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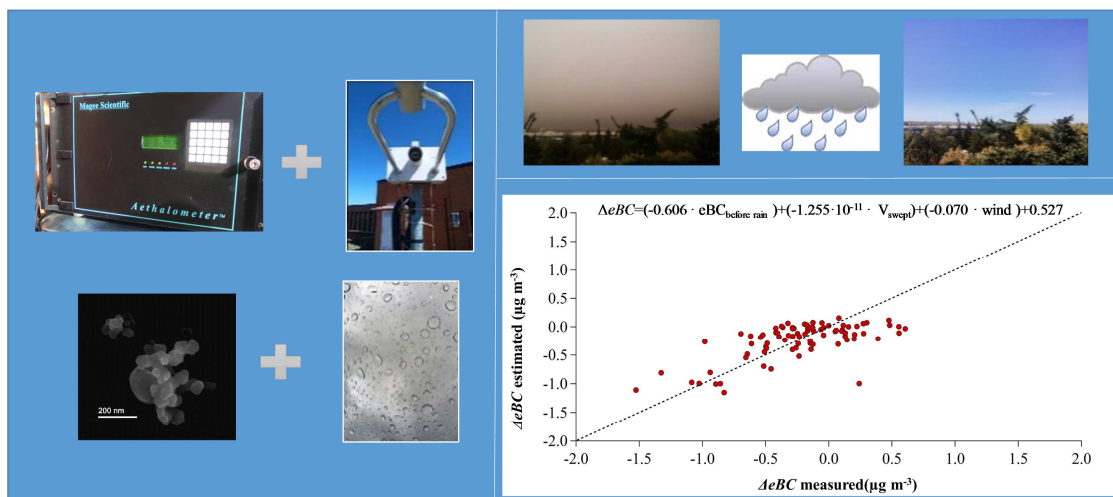
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ACCEPTED MANUSCRIPT

Quantification of source specific black carbon scavenging using an aethalometer and a disdrometer

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ABSTRACT

Aerosol black carbon (BC) is the second strongest contributor to global warming, after CO₂, and it is linked to many adverse health effects. A sampling campaign of 15 months was carried out in León (Spain) in order to evaluate the scavenging of BC with an ensemble aethalometer-disdrometer. The aethalometer provides the concentration of equivalent black carbon (eBC), and the disdrometer, the raindrop size distribution. A total of seventy-five rain events were studied and in 73% of them there was an effective (eBC_{initial} > eBC_{final}) scavenging, with a mean decrease of 48 ± 37% in long rain events (>8 h) and 39 ± 38% in short rain events. The scavenging of BC is strongly related to its source. Thus, the scavenging coefficient (SC) mean value of the BC from fossil fuel (eBC_{ff}) for short and long rain events was 5.1 10⁻⁵ and 1.3 10⁻⁵ s⁻¹, respectively. For the BC from biomass burning (eBC_{bb}), the SC values were 1.6 10⁻⁴ and 2.8 10⁻⁵ s⁻¹ in short and long events, respectively. There was a significant positive correlation between the SC and the number of drops with diameters between 0.375 and 2.5 mm. Rain scavenging of eBC was analyzed depending on the air mass origin obtaining an effective scavenging for air masses from Atlantic, Arctic and Africa. A linear model (R²=0.72) was built to estimate the ΔeBC values with variables from an aethalometer, a disdrometer and a weather station: eBC concentration before rain, swept volume and precipitation accumulated. A Kolmogorov-Smirnov statistical test confirmed the goodness of fit of the model to the measured data.

KEYWORDS: Black carbon, BC scavenging estimation model, raindrop diameter, rainfall, scavenging coefficient, wet deposition.

CAPSULE:

The combination of aethalometer and disdrometer measurements reveals that the scavenging of BC depends strongly on the BC source (biomass burning or fossil fuel combustion).

1. INTRODUCTION

Atmospheric aerosols, both natural and anthropogenic, affect human health (Apte et al., 2015; Fröhlich-Nowoisky et al., 2016; HEI Review Panel, 2013; Pöschl, 2005; WHO, 2007) and climatic change because of the significant contribution to the Earth's radiation budget either directly and indirectly (Andreae and Ramanathan, 2013; ICCP, 2014; Menon et al., 2002; Pöschl, 2005). In Europe, more than 90% of city dwellers are exposed to PM_{2.5} levels that exceed the reference value set by WHO (EEA, 2013).

One of the major atmospheric aerosol pollutants is black carbon (BC) which is emitted during incomplete combustion of fossil fuel or biomass. It is a carbonaceous material formed primarily in flames and directly emitted to the atmosphere, with some particular physical properties: it strongly absorbs visible light and is refractory with a vaporization temperature of around 3700 °C (Bond et al., 2013). It is noteworthy that BC particles are insoluble in water and organic solvents and, therefore, they are not scavenged directly from the atmosphere due to wet deposition (AMAP, 2011). Some studies (e.g. Granat et al., 2010) indicate that BC keeps its hydrophobic property even after being in the atmosphere for several days. However, the BC morphology (organized as fractal-like aggregates) facilitates sorption of other species (Petzold et al., 2013) and it becomes hydrophilic and accessible for wet deposition. The BC cycle is controlled by emissions, transport and deposition. Among these factors, wet deposition is the most complex (Mori et al., 2014).

According to Bond et al., (2013), the main sources of BC are: i) diesel engines used for transport, ii) residential solid fuels (wood and coal), iii) forest fires, and iv) industrial processes. In these burning processes, small carbon spherules are formed, with diameters between 10 and 50 nm and, subsequently, accumulated in aggregates. This aggregates formation starts at or shortly after emission and the aggregates get internally or externally mixed. Their size distribution depends on the formation mechanism and atmospheric processes during transport, while they grow via coagulation. Regarding to aggregate sizes, these particles belong to the Aitken mode (30-100 nm) but due to subsequent coagulation and condensation of inorganic and organic secondary molecules, they can grow to sizes in the accumulation mode (100-1000 nm) (Conrady et al., 2013).

BC also plays an important role in the environment, for example in the formation of acid precipitation through the catalytic oxidation of sulfur dioxide to sulfate (Novakov, 1984; Singh et al., 2016). The main BC aerosol effect on climate is due to the strong ability to absorb solar radiation. The impact of BC on climate change remains largely uncertain (Hienola et al., 2013). Nevertheless, the global mean radiative forcing caused by BC was estimated to be from 0.4 to 1.2 W m⁻², becoming the second strongest contributor to global warming, after CO₂ (Bond et al., 2013; ICCP, 2014). Several studies (AMAP, 2011; Righi et al., 2011) have found that BC affects the Earth's radiation budget in three different ways: i) aerosol direct effect (absorption or scattering of shortwave radiation), ii) aerosol indirect effect (interaction with clouds) and iii) semi-direct effects (BC deposition to ice/snow enhances the absorption of shortwave radiation inducing melting process).

BC may be a pollutant toxic to human health, linked to many illnesses: respiratory (such as adverse effects on lung function and increase cancer risks) and cardiovascular diseases (Janssen et al., 2011; WHO, 2012), and it causes an increase in population morbidity and mortality (Silverman et al., 2012; Suglia et al., 2008), mainly affecting the children's health (UNICEF, 2016) (due to the immature host defense system) and people with chronic respiratory diseases

47 (Jansen et al., 2005). The set of BC and organic carbon is estimated that producing annually
48 around 3 million premature deaths (Apte et al., 2015; Bond et al., 2013; Lelieveld et al., 2015;
49 WHO, 2012). Hence, the study of black carbon concentration is crucial due to its effects on
50 multiple essential policy objectives like climate, air quality or public health (EEA, 2016; Font
51 and Fuller, 2016; Kinney, 2008; Tong et al., 2017, 2016).

52 To measure BC, the use of aethalometer (Hansen et al., 1984) has been common in last years.
53 The aethalometer provides the concentration of equivalent black carbon (eBC) (carbon mass
54 derived from the light attenuation coefficient). The multi-wavelength aethalometer data may be
55 used to separate the fossil fuel (eBC_{ff}) and the biomass burning (eBC_{bb}) contributions to eBC
56 through the “aethalometer model” (Becerril-Valle et al., 2017; Harrison et al., 2013; Sandradewi
57 et al., 2008b; Zotter et al., 2017).

58 Wet and dry deposition is the only important sink of BC, due to its stability. The atmospheric
59 lifetime of BC aerosols ranges from days to weeks depending on the local meteorology (Begam
60 et al., 2016). Dry deposition is able to eliminate larger particles in several days, and sub-micron
61 fraction in several weeks, hence this is a slow process. However, observations have shown that
62 wet deposition represents 70–85% of the tropospheric sink for the carbonaceous aerosols (Pöschl,
63 2005). Therefore, wet deposition is the main process to mitigate the effects of BC on the climate,
64 human health and ecosystems (Cerqueira et al., 2010) in a brief lapse of time, but this process is
65 not yet well explored. Wet deposition is considered one of the most uncertain processes in
66 models (Textor et al., 2006).

67 Below-cloud scavenging (BCS) linked to wet deposition, constitutes an important sink of
68 aerosol particles, including BC (Chate, 2005; Latha et al., 2005; Sportisse, 2007; Tost et al.,
69 2006; Zhao et al., 2015). However, the study of BC aerosol-precipitation interaction does not
70 constitute an easy task. The complexity in the characterization of BCS lies in the dependence on
71 several parameters, for example concentration, chemical composition or electric charges
72 (Ladino et al., 2011). Many researchers have tried to quantify the wet scavenging effect on
73 aerosols and gases (Chate et al., 2003; Laakso et al., 2003; Maria and Russell, 2005; Olszowski,
74 2015), but there are few studies on BC scavenging by rainfall to date. Furthermore, most of
75 them are based on the determination of elemental carbon (EC) concentration by thermo-optical
76 methods (Armalis, 1999; Budhavant et al., 2016; Cerqueira et al., 2010; Custódio et al., 2014;
77 Granat et al., 2010). These studies are carried out in time intervals (sampling period of one day
78 or several hours) that are long compared to the rain event duration. A simultaneous and
79 continuous study of BC concentration and raindrop sizes would provide a deeper knowledge
80 about the BC-precipitation interaction.

81 A 15 months sampling campaign for measuring BC concentration (with an aethalometer)
82 and raindrop physical characteristics (with a disdrometer) has been carried out in León (NW
83 Spain). An estimation of below-cloud scavenging on BC according to rainfall characteristics
84 and origin of air masses has also been accomplished. As far as we know, this type of studies is
85 unprecedented.

86

87 2. MATERIAL AND METHODS

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89 2.1. Sampling site and measurements

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91 The sampling site is located in the campus of the University of León (Fig. S1), a city situated
92 in the NW of the Iberian Peninsula (42° 36' N, 05° 35' W) at 838 m above sea level and with a
93 population of about 200,000 including the metropolitan area. According to Oduber et al. (2018)
94 the main source of particulate emissions is traffic and domestic heating, due to the absence of

95 large emitting industries. It is worth noting that, nowadays, the use of coal combustion in heating
96 devices is still usual in León. León features a Mediterranean climate with continental features and
97 tempered by the proximity of the Cantabrian Mountain Range with an annual mean precipitation
98 of 515 mm (Castro et al., 2010).

99 The sampling was carried out between 12 February 2016 and 31 August 2017 (except spring
100 2017, due to technical issues). The instruments were located on the terrace of the Faculty of
101 Veterinary Medicine of the University of León.

102 The following season distribution along the year was considered: winter from 15 December
103 to 14 March, spring from 15 March to 14 June, summer from 15 June to 14 September and
104 autumn from 15 September to 14 December.

106 2.2. Black carbon data

107
108 An aethalometer model AE-31 (Magee Scientific, USA) with a time resolution of 2 minutes
109 and a precision of $10^{-3} \mu\text{g m}^{-3}$ was used for BC concentration determination. It continuously
110 measures attenuation of light due to the deposition of ambient aerosol on the instrument's filter
111 tape at seven wavelengths: 370, 470, 520, 590, 660, 880 and 950 nm. The sample flow rate was
112 set to be 4 L min^{-1} and verified with Gilibrator measurements. To determine the eBC
113 concentration in the sampled air, the aethalometer uses a differential radiometric optical
114 transmission technique. The instrument operation was described in detail by Hansen (2005) and
115 Virkkula et al. (2007).

116 In order to avoid possible bias in measurements due to the fast changes in relative humidity,
117 the aethalometer data recorded during rain was not taken into account: only data before and
118 after rain were considered. To decrease uncertainties derived from detector response,
119 instrumental noise, flow rate and filter spot area (Corrigan et al., 2006), the 2-min eBC data has
120 been averaged at a resolution of 1 h, comparable to the rain events duration.

