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# Survival in Common Snapping Turtles, Chelydra serpentina (Testudines: Chelydridae), in western Nebraska 

John B. Iverson ${ }^{1, *}$ and Geoffrey R. Smith ${ }^{2}$


#### Abstract

Annual estimates of survival for Common Snapping Turtles (Chelydra serpentina) in western Nebraska USA were generated from mark-recapture data from nesting females encountered in 2005-2017. Our population models suggested no annual variation in either adult annual survival ( $0.947 \pm 0.017 \mathrm{SE}$ ) or annual capture probability ( $0.294 \pm 0.027 \mathrm{SE}$ ). However, there was a tendency toward higher survival in larger females. High annual survival (e.g. $>90 \%$ ) characterises populations of Chelydra from Ontario to Texas.


Key words: Chelydra serpentina, Chelydridae, Nebraska, survival

## Introduction

The Common Snapping Turtle, Chelydra serpentina (Linnaeus 1758), is the most widely distributed turtle species in North America and the fourth most widely distributed in the world, even excluding introduced populations (Turtle Taxonomy Working Group, 2021). It is also one of the most frequently encountered turtles in its range (Ernst and Lovich, 2009), and one of the world's best-studied turtle species (among the top five; Lovich and Ennen, 2013).
One of the well-studied populations of C. serpentina is that on the Crescent Lake National Wildlife Refuge (CLNWR) in Nebraska USA, where our research was on-going from 1981 through 2018. We have studied the reproductive ecology of this species (Iverson et al., 1997), including correlates of its reproductive output (Hedrick et al., 2018; Iverson and Hedrick, 2018), examined climate effects on its nesting phenology (Hedrick et al., 2021), quantified sex ratio and density (Iverson and Smith, 2010), growth and longevity (Iverson and Lewis, 2019), and diet (Lewis and Iverson, 2018). However, we have not previously reported on survival of adult females, perhaps the most important determinant of long-term population viability in turtles (e.g., Heppell,

[^0]1998). Here we add that survival information, examine the effect of body size on survival, and compare our survival data with those published by others.

## Materials and Methods

Our field site was located in the western Sandhills region of Nebraska on the CLNWR, Garden County, adjacent to the Gimlet Lake wetland complex $\left(41.7651^{\circ} \mathrm{N}, 102.4367^{\circ} \mathrm{W}\right)$. Our field methods were described in Iverson et al. (1997). We restricted our survival analysis to females captured during nesting forays in the years 2005-2017 (excluding 2016, when we did not sample), during which we consistently and rigorously surveyed potential nesting areas on Gimlet Lake, rather than sporadic surveying prior to 2005. During this period we captured 94 adult nesting females 200 times. All turtles were measured (e.g., maximum carapace length [CL] in mm), weighed, marked, and released immediately.
We used Program MARK to estimate annualised survival estimates and capture probabilities (White and Burnham, 1999). In addition, to test for a body size effect on survival and capture probability, we conducted separate MARK analyses with 42 smaller females ( $<300$ mm CL; estimated age 11-17 years; Iverson and Lewis, 2019), and 52 larger females (> 300 mm CL ). Only six females passed 300 mm CL during their recapture history. One was initially 293 mm but was over 300 for her four subsequent recaptures, and was considered a large female. Five others were 307 to 312 mm CL at their last capture, but had spent most of their capture history below 300 mm , and hence were considered small females for this analysis.

## Results

The best population model for our full 2005-2017 sample, based on the lowest Akaike Information Criterion value (598) and model weight ( $97 \%$, Table 1), suggested no annual variation in either adult annual survival ( $0.947 \pm 0.017 \mathrm{SE}$ ) or annual capture probability $(0.294 \pm 0.027$ SE). Estimated population sizes from the fully parameterised model (i.e., including annual estimates of survival and capture probability) suggested a stable population of about $59 \pm 14$ (SD) females (excluding first and last year estimates; Table 2).
For only those Nebraska females $<300 \mathrm{~mm}$ CL, the best model ( $\mathrm{AICc}=182$; weight $=99.9 \%$ ) again suggested no annual variation in survival ( $0.878 \pm$ $0.043 \mathrm{SE} ; 95 \%$ Confidence Interval, $0.77-0.94$ ) or capture probability $(0.363 \pm 0.061 \mathrm{SE})$, as did the best model for females $>300 \mathrm{~mm}$ CL $($ AICc $=395$; weight $=99.3 \%$; survival $=0.954 \pm 0.019 \mathrm{SE}, 95 \%$ CI 0.90 -0.98 ; capture probability $=0.285 \pm 0.033 \mathrm{SE})$. These restricted data sets suggest that small (younger) females are more reliably captured each year, but may suffer higher mortality than larger females.

