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# Developing an indicator of the abundance, extent and impact of invasive non-native species 

## Final report

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## Summary

This report describes an indicator of the abundance, extent and impact of invasive non-native species in Great Britain. The main ideas and options for the abundance indicator and for the impact indicator are considered. A third type of indicator, the annual rate of establishment of new non-native species, is outlined, with provisional data presented only for England.

Most of the options for the abundance indicator do not in fact measure abundance, but use either frequency in samples or frequency in recording scheme data as a substitute. An exception is the Breeding Bird Survey, for which numbers of individuals are counted. Several well-recorded groups of organisms have no non-native species (e.g. butterflies and lichens) or exceedingly few non-native species (macro-moths).

Datasets selected for the abundance indicator were the Breeding Bird Survey (birds and mammals), Countryside Survey (vascular plants), British Bryological Society data (bryophytes) and Marine Biological Society data (marine organisms). From samples of records in each species group, the non-native component was calculated as proportion of all species sampled. This provided a temporal trend in non-native proportions, which were calculated separately for England, Scotland and Wales. The GB trend was derived by combining the trends for each component country, weighted by the area of each. Finally, the overall trend was calculated as a weighted geometric mean of trends for each species group, converted to an index by dividing by a constant to start at 1 in the baseline year 1990. The weights applied were birds $20 \%$, mammals $20 \%$, vascular plants $30 \%$, bryophytes $10 \%$ and marine organisms $20 \%$. There were no suitable datasets from the freshwater environment.

No direct measure of impact could feasibly be calculated for all invasive species in Great Britain. As a substitute, an indicator based on the extent of occupation by invasive species was adopted. The methodology for the indicator was based on a scheme developed by the Belgian Forum on Invasive Species. First, a list of the most invasive species was compiled, using a simplified environmental impact assessment protocol to assign species to threat categories. Then the extent of each invasive species was scored for 1960, 1970, 1980, 1990, 2000 and 2007, on a 5 -point scale ranging from 0 (absent) to 4 (present in more than half the territory). Extent scores were added to obtain the indicator.

Over the period 1990-2007, the mean indexed proportion of records of nonnative species in samples of birds, mammals, plants and marine life rose by $23 \%$. Except for mammals, the absolute proportion was still only about $1 \%$ of the total.

The assessment protocol assigned 49 species in Great Britain to the highest threat category. There were 3 marine plants, 16 marine animals, 4 freshwater plants, 8 freshwater animals, 8 terrestrial plants and 10 terrestrial animals. Over the period 1990-2007, the summed extent scores of these invasive non-native species rose by $40 \%$. The increase of invasive species was particularly large in the freshwater and marine environments.

Although non-native species are a potential threat, they are still only a small proportion of the animals and plants to be found in most of the land area and coasts of Great Britain. Vertebrates stand out as the most invasive group.

For all groups of organisms reported here, England was the country most affected by non-native species. Scotland was the least affected. Wales was intermediate. In 2008, values of the impact indicator for the three countries were respectively 135, 73 and 95 . Most species groups showed a trend over time towards an increasingly non-native biota.

If the indicator is to be developed further, the main priority is to include freshwater species in the abundance component. Because the list of invasive species depends on expert judgement, it needs to be reviewed and if necessary updated at regular intervals. Further analytical work is desirable, to improve the signal obtained from recording scheme data.

## Introduction

## Background and timescale of the project

Biological invasions by non-native species are recognized as a significant component of global environmental change. Under the CBD, the United Kingdom has an international obligation to address the impacts of invasive non-native species. Progress in fulfilling this obligation can be measured by an indicator. Defra, on behalf of the UK Government, invited tenders for a project to develop and present an indicator of the abundance, extent and impact of invasive non-native species. A consortium composed of the Centre for Ecology and Hydrology (CEH; lead partner), the British Trust for Ornithology, the Central Science Laboratory and the Marine Biological Association successfully tendered for the project.

Work began with a meeting of the project steering group on 13 August 2008. A consortium meeting was held at CEH Wallingford on 10 September. This identified the approach that was eventually adopted for the indicator of invasive non-native species and suggested datasets that could be used for the abundance component. A proposed methodology was presented to the UK Biodiversity Indicators Steering Group on 23 September and discussed more fully at a workshop for the project steering group, held at Nobel House on 25 November. The project was originally due to report on 8 December 2008. Sourcing and analysis of data took longer than had originally been hoped. Reporting was delayed by two months.

## Objectives

The project as originally set up had two main objectives, namely to develop:

1. A measure of the abundance of non-native species in Great Britain and, where possible, the proportion which are invasive.
2. A measure of the impacts from the sub-set of non-native species that are invasive.
In addition to the two main indicators for GB, analogous indicators were required for England, Scotland and Wales.

## Approach to the problem

The European Environment Agency (2007) proposed that the headline indicator on trends in invasive alien species in Europe should be composed of two parts. The first part was the cumulative number of alien species in Europe since 1900. The authors of the EEA report assert that there is a direct correlation between this number and the number of species that may become invasive. The second part of the headline indicator, which had not been calculated, was the cost of invasive alien species.

The EEA proposal, as well as others, were considered in the light of data availability. Given the short timescale of the project, the indicator could be calculated only if data could be interrogated quickly. Several guiding principles were set out at an early stage.

1. Crop species should not be included in those locations where they are cultivated; descendents of crops should be excluded unless they are part of a self-sustaining population.
2. Non-breeding species should be excluded unless they have effective clonal spread (which is for this purpose treated as reproduction).
3. Organisms that occur only in man-made, as opposed to semi-natural or natural habitats, should excluded or down-weighted.
4. A newly-arrived non-native should affect the indicator additively.
5. Indicators should reflect the current situation and not depend on what happened in the past.

## Definition of non-native and invasive species

The definition of a non-native species in the Defra Review of Non-native Species Policy (Anon., 2003) is:-

- A species introduced (i.e. by human action) outside its natural past or present distribution.
Stated thus, the definition would include species that are native to certain parts of Britain but are outside their natural range in others - for example Welsh Poppy Meconopsis cambrica in Scotland, Scots Pine Pinus sylvestris in England and Sea-buckthorn Hippophae rhamnoides in Wales. For the indicator reported here, species that are native in any part of Britain are excluded. Also excluded are species such as Little Egret Egretta garzetta that have newly arrived by natural means.

The UK Government's Invasive Non-native Species Framework Strategy (UK Government, 2008) defines invasive non-native species thus:-

- Invasive non-native species are those whose introduction and/or spread threaten biological diversity or have other unforeseen impacts.
The GB indicator is required mainly for biodiversity reporting under the CBD. The definition used here therefore ignores other impacts such as effects on human health, harm to crops or damage to property.


## The non-native biota of Britain

An enumeration of the non-native biota from a previous report to Defra (Hill et al., 2008) provided a sound basis on which to build. If fungi, bacteria and protists are omitted, the number of non-native species in Britain is 3473 (Table 1). The grouping was adjusted from that used by Hill et al. (2008) to remove those vascular plants that lived in marine or freshwater environments to their proper categories. Amphibians were assigned to the freshwater category, on the grounds that they reproduce in freshwater. Insects were not reviewed; all have been treated as terrestrial animals, regardless of the medium in which they reproduce.

## Methods

## Possible measures of abundance

Several possible measures of abundance were considered, most of which are based on the proportion of non-native species (Table 2).

| Group of species | No. | Group of species | No. |
| :--- | ---: | :--- | ---: |
| (a) Marine plants | 49 | (d) Freshwater animals | 99 |
| Chromista (Haptophyta) | 2 | Nematodes and leeches | 7 |
| Chrysomonada | 1 | Flatworms (Platyhelminthes) | 14 |
| Dinoflagellates (Dinophyta) | 6 | Crustaceans (Crustacea) | 19 |
| Flowering plants | 4 | Bryozoans (Ectoprocta) | 1 |
| Green algae (Chlorophyta) | 2 | Bony fishes (Osteichthyes) | 39 |
| Other algae (Bacillariophyta) | 2 | Amphibians (Amphibia) | 10 |
| Brown algae (Ochrophyta) | 15 | Cnidarians (Cnidaria) | 1 |
| Red algae (Rhodophyta) | 17 | Myxozoa | 1 |
|  |  | Molluscs (Mollusca) | 7 |
| (b) Marine animals | 99 |  |  |
| Segmented worms (Annelida) | 17 | (e) Terrestrial plants | 1851 |
| Crustaceans (Crustacea) | 34 | Flowering plants | 1784 |
| Bryozoans (Ectoprocta) | 4 | Gymnosperms (conifers) | 39 |
| Bony fishes (Osteichthyes) | 2 | Ferns (Pteridophyta) | 10 |
| Sea squirts (Tunicata) | 8 | Mosses and liverworts | 18 |
| Cnidarians (Cnidaria) | 6 |  |  |
| Bryozoans (Ectoprocta) | 1 | (f) Terrestrial animals | 1353 |
| Goblet worms (Entoprocta) | 2 | Insects (Insecta) | 882 |
| Molluscs (Mollusca) | 22 | Other arthropods | 75 |
| Flatworms (Platyhelminthes) | 1 | Molluscs (Mollusca) | 26 |
| Sponges (Porifera) | 2 | Nematodes and Nemertea | 5 |
|  |  | Flatworms (Platyhelminthes) | 5 |
| (c) Freshwater plants | 22 | Birds (Aves) | 323 |
| Flowering plants | 21 | Mammals (Mammalia) | 31 |
| Ferns (Pteridophyta) | 1 | Reptiles (Reptilia) | 6 |

Table 1. Species groups included in the non-native biota of Great Britain for the purposes of the indicator; species that were introduced before 1500 are omitted from the counts

| Type | Measure of abundance | Comment |
| :---: | :---: | :---: |
| 1 | Biomass proportion of non-native species | Data are not available for most groups; for birds one could multiply estimated density by individual mass |
| 2 | Population of non-native species as a proportion of the total | Unlikely to be available except for some vertebrates |
| 3 | Frequency as a proportion in fullyrecorded units | Examples are the proportion of non-natives in quadrats in Countryside Survey or in the Rothamsted Insect Survey |
| 4 | Frequency as a proportion in incompletely recorded units | Such data typically arise from national recording schemes that do not cover the country in a single campaign |
| 5 | Geometric mean of population size for non-native species, converted to an index | Not a relative measure; geometric means are difficult to apply to rare species and newly-appearing species |
| 6 | Count of all breeding non-native species | A measure of diversity rather than abundance; it can be relativized by dividing by the total number of species |

Table 2. Possible measures of abundance or frequency

The following points amplify the comments in Table 2.

1. The biomass proportion of non-native species is a measure that could in principle be applied without regard to taxonomic group. In practice, biomass data for most groups are not available and cannot readily be calculated. Furthermore, because of the pyramid of numbers, the biomass proportion would grossly favour plants over animals, even though animals may have a larger impact. Where population estimates are available, biomass can be calculated by multiplying by individual mass by the density of individuals per unit area.
2. The population of non-native species as a proportion of the total is not easily applied to plants, because many of them, including the famously invasive Japanese Knotweed Fallopia japonica are clonal. Indeed, Japanese Knotweed appears to be represented by a single clone in Britain and could therefore be counted as a single individual (Hollingsworth and Bailey, 2000). On the other hand, population estimates are available for some vertebrates, such as birds, so that the non-native proportion could be estimated.
3. Frequency as a proportion in fully-recorded units is available for plants in quadrats recorded by Countryside Survey (Countryside Survey Partnership, 2008). Likewise, it is available for well-recorded groups such as birds and butterflies at the hectad (10-km square) scale.
4. Frequency as a proportion in incompletely recorded units is all that is normally available from standard volunteer recording schemes. The character of volunteer data can fluctuate widely from year to year, depending on where recorders are most active at any one time. Such fluctuations can be partially corrected by giving a standard weighting to the contribution from different regions. A correction process of this type is described below.
5. Indicators based on geometric means are routinely used for reporting bird populations (British Trust for Ornithology, 2006). Noble, Newson \& Gregory (2004) show how to set thresholds for new species arriving or others declining to very low numbers. There are severe difficulties in using this methodology directly for non-native species, because most have very small populations (Appendix 3).
6. A count of all breeding non-native species is in fact a measure of diversity rather than abundance. It is very close to the measure proposed by the EEA (2007), and differs only in subtracting the species that have died out. When the non-native total is calculated as a proportion of the total biota, it is analogous to a frequency proportion (i.e. a measure of type 3), where the unit sampled is the whole territory.

## Choice of abundance indicator

The choice of abundance indicator was guided by the principles set out at the beginning but also depended crucially on the availability of data. No data for biomass or total population counts were readily available. Measures of type 3, based on fully-recorded units, were preferred if
possible, on the grounds that calculated proportions do not depend on the current activity of recorders. Measures of type 4, based on national recording schemes that may not make full species inventories, were used where these were suitable. Data held by the Biological Records Centre for most national recording schemes were unsuitable, either because the scheme records a group that has very few non-native species or because the data have poor temporal coverage. Although non-native insects appear as the largest group other than flowering plants in Table 1, butterflies and macro-moths, which have the most active recording schemes are remarkably lacking in non-native species. There are no non-native butterflies. Only one definitely non-native macro-moth has colonized Britain since 1900.