121 The contribution from fossil fuel (eBC_{ff}) and biomass burning (eBC_{bb}) was estimated through
122 the application of the aethalometer model (Sandradewi et al., 2008a). Likewise, hourly eBC,
123 eBC_{ff}, eBC_{bb} concentrations and Absorption Ångström Exponent (AAE) were determined. Light
124 absorption measurements at $\lambda_1=470 \text{ nm}$ and $\lambda_2=950 \text{ nm}$ (Becerril-Valle et al., 2017; Harrison et
125 al., 2013; Sandradewi et al., 2008b) have been used in this work
126 ($\text{AAE}(\lambda_1, \lambda_2) = -\frac{\ln(b_{\text{abs}}(\lambda_1)/b_{\text{abs}}(\lambda_2))}{\ln(\lambda_1/\lambda_2)}$). To obtain the AAE values, the shorter wavelength of
127 470 nm has been used rather than the 370 nm one, because the latter is influenced by the varying
128 presence of secondary organic aerosol (SOA) with highly variable optical properties (Zotter et
129 al., 2017). The source specific AAE values used in the aethalometer model to estimate the
130 biomass burning and fossil fuel contributions are $\text{AAE}_{\text{bb}}=1.68$, according to Zotter et al. (2017)
131 and $\text{AAE}_{\text{ff}}=0.95$ derived from measurements in a traffic hotspot in the León center in May
132 during morning rush hours. The mean temperatures were high, so traffic can be considered like
133 the only source at this point. In Supplementary material, the Aethalometer model equations are
134 shown.

135 The AAE values below 0.7 and above 4 (less than 2% along the sampling) were eliminated
136 from the database because these measurements could be affected by instrumental noise, detector
137 response or meteorological conditions (Corrigan et al., 2006).

138 The eBC data recorded during the sampling period were treated following Aerosol, Clouds,
139 and Trace gases Research InfraStructure Network (ACTRIS) guidelines (Virkkula et al., 2007).

141 2.3. Disdrometer and meteorological data

142

143 The raindrop size spectrum has been obtained using a disdrometer, Laser Precipitation
144 Monitor (LPM) of *Thies Clima*, which registered drops between 0.125 and 8 mm in 21 drop
145 size ranges. A detailed description can be found in Fernández-Raga et al. (2009). From the data
146 provided by the LPM, the following rainfall variables were obtained every minute: precipitation
147 intensity, accumulated precipitation, number of drops in 21 channels, volume swept by falling
148 drops ($\text{mm}^{-3} \text{m}^{-3}$), mean and standard deviation of raindrop sizes. The 1-min data have been
149 averaged at a resolution of 1 h, like eBC data.

150 Next to disdrometer, a weather station was installed for continuously registering the
151 temperature, humidity, wind intensity and direction. Univariate analysis (i.e. mean, median,
152 minimum, maximum, quartiles and standard deviation) was used to calculate hourly BC
153 concentration, rainfall variables and meteorological data.

154

155 2.4. Air mass trajectories

156

157 Through the application of HYSPLIT4 (Hybrid Single Particle Lagrangian Integrated
158 Trajectory) (Draxler and Rolph, 2012) the air masses trajectories were analyzed in order to
159 determine sources and origin of the air masses present in León during rain events. This model
160 was used to compute four-days back trajectories at every rain event, at arrival altitudes of 500
161 and 1000 m a.g.l. over León (Custódio et al., 2014). The average altitude of 1000 m is
162 representative of the diurnal mixing layer thickness at León (Calles et al., 2018).

163 The model was run with meteorological data from the Global Data Assimilation System
164 (GDAS) archives (http://ready.arl.noaa.gov/HYSPLIT_traj.php).

165 Based on back trajectories at 1000 m a.g.l. over sampling point, a six groups classification of
166 the prevailing air mass origin and transport pathway have been made. The group assigned
167 coincides with the sector on which the air mass spent most of the time. The regions (Fig. S2)
168 are:

169

- 169 – Group I: Arctic
- 170 – Group II: Atlantic
- 171 – Group III: Continental
- 172 – Group IV: North America
- 173 – Group V: North Atlantic
- 174 – Group VI: Saharan

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176 Besides, a Circulation Weather Types (CWTs) classification was carried out based on Lamb
177 (1972), to identify the type of weather related with a specific synoptic situation. This method
178 has been previously used in the Iberian Peninsula (Calvo et al., 2012; Russo et al., 2014). The
179 direction and vorticity of the geostrophic flow, obtained for 16 grid points distributed over the
180 Iberian Peninsula (Trigo and DaCamara, 2000), have been used to establish each of the 26
181 different CWTs. Eight weather types are identified as “pure” and are characterized by a specific
182 predominant wind component, regardless of their intensity: N, S, E, W, NW, SW, SE and NE.
183 Other two of them are the so-called “non-directional”: anticyclonic (A) and cyclonic (C). As a
184 result of the combination of “non-directional” with “pure” types, other 16 CWTs, so-called
185 “hybrid” types, are obtained. A detailed explication of this classification may be found in Trigo
186 and DaCamara (2000).

187

188 2.5. Selecting data criteria

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The rain events considered in this study are those that fulfill the following conditions:

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a) an accumulated precipitation greater than 0.2 mm was registered in one hour (minimum rain amount considered significant),

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b) there was a minimum of 2 h without rain between events,

194

c) there were eBC and eBC_{ff} concentrations greater than 0.5 µg m⁻³ and greater than 0.1 µg m⁻³ for eBC_{bb},

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d) rain duration was shorter than 24 h.

197

e) a maximum difference of 2 m s⁻¹ in wind speed and 50° in wind direction during rain event has been registered (Kyrö et al., 2009; Paramonov et al., 2011) to avoid changes in eBC concentration due to advection (Laakso et al., 2003).

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Furthermore, in order to minimize the eventual interference of coal combustion in the eBC_{bb} estimated by the Aethalometer model, the events in which coal combustion tracers (As and Se) (Vejahati et al., 2010; Wang et al., 2018) were registered, have been removed from eBC_{bb} analysis. Levoglucosan (biomass burning tracer (Gonçalves et al., 2011)) or ¹⁴C measurements (as in Zotter et al. 2017) were not available.

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For each rain event, the eBC, eBC_{ff} and eBC_{bb} concentration one hour before and one hour after precipitation have been analyzed. In order to determine the eBC concentration change between time intervals t_1 and t_2 , with eBC concentrations c_1 and c_2 , respectively, the following parameter was calculated:

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$$\Delta eBC_{rel} = -100 \frac{\Delta eBC}{c_1} \quad (1)$$

213

where $\Delta eBC = c_2 - c_1$ and the minus sign before 100 has been introduced in order to get a positive value of ΔeBC_{rel} when the eBC concentration decreases (effective scavenging). Furthermore, for a group of events, a new parameter has been defined as the concentration-weighted average ($\Delta eBC\% = \frac{\sum(\Delta eBC_{rel} \cdot c_1)}{\sum c_1}$).

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A global analysis of all the events that meet the previously cited requirements was carried out. The average length of rain events in León was 3:28 h with a standard deviation of 4:07 h. The episodes of extreme duration could be referred to as “long events”. A possible quantification may be the 10 % of the total events or the average plus one standard deviation. Both criteria conducted to the same threshold of about 8 hours. Consequently, events exceeding 8 h were called long events, and the rest short events.

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The equation used to estimate eBC scavenging coefficient, was the same often used to calculate scavenging coefficient (SC) from the concentration change c of aerosol particles and other elements like sulphates (Chate et al., 2003; Laakso et al., 2003; Maria and Russell, 2005; Olszowski, 2015). In this study, c is the eBC concentration. The scavenging coefficient SC (s⁻¹) is defined as a rate of aerosols washout by precipitation.

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$$SC = -\frac{1}{c} \frac{dc}{dt} \quad (2)$$

232

233

The integration of this equation between t_1 and t_2 with concentrations c_1 and c_2 gives:

234

234

$$SC = -\frac{1}{t_1 - t_2} \ln \frac{c_2}{c_1} \quad (3)$$

236

237 The scavenging coefficient SC varies with collection efficiency, particles and raindrop size
 238 distributions and their terminal velocities, for different particle sizes (Seinfeld and Pandis, 2006).
 239 Although the scavenging coefficient is usually called λ , it is represented in this study by SC , as λ
 240 refers here to wavelength.

241 When speaking about an effective scavenging, we will refer to the positive values of $\Delta eBC\%$
 242 and SC . This decrease of the BC concentration can also be influenced by vertical mixing
 243 changes or even by advection.

244

245 2.6. Diurnal pattern normalization

246

247 We eliminated the important effect of the diurnal cycle on the eBC, eBC_{ff} and eBC_{bb}
 248 concentrations by normalizing the daily concentrations to the daily average. For no-rain days
 249 during the sampling period, daily patterns of eBC, eBC_{ff} and eBC_{bb} hourly concentration were
 250 obtained. The eBC concentration of all days without rain between March 2016 and August 2017
 251 were used and the seasonal daily patterns were calculated in the following way: first of all, for
 252 each day, the ratio between the hourly eBC concentration and the mean eBC concentration of
 253 that day was calculated. With the daily normalized values, the average diurnal pattern was
 254 determined. In other words, we have calculated the hourly ratio ($R_{i,j}$) of the diurnal pattern for
 255 the hour i of the day j through the following formula:

$$R_{i,j} = \frac{c_{i,j}}{\bar{c}_j} \quad (4)$$

257 where $\bar{c}_j = \frac{\sum_i c_{i,j}}{n}$, n being the number of hours in one day (24). We have subsequently
 258 determined the average R_i mean for each season:

259

$$R_i = \frac{\sum_j R_{i,j}}{d} \quad (5)$$

260

261 d being the number of complete sampling days without rain for each season (22, 52, 81, 70, 81
 262 and 62 days for winter 2016, spring 2016, summer 2016, autumn 2016, winter 2017 and
 263 summer 2017, respectively).

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3. RESULTS AND DISCUSSION

268

269 The seasonal eBC values in León, during sampling are presented in Table S1. eBC_{ff} values
 270 registered during autumn 2016 and winter 2017 are higher probably due to the increase in road
 271 traffic and the use of heating devices. In all seasons, there were two peaks throughout the day
 272 (0600-0800 and 1600-2000 UTC), mainly during rush hours' traffic (Fig. S3). In the afternoons,
 273 heating emissions add to the afternoon peak. Likewise, eBC_{bb} values in winter and autumn were
 274 higher, because of the use of biomass for heating. Other cities like Beijing, Leicester, Hefei or
 275 Kadapa, have shown the same pattern (Begam et al., 2016; Cheng et al., 2014; Hama et al.,

276 2017; Zhang et al., 2015, 2017) caused by traffic and domestic heating.

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279 3.1. Rain effect on equivalent black carbon concentration

280

281 Seventy-five rain events were observed during the sampling period. The rain events are
282 concentrated in winter and spring, showing a clear decrease in summer (Fig. S4).

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285 3.1.1. Rain events characteristics

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287 The main characteristics of short and long events are shown in Table 1. Seven long events
288 were observed during the sampling period. There was an effective eBC scavenging in 6 events
289 (86% of total), with a mean decrease of $48 \pm 37\%$. Long events featured the mean raindrop
290 diameter of 0.34 ± 0.19 mm, with a mean swept volume per event of $1.9 \cdot 10^{10} \text{ mm}^3 \text{ m}^{-3}$ and a
291 mean number of drops of $8.8 \cdot 10^7 \text{ m}^{-2}$. Of the 75 events registered, 68 rain events were short.
292 There was an effective eBC scavenging in 49 of the short events (73% of total), with a mean
293 decrease of $39 \pm 38\%$. The mean raindrop diameter for short events was 0.34 ± 0.20 mm, with a
294 mean swept volume per event of $2.5 \cdot 10^9 \text{ mm}^3 \text{ m}^{-3}$ and a mean number of drops of $9.0 \cdot 10^6 \text{ m}^{-2}$.

295 The decrease in eBC_{bb} concentration in both short and long events is significantly higher
296 compared to eBC_{ff}. This fact is probably due to the higher fraction of organics from biomass
297 burning, and also a higher degree of oxygenation (O:C) of biomass burning organic aerosol
298 compared to traffic organic aerosol, thus increasing its hydrophilic property (Cerully et al.,
299 2011; Safai et al., 2014; Zheng et al., 2017). In addition, the size including the organics for
300 wood burning being larger than for traffic emissions may play a role (Blanco-Alegre et al.,
301 2018). Following the criteria given in section 2.5 (more specifically, the criterium c), the events
302 considered to determine the eBC, eBC_{ff} and eBC_{bb} values were different. Therefore, the eBC
303 value may not lie between the eBC_{ff} and eBC_{bb} specific ones.

304

305 Table 1. Number of rain events (*N*), percentage of rain events with effective scavenging (% of events) and mean
306 decrease (and standard deviation) in BC concentration ($\Delta eBC\%$).

	Short events (< 8 h)			Long events (> 8 h)		
	<i>N</i>	% of events	$\Delta eBC\%$	<i>N</i>	% of events	$\Delta eBC\%$
eBC	68	73	39 ± 38	7	86	48 ± 37
eBC _{ff}	60	75	40 ± 33	7	71	46 ± 30
eBC _{bb}	27	85	75 ± 46	3	67	96 ± 5

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309 3.1.2. Extreme scavenging events

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311 According to the scavenging efficiency, a classification of rain events in three groups has
312 been carried out: type I (ΔeBC_{rel} higher than 50%), type II (ΔeBC_{rel} between 0-50%) and type III
313 (non-effective scavenging) (Table 2). It can be seen that events with more effective scavenging
314 were characterized by: i) a longer duration of events, ii) a greater precipitation accumulated and
315 iii) a higher wind intensity. The events with no scavenging effect featured somewhat lower eBC
316 concentration before rain. However, there were no major differences in the rainfall parameters
317 analyzed between types II and III.

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Table 2. Mean and standard deviation values of rain parameters for the three types of events established according to the scavenging efficiencies.

