitute 27

the preferred feeding and resting trees of porcupines (Dodge, 1967; Griesemer, 1995;

- 2 Speer and Dilworth, 1978). We observed that feeding by porcupines on oaks was
- 3 restricted to leaves and acorns in the canopy, whereas consumption of bark was apparent
- 4 only in pinyons. Pine bark and cambium is easier to remove than oak bark and is
- 5 generally higher in fats and water content. Conversely, oak leaves and acorns are higher
- 6 in protein than pine bark and cambium (Stricklan et al., 1995). Live oaks on site allowed
- 7 porcupines to supplement and balance winter nutritional needs with foliage, precluding a
- 8 diet restricted to bark.
- Paired sampling on transects allowed us to eliminate site favorability as a cause
- for tree selection by porcupines. Significant correlations of height, crown diameter, bark
- thickness, and resin flow with diameter at root collar indicate that size is the dominant
- factor in intraspecific selection of trees. Similar findings of size-related selectivity by
- porcupines have been reported (Krefting et al., 1962; Tenneson and Oring, 1985; Sullivan
- et al., 1986). Trees colonized by beetles also were larger than nontarget trees.
- 15 Preponderance of horizontal limb structure in target trees was indicative of the habit of
- porcupines to rest on branches and then feed and remove bark within comfortable reach
- 17 (Spencer, 1964). These data validate the contention that morphological selection of trees
- by porcupines reflects foraging optimization (Roze, 1989).
- We reject alternative explanations regarding the direction of the indirect
- interaction. First, it is unlikely that beetles are predisposing subsequent feeding by
- 21 porcupines because beetles are facultative colonizers, usually requiring stressed hosts
- 22 (but see Santoro et al., 2001), whereas porcupines are not limited by host health.
- Facilitation of beetle attack by porcupine feeding exemplifies an asymmetrical process.
- 24 Porcupines, because of their size, sharp incisors, and large claws, select large, healthy
- 25 trees from which they can easily remove bark. *Ips*, however, is a facultative parasite that
- 26 generally attacks stressed or injured hosts (Raffa et al., 1993; but see Santoro et al.,
- 27 2001). Unlike the porcupine, which is not dependent on a single host tree for survival, the

pine engraver spends much of its life cycle within the inner bark of the host, leaving only 1 to disperse to a new host. Furthermore, successful brood production by all scolytid 2 beetles requires recently dead tissue (Raffa et al., 1993). Hence, the engraving pattern 3 4 typical in the course of colonization by *Ips* beetles results in death of the colonized host limb/whole tree, rendering it unavailable for use by porcupines. 5 We also reject the alternative explanation that the relationship we observed was a 6 result of similar tree preferences by the 2 dendrophages in our study system. This 7 explanation was similarly discussed and ultimately rejected for a budworm 8 (Choristoneura pinus pinus)—bark beetle (Ips grandicollis)—woodborer (Monochamus 9 carolinensis) system (Wallin and Raffa, 2001). Porcupines and beetles exhibited several 10 11 differences in tree selection, especially physiologically. Our data indicate that 12 biochemical variability may influence porcupine diet selection, but minimally. We found no correlation between feeding activity of porcupines and levels of limonene, in contrast 13 to Snyder and Linhart (1997). Target and nontarget trees of porcupines differed for only 14 1 physiological measure, myrcene concentration. On the other hand, as discussed below, 15 the biochemical composition of trees colonized by beetles differed from uncolonized 16 trees for several terpenes and carbohydrates. Association of increased beetle colonization 17 on trees with higher levels of glucose and fructose was not surprising given nutritional 18 19 requirements of the pine engraver (Haack and Slansky, 1987). We propose that the second mechanism to explain the directional nature of the 20 association between beetle colonization and porcupine feeding activity is that damage by 21 22 porcupines causes release of volatile terpenes, thereby cueing pine engraver beetles to 23 presence of toxic substances and availability of potential pheromone precursors. Such a 24 mechanism parallels that observed for defoliator-conifer-bark beetle models of hostherbivore interactions (Wallin and Raffa, 1999; Erbilgin et al., 2003). In those systems, 25 variation in monoterpene content and composition of the induced defenses of a host tree 26 led to variation in the relative aggregation or inhibition of bark beetles (*Ips pini*; Erbilgin 27

