

Journal Pre-proof

Deciphering the functional importance of comammox vs. canonical ammonia oxidisers in nitrification and N₂O emissions in acidic agricultural soils

Che Tan, Chang Yin, Lei Zhang, Yu Zeng, Cécile Gubry-Rangin, Hao Chen, Zixiang Gao, Hongyun Peng, Tingqiang Li, Yongchao Liang

PII: S0038-0717(24)00104-4

DOI: <https://doi.org/10.1016/j.soilbio.2024.109415>

Reference: SBB 109415

To appear in: *Soil Biology and Biochemistry*

Received Date: 19 November 2023

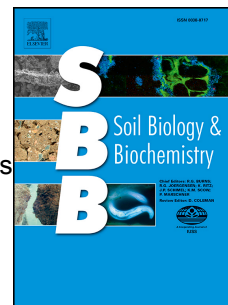
Revised Date: 22 February 2024

Accepted Date: 24 March 2024

Please cite this article as: Tan, C., Yin, C., Zhang, L., Zeng, Y., Gubry-Rangin, Cé., Chen, H., Gao, Z., Peng, H., Li, T., Liang, Y., Deciphering the functional importance of comammox vs. canonical ammonia oxidisers in nitrification and N₂O emissions in acidic agricultural soils, *Soil Biology and Biochemistry* (2024), doi: <https://doi.org/10.1016/j.soilbio.2024.109415>.

This is a PDF file of an article that has undergone enhancements after acceptance, such as the addition of a cover page and metadata, and formatting for readability, but it is not yet the definitive version of record. This version will undergo additional copyediting, typesetting and review before it is published in its final form, but we are providing this version to give early visibility of the article. Please note that, during the production process, errors may be discovered which could affect the content, and all legal disclaimers that apply to the journal pertain.

© 2024 Published by Elsevier Ltd.



1 **Deciphering the functional importance of comammox vs. canonical ammonia oxidisers in**
2 **nitrification and N₂O emissions in acidic agricultural soils**

3

4 Che Tan¹, Chang Yin², Lei Zhang¹, Yu Zeng¹, Cécile Gubry-Rangin³, Hao Chen¹, Zixiang Gao¹,
5 Hongyun Peng¹, Tingqiang Li¹, Yongchao Liang^{1,*}

6

7 ¹ Key Laboratory of Environment Remediation and Ecological Health, Ministry of Education,
8 College of Environmental & Resource Sciences, Zhejiang University, Hangzhou 310058, China

9 ² Institute of Environment, Resource, Soil and Fertilizer, Zhejiang Academy of Agricultural
10 Sciences, Hangzhou 310021, China

11 ³ School of Biological Sciences, University of Aberdeen, Cruickshank Building, Aberdeen AB24
12 3UU, UK

13 *Corresponding author: Prof. Yongchao Liang (Orcid ID: 0000-0002-0501-2823)

14 Address: College of Environmental and Resource Sciences, Zhejiang University, 866 Yuhangtang
15 Road, Hangzhou, Zhejiang, China

16 Tel: +86-57188982073

17 Fax: +86-57188982907

18 E-mail: ycliang@zju.edu.cn

19 **Abstract**

20 The discovery of comammox *Nitrospira* has altered our perception of the nitrogen biogeochemical
21 cycle. However, their functional importance compared to canonical ammonia oxidisers (i.e.,
22 ammonia-oxidising bacteria (AOB) and archaea (AOA)) in agricultural soils remains elusive,
23 especially in acidic soils. Here, we assessed the functional importance of these functional guilds
24 in nitrification and nitrous oxide (N₂O) emissions in three acidic agricultural soils by using a range
25 of nitrification inhibitors (acetylene, 3,4-dimethylpyrazole phosphate (DMPP) and different
26 concentrations of 1-octyne) and monitored their community assemblage and population dynamics.
27 The sensitivity of comammox *Nitrospira* clade A to 1-octyne varied across soils, highlighting that
28 the inappropriate use of 1-octyne can lead to misestimation of comammox activity. AOA were key
29 NH₃ oxidisers in the three soils, while AOB also contributed significantly to nitrification in one
30 soil. In contrast, comammox *Nitrospira* always played a minor role in ammonia oxidation and N₂O
31 emissions, likely due to their low abundances, restricted cellular kinetic properties and N₂O
32 production mechanisms. Together, this study demonstrates that comammox *Nitrospira* play a less
33 important role in ammonia oxidation and N₂O production in acidic agricultural soils than AOA
34 and AOB, thereby providing important novel insights into the mitigation of nitrogen fertiliser loss
35 and N₂O emissions.

36

37 **Keywords:** *Comammox Nitrospira*; *Nitrous oxide*; *1-Octyne*; *3,4-Dimethylpyrazole phosphate*
38 *(DMPP)*; *Acid soils*.

39 1. Introduction

40 The ongoing discovery of microorganisms has been refining our knowledge of the nitrogen
41 biogeochemical cycle (Kuypers et al., 2018; Wu et al., 2021). Among these, comammox *Nitrospira*
42 are of particular interest since they are equipped with all the enzymes needed for ammonia and
43 nitrite oxidation, enabling the oxidation of ammonia (NH_3) to nitrate (NO_3^-) in a single cell (Daims
44 et al., 2015; van Kessel et al., 2015). This significantly differs from their canonical counterparts
45 (i.e., ammonia-oxidising archaea (AOA) and ammonia-oxidising bacteria (AOB)) in that the latter
46 two guilds catalyse the conversion of NH_3 to nitrite (NO_2^-), the first step of nitrification.
47 Meanwhile, mounting research suggests that these three guilds generally coexist in agricultural
48 soils (Pjevac et al., 2017; Orellana et al., 2018; Xu et al., 2020), actively contributing to
49 nitrification and likely nitrous oxide (N_2O) emissions (Hink et al., 2017; Wang et al., 2019b; Huang
50 et al., 2021), a long-lived and ozone-depleting greenhouse gas (Ravishankara et al., 2009; Prather
51 et al., 2015). In many cases, comammox *Nitrospira* outnumbered AOA and AOB (Li et al., 2019;
52 Hu et al., 2021). These findings raise the speculation that comammox *Nitrospira* have an
53 underappreciated role in the nitrogen cycling of agricultural soils. However, the ecological
54 significance of comammox vs. canonical NH_3 oxidisers for NH_3 oxidation and N_2O emissions in
55 agricultural soils is rarely empirically evidenced (Tan et al., 2022; Jiang et al., 2023).

56 The ecological importance of comammox vs. canonical NH_3 oxidisers in nitrification is
57 expected to depend on the selected soils, as these guilds have been found to show distinct niche
58 preferences (i.e., niche differentiation) for edaphic properties such as moisture, organic matter,
59 nitrogen availability, and pH (Hatzenpichler, 2012; Prosser and Nicol, 2012). Of these, the
60 overarching role of soil pH in driving the assemblage and activities of AOA and AOB has been

61 extensively studied. AOA often prefer to flourish in oligotrophic habitats with low pH, while AOB
62 normally dominate nitrification in nitrogen-rich neutral and alkaline environments (Nicol et al.,
63 2008; He et al., 2012), even if several studies reported the dominance of AOB in nitrification of
64 some acidic soils (Dai et al., 2017; Huang et al., 2018; Lin et al., 2018; Zhang et al., 2019). In
65 contrast, the pH preference of comammox *Nitrospira* is still inconclusive, and multiple lines of
66 evidence support its activity in both high and low pH soils (Wang et al., 2019b; Zhao et al., 2020b;
67 Hu et al., 2021; Yuan et al., 2021; Tan et al., 2022). These observations hint at a complex scenario
68 in acidic soils, where comammox *Nitrospira* and AOB may play substantial roles in nitrification
69 and associated N₂O emissions. Consequently, there is an urgent need to reassess the relative
70 contribution of NH₃ oxidisers, particularly comammox *Nitrospira* and AOB, to nitrification and
71 N₂O production in diverse acidic soils.

72 The niche differentiation among AOA, AOB and comammox *Nitrospira* is primarily driven
73 by their distinct metabolic and physiological characteristics. AOB typically exhibit lower affinity
74 for NH₃ and higher NH₃ tolerance and maximum oxidation rate (V_{\max}) than AOA and comammox
75 *Nitrospira* (Jung et al., 2022), reflecting a copiotrophic lifestyle. AOA possess a surprising
76 variability of cellular kinetic properties (Jung et al., 2022), while comammox *Nitrospira* display
77 extremely high NH₃ affinity, high growth yield, and low maximum oxidation rate (Kits et al., 2017;
78 Sakoula et al., 2021), which are typical characteristics of an oligotrophic lifestyle. Recently,
79 genome and *in vitro* analyses further indicated that comammox *Nitrospira* may have a broad
80 ecological niche and thrive in high N-input agricultural soils (Li et al., 2023). In addition, the
81 mechanisms of N₂O production differ substantially among these guilds. In comparison with the
82 enzymatic pathways in AOB (Shaw et al., 2006; Zhu et al., 2013; Prosser et al., 2020), the

83 comammox bacterium *Nitrospira inopinata* only produces N₂O via the abiotic conversion of
84 hydroxylamine (NH₂OH) (Kits et al., 2019), while AOA-associated N₂O production is derived
85 from hybrid formation and NH₂OH oxidation (Stieglmeier et al., 2014; Kozłowski et al., 2016;
86 Wan et al., 2023). Consequently, the N₂O yields of comammox *Nitrospira* and AOA are much
87 lower than those of AOB (Hink et al., 2017; Kits et al., 2019; Han et al., 2021). Altogether, we
88 hypothesised that comammox *Nitrospira* and AOA dominate NH₃ oxidation in acidic soils but
89 produce less N₂O than AOB.

90 The ecological significance of these functional guilds was rarely explicitly assessed until the
91 emergence of specific inhibitors such as 1-octyne, 2-phenyl-4,4,5,5-tetramethylimidazoline-1-
92 oxyl-3-oxide (PTIO), 3,4-dimethylpyrazole phosphate (DMPP), simvastatin, and chlorate (Taylor
93 et al., 2013; Martens-Habbena et al., 2015; Papadopoulou et al., 2020; Zhao et al., 2020a; Wang et
94 al., 2021). Recently, based on the differential response patterns of NH₃ oxidisers to 1-octyne and
95 DMPP, we developed and tested a novel method to parse out the contribution of comammox to
96 nitrification and N₂O production (Tan et al., 2022), demonstrating that comammox played a minor
97 role in N₂O production and nitrification in an alkaline soil. In contrast, comammox were found to
98 cause significant NH₃ fertiliser loss comparable in magnitude to AOB in a global-scale survey
99 using the same approach (Jiang et al., 2023). The specific inhibition efficacy of 1-octyne has been
100 recently questioned, with several studies noting that 1-octyne can also inhibit comammox
101 *Nitrospira* (Taylor et al., 2017; Lin et al., 2023). In addition, the ability of 1-octyne to inhibit AOB
102 growth may be limited due to restricted diffusion in some soils (Yin et al., 2021). To prevent future
103 misevaluation of the functional importance of NH₃ oxidisers, it is imperative to re-evaluate the
104 inhibitory effect of 1-octyne on comammox *Nitrospira* and AOB in a range of soils, especially in

105 acidic soils where comammox *Nitrospira* are active.

