Contents lists available at ScienceDirect



Ecological Indicators



journal homepage: www.elsevier.com/locate/ecolind

Ground-dwelling arthropods as biodiversity indicators in maize agroecosystems of Northern Italy

Francesco Lami^{a,*}, Giovanni Burgio^a, Serena Magagnoli^a, Daniele Sommaggio^a, Roland Horváth^{b,c}, Dávid D. Nagy^d, Antonio Masetti^a

^a DISTAL-Department of Agricultural and Food Sciences, University of Bologna, 40127 Bologna, Italy

^b Department of Ecology, University of Debrecen, H-4032 Debrecen, Hungary

^c ELKH-DE Anthropocene Ecology Research Group, University of Debrecen, Debrecen, Hungary

^d MTA-DE Biodiversity and Ecosystem Services Research Group, Debrecen, Hungary

ARTICLE INFO

Keywords: Carabidae Staphylinidae Araneae Surrogate taxa Species turnover Species co-occurrence

ABSTRACT

Reliable monitoring of arthropod diversity in a given agroecosystem is essential for the conservation of the related ecosystem services, such as biological control. The often daunting complexity of arthropod collection and identification, however, highlights the need for surrogate taxa that can be easily sampled and be representative of a number of other taxa in term of diversity, general community features and specific composition.

In this study, we used pitfall traps to sample three ground-dwelling arthropod taxa important as biocontrol agents (ground beetles, rove beetles and spiders) in 9 conventionally managed maize agroecosystems of Northern Italy over the course of two years, with the goal of characterizing their assemblages and evaluating their reciprocal potential as indicators of activity density, species richness, community turnover and species co-occurrence.

Although dominated by few generalist species, sampled arthropod communities were relatively species-rich, and included the first Italian record of the spider *Zelotes metellus* (Roewer) (Araneae: Gnaphosidae). Ground beetles as a group were confirmed as promising indicators for the species richness and community composition turnover of rove beetles and spiders. Additionally, several abundant arthropod species acted as indicators of the species richness of their respective groups, and the ground beetle *Pterostichus macer* (Marsham) also worked as an indicator of overall rove beetle activity density. While the co-occurrence of individual arthropod species was limited for the studied taxa, a few species such as the ground beetle *Parophonus maculicornis* (Duftschmid) did show promise as species-specific bioindicators. Our results could be useful in improving the monitoring and management of these important natural enemies in maize-growing regions.

1. Introduction

Accurately monitoring and mapping biodiversity is essential for planning conservation actions (Niemelä, 2000). This becomes especially important in agroecosystems, as biodiversity is linked with several ecosystem services crucial for agriculture itself (Altieri, 1999; Mace et al., 2012; Swinton et al., 2007). Therefore, agricultural practices need to be carefully planned in order to minimize the negative impacts on beneficial organisms (Garbach et al., 2014; Moonen and Bàrberi, 2008).

As one of the most abundant and diversified group of terrestrial animals (Santos et al., 2021), arthropods are in fact responsible for the provision of a wide array of important ecosystem services, ranging from pollination to nutrient cycling acceleration (Dangles and Casas, 2019; Kremen et al., 1993; Losey and Vaughan, 2006). The biological control of harmful organisms through the conservation and enhancement of predatory and parasitoid arthropod communities (conservation biological control) has gained impetus in the last few decades as a promising, environmentally friendly way of protecting crops (Begg et al., 2017; Fiedler et al., 2008). Ground-dwelling arthropods such as ground beetles (Coleoptera: Carabidae) are included among the most important natural enemies of pests and weeds (Kromp, 1999; Lami et al., 2020; Lövei and Sunderland, 1996). Other groups such as rove beetles (Coleoptera Staphylinidae) and ground-dwelling spiders (Araneae), comparatively less studied, have been attracting much interest in recent years owing to

https://doi.org/10.1016/j.ecolind.2023.110352

Received 29 November 2022; Received in revised form 3 May 2023; Accepted 10 May 2023 Available online 19 May 2023

1470-160X/© 2023 Published by Elsevier Ltd. This is an open access article under the CC BY-NC-ND license (http://creativecommons.org/licenses/by-nc-nd/4.0/).

^{*} Corresponding author. E-mail address: francesco.lami2@unibo.it (F. Lami).

an increasing body of evidence highlighting their biological control potential (Albertini et al., 2018; Betz et al., 2018; Michalko et al., 2019b, 2019a). Ground beetles and ground-dwelling arthropods in general are also studied for their potential as bioindicators of environmental impacts caused by agricultural practices and other human activities (Hågvar and Klanderud, 2009; Magagnoli et al., 2021; Pearce and Venier, 2006; Rainio and Niemelä, 2003).

The conservation of biodiversity, including the diversity of arthropods, is greatly improved when based on up-to-date biodiversity inventories that can be used as a reference to detect changes and compare biotic communities (Balmford and Gaston, 1999; Stephenson and Stengel, 2020), underlining the importance of faunistic surveys (Ejsmont-Karabin, 2019). The unparalleled diversity of arthropods coupled with their complex and ever-shifting taxonomy informed by advances in systematics, however, represents an enormous challenge when undertaking such efforts (Lovell et al., 2007; Samways, 2015). These difficulties are exacerbated by the currently dwindling numbers of expert taxonomists able to identify arthropod taxa to the species level (Ebach et al., 2011; Hopkins and Freckleton, 2002). As a consequence, an increasing number of studies highlights the need for the identification of surrogate taxa that can be used as indicators of the overall biodiversity in a given area or, more realistically, of the biodiversity of a number of other important taxa, or as indicators of ecosystem services (Birkhofer et al., 2018; Harry et al., 2019; Lovell et al., 2007). This line of research has led, until now, to mixed successes (Heino, 2010), with sometimes contradictory results regarding the potential value of certain taxa as surrogates (Lewandowski et al., 2010). Vascular plants, for instance, are frequently proposed as biodiversity indicators of various arthropod groups, but while significant correlations are often reported with phytophagous and plant-associated taxa (Bucher et al., 2019; Larrieu et al., 2019), relations with groups such as ground-dwelling taxa are less consistent (Harry et al., 2019; Schoeman et al., 2020; Uboni et al., 2019). Recent studies indicated that the habitat context can greatly affect the potential of a given taxon as a biodiversity indicator (Corcos et al., 2021; Yong et al., 2020). This suggests that each habitat type might require different indicators, implying that research should prioritize the most abundant and ecologically important natural and agricultural habitats in a given area.

Ideally, a good biodiversity indicator group should be relatively easy to collect and identify, and of course its diversity should co-vary with that of other taxa (Mandelik et al., 2010; Oberprieler et al., 2020; Westgate et al., 2017). The usefulness of a certain diversity index in bioindication, from simple abundance or species richness to more complex metrics based on community composition, is never universal, and different indices might be suitable for different taxa or habitat contexts (Corcos et al., 2021; Gao et al., 2015). The choice of a certain index might also depend on the conservation goals (Duelli and Obrist, 2003; McQuatters-Gollop et al., 2019). If the aim is to preserve a high level of biodiversity in general, then indices that provide general information on bioindicator and target communities, such as species richness or evenness (Magurran, 1988), are adequate. However, different species in a same taxon might play different ecological role (Cane and Payne, 1993; Harvey et al., 2008); thus, if the aim is to gather data useful for preserving specific ecological functions, indices based on functional diversity and ecological traits (Teresa and Casatti, 2017) or on community composition, turnover and species identity (Carvalho et al., 2013; Veech, 2013) might be more suitable. Different species or taxa are more likely to co-occur (and thus act as biodiversity indicators) if they depend on each other, for instance through mutualism, parasitism or predation (Aubier and Elias, 2020; Bell et al., 2010), or if they share the same environmental requirements without strongly competing for important resources (Araújo et al., 2011; Sfenthourakis et al., 2006). In case of strong competition or very different environmental requirements, on the other hand, two species or taxa will tend not to occur simultaneously (Gotelli and McCabe, 2002).

on biodiversity by characterizing arthropod assemblages and ii) evaluate the potential of three important groups of ground-dwelling predatory arthropods (ground beetles, rove beetles and spiders) as biodiversity indicators of each other in maize agroecosystems of Northern Italy. Conventionally-managed maize (Zea mays L.) was chosen as it is one of the most widespread and ecologically important crops in the study area (6th Census of Agricultural Holdings - ISTAT, Istituto Nazionale di Statistica). Given the impacts on biodiversity associated with maize cropping (Chmelíková and Wolfrum, 2019; Norris et al., 2016) and the projected widespread positive changes that would derive from large-scale application of sound agroecological principles to it (Triquet et al., 2022; von Redwitz et al., 2019), it is particularly important to draw a clear picture of beneficial arthropod diversity and protect it in such agroecosystems. Based on methodologies from recent relevant literature (Corcos et al., 2021; Griffith et al., 2016; Zara et al., 2021), the potential of biodiversity indicators of the studied groups was assessed in terms of i) general community features (activity density and species richness) ii) community turnover, iii) specific species relations to community features and iv) species co-occurrence.

2. Materials and methods

2.1. Sampling sites

The sampling was carried out in the spring and summer months of 2013 and 2014 in the Bologna province in Northern Italy. The study area is dominated by arable crops, with a mean annual temperature of $14 \,^{\circ}$ C and mean annual precipitation of approximately 800 mm. We sampled a total of 9 sites (maize fields) in an area of roughly 70 km². In order to adapt to the local crop rotations, 4 of the sites were sampled only in 2013, 4 only in 2014 and one in both years (Table A.1). For the sake of bioindicator potential analyses, the two sampling years at the same site were considered as two separate sites (10 total sites). Distance between fields sampled in the same year ranged between 0.4 km and 15.5 km.