et al., 2003). Defoliation intensity, which is analogous to amount of bark removed in our 1 study, directly altered monoterpene composition and was inversely related to 2 monoterpene concentration over long time frames (≥ 12 months; Wallin and Raffa, 1999, 3 4 2001). Beetles, in contrast to the porcupine, appeared to select trees based on 5 monoterpene characteristics of resin. High levels of limonene are toxic to many bark 6 beetles (Harborne, 1993) and explain increased beetle activity on trees in our study area 7 exhibiting low proportions of this monoterpene. Loreto et al. (2000) reported increased 8 emissions of limonene and α-pinene in artificially wounded needles of the Mediterranean 9 pine (*Pinus pinea*). Hence, bark removal by the porcupine and resultant vaporization of 10 11 limonene may alert the beetle to levels of this hydrocarbon. Similar responses to 12 herbivore-induced volatile plant terpenes and other plant chemicals have been reported for parasitic wasps (Turlings et al., 1990) and anthocorid predators (Drukker et al., 2000). 13 14 Chemical reactions between host monoterpenes and pheromone production are complex and have not been identified completely for *I. hoppingi*. However, myrcene and 15 α -pinene serve as pheromone precursors in other species of pine engraver beetles 16 (Hughes, 1974; Renwick et al., 1976; Hughes and Renwick, 1977; Byers et al., 1979). 17 Cane et al. (1990) reported a lack of pheromone specificity occurring between *I. confusus* 18 19 and *I. hoppingi*, which are closely related to *I. paraconfusus* but are host-specific to pinyon pines. They attributed this lack of specificity to recent phylogenetic divergence of 20 these beetles. Based on the close phylogenetic relationship among these species of *Ips*, 21 22 we speculate that myrcene and α -pinene may be suitable pheromone precursors for *I*. hoppingi. 23 24 The overwhelming dominance of monoterpene composition by α -pinene in all groups of trees may have masked an association between beetle colonization and this 25 terpene. Alpha-pinene, a major constituent of pines and other conifers, is particularly 26 dominant in pinyon pines. Percent composition of this monoterpene ranges from about 27

10% in ponderosa pine (Sturgeon, 1979; Snyder, 1992) to > 90% in populations of 1 Mexican pinyon pine (*Pinus cembroides*) in the Big Bend region of Texas (Zavarin and 2 Snajberek, 1985). We postulate that *I. hoppingi* is either an α-pinene obligate species or 3 4 harbors Bacillus cereus and is therefore able to biosynthesize this monoterpene, as in Ips paraconfusus (Brand et al. 1975). Beetles attacked or colonized trees with lower levels 5 of sabinene and terpinolene, which were terpenes that were correlated negatively with 6 concentrations of α -pinene. 7 We acknowledge that a stronger test of our hypothesis would have included 8 experimental manipulation of trees or demonstrating the temporal sequence of 9 dendrophagy. Alternatively, long-term temporal observation across a quantified range of 10 11 bark removal by porcupines (sensu Wallin and Raffa, 2001) also would also strengthen 12 our inference. However, we were constrained by the limited number of pinyon pines of comparable size and possessing physiological characteristics preferred by porcupines that 13 14 were also free of porcupine and/or beetle damage (Ilse, 2001). Conservation concerns on the study area also precluded the ability to artificially inflict damage requisite to 15 experimental manipulation. Therefore, we were unable to adequately test our hypothesis 16 with an experimental manipulation at our study site. However, our detailed observational 17 data supports the hypothesis and rejects alternative explanations. 18 19

The novel aspect of our work is the focus on the indirect effects of a vertebrate dendrophage on a insect dendrophage with the same host plant, as opposed to the large body of work exploring effects of insect defoliators on host plant suitability to other insect herbivores. However, the proposed mechanisms for the association between porcupines and bark beetles fit within the framework of integrated theories explaining host plant-herbivore interactions. First, the porcupine feeding process can be generalized to a mechanical stress to the host tree, rendering it susceptible to invertebrate infestation. The outcome of the invertebrate attack likely depends on the interaction among seasonal timing of initial damage by porcupines, attack magnitude by beetles, and the induced-

