

Groundwater is a hidden global keystone ecosystem

Mattia Saccò¹, Stefano Mammola^{2,3,4}, Florian Altermatt^{5,6}, Roman Alther^{5,6}, Rossano Bolpagni⁷, Anton Brancelj^{8,9}, David Brankovits², Cene Fišer¹⁰, Vasilis Gerovasileiou^{11,12}, Christian Griebler¹³, Simone Guareschi¹⁴, Grant C Hose¹⁵, Kathryn Korbel¹⁵, Elisabeth Lictevout¹⁶, Florian Malard¹⁷, Alejandro Martínez², Matthew L Niemiller¹⁸, Anne Robertson¹⁹, Krizler C Tanalgo²⁰, Maria Elina Bichuette²¹, Špela Borko¹⁰, Traian Brad²², Matthew A Campbell²³, Pedro Cardoso^{3,24}, Fulvio Celico⁷, Steven J B Cooper²⁵, David Culver²⁶, Tiziana Di Lorenzo^{4,27}, Diana M P Galassi²⁸, Michelle T Guzik²⁹, Adam Hartland³⁰, William F Humphreys³¹, Rodrigo Lopes Ferreira³², Enrico Lunghi²⁸, Daniele Nizzoli⁷, Giulia Perina¹, Rajeev Raghavan³³, Zoe Richards³⁴, Ana Sofia P S Reboleira²⁴, Melissa M Rohde³⁵, David Sánchez Fernández³⁶, Susanne I Schmidt³⁷, Mieke Van Der Heyde¹, Louise Weaver³⁸, Nicole E White¹, Maja Zgamajster¹⁰, Ian Hogg³⁹, Albert Ruhi⁴⁰, Marthe M Gagnon⁴¹, Morten E Allentoft²³, and Robert Reinecke⁴²

¹School of Molecular and Life Sciences, Subterranean Research and Groundwater Ecology (SuRGE) Group, Trace and Environmental DNA (TrEnD) Lab, Curtin University

²Molecular Ecology Group (MEG), Water Research Institute (CNR-IRSA), National Research Council

³Finnish Museum of Natural History (LUOMUS), Laboratory for Integrative Biodiversity Research (LIBRe), University of Helsinki

⁴National Biodiversity Future Center

⁵Department of Aquatic Ecology, Swiss Federal Institute of Aquatic Science and Technology

⁶Department of Evolutionary Biology and Environmental Studies, University of Zurich

⁷Department of Chemistry, Life Sciences and Environmental Sustainability, University of Parma

⁸Department of Organisms and Ecosystems Research, National Institute of Biology

⁹Department for Environmental Science, University of Nova Gorica, Nova Gorica

¹⁰Biotechnical Faculty, Department of Biology, SubBio Lab, University of Ljubljana

¹¹Department of Environment, Faculty of Environment, Ionian University

¹²Institute of Marine Biology, Biotechnology and Aquaculture (IMBBC), Hellenic Centre for Marine Research (HCMR)

¹³Department of Functional & Evolutionary Ecology, University of Vienna

¹⁴Estación Biológica de Doñana (EBD-CSIC)

¹⁵School of Natural Sciences, Macquarie University

¹⁶International Groundwater Resources Assessment Center (IGRAC)

¹⁷Univ Lyon, Université Claude Bernard Lyon 1, CNRS, ENTPE, UMR 5023 LEHNA, Villeurbanne, France

¹⁸Department of Biological Sciences, The University of Alabama in Huntsville

¹⁹School of Life and Health Sciences, Roehampton University, Holybourne Avenue, London SW15 4JD, UK

- ²⁰Ecology and Conservation Research Laboratory (Eco/Con Lab), Department of Biological Sciences, College of Science and Mathematics, University of Southern Mindanao, Kabacan, Cotabato, Philippines
- ²¹Laboratory of Subterranean Studies (LES), Department of Ecology and Evolutionary Biology, Federal University of São Carlos, São Carlos, Brazil
- ²²Department of Cluj-Napoca, Emil Racovita Institute of Speleology, Cluj-Napoca, Romania
- ²³Trace and Environmental DNA (TrEnD) Lab, School of Molecular and Life Sciences, Curtin University, Perth, Western Australia, Australia
- ²⁴Departamento de Biologia Animal, and Centre for Ecology, Evolution and Environmental Changes & CHANGE – Global Change and Sustainability Institute, Faculdade de Ciências, Universidade de Lisboa, Campo Grande, 1749-016, Lisbon, Portugal
- ²⁵South Australian Museum, North Terrace, Adelaide, South Australia 5000, Australia
- ²⁶Department of Environmental Science, American University, 4400 Massachusetts Ave NW, Washington, DC 20016, USA
- ²⁷Research Institute on Terrestrial Ecosystems of the National Research Council of Italy (IRET CNR), Florence, Italy
- ²⁸Department of Life, Health and Environmental Sciences (MESVA), University of L'Aquila, L'Aquila, Italy
- ²⁹School of Biological Sciences, The University of Adelaide, Adelaide, South Australia, Australia
- ³⁰Lincoln Agritech Ltd, Ruakura, Kirikiriroa, 3216, Aotearoa New Zealand
- ³¹Western Australian Museum, Kew Street, Welshpool, WA
- ³²Centro de Estudos em Biologia Subterrânea, Departamento de Ecologia e Conservação, Instituto de Ciências Naturais, Universidade Federal de Lavras, Cx Postal 3037, Campus Universitário, CEP 37200-000, Lavras, Minas Gerais, Brazil
- ³³Department of Fisheries Resource Management, Kerala University of Fisheries and Ocean Studies, Kochi, India 682 506
- ³⁴36Coral Conservation and Research Group, Trace and Environmental DNA (TrEnD) Lab, School of Molecular and Life Sciences, Curtin University, Australia
- ³⁵Rohde Environmental Consulting, LLC, Seattle, Washington, USA
- ³⁶Department of Ecology and Hydrology, University of Murcia, 30100, Murcia, Spain
- ³⁷Department of Lake Research, Helmholtz Centre for Environmental Research, 39114 Magdeburg, Germany
- ³⁸Water & Environment Group, Institute of Environmental Science & Research Ltd., Christchurch, New Zealand
- ³⁹School of Science, University of Waikato, Hamilton, New Zealand
- ⁴⁰Department of Environmental Science, Policy & Management, University of California, Berkeley, CA, 94702, USA
- ⁴¹School of Molecular and Life Sciences, Curtin University, Brand Drive, Bentley, WA 6102, Australia
- ⁴²Institute of Geography, University of Mainz, Mainz, Germany

August 17, 2023

Groundwater is a hidden global keystone ecosystem

Mattia Saccò^{1,*}, Stefano Mammola^{2,3,4^}, Florian Altermatt^{5,6}, Roman Alther^{5,6}, Rossano Bolpagni⁷, Anton Brancelj^{8,9}, David Brankovits², Cene Fišer¹⁰, Vasilis Gerovasileiou^{11,12}, Christian Griebler¹³, Simone Guareschi¹⁴, Grant C. Hose¹⁵, Kathryn Korbel¹⁵, Elisabeth Lictévout¹⁶, Florian Malard¹⁷, Alejandro Martínez², Matthew L. Niemiller¹⁸, Anne Robertson¹⁹, Krizler C. Tanalgo²⁰, Maria Elina Bichuette²¹, Špela Borko¹⁰, Traian Brad²², Matthew A. Campbell²³, Pedro Cardoso^{3,24}, Fulvio Celico⁷, Steven J. B. Cooper^{25,26}, David Culver²⁷, Tiziana Di Lorenzo^{4,28}, Diana M. P. Galassi²⁹, Michelle T. Guzik³⁰, Adam Hartland³¹, William F. Humphreys^{32,33}, Rodrigo Lopes Ferreira³⁴, Enrico Lunghi²⁹, Daniele Nizzoli⁷, Giulia Perina¹, Rajeev Raghavan³⁵, Zoe Richards^{34,36}, Ana Sofia P. S. Reboleira²⁴, Melissa M. Rohde^{37,38}, David Sánchez Fernández³⁹, Susanne I. Schmidt⁴⁰, Mieke van der Heyde¹, Louise Weaver⁴¹, Nicole E. White¹, Maja Zgmajster¹⁰, Ian Hogg^{42,43}, Albert Ruhi⁴⁴, Marthe M. Gagnon⁴⁵, Morten E. Allentoft^{23,46}, Robert Reinecke⁴⁷

* Corresponding author: mattia.sacco@curtin.edu.au

^ Shared first author

AFFILIATIONS

¹Subterranean Research and Groundwater Ecology (SuRGE) Group, Trace and Environmental DNA (TrEnD) Lab, School of Molecular and Life Sciences, Curtin University, Perth, Western Australia, Australia

²Molecular Ecology Group (MEG), Water Research Institute (CNR-IRSA), National Research Council, Verbania Pallanza, Italy

³Laboratory for Integrative Biodiversity Research (LIBRe), Finnish Museum of Natural History (LUOMUS), University of Helsinki, Helsinki, Finland

⁴National Biodiversity Future Center, Palermo, Italy

⁵Eawag, Swiss Federal Institute of Aquatic Science and Technology, Department of Aquatic Ecology, Dübendorf, Switzerland

⁶University of Zurich, Department of Evolutionary Biology and Environmental Studies, Zürich, Switzerland

⁷Department of Chemistry, Life Sciences and Environmental Sustainability, University of Parma, Parco Area delle Scienze 11/A, 43124 Parma, Italy

⁸National Institute of Biology, Department of Organisms and Ecosystems Research, Ljubljana, Slovenia

⁹University of Nova Gorica, Department for Environmental Science, Nova Gorica, Slovenia

¹⁰University of Ljubljana, Biotechnical Faculty, Department of Biology, SubBio Lab, Slovenia

¹¹Department of Environment, Faculty of Environment, Ionian University, Zakynthos, Greece

¹²Hellenic Centre for Marine Research (HCMR), Institute of Marine Biology, Biotechnology and Aquaculture (IMBBC), Thalassocosmos, Heraklion, Greece

¹³University of Vienna, Department of Functional & Evolutionary Ecology, Vienna, Austria.

¹⁴Estación Biológica de Doñana (EBD-CSIC), Seville, Spain

¹⁵School of Natural Sciences, Macquarie University, Sydney, New South Wales, Australia

¹⁶International Groundwater Resources Assessment Center (IGRAC)

¹⁷Univ Lyon, Université Claude Bernard Lyon 1, CNRS, ENTPE, UMR 5023 LEHNA, Villeurbanne, France

¹⁸Department of Biological Sciences, The University of Alabama in Huntsville, Huntsville, Alabama, USA

¹⁹School of Life and Health Sciences, Roehampton University, Holybourne Avenue, London SW15 4JD, UK

²⁰Ecology and Conservation Research Laboratory (Eco/Con Lab), Department of Biological Sciences, College of Science and Mathematics, University of Southern Mindanao, Kabacan, Cotabato, Philippines

²¹Laboratory of Subterranean Studies (LES), Department of Ecology and Evolutionary Biology, Federal University of São Carlos, São Carlos, Brazil

²²Department of Cluj-Napoca, Emil Racovita Institute of Speleology, Cluj-Napoca, Romania

²³Trace and Environmental DNA (TrEnD) Lab, School of Molecular and Life Sciences, Curtin University, Perth, Western Australia, Australia

²⁴Departamento de Biologia Animal, and Centre for Ecology, Evolution and Environmental Changes & CHANGE – Global Change and Sustainability Institute, Faculdade de Ciências, Universidade de Lisboa, Campo Grande, 1749-016, Lisbon, Portugal

²⁵South Australian Museum, North Terrace, Adelaide, South Australia 5000, Australia

²⁶Department of Ecology and Evolutionary Biology, School of Biological Sciences and Environment Institute, The University of Adelaide, Adelaide, South Australia 5005, Australia

²⁷Department of Environmental Science, American University, 4400 Massachusetts Ave NW, Washington, DC 20016, USA

²⁸Research Institute on Terrestrial Ecosystems of the National Research Council of Italy (IRET CNR), Florence, Italy

²⁹Department of Life, Health and Environmental Sciences (MESVA), University of L'Aquila, L'Aquila, Italy

³⁰School of Biological Sciences, The University of Adelaide, Adelaide, South Australia, Australia

³¹Lincoln Agritech Ltd, Ruakura, Kirikiriroa, 3216, Aotearoa New Zealand

³²School of Biological Sciences, University of Western Australia, Crawley, WA.