Taking these considerations into account, the abundance indicator was based on six major components (Table 3), comprising birds, mammals, vascular plants (in two types of location), bryophytes and marine organisms.

| Species group | Weight | Observation | Index calculated |
| :--- | :--- | :--- | :--- |
| Birds (Breeding Bird <br> Survey) | 0.2 | Counts of individuals <br> per transect in 1-km <br> square | Non-native count as proportion <br> of total count |
| Mammals (Breeding <br> Bird Survey) | 0.2 | Presence in Breeding <br> Bird Survey (BBS) <br> square | Mean occurrence frequency of <br> 6 species of non-native <br> mammals in those BBS <br> squares with mammal records |
| Vascular plants, <br> random plots <br> (Countryside Survey) | 0.15 | Presence in random <br> 14 m square quadrat | Frequency of non-native non- <br> crop species as a proportion of <br> all species recorded |
| Vascular plants, <br> streamside plots <br> (Countryside Survey) | 0.15 | Presence in quadrat <br> by watercourse ('S' <br> and 'W' plots) | Frequency of non-native non- <br> crop species as a proportion of <br> all species recorded |
| Bryophytes (BRC <br> dataset) | 0.1 | Presence in hectad <br> (10 km square) | Frequency of non-native <br> hectad records as a proportion <br> of all hectad records |
| Marine (Marine Life <br> Information Network) | 0.2 | Record in database | Frequency of non-native <br> records as a proportion of total |

Table 3. Components of indicator of non-native abundance

In terms of the measures outlined in Table 2, that for birds approximates to a measure of type 2, but uses counts per transect rather than an estimate of population size. The measure for mammals is based on signs and sightings of Brown Rat, Grey Squirrel, American Mink, Reeves' Muntjac, Sika Deer and Chinese Water Deer. It is not divided by total mammal frequency. The reason for this is that most mammals, especially small mammals and (by day) bats, are cryptic, so that a reliable denominator of the total frequency of mammals is not available. The measure for vascular plants is a measure of type 3. Measures of type 4 are used for bryophytes and marine organisms. Measures of type 5 were not used directly, but the component indices were
combined by a weighted geometric mean. Specifically, let the six component indices be $x_{1}, x_{2}, x_{3}, x_{4}, x_{5}$ and $x_{6}$, with weightings (Table 3) $w_{1}$, $w_{2}, w_{3}, w_{4}, w_{5}$ and $w_{6}$. Then the abundance indicator $A$ is defined as

$$
A=\text { antilog }\left(w_{1} \log x_{1}+w_{2} \log x_{2}+w_{3} \log x_{3}+w_{4} \log x_{4}+w_{5} \log x_{5}+w_{6} \log x_{6}\right)
$$

For presentation purposes, the indicator and component indices are relativized to 1 in 1990, which is taken as the starting date.

## Measures of impact

Several possible measures of impact were considered (Table 4). In the event, all but one of them, based on a simplified environmental impact assessment protocol developed by the Belgian Forum on Invasive Species (Branquart, 2007), proved to have insuperable difficulties. This was therefore selected.

| Measure of impact | Comment |
| :--- | :--- |
| 1. Damage to ecosystems or species | Very hard to quantify; few potentially damaging <br> non-natives are listed as threats to BAP species |
| 2. Impact of 20 most invasive species | Measure of impact difficult to define; likely to be <br> too subjective |
| 3. Extent of native species impacted upon <br> by non-native species | Not easily measurable; the degree of impact is <br> equally important |
| 4. Cost of control measures | Data could not be updated regularly <br> 5. Economic damage by species <br> 6. Modified Belgian system for rating <br> environmental risk and species extent |

Table 4. Possible measures of impact

The following points amplify the comments in the table.

1. Damage to ecosystems or species is exceedingly hard to quantify. For example, even if it were possible to measure the number of indigenous White-clawed Crayfish Austropotamobius pallipes that are absent as a result of the alien Signal Crayfish Pacifastacus leniusculus, this impact would need to be added to that of Rhododendron Rhododendron ponticum in obliterating moorland and replacing it by scrub. The units of measurement are simply not comparable, and incompatible units would be found across a wide range of impacts.
2. The impact of the twenty most invasive species suffers from the same problems as the more general measure of damage, considered above.
3. The extent of native species impacted upon by non-native species can in principle be measured. However, the degree of impact is important. For example, Sycamore Acer pseudoplatanus occurs widely in mixtures with native Ash Fraxinus excelsior, which is
reduced in quantity but almost never eliminated. This particular impact is therefore of much lower importance than that of the nonnative crayfish.
4. The cost of control measures has the merit of using a common system of measurement, for example pounds sterling per year. However, the current cost of control measures does not necessarily reflect the damage being done to biodiversity. Indeed, some very damaging species may at present incur little or no cost because they cannot be controlled. For those species that are currently being controlled, control costs are mostly not collected.
5. Economic damage may not be relevant to biodiversity. At present, the highest control costs in Great Britain are for Japanese Knotweed Fallopia japonica, mainly because of economic damage in the urban environment.
6. The Belgian Forum on Invasive Species developed a system for scoring non-native species on simple numerical scales to measure environmental risk and species extent. This system is used here to define a measure of impact, based on invasiveness and extent of spread, described below.

## Distinguishing invasive species

Invasive non-native species were selected by means of a short risk assessment, based on that developed by the Belgian. This was applied to the species signified as having high impact in an audit for English Nature (Hill et al., 2005), as well as to the 100 European high-impact species identified by DAISIE (DAISIE, 2009).

The modified Belgian system starts by selecting the most invasive species. Each species is rated on a three-point scale in four categories:

1. Dispersal potential
2. Colonization of natural and semi-natural habitats
3. Adverse impacts on native species
4. Alteration of ecosystem function.

In each category, the species is rated on a scale of 1 (low risk) to 3 (high risk). Full definitions of these risk scales for each of the four categories are given in Appendix 1. For categories 3 and 4, the risk is divided into subcategories (Table 5). The contribution of each of these categories is calculated as the maximum risk score of its subcategories. Additional numerical scores are given to three other types of entry, namely DD (data deficient) scoring 0 , unlikely (adverse impact unlikely) scoring 1, and likely (adverse impact likely) scoring 2.

In the example given, Sycamore scores 3 for dispersal potential and natural habitats. However, as pointed out above, it does not eliminate native species but mixes with them, and it does not produce any physical alteration that might not be produced by a native tree. On this basis it is seen as a potential risk but does not qualify as invasive.

Of 152 species scored in this way, 49 were in the highest threat category and therefore deemed to be invasive (Table 6). Many other species were
considered informally, but were omitted because they were clearly not in the highest threat category.

| Major category | Marine invertebrate | Freshwater vertebrate | Terrestrial plant | Terrestrial vertebrate |
| :---: | :---: | :---: | :---: | :---: |
| Scientific name | Eriocheir sinensis | Alytes obstetricans | Acer pseudoplatanus | Myocastor coypus |
| English name | Chinese Mitten Crab | Midwife Toad | Sycamore | Coypu |
| 1 Dispersal potential | 3 | 1 | 3 | 3 |
| 2 Natural habitats | 3 | 3 | 3 | 3 |
| 3.1 Predation, herbivory | 3 | 1 | - | 2 |
| 3.2 Competition | 3 | 1 | 2 | unlikely |
| 3.3 Disease vector | DD | 2 | 1 | likely |
| 3.4 Genetic pollution | 1 | 1 | 1 | 1 |
| 3 Max Species impact | 3 | 2 | 2 | 2 |
| 4.1 Nutrient cycling | DD | 1 | 1 | 3 |
| 4.2 Physical alteration | 3 | 1 | 2 | 3 |
| 4.3 Successions | 1 | 1 | 2 | 2 |
| 4.4 Food webs | 2 | 2 | 2 | likely |
| 4 Max Ecosyst impact | 3 | 2 | 2 | 3 |
| Sum of 1 to 4 | 12 | 8 | 10 | 11 |
| Risk category | A | C | B | A |

Table 5. Calculation of risk category using the Modified Belgian system; species with combined score 11 or 12 are rated as invasive (A), those scoring 9 or 10 are rated as a potential risk (B) and those with lower scores are not thought likely to become invasive (C)

## Extent values of invasive species summed for impact indicator

The original Belgian system of scoring species for extent used just three categories, together with zero for absence. For a larger territory such as Britain, this number was judged to be too small. Four extent categories were used for species present in the territory (Table 7). The indicator of non-native impact was then derived by summing extent values for those species that were deemed to be invasive.

## Data

## Breeding Bird Survey (birds and mammals)

The Breeding Bird Survey (BBS) http://www.bto.org/bbs/ is a line-transect survey based on randomly selected 1-km squares. Squares are chosen according to a stratified random design, with more squares in areas with more potential volunteers. Each BBS observer makes two early morning visits during the April-June survey period to count all birds encountered while walking two $1-\mathrm{km}$ transects across their square. Birds are recorded in three distance categories, or as 'in flight'. Observers also record the habitat along the transects, and record any mammals seen during the survey. In 2007, the BBS recorded 2,783 squares in England, 409 squares in Scotland
and 263 squares in Wales. BBS data are available for birds from 1994 and for mammals from 1995.

| (a) Marine plants | (d) Freshwater animals |
| :---: | :---: |
| Sargassum muticum (Jap Weed, Wire Weed) | Pacifastacus leniusculus (Signal Crayfish) |
| Undaria pinnatifida (Japanese Kelp, Wakame) | Procambarus clarkii (Red Swamp Crayfish) |
| Codium fragile ssp. tomentosoides (Green Sea Fingers) | Corbicula fluminea (Asian Clam) |
|  | Dreissena polymorpha (Zebra Mussel) |
|  | Pseudorasbora parva (Topmouth Gudgeon) |
| (b) Marine animals | Sander lucioperca (Pikeperch, Zander) |
| Tricellaria inopinata (a bryozoan) | Lithobates catesbeianus (American Bullfrog) |
| Watersipora subtorquata (a bryozoan) | Trachemys scripta (Common Slider Turtle) |
| Corophium sextonae (an amphipod) |  |
| Gammarus tigrinus (an amphipod) | (e) Terrestrial plants |
| Elminius modestus (an acorn barnacle) | Carpobrotus edulis (Hottentot Fig) |
| Solidobalanus fallax (a barnacle) | Disphyma crassifolium (Purple Dewplant) |
| Eriocheir sinensis (Chinese Mitten Crab) | Fallopia japonica (Japanese Knotweed) |
| Rhithropanopeus harrisii (Dwarf Crab) | Heracleum mantegazzianum (Giant Hogweed) |
| Crassostrea gigas (Pacific Oyster) | Impatiens glandulifera (Himalayan Balsam) |
| Crepidula fornicata (Slipper Limpet) | Quercus ilex (Evergreen Oak) |
| Rapana venosa (Rapa Whelk) | Rhododendron ponticum (Rhododendron) |
| Anguillicola crassus (Swim-bladder Nematode) | Rosa rugosa (Japanese Rose) |
| Botrylloides violaceus (a tunicate) |  |
| Corella eumyota (a tunicate) | (f) Terrestrial animals |
| Didemnum vexillum (a tunicate) | Arthurdendyus triangulata (New Zealand Flatworm) |
| Styela clava (Leathery Sea Squirt) | Harmonia axyridis (Harlequin Ladybird) |
|  | Branta canadensis (Canada Goose) |
|  | Oxyura jamaicensis (Ruddy Duck) |
| (c) Freshwater plants | Cervus nippon (Sika Deer) |
| Crassula helmsii (New Zealand Pigmyweed) | Muntiacus reevesi (Reeves' Muntjac) |
| Hydrocotyle ranunculoides (Floating Pennywort) | Mustela vison (American Mink) |
| Ludwigia grandiflora (Uruguayan Hampshire-purslane) | Myocastor coypus (Coypu) |
| Myriophyllum aquaticum (Parrot's-feather) | Rattus norvegicus (Brown Rat) |
|  | Sciurus carolinensis (Grey Squirrel) |

Table 6. Invasive species in the highest threat category in Great Britain; details of how these were selected are given in Appendix 2

| Definition | Interpretation | Extent |
| :---: | :---: | :---: |
| Not present in territory | Absent | 0 |
| Present in territory and either not established or with established populations that have not spread more than 10 km from their source | Not or scarcely established | 1 |
| Established populations present less than $10 \%$ of territory, with some having arrived from further than 10 km from their source; or if more widespread then populations scattered and sparse | Established but still generally absent or at most occasional | 2 |
| Established populations present in 10-50\% of the territory | Established and frequent in part of the territory | 3 |
| Established in more than $50 \%$ of the territory | Widespread | 4 |

Table 7. Invasion extent of non-native species

Species introduced before 1500 were treated as native. Geese presented a problem, because several species are non-native breeders but have large wild wintering populations that are sometimes present till April and may be recorded on early BBS visits. Such birds are all treated as native, except for Barnacle Geese, which, although treated as native in Scotland, were treated as non-native in England and Wales because of their introduced breeding populations in those countries and the scarcity of wild birds there during the BBS season.

