

# Biogeochemical accumulation of trace elements in soils and plants of the Russian Far East

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**Abstract.** The accumulation of total mercury and other elements in soils and plants growing on them (legume and valerian families) selected in the Khabarovsk and Primorsky Territories, Buryatia, and the Amur Region was studied. The possibility of determining mercury in natural objects along with the natural atomic absorption method is also shown by X-ray fluorescence spectroscopy simultaneously with other trace elements. Analysis of the data obtained indicates that the main sources of trace elements at the test sites are, apparently, sources of natural origin – soils, underlying rocks, underground and surface waters, etc.

## 1 Introduction

One of the most important problems of geochemical ecology is the study of the chemical elemental composition of plants in conditions of various biogeochemical provinces [1,2], formed under the influence of both natural (soil-forming rocks and processes, ore anomalies, etc.) and anthropogenic (high-temperature emissions of various industries, mining, fertilization and pesticides, etc.) factors. The actual problems of ecological expertise of technogenic flows of heavy metals scattering are both the locations and functioning of industrial production, and the definition of background territories remote from large industrial centers [2, 3]. The content of elements in the profile of virgin soils, and, accordingly, in plants growing on these soils, is inherited mainly from the parent breed. Plants are the most important intermediate link through which chemical elements pass from soil, water and air into the body of animals and humans. Therefore, in conditions of constant anthropogenic pollution of the environment, it is possible to include a number of chemical elements in the human food chain, which poses a potential health hazard [4, 5].

Plants, reflecting the specific features of the accumulation and content of trace elements, nevertheless bear a local "coloring" in the form of a predominance in their composition of elements that are rich in soils in the area of growth, both vital and toxic at certain concentrations. In particular, the accumulation of heavy metals in agricultural plants causes human diseases at the following contents (one part of metal per million parts of dry plant mass): Cd – 0.1–6.4; Pb – 1.8–50; Ni – 1.5–36; Hg – 0.02–0.03 [6].

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Mercury is one of the most toxic chemical elements. Mercury occurs in nature both in its native state and in compounds (cinnabar, timanite, etc.), as well as as an isomorphic or mechanical impurity, it is part of molybdenum, pyrite, pyrrhotite, sphalerite, coal, and during the processing of these minerals it enters the environment in significant quantities. Technogenic mercury was also dumped into the natural environment in gold mining areas, until the technology of extracting loose gold by amalgamation was banned in 2018, and continues to enter all ecosystems of the biosphere as part of pesticides.

From inorganic and organic compounds of mercury in the biosphere, its most toxic methyl, ethyl and other organic derivatives are formed as a result of the process of biological methylation [6, 7]. Entering the biosphere, mercury and its compounds, capable of being in different phase states, easily migrate in all its components, accumulating, like other trace elements, in plant, living organisms, and displacing vital elements from biomacromolecules.

As a result of metamorphic and biogenic redistribution, extensive mercury scattering halos are created both in areas of mining, metallurgical and other enterprises, and in places remote from intensive production activities, for example, at various types of sulfide deposits. Moreover, the forms of mercury in halos are different: in the form of independent mercury minerals, isomorphic impurities, in the sorbed state [8]. In soil conditions, mercury is in elementary form or in the form of cationic or anionic complexes, is retained by the soil mainly in the form of complex weakly mobile organic and inorganic complexes, methylation and biological destruction of which, with subsequent dissolution, leads to the accumulation of mobile forms easily accessible and easily digestible by plants in the upper soil layer [6-10]. It is assumed that all mercury in the soil eventually gets into terrestrial plants [9].

Biogeochemical studies of plants and soil cover, both in areas with technogenic pollution, and where it is absent or minimal, are necessary for several reasons: for a one-time assessment of the state of the terrestrial ecosystem, for background monitoring to identify the content of trace elements in plants of natural ecosystems and establish a connection of the elemental composition with the geochemical environment, for comparison with the similar data for polluted man-made systems and, finally, as a result – for biogeochemical examination of a constantly changing environment and solving other geoecological problems, for example, with metallometric methods for detecting ore deposits [1,3]. In particular, when searching for sulfide deposits, the study of the distribution of mercury in scattering halos is preferable due to the significant size of these halos. It should be noted that when determining the mercury content for solving the tasks outlined above, it is important to choose a technique with one or another sensitivity, which should be higher than  $(3-5) \cdot 10^{-5}\%$ , otherwise mercury scattering halos will not be detected [5,6,9,11,12]. Obviously, such an approach should also take place to the choice of a method for determining other trace elements in environmental objects.