106 Therefore, the main objectives of this study were to (i) decipher the contribution of
107 comammox *Nitrospira* vs. AOA and AOB to NH₃ oxidation and N₂O emissions in acidic
108 agricultural soils and (ii) reassess the inhibitory effect of different concentrations of 1-octyne on
109 NH₃ oxidisers. For this, three different types of acidic agricultural soils were selected, and soil
110 microcosms were conducted using a combined inhibitor method (acetylene, 1-octyne, and DMPP).
111 The shift in the *amoA* gene abundance of NH₃ oxidisers was monitored by qPCR, while the
112 community composition of canonical NH₃ oxidisers (AOA and AOB) and comammox *Nitrospira*
113 was assessed by Illumina MiSeq sequencing.

114 **2. Materials and methods**

115 **2.1 Study sites and soil sampling**

116 The sampling sites, located in the southern region of China, were representative agricultural
117 fields with distinct soil characteristics (Fig. S1). The specific locations were Xiantao city (30°32'N,
118 113°46'E) in Hubei Province (HB), Lingao County (19°94'N, 109°74'E) in Hainan Province (HN),
119 and Guiyang city (26°44'N, 106°53'E) in Guizhou Province (GZ). In May 2022, three 100 m² plots
120 were randomly selected at each site. Five soil cores were randomly taken from each plot's top layer
121 (0–20 cm in depth) and combined. The samples were sieved (2 mm) to remove plant debris and
122 stones and then mixed to generate each site's final composite soil samples. The collected soils were
123 immediately transported to the laboratory on ice. The composite samples were divided into two
124 parts: one subsample was used for microcosm experiments to determine microbial activity, and the
125 other was air-dried for physicochemical analysis. Climate variables for each site, including mean

126 annual temperature (MAT) and mean annual precipitation (MAP), were obtained from the
127 WorldClim database (www.worldclim.org). Details of the study sites and soil physicochemical
128 properties are provided in Table S1.

129 ***2.2 Microcosm incubation using the optimised combined inhibitor method***

130 Preliminary experiments showed a decrease in N₂O emissions in 1-octyne-inhibited
131 microcosms as the concentration of 1-octyne increased. We proposed two possible explanations
132 for this phenomenon: 1) Low concentrations of 1-octyne only partially inhibited AOB activity, and
133 as the concentration of 1-octyne increased, AOB activity was completely inhibited; 2) High
134 concentrations of 1-octyne partially inhibited the activity of comammox *Nitrospira*. The
135 illustration is presented in Fig. S2. To test the above hypotheses, we incorporated a range of 1-
136 octyne concentrations (see below) into the previously established combined inhibitor method (Tan
137 et al., 2022).

138 Soil microcosms were established in 125-ml serum bottles filled with ~18 g of homogeneous
139 fresh soil (equivalent to 15 g of dry soil). The bottles were sealed with butyl rubber stoppers and
140 aluminium caps and preincubated at 25 °C in the dark for seven days with ventilation every three
141 days. Following preincubation, the soil moisture content was adjusted to 50% WHC (water holding
142 capacity) using an ammonium sulfate solution (150 µg (NH₄)₂SO₄-N g⁻¹ soil_{dw}) in the absence or
143 presence of inhibitors acetylene (0.1%, v/v), 1-octyne (0.01%, 0.015%, 0.02%, 0.03%, 0.04% and
144 0.05%, v/v), or DMPP (1.5% of added NH₄⁺-N). Microcosms were then incubated in the dark at
145 25 °C for 21 days, with aeration on Days 2, 5, 8, 11, 14, 17 and 21 to maintain oxic conditions.
146 After resealing, microcosms were evacuated using a vacuum pump and replenished with 125 ml

147 of air, air-diluted acetylene (0.1%, v/v), or air-diluted 1-octyne (0.01%, 0.015%, 0.02%, 0.03%,
148 0.04% and 0.05%, v/v) to maintain the inhibitor concentrations. Gas samples (12 ml) were taken
149 before opening the microcosms using a syringe and injected into pre-evacuated 12-ml glass vials
150 (Labco, Lampeter, UK) to determine N₂O and CO₂ concentrations. Triplicate microcosms were
151 destructively sampled on Days 0, 5, 11 and 21, and the soil samples were immediately divided into
152 two parts and stored at -20 °C for chemical analysis (storage ≤ 2 weeks) and at -80 °C for
153 molecular analysis (storage ≤ 4 weeks).

154 ***2.3 Physicochemical analysis of soil and gas samples***

155 Soil texture was measured using the laser diffraction method with a Mastersizer 2000
156 (Malvern, Worcestershire, UK). Soil moisture content was determined by measuring the weight
157 loss after drying at 105 °C for 24 h. Soil pH was measured using an FE28-Standard pH meter
158 (Mettler Toledo, Shanghai, China) with a soil-to-water ratio of 1:2.5 (w/v). Extractable NH₄⁺ and
159 combined NO₂⁻ and NO₃⁻ were measured using the indophenol blue and VCl₃/Griess methods
160 (Kandeler and Gerber, 1988; Hink et al., 2018), respectively, after 6 g of wet soil was extracted
161 with 30 ml of 2 M KCl. The detailed procedures were described elsewhere (Yin et al., 2021).
162 Results of combined NO₂⁻ and NO₃⁻ analysis are expressed as NO₃⁻ concentration, as NO₂⁻ was
163 below the detection level (0.15 µg g⁻¹) in all samples. Soil organic carbon (SOC) was measured
164 by dichromate oxidation. Soil available phosphorus (AP) was determined by the molybdenum blue
165 colorimetry after extraction with 0.5 M NaHCO₃. Soil available potassium (AK) was extracted
166 with 1 M CH₃COONH₄ (pH 7.0) and determined with a flame photometer (Inesa Instrument,
167 Shanghai, China). The N₂O concentrations in the gas samples were measured using Trace1300 and

168 1310 gas chromatographs equipped with a flame ionization detector (FID) and an electron capture
169 detector (ECD) (Thermo Fisher Scientific, Rodano, Italy).

170 **2.4 Rates, relative contributions and N₂O yield of ammonia oxidisers**

171 Potential NH₃ oxidation rates for each treatment were calculated by performing a linear
172 regression of NH₄⁺ consumption over time, with a coefficient of determination (R^2) > 0.89 for all
173 microcosms except those treated with acetylene. Potential N₂O production rates were determined
174 by conducting a linear regression of N₂O accumulation over time, with R^2 > 0.75 for all
175 microcosms.

176 As 1-octyne specifically inhibits the activity of AOB at specific concentrations (Taylor et al.,
177 2013; Li et al., 2019), the AOB contribution can be assessed by subtracting the NH₃ oxidation (or
178 N₂O production) rate of the 1-octyne-treated microcosms from the rate of non-inhibited
179 microcosms. Since DMPP inhibits the activity of both AOB and comammox *Nitrospira*
180 (Papadopoulou et al., 2020; Zhou et al., 2020), AOA is the only NH₃ oxidiser consuming NH₄⁺ or
181 producing N₂O in DMPP-treated microcosms. Therefore, the AOA rate can be obtained by
182 subtracting the rate of acetylene-treated microcosms, which represents the non-NH₃ oxidation
183 process rate, from the rate of DMPP-treated microcosms. The comammox rate can be calculated
184 by subtracting the rate of DMPP-treated microcosms from the rate of 1-octyne-treated microcosms.
185 We only selected the 1-octyne concentrations that achieved the ideal inhibitory effect (completely
186 inhibiting AOB but not inhibiting comammox *Nitrospira* clade A growth) and calculated the
187 average rate of NH₃ oxidisers and their relative contribution to total NH₃ oxidation and NH₃
188 oxidation-derived N₂O production.

189 The relative contributions of NH₃ oxidisers to NH₄⁺ consumption or N₂O production were
 190 calculated by dividing the potential rates of NH₃ oxidisers by the total NH₃ oxidation rate or the
 191 total NH₃ oxidation-related N₂O production rate.

$$192 \quad Contribution_{NH_4^+ \text{ or } N_2O} (Comammox \text{ or } AOA \text{ or } AOB) = \frac{Rate_{(Comammox \text{ or } AOA \text{ or } AOB)}}{Rate_{(Comammox + AOA + AOB)}} \times 100\%$$

193 The N₂O yields of AOA, AOB, and comammox *Nitrospira* were calculated as the amount of
 194 N₂O-N produced per NH₄⁺-N consumed, as shown in the following equation:

$$195 \quad N_2O \text{ yield } (\%) = \frac{Rate_{N_2O \text{ production}}}{Rate_{NH_4^+ \text{ consumption}}} \times 100\%$$

196 **2.5 Nucleic acid extraction and quantitative PCR**

197 Nucleic acid was extracted from 0.5 g of wet soil, equivalent to 0.39 g (HB), 0.36 g (HN),
 198 and 0.35 g (GZ) of dry soil, using the Fast DNA SPIN Kit for Soil (MP Biomedicals, OH, USA)
 199 according to the manufacturer's instructions. DNA quality and quantity were assessed using a
 200 Colibri Microvolume Spectrometer (Titertek Berthold, Pforzheim, Germany). The abundances of
 201 AOA, AOB, comammox *Nitrospira* clade A and B *amoA* genes were quantified by specific qPCR
 202 assays on a LightCycler® 480II (Roche Diagnostics, Rotkreuz, Switzerland). Details of the primer
 203 sets and qPCR conditions are provided in Table S2. The *R*² values were >0.99 for all genes, and
 204 the amplification efficiencies were 79.4 to 91.0% for AOA, 90.4 to 96.5% for AOB, 79.8 to 86.9%
 205 for comammox *Nitrospira* clade A and 78.6% for comammox *Nitrospira* clade B. The
 206 amplification specificity was verified by melting curve analysis and agarose gel electrophoresis.
 207 As we could not amplify the comammox *Nitrospira* clade B *amoA* gene in the HN and GZ soil
 208 samples, we excluded clade B from the downstream analysis of these samples.

209 **2.6 Illumina sequencing and phylogenetic analysis**

210 DNA samples were sent to the Illumina MiSeq sequencing platform (Shanghai Majorbio Bio-
211 Pharm Technology Co., Ltd., China) for high-throughput sequencing. The V4 region of the
212 archaeal and bacterial 16S rRNA genes was amplified using the barcoded primer pair
213 515FmodF/806RmodR (Walters et al., 2016). In addition, the comammox *Nitrospira* clade A and
214 clade B *amoA* genes were amplified using the primer pairs CA377f/C576r and CB377f/C576r
215 (Jiang et al., 2020), respectively. The obtained raw sequences were quality-filtered and checked
216 for chimaeras using Quantitative Insights into Microbial Ecology (QIIME) 2 with default criteria
217 and UCHIME 4.2 (Edgar et al., 2011; Bolyen et al., 2019). The high-quality sequences were then
218 classified into operational taxonomic units (OTUs) based on 97% similarity using Uparse 7.0.1090
219 (Edgar, 2013). The 16S rRNA OTUs were classified using RDP Classifier v. 2.11 against the
220 SILVA database v. 138 (Quast et al., 2012) and comammox *Nitrospira* OTUs were classified using
221 the nucleotide database (nt_v20210917) of the NCBI database. Representative AOA, AOB, and
222 comammox *Nitrospira* sequences with identity values <90% from BLAST alignment or relative
223 abundance <0.05% were excluded from the downstream analysis.