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1 defense response of the stressed host (Raffa and Berryman, 1983), which is a trade-off best explained by the growth-differentiation model of plant resistance (Lorio, 1986; 2 Herms and Mattson, 1992). Second, the putative mechanism linking the 2 dendrophages 3 4 involves release of chemical cues (e.g., monoterpenes) from the host tree that are triggered by the mechanical damage. The specific composition and concentration of these 5 cues can either trigger beetle aggregation or inhibit beetle attraction (Erbilgin et al., 6 2003). We suggest that the differential nature of the induced response by host plants to 7 varying types of wounding (e.g., defoliation, root pathogenic infection, bark removal) be 8 examined. 9 10 11 12 Acknowledgements 13 14 We are indebted to D. Stuart and D. Klepper for invaluable logistical, technical, 15 and field assistance. D. M. Leslie, Jr., R. Masters, and W. Ross provided valuable 16 research suggestions, and E. M. Schauber reviewed the manuscript. T. Reeder and the M. 17 Lacy family assisted in capturing and handling of porcupines. B. A. Kimball of the U.S. 18 19 Department of Agriculture and D. Chrz of the Department of Horticulture at Oklahoma State University were instrumental in GC-MS and HPLC analyses of terpenes and 20 carbohydrates. Funding for this research was provided by the Rob and Bessie Welder 21 22 Wildlife Foundation, Max McGraw Wildlife Foundation, Theodore Roosevelt Memorial 23 Fund, Oklahoma Cooperative Fish and Wildlife Research Unit (Oklahoma State

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1 2 References 3 Aebischer, N. J., Robertson, P. A., Kenward, R. E. 1993. Compositional analysis of 4 5 habitat use from animal radio-tracking data. Ecology 74:1313-1325. Berryman, A.A. 1972. Resistance of conifers to invasion by bark beetles-fungus 6 associations. Bioscience. 22, 598-602. 7 Blanche, C.A., Hodges, J.D., Nebeker, T.E.. 1985. Changes in bark beetle 8 susceptibility: indicators in a lightning-struck pine. Can. J. For. Res. 15, 397-399. 9 Brand, J. M., Bracke, J.W., Markowitz, A.J., Wood, D.L., Browne, L.E.. 1975. 10 Production of verbenol pheromone by a bacterium isolated from bark beetles. 11 12 Nature 254, 136-137. 13 Byers, J. A., Wood, D.L., Browne, L.E., Fish, R.H., Piatek, B., Hendry, L.B. 1979. Relationship between a host plant compound, myrcene and pheromone production 14 in the bark beetle *Ips paraconfusus*. J. Insect Physiol. 25, 477-482. 15 Cane, J. H., Merrill, L.D., Wood, D.L. 1990. Attraction of pinyon pine bark beetle, *Ips* 16 hoppingi to conspecific and I. confusus pheromones (Coleoptera: Scolytidae). J. 17 18 Chem. Ecol. 16, 2791-2798. Cates, R.G., Alexander, H. 1982. Host resistance and susceptibility. In: Mitton, J.B., 19 Sturgeon, K.B. (Eds.), Bark beetles in North American Conifers: A System for the 20 Study of Evolutionary Biology. University of Texas Press, Austin, USA, pp. 212-21 263. 22 23 Christiansen, E., Waring, R.H., Ferryman, A.A. 1987. Resistance of conifers to bark beetles attack: searching for general relationships. For. Ecol. Manage. 22, 89-24 106. 25 Christensen, K.M., Whitham, T.G. 1993. Impact of insect herbivores on competition 26 between birds and mammals for pinyon pine seeds. Ecology 74, 2270-2278. 27