	N	Duration (h:min)	Rain (mm)	Rainfall intensity (mm h ⁻¹)	Wind speed (m s ⁻¹)	Raindrop diameter (mm)	eBC before rain (µg m ⁻³)	Swept volume (mm ³ m ⁻³)
Type I ($\Delta eBC_{rel} > 50\%$)	18	4:24 ± 3:30	8.2 ± 12.5	1.2 ± 1.3	1.9 ± 1.3	0.31 ± 0.09	1.1 ± 0.7	9.6 10 ⁹
Type II ($\Delta eBC_{rel} 0-50\%$)	39	2:47 ± 2:32	2.4 ± 3.4	0.7 ± 0.8	1.1 ± 1.3	0.36 ± 0.09	1.1 ± 0.5	3.0 10 ⁹
Type III ($\Delta eBC_{rel} < 0\%$)	15	2:36 ± 2:08	1.8 ± 3.0	0.5 ± 0.4	1.2 ± 1.0	0.35 ± 0.06	0.9 ± 0.4	2.9 10 ⁹

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322 The event with the highest scavenging efficiency occurred on 21st March 2017 with an
323 accumulated rainfall of 3.05 mm and a mean rainfall intensity of 0.57 mm h⁻¹. The maximum
324 intensity was 0.05 mm min⁻¹. The rain started at 0800 UTC and lasted 228 min. The eBC
325 concentration was 2.62 µg m⁻³ and 0.22 µg m⁻³ two hours before and two hours after rain,
326 respectively ($\Delta eBC_{rel} = 88\%$). The mean raindrop size was 0.44 ± 0.22 mm and the mode of
327 raindrop size was 0.31 mm. The swept volume was 6.2 10⁹ mm³ m⁻³.

328 On the other hand, the event with the highest eBC percentage increase occurred on 24th
329 February 2016 with an accumulated rainfall of 1.92 mm and a mean rainfall intensity of 3.96
330 mm h⁻¹. The maximum intensity was 0.06 mm min⁻¹. A eBC increase was registered, with 1.33
331 µg m⁻³ and 1.98 µg m⁻³ two hours before and two hours after rain, respectively
332 ($\Delta eBC_{rel} = -34\%$). The rain started at 1149 UTC and lasted 173 min. The mean raindrop size
333 was 0.32 ± 0.19 mm and the mode of raindrop sizes was 0.27 mm. The swept volume was
334 2.8 10⁹ mm³ m⁻³. An explanation could be the change from a deep mixing layer to a shallow
335 one, as proposed by Talukdar et al. (2015) for Calcutta. Thus, the change in surface and air
336 temperature reduced the mixing layer height and the eBC concentration increased.

337 Besides, another contribution to the differences in scavenging efficiency may be the higher
338 swept volume recorded in the event with effective scavenging, caused by a longer event
339 duration and a higher raindrop diameter.

340

341 3.2. Air masses origin and weather types

342

343 The four-days back trajectories of air masses arriving in León at 1000 m a.g.l. were obtained
344 from HYSPLIT model. In Fig. S5 the seasonal back trajectories for rain events are shown.

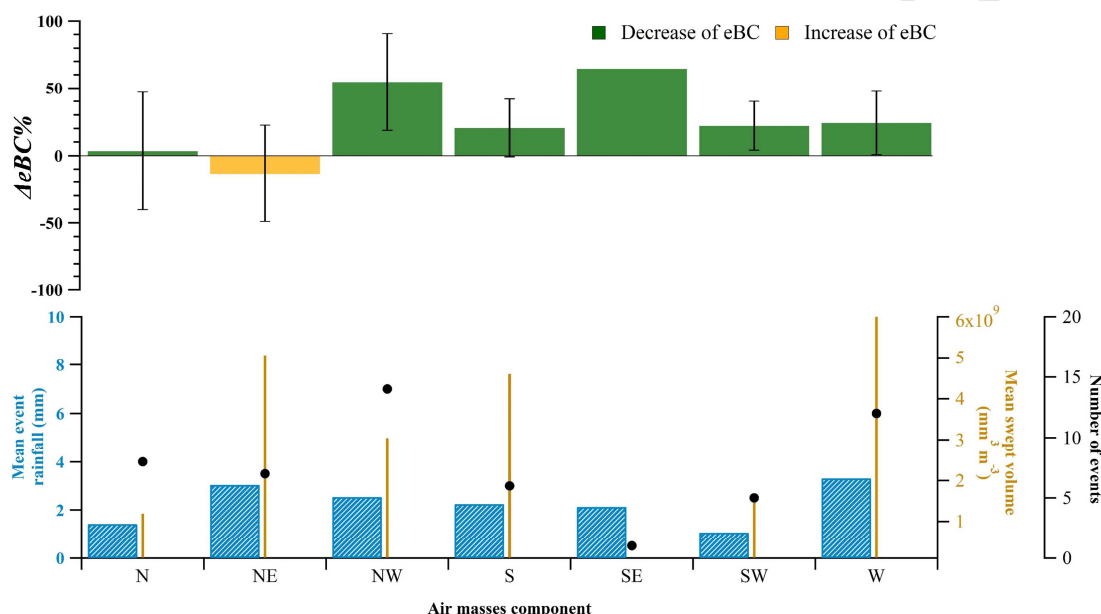
345 The main origins of air masses during rain events were Atlantic (33%), Continental (20%)
346 and North Atlantic (19%) (Table S2). In winter 2017, there was a prevalence of air mass
347 coming from North Atlantic (56%), and in summer 2017 there was a high frequency of air
348 masses with Saharan origin (56%). The highest values of eBC concentrations were recorded in
349 air masses coming from North America (during winter 2017) and in intrusions from Sahara
350 (during summer 2017). No significant statistical differences were observed in eBC
351 concentration between air masses during rainy days.

352 The air masses presented the following effective ($eBC_{initial} > eBC_{final}$) scavenging values:
353 Arctic (29 ± 21%), North Atlantic (35 ± 31%) and Saharan (18 ± 21%). However, Continental
354 (6 ± 39%), North America (6 ± 7%) and Atlantic (4 ± 39%) masses present a lower mean eBC
355 decrease (Table S3), maybe due to the high eBC load transported by these air masses.

356 The most frequent weather types during rain events were C (23 cases) and NW (9 cases).
357 There was a decrease in eBC concentration in all the weather types except for NE and AS. NE

358 type is noteworthy, since it comprises many events. The two AS events were characterized by a
 359 stagnation of the air mass coming from Sahara. On the other hand, during NW rain events (9
 360 cases), a $\Delta eBC\%$ of 45% was registered. Other weather types with W component, like W, SW,
 361 ANW, CW and AW present a clear decrease (Table S3). Figure 1 shows the $\Delta eBC\%$ regarding
 362 the origin of air (according to Weather Lamb type). There was an effective scavenging in all
 363 components except NE, indicative of an eBC source in these directions. This increase coincides
 364 with that observed in Continental masses based on HYSPLIT trajectories.

365 During rain events of pure anticyclonic type, there was an increase in eBC concentration of
 366 27%, and a mean raindrop diameter of 0.30 ± 0.09 mm. On the contrary, when the type was
 367 pure cyclonic, a decrease in eBC concentration of 1% was recorded, with a mean raindrop
 368 diameter of 0.36 ± 0.07 mm.
 369



370
 371 Fig. 1. $\Delta eBC\%$ between before and after rain according to air mass origin of Circulation Weather Types (CWTs)
 372 during rain events. Black dots indicate the number of rain events, striped boxes indicate the mean rainfall (mm) and
 373 vertical lines indicate the mean swept ($\text{mm}^3 \text{m}^{-3}$) per event.
 374

375 3.3. Scavenging coefficients

376
 377 During the 68 rain short events (with a mean intensity of 0.82 mm h^{-1}), SC for eBC showed a
 378 mean value of $4.7 \cdot 10^{-5} \text{ s}^{-1}$ (with a standard deviation of $8.4 \cdot 10^{-5} \text{ s}^{-1}$) and a median of $4.0 \cdot 10^{-5} \text{ s}^{-1}$
 379 (interquartile range: $-8.8 \cdot 10^{-6}$ to $7.92 \cdot 10^{-5} \text{ s}^{-1}$). Positive values of SC are indicative of effective
 380 scavenging (Table 3).

381 In the 7 rain long events (with a mean intensity of 1.42 mm h^{-1}), SC for eBC indicated a
 382 mean value of $1.0 \cdot 10^{-5} \text{ s}^{-1}$ (with a standard deviation of $1.1 \cdot 10^{-5} \text{ s}^{-1}$) and a median of $1.1 \cdot 10^{-5} \text{ s}^{-1}$
 383 (interquartile range: $3.7 \cdot 10^{-6}$ to $1.9 \cdot 10^{-5} \text{ s}^{-1}$). It should be noted that SC for eBC_{bb} in short events
 384 was around three to four times higher than SC of eBC and eBC_{ff} . The SC for eBC in short
 385 events was about four times the value obtained for long events. However, $\Delta eBC\%$ was greater
 386 in long events (Table 1). Thus, the greater decreased in long events may be caused by the great
 387 long duration of events. To estimate the effect of rain duration on the $\Delta eBC\%$, the correlation
 388 between both variables was analyzed. A correlation coefficient R^2 of 0.14 was obtained, which
 389 means that the rain duration explains a 14% of the total variance of $\Delta eBC\%$.
 390

391 Table 3. Scavenging coefficients (SC) in s^{-1} of eBC, eBC_{ff} and eBC_{bb} in short and long events. Number of events are
 392 in brackets.

	SHORT EVENTS (68)			LONG EVENTS (7)		
	SC (eBC) (68)	SC (eBC _{ff}) (68)	SC (eBC _{bb}) (27)	SC (eBC) (7)	SC (eBC _{ff}) (7)	SC (eBC _{bb}) (3)
Min	$-1.3 \cdot 10^{-4}$	$-1.0 \cdot 10^{-4}$	$-8.0 \cdot 10^{-5}$	$-6.7 \cdot 10^{-6}$	$-6.6 \cdot 10^{-6}$	$-9.5 \cdot 10^{-6}$
Q1	$-8.8 \cdot 10^{-6}$	$-9.2 \cdot 10^{-6}$	$2.2 \cdot 10^{-6}$	$3.7 \cdot 10^{-6}$	$-3.4 \cdot 10^{-5}$	$1.1 \cdot 10^{-5}$
Median	$4.0 \cdot 10^{-5}$	$3.2 \cdot 10^{-5}$	$6.6 \cdot 10^{-5}$	$1.1 \cdot 10^{-5}$	$1.3 \cdot 10^{-5}$	$3.1 \cdot 10^{-5}$
Q3	$7.9 \cdot 10^{-5}$	$9.4 \cdot 10^{-5}$	$2.5 \cdot 10^{-4}$	$1.9 \cdot 10^{-5}$	$2.9 \cdot 10^{-5}$	$4.7 \cdot 10^{-5}$
Max	$2.5 \cdot 10^{-4}$	$2.8 \cdot 10^{-4}$	$1.1 \cdot 10^{-3}$	$2.3 \cdot 10^{-5}$	$3.4 \cdot 10^{-5}$	$6.4 \cdot 10^{-5}$
Mean	$4.7 \cdot 10^{-5}$	$5.1 \cdot 10^{-5}$	$1.6 \cdot 10^{-4}$	$1.0 \cdot 10^{-5}$	$1.3 \cdot 10^{-5}$	$2.8 \cdot 10^{-5}$
Desvest	$8.4 \cdot 10^{-5}$	$8.5 \cdot 10^{-5}$	$2.6 \cdot 10^{-4}$	$1.1 \cdot 10^{-5}$	$1.8 \cdot 10^{-5}$	$3.7 \cdot 10^{-5}$

393
 394 Current information on BC scavenging ratios available in the literature is still scarce.
 395 Nevertheless, the values obtained were similar to those indicated by Latha et al. (2005)
 396 ($1.64 \cdot 10^{-5} s^{-1}$) in India, although their value was obtained using a different method. The
 397 observed differences could be due to the lower BC concentrations in León (8 times lower than
 398 in its location, Hyderabad and Secunderabad, India).

399 A global analysis of rain events gives positive values of SC for eBC (indicative of an
 400 effective scavenging), higher for short events. Furthermore, these SC values are in the same
 401 order of magnitude that those obtained in other studies about fine and ultrafine aerosols (Laakso
 402 et al., 2003; Zikova and Zdimal, 2016).

403
 404
 405

3.4. Relation between SC and meteorological parameters

406 For each rain event, the SC value has been related to the wind intensity and the rainfall
 407 intensity in order to observe the influence of these meteorological variables on eBC
 408 concentration. In Fig. 2, it can be observed that most of the events presented positive values
 409 (effective scavenging). The rain events with intensities greater than 2 mm h^{-1} were scarce but
 410 they always presented an effective eBC scavenging. Events with rain intensities less than
 411 2 mm h^{-1} caused an effective eBC scavenging in 70% of cases.

412 It should be noted that there was no clear influence of the wind speed during rain, although in
 413 the events with no effective scavenging, low wind speed dominated. Nevertheless, the wind and
 414 the eBC concentration two hours before rain presented a statistically significant negative
 415 correlation. The wind causes a higher dispersion of eBC, hindering the scavenging by rain.