## Countryside Survey (vascular plants)

Countryside survey http://www.countrysidesurvey.org.uk/ provides a unique record of plants growing in the British countryside from 1978 to the present day. In a recent report (Countryside Survey Partnership, 2008), the authors observe that "The number of non-native or "alien" plant species recorded in Great Britain has increased greatly in the past sixty years. Most non-native species remain relatively scarce in the CS sampling plots (over 14,000 plots in this analysis), although locally they can be very abundant e.g. Rhododendron and Japanese Knotweed. Together, non-native species now account for nearly $2 \%$ of the vegetation cover of the British countryside'.

In the analysis of the 2007 Countryside Survey, crop plants were included along with other non-native species. For the purpose of the relative proportion indicator, however, crop plants (Table 8) were omitted. The argument for this is that the majority of occurrences of crop species are either as a standing crop or as first-generation descendents of such a crop. Some crop species, especially forest trees, may in future become invasive. Indeed, Corsican Pine is already invasive locally on sand dunes. However, because the great majority of its occurrences are as planted trees, it is retained in the crop category. In the other direction, sycamore is sometimes
grown as a crop, but is much more often naturalized. It is therefore not treated as a crop for the indicator.

| Field crops (major) | Field crops (minor) | Tree crops |
| :--- | :--- | :--- |
| Barley | Buckwheat | Austrian Pine |
| Beet | Carrot | Corsican Pine |
| Cabbage etc. | Flax | Douglas Fir |
| Field bean | Garden Strawberry | European Silver-fir |
| Italian Rye-grass | Millet | Giant Fir |
| Lucerne | Pea | Grey Poplar |
| Maize | Phacelia | Larch (all) |
| Oats | Quinoa | Lawson's Cypress |
| Potato | Rye | Lodgepole Pine |
| Rape | Salsify | Noble Fir |
| Swede, turnip | Sunflower | Pear |
| Wheat | Tomato | Spruce (all) |

Table 8. Crop species omitted from non-native totals in calculating a nonnative indicator based on Countryside Survey

In the course of data analysis, an anomaly emerged with Pineappleweed Matricaria discoidea in 1978. In that year, it was recorded from 14.7\% of the random (X) plots, whereas in 1990 it was found in $3.5 \%$ of these plots. It may have been confused, probably not by the original recorders, with Scentless Mayweed Tripleurospermum inodorum, which is unexpectedly scarce in the data for that year. There was insufficient time for us to consult the original field sheets. In an effort to correct for the discrepancy, Pineappleweed was downweighted in 1978 by assuming that it made up the same proportion of the non-natives as in 1990. On this basis records were treated as $51 \%$ native and $49 \%$ non-native. With the downweighting, it made up $27 \%$ of the 1978 non-native total.

| Region | Definition | Area $\left(\mathrm{km}^{2}\right)$ | Proportion |
| :--- | :--- | ---: | ---: |
| Southern England | Vice-counties 1-34 | 66966 | $29 \%$ |
| Northern England | Vice-counties 36-40, 53-70 | 65249 | $28 \%$ |
| Scotland | Vice-counties 72-112 | 79972 | $34 \%$ |
| Wales | Vice-counties 35, 41-52 | 21161 | $9 \%$ |
| Great Britain | Vice-counties 1-70, 72-112 | 233348 | $100 \%$ |

Table 9. Regions of Great Britain and their area; vice-counties are widely used for biological recording

Plots were weighted so that the contribution of each region was proportional to the area of the region. Southern England (Table 9) was for this purpose defined by Watsonian vice-counties (NBN, 2008), comprising England south of the Severn-Wash line, together with Gloucestershire. Northern England, defined as the rest of England, has a very similar area.

For example, in southern England in 2007, there were 608 random plots, in which 184 non-native occurrences were reported. Thus the mean number of non-natives was on average 184/608 = 0.303 per plot. In calculating the weighted total for all Britain, the contribution of southern England was adjusted for the area of the territory.

Southern England contribution $=0.303 * 66966 / 233348=0.0868$.
The weighted total of non-natives for all regions is the sum of the contributions for Southern England, Northern England, Scotland and Wales. Finally, the proportion of non-natives was calculated as the weighted total of non-natives by the weighted total species counts.

## Non-native bryophytes in the BRC database

Bryophyte recording by the British Bryological Society started in 1960 with the launch of the Society's Mapping Scheme. From then onwards, systematic recording of bryophytes has provided a steady stream of records. For the purposes of calculating non-native frequency, a bryophyte record was taken to be the occurrence of a species in a 10-km square (hectad) in a year. If the same species was found twice in the hectad in a year, the occurrence is counted only once.

As with vascular plants in Countryside Survey, Great Britain was divided into four regions, each of which was weighted in proportion to its area. In each of the component regions (Southern England, Northern England, Scotland, Wales) counts of non-native records and of total records were summed over five years. Calculated proportions are therefore 5-year running means. The non-native proportion for Great Britain was calculated as the weighted mean of the proportions for component regions. Specifically, let $A_{\mathrm{SE}}, A_{\mathrm{NE}}, A_{\mathrm{S}}, A_{\mathrm{w}}$ be the areas of Southern England, Northern England, Scotland and Wales, and let $p_{\mathrm{SE}}, p_{\mathrm{NE}}, p_{\mathrm{S}}, p_{\mathrm{W}}$ be the corresponding proportions of non-native records. Then the non-native proportion for Great Britain is calculated as

$$
p_{\mathrm{GB}}=\left(p_{\mathrm{SE}} A_{\mathrm{SE}}+p_{\mathrm{NE}} A_{\mathrm{NE}}+p_{\mathrm{S}} A_{\mathrm{S}}+p_{\mathrm{W}} A_{\mathrm{W}}\right) / A_{\mathrm{GB}}
$$

where $A_{\mathrm{GB}}$ is the area of Great Britain and

$$
A_{G B}=A_{\mathrm{SE}}+A_{\mathrm{NE}}+A_{\mathrm{S}}+A_{\mathrm{W}}
$$

| Taxonomic group | Dataset | Custodian |
| :---: | :---: | :---: |
| (a) Datasets used for indicator |  |  |
| Marine shore and seabed species | Data Archive for Seabed Species and Habitats (DASSH) | MBA |
| Vascular plants | Countryside Survey | CEH |
| Bryophytes | British Bryological Society national recording scheme | BRC |
| Birds | Breeding Bird Survey | BTO |
| Mammals | Breeding Bird Survey | BTO |
| (b) Other datasets |  |  |
| Marine plankton | Continuous Plankton Recorder Database (Sir Alister Hardy Foundation for Ocean Science) | SAHFOS |
| Crayfish | No single compilation | EA and others |
| Fish | No regular recording; EA have data for a few, including Topmouth Gudgeon | EA and others |
| Amphibians and reptiles | National Amphibian and Reptile Recording Scheme (NARRS) | Herpetological Conservation Trust |
| Aquatic plants | BSBI vascular plant database | BRC |
| Riparian plants | River Habitat Survey | EA |
| Butterflies | Butterflies for the New Millennium | BC |
| Macro-moths | The Rothamsted Insect Survey | Rothamsted Research |
| Macro-moths | Moths Count recording scheme | BC |
| Orthoptera and allies | Orthoptera recording scheme | BRC |
| Aphids | The Rothamsted Insect Survey | Rothamsted Research |
| Terrestrial vascular plants | BSBI vascular plant database | BRC |

Table 10. Datasets with information on non-native species

## Marine organisms

Data from biological survey of Britain's seas and coasts are relatively sparse. Many areas of the coast and seabed are under-surveyed. DASSH (Data Archive for Seabed Species and Habitats) and the NBN Gateway hold marine life datasets for Great Britain. For the frequency indicator, data from NBN Gateway and DASSH within 12 nautical miles ( 22 km ) of the British coast were combined. Most of the data on the NBN Gateway originated from the Marine Nature Conservation Review. There was no discrimination between taxonomic groups but some groups of species were not represented. Records were mainly of benthic and intertidal plants and animals. Records of absence were excluded.

Starting with a fuller set of data, including records from further out to sea, the 12 nautical mile limit was imposed using ESRI ArcGIS software.

Records were assigned to England, Scotland and Wales, using national boundaries supplied by Defra. The data were then transferred from an ArcGIS geodatabase to a series of MySQL database tables and queried via a web-interface using a PHP-based script. The script extracted the number of records and the number of non-natives recorded for each given time period for England, Scotland, Wales and Great Britain.

The list of non-native marine species was derived from the DAISIE list of marine aquatic species for Great Britain, with some additions based on expert knowledge and the scientific literature. The list included 96 marine, diadromous and brackish species from 11 phyla.

## Other datasets

The search for data revealed several national datasets that could in principle be used (Table 10). Those that were eventually selected were chosen because they covered all of Great Britain and were readily available to the consortium. The Environment Agency, Rothamsted Research and the Herpetological Conservation Trust were approached, and expressed a willingness in principle to supply data. However, the short time-scale of the project meant that data could not be used unless fully compiled in an easilysearchable database. This was not the case with the first two sources. An additional problem with the Environment Agency data is that Scotland is excluded. Herpetological Conservation Trust data may well have been suitable, but there was insufficient time to pursue the matter more fully with them. Butterflies and macro-moths are well recorded groups. However, there are no established non-native butterflies, and very few non-native macro-moths. The newly-introduced Oak Processionary Moth Thaumetopoea processionea and the reintroduced Gypsy Moth Lymantria dispar (which may by now have been exterminated) are possibly the only introduced species to have become established since 1900.

The NBN Gateway was a particularly useful resource for checking the spread of invasive species.

## Results

## Non-native birds in Breeding Bird Survey

The graphs in this section show the proportion of birds counted on BBS squares that were of non-native species. Values for 2008 are based on roughly two thirds of the expected final sample.

Proportions of non-native birds in England were first assessed separately for the nine Government Office regions. The proportion of non-natives in London has been rising fast, and since 2003 has exceeded that in other regions. In 2007-8 it was about 3\%, fully twice the average for all England. Many of the London birds were Rose-ringed Parakeets, which are not widely distributed elsewhere. The lowest proportion of non-natives was in the North East, where values were similar to those in Scotland and Wales.

Because of regional differences in the density of BBS squares, trends for England and GB were compiled by first weighting the count values for each region by the reciprocal of the density of BBS squares surveyed there. Weighted counts were added, and the proportion calculated from these.

The proportion of non-native birds in England was roughly three times that in Wales. Scotland had the lowest non-native proportion (Figure 1). The trend was upward in all three countries, with England passing 1\% nonnative in 2000. GB reached the 1\% level in 2007 but, on provisional 2008 data, has subsequently fallen.


Figure 1. Proportion of non-native birds in countries of Great Britain

## Non-native mammals in Breeding Bird Survey

BBS data were used to investigate trends for six of the seven mammal species deemed to be invasive in Britain. The seventh, Coypu, was exterminated during the 1980s, before BBS mammal recording began in 1995.

More than $80 \%$ of BBS observers record mammals, despite this being an optional part of the survey. Most make no special efforts, however, and record only what is noted on bird-counting visits to the square. Many BBS mammal records refer to presence of field signs or to local knowledge that a species was present, rather than to counts of live animals.

All species are under-recorded, but to differing extents. It is not therefore possible to compare native and non-native species, as was done for birds.

Instead, the proportion of sites (1-km squares) occupied was investigated over time, separately for each of the six invasive species. A square was registered as occupied on the strength of BBS evidence from any source: counts on BBS visits, other sightings of live or dead animals, field signs, or local knowledge.

Trends were first assessed separately for the nine Government Office regions. The 'rough and ready' nature of the data did not warrant any weighting procedure when the data were combined into larger regions. Data for 2008 were drawn from less than $20 \%$ of the expected final sample. Proportions for 2001 are biased because of access restrictions following a major outbreak of Foot \& Mouth Disease. In that year, much farmland was inaccessible. In the GB indicator, value of the mammal index for 2001 was interpolated as the average of those for 2000 and 2002.


Figure 2. Mean frequency of six non-native mammal species in Breeding Bird Scheme squares; the peak in 2001 is due to many rural sites being inaccessible in England and Wales during an outbreak of Foot and Mouth Disease

Trends of most species are upward. The spread of Muntjac is shown clearly in data for the East of England, but is less clear in other regions and on broader scales. Other analyses of BBS mammal data have used different approaches (British Trust for Ornithology, 2008).

## Non-native plants in Countryside Survey

Taken over all time periods, the top 10 non-native plants in Countryside Survey account for $82 \%$ of all 2443 non-native occurrences in the random
(X) and stream (S and W) plots in the data (Table 11). Japanese Knotweed is not even one of the top ten in the countryside; it is in 11th place.

| Scientific name | English name | Total | Random | Stream |
| :--- | :--- | ---: | ---: | ---: |
| Acer pseudoplatanus | Sycamore | 644 | 263 | 381 |
| Matricaria discoidea | Pineappleweed | 459 | 415 | 44 |
| Veronica persica | Common Field-speedwell | 432 | 416 | 16 |
| Impatiens glandulifera | Indian Balsam | 122 | 5 | 117 |
| Epilobium brunnescens | New Zealand Willowherb | 97 | 11 | 86 |
| Mimulus guttatus | Monkeyflower | 70 | 1 | 69 |
| Rhododendron ponticum | Rhododendron | 65 | 33 | 32 |
| Epilobium ciliatum | American Willowherb | 63 | 20 | 43 |
| Claytonia sibirica | Pink Purslane | 32 | 1 | 31 |
| Aesculus hippocastanum | Horse-chestnut | 27 | 17 | 10 |
| Total of top 10 species |  | 2011 | 1182 | 829 |
| Total of all NNNCS |  | 2443 | 1426 | 1017 |
| Top 10 as \% of all NNNCS |  | $82.3 \%$ | $82.9 \%$ | $81.5 \%$ |
| Total no. of plots | 16884 | 9513 | 7371 |  |
| Mean no. of NNNCS per plot |  | 0.14 | 0.15 | 0.14 |

Table 11. Counts of occurrence in sample plots for the 10 most frequent non-native non-crop species (NNNCS) in Countryside Survey (all Countryside Surveys 1978-2007).