As the landscape context (Lami et al., 2021) and the nature of field margins (Maas et al., 2021; Marshall and Moonen, 2002) can deeply influence arthropod communities, we characterized landscape features in a 1000 m radius around each field using ArcGIS Desktop v10.8.2 and a 2011 CORINE Land Cover map (Büttner et al., 2002) of the Bologna province publicly available at the website of Regione Emilia-Romagna (https://geoportale.regione.emilia-romagna.it/). In general, landscapes around fields were dominated by farmland, with little, if any, natural habitats (Table A.2). Field margins were represented by spontaneous grassy strips in all cases.

Agronomic practices (and especially soil management for grounddwelling arthropods) are other important factors that can impact arthropod communities (Gallé et al., 2020; Gayer et al., 2019; Lami et al., 2020; Rusch et al., 2016). In our case, all fields were managed conventionally and were similar in terms of tillage (conventional tillage), cover crop usage (no cover crops) and crop that preceded maize (wheat). Additional details about soil insecticides, herbicides and fertilization are reported in Table A.2.

2.2. Arthropod sampling and identification

Ground-dwelling arthropods were collected using pitfall traps (Brown and Matthews, 2016). Each trap consisted of 2 plastic cups (600 ml, 10 cm in diameter) flushed with soil surface, placed at 1 m from each other and connected with a 10 cm high plastic barrier used to intercept arthropods and direct them towards the collection cups. A plastic cover was placed above the cups to protect them from rain. During activation, traps contained about 200 ml of 40% propylene glycol per cup.

The number of traps in each sampled field was proportional to the size of the field, ranging from 10 to 20 traps and maintaining a density of roughly 1 trap per 0.05 ha, as field size ranged from 0.5 to 1 ha. Traps were placed 10 m from each other, either in single or multiple rows

depending on the size and shape of the field. In two cases, two neighboring maize fields were sampled together and considered as a single site given their small size and that they were managed in the exact same way (Table A.1–A.2).

The sampling timing was partially dictated by the protocol of the funding project (see Acknowledgements), and it took place monthly from tasseling to harvest, which resulted in 4 sampling rounds (from June to September) in 2013 and 5 sampling rounds (from May to September) in 2014, with each sampling round lasting 7–9 consecutive days. The total number of sampling days per site is reported in Table A.1. The temporal window of our sampling encompassed the expected peak activity periods of both spring/early summer and late summer/early autumn of many ground-dwelling arthropod groups in Italian agroecosystems (Lami et al., 2021; Lövei and Sunderland, 1996; Nardi et al., 2019; Rossi et al., 2019).

The collected ground beetles, rove beetles and spiders were counted and identified to species level using morphological characters. A reason to focus on these three groups was their previously mentioned and widely reported potential as biological control agents and bioindicators. Moreover, they were among the most abundant groups collected during the study, and experts were available for their identification. Specifically, authors carried out the identification process using relevant literature (Assing and Schülke, 2012; Freude et al., 1974; Netwig et al., 2023; Pesarini and Monzini, 2011, 2010; Trautner and Geigennueller, 1987). In the case of rove beetles, 8 specimens (2.1% of the total) could not be identified to species level and were assigned to 7 morphospecies.

2.3. Data analysis

2.3.1. General community features, diversity and species turnover

We calculated the overall activity density and species richness of ground beetles, rove beetles and spiders in each site, with activity density being the average number of captures of a single trap station per day. Pearson correlations between groups were computed for both metrics. Analyses were carried out in R v3.6.2 (R Core Team, 2016).

In order to evaluate cross-taxon congruence in community composition spatial turnover (Corcos et al., 2021), we first calculated Jaccard's dissimilarity index (Carvalho et al., 2013) based on species presence/ absence for each taxon. We then used a Mantel test (Mantel, 1967) with 9999 randomized permutations to evaluate pairwise correlations between the dissimilarity matrices of the different arthropod groups.

We pooled all the traps for each maize field rather than analyzing the data divided by trap, as our aim was to test the correlation between community features and the cooccurrence of species within habitats, and because the distance between traps within a field was vastly inferior to the average dispersal abilities of ground-dwelling macroarthropods (Bertrand et al., 2016; Schmidt et al., 2008). The individual traps within a field were thus highly unlikely to represent independent communities or micro-habitat patches.

2.3.2. Species-level bioindication and co-occurrence

The potential of individual species as bioindicators of entire community features was tested using the activity density of the species as an indication parameter. We focused on the most abundant species as they were the most likely to occur in a high number of sites with varying levels of activity density, also meaning that they would likely be easier to sample and provide useful data in future real world applications. Specifically, we selected species with at least 100 sampled individuals. As no rove beetle species occurred with more than 100 individuals, to include representatives of this family we decided to select rove beetle species that were sampled with at least 50 individuals and found in more than half of the sites. We then calculated Pearson correlations between the activity density of each species and the activity density and species richness of each group.

Second, we wanted to test specific species co-occurrence. As the analysis is based on presence-absence data rather than abundance, we

included a higher number of species, but we nonetheless removed very rare species (<10 total individuals) from the dataset, as estimation of cooccurrence and other ecological characters for rare species is problematic (Barlow et al., 2010; Plowman et al., 2020). We used Veech's pairwise method (Veech, 2013) to identify species pairs which showed significant positive co-occurrence (i.e. the two species tended to occur simultaneously in the studied sites) or negative co-occurrence (i.e. the two species tended to avoid each other). Co-occurrence analysis was performed using the "cooccur" v1.3 package (Griffith et al., 2016) in R v3.6.2. As the number of negative co-occurrences was very low (13 species pairs, 0.9% of the total), we focused on the more abundant positive co-occurrences (65 species pairs, 4.5% of the total). Our response variable was thus binary (1 in the case of positive significant co-occurrence or 0 in the other cases). We therefore fitted two logit regression models: one testing the proportion of positive species cooccurrences over the entire dataset (using taxon as an explanatory variable), and the other testing the proportion of positive co-occurrences in each possible taxon pairing (Ground beetle - Ground beetle, Rove beetle - Rove beetle, Spider - Spider, Ground beetle - Rove beetle, Ground beetle - Spider, Rove beetle - Spider). We then conducted a posthoc analysis by calculating pairwise comparisons with a Tukey adjustment with the "emmeans" v1.4.4 package (Lenth, 2018).

3. Results

3.1. General community features and faunistic notes

We collected a total of 3447 ground beetles (Table A.3), 386 rove beetles (Table A.4) and 2210 spiders (Table A.5) belonging to 45, 43 and 91 species respectively. Spiders were representative of 21 families. Most species were infrequent, with only 17 ground beetle species, 10 rove beetle species and 28 spider species being represented by at least 10 individuals.

The dominant ground beetle species was *Harpalus rufipes* (De Geer) with 1951 individuals (56.6% of the collected ground beetles). The most abundant rove beetle was *Dinaraea angustula* (Gyllenhal) with 89 individuals (23.1% of the collected rove beetle). Finally, the most abundant spider family was Titanoecidae, and the dominant spider species was its representative *Nurscia albomaculata* (Lucas) with 781 individuals (35.3% of the collected spiders). An interesting faunistic note is represented by the first Italian record of the spider *Zelotes metellus* (Roewer) (Araneae: Gnaphosidae), which was found in 8 out of 9 studied fields with a total of 41 individuals.

3.2. Bioindicator potential

We did not detect any significant relationships between the activity densities of the studied groups. Conversely, we found significant positive correlations among the species richness of all arthropod groups (Table 1). Pairwise correlations in community turnover among all groups were also positive and significant (Table 1).

The most abundant species, which were tested for their potential as indicators of group-level features, included the ground beetles *H. rufipes*, *Pterostichus macer* (Marsham) and *Pterostichus melas* (Creutzer), the spiders *N. albomaculata*, *Pardosa cribrata* Simon and *Pardosa agrestis*

Table 1

Pearson correlations (for activity density and species richness) and Mantel correlations (for community turnover) between the studied arthropod groups.

| | Activity density | | Species richness | | Community turnover | |
|---|-----------------------|-------------------------|----------------------|-------------------------|-----------------------|-------------------------|
| Pair type | r _p | р | r _p | р | r _m | р |
| Carabidae - Staphylinidae Carabidae - Araneae Araneae - Staphylinidae | 0.51 -0.14 0.29 | 0.132 0.703 0.423 | 0.78 0.71 0.71 | 0.008 0.023 0.020 | 0.39 0.29 0.31 | 0.013 0.041 0.032 |

(Westring) and the rove beetle *D. angustula*. The activity densities of the tested ground and rove beetle species always had significant positive correlations with the overall activity density and species richness of their respective groups (Table 2). Moreover, the ground beetle *P. macer* was positively correlated with rove beetle activity density, while the rove beetle *D. angustula* was positively correlated with both activity density and species richness of ground beetles. The only significant correlations for spiders were between *N. albomaculata* and overall spider activity density, and between *P. cribrata* and overall spider species richness.

Considering the whole arthropod dataset, rove beetle species showed a significantly higher proportion of positive co-occurrences with other species in the 3 groups (Fig. 1a). When dividing the dataset by arthropod taxon pairing, the proportionally highest number of co-occurrences was among rove beetle species, and the lowest was among ground beetle and spider species, or between different spider species, with the proportion of co-occurrences between ground and rove beetle species at an intermediate level (Fig. 1b).