416 There was a statistically significant negative correlation between event duration and ΔeBC_{rel} .
 417 This supports the influence of rain duration over eBC concentration (see Fig. S6). Other studies
 418 have also shown that low intensities and large duration of rain events produce a higher effective
 419 scavenging (Chatterjee et al., 2012). Besides, vertical mixing and horizontal advection could also
 420 be related with ΔeBC (Joshi et al., 2016), mainly in autumn and winter.

421
 422

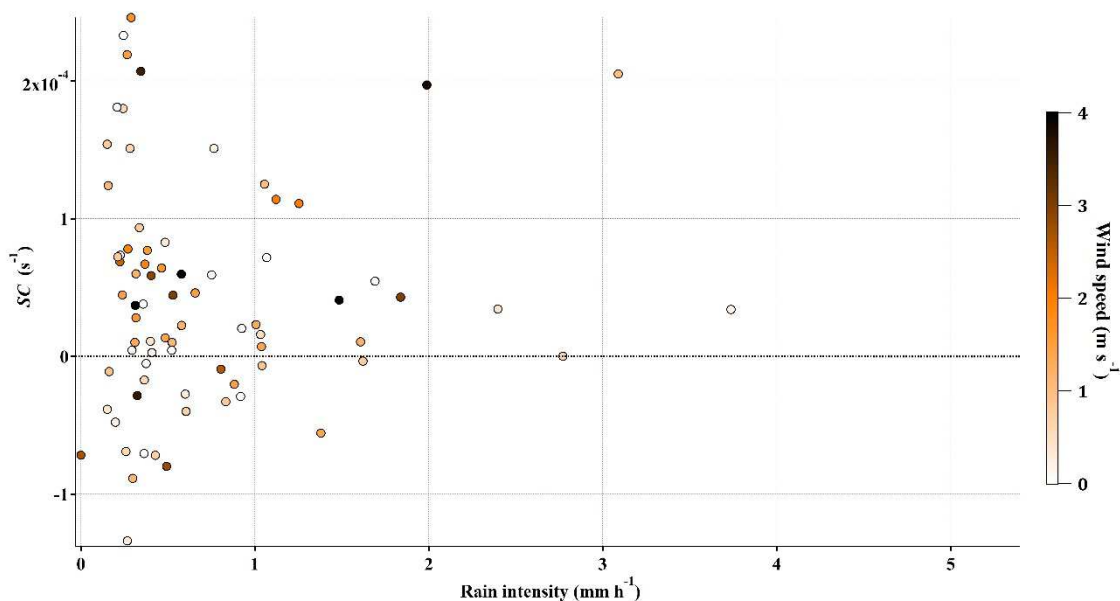


Fig. 2. Relationship between black carbon scavenging coefficients (SC) and rainfall intensity. Wind speed is indicated by the colour scale inside the dots.

3.5. Model for ΔeBC

To parameterize the ΔeBC , a linear model has been built using aethalometer and disdrometer variables. What variables can be the most appropriate to build this model? Previously, the correlations between ΔeBC and variables like wind speed, eBC concentration before rain, event duration, precipitation accumulated, mean rainfall intensity, mean raindrop diameter and sum of volume swept by falling drops have already been analyzed. Furthermore, the raindrop size could also be connected to ΔeBC . In fact, the data obtained from 21 drop size ranges of LPM (between 0.125 and 8 mm) have been used to check the relationship between the number of raindrops in each size channel and the SC in events with efficient scavenging (Table S4). As a result, a significant correlation between the SC and the number of drops with diameters between 0.375 and 2.5 mm has been found. The rest of the raindrop diameters did not present significant correlations with SC .

Consequently, the variables tested to build the model have been the following: wind speed, eBC concentration before rain, event duration, precipitation accumulated, mean rainfall intensity, mean raindrop diameter, change in the mixing layer height, sum of volume swept by falling drops and number of drops with diameters between 0.375 and 2.5 mm. Besides, we have included two variables related to the mixing layer height that could affect the rain scavenging: the mean height during rain and the change in height throughout the rain event. The application of an automatic linear modelling (IBM SPSS Statistics 24) by stepwise, with an entry probability of 0.05 and removal probability of 0.10 has been used.

With this methodology, a model has been built from a random sample including the 75% of the total data set. This model has been applied to the remaining 25%. Subsequently, a Kolmogorov-Smirnov statistical test has been carried out in order to check the goodness of fit of the model to the measured data. This process has been repeated ten times and the results are shown in Table S5. The significant values obtained ($\alpha > 0.05$) shows that null hypothesis is confirmed, measured and predicted data are related and, consequently, the model created may be enforceable.

All ten repetitions of the model include the following variables: eBC concentration before

456 rain, swept volume and precipitation accumulated. Besides, two models also include the mean
457 raindrop diameter. This last variable shows the greater variability.

458 Finally, a multi-linear regression model has been established ($R^2=0.72$) based on the whole
459 data set. As expected, the model includes the three aforementioned variables, as follows:
460

$$\Delta eBC = (k_1 \cdot eBC_{\text{before rain}}) + (k_2 \cdot \text{Precipitation accumulated}) + (k_3 \cdot V_{\text{swept}}) + k_4$$

461

462 The coefficients presented the following values (\pm standard deviations): $k_1 = -0.557$ (\pm
463 0.063), $k_2 = -0.0741$ (± 0.007) $\mu\text{g m}^{-3} \text{mm}^{-1}$, $k_3 = -3.37 \cdot 10^{-11}$ ($\pm 8.12 \cdot 10^{-12}$) $\mu\text{g mm}^{-3}$ and $k_4 =$
464 0.210 (± 0.083) $\mu\text{g m}^{-3}$.

465 Furthermore, a multi-linear regression model has been established for ΔeBC_{ff} ($R^2=0.81$) and
466 ΔeBC_{bb} ($R^2=0.88$).

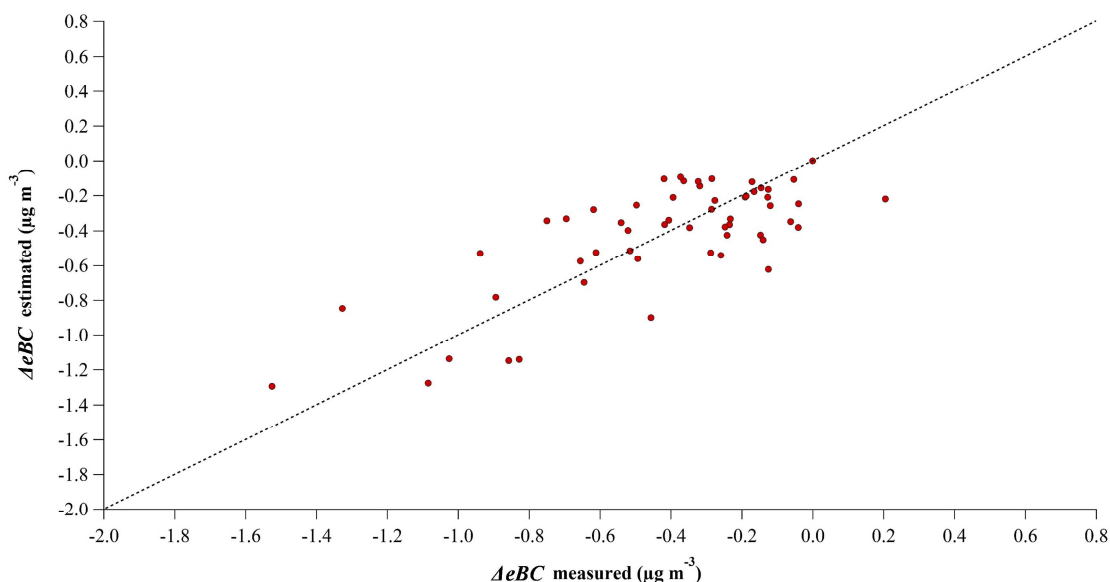
467

$$\begin{aligned} \Delta eBC_{ff} &= (k_{1f} \cdot eBC_{\text{before rain}}) + (k_{2f} \cdot \text{Precipitation accumulated}) + (k_{3f} \cdot V_{\text{swept}}) \\ &\quad + (k_{4f} \cdot \text{duration}) + k_{5f} \\ \Delta eBC_{bb} &= (k_{1b} \cdot eBC_{\text{before rain}}) + (k_{2b} \cdot \phi_{\text{raindrop}}) + k_{3b} \end{aligned}$$

468

469 The coefficients presented the following values for ΔeBC_{ff} (\pm standard deviations):
470 $k_{1f} = -0.598$ (± 0.052), $k_{2f} = 0.044$ (± 0.007) $\mu\text{g m}^{-3} \text{mm}^{-1}$, $k_{3f} = -6.44 \cdot 10^{-11}$ ($\pm 1.59 \cdot 10^{-11}$)
471 $\mu\text{g mm}^{-3}$, $k_{4f} = 0.056$ (± 0.02) $\mu\text{g h}^{-1} \text{m}^{-3}$ and $k_{5f} = 0.153$ (± 0.068) $\mu\text{g m}^{-3}$. For ΔeBC_{bb} the values
472 were: $k_{1b} = -0.980$ (± 0.070), $k_{2b} = 0.924$ (± 0.379) $\mu\text{g m}^{-3} \text{mm}^{-1}$ and $k_{3b} = -0.194$ (± 0.145)
473 $\mu\text{g m}^{-3}$. The ΔeBC_{bb} depends more on eBC concentration before rain than ΔeBC . The ΔeBC_{ff}
474 model includes similar variables to ΔeBC model, also incorporating the duration variable.

475 The model estimates the reduction in eBC well compared to the measured data (Fig. 3)
476 becoming a valuable tool to predict the eBC behavior during rain events. Analogous
477 representations for eBC_{ff} (Fig. S7) and eBC_{bb} (Fig. S8) are shown in Supplementary material.
478



479

480

Fig. 3. ΔeBC estimated by the model vs ΔeBC measured. The dashed line ($y=x$) shows the perfect estimation.

481

482

4. CONCLUSIONS

483

484

We investigated BC scavenging during rain events in an urban background environment,
485 with these main conclusions:

485

- 486 1. In 73% of rain events there was an effective scavenging, with a mean eBC decrease
487 $48 \pm 37\%$ in long rain events (>8 h) and $39 \pm 38\%$ in short rain events.
- 488 2. Rain scavenging was analyzed depending on air mass origin. The air masses with a clear
489 efficient scavenging came from Arctic ($\Delta eBC\% = 29 \pm 21\%$), North Atlantic ($35 \pm 31\%$)
490 and Saharan ($18 \pm 21\%$) areas.
- 491 3. Concerning the BC sources, the scavenging of eBC_{bb} is significantly higher compared to
492 eBC_{ff} , probably due to the increase of its hydrophilicity because of the higher fraction of
493 organics and a higher degree of oxygenation of biomass burning organic aerosol
494 compared to traffic organic aerosol. Concretely, the scavenging coefficient (SC) mean
495 value of eBC_{ff} for short and long rain events was $5.1 \cdot 10^{-5}$ and $1.3 \cdot 10^{-5} \text{ s}^{-1}$, respectively.
496 For eBC_{bb} , the SC values were $1.6 \cdot 10^{-4}$ and $2.8 \cdot 10^{-5} \text{ s}^{-1}$ in short and long events,
497 respectively.
- 498 4. Events with intensities higher than 2 mm h^{-1} (the maximum intensity registered in the
499 sampling campaign was 7 mm h^{-1}) always presented an effective scavenging on eBC.
500 However, the highest values of SC were recorded for rain events characterized by a low
501 rainfall intensity and long duration. Furthermore, there was a significant positive
502 correlation between the SC and the number of drops with diameters between 0.375 and 2.5
503 μm .
- 504 5. A linear model ($R^2=0.72$) was built to estimate the ΔeBC values with variables from
505 weather station, aethalometer and disdrometer: eBC concentration before rain, swept
506 volume and precipitation accumulated. A Kolmogorov-Smirnov statistical test has
507 confirmed the goodness of fit of the model to the measured data. Finally, two other
508 models have also been proposed to estimate the ΔeBC values from biomass burning and
509 fossil fuel combustion.

510

511 One of the most important outcomes of the project is the finding that the scavenging of BC
512 depends on the BC source (biomass burning and fossil fuel combustion). This fact could be a key
513 factor to include in the climate models to account both the aerosol-cloud interaction and the
514 radiative forcing caused by clouds and BC.

515 The combination of aethalometer-disdrometer measurements has proved to be a valuable tool
516 for the quantification of the eBC scavenging, allowing identifying the raindrop variables that
517 contribute to an effective scavenging of this pollutant.