The proportion of non-native species in England, and therefore in Britain, fell markedly between 1978 and 1990 (Figs 3 and 4). While this is at first sight a surprising result, it is not implausible, because the main contributors to the non-native total in random plots are arable weeds. If crops were being more effectively weeded with selective herbicides after 1978, then the non-native proportion could have fallen.

The lower proportion of non-natives in streamside plots than in random plots applies only in England. In Wales and Scotland, streamside plots had a slightly higher proportions of non-natives than random plots (Figs 4, 5). The discrepancy is particularly marked in southern England, where about 2\% of the flora was non-native in random plots, compared with about $1 \%$ in streamside plots.

## Non-native bryophytes in the BRC database

Scotland had overall a much lower proportion of non-native species than England (Table 12; Fig. 6). Northern England, on the other hand, had a higher non-native proportion than southern England. This probably reflects that fact that the top three non-native bryophytes (Table 13) are all calcifuges, which are relatively scarce in calcareous or intensively farmed countryside.


Figure 3. Great Britain proportions of non-native vascular plant species in random ( X ) and streamside ( S and W ) plots in Countryside Survey


Figure 4. Country proportions of non-native vascular plant species in random (X) plots in Countryside Survey; southern England is defined as in Table 9


Figure 5. Country proportions of non-native vascular plant species in streamside ( S and W ) plots in Countryside Survey; southern England is defined in Table 9

The early phase of increasing non-native proportion, to 1976 (Fig. 6), corresponds to the time when Campylopus introflexus was still increasing. Since then, two liverworts from the Southern Hemisphere, Lophocolea bispinosa and L. semiteres, have begun to spread rapidly. If they soon become widespread throughout Great Britain, then the overall non-native proportion will show a second period of increase in the near future.

| Region | Area <br> $\left(\mathrm{km}^{2}\right)$ | Total <br> NNS | Total <br> all spp. | Proportion <br> NNS | Proportion <br> $1960-89$ | Proportion <br> $1990-2008$ |
| :--- | ---: | ---: | ---: | :---: | :---: | :---: |
| Southern England | 66966 | 2702 | 235043 | $1.15 \%$ | $0.87 \%$ | $1.42 \%$ |
| Northern England | 65249 | 2248 | 155746 | $1.44 \%$ | $1.10 \%$ | $1.83 \%$ |
| Scotland | 79972 | 680 | 234632 | $0.29 \%$ | $0.23 \%$ | $0.43 \%$ |
| Wales | 21161 | 752 | 103952 | $0.72 \%$ | $0.65 \%$ | $0.76 \%$ |
| Total GB | 233348 | 6382 | 729373 | $0.87 \%$ |  |  |
| Weighted GB |  |  |  | $0.90 \%$ | $0.70 \%$ | $1.14 \%$ |
| proportion |  |  |  |  |  |  |

Table 12. Counts of occurrence of non-native bryophyte species in $10-\mathrm{km}$ squares (hectads) of the National Grid, 1960-2008; weighted GB proportions are calculated from regional proportions, weighted by the areas in Table 9

| Non-native Bryophyte | Total | 1960-1990 | 1990-2008 |
| :--- | ---: | ---: | ---: |
| Campylopus introflexus | 3209 | 1029 | 2180 |
| Orthodontium lineare | 2436 | 1093 | 1343 |
| Atrichum crispum | 228 | 139 | 89 |
| Hennediella stanfordensis | 158 | 95 | 63 |
| Hennediella macrophylla | 107 | 80 | 27 |
| Lophocolea semiteres | 95 | 17 | 78 |
| Tortula freibergii | 45 | 19 | 26 |
| Lophocolea bispinosa | 33 | 15 | 18 |
| Riccia crystallina | 17 | 8 | 9 |
| Riccia rhenana | 16 | 12 | 4 |
| Chenia leptophylla | 14 | 6 | 8 |
| Telaranea murphyae | 11 | 8 | 3 |
| Calyptrochaeta apiculata | 9 | 6 | 3 |
| Telaranea tetradactyla | 8 | 6 | 2 |
| Syntrichia amplexa | 6 | 1 | 5 |

Table 13. Counts of non-native bryophyte records; each occurrence in a distinct $10-\mathrm{km}$ square (hectad) and year counts as a separate record


Figure 6. Proportion of bryophyte records that are of non-native species (running means based on 5 -year totals); values for GB are derived from those from the separate countries, weighted by their land area.

## Marine life

Marine organisms are not only far more diverse than any of the terrestrial groups, but marine records came from a wider range of habitats. They were not collected as part of a recording scheme, and therefore also show much larger year-to-year variation in numbers (Table 14). During the 1980s, most
marine records came from Scotland. During the early 1990s, English records were the most numerous. During the late 1990s and early 2000s, Welsh records were most numerous. From 2005 records have come more evenly from England and Wales, but Scottish records were much fewer.

The proportion of non-native records (Figure 7) was much more consistent than the intensity of recording, at least after 1983. Both in England and Scotland, it has fallen since the mid 1980s. The Welsh proportion has varied but does not show this trend.

| Period | England | Scotland | Wales |
| :--- | ---: | ---: | ---: |
| $1980-84$ | 3838 | 8547 | 4658 |
| $1985-89$ | 12546 | 20360 | 2625 |
| $1990-94$ | 25978 | 18159 | 3841 |
| $1995-99$ | 9335 | 12506 | 15499 |
| $2000-04$ | 7450 | 7084 | 10034 |
| $2005-07$ | 12912 | 5808 | 10729 |

Table 14. Average numbers of marine records per year in England, Scotland and Wales, 1980-2007


Figure 7. Proportion of marine records that are of non-native species (running means based on 5 -year running totals); values for GB are derived from those from the separate countries, weighted by the square root of their land area

## Summed extent of invasive non-native species

The summed extent of invasive species (Fig. 8) differed between countries in a similar way to the proportions of non-natives, discussed above. There
were marked differences between the three environments, with marine and freshwater species increasing faster than those on land. Trends for Scotland and Wales were very similar except for marine species, for which Wales was intermediate between Scotland and England. The difference between Scotland and England is particularly marked in the freshwater environment. It should be noted, however, that many English freshwater fish are considered to be invasive in Scotland.





Figure 8. Summed extent of invasive non-native species in countries within Great Britain; one diagram shows totals summed over all environments, the others show totals for marine, freshwater and terrestrial species separately.

## Discussion

## The non-native biota of Great Britain

The non-native biota of Great Britain is moderately well known, but there is also a steady turn-over of species. Except for birds, breeding in the wild is not always well documented. Many non-native birds that breed in the wild fall in category E of the British Ornithologists' Union, namely 'Species that have been recorded as introductions, human-assisted transportees or escapees from captivity, and whose breeding populations (if any) are
thought not to be self-sustaining'. The difficulties of constructing an index from rare and possibly non-established non-native birds are discussed in Appendix 3.

Plants present difficulties of a different kind, in that many species escape from gardens or occur as casuals by docks, but most of them do not become properly established. They may persist at a site for a long time or may be repeatedly reintroduced. Moreover, in field surveys, the origin of shrubs and trees planted in the countryside may not be at all obvious.

These difficulties apply mainly to species that are scattered and rare. In most surveys, the identity of the native flora and fauna is well known, so that non-native species can be separated from others in reported data. Furthermore, good lists of non-native species are available from DAISIE (2009) and earlier enumerations by British authors (Hill et al., 2008, Hill et al., 2005).

It is possible to use the English audit (Hill et al., 2005) to show rates of arrival for major groupings over time (Fig. 9). The dates in the English audit are mostly in fact British first records, because separate dates for establishment in England were not available. The low value for plant arrivals in the 1850s followed a period of very high rates of arrival. The trend for all groups except plants is for increasing rates of arrival in the late 20th century. A similar result for Europe is shown by Hulme et al. (2009).


Figure 9. Rate of arrival of non-native species that subsequently became established in England; terrestrial plants include bryophytes

## Defining the invasive species

The definition of invasive species employed here is that used in the Convention on Biological Diversity, namely 'Invasive alien species (IAS) are species whose introduction and/or spread outside their natural past or present distribution threatens biological diversity'. This is not the same as the definition used by McGeoch, Chown \& Kalwij (2006), who define invasive merely as 'naturalized species that produce reproductive offspring in very large numbers and are able to spread over a considerable area'. They use the term 'transformers' for the subset that also cause harm.

|  | GB non-native risk assessment mechanism | Modified Belgian system |
| :---: | :---: | :---: |
| Web source | http://www.nonnativespecies.org/ | http://ias.biodiversity.be/ |
| Basis of risk assessment | EPPO risk assessment method | Belgian Forum environmental impact assessment |
| Impacts considered | Environmental, Economic and Social | Environmental |
| Impact scoring | Impact 1 (low) to 5 (high) | Impact 1 (low) to 3 (high) |
| Uncertainty | Low (0), Medium (1) or High (2) | Not assessed |
| Format | Excel spreadsheet (soon to be web-based format) | Excel spreadsheet |
| Speed of assessment | Slow | Rapid |
| Risk assessor | Expert opinion coupled with risk analysis panel | Expert opinion |
| Information sources listed | Yes | No (but details available) |
| Consultation and peer review | Yes | No |

Table 15. Comparison of the methods used by the GB non-native risk assessment mechanism and the modified Belgian system used here.

The 'modified Belgian system' for selecting invasive species is, in effect, a risk assessment mechanism, and can be compared with the full GB risk assessment scheme (Table 15). In January 2009, seven species had been assessed by the GB non-native risk assessment mechanism. The results were compared (Table 15).

The GB risk assessment scheme (Baker et al., 2008) is divided into two sections. In the first section (A) the species is screened by means of 14 questions to decide whether a detailed risk assessment is required. Highrisk organisms, capable of causing unacceptable economic, environmental or social impacts, are considered further in section B. Section B contains 51 questions. There are six key questions relating to impact and if these score as 'massive impact' or 'very likely impact' then evaluation of subsidiary questions is not necessary. Every score in section $B$ is linked to a written
comment. The risk analysis is in Excel format. Assessments are peerreviewed.

The two systems produced comparable final scores and highlighted the same species as high risk (Table 16). It is possible that a few invasive species were missed by the screening process used for the indicator. A system for picking them out in future is required.

The modified Belgian system as applied here addresses impacts only on biodiversity. The GB non-native risk assessment mechanism is comprehensive and rigorous. It is vastly more time-consuming. The modified Belgian system was the best available for screening large numbers of species in a short project.

| Species | Risk (GB) | Certainty (GB) | Risk (Belgian) |
| :---: | :---: | :---: | :---: |
| Chinese Mitten Crab Eriocheir sinensis | High | Low | A (high) |
| Chipmunk <br> Tamias sibiricus | Medium | Medium | B (medium) |
| Monk Parakeet <br> Myiopsitta monachus | Medium | Low | B (medium) |
| Red Swamp Crayfish Procambarus clarkii | High | Low | A (high) |
| Rose-ring Parakeet Psittacula krameri | Medium | Low | B (medium) |
| Signal Crayfish Pacifastacus leniusculus | High | Low | A (high) |
| Spiny-cheek Crayfish Orconectes limosus | Medium | Medium | B (medium) |

Table 16: Comparison of results obtained from the GB non-native risk assessment mechanism and the modified Belgian system.

## Other possible indicators

According to McGeoch et al. (2006), the only regional indicator developed at that time was the cumulative number of alien species in Europe since 1900. Such an indicator takes no account of the fact that many non-natives are not persistent.

A more recent major overview of ecosystem indicators in the United States (The H. John Heinz III Center for Science, 2008) suggests using established non-native species as one of 13 core national indicators. Within that category, they propose ( $p .52$ ) three possibilities:

1. The number of new non-native species that become established over time;
2. The area with different numbers of established non-native species; and
3. The area with different proportions of established non-native species, as a percentage of total species.

The first of these has already been discussed, and is plotted above for England (Fig. 9). It is essentially the rate of change of the regional indicator proposed by McGeoch et al. (2006).

The second of the Heinz Center's proposals amounts to charting the nonnative species density for selected areas. They give an example (p. 182) for non-native fish, where the number of non-native species per catchment is mapped, without dividing by the total number of species. A measure of this type could be constructed for British fish using 10-km squares of the National Grid and the data in Davies et al. (2004). Such a measure is similar to those of type 4 in Table 1, and could be applied to any group for which species mapping is considered to be complete. It does not, however, result in an overall summary of trend.