The species showing the highest number of co-occurrences across the entire dataset were the rove beetles *Anotylus sculpturatus* (Gravenhorst) and *Tachyporus hypnorum* (Fabricius) (Table A.4) and the spider *Trachyzelotes huberti* Platnick & Murphy with 13 co-occurrences each (24.1% of the dataset) (Table A.5). *Anotylus sculpturatus* and *T. hypnorum* also had the highest proportion of co-occurrences with ground beetle species (5 species for each, 29.4% of the ground beetle dataset) and with spider species (5 species each, 17.9% of the spider dataset). On the other hand, the ground beetle *Parophonus maculicornis* (Duftschmid) and the spider *Robertus arundineti* (O. Pickard-Cambridge) had the highest proportion of co-occurrences with rove beetle species (6 species each, 60% of the rove beetle dataset).

4. Discussion

4.1. Community features and faunistic notes

Our data shed some light on the ground beetle, rove beetle and spider fauna of Italian maize agroecosystems, highlighting the local presence of several species considered important or potentially important as biocontrol agents (Carbonne et al., 2020; Klimaszewski et al., 2018; Kuusk and Ekbom, 2010; Renkema et al., 2012). In spite of their lower abundance (6 and 9 fold lower than spiders and ground beetles, respectively), rove beetles too were recorded with a comparatively high number of species, a pattern that had been found before in agricultural landscapes (Corcos et al., 2021). Assemblages of all arthropod groups (and especially ground beetles, the most abundant) were dominated by a



Fig. 1. (a) Effect of arthropod group on the proportion of positive species cooccurrences of each group over the entire dataset. (b) Effect of arthropod group pairing on the proportion of positive species co-occurrences in the pairing. The results of the relative logit regression models, complete with pvalues, are reported on the plots. Different letters indicate statistically significant differences (p < 0.05) according to the pairwise comparison test with Tukey adjustment (p < 0.05).

Table 2

Pearson correlations between the activity density (AD) of the most abundant arthropod species and the activity density and species richness of the 3 arthropod groups.

| | | Carabidae AD | | Staphylinidae AD | | Araneae AD | |
|--|--|--|---|--|---|---|---|
| Group | Species | r _p | р | r _p | р | r _p | р |
| Carabidae | Harpalus rufipes (Duftschmid) | 0.98 | < 0.001 | 0.39 | 0.254 | -0.23 | 0.515 |
| | Pterostichus macer (Marsham) | 0.89 | < 0.001 | 0.64 | 0.048 | 0.12 | 0.748 |
| | Pterostichus melas (Creutzer) | 0.97 | < 0.001 | 0.39 | 0.254 | -0.29 | 0.413 |
| Staphylinidae | Dinarea angustula (Gyllenhal) | 0.69 | 0.026 | 0.83 | 0.003 | 0.12 | 0.749 |
| Araneae | Nurscia albomaculata (Lucas) | -0.54 | 0.110 | -0.15 | 0.686 | 0.74 | 0.015 |
| | Pardosa cribrata Simon | 0.34 | 0.332 | 0.36 | 0.305 | 0.15 | 0.674 |
| | Pardosa agrestis (Westring) | 0.39 | 0.258 | 0.57 | 0.085 | 0.05 | 0.882 |
| | | Carabidae richness | | | | | |
| | | Carabidae ri | chness | Staphylinida | e richness | Araneae rich | ness |
| Group | Species | Carabidae rie r _p | chness | Staphylinida r _p | e richness | Araneae rich: r _p | ness p |
| Group Carabidae | Species Harpalus rufipes (Duftschmid) | Carabidae rie r _p 0.73 | <u>p</u> 0.017 | Staphylinida r _p 0.42 | e richness p 0.230 | Araneae rich r _p 0.57 | ness <u>p</u> 0.086 |
| Group Carabidae | Species Harpalus rufipes (Duftschmid) Pterostichus macer (Marsham) | Carabidae rio r _p 0.73 0.81 | p 0.017 0.005 | Staphylinida r _p 0.42 0.54 | e richness | Araneae rich r _p 0.57 0.57 | p 0.086 0.085 |
| Group Carabidae | Species Harpalus rufipes (Duftschmid) Pterostichus macer (Marsham) Pterostichus melas (Creutzer) | Carabidae rie r _p 0.73 0.81 0.71 | chness p 0.017 0.005 0.021 | Staphylinida r _p 0.42 0.54 0.46 | e richness p 0.230 0.104 0.178 | Araneae rich r _p 0.57 0.57 0.57 0.57 | ness <i>p</i> 0.086 0.085 0.083 |
| Group Carabidae Staphylinidae | Species Harpalus rufipes (Duftschmid) Pterostichus macer (Marsham) Pterostichus melas (Creutzer) Dinarea angustula (Gyllenhal) | Carabidae rie r _p 0.73 0.81 0.71 0.75 | chness p 0.017 0.005 0.021 0.013 | Staphylinida r _p 0.42 0.54 0.46 0.67 | e richness <u>p</u> 0.230 0.104 0.178 0.034 | Araneae rich r _p 0.57 0.57 0.57 0.49 | ness p 0.086 0.085 0.083 0.155 |
| Group Carabidae Staphylinidae Araneae | Species Harpalus rufipes (Duftschmid) Pterostichus macer (Marsham) Pterostichus melas (Creutzer) Dinarea angustula (Gyllenhal) Nurscia albomaculata (Lucas) | Carabidae rie r _p 0.73 0.81 0.71 0.75 -0.28 | chness p 0.017 0.005 0.021 0.013 0.434 | Staphylinida rp 0.42 0.54 0.46 0.67 -0.23 | e richness p 0.230 0.104 0.178 0.034 0.526 | Araneae rich r _p 0.57 0.57 0.57 0.57 0.49 -0.16 | ness p 0.086 0.085 0.083 0.155 0.650 |
| Group Carabidae Staphylinidae Araneae | Species Harpalus rufipes (Duftschmid) Pterostichus macer (Marsham) Pterostichus melas (Creutzer) Dinarea angustula (Gyllenhal) Nurscia albomaculata (Lucas) Pardosa cribrata Simon | Carabidae rie rp 0.73 0.81 0.71 0.75 -0.28 0.33 | p 0.017 0.005 0.021 0.013 0.434 0.348 | Staphylinida r _p 0.42 0.54 0.46 0.67 -0.23 0.45 | e richness p 0.230 0.104 0.178 0.034 0.526 0.193 | Araneae rich rp 0.57 0.57 0.57 0.49 -0.16 0.73 | ness p 0.086 0.085 0.083 0.155 0.650 0.016 |

limited number of agrobiont species adaptable to heavily intensified agricultural contexts.

The prevalent ground beetle *H. rufipes*, in particular, is often very common in agroecosystems (Labruyere et al., 2016), were it has been reported as an important omnivorous natural enemy of both weeds and pest insects (Carbonne et al., 2020; Monzó et al., 2011). Roughly two thirds of the remaining ground beetle individuals belonged to the second and third most abundant species of the study, the opportunist predators *P. macer* and *P. melas* (Giglio et al., 2021; Nourmohammadpour-amiri et al., 2022), which are both habitat generalists and common in agricultural landscapes (Corcos et al., 2021; Lövei et al., 2006; Vician et al., 2015). The dominant rove beetle species *D. angustula* is also a common agrobiont (Balog et al., 2009). Assemblages dominated by such species are frequently found in disturbed agricultural ecosystems (Lami et al., 2021; Shearin et al., 2008).

Nonetheless, several less common and more specialized species were sampled as well; for instance, a single individual of the ground beetle Calosoma sycophanta (L.) was collected. This was an unusual finding given the specialization of this species for forest habitats (Burgess, 1911), which are extremely scarce in our study areas and completely lacking in the landscape surrounding the field (CAS2) in which the species was found (Table A.2). We even recorded in most sites the presence of the spider Z. metellus, which was previously known from Greece, Iran, Israel, Russia and recently France (Mazzia et al., 2018), and which had never been reported in Italy - a reminder of the fact that our knowledge of the fauna of even these supposedly familiar and simplified agroecosystems is often lacking. Baseline faunistic datasets such as these are decisive for the organization of conservation efforts and for monitoring the long term effects of anthropogenic impacts (Ejsmont-Karabin, 2019; Girardello et al., 2018; Valdecasas and Camacho, 2003) including climate change and agricultural practices such as tillage, pesticide use and genetically modified plant cultivation (Arpaia et al., 2018; Lami et al., 2016; Lener et al., 2013).

4.2. Community-level bioindication potential

Before discussing bioindication potential, it is worth mentioning that the relatively limited spatial and temporal extent of the sampling might represent a drawback. Even though the conditions of the studied sites are very common in maize agroecosystems of Northern Italy, thus improving the likelihood that the conclusions of our study could find application in a wider area, confident extrapolation of general patterns (even if just at the regional scale) should definitely be based on additional research in similar environmental contexts. That being said, some interesting considerations can be made about our results.

Regarding community-level bioindication, activity density showed little potential as a biodiversity indication parameter in the studied area and habitat type, as correlations between groups were never significant. Species richness and community turnover, however, showed significant co-variation among all three groups. The correlations involving species richness were particularly strong, with r values always higher than 0.7 a value that has been often proposed as a threshold to identify truly informative biodiversity indicators (Harry et al., 2019; Heino, 2010; Lovell et al., 2007). Such strong correlations are rarely found in real world scenarios (Burrascano et al., 2018; Larrieu et al., 2019; Oberprieler et al., 2020), a fact that is often attributed to the confounding effects of different environmental and habitat factors (Filgueiras et al., 2019; Schalkwyk et al., 2019). Our findings, therefore, support the appropriateness of focusing on a specific habitat type, and the fact that monitoring biodiversity through surrogate taxa might be more feasible in maize agroecosystems than in other contexts.