518

519

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520

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527

528

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529

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534

535

5. REFERENCES

- 536 Andreae, M., Ramanathan, V., 2013. Climate's dark forcings. *Science* (80-.). 340, 280–281.
- 537 Apte, J.S., Marshall, J.D., Cohen, A.J., Brauer, M., 2015. Addressing Global Mortality from
538 Ambient PM_{2.5}. *Environ. Sci. Technol.* 49, 8057–8066. doi:10.1021/acs.est.5b01236
- 539 Armalis, S., 1999. Wet deposition of elemental carbon in Lithuania. *Sci. Total Environ.* 239,
540 89–93. doi:10.1016/S0048-9697(99)00288-0
- 541 Becerril-Valle, M., Coz, E., Prévôt, A.S.H., Močnik, G., Pandis, S.N., Sánchez de la Campa,
542 A.M., Alastuey, A., Díaz, E., Pérez, R.M., Artíñano, B., 2017. Characterization of
543 atmospheric black carbon and co-pollutants in urban and rural areas of Spain. *Atmos.*
544 *Environ.* 169. doi:10.1016/j.atmosenv.2017.09.014
- 545 Begam, G.R., Vachaspati, C.V., Ahammed, Y.N., Kumar, K.R., Babu, S.S., Reddy, R.R., 2016.
546 Measurement and analysis of black carbon aerosols over a tropical semi-arid station in
547 Kadapa, India. *Atmos. Res.* 171, 77–91. doi:10.1016/j.atmosres.2015.12.014
- 548 Bond, T.C., Doherty, S.J., Fahey, D.W., Forster, P.M., Berntsen, T., Deangelo, B.J., Flanner,
549 M.G., Ghan, S., Kürcher, B., Koch, D., Kinne, S., Kondo, Y., Quinn, P.K., Sarofim, M.C.,
550 Schultz, M.G., Schulz, M., Venkataraman, C., Zhang, H., Zhang, S., Bellouin, N.,
551 Guttikunda, S.K., Hopke, P.K., Jacobson, M.Z., Kaiser, J.W., Klimont, Z., Lohmann, U.,
552 Schwarz, J.P., Shindell, D., Storelvmo, T., Warren, S.G., Zender, C.S., 2013. Bounding
553 the role of black carbon in the climate system: A scientific assessment. *J. Geophys. Res.*
554 *Atmos.* 118, 5380–5552. doi:10.1002/jgrd.50171
- 555 Budhavant, K.B., Rao, P.S.P., Safai, P.D., Leck, C., Rodhe, H., 2016. Black carbon in cloud-
556 water and rain water during monsoon season at a high altitude station in India. *Atmos.*
557 *Environ.* 129, 256–264. doi:10.1016/j.atmosenv.2016.01.028
- 558 Calvo, A.I., Pont, V., Olmo, F.J., Castro, A., Alados-Arboledas, L., Vicente, A.M., Fernández-
559 Raga, M., Fraile, R., 2012. Air masses and weather types: A useful tool for characterizing
560 precipitation chemistry and wet deposition. *Aerosol Air Qual. Res.* 12, 856–878.
561 doi:10.4209/aaqr.2012.03.0068
- 562 Castro, A., Alonso-Blanco, E., González-Colino, M., Calvo, A.I., Fernández-Raga, M., Fraile,
563 R., 2010. Aerosol size distribution in precipitation events in León, Spain. *Atmos. Res.* 96,
564 421–435. doi:10.1016/j.atmosres.2010.01.014
- 565 Cerqueira, M., Pio, C., Legrand, M., Puxbaum, H., Kasper-Giebl, A., Afonso, J., Preunkert, S.,
566 Gelencsér, A., Fialho, P., 2010. Particulate carbon in precipitation at European
567 background sites. *J. Aerosol Sci.* 41, 51–61. doi:10.1016/j.jaerosci.2009.08.002
- 568 Cerully, K.M., Raatikainen, T., Lance, S., Tkacik, D., Tiitta, P., Petäjä, T., Ehn, M., Kulmala,
569 M., Worsnop, D.R., Laaksonen, A., Smith, J.N., Nenes, A., 2011. Aerosol hygroscopicity
570 and CCN activation kinetics in a boreal forest environment during the 2007 EUCAARI
571 campaign. *Atmos. Chem. Phys.* 11, 12369–12386. doi:10.5194/acp-11-12369-2011
- 572 Chate, D.M., 2005. Study of scavenging of submicron-sized aerosol particles by thunderstorm
573 rain events. *Atmos. Environ.* 39, 6608–6619. doi:10.1016/j.atmosenv.2005.07.063
- 574 Chate, D.M., Rao, P.S.P., Naik, M.S., Momin, G.A., Safai, P.D., Ali, K., 2003. Scavenging of
575 aerosols and their chemical species by rain. *Atmos. Environ.* 37, 2477–2484.
576 doi:10.1016/S1352-2310(03)00162-6
- 577 Chatterjee, A., Das, S., Ghosh, S., Raha, S., 2012. Wet scavenging of black carbon and sulphate
578 depends on the nature of the rain; effect on the climate and global change, in: EGU
579 General Assembly Conference Abstracts. p. 8392.
- 580 Cheng, Y.H., Liao, C.W., Liu, Z.S., Tsai, C.J., Hsi, H.C., 2014. A size-segregation method for
581 monitoring the diurnal characteristics of atmospheric black carbon size distribution at
582 urban traffic sites. *Atmos. Environ.* 90, 78–86. doi:10.1016/j.atmosenv.2014.03.023
- 583 Conrady, K., Heinke Schlünzen, K., Herber, A., 2013. Atmospheric concentration of black
584 carbon in the western Arctic. Universität Hamburg.
- 585 Corrigan, C.E., Ramanathan, V., Schauer, J.J., 2006. Impact of monsoon transitions on the
586 physical and optical properties of aerosols. *J. Geophys. Res. Atmos.* 111, 1–15.
587 doi:10.1029/2005JD006370
- 588 Custódio, D., Cerqueira, M., Fialho, P., Nunes, T., Pio, C., Henriques, D., 2014. Wet deposition

- 589 of particulate carbon to the Central North Atlantic Ocean. *Sci. Total Environ.* 496, 92–99.
590 doi:10.1016/j.scitotenv.2014.06.103
- 591 Draxler, R., Rolph, G., 2012. HYSPLIT (Hybrid Single-Particle Lagrangian Integrated
592 Trajectory). Silver Spring, MD NOAA Air Resour. Lab.
- 593 EEA, 2016. Air quality in Europe - 2016 Report. doi:10.2800/413142
- 594 EEA, 2013. Air Quality in Europe - 2013 Report. Denmark. doi:10.2800/92843
- 595 Fernández-Raga, M., Castro, A., Palencia, C., Calvo, A.I., Fraile, R., 2009. Rain events on 22
596 October 2006 in León (Spain): Drop size spectra. *Atmos. Res.* 93, 619–635.
597 doi:10.1016/j.atmosres.2008.09.035
- 598 Font, A., Fuller, G.W., 2016. Did policies to abate atmospheric emissions from traffic have a
599 positive effect in London? *Environ. Pollut.* 218, 463–474.
600 doi:10.1016/j.envpol.2016.07.026
- 601 Fröhlich-Nowoisky, J., Kampf, C.J., Weber, B., Huffman, J.A., Pöhlker, C., Andreae, M.O.,
602 Lang-Yona, N., Burrows, S.M., Gunthe, S.S., Elbert, W., Su, H., Hoor, P., Thines, E.,
603 Hoffmann, T., Després, V.R., Pöschl, U., 2016. Bioaerosols in the Earth system: Climate,
604 health, and ecosystem interactions. *Atmos. Res.* 182, 346–376.
605 doi:10.1016/j.atmosres.2016.07.018
- 606 Gonçalves, C., Alves, C., Fernandes, A.P., Monteiro, C., Tarelho, L., Evtugina, M., Pio, C.,
607 2011. Organic compounds in PM_{2.5} emitted from fireplace and woodstove combustion of
608 typical Portuguese wood species. *Atmos. Environ.* 45, 4533–4545.
609 doi:10.1016/j.atmosenv.2011.05.071
- 610 Granat, L., Engström, J.E., Praveen, S., Rodhe, H., 2010. Light absorbing material (soot) in
611 rainwater and in aerosol particles in the Maldives. *J. Geophys. Res. Atmos.* 115, 1–12.
612 doi:10.1029/2009JD013768
- 613 Hama, S.M.L., Cordell, R.L., Kos, G.P.A., Weijers, E.P., Monks, P.S., 2017. Sub-micron
614 particle number size distribution characteristics at two urban locations in Leicester.
615 *Atmos. Res.* 194, 1–16. doi:10.1016/j.atmosres.2017.04.021
- 616 Hansen, A.D.A., 2005. Aethalometer Magee Scientific.
- 617 Hansen, A.D.A., Rosen, H., Novakov, T., 1984. The aethalometer - An instrument for the real-
618 time measurement of optical absorption by aerosol particles. *Sci. Total Environ.* 36, 191–
619 196. doi:10.1016/0048-9697(84)90265-1
- 620 Harrison, R.M., Beddows, D.C.S., Jones, A.M., Calvo, A., Alves, C., Pio, C., 2013. An
621 evaluation of some issues regarding the use of aethalometers to measure woodsmoke
622 concentrations. *Atmos. Environ.* 80, 540–548. doi:10.1016/j.atmosenv.2013.08.026
- 623 HEI Review Panel, 2013. Understanding the Health Effects of Ambient Ultrafine Particles.
624 *Heal. Eff. Inst.* 122.
- 625 Hienola, A.I., Pietikäinen, J.P., Jacob, D., Pozdun, R., Petäjä, T., Hyvärinen, A.P., Sogacheva,
626 L., Kerminen, V.M., Kulmala, M., Laaksonen, A., 2013. Black carbon concentration and
627 deposition estimations in Finland by the regional aerosol-climate model REMO-HAM.
628 *Atmos. Chem. Phys.* 13, 4033–4055. doi:10.5194/acp-13-4033-2013
- 629 ICCP, 2014. Climate Change 2014 Impacts, Adaptation, and Vulnerability Part B: Regional
630 Aspects, *Igarss 2014*. doi:10.1007/s13398-014-0173-7.2
- 631 Jansen, K.L., Larson, T. V., Koenig, J.Q., Mar, T.F., Fields, C., Stewart, J., Lippmann, M.,
632 2005. Associations between health effects and particulate matter and black carbon in
633 subjects with respiratory disease. *Environ. Health Perspect.* 113, 1741–1746.
634 doi:10.1289/ehp.8153
- 635 Janssen, N.A.H., Hoek, G., Simic-Lawson, M., Fischer, P., van Bree, L., ten Brink, H., Keuken,
636 M., Atkinson, R.W., Anderson, H.R., Brunekreef, B., Cassee, F.R., 2011. Black Carbon as
637 an Additional Indicator of the Adverse Health Effects of Airborne Particles Compared
638 with PM₁₀ and PM_{2.5}. *Environ. Health Perspect.* 119, 1691–1699.
639 doi:10.1289/ehp.1003369
- 640 Kinney, P.L., 2008. Climate Change, Air Quality, and Human Health. *Am. J. Prev. Med.* 35,
641 459–467. doi:10.1016/j.amepre.2008.08.025
- 642 Kyrö, E.M., Grönholm, T., Vuollekoski, H., Virkkula, A., Kulmala, M., Laakso, L., 2009. Snow
643 scavenging of ultrafine particles: Field measurements and parameterization. *Boreal*

- 644 Environ. Res. 14, 527–538.
- 645 Laakso, L., Grönholm, T., Rannik, Ü., Kosmale, M., Fiedler, V., Vehkamäki, H., Kulmala, M.,
646 2003. Ultrafine particle scavenging coefficients calculated from 6 years field
647 measurements. *Atmos. Environ.* 37, 3605–3613. doi:10.1016/S1352-2310(03)00326-1
- 648 Ladino, L., Stetzer, O., Hattendorf, B., Günther, D., Croft, B., Lohmann, U., 2011.
649 Experimental Study of Collection Efficiencies between Submicron Aerosols and Cloud
650 Droplets. *J. Atmos. Sci.* 68, 1853–1864. doi:10.1175/JAS-D-11-012.1
- 651 Lamb, H.H., 1972. British Isles Weather Types and a Register of the Daily Sequence of
652 Circulation Patterns 1861-1971, Geophysical memoirs. H.M. Stationery Office.
- 653 Latha, K., Badarinath, K.V.S., Manikya Reddy, P., 2005. Scavenging efficiency of rainfall on
654 black carbon aerosols over an urban environment. *Atmos. Sci. Lett.* 6, 148–151.
655 doi:10.1002/asl.108
- 656 Lelieveld, J., Evans, J.S., Fnais, M., Giannadaki, D., Pozzer, A., 2015. The contribution of
657 outdoor air pollution sources to premature mortality on a global scale. *Nature* 525, 367–
658 371. doi:10.1038/nature15371
- 659 Maria, S.F., Russell, L.M., 2005. Organic and inorganic aerosol below-cloud scavenging by
660 suburban New Jersey precipitation. *Environ. Sci. Technol.* 39, 4793–4800.
661 doi:10.1021/es0491679
- 662 Menon, S., Hansen, J., Nazarenko, L., Luo, U., 2002. Climate Effects of Black Carbon Aerosols
663 in China and India. *Science* (80-.). 297, 2250–2253. doi:10.1126/science.1075159
- 664 Mori, T., Kondo, Y., Ohata, S., Moteki, N., Matsui, H., Oshima, N., Iwasaki, A., 2014. Wet
665 deposition of black carbon at a remote site in the East China Sea 1–25.
666 doi:10.1002/2013JD020968.Received
- 667 Novakov, T., 1984. The role of soot and primary oxidants in atmospheric chemistry. *Sci. Total*
668 *Environ.* 36, 1–10. doi:10.1016/0048-9697(84)90241-9
- 669 Oduber, F., Castro, A., Calvo, A.I., Blanco-Alegre, C., Alonso-Blanco, E., Belmonte, P., Fraile,
670 R., 2018. Summer-autumn air pollution in León, Spain: changes in aerosol size
671 distribution and expected effects on the respiratory tract. *Air Qual. Atmos. Heal.*
672 doi:10.1007/s11869-018-0556-6
- 673 Olszowski, T., 2015. Concentration Changes Of PM10 Under Liquid Precipitation Conditions.
674 *Ecol. Chem. Eng. S* 22, 363–378. doi:10.1515/eces-2015-0019
- 675 Paramonov, M., Grönholm, T., Virkkula, A., 2011. Below-cloud scavenging of aerosol particles
676 by snow at an urban site in Finland. *Boreal Environ. Res.* 16, 304–320.
677 doi:10.1017/CBO9781107415324.004
- 678 Petzold, A., Ogren, J.A., Fiebig, M., Laj, P., Li, S.M., Baltensperger, U., Holzer-Popp, T.,
679 Kinne, S., Pappalardo, G., Sugimoto, N., Wehrli, C., Wiedensohler, A., Zhang, X.Y.,
680 2013. Recommendations for reporting black carbon measurements. *Atmos. Chem. Phys.*
681 13, 8365–8379. doi:10.5194/acp-13-8365-2013
- 682 Pöschl, U., 2005. Atmospheric aerosols: Composition, transformation, climate and health
683 effects. *Angew. Chemie - Int. Ed.* 44, 7520–7540. doi:10.1002/anie.200501122
- 684 Quinn, P., Stohl, A., Arneth, A., Berntsen, T., 2011. AMAP, 2011. The impact of black carbon
685 on Arctic climate., Arctic Monitoring and Assessment Programme (AMAP), Oslo.
- 686 Righi, M., Klinger, C., Eyring, V., Hendricks, J., Lauer, A., Petzold, A., 2011. Climate impact
687 of biofuels in shipping: Global model studies of the aerosol indirect effect. *Environ. Sci.*
688 *Technol.* 45, 3519–3525. doi:10.1021/es1036157
- 689 Russo, A., Trigo, R.M., Martins, H., Mendes, M.T., 2014. NO₂, PM₁₀ and O₃ urban
690 concentrations and its association with circulation weather types in Portugal. *Atmos.*
691 *Environ.* 89, 768–785. doi:10.1016/j.atmosenv.2014.02.010
- 692 Safai, P.D., Raju, M.P., Rao, P.S.P., Pandithurai, G., 2014. Characterization of carbonaceous
693 aerosols over the urban tropical location and a new approach to evaluate their climatic
694 importance. *Atmos. Environ.* 92, 493–500. doi:10.1016/j.atmosenv.2014.04.055
- 695 Sandradewi, J., Prévôt, A.S.H., Szidat, S., Perron, N., Alfarra, M.R., Lanz, V.A., Weingartner,
696 E., Baltensperger, U.R.S., 2008a. Using aerosol light absorption measurements for the
697 quantitative determination of wood burning and traffic emission contribution to particulate
698 matter. *Environ. Sci. Technol.* 42, 3316–3323. doi:10.1021/es702253m