- 1 Clarke, S.H., Brander, R.B. 1973. Radiometric determination of porcupine surface
- temperature under two conditions of overhead cover. J. Physiol. Zool. 46, 230-
- 3 237.
- 4 Cognac, D.C. 1996. Estimating diameter growth for pinyon and juniper trees in Arizona
- 5 and New Mexico. USDA Forest Service Research Note INT-RN-429.
- 6 Connell, J.H. 1983. On the prevalence and relative importance of interspecific
- 7 competition: evidence from field experiments. Am. Nat. 122, 81-116.
- 8 Conner, R.N., Rudolph, D.C. 1995. Losses of red-cockaded woodpecker cavity trees to
- 9 southern pine beetles. Wilson Bull. 107, 81-92.
- Davidson, D.W., Inouye, R.S., Brown, J.H. 1984. Granivory in a desert ecosystem:
- experimental evidence for indirect facilitation of ants by rodents. Ecology 65,
- 12 1780-1786.
- Dodge, W.E. 1967. The biology and life history of the porcupine, (*Erethizon dorsatum*)
- in western Massachusetts. Ph.D. Dissertation, University of Massachusetts,
- 15 Amherst.
- Dunn, J.P., Lorio, Jr., P.L. 1992. Effects of bark girdling on carbohydrate supply and
- resistance of loblolly pine to southern pine beetle (*Dendroctonus frontalis* Zimm.)
- 18 attack. For. Ecol. Manage. 50, 317-330.
- 19 Drukker, B., Bruin, J., Sabelis. M.W. 2000. Anthocorid predators learn to associate
- 20 herbivore-induced plant volatiles with presence or absence of prey. Physiol.
- 21 Entomol. 25, 260-265.
- 22 Elkinton, J.S., Healy, W.M., Buonccorsi, J.P., Boettner, G.H., Hazzard, A.M., Smith,
- 23 H.R., Liebhold, A.M. 1996. Interactions among gypsy moths, white-footed mice,
- and acorns. Ecology 77, 2332-2342.
- 25 Erbilgin, N., Powell, J.S., Raffa, K.F. 2003. Effect of varying monoterpene
- 26 concentrations on the responses of *Ips pini* (Coleoptera: Scolytidae) to its

aggregation pheromone: implications for pest management and ecology of bark

- beetles. Agric. For. Entomol. 5, 269-274.
- 3 Griesemer, S.J. 1995. Habitat selection by porcupines in central Massachusetts. MSc
- 4 Thesis, University of Massachusetts, Amherst.
- 5 Haack, R.A., Slansky, Jr., F. 1987. Nutritional ecology of wood-feeding coleoptera,
- 6 lepidoptera, and hymenoptera. In: Slansky, Jr., F., G. Rodriguez, J.G. (Eds.),
- Nutritional Ecology of Insects, Mites, Spiders, and Related Invertebrates. John
- 8 Wiley and Sons, New York, USA, pp.449-486.
- 9 Hale, M.B., Fuller, T.K. 1996. Porcupine (*Erethizon dorsatum*) demography in central
- 10 Massachusetts. Can. J. Zool. 74, 480-484.
- Hale, M.B., Griesemer, S.J., Fuller, T.K. 1994. Immobilization of porcupines with
- tiletamine hydrochloride and zolazepam hydrochloride (Telazol®). J. Wildl. Dis.
- 13 30, 429-431.
- Harborne, J.B. 1993. Hormonal interactions between plants and animals. Ecological
- Biochemistry, Academic Press, New York, USA, pp.104-127.
- Herms, D.A., Mattson, D.J. 1992. The dilemma of plants: to grow or to defend. Quart.
- 17 Rev. Biol. 67, 283-355.
- Hodges, J.D., Elam, W.W., Watson, W.F., Nebeker, T.E. 1979. Oleoresin characteristics
- and susceptibility of four southern pines to southern pine beetle
- 20 (Coleoptera: Scolytidae) attacks. Can. Entomol. 11, 889-896.
- Hodges, J.D., Lorio, Jr., P.L. 1975. Moisture stress and composition of xylem oleoresin
- in loblolly pine. For. Sci. 21, 283-290.
- Hughes, P.R. 1974. Myrcene: a precursor of pheromones in *Ips* beetles. J. Insect
- 24 Physiol. 20, 1271-1275.
- 25 Hughes, P.R., Renwick, J.A.A. 1977. Neural and hormonal control of pheromones in the
- bark beetle *Ips paraconfusus*. Physiol. Entomol. 2, 117-123.

Huntly, N. 1991. Herbivores and the dynamics of communities and ecosystems. Ann.