224 **2.7 Statistical analysis**

225 Statistical analysis was performed using R (R Core Team, 2021) by comparing the means
226 from triplicate samples. Differences in *amoA* gene abundance were tested using a factorial two-
227 way ANOVA with inhibitors and incubation time as fixed factors, followed by a Tukey HSD *post*
228 *hoc* test. Temporal differences in NH_4^+ , NO_3^- and N_2O concentrations were determined by
229 comparing the slopes of the linear models using one-way ANOVA, followed by a Tukey HSD *post*

230 *hoc* test. Differences between the N₂O yields of AOA, AOB, and comammox *Nitrospira* were
231 tested by one-way ANOVA. The absolute abundance of comammox *Nitrospira* OTUs was
232 calculated as the comammox *Nitrospira* qPCR values multiplied by the corresponding relative
233 abundance obtained from MiSeq sequencing. Differences in the relative and absolute abundance
234 of OTUs between treatments were determined using one-way ANOVA.

235 **3. Results**

236 **3.1 Sensitivity of ammonia oxidisers to inhibitors in different soils**

237 In the absence of inhibitor, the *amoA* gene abundance of AOA, AOB, and comammox
238 *Nitrospira* increased over time in response to NH₄⁺ addition in HB and HN soils (Fig. 1a and b).
239 In GZ soil, NH₄⁺ amendment without inhibitor supplementation stimulated AOA growth but led
240 to a decreased abundance of AOB ($P < 0.001$) and comammox *Nitrospira* clade A ($P < 0.001$) over
241 time (Fig. 1c).

242 Acetylene inhibited the growth of AOA, AOB, and comammox *Nitrospira* in all soils during
243 incubation (compared with initial abundance, two-way ANOVA; Table S4). In HN and GZ soils,
244 AOA *amoA* gene abundance was significantly greater in the presence of 1-octyne or DMPP than
245 in non-inhibited microcosms at the end of incubation (Table S5). 1-Octyne and DMPP application
246 prevented the growth of AOB in all soils. DMPP inhibited the growth of comammox *Nitrospira*
247 clade A, and different concentrations of 1-octyne differentially affected its growth in different soils.
248 In HB soil, the growth of comammox *Nitrospira* clade A was unaffected by the presence of 0.01%
249 and 0.015% 1-octyne, but higher concentrations of 1-octyne inhibited their growth (Fig. 1a; Table
250 S6). In HN soil, none of the 1-octyne concentrations tested (0.01–0.05%) inhibited the growth of

251 comammox *Nitrospira* clade A. Instead, their growth was stimulated in the presence of 0.01% and
252 0.04% 1-octyne compared with that in non-inhibited microcosms (Fig. 1b; Table S6). In GZ soil,
253 the abundance of comammox *Nitrospira* clade A decreased significantly ($P < 0.001$) in the
254 presence of 0.01–0.05% 1-octyne, but that of the *amoA* gene at the last sampling point was greater
255 than that in non-inhibited microcosms (Fig. 1c; Table S6).

256 **3.2 Dynamics of nitrification and cumulative N₂O emissions**

257 The differential growth patterns of NH₃ oxidisers in the absence or presence of inhibitors
258 were ultimately reflected in the nitrification dynamics and N₂O emissions. In the absence of
259 inhibitors, NH₄⁺ was oxidised into NO₃⁻ at rates of ~5.4, 7.6 and 6.8 μg N g⁻¹ soil_{dw} d⁻¹ in HB,
260 HN and GZ soils, respectively (Fig. 2). In acetylene-treated microcosms, mineralisation-derived
261 NH₄⁺ did not accumulate over time, but NO₃⁻ concentrations decreased in HB and HN soils (Fig.
262 2a and b, $P < 0.05$), which may have resulted from microbial assimilation or denitrification.
263 Nevertheless, N₂O produced via denitrification was negligible under oxic conditions, as NO₃⁻
264 addition did not affect N₂O emissions compared with those in acetylene-treated microcosms (Fig.
265 S3).

266 Inhibition of AOB and comammox *Nitrospira* activity by DMPP reduced the nitrification
267 rates by ~20%, 50% and 30% in HB, HN and GZ soils, respectively, and N₂O accumulation rates
268 were also reduced by ~50%, 80% and 40% in HB, HN and GZ soils (Fig. 2). This suggests that
269 AOA dominated NH₃ oxidation and associated N₂O production in HB and GZ soils, while N₂O
270 emissions in HN soil resulted mainly from AOB and comammox *Nitrospira* activities. No
271 significant difference in the nitrification and N₂O emissions rates was observed between the

272 different 1-octyne treatments (0.01–0.05%) in any of the soils tested (Fig. 2).

273 **3.3 Ammonia oxidation and N₂O production rates, relative contributions and N₂O yields of**
 274 **ammonia oxidisers**

275 In HN and GZ soils, the application of 1-octyne and DMPP did not significantly affect the
 276 growth of comammox *Nitrospira* and AOA (increases in *amoA* gene abundance) compared to non-
 277 inhibited microcosms during incubation for 11 days (Table S5–6), suggesting that NH₃ oxidation
 278 and N₂O production were also unaffected; thus, we calculated the NH₃ oxidation and N₂O
 279 production rates of NH₃ oxidisers within the initial 11 days of incubation. A similar assumption
 280 was made for HB soil after 21 days of incubation, enabling calculation of the NH₃ oxidation and
 281 N₂O production rates of NH₃ oxidisers throughout the entire period of incubation (21 days).

282 Comammox *Nitrospira* contributed to a small extent to NH₃ oxidation in the three acidic
 283 agricultural soils (3.3 to 8.5% of the total NH₃ oxidation rate) (Fig. 3a). The NH₃ oxidation rate of
 284 comammox was not significantly different among HB ($0.17 \pm 0.39 \mu\text{g N g}^{-1} \text{ soil}_{\text{dw}} \text{ d}^{-1}$), HN (0.39
 285 $\pm 0.29 \mu\text{g N g}^{-1} \text{ soil}_{\text{dw}} \text{ d}^{-1}$) and GZ soils ($0.50 \pm 0.34 \mu\text{g N g}^{-1} \text{ soil}_{\text{dw}} \text{ d}^{-1}$). AOA dominated NH₃
 286 oxidation in HB ($4.41 \pm 0.29 \mu\text{g N g}^{-1} \text{ soil}_{\text{dw}} \text{ d}^{-1}$, ~83.2%) and GZ soils ($3.99 \pm 0.34 \mu\text{g N g}^{-1}$
 287 $\text{soil}_{\text{dw}} \text{ d}^{-1}$, ~67.4%). The NH₃ oxidation rate of AOB varied significantly between soils ($P < 0.001$);
 288 it was similar to the AOA rate in HN soil (AOB: $3.29 \pm 0.34 \mu\text{g N g}^{-1} \text{ soil}_{\text{dw}} \text{ d}^{-1}$, ~47.0% and AOA:
 289 $3.32 \pm 0.29 \mu\text{g N g}^{-1} \text{ soil}_{\text{dw}} \text{ d}^{-1}$, ~47.5%) but was significantly lower than the AOA rate in HB
 290 ($0.71 \pm 0.43 \mu\text{g N g}^{-1} \text{ soil}_{\text{dw}} \text{ d}^{-1}$, $P < 0.001$) and GZ ($1.43 \pm 0.34 \mu\text{g N g}^{-1} \text{ soil}_{\text{dw}} \text{ d}^{-1}$, $P < 0.001$)
 291 soils. These results indicate that AOA are key NH₃ oxidisers in all three acidic agricultural soils
 292 tested, with AOB also contributing significantly in one of these soils, while comammox *Nitrospira*

293 only play a limited role.

294 Consistent with their limited contribution to NH₃ oxidation, comammox *Nitrospira*
 295 contributed the lowest N₂O production among the three NH₃ oxidiser groups, producing 3.1 to
 296 13.0% of NH₃ oxidation-associated N₂O emissions across the three soils (Fig. 3b). The N₂O
 297 production rates of comammox in HB ($0.21 \pm 0.48 \text{ ng N g}^{-1} \text{ soil}_{\text{dw}} \text{ d}^{-1}$), HN ($0.76 \pm 0.04 \text{ ng N g}^{-1}$
 298 $\text{soil}_{\text{dw}} \text{ d}^{-1}$) and GZ soils ($0.76 \pm 0.19 \text{ ng N g}^{-1} \text{ soil}_{\text{dw}} \text{ d}^{-1}$) were significantly lower than those of
 299 AOB (HB, $3.27 \pm 0.57 \text{ ng N g}^{-1} \text{ soil}_{\text{dw}} \text{ d}^{-1}$, $P < 0.01$; HN, $4.32 \pm 0.19 \text{ ng N g}^{-1} \text{ soil}_{\text{dw}} \text{ d}^{-1}$, $P <$
 300 0.001 ; GZ, $2.93 \pm 0.19 \text{ ng N g}^{-1} \text{ soil}_{\text{dw}} \text{ d}^{-1}$, $P < 0.001$). AOB played a key role in N₂O production
 301 (32.8 to 73.5%), especially in HN soil, where AOB dominated N₂O production, producing ~73.5%
 302 of the NH₃ oxidation-associated N₂O generated (Fig. 3b). The AOA-associated N₂O production
 303 rate varied significantly among the soils ($P < 0.001$). In HN soil, the AOA rate ($0.79 \pm 0.02 \text{ ng N}$
 304 $\text{g}^{-1} \text{ soil}_{\text{dw}} \text{ d}^{-1}$) was similar to the comammox rate, but it increased to $3.33 \pm 0.33 \text{ ng N g}^{-1} \text{ soil}_{\text{dw}}$
 305 d^{-1} in HB soil and $5.25 \pm 0.14 \text{ ng N g}^{-1} \text{ soil}_{\text{dw}} \text{ d}^{-1}$ in GZ soil.

306 N₂O yield, which incorporates NH₃ oxidation activity and N₂O production, enabled the
 307 assessment of the N₂O footprint of the three guilds of NH₃ oxidisers. N₂O yields for AOA ($P <$
 308 0.001), AOB ($P < 0.001$) and comammox *Nitrospira* ($P < 0.05$) varied significantly between soil
 309 types. N₂O yields for AOB (0.13 to 0.46%) were higher than those for AOA (0.02 to 0.13%) and
 310 comammox *Nitrospira* (0.12 to 0.22%) (Fig. 3c).

311 **3.4 Temporal changes in ammonia oxidiser community composition**

312 The AOA community in HB and HN soils was dominated by OTUs within *Nitrosocosmicus*
 313 and *Nitrososphaera*, while GZ soil also contained a large proportion of *Ca. Nitrosotaleales* OTUs

314 (Fig. 4a). Most of the abundant OTUs responded to the soil treatments, with different relative
315 abundances under the different inhibitor conditions. However, the common OTUs across the three
316 soils (such as OTU8971 or OTU9644) did not respond similarly to the treatments across the three
317 soils.

