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# Differential metabolic responses in bold and shy sea anemones during a simulated heatwave

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## Summary Statement

Plastic changes in metabolic rates are associated with boldness in beadlet anemones, such that different personalities show opposite metabolic patterns under a simulated heatwave, as compared to a non-stressful temperature.

## Abstract

As climate change-induced heatwaves become more common, phenotypic plasticity at multiple levels is a key mitigation strategy by which organisms can optimise selective outcomes. In ectotherms, changes to both metabolism and behaviour can help alleviate thermal stress. Nonetheless, no study in any ectotherm has yet empirically investigated how changing temperatures affect among-individual differences in the associations between these traits. Using the beadlet anemone (*Actinia equina*), an intertidal species from a thermally heterogeneous environment, we investigated how individual metabolic rates, linked to morphotypic differences in *A. equina*, and boldness were related across changing temperatures. A crossed-over design and a temporal control was used to test the same individuals at a non-stressful temperature, 13°C, and under a simulated heatwave at 21°C. At each temperature, short-term repeated measurements of routine metabolic rate (RMR) and a single measurement of a repeatable

boldness-related behaviour, immersion response-time (IRT), were made. Individual differences, but not morphotypic differences, were highly predictive of metabolic plasticity, and the plasticity of RMR was associated with IRT. At 13°C, shy animals had the highest metabolic rates, while at 21°C this relationship was reversed. Individuals that were bold at 13°C also exhibited the highest metabolic rates at 21°C. Additional metabolic challenges during heatwaves could be detrimental to fitness in bold individuals. Equally, lower metabolic rates at non-stressful temperatures could be necessary for optimal survival as heatwaves become more common. These results provide novel insight into the relationship between metabolic and behavioural plasticity, and its adaptive implications in a changing climate.

## Introduction

The pace-of-life syndrome hypothesis (POLS; Réale et al., 2010) posits that organisms must trade-off long-term survival against short term reproductive success, leading to predictable correlations between risk-related or metabolically costly behaviour and energetic physiology. POLS predicts that individuals with more risk-prone, “bolder” personalities (consistent behavioural differences among individuals; Sih et al., 2004), should show “faster” physiological characteristics, such as a high metabolic rate, that are conducive to higher early fecundity and growth, but not to long-term survival (Montiglio et al., 2018; Réale et al., 2010). It is well established that life-history trade-offs can drive the maintenance of personality-types (Wolf & McNamara, 2012; Wolf & Weissing, 2010); bold individuals often place themselves at greater risk of mortality than shy conspecifics, but benefit from riskier lifestyles via increased foraging opportunities and more energetic scope for reproduction (Smith & Blumstein, 2008). POLS simply extends these trade-offs by predicting that the physiology of bolder individuals should be primed to maintain high levels of risky activity, exploration, and foraging, even under periods of heightened environmental stress (Montiglio et al., 2018; Réale et al., 2010). As such, according to POLS, bolder animals should exhibit reduced physiological stress reactivity, faster growth, and higher metabolic rates than shyer conspecifics (Biro & Stamps, 2010). The relationships proposed by POLS could be of great value to conservationists, indicating the life-history strategies present in a given population and informing whether some behavioural phenotypes might be more susceptible to novel selective pressures than others.

Evidence for POLS is mixed, with studies finding positive (Polverino et al., 2018), negative (Le Galliard et al., 2013), and inconclusive associations (Killen et al., 2012) between metabolic rate and boldness. As such, a recent meta-analysis concluded that empirical evidence for POLS is weak (Royauté et al., 2018). However, no studies have yet investigated how individual variation in the

scope and nature of phenotypic plasticity to environmental fluctuation might affect relationships between bioenergetics and boldness (Killen et al., 2013; Montiglio et al., 2018). “Activational” phenotypic plasticity (Snell-Rood, 2013), an animal’s ability to rapidly alter aspects of its phenotype in response to acute environmental change (Seebacher et al., 2015; Stamps, 2016), could be crucial to understanding these relationships under a rapidly changing climate, where extreme weather events are becoming increasingly common (IPCC, 2013, 2018). Activational plasticity can have selective advantages, helping many species remain robust to fluctuating environments (Snell-Rood, 2013), and disadvantages, as maintaining scope for short-term plasticity is energetically costly (Dall et al., 2004). This leads to variation within populations in the abilities of different individuals to deal with acute environmental changes (Dingemanse & Wolf, 2013) and could thus influence relationships between metabolism and boldness, and their selective implications under different environmental scenarios.

Although activational plasticity does not itself form part of POLS (Montiglio et al., 2018), the relationship between metabolism and behaviour is likely to be both species and context-dependent (Killen et al., 2012). In mosquitofish, for example, lineages from more stochastic environments grow faster, but also exhibit lower metabolic rates and are shyer than those from stable environments (Polverino et al., 2018). Among-individual variation in plasticity may be an especially important driver of relationships between behaviour and metabolism in heterogeneous environments, where the coexistence of different plastic strategies can be driven by spatial and temporal variation in both normal environmental conditions and how much those conditions vary (Dingemanse & Wolf, 2013; Wolf & McNamara, 2012). Despite the clear need to incorporate activational plasticity into investigations of POLS and its selective implications, we currently do not have any empirical information on how short-term environmental changes, and the overall variability of an animal’s environment, might affect the relationships between metabolism and behaviour.

When investigating how acute environmental changes affect these relationships, temperature is a particularly important environmental variable to consider. As climate change leads to higher average temperatures and increased frequency of heatwaves (IPCC, 2013, 2018), animal populations are being placed under substantial, novel selective pressures (Parmesan, 2006). Thermal physiology and thermal preferences are especially important drivers of associations between metabolism and behaviour in ectotherms (Abram et al., 2017), where many physiological processes, including metabolic rate, are intrinsically linked to environmental temperatures (Seebacher et al., 2015). In ants, for example, colonies from warmer environments are more likely to show the positive associations among energetically costly or risk-related behaviours that are predicted by POLS (Segev et al., 2017). Adult parasitic wasps with a preference for low temperatures are another example,

exhibiting high metabolic rates, low foraging efficiency, and short lifespans when exposed to high temperatures (Le Lann et al., 2011). As an acute thermal stressor to which they are unable to immediately physiologically adjust, climate change-induced heatwaves present ectotherms with a different set of ecological challenges than chronic temperature change (Vajedsamiei et al., 2021), and many species rely heavily on behavioural plasticity to address these challenges (Abram et al., 2017). Nonetheless, the effect of acute temperature rises on the relationship between ectothermic physiology and behaviour has yet to be investigated.

The beadlet anemone, *Actinia equina*, lives across a gradient of shore-heights in the spatially and temporally heterogeneous intertidal zone (Allcock et al., 1998). Although *A. equina* is particularly robust to temperature stress (Griffiths, 1977), making it an ideal organism with which to investigate the effects of thermal perturbation without risk of mortality, the heterogeneous nature of its environment leads it to show ranges of thermal preferences within single populations (Navarro et al., 1981). Anemones living further from the low-tide line (higher up the shore), experience increased exposure to stressful temperatures (Brahim et al., 2019) and greater thermal variation than those living lower down the shore (Chappon et al., 2016). Multiple genetically distinct morphotypes have developed which favour different shore-heights; the red morphotype is associated with the high-shore, and is thus predicted to deal more effectively with temperature extremes than the green morphotype, which is associated with the low-shore (Allcock et al., 1998). Morphotypes further display consistent differences in boldness-related behaviours. Of particular relevance could be immersion response-time (IRT), which not only displays among-individual variation within and across environments, but may be closely associated with an anemone's metabolic rate (Maskrey et al., 2020, 2021). IRT is a risk-related behaviour, which measures how long after a simulated tide-cycle it takes an anemone to re-extend its feeding tentacles and resume foraging. Because *A. equina* is unable to feed when the tide is out, more protracted tentacle retraction upon immersion will lead to an animal having less time to forage and thus less energy for metabolic processes. Further, tentacle extension in *A. equina* is inherently linked to respiratory efficiency and gas exchange (although anemones do continue to respire with their tentacles retracted or when exposed; Griffiths, 1977; Navarro et al., 1981), so spending longer with tentacles retracted can be highly costly in the face of metabolic challenges (Griffiths, 1977). Conversely, while its primary function is to support essential metabolic processes and gas exchange, tentacle extension is also risky, placing anemones under increased threat of predation (Edmunds et al., 1974, 1976). Given that anemones are still able to respire with their tentacles retracted (Griffiths, 1977; Navarro et al., 1981), and thus tentacle extension is not perfectly correlated with RMR, tentacle extension behaviour should represent a trade-off. If an individual is more cautious to predation risk it is likely

to face the costs of reduced efficiency of gas exchange and reduced foraging opportunities but may survive longer in the face of predators. As such general tentacle extension behaviour, and in turn IRT, should act as a useful gauge of boldness. This trade-off should be particularly pronounced at high temperatures, which drive increases in metabolic demand (Abram et al., 2017) but are also likely to place anemones under greater risk of predation from other ectotherms (e.g. Twardochleb et al., 2020; Yamane & Gilman, 2009). Relationships between IRT and metabolic rate at different temperatures should thus provide clear indications of whether *A. equina* follows POLS and enable prediction of the metabolic responses of different individuals to thermal extremes.

In this study, we investigated for the first time how relationships between behavioural phenotypes and metabolic rates were influenced by different temperatures, both within and among individuals. We measured the associations between the routine metabolic rates (RMR; Metcalfe et al., 2016; Velasque & Briffa, 2016), which is the metabolic rate of an animal undergoing normal activity, of two morphotypes of *A. equina* and their IRTs at two temperatures using a crossed-over temperature design, where one temperature was non-stressful and the other simulated a heatwave. We also incorporated a temporal control where the temperature remained non-stressful throughout. Importantly, all experimental individuals had their IRTs and RMRs measured at both temperatures, allowing us to use Bayesian multivariate methodologies to investigate associations between IRT and RMR both within and across environments. We expected that RMR would show a general increase across all individuals at high temperatures and that bolder animals, exhibiting shorter IRTs, would have higher RMRs at both temperatures, as predicted by POLS. Since both metabolic demand and predation risk should be exacerbated at high temperatures, which should theoretically intensify any trade-off between the efficiency of gas exchange and foraging, and risk-taking, we further predicted that this relationship would be stronger at high temperatures. We also predicted that the degree of change in RMR between temperatures would vary at individual and morphotypic levels, with discrepancies being influenced by the red morphotype's tolerance for higher, more variable temperatures than the green.

## Materials and Methods

### ***Section 1: Experimental Schedule***

Data collection was carried out between January and March 2020 over the course of four, nine-day data collection blocks. Each block began with three days of anemone (*Actinia equina* Linnaeus) collection from rocky sea defences at the top of the beach at New Brighton, UK (lat:53.4400, long:-3.0565). Five animals were collected on each day, split between the red and

green morphotypes, such that three of one morphotype and two of the other were collected. This uneven split, necessary due to 15 being the maximum number of individuals that could be investigated per block in our respirometry apparatus, was randomised and evened out over the course of the study. Overall, 30 anemones of each morphotype were used. Collection and identification of morphotypes was carried out using previously described methods (Maskrey et al., 2020). Anemones were size-matched at the point of collection as far as was possible by measuring pedal disc diameters (PDD). Because the raw routine metabolic rate (RMR; Metcalfe et al., 2016; Velasque & Briffa, 2016) of larger individuals of all species is higher than that of smaller individuals (Glazier, 2005), and thus easier to accurately detect, only anemones with a PDD of over 20mm, the largest individuals in the population, were selected. Further, because pilot work found the measurement of wet-weight to be highly invasive and cause significant damage to anemones, and because the volume of *A. equina* is not static (Griffiths, 1977) so cannot be reliably measured, this size-matching was also used to minimise any effect of anemone volume on metabolic results.

Each group of five anemones was returned to the laboratory, again using previously described methods (Maskrey et al., 2020), and left to acclimate to their surroundings for 48 hours. Anemones were housed in a  $13 \pm 1^{\circ}\text{C}$  temperature-controlled room on a 12h:12h day-night cycle. Each anemone was placed in a 7x15cm plastic cup containing drainage holes and a rock to which the anemone could adhere. Cups were situated in one of two 80x45x40cm tanks filled to ~12cm with artificial seawater, at a salinity of  $34 \pm 1$  ppt (RO water with Tropic Marin Pro Reef Salt, Germany). One of these tanks, which housed 10 individuals, was maintained at the ambient temperature of  $13^{\circ}\text{C}$  ( $\pm 1^{\circ}\text{C}$ ), while the other, which housed 5 individuals, was maintained at  $21^{\circ}\text{C}$  ( $\pm 1^{\circ}\text{C}$ ).  $13^{\circ}\text{C}$  was chosen as a suitable non-stressful temperature because it is well within the normal range of ambient temperatures *A. equina* would experience on the shore at Llandudno (Maskrey et al., 2021).  $21^{\circ}\text{C}$ , meanwhile, falls near the thermal maximum for this species and is well above the normal temperature range experienced on the Llandudno seashore (Maskrey et al., 2020). As such, exposure to three consecutive days at this temperature was deemed an appropriate method by which to simulate the effects of an atmospheric heatwave on the intertidal zone (Perkins & Alexander, 2013). Tanks were situated in two flow-through systems, which each contained three tanks. Every tank contained anemones at least once and the order in which tanks were used was randomised across blocks. Across the four blocks, each system was maintained at each temperature twice, and the order of this was randomised. Tanks were retained at a single temperature within blocks. Water quality (salinity, pH, nitrate, nitrite, ammonia) and water temperature in holding tanks were regularly recorded and full water changes in each system were carried out before each block.