- 699 Sandradewi, J., Prévôt, A.S.H., Weingartner, E., Schmidhauser, R., Gysel, M., Baltensperger,
700 U., 2008b. A study of wood burning and traffic aerosols in an Alpine valley using a multi-
701 wavelength Aethalometer. *Atmos. Environ.* 42, 101–112.
702 doi:10.1016/j.atmosenv.2007.09.034
- 703 Seinfeld, J.H., Pandis, S., 2006. Atmospheric from Air Pollution to Climate Change. *Atmos.*
704 *Chem. Phys.* 51, 1248. doi:10.1016/0016-7037(87)90252-3
- 705 Silverman, D.T., Samanic, C.M., Lubin, J.H., Blair, A.E., Stewart, P.A., Vermeulen, R., Coble,
706 J.B., Rothman, N., Schleiff, P.L., Travis, W.D., Ziegler, R.G., Wacholder, S., Attfield,
707 M.D., 2012. The Diesel Exhaust in Miners Study: A Nested Case-Control Study of Lung
708 Cancer and Diesel Exhaust. *JNCI J. Natl. Cancer Inst.* 104, 855–868.
709 doi:10.1093/jnci/djs034
- 710 Singh, S., Elumalai, S.P., Pal, A.K., 2016. Rain pH estimation based on the particulate matter
711 pollutants and wet deposition study. *Sci. Total Environ.* 563–564, 293–301.
712 doi:10.1016/j.scitotenv.2016.04.066
- 713 Sportisse, B., 2007. A review of parameterizations for modelling dry deposition and scavenging
714 of radionuclides. *Atmos. Environ.* 41, 2683–2698. doi:10.1016/j.atmosenv.2006.11.057
- 715 Suglia, S.F., Gryparis, A., Schwartz, J., Wright, R.J., 2008. Association between traffic-related
716 black carbon exposure and lung function among urban women. *Environ. Health Perspect.*
717 116, 1333–1337. doi:10.1289/ehp.11223
- 718 Talukdar, S., Jana, S., Maitra, A., Gogoi, M.M., 2015. Characteristics of black carbon
719 concentration at a metropolitan city located near land-ocean boundary in Eastern India.
720 *Atmos. Res.* 153, 526–534. doi:10.1016/j.atmosres.2014.10.014
- 721 Textor, C., Schulz, M., Guibert, S., Kinne, S., Balkanski, Y., Bauer, S., Berntsen, T., Berglen,
722 T., Boucher, O., Chin, M., Dentener, F., Diehl, T., Easter, R., Feichter, H., Fillmore, D.,
723 Ghan, S., Ginoux, P., Gong, S., Grini, a., Hendricks, J., Horowitz, L., Huang, P., Isaksen,
724 I., Iversen, T., Kloster, S., Koch, D., Kirkevåg, a., Kristjansson, J.E., Krol, M., Lauer, a.,
725 Lamarque, J.F., Liu, X., Montanaro, V., Myhre, G., Penner, J., Pitari, G., Lamarque, J.F.,
726 Liu, X., Montanaro, V., Myhre, G., Penner, J., Pitari, G., Reddy, S., Seland, Ø., Stier, P.,
727 Takemura, T., Tie, X., 2006. Analysis and quantification of the diversities of aerosol life
728 cycles within AeroCom. *Atmos. Chem. Phys.* 6, 1777–1813. doi:10.5194/acpd-5-8331-
729 2005
- 730 Tong, Z., Chen, Y., Malkawi, A., Adamkiewicz, G., Spengler, J.D., 2016. Quantifying the
731 impact of traffic-related air pollution on the indoor air quality of a naturally ventilated
732 building. *Environ. Int.* 89–90, 138–146. doi:10.1016/j.envint.2016.01.016
- 733 Tong, Z., Yang, B., Hopke, P.K., Zhang, K.M., 2017. Microenvironmental air quality impact of
734 a commercial-scale biomass heating system. *Environ. Pollut.* 220, 1112–1120.
735 doi:10.1016/j.envpol.2016.11.025
- 736 Tost, H., Jöckel, P., Kerkweg, A., Sander, R., Lelieveld, J., 2006. Technical note: A new
737 comprehensive scavenging submodel for global atmospheric chemistry modelling. *Atmos.*
738 *Chem. Phys. Discuss.* 5, 11157–11181. doi:10.5194/acpd-5-11157-2005
- 739 Trigo, R.M., DaCamara, C.C., 2000. Circulation weather types y their influence on the
740 precipitation regime in Portugal. *Int. J. Clim.* 20, 1559–1581. doi:10.1002/1097-
741 0088(20001115)20
- 742 UNICEF, 2016. Clear the air for children. The impact of air pollution on children. [WWW
743 Document]. URL https://www.unicef.org/publications/index_92957.html (accessed
744 11.11.16).
- 745 Vejehati, F., Xu, Z., Gupta, R., 2010. Trace elements in coal: Associations with coal and
746 minerals and their behavior during coal utilization - A review. *Fuel* 89, 904–911.
747 doi:10.1016/j.fuel.2009.06.013
- 748 Virkkula, A., Mäkelä, T., Hillamo, R., Yli-Tuomi, T., Hirsikko, A., Hämeri, K., Koponen, I.K.,
749 2007. A Simple Procedure for Correcting Loading Effects of Aethalometer Data. *J. Air*
750 *Waste Manage. Assoc.* 57, 1214–1222. doi:10.3155/1047-3289.57.10.1214
- 751 Wang, C., Liu, H., Zhang, Y., Zou, C., Anthony, E.J., 2018. Review of arsenic behavior during
752 coal combustion: Volatilization, transformation, emission and removal technologies.
753 *Prog. Energy Combust. Sci.* 68, 1–28. doi:10.1016/j.peccs.2018.04.001

- 754 WHO, 2012. Health effects of Black carbon. Copenhagen, Denmark.
- 755 Zhang, X., Rao, R., Huang, Y., Mao, M., Berg, M.J., Sun, W., 2015. Black carbon aerosols in
756 urban central China. *J. Quant. Spectrosc. Radiat. Transf.* 150, 3–11.
757 doi:10.1016/j.jqsrt.2014.03.006
- 758 Zhang, Y., Chen, J., Yang, H., Li, R., Yu, Q., 2017. Seasonal variation and potential source
759 regions of PM_{2.5}-bound PAHs in the megacity Beijing, China: Impact of regional
760 transport. *Environ. Pollut.* 231, 329–338. doi:10.1016/j.envpol.2017.08.025
- 761 Zhao, S., Yu, Y., He, J., Yin, D., Wang, B., 2015. Below-cloud scavenging of aerosol particles
762 by precipitation in a typical valley city, northwestern China. *Atmos. Environ.* 102, 70–78.
763 doi:10.1016/j.atmosenv.2014.11.051
- 764 Zheng, X., Zhang, S., Wu, Y., Zhang, K.M., Wu, X., Li, Z., Hao, J., 2017. Characteristics of
765 black carbon emissions from in-use light-duty passenger vehicles. *Environ. Pollut.* 231,
766 348–356. doi:10.1016/j.envpol.2017.08.002
- 767 Zikova, N., Zdimal, V., 2016. Precipitation scavenging of aerosol particles at a rural site in the
768 Czech Republic. *Tellus B* 68, 1–14. doi:10.3402/tellusb.v68.27343
- 769 Zotter, P., Herich, H., Gysel, M., El-Haddad, I., Zhang, Y., Močnik, G., Hüglin, C.,
770 Baltensperger, U., Szidat, S., Prévôt, A.S.H., 2017. Evaluation of the absorption Ångström
771 exponents for traffic and wood burning in the Aethalometer based source apportionment
772 using radiocarbon measurements of ambient aerosol. *Atmos. Chem. Phys.* 1–29.
773 doi:10.5194/acp-17-4229-2017
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777 **Quantification of source specific black carbon scavenging using an**
778 **aethalometer and a disdrometer**

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792 **SUPPLEMENTARY MATERIAL**

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795 1. Aethalometer model

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797 Assuming that only these two sources exist, the total absorption coefficient $b_{abs,total}(\lambda)$ at
798 wavelength λ is:

$$799 \quad b_{abs,total}(\lambda) = b_{abs,ff}(\lambda) + b_{abs,bb}(\lambda)$$

800
801 where $b_{abs,ff}(\lambda)$, $b_{abs,bb}(\lambda)$ are the absorption coefficients of fossil fuel combustion and biomass
802 burning, respectively.
803

804
805 Source apportionment of fossil fuel has been estimated through the following equations:

$$806 \quad \frac{b_{abs,ff}(\lambda_1)}{b_{abs,ff}(\lambda_2)} = \left(\frac{\lambda_1}{\lambda_2}\right)^{-AAE_{ff}}$$

$$807 \quad eBC_{ff} = \frac{b_{abs,ff}(\lambda_2)}{b_{abs,total}(\lambda_2)} \cdot eBC_{total}(\lambda_2)$$

808
809 where AAE_{ff} is the absorption Ångström exponent for eBC fossil fuel.

810
811 2. Tables

812
813 Table S1. Seasonal mean values of eBC ($\mu\text{g m}^{-3}$), eBC_{ff} ($\mu\text{g m}^{-3}$) and eBC_{bb} ($\mu\text{g m}^{-3}$) measured at León during
814 sampling (January 2016–August 2017). The values next to the mean correspond to the standard deviation.

SEASON	eBC ($\mu\text{g m}^{-3}$)	eBC_{ff} ($\mu\text{g m}^{-3}$)	eBC_{bb} ($\mu\text{g m}^{-3}$)
Winter 2016	1.0 ± 0.9	0.8 ± 0.8	0.2 ± 0.3
Spring 2016	0.6 ± 0.5	0.5 ± 0.4	0.1 ± 0.2
Summer 2016	0.6 ± 0.5	0.6 ± 0.4	0.1 ± 0.2
Autumn 2016	1.2 ± 1.1	1.0 ± 1.0	0.2 ± 0.4
Winter 2017	1.3 ± 1.3	1.0 ± 1.2	0.4 ± 0.5
Summer 2017	0.6 ± 0.5	0.5 ± 0.3	0.1 ± 0.2

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Table S2. Seasonal percentage of air masses origin (four-days before rain events).