In addition to that, ground beetles are taxonomically well-known, and usually easier to identify than the other two studied groups (Kotze et al., 2011). Given their importance as biocontrol agents of pest invertebrates and weeds (Honek et al., 2003; Lövei and Sunderland, 1996) and their use as environmental indicators (Piano et al., 2020; Pizzolotto et al., 2018; Rainio and Niemelä, 2003) there is great interest in ground beetle conservation. For the above reasons, ground beetles have been widely studied, and a relatively high number of experts are available for their identification (Kotze et al., 2011; Magura and Lövei, 2021; Niemelä, 1995). Ground beetles thus have all the necessary features to be considered potentially useful bioindicators of the species richness and community turnover of rove beetle and spider communities in maize agroecosystems of Northern Italy. Future research will have to assess the validity of our findings for other geographical areas, as cross-taxon congruence involving ground beetles has been mostly investigated in relation to different organisms such as plants or in a variety of habitat types, not focusing specifically on maize (Oberprieler et al., 2020; Uboni et al., 2019; Zara et al., 2021). Nonetheless, some existing sources concerning general bioindication potential in maize seem to draw an encouraging picture (Albajes et al., 2013; Lee and Albajes, 2016).

4.3. Abundant species as bioindicators of community features

While tested for the sake of completeness, the fact that the activity density of the most abundant species was often linked with the activity density of their respective groups is hardly surprising and scarcely informative. Skewed species-abundance distributions, in which one or few species are disproportionally influential on community size and function because of their numerical dominance, are very common in nature (McGill et al., 2007; Winfree et al., 2015). Notwithstanding, the case of the two species that have significant correlations with different groups (the ground beetle *P. macer* with rove beetles, and the rove beetle D. angustula with ground beetles) is potentially important for bioindication. Pterostichus macer seems particularly promising, being a significantly larger ground beetle (Klimaszewski et al., 2013; Magura et al., 2006), and thus comparatively easier to confidently identify, and being more common than D. angustula, at least in our study. Pterostichus macer might thus be a useful, easily sampled indicator that simplifies the estimation of rove beetle abundance and activity, making up for the lack of significant correlation between overall ground beetle activity density and rove beetle activity density.

Perhaps even more interesting are the cases of species whose activity density is significantly correlated with the species richness of their respective groups. In literature it is often reported that increasing arthropod community abundance, while usually disproportionately driven by common species, is also linked with higher species richness (Hallmann et al., 2021; McArt et al., 2012). This might be explained by the more-individuals hypothesis (Gaston, 2000), which postulates that habitats with a higher availability of resources can support a higher number of individuals of each species (including dominant ones), which in turn contributes at least partially to a higher species richness, as larger populations have a lower chance of extinction (Storch et al., 2018). The practical consequence is that the activity density of H. rufipes, P. macer and P. melas could be used as an indicator for overall ground beetle species richness in maize agroecosystems of Northern Italy, just as the activity density of D. angustula and P. cribrata (alongside ground beetle richness) could be used as indicators of rove beetle and spider richness respectively. This could potentially greatly reduce the taxonomic burden of monitoring the diversity of these groups. It is also worth noting that D. angustula activity density could be an indicator of ground beetle richness, although for the reasons discussed above it would be likely more convenient to use one of the more abundant and easily identifiable ground beetle species, rather than this rarer and more difficult to identify rove beetle.

4.4. Species co-occurrence

The situation is more complex when examining species-specific bioindication (co-occurrence). Available literature reports varying levels of positive ("true") co-occurrence and avoidance among grounddwelling arthropod species, once again often indicating environmental features as key drivers of these patterns (Fernandes et al., 2020; Tsafack et al., 2021; Ulrich et al., 2010), as the availability of space, food and other resources, as well as the level of disturbance, heavily influences the strength of interspecific competition. The very limited number of negative co-occurrences (i. e. species avoidances) suggested that strong competition or strong differences in general habitat requirements are extremely rare among the species in the studied context. Positive cooccurrences were more common, but they still represented a minority of the dataset. Agricultural landscapes, and maize agroecosystems in particular, are highly disturbed and environmentally homogeneous (Chmelíková and Wolfrum, 2019; Ponisio et al., 2016), so the low proportion of positive and negative relations is coherent with literature highlighting disturbance as a factor disrupting segregated arthropod communities (Pitzalis et al., 2010; Ulrich et al., 2010), and habitat homogeneity as a factor disrupting both segregation and co-occurrence (Tsafack et al., 2021). More heterogeneous landscapes with a high proportion of semi-natural habitats and margins might have yielded different results and thus influenced the implications for bioindication, given the importance that these elements have as habitat resources for arthropods (Bertrand et al., 2016; Maas et al., 2021; Marshall and Moonen, 2002). In our case, the highest proportion of positive cooccurrences were found between rove beetle species and other arthropod species (especially other rove beetles), with spiders showing the lowest proportion of co-occurrences with other species and ground beetles not faring much better than spiders in that regard. The most common rove beetle species in the studied habitats are thus likely to share the same environmental needs with each other in a much stronger fashion than either spiders or ground beetles.

It follows that most ground beetle species will be scarcely useful as bioindicators of other predatory arthropod species in the studied context, with some exceptions. Among these promising species-specific bioindicator candidates we can mention P. maculicornis - especially as a rove beetle bioindicator, given its co-occurrence with a sizeable portion of the most common rove beetle species of the studied sites. A possible limitation that can be mentioned in the case of this species is the fact that, at least in our study, it was a scarce species (11 individuals), meaning that detecting its presence in an area might require an intensive sampling effort. Many rove beetles, such as A. sculpturatus and T. hypnorum, and even some spiders, such as T. huberti and R. arundineti, also showed potential as species-specific biodiversity indicators for one or more arthropod groups because of their relatively high proportion of co-occurrences with other species – with A. sculpturatus also having the advantage of being relatively abundant (74 individuals) when compared with the other species. However, it must be considered that the taxonomy and identification of rove beetles and spiders is often complex (Bohac, 1999; Pearce and Venier, 2006) and this might limit the usefulness of these species in biodiversity indication. It is thus advisable, while further investigating the potential of these promising species, to also continue the search for other surrogate taxa that can act as speciesspecific bioindicators for ground-dwelling predatory arthropods of maize.

5. Conclusions

In this study, we characterized the ground-dwelling predatory arthropod fauna in maize agroecosystems of Northern Italy, focusing on three groups potentially very important for biological control – ground beetles, rove beetles and spiders. Positive correlations were found among the species richness and community turnover of all arthropod groups, whereas activity density correlations were non-significant. In particular, we identified ground beetles as useful bioindicators for the species richness and community turnover of the other two groups. Additionally, several abundant arthropod species worked as bioindicators of the species richness of their respective groups, and the ground beetle *P. macer* also worked as a bioindicator of overall rove beetle activity density. We also showed that the potential for species specific bioindication was in general limited in the three studied groups, highlighting the need to search for other potential surrogate taxa to fill this gap. Nonetheless, we did find some species showing promise in this sense, such as the ground beetle *P. maculicornis* as a rove beetle bioindicator, which might deserve further evaluation.

The results of this work could improve the monitoring and management of these important natural enemies in maize-rich regions, speeding up biodiversity assessments and thus facilitating the understanding of the relation between biodiversity and environmental/agricultural factors (Dudley and Alexander, 2017), or between biodiversity and ecosystem services provision (Duncan et al., 2015). Another important application of these biodiversity indicators would be facilitating the identification of biodiversity-rich cropland that should be maintained in its current state and of biodiversity-poor cropland to be restored (Barral et al., 2015). Finally, our study provides baseline faunistic datasets that could prove useful as a reference to detect and evaluate the effects of climate change, land use change and agricultural practices on soil biodiversity.

CRediT authorship contribution statement

Francesco Lami: Conceptualization, Methodology, Investigation, Data curation, Formal analysis, Visualization, Writing – original draft, Writing – review & editing. **Giovanni Burgio:** Supervision, Funding acquisition, Writing – review & editing. **Serena Magagnoli:** Investigation, Writing – review & editing. **Daniele Sommaggio:** Investigation, Writing – review & editing. **Roland Horváth:** Investigation, Writing – review & editing. **Dávid D. Nagy:** Investigation, Writing – review & editing. **Antonio Masetti:** Conceptualization, Investigation, Data curation, Supervision, Writing – review & editing.

Declaration of Competing Interest

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests: Giovanni Burgio reports financial support was provided by European Commission Seventh Framework Programme for Research and Technological Development.

Data availability

Data will be made available on request.

Acknowledgements

This study was funded by the FP7 European project AMIGA – "Assessing and Monitoring the Impacts of Genetically modified plants on Agro-ecosystems" (grant agreement ID: 289706). Spider vector art should be credited to Lafage and ground beetle vector art should be credited to T. Michael Keesey (vectorization); Thorsten Assmann, Jörn Buse, Claudia Drees, Ariel-Leib-Leonid Friedman, Tal Levanony, Andrea Matern, Anika Timm, and David W. Wrase (photography) (https://cre ativecommons.org/licenses/by/3.0/). We are grateful to Agata Morelli (DISTAL – University of Bologna) for the English language and to Irene Diti (DISTAL – University of Bologna) for the land-scape analysis.