After 24 hours of the 48-hour acclimation period, anemones were fed *ad libitum* on defrosted *Artemia* (Monkfield Nutrition, Ely, UK). This gave anemones time to bind to substrate and limited variation among individuals in levels of satiation, whilst also allowing a period of 24 hours without food before metabolic experimentation commenced. After the full 48 hours, each group of five anemones was first subject to a ~5pm immersion response-time (IRT) measurement, using anemones' natural response to changes in tides to measure boldness. IRTs have previously been shown to be repeatable within contexts ( $r = 0.26-0.4$ ), and show significant among-individual variation in how they change across temperatures (Maskrey et al., 2020, 2021). Similarly, risk-related tentacle extension behaviour has been shown to exhibit both repeatability and among-individual variation in plasticity in *A. equina* and other anemone species (Briffa & Greenaway, 2011; Hensley et al., 2012; Rudin & Briffa, 2012). As such, although not taking repeated measurements of a trait can give rise to statistical concerns in some cases (Brommer, 2013), a single measurement of IRT was taken before the commencement of metabolic testing, due both to the weight of previous evidence indicating that variation in IRT is associated with among-individual differences, and logistical constraints. Cups containing anemones were removed from holding tanks and drained. Anemones were left emersed for half an hour, before being re-immersed in their holding tank. After re-immersion, IRT, defined as the length of time to re-extend feeding tentacles fully (Maskrey et al., 2020), was recorded by two GoPro Hero 4 (GoPro Inc., San Mateo, CA) cameras mounted directly above the tank, taking time-lapse photographs every 30 seconds. 50 minutes of footage was recorded, of which the 45 minutes immediately after each individual's immersion was used to determine IRT. The number of photos an anemone took to re-extend its tentacles was recorded by DKM (blinded to treatment) and converted into seconds. Anemones which showed no complete response within 45 minutes were given a maximum value of 2700 seconds (Maskrey et al., 2020).

After their evening IRT measurement, anemones were gently separated from their pebbles and transferred to one of six freshly sterilised (with bleach-water and rinsed with fresh water) 14x6cm 425ml glass intermittent-flow respirometry chambers sealed with polypropylene lids (IKEA, Älmhult, Sweden) and each containing a 2.5g magnetic stir bar with which anemones could not interact. One of these chambers was designated as the blank, and thus remained empty. For identification purposes, anemones of the same morphotype were grouped in adjacent chambers, but where the morphotypes were, and which chamber was designated as the blank, was randomised. Chambers were situated in a temperature-controlled respirometry tank (67x46x38cm) filled to a depth of 20cm with freshly made artificial seawater (salinity  $34 \pm 1$  ppt). The seawater was heated or chilled to the appropriate temperature ( $\pm 0.3^\circ\text{C}$ ) prior to anemone introduction. Chambers were also subject to blank measurements prior to the introduction of anemones to ensure that only



negligible background respiration was present; where background was detected, chambers were re-sterilised to ensure minimal microbial activity. After their introduction, anemones were left for 12 hours overnight to acclimate and attach to their chamber. We used automated intermittent-flow respirometry (Svendsen et al., 2016) to measure aquatic RMR, so this 12-hour period also served to acclimate anemones to the intermittent-flow cycle. Measurements were unable to be conducted throughout the night due to difficulties with having the magnetic stir-bars move continuously for 12 hours. The intermittent-flow cycle consisted of a three-minute water flush period and a 38-minute closed period where oxygen consumption was measured. This measurement period was shorter than those used in previous *A. equina* respirometry studies (Navarro et al., 1981), but with this volume of chamber, pilot investigations found it produced sufficient data with which to measure oxygen consumption, while allowing more repeated metabolic measurements to be taken each day. After this 12-hour overnight acclimation period, experimentation was carried out for four hours in the morning and four hours in the afternoon, with a brief (~10 mins) break period between the two where any small bubbles that had formed were removed from chambers. During trials, aquatic oxygen concentration within chambers never fell below 80% saturation. Four metabolic slopes of oxygen consumption were produced for each anemone at each time of day (for full descriptions and pictures of metabolic apparatus, see Supplementary Information 1.1; Killen et al., 2021).

At the end of their first day of testing, groups were transferred back to one of the two holding tanks and the whole cycle was repeated, beginning with another 48h acclimation period. To account for temporal variation and treatment order effects we utilised a crossed-over design (Briffa et al., 2013) and a temporal control (White & Briffa, 2017). Within blocks, each group of five anemones was designated to one of three treatment groups, which were subjected to repeated RMR measurements and a single IRT measurement at each temperature. Each treatment had an overall sample size of 10 red and 10 green individuals:

- The temporal control (L-L) group was housed and tested at 13°C throughout.
- The low temperature to high temperature group (L-H) was first housed and tested at 13°C, then at 21°C.
- The high temperature to low temperature group (H-L) was first housed and tested at 21°C, then at 13°C.

After the second round of metabolic testing, anemones were placed directly into a -20°C freezer in individually labelled plastic bags, to be stored for later drying and weighing. This method of euthanasia added an extra step compared with previous studies of *A. equina*, where anemones have been dried from live (e.g. Navarro et al., 1981; Rudin & Briffa, 2012). Figure 1 provides a visualisation of this schedule.

## ***Section 2: Morphological Measurement***

To measure the dry weight of each individual, frozen anemones were placed in a Carbolite (Hope, UK) CWF100 muffle furnace, maintained at 110°C for 90 hours. Dry anemones were then weighed using a Sartorius (Stonehouse, UK) R2000 balance scale, accurate to the nearest hundredth of a milligram.

## ***Section 3: Statistical analysis***

The question of how behavioural and other phenotypic traits covary with one-another within and across environments is multivariate by its nature. Multiple traits can be influenced by the same set of fixed effects, including environmental variation, while simultaneously covarying with one-another. While univariate modelling approaches facilitate investigating single traits (Nakagawa & Schielzeth, 2010), they are not appropriate for investigating the associations between different traits (Houslay & Wilson, 2017), giving rise to anti-conservative correlation estimates and spurious conclusions. Simple correlational analyses using raw, rather than modelled, data are also of limited value when investigating relationships between phenotypic traits as they do not account for the complexity of the relationships in question (Nakagawa & Schielzeth, 2010). Many researchers in behavioural ecology and quantitative genetics have thus turned to multivariate mixed effects modelling to better describe these multivariate questions (Cleasby et al., 2015; Houslay & Wilson, 2017; Mitchell et al., 2016). While it is possible to fit multivariate models using a frequentist framework, methods utilising free software or statistical packages lack flexibility (e.g. lme4; Bates et al., 2015), and relatively more flexible frequentist packages require paying a subscription (e.g. ASReml-r; Butler et al., 2018). By contrast, multiple free, flexible Bayesian packages and softwares exist for fitting mixed effects models across multiple response variables or across multiple levels of variance (e.g. MCMCglmm; Hadfield, 2009; JAGS; Plummer, 2003; Stan; Carpenter et al., 2017). By fitting Bayesian models over uninformative priors as we do in this study, study data drives modelled posterior distributions, meaning that the information and estimates contained in those distributions are functionally the same as those contained in frequentist models (Cleasby et al., 2015; Houslay & Wilson, 2017). Moreover, Bayesian methodologies are more robust to small sample sizes, or violations of the assumptions of frequentist modelling approaches (McNeish, 2016). Bayesian modelling has become standard practice when investigating associations between multiple phenotypic traits within or across environments (e.g. Briffa et al., 2013; Jolles et al., 2019; Maskrey et al., 2021; Mitchell et al., 2020), and current advancements in these approaches use predominantly Bayesian frameworks (Mitchell et al., 2016; Mitchell & Houslay, 2021; Prentice et al.,

2020). We utilise these methods, exploring associations between traits at multiple phenotypic levels (i.e. behavioural, as IRT, and physiological, as RMR), here.

### *Calculation of RMR*

Metabolic data were analysed using the respR package (Harianto et al., 2019) in R version 3.6.2 or later (R Core Team, 2020). A minimum of two slopes were extracted for every chamber for both morning and afternoon sets of readings, providing up to six measurements of oxygen consumption per chamber, per day. The first 41 minute cycles in both the morning and afternoon were always excluded, as preliminary data indicated that RMRs settled during the first cycle, such that they were different from subsequent slopes in a sampling repeat. After a wait period before recording of at least a minute to ensure only the linear portion of slopes were measured (Svendsen et al., 2016), the  $r^2$  of recorded oxygen consumption slopes for all anemones fell above a threshold of 0.9. For each repeated measure, slopes from blank chambers were used to calculate background-adjusted whole-organism oxygen consumption (RMR) by subtracting blank oxygen consumption values from the values of each experimental chamber. Raw blank-adjusted measurements were converted to provide final whole-organism short-term RMR estimates in milligrams of oxygen per hour ( $\text{mgO}_2/\text{h}$ ; Supplementary Information 1.2). Each of these slopes was treated as a repeated measure in subsequent analyses.

### *Bayesian Analyses*

All individual-level models were run within a Bayesian Markov Chain Monte-Carlo framework using the R package MCMCglmm (Hadfield, 2009). See Supplementary Information 1.3 for details of general model specification and convergence checks, additional analytical details, and discussion of a confirmatory analysis into the relationship between RMR and dry weight.

### *Repeatability and Plasticity of RMR*

Before investigating relationships between RMR and IRT across temperatures, we first sought to confirm the presence of among-individual in RMR and in its plasticity, a pattern previous research has already shown in IRT (Maskrey et al., 2020, 2021). We also sought to investigate the influence of morphotypic differences on RMR. To achieve this, a random slopes model was run (see Supplementary Information 1.3 for details of comparison with the equivalent random intercepts model). RMR (z-transformed) was the response variable and morphotype, temperature, treatment, dry weight (z-transformed), sampling occasion, sampling day (i.e. whether it was an individual's first or second day of metabolic testing), and data collection block (z-transformed) were all included as

fixed effects. Because data collection block was temporally directional, and to improve model fits and simplify outputs, it was included in this and subsequent models as a continuous variable. This was denoted by the number of days between the start of the first block and the start of the block in question (0, 12, 28, and 41 respectively). Z-transformation of continuous variables, which centres data around zero based on how many standard deviations away from the mean each data point falls, was carried out in this and subsequent models to improve model convergence by placing continuous variables on the same scale (Houslay & Wilson, 2017). Temperature was allowed to interact, separately, with morphotype and treatment-order. Previous iterations of univariate random regressions showed no significant interaction between temperature and dry weight in relation to RMR ( $p=0.111$ ). This interaction was thus removed from the models as it was not of direct interest, unlike those between temperature and morphotype or temperature and treatment-order. A random slope effect, accounting for individual variation and the interaction between individual identity and metabolic plasticity to temperature (IxE), was specified as the random effect. The inclusion of temporal control data in the model, alongside including sampling occasion as a fixed effect, allowed robust control for both treatment-order and time-related effects.

To investigate whether individual differences in metabolic rate were consistent across the two temperatures, adjusted cross-context repeatability was calculated. To test the explanatory importance of individual variation in determining metabolic plasticity to temperature, the temperature-driven individual\*environment (IxE), or random slope effect estimate, was extracted. Finally, to explore how an individual's metabolic rate at 13°C related to their metabolic plasticity to temperature, the correlation between individual intercepts at 13°C and their random slopes between 13°C and 21°C was calculated. Statistical significance was inferred for fixed effects and covariance terms where 95% credible intervals did not cross zero. For repeatability and random slope effect estimates, significance was inferred where 95% credible intervals were not close to zero and histograms of the term's posterior estimates were not pushed up against zero (Maskrey et al., 2020).