Origin	Winter 2016	Spring 2016	Summer 2016	Autumn 2016	Winter 2017	Summer 2017	Total
Arctic	17	11	13	8	0	0	7
Atlantic	17	44	25	50	13	22	35
Continental	0	22	25	21	19	22	19
North America	0	0	0	8	6	0	3
North Atlantic	67	11	0	0	56	0	19
Saharan	0	11	38	13	6	56	17

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820

Table S3. Rainfall characteristics and eBC variation based on Circulation Weather Types (CWTs) and air mass origin with HYSPLIT during rain events measured at León during sampling campaign. The values next to the mean correspond to the standard deviation

	N	Duration (min)	Rain (mm)	Rainfall intensity (mm h ⁻¹)	Wind speed (m s ⁻¹)	Raindrop diameter (mm)	eBC before rain (µg m ⁻³)	Swept volume (mm ³ m ⁻³)	ΔeBC%	
Lamb Weather type	C	23	313	5 ± 6.8	1.52	0.9 ± 0.7	0.4 ± 0.1	0.9 ± 0.4	7.2 10 ⁹	-1
	NW	9	173	3.4 ± 4.6	1.07	2.1 ± 1	0.3 ± 0.1	1.1 ± 0.6	3.7 10 ⁹	-45
	N	8	98	1.4 ± 1.2	0.9	1.2 ± 0.8	0.4 ± 0.1	1.3 ± 0.6	1.4 10 ⁹	-3
	NE	6	180	3.5 ± 4.4	1.36	0.9 ± 0.7	0.3 ± 0.2	0.9 ± 0.3	3.4 10 ⁹	22
	W	7	266	3.6 ± 5.3	0.59	2.1 ± 1.4	0.3 ± 0.1	1.1 ± 0.4	6.5 10 ⁹	-27
	A	5	72	0.3 ± 0.2	0.25	0.4 ± 0.3	0.3 ± 0.1	1.2 ± 0.7	5.6 10 ⁸	27
	SW	4	120	1.2 ± 1.7	0.47	1.4 ± 0.5	0.3 ± 0.1	0.9 ± 0.4	1.5 10 ⁹	-29
	ANW	5	108	0.9 ± 1.2	0.45	1.5 ± 1.9	0.4 ± 0.1	1.5 ± 1	1.7 10 ⁹	-71
	CS	3	140	1.5 ± 0.2	0.92	0.5 ± 0.7	0.4 ± 0.1	1.4 ± 0.9	1.7 10 ⁹	-38
	CW	3	240	3.4 ± 4.3	1.07	2.2 ± 2.1	0.3 ± 0.1	0.8 ± 0.2	4.7 10 ⁹	-26
	AS	2	90	0.8 ± 0.3	0.68	1 ± 0	0.4 ± 0.1	0.6 ± 0.2	1.4 10 ⁹	10
	AW	2	270	2.2 ± 2.6	0.38	0.8 ± 0.3	0.3 ± 0	0.9 ± 0.4	5.2 10 ⁹	-9
	S	1	480	7.4	0.92	0.1	0.3	0.6	1.3 10 ¹⁰	-29
	ANE	1	120	0.5	0.23	0	0.3	1	7.4 10 ⁸	-42
	CSE	1	120	2.1	1.07	0	0.4	1.6	3.1 10 ⁹	-65
	CSW	1	60	0.6	0.6	0.7	0.4	0.5	7.5 10 ⁸	8
	A component	13	120	0.9 ± 1.2	0.37	0.7 ± 1.0	0.35 ± 0.08	1.2 ± 0.8	1.8 10 ⁹	-20
C component	30	282	4.4 ± 6.2	1.38	0.9 ± 1.0	0.36 ± 0.08	1 ± 0.5	6.3 10 ⁹	-8	
HYSPLIT model	Arctic	6	110	2.8 ± 5.9	0.86	1 ± 1.4	0.4 ± 0.1	1.1 ± 0.6	2.2 10 ⁹	-29
	Atlantic	27	180	2.8 ± 3.5	1.18	1.3 ± 1.2	0.3 ± 0.1	1 ± 0.5	4.1 10 ⁹	-4
	Continental	16	281	4 ± 7.5	1.16	0.9 ± 0.6	0.4 ± 0.1	1 ± 0.4	5.8 10 ⁹	-6
	North America	3	120	1.3 ± 0.4	0.94	0.7 ± 0.6	0.4 ± 0	0.9 ± 0.2	1.4 10 ⁹	-6
	North Atlantic	15	212	2.5 ± 3.8	0.55	1.7 ± 1.4	0.3 ± 0.1	1 ± 0.5	3.8 10 ⁹	-35
	Saharan	14	210	3.4 ± 3.9	0.97	0.9 ± 0.8	0.3 ± 0.1	1.3 ± 0.8	4.6 10 ⁹	-18

Table S4. Pearson correlation values between $\Delta eBC\%$ and rain event characteristics.

	Variable	$\Delta eBC\%$
Rain event characteristics	Wind intensity (2 hours before rain)	-0.348
	Accumulated rain	-0.225
	Rain intensity	-0.227
	Event duration	-0.242
	Swept volume	-0.230
Raindrop diameter (mm)	0.125	-0.251
	0.25	-0.223
	0.375	-0.257
	0.5	-0.287
	0.75	-0.284
	1	-0.281
	1.25	-0.278
	1.5	-0.276
	1.75	-0.269
	2	-0.278
	2.5	-0.255
	3	-0.209
	3.5	-0.162
4	-0.124	

*Bold font indicates that the correlation is significant at 95% level.

Table S5. Verification of the model performed (N=10), variables obtained in each N model and Kolmogorov-Smirnov (K-S) statistical test carried out between the predicted and the measured values. In N=0 was represented the model obtained with 100% of data.

N	$eBC_{\text{before rain}}$	V_{swept}	Precipitation accumulated	$\varnothing_{\text{raindrop}}$	Intercept	K-S (α)
1	-0.559	$-3.072 \cdot 10^{-11}$	0.040	-	0.217	0.490
2	-0.559	$-3.802 \cdot 10^{-11}$	0.044	1.126	-0.142	0.538
3	-0.636	$-2.979 \cdot 10^{-11}$	0.036	-	0.244	0.456
4	-0.587	$-3.443 \cdot 10^{-11}$	0.041	-	0.249	0.417
5	-0.639	$-2.913 \cdot 10^{-11}$	0.039	-	0.261	0.522
6	-0.559	$-3.072 \cdot 10^{-11}$	0.040	-	0.217	0.490
7	-0.594	$-4.228 \cdot 10^{-11}$	0.045	-	0.276	0.569
8	-0.481	$-3.198 \cdot 10^{-11}$	0.041	0.844	-0.112	0.621
9	-0.587	$-3.443 \cdot 10^{-11}$	0.041	-	0.249	0.373
10	-0.577	$-3.381 \cdot 10^{-11}$	0.041	-	0.246	0.500
0	-0.557	$-3.370 \cdot 10^{-11}$	0.041	-	0.210	0.468

3. Figures

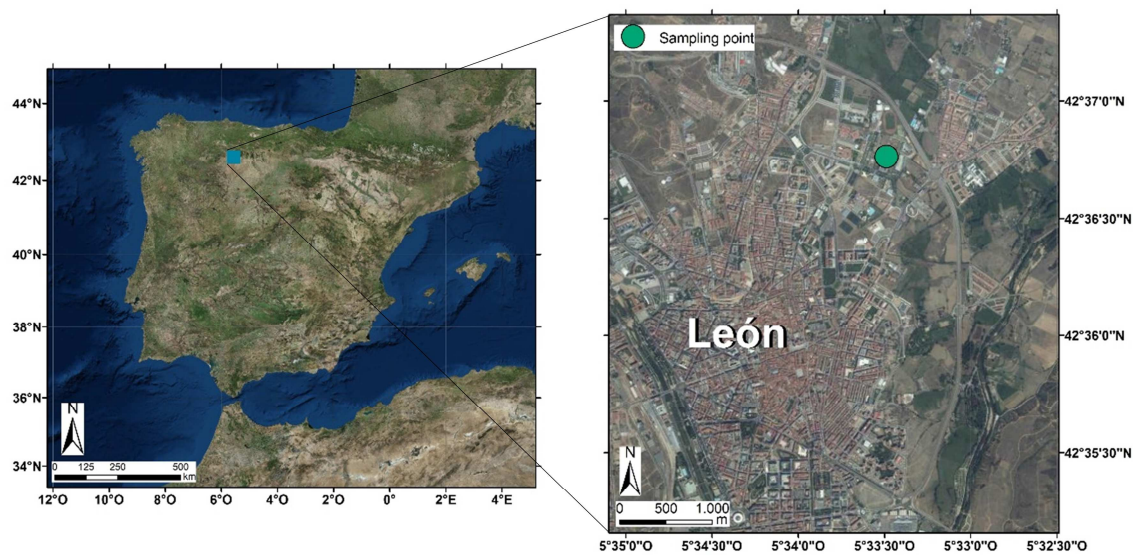


Fig. S1. León city in the NW Iberian Peninsula and the surroundings of the sampling site. Source: Earthstar Geographics, ESRI.

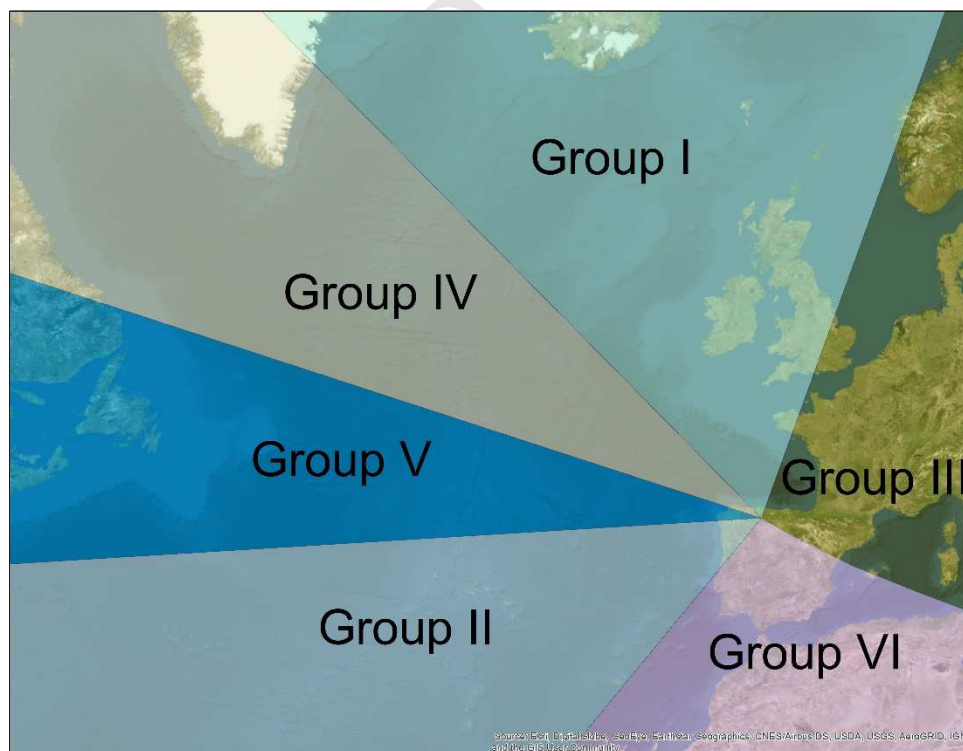


Fig. S2. Delimitation of regions for the determination of the air masses origin: I: Arctic; II: Atlantic; III: Continental; IV: North America; V: North Atlantic, and VI: Saharan.

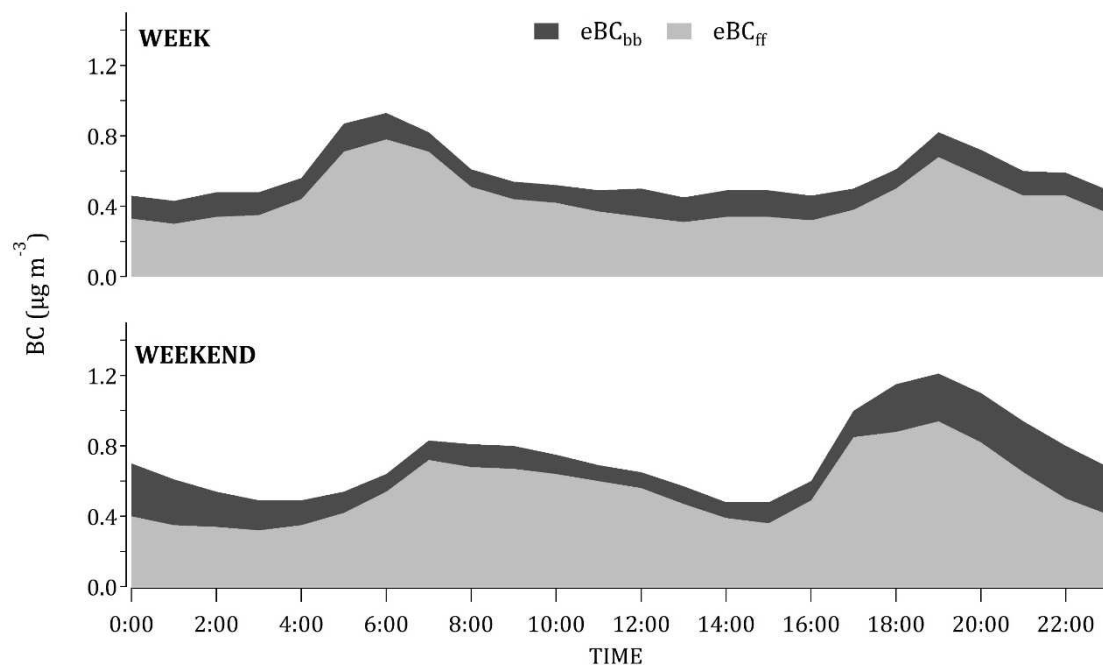
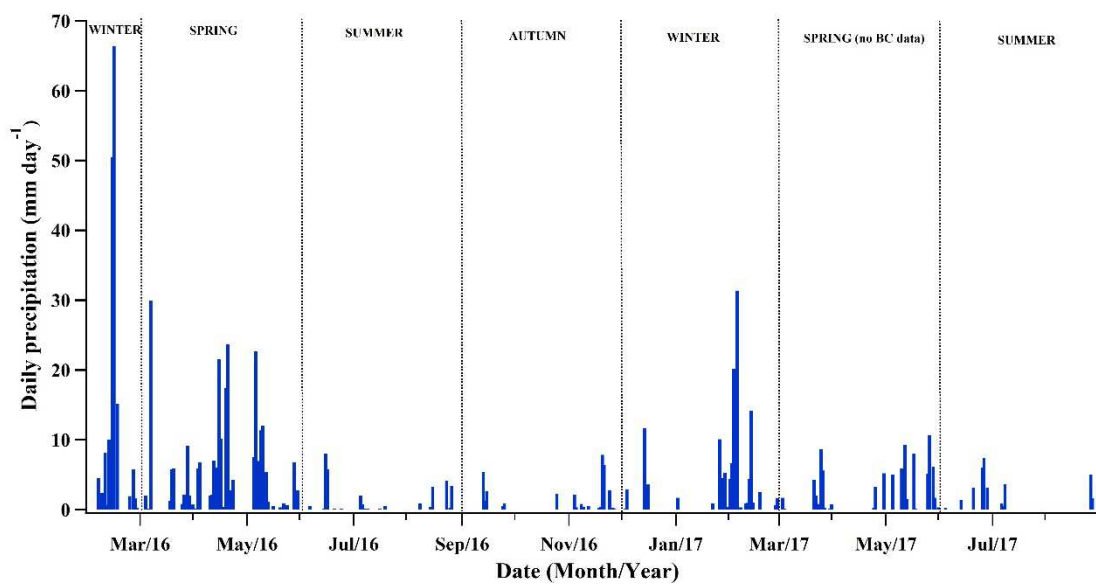
Fig. S3. Diurnal pattern of eBC_{ff} and eBC_{bb} during summer 2017.