The third of the Heinz Center's indicators is a relativized value, like those used above for the abundance indicator. A map showing this proportion for vascular plants is presented by Preston, Pearman \& Dynes (2002) and shows a strong increase from north to south, except for the Scottish lowlands between Glasgow and Edinburgh, where the proportions are similar to those in southern England.

Various possible measures of impact are discussed in the methods section above. The Heinz Center report (p.52) also suggested the area of forests damaged by non-native pests, and the amount of native vegetation displaced by non-native plants. McGeoch et al. (2006) recommend calculating the number of invasive (in their sense) and transformer species. The summed extent of invasives, used here, is similar in principle to the summed number of invasives. If the 4-point extent scale for extent were reduced to a 1-point scale, then the two would be the same.

A separate point, which is not addressed at all in this report, is that some species such as Hottentot Fig Carpobrotus edulis are invasive in southwest England, but are not invasive in Scotland. Likewise, many species such as Acacia dealbata that are invasive in southern Europe, are cultivated as tender garden plants in England where they show no tendency to spread.

## Choice of datasets and weighting

The proportion of non-natives depends greatly on the scale of sampling. It rises as the plot or sample size increases, because non-native species are mostly rarer than natives. Sampling for the indicator was at widely differing scales, including individual records (marine life), quadrats with area $200 \mathrm{~m}^{2}$ (vascular plants) encounters in 1-km squares (birds and mammals) and frequency in records from $10-\mathrm{km}$ squares (bryophytes). These scales were defined not so much by consideration of what would be best for the indicator, but by the availability of data. However, the fact that non-native proportions were about $1 \%$ in all of these meant that they could be combined in an index based on broadly comparable values. Mammals were an exception. In 2006, the three commonest non-native mammals, Brown Rat, Grey Squirrel and Reeves's Muntjac made up 9\% of counted
individuals and 18\% of the summed frequency of mammals in BBS squares (British Trust for Ornithology, 2008).

Although we are fortunate in Britain to have a wealth of data from biological recording, many existing datasets could not be used. The most frequent reason was that they were for a group such as butterflies, lichens and macro-moths in which the non-native species are either absent or too few to be suitable for inclusion in an indicator. With other groups such as aphids, national data are collected but are not yet in a form where they can rapidly be interrogated.

Weights (Table 2) for the geometric mean were selected to give equal weighting ( 0.4 and 0.4 ) to terrestrial plants and animals and to give a moderate weighting (0.2) to marine life. This is arbitrary, but was intended to be equitable between plants and animals, while at the same time including marine organisms. If comparable data had been available for the freshwater environment, the freshwater component also would have been given weight 0.2

Weights (Table 8) for the regions of Great Britain were used as a means to ensure that the whole territory was evenly represented. This was particularly important for bryophytes and marine organisms, for which recording was in some years very unevenly distributed between England and Scotland. As the proportion of non-natives was markedly different in these two countries, the overall proportion for Great Britain was found to be very unstable without such a weighting. Marine data were weighted not by the area of each region, but by the square root of its area. This was intended as a crude measure of the amount of coastline.

## Trends and differences between countries

For all groups of organisms reported here, England was the country most affected by non-native species. Scotland was the least affected. Wales was intermediate. Most species groups showed a trend over time towards an increasingly non-native biota.

A remarkable exception was vascular plants in Countryside Survey, which had a higher non-native proportion in 1978 than at any later time. The likely explanation is that in 1978 arable fields were weedier than they were in later years. Because two of the commonest non-native plants are arable weeds, this resulted in a decrease in the value of the indicator.

Equally remarkable is the decreasing proportion of non-native marine records over the period since 1983. This may not reflect a genuine decrease in non-native species. One possible explanation is that the later records were based on a larger number of taxonomic groups, including those groups that lacked non-native taxa. Another possible explanation is that earlier records were concentrated near ports and harbours, and that they therefore exaggerated the representation of non-natives. It is also possible that in later years, there was less interest in recording non-native
species, though this seems unlikely. These hypotheses could readily be tested, but this was not possible in the time-scale of the project.

## Conclusions and further developments

Over the period 1990-2007, the mean indexed proportion of records of nonnative species in samples of birds, mammals, plants and marine life rose by $23 \%$ (Fig. 10). Except for mammals, the absolute proportion was still only about $1 \%$ of the total.


Figure 10. Trends in the proportion of non-native species in samples of birds, mammals, plants and marine organisms, 1990-2007

Over the same period, the summed extent of invasive non-native species rose by $40 \%$ (Fig. 11). The increase of invasive species was particularly large in the freshwater and marine environments.

Although non-native species are a potential threat, they are still only a small proportion of the animals and plants to be found in most of the land area and coasts of Great Britain. Vertebrates stand out as the most invasive group.

There are three main areas that deserve further development. The first is to secure sources of additional data, especially freshwater data - e.g. on fish. The second is to make a regular review of the list of invasive species, checking for new arrivals, omissions and species that have lost their harmful character. The third is to improve the analysis of species from recording schemes that do not make a full survey. This is particularly desirable for marine life, where the falling proportion of non-native species in the records did not accord with expectations, and needs confirmation.


Figure 11. Trends in the summed extent of invasive non-native species in marine, freshwater and terrestrial environments, 1960-2008

Updating of the abundance indicator will be facilitated in future by the existence of a central data repository (CDR) for non-native species in Great Britain. This will increase the flow of records from the marine environment. It will establish a flow of records for species that breed in freshwater, such as amphibians, fish and crayfish. It will maintain the flow of records for terrestrial vertebrates. It will not allow for more rapid recording of vascular plants through Countryside Survey, but may increase the capacity of voluntary sector plant recording to the point where an alternative vascular plant indicator could be introduced. Bryophyte recording will not be enhanced by the CDR, but the flow of bryophyte records is already adequate. Annual updates for birds, mammals and bryophytes should be available without a large amount of work.

The CDR should allow for more-or-less automatic updates of the extent to which the high-risk invasive species have spread. The risk register will need to be revised at intervals, both to account for species that have newly arrived or become invasive, and for species that may have lost their invasive character. Such a review could be undertaken in 2010, towards the end of the process of establishing the CDR, and would probably need about 12 days of staff time, at a cost of about $£ 5,000$. It is not included in current costings for the CDR. The year 2010 would be timely, as the impacts indicator was calculated at decadal intervals except for 2008.

If the abundance indicator were extended to include other taxonomic groups, or to allow for substantial reanalysis of data for existing groups, the costs would be more substantial. This is partly because the CDR will record
the presence of non-native species, but does not allow for recording of native species. All but the mammal component of the abundance indicator were derived by dividing non-native totals by native totals. Costs of reanalysing existing data, for example to find the cause of the paradoxical fall in proportion of non-native species in marine data, would also be appreciable. Worthwhile extensions to the existing indicator along these lines would probably cost about $£ 10,000$ to set up, but thereafter the running costs might be quite low.

Given that the establishment and spread of non-native species are relatively slow processes, annual updates to the indicator are unlikely to show much change. Intervals of about five years would be sufficient.

Finally, an indicator based on the rate of establishment of non-native species deserves further consideration. Such an indicator could be reported in other European countries. It would not be so dependent on extensive data sources as the indicators developed here. Even so, it would require substantial further work to determine which species are in fact established.

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## References

Anon. (2003) Review of non-native species policy, report of the working group. Defra Publications, London.
Baker, R.H.A., Black, R., Copp, G.H., Haysom, K.A., Hulme, P.E., Thomas, M.B., Brown, A., Brown, M., Cannon, R.J., Ellis, J., Ellis, M., Ferris, R., Glaves, P., Gozlan, R.E., Holt, J., Howe, L., Knight, J.D., MacLeod, A., Moore, N.P., Mumford, J.D., Murphy, S.T., Parrott, D., Sansford, C.E., Smith, G.C., St-Hilaire, S. \& Ward, N.L. (2008) The UK risk assessment scheme for all non-native species. Biological Invasions - from Ecology to Conservation. (Neobiota, Vol. 7, eds W. Rabitsch, F. Essl \& F. Klingenstein), pp. 46-57. Berlin.
BOU (2009) The British List. BOU species categories. Web page http://www.bou.org.uk/reccats.html.
Branquart, E. (ed.) (2007) Guidelines for environmental impact assessment and list classification of non-native organisms in Belgium. Web publication by Belgian Forum on Invasive Species http://ias.biodiversity.be/ias/documents/ISEIA protocol.pdf
British Trust for Ornithology (2006) Indicators of wild bird populations. Web page http://www.bto.org/research/indicators/index.htm BTO, Thetford.
British Trust for Ornithology (2008) Mammal monitoring through the BBS. http://www.bto.org/bbs/results/mammal monitoring.htm. BTO, Thetford.
Countryside Survey Partnership (2008) Countryside Survey: UK Results from 2007. 2. The national picture. Web publication. http://www.countrysidesurvey.org.uk/.
DAISIE (2009) Handbook of alien species in Europe. Springer, Dordrecht.
Davies, C., Shelley, J., Harding, P., McLean, I., Gardiner, R. \& Peirson, G. (2004) Freshwater fishes in Britain - the species and their distribution. Harley Books, Colchester.
European Environment Agency (2007) Halting the loss of biodiversity by 2010: proposal for a first set of indicators to monitor progress in Europe. EEA, Copenhagen.
Hill, M., Baker, R., Broad, G., Chandler, P.J., Copp, G.H., Ellis, J., Jones, D., Hoyland, C., Laing, I., Longshaw, M., Moore, N., Parrott, D., Pearman, D., Preston, C., Smith, R.M. \& Waters, R. (2005) Audit of non-native species in England. English Nature Research Reports, 662, 1-81.

Hill, M.O., Ames, S., Bacon, J., Botham, M.S., Lindsley-Leake, S., Marchant, J.H., Preston, C.D., Rehfisch, M.M., Roy, D.B., Roy, H.E., Tyler-Walters, H. \& Wright, L. (2008) Non-native Species Portal Feasibility Study. Report to Defra. CEH Monks Wood, Huntingdon.
Hollingsworth, M.L. \& Bailey, J.P. (2000) Hybridisation and clonal diversity in some introduced Fallopia species (Polygonaceae). Watsonia, 23, 111-121.
Hulme, P.E., Roy, D.B., Cunha, T. \& Larsson, T.-B. (2009) A pan-European inventory of alien species: rationale, implementation and implications for managing biological invasions. Handbook of alien species in Europe (ed DAISIE), pp. 1-14. Springer, Dordrecht
McGeoch, M.A., Chown, S.L. \& Kalwij, J.M. (2006) A global indicator for biological invasion. Conservation Biology, 20, 1635-1646.
NBN (2008) Watsonian Vice County Boundaries http://www.nbn.org.uk/Useful-things/Mapping/Watsonian-Boundary-CD-(1).aspx. National Biodiversity Network.
Noble, D.G., Newson, S.E. \& Gregory, R.D. (2004) Approaches to dealing with disappearing and invasive species in the UK's indicators of wild bird populations. Report to Defra by BTO \& RSPB. http://www.defra.gov.uk/environment/statistics/wildlife/download/pdf/WildBirdIndicators20041 0. pdf.

Preston, C.D., Pearman, D.A. \& Dines, T.D. (2002) New atlas of the British and Irish Flora. Oxford University Press, Oxford.
The H. John Heinz III Center for Science, Economics and the Environment. (2008) The state of the nation's ecosystems 2008. Island Press, Washington, DC.
UK Government (2008) The invasive non-native species framework strategy for Great Britain. Department for Environment, Food and Rural Affairs, London.

## List of abbreviations

| BBS | Breeding Bird Survey |
| :--- | :--- |
| BOU | British Ornithologists' Union |
| BRC | Biological Records Centre |
| BTO | British Trust for Ornithology |
| CBD | Convention on Biological Diversity |
| CEH | Centre for Ecology and Hydrology |
| CS | Countryside Survey |
| DAISIE | Delivering Alien Invasive Species Inventories for Europe |
| DASSH | Data Archive for Seabed Species and Habitats |
| EA | Environment Agency |
| EEA | European Environment Agency |
| EPPO | European and Mediterranean Plant Protection Organization |
| GB | Great Britain |
| IAS | Invasive Alien Species |
| JNCC | Joint Nature Conservation Committee |
| MBA | Marine Biological Association |
| NBN | National Biodiversity Network |
| NNS | Non-native Species |
| RBBP | Rare Breeding Birds Panel |
| SAHFOS | Sir Alister Hardy Foundation for Ocean Science |

## Appendix 1. Definitions for impact scoring

Invasion potential is rated in 4 categories, namely Dispersal potential, Colonization of natural and semi-natural habitats, Adverse impacts on native species and Alteration of ecosystem function. As explained in Appendix 2, two of the categories are subdivided, and the maximum impact for each of the subdivisions is taken as the score for the category.

