This is a peer reviewed version of an article published by Taylor & Francis. The published version of Jesús Balado, Celia Olabarria, Joaquín Martínez-Sánchez, José R. Rodríguez-Pérez & Arias Pedro (2021) Semantic segmentation of major macroalgae in coastal environments using high-resolution ground imagery and deep learning, International Journal of Remote Sensing, 42:5, 1785-1800 is available at https://doi.org/10.1080/01431161.2020.1842543

1	Semantic segmentation of major macroalgae in coastal
2	environments using high-resolution ground imagery
3	and deep learning
4	Jesús Balado ^{a*} , Celia Olabarria ^{b,c} , Joaquín Martínez-Sánchez ^a , José R.
5	Rodríguez-Pérez ^d , Pedro Arias ^a
6	^a Universidade de Vigo, CINTECX, Campus universitario de Vigo, As Lagoas, Marcosende 36310 Vigo, Spain

^b Universidade de Vigo, Departamento de Ecoloxía e Bioloxía animal, Facultade de Ciencias do Mar. Campus Lagoas-Marcosende, s/n, 36310 Vigo, Spain

> ^c Universidade de Vigo. CIM, Centro de Investigación Mariña. Illa de Toralla, s/n, 36331 Vigo, Spain ^dUniversidad de León. GEOINCA. Avenida de Astorga, s/n, 24401 Ponferrada, Spain

- 12 * Correspondig autor.
- 13 E-mail address: jbalado@uvigo.es (J. Balado)

14 Abstract

7

8

9

10

11

Macroalgae are a fundamental component of coastal ecosystems and play a key role in shaping 15 16 community structure and functioning. Macroalgae are currently threatened by diverse stressors, 17 particularly climate change and invasive species, but they do not all respond in the same way to the 18 stressors. Effective methods of collecting qualitative and quantitative information are essential to enable 19 better, more efficient management of macroalgae. Acquisition of high-resolution images, in which 20 macroalgae can be distinguished on the basis of their texture and colour, and the automated processing of 21 these images are thus essential. Although ground images are useful, labelling is tedious. This study 22 focuses on the semantic segmentation of five macroalgal species in high-resolution ground images taken 23 in 0.5 x 0.5 m quadrats placed along an intertidal rocky shore at low tide. The target species, Bifurcaria 24 bifurcata, Cystoseira tamariscifolia, Sargassum muticum, Sacchoriza polyschides and Codium spp., 25 which predominate on intertidal shores, belong to different morpho-functional groups. The study explains 26 how to convert vector-labelled data to raster-labelled data for adaptation to convolutional neural network 27 (CNN) input. Three CNNs (MobileNetV2, Resnet18, Xception) were compared, and ResNet18 yielded the highest accuracy (91.9%). The macroalgae were correctly segmented, and the main confusion 28 29 occurred at the borders between different macroalgal species, a problem derived from labelling errors. In 30 addition, the interior and exterior of the quadrats were correctly delimited by the CNNs. The results were 31 obtained from only one hundred labelled images and can be performed on personal computers, without 32 the need to resort to external servers. The proposed method helps automation of the labelling process.

33

Keywords: Macroalgae; intertidal rocky shore; convolutional neural networks; image processing;
 semantic segmentation

36

37 **1. Introduction**

Macroalgae are important primary producers on subtidal and intertidal rocky shores worldwide (Jenkins et al., 2008) and make a substantial contribution to carbon sequestration, nutrient cycling and global oxygen production (Bañolas et al., 2020; Macreadie et al., 2017). As ecosystem engineers, they modify habitat conditions, facilitating the existence and survival of other intertidal species and thus strongly influencing the structure and functioning of coastal ecosystems (Purvaja et al., 2018). They also provide
 food, shelter and nursery grounds for many invertebrate and vertebrate species, including commercially

44 important species (Smale et al., 2013).

45 Macroalgae also provide various, highly valuable ecosystem services to humans (Beaumont et al., 2008; 46 Quiros et al., 2018). They are highly nutritious and produce bioactive compounds used in fertilizers, food 47 and medical and cosmetic products (Rebours et al., 2014). Macroalgae are also beginning to be 48 considered as a renewable energy source and as an alternative to fossil fuels (algal biofuel) (Adeniyi et 49 al., 2018; Debiagi et al., 2017). The main benefit of algal biofuel is that it is CO_2 neutral as the CO_2 50 emitted to the atmosphere during biofuel combustion is equivalent to CO_2 needed by the algae to grow 51 and be converted into biofuel. In addition, compounds extracted from macroalgae can be used to 52 neutralize harmful substances from water, at a much smaller cost than other methods. Some authors have 53 reported that algal biofuel can be generated during the use of algae in wastewater treatment (Park et al., 54 2011; Pittman et al., 2011).

55 Coastal and nearshore loss of biodiversity is occurring as a consequence of diverse stressors, including 56 climate change, habitat loss, eutrophication, overfishing, pollution and the introduction of non-native 57 species (Griffiths et al., 2020; Hawkins et al., 2009). Intertidal species of macroalgae are vulnerable to these stressors because they are already close to their physiological tolerance thresholds during 58 59 consecutive periods of emersion and immersion (Helmuth et al., 2006). Shifts in the distributional range 60 of diverse intertidal macroalgae due to increased air and sea surface temperatures (SSTs) on the Atlantic 61 shores of the Iberian Peninsula have been documented (Duarte et al., 2013; Lamela-Silvarrey et al., 2012; 62 Lima et al., 2007).

63 Given the ecological importance of macroalgae and the large number of uses that have been identified, 64 obtaining information about their distribution and abundance is important for monitoring, managing and 65 understanding coastal ecosystems, particularly in the context of global change, in which multiple stressors 66 act together (Floor et al., 2018). Video and photographic monitoring have proven valuable ground-based 67 and remote-sensing techniques for evaluating the cover and distribution of coastal organisms with high 68 spatial and temporal resolution, by using satellites (Li et al., 2012; Sagawa et al., 2012; Topouzelis et al., 69 2016; Wang et al., 2018; Wilson et al., 2019), UAVs (Duffy et al., 2018; Taddia et al., 2019; Tamondong 70 et al., 2018; Ventura et al., 2018; Wang et al., 2019) or underwater drones (Kellaris et al., 2019; Martin-71 Abadal et al., 2018; Moniruzzaman et al., 2019; Rahnemoonfar and Dobbs, 2019). Satellite monitoring is 72 useful for mapping large areas as it provides wide coverage, spatio-temporal refreshment of a few days, 73 often at visible and infrared wavelengths, and requires less input of time and labour than traditional 74 surveys (Schroeder et al., 2019). However, because of the patchiness of macroalgae, especially on 75 intertidal shores (Matias et al., 2015), many species are not recognizable at the resolution of satellite 76 images, unless they cover large areas and are different colours (Brodie et al., 2018). Thus, studies based 77 on aerial data only can monitor biomass or blooms of one species (Xing et al., 2019). Due to the rapid 78 progress of these technologies, the spatial resolution of colour images and the size of image archives are 79 increasing yearly. More sophisticated and efficient image processing algorithms and methods are 80 therefore urgently needed.