Fig. S4. Time series of rain events at León during the sampling period.

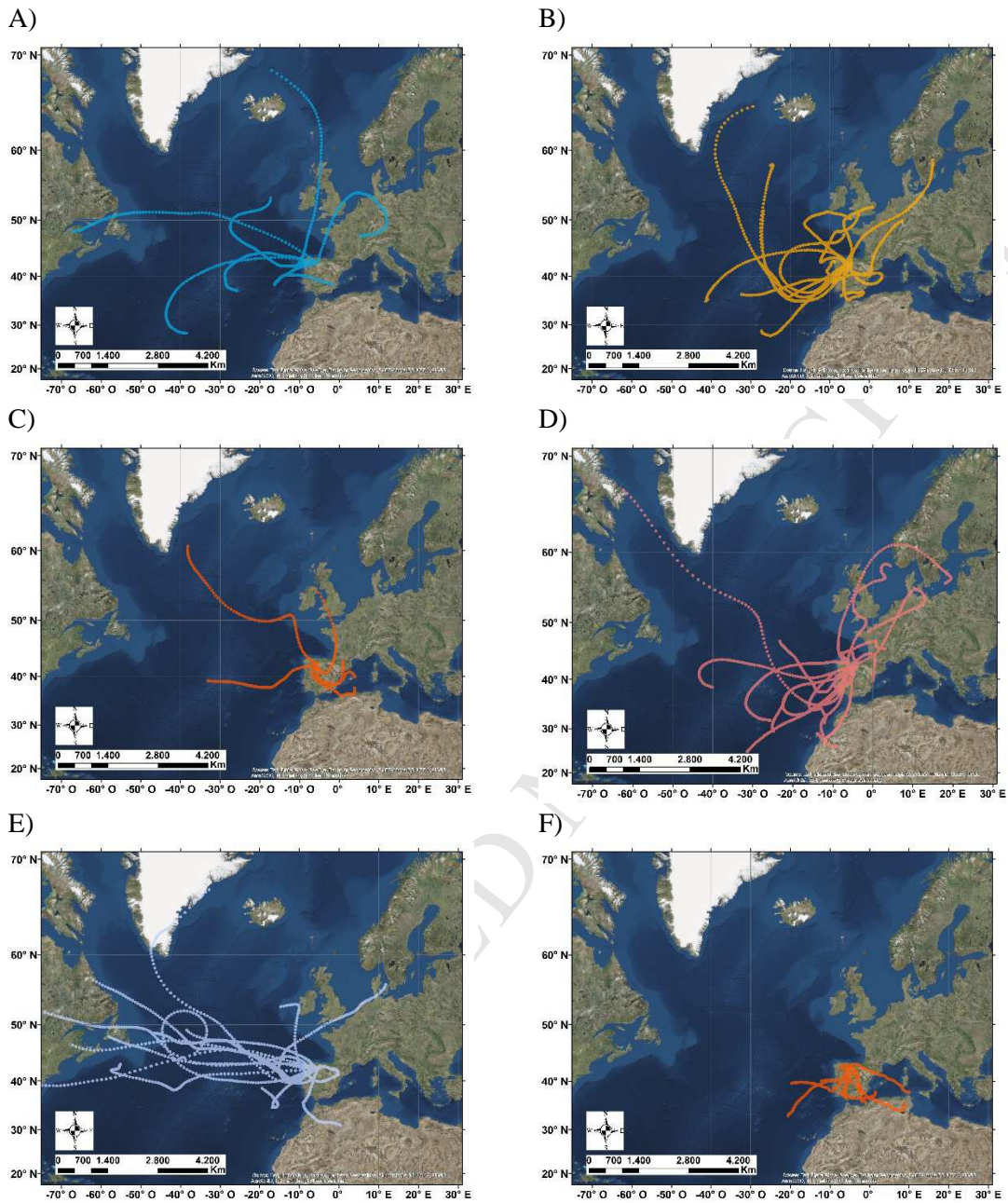


Fig. S5. Four-days back trajectories arriving at 1000 m a.g.l during rain events in León. A) Winter 2016; B) Spring 2016; C) Summer 2016; D) Autumn 2016; E) Winter 2017; F) Summer 2017.

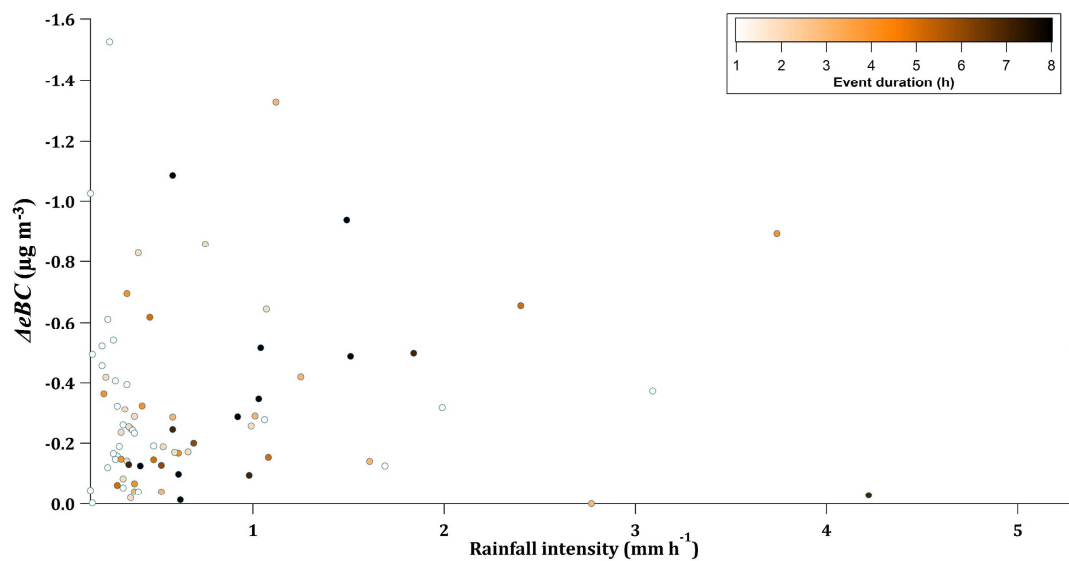


Fig. S6. ΔeBC as a function of rainfall intensity in events with effective scavenging. Colour dots indicate event duration (h).

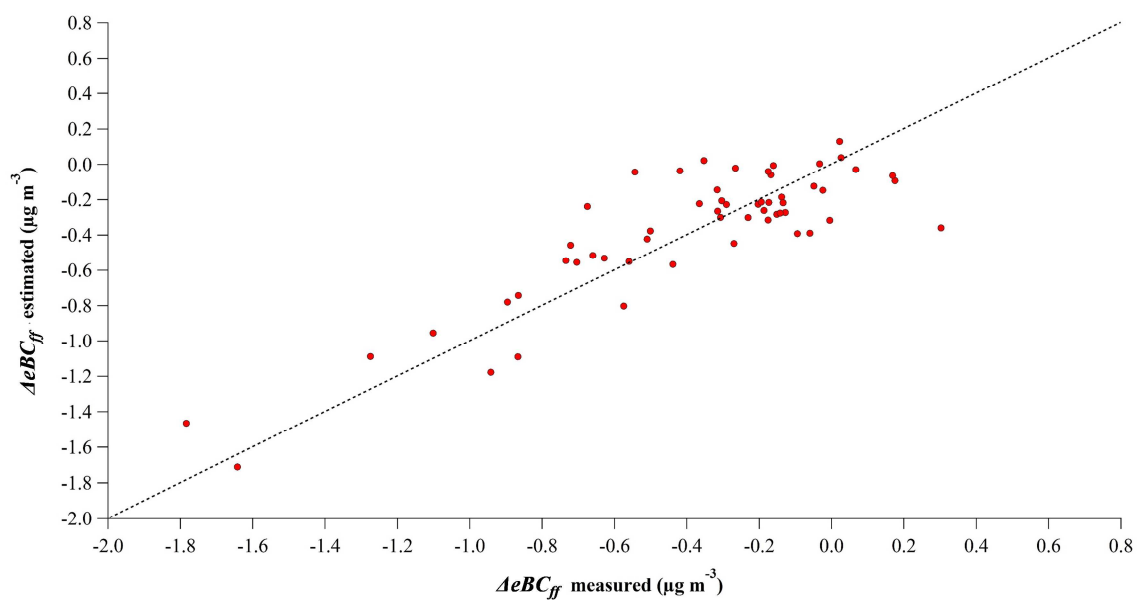


Fig. S7. ΔeBC_{ff} estimated by the model vs ΔeBC_{ff} measured. The dashed line ($y=x$) shows the perfect estimation.

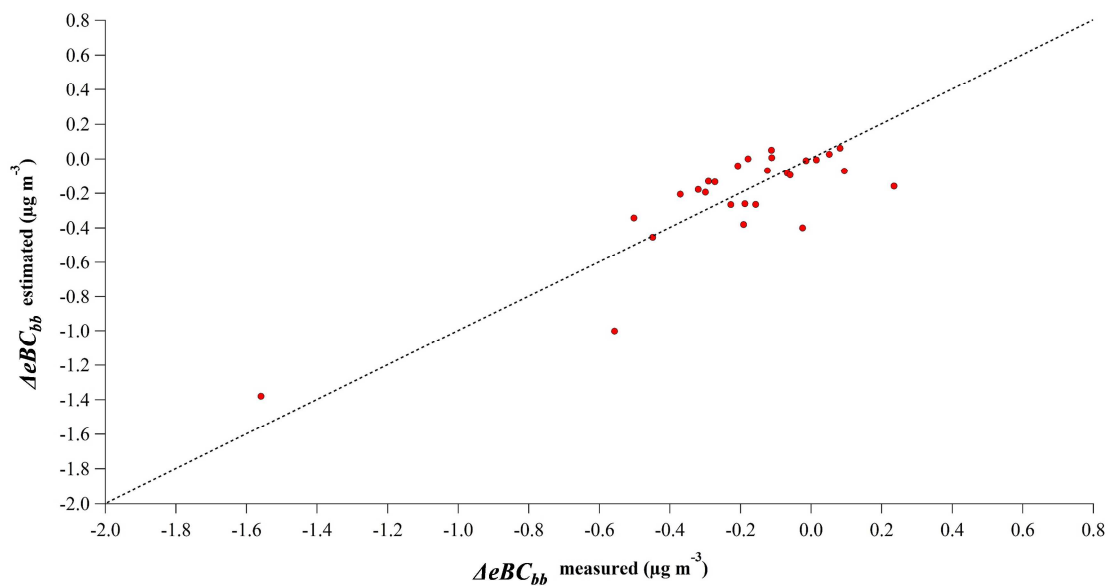


Fig. S8. ΔeBC_{bb} estimated by the model vs ΔeBC_{bb} measured. The dashed line ($y=x$) shows the perfect estimation.

HIGHLIGHTS

- Hourly measures were taken of black carbon and rain during 15 months in León, Spain
- 70% of the rain events showed an effective scavenging (37% eBC decrease)
- The scavenging coefficient for short rain events is 3 times the one for long events
- The scavenging was different according to specific BC source
- A model for rain scavenging of BC with aethalometer and disdrometer data was built