³³Western Australian Museum, Kew Street, Welshpool, WA.

³⁴Centro de Estudos em Biologia Subterrânea, Departamento de Ecologia e Conservação, Instituto de Ciências Naturais, Universidade Federal de Lavras, Cx Postal 3037, Campus Universitário, CEP 37200-000, Lavras, Minas Gerais, Brazil.

³⁵Department of Fisheries Resource Management, Kerala University of Fisheries and Ocean Studies, Kochi, India 682 506

³⁶Coral Conservation and Research Group, Trace and Environmental DNA (TrEnD) Lab, School of Molecular and Life Sciences, Curtin University, Australia

³⁷Rohde Environmental Consulting, LLC, Seattle, Washington, USA

³⁸Graduate Program in Environmental Science, State University of New York College of Environmental Science and Forestry, Syracuse, NY, 13210, USA

³⁹Department of Ecology and Hydrology, University of Murcia, 30100, Murcia, Spain

⁴⁰Department of Lake Research, Helmholtz Centre for Environmental Research, 39114 Magdeburg, Germany

⁴¹Water & Environment Group, Institute of Environmental Science & Research Ltd., Christchurch, New Zealand

⁴²School of Science, University of Waikato, Hamilton, New Zealand

⁴³Canadian High Arctic Research Station, Polar Knowledge Canada, Cambridge Bay, NU, Canada

⁴⁴Department of Environmental Science, Policy & Management, University of California, Berkeley, CA, 94702, USA

⁴⁵School of Molecular and Life Sciences, Curtin University, Brand Drive, Bentley, WA 6102, Australia

⁴⁶Lundbeck Foundation GeoGenetics Centre, Globe Institute, University of Copenhagen, Copenhagen, Denmark

⁴⁷Institute of Geography, University of Mainz, Mainz, Germany

ORCID

Mattia Saccò - <https://orcid.org/0000-0001-6535-764X>
Stefano Mammola - <https://orcid.org/0000-0002-4471-9055>
Florian Altermatt - <https://orcid.org/0000-0002-4831-6958>
Roman Alther - <https://orcid.org/0000-0001-7582-3966>
Rossano Bolpagni - <https://orcid.org/0000-0001-9283-2821>
Anton Brancelj - <https://orcid.org/0000-0002-8767-3894>
David Brankovits - <https://orcid.org/0000-0001-9195-8115>
Cene Fišer - <https://orcid.org/0000-0003-1982-8724>
Vasilis Gerovasileiou - <https://orcid.org/0000-0002-9143-7480>
Christian Griebler - <https://orcid.org/0000-0002-8602-581X>
Simone Guareschi - <https://orcid.org/0000-0003-2962-0863>
Grant C. Hose - <https://orcid.org/0000-0003-2106-5543>
Kathryn Korbel - <https://orcid.org/0000-0003-4376-787X>
Elisabeth Lictévout - <https://orcid.org/0000-0003-4983-5650>
Florian Malard - <https://orcid.org/0000-0001-8037-4464>
Alejandro Martínez - <https://orcid.org/0000-0003-0073-3688>
Matthew L. Niemiller - <https://orcid.org/0000-0001-6353-8797>
Anne Robertson - <https://orcid.org/0000-0001-8398-3556>
Krizler C. Tanalgo - <https://orcid.org/0000-0003-4140-336X>
Maria Elina Bichuette - <https://orcid.org/0000-0002-9515-4832>
Špela Borko - <https://orcid.org/0000-0002-8383-8778>
Traian Brad - <https://orcid.org/0000-0002-6749-4338>
Matthew A. Campbell - <https://orcid.org/0000-0002-0353-8389>
Pedro Cardoso - <https://orcid.org/0000-0001-8119-9960>
Fulvio Celico - <https://orcid.org/0000-0003-4666-5924>
Steven J. B. Cooper - <https://orcid.org/0000-0002-7843-8438>
David Culver - <https://orcid.org/0000-0002-8866-9053>
Tiziana Di Lorenzo - <https://orcid.org/0000-0002-3131-7049>
Diana Paola Maria Galassi - <https://orcid.org/0000-0002-6448-2710>
Michelle T. Guzik - <https://orcid.org/0000-0002-4947-9353>
Adam Hartland - <https://orcid.org/0000-0002-1864-5144>
William F. Humphreys - <https://orcid.org/0000-0002-8998-9323>
Rodrigo Lopes Ferreira - <https://orcid.org/0000-0003-3288-4405>
Enrico Lunghi - <https://orcid.org/0000-0002-4228-2750>
Daniele Nizzoli - <https://orcid.org/0000-0003-4731-9804>
Giulia Pierina - <https://orcid.org/0000-0002-0349-3803>
Rajeev Raghavan - <https://orcid.org/0000-0002-0610-261X>
Zoe Richards - <https://orcid.org/0000-0002-8947-8996>
Ana Sofia P. S. Reboleira - <https://orcid.org/0000-0002-4756-7034>
Melissa M. Rohde - <https://orcid.org/0000-0002-1252-0711>
David Sánchez Fernández - <https://orcid.org/0000-0003-1766-0761>
Susanne I. Schmidt - <https://orcid.org/0000-0003-0051-6480>
Mieke van der Heyde - <https://orcid.org/0000-0002-1658-9927>
Louise Weaver - <https://orcid.org/0000-0002-3750-868X>

Nicole E. White - <https://orcid.org/0000-0002-0068-6693>

Maja Zagnajster - <https://orcid.org/0000-0003-1323-9937>

Ian Hogg - <https://orcid.org/0000-0002-6685-0089>

Albert Ruhi - <https://orcid.org/0000-0003-4011-6457>

Marthe M. Gagnon - <https://orcid.org/0000-0002-3190-5094>

Morten E. Allentoft - <https://orcid.org/0000-0003-4424-3568>

Robert Reinecke - <https://orcid.org/0000-0001-5699-8584>

Preface

Groundwater is a vital ecosystem of the global water cycle, hosting unique biodiversity and providing essential ecosystem services to societies. Despite being the largest unfrozen freshwater resource, in a period of depletion by extraction and pollution, groundwater environments have been repeatedly overlooked in global biodiversity conservation agendas. Disregarding the importance of groundwater as an ecosystem ignores its critical role in preserving surface biomes. To foster timely global conservation of groundwater, we propose elevating the concept of keystone species into the realm of ecosystems, claiming groundwater as a keystone ecosystem that influences the integrity of many dependent ecosystems. Our global analysis shows that over half of land surface areas (52.6%) has a medium-to-high interaction with groundwater, reaching up to 75% when deserts and high mountains are excluded. We postulate that the intrinsic transboundary features of groundwater are critical for shifting perspectives toward more holistic approaches in aquatic ecology and beyond. Furthermore, we propose eight key themes to develop a science-policy integrated groundwater conservation agenda. Given ecosystems above and below the ground intersect at many levels, considering groundwater as an essential component of planetary health it is pivotal to the reduction of biodiversity loss and buffering against climate change.

Introduction

Groundwater is the most extensive unfrozen continental reserve of freshwater on Earth^{1,2}. From deep karstic aquifers to shallow alluvial sediments, groundwater is globally ubiquitous and functionally connected to surficial aquatic and terrestrial groundwater-dependent ecosystems (GDEs). Groundwater interacts with the five global surface aquatic biomes (Fig. 1) and, together with oceans and the atmosphere, is the backbone of the global water cycle³. While often exclusively regarded as an economic resource, providing drinking water and water for irrigation and industrial uses⁴, groundwater is also an ecosystem. It hosts a vast diversity of microbial and metazoan species sustaining essential functions and processes^{5,6}, many of which are endemic and highly specialised to a life in permanent darkness⁷. Together, these specialised organisms account for a unique share of the global taxonomic, phylogenetic, and functional diversity⁸, with recent research estimating that more than 25,000 aquatic metazoan species exist in freshwater and saline groundwaters worldwide⁹.

The groundwater ecosystem is facing mounting anthropogenic pressure¹⁰. Water depletion driven by urbanisation, industry, agriculture, and exacerbated by climate change, has been documented on both regional and global scales¹¹. According to estimations, nearly 50% of the world's urban population depends on groundwater resources⁴, with the human demand currently being about 3.5 times the actual volume of aquifers¹². Predictably, this situation is likely to further deteriorate. As the intensification of drought and flood events induced by climate and land use change increases, the demand and dependence on groundwater for human consumption, agricultural irrigation, and environmental water needs will also escalate^{13,14}. Furthermore, salinization and contamination of groundwaters by persistent organic pollutants such as nitrate, heavy metals, oil, and microplastics is a major threat to diverse subterranean ecosystems and, in turn, to the integrity of the global water cycle¹⁰. Subterranean waters are often old: once meteoric waters enter subterranean systems, it may take months, years, and sometimes millennia before they resurface¹⁵. Hence, there is often a generational lag between contamination event and effect, and even major conservation efforts might take an epoch before these ecosystems recover if groundwater quality deterioration is not urgently considered. Ultimately, we risk compromising the insurance policy of life on Earth: the largest body of liquid freshwater.

Despite growing concerns over global groundwater depletion and degradation, and the feedback effect on diverse surface ecosystems, subterranean ecosystems remain the dark exotic siblings of surface water bodies when it comes to conservation⁷. Indeed, groundwaters have so

far been largely overlooked in global conservation policies, and biodiversity and climate change agendas for water resources^{16,17,18}. For example, as many as 85% of protected areas with GDEs have groundwater-sheds (or catchments) that are unprotected¹⁹. Foremost, this is because of the still incomplete knowledge about the spatial distribution, biodiversity, vulnerability, and biochemical processes and services of groundwater ecosystems^{17,20,21}. While divers can physically explore submerged caves and cenotes, the vast majority of subterranean water bodies are inaccessible to humans unless by indirect means^{22,23,24}. Indeed, access to groundwater organisms is often restricted to caves, wells, and springs that serve as windows to the subterranean world⁸. The real extent of groundwater ecosystems is therefore roughly estimated (between 22.6 and 23.6 million km³ in the upper 2 km of continental crust^{1,2}) and we have only a partial understanding of their three-dimensionality and verticality—i.e., structural diversity²⁵. Furthermore, as the adage “*out of sight, out of mind*” goes, there is generally poor awareness about the importance of the groundwater biodiversity and ecosystem services across policymakers, stakeholders, and the general public alike (Supplementary Section 1). This lack of awareness reflects the conservation status of groundwaters: in many areas of the world, groundwater ecosystem protection is confined to aquifers with economic value or the unplanned overlap between valuable groundwater ecosystems and protected areas established for surface ecosystems^{16,26}.