- 2 Rev. Ecol. Syst. 22, 477-503.
- 3 Ilse, L. M. 2001. Porcupines, pinyon pines and pine engravers: interspecific
- 4 relationships in pinyon-juniper woodlands. PhD Thesis, Oklahoma State
- 5 University, Stillwater.
- 6 Ilse, L.M., Hellgren, E.C. 2001. Demographic and behavioral characteristics of North
- 7 American porcupines (*Erethizon dorsatum*) in pinyon-juniper woodlands of
- 8 Texas. Am. Midl. Nat. 146, 329-338.
- 9 Krefting, L.W., Stoeckeler, J.H., Bradle, B.J., Fitzwater, W.D. 1962. Porcupine-timber
- relationships in the Lake States. J. For. 60, 325-330.
- Lombardero, M.J., Ayes, M.P., Lorio, Jr., P.L., Ruel, J.J. 2000. Environmental effects
- on constitutive and inducible resin defences of *Pinus taeda*. Ecol. Letters 3, 329-
- 13 339.
- Loreto, F., Nascetti, P., Graverini, A., Manozzi, M. 2000. Emission and content of
- monoterpenes in intact and wounded needles of the Mediterranean Pine *Pinus*
- 16 pinea. Funct. Ecol. 14, 589-596.
- Lorio, P.L. Jr. 1986. Growth-differentiation balance: a basis for understanding southern
- pine beetle-tree interactions. For. Ecol. Manage. 14, 259-273.
- Lorio, P.L. Jr. 1993. Environmental stress and whole-tree physiology. In: Schwatter,
- T.J., Filip, G.M. (Eds.), Beetle-Pathogen Interactions in Conifer Forests.
- Academic Press, Ltd., London, UK, pp. 88-101.
- Martin, T.E. 1988. Processes organizing open-nesting bird assemblages: competition or
- 23 nest predation? Evol. Ecol. 2, 37-50.
- Martinsen, G.D., Driebe, E.M., Whitham, T.G. 1998. Indirect interactions mediated by
- 25 changing plant chemistry: beaver browsing benefits beetles. Ecology 79, 192-
- 26 200.

1 Mattson, W. J., Haack, R.A. 1987. The role of drought in outbreaks of plant-eating

- 2 insects. Bioscience 37, 110-118.
- 3 McCready, R.M., Guggolz, J., Silviera, V., Owens, H.S. 1950. Determination of starch
- and amylase in vegetables. Analyt. Chem. 22, 1156-1158
- 5 National Oceanographic and Atmospheric Administration. 1999. Texas Climatological
- 6 Summary. National Climatic Data Center, Asheville, North Carolina, USA
- 7 Nebeker, T.E., Hodges, J.D. 1985. Thinning and harvesting practices to minimize stand
- 8 disturbance and susceptibility to bark beetle and disease attacks. In: Branham,
- 9 S.J., Thatcher, R.C. (Eds.), Proceedings of the Integrated Pest Management
- Research Symposium. USDA Forest Service General Technical Report SOH-56,
- pp. 263-271.
- Olofsson, J., Strengbom, J. 2000. Response of galling invertebrates on Salix lanata to
- reindeer herbivory. Oikos 91, 493-498.
- Paine, T.D., Baker, F.A. 1993. Abiotic and biotic predisposition. In: T. D. Schwatter,
- T.D., Filips, G.M. (Eds.) Beetle-Pathogen Interactions in Conifer Forests.
- Academic Press, Ltd., London, UK, pp. 61-79.
- 17 Raffa, K.F. 1991. Induced defensive reactions in conifer-bark beetle systems. In:
- Tallamy, D.W., Raup, M.J. (Eds.), Phytochemical induction by herbivores. John
- Wiley and Sons, Inc., New York, USA, pp. 245-276.
- 20 Raffa, K.F., Berryman, A.A. 1983. The role of host plant resistance in the colonization
- behavior and ecology of bark beetles (Coleoptera: Scolytidae). Ecol. Monogr. 53,
- 22 27-49.
- Raffa, K.F., Phillips, T.W., Salom, S.M. 1993. Strategies and mechanisms of host
- colonization by bark beetles. In: Schwatter, T.J., Filips, G.M. (Eds.) Beetle-
- Pathogen Interactions in Conifer Forests. Academic Press, Ltd., London, UK, pp.
- 26 103-128.