Table A1. Dispersion potential

| Definition | Mnemonic | Score |
| :--- | :--- | :---: |
| Low risk. The species does not spread in the <br> environment because of poor dispersal capacity <br> or low reproductive potential. | Species does not spread <br> naturally | 1 |
| Medium risk. Unlikely to colonize remote places. <br> Natural dispersal rarely exceeds more than 1 km <br> per year, and species is not regularly transported <br> within the territory by humans. The species can <br> however become locally invasive because of a <br> strong reproductive potential. | Species spreads up to 1 km <br> per year | 2 |
| High risk. The species is highly fecund, can easily <br> disperse by active or passive means over <br> distances >1 km per year and initiate new <br> populations. Means of dispersal include wind, <br> water, animal movements, translocation by <br> humans or accidental transport by human agency. | Species spreads naturally or <br> by human agency >1 km per <br> year | 3 |

Table A2. Colonization of natural and semi-natural habitats

| Definition | Mnemonic | Score |
| :--- | :--- | :---: |
| Low risk. Populations of the non-native species <br> are restricted to man-made habitats (low <br> conservation value). | Restricted to man-made <br> habitats | 1 |
| Medium risk. Populations of the non-native <br> species are usually confined to habitats with low <br> or medium conservation value and may <br> occasionally colonize natural and semi-natural <br> habitats. | May occasionally colonize <br> natural and semi-natural <br> habitats | 2 |
| High risk. The non-native species often colonizes <br> natural or semi-natural habitats (i.e. many sites of <br> at least one natural or semi-natural habitat could <br> be readily colonized by the species when source <br> populations are present in the vicinity). | Often colonizes natural or <br> semi-natural habitats | 3 |

Table A3. Adverse impacts on native species

| Definition | Mnemonic | Score |
| :--- | :--- | :---: |
| Low risk. Data from invasion histories suggest that <br> any negative impact on native populations is <br> negligible. | Negative impact on native <br> populations negligible | 1 |
| Medium risk. The non-native species is known to <br> cause local changes (< 80\%) in population <br> abundance, growth or distribution of one or <br> several native species. This effect is thought to be <br> reversible, or the affected species are either <br> common or ruderal or both. | Causes reversible changes of <br> $<80 \%$ in population <br> abundance of native species | 2 |
| High risk. The non-native species can cause local <br> severe (> 80\%) population declines of valued or <br> rare species, or may reduce local species <br> richness. At a regional scale, it may cause <br> species decline. When such non-native species <br> are established, their impacts on native <br> biodiversity are unlikely to be reversible. | Often causes changes of <br> $>80 \%$ in population <br> abundance of native species, <br> including rare species | 3 |

Table A4. Alteration of ecosystem function

| Definition | Mnemonic | Score |
| :--- | :--- | :---: |
| Low risk. The impact on ecosystem processes <br> and structures is considered as negligible. | Ecosystem impact negligible | 1 |
| Medium risk. The impact on ecosystem processes <br> and structures is moderate and considered as <br> easily reversible. | Ecosystem impact moderate, <br> reversible | 2 |
| High risk. The impact on ecosystem processes <br> and structures is strong and difficult to reverse. | Ecosystem impact strong, <br> difficult to reverse | 3 |

Note: When the impact is variable, depending on the type of ecosystem, one should consider the effect on those ecosystems that are most vulnerable.

## Appendix 2. Scoring species for invasiveness

Species are rated in the categories specified in Table 3, using the scoring system outlined in Appendix 1. In addition to the standard scores, the following conversions were used:

D data deficient
U adverse impact unlikely
L adverse impact likely
score 0
score 1
score 2.

Column headings are shown as numbers as follows:
1 Dispersal potential
2 Colonization of natural and semi-natural habitats
3.1 Adverse impacts on native species: predation and herbivory
3.2 Adverse impacts on native species: competition
3.3 Adverse impacts on native species: disease vector
3.4 Adverse impacts on native species: genetic pollution

3 Adverse impacts on native species: maximum of 3.1, 3.2, 3.3, 3.4
4.1 Alteration of ecosystem function: nutrient cycling
4.2 Alteration of ecosystem function: physical alteration
4.3 Alteration of ecosystem function: successions
4.4 Alteration of ecosystem function: food webs

4 Alteration of ecosystem function: maximum of 4.1, 4.2, 4.3, 4.4
The column labelled Sum is the sum of columns 1, 2, 3 and 4. The column labelled Category assigns a threat category in accordance with the sum:

A invasive
B potential threat
C low threat

Sum = 11 or 12
Sum = 9 or 10
Sum = 8 or less.

Table A5. Scoring for invasiveness: marine organisms

| Scientific name | English name | Group | 1 | 2 | 3.1 | 3.2 | 3.3 | 3.4 | 3 | 4.1 | 4.2 | 4.3 | 4.4 | 4 | Sum | Category |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (a) Marine plants |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Sargassum muticum | Jap Weed/ Wire Weed | brown alga | 3 | 3 | - | 3 | D | 1 | 3 | 2 | 3 | 3 | 2 | 3 | 12 | A |
| Undaria pinnatifida | Japanese Kelp/ Wakame | brown alga | 3 | 2 | - | 3 | D | 1 | 3 | 2 | 2 | 3 | 2 | 3 | 11 | A |
| Codium fragile ssp. tomentosoides | Green Sea Fingers | green alga | 3 | 3 | - | 3 | D | D | 3 | 2 | 3 | 3 | 2 | 3 | 12 | A |
| (b) Marine animals |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Tricellaria inopinata | (a bryozoan) | bryozoan | 3 | 2 | 1 | 3 | 1 | D | 3 | 2 | 2 | 2 | 3 | 3 | 11 | A |
| Watersipora subtorquata | (a bryozoan) | bryozoan | 3 | 2 | 1 | 3 | 1 | D | 3 | 2 | 3 | 3 | 2 | 3 | 11 | A |
| Corophium sextonae | (an amphipod) | crustacean | 3 | 3 | D | 3 | D | D | 3 | 2 | 2 | D | D | 2 | 11 | A |
| Elminius modestus | (an acorn barnacle) | crustacean | 3 | 3 | 1 | 3 | D | 1 | 3 | 1 | 2 | 2 | 2 | 2 | 11 | A |
| Eriocheir sinensis | Chinese Mitten Crab | crustacean | 3 | 3 | 3 | 3 | D | 1 | 3 | D | 3 | 1 | 2 | 3 | 12 | A |
| Gammarus tigrinus | (an amphipod) | crustacean | 3 | 3 | 2 | 3 | D | D | 3 | D | 1 | 1 | 2 | 2 | 11 | A |
| Rhithropanopeus harrisii | Dwarf Crab | crustacean | 3 | 3 | 2 | 2 | 3 | D | 3 | D | 2 | 2 | 2 | 2 | 11 | A |
| Solidobalanus fallax | (a barnacle) | crustacean | 3 | 3 | 1 | 3 | D | D | 3 | D | 2 | 2 | D | 2 | 11 | A |
| Crassostrea gigas | Pacific Oyster | mollusc | 3 | 3 | 1 | 2 | D | 1 | 2 | 2 | 3 | 3 | 3 | 3 | 11 | A |
| Crepidula fornicata | Slipper Limpet | mollusc | 3 | 3 | 1 | 3 | D | 1 | 3 | 2 | 3 | 2 | 3 | 3 | 12 | A |
| Rapana venosa | Rapa Whelk | mollusc | 3 | 2 | 3 | 2 | 2 | 1 | 3 | 1 | 1 | 1 | 3 | 3 | 11 | A |
| Anguillicola crassus | Swim-Bladder Nematode | nematode | 3 | 3 | 3 | 1 | 3 | D | 3 | D | 1 | 1 | 2 | 2 | 11 | A |
| Botrylloides violaceus | (a tunicate) | tunicate | 3 | 3 | 1 | 3 | D | 1 | 3 | 2 | 2 | 2 | 2 | 2 | 11 | A |
| Corella eumyota | (a tunicate) | tunicate | 3 | 3 | 1 | 3 | D | 1 | 3 | 2 | 2 | 2 | 3 | 3 | 12 | A |
| Didemnum vexillum | (a tunicate) | tunicate | 3 | 3 | 1 | 3 | D | 1 | 3 | 2 | 2 | 3 | 3 | 3 | 12 | A |
| Styela clava | Leathery Sea Squirt | tunicate | 3 | 2 | 1 | 3 | D | 1 | 3 | 2 | 3 | 2 | 2 | 3 | 11 | A |

Table A6. Scoring for invasiveness: freshwater organisms

| Scientific name | English name | Group | 1 | 2 | 3.1 | 3.2 | 3.3 | 3.4 | 3 | 4.1 | 4.2 | 4.3 | 4.4 | 4 | Sum | Category |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (c) Freshwater plants |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Azolla filiculoides | Water Fern | plant | 3 | 2 | - | 2 | 1 | 1 | 2 | 2 | 3 | 3 | 3 | 3 | 10 | B |
| Crassula helmsii | New Zealand Pigmyweed | plant | 3 | 3 | - | 3 | 1 | 1 | 3 | 2 | 2 | 2 | 2 | 2 | 11 | A |
| Elodea canadensis | Canadian Waterweed | plant | 3 | 3 | - | 2 | 1 | 1 | 2 | 2 | 2 | 2 | 2 | 2 | 10 | B |
| Elodea nuttallii | Nuttall's Waterweed | plant | 3 | 3 | - | 2 | 1 | 1 | 2 | 2 | 2 | 2 | 2 | 2 | 10 | B |
| Hydrocotyle ranunculoides | Floating Pennywort | plant | 3 | 2 | - | 3 | 1 | 1 | 3 | 2 | 3 | 2 | 2 | 3 | 11 | A |
| Lemna minuta | Least Duckweed | plant | 3 | 3 | - | 2 | 1 | 1 | 2 | 1 | 2 | 2 | 2 | 2 | 10 | B |
| Ludwigia grandiflora | Uruguayan Hampshire-Purslane | plant | 3 | 3 | 1 | 3 | 1 | 1 | 3 | 3 | 3 | 3 | 1 | 3 | 12 | A |
| (d) Freshwater animals |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Orconectes limosus | Spiny Cheeked Crayfish | crustacean | 3 | 3 | 1 | 1 | 2 | 1 | 2 | 1 | 2 | 1 | 2 | 2 | 10 | B |
| Pacifastacus leniusculus | Signal Crayfish | crustacean | 3 | 3 | 2 | 2 | 3 | 1 | 3 | 1 | 3 | 2 | 1 | 3 | 12 | A |
| Procambarus clarkii | Red Swamp Crayfish | crustacean | 3 | 3 | 1 | 3 | 3 | 1 | 3 | 1 | 3 | 1 | 1 | 3 | 12 | A |
| Corbicula fluminea | Asian Clam | mollusc | 3 | 3 | 1 | 2 | 1 | 1 | 2 | 1 | 2 | 1 | 3 | 3 | 11 | A |
| Dreissena polymorpha | Zebra Mussel | mollusc | 3 | 3 | 1 | 3 | 1 | 1 | 3 | 2 | 3 | 2 | 2 | 3 | 12 | A |
| Carassius auratus | Goldfish/Prussian Carp | fish | 2 | 2 | 1 | 3 | 3 | 3 | 3 | 1 | 1 | 1 | 2 | 2 | 9 | B |
| Lepomis gibbosus | Pumpkinseed | fish | 2 | 3 | 3 | 2 | 1 | 1 | 3 | 1 | 1 | 1 | 1 | 1 | 9 | B |
| Misgurnus fossilis | European Weatherfish | fish | 2 | 2 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 6 | C |
| Pseudorasbora parva | Topmouth Gudgeon | fish | 2 | 3 | 3 | 3 | 3 | 1 | 3 | 1 | 1 | 1 | 3 | 3 | 11 | A |
| Salvelinus fontinalis | Brook Trout | fish | 1 | 3 | 3 | 3 | 1 | 3 | 3 | 2 | 2 | 2 | 2 | 2 | 9 | B |
| Sander lucioperca | Pikeperch (Zander) | fish | 3 | 3 | 3 | 3 | 1 | 1 | 3 | 1 | 1 | 1 | 2 | 2 | 11 | A |
| Silurus glanis | European (Wels) Catfish | fish | 2 | 2 | 3 | 3 | 3 | 1 | 3 | 1 | 1 | 1 | 3 | 3 | 10 | B |
| Alytes obstetricans | Midwife Toad | amphibian | 1 | 3 | 1 | 1 | 2 | 1 | 2 | 1 | 1 | 1 | 2 | 2 | 8 | C |
| Lithobates catesbeianus | American Bullfrog | amphibian | 2 | 3 | 3 | 3 | 3 | 1 | 3 | 1 | 1 | 1 | 3 | 3 | 11 | A |
| Triturus carnifex | Italian Crested Newt | amphibian | 2 | 3 | 1 | 1 | 1 | 3 | 3 | 1 | 1 | 1 | 1 | 1 | 9 | B |
| Xenopus laevis | African Clawed Frog | amphibian | 1 | 2 | 3 | 2 | 2 | 1 | 3 | 1 | 1 | 1 | 2 | 2 | 8 | C |
| Trachemys scripta | Common Slider Turtle | reptile | 2 | 3 | 3 | 3 | 2 | 1 | 3 | 1 | 1 | 1 | 3 | 3 | 11 | A |
| Trechemys scripta elegans | Red Eared Terrapin | reptile | 2 | 3 | 1 | D | 2 | 1 | 2 | 1 | 1 | 1 | 1 | 1 | 8 | C |