As a result, a global approach to policy that incorporates the value of groundwater ecosystems and their services is required to protect these precious resources. With this in mind, we propose the application of the keystone ecosystem concept to groundwater, as this approach has proven to be extremely valuable in nature conservation²⁷. By mapping predicted groundwater biodiversity and its overlap with surface biodiversity at global scale, we provide both conceptual and empirical evidence that this focus is scientifically sound, timely, and beneficial for the broader context of groundwater conservation. Following the GDEs categorization proposed by Eamus et al.²⁸, we focus on the ecological and functional links between groundwater ecosystems (e.g., aquifers and caves where aquatic subterranean biota reside; GDE class I) and GDEs requiring the surface expression of groundwater (e.g., wetlands and rivers; GDE class II) or GDEs dependent on groundwater availability for their biodiversity, growth and productivity (e.g., forests, GDE class III).

With the goal of taking a step further towards inter-realm approaches, we also highlight eight directions – spanning from biomonitoring to transboundary policies – to advance conservation of groundwater and groundwater-dependent ecosystems over two interlinked axes of science and policy. A much stronger focus on groundwater conservation is needed in the

face of accelerating global climate change and uncontrolled biodiversity loss, and we advocate that such a change in perspective and management strategies will consistently increase the efficacy of our global conservation strategies.

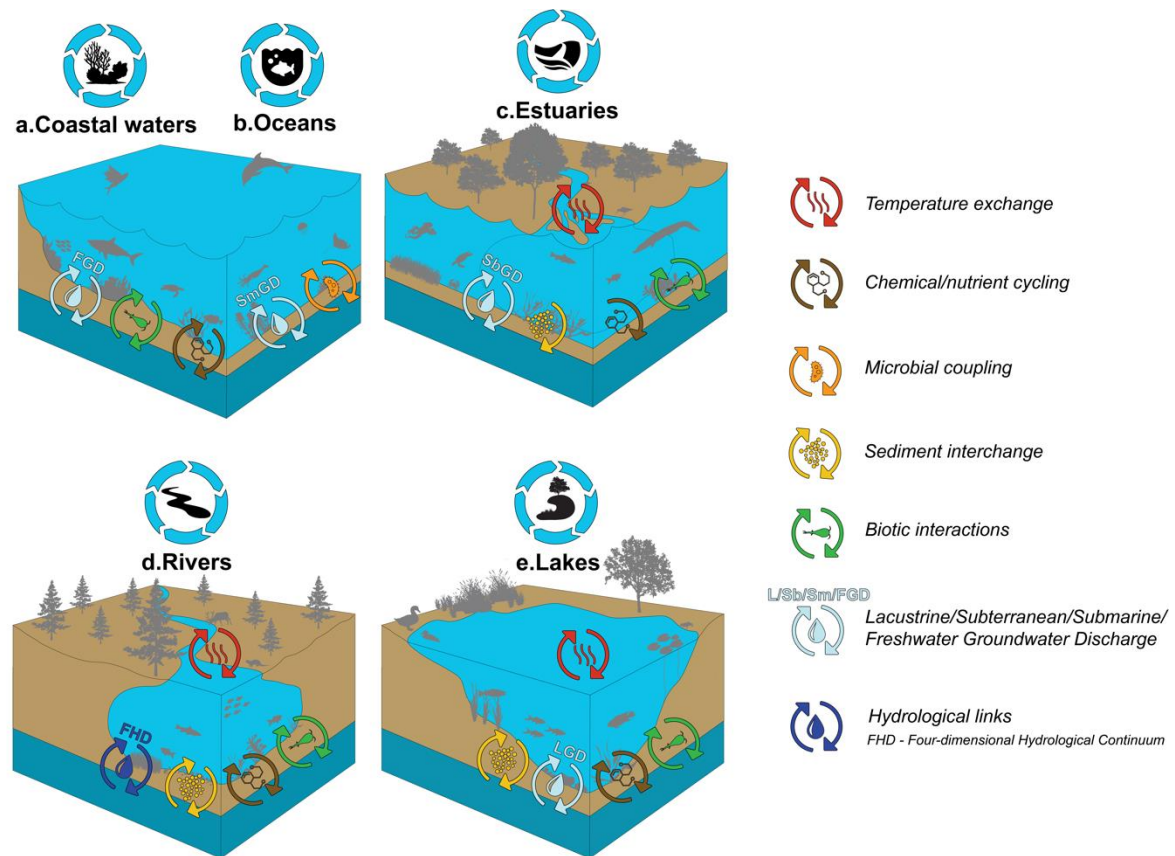


Fig. 1|Schematic representation of interactions and functional links of groundwater ecosystems (in dark blue) with the five unfrozen surface water biomes (marine and freshwater) composing the global water cycle (in light blue). See Supplementary Section 2 for a detailed description of the ecological and hydrological connections between them. For conciseness, anthropogenic impacts are not illustrated; gaps between groundwater environments and the five unfrozen surface water biomes have been added for illustrative purposes.

1. Current conservation efforts of groundwater ecosystems: the challenge of protecting the “unknown”

Comprehensive protection of groundwaters, whether direct or indirect *via* conservation of GDEs, is lacking or not implemented in most regions^{29,30}. Globally, there are only a few examples of direct conservation measures for subterranean habitats or groundwater species³⁰. Global treaties on biodiversity or conservation frequently fail to recognise groundwaters^{7,31} or

are hindered by the limited taxonomic description of most groundwater biota³⁰. The application of direct conservation measures is complicated by inconsistencies between conservation and natural resource legislation³² and often the boundaries of aquifers transcend those of jurisdictions or surface catchments that are the typical focus of land and water management¹⁹.

Until recently, direct protection and conservation measures for groundwater ecosystems have focused on protecting rare, iconic species or habitats^{21,30,33}, being generally informed by habitat mapping³⁴ and species-occurrence databases³⁵. This focus has enabled the conservation of globally significant areas^{32,36}, but is ineffective in areas where the knowledge of habitats is limited and biota are unknown or undescribed^{37,38}. Phylogenetic or functional diversity can be used to prioritise conservation sites when taxonomic information is lacking³⁹; conservation biogeography and species distribution modelling approaches also have potential as management tools⁴⁰ but are challenged by a lack of robust theoretical models to explain the distribution of biota at relevant spatial and temporal scales³⁰ and the high endemism typical of groundwater fauna⁴⁰.

The sustainable management of groundwater resources has been insufficient in protecting groundwater ecosystems, partly because its primary focus is the availability of water for humans rather than the ecological needs of the organisms therein. Although limiting groundwater allocations indirectly benefits groundwater ecosystems, this anthropocentric focus often ignores the quality and quantity of water needed for maintaining ecosystem processes^{41,42}. Groundwater vulnerability mapping⁴³ has promise as a means for assessing and managing risks to groundwaters but is generally more focused on a single resource protection than ecosystem protection. This is problematic because only through the preservation of healthy groundwater biota, including both microbes and metazoans, can we ensure the maintenance of key ecological processes and the functional links with surface water ecosystems (Fig. 1).

Ultimately, groundwater and connected GDEs should be managed and conserved together, under a “one water” framework^{44,45}. However, human needs often triumph over environmental water needs where knowledge is limited⁴⁶, rendering this an unrealistic option for conservation. As a result, other approaches must be explored and implemented to ensure the preservation of a healthy groundwater ecosystem. Like climate change more broadly, current inaction (“too little”) is not only generating increased contamination, habitat fragmentation, and higher rates of biodiversity loss, but also risks compromising the efficacy of our future actions (“too late”) because they will be implemented on already deteriorated groundwater ecosystems.

2. Shaping groundwater as a keystone ecosystem

Assessment, monitoring and management of biodiversity frequently relies on the use of community representatives such as flagship, umbrella and keystone species, whose protection effectively preserves many other species⁴⁷. While all these proxy species approaches are constantly constrained by their intrinsic metaphorical nature⁴⁸, the emphasis of the keystone species on links among species has been raised as an “appropriate target for management”, given the implementation of this approach can provide a good compromise between species-oriented and ecosystem function-oriented conservation strategies⁴⁹.

Initially coined by Robert T. Paine (1933–2016), the term “keystone species” was intended for species of high trophic status, whose activities exert disproportionate influence on the structure and function of biological communities^{50,51}. This concept argues that a single top predator indirectly controls resource-use at lower trophic levels. Upon its removal, one species would monopolize resources, exclude competitor species, and cause a decline of biodiversity⁵². The use of keystone or any other proxy species in nature conservation is frequently advocated for systems where the number of species being protected or monitored is uncertain⁵³, such as groundwater⁵⁴. However, while keystone species appear to be a promising approach for protection and monitoring of groundwater ecosystems, its implementation is hindered by conceptual and applied issues (Box 1).

BOX 1: Keystone species in groundwater ecosystems: an impossible task?

There are many obstacles to the implementation of the concept of keystone species in groundwater ecosystems, emphasizing the need to adopt a “keystone ecosystem” approach. The first, main challenge lies in the identification of appropriate keystone species. The term “keystone” has been broadly debated^{55,56} and refined such that it could apply to all species from any trophic level. The ultimate recognition of keystone species, however, remains a two-step procedure that first applies operational criteria to identify keystone candidates, and then empirically tests how their removal impacts species diversity in a community⁵⁶. Nonetheless, the application of this procedure to groundwater is theoretically questionable and technically challenging because a clear picture of trophic structure for all GDEs is missing. For example, until recently, groundwater was considered a bottom-truncated ecosystem, with no primary producers and few specialized top predators⁵⁷. Since then, some evidence for trophic specialization within trophic levels has been identified^{58,59,60}, as well as for multiple trophic levels within species-rich groundwater communities^{61,62,63,64} making it difficult to identify suitable keystone species in most cases.

Second, there is a remarkably high frequency of narrow range endemics among groundwater species⁶⁵. High spatial turnover in groundwater species composition at larger geographical scales emerges as a consequence of the dominance of species with small distributional ranges^{66,67}. Identifying keystone species on a scale of some ten kilometres is often an impossible task.

Third, the vertical dimension of groundwater exacerbates the aforementioned issues. Groundwater is not a homogenous habitat, but an array of interconnected habitats^{68,69}. In groundwater ecosystems, life has evolved to use space in three dimensions. In karstic massifs alone, at the same geographic point, species from fissure systems in the unsaturated zone live under different environmental conditions to species from the permanently flooded zone⁷⁰, leading to vertically stratified communities. Such vertically distributed communities may be only weakly connected functionally, with predators in lower zones hardly influencing dynamics in upper zones.

The extension of the keystone concept to communities or ecosystems⁷¹ is a plausible area to explore for easing some of the current roadblocks in groundwater conservation efforts (Supplementary Section 1). Since the early 1990s, conservation strategies across the globe have shifted their focus from species- to habitat/ecosystem-level⁷². Complementarity between both approaches has been recognized as beneficial⁷², but overall, the increased cost-effectiveness and elaboration of more effective management guidelines are reported for the ecosystem-level

focus⁷³, as well as reducing funding bias⁷⁴. The value of this approach is enhanced when applied to groundwater habitats, where biodiversity is still mostly spared from macro-organismal invasive species possibly due to the selective conditions and isolation of these environments⁷⁵. As a result, compared to other surface counterparts such as rivers and lakes, groundwaters can be broadly considered less biologically degraded (even if still mostly unprotected worldwide) ecosystems, a common prerogative for conservational purposes through keystone ecosystems approaches⁷¹.

Concurrently, recent investigations into GDEs (class II and III according to Eamus et al.²⁸) indicate that they are widely distributed in dry climate zones (accounting for almost a third of the total global surface area⁷⁶), and groundwater supports riparian and floodplain vegetation in tropical and temperate zones⁷⁷. Globally, groundwater has strong physical/ecological relationships with surface water (e.g., intermittent streams), and the presence of surface water in some geographic areas is highly related (at least in some periods of the year) to groundwater level (e.g., groundwater-fed streams in semi-arid areas)²⁸. For instance, shallow groundwater influences 22 to 32% of global land area, and 15% of groundwater-fed surface water features and plant rooting zones⁷⁸.