Appendix A. Supplementary data

Supplementary data to this article can be found online at https://doi.org/10.1016/j.ecolind.2023.110352.

References

- Albajes, R., Lumbierres, B., Pons, X., Comas, J., 2013. Representative taxa in field trials for environmental risk assessment of genetically modified maize. Bull. Entomol. Res. 103, 724–733. https://doi.org/10.1017/S0007485313000473.
- Albertini, A., Marchi, S., Ratti, C., Burgio, G., Petacchi, R., Magagnoli, S., 2018. Bactrocera oleae pupae predation by Ocypus olens detected by molecular gut content analysis. BioControl 63, 227–239. https://doi.org/10.1007/s10526-017-9860-6.
- Altieri, M.A., 1999. The ecological role of biodiversity in agroecosystems. Agric. Ecosyst. Environ. 74, 19–31. https://doi.org/10.1016/S0167-8809(99)00028-6.
- Araújo, M.B., Rozenfeld, A., Rahbek, C., Marquet, P.A., 2011. Using species cooccurrence networks to assess the impacts of climate change. Ecography (Cop.) 34, 897–908. https://doi.org/10.1111/j.1600-0587.2011.06919.x.
- Arpaia, S., Baldacchino, F., Bosi, S., Burgio, G., Errico, S., Magarelli, R.A., Masetti, A., Santorsola, S., 2018. Evaluation of the potential exposure of butterflies to genetically modified maize pollen in protected areas in Italy. Insect Sci. 25, 549–561. https:// doi.org/10.1111/1744-7917.12591.
- Assing, V., Schülke, M., 2012. Die Käfer Mitteleuropas. Band 4. Staphylinidae I. Zweite neubearbeitete Auflage. Spektrum Akademischer Verlag, Berlin/Heidelberg.
- Aubier, T.G., Elias, M., 2020. Positive and negative interactions jointly determine the structure of Müllerian mimetic communities. Oikos 129, 983–997. https://doi.org/ 10.1111/oik.06789.
- Balmford, A., Gaston, K.J., 1999. Why biodivdersity surveys are good value 398, 204–205. 18339.
- Balog, A., Markó, V., Imre, A., 2009. Farming system and habitat structure effects on rove beetles (Coleoptera: Staphylinidae) assembly in Central European apple and pear orchards. Biologia (Bratisl). 64, 343–349. https://doi.org/10.2478/s11756-009-0045-3.
- Barlow, J., Gardner, T.A., Louzada, J., Peres, C.A., 2010. Measuring the conservation value of tropical primary forests: The effect of occasional species on estimates of biodiversity uniqueness. PLoS One 5. https://doi.org/10.1371/journal. pone.0009609.
- Barral, M.P., Rey Benayas, J.M., Meli, P., Maceira, N.O., 2015. Quantifying the impacts of ecological restoration on biodiversity and ecosystem services in agroecosystems: A global meta-analysis. Agric. Ecosyst. Environ. 202, 223–231. https://doi.org/ 10.1016/j.agee.2015.01.009.
- Begg, G.S., Cook, S.M., Dye, R., Ferrante, M., Franck, P., Lavigne, C., Lövei, G.L., Mansion-Vaquie, A., Pell, J.K., Petit, S., Quesada, N., Ricci, B., Wratten, S.D., Birch, A.N.E., 2017. A functional overview of conservation biological control. Crop Prot. 97, 145–158. https://doi.org/10.1016/j.cropro.2016.11.008.
- Bell, J.R., King, R.A., Bohan, D.A., Symondson, W.O.C., 2010. Spatial co-occurrence networks predict the feeding histories of polyphagous arthropod predators at field scales. Ecography (Cop.) 33, 64–72. https://doi.org/10.1111/j.1600-0587 2009 06046 x
- Bertrand, C., Burel, F., Baudry, J., 2016. Spatial and temporal heterogeneity of the crop mosaic influences carabid beetles in agricultural landscapes. Landsc. Ecol. 31, 451–466. https://doi.org/10.1007/s10980-015-0259-4.
- Betz, O., Irmler, U., Klimaszewski, J., 2018. Biology of rove beetles (Staphylinidae), Biology of Rove Beetles (Staphylinidae). Springer International Publishing, Cham. 10.1007/978-3-319-70257-5.
- Birkhofer, K., Rusch, A., Andersson, G.K.S., Bommarco, R., Dänhardt, J., Ekbom, B., Jönsson, A., Lindborg, R., Olsson, O., Rader, R., Stjernman, M., Williams, A., Hedlund, K., Smith, H.G., 2018. A framework to identify indicator species for ecosystem services in agricultural landscapes. Ecol. Indic. 91, 278–286. https://doi. org/10.1016/j.ecolind.2018.04.018.
- Bohac, J., 1999. Staphylinid beetles as bioindicators. Agric. Ecosyst. Environ. 74, 357–372. https://doi.org/10.1016/S0167-8809(99)00043-2.
- Brown, G.R., Matthews, I.M., 2016. A review of extensive variation in the design of pitfall traps and a proposal for a standard pitfall trap design for monitoring ground-active arthropod biodiversity. Ecol. Evol. 6, 3953–3964. https://doi.org/10.1002/ ecc3.2176.
- Bucher, R., Nickel, H., Kaib, S., Will, M., Carchi, J., Farwig, N., Schabo, D.G., 2019. Birds and plants as indicators of arthropod species richness in temperate farmland. Ecol. Indic. 103, 272–279. https://doi.org/10.1016/j.ecolind.2019.04.011.
- Burgess, A.F., 1911. Calosoma Sycophanta: Its Life History, Behavior, and Successful Colonization in New England. US Department of Agriculture, Bureau of Entomology.
- Burrascano, S., de Andrade, R.B., Paillet, Y., Ódor, P., Antonini, G., Bouget, C., Campagnaro, T., Gosselin, F., Janssen, P., Persiani, A.M., Nascimbene, J., Sabatini, F. M., Sitzia, T., Blasi, C., 2018. Congruence across taxa and spatial scales: Are we asking too much of species data? Glob. Ecol. Biogeogr. 27, 980–990. https://doi.org/ 10.1111/geb.12766.
- Büttner, G., Feranec, J., Jaffrain, J., 2002. CORINE Land Cover Update 2000. Technical Guidelines. European Environment Agency.
- Cane, J.H., Payne, J.A., 1993. Regional, annual, and seasonal variation in pollinator guilds: Intrinsic traits of bees (Hymenoptera: Apoidea) underlie their patterns of abundance at Vaccinium ashei (Ericaceae). Ann. Entomol. Soc. Am. 86, 577–588. https://doi.org/10.1093/aesa/86.5.577.
- Carbonne, B., Bohan, D.A., Petit, S., 2020. Key carabid species drive spring weed seed predation of Viola arvensis. Biol. Control 141, 104148. https://doi.org/10.1016/j. biocontrol.2019.104148.
- Carvalho, J.C., Cardoso, P., Borges, P.A.V., Schmera, D., Podani, J., 2013. Measuring fractions of beta diversity and their relationships to nestedness: a theoretical and empirical comparison of novel approaches. Oikos 122, 825–834. https://doi.org/ 10.1111/j.1600-0706.2012.20980.x.