Redman, A.M., Scriber, J.M. 2000. Competition between the gypsy moth, *Lymantia*

- 2 dispar, and the northern tiger swallowtail, Papilio canadensis: interactions
- mediated by host plant chemistry, pathogens, and parasitoids. Oecologia 125,
- 4 218-228.
- 5 Reid, M.L., Robb, T. 1999. Death of vigorous trees benefits bark beetles. Oecologia
- 6 120, 555-562.
- 7 Ritchie, G.A., Hinckley, T.M. 1975. The pressure chamber as an instrument for
- 8 ecological research. Adv. Ecol. Res. 9, 165-254.
- 9 Roininen, H., Price, P.W., Bryant, J.P. 1997. Response of galling insects to natural
- browsing by mammals in Alaska. Oikos 80, 481-486.
- Roze, U. 1989. The North American porcupine. Smithsonian Institution Press,
- Washington, D. C., USA.
- Roze, U., Ilse, L.M. 2003. Porcupine (Erethizon dorsatum). In: Feldhamer, G.A.,
- Thompson, B.C., Chapman, J.A. (Eds.), Wild Mammals of North America:
- Biology, Management and Conservation. Johns Hopkins Press, Baltimore, MD,
- 16 USA, pp. 371-380.
- Russo, V.M., Williamson, J, Roberts, K., Wright, J.R., Maness, N. 1998. 13-nmr
- spectroscopy to monitor sugars in pith of internodes of a *sh2* corn at
- developmental stages. HortScience 33, 980-983.
- Santoro, A.E., Lombardero, M.J., Ayres, M.P., Ruel, J.J. 2001. Interactions between fire
- and bark beetles in an old growth pine forest. For. Ecol. Manage. 144, 245-254.
- Schoener, T.W. 1983. Field experiments on interspecific competition. Am. Nat. 122,
- 23 240-285.
- Sharpe, G.W., Hendee, C.W., Sharpe, W.F., Hendee, J.C.. 1995. Insects and mammals.
- 25 Pages 141-161 In: Introduction to Forests and Renewable Resources. McGraw-
- 26 Hill, New York. USA, pp. 141-161.

- Sih, A., Crowley, P., McPeek, M., Petranka, J., Strohmeier, K. 1985. Predation,
- competition and prey communities: a review of field experiments. Ann. Rev.
- 3 Ecol. Syst. 16, 269-311.
- 4 Smith, R.H. 1977. Monoterpenes of ponderosa pine xylem resin in western United
- 5 States. USDA Forest Service General Technical Report PSW-1.
- 6 Snyder, M.A. 1992. Selective herbivory by Abert's squirrel mediated by chemical
- variability in ponderosa pine. Ecology 73, 1730-1741.
- 8 Snyder, M.A., Linhart, Y.B. 1997. Porcupine feeding patterns: selectivity by a generalist
- 9 herbivore? Can. J. Zool. 7, 2107-2111.
- Speer, R.J., Dilworth, T.G. 1978. Porcupine winter foods and habitat utilization in
- central New Brunswick. Can. Field Nat. 92, 271-274.
- 12 Spencer, D.A. 1964. Porcupine fluctuations in past centuries revealed by
- dendrochronology. J. Appl. Ecol. 1, 127-149.
- 14 Stark, R.W. 1982. Generalized ecology and life cycle of bark beetles. In: Mitton, J.B.,
- and K. B. Sturgen, K.B. (Eds.), Bark Beetles in North American Conifers: A
- System for the Study of Evolutionary Biology. University of Texas Press, Austin,
- 17 USA, pp. 21-45.
- Stricklan, D.J., Flinders, J.T., Cates, R.G. 1995. Factors affecting selection of winter
- food and roosting resources by porcupines in Utah. Great Basin Nat. 55, 29-36.
- Sturgeon, K.B. 1979. Monoterpene variation in Ponderosa pine xylem resin related to
- western pine beetle predation. Evolution 33, 803-814.
- Sullivan, T. P., Jackson, W.T., Pojar, J., Banner, A. 1986. Impact of feeding damage by
- 23 the porcupine on western hemlock-Sitka spruce forests of north-coastal British
- 24 Columbia. Can. J. For. Res. 16, 642-647.
- Tallamy, D.W., Raupp, M.J. 1991. Phytochemical Induction by Herbivores. John Wiley
- and Sons Inc., New York, USA.

Tenneson, C., Oring, W. 1985. Winter food preferences of porcupines. J. Wildl.