#### *The relationship between RMR and immersion response-time (IRT)*

We explored several approaches to investigate the relationship between RMR and IRT, how it might relate to POLS at different temperatures, and whether boldness at 13°C or 21°C could be used to predict RMR at 21°C, whilst accounting for censored IRT data. Initially, we used a bivariate analysis with the brms package in Stan, using a Gaussian distribution for rate measures and a Weibull distribution for the censored IRT data; however, the software was unable to fit the appropriate model. We were also unable to develop a model in MCMCglmm that allowed for both a

multi-variable response as well as censored observations. As such, we have used a Gaussian distribution for IRT without censoring; whilst there are problems associated with this approach, described below, current approaches struggle to deal with these data (as explored in Stamps et al., 2012), and the approach used had been previously justified (see Maskrey et al., 2021).

Two separate datasets were used. The '13°C dataset' included all values measured at 13°C, including 13°C values from both the control and crossed-over temperature treatments. The 'crossover dataset' included all values measured in the crossed-over temperature treatments at either temperature. A bivariate mixed effects model was run on the 13°C dataset to determine the covariance between the two traits at 13°C, with RMR and IRT (both z-transformed) set as response variables. For the crossover dataset, a multivariate model was run where the four response variables were z-transformed RMR and IRT at each temperature. For both models, morphotype, dry weight (z-transformed), treatment, sampling occasion and data collection block were all fixed effects, with individual identity included as a random effect. Residual (within-individual) covariance was not identifiable, both because IRT and RMR were not measured at the same time in any treatment, and because only one IRT measurement per individual was modelled within each temperature for crossed over treatments. As such, values within the residual variance-covariance matrices of each model were therefore fixed at 1. After the appropriate model checks had been carried out, among-individual correlations between IRT and RMR within each temperature and between temperatures, alongside associated 95% credible intervals, were extracted. Statistical significance was inferred where 95% credible intervals did not cross zero. This analysis was split into a bivariate and a multivariate component to allow the incorporation of the control data into the 13°C correlation estimate. Including control data in the bivariate model was especially important to account for temporal effects, as 13°C experimental data only covered either the first or second sampling day, depending on treatment. Conversely, while control data was not included in the multivariate model, the nature of the crossover dataset, incorporating both sampling days for both treatments, means that the multivariate model should still have been robust to both treatment-order and time-related effects.

### ***Ethical Note***

No licenses or ethical approval were required to carry out this study as this species is not currently protected under UK legislation. Nonetheless, anemones were treated humanely, temperatures to which anemones were exposed fell within their tolerance ranges (Maskrey et al., 2020), and no mortality occurred during the experimental periods.

## Results

### *Repeatability and Plasticity of RMR*

There was a general trend across almost all animals for increased RMR at 21°C such that sample mean RMR increased by over 50% from 0.12 mgO<sub>2</sub>/h (SD = 0.04) at 13°C to 0.19 mgO<sub>2</sub>/h (SD = 0.06) at 21°C. This translated into a significant modelled temperature effect on RMR, but the size of this effect was estimated with a large degree of uncertainty (estimate = 0.98, 95%CI = 0.05, 1.95). Both morphotype and dry weight were also significant predictors of RMR, and neither of these relationships were affected by temperature. The green morphotype exhibited significantly higher mean RMRs than the red morphotype at both temperatures (25% higher at 13°C and 21% higher at 21°C; estimate = -0.34, 95%CI = -0.59, -0.10), although figure 2 does show a wider range of RMR values for both morphotypes at 21°C. Heavier individuals exhibited significantly higher RMRs than lighter individuals (estimate = 0.41, 95%CI = 0.29, 0.53). The relationship between data collection block and RMR was also on the verge of significance (estimate = -0.12, 95%CI = -0.24, -0.01), such that animals in later blocks had lower RMRs than those in earlier blocks.

At the individual level, repeatability estimates indicated that RMR varied consistently among individuals and that some of this variation was retained across temperatures, such that the rank-order of individual RMRs at 21°C remained partially similar to the rank order at 13°C. As such, adjusted cross-context repeatability was low, but still significant ( $R_{adj} = 0.18$ , 95%CI = 0.11, 0.25). This result should be taken with some caution, given that while individuals did show variation in the magnitude of their metabolic slopes within temperatures, within-temperature measurements may have been subject to some level of temporal autocorrelation (Mitchell et al., 2020), leading to a lower cross-context repeatability estimate. Nonetheless, in these data, much of the among-individual variation in repeated RMR measurements was explained by IxE, translating into a highly significant, strong random slope effect (estimate = 0.70, 95%CI = 0.60, 0.80). Figure 3 shows that individuals differed greatly in the degree to which their short-term RMR increased as the temperature was raised. 2/40 exhibited lower RMRs at 21°C than at 13°C and 2/40 showed minimal change between the two temperatures. How the RMR of different individuals changed between temperatures was associated with their RMR at 13°C, with individuals that exhibited lower RMRs at 13°C showing larger increases at 21°C than those that exhibited higher RMRs at 13°C. This translated into a significant negative correlation between individual RMRs at 13°C and the gradient of individual random slopes between the two temperatures ( $r = -0.53$ , 95%CI = -0.70, -0.21).

### *The relationship between RMR and immersion response-time (IRT)*

RMR was correlated with IRT, and the nature of this correlation was related to environmental temperature such that it did not always follow the assumptions of POLS. At 13°C, shyer individuals (exhibiting longer IRTs) showed higher RMRs than bolder individuals, translating into a moderate, significant positive correlation between IRT and RMR at that temperature (Table 1; Figure 4a). At 21°C, the pattern of this correlation swapped, such that bolder individuals (with shorter IRTs) at either temperature exhibited higher RMRs at 21°C than individuals that were shyer at either temperature (Figures 4b, 4d). This correlation, while still of interest, was not significant between IRT and RMR both measured at 21°C (Table 1; Figure 4b). It was, however, at the bound of significance between IRT at 13°C and RMR at 21°C, as repeated runs of the model found the lower bound of the 95% credible interval to hover within 0.01 units either side of 0 (Table 1, which shows the results of one such run; Figure 4d). This indicates that individuals' IRTs at 13°C were moderately predictive of their RMRs at 21°C. It is worth noting that this significance may have been clearer but for the smaller effective sample size of the crossover dataset (n=40). There was no relationship between RMR at 13°C and IRT at 21°C (Table 1; Figure 4c).

### **Discussion**

As the climate warms and marine invertebrates experience extreme temperatures with increasing regularity (IPCC, 2013), the survival of individuals under the greatest metabolic demand at those temperatures may be placed in jeopardy (Montiglio et al., 2018). The pace-of-life syndrome hypothesis (POLS; Réale et al., 2010), which predicts positive correlations between boldness and metabolic rate, could be used to help indicate those individuals most vulnerable to warming, but how these correlations change when individuals are exposed to ecologically relevant heat stress has not been tested. In this study we found evidence in beadlet sea anemones for a complex association between routine metabolic rate (RMR) and immersion response-time (IRT) across temperatures, such that the correlations between the two did not follow the assumptions of POLS at a non-stressful temperature of 13°C but did during a simulated heatwave at 21°C. We further showed that individual differences explained most of the variance in how RMR changed between temperatures. If individuals of certain personality-types are behaviourally or physiologically more sensitive to temperature stress, using more energy or experiencing reduced foraging opportunities or increased predation risk, then more regular heatwaves could put those individuals at a selective disadvantage.

At a non-stressful temperature, shyer individuals exhibited higher RMRs, a result entirely at odds with POLS. One intuitive explanation for this could be that, at lower temperatures, shyer

individuals were investing more than bolder individuals in metabolically costly maintenance processes such as growth, as suggested by the allocation model of resource management (Biro et al., 2020; Careau et al., 2019). Although population-wide studies tend to find faster growth at warmer temperatures in marine invertebrates (reviewed in: Angilletta et al., 2004), this rule is not universal (Angilletta & Dunham, 2003; Biro et al., 2020; Careau et al., 2019). In *A. equina*, even very regular feeding regimens are not enough to counteract a loss of body-mass at high temperatures in laboratory environments (Chomsky et al., 2004b). At low temperatures, meanwhile, *A. equina* is able to grow even when fed only once or twice a week (Chomsky et al., 2004a). As such, while 24 hours of fasting should have eliminated the costs of digestion (Navarro et al., 1981), individuals likely still had flexibility in how they expended their energy at 13°C. This flexibility could have provided the scope for different personality-types to employ different resource allocation strategies and for RMR to be positively correlated with growth (Killen, 2014). Shy individuals may thus have prioritised growth to a greater degree than bold individuals at low temperatures, both because of the greater flexibility in resource allocation that those temperatures afforded, and because low risks of predation from ectothermic predators at those temperatures (e.g. Twardochleb et al., 2020; Yamane & Gilman, 2009) should have allowed them to forage more freely. Alternatively, the move from holding tanks to respirometry chambers, or the slight disturbance in the chambers caused by the intermittent-flow cycle itself, might have caused disproportionate stress-responses in shyer animals as compared with their bolder conspecifics, driving their higher RMRs at 13°C (e.g. Martins et al., 2011). In turn, shy individuals may already have been maximally stressed at that temperature and thus had little capacity to further increase their RMR further at higher temperatures. Future studies could investigate the relationships between growth-rate, dry weight and IRT to see whether shy individuals do grow faster than bold individuals at non-stressful temperatures, and what implications this has for selection.

In contrast to 13°C, individuals that were bolder at either temperature exhibited higher RMRs when under thermal stress than those that were shyer. One possible mechanistic underpinning for this relationship could have involved shy individuals suppressing their RMRs to a greater degree than bold individuals at stressful high temperatures, in preparation for cooler “recovery periods” in the future (Schulte et al., 2011). Temperature-driven adaptive metabolic suppression (either limiting the increase or even decreasing metabolic rates rapidly by curbing processes with high energy requirements; Schulte et al., 2011) appears to be specific to intertidal and shallow-water invertebrates, having been so far documented in gastropods (McMahon et al., 1995), molluscs (Vajedsamiei et al., 2021), and freshwater crustaceans (Glazier et al., 2020). Animals employing this strategy may be equipped during heatwaves to avoid their energetic demands



exceeding their available energy supply but, as heatwaves become more frequent, suppression strategies could place individuals at a selective disadvantage by creating energetic deficits which they are unable to fulfil during progressively shorter recovery periods (Pörtner, 2012). Although they were of varying personality-types, those individuals suppressing their RMRs most in our sample could have been the four that failed to increase their RMRs at all at 21°C (e.g. Vajedsamiei et al., 2021). It is also possible that these individuals had surpassed their metabolic thermal maximum (Schulte et al., 2011). Although the high RMRs of all four of these individuals at 13°C might suggest this is plausible, it would seem unlikely that any of them had surpassed this threshold given that lethal temperatures for other UK populations of *A. equina* are at least two degrees higher than the 21°C used in this experiment (Maskrey et al., 2020). Future studies could investigate molecular changes associated with metabolic suppression (Richier et al., 2008; Tomanek & Somero, 1999) to address these two possibilities directly.

The effects of individual and morphotypic variation on the plasticity of RMR could shed further light on its relationship with IRT. There was a general trend towards higher RMRs at 21°C, which remained unexpectedly consistent between the two morphotypes. Importantly, however, as predicted, this response was not ubiquitous, possibly indicating variation among individuals in strategies for dealing with extreme high temperatures. One potential explanation is that some individuals may compensate for increased RMR induced by higher temperatures by reducing activity. This could have led to the observed negative correlation between IRT and RMR at 21°C. After considering the available evidence, however, we believe that such a compensatory reduction in activity may be unlikely. Foraging tentacle expansion is the main energetically costly activity these anemones undertake (Griffiths, 1977) and IRT is thus not only an excellent proxy for boldness but should also act as a proxy for activity. As such, if a change in activity was driving changing RMR, IRT and RMR at 21°C should have been closely correlated; a pattern that is not borne out by the data presented here. Indeed, RMR at 21°C was more closely correlated with IRT at 13°C than at 21°C, which does not suggest a close association between an individual's current level of activity and its RMR. Previous research also shows differences in how morphotypes change their IRTs, and thus their activity, across temperatures (Maskrey et al., 2020). These discrepancies were, unexpectedly, not reflected in morphotypic patterns of RMR in these data, further indicating that changing activity was not driving changing RMR. Instead, it seems more likely that relationships between IRT and RMR were at least partly driven by boldness-associated physiological differences, either via varied growth strategies at non-stressful temperatures and RMR suppression at high temperatures as we speculate, or by other mechanisms.