318 Two of the soils (HB and HN) showed a diverse AOB community containing OTUs affiliated
319 with *Nitrosomonas*, *Nitrospira*, the MND1 cluster and *Nitrosococcus*, while the GZ soil was less
320 diverse (no *Nitrosomonas* and *Nitrosococcus* OTUs) (Fig. 4b). In contrast to AOA, many OTUs
321 did not respond to the different treatments. Nevertheless, the relative abundance of some OTUs
322 (e.g., *Nitrospira* or *Nitrosomonas*) increased significantly in the absence of inhibitor in HB and
323 HN soils ($P < 0.01$).

324 Four major phylogenetic clusters of comammox *Nitrospira* were detected in the tested soils,
325 including clade A.2.2, A.2.1, A.3 and B (Fig. 4c). Clade A.2.2 was the most abundant cluster in all
326 soils, accounting for 76.8%, 88.9% and 100% of the clade A sequences, respectively (Fig. 4c).
327 Within each soil, the inhibitor treatments only affected the absolute abundance of a restricted
328 number of OTUs. Similar to the other nitrifying guilds, the common OTUs across all soils did not
329 respond consistently to the inhibitor treatments. Specifically, OTU31 grew in 1-octyne-treated
330 microcosms of HB and HN soils but not in GZ soil; OTU10 grew in 1-octyne-treated microcosms
331 of HN soil but not in HB and GZ soils (Fig. S4). These results indicate that in addition to the
332 sensitivity to inhibitors, there are other factors affecting the growth of NH_3 oxidisers.

333 **4. Discussion**

334 This study demonstrates that comammox *Nitrospira* play a minor role in NH_3 oxidation and

335 N₂O emissions in all three fertilised acidic agricultural soils tested. Our results were consistent
336 with global-scale results, which show that comammox *Nitrospira* contribute ~11.7% to NH₃
337 oxidation-associated N₂O emissions (Jiang et al., 2023), indicating that comammox *Nitrospira* are
338 a minor but nonnegligible N₂O source in acidic agricultural soils. Comammox *Nitrospira* (0.12 to
339 0.22%) and AOA (0.02 to 0.13%) produced lower yields of N₂O than AOB (0.13 to 0.46%). These
340 results are generally consistent with previous pure culture, enrichment and microcosm studies
341 where yields of 0.07 to 0.20% and 0.05 to 0.09% were recorded for comammox *Nitrospira* and
342 AOA, respectively, while AOB were able to reach 0.51% (Stieglmeier et al., 2014; Hink et al.,
343 2017; Kits et al., 2019; Han et al., 2021). They are also aligned with current knowledge of the
344 mechanism of N₂O production in comammox *Nitrospira* and AOA, where production is restricted
345 to abiotic hybrid formation and NH₂OH oxidation (Stieglmeier et al., 2014; Kozłowski et al., 2016;
346 Kits et al., 2019; Wan et al., 2023), and in AOB, which can additionally produce N₂O via nitrifier
347 denitrification (Shaw et al., 2006). It should be noted that the mechanism for N₂O production in
348 comammox *Nitrospira* was derived from a single pure culture obtained from aquatic environments,
349 despite a vast phylogenetic diversity within comammox *Nitrospira* (Palomo et al., 2022). In the
350 tested soils, comammox *Nitrospira* exhibited higher N₂O yields compared to previous pure culture
351 studies (Kits et al., 2019; Han et al., 2021), suggesting that comammox *Nitrospira* may possess
352 other pathways for N₂O production.

353 The limited contribution of comammox *Nitrospira* to NH₃ oxidation and N₂O emissions
354 may result from their low abundance and cellular kinetic properties. In HB, HN, and GZ soils,
355 comammox *Nitrospira amoA* genes were 61-, 30-, and 6-fold lower than AOA *amoA* genes,
356 respectively. Despite AOB having a comparable or even lower abundance than comammox

357 *Nitrospira*, they contributed more to NH_3 oxidation and N_2O production. In addition to the low
358 abundance, comammox *Nitrospira* had lower cell-specific NH_3 oxidation rates (0.44 to 0.58 fmol
359 $\text{NH}_4^+\text{-N cell}^{-1} \text{ h}^{-1}$) than AOA (0.15 to 0.35 fmol $\text{NH}_4^+\text{-N cell}^{-1} \text{ h}^{-1}$) or AOB (0.36 to 15.80 fmol
360 $\text{NH}_4^+\text{-N cell}^{-1} \text{ h}^{-1}$). These cell-specific rates are in the same range as those previously published
361 for soils (Table S3), but comparisons of the cell-specific rates should be made with caution due to
362 different methodologies. Consistently, the cell-specific NH_3 oxidation rate of AOB can be >10
363 times greater than that of AOA and comammox *Nitrospira* (Prosser and Nicol, 2012). Comammox
364 *Nitrospira* were characterised as k-strategists adapted to slow growth in oligotrophic habitats
365 (Costa et al., 2006; Daims et al., 2015). They exhibited low nutrient and energy requirements at
366 low maximum growth rates (G_{max}) (Flynn et al., 2018), thus avoiding the need for rapid NH_3
367 oxidation. This is consistent with the lower V_{max} value for the comammox strain *N. inopinata* (~12
368 $\mu\text{mol N mg protein}^{-1} \text{ h}^{-1}$) compared with that for the AOB strains (38–84.2 $\mu\text{mol N mg protein}^{-1}$
369 h^{-1}) (Jung et al., 2022).

370 Our demonstration that the growth of comammox *Nitrospira* clade A was inhibited by higher
371 concentrations of 1-octyne in HB and GZ soils but stimulated in HN soil indicates that the
372 sensitivity of comammox *Nitrospira* to 1-octyne varies among soils. This can be attributed to the
373 differences in sensitivity among different comammox *Nitrospira* clade A phylotypes, given the
374 distinct comammox *Nitrospira* community composition in these soils. This hypothesis is
375 reinforced by the growth of comammox *Nitrospira* in neutral or alkaline soils but inhibition in
376 acidic soils, all supplemented with the same concentration of 1-octyne (Li et al., 2019; Tan et al.,
377 2022; Lin et al., 2023). Earlier works on AOA and AOB pure cultures show that different
378 phylotypes of AOA and AOB have different sensitivities to 1-octyne (Taylor et al., 2013; Taylor et

379 al., 2015), but there is currently limited information on the sensitivity of comammox pure cultures
380 to 1-octyne (Taylor et al., 2017). An alternative explanation, presumably, is the varying
381 degradation of 1-octyne in these soils. However, the stable and complete inhibition of AOB growth
382 by 1-octyne in all soils, along with the regular replenishment of 1-octyne, question the strength of
383 this argument. Although the inhibition of comammox *Nitrospira* growth by $\geq 0.02\%$ 1-octyne
384 in HB soil did not result in underestimation of the comammox rates (Fig. S5), possibly due to the
385 low rate hindering the detection of changes, the results might be different in some habitats with
386 high comammox activity. Therefore, when applying the combined inhibitor method in various soils,
387 it is essential to evaluate the effectiveness of inhibitors on NH_3 oxidisers by measuring the growth
388 of NH_3 oxidisers through their change in absolute abundance over time, as performed in some
389 previous studies (Hink et al., 2017; Tan et al., 2022; Krüger et al., 2023).

390 This study confirms that AOA dominate NH_3 oxidation in fertilised acidic agricultural soils,
391 while AOB also contribute significantly to NH_3 oxidation and NH_3 oxidation-related N_2O
392 production in one of the soils tested. Chemolithotrophic NH_3 oxidisers were the primary drivers
393 of NO_3^- and N_2O production in the tested soils, and the contribution of heterotrophic denitrifiers
394 and heterotrophic nitrifiers was found to be negligible (Fig. S3), likely due to the oxic conditions
395 and the limited availability of easily decomposable organic substrates (Bateman and Baggs, 2005;
396 Martikainen, 2022). One previous hypothesis explaining the dominance of AOA in acidic soils
397 was their high substrate affinity for NH_3 (Martens-Habbena et al., 2009; Jung et al., 2011), which
398 confers a competitive advantage in acidic environments with low NH_3 availability, but this
399 argument was questioned (Qin et al., 2024). In addition, most AOA sequences in HB and HN soils
400 were affiliated with the *Nitrosocosmicus* cluster, and several *Nitrosocosmicus* species were

401 previously demonstrated to show similar substrate affinity to AOB (Jung et al., 2022). The function
402 of V-type ATPases as proton pumps has been proposed, among others, as an adaptive mechanism
403 in acidic conditions. *Nitrosocosmicus* species encode V-type ATPase to thrive at low pH (Wang et
404 al., 2019a). This study also provides another piece of evidence that AOB can be important NH₃
405 oxidisers in acidic soils (Lin et al., 2021), contributing ~47.0% to NH₃ oxidation in HN soil,
406 potentially due to the activity of acidic AOB organisms (Aigle et al., 2019). *Nitrosomonas* and
407 *Nitrospira* represented two AOB phylotypes promoted by NH₄⁺ amendment in HN soil,
408 suggesting an important role of these phylotypes in nitrification. This is consistent with earlier
409 studies reporting that *Nitrosomonas communis* and *Nitrospira multiformis* drive nitrification in
410 acidic agricultural and forest soils, respectively (Huang et al., 2018; Yin et al., 2021).

411 5. Conclusion

412 Our study demonstrates that AOA dominate nitrifying activities in fertilised acidic
413 agricultural soils, followed by AOB, while comammox *Nitrospira* only play a minor role in NH₃
414 oxidation and N₂O production. It also robustly illustrates that the inhibitory effect of 1-octyne on
415 comammox *Nitrospira* clade A varies across soils, emphasising that the inappropriate use of 1-
416 octyne may lead to misestimation of NH₃ oxidiser activity. This is fundamental for future studies
417 using the combined inhibitor method to assess the functional importance of NH₃ oxidisers in
418 diverse soils. Therefore, this work opens avenues of research to reduce nitrogen fertiliser loss and
419 N₂O emissions by manipulating the role of comammox *Nitrospira* in cropland nitrification.

420 **Data availability**

421 The 16S rRNA sequence data obtained in this study have been deposited in the National Center
422 for Biotechnology Information (NCBI) database under Bioproject accession number
423 PRJNA975883. The *amoA* gene sequences of comammox *Nitrospira* have been uploaded to
424 Genbank with accession numbers OR062097–OR062210. The shapefile of China is publicly
425 obtained from <https://www.csgpc.org/list/254.html> (drawing approval no. GS(2020)4632). The
426 cropland cover map of China is provided by Yang and Huang (2021)
427 (<https://zenodo.org/record/5816591#.ZAWM3BVBy5c>).

428 **Acknowledgments**

429 This work was jointly supported by grants from the National Key Research and Development
430 Program of China (2018YFD0800202), the National Key Research and Development Program of
431 China (2017YFD0200707 & 2017YFD0200102), the Fundamental Research Funds for the Central
432 Universities (226-2023-00077) and Zhejiang University-Julong Ecological Environment R&D
433 Centre (2019-KYY-514106-0006).