- Chmelíková, L., Wolfrum, S., 2019. Mitigating the biodiversity footprint of energy crops – A case study on arthropod diversity. Biomass Bioenergy 125, 180–187. https://doi. org/10.1016/j.biombioe.2019.04.023.
- Corcos, D., Lami, F., Nardi, D., Boscutti, F., Sigura, M., Giannone, F., Pantini, P., Tagliapietra, A., Busato, F., Sibella, R., Marini, L., 2021. Cross-taxon congruence between predatory arthropods and plants across Mediterranean agricultural landscapes. Ecol. Indic. 123, 107366 https://doi.org/10.1016/j. ecolind.2021.107366.
- Dangles, O., Casas, J., 2019. Ecosystem services provided by insects for achieving sustainable development goals. Ecosyst. Serv. 35, 109–115. https://doi.org/ 10.1016/j.ecoser.2018.12.002.
- Dudley, N., Alexander, S., 2017. Agriculture and biodiversity: a review. Biodiversity 18, 45–49. https://doi.org/10.1080/14888386.2017.1351892.
- Duelli, P., Obrist, M.K., 2003. Biodiversity indicators: The choice of values and measures. Agric. Ecosyst. Environ. 98, 87–98. https://doi.org/10.1016/S0167-8809(03)00072-0.
- Duncan, C., Thompson, J.R., Pettorelli, N., 2015. The quest for a mechanistic understanding of biodiversity–Ecosystem services relationships. Proc. R. Soc. B Biol. Sci. 282 https://doi.org/10.1098/rspb.2015.1348.
- Ebach, M.C., Valdecasas, A.G., Wheeler, Q.D., 2011. Impediments to taxonomy and users of taxonomy: Accessibility and impact evaluation. Cladistics 27, 550–557. https:// doi.org/10.1111/j.1096-0031.2011.00348.x.
- Ejsmont-Karabin, J., 2019. Does the world need faunists? Based on rotifer (Rotifera) occurrence reflections on the role of faunistic research in ecology. Int. Rev. Hydrobiol. 104, 49–56. https://doi.org/10.1002/iroh.201901991.
- Fernandes, T.T., Dáttilo, W., Silva, R.R., Luna, P., Braz, A.B., Morini, M.S.C., 2020. Cohabitation and niche overlap in the occupation of twigs by arthropods in the leaf litter of Brazilian Atlantic Forest. Insectes Soc. 67, 239–247. https://doi.org/ 10.1007/s00040-020-00753-w.
- Fiedler, A.K., Landis, D.A., Wratten, S.D., 2008. Maximizing ecosystem services from conservation biological control: The role of habitat management. Biol. Control 45, 254–271. https://doi.org/10.1016/j.biocontrol.2007.12.009.
- Filgueiras, B.K.C., Melo, D.H.A., Andersen, A.N., Tabarelli, M., Leal, I.R., 2019. Crosstaxon congruence in insect responses to fragmentation of Brazilian Atlantic forest. Ecol. Indic. 98, 523–530. https://doi.org/10.1016/j.ecolind.2018.11.036.
- Freude, H., Harde, K.W., Lohse, G.A., 1974. Staphylinidae II (Hypocyphtinae und Aleocharinae) Pselaphidae, Band 5. Die K\u00e4fer Mitteleuropas. Goecke & Evers Verlag, Krefeld.
- Gallé, R., Geppert, C., Földesi, R., Tscharntke, T., Batáry, P., 2020. Arthropod functional traits shaped by landscape-scale field size, local agri-environment schemes and edge effects. Basic Appl. Ecol. 48, 102–111. https://doi.org/10.1016/j.baae.2020.09.006.
- Gao, T., Nielsen, A.B., Hedblom, M., 2015. Reviewing the strength of evidence of biodiversity indicators for forest ecosystems in Europe. Ecol. Indic. 57, 420–434. https://doi.org/10.1016/j.ecolind.2015.05.028.
- Garbach, K., Milder, J.C., Montenegro, M., Karp, D.S., DeClerck, F.A.J., 2014. Biodiversity and Ecosystem Services in Agroecosystems. Encycl. Agric. Food Syst. 2, 21-40. https://doi.org/10.1016/B978-0-444-52512-3.00013-9.
- Gaston, K.J., 2000. Global patterns in biodiversity. Nature 405, 220-227.
- Gaver, C., Lövei, G.L., Magura, T., Dieterich, M., Batáry, P., 2019. Carabid functional diversity is enhanced by conventional flowering fields, organic winter cereals and edge habitats. Agric. Ecosyst. Environ. 284, 106579 https://doi.org/10.1016/j. agee.2019.106579.
- Giglio, A., Vommaro, M.L., Gionechetti, F., Pallavicini, A., 2021. Gut microbial community response to herbicide exposure in a ground beetle. J. Appl. Entomol. 145, 986–1000. https://doi.org/10.1111/jen.12919.
- Girardello, M., Martellos, S., Pardo, A., Bertolino, S., 2018. Gaps in biodiversity occurrence information may hamper the achievement of international biodiversity targets: Insights from a cross-Taxon analysis. Environ. Conserv. 45, 370–377. https://doi.org/10.1017/S0376892918000115.
- Gotelli, N.J., McCabe, D.J., 2002. Species co-occurrence: A meta-analysis of J. M. Diamond's assembly rules model. Ecology 83, 2091–2096. https://doi.org/10.1890/ 0012-9658(2002)083[2091:SCOAMA]2.0.CO;2.
- Griffith, D.M., Veech, J.A., Marsh, C.J., 2016. Cooccur: Probabilistic species cooccurrence analysis in R. J. Stat. Softw. 69, 1–17. 10.18637/jss.v069.c02.
- Hågvar, S., Klanderud, K., 2009. Effect of simulated environmental change on alpine soil arthropods. Glob. Chang. Biol. 15, 2972–2980. https://doi.org/10.1111/j.1365-2486.2009.01926.x.
- Hallmann, C.A., Ssymank, A., Sorg, M., de Kroon, H., Jongejans, E., 2021. Insect biomass decline scaled to species diversity: General patterns derived from a hoverfly community. Proc. Natl. Acad. Sci. U. S. A. 118, 1–8. https://doi.org/10.1073/ PNAS.2002554117.
- Harry, I., Höfer, H., Schielzeth, H., Assmann, T., 2019. Protected habitats of Natura 2000 do not coincide with important diversity hotspots of arthropods in mountain grasslands. Insect Conserv. Divers. 12, icad.12349. https://doi.org/10.1111/ icad.12349.
- Harvey, J.A., van der Putten, W.H., Turin, H., Wagenaar, R., Bezemer, T.M., 2008. Effects of changes in plant species richness and community traits on carabid assemblages and feeding guilds. Agric. Ecosyst. Environ. 127, 100–106. https://doi.org/10.1016/ j.agee.2008.03.006.
- Heino, J., 2010. Are indicator groups and cross-taxon congruence useful for predicting biodiversity in aquatic ecosystems? Ecol. Indic. 10, 112–117. https://doi.org/ 10.1016/j.ecolind.2009.04.013.
- Honek, A., Martinkova, Z., Jarosik, V., 2003. Ground beetles (Carabidae) as seed predators. Eur. J. Entomol. 100, 531–544. 10.14411/eje.2003.081.

- Hopkins, G.W., Freckleton, R.P., 2002. Declines in the numbers of amateur and professional taxonomists: Implications for conservation. Anim. Conserv. 5, 245–249. https://doi.org/10.1017/S1367943002002299.
- Klimaszewski, J., Brunke, A.J., Work, T.T., Venier, L., 2018. Rove beetles (Coleoptera, Staphylinidae) as bioindicators of change in boreal forests and their biological control services in agroecosystems: Canadian case studies. In: Biology of Rove Beetles (Staphylinidae). Springer International Publishing, Cham, pp. 161–181. 10.1007/978-3-319-70257-5_9.
- Klimaszewski, J., Webster, R.P., Langor, D.W., Bourdon, C., Jacobs, J., 2013. Review of Canadian species of the genus Dinaraea Thomson, with descriptions of six new species (Coleoptera, Staphylinidae, Aleocharinae, Athetini). Zookeys 327, 65–101. https://doi.org/10.3897/zookeys.327.5908.
- Kotze, D.J., Brandmayr, P., Casale, A., Dauffy-Richard, E., Dekoninck, W., Koivula, M.J., Lövei, G.L., Mossakowski, D., Noordijk, J., Paarmann, W., Pizzolotto, R., Saska, P., Schwerk, A., Serrano, J., Szyszko, J., Taboada, A., Turin, H., Venn, S., Vermeulen, R., Zetto, T., 2011. Forty years of carabid beetle research in Europe - from taxonomy, biology, ecology and population studies to bioindication, habitat assessment and conservation. Zookeys 100, 55–148. https://doi.org/10.3897/zookeys.100.1523.
- Kremen, C., Colwell, R.K., Erwin, T.L., Murphy, D.D., Noss, R.F., Sanjayan, M.A., 1993. Terrestrial arthropod assemblages: their use in conservation planning. Conserv. Biol. 7, 796–808. https://doi.org/10.1046/j.1523-1739.1993.740796.x.
- Kromp, B., 1999. Carabid beetles in sustainable agriculture: a review on pest control efficacy, cultivation impacts and enhancement. Agric. Ecosyst. Environ. 74, 187–228. https://doi.org/10.1016/b978-0-444-50019-9.50014-5.
- Kuusk, A.K., Ekbom, B., 2010. Lycosid spiders and alternative food: Feeding behavior and implications for biological control. Biol. Control 55, 20–26. https://doi.org/ 10.1016/j.biocontrol.2010.06.009.
- Labruyere, S., Ricci, B., Lubac, A., Petit, S., 2016. Crop type, crop management and grass margins affect the abundance and the nutritional state of seed-eating carabid species in arable landscapes. Agric. Ecosyst. Environ. 231, 183–192. https://doi.org/ 10.1016/j.agee.2016.06.037.
- Lami, F., Masetti, A., Neri, U., Lener, M., Staiano, G., Arpaia, S., Burgio, G., 2016. Diversity of Coccinellidae in ecological compensation areas of Italy and overlap with maize pollen shedding period. Bull. Insectol. 69, 49–57.
- Lami, F., Boscutti, F., Masin, R., Sigura, M., Marini, L., 2020. Seed predation intensity and stability in agro-ecosystems: Role of predator diversity and soil disturbance. Agric. Ecosyst. Environ. 288, 106720 https://doi.org/10.1016/j.agee.2019.106720.
- Lami, F., Bartomeus, I., Nardi, D., Beduschi, T., Boscutti, F., Pantini, P., Santoiemma, G., Scherber, C., Tscharntke, T., Marini, L., 2021. Species-habitat networks elucidate landscape effects on habitat specialisation of natural enemies and pollinators. Ecol. Lett. 24, 288–297. https://doi.org/10.1111/ele.13642.
- Larrieu, L., Gosselin, F., Archaux, F., Chevalier, R., Corriol, G., Dauffy-Richard, E., Deconchat, M., Gosselin, M., Ladet, S., Savoie, J.M., Tillon, L., Bouget, C., 2019. Assessing the potential of routine stand variables from multi-taxon data as habitat surrogates in European temperate forests. Ecol. Indic. 104, 116–126. https://doi. org/10.1016/j.ecolind.2019.04.085.
- Lee, M.S., Albajes, R., 2016. Monitoring carabid indicators could reveal environmental impacts of genetically modified maize. Agric. For. Entomol. 18, 238–249. https:// doi.org/10.1111/afe.12156.
- Lener, M., Giovannelli, V., Arpaia, S., Baldacchino, F., Benedetti, A., Burgio, G., Canfora, L., Dinelli, G., Manachini, B., Marotti, I., Masetti, A., Sbrana, C., Rastelli, V., Staiano, G., 2013. Applying an operating model for the environmental risk assessment in Italian Sites of Community Importance (SCI) of the European Commission Habitats Directive (92/43/EEC). Bull. Insectol. 66, 257–267.
- Lenth, R., 2018. Package 'emmeans'. Version 1.4.4. 10.1080/ 00031305.1980.10483031>.License.
- Lewandowski, A.S., Noss, R.F., Parsons, D.R., 2010. The effectiveness of surrogate taxa for the representation of biodiversity. Conserv. Biol. 24, 1367–1377. https://doi. org/10.1111/j.1523-1739.2010.01513.x.
- Losey, J.E., Vaughan, M., 2006. The economic value of ecological services provided by insects. Bioscience 56, 311–323. https://doi.org/10.1641/0006-3568(2006)56[311: TEVOES]2.0.CO;2.
- Lövei, G.L., Magura, T., Tóthmérész, B., Ködöböcz, V., 2006. The influence of matrix and edges on species richness patterns of ground beetles (Coleoptera: Carabidae) in habitat islands. Glob. Ecol. Biogeogr. 15, 283–289. https://doi.org/10.1111/j.1466-822X.2005.00221.x.
- Lövei, G.L., Sunderland, K.D., 1996. Ecology and behavior of ground beetles. Annu. Rev. Entomol. 41, 231–256. https://doi.org/10.1146/annurev.en.41.010196.001311.
- Lovell, S., Hamer, M., Slotow, R., Herbert, D., 2007. Assessment of congruency across invertebrate taxa and taxonomic levels to identify potential surrogates. Biol. Conserv. 139, 113–125. https://doi.org/10.1016/j.biocon.2007.06.008.
- Maas, B., Brandl, M., Hussain, R.I., Frank, T., Zulka, K.P., Rabl, D., Walcher, R., Moser, D., 2021. Functional traits driving pollinator and predator responses to newly established grassland strips in agricultural landscapes. J. Appl. Ecol. 58, 1728–1737. https://doi.org/10.1111/1365-2664.13892.
- Mace, G.M., Norris, K., Fitter, A.H., 2012. Biodiversity and ecosystem services: A multilayered relationship. Trends Ecol. Evol. 27, 19–26. https://doi.org/10.1016/j. tree.2011.08.006.
- Magagnoli, S., Lanzoni, A., Masetti, A., Depalo, L., Albertini, M., Ferrari, R., Spadola, G., Degola, F., Restivo, F.M., Burgio, G., 2021. Sustainability of strategies for Ostrinia nubilalis management in Northern Italy: Potential impact on beneficial arthropods and aflatoxin contamination in years with different meteorological conditions. Crop Prot. 142, 105529 https://doi.org/10.1016/j.cropro.2020.105529.
- Magura, T., Lövei, G.L., 2021. Consequences of urban living: urbanization and ground beetles. Curr. Landsc. Ecol. Reports 6, 9–21. https://doi.org/10.1007/s40823-020-00060-x.