The patterns elucidated here may have important selective implications for *A. equina* as the frequency of heatwaves continues to increase (IPCC, 2013, 2018). At the broader genotypic scale, the red morphotype, associated with the high-shore, may be at a selective advantage as predicted. Not only do individuals of the red morphotype, in contrast to their green counterparts, lower their IRTs (Maskrey et al., 2020, 2021) at high temperatures, which should provide better mitigation against metabolic challenges under higher energetic demand (Griffiths, 1977; Navarro et al., 1981), but these data show that they also exhibit lower overall RMRs than green individuals, potentially further accentuating this increased efficiency. At the level of individuals the directionality of any selective implications is less clear. It is possible that shy animals exhibit a more robust strategy than bold under current environmental conditions, up-regulating their metabolism when not under heat-stress to invest in increased body mass, and conserving energy when they are. Bold individuals, meanwhile, may be at greater risk of perishing during heatwaves, investing less in growth at lower temperatures and exposing themselves to greater metabolic requirements and increased risks of predation when under heat stress. Equally, shy animals could be at a selective disadvantage if unable to recoup any energetic deficit incurred during metabolically suppressed periods, which could become increasingly likely as the frequency of heatwaves increases and recovery periods become more infrequent. Future work should investigate these possibilities directly, and incorporate life-history characteristics such as survival or fecundity. This could not only provide clearer information as to the vulnerability of different *A. equina* individuals and genotypes to heatwaves, but also that of other species, as the heterogeneous nature of the seashore leads many intertidal ectotherms to display intraspecific variation in behavioural and physiological responses to stressful high temperatures (Briffa et al., 2013; Chapperon et al., 2016; Dias et al., 2018; Maskrey et al., 2021).

Our measures of IRT resulted in censored data, given that there was a maximum cut-off time for shy individuals. Such data are common among behavioural studies (e.g. Stamps et al., 2012), but nevertheless they complicated our analysis of the relationships between RMR and IRT. We attempted a variety of approaches but the ability of current statistical packages we used to deal with these sorts of data were limited (although future developments may provide ways forward). Though our final statistical design was able to produce appropriate models, the likes of which are utilised in behavioural studies across the literature, there are concerns over its ability to deal with censored values (Stamps et al., 2012). As such, these results may need to be considered with care but we are confident that they are reflective of the biology of these organisms.

This study shows that associations between RMR and boldness in *A. equina* are highly temperature-dependent, and that variation in trait plasticity could be part of the reason why evidence for POLS is so inconsistent. Individual differences were key drivers of how RMR changed

between temperatures in this study, while morphotypic differences remained consistent, suggesting changing associations between RMR and boldness were being driven by individual-level plasticity. By measuring the boldness and metabolism of animals living in heterogeneous environments under different contexts, and adding a life-history component, future work could draw a clearer picture of POLS, gain insights into how different populations might respond to more regular climate change-induced heatwaves, and facilitate a more comprehensive understanding of the adaptive value of different personality-types under climate change.

### **Author Contributions**

DKM, SSK, LUS, KEA, and JST formulated the study. Laboratory set-up and animal husbandry was carried out by DKM and DCCW. DKM carried out experimental work and analyses, with assistance from all authors. All authors contributed extensively to the manuscript.

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### **Competing Interests**

No competing interests declared

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### **Data Availability Statement**

Upon acceptance, data and analytical code will be made available on Figshare provided by DKM.

## References

- Abram, P. K., Boivin, G., Moiroux, J., & Brodeur, J. (2017). Behavioural effects of temperature on ectothermic animals: unifying thermal physiology and behavioural plasticity. *Biological Reviews*, *92*(4), 1859–1876. <https://doi.org/10.1111/brv.12312>
- Allcock, A. L., Watts, P. C., & Thorpe, J. P. (1998). Divergence of Nematocysts in Two Colour Morphs of the Intertidal Beadlet Anemone *Actinia Equina*. *Journal of the Marine Biological Association of the United Kingdom*, *78*(3), 821–828. <https://doi.org/10.1017/s0025315400044805>
- Angilletta, M. J., & Dunham, A. E. (2003). The Temperature-Size Rule in Ectotherms: Simple Evolutionary Explanations May Not Be General. *American Naturalist*, *162*(3), 332–342. <https://doi.org/10.1086/377187>
- Angilletta, M. J., Steury, T. D., & Sears, M. W. (2004). Temperature, growth rate, and body size in ectotherms: Fitting pieces of a life-history puzzle. *Integrative and Comparative Biology*, *44*(6), 498–509. <https://doi.org/10.1093/icb/44.6.498>
- Bates, D., Mächler, M., Bolker, B. M., & Walker, S. C. (2015). Fitting linear mixed-effects models using lme4. *Journal of Statistical Software*, *67*(1), 1–48. <https://doi.org/10.18637/jss.v067.i01>
- Biro, P. A., & Stamps, J. A. (2010). Do consistent individual differences in metabolic rate promote consistent individual differences in behavior? *Trends in Ecology and Evolution*, *25*(11), 653–659. <https://doi.org/10.1016/j.tree.2010.08.003>
- Biro, P. A., Thomas, F., Ujvari, B., Adriaenssens, B., & Beckmann, C. (2020). Spontaneous activity rates and resting metabolism: Support for the allocation model of energy management at the among-individual level. *Ethology*, *126*(1), 32–39. <https://doi.org/10.1111/eth.12957>

- Brahim, A., Mustapha, N., & Marshall, D. J. (2019). Non-reversible and reversible heat tolerance plasticity in tropical intertidal animals: Responding to habitat temperature heterogeneity. *Frontiers in Physiology*, *9*, 1909. <https://doi.org/10.3389/fphys.2018.01909>
- Briffa, M., Bridger, D., & Biro, P. A. (2013). How does temperature affect behaviour? Multilevel analysis of plasticity, personality and predictability in hermit crabs. *Animal Behaviour*, *86*(1), 47–54. <https://doi.org/10.1016/j.anbehav.2013.04.009>
- Briffa, M., & Greenaway, J. (2011). High in situ repeatability of behaviour indicates animal personality in the beadlet anemone *Actinia equina* (Cnidaria). *PLoS ONE*, *6*(7), e21963. <https://doi.org/10.1371/journal.pone.0021963>
- Brommer, J. E. (2013). On between-individual and residual (co)variances in the study of animal personality: Are you willing to take the “individual gambit”? *Behavioral Ecology and Sociobiology*, *67*(6), 1027–1032. <https://doi.org/10.1007/s00265-013-1527-4>
- Butler, D. G., Cullis, B. R., Gilmour, A. R., Gogel, B. J., & Thompson, R. (2018). ASReml-R Reference Manual Version 4. In *ASReml-R Reference Manual* (p. 176). <http://www.homepages.ed.ac.uk/iwhite/asreml/uop>.
- Careau, V., Beauchamp, P. P., Bouchard, S., & Morand-Ferron, J. (2019). Energy metabolism and personality in wild-caught fall field crickets. *Physiology and Behavior*, *199*(November 2018), 173–181. <https://doi.org/10.1016/j.physbeh.2018.11.023>
- Carpenter, B., Gelman, A., Hoffman, M. D., Lee, D., Goodrich, B., Betancourt, M., Brubaker, M. A., Guo, J., Li, P., & Riddell, A. (2017). Stan: A probabilistic programming language. *Journal of Statistical Software*, *76*(1). <https://doi.org/10.18637/jss.v076.i01>
- Chapperon, C., Volkenborn, N., Clavier, J., Séité, S., Seabra, R., & Lima, F. P. (2016). Exposure to solar radiation drives organismal vulnerability to climate: Evidence from an intertidal limpet. *Journal of Thermal Biology*, *57*, 92–100. <https://doi.org/10.1016/j.jtherbio.2016.03.002>

- Chomsky, O., Kamenir, Y., Hyams, M., Dubinsky, Z., & Chadwick-Furman, N. E. (2004a). Effects of feeding regime on growth rate in the Mediterranean Sea anemone *Actinia equina* (Linnaeus). *Journal of Experimental Marine Biology and Ecology*, *299*(2), 217–229. <https://doi.org/10.1016/j.jembe.2003.09.009>
- Chomsky, O., Kamenir, Y., Hyams, M., Dubinsky, Z., & Chadwick-Furman, N. E. (2004b). Effects of temperature on growth rate and body size in the Mediterranean Sea anemone *Actinia equina*. *Journal of Experimental Marine Biology and Ecology*, *313*(1), 63–73. <https://doi.org/10.1016/j.jembe.2004.07.017>
- Cleasby, I. R., Nakagawa, S., & Schielzeth, H. (2015). Quantifying the predictability of behaviour: Statistical approaches for the study of between-individual variation in the within-individual variance. *Methods in Ecology and Evolution*, *6*(1), 27–37. <https://doi.org/10.1111/2041-210X.12281>
- Dall, S. R. X., Houston, A. I., & McNamara, J. M. (2004). The behavioural ecology of personality: Consistent individual differences from an adaptive perspective. *Ecology Letters*, *7*(8), 734–739. <https://doi.org/10.1111/j.1461-0248.2004.00618.x>
- Dias, G. M., Christofoletti, R. A., Kitazawa, K., & Jenkins, S. R. (2018). Environmental heterogeneity at small spatial scales affects population and community dynamics on intertidal rocky shores of a threatened bay system. *Ocean and Coastal Management*, *164*, 52–59. <https://doi.org/10.1016/j.ocecoaman.2017.12.001>
- Dingemanse, N. J., & Wolf, M. (2013). Between-individual differences in behavioural plasticity within populations: Causes and consequences. *Animal Behaviour*, *85*(5), 1031–1039. <https://doi.org/10.1016/j.anbehav.2012.12.032>
- Edmunds, M., Potts, G. W., Swinfen, R. C., & Waters, V. L. (1974). The feeding preferences of *Aeolidia papillosa* (L.) (Mollusca, Nudibranchia). *Journal of the Marine Biological Association of the United Kingdom*, *54*(4), 939–947. <https://doi.org/10.1017/S0025315400057660>

- Edmunds, M., Potts, G. W., Swinfen, R. C., & Waters, V. L. (1976). Defensive behaviour of sea anemones in response to predation by the opisthobranch mollusc *Aeolidia papillosa* (L.). *Journal of the Marine Biological Association of the United Kingdom*, *56*(1), 65–83. <https://doi.org/10.1017/s0025315400020440>
- Glazier, D. S. (2005). Beyond the “3/4-power law”: Variation in the intra- and interspecific scaling of metabolic rate in animals. *Biological Reviews of the Cambridge Philosophical Society*, *80*(4), 611–662. <https://doi.org/10.1017/S1464793105006834>
- Glazier, D. S., Borrelli, J. J., & Hoffman, C. L. (2020). Effects of fish predators on the mass-related energetics of a keystone freshwater crustacean. *Biology*, *9*(3). <https://doi.org/10.3390/biology9030040>
- Griffiths, R. J. (1977). Thermal stress and the biology of *Actinia equina* L. (Anthozoa). *Journal of Experimental Marine Biology and Ecology*, *27*(2), 141–154. [https://doi.org/10.1016/0022-0981\(77\)90134-4](https://doi.org/10.1016/0022-0981(77)90134-4)
- Hadfield, J. D. (2009). MCMCgmm: Markov chain Monte Carlo methods for Generalised Linear Mixed Models. *Tutorial for MCMCgmm Package in R*.
- Hariato, J., Carey, N., & Byrne, M. (2019). respR—An R package for the manipulation and analysis of respirometry data. *Methods in Ecology and Evolution*, *10*(6), 912–920. <https://doi.org/10.1111/2041-210X.13162>
- Hensley, N. M., Cook, T. C., Lang, M., Petelle, M. B., & Blumstein, D. T. (2012). Personality and habitat segregation in giant sea anemones (*Condylactis gigantea*). *Journal of Experimental Marine Biology and Ecology*, *426–427*, 1–4. <https://doi.org/10.1016/j.jembe.2012.05.011>
- Houslay, T. M., & Wilson, A. J. (2017). Avoiding the misuse of BLUP in behavioural ecology. *Behavioral Ecology*, *28*(4), 948–952. <https://doi.org/10.1093/beheco/arx023>

- IPCC. (2013). Summary for Policymakers. In Intergovernmental Panel on Climate Change (Ed.), *Climate Change 2013 - The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*. Cambridge University Press. <https://doi.org/10.1017/CBO9781107415324>
- IPCC. (2018). Summary for Policymakers. In Intergovernmental Panel on Climate Change (Ed.), *Global warming of 1.5°C. An IPCC Special Report on the impacts of global warming of 1.5°C above pre-industrial levels*. Cambridge University Press. <https://apps.ipcc.ch/outreach/documents/451/1551801374.pdf>
- Jolles, J. W., Briggs, H. D., Araya-Ajoy, Y. G., & Boogert, N. J. (2019). Personality, plasticity and predictability in sticklebacks: bold fish are less plastic and more predictable than shy fish. *Animal Behaviour*, *154*, 193–202. <https://doi.org/10.1016/j.anbehav.2019.06.022>
- Killen, S. S. (2014). Growth trajectory influences temperature preference in fish through an effect on metabolic rate. *Journal of Animal Ecology*, *83*(6), 1513–1522. <https://doi.org/10.1111/1365-2656.12244>
- Killen, S. S., Christensen, E. A. F., Cortese, D., Závorka, L., Norin, T., Cotgrove, L., Crespel, A., Munson, A., Nati, J. J. H., Papatheodoulou, M., & McKenzie, D. J. (2021). Guidelines for reporting methods to estimate metabolic rates by aquatic intermittent-flow respirometry. *Journal of Experimental Biology*, *224*(18), jeb242522. <https://doi.org/10.1242/jeb.242522>
- Killen, S. S., Marras, S., Metcalfe, N. B., McKenzie, D. J., & Domenici, P. (2013). Environmental stressors alter relationships between physiology and behaviour. *Trends in Ecology and Evolution*, *28*(11), 651–658. <https://doi.org/10.1016/j.tree.2013.05.005>
- Killen, S. S., Marras, S., Ryan, M. R., Domenici, P., & McKenzie, D. J. (2012). A relationship between metabolic rate and risk-taking behaviour is revealed during hypoxia in juvenile European sea bass. *Functional Ecology*, *26*(1), 134–143. <https://doi.org/10.1111/j.1365-2435.2011.01920.x>