81 The use of photograph-based technology for monitoring coastal organisms, including macroalgae and 82 seagrasses, is increasingly being reported. For instance, seaweeds and seagrasses have been mapped using 83 optical processing techniques and textures (Kakuta et al., 2016). Local Binary Patterns (LBP) and 84 Histogram of Oriented Gradients (HOG) have also been implemented as a feature extractor for 85 segmentation of plant species in wetlands (Wang et al., 2018) or in underwater environments (Reus et al., 86 2018). Calculation of indices such as the Normalized Difference Vegetation Index (NDVI), the Floating 87 Algae Index (FAI) and the Seaweed Enhancing Index (SEI) can be also used for feature extraction 88 (Siddiqui et al., 2019).

89 Traditional image processing techniques for detecting and classifying macroalgae are being displaced by 90 machine learning and deep learning methods, which provide more accurate results (O'Byrne et al., 2018; 91 Reus et al., 2018). Deep learning based techniques are limited by the large numbers of samples needed to train the classifier. To differentiate species of macroalgae, samples must be acquired using traditional approaches based on field data, such as diving or intertidal sampling, with in situ quadrats or line transects, which provide high accuracy and resolution, but which are time consuming and limited to small areas (Casal et al., 2013).

96 The aim of this study was to automate the process of labelling high-resolution images to differentiate five 97 macroalage: Bifurcaria bifurcata Linnaeus, Cystoseira tamariscifolia (Hudson) Papenfuss, Sargassum 98 muticum (Yendo) Fensholt, Sacchoriza polyschides (Lightfoot) Batters, and Codium spp. The process was 99 automated using semantic segmentation and convolutional neural networks (CNNs). To our best 100 knowledge, no other works have previously addressed semantic segmentation of five macroalgal species 101 at once from ground, aerial or satellite images. This paper presents a new method of converting labels 102 (from polygons to raster images) and compares the results obtained with three different CNNs 103 (MobileNetV2, Resnet18, Xception). This work is part of ALGANAT2000 project, which aims to 104 monitor the spatio-temporal distribution of macroalgae in an intertidal coastal area within a marine 105 protected area in Galicia (NW Spain) during 2019.

106

107 2. Material and Methods

108 *2.1. Study area*

The study was conducted in the Atlantic Islands National Park (Galicia, NW Spain), a terrestrial and marine reserve formed by four main archipelagos. The exposed intertidal area of Bufardo on the Illa de Monteagudo (area surrounding coordinates 42.23551° N, 8.89956 °W) belonging to the Illas Cíes archipelago was the selected as the sampling location (Figure 1). The location is a gently sloping rocky platform with the upper intertidal dominated by *Pelvetia canaliculata* Decaisne and Thuret, and the mid and low intertidal areas are dominated by conspicuous red, green and brown macroalgae, such as *Asparagopsis armata* Harvey, *B. bifurcata*, *C. tamariscifolia*, *S. polyschides* and *Codium* spp.

116



118 Figure 1. Location of the Bufardo rocky shore in the Atlantic Islands National Park (Galicia, Spain).

119 The images of the five species (Figure 2) were acquired from 0.5 x 0.5 m quadrats placed on the low

120 intertidal shore during low spring tides in July, August and September 2019. The area is emerged, and

thus exposed to the air, for about 3-3.5 hours during low spring tides. Depending on the altitude at which

- 122 the macroalgae occur on the shore, they experience different conditions of solar radiation and desiccation.
- 123 Thus, macroalgae living on the upper shore are drier and absorb more heat than macroalgae inhabiting the
- lower shore.



125 126

Figure 2. Target macroalgae used in the semantic segmentation.

127 The images were acquired with a Fujifilm FinePix JV200 camera mounted on a tripod, with a top view

- perspective 0.7 m above the ground (see Figure 3). Because of the shape of the tripod, the base was also
- 129 captured in each image. The square base delimited the labelling area, and the area outside of the base was
- 130 labelled as the "out" class.



131 132

Figure 3. Quadrat used for image acquisition.

133 2.2. Methods

134 Labelled data (in the form of manually digitized georeferenced vector polygons) and high-resolution 135 images of macroalgae were used for semantic segmentation. Vector labelled data were adapted to raster 136 labelled data following convolutional neural network (CNN) standards. The choice of CNN was justified 137 by the higher success rate than those of other traditional methods based on texture analysis (O'Byrne et 138 al., 2018), histogram of oriented gradients (HOG), local binary patterns (LBP) (Reus et al., 2018). In 139 addition, pre-trained CNNs also enable more efficient feature extraction, with a faster design process than 140 traditional techniques (assembly of successive masks with successive tests). The analysis was approached 141 from the perspective of data analysis, and the system improved as new CNN architectures become 142 available without the entire work process having to be redesigned. The workflow of the method is 143 represented in Figure 4.



145

Figure 4. Workflow

146 2.2.1. Label adaptation for CNN

147 Conventional labelling of vector objects with a geographic information system (GIS) is not suitable for 148 CNN-based semantic segmentation. The conventional method consists of creating a vector file layer with 149 polygons labelled with the class. These polygons are delimited, by an expert, on the background image 150 collected in the field. Apart from their topological attributes, the only condition for labelling is that the 151 polygons must contain only one class of pixels. The expert selects which pixels to polygonise (for training 152 the subsequent class). As a result, the expert classification is considered ground truth and consists of a 153 labelled vector layer.