Similar to the transition from species- to ecosystem-level conservation agendas, the shift from local to regional and continental studies in groundwater ecology has been undoubtedly enabled by the increased availability of data, combined with the enhanced awareness of the importance of groundwater at global scale¹⁹. As a result of all these observations, groundwater provides a uniquely valid conceptual candidate to be a keystone ecosystem, defined as ecological structures “*providing resources, shelter or ‘goods and services’ crucial for other species*”²⁷.

Partially due to the lack of groundwater accessibility and the resultant lack of subterranean spatial analysis, data sources for environmental parameters driving groundwater biodiversity patterns on a global scale are currently limited to estimates of water quantity (e.g., groundwater recharge and water table depth). To evaluate the potential of groundwater ecosystems as keystone ecosystems, we modelled available data to map the biodiversity of groundwater ecosystems in combination with groundwater interaction with the surface (Fig. 2). This analysis is based on an indicator composed by four proxies: three proxies that are positively associated with groundwater ecosystem biodiversity, (i) groundwater recharge⁷⁹, proxy for high biodiversity because groundwater recharge regimes are associated with the inflow of nutrients, replenishment of water, and oxygen regeneration; (ii) existence of karst⁸⁰, proxy for habitat availability and connectivity (iii) interaction between groundwater and

surface water⁸¹, another proxy for high biodiversity being a key factor in enriching oligotrophic groundwater environments with carbon loads and fresher water resources; and (iv) groundwater water table depth as negatively associated proxy to the same biodiversity factor⁷⁸ (see Supplementary Section 3 for further information).

Globally, 7.1% of the land area shows a high degree of groundwater biodiversity (90th percentile globally) and high interconnectivity to surface water bodies (90th percentile globally). 52.6% of global areas have medium to high interactions, independent to the modelled groundwater biodiversity considered. In almost a third of the global area (29.8%) there is only low (10th percentile) predicted subsurface biodiversity coupled with groundwater - surface water interaction. Within this category, a vast portion is occupied by deserts (e.g., Sahara Desert covering 8% of total global area) and high mountains, regions where the water table can be very deep (e.g., Andes), the recharge rates are very low (e.g., Arabian Desert), and/or surface environments host low biodiversity (e.g., Kalahari Desert). Once those areas with modelled low biodiversity and low interactions are removed from the global analysis, the proportion of areas with medium to high interactions jumps to 75%. Nonetheless, within these broad regions categorised as low biodiverse, important pockets of groundwater biodiversity do exist. For instance, the Pilbara in Australia is considered a major subterranean biodiversity hotspot globally⁸², and the seemingly inhospitable Sahara Desert hosts endemic species of copepods in its groundwater ecosystems⁸³. An in-depth global analysis on these “*islands under the desert*”⁸⁴ would shed further light on the understanding of functional groundwater-surface water interactions, and will only be possible once further data is gathered.

Having mapped where groundwater biodiversity is potentially high and connected to the surface, we incorporated the occurrence of surface ecosystems into the analysis (Fig. 3a-b). We combined the previous map (Fig. 2) with an indicator for surface ecosystem biodiversity (consisting of the integration of four proxies: soil bacteria, plant diversity, macrophyte occurrence and riverine species richness; Supplementary Section 3). Our goal was to estimate the overlaps and interdependence between groundwater and surficial ecosystems' biodiversity patterns. Therefore, we excluded higher-order biodiversity indicators such as avian or mammalian diversity, given that these taxa are not necessarily associated with the interlinked groundwater-surface ecosystems at a global scale. Indeed, an analysis involving groups such as marine animals⁸⁵ or reptiles⁸⁶, and modelling their degree of direct or indirect dependency/functional links with groundwater resources could be of much interest, but it lies outside of the scope of current work.

Globally, for 10.1% of the land area there is an overlap between predicted high groundwater biodiversity and interactions (90th percentile globally), and predicted high surface biodiversity (90th percentile globally). Half of global surficial area (50.0%) has high biodiversity with some extent of groundwater interactions, reaching up to 71.7% when groundwater-sheds¹⁹ are considered (see Supplementary Section 3). For all the three surface biodiversity categories (low, medium and high), the areas with the lowest groundwater biodiversity and interactions (10th percentiles) were the most abundant (8.4%, 32.1% and 23.9%, respectively). However, the choice of aggregation of Fig. 2 (compare Fig. S11) influences this outcome towards more areas with low biodiversity and interaction.

Overall, our findings suggest that global groundwater biodiversity and interactions can be considered as a first order estimator for surface biodiversity (Fig. 3a). For example, when we focussed into the Po (North Italy) (Fig. 2b and Fig. S14b) and Mekong (Southeast Asia) (Fig. 2c and Fig. S16b) river basins, two areas that in 2022 experienced the worst droughts in 70 years^{87,88}, distinctive patterns emerged. The Po basin shows a high groundwater ecosystem biodiversity close to the Alps and the Mediterranean Sea with medium interconnectivity to surface waters compared to other global systems (Supplementary Section 3). On the other hand, the Mekong shows a high groundwater ecosystem biodiversity and interconnection between groundwater and surface water. When surface biodiversity is incorporated in the modelling, the Po basin (Fig. S14b) shows hotspots of groundwater ecosystem biodiversity and surface ecosystem biodiversity closer to the delta and the pre-Alp areas. In contrast, hotspots of interconnectivity remain as in Fig. 2b. The Mekong shows extensive areas of high surface and subsurface ecosystem biodiversity together with a highly interconnected system (Fig. S16).

Groundwater and surface systems are often inter-connected, and focusing only on one, limits the effectiveness of conservation efforts. Only a holistic view that includes groundwater ecosystems will enable us to understand how excessive groundwater extraction will also affect surface ecosystems⁸⁹ and how land cover changes, e.g., deforestation, agricultural use or effect of river incision, will affect the groundwater quantity and quality and, in turn, the connected ecosystems. Without further research, the global role of groundwater in the carbon cycle remains unclear. When prioritizing areas for biodiversity conservation, integrating surface and groundwater biodiversity is more effective⁹⁰. Combined protection of surface and subsurface areas is most efficient in terms of costs, available space, and societal awareness. Recognizing groundwater as a keystone ecosystem highlights the cascading effects that would be triggered if we further contaminate and/or deplete groundwater. While some authors have already discussed the hydrological transboundary role of groundwater at global scale⁹¹, to the best of

our knowledge, this is the first ecological quantification of groundwater ecosystems' relevance for the Earth system.

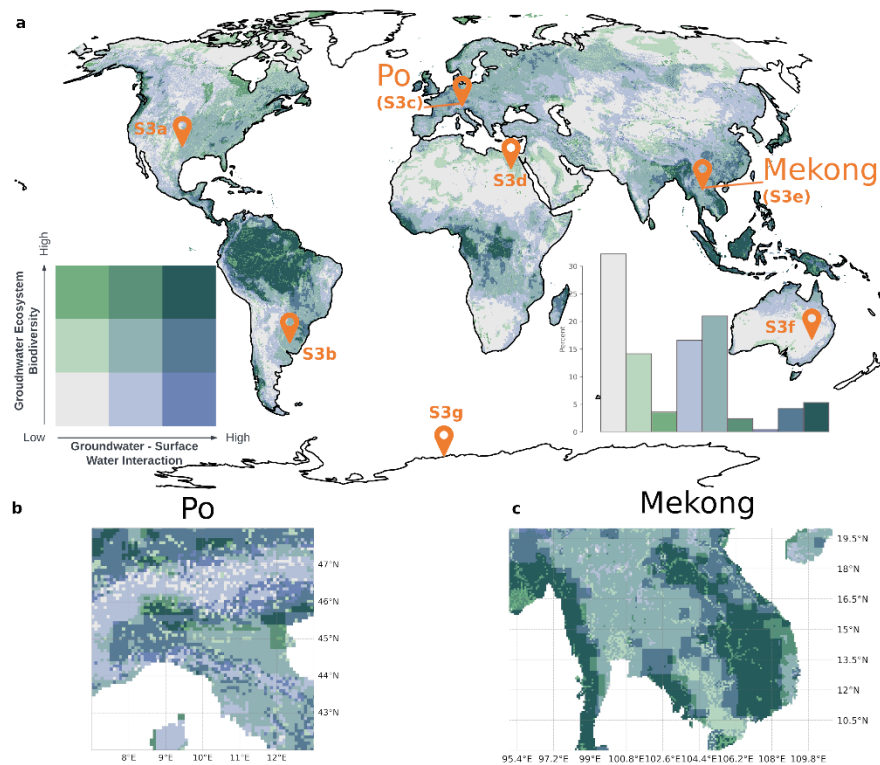


Fig. 2|Linkages between predicted groundwater ecosystem biodiversity and groundwater-surface water exchange fluxes. Dark green areas indicate a high groundwater ecosystem biodiversity and a high interaction between groundwater and surface water. Light green indicates areas with high groundwater biodiversity but low interactions, blue indicates high interactions (in both directions) between surface water and groundwater but low groundwater biodiversity. Groundwater ecosystem biodiversity is approximated by groundwater recharge, karst and water table depth. The interactions between groundwater and surface water are based on a global groundwater model. The categories of biodiversity and exchange fluxes are based on quantiles of normalized data. Orange markers identify focus regions used to evaluate the map. (See Supplementary Section 3 for an in-depth development and discussion of this figure).

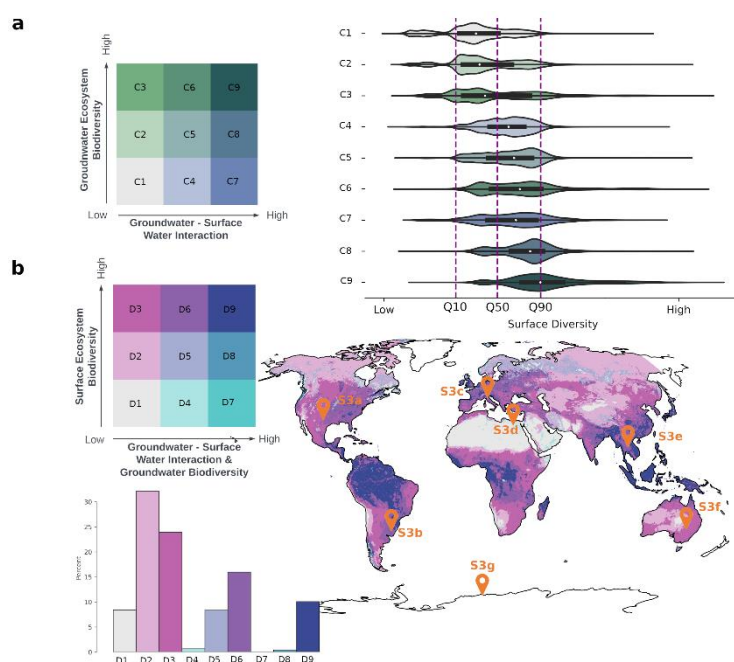


Fig. 3|Linkages between predicted surface ecosystem biodiversity and connected groundwater biodiversity. Here we show how categories of groundwater biodiversity and interaction (Fig. 2) relate to surface ecosystem biodiversity. With higher groundwater ecosystem biodiversity and interaction (C1 = lowest; C9 = highest), surface ecosystem biodiversity increases as well a). This relationship is mapped into nine new categories of surface ecosystem biodiversity and groundwater ecosystem biodiversity and interaction (D1 = lowest; D9 = highest) shown on a global map b). Dark blue in b) indicates areas of high ecosystem biodiversity, high groundwater ecosystem biodiversity, and high interactions between groundwater and surface water. Pink areas indicate only a high surface biodiversity, and turquoise, areas without large surface ecosystem biodiversity. Groundwater-surface interactions and groundwater ecosystem biodiversity are based on Figure 2. Surface ecosystem biodiversity is based on soil bacteria, fish diversity, macrophyte diversity, and vascular plant diversity, and the biodiversity categories are based on normalized data quantiles (see Supplementary Section 3).