- Le Galliard, J. F., Paquet, M., Cisel, M., & Montes-Poloni, L. (2013). Personality and the pace-of-life syndrome: Variation and selection on exploration, metabolism and locomotor performances. *Functional Ecology*, *27*(1), 136–144. <https://doi.org/10.1111/1365-2435.12017>
- Le Lann, C., Wardziak, T., van Baaren, J., & van Alphen, J. J. M. (2011). Thermal plasticity of metabolic rates linked to life-history traits and foraging behaviour in a parasitic wasp. *Functional Ecology*, *25*(3), 641–651. <https://doi.org/10.1111/j.1365-2435.2010.01813.x>
- Martins, C. I. M., Castanheira, M. F., Engrola, S., Costas, B., & Conceição, L. E. C. (2011). Individual differences in metabolism predict coping styles in fish. *Applied Animal Behaviour Science*, *130*(3–4), 135–143. <https://doi.org/10.1016/j.applanim.2010.12.007>
- Maskrey, D. K., Sneddon, L. U., Arnold, K. E., Wolfenden, D. C. C., & Thomson, J. S. (2020). The impact of personality, morphotype and shore height on temperature-mediated behavioural responses in the beadlet anemone *Actinia equina*. *Journal of Animal Ecology*, *89*(10), 2311–2324. <https://doi.org/10.1111/1365-2656.13301>
- Maskrey, D. K., Sneddon, L. U., Arnold, K. E., Wolfenden, D. C. C., & Thomson, J. S. (2021). Temperature-driven changes in behavioural unpredictability and personality in the beadlet sea anemone, *Actinia equina*. *Animal Behaviour*, *181*, 13–27. <https://doi.org/10.1016/j.anbehav.2021.08.022>
- McMahon, R. F., Russell-Hunter, W. D., & Aldridge, D. W. (1995). Lack of metabolic temperature compensation in the intertidal gastropods, *Littorina saxatilis* (Olivi) and *L. obtusata* (L.). In *Advances in Littorinid Biology* (pp. 89–100). Springer Netherlands. [https://doi.org/10.1007/978-94-011-0435-7\\_9](https://doi.org/10.1007/978-94-011-0435-7_9)
- McNeish, D. (2016). On Using Bayesian Methods to Address Small Sample Problems. *Structural Equation Modeling*, *23*(5), 750–773. <https://doi.org/10.1080/10705511.2016.1186549>

- Metcalfe, N. B., Van Leeuwen, T. E., & Killen, S. S. (2016). Does individual variation in metabolic phenotype predict fish behaviour and performance? *Journal of Fish Biology*, *88*(1), 298–321. <https://doi.org/10.1111/jfb.12699>
- Mitchell, D. J., Dujon, A. M., Beckmann, C., & Biro, P. A. (2020). Temporal autocorrelation: A neglected factor in the study of behavioral repeatability and plasticity. *Behavioral Ecology*, *31*(1), 222–231. <https://doi.org/10.1093/beheco/arz180>
- Mitchell, D. J., Fanson, B. G., Beckmann, C., & Biro, P. A. (2016). Towards powerful experimental and statistical approaches to study intraindividual variability in labile traits. *Royal Society Open Science*, *3*(10). <https://doi.org/10.1098/rsos.160352>
- Mitchell, D. J., & Houslay, T. M. (2021). Context-dependent trait covariances: how plasticity shapes behavioral syndromes. *Behavioral Ecology*, *32*(1), 25–29. <https://doi.org/10.1093/beheco/araa115>
- Montiglio, P. O., Dammhahn, M., Dubuc Messier, G., & Réale, D. (2018). The pace-of-life syndrome revisited: the role of ecological conditions and natural history on the slow-fast continuum. *Behavioral Ecology and Sociobiology*, *72*(7), 1–9. <https://doi.org/10.1007/s00265-018-2526-2>
- Nakagawa, S., & Schielzeth, H. (2010). Repeatability for Gaussian and non-Gaussian data: A practical guide for biologists. In *Biological Reviews* (Vol. 85, Issue 4, pp. 935–956). <https://doi.org/10.1111/j.1469-185X.2010.00141.x>
- Navarro, E., Ortega, M. M., & Madariaga, J. M. (1981). Effect of body size, temperature and shore level on aquatic and aerial respiration of *Actinia equina* (L.) (Anthozoa). *Journal of Experimental Marine Biology and Ecology*, *53*(2–3), 153–162. [https://doi.org/10.1016/0022-0981\(81\)90016-2](https://doi.org/10.1016/0022-0981(81)90016-2)
- Parmesan, C. (2006). Ecological and Evolutionary Responses to Recent Climate Change. *Annual Review of Ecology, Evolution, and Systematics*, *37*(1), 637–669. <https://doi.org/10.1146/annurev.ecolsys.37.091305.110100>

- Perkins, S. E., & Alexander, L. V. (2013). On the measurement of heat waves. *Journal of Climate*, 26(13), 4500–4517. <https://doi.org/10.1175/JCLI-D-12-00383.1>
- Plummer, M. (2003). DSC 2003 Working Papers JAGS: A program for analysis of Bayesian graphical models using Gibbs sampling. *DSC 2003 Working Papers*. <http://www.ci.tuwien.ac.at/Conferences/DSC-2003/>
- Polverino, G., Santostefano, F., Díaz-Gil, C., & Mehner, T. (2018). Ecological conditions drive pace-of-life syndromes by shaping relationships between life history, physiology and behaviour in two populations of Eastern mosquitofish. *Scientific Reports*, 8, 14673. <https://doi.org/10.1038/s41598-018-33047-0>
- Pörtner, H. O. (2012). Integrating climate-related stressor effects on marine organisms: Unifying principles linking molecule to ecosystem-level changes. *Marine Ecology Progress Series*, 470, 273–290. <https://doi.org/10.3354/meps10123>
- Prentice, P. M., Houslay, T. M., Martin, J. G. A., & Wilson, A. J. (2020). Genetic variance for behavioural ‘predictability’ of stress response. *Journal of Evolutionary Biology*, 33(5), 642–652. <https://doi.org/10.1111/jeb.13601>
- R Core Team. (2020). *R: A language and environment for statistical computing*. R Foundation for Statistical Computing. <http://www.r-project.org/>
- Réale, D., Garant, D., Humphries, M. M., Bergeron, P., Careau, V., & Montiglio, P. O. (2010). Personality and the emergence of the pace-of-life syndrome concept at the population level. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 365(1560), 4051–4063. <https://doi.org/10.1098/rstb.2010.0208>
- Richier, S., Rodriguez-Lanetty, M., Schnitzler, C. E., & Weis, V. M. (2008). Response of the symbiotic cnidarian *Anthopleura elegantissima* transcriptome to temperature and UV increase. *Comparative Biochemistry and Physiology - Part D: Genomics and Proteomics*, 3(4), 283–289. <https://doi.org/10.1016/j.cbd.2008.08.001>

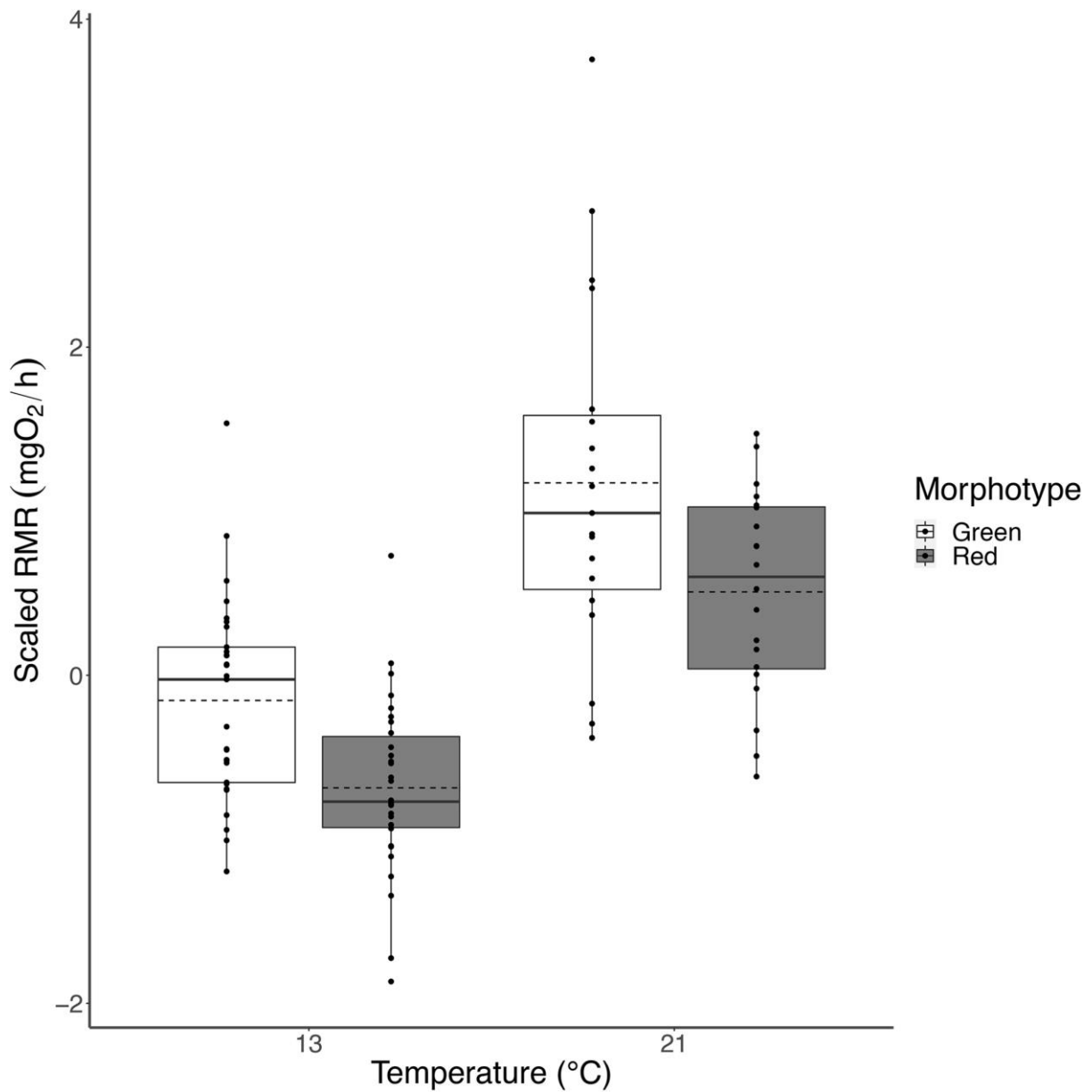
- Royauté, R., Berdal, M. A., Garrison, C. R., & Dochtermann, N. A. (2018). PACELESS life? A meta-analysis of the pace-of-life syndrome hypothesis. *Behavioral Ecology and Sociobiology*, 72(3), 1–10. <https://doi.org/10.1007/s00265-018-2472-z>
- Rudin, F. S., & Briffa, M. (2012). Is boldness a resource-holding potential trait? Fighting prowess and changes in startle response in the sea anemone, *Actinia equina*. *Proceedings of the Royal Society B: Biological Sciences*, 279(1735), 1904–1910. <https://doi.org/10.1098/rspb.2011.2418>
- Schulte, P. M., Healy, T. M., & Fangué, N. A. (2011). Thermal performance curves, phenotypic plasticity, and the time scales of temperature exposure. *Integrative and Comparative Biology*, 51(5), 691–702. <https://doi.org/10.1093/icb/icr097>
- Seebacher, F., White, C. R., & Franklin, C. E. (2015). Physiological plasticity increases resilience of ectothermic animals to climate change. *Nature Climate Change*, 5(1), 61–66. <https://doi.org/10.1038/nclimate2457>
- Segev, U., Burkert, L., Feldmeyer, B., & Foitzik, S. (2017). Pace-of-life in a social insect: behavioral syndromes in ants shift along a climatic gradient. *Behavioral Ecology*, 28(4), 1149–1159. <https://doi.org/10.1093/beheco/arx079>
- Sih, A., Bell, A., & Johnson, J. C. (2004). Behavioral syndromes: An ecological and evolutionary overview. *Trends in Ecology and Evolution*, 19(7), 372–378. <https://doi.org/10.1016/j.tree.2004.04.009>
- Smith, B. R., & Blumstein, D. T. (2008). Fitness consequences of personality: A meta-analysis. *Behavioral Ecology*, 19(2), 448–455. <https://doi.org/10.1093/beheco/arm144>
- Snell-Rood, E. C. (2013). An overview of the evolutionary causes and consequences of behavioural plasticity. *Animal Behaviour*, 85(5), 1004–1011. <https://doi.org/10.1016/j.anbehav.2012.12.031>