## Discussion

Despite the extensive distribution of Chelydra, there is apparently little geographic variation in annual adult female survival, although there are still large areas of the species' range that lack data (Table 3). However, despite differences in sample sizes, population body sizes, sample periods, season lengths, and capture probabilities among studies, survival is consistently

Table 1. Model comparison for mark-recapture data for Chelydra serpentina in western Nebraska, analysing the effects of time (t, in years) on annual survival (phi) and capture probability (p). The first model (no annual variation in either adult survival or capture probability) is by far the most strongly supported $($ Akaike weight $=0.97)$.

|  | AICc | delta AICc | Weight | Likelihood | \# parameters |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Phi(.) p(.) | 597.76 | 0.00 | 0.97 | 1.00 | 2.00 |
| Phi(.) p(t) | 604.91 | 7.16 | 0.03 | 0.08 | 12.00 |
| Phi(t) p(.) | 614.80 | 17.04 | 0.00 | 0.00 | 12.00 |
| Phi(t) p(t) | 623.37 | 25.61 | 0.00 | 0.00 | 21.00 |

very high (average from Table $3=94 \%$ ), which no doubt explains the success of this species when so many others are in decline. Only two studies have reported survival estimates for both males and females (Eskew et al., 2010; Rose and Small, 2014), both of which suggest higher survival in males than females. Additional studies are needed to confirm this possible pattern.
Our ability to detect younger females more easily may also apply to their predators. In the only other study to examine the effect of body size on survival, Armstrong et al. (2017) also found annual survival to increase with body size.
It is perhaps not surprising that such a large, dangerous turtle would have high rates of adult survival. Nevertheless, despite those high rates across its range (Table 3), snapping turtle populations apparently have great difficulty recovering from catastrophic mortality events. Brooks et al. (1991) documented the loss of approximately $40 \%$ of the adults to River Otters (Lutra

Table 2. Fully parameterised model by year based on mark-recapture data for Chelydra serpentina in western Nebraska, analysing the effects of time ( t , in years) on annual survival (phi) and capture probability (p). $\mathrm{N}=$ number of captures; N estimate $=$ female population size estimate; $\mathrm{SE}=$ standard error.

| Data year | N | phi estimate | SE | p estimate | SE | N estimate |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| $2005-2006$ | 6 | 0.882 | 0.176 | 0.200 | 0.109 | 30.0 |
| $2006-2007$ | 14 | 0.912 | 0.172 | 0.365 | 0.123 | 38.4 |
| $2007-2008$ | 18 | 1.000 | 0.000 | 0.205 | 0.083 | 87.8 |
| $2008-2009$ | 24 | 0.963 | 0.105 | 0.416 | 0.091 | 57.7 |
| $2009-2010$ | 21 | 0.869 | 0.110 | 0.400 | 0.084 | 52.5 |
| $2010-2011$ | 10 | 1.000 | 0.000 | 0.130 | 0.051 | 46.9 |
| $2011-2012$ | 21 | 0.948 | 0.115 | 0.337 | 0.076 | 62.3 |
| $2012-2013$ | 19 | 1.000 | 0.000 | 0.299 | 0.068 | 63.5 |
| $2013-2014$ | 18 | 1.000 | 0.000 | 0.301 | 0.066 | 59.8 |
| $2014-2015$ | 24 | 0.740 | 0.241 | 0.364 | 0.133 | 65.9 |
| $2015-2017$ | 25 | 0.648 | 0.000 | 0.596 | 0.000 | 41.9 |

Table 3. Annual survival estimates for Snapping Turtles (Chelydra serpentina) from the literature arranged in order of declining latitude. Sample restrictions are indicated; abbreviations include males (M), females (F), and carapace length (CL in cm ). Note that the first five studies refer to the same population.

| State | Latitude | Sample | Survival | Source |
| :--- | :---: | :---: | :---: | :--- |
| Ontario | 45.5 | F | 0.97 | Galbraith and Brooks, 1987 |
| Ontario | 45.5 | F | 0.93 | Galbraith and Brooks, 1987 |
| Ontario | 45.5 | F | 0.966 | Cunnington and Brooks, 1996 |
| Ontario | 45.5 | 25 cm CL F | 0.935 | Armstrong et al., 2017 |
| Ontario | 45.5 | 30 cm CL F | 0.962 | Armstrong et al., 2017 |
| Ontario | 45.5 | 35 cm CL F | 0.976 | Armstrong et al., 2017 |
| Wisconsin | 43.7 | $\mathrm{M}, \mathrm{F}$ | 0.963 | Paisley et al., 2009 |
| Minnesota | 43.7 | $\mathrm{M}, \mathrm{F}$ | 0.939 | Paisley et al., 2009 |
| Michigan | 42.5 | F | 0.88 | Congdon et al., 1994 |
| Nebraska | 41.8 | $<30 \mathrm{~cm} \mathrm{CL} \mathrm{F}$ | 0.878 | This paper |
| Nebraska | 41.8 | $>30 \mathrm{~cm} \mathrm{CL} \mathrm{F}$ | 0.954 | This paper |
| W Virginia | 39.1 | $\mathrm{M}, \mathrm{F}$ | 0.97 | Flaherty et al., 2008 |
| Virginia | 37.6 | $\mathrm{M}, \mathrm{F}$ | 0.91 | Colteaux and Johnson, in Colteaux, 2017 |
| N Carolina | 35.2 | M | 0.994 | Eskew et al., 2010 |
| N Carolina | 35.2 | F | 0.914 | Eskew et al., 2010 |
| Texas | 29.9 | M | 0.94 | Rose and Small, 2014 |
| Texas | 29.9 | F | 0.93 | Rose and Small, 2014 |

canadensis) in 1986-1989 at the northern limit of the range in Ontario, Canada. However, despite high annual survival (97\%) since that event, the population had not recovered after 23 years (Keevil et al., 2018), and females remained at only ca. $60 \%$ of pre-catastrophe numbers. Whether more southern populations (with more benign climates) would recover more quickly is unknown. In any case, the high annual adult survival rates known for this species no doubt contributes to their abundance in diverse habitats across the largest range of any North American turtle.

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