154 CNN for semantic segmentation only allows images (raster data) as input for both ground truth and 155 labelled data. In addition to the data type, the labelling content is distinct and must obey the following 156 rules:

- All pixels in the image must be labelled. Unlabelled pixels are assigned as "others". The "others"
 class may include macroalgae that are not of interest for the study, e.g. sand, rocks and
 unidentified objects.
- All pixels of the objects must be labelled in their corresponding class. The "others" class must not include pixels representing objects belonging to any class.

In order to fulfil these requirements for CNN training, each image was re-labelled accordingly. The re labelling procedure depended on the following scenarios that can occur in each image-labelled data:

- The expert only polygonised the largest or most relevant macroalgae. In this option, the other pixels must be manually analysed and assigned to the corresponding class (whether macroalgae or "others").
- The expert polygonised both large and small objects. In this case, only the data corresponding to ground and no relevant species need to be labelled *"others"*. In this option, the process could be performed automatically by rasterizing the polygons.

170 Considering that the option used for each image was not known, the re-labelling process was performed 171 manually. In addition, due to the image acquisition method used, macroalgae were labelled exclusively in 172 a region of interest (ROI) in the images. In this case, the ROI in each image was the area enclosed by the quadrat, and the area outside of the quadrat was labelled "out", regardless of whether it included 173 174 macroalgae or not. In the example, from the picture acquired (Figure 5.a), only four polygons 175 corresponding to three different classes were labelled when the expert polygonised the largest relevant 176 macroalgae (Figure 5b). In this case the polygons did not cover all pixels corresponding to each macroalga. A label was then assigned to each pixel (Figure 5c). The contours of the macroalgae were 177 more detailed than those of the respective polygons. 178



Figure 5. Types of labelled data: a) geo-referenced and acquired data, b) polygons labelled and selected
 for training in GIS software, c) raster data for training a semantic segmentation CNN. Classes are
 represented by different colours.

183 2.2.2. Semantic segmentation

184 Semantic segmentation and object detection are classification methods that can be applied to image 185 segmentation and labelling (Ruiz-Santaquiteria et al., 2020). Although both methods are based on deep 186 learning, semantic segmentation aims to assign classes to each pixel of the image while the detector 187 frames the detected objects in a bounding box. This bounding box is defined by a fixed number of 188 vertices that frequently cover several pixels that do not correspond to the class detected. Such 189 misclassification is usual in complex scenarios with contiguous classes, as in the case of distribution of 190 macroalgae. Semantic segmentation delineates the classes more precisely, because it is a pixel-based 191 classification, and it was therefore selected as the classification method for this research.

192 In this paper, we compared the performance of three CNNs in relation to semantic segmentation: 193 MobileNetV2, Resnet18, Xception. These networks each represent different architectures and perform 194 well in segmentation/classification problems. In addition, the training cost of all three CNN is low, both 195 in computational cost and number of labelled samples, as indicated by the number of hidden layers and 196 adjustable parameters. Labelling a large number of samples is a tedious manual task that requires time 197 from biologists familiar with differentiation of macroalgal species. In addition, many laboratories and 198 professionals do not have access to expensive servers to train more complex neural networks, and they are limited to employ personal computers. The characteristics of the different CNNs are summarised below: 199

- MobilNetv2. This CNN is specially designed for operating on mobile devices, and the ratio between accuracy and cost of training is therefore particularly high. It consists of 53 layers and only 3.5 million adjustable parameters and is based on an inverted residual structure in which the shortcut connections are between the thin bottleneck layers (Sandler et al., 2018).
- ResNet18. This is the shallowest of the Deep Residual Networks. It has 18 layers and 11.7 million adjustable parameters. The most important aspect of this CNN is that, during training, it can skip layers if it considers that feature extraction does not contribute relevant information (He et al., 2016).
- Xception. This is an evolution of Inception architecture. It has 71 layers and 22.9 million adjustable parameters, and is thus the deepest of the networks used in this study. This CNN is based entirely on depth-wise separable convolution layers (Chollet, 2017).

Images for semantic segmentation with CNN must have minimum dimensions according to the feature 211 extractor (224x224x3 pixels for MobilNetv2 and ResNet18, and 299x299x3 pixels for Xception). In the 212 213 present study, the dimensions of the acquired images were 4288 x 3216 x 3 pixels. Given this high 214 resolution, the images contained a great deal of detail, facilitating manual labelling by experts. 215 Unfortunately, the amount of computer resources that must be allocated for network training increases 216 with the image size. In order to train the networks on a conventional computer, the images were re-sized maintaining the aspect ratio to 1000 x 750 x 3 pixels. This resolution still retained a high level of detail in 217 218 the images. CNNs were adapted from image classification for semantic segmentation using DeepLabV3 219 (Chen et al., 2018) in Matlab.

220 2.2.3. Data augmentation and distribution

The data set was also modified by data augmentation. Data augmentation allows the training set to be extended to automatically generate new samples through small modifications of the original set. In this case, data augmentation was applied from reflections on the X and Y axes, 20-pixel translations in both axes and rotations with angles less than 25°.

225 One aim of the data acquisition process was to obtain a representative number of images of each species. 226 Nevertheless, the percentage occupation of each image was unbalanced (Table 1), which is a typical 227 problem in semantic segmentation. Although almost all classes were of the same order of magnitude, relative to other species, there were very few samples of S. muticum. The "out" class included a larger 228 229 number of pixels, as it appeared in all images outside the ROI. The imbalance between classes can be 230 minimized by assigning weights to the pixels according to the quantity in the training set. For the 231 validation and testing sets, balanced sample sets were chosen so that the results were as balanced as 232 possible. In total, 130 images were labelled and distributed as follows: 90 images for training, 10 images 233 for validation and 30 images for testing.

Class	Total number of pixels (10 ⁶)	Training number of pixels (10 ⁶)	Validation number of pixels (10 ⁶)	Testing number of pixels (10 ⁶)
"out"	58.93	41.16	4.47	13.29
B. bifurcata	6.28	4.46	0.47	1.34
C. tamariscifolia	4.96	3.50	0.68	0.77
S. muticum	1.36	0.59	0.30	0.47
S. polyschides	8.33	6.10	0.59	1.63
Codium spp	5.89	3.43	0.46	2.00
"others"	11.76	8.25	0.52	2.99

Table 1. Number of pixels per class.