- Magura, T., Tóthmérész, B., Lövei, G.L., 2006. Body size inequality of carabids along an urbanisation gradient. Basic Appl. Ecol. 7, 472–482. https://doi.org/10.1016/j. baae.2005.08.005.
- Magurran, A.E., 1988. Ecological Diversity and its Measurements. Princeton University Press, Princeton.
- Mandelik, Y., Roll, U., Fleischer, A., 2010. Cost-efficiency of biodiversity indicators for Mediterranean ecosystems and the effects of socio-economic factors. J. Appl. Ecol. 47, 1179–1188. https://doi.org/10.1111/j.1365-2664.2010.01864.x.
- Mantel, N., 1967. The detection of disease clustering and a generalized regression approach. Cancer Res. 27, 209–220.
- Marshall, E.J.P., Moonen, A.C., 2002. Field margins in northern Europe: Their functions and interactions with agriculture. Agric. Ecosyst. Environ. 89, 5–21. https://doi.org/ 10.1016/S0167-8809(01)00315-2.

Mazzia, C., Cornic, J., Capowiez, Y., Bounias-delacour, A., 2018. Zelotes metellus Roewer, 1928, nouvelle espèce pour la faune de France. Rev. Arachnol. 2, 14–17.

- McArt, S.H., Cook-Patton, S.C., Thaler, J.S., 2012. Relationships between arthropod richness, evenness, and diversity are altered by complementarity among plant genotypes. Oecologia 168, 1013–1021. https://doi.org/10.1007/s00442-011-2150-6
- McGill, B.J., Etienne, R.S., Gray, J.S., Alonso, D., Anderson, M.J., Benecha, H.K., Dornelas, M., Enquist, B.J., Green, J.L., He, F., Hurlbert, A.H., Magurran, A.E., Marquet, P.A., Maurer, B.A., Ostling, A., Soykan, C.U., Ugland, K.I., White, E.P., 2007. Species abundance distributions: Moving beyond single prediction theories to integration within an ecological framework. Ecol. Lett. 10, 995–1015. https://doi. org/10.1111/j.1461-0248.2007.01094.x.
- McQuatters-Gollop, A., Mitchell, I., Vina-Herbon, C., Bedford, J., Addison, P.F.E., Lynam, C.P., Geetha, P.N., Vermeulan, E.A., Smit, K., Bayley, D.T.I., Morris-Webb, E., Niner, H.J., Otto, S.A., 2019. From science to evidence - how biodiversity indicators can be used for effective marine conservation policy and management. Front. Mar. Sci. 6, 1–16. https://doi.org/10.3389/fmars.2019.00109.
- Michalko, R., Pekár, S., Dul'a, M., Entling, M.H., 2019a. Global patterns in the biocontrol efficacy of spiders: A meta-analysis. Glob. Ecol. Biogeogr. 28, 1366–1378. https:// doi.org/10.1111/geb.12927.
- Michalko, R., Pekár, S., Entling, M.H., 2019b. An updated perspective on spiders as generalist predators in biological control. Oecologia 189, 21–36. https://doi.org/ 10.1007/s00442-018-4313-1.
- Monzó, C., Sabater-Muñoz, B., Urbaneja, A., Castañera, P., 2011. The ground beetle Pseudophonus rufipes revealed as predator of Ceratitis capitata in citrus orchards. Biol. Control 56, 17–21. https://doi.org/10.1016/j.biocontrol.2010.09.004.
- Moonen, A.C., Bàrberi, P., 2008. Functional biodiversity: An agroecosystem approach. Agric. Ecosyst. Environ. 127, 7–21. https://doi.org/10.1016/j.agee.2008.02.013.
- Nardi, D., Lami, F., Pantini, P., Marini, L., 2019. Using species-habitat networks to inform agricultural landscape management for spiders. Biol. Conserv. 239, 108275 https:// doi.org/10.1016/j.biocon.2019.108275.
- Netwig, W., Blick, T., Gloor, D., Hänggi, A., Kropf, C., 2023. Spiders of Europe. 10.24436/1.
- Niemelä, J., 1995. Preface: From systematics to conservation carabidologists do it all. Ann. Zool. Fennici 33, 1–4.
- Niemelä, J., 2000. Biodiversity monitoring for decision-making. Ann. Zool. Fennici 37, 307–317.
- Norris, S.L., Blackshaw, R.P., Dunn, R.M., Critchley, N.R., Smith, K.E., Williams, J.R., Randall, N.P., Murray, P.J., 2016. Improving above and below-ground arthropod biodiversity in maize cultivation systems. Appl. Soil Ecol. 108, 25–46. https://doi. org/10.1016/j.apsoil.2016.07.015.
- Nourmohammadpour-amiri, M., Shayanmehr, M., Amiri-besheli, B., 2022. Influence of ground beetles (Carabidae) as biological agent to control of the Mediterranean fruit fly pupae, Ceratitis capitata, in Iranian citrus orchards. J. Asia. Pac. Entomol. 25, 101986 https://doi.org/10.1016/j.aspen.2022.101986.

Oberprieler, S.K., Andersen, A.N., Yeates, D.K., 2020. Selecting complementary target taxa for representing terrestrial invertebrate diversity in the Australian seasonal tropics. Ecol. Indic. 109, 105836 https://doi.org/10.1016/j.ecolind.2019.105836.

- Pearce, J.L., Venier, L.A., 2006. The use of ground beetles (Coleoptera: Carabidae) and spiders (Araneae) as bioindicators of sustainable forest management: A review. Ecol. Indic. 6, 780–793. https://doi.org/10.1016/j.ecolind.2005.03.005.
- Pesarini, C., Monzini, V., 2010. Insetti della fauna italiana: Coleotteri Carabidi I. Società Italiana di Scienze Naturali e Museo Civico di Storia Naturale, Milano.
- Pesarini, C., Monzini, V., 2011. Insetti della fauna italiana: Coleotteri Carabidi II. Società Italiana di Scienze Naturali e Museo Civico di Storia Naturale, Milano.
- Piano, E., Souffreau, C., Merckx, T., Baardsen, L.F., Backeljau, T., Bonte, D., Brans, K.I., Cours, M., Dahirel, M., Debortoli, N., Decaestecker, E., De Wolf, K., Engelen, J.M.T., Fontaneto, D., Gianuca, A.T., Govaert, L., Hanashiro, F.T.T., Higuti, J., Lens, L., Martens, K., Matheve, H., Matthysen, E., Pinseel, E., Sablon, R., Schön, I., Stoks, R., Van Doninck, K., Van Dyck, H., Vanormelingen, P., Van Wichelen, J., Vyverman, W., De Meester, L., Hendrickx, F., 2020. Urbanization drives cross-taxon declines in abundance and diversity at multiple spatial scales. Glob. Chang. Biol. 26, 1196–1211. https://doi.org/10.1111/gcb.14934.
- Pitzalis, M., Luiselli, L., Bologna, M.A., 2010. Co-occurrence analyses show that nonrandom community structure is disrupted by fire in two groups of soil arthropods (Isopoda Oniscidea and Collembola). Acta Oecol. 36, 100–106. https://doi.org/ 10.1016/j.actao.2009.10.009.
- Pizzolotto, R., Mazzei, A., Bonacci, T., Scalercio, S., Iannotta, N., Brandmayr, P., 2018. Ground beetles in mediterranean olive agroecosystems: Their significance and functional role as bioindicators (Coleoptera, Carabidae). PLoS One 13, 1–18. https:// doi.org/10.1371/journal.pone.0194551.
- Plowman, N.S., Mottl, O., Novotny, V., Idigel, C., Philip, F.J., Rimandai, M., Klimes, P., 2020. Nest microhabitats and tree size mediate shifts in ant community structure

F. Lami et al.

across elevation in tropical rainforest canopies. Ecography (Cop.) 43, 431–442. https://doi.org/10.1111/ecog.04730.