434 **6. References**

- 435 Aigle, A., Prosser, J.I., Gubry-Rangin, C., 2019. The application of high-throughput sequencing technology to analysis
436 of *amoA* phylogeny and environmental niche specialisation of terrestrial bacterial ammonia-oxidisers.
437 *Environmental Microbiome* 14, 3.
- 438 Bateman, E.J., Baggs, E.M., 2005. Contributions of nitrification and denitrification to N₂O emissions from soils at
439 different water-filled pore space. *Biology and Fertility of Soils* 41, 379-388.
- 440 Bolyen, E., Rideout, J.R., Dillon, M.R., Bokulich, N.A., Abnet, C.C., Al-Ghalith, G.A., Alexander, H., Alm, E.J.,
441 Arumugam, M., Asnicar, F., Bai, Y., Bisanz, J.E., Bittinger, K., Brejnrod, A., Brislawn, C.J., Brown, C.T.,
442 Callahan, B.J., Caraballo-Rodríguez, A.M., Chase, J., Cope, E.K., Da Silva, R., Diener, C., Dorrestein, P.C.,
443 Douglas, G.M., Durall, D.M., Duvallet, C., Edwardson, C.F., Ernst, M., Estaki, M., Fouquier, J., Gauglitz,

- 444 J.M., Gibbons, S.M., Gibson, D.L., Gonzalez, A., Gorlick, K., Guo, J., Hillmann, B., Holmes, S., Holste, H.,
 445 Huttenhower, C., Huttley, G.A., Janssen, S., Jarmusch, A.K., Jiang, L., Kaehler, B.D., Kang, K.B., Keefe,
 446 C.R., Keim, P., Kelley, S.T., Knights, D., Koester, I., Kosciulek, T., Kreps, J., Langille, M.G.I., Lee, J., Ley,
 447 R., Liu, Y.X., Loftfield, E., Lozupone, C., Maher, M., Marotz, C., Martin, B.D., McDonald, D., McIver, L.J.,
 448 Melnik, A.V., Metcalf, J.L., Morgan, S.C., Morton, J.T., Naimey, A.T., Navas-Molina, J.A., Nothias, L.F.,
 449 Orchanian, S.B., Pearson, T., Peoples, S.L., Petras, D., Preuss, M.L., Pruesse, E., Rasmussen, L.B., Rivers,
 450 A., Robeson, M.S., Rosenthal, P., Segata, N., Shaffer, M., Shiffer, A., Sinha, R., Song, S.J., Spear, J.R.,
 451 Swafford, A.D., Thompson, L.R., Torres, P.J., Trinh, P., Tripathi, A., Turnbaugh, P.J., Ul-Hasan, S., van der
 452 Hooft, J.J.J., Vargas, F., Vázquez-Baeza, Y., Vogtmann, E., von Hippel, M., Walters, W., Wan, Y., Wang, M.,
 453 Warren, J., Weber, K.C., Williamson, C.H.D., Willis, A.D., Xu, Z.Z., Zaneveld, J.R., Zhang, Y., Zhu, Q.,
 454 Knight, R., Caporaso, J.G., 2019. Reproducible, interactive, scalable and extensible microbiome data science
 455 using QIIME 2. *Nature Biotechnology* 37, 852-857.
- 456 Costa, E., Pérez, J., Kreft, J.U., 2006. Why is metabolic labour divided in nitrification? *Trends in Microbiology* 14,
 457 213-219.
- 458 Dai, S., Liu, Q., Zhao, J., Zhang, J., 2017. Ecological niche differentiation of ammonia-oxidising archaea and bacteria
 459 in acidic soils due to land use change. *Soil Research* 56, 71-79.
- 460 Daims, H., Lebedeva, E.V., Pjevac, P., Han, P., Herbold, C., Albertsen, M., Jehmlich, N., Palatinszky, M., Vierheilig,
 461 J., Bulaev, A., Kirkegaard, R.H., von Bergen, M., Rattei, T., Bendinger, B., Nielsen, P.H., Wagner, M., 2015.
 462 Complete nitrification by *Nitrospira* bacteria. *Nature* 528, 504-509.
- 463 Edgar, R.C., 2013. UPARSE: highly accurate OTU sequences from microbial amplicon reads. *Nature Methods* 10,
 464 996-998.
- 465 Edgar, R.C., Haas, B.J., Clemente, J.C., Quince, C., Knight, R., 2011. UCHIME improves sensitivity and speed of
 466 chimera detection. *Bioinformatics* 27, 2194-2200.
- 467 Flynn, K.J., Skibinski, D.O.F., Lindemann, C., 2018. Effects of growth rate, cell size, motion, and elemental
 468 stoichiometry on nutrient transport kinetics. *PLoS Computational Biology* 14, e1006118.
- 469 Han, P., Wu, D., Sun, D., Zhao, M., Wang, M., Wen, T., Zhang, J., Hou, L., Liu, M., Klümper, U., Zheng, Y., Dong,
 470 H.P., Liang, X., Yin, G., 2021. N₂O and NO_y production by the comammox bacterium *Nitrospira inopinata*
 471 in comparison with canonical ammonia oxidizers. *Water Research* 190, 116728.
- 472 Hatzenpichler, R., 2012. Diversity, Physiology, and Niche Differentiation of Ammonia-Oxidizing Archaea. *Applied*
 473 *and Environmental Microbiology* 78, 7501-7510.
- 474 He, J.Z., Hu, H.W., Zhang, L.M., 2012. Current insights into the autotrophic thaumarchaeal ammonia oxidation in
 475 acidic soils. *Soil Biology and Biochemistry* 55, 146-154.
- 476 Hink, L., Gubry-Rangin, C., Nicol, G.W., Prosser, J.I., 2018. The consequences of niche and physiological
 477 differentiation of archaeal and bacterial ammonia oxidisers for nitrous oxide emissions. *The ISME Journal*
 478 12, 1084-1093.
- 479 Hink, L., Nicol, G.W., Prosser, J.I., 2017. Archaea produce lower yields of N₂O than bacteria during aerobic ammonia
 480 oxidation in soil. *Environmental Microbiology* 19, 4829-4837.
- 481 Hu, J., Zhao, Y., Yao, X., Wang, J., Zheng, P., Xi, C., Hu, B., 2021. Dominance of comammox *Nitrospira* in soil
 482 nitrification. *Science of the Total Environment* 780, 146558.
- 483 Huang, L., Chakrabarti, S., Cooper, J., Perez, A., John, S.M., Daroub, S.H., Martens-Habbena, W., 2021. Ammonia-
 484 oxidizing archaea are integral to nitrogen cycling in a highly fertile agricultural soil. *ISME Communications*
 485 1, 19.
- 486 Huang, X., Zhao, J., Su, J., Jia, Z., Shi, X., Wright, A.L., Zhu-Barker, X., Jiang, X., 2018. Neutrophilic bacteria are
 487 responsible for autotrophic ammonia oxidation in an acidic forest soil. *Soil Biology and Biochemistry* 119,

- 488 83-89.
- 489 Jiang, L., Yu, J., Wang, S., Wang, X., Schwark, L., Zhu, G., 2023. Complete ammonia oxidization in agricultural soils:
490 High ammonia fertilizer loss but low N₂O production. *Global Change Biology* 29, 1984-1997.
- 491 Jiang, R., Wang, J.G., Zhu, T., Zou, B., Wang, D.Q., Rhee, S.K., An, D., Ji, Z.Y., Quan, Z.X., Stams Alfons, J.M.,
492 2020. Use of Newly Designed Primers for Quantification of Complete Ammonia-Oxidizing (Comammox)
493 Bacterial Clades and Strict Nitrite Oxidizers in the Genus *Nitrospira*. *Applied and Environmental*
494 *Microbiology* 86, e01775-01720.
- 495 Jung, M.Y., Park, S.J., Min, D., Kim, J.S., Rijpstra, W.I.C., Sinninghe Damsté Jaap, S., Kim, G.J., Madsen Eugene,
496 L., Rhee, S.K., 2011. Enrichment and Characterization of an Autotrophic Ammonia-Oxidizing Archaeon of
497 Mesophilic Crenarchaeal Group I.1a from an Agricultural Soil. *Applied and Environmental Microbiology* 77,
498 8635-8647.
- 499 Jung, M.Y., Sedlacek, C.J., Kits, K.D., Mueller, A.J., Rhee, S.K., Hink, L., Nicol, G.W., Bayer, B., Lehtovirta-Morley,
500 L., Wright, C., de la Torre, J.R., Herbold, C.W., Pjevac, P., Daims, H., Wagner, M., 2022. Ammonia-oxidizing
501 archaea possess a wide range of cellular ammonia affinities. *The ISME Journal* 16, 272-283.
- 502 Kandeler, E., Gerber, H., 1988. Short-term assay of soil urease activity using colorimetric determination of ammonium.
503 *Biology and Fertility of Soils* 6, 68-72.
- 504 Kits, K.D., Jung, M.Y., Vierheilig, J., Pjevac, P., Sedlacek, C.J., Liu, S., Herbold, C., Stein, L.Y., Richter, A., Wissel,
505 H., Brüggemann, N., Wagner, M., Daims, H., 2019. Low yield and abiotic origin of N₂O formed by the
506 complete nitrifier *Nitrospira inopinata*. *Nature Communications* 10, 1836.
- 507 Kits, K.D., Sedlacek, C.J., Lebedeva, E.V., Han, P., Bulaev, A., Pjevac, P., Daebeler, A., Romano, S., Albertsen, M.,
508 Stein, L.Y., Daims, H., Wagner, M., 2017. Kinetic analysis of a complete nitrifier reveals an oligotrophic
509 lifestyle. *Nature* 549, 269-272.
- 510 Kozłowski, J.A., Stieglmeier, M., Schleper, C., Klotz, M.G., Stein, L.Y., 2016. Pathways and key intermediates
511 required for obligate aerobic ammonia-dependent chemolithotrophy in bacteria and Thaumarchaeota. *The*
512 *ISME Journal* 10, 1836-1845.
- 513 Krüger, M., Chaudhari, N., Thamdrup, B., Overholt, W.A., Bristow, L.A., Taubert, M., Küsel, K., Jehmlich, N., von
514 Bergen, M., Herrmann, M., 2023. Differential contribution of nitrifying prokaryotes to groundwater
515 nitrification. *The ISME Journal* 17, 1601-1611.
- 516 Kuypers, M.M.M., Marchant, H.K., Kartal, B., 2018. The microbial nitrogen-cycling network. *Nature Reviews*
517 *Microbiology* 16, 263-276.
- 518 Li, C., He, Z.Y., Hu, H.W., He, J.Z., 2023. Niche specialization of comammox *Nitrospira* in terrestrial ecosystems:
519 Oligotrophic or copiotrophic? *Critical Reviews in Environmental Science and Technology* 53, 161-176.
- 520 Li, C., Hu, H.W., Chen, Q.L., Chen, D., He, J.Z., 2019. Comammox *Nitrospira* play an active role in nitrification of
521 agricultural soils amended with nitrogen fertilizers. *Soil Biology and Biochemistry* 138, 107609.
- 522 Lin, Y., Duan, C., Fan, J., Hu, H.W., He, Z.Y., Ye, G., He, J.Z., 2023. Nitrification inhibitor 1-octyne inhibits growth
523 of comammox *Nitrospira* but does not alter their community structure in an acidic soil. *Journal of Soils and*
524 *Sediments* 23, 989-997.
- 525 Lin, Y., Hu, H.W., Ye, G., Fan, J., Ding, W., He, Z.Y., Zheng, Y., He, J.Z., 2021. Ammonia-oxidizing bacteria play an
526 important role in nitrification of acidic soils: A meta-analysis. *Geoderma* 404, 115395.
- 527 Lin, Y., Ye, G., Luo, J., Di, H.J., Liu, D., Fan, J., Ding, W., 2018. *Nitrospira* Cluster 8a Plays a Predominant Role
528 in the Nitrification Process of a Subtropical Ultisol under Long-Term Inorganic and Organic Fertilization.
529 *Applied and Environmental Microbiology* 84, e01031-01018.
- 530 Martens-Habbena, W., Berube, P.M., Urakawa, H., de la Torre, J.R., Stahl, D.A., 2009. Ammonia oxidation kinetics
531 determine niche separation of nitrifying Archaea and Bacteria. *Nature* 461, 976-979.