235

236 2.2.4. Training

The network was trained on a laptop computer (GPU NVIDIA GTX1050 4GB GDDR5, CPU i7-7700HQ 2.8Ghz and 16GB RAM DDR4). The hyperparameters were chosen experimentally through several tests, maximizing the performance and minimizing the overfitting. The hyperparameters that obtained a better result for training were as follows: optimization method, *sgdm*; learning rate, 0.001; momentum, 0.9; L2 regularization, 0.005; and max epochs, 15. The mini batch size was set at 4 limited by the amount of memory of the graphic card. The time consumed by each training was around 150 min. The programming language used was *Matlab*. All training sessions converged satisfactorily (Figure 6).



Figure 6. Variation in the loss during the training process

246 **3. Results**

Table 2. Overall accuracy obtained on training, validation and test sets.

	Train	Val	Test
ResNet18	93.7%	90.6%	91.9%
MobileNetV2	91.4%	85.0%	88.4%
Xception	90.9%	84.7%	87.3%

247

249 Overall accuracy obtained on training, validation and test sets are shown in Table 2. Overfitting was 250 detected among the training and validation sets, although it was reduced in the testing set. The difference 251 between the validation and test sets is due to the difference in the number of pixels per class. Although 252 the overfitting was reduced with the adjustment of the hyperparameters, it was not completely eliminated. 253 The remaining overfitting was considered acceptable in view of the results, both qualitative and 254 quantitative. The best result was obtained with Resnet18, although the performance was not the same for 255 all classes. The confusion matrices for training the three different CNNs are shown in Tables 3 to 5. The 256 best result was obtained with Resnet18, although the performance was not the same for all classes. 257 ResNet18 produced better segmentation of the B. bifurcata, S. muticum, S. polyschides, and "out" classes. 258 Accurate identification of the "out" class led to good ROI delimitation. MobileNetV2 produced better 259 segmentation of the Codium spp and "others" (mainly composed of sand) classes, but produced very similar results to ResNet18. Xception produced by far the best results for the C. tamariscifolia class. In 260 261 the confusion matrices, the success rates were lowest for the C. tamariscifolia and S. muticum classes, 262 which yielded more confusion than the other classes. Specifically, C. tamariscifolia was confused with 263 the "others" classes by 0.296, and S. muticum was confused with the "others" by 0.197, and with C. 264 tamariscifolia by 0.141. The colours of these classes were similar; in addition, very few samples of the S. muticum class were available for training. Good success rates were obtained for the remaining classes, 265 266 and the confusion between them was minimal.

The most notable results for semantic segmentation with ResNet18 were the areas classified as macroalgae outside the ROI (Figure 7). However, these areas corresponded to macroalgae that were well classified and with continuous macroalgae within the ROI. In addition, although the centres of the macroalgae were well defined, the borders were quite irregular and not well defined. The borders did not fit properly, mainly in dark areas, overlapping areas between macroalgae or when a small macroalga was surrounded by another macroalga.

Resnet18 produced the best segmentation of macroalgae. It was foreseeable that MobileNetv2 would not perform particularly well, given the fewer configurable parameters. However, Xception did not produce better results, despite being a much deeper CNN with the capacity to extract more complex features. The Xception network only outperformed the other CNNs in the accuracy of segmenting the *C. tamariscifolia* class (one of the classes for which ResNet18 produced the least accurate results), but at the cost of increasing confusion about the *S. muticum* class, for which relatively poor results were obtained.

Very high success rates were obtained for the segmentation of most classes (including three different macroalgal species). ResNet18 learned the texture and colour patterns of different species, regardless of factors that led to changes, such as the time out of water between acquisitions. Although the *B. bifurcata* and *Codium spp* classes were of similar texture, they were easily distinguished by their colour. The *S. polyschides* class did not coincide in colour or texture with any of the other classes. Low success rates (of around 60%), were only obtained for the *C. tamariscifolia* and *S. muticum* classes, possibly due to the similar colour and texture of these species.

287 Table 3. Confusion matrix for ResNet18.

ref\pred	"out"	B. bifurcata	C. tamariscifolia	S. muticum	S. polyschides	<i>Codium</i> spp	"others"
"out"	0.962	0.004	0.001	0.001	0.012	0.010	0.009
B. bifurcata	0.003	0.921	0.002	0.009	0.001	0.007	0.058

C. tamariscifolia	0.004	0.020	0.590	0.042	0.017	0.032	0.296
S. muticum	0.015	0.008	0.197	0.618	0.018	0.004	0.141
S. polyschides	0.008	0.003	0.004	0.000	0.939	0.019	0.026
Codium spp	0.003	0.010	0.012	0.002	0.026	0.912	0.034
"others"	0.008	0.025	0.032	0.004	0.025	0.052	0.854

289 Table 4. Confusion matrix for MobileNetV2.

ref\pred	"out"	B. bifurcata	C. tamariscifolia	S. muticum	S. polyschides	<i>Codium</i> spp	"others"
"out"	0.902	0.008	0.005	0.004	0.023	0.027	0.032
B. bifurcata	0.007	0.907	0.006	0.007	0.002	0.014	0.058
C. tamariscifolia	0.003	0.016	0.606	0.039	0.009	0.013	0.314
S. muticum	0.031	0.006	0.121	0.601	0.005	0.008	0.229
S. polyschides	0.016	0.003	0.007	0.000	0.899	0.027	0.048
Codium spp	0.007	0.007	0.009	0.003	0.020	0.926	0.028
"others"	0.012	0.016	0.034	0.004	0.018	0.044	0.871

291 Table 5. Confusion matrix for Xception.

ref\pred	out	B. bifurcata	C. tamariscifolia	S. muticum	S. polyschides	<i>Codium</i> spp	"others"
"out"	0.908	0.012	0.008	0.001	0.019	0.021	0.031
B. bifurcata	0.004	0.884	0.020	0.000	0.001	0.014	0.077
C. tamariscifolia	0.006	0.019	0.725	0.003	0.009	0.031	0.206
S. muticum	0.017	0.014	0.500	0.266	0.003	0.010	0.190
S. polyschides	0.016	0.011	0.008	0.000	0.878	0.026	0.060
Codium spp	0.009	0.010	0.027	0.001	0.019	0.899	0.036
"others"	0.011	0.023	0.076	0.002	0.017	0.042	0.829



293 294

Figure 7. Results of semantic segmentation with ResNet18. The acquired image was superimposed on colours representing each class.