Ponisio, L.C., M'Gonigle, L.K., Kremen, C., 2016. On-farm habitat restoration counters biotic homogenization in intensively managed agriculture. Glob. Chang. Biol. 22, 704–715. https://doi.org/10.1111/gcb.13117.

R Core Team, 2016. R: A Language and Environment for Statistical Computing. R Found. Stat. Comput. Vienna, Austria. 10.1017/CBO9781107415324.004.

Rainio, J., Niemelä, J.K., 2003. Ground beetles (Coleoptera : Carabidae) as bioindicators. Biodivers. Conserv. 12, 487–506. https://doi.org/10.1023/A:1022412617568.

Renkema, J.M., Lynch, D.H., Cutler, G.C., MacKenzie, K., Walde, S.J., 2012. Predation by Pterostichus melanarius (Illiger) (Coleoptera: Carabidae) on immature Rhagoletis mendax Curran (Diptera: Tephritidae) in semi-field and field conditions. Biol. Control 60, 46–53. https://doi.org/10.1016/j.biocontrol.2011.10.004.

Rossi, E., Antichi, D., Loni, A., Canovai, R., Sbrana, M., Mazzoncini, M., 2019. Ground beetle (Coleoptera: Carabidae) assemblages and slug abundance in agricultural fields under organic and low-input conventional management within a long-term agronomic trial in Central Italy. Environ. Entomol. 48, 1377–1387. https://doi.org/ 10.1093/ee/nv2119.

Rusch, A., Binet, D., Delbac, L., Thiéry, D., 2016. Local and landscape effects of agricultural intensification on Carabid community structure and weed seed predation in a perennial cropping system. Landsc. Ecol. 31, 2163–2174. https://doi. org/10.1007/s10980-016-0390-x.

Samways, M.J., 2015. Future-proofing insect diversity. Curr. Opin. Insect Sci. 12, 71–78. https://doi.org/10.1016/j.cois.2015.09.008.

Santos, J.C., de Almeida, W.R., Fernandes, G.W., 2021. Arthropods: Why it is so crucial to know their biodiversity?. In: Measuring Arthropod Biodiversity. Springer International Publishing, Cham, pp. 3–11. https://doi.org/10.1007/978-3-030-53226-0_1.

Schalkwyk, J., Pryke, J.S., Samways, M.J., Gaigher, R., 2019. Congruence between arthropod and plant diversity in a biodiversity hotspot largely driven by underlying abiotic factors. Ecol. Appl. 29, e01883.

Schmidt, M.H., Thies, C., Nentwig, W., Tscharntke, T., 2008. Contrasting responses of arable spiders to the landscape matrix at different spatial scales. J. Biogeogr. 35, 157–166. https://doi.org/10.1111/j.1365-2699.2007.01774.x.

Schoeman, C.S., Hahn, N., Hamer, M., Foord, S.H., 2020. Regional invertebrate crossand within-taxon surrogacy are scale and taxon dependent. Trans. R. Soc. South Africa 75, 23–32. https://doi.org/10.1080/0035919X.2019.1658656.

Sfenthourakis, S., Tzanatos, E., Giokas, S., 2006. Species co-occurrence: The case of congeneric species and a causal approach to patterns of species association. Glob. Ecol. Biogeogr. 15, 39–49. https://doi.org/10.1111/j.1466-822X.2005.00192.x.

Shearin, A.F., Reberg-horton, S.C., Gallandt, E.R., 2008. Cover crop effects on the activity-density of the weed seed predator Harpalus rufipes (Coleoptera : Carabidae). Weed Sci. 56, 442–450.

Stephenson, P.J., Stengel, C., 2020. An inventory of biodiversity data sources for conservation monitoring. PLoS One 15, 1–14. https://doi.org/10.1371/journal. pone.0242923.

Storch, D., Bohdalková, E., Okie, J., 2018. The more-individuals hypothesis revisited: the role of community abundance in species richness regulation and the productivitydiversity relationship. Ecol. Lett. 21, 920–937. https://doi.org/10.1111/ele.12941.

- Swinton, S.M., Lupi, F., Robertson, G.P., Hamilton, S.K., 2007. Ecosystem services and agriculture: Cultivating agricultural ecosystems for diverse benefits. Ecol. Econ. 64, 245–252. https://doi.org/10.1016/j.ecolecon.2007.09.020.
- Teresa, F.B., Casatti, L., 2017. Trait-based metrics as bioindicators: Responses of stream fish assemblages to a gradient of environmental degradation. Ecol. Indic. 75, 249–258. https://doi.org/10.1016/j.ecolind.2016.12.041.

Trautner, J., Geigenmueller, K., 1987. Tiger Beetles, Ground Beetles: Illustrated Key to the Cicindelidae and Carabidae of Europe. J. Margraf Publishing, Aichtal.

Triquet, C., Roume, A., Tolon, V., Wezel, A., Ferrer, A., 2022. Undestroyed winter cover crop strip in maize fields supports ground-dwelling arthropods and predation. Agric. Ecosyst. Environ. 326 https://doi.org/10.1016/j.agee.2021.107783.

Tsafack, N., Wang, X., Xie, Y., Fattorini, S., 2021. Niche overlap and species cooccurrence patterns in carabid communities of the northern Chinese steppes. Zookeys 2021, 929–949. https://doi.org/10.3897/zookeys.1044.62478.

Uboni, C., Tordoni, E., Brandmayr, P., Battistella, S., Bragato, G., Castello, M., Colombetta, G., Poldini, L., Bacaro, G., 2019. Exploring cross-taxon congruence between carabid beetles (Coleoptera: Carabidae) and vascular plants in sites invaded by Ailanthus altissima versus non-invaded sites: The explicative power of biotic and abiotic factors. Ecol. Indic. 103, 145–155. https://doi.org/10.1016/j. ecolind.2019.03.052.

Ulrich, W., Zalewski, M., Hajdamowicz, I., Stańska, M., Ciurzycki, W., Tykarski, P., 2010. Tourism disassembles patterns of co-occurrence and weakens responses to environmental conditions of spider communities on small lake islands. Community Ecol. 11, 5–12. https://doi.org/10.1556/ComEc.11.2010.1.2.

Valdecasas, A.G., Camacho, A.I., 2003. Conservation to the rescue of taxonomy. Biodivers. Conserv. 12, 1113–1117.

Veech, J.A., 2013. A probabilistic model for analysing species co-occurrence. Glob. Ecol. Biogeogr. 22, 252–260. https://doi.org/10.1111/j.1466-8238.2012.00789.x.

- Vician, V., Svitok, M., Kočík, K., Stašiov, S., 2015. The influence of agricultural management on the structure of ground beetle (Coleoptera: Carabidae) assemblages. Biol. 70, 240–251. https://doi.org/10.1515/biolog-2015-0028.
- von Redwitz, C., Glemnitz, M., Hoffmann, J., Brose, R., Verch, G., Barkusky, D., Saure, C., Berger, G., Bellingrath-Kimura, S., 2019. Microsegregation in maize cropping—a chance to improve farmland biodiversity. Gesunde Pflanz. 71, 87–102. https://doi. org/10.1007/s10343-019-00457-7.

Westgate, M.J., Tulloch, A.I.T., Barton, P.S., Pierson, J.C., Lindenmayer, D.B., 2017. Optimal taxonomic groups for biodiversity assessment: a meta-analytic approach. Ecography (Cop.) 40, 539–548. https://doi.org/10.1111/ecog.02318.

Winfree, R., Fox, J.W., Williams, N.M., Reilly, J.R., Cariveau, D.P., 2015. Abundance of common species, not species richness, drives delivery of a real-world ecosystem service. Ecol. Lett. 18, 626–635. https://doi.org/10.1111/ele.12424.

Yong, D.L., Barton, P.S., Okada, S., Crane, M., Cunningham, S.A., Lindenmayer, D.B., 2020. Conserving focal insect groups in woodland remnants: The role of landscape context and habitat structure on cross-taxonomic congruence. Ecol. Indic. 115, 106391 https://doi.org/10.1016/j.ecolind.2020.106391.

Zara, L., Tordoni, E., Castro-Delgado, S., Colla, A., Maccherini, S., Marignani, M., Panepinto, F., Trittoni, M., Bacaro, G., 2021. Cross-taxon relationships in Mediterranean urban ecosystem: A case study from the city of Trieste. Ecol. Indic. 125, 107538 https://doi.org/10.1016/j.ecolind.2021.107538.