- Stamps, J. A., Briffa, M., & Biro, P. A. (2012). Unpredictable animals: Individual differences in intraindividual variability (IIV). *Animal Behaviour*, *83*(6), 1325–1334. <https://doi.org/10.1016/j.anbehav.2012.02.017>
- Stamps, J. A. (2016). Individual differences in behavioural plasticities. *Biological Reviews*, *91*(2), 534–567. <https://doi.org/10.1111/brv.12186>
- Tomanek, L., & Somero, G. N. (1999). Evolutionary and acclimation-induced variation in the heat-shock responses of congeneric marine snails (genus *Tegula*) from different thermal habitats: implications for limits of thermotolerance and biogeography. *Journal of Experimental Biology*, *202*(21), 2925–2936. <https://doi.org/10.1242/jeb.202.21.2925>
- Twardochleb, L. A., Treakle, T. C., & Zarnetske, P. L. (2020). Foraging strategy mediates ectotherm predator–prey responses to climate warming. *Ecology*, *101*(11), 1–11. <https://doi.org/10.1002/ecy.3146>
- Vajedsamiei, J., Wahl, M., Schmidt, A. L., Yazdanpanahan, M., & Pansch, C. (2021). The Higher the Needs, the Lower the Tolerance: Extreme Events May Select Ectotherm Recruits With Lower Metabolic Demand and Heat Sensitivity. *Frontiers in Marine Science*, *8*, 264. <https://doi.org/10.3389/fmars.2021.660427>
- Velasque, M., & Briffa, M. (2016). The opposite effects of routine metabolic rate and metabolic rate during startle responses on variation in the predictability of behaviour in hermit crabs. *Behaviour*, *153*(13–14), 1545–1566. <https://doi.org/10.1163/1568539X-00003371>
- White, S. J., & Briffa, M. (2017). How do anthropogenic contaminants (ACs) affect behaviour? Multi-level analysis of the effects of copper on boldness in hermit crabs. *Oecologia*, *183*(2), 391–400. <https://doi.org/10.1007/s00442-016-3777-0>
- Wolf, M., & McNamara, J. M. (2012). On the Evolution of Personalities via Frequency-Dependent Selection. *The American Naturalist*, *179*(6), 679–692. <https://doi.org/10.1086/665656>

Wolf, M., & Weissing, F. J. (2010). An explanatory framework for adaptive personality differences. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 365(1560), 3959–3968. <https://doi.org/10.1098/rstb.2010.0215>

Yamane, L., & Gilman, S. E. (2009). Opposite responses by an intertidal predator to increasing aquatic and aerial temperatures. *Marine Ecology Progress Series*, 393, 27–36. <https://doi.org/10.3354/meps08276>

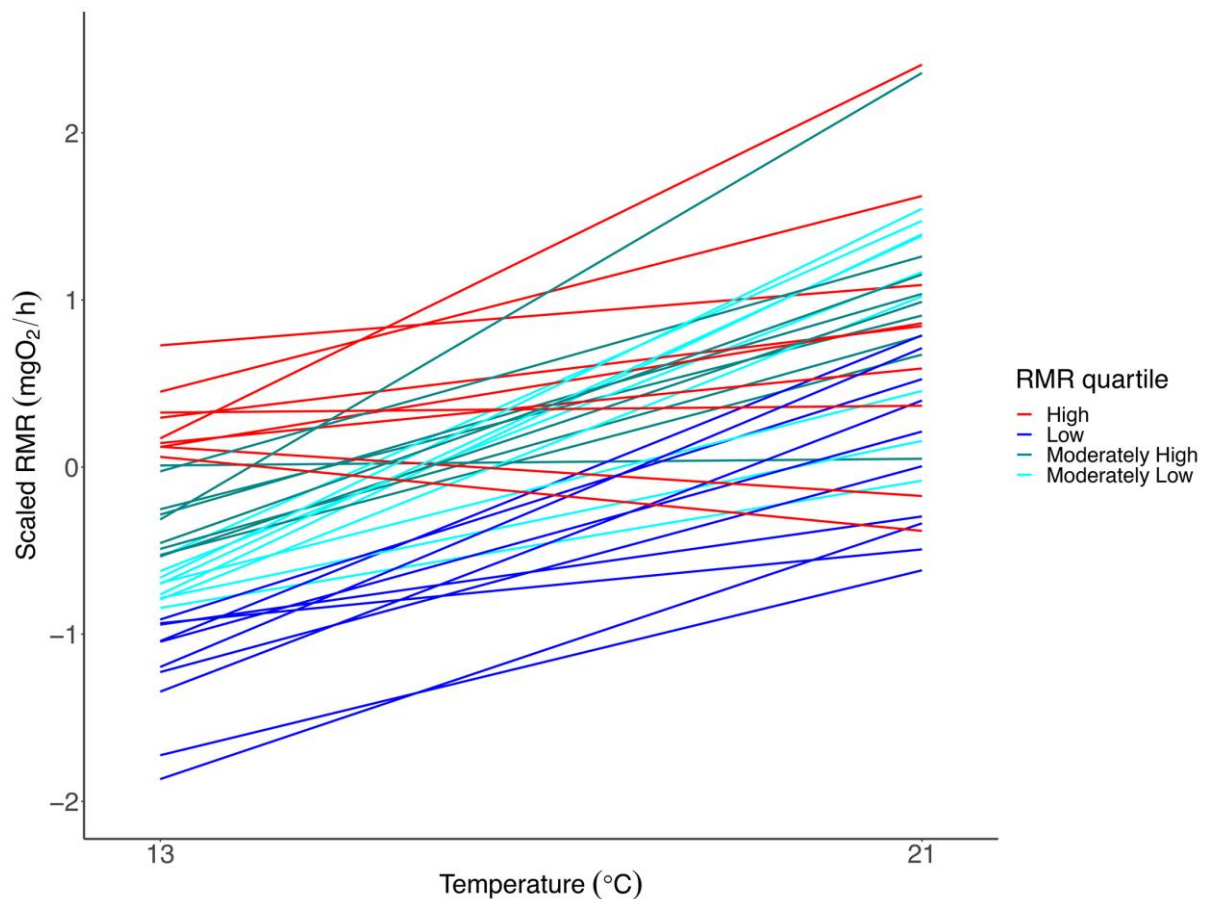




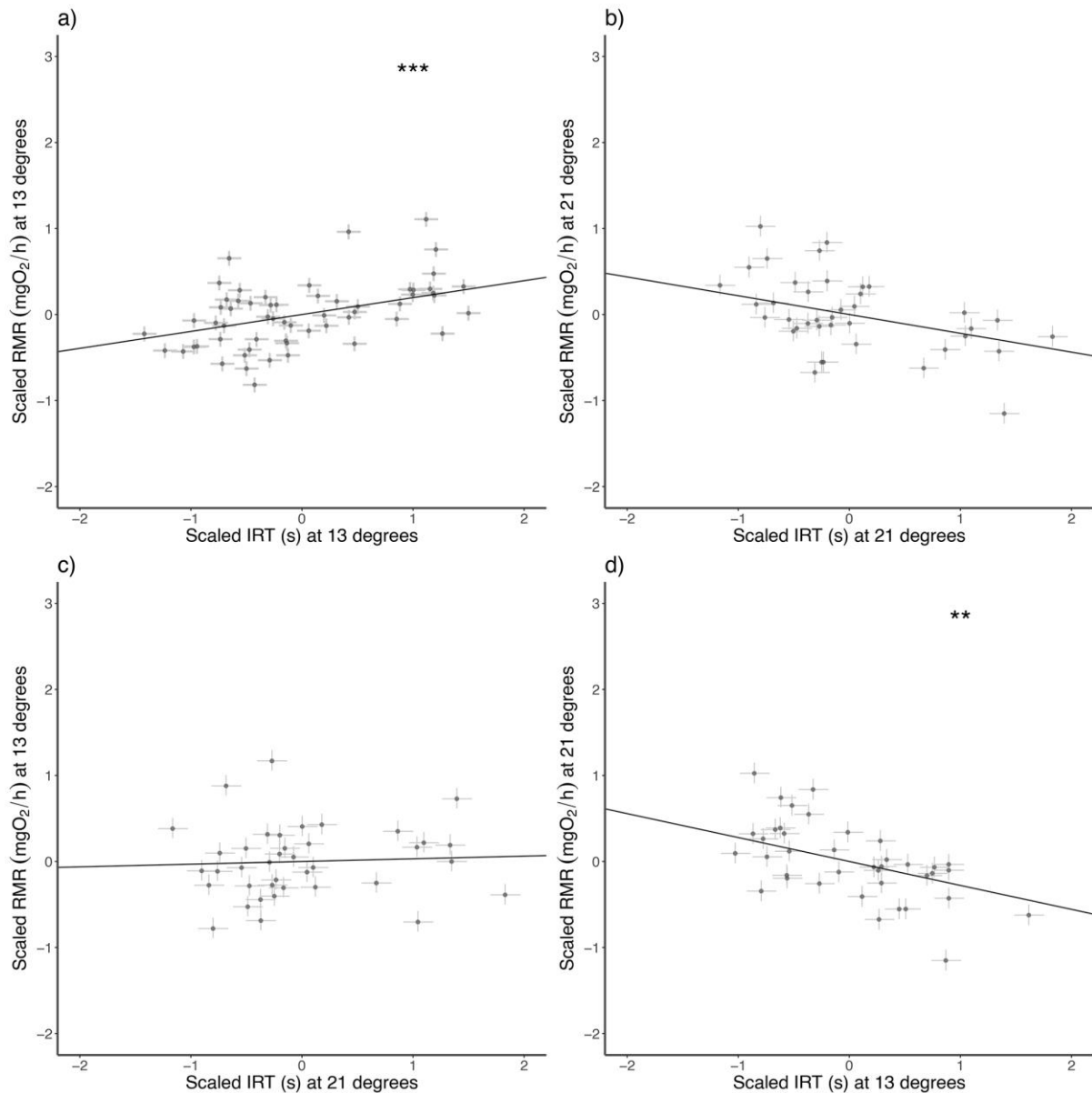
**Fig. 2. Variation between red and green morphotypes predicted mean routine metabolic rate (RMR), scaled to standard deviation units and centred on the mean at zero, at 13°C and 21°C for all individuals (n = 60), derived from a Bayesian random slopes analysis.**

Boxes denote the median value, with the first and third quartiles forming the box limits, dotted lines denote the mean. Whiskers extend to encompass all data or 1.5 times the interquartile range. Any point falling outside the whiskers can be deemed to be an outlier.





**Fig. 3.** The change in individual-level posterior mean predictions, scaled to standard deviation units and centred on the mean at zero, for experimental individuals (n=40) for routine metabolic rate (RMR) between 13°C and 21°C. Each line corresponds to a single individual's predicted RMR at each temperature. Individuals are coloured based on their RMR quartiles, compared with other experimental individuals, at 13°C. Predictions are derived from a random regression model across 13°C and 21°C run using the full dataset incorporating control individuals (n = 60). For visualisation purposes, one individual that exhibited especially high scaled estimates at both temperatures (13°C = 1.54, 21°C = 3.76), is excluded from the plot.



**Fig. 4. The relationship between individual-level posterior mode estimates (Bayesian BLUPS) for routine metabolic rate (RMR) and immersion response-time (IRT), both scaled to their respective standard deviation units and centred on their respective means at zero, at A) 13°C for both traits, B) 21°C for both traits, C) 13°C for RMR and 21°C for IRT and D) 21°C for RMR and 13°C for IRT. “\*\*\*” denotes a statistically significant relationship between the two traits. “\*\*” denotes a relationship falling at the bound of statistical significance. Estimates are derived from bivariate (panel a) and multivariate (panels b, c and d) Bayesian mixed effects models. Error bars denote 95% credible intervals around each of the traits at each temperature.**

**Table 1.** Bivariate relationships between immersion response-time (IRT) and routine metabolic rate (RMR) across different temperatures. Significant relationships, and those falling on the bound of significance, are denoted by bold text. Relationships were extracted from one bivariate and one multivariate Bayesian model.