- 532 Martens-Habbena, W., Qin, W., Horak, R.E.A., Urakawa, H., Schauer, A.J., Moffett, J.W., Armbrust, E.V., Ingalls,
 533 A.E., Devol, A.H., Stahl, D.A., 2015. The production of nitric oxide by marine ammonia-oxidizing archaea
 534 and inhibition of archaeal ammonia oxidation by a nitric oxide scavenger. *Environmental Microbiology* 17,
 535 2261-2274.
- 536 Martikainen, P.J., 2022. Heterotrophic nitrification – An eternal mystery in the nitrogen cycle. *Soil Biology and*
 537 *Biochemistry* 168, 108611.
- 538 Nicol, G.W., Leininger, S., Schleper, C., Prosser, J.I., 2008. The influence of soil pH on the diversity, abundance and
 539 transcriptional activity of ammonia oxidizing archaea and bacteria. *Environmental Microbiology* 10, 2966-
 540 2978.
- 541 Orellana, L.H., Chee-Sanford, J.C., Sanford, R.A., Löffler, F.E., Konstantinidis, K.T., 2018. Year-Round Shotgun
 542 Metagenomes Reveal Stable Microbial Communities in Agricultural Soils and Novel Ammonia Oxidizers
 543 Responding to Fertilization. *Applied and Environmental Microbiology* 84, e01646-01617.
- 544 Palomo, A., Dechesne, A., Pedersen, A.G., Smets, B.F., 2022. Genomic profiling of *Nitrospira* species reveals
 545 ecological success of comammox *Nitrospira*. *Microbiome* 10, 204.
- 546 Papadopoulou, E.S., Bachtsevani, E., Lampronikou, E., Adamou, E., Katsaouni, A., Vasileiadis, S., Thion, C.,
 547 Menkissoglu-Spiroudi, U., Nicol, G.W., Karpouzias, D.G., 2020. Comparison of Novel and Established
 548 Nitrification Inhibitors Relevant to Agriculture on Soil Ammonia- and Nitrite-Oxidizing Isolates. *Frontiers*
 549 *in Microbiology* 11, 581283.
- 550 Pjevac, P., Schaubberger, C., Poghosyan, L., Herbold, C.W., van Kessel, M.A.H.J., Daebeler, A., Steinberger, M., Jetten,
 551 M.S.M., Lückner, S., Wagner, M., Daims, H., 2017. *AmoA*-Targeted Polymerase Chain Reaction Primers for
 552 the Specific Detection and Quantification of Comammox *Nitrospira* in the Environment. *Frontiers in*
 553 *Microbiology* 8, 1508.
- 554 Prather, M.J., Hsu, J., DeLuca, N.M., Jackman, C.H., Oman, L.D., Douglass, A.R., Fleming, E.L., Strahan, S.E.,
 555 Steenrod, S.D., Søvde, O.A., Isaksen, I.S.A., Froidevaux, L., Funke, B., 2015. Measuring and modeling the
 556 lifetime of nitrous oxide including its variability. *Journal of Geophysical Research: Atmospheres* 120, 5693-
 557 5705.
- 558 Prosser, J.I., Hink, L., Gubry-Rangin, C., Nicol, G.W., 2020. Nitrous oxide production by ammonia oxidizers:
 559 Physiological diversity, niche differentiation and potential mitigation strategies. *Global Change Biology* 26,
 560 103-118.
- 561 Prosser, J.I., Nicol, G.W., 2012. Archaeal and bacterial ammonia-oxidisers in soil: the quest for niche specialisation
 562 and differentiation. *Trends in Microbiology* 20, 523-531.
- 563 Qin, W., Wei, S.P., Zheng, Y., Choi, E., Li, X., Johnston, J., Wan, X., Abrahamson, B., Flinkstrom, Z., Wang, B., Li,
 564 H., Hou, L., Tao, Q., Chlouber, W.W., Sun, X., Wells, M., Ngo, L., Hunt, K.A., Urakawa, H., Tao, X., Wang,
 565 D., Yan, X., Wang, D., Pan, C., Weber, P.K., Jiang, J., Zhou, J., Zhang, Y., Stahl, D.A., Ward, B.B., Mayali,
 566 X., Martens-Habbena, W., Winkler, M.K.H., 2024. Ammonia-oxidizing bacteria and archaea exhibit
 567 differential nitrogen source preferences. *Nature Microbiology* 9, 524-536.
- 568 Quast, C., Pruesse, E., Yilmaz, P., Gerken, J., Schweer, T., Yarza, P., Peplies, J., Glöckner, F.O., 2012. The SILVA
 569 ribosomal RNA gene database project: improved data processing and web-based tools. *Nucleic Acids*
 570 *Research* 41, D590-D596.
- 571 R Core Team, 2021. R: A language and environment for statistical computing. R Foundation for Statistical Computing,
 572 Vienna, Austria.
- 573 Ravishankara, A.R., Daniel, J.S., Portmann, R.W., 2009. Nitrous Oxide (N₂O): The Dominant Ozone-Depleting
 574 Substance Emitted in the 21st Century. *Science* 326, 123-125.
- 575 Sakoula, D., Koch, H., Frank, J., Jetten, M.S.M., van Kessel, M.A.H.J., Lückner, S., 2021. Enrichment and

- 576 physiological characterization of a novel comammox *Nitrospira* indicates ammonium inhibition of complete
 577 nitrification. *The ISME Journal* 15, 1010-1024.
- 578 Shaw, L.J., Nicol, G.W., Smith, Z., Fear, J., Prosser, J.I., Baggs, E.M., 2006. *Nitrosospira* spp. can produce nitrous
 579 oxide via a nitrifier denitrification pathway. *Environmental Microbiology* 8, 214-222.
- 580 Stieglmeier, M., Mooshammer, M., Kitzler, B., Wanek, W., Zechmeister-Boltenstern, S., Richter, A., Schleper, C.,
 581 2014. Aerobic nitrous oxide production through N-nitrosating hybrid formation in ammonia-oxidizing
 582 archaea. *The ISME Journal* 8, 1135-1146.
- 583 Tan, C., Yin, C., Li, W., Fan, X., Jiang, Y., Liang, Y., 2022. Comammox *Nitrospira* play a minor role in N₂O emissions
 584 from an alkaline arable soil. *Soil Biology and Biochemistry* 171, 108720.
- 585 Taylor, A.E., Giguere, A.T., Zoebelin, C.M., Myrold, D.D., Bottomley, P.J., 2017. Modeling of soil nitrification
 586 responses to temperature reveals thermodynamic differences between ammonia-oxidizing activity of archaea
 587 and bacteria. *The ISME Journal* 11, 896-908.
- 588 Taylor, A.E., Taylor, K., Tennigkeit, B., Palatinszky, M., Stieglmeier, M., Myrold, D.D., Schleper, C., Wagner, M.,
 589 Bottomley, P.J., 2015. Inhibitory Effects of C₂ to C₁₀ 1-Alkynes on Ammonia Oxidation in Two
 590 *Nitrososphaera* Species. *Applied and Environmental Microbiology* 81, 1942-1948.
- 591 Taylor, A.E., Vajrala, N., Giguere, A.T., Gitelman, A.I., Arp, D.J., Myrold, D.D., Sayavedra-Soto, L., Bottomley, P.J.,
 592 2013. Use of Aliphatic *n*-Alkynes To Discriminate Soil Nitrification Activities of Ammonia-Oxidizing
 593 Thaumarchaea and Bacteria. *Applied and Environmental Microbiology* 79, 6544-6551.
- 594 van Kessel, M.A.H.J., Speth, D.R., Albertsen, M., Nielsen, P.H., Op den Camp, H.J.M., Kartal, B., Jetten, M.S.M.,
 595 Lückner, S., 2015. Complete nitrification by a single microorganism. *Nature* 528, 555-559.
- 596 Walters, W., Hyde, E.R., Berg-Lyons, D., Ackermann, G., Humphrey, G., Parada, A., Gilbert, J.A., Jansson, J.K.,
 597 Caporaso, J.G., Fuhrman, J.A., Apprill, A., Knight, R., 2016. Improved Bacterial 16S rRNA Gene (V4 and
 598 V4-5) and Fungal Internal Transcribed Spacer Marker Gene Primers for Microbial Community Surveys.
 599 *mSystems* 1, e00009-00015.
- 600 Wan, X.S., Hou, L., Kao, S.J., Zhang, Y., Sheng, H.X., Shen, H., Tong, S., Qin, W., Ward, B.B., 2023. Pathways of
 601 N₂O production by marine ammonia-oxidizing archaea determined from dual-isotope labeling. *Proceedings*
 602 *of the National Academy of Sciences* 120, e2220697120.
- 603 Wang, B., Qin, W., Ren, Y., Zhou, X., Jung, M.Y., Han, P., Eloë-Fadrosch, E.A., Li, M., Zheng, Y., Lu, L., Yan, X., Ji,
 604 J., Liu, Y., Liu, L., Heiner, C., Hall, R., Martens-Habbena, W., Herbold, C.W., Rhee, S.k., Bartlett, D.H.,
 605 Huang, L., Ingalls, A.E., Wagner, M., Stahl, D.A., Jia, Z., 2019a. Expansion of Thaumarchaeota habitat range
 606 is correlated with horizontal transfer of ATPase operons. *The ISME Journal* 13, 3067-3079.
- 607 Wang, S., Wang, X., Jiang, Y., Han, C., Jetten, M.S.M., Schwark, L., Zhu, G., 2021. Abundance and Functional
 608 Importance of Complete Ammonia Oxidizers and Other Nitrifiers in a Riparian Ecosystem. *Environmental*
 609 *Science & Technology* 55, 4573-4584.
- 610 Wang, Z., Cao, Y., Zhu-Barker, X., Nicol, G.W., Wright, A.L., Jia, Z., Jiang, X., 2019b. Comammox *Nitrospira* clade
 611 B contributes to nitrification in soil. *Soil Biology and Biochemistry* 135, 392-395.
- 612 Wu, M.R., Hou, T.T., Liu, Y., Miao, L.L., Ai, G.M., Ma, L., Zhu, H.Z., Zhu, Y.X., Gao, X.Y., Herbold, C.W., Wagner,
 613 M., Li, D.F., Liu, Z.P., Liu, S.J., 2021. Novel *Alcaligenes ammonioxydans* sp. nov. from wastewater treatment
 614 sludge oxidizes ammonia to N₂ with a previously unknown pathway. *Environmental Microbiology* 23, 6965-
 615 6980.
- 616 Xu, S., Wang, B., Li, Y., Jiang, D., Zhou, Y., Ding, A., Zong, Y., Ling, X., Zhang, S., Lu, H., 2020. Ubiquity, diversity,
 617 and activity of comammox *Nitrospira* in agricultural soils. *Science of the Total Environment* 706, 135684.
- 618 Yang, J., Huang, X., 2021. The 30 m annual land cover dataset and its dynamics in China from 1990 to 2019. *Earth*
 619 *System Science Data* 13, 3907-3925.