3. Setting the ground(water) for a more effective protection of aquatic subterranean ecosystems

The success of groundwater conservation in the 21st century will be contingent on our ability to limit climate change⁹², minimize contamination⁴, and reduce overexploitation of natural resources⁹³. However, the magnitude of the challenge ahead is in stark contrast with ongoing

conservation inaction^{16,21,37}. Amidst an increasingly unpredictable climate, widespread aridification, and scattered rainfall events⁹⁴, many rivers and lakes are transitioning from permanent to intermittent⁹⁵, glaciers and snowfields are melting away, and thus two major freshwater sources are rapidly disappearing across several regions⁹⁶. As a result, the reliance of surficial watersheds on aquifers is increasing, with aquifers providing the only permanent (if replenished) freshwater resource available for many areas worldwide. Given the uneven distribution of global groundwater⁹⁷, inequitable access, and the limited replenishment of ancient global groundwater reserves, shifts in the dependence of ecosystems from surface to groundwater will be spatially variable⁹⁸. Therefore, effective groundwater governance will be a crucial aspect to mitigate the impact of droughts on economies, societies, and diverse environments⁹⁹.

Recent research has demonstrated that groundwater ecosystems and their biota actively assimilate terrigenous carbon¹⁰⁰, acting as carbon sinks¹⁰¹ analogous to freshwater wetlands. Hence, maintaining the carbon assimilation capacity of groundwater ecosystems is essential to maximise the terrestrial carbon sink and minimise climate change effects. Aquifers are also crucial for maintaining surface environments¹⁰², including their biodiversity, within natural and anthropogenic contexts¹⁰³ (Fig. 4). However, current lack of implementation of effective groundwater management strategies is hindering also the preservation of associated GDEs. The development of biodiversity indices for groundwater ecosystems, similarly to biodiversity variables proposed to monitor biodiversity at global levels^{104,105} and for discrete targeted purposes¹⁰⁶, could provide a solution to overcome this roadblock. By initially targeting well-studied regions with comprehensive diversity datasets (e.g., the Krim region in Slovenia¹⁰⁷ or the Pilbara in Western Australia⁸⁹) regional biodiversity indices can be designed, with the goal to expand the foci as groundwater biodiversity data from less studied systems become available.

Overall, our analysis emphasizes the high interconnectedness between groundwater and surface systems, and demonstrates how focusing only on one compartment limits the effectiveness, scope, and comprehensiveness of conservation efforts. To achieve more holistic conservation strategies, we will need to find effective strategies able to overcome the surface-subterranean divide. With this in mind, we advocate for a two-tiered approach for the conservation of groundwaters, composed by science and policy, and we propose eight key focal areas to develop an effective global strategy.

(i) Create standardized global datasets. Global dataset's record information on groundwater fauna is abundant, but generally scattered across myriad databases, publications, and personal

datasets, often not openly accessible and lacking inter-operability due to different data standards and vocabularies. Two ongoing ambitious projects, the World Register for marine Cave Species (WoRCS)¹⁰⁸ and *Stygofauna Mundi*⁹, aim to create a centralized, openly available, and comprehensive taxonomic and ecological database of all groundwater organisms. If successful, this will break a major barrier hampering conservation, offering much-needed data for accurate assessments of global groundwater biodiversity and providing information for evidence-based conservation²¹. Similar to rivers and lakes, integration of this information with available hydrogeological data will directly enhance the quality of groundwater environmental assessments. At transboundary ecosystem levels, published global data on the distribution of GDEs are not available to date. However, successful initiatives such as the Australian GDE Atlas¹⁰⁹ provide a promising initial step towards the creation of a scientifically sound global GDE map. Like in other disciplines, application of FAIR Data Principles¹¹⁰ to all global groundwater-based generated data should be ensured, assuring effective findability, accessibility, interoperability, and reuse of these digital assets.

(ii) Test and apply novel biomonitoring approaches. Novel biomonitoring of groundwater and its typical biota is a crucial aspect of environmental management, as many ecosystem services are dependent on a healthy environment and diversity of species that, despite being almost invariably overlooked, are irreplaceable⁵. While monitoring of physical-chemical properties or chemical pollutants in groundwater is a regular practice across the world, the biota are often overlooked if not in connection with pollutant contamination. Therefore, novel tools are required to monitor these ecosystems. Particularly promising is the use of DNA extracted from environmental samples (environmental DNA or eDNA)¹¹¹ to assess diversity of, and map the distributions of, species¹¹². First applications of eDNA to groundwater systems have been promising, recovering vast biodiversity hitherto mostly undocumented^{113,114,115,116}. For selected taxa such as subterranean salamanders and cavefish, bioacoustics, the study of animal sounds, can be used to not only detect species, but also inform on their welfare and behaviour^{117,118}.

(iii) Advance science to better understand ecosystem function. Capturing the entire diversity of subterranean species is currently not logistically feasible. For instance, it is estimated that 80% of the world's biggest subterranean biodiversity hotspot region, Western Australia, is undescribed¹¹⁹. Therefore, traditional diversity metrics may not provide a mechanistic understanding of disturbance effects¹²⁰. To circumvent this, the use of trait-based (functional) methods is gaining ground in recent ecological studies. This approach highlights how functional traits (intended, in a broad sense, as morphological, ecological, physiological, behavioural features measured at the species level¹²¹) mediate a species' ability to respond to

changes in their environment^{122,123}. However, functional studies targeting groundwater ecosystems are still rare^{89,124}. At a global level, an in-depth and groundwater-specific functional characterisation proposed by Keith et al.¹²⁵ could be informative. Microbes and aquatic invertebrates are essential for subterranean ecosystem functioning, contributing to nutrient cycling, energy flow, water filtration, and biodiversity^{8,126,127}. Therefore, targeting these components of underground aquatic ecosystems unveils crucial aspects of functioning and resilience.

(iv) Involve interdisciplinary approaches: A cross-pollination of ideas among researchers from different scientific backgrounds — e.g., hydrologists, hydrogeologists, climatologists, geochemists, ecologists and taxonomists — and operating both above and below the ground would enhance the implementation of conservation interventions able to embrace the entirety of the surface-subterranean continuum. Some possible ways forward to break the artificial divide between surface- and subterranean-based scientists and foster cooperation could include: a) limiting discipline-specific jargon in communication¹²⁸; b) broadening reading habits outside one's own niche expertise; c) seeking active collaboration by exposing oneself to different scientific cultures (e.g., by attending scientific meeting outside one's own expertise) and d) fostering open data policies to ensure data exchange among researchers, groups, and companies as well as data availability for future generations.

(v) Implement global policies to protect transboundary waters. Conservation of biodiversity often requires operating across country boundaries¹²⁹, an endeavour often complicated by bureaucracy and geopolitical instability^{130,131}. Worldwide, 468 transboundary aquifers (namely aquifers crossing multiple states¹³²) have been delineated¹³³, several of which are subject to mounting human pressure¹³⁴. However, there is currently no specific global convention or law for the management of transboundary aquifers. Today, transboundary aquifers are still governed by the 1997 UN Watercourses Convention which applies to groundwater systems, “[...] *but only to the extent that an aquifer is connected hydrologically to a system of surface waters, parts of which are situated in different States*”¹³⁵. Transboundary aquifers cooperation is still lagging as it is directly related to the capacity of the States to understand and value the groundwater systems and associated ecosystems they depend upon. Efforts should be made on valuing groundwater as a shared resource beyond frontiers—e.g., by reporting evidence of anthropogenic impact on transboundary groundwater ecosystems to showcase and boost transboundary aquifers' cooperation¹³⁶.

(vi) Improve water management and governance. It is essential to achieve a more balanced effort (both financial and conservational) to the management of the different components of

the hydrosphere and biosphere. The historical focus on surface water in freshwater management⁹³, in part reflects knowledge deficits on the role of groundwater ecosystems at the time when the main freshwater policies were set up¹³⁷ and the lack of ability in updating and adjusting strategies as scientific research progresses¹³⁸ (Supplementary Section 1). Now, thirty years after the publication of the cornerstone book “The Freshwater imperative”¹³⁹, inter-realm monitoring and management are more imperative than ever¹⁴⁰. It is just a matter of treasure lessons learned, expanding views, and being ambitious¹⁴¹. Most ecosystems will benefit from this timely (almost overdue) shift in perspectives.

(vii) Develop restoration and monitoring programs. Hydrogeological restoration of aquifers¹⁴² and surface-groundwater interactions¹⁴³ have been the focus of extensive research over the last three decades, yet studies on the ecological restoration of groundwater ecosystems are still rare¹⁴⁴. As data on groundwater biodiversity and resilience to contamination and climate change are gathered, integration of comprehensive biotic-driven restoration guidelines is essential for the effective management of groundwater pollution both in natural and anthropogenic contexts³.

(viii) Encourage participatory approaches. The value of a natural resource is only acknowledged when citizens are involved¹⁴⁵. Alther et al.¹¹³ and Raghavan et al.¹⁴⁶ employed participatory approaches to raise awareness on the importance of aquatic subterranean fauna, in projects that also led to the discovery of new species (amphipod genus *Niphargus* and catfish *Horaglanis populi*). Extension and upscaling of such an initiative to other regions, countries, and continents can provide a highly effective tool to increase societal awareness and advance science. Concurrently, the incorporation of local indigenous knowledge into ecological science harbours enormous potential to increase the efficacy of conservation and management strategies¹⁴⁷. For instance, by harnessing the power of local knowledge through participatory science programs, the opportunity exists to build up a database of active and inactive global spring locations¹⁴⁸. Such community-led monitoring programs could also provide information about groundwater quality (levels of eutrophication and contamination) and provide the catalyst to building a groundswell of support for rehabilitating and restoration of inactive springs to benefit surface and subsurface biodiversity.