Table A7. Scoring for invasiveness: terrestrial plants

| Scientific name | English name | Group | 1 | 2 | 3.1 | 3.2 | 3.3 | 3.4 | 3 | 4.1 | 4.2 | 4.3 | 4.4 | 4 | Sum | Category |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (e) Terrestrial plants |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Acaena novae-zelandiae | Pirri-Pirri-Bur | plant | 2 | 3 | - | 3 | 1 | 1 | 3 | 2 | 2 | 1 | 1 | 2 | 10 | B |
| Acer pseudoplatanus | Sycamore | plant | 3 | 3 | - | 2 | 1 | 1 | 2 | 1 | 2 | 2 | 2 | 2 | 10 | B |
| Ailanthus altissima | Tree Of Heaven | plant | 2 | 3 | - | 3 | 1 | 1 | 3 | 2 | 2 | 2 | 2 | 2 | 10 | B |
| Allium paradoxum | Few-Flowered Garlic | plant | 3 | 2 | - | 2 | 1 | 1 | 2 | 1 | 1 | 2 | 1 | 2 | 9 | B |
| Allium triquetrum | Three-Cornered Garlic | plant | 3 | 2 | - | 2 | 1 | 1 | 2 | 1 | 1 | 2 | 1 | 2 | 9 | B |
| Ambrosia artemisiifolia | Common Ragweed | plant | 2 | 1 | - | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 5 | C |
| Buddleja davidii | Butterfly-Bush | plant | 3 | 2 | - | 2 | 1 | 1 | 2 | 1 | 2 | 2 | 2 | 2 | 9 | B |
| Carpobrotus edulis | Hottentot Fig | plant | 3 | 3 | - | 3 | 1 | 1 | 3 | 3 | 3 | 3 | 2 | 3 | 12 | A |
| Cornus sericea | Red-Osier Dogwood | plant | 2 | 2 | - | 3 | 1 | 1 | 3 | 2 | 2 | 2 | 2 | 2 | 9 | B |
| Cortaderia selloana | Pampas Grass | plant | 3 | 3 | - | 2 | 1 | 1 | 2 | 2 | 2 | 2 | 2 | 2 | 10 | B |
| C. $x$ crocosmiiflora | Montbretia | plant | 3 | 2 | - | 2 | 1 | 1 | 2 | 1 | 1 | 2 | 1 | 2 | 9 | B |
| Disphyma crassifolium | Purple Dewplant | plant | 3 | 3 | - | 3 | 1 | 1 | 3 | 2 | 2 | 2 | 2 | 2 | 11 | A |
| Fallopia japonica | Japanese Knotweed S.L. | plant | 3 | 2 | - | 3 | 1 | 2 | 3 | 2 | 3 | 2 | 2 | 3 | 11 | A |
| Gaultheria shallon | Shallon | plant | 2 | 3 | - | 3 | 1 | 1 | 3 | 1 | 1 | 2 | 1 | 2 | 10 | B |
| Heracleum mantegazzianum | Giant Hogweed | plant | 3 | 3 | - | 3 | 1 | 2 | 3 | 2 | 2 | 2 | 2 | 2 | 11 | A |
| Hyacinthoides hispanica | Spanish Bluebell | plant | 2 | 3 | - | 2 | 1 | 3 | 3 | 1 | 1 | 1 | 1 | 1 | 9 | B |
| Impatiens glandulifera | Himalayan Balsam | plant | 3 | 3 | - | 2 | 1 | 1 | 2 | 2 | 3 | 2 | 2 | 3 | 11 | A |
| Myriophyllum aquaticum | Parrot's-Feather | plant | 3 | 2 | - | 3 | 1 | 1 | 3 | 2 | 3 | 2 | 2 | 3 | 11 | A |
| Oxalis pes-caprae | Bermuda Buttercup | plant | 3 | 2 | - | 2 | 1 | 1 | 2 | 1 | 1 | 1 | 2 | 2 | 9 | B |
| Paspalum paspalodes | Knotgrass | plant | 2 | 2 | - | 2 | 1 | 1 | 2 | 2 | 2 | 2 | 2 | 2 | 8 | C |
| Persicaria wallichii | Himalayan Knotweed | plant | 3 | 2 | - | 3 | 1 | 1 | 3 | 2 | 2 | 2 | 2 | 2 | 10 | B |
| Petasites fragrans | Winter Heliotrope | plant | 2 | 2 | - | 3 | 1 | 1 | 3 | 2 | 2 | 2 | 2 | 2 | 9 | B |
| Prunus serotina | Black Cherry | plant | 2 | 2 | - | 2 | 1 | 1 | 2 | 2 | 2 | 3 | 2 | 3 | 9 | B |
| Quercus cerris | Turkey Oak | plant | 3 | 3 | - | 2 | 2 | 1 | 2 | 2 | 2 | 2 | 2 | 2 | 10 | B |
| Quercus ilex | Evergreen Oak | plant | 3 | 3 | - | 3 | 1 | 1 | 3 | 1 | 2 | 2 | 2 | 2 | 11 | A |
| Rhododendron ponticum | Rhododendron | plant | 2 | 3 | - | 3 | 1 | 1 | 3 | 3 | 3 | 3 | 2 | 3 | 11 | A |
| Rosa rugosa | Japanese Rose | plant | 2 | 3 | - | 3 | 1 | 1 | 3 | 2 | 3 | 2 | 2 | 3 | 11 | A |

Table A8. Scoring for invasiveness: terrestrial animals

| Scientific name | English name | Group | 1 | 2 | 3.1 | 3.2 | 3.3 | 3.4 | 3 | 4.1 | 4.2 | 4.3 | 4.4 | 4 | Sum | Category |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (f) Terrestrial animals |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Arthurdendyus triangulata | New Zealand Flatworm | flatworm | 2 | 3 | 3 | D | U | U | 3 | L | L | 1 | 3 | 3 | 11 | A |
| Anoplophora chinensis | Citrus Longhorn Beetle | insect | 2 | 2 | 3 | D | U | U | 3 | 1 | L | 1 | 1 | 2 | 9 | B |
| Harmonia axyridis | Harlequin Ladybird | insect | 3 | 3 | 3 | 3 | 1 | 1 | 3 | 1 | 1 | 1 | L | 2 | 11 | A |
| Thaumetopoea processionea | Oak Processionary Moth | insect | 2 | 3 | 3 | D | U | U | 3 | 1 | 1 | 1 | 1 | 1 | 9 | B |
| Agapornis roseicollis | Rosy-Faced Lovebird | bird | 3 | 3 | U | L | D | D | 2 | U | 1 | U | U | 1 | 9 | B |
| Aix galericulata | Mandarin Duck | bird | 3 | 3 | 1 | L | D | 1 | 2 | 1 | 1 | 1 | 1 | 1 | 9 | B |
| Aix sponsa | Wood Duck | bird | 3 | 3 | U | L | D | 1 | 2 | 1 | 1 | 1 | 1 | 1 | 9 | B |
| Alectoris chukar | Chukar Partridge | bird | 3 | 2 | 2 | L | D | 1 | 2 | 1 | 1 | 1 | 1 | 1 | 8 | C |
| Alectoris rufa | Red-Legged Partridge | bird | 3 | 2 | 2 | L | D | 1 | 2 | 1 | 1 | 1 | 1 | 1 | 8 | C |
| Alopochen aegyptiaca | Egyptian Goose | bird | 3 | 3 | 1 | 2 | D | 1 | 2 | 2 | 2 | 1 | L | 2 | 10 | B |
| Anas cyanoptera | Cinnamon Teal | bird | 3 | 3 | U | U | D | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 8 | C |
| Anas discors | Blue-Winged Teal | bird | 3 | 3 | U | U | D | 2 | 2 | 1 | 1 | 1 | 1 | 1 | 9 | B |
| Anser albifrons | Greater White-Fronted Goose | bird | 3 | 3 | 1 | 1 | D | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 8 | C |
| Anser brachyrhynchus | Pink-Footed Goose | bird | 3 | 3 | 1 | 1 | D | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 8 | C |
| Anser cygnoides | Swan Goose | bird | 3 | 3 | D | 2 | D | 2 | 2 | D | D | D | D | 0 | 8 | C |
| Anser erythropus | Lesser White-Fronted Goose | bird | 3 | 3 | 1 | 1 | D | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 8 | C |
| Anser fabalis | Bean Goose | bird | 3 | 3 | 1 | 1 | D | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 8 | C |
| Anser indicus | Bar-Headed Goose | bird | 3 | 3 | 1 | L | D | 1 | 2 | D | D | D | D | 0 | 8 | C |
| Aratinga acuticaudata | Blue-Crowned Parakeet | bird | 3 | 3 | L | L | 1 | U | 2 | 1 | L | 1 | L | 2 | 10 | B |
| Athene noctua | Little Owl | bird | 3 | 3 | 2 | 2 | D | 1 | 2 | 1 | 1 | 1 | 2 | 2 | 10 | B |
| Branta canadensis | Canada Goose | bird | 3 | 3 | 2 | 2 | L | 2 | 2 | 3 | 2 | L | L | 3 | 11 | A |
| Branta leucopsis | Barnacle Goose | bird | 3 | 3 | 2 | 1 | D | 2 | 2 | L | 1 | 1 | 1 | 2 | 10 | B |
| Branta ruficollis | Red-Breasted Goose | bird | 3 | 3 | 1 | 1 | D | 1 | 1 | D | D | D | D | 0 | 7 | C |
| Bubo bubo | Eurasian Eagle-Owl | bird | 3 | 3 | 2 | 2 | D | 1 | 2 | 1 | 1 | 1 | L | 2 | 10 | B |
| Buteo jamaicensis | Red-Tailed Hawk | bird | 3 | 3 | L | L | D | 2 | 2 | 1 | 1 | 1 | 1 | 1 | 9 | B |
| Cairina moschata | Muscovy Duck | bird | 3 | 2 | 1 | L | D | 1 | 2 | 1 | 1 | 1 | 1 | 1 | 8 | C |
| Chen caerulescens | Snow Goose | bird | 3 | 3 | 1 | L | D | 1 | 2 | D | D | D | D | 0 | 8 | C |
| Chen canagica | Emperor Goose | bird | 3 | 3 | 1 | D | D | 1 | 1 | D | D | D | D | 0 | 7 | C |
| Chen rossii | Ross's Goose | bird | 3 | 3 | 1 | L | D | 1 | 2 | D | D | D | D | 0 | 8 | C |

Table A8 (continued). Scoring for invasiveness: terrestrial animals

| Scientific name | English name | Group | 1 | 2 | 3.1 | 3.2 | 3.3 | 3.4 | 3 | 4.1 | 4.2 | 4.3 | 4.4 | 4 | Sum | Category |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Chloephaga picta | Upland Goose | bird | 3 | 3 | 1 | L | D | 1 | 2 | 1 | 1 | 1 | 1 | 1 | 9 | B |
| Chrysolophus amherstiae | Lady Amherst's Pheasant | bird | 3 | 3 | 2 | 1 | D | 1 | 2 | 1 | 2 | 1 | 1 | 2 | 10 | B |
| Chrysolophus pictus | Golden Pheasant | bird | 3 | 3 | 2 | 1 | D | 1 | 2 | 1 | 2 | 1 | 1 | 2 | 10 | B |
| Colinus virginianus | Northern Bobwhite | bird | 3 | 2 | 1 | 1 | D | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 7 | C |
| Cygnus atratus | Black Swan | bird | 3 | 3 | D | L | D | 2 | 2 | 2 | 1 | 1 | 1 | 2 | 10 | B |
| Cygnus buccinator | Trumpeter Swan | bird | 3 | 3 | D | L | D | L | 2 | 1 | 1 | 1 | 1 | 1 | 9 | B |
| Gallus gallus | Red Junglefowl | bird | 3 | 3 | 1 | 1 | D | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 8 | C |
| Garrulax formosus | Red-Winged Laughingthrush | bird | 3 | 3 | 1 | L | D | 1 | 2 | 1 | 1 | 1 | L | 2 | 10 | B |
| Lophura nycthemera | Silver Pheasant | bird | 3 | 3 | 2 | 1 | D | 1 | 2 | 1 | 2 | 1 | 1 | 2 | 10 | B |
| Melopsittacus undulatus | Budgerigar | bird | 3 | 3 | 2 | 2 | D | 1 | 2 | 1 | 1 | 1 | L | 2 | 10 | B |
| Myiopsitta monachus | Monk Parakeet | bird | 3 | 3 | L | L | 1 | U | 2 | 1 | L | 1 | L | 2 | 10 | B |
| Netta rufina | Red-Crested Pochard | bird | 3 | 3 | 1 | 1 | D | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 8 | C |
| Numida meleagris | Helmeted Guineafowl | bird | 3 | 2 | 1 | 1 | D | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 7 | C |
| Nycticorax nycticorax | Black-Crowned Night Heron | bird | 3 | 3 | U | U | D | 1 | 1 | D | D | D | D | 0 | 7 | C |
| Oxyura jamaicensis | Ruddy Duck | bird | 3 | 3 | 1 | 1 | D | 3 | 3 | 1 | 1 | 1 | L | 2 | 11 | A |
| Pavo cristatus | Indian Peafowl | bird | 3 | 3 | 1 | 1 | D | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 8 | C |
| Psittacula eupatria | Alexandrine Parakeet | bird | 3 | 3 | L | L | 1 | U | 2 | 1 | L | 1 | L | 2 | 10 | B |
| Psittacula krameri | Rose-Ringed Parakeet | bird | 3 | 3 | 2 | 2 | 2 | U | 2 | 1 | 2 | 1 | 2 | 2 | 10 | B |
| Syrmaticus reevesii | Reeves's Pheasant | bird | 3 | 3 | L | 1 | D | 1 | 2 | 1 | 1 | 1 | 1 | 1 | 9 | B |
| Tadorna ferruginea | Ruddy Shelduck | bird | 3 | 3 | 1 | L | D | 1 | 2 | D | D | D | D | 0 | 8 | C |
| Alopex lagopus | Arctic Fox | mammal | 2 | 3 | 3 | L | L | 2 | 3 | 1 | 1 | 1 | L | 2 | 10 | B |
| Capra ibex | Ibex | mammal | 1 | 2 | 2 | L | L | 2 | 2 | 1 | 1 | 2 | L | 2 | 7 | C |
| Cervus canadensis | Wapiti | mammal | 2 | 3 | 3 | L | L | 3 | 3 | 2 | 2 | 2 | L | 2 | 10 | B |
| Cervus nippon | Sika Deer | mammal | 3 | 3 | 3 | L | L | 3 | 3 | 2 | 2 | 2 | L | 2 | 11 | A |
| Cynomys ludovicianus | Black-Tailed Prairie Dog | mammal | 1 | 2 | 3 | L | L | 1 | 3 | 2 | 2 | 2 | L | 2 | 8 | C |
| Glaucomys sp. | Flying Squirrel | mammal | 2 | 2 | 2 | L | L | 1 | 2 | 1 | 1 | 1 | L | 2 | 8 | C |
| Glis glis | Fat Dormouse | mammal | 2 | 2 | 2 | L | L | 1 | 2 | 1 | 1 | 1 | U | 1 | 7 | C |