Relationship	Correlation Estimate (r)	Lower 95% CI	Upper 95% CI
IRT 13°C & RMR 13°C	<b>0.34</b>	<b>0.02</b>	<b>0.63</b>
IRT 13°C & RMR 21°C	<b>-0.35</b>	<b>-0.68</b>	<b>-0.01</b>
IRT 21°C & RMR 13°C	0.04	-0.33	0.45
IRT 21°C & RMR 21°C	-0.30	-0.65	0.06

## **Supplementary Materials and Methods**

### **1.1 Metabolic apparatus and additional measurement detail**

Automated intermittent-flow cycles were precisely controlled by a National Instruments C Series Counter Input Module (National Instruments, Austin, USA). Water in the respirometry tank was fully aerated throughout testing by a Tetra Second Nature Whisper 800 air pump (Tetra GMBH, Melle, Germany). Each chamber contained a central 6 x 6 x 6 cm polypropylene mesh cage containing a magnetic stir bar, which was controlled by an IKAMAG multi-position magnetic stirrer plate (IKA England LTD, Oxford, UK) positioned under the respirometry tank. For the ambient 13°C measurements, the temperature within the respirometry tank was maintained to a precision of  $\pm 0.2^\circ\text{C}$  by the laboratory's temperature control. At 21°C, the temperature of the tank was maintained to a precision of  $\pm 0.3^\circ\text{C}$ ; a waterproof DS1820B temperature sensor (Maxim Integrated Products, Sunnyvale, USA) was connected to an Arduino Uno (Arduino LLC, Boston, USA) and used to control a LightwaveRF wireless control plug socket (LightwaveRF, Birmingham, UK) which was itself connected to an Eheim thermocontrol 300W heater (Eheim, Deizisau, Germany) submerged in the tank. During experimentation, the temperature in the respirometry tank was recorded at half hour intervals, taking measurements directly from the temperature sensor using the serial monitor within the Arduino software. Oxygen concentration, and by extension, oxygen consumption, was measured at second intervals using two PyroScience Firesting O<sub>2</sub> dissolved oxygen sensors (PyroScience GmbH, Aachen, Germany) each connected to three PyroScience 3mm diameter robust oxygen probes fed directly into chambers. Probes were calibrated at the start of each day of metabolic testing.

**Table S1.** Full checklist of essential information on our aquatic respirometry methods (from: Killen et al., 2021)

Number	Criterion and Category	Response	Value (where required)	Units
<b>EQUIPMENT, MATERIALS, AND SETUP</b>				
1	Body mass of animals at time of respirometry	Pedal disc diameter (PDD) threshold used as no reliable, non-invasive measure of anemone volume	20	mm
2	Volume of empty respirometers	422.5ml		
3	How chamber mixing was achieved	Magnetic stirrer		
4	Ratio of net respirometer volume (plus any associated tubing in mixing circuit) to animal body mass	N/a, no reliable measure of live anemone volume so threshold PDD used		
5	Material of tubing used in any mixing circuit	PVC		
6	Volume of tubing in any mixing circuit	N/a		
7	Confirm volume of tubing in any mixing circuit was included in calculations of oxygen uptake	N/a		
8	Material of respirometer (e.g. glass, acrylic, etc.)	Glass, Polypropylene lid		
9	Type of oxygen probe and data recording	Pyroscience Firesting O <sub>2</sub>		
10	Sampling frequency of water dissolved oxygen	1s		
11	Describe placement of oxygen probe (in mixing circuit or directly in chamber)	Directly		
12	Flow rate during flushing and recirculation, or confirm that chamber returned to normoxia during flushing	Chamber returned to normoxia (Appendix 1.3)		
13	Timing of flush/closed cycles			
14	Wait (delay) time excluded from closed measurement cycles	>60s		
15	Frequency and method of probe calibration (for both 0 and 100% calibrations)	Daily, calibrated to 100% dissolved oxygen saturation		
16	State whether software temperature compensation was used during recording of water oxygen concentration	Yes, standardised to temperature of treatment, not to real-time fluctuation		
<b>MEASUREMENT CONDITIONS</b>				
17	Temperature during respirometry	Variable	13/21	°C
18	How temperature was controlled	Arduino	± 0.3	°C
19	Photoperiod during respirometry	12:12h		
20	If (and how) ambient water bath was cleaned and aerated during measurement of oxygen uptake (e.g. filtration, periodic or continuous water changes)	Periodic water changes and bleach sterilisation. Air stone for aeration.		

21	Total volume of ambient water bath and any associated reservoirs	61640cm <sup>3</sup> (67cmx46cmx20cm)
22	Minimum water oxygen dissolved oxygen reached during closed phases	>80%
23	State whether chambers were visually shielded from external disturbance	Yes, opaque shield around water bath
24	How many animals were measured during a given respirometry trial (i.e. how many animals were in the same water bath)	Five
25	If multiple animals were measured simultaneously, state whether they were able to see each other during measurements	N/a
26	Duration of animal fasting before placement in respirometer	>24h
27	Duration of all trials combined (number of days to measure all animals in the study)	24 days
28	Acclimation time to the laboratory (or time since capture for field studies) before respirometry measurements	48h x 2
<b>BACKGROUND RESPIRATION</b>		
29	Whether background microbial respiration was measured and accounted for, and if so, method used (e.g. parallel measures with empty respirometry chamber, measurements before and after for all chambers while empty, both)	Yes, parallel measures
30	If background respiration was measured at beginning and/or end, state how many slopes and for what duration	N/a
31	How changes in background respiration were modelled over time (e.g. linear, exponential, parallel measures)	Parallel measures
32	Level of background respiration (e.g. as a percentage of SMR)	Temperature and trial-dependent
33	Method and frequency of system cleaning (e.g. system bleached between each trial, UV lamp)	System and chambers bleached between each trial
<b>STANDARD OR ROUTINE METABOLIC RATE</b>		
34	Acclimation time after transfer to chamber, or alternatively, time to reach beginning of metabolic rate measurements after introduction to chamber	12h
35	Duration over which metabolic rate was estimated	38 mins x 6
36	Value taken as SMR/RMR (e.g. quantile, mean of lowest 10 percent, mean of all values)	N/a, individual slopes recorded for analyses
37	Total number of slopes measured and used to derive metabolic rate (e.g. how much data were used to calculate quantiles)	N/a, individual slopes recorded for analyses

38	Whether any time periods were removed from calculations of SMR/RMR (e.g. data during acclimation, periods of high activity [e.g. daytime])	No data were recorded during acclimation / attachment
39	r <sup>2</sup> threshold for slopes used for SMR/RMR (or mean)	0.9
40	Proportion of data removed due to being outliers below r-squared threshold	0%
<b>MAXIMUM METABOLIC RATE</b>		
41	When MMR was measured in relation to SMR (i.e. before or after)	N/a
42	Method used (e.g. critical swimming speed respirometry, swim to exhaustion in swim tunnel, or chase to exhaustion)	N/a
43	Value taken as MMR (e.g. the highest rate of oxygen uptake value after transfer, average of highest values)	N/a
44	If MMR measured post-exhaustion, length of activity challenge or chase (e.g. 2 min, until exhaustion, etc.)	N/a
45	If MMR measured post-exhaustion, state whether further air-exposure was added after exercise	N/a
46	If MMR measured post-exhaustion, time until transfer to chamber after exhaustion or time to start of oxygen uptake recording	N/a
47	Duration of slopes used to calculate MMR (e.g. 1 min, 5 min, etc.)	N/a
48	Slope estimation method for MMR (e.g. rolling regression, sequential discrete time frames)	N/a
49	How absolute aerobic scope and/or factorial aerobic scope is calculated (i.e. using raw SMR and MMR, allometrically mass-adjusted SMR and MMR, or allometrically mass-adjusting aerobic scope itself)	N/a
<b>DATA HANDLING AND STATISTICS</b>		
50	Sample size	60 overall, 20 control, 40 crossed-over
51	How oxygen uptake rates were calculated (software or script, equation, units, etc.)	respr R package followed by conversions to mgO <sub>2</sub> /h
52	Confirm that volume (mass) of animal was subtracted from respirometer volume when calculating oxygen uptake rates	No, size was standardised to PDD as far as possible
53	State whether analyses accounted for variation in body mass and describe any allometric mass-corrections or adjustments	Yes, all analyses incorporated dry-weight as a fixed effect

## **1.2 Example Slopes and conversion**

Raw oxygen concentration was measured in % O<sub>2</sub> concentration and chamber volume less the volume of the magnetic stirrers, was 422.5ml. Because wet-weight and anemone volume could not be reliably measured, the volume of anemones was not corrected for. Initial slopes were thus measured in %O<sub>2</sub>/s/422.5ml. These were first converted, using temperature-specific concentration conversion tables (Boyd & Pillai, 1985), to mgO<sub>2</sub>/s/422.5ml. These values were then multiplied by 3600 to give the slopes in mgO<sub>2</sub>/h/422.5ml, before being further multiplied by 0.4225 to give the values in mgO<sub>2</sub>/ h.

In the majority of measured slopes of oxygen consumption a steady, linear decline in oxygen concentration was observed after an initial wait period of at least one minute (Svendsen et al., 2016; Figure S1), but there were some instances where slopes showed more noise. The respR package provides mitigation for this by calculating rolling regressions (Harianto et al., 2019), allowing accurate slope estimation even with a substantial degree of noise in the data around the gradient of the slope (Figure S2). In some instances, there were periods of noise which influenced the results of the rolling regression (Figure S3). In these instances, where possible, the longest, most stable, periods of linear decline of oxygen concentration were used in slope calculations (White et al., 2016). Where this was not possible, the slope was discounted from the dataset. One individual, from the H-L treatment and of the green morphotype, was discounted from all analyses as it showed no measurable slopes at 21°C. With this individual removed, of 826 oxygen consumption measurements (including blanks), 42 were measured using substantially truncated slopes. Of 708 anemone consumption measurements, nine were discounted from analyses.



### **1.3 Statistical detail**

#### *Model Specification*

Univariate models were fitted over uninformative Inverse-Wishart priors and run for 420000 iterations with a 20000-iteration initial burn-in period and a thinning interval of 100. Bivariate and multivariate models were fit similarly, but over 1000000 iterations with a 50000-iteration burn-in. To ensure successful convergence visual inspection of convergence and autocorrelation plots was carried out. To further confirm this, Heidelberger-Welch and Gelman-Rubin diagnostic tests were run on all models. Comparative models were also run over alternative, parameter-expanded priors, to ensure robustness to different prior specifications. Model estimates and deviance-information criteria (DIC; Spiegelhalter et al., 2002) did not differ meaningfully between the different specifications.

#### *Model choice random slopes*

Two models were initially run on our full dataset. A random intercepts model, not taking into account individual-level variation in metabolic plasticity to temperature, and a random slopes model. Once assumptions and convergence of both models had been checked, the two models' DIC values (Spiegelhalter et al., 2002) were compared. The DIC of the random slopes model (DIC = 666) was dramatically lower than that of the random intercepts model (DIC = 1110), indicating very clearly that the former, incorporating a random slope effect, was a better fit to the data than the latter.