- 620 Yin, C., Fan, X., Chen, H., Jiang, Y., Ye, M., Yan, G., Peng, H., Wakelin, S.A., Liang, Y., 2021. 3, 4-Dimethylpyrazole
 621 phosphate is an effective and specific inhibitor of soil ammonia-oxidizing bacteria. *Biology and Fertility of*
 622 *Soils* 57, 753-766.
- 623 Yuan, D., Zheng, L., Tan, Q., Wang, X., Xing, Y., Wang, H., Wang, S., Zhu, G., 2021. Comammox activity dominates
 624 nitrification process in the sediments of plateau wetland. *Water Research* 206, 117774.
- 625 Zhang, Q., Li, Y., He, Y., Liu, H., Dumont, M.G., Brookes, P.C., Xu, J., 2019. *Nitrospira* cluster 3-like bacterial
 626 ammonia oxidizers and *Nitrospira*-like nitrite oxidizers dominate nitrification activity in acidic terrace paddy
 627 soils. *Soil Biology and Biochemistry* 131, 229-237.
- 628 Zhao, J., Bello, M.O., Meng, Y., Prosser, J.I., Gubry-Rangin, C., 2020a. Selective inhibition of ammonia oxidising
 629 archaea by simvastatin stimulates growth of ammonia oxidising bacteria. *Soil Biology and Biochemistry* 141,
 630 107673.
- 631 Zhao, J., Meng, Y., Drewer, J., Skiba, U.M., Prosser, J.I., Gubry-Rangin, C., 2020b. Differential Ecosystem Function
 632 Stability of Ammonia-Oxidizing Archaea and Bacteria following Short-Term Environmental Perturbation.
 633 *mSystems* 5, e00309-00320.
- 634 Zhou, X., Wang, S., Ma, S., Zheng, X., Wang, Z., Lu, C., 2020. Effects of commonly used nitrification inhibitors—
 635 dicyandiamide (DCD), 3,4-dimethylpyrazole phosphate (DMPP), and nitrapyrin—on soil nitrogen dynamics
 636 and nitrifiers in three typical paddy soils. *Geoderma* 380, 114637.
- 637 Zhu, X., Burger, M., Doane, T.A., Horwath, W.R., 2013. Ammonia oxidation pathways and nitrifier denitrification are
 638 significant sources of N₂O and NO under low oxygen availability. *Proceedings of the National Academy of*
 639 *Sciences* 110, 6328-6333.

640

641 **Figure legends**

642 **Figure 1** Changes in the *amoA* gene copies of AOA, AOB and comammox *Nitrospira* clades A
 643 and B during the 21-day incubation period of HB (Panels a), HN (Panels b) and GZ (Panels c)
 644 soils. Soil microcosms were amended with NH₄⁺ in the absence or presence of inhibitors (acetylene,
 645 0.01 to 0.05% 1-octyne, or DMPP). Error bars represent the standard errors of the mean (*n* = 3).

646 **Figure 2** Temporal dynamics of NH₄⁺, NO₃⁻ and cumulative N₂O emissions during the 21-day
 647 incubation period of HB (Panels a), HN (Panels b) and GZ (Panels c) soils. Soil microcosms were
 648 amended with NH₄⁺ in the absence or presence of inhibitors (acetylene, 0.01 to 0.05% 1-octyne,
 649 or DMPP). Error bars represent the standard errors of the mean (*n* = 3).

650 **Figure 3** NH₃ oxidation and N₂O production driven by NH₃ oxidisers in HB, HN and GZ soils.
 651 Average NH₃ oxidation (Panels a) and N₂O production (Panels b) rates and relative contributions

652 of AOA (orange), AOB (gray), and comammox *Nitrospira* (CMX, green) in HB, HN, and GZ soils.
653 The contribution of each group is displayed under the columns with the corresponding colours. **c**
654 The N₂O yield of AOA (orange), AOB (gray), and comammox *Nitrospira* (CMX, green) in
655 different soils. Error bars indicate the standard errors of triplicate microcosms. The average rates
656 and relative contributions in HB soil were calculated as the mean values based on the 0.01% and
657 0.015% 1-octyne-treated microcosms. The average rates and relative contributions in HN and GZ
658 soil were calculated as the mean values based on the 0.01 to 0.05% 1-octyne-treated microcosms.

659 **Figure 4** Ammonia oxidiser community analysis in HB, HN and GZ soils incubated in the initial
660 sample (Original), in the absence (N) or presence of acetylene (CH), DMPP (DP), or 0.01% to
661 0.05% 1-octyne (O1 to O5) for 21 days. **a** Relative abundance of AOA OTUs (based on MiSeq
662 sequencing using universal 16S rRNA gene primers). **b** Relative abundance of AOB OTUs (based
663 on MiSeq sequencing using universal 16S rRNA gene primers). **c** Relative abundance of the
664 comammox *Nitrospira* OTUs (relative abundance > 5%). Data are presented as the mean and
665 standard error of triplicate microcosms for each treatment. An asterisk (*) next to the OTU
666 indicates a significant difference in abundance between treatments ($p < 0.05$), based on a one-way
667 ANOVA.
668

669 **Supporting Information**

670 **Supplementary Tables S1–S6**

671 **Table S1** Locations, climatic parameters of the study sites, and physicochemical properties of the
672 soils.

673 **Table S2** Reaction systems and thermal cycling protocols used for target gene amplification in
674 qPCR.

675 **Table S3** Cell-specific NH₃ oxidation rates in soils. It is assumed that each cell of AOB, AOA and
676 comammox *Nitrospira* contains 2.5, 1.0 and 1.0 *amoA* gene copies, respectively (Huang et al.,
677 2021).

678 **Table S4** The *amoA* gene abundance of AOA, AOB and comammox *Nitrospira* in HB, HN and
679 GZ soils in the presence of acetylene (CH) statistical analyses (Figure 1).

680 **Table S5** The *amoA* gene abundance of AOA in HB, HN and GZ soils in the absence (N) or
681 presence of 1-octyne (OC) and DMPP (DP) statistical analyses (Figure 1).

682 **Table S6** The comammox *Nitrospira* clade A *amoA* gene abundance in HB, HN and GZ soils in
683 the absence (N) or presence of 1-octyne (OC) and DMPP (DP) statistical analyses (Figure 1).

684

685 **Supplementary Figures S1–S5**

686 **Figure S1** Geographical location of the sampling sites for agricultural soils across southern China.

687 **Figure S2** Schematic diagram of the inhibitory effects of inhibitors. Acetylene completely
688 inactivates AOA, AOB and comammox *Nitrospira* (CMX). 1-Octyne specifically inhibits AOB at
689 certain concentrations. Low concentrations of 1-octyne result in partial inhibition of AOB activity,
690 and high concentrations of 1-octyne partially inhibit the activity of comammox *Nitrospira*. Both
691 AOB and comammox *Nitrospira* are inhibited by DMPP.

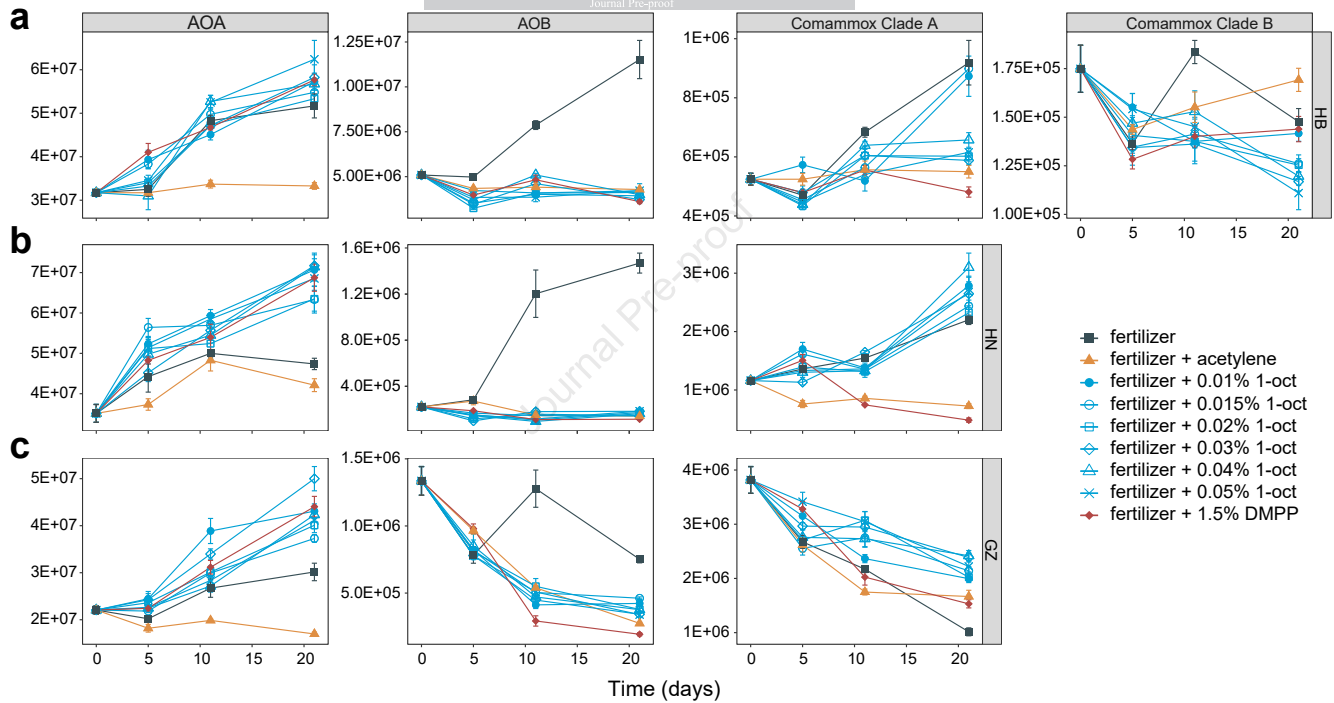
692 **Figure S3** Temporal dynamics of cumulative N₂O emissions during the 5-day incubation period
693 of HB, HN and GZ soils. Soil microcosms were amended with NH₄⁺ (150 μg N g⁻¹ soil_{dw}) in the
694 absence of acetylene (0.01%, v/v) or amended with NO₃⁻ (150 μg N g⁻¹ soil_{dw}) in the presence of
695 acetylene (0.01%, v/v). Error bars represent the standard errors of the mean (*n* = 3).