- 2 Manage. 49, 28-33.
- Tomlin, E. S., Antonejevic, E., Alfaro, R.I., J. H. Borden, J.H. 2000. Changes in
- 4 volatile terpene and diterpene resin acid composition of resistant and susceptible
- 5 white spruce leaders exposed to simulated white pine weevil damage. Tree
- 6 Physiol. 20, 1087-1095.
- 7 Turlings, C. J., Tumlinson, J.H., Lewis, W.J. 1990. Exploitation of herbivore-induced
- plant odors by host-seeking parasitic wasps. Science 250, 1251-1253.
- 9 Wallin, K.F., Raffa, K.F. 2001. Effects of folivory on subcortical plant defenses: can
- defense theories predict interguild processes? Ecology 82, 1387-1400.
- Wallin, K. F., Raffa, K.F. 1999. Altered constitutive and inducible phloem
- monoterpenes following natural defoliation of jack pine: implications to host-
- mediated interguild interactions and plant defense theories. J. Chemical Ecol. 25,
- 14 861-880.
- Waring, R.H., Cleary, B.D. 1967. Plant moisture stress: evaluation by pressure bomb.
- Science 155, 1248-1254.
- Wood, B.W., McMeans, J.L. 1981. Carbohydrate changes in various organs of bearing
- and non-bearing pecan trees. J. Am. Soc. Hort. Sci. 106, 758-761.
- 19 Wood, D.L. 1963. Studies on host selection by *Ips confusus* (Le Conte)
- 20 (Coleoptera: Scolytidae). California University Publication in Entomology 27,
- 21 241-282.
- 22 Wood, D.L. 1982. The role of pheromones, kairomones, and allomones in the host
- selection and colonization behavior of bark beetles. Ann. Rev. Entomol. 27, 411-
- 24 446.
- Wooton, J.T. 1992. Indirect effects, prey susceptibility, and habitat selection: impacts
- of birds on limpets and algae. Ecology 73, 981-991.

Wooton, J.T. 1994. The nature and consequences of indirect effects in ecological
communities. Ann. Rev. Ecol. Syst. 25, 413-466.
Zavarin E., Snajberk, K. 1985. Monoterpenoid and morphological differentiation within
Pinus cembroides. Biochem. Syst. Ecol. 13, 89-104.

Table 1. Availability and use of dominant tree species by porcupines on Kickapoo Caverns State Park, southwestern Texas, USA, 1996-1999.

	Availability	$y^a (n = 965)$	Use $(n = 1092)$		
Tree species	\overline{n}	%	n	%	
Ashe juniper	350	36	155	14	
Oak species	253	26	393	35	
Texas persimmon	216	22	2	0.2	
Papershell pinyon pine	146	15	542	50	

^a Based on 300, 0.04 ha fixed-radius plots.

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Table 2. Morphological and physiological characteristics of pinyon pines differing with regard to porcupine and bark beetle activity on the

3 Kickapoo Caverns State Park, southwestern Texas, USA 1996-1999. Values represent average of 20 transect means.

	Porcupines					Beetles ^a							
	Target Nor		Nont	ntarget		None		Attacked		Colonized			
Characteristic	$\frac{}{x}$	SE	$\frac{}{x}$	SE	P^{b}	\overline{x}	SE	\overline{x}	SE	\overline{x}	SE	P^{c}	
DRC (cm) ^d	22.3	1.1	12.4	0.5	< 0.001	14.8 A	0.6	15.2 A	0.9	23.6 B	0.9	< 0.001	
Height (m)	4.6	0.2	3.2	0.1	< 0.001	4.4 A	0.2	4.6 A	0.3	7.1 B	0.3	< 0.001	
Crown diameter (m)	4.1	0.2	2.3	0.1	< 0.001	2.7 A	0.2	2.7 A	0.2	4.1 B	0.2	< 0.001	
Crown density (%)	88.7	1.7	88.2	1.7	0.755	88.1AB	0.2	91.8A	0.2	84.3B	0.2	0.147	
Bark thickness (cm)	0.8	0.03	0.4	0.02	< 0.001	0.4 A	0.03	0.4 A	0.0	0.8 B	0.0	< 0.001	
Basal area (m²/ha)e	3.1	0.2	1.5	0.2	0.424	1.5 A	0.3	1.3 A	0.3	2.0 B	0.4	0.04	
24-hr Resin flow (mL)	2.6	0.3	1.4	0.2	< 0.001	1.9	0.2	2.0	0.3	2.6	0.3	0.09	

^a Means for each characteristic with the same letter are not different at $\alpha = 0.05$ level of significance.