This paper presents data from a selective geochemical study of the accumulation of total mercury and other elements in soils and plants selected in several territories of the Far East – in the Khabarovsk and Primorsky Territories, in Buryatia and the Amur Region, taking into account the corresponding clarks and existing background concentrations. One of the objectives of the study was to assess the possibility of determining total mercury from one sample with other trace elements in natural objects by X-ray fluorescence spectroscopy. Since the chemical composition of plants is formed under the influence of a large number of both internal physiological and external environmental factors, and the studied territories differ from each other in ecological terms, in order to identify any patterns, it was necessary to minimize the influence of these factors as much as possible. Firstly, landfills remote from large industrial centers and experiencing insignificant anthropogenic load were selected for research in these territories, and secondly, samples of soils and plants growing on them from only two families were selected for testing: legumes (Amur maakia, Daurianlespedets, kopechnovaya, plotnokistevaya; small-leaved and Ussurikaragany) and valerian

(scabiosolic and rocky patrinias). Samples of soils and these plants were taken in the following areas: Buryatia, 10 km southeast of the village of Mukhorshibir, right-bank part of the Khilok River valley, slope 50° (sampling point 1); Amur Region, 10 km from the village of Mukhinka, right-bank part of the valley of the Zeya River (2); Khabarovsk Territory, 12 km northwest of the village of Polyarny (3) and the right bank of the Amur River, slope 60° (4); Primorsky Krai, 5 km northwest of the village of Ariadne, the basin of the Pad Todokhov river (5), 10 km northeast of the village. Wrangel, the coastal location station of Tazgou Bay (6), southeastern slope 700 (testing point 7).

The soils of the sites selected for study are medium- and slightly podzolic. At each point where the studied species form thickets, 10 accounting sites of 1 m<sup>2</sup> were laid, from which 10 model specimens of the same species were selected. Samples were taken in the first or second decade of August. A simultaneous biogeochemical sampling of averaged dried samples of plant parts of the same species collected at the test sites and soil samples corresponding to the soil profile from the root system location zone at the depth of the humus horizon was carried out.

## 2 Methods

Determination of the content of mercury and other trace elements for research purposes in environmental objects, including soils and plants, is carried out by various methods: nephelometric, emission spectral, flame-free atomic absorption [13-15]. In the laboratories of Rospotrebnadzor of Russia and the Federal Service of Russia for Hydrometeorology and Environmental Monitoring, the certified atomic absorption method of "cold steam" is used to determine mercury in natural water, soil, plants and living organisms, as well as in their processed products. In general, when carrying out mass measurements, it is recommended that when preparing samples of soils and other environmental objects, they should be wet-salted with mixtures of strong acids in the presence of oxidants [2, 16]. When determining mercury, it is necessary to avoid its losses at all stages of sample preparation for analysis due to the volatility of its compounds, especially chloride and methyl chloride.

Elemental analysis of the studied samples of soils and plants prepared in accordance with the recommendations [16,17], as well as flour 4.1.005-4.1.008-94 Rospotrebnadzor, performed by X-ray fluorescence method with total external reflection (TXRF) on a TXRF 8030 C device. The measurement time is 500 s, the excitation sources are  $Mok_{\alpha}$  and  $WL_{\beta}$ . The internal standard is Y with a concentration of 5 micrograms/ml. The detection limit varies for different elements from 10<sup>-7</sup> to 10<sup>-10</sup> %. A sample with a volume of 10 μl was applied to an organic glass substrate. The possibility of simultaneous determination of mercury from a single sample with other heavy metals was tested on standard samples prepared similarly to the studied samples by wet decomposition of precise oxide attachments. In addition, in parallel, the determination of mercury in the samples was carried out in accordance with the MUK atomic absorption method on the mercury analyzer "Julia-2". A good convergence of the results of mercury determination by these methods is shown. The data on the content of mercury and other trace elements are given in Tables 1, 2. For comparison, clarks were used in sedimentary rocks and soils according to A.P. Vinogradov and in vegetation by V.V. Dobrovolsky [3, 6, 7, 18]. It is obvious that global average concentrations of trace elements in soil and vegetation cannot be accurately determined [7, 9], since they are determined by many factors, the most important of which are fractional heterogeneity and the prevalence of different types of soils.