297 4. Discussion

298 In this paper, we report a CNN-based segmentation procedure for macroalgae, which yielded a success 299 rate > 90% for all three CNNs tested. One of the key reasons for the high success rate in classifying the 300 species was the use of high-resolution ground images collected in the field. The images were reduced in 301 order to save time and computational resources, to a final resolution of 1000 x 750 pixels, which was high 302 relative to the examples reported in the literature. By contrast, underwater images used to segment 303 seagrass coverage were reduced to 512 x 256 pixels in previous studies (Weidmann et al., 2019). The study findings show that the proposed resolution satisfactory differentiated the five species and the 304 305 interior/exterior zones of each quadrat. At the same time, a laptop workstation was adequate for training 306 the CNNs at this resolution, and computational resources from external servers were not required.

307 Although the accuracy rate was similar to that obtained in other studies with satellite, aerial and 308 submarine sources, the present study aimed to differentiate five different macroalgae and it is, therefore, 309 not generally comparable to other studies concerning the detection of single species. The accuracy 310 achieved in number of CNN-based studies is very variable: 99.4% (Zhou et al., 2019), 97.0% (Wang et 311 al., 2019), 95.8% (Rahnemoonfar and Dobbs, 2019), 95.0% (Martin-Abadal et al., 2018) and 90.1% 312 (Arellano-Verdejo et al., 2018). These studies usually based on satellite and airborne data have only 313 focused on detecting the predominant macroalgal species given the low resolution relative to ground data. 314 Their technical complexity is considerably lesser than the presented in this work. They segmented of one 315 macroalgae class from the bottom, often sand or water, without differentiation between macroalgal 316 species. Differentiating between different macroalgal species is feasible when broad taxonomic groups

317 are considered, e.g. green, red and brown algae (Andrefouet et al., 2004; Kotta et al., 2018). The findings 318 of the present study showed that colour attributes were not sufficient for correct classification, as C. 319 tamariscifolia and S. muticum are very similar in colour and only differ in texture. Depending on the 320 classification scale, the texture feature is not extractable from satellite and aerial images due to the lower 321 resolution of these. In addition, the same species can display notable differences in colour depending on 322 the morphology, thickness of thalli and cellular architecture, which determine pigment densities, 323 absorption, and thus reflectance spectra (Vogelmann and Björn, 1986). Environmental conditions, such as 324 the emersion time and intensity of solar radiation, also contribute to differences in pigmentation within 325 and between macroalgal species (Dieter et al., 2003). The resolution of images obtained by underwater 326 drones is higher than that of airborne data, and the colour is modified relative to images taken outside the 327 water; however, the modification affects all species equally (O'Byrne et al., 2018). Macroalgal species 328 should be able to be differentiated by these characteristics (resolution and colour), and therefore texture, 329 in underwater images. Nevertheless, most studies based on underwater images and also studies based on 330 satellite and aerial images have only focused on detecting single macroalgal species (Gonzalez-Cid et al., 331 2017; Moniruzzaman et al., 2019).

332 The features extracted from one species were easier to obtain and learn with a classifier based on artificial 333 intelligence, as the macroalgal classes shared more features with each other than with non-macroalgal 334 classes such as rock, sand and seawater. However, when the macroalgal class no longer corresponded to 335 one species (as in this study) and was divided into five classes, the difficulty for the algorithm increased 336 in relation to both finding and in extracting distinctive features. More advanced techniques than simple 337 Support Vector Machines or Artificial Neural Networks were thus required. Classic image processing 338 techniques such as LBP lack the ability to learn complex features and thus produce poorer results than 339 those obtained with CNNs. The accuracy of segmentation of single seagrass species only reaches 85.0% 340 with LBP techniques, but increases to 93.4% with CNN (Wang et al., 2018; Reus et al., 2018).

341 The CNNs under study proved very useful for segmentation of the five macroalgae considered. Although, 342 in theory, a large number of labelled images was required, in practice the number of pixels was more 343 important for semantic segmentation. Given the high resolution of the images used, the number of pixels 344 was sufficient to train a CNN with only 100 images, which can be obtained quickly. From these 100 345 labelled images, and after training, infinite images can be labelled without further human intervention. 346 However, the labelling process for training must be conducted carefully, as CNNs can learn labelling 347 errors. Confusion at the borders of macroalgae (Figure 7) was due to labelling errors of the images 348 (Vogelmann and Björn, 1986). For human observers, the centre of the algae is easy to segment and label 349 manually, as with CNN segmentation. However, the borders of many algae overlap and it is not easy to 350 define outlines to separate them. Because of these errors in the labelled images, the errors were also learnt 351 by the CNN and replicated in the segmentation. These types of errors tend to be minimized when the data 352 are tagged by different people.

353 The acquisition time is slower with ground imaging than in satellite and drone-based methods, because 354 the quadrat has constantly to be moved to a new location for each new image. In addition, the area 355 covered by each image was only 0.25 m². Nevertheless, ground images are required in order to provide 356 reference data to train models and map data obtained with other automated instruments. This study 357 focused on the exclusive use of RGB photographic images to minimize pre-processing time by fusing 358 information and acquisition efforts. Because of the high success rates obtained, other types of data, such 359 as spectroradiometer data (Hu, 2009), multispectral-hyperspectral images (Fauzan et al., 2017; Li et al., 2012; Taddia et al., 2019; Zacharias et al., 1992) and environmental data (De Oliveira et al., 2006), were 360 not included. 361

362

363 5. Conclusion and future work

This study involved a CNN-based semantic segmentation of high-resolution ground images of five different macroalgae inhabiting rocky shores. The study findings demonstrate that vector-labelled samples can be adapted for use with CNNs. Of the three different CNNs compared, ResNet18 produced the best results, i.e. 91.9% accuracy. Most of the samples were correctly labelled, although there was a tendency for some macroalgae outside the ROI to be labelled, and the borders between macroalgal species were diffuse. Although theoretically considered an error, in practice segmentation of macroalgae outside the ROI is not problematical, as long as the classification is correct, as in this case. Definition of borders is also a problem experienced by human observers. The proposed method is therefore considered a suitable alternative for the automation of sample labelling.

373 The study findings demonstrated that automation of the labelling process is possible with only 100 high-374 resolution images obtained in the field, without the need for other types of data. A further step will be to 375 apply the method to data obtained from UAVs. Nevertheless, further research is required to determine the 376 minimum resolution needed to guarantee correct results and for transfer to learning between UAV and ground images, which differ in resolution and, therefore, in texture and colour. The use of UAVs, together 377 378 with the findings presented here, will facilitate the rapid acquisition and mapping of macroalgal cover on 379 intertidal rocky shores, with a high degree of automation. Use of these methods could greatly improve the 380 management of coastal areas.