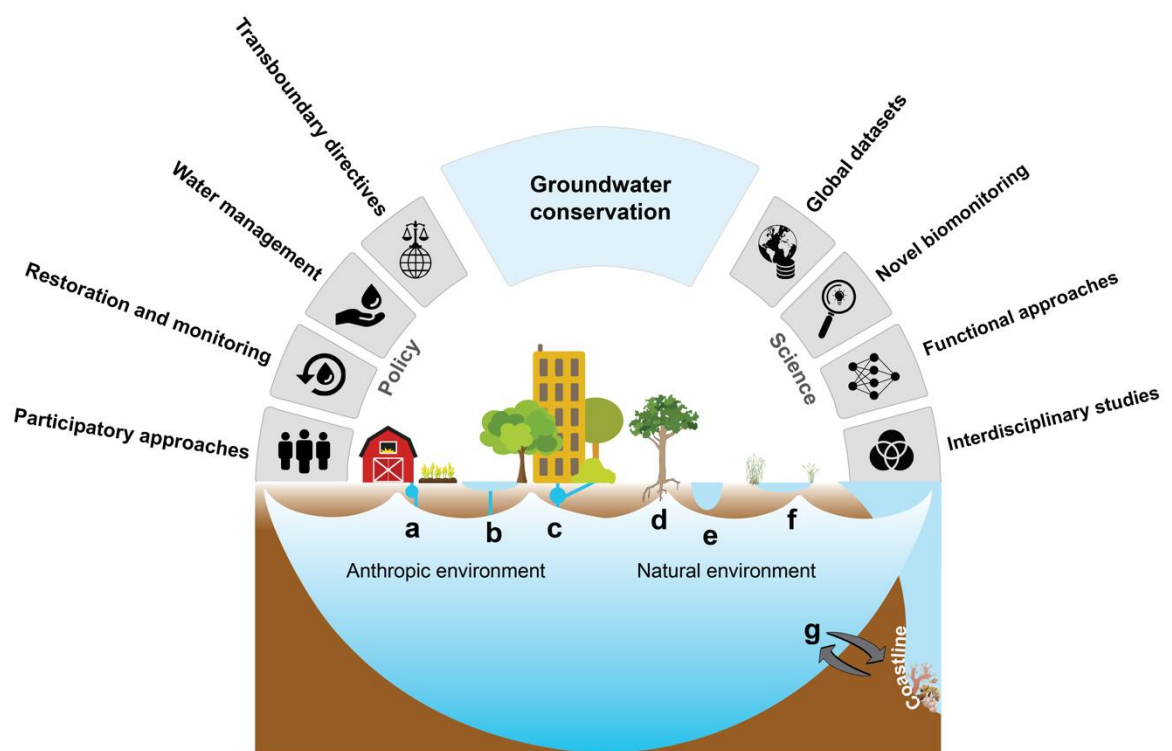


Fig. 4|Examples of groundwater ecosystem services within anthropic (a, b, c) and natural (d, e, f, g) frameworks and recommended guidelines for groundwater conservation in terms of scientific advancements (top right) and policy developments (top left). Anthropic environment: a, clean groundwater plays a key role in maintaining the agrobiodiversity¹⁴⁹; b, interchanges between urban wetlands and groundwater can maintain the diversity of aquatic species and the functional integrity of urban wetlands¹⁵⁰; c, water for urbanisation can also supply a key resource for the maintenance of urban vegetation¹⁵¹; natural environment: d, terrestrial vegetation groundwater dependent ecosystem (GDE)¹⁵²; e, lotic GDE¹⁵³; f, lentic GDE¹⁵⁴; g, coastal GDE¹⁵⁵.

Conclusions

Water is the basis of life on Earth: by overlooking the ecological integrity of groundwater, we are threatening the long-term prospects of entire ecosystems and ultimately of humanity itself. Too often, current conservation efforts consider groundwater as disjoint from the rest of the components of the global water cycle, despite the multiple functional interlinks between the subterranean, surface and atmospheric realms. The application of the keystone ecosystem concept to groundwater would enable breaking the conceptual and mechanistic barriers still existing in water science and policy. We provide evidence that most habitable global areas

(75%) have a medium to high level of ecological interactions with groundwater. We also provide the first indication that groundwater biodiversity and interconnections can represent an ecological estimator for global surface biodiversity patterns. Given this foundation, conservation and water resource policies are pivotal to assure the maintenance of the essential ecosystem services provided by groundwater ecosystems worldwide. We argue that the overall benefits of this approach extend beyond the dark underworld, allowing the preservation of diverse terrestrial and aquatic ecosystems. This is urgent for wise water management plans within the current climate change scenario, considering that many regions across the globe are already experiencing a water crisis.

Acknowledgments

The authors would like to acknowledge the broader groundwater ecology community for its persistence in raising awareness of the importance of groundwater ecosystems for life on Earth.

Author contributions

This study was conceived by M.S., S.M. and R.R. The methods were developed by M.S. and R.R.

Analysis was performed by R.R. Figures were developed by M.S. and R.R. Paper writing was led by M.S. with input from all authors. All authors discussed the results and edited the paper at multiple stages.

Competing interests

The authors declare no competing interests.

Data availability

This study uses data from multiple open-access datasets. Source data are documented in Supplementary Section 2 and can be downloaded from the persistent web-links provided. Data produced in this study have been deposited on Zenodo at the address doi: 10.5281/zenodo.7924816.

Funding

This project received support from BHP-Curtin alliance within the framework of the “eDNA for Global Environment Studies (eDGES)” programme, and Biodiversa+, the European Biodiversity Partnership under the 2021-2022 BiodivProtect joint call for research proposals,

co-funded by the European Commission (GA N°101052342) and with the funding organisations Ministry of Universities and Research (Italy), Agencia Estatal de Investigación – Fundación Biodiversidad (Spain), Fundo Regional para a Ciência e Tecnologia (Portugal), Suomen Akatemia – Ministry of the Environment (Finland), Belgian Science Policy Office (Belgium), Agence Nationale de la Recherche (France), Deutsche Forschungsgemeinschaft e.V. – BMBF-VDI/VDE INNOVATION + TECHNIK GMBH (Germany), Schweizerischer Nationalfonds zur Forderung der Wissenschaftlichen Forschung (Switzerland), Fonds zur Förderung der Wissenschaftlichen Forschung (Austria), Ministry of Higher Education, Science and Innovation (Slovenia), and the Executive Agency for Higher Education, Research, Development and Innovation Funding (Romania). S.M. and T.D.L. acknowledge the support of NBFC to CNR, funded by the Italian Ministry of University and Research, P.N.R.R., Missione 4 Componente 2, “Dalla ricerca all’impresa”, Investimento 1.4, Project CN00000033. S.M. was further supported by the PRIN DEEP CHANGE (2022MJSYF8), funded by the Italian Ministry of Education, University and Research. A.S.P.S.R was supported by the VILLUM FONDEN (research grant 15471) and by Portuguese National Funds through “Fundação para a Ciência e a Tecnologia” (FCT) within the cE3c Unit funding UIDB/00329/2020. A.B. was supported by national Slovenian program P1-0255 financed by the Slovenian Research Agency (ARRS). F.A. has been funded by the University of Zurich Research Priority Programme on Global Change and Biodiversity and the Swiss National Science Foundation Grant. Š.B., M.Z. and C.F. were funded by Slovenian Research Agency through core program P1-0184 and project J1-2464. R.L.F. was supported by the CNPq (National Council for Scientific and Technological Development, grant n. 302925/2022-8). E.L. was funded by the Water Development and Partnership Programme of the Dutch Ministry of Foreign Affairs. K.K. was supported by Australian Research Council (ARC) grant LP190100927. S.J.B.C., M.T.G. and W.F.H. were funded by ARC grants LP190100555, DP180103851 and DP230100731. F.M. was supported by the French National Research Agency projects CONVERGENOMICS (ANR-15-CE32-0005) and EUR H20’Lyon project (ANR-17-EURE-0018).

References

1. Gleeson, T., et al. The global volume and distribution of modern groundwater. *Nature Geoscience* **9.2**, 161–167 (2016).
2. Ferguson, G., et al. Crustal groundwater volumes greater than previously thought. *Geophysical Research Letters* **4**, e2021GL093549 (2021).
3. Scanlon, B. R., et al. Global water resources and the role of groundwater in a resilient water future. *Nature Reviews Earth & Environment* **4**, 87–101 (2023).

Work highlighting dire declining in groundwater levels and proposing new approaches

4. United Nations. The United Nations World Water Development Report 2022: groundwater: making the invisible visible (UNESCO, Paris, 2022).
5. Griebler, C. & Avramov, M. Groundwater ecosystem services: a review. *Freshwater Science* **34**, 355–367 (2015).

Review compiling global groundwater services and pinpointing crucial research gaps

6. Canedoli, C., et al. Integrating landscape ecology and the assessment of ecosystem services in the study of karst areas. *Landscape Ecology* 1–19 (2022).
7. Howarth, F. G. & Moldovan, O. T. The ecological classification of cave animals and their adaptations. *Cave ecology* 41–67 (2018).
8. Malard, F., Griebler, C. & Retaux, S. Groundwater Ecology & Evolution, 2nd Edition, Elsevier (2023).
9. Martinez, A., et al. A new insight into the Stygofauna Mundi: assembling a global dataset for aquatic fauna in subterranean environments. *ARPHA conference abstracts*. Vol. 1. Pensoft Publishers, 2018.
10. Castaño-Sánchez, A., Hose G. C. & Reboleira, A. S. P. S. Ecotoxicological effects of anthropogenic stressors in subterranean organisms: A review. *Chemosphere* **244**, 125422 (2020).
11. Wada, Y., et al. Global depletion of groundwater resources. *Geophysical research letters* **37**, 20 (2010).
12. Gleeson, T., Wada, Y., Bierkens, M. F. & Van Beek, L. P. Water balance of global aquifers revealed by groundwater footprint. *Nature* **488**, 197–200 (2012).
13. Condon, L. E., Atchley, A. L. & Maxwell, R. M. Evapotranspiration depletes groundwater under warming over the contiguous United States. *Nature communications* **11**, 873 (2020).
14. Wu, W-Y., et al. Divergent effects of climate change on future groundwater availability in key mid-latitude aquifers. *Nature communications* **11**, 3710 (2020).

15. Jasechko, S., et al. The pronounced seasonality of global groundwater recharge." *Water Resources Research* **50**, 8845–8867.

16. Sánchez-Fernández, D., Galassi, D. M., Wynne, J. J., Cardoso, P. & Mammola, S., Don't forget subterranean ecosystems in climate change agendas. *Nature Climate Change* **11**, 458–459 (2021).

Correspondence on the importance of groundwater under current climate change

17. Wynne, J. J., et al. A conservation roadmap for the subterranean biome. *Conservation Letters* **14**, e12834 (2021).

18. Fišer, C., et al. The European green deal misses Europe's subterranean biodiversity hotspots. *Nature Ecology & Evolution* **6**, 1403–1404 (2022).

19. Huggins, X, et al. Overlooked risks and opportunities in groundwatersheds of the world's protected areas. *Nature Sustainability* 1–10 (2023).

Work quantifying the lack of protection for groundwatersheds and depending habitats

20. Gerovasileiou, V. & Bianchi, C. N., Mediterranean marine caves: a synthesis of current knowledge. *Oceanography and Marine Biology: An Annual Review* **59**, 1–88 (2021).

21. Mammola, S., et al. Towards evidence-based conservation of subterranean ecosystems. *Biological reviews* **97**, 1476–1510 (2022).

Review on quantitative approaches for subterranean conservation and future avenues

22. Ficetola, G. F., Canedoli, C. & Stoch, F. The Racovitzan impediment and the hidden biodiversity of unexplored environments. *Conservation Biology* **33**, 214–216 (2019).

23. Navarro-Barranco, C., et al. Conservation of dark habitats. *Coastal Habitat Conservation*. Academic Press, 147-170 (2023).

24. Saccò, M., et al. Stygofaunal diversity and ecological sustainability of coastal groundwater ecosystems in a changing climate: The Australian paradigm. *Freshwater Biology* **67**, 2007–2023 (2022).

25. Fei, S., LaRue, E. A., Hardiman, B. S. & Dahlin, K. M. Structural diversity: a digital revolution. *Frontiers in Ecology and the Environment* **21**, 3–3 (2023).

26. Giakoumi, S., et al. Ecoregion-based conservation planning in the Mediterranean: dealing with large-scale heterogeneity. *PloS one* **8**, e76449 (2013).

27. Tews, J., et al. Animal species diversity driven by habitat heterogeneity/diversity: the importance of keystone structures. *Journal of biogeography* **31**, 79–92 (2004).

28. Eamus, D., Froend, R., Loombes, R., Hose G. C. & Murray B. R. A functional methodology for determining the groundwater regime needed to maintain health of groundwater dependent ecosystems. *Australian Journal of Botany* **54**, 97-114 (2006).