Table A8 (continued). Scoring for invasiveness: terrestrial animals

| Scientific name | English name | Group | 1 | 2 | 3.1 | 3.2 | 3.3 | 3.4 | 3 | 4.1 | 4.2 | 4.3 | 4.4 | 4 | Sum | Category |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Hemitragus jemlahicus | Himalayan Tahr | mammal | 1 | 2 | 2 | L | L | 1 | 2 | 1 | 1 | 2 | L | 2 | 7 | C |
| Hydropotes inermis | Chinese Water Deer | mammal | 3 | 3 | 2 | L | L | 1 | 2 | 1 | 2 | 2 | L | 2 | 10 | B |
| Hystrix brachyura | Hialayan Porcupine | mammal | 1 | 2 | 2 | L | L | 1 | 2 | 1 | 2 | 1 | L | 2 | 7 | C |
| Hystrix cristata | Crested Porcupine | mammal | 1 | 2 | 2 | L | L | 1 | 2 | 1 | 2 | 1 | L | 2 | 7 | C |
| Lepus californicus | Black-Tailed Jack Rabbit | mammal | 2 | 2 | 2 | L | L | 1 | 2 | 1 | 2 | 2 | L | 2 | 8 | C |
| Macropus rufogriseus | Red Necked Wallaby | mammal | 2 | 2 | 1 | L | U | 1 | 2 | 1 | 1 | 1 | L | 2 | 8 | C |
| Meriones unguiculatus | Mongolian Gerbil | mammal | 2 | 1 | 1 | U | L | 1 | 2 | 1 | 1 | 1 | U | 1 | 6 | C |
| Mesocricetus auratus | Golden Hamster | mammal | 2 | 2 | 1 | U | L | 1 | 2 | 1 | 1 | 1 | U | 1 | 7 | C |
| Moschus moschiferus | Siberian Musk Deer | mammal | 1 | 3 | 2 | L | L | 1 | 2 | 1 | 1 | 2 | L | 2 | 8 | C |
| Muntiacus muntjak | Common Muntjac | mammal | 2 | 3 | 2 | L | L | 1 | 2 | 1 | 1 | 2 | L | 2 | 9 | B |
| Muntiacus reevesi | Reeve's Muntjac | mammal | 3 | 3 | 2 | L | L | 1 | 2 | 1 | 1 | 3 | 3 | 3 | 11 | A |
| Mustela vison | American Mink | mammal | 3 | 3 | 3 | L | L | 1 | 3 | 1 | 1 | 1 | L | 2 | 11 | A |
| Myocastor coypus | Coypu | mammal | 3 | 3 | 2 | U | L | 1 | 2 | 3 | 3 | 2 | L | 3 | 11 | A |
| Nyctereutes procyonoides | Raccoon Dog | mammal | 1 | 3 | 2 | L | L | 1 | 2 | 1 | 1 | 1 | L | 2 | 8 | C |
| Odocoileus virginianus | White Tailed Deer | mammal | 1 | 3 | 2 | L | L | 1 | 2 | 1 | 1 | 2 | L | 2 | 8 | C |
| Ondatra zibethicus | Musk Rat | mammal | 2 | 3 | 2 | L | L | 1 | 2 | 3 | 3 | 3 | L | 3 | 10 | B |
| Pecari tajacu | Collared Peccary | mammal | 1 | 2 | 3 | L | L | 1 | 3 | 1 | 3 | 2 | L | 3 | 9 | B |
| Procyon lotor | Raccoon | mammal | 2 | 3 | 3 | L | L | 1 | 3 | 1 | 1 | 1 | L | 2 | 10 | B |
| Rangifer tarandus | Reindeer | mammal | 1 | 2 | 2 | L | L | 1 | 2 | 1 | 1 | 1 | U | 1 | 6 | C |
| Rattus norvegicus | Brown Rat | mammal | 3 | 3 | 3 | L | L | 1 | 3 | 1 | 2 | 1 | L | 2 | 11 | A |
| Sciurus carolinensis | Grey Squirrel | mammal | 3 | 3 | 3 | L | L | 1 | 3 | 1 | 1 | 1 | L | 2 | 11 | A |
| Sigmodon hispidus | Cotton Rat | mammal | 2 | 2 | 2 | L | L | 1 | 2 | 1 | 1 | 1 | U | 1 | 7 | C |
| Sylvilagus floridanus | Easter Cottontail | mammal | 2 | 2 | 2 | L | L | 1 | 2 | 1 | 2 | 2 | L | 2 | 8 | C |
| Tamias sibiricus | Siberian Chipmunk | mammal | 3 | 3 | 2 | L | L | 1 | 2 | 1 | 1 | 1 | L | 2 | 10 | B |
| Tamias striatus | Eastern Chipmunk | mammal | 2 | 2 | 2 | L | L | 1 | 2 | 1 | 1 | 1 | L | 2 | 8 | C |

## Appendix 3. Possible indicators for birds

## Species to be included

About 350 species of birds have been noted as introductions to Great Britain, mostly as occasional escapes of individual birds from captivity. For the purpose of producing an indicator this list of species can be substantially reduced, using the following criteria.

- Birds that have not laid eggs in the wild in Britain are excluded.
- Birds that are native anywhere in the UK are excluded.
- Birds noted as non-native breeding birds prior to 1500 are excluded (Mute Swan, Common Pheasant).
- Reintroduced species (Capercaillie, White-tailed Eagle, Goshawk, Great Bustard) are excluded.
Species that have formed mixed pairings are listed if they have paired with a native species (Blue-winged Teal, Red-tailed Hawk).

There are 45 species that meet these criteria (Table A9). Canada Goose and Redlegged Partridge have been breeding since the late 1600s and are the earliest post1500 introductions. Several of these species have bred only sporadically, or were once established but now extinct. Northern Bobwhite is not listed by the BOU, but this is presumably an oversight, since there were established populations for a number of decades, although these are now extinct. (Budgerigar, listed by the BOU, is a comparable example).

## Approaches to building an indicator - parametric

Most indicators developed so far for birds draw on population indices that are set to the same base year. The geometric mean index value is then taken, across the species that form the indicator, year by year, to create an indicator from the average index values. A similar numerical approach would be possible for non-native birds, although only for ten species for which suitable numerical data are available (Fig A1). The main source for these data is the Breeding Bird Survey, which started in 1994.

Where BBS index values are available, no further processing is needed. For a few species, a better measure of population change can be obtained by calculating the total number of birds counted on BBS squares in a year and dividing by the total number of squares counted. The resulting data may contain zeros that are not compatible with taking a geometric mean.

| Species | Scientific name | BOU categories |
| :---: | :---: | :---: |
| Helmeted Guineafowl | Numida meleagris | E* |
| Northern Bobwhite | Colinus virginianus |  |
| Chukar Partridge | Alectoris chukar | E* |
| Red-legged Partridge | Alectoris rufa | C1E* |
| Red Junglefowl | Gallus gallus | E |
| Silver Pheasant | Lophura nycthemera | E* |
| Reeves's Pheasant | Syrmaticus reevesii | E* |
| Golden Pheasant | Chrysolophus pictus | C1E* |
| Lady Amherst's Pheasant | Chrysolophus amherstiae | C6E* |
| Indian Peafowl | Pavo cristatus | E* |
| Black Swan | Cygnus atratus | E* |
| Trumpeter Swan | Cygnus buccinator | E |
| Swan Goose | Anser cygnoides | E* |
| Pink-footed Goose | Anser brachyrhynchus | AE* |
| Bean Goose | Anser fabalis | AE |
| Greater White-fronted Goose | Anser albifrons | AE* |
| Lesser White-fronted Goose | Anser erythropus | $A E^{*}$ |
| Bar-headed Goose | Anser indicus | E* |
| Snow Goose | Chen caerulescens | AC2E* |
| Ross's Goose | Chen rossii | E* |
| Emperor Goose | Chen canagica | E* |
| Canada Goose | Branta canadensis | C2E* |
| Barnacle Goose | Branta leucopsis | AC2E* |
| Red-breasted Goose | Branta ruficollis | AE* |
| Upland Goose | Chloephaga picta | E* |
| Egyptian Goose | Alopochen aegyptiaca | C1E* |
| Ruddy Shelduck | Tadorna ferruginea | BDE* |
| Muscovy Duck | Cairina moschata | E* |
| Wood Duck | Aix sponsa | E* |
| Mandarin Duck | Aix galericulata | C1E* |
| Blue-winged Teal | Anas discors | AE |
| Cinnamon Teal | Anas cyanoptera | E* |
| Red-crested Pochard | Netta rufina | AC2E* |
| Ruddy Duck | Oxyura jamaicensis | C1E* |
| Black-crowned Night Heron | Nycticorax nycticorax | AE* |
| Red-tailed Hawk | Buteo jamaicensis | E |
| Budgerigar | Melopsittacus undulatus | E* |
| Rosy-faced Lovebird | Agapornis roseicollis | E |
| Alexandrine Parakeet | Psittacula eupatria | E* |
| Rose-ringed Parakeet | Psittacula krameri | C1E* |
| Blue-crowned Parakeet | Aratinga acuticaudata | E* |
| Monk Parakeet | Myiopsitta monachus | E |
| Eurasian Eagle-Owl | Bubo bubo | E* |
| Little Owl | Athene noctua | C1 |
| Red-winged Laughingthrush | Garrulax formosus | E* |

Table A9. Breeding non-native bird species in Great Britain; BOU categories (BOU, 2009) are as follows:- A naturally occurring since 1950; B naturally occurring but not since 1950; C established introductions; D doubtful; E non-native, not established.


Figure A1. Indexed geometric mean frequencies of ten common non-native breeding birds in Great Britain

A further 25 or so species can also be given a numerical treatment, based not on population index values but on actual estimates of breeding numbers assembled by the Rare Breeding Birds Panel. These data cover the years 1996 to 2005, with figures for later years not yet published. The species that can be included are those for which RBBP recorded nesting at least once during their period of recording, but for which no better trend estimates are available (e.g. through BBS). In general they are species that nested either sporadically or in numbers less than ten pairs. The approach taken is to sum all the pairs of the relevant species across each year, and then to scale these totals to 1 in an arbitrary datum year - this ensures that every species has equal weight in the indicator. Results are shown in Fig A2.


Figure A2. Indexed total of non-native breeding birds in Great Britain, drawn from annual estimates of population size (mostly from RBBP)

Species that did not breed during the timespan of the indicator and those are data deficient (Red Junglefowl) are necessarily excluded. The high initial value of this indicator is due to high counts of Snow Goose, Red-crested Pochard and Blackcrowned Night-heron in 1996.

Reliability of data varies considerably between species. Any species for which data were considered unreliable would be excluded from the indicator, but the species to be excluded on this criterion might change between years, making it difficult to maintain consistency over time.

## Approaches to building an indicator - non-parametric

A variety of indicators are possible that demand less of the data, for example the total of species known to have bred in each year, or each decade. This would use data from the full list of 46 species (together with any additions in future years), but would not reflect population change (Fig. A3). Such an indicator could span many decades. It would show relatively low annual values before 1996, when RBBP began reporting records of non-native breeding species, but decadal values would be less affected.


Figure A3. Index of species richness of rare breeding non-native birds in Great Britain, 1996-2005