381

382 Acknowledgements

This work was supported by the Fundación Biodiversidad, the Ministerio para la Transición Ecológica y el Reto Demográfico through the Pleamar program, co-funded by the European Maritime and Fisheries Fund (EMFF), call 2018. It was also partly funded through grants awarded by the Xunta de Galicia for human resources and competitive reference groups (ED481B-2019-061 and ED431C 2016-038) and by the Ministerio de Ciencia, Innovación y Universidades -Gobierno de España (RTI2018-095893-B-C21). This document only reflects the views of the authors, and the statements made herein are solely the responsibility of the authors.

390

391 References and Notes

- Adeniyi, O.M., Azimov, U., Burluka, A., 2018. Algae biofuel: Current status and future applications.
 Renewable and Sustainable Energy Reviews 90, 316–335.
 https://doi.org/https://doi.org/10.1016/j.rser.2018.03.067
- Andrefouet, S., Payri, C., Hochberg, E., Hu, C., Atkinson, M., Muller-Karger, F., 2004. Use of In Situ
 and Airborne Reflectance for Scaling-Up Spectral Discrimination of Coral Reef Macroalgae from
 Species to Communities. Marine Ecology-progress Series MAR ECOL-PROGR SER 283, 161–
 177. https://doi.org/10.3354/meps283161
- Arellano-Verdejo, J., Lazcano, H., Cabanillas-Terán, N., 2018. ERISNet: Deep learning network for
 Sargassum detection along the coastline of the Mexican Caribbean.
 https://doi.org/10.7287/peerj.preprints.27445
- 402Bañolas, G., Fernández, S., Espino, F., Haroun, R., Tuya, F., 2020. Evaluation of carbon sinks by the403seagrass Cymodocea nodosa at an oceanic island: Spatial variation and economic valuation. Ocean404& Coastal405https://doi.org/https://doi.org/10.1016/j.ocecoaman.2020.105112
- 406Beaumont, N., Austen, M., Mangi, S., Townsend, M., 2008. Economic valuation for the conservation of407marinebiodiversity.Marinepollutionbulletin56,386–396.408https://doi.org/10.1016/j.marpolbul.2007.11.013
- Brodie, J., Ash, L., Tittley, I., Yesson, C., 2018. A comparison of multispectral aerial and satellite
 imagery for mapping intertidal seaweed communities. Aquatic Conservation: Marine and
 Freshwater Ecosystems 28. https://doi.org/10.1002/aqc.2905
- Casal, G., Kutser, T., Gómez, J., Sánchez-Carnero, N., Freire, J., 2013. Assessment of the hyperspectral
 sensor CASI-2 for macroalgal discrimination on the Ría de Vigo coast (NW Spain) using field
 spectroscopy and modelled spectral libraries. Continental Shelf Research 55, 129–140.
 https://doi.org/10.1016/j.csr.2013.01.010

- Chen, L.-C., Zhu, Y., Papandreou, G., Schroff, F., Adam, H., 2018. Encoder-decoder with atrous
 separable convolution for semantic image segmentation, in: Proceedings of the European
 Conference on Computer Vision (ECCV). pp. 801–818.
- Chollet, F., 2017. Xception: Deep Learning with Depthwise Separable Convolutions, in: 2017 IEEE
 Conference on Computer Vision and Pattern Recognition (CVPR). pp. 1800–1807.
 https://doi.org/10.1109/CVPR.2017.195
- 422 De Oliveira Eric Populus Jacques, G.B., 2006. Predictive modelling of coastal habitats using remote
 423 sensing data and fuzzy logic: A case for seaweed in Brittany (France). EARSeL eProceedings.
- Debiagi, P., Trinchera, M., Frassoldati, A., Ranzi, E., Faravelli, T., 2017. Third Generation Biomass.
 Classification and Characterization of Algae Fuels.
- Dieter, H., Wiencke, C., Bischof, K., 2004. Photosynthesis in Marine Macroalgae. pp. 413–435.
 https://doi.org/10.1007/978-94-007-1038-2_18
- 428Duarte, L., Viejo, R.M., Martínez, B., deCastro, M., Gómez-Gesteira, M., Gallardo, T., 2013. Recent and429historical range shifts of two canopy-forming seaweeds in North Spain and the link with trends in430sea431https://doi.org/https://doi.org/10.1016/j.actao.2013.05.002
- Duffy, J.P., Pratt, L., Anderson, K., Land, P.E., Shutler, J.D., 2018. Spatial assessment of intertidal
 seagrass meadows using optical imaging systems and a lightweight drone. Estuarine, Coastal and
 Shelf Science 200, 169–180. https://doi.org/https://doi.org/10.1016/j.ecss.2017.11.001
- Fauzan, M.A., Kumara, I.S.W., Yogyantoro, R., Suwardana, S., Fadhilah, N., Nurmalasari, I., Apriyani,
 S., Wicaksono, P., 2017. Assessing the Capability of Sentinel-2A Data for Mapping Seagrass
 Percent Cover in Jerowaru, East Lombok. The Indonesian Journal of Geography 49, 195–203.
- Floor, J.R., van Koppen, C.S.A. (Kris), van Tatenhove, J.P.M., 2018. Science, uncertainty and changing
 storylines in nature restoration: The case of seagrass restoration in the Dutch Wadden Sea. Ocean &
 Coastal Management 157, 227–236.
 https://doi.org/https://doi.org/10.1016/j.ocecoaman.2018.02.016
- Gonzalez-Cid, Y., Burguera, A., Bonin-Font, F., Matamoros, A., 2017. Machine learning and deep
 learning strategies to identify Posidonia meadows in underwater images, in: OCEANS 2017 Aberdeen. pp. 1–5. https://doi.org/10.1109/OCEANSE.2017.8084991
- Griffiths, L.L., Connolly, R.M., Brown, C.J., 2020. Critical gaps in seagrass protection reveal the need to
 address multiple pressures and cumulative impacts. Ocean & Coastal Management 183, 104946.
 https://doi.org/https://doi.org/10.1016/j.ocecoaman.2019.104946
- Hawkins, S., Sugden, H., Mieszkowska, N., Moore, P., Poloczanska, E., Leaper, R., Herbert, R., Genner,
 M., Moschella, P.S., Thompson, R.C., Jenkins, S.R., Southward, A.J., Burrows, M., 2009.
 Consequences of climate-driven biodiversity changes for ecosystem functioning of North European
 rocky shores. Marine Ecology Progress Series 396.
- He, K., Zhang, X., Ren, S., Sun, J., 2016. Deep residual learning for image recognition, in: Proceedings of
 the IEEE Conference on Computer Vision and Pattern Recognition. pp. 770–778.
- Helmuth, B., Mieszkowska, N., Moore, P., Hawkins, S.J., 2006. Living on the Edge of Two Changing
 Worlds: Forecasting the Responses of Rocky Intertidal Ecosystems to Climate Change. Annual
 Review of Ecology, Evolution, and Systematics 37, 373–404.
 https://doi.org/10.1146/annurev.ecolsys.37.091305.110149
- Hu, C., 2009. A novel ocean color index to detect floating algae in the global oceans. Remote Sensing of
 Environment 113, 2118–2129. https://doi.org/https://doi.org/10.1016/j.rse.2009.05.012
- Jenkins, S.R., Moore, P., Burrows, M.T., Garbary, D.J., Hawkins, S.J., Ingólfsson, A., Sebens, K.P.,
 Snelgrove, P.V.R., Wethey, D.S., Woodin, S.A., 2008. Comparative ecology of North Atlantic
 shores: Do differences in players matter for process? Ecology 89, S3–S23.
 https://doi.org/10.1890/07-1155.1
- Kellaris, A., Gil, A., Faria, J., Amaral, R., Moreu-Badia, I., Neto, A., Yesson, C., 2019. Using low-cost drones to monitor heterogeneous submerged seaweed habitats: A case study in the Azores. Aquatic Conservation: Marine and Freshwater Ecosystems 29, 1909–1922. https://doi.org/10.1002/aqc.3189
- 467 Kotta, J., Valdivia, N., Kutser, T., Toming, K., Rätsep, M., Orav-Kotta, H., 2018. Predicting the cover