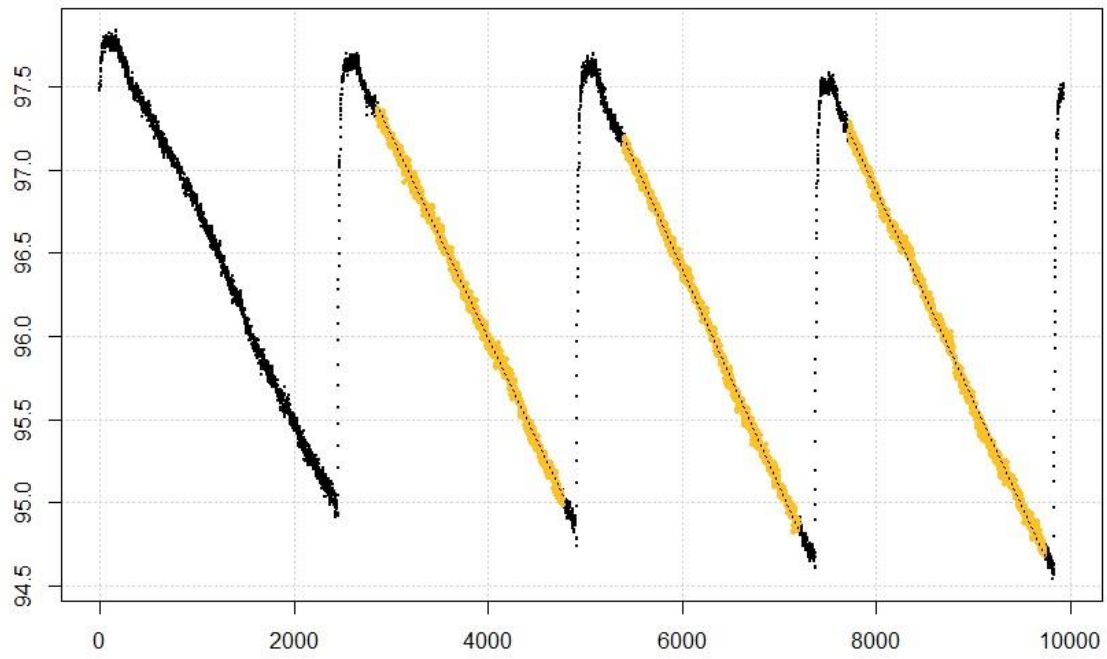
#### *The relationship between RMR and dry weight*

To explore the relationship between size and RMR, a further bivariate model was run where RMR and dry weight were set as the response variables. Fixed and random effects were the same as the bivariate model above but with temperature set as an extra fixed effect. This confirmed a strong positive correlation between RMR and dry weight ( $r = 0.72$ , 95%CI = 0.57, 0.86; Figure S4). The size

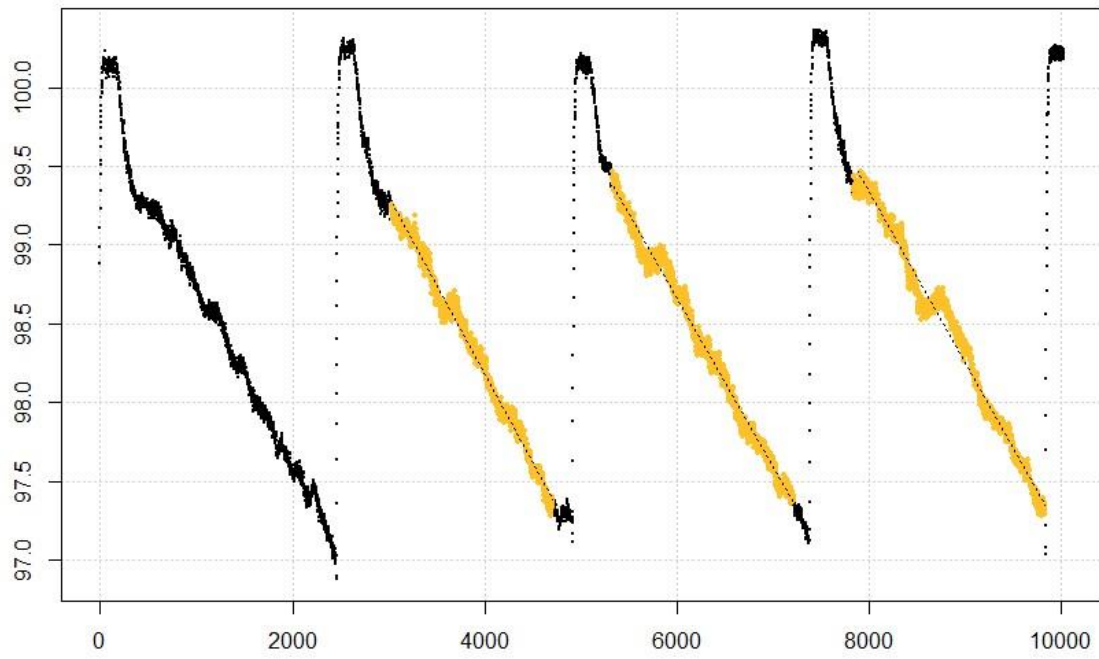
of the correlation found here is in line with previous work in *A. equina* (Navarro et al., 1981) and also indicates that RMR in this species, when investigated in relation to dry weight, follows a scaling pattern roughly in line with the  $\frac{3}{4}$  power relationship discussed in Brown et al (2004).

## SI: References

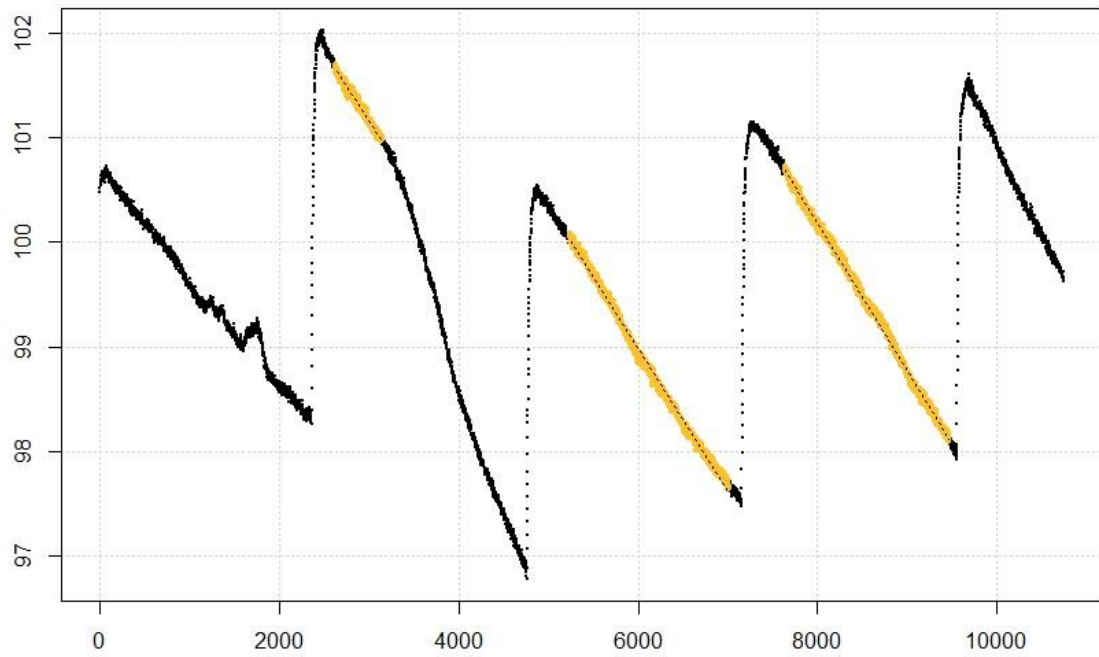
- Boyd, C. E., & Pillai, V. K. (1985). Water Quality Management in Aquaculture. *CMFRI Special Publication 22*, 1–44.
- Brown, J. H., Gillooly, J. F., Allen, A. P., Savage, V. M., & West, G. B. (2004). Toward a metabolic theory of ecology. *Ecology*, *85*(7), 1771–1789. <https://doi.org/10.1890/03-9000>
- Hariato, J., Carey, N., & Byrne, M. (2019). respR—An R package for the manipulation and analysis of respirometry data. *Methods in Ecology and Evolution*, *10*(6), 912–920. <https://doi.org/10.1111/2041-210X.13162>
- Killen, S. S., Christensen, E. A. F., Cortese, D., Závorka, L., Norin, T., Cotgrove, L., Crespel, A., Munson, A., Nati, J. J. H., Papatheodoulou, M., & McKenzie, D. J. (2021). Guidelines for reporting methods to estimate metabolic rates by aquatic intermittent-flow respirometry. *Journal of Experimental Biology*, *224*(18), jeb242522. <https://doi.org/10.1242/jeb.242522>
- Navarro, E., Ortega, M. M., & Madariaga, J. M. (1981). Effect of body size, temperature and shore level on aquatic and aerial respiration of *Actinia equina* (L.) (Anthozoa). *Journal of Experimental Marine Biology and Ecology*, *53*(2–3), 153–162. [https://doi.org/10.1016/0022-0981\(81\)90016-2](https://doi.org/10.1016/0022-0981(81)90016-2)
- Spiegelhalter, D. J., Best, N. G., Carlin, B. P., & Van Der Linde, A. (2002). Bayesian measures of model complexity and fit. *Journal of the Royal Statistical Society. Series B: Statistical Methodology*, *64*(4), 583–616. <https://doi.org/10.1111/1467-9868.00353>
- Svendsen, M. B. S., Bushnell, P. G., & Steffensen, J. F. (2016). Design and setup of intermittent-flow respirometry system for aquatic organisms. *Journal of Fish Biology*, *88*(1), 26–50. <https://doi.org/10.1111/jfb.12797>
- White, S. J., Kells, T. J., & Wilson, A. J. (2016). Metabolism, personality and pace of life in the Trinidadian guppy, *Poecilia reticulata*. *Behaviour*, *153*(13–14), 1517–1543. <https://doi.org/10.1163/1568539x-00003375>



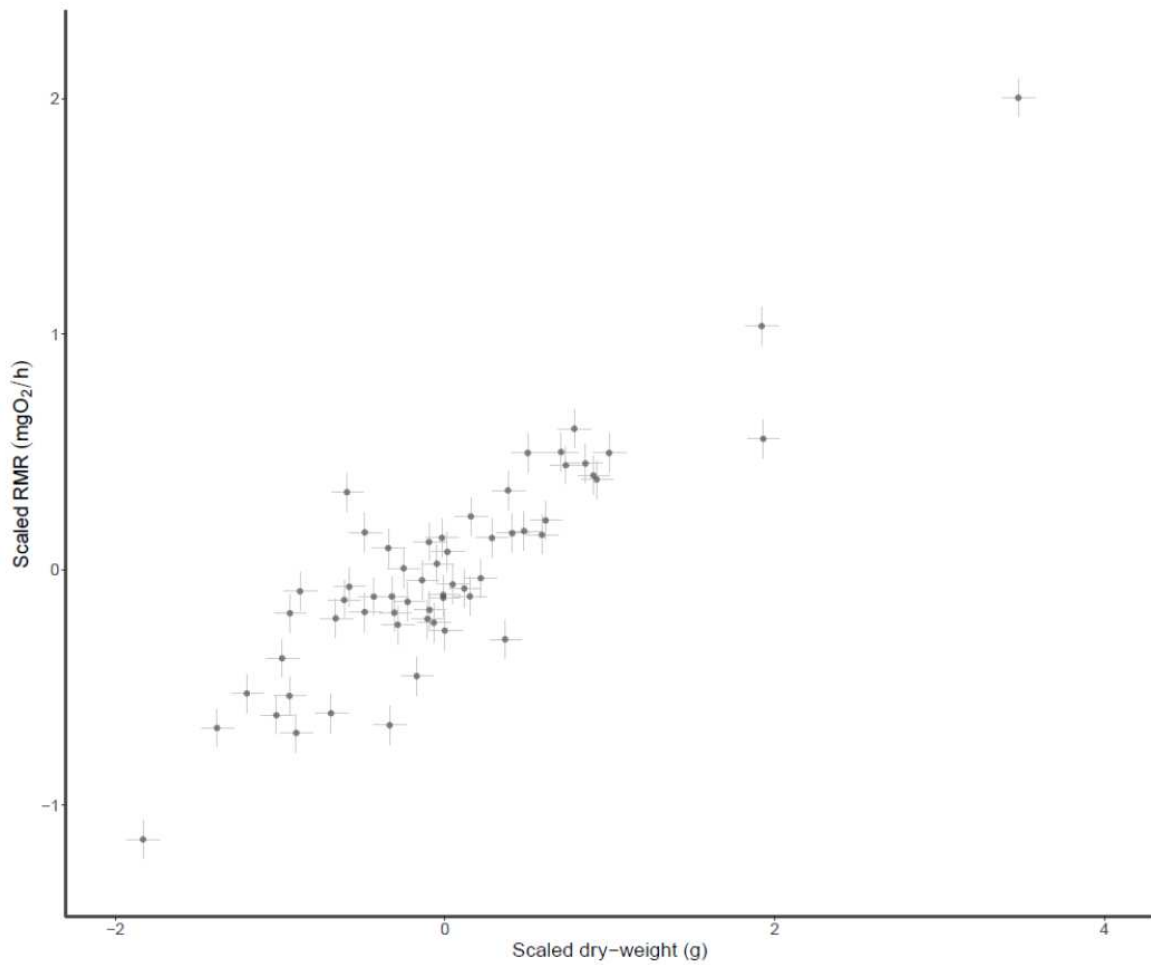
**Fig. S1. Example of metabolic slopes with no fluctuation or truncation.**



**Fig. S2.** Example of metabolic slopes where fluctuations could be dealt with by rolling regression.



**Fig. S3.** Example of a metabolic slope where a large degree of truncation was required.



**Fig. S4.** The relationship between the posterior mode estimates (Bayesian BLUPS) of individual RMR and individual dry weight, derived from a bivariate mixed effects model run on the full dataset (n=60).