29. Famiglietti, J. S. The global groundwater crisis. *Nature Climate Change* **4**, 945–948 (2014).

Work highlighting global rapid rates of groundwater depletion and analysis of factors

30. Boulton, A. J., et al. Recent concepts and approaches for conserving groundwater biodiversity. *Groundwater Ecology and Evolution* 525–550 (2023).

31. Iannella, M., et al. Getting the ‘most out of the hotspot’ for practical conservation of groundwater biodiversity. *Global Ecology and Conservation* **31**, e01844 (2021).

32. Devitt, T. J., Wright, A. M., Cannatella, D. C. & Hillis, D. M. Species delimitation in endangered groundwater salamanders: Implications for aquifer management and biodiversity conservation. *Proceedings of the National Academy of Sciences* **116**, 2624–2633 (2019).

33. Moldovan, O. T. Cave protection in Romania. In: Ponta, G.M., Onac, B.P. (Eds.), *Cave and Karst Systems of Romania*. Springer, Cham, 537e541 2019.

34. Cornu, J-F., Eme D., & Malard F. The distribution of groundwater habitats in Europe. *Hydrogeology Journal* **21**, 949 (2013).

35. Zagnajster, M., et al. Geographic variation in range size and beta diversity of groundwater crustaceans: insights from habitats with low thermal seasonality. *Global ecology and biogeography* **23**, 1135–1145 (2014).

36. Iannella, M., et al. Jumping into the grids: mapping biodiversity hotspots in groundwater habitat types across Europe. *Ecography* **43**, 1825–1841 (2020).

37. Mammola, S., et al. Scientists' warning on the conservation of subterranean ecosystems. *BioScience* **69**, 641–65 (2019).

38. Raghavan, R., Britz, R. & Dahanukar, N. Poor groundwater governance threatens ancient subterranean fishes. *Trends in Ecology & Evolution* **36**, 875–878 (2021).

39. Asmyhr M. G., Linke, S., Hose G. C., & Nipperess, D. A. Systematic conservation planning for groundwater ecosystems using phylogenetic diversity. *Plos One* **9**, e115132 2014.

40. Mammola, S. & Leroy, B. Applying species distribution models to caves and other subterranean habitats. *Ecography* **41**, 1194–1208 (2018).

41. Howard, J. K., Dooley, K., Brauman, K. A., Klausmeyer, K. R. & Rohde, M. M. Ecosystem services produced by groundwater dependent ecosystems: a framework and case study in California. *Frontiers in Water* **5**, 1115416 (2023).

42. Korbel, K. L. & Hose, G. C. A tiered framework for assessing groundwater ecosystem health. *Hydrobiologia* **661**, 329–349 (2011).

43. Machiwal, D., Jha, M. K., Singh, V. P. & Mohan, C. Assessment and mapping of groundwater vulnerability to pollution: current status and challenges. *Earth-Science Reviews* **185**, 901e927 2018.

44. McNutt, M. The drought you can't see. *Science* **345**, 1543–1543 (2014).
45. Linke, S., Turak, E., Asmyhr, M. & Hose, G. C. 3D conservation planning: Including aquifer protection in freshwater plans refines priorities without much additional effort. *Aquatic Conservation: Marine and Freshwater Ecosystems* (2019).
46. Rohde, M. M., Froend, R. & Howard, J. A Global Synthesis of Managing Groundwater Dependent Ecosystems Under Sustainable Groundwater Policy. *Groundwater* **55**, 293–301 (2017).

Study on adaptive management approaches of GDE, challenges and opportunities

47. Caro, T. Conservation by proxy. Indicator, Umbrella, Keystone, Flagship and Other Surrogate Species. Island Press, Washington, Covelo, London, 374 pp (2010).
48. Barua, M. Mobilizing metaphors: the popular use of keystone, flagship and umbrella species concepts. *Biodiversity and Conservation* **20**, 1427–1440 (2011).
49. Simberloff, D. Flagships, umbrellas, and keystones: Is single-species management passe in the landscape era? *Biological Conservation* **83**, 247–257 (1998).
50. Paine, R. T. A note on trophic complexity and community stability. *The American Naturalist* **103**, 91–93 (1969a).
51. Paine, R. T. The Pisaster-Tegula Interaction: Prey Patches, Predator Food Preference, and Intertidal Community Structure. *Ecology* **50**, 950–961 (1969b).
52. Paine, R. T. Food Web Complexity and Species Diversity. *The American Naturalist* **100**, 65–75 (1966).
53. Wiens, J. A., Hayward, G. D., Holthausen, R. S. & Wisdom, M. J. Using surrogate species and groups for conservation planning and management. *BioScience* **58**, 241–252 (2008).
54. Larned, S. T. Phreatic groundwater ecosystems: research frontiers for freshwater ecology. *Freshwater Biology* **57**, 885–906 (2012).
55. Mills, L. S. & Doak, D. F. The Keystone-Species Concept in Ecology and Conservation. *BioScience* **43**, 219–224 (1993).
56. Davic, R. D. Linking keystone species and functional groups: A new operational definition of the keystone species concept. *Ecology and Society* **7** (2003).
57. Gibert, J. & Deharveng, L. Subterranean Ecosystems: A Truncated Functional Biodiversity. *BioScience* **52**, 473 (2002).
58. Francois, C. M., et al. Trophic selectivity in aquatic isopods increases with the availability of resources. *Functional Ecology* **34**, 1078–1090 (2020).
59. Francois, C. M., et al. Trophic ecology of groundwater species reveals specialization in a low-productivity environment. *Functional Ecology* **30**, 262–273 (2016).

60. Ercoli F, et al., (2019) Differing trophic niches of three French stygobionts and their implications for conservation of endemic stygofauna. *Aquatic Conservation: Marine and Freshwater Ecosystems* **29**, 2193-2203.
61. Hutchins, B. T., Summers E. A., Nowlin, W. H. & Schwartz B. F. Chemolithoautotrophy supports macroinvertebrate food webs and affects diversity and stability in groundwater communities. *Ecology* **97**, 1530–1542 (2016).
62. Saccò M., et al. Elucidating stygofaunal trophic web interactions via isotopic ecology. *PLoS One* **14**, e0223982 (2019).
63. Saccò M., et al. Refining trophic dynamics through multi-factor Bayesian mixing models : A case study of subterranean beetles. *Ecology and Evolution* **10**, 8815–8826 (2020).
64. Premate E., et al. (2021) Cave amphipods reveal co-variation between morphology and trophic niche in a low-productivity environment. *Freshwater Biology* **66**, 1876–1888 (2021):
65. Malard, F., et al. Diversity patterns of stygobiotic crustaceans across multiple spatial scales in Europe. *Freshwater Biology* **54**, 756–776 (2009).
66. Trontelj, P., et al. A molecular test for cryptic diversity in ground water: How large are the ranges of macro-stygobionts? *Freshwater Biology* **54**, 727–744 (2009).
67. Bregović, P., Fišer, C. & Zagamajster, M. Contribution of rare and common species to subterranean species richness patterns. *Ecology and Evolution* **9**, 11606–11618 (2019).
68. Culver, D. C. & Pipan, T. *Shallow Subterranean Habitats: Ecology, Evolution, and Conservation*. 1st ed. Oxford University Press, 288 pp (2014).
69. Fišer, C., Pipan, T. & Culver, D. C. The Vertical Extent of Groundwater Metazoans: An Ecological and Evolutionary Perspective. *BioScience* **64**, 971–979 (2014).
70. Culver, D. C. & Pipan, T. *The Biology of Caves and Other Subterranean Habitats*. Oxford University Press, 336 pp (2019).
71. Mouquet, N., Gravel, D., Massol, F. & Calcagno, V. Extending the concept of keystone species to communities and ecosystems. *Ecology Letters* **16**, 1–8 (2013).

Dissertation on the application of keystone ecosystems for biodiversity management

72. Lindenmayer, D., et al. The complementarity of single-species and ecosystem-oriented research in conservation research. *Oikos* **116**, 1220–1226 (2007).
73. Walker, B. & Salt, D. *Resilience thinking: sustaining ecosystems and people in a changing world*. Island press. (2012).
74. Adamo, M., et al. Dimension and impact of biases in funding for species and habitat conservation. *Biological Conservation* **272**, 109636 (2022).

75. Nicolosi, G., Mammola, S., Verbrugge, L. & Isaia, M. Aliens in caves: the global dimension of biological invasions in subterranean ecosystems. *Biological Reviews* **98**, 849–867 (2023).
76. Salem, B. B. Arid zone forestry: a guide for field technicians. No. 20. Food and Agriculture Organization (FAO) (1989).
77. Glanville, K., Sheldon, F., Butler, D. & Capon, S. Effects and significance of groundwater for vegetation: A systematic review. *Science of The Total Environment* 162577 (2023).
78. Fan, Y., Li, H & Miguez-Macho, G. Global patterns of groundwater table depth. *Science* **339**, 940–943 (2013).
79. Moeck, C., et al. A global-scale dataset of direct natural groundwater recharge rates: A review of variables, processes and relationships. *Science of the total environment* **717**, 137042 (2020).
80. Zigmajster, M., Malard, F., Eme, D. & Culver, D. C. Subterranean biodiversity patterns from global to regional scales. *Cave ecology* 195–227 (2018).
81. Danielopol, D. L., Pospisil, P. & Rouch, R. Biodiversity in groundwater: a large-scale view. *Trends in Ecology & Evolution* **15**, 223–224 (2000).
82. Saccò, M., et al. New light in the dark—a proposed multidisciplinary framework for studying functional ecology of groundwater fauna. *Science of the Total Environment* **662**, 963–977 (2019).
83. Brancelj, A. Two new stygobiotic copepod species from the Tibesti area (Northern Chad) and a re-description of *Pilocamptus schroederi* (van Douwe, 1915). *Zootaxa* **3994**, 531–555 (2015).
84. Cooper, S. J. B., Hinze, S., Leys, R., Watts, C. H. S. & Humphreys, W. F. Islands under the desert: molecular systematics and evolutionary origins of stygobitic water beetles (Coleoptera: Dytiscidae) from central Western Australia. *Invertebrate systematics* **16**, 589–590 (2002).
85. Lecher, A. L. & Mackey, K. R. Synthesizing the effects of submarine groundwater discharge on marine biota. *Hydrology* **5**, 60 (2018).
86. Bateman, H. L. & Merritt, D. M. Complex riparian habitats predict reptile and amphibian diversity. *Global ecology and conservation* **22**, e00957 (2020).
87. Bonaldo, D., et al. The summer 2022 drought: a taste of future climate for the Po valley (Italy)? *Regional Environmental Change* **23**, 1 (2023).
88. Kang, H., Sridhar, V. & Ali, S. A. Climate change impacts on conventional and flash droughts in the Mekong River Basin. *Science of The Total Environment* **838**, 155845 (2022).
89. Uhl, A., et al. Making waves: Pulling the plug—Climate change effects will turn gaining into losing streams with detrimental effects on groundwater quality. *Water Research* **220**, 118649 (2022).