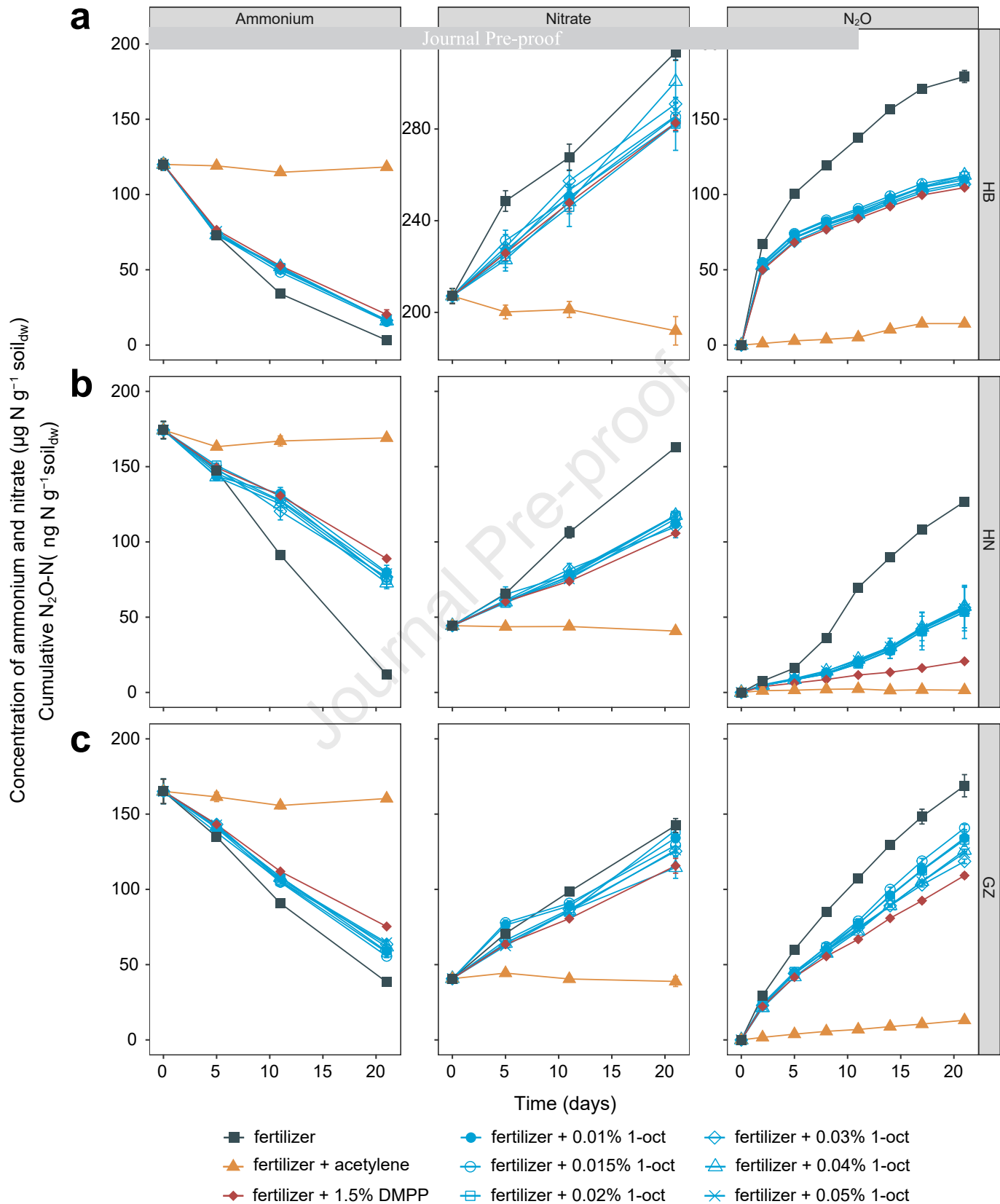
696 **Figure S4** Absolute abundance of comammox *Nitrospira* OTUs (relative abundance > 5%)
697 retrieved in HB, HN and GZ soils incubated in the initial sample (Original), in the absence (N) or
698 presence of acetylene (CH), DMPP (DP), or 0.01% to 0.05% 1-octyne (O1 to O5) for 21 days. The
699 absolute abundance of comammox *Nitrospira* OTUs was calculated as the product of comammox
700 *Nitrospira amoA* gene numbers (obtained by qPCR) and the relative abundance of the
701 corresponding OTUs (obtained by MiSeq sequencing using *amoA* gene primers). Data are
702 presented as the mean and standard error of triplicate microcosms for each treatment. An asterisk
703 (*) next to the OTU indicates a significant difference in abundance between treatments (*p* < 0.05),
704 based on a one-way ANOVA.

705 **Figure S5** Deviation of NH₃ oxidiser rates in microcosms treated with different concentrations of
706 1-octyne. **a** NH₃ oxidation rates of AOA (orange), AOB (gray), and comammox *Nitrospira* (CMX,

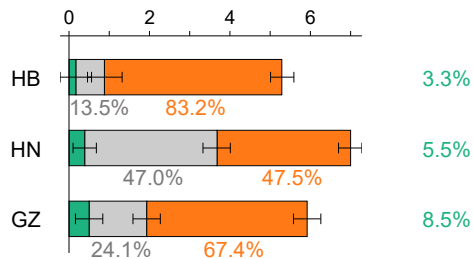
707 green) calculated from all 1-octyne concentrations tested (0.01%–0.05%). **b** N₂O production rates
708 of AOA, AOB, and comammox *Nitrospira* calculated from all 1-octyne concentrations tested
709 (0.01%–0.05%).

Journal Pre-proof

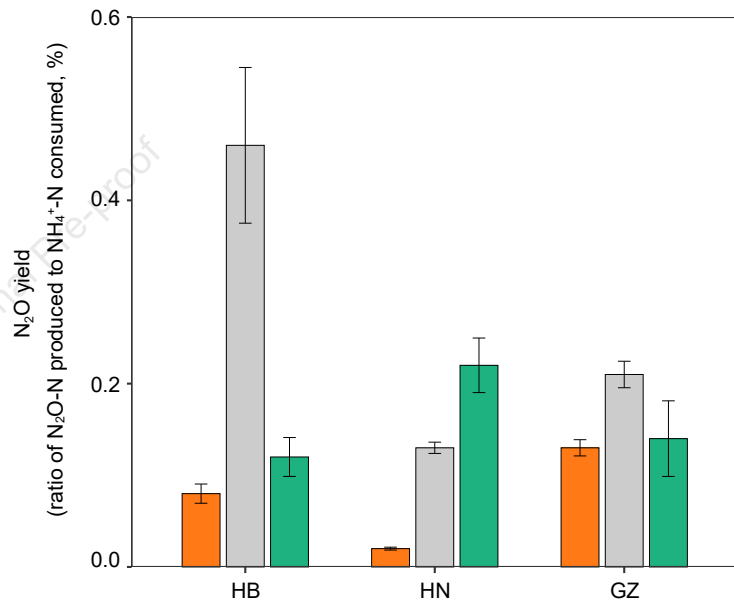
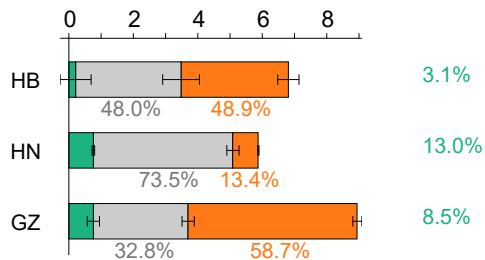




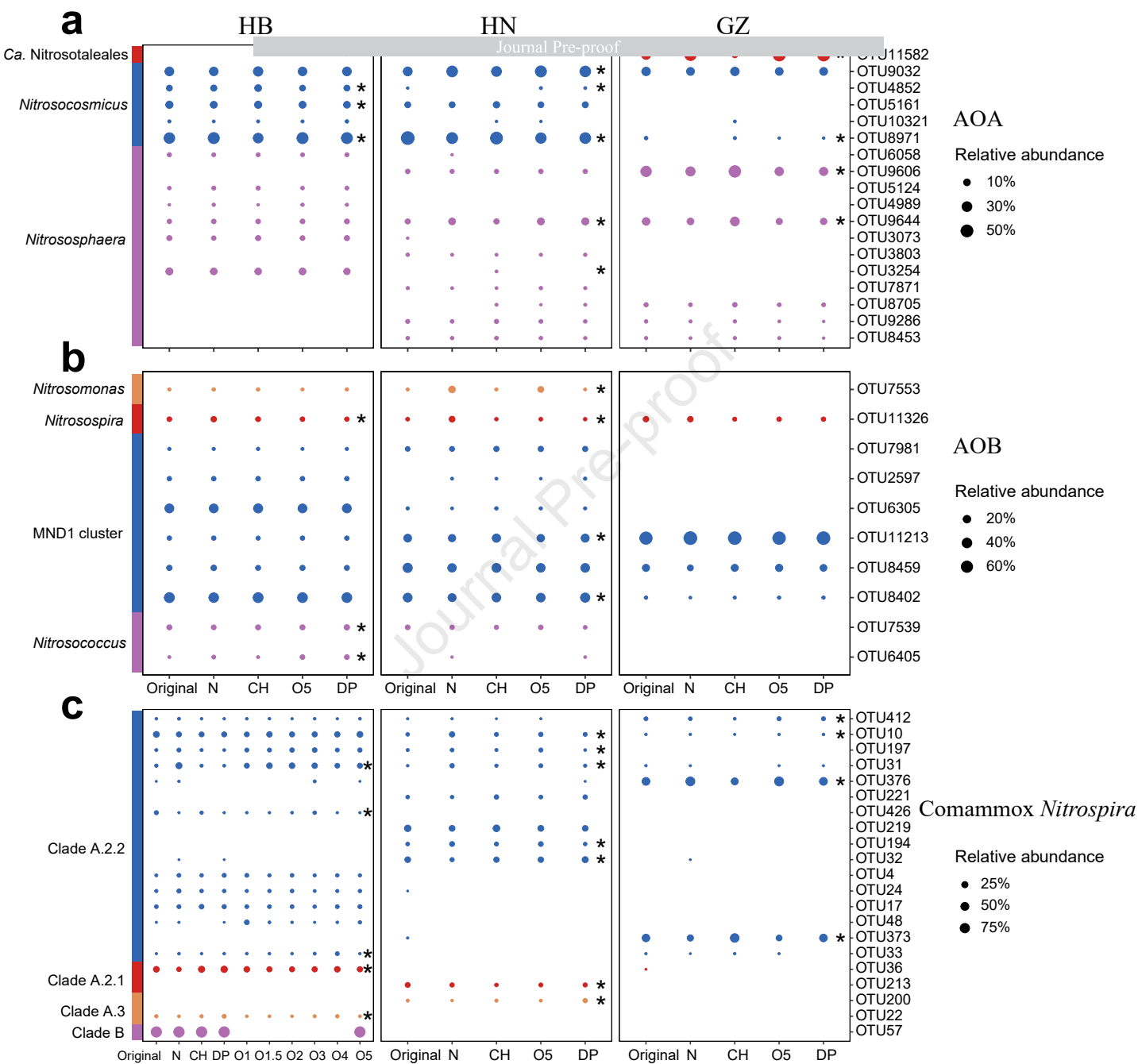
a NH₃ oxidation rate ($\mu\text{g N g}^{-1} \text{d}^{-1}$) CMX contribution (%)



b N₂O production rate ($\text{ng N g}^{-1} \text{d}^{-1}$)



AOA AOB CMX



Highlights

- Comammox play a minor role in NH_3 oxidation and N_2O emissions in acidic arable soils;
- AOB are functionally active in acidic agricultural soils;
- The sensitivity of comammox *Nitrospira* clade A to 1-octyne varies across soils.

Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Journal Pre-proof