- and richness of intertidal macroalgae in remote areas: a case study in the Antarctic Peninsula.
 Ecology and Evolution 8, 9086–9094. https://doi.org/10.1002/ece3.4463
- Lamela-Silvarrey, C., Fernández, C., Anadón, R., Arrontes, J., 2012. Fucoid assemblages on the north
 coast of Spain: past and present (1977–2007). Botanica Marina. https://doi.org/10.1515/bot-2011 0081
- Li, R., Liu, J.-K., Sukcharoenpong, A., Yuan, J., Zhu, H., Zhang, S., 2012. A Systematic Approach toward Detection of Seagrass Patches from Hyperspectral Imagery. Marine Geodesy 35, 271–286. https://doi.org/10.1080/01490419.2012.699019
- Lima, F.P., Ribeiro, P.A., Queiroz, N., Hawkings, S.J., Santos, A.M., 2007. Do distributional shifts of
 northern and southern species of algae match the warming pattern? Global Change Biology 13,
 2592–2604. https://doi.org/10.1111/j.1365-2486.2007.01451.x
- Macreadie, P.I., Jarvis, J., Trevathan-Tackett, S.M., Bellgrove, A., 2017. Seagrasses and Macroalgae:
 Importance, Vulnerability and Impacts. Climate Change Impacts on Fisheries and Aquaculture,
 Wiley Online Books. https://doi.org/doi:10.1002/9781119154051.ch22
- Martin-Abadal, M., Guerrero-Font, E., Bonin-Font, F., Gonzalez-Cid, Y., 2018. Deep Semantic
 Segmentation in an AUV for Online Posidonia Oceanica Meadows Identification. IEEE Access 6,
 60956–60967. https://doi.org/10.1109/ACCESS.2018.2875412
- Matias, M.G., Arenas, F., Rubal, M., Pinto, I.S., 2015. Macroalgal Composition Determines the Structure
 of Benthic Assemblages Colonizing Fragmented Habitats. PloS one 10, e0142289–e0142289.
 https://doi.org/10.1371/journal.pone.0142289
- Moniruzzaman, M., Islam, S.M.S., Lavery, P., Bennamoun, M., 2019. Faster R-CNN Based Deep
 Learning for Seagrass Detection from Underwater Digital Images, in: 2019 Digital Image
 Computing: Techniques and Applications (DICTA). pp. 1–7.
 https://doi.org/10.1109/DICTA47822.2019.8946048
- 492 O'Byrne, M., Pakrashi, V., Schoefs, F., Ghosh, B., 2018. Semantic Segmentation of Underwater Imagery
 493 Using Deep Networks Trained on Synthetic Imagery. Journal of Marine Science and Engineering .
 494 https://doi.org/10.3390/jmse6030093
- 495Park, J.B.K., Craggs, R.J., Shilton, A.N., 2011. Wastewater treatment high rate algal ponds for biofuel496production.BioresourceTechnology102,35-42.497https://doi.org/https://doi.org/10.1016/j.biortech.2010.06.158
- Pittman, J.K., Dean, A.P., Osundeko, O., 2011. The potential of sustainable algal biofuel production using
 wastewater resources. Bioresource Technology 102, 17–25.
 https://doi.org/https://doi.org/10.1016/j.biortech.2010.06.035
- Purvaja, R., Robin, R.S., Ganguly, D., Hariharan, G., Singh, G., Raghuraman, R., Ramesh, R., 2018.
 Seagrass meadows as proxy for assessment of ecosystem health. Ocean & Coastal Management
 159, 34–45. https://doi.org/https://doi.org/10.1016/j.ocecoaman.2017.11.026
- Quiros, T.E.A.L., Beck, M.W., Araw, A., Croll, D.A., Tershy, B., 2018. Small-scale seagrass fisheries
 can reduce social vulnerability: a comparative case study. Ocean & Coastal Management 157, 56–
 67. https://doi.org/10.1016/j.ocecoaman.2018.02.003
- Rahnemoonfar, M., Dobbs, D., 2019. SEMANTIC SEGMENTATION OF UNDERWATER SONAR
 IMAGERY WITH DEEP LEARNING. https://doi.org/10.13140/RG.2.2.32343.52644
- Rebours, C., Marinho-Soriano, E., Zertuche-González, J.A., Hayashi, L., Vásquez, J.A., Kradolfer, P.,
 Soriano, G., Ugarte, R., Abreu, M.H., Bay-Larsen, I., others, 2014. Seaweeds: an opportunity for
 wealth and sustainable livelihood for coastal communities. Journal of applied phycology 26, 1939–
 1951.
- Reus, G., Moller, T., Jager, J., Schultz, S., Kruschel, C., Hasenauer, J., Wolff, V., Fricke-Neuderth, K.,
 2018. Looking for Seagrass: Deep Learning for Visual Coverage Estimation.
 https://doi.org/10.1109/OCEANSKOBE.2018.8559302
- Ruiz-Santaquiteria, J., Bueno, G., Deniz, O., Vallez, N., Cristobal, G., 2020. Semantic versus instance
 segmentation in microscopic algae detection. Engineering Applications of Artificial Intelligence 87,
 103271. https://doi.org/https://doi.org/10.1016/j.engappai.2019.103271
- 519 Sagawa, T., Mikami, A., Aoki, M., Komatsu, T., 2012. Mapping seaweed forests with IKONOS image