90. Rohde, M. M., Sweet, S. B., Ulrich, C. & Howard, J. A transdisciplinary approach to characterize hydrological controls on groundwater-dependent ecosystem health. *Frontiers in Environmental Science* **7**, 175 (2019).
91. Gleeson, T., Cuthbert, M., Ferguson, G. & Perrone, D. Global groundwater sustainability, resources, and systems in the Anthropocene. *Annual review of earth and planetary sciences* **48**, 431–463 (2020).
92. Amanambu, et al. Groundwater system and climate change: Present status and future considerations. *Journal of Hydrology* **589**, 125163 (2020).
93. Foster, S., Chilton, J., Nijsten, G. J. & Richts, A. Groundwater—a global focus on the ‘local resource’. *Current opinion in environmental sustainability* **5**, 685–695 (2013).
94. IPCC, 2022: Climate Change 2022: Mitigation of Climate Change. Contribution of Working Group III to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change [P.R. Shukla, J. Skea, R. Slade, A. Al Khourdajie, R. van Diemen, D. McCollum, M. Pathak, S. Some, P. Vyas, R. Fradera, M. Belkacemi, A. Hasija, G. Lisboa, S. Luz, J. Malley, (eds.)]. Cambridge University Press, Cambridge, UK and New York, NY, USA.
95. Messenger, M. L., et al. Global prevalence of non-perennial rivers and streams. *Nature* **594**, 391–397 (2021).
96. Peterson, T. J., Saft, M., Peel, M. C. & John, A. Watersheds may not recover from drought. *Science* **372**, 745–749 (2021).
97. Kretschmer, D., Wachholz, A. & Reinecke, R. Global groundwater in the Anthropocene. In *Groundwater Ecology and Evolution* (pp. 483-500). Academic Press (2023).
98. López-Corona, O., et al. Playing with models and optimization to overcome the tragedy of the commons in groundwater. *Complexity* **19**, 9–21 (2013).
99. Petersen-Perlman, J. D., Aguilar-Barajas, I. & Megdal, S. B. Drought and Groundwater Management: Interconnections, challenges, and policy responses. *Current Opinion in Environmental Science & Health* 100364 (2022).

Work on groundwater policy and governance under current global drought trends

100. Hartland, A., Bury, S. & Fenwick, G. Tracing sewage-derived organic matter into a shallow groundwater food web using stable isotope and fluorescence signatures. *Marine and Freshwater Research* **62**, 119–129 (2011).
101. Chen, L., et al. Karst carbon sink processes and effects: A review. *Quaternary International* **652**, 63-73 (2023).
102. Boulton, A. J., Datry, T., Kasahara, T., Mutz, M. & Stanford, J. A., Ecology and management of the hyporheic zone: stream–groundwater interactions of running waters and their floodplains. *Journal of the North American Benthological Society* **29**, 26–40 2010.

103. Becher, J., Englisch, C., Griebler, C. & Bayer, P. Groundwater fauna downtown—Drivers, impacts and implications for subsurface ecosystems in urban areas. *Journal of Contaminant Hydrology* **248**, 104021 (2022).
104. Pereira, H. M., et al. Essential biodiversity variables. *Science* **339**, 277–278 (2013).
105. Jetz, W., et al. Essential biodiversity variables for mapping and monitoring species populations. *Nature ecology & evolution* **3**, 539–551 (2019).
106. Guerra, C.A., et al. Tracking, targeting, and conserving soil biodiversity. *Science* **371**, 239–241 (2021).
107. Sket, B., Paragamian, K. & Trontelj, P. A census of the obligate subterranean fauna of the Balkan Peninsula (pp. 309-322). Springer Netherlands (2004).
108. Gerovasileiou V, et al. World Register of marine Cave Species (WoRCS): a new Thematic Species Database for marine and anchialine cave biodiversity. *Research Ideas and Outcomes* **2**, e10451 (2016).
109. Doody, T. M., et al. Continental mapping of groundwater dependent ecosystems: A methodological framework to integrate diverse data and expert opinion. *Journal of Hydrology: Regional Studies* **10**, 61–81 (2017).
110. Wilkinson, M. D., et al. The FAIR Guiding Principles for scientific data management and stewardship. *Scientific data* **3**, 1–9 (2016).
111. Pawlowski, J., Apothéoz-Perret-Gentil L. & Altermatt, F. Environmental DNA: What's behind the term? Clarifying the terminology and recommendations for its future use in biomonitoring. *Molecular Ecology* **29**, 4258–4264 (2020).
112. Takahashi, M., et al. Aquatic environmental DNA: A review of the macro-organismal biomonitoring revolution. *Science of the Total Environment* **873**, 162322 (2023).
113. Alther, R., Bongni, N., Borko, Š., Fišer, C. & Altermatt, F. Citizen science approach reveals groundwater fauna in Switzerland and a new species of *Niphargus* (Amphipoda, Niphargidae). *Subterranean Biology* **39**, 1–31 (2021).
114. Saccò, M., et al. eDNA in subterranean ecosystems: Applications, technical aspects, and future prospects. *Science of the Total Environment* **820**, 153223 (2022).
115. Couton, M., Hürlemann, S., Studer, A., Alther, R. & Altermatt, F. Groundwater environmental DNA metabarcoding reveals hidden diversity and reflects land-use and geology. *Molecular Ecology* In press (2023).
116. van der Heyde, M., et al. Taking eDNA underground: factors affecting eDNA detection of subterranean fauna in groundwater. *Molecular Ecology Resources* In press (2023).
117. Hyacinthe, C., Attia, J. & Rétaux, S. Evolution of acoustic communication in blind cavefish. *Nature Communications* **10**, 4231 (2019).

118. Mcloughlin, M. P., Stewart, R. & McElligott, A. G. Automated bioacoustics: methods in ecology and conservation and their potential for animal welfare monitoring. *Journal of the Royal Society Interface* **16**, 20190225 (2019).
119. Guzik, M. T., et al. Is the Australian subterranean fauna uniquely diverse? *Invertebrate Systematics* **24**, 407–418 (2011).
120. Li, W., He, S., Cheng, X. & Zhang, M. Functional diversity outperforms taxonomic diversity in revealing short-term trampling effects. *Scientific Reports* **11**, 1 (2021).
121. Toussaint, A., et al. Extinction of threatened vertebrates will lead to idiosyncratic changes in functional diversity across the world. *Nature communications* **12**, 5162 (2021).
122. Palacio, F. X., et al. A protocol for reproducible functional diversity analyses. *Ecography* **11**, e06287 (2022).
123. Green, S. J., Brookson, C. B., Hardy, N. A. & Crowder, L. B. Trait-based approaches to global change ecology: Moving from description to prediction. *Proceedings of the Royal Society B: Biological Sciences* **289**, 20220071 (2022).
124. Hose, G. C., et al. Invertebrate traits, diversity and the vulnerability of groundwater ecosystems. *Functional Ecology* **36**, 2200–2214 (2022).
125. Keith, D.A., et al. A function-based typology for Earth's ecosystems. *Nature* **610**, 513–518 (2022).
126. Saccò, M., et al. Tracking down carbon inputs underground from an arid zone Australian calcrete. *Plos One* **15**, e0237730 (2020).
127. Saccò, M., Blyth, A. J., Venarsky, M. & Humphreys, W. F. Trophic interactions in subterranean environments. In *Encyclopedia of Inland Waters*, Second Edition (pp. 537-547). Elsevier (2022).
128. Martínez, A. & Mammola, S. Specialized terminology reduces the number of citations of scientific papers. *Proceedings of the Royal Society B* **288**, 20202581 (2021).
129. Liu, J., Yong, D. L., Choi, C. Y. & Gibson, L. Transboundary frontiers: an emerging priority for biodiversity conservation. *Trends in Ecology & Evolution* **35**, 679–690 (2020).
130. Sousa, R. G., Silva, J. P. D., Douda, K. & Mammola, S. The cost of war for biodiversity: a potential ecocide in Ukraine. *Frontiers in ecology and the environment* (2022).
131. Allan, J. R., et al. Navigating the complexities of coordinated conservation along the river Nile. *Science Advances* **5**, eaau7668 (2019).
132. Stephan, R. M. Transboundary aquifers: Managing a vital resource: The UNILC draft articles on the law of transboundary aquifers. Unesco (2009).
133. IGRAC Transboundary Aquifers of the World, Update 2021, Scale 1:50 000 000 (2021).

134. Wada, Y. & Heinrich L. Assessment of transboundary aquifers of the world—vulnerability arising from human water use. *Environmental Research Letters* **8**, 024003 (2013).
135. United Nation. Convention on the Law of the Non-navigational Uses of International Watercourses (1997).
136. Brancelj, A., Mori, N., Treu, F. & Stoch, F. The groundwater fauna of the Classical Karst: hydrogeological indicators and descriptors. *Aquatic ecology* **54**, 205–224 (2020).
137. EC-GWD, 2006. Directive 2006/118/EC of the European Parliament and of the Council of 12 December 2006 on the protection of groundwater against pollution and deterioration. Official Journal of the European Union L 372 (19).
138. Backhaus, T. Commentary on the EU Commission's proposal for amending the Water Framework Directive, the Groundwater Directive, and the Directive on Environmental Quality Standards. *Environmental Sciences Europe* **35**, 1–7 (2023).
139. Naiman, R. J., Magnuson, J. J., McKnight D. M. & Stanford J. A. The freshwater imperative. Island Press (1995a).
140. Bugnot, A. B., et al. Urban impacts across realms: making the case for inter-realm monitoring and management. *Science of the Total Environment* **648**, 711–719 (2019).
141. Saito, L., et al. Managing groundwater to ensure ecosystem function. *Groundwater* **59**, 322–333 (2021).

Holistic study on ecologically-based approaches to enhance the resilience of GDEs

142. Kresic, N. Groundwater resources: Sustainability, management, and restoration. McGraw-Hill Education (2009).
143. Kasahara, T., Datry, T., Mutz, M. & Boulton, A. J. Treating causes not symptoms: restoration of surface–groundwater interactions in rivers. *Marine and freshwater research* **60**, 976–981 (2009).
144. Liu, Q. & Mou, X. Interactions between surface water and groundwater: key processes in ecological restoration of degraded coastal wetlands caused by reclamation. *Wetlands* **36**, 95–102 (2016).
145. Kobori, H., et al. Citizen science: a new approach to advance ecology, education, and conservation. *Ecological research* **31**, 1–19 (2016).
146. Raghavan, R., Sundar, R. L., Arjun, C. P., Britz, R. & Dahanukar, N. Evolution in the dark: Unexpected genetic diversity and morphological stasis in the blind, aquifer-dwelling catfish *Horaglanis*. *Vertebrate Zoology* **73**, 57–74 (2023).
147. Ban, N. C., et al. Incorporate Indigenous perspectives for impactful research and effective management. *Nature Ecology & Evolution* **2**, 1680–1683 (2018).

148. Goodall, H. Riding the tide: Indigenous knowledge, history and water in a changing Australia. *Environment and history* **14**, 355–384 (2008).
149. Trajkova, F., Arsov, S. & Gudeva, L. K. The role and importance of agrobiodiversity for agriculture. *Journal of Agriculture and Plant Sciences* **19**, 47–64 (2021).
150. Ameli, A. A. & Creed, I. F. Groundwaters at risk: wetland loss changes sources, lengthens pathways, and decelerates rejuvenation of groundwater resources. *JAWRA Journal of the American Water Resources Association* **55**, 294–306 (2019).
151. Marchionni, V., et al. Groundwater buffers drought effects and climate variability in urban reserves. *Water Resources Research* **56**, e2019WR026192 (2020).
152. Shukla, J., Dhyani, S., Pujari, P. & Verma, P. Groundwater-Dependent Vegetation to Address the Loss of Ecosystems Dependent on Groundwater Resources. In *Forest Dynamics and Conservation: Science, Innovations and Policies* (pp. 263–278). Singapore: Springer Nature Singapore (2022).
153. Erostate, M., et al. Groundwater dependent ecosystems in coastal Mediterranean regions: Characterization, challenges and management for their protection. *Water research* **172**, 115461 (2020).
154. Wu, X., Ma, T. & Wang, Y. Surface water and groundwater interactions in wetlands. *Journal of Earth Science* **31**, 1016–1028 (2020b).
155. Santos, I. R., et al. Submarine groundwater discharge impacts on coastal nutrient biogeochemistry. *Nature Reviews Earth & Environments* **2**, 307-323 (2021).