### 3 Results and Discussion

It is known that the accumulation of mercury in soils is generally associated with the level of organic carbon and sulfur [9], and its concentration in the upper soil layer is several times higher than in the subsurface horizons, although due to the limited participation of this metal in the soil migration processes of the substance, its concentration in the surface layer grows quite slowly even in the case of a small anthropogenic contribution. Practically all tested soils in connection with the study of the elemental composition of medicinal plants were selected by us in relatively unpolluted territories. Mercury was not detected in the samples of the soil layer of the studied polygons of the Primorsky Territory and the Amur Region (Table 1), which may be related both to concentrations below the detection limit by the methods used, i.e. less than  $10^{-7}$  %, and to the formation of easily mobile forms of mercury compounds under the influence of specific soil conditions and its subsequent losses from the soil in the form of vapors, as well as easy biological accessibility for plants [1, 9].

**Table 1.** Content, % / clarke concentration of trace elements in the soil.

Element	Testing point					Clark [3]
	1	2	3	4	5	
Ti	0.0504/0.1	0.0217/0.05	0.0693/0.15	0.0863/0.2	0.0118/0.05	0.45
Cr	0.0019/0.1	0.0019/0.1	0.0011/0.06	0.0027/0.14	0.0002/0.01	0.02
Mn	0.0413/0.5	0.0333/0.4	0.0254/0.3	0.0226/0.3	0.0185/0.2	0.085
Ni	0.0029/0.7	0.0017/0.4	0.0012/0.3	0.0010/0.25	0.003/0.08	0.004
Cu	0.0019/1.0	0.0003/0.15	0.0035/1.8	0.0035/1.9	0.0183/9.2	0.002
Zn	0.0045/0.9	0.0029/0.6	0.0025/0.5	0.0028/0.6	0.0022/0.4	0.005
As	0.0008/1.2	0.0006/0.9	0.0004/0.6	0.0006/0.9	-/-	0.0007
Sr	0.0016/0.04	0.0027/0.06	0.0032/0.07	0.0020/0.04	0.0002/0.3	0.045
Pt	0.0002/40.0	-/-	0.0002/40.0	-/-	-/-	$5 \cdot 10^{-6}$
Au	-/-	-/-	0.0003/60.0	-/-	-/-	$5 \cdot 10^{-6}$
Hg	0.0001/25.0	-/-	0.0004/100.0	0.0001/25.0	-/-	$4 \cdot 10^{-6}$
Pb	0.0007/0.7	0.0006/0.6	0.0003/0.3	-/-	0.0023/2.3	0.002
Bi	-/-	0.0003/3.0	-/-	0.0003/3.0	-/-	0.0001*
Th	-/-	-/-	-/-	-/-	0.0012/1.1	0.0011*
Ga	-/-	-/-	-/-	-/-	0.0001/0.03	0.0004*

Note. Dash - not found.

\* - clarke in rocks according to A.P. Vinogradov [3].

Indeed, despite the fact that mercury has not been detected in the soils of a number of territories, mercury accumulates in amounts from  $10^{-5}$  to  $10^{-4}$  % in the tested plants of these territories (Table 2), which is higher than Clark in plants according to S.M. Tkalich [1].

In the soils of the remaining tested territories, the mercury content significantly exceeds normal. The concentration of mercury in these samples of the upper soil layer exceeds Clark in sedimentary rocks and soils according to A.P. Vinogradov, estimated respectively 0.0 n–0, n mg/kg [9],  $4 \cdot 10^{-5}$  % [3] and  $n \cdot 10^{-6}$  % [6], by an order of magnitude (Table 1). Estimate the level of chemical contamination of soils can be determined by the concentration coefficient, which is calculated as the ratio of the actual content of an element in the soil to its background content [11]. Numerically estimating background mercury levels in soils is quite difficult due to widespread and increasing anthropogenic pollution. According to [9], the background mercury content in the surface layer of soils ranges from 0.01 to 0.4 mg/kg of dry weight, not exceeding, as a rule, the latter value. Calculated even relative to the maximum background content equal to 0.4 mg/kg, the values of the mercury concentration coefficient for the tested soils are: the village of Polyarny, Khabarovsk Territory – 10.1; the village of Wrangel, Primorsky Territory, coastal radar station – 7.9; Khabarovsk, the Amur

River bank – 3.72; Buryatia – 3.48, which indicates a significant excess over the background and, taking into account the content of other metals (Table 1), allows these soils to be classified as moderately polluted [11]. Moreover, taking into account the fact that sampling sites are not subject to significant anthropogenic load, the increased content of mercury, as well as other trace elements, in soils, as well as in tested plants, can be explained more likely by the influence of natural rather than man-made factors and, apparently, is associated with the content of mercury, yes and most other trace elements in the underlying rocks [6].