⁶ b Main effect of porcupine damage (present or absent).

⁷ c Main effect of level of beetle activity.

⁸ d Diameter at root collar.

⁹ e See text for discussion of interactive effects.

Table 3. Percent (± SE) composition of seven monoterpenes occurring in papershell pinyon pines used by porcupines and pine engraver beetles on

the Kickapoo Caverns State Park, southwestern Texas, USA, 1999. Values represent average of 20 transects (3 pairs of trees/transect).

		Porcupines		Beetles ^a					
Monoterpene %	Target	Nontarget	P^{b}	None	Attacked	Colonized	P°		
α-pinene	86.15 <u>+</u> 1.83	82.63 <u>+</u> 2.16	0.1187	80.14 ± 2.28	86.30 <u>+</u> 3.03	88.9 <u>+</u> 1.96	0.1201		
Camphene d	0.39 ± 0.06	0.56 <u>+</u> 0.10	0.3864	0.37 ± 0.07 A	0.38 ± 0.10 A	0.67 <u>+</u> 0.28B	0.0223		
Sabinene	0.60 <u>+</u> 0.21	0.79 <u>+</u> 0.25	0.4483	0.62 <u>+</u> 0.17A	1.65 <u>+</u> 0.60B	0.15 ± 0.08 A	0.0218		
β-pinene	5.04 <u>+</u> 0.91	6.65 <u>+</u> 0.97	0.1298	7.64 <u>+</u> 1.06	4.8 <u>+</u> 1.27	4.11 <u>+</u> 1.09	0.2111		
Myrcene	0.63 <u>+</u> 0.11	1.08 <u>+</u> 0.14	0.0084	1.03 <u>+</u> 0.15	0.86 <u>+</u> 0.19	0.61 <u>+</u> 0.14	0.4413		
Limonene	6.79 <u>+</u> 0.98	7.89 <u>+</u> 1.11	0.2957	9.82 <u>+</u> 1.25B	5.26 <u>+</u> 1.40A	5.38 ± 0.93 A	0.0247		
Terpinolene	0.37 <u>+</u> 0.11	0.44 <u>+</u> 0.11	0.6043	0.34 ± 0.09 A	0.91 <u>+</u> 0.28B	0.17 <u>+</u> 0.05A	0.0138		

^a Means for each monoterpene with the same letter are not different at $\alpha = 0.05$ level of significance.

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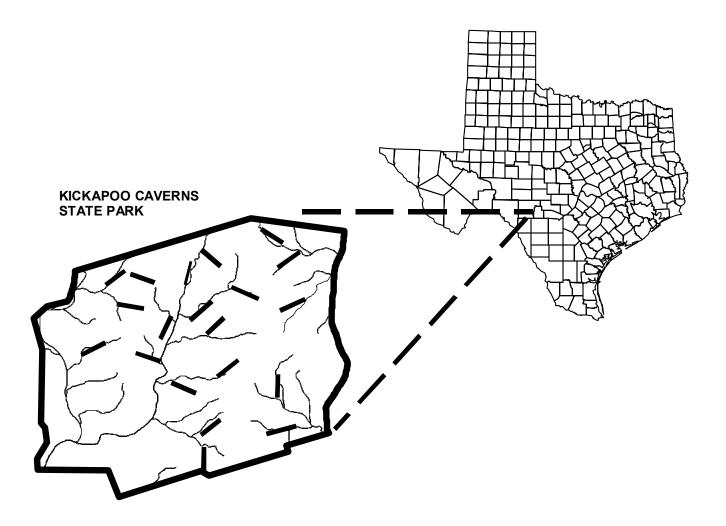
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^b Main effect of porcupine damage (present or absent).

^c Main effect of level of beetle activity.

d See text for discussion of interactive effects.

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5	Figure Captions
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7	Fig. 1. Location of study site, Kickapoo Caverns State Park, in southwestern Texas, USA with
8	solid bars delineating vegetation transects.
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10	Fig. 2. Association of bark beetle activity on target (porcupine-damaged) and nontarget (non-
11	damaged) pinyon pines on Kickapoo Caverns State , in southwestern Texas, USA. $N = no$ beetle
12	activity; A = beetle attack only, C = beetle colonization.
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☐ Porcupine-damaged Ѿ Undamaged