- based on bottom surface reflectance, Proceedings of SPIE The International Society for Optical
 Engineering. https://doi.org/10.1117/12.975678
- Sandler, M., Howard, A., Zhu, M., Zhmoginov, A., Chen, L., 2018. MobileNetV2: Inverted Residuals and
 Linear Bottlenecks, in: 2018 IEEE/CVF Conference on Computer Vision and Pattern Recognition.
 pp. 4510–4520. https://doi.org/10.1109/CVPR.2018.00474
- Schroeder, S.B., Dupont, C., Boyer, L., Juanes, F., Costa, M., 2019. Passive remote sensing technology
 for mapping bull kelp (Nereocystis luetkeana): A review of techniques and regional case study.
 Global Ecology and Conservation 19, e00683.
 https://doi.org/https://doi.org/10.1016/j.gecco.2019.e00683
- Siddiqui, D.M., Zaidi, Z.A., Abdullah, M., 2019. Performance Evaluation of Newly Proposed Seaweed
 Enhancing Index (SEI). Remote Sensing . https://doi.org/10.3390/rs11121434
- Smale, D.A., Burrows, M.T., Moore, P., O'Connor, N., Hawkins, S.J., 2013. Threats and knowledge gaps
 for ecosystem services provided by kelp forests: a northeast Atlantic perspective. Ecology and
 Evolution 3, 4016–4038. https://doi.org/10.1002/ece3.774
- Taddia, Y., Russo, P., Lovo, S., Pellegrinelli, A., 2019. Multispectral UAV monitoring of submerged
 seaweed in shallow water. Applied Geomatics. https://doi.org/10.1007/s12518-019-00270-x
- Tamondong, A., Cruz, C., Guihawan, J., Garcia, M., Quides, R.R., Cruz, J., Blanco, A., 2018. Remote
 sensing-based estimation of seagrass percent cover and LAI for above ground carbon sequestration
 mapping. https://doi.org/10.1117/12.2324695
- Topouzelis, K., Spondylidis, S.C., Papakonstantinou, A., Soulakellis, N., 2016. The use of Sentinel-2
 imagery for seagrass mapping: Kalloni Gulf (Lesvos Island, Greece) case study, in: Proc.SPIE.
- Ventura, D., Bonifazi, A., Gravina, F.M., Belluscio, A., Ardizzone, G., 2018. Mapping and Classification
 of Ecologically Sensitive Marine Habitats Using Unmanned Aerial Vehicle (UAV) Imagery and
 Object-Based Image Analysis (OBIA). Remote Sensing . https://doi.org/10.3390/rs10091331
- Vogelmann, T.C., Björn, L.O., 1986. Plants as light traps. Physiologia Plantarum 68, 704–708.
 https://doi.org/10.1111/j.1399-3054.1986.tb03421.x
- Wang, M., Fei, X., Zhang, Y., Chen, Z., Wang, X., Tsou, Y.J., Liu, D., Lu, X., 2018. Assessing Texture
 Features to Classify Coastal Wetland Vegetation from High Spatial Resolution Imagery Using
 Completed Local Binary Patterns (CLBP). Remote Sensing . https://doi.org/10.3390/rs10050778
- Wang, S., Liu, L., Qu, L., Yu, C., Sun, Y., Gao, F., Dong, J., 2019. Accurate Ulva prolifera regions
 extraction of UAV images with superpixel and CNNs for ocean environment monitoring.
 Neurocomputing 348, 158–168. https://doi.org/https://doi.org/10.1016/j.neucom.2018.06.088
- Weidmann, F., Jäger, J., Reus, G., Schultz, S.T., Kruschel, C., Wolff, V., Fricke-Neuderth, K., 2019. A
 Closer Look at Seagrass Meadows: Semantic Segmentation for Visual Coverage Estimation, in:
 OCEANS 2019 Marseille. pp. 1–6. https://doi.org/10.1109/OCEANSE.2019.8867064
- Wilson, K.L., Skinner, M.A., Lotze, H.K., 2019. Eelgrass (Zostera marina) and benthic habitat mapping
 in Atlantic Canada using high-resolution SPOT 6/7 satellite imagery. Estuarine, Coastal and Shelf
 Science 226, 106292. https://doi.org/10.1016/j.ecss.2019.106292
- Xing, Q., An, D., Zheng, X., Wei, Z., Wang, X., Li, L., Tian, L., Chen, J., 2019. Monitoring seaweed
 aquaculture in the Yellow Sea with multiple sensors for managing the disaster of macroalgal
 blooms. Remote Sensing of Environment 231, 111279.
 https://doi.org/https://doi.org/10.1016/j.rse.2019.111279
- Zacharias, M., Niemann, O., Borstad, G., 1992. An Assessment and Classification of a Multispectral
 Bandset for the Remote Sensing of Intertidal Seaweeds. Canadian Journal of Remote Sensing 18,
 263–274. https://doi.org/10.1080/07038992.1992.10855331
- Zhou, Y., Wang, J., Li, B., Meng, Q., Rocco, E., Saiani, A., 2019. Underwater scene segmentation by deep neural network.



This is Accepted Manuscript version is deposited under the terms of the Creative Commons Attribution-NonCommercial-NoDerivatives License (http://creativecommons.org/licenses/by-nc-nd/4.0/), which permits non-commercial re-use, distribution, and reproduction in any medium, provided the original work is properly cited, and is not altered, transformed, or built upon in any way.