Assessing the levels of accumulation of scattered elements, the concentrations ( $K_K$ ) of trace elements were calculated for all tested soils and plants (Tab. 1), as well as the coefficients of biological absorption ( $K_b$ ) of trace elements by plants (Table 2), defined respectively as the ratio of the actual content of trace elements in the soil or plant to the corresponding clark and as the ratio of the actual concentrations of trace elements in the plant to the actual content of this trace element in the corresponding soil sample [1, 6, 7, 19]. For comparison in the table 2 the global values of the  $K_b$  vegetation of the continents given in [7] are given.

**Table 2.** The content of microelements, % /coefficient of biological absorption ( $K_b$ ) of tested plants.

Element	Sample point				Global value of $K_b$ vegetation of continents [b].
	Buryatia	Amur region	Khabarovsk Krai	Primorsky Krai	
Ti	0.0059/0.1	0.0083/0.4	0.0148/0.2	0.0017/0.4	0.2
Cr	0.0012/0.6	0.00094/0.5	0.00197/0.7	0.00025/0.2	1.0
Mn	0.0085/0.2	0.0270/0.8	0.0083/0.4	0.0084/0.2	6.7
Ni	0.0035/1.2	0.0016/0.9	0.0030/3.0	0.00095/0.9	1.5
Cu	0.0023/1.2	0.0025/8.3	0.0011/0.3	0.0016/0.3	9.1
Zn	0.0028/0.6	0.0026/0.9	0.0030/1.1	0.0078/3.1	19.6
As	0.0024/0.3	0.00016/0.3	0.00015/0.3	0.000125/0.03	1.6
Sr	0.00129/8.1	0.0076/2.8	0.0059/3.0	0.0065/4.3	3.5
Pt	0.00025/1.1	0.0076/2.8	-/-	-/-	-
Au	0.0005/-	-/-	-/-	-/-	-
Hg	0.00016/1.6	0.000072/7.2	0.0001/1.0	0.00012/1.2	7.6
Pb	0.00038/0.5	0.00026/0.4	0.0004/-	0.0005/0.4	3.7
Bi	0.00064/-	0.00022/0.7	-/-	-/-	-
Th	-/-	-/-	-/-	-/-	-
Ga	-/-	-/-	-/-	-/-	0.05

For mercury, the accumulation level in all tested soils, where it is found, is very high: the  $K_K$  ranges from 25 to 100. Similar high mercury concentrations have already been observed for some territories of the Far East [1]. The comparison also indicates a significant increase in the platinum content ( $K_K = 40$ ) in the tested soils of Buryatia and Khabarovsk Krai (Polyarny settlement), and in the latter case there is an increase in the gold content ( $K_K = 60$ ). Apparently, the correlation of mercury, gold and platinum  $K_K$  observed for these soils should be noted.

In the studied soils of several territories, there is an increased content of bismuth, copper, lead (Table 1). The highest concentrations of copper ( $K_K = 9.2$ ) and lead ( $K_K = 2.3$ ) were found in the soil layer of the village Ariadnoye (Primorsky Krai), where gallium and thorium are also present ( $K_K = 1.1$ ), not found in any other studied soil sample. However, in all other soils, except claim 5, the presence of highly toxic arsenic for biota was found in amounts from 0.0004 to 0.0008 %, which is comparable or slightly higher than the average content in sedimentary rocks [3]. The content of other trace elements in the studied soils does not exceed Clark according to A.P. Vinogradov.

The content of some trace elements, including mercury, in the studied plants varies widely depending on the collection area (Table 2). At the same time, the concentration level of some trace elements in the studied plants does not always directly correlate with their content in the corresponding soils, but the general tendency described in [7-9,18] is greater or less accumulation by plants individual trace elements, depending on their role in metabolism and at the same time availability in the form of mobile forms, are preserved. Thus, the analysis of mercury content data in soils and plants indicates the multifactorial nature of the accumulation of this phytotoxic trace element, which was noted earlier [9]. Plants of those territories where there is an increased content of mercury in the soil ( $KK = 25-75$ ) can either accumulate mercury in significant quantities ( $K_b = 2-4$ ), or do not extract it from the soil. Moreover, the same plant species selected in different territories exhibit different abilities to deposit mercury. At the same time, mercury in plants taken from the territory in the soil layer of which it is not found accumulates in the same concentrations –  $0-10^{-4}\%$ , as from territories with a high content of trace elements. Such patterns of biological absorption of mercury are quite understandable, given the diversity of both its forms of existence in nature and ways of entry into plant communities. [7, 9, 20, 21].

Analysis of both absolute and average values of the content of trace elements in the studied plants (Table 2) showed that to the greatest extent all tested plants, regardless of the place of growth, tend to accumulate strontium: with an average  $KK$  of 0.04,  $K_b$  is in the range of 2.8–8.0 and corresponds to or exceeds the global values of  $K_b$  of vegetation calculated in [7] by 2 times. Mercury, as noted above, is differentially absorbed by plants, and in some cases quite significantly, however, the calculated value of  $K_b$  according to the data obtained is 3 times less than the rather high global value of  $K_b = 7.58$  for all polygons except the Amur region, which is consistent with the data of the study of mercury content in this region [10]. Toxic arsenic with a relatively high content in soils ( $KK = 0.9-1.2$ ) is not absorbed, as shown by the calculations of  $K_b$ , tested plants in significant quantities. In large quantities, plants accumulate copper and zinc, which is explained by their important biochemical function. The contents of copper and zinc in plants are either comparable to their content in the corresponding soil layer, or exceed it several times.

The corresponding values of the  $K_b$  of these elements illustrate the well-known antagonistic nature of their interaction when consumed by vegetation: with an increase in the  $K_b$  of copper, the  $K_b$  of zinc decreases by an order of magnitude in all cases, and vice versa [8,21]. It should be noted that at the values of the  $KK$  of nickel in soils 0.2–0.7, the biological absorption coefficient is 0.9–3.0, which may indicate a directed extraction of nickel by vegetation, as in the previously observed case with tin [1]. For all other trace elements in the tested plants, the calculated values of  $K_b$  are less or comparable to those given in [20, 22] for vegetation in some regions of the Far East and significantly less than the known global values of this coefficient [8], which indicates a low intensity of their biological absorption by plants under these conditions. Lead is found in plants in very small amounts – an order of magnitude or two less than Clark in vegetation [1, 7] – and even with an increased content in soils, it does not accumulate in plants (Tables 1, 2). For titanium and chromium, in some cases, the same correspondence is observed in the Clark values of the concentration in the soil and the biological absorption coefficient as for nickel:  $K_b > KK$ . At the same time, manganese in all tested plants, regardless of the place of collection, is characterized by lower values of  $K_b$  in comparison with clarks of concentration in the soil layer. It should be noted that the calculation of concentration clarks for these trace elements did not reveal a significant accumulation of them in the soils of the studied territories (Table 1).

## 4 Conclusion

Thus, we have studied the accumulation of mercury and some trace elements in the soils of several polygons remote from large industrial centers, in the Khabarovsk and Primorsky Territories, Buryatia and the Amur Region, as well as in plants growing on them. Using the example of elemental analysis of natural objects – soils and plants – the possibility of determining total mercury along with the well-known atomic absorption method is also shown by the method of X-ray fluorescence spectroscopy simultaneously from one sample with other trace elements. All tested plants, regardless of the place of growth, accumulate strontium to the greatest extent, as well as biochemically important copper and zinc. On the contrary, phytotoxic mercury, arsenic and lead are not absorbed by the tested plants in significant quantities [15, 23]. The analysis of the data on the biogeochemical accumulation of total mercury and some trace elements showed, in accordance with the quantitative characteristics of biological absorption, not exceeding the known average values for vegetation of continents (Clark,  $K_b$ ), that the studied polygons really avoided anthropogenic pollution, and the main sources of trace elements there of natural origin – soils, underlying rocks, underground and surface waters, etc. [1, 6, 7, 9, 18, 24].

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