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4 **Field comparison of ecophysiological traits between an invader and a native**
5 **species in a Mediterranean coastal dune**

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14
15 **ABSTRACT**

16 Photochemical efficiency, gas exchange, leaf water potential and pigment and free proline
17 content of *Oenothera drummondii* subsp. *drummondii* Hook (an invasive non-native
18 species) and *Achillea maritima* (L.) Ehrend. & Y.P. Guo, (an ecologically similar native
19 species) were explored to understand the success of invasive non-native species in
20 Mediterranean coastal dunes of southwest Spain. We have conducted a field study during
21 a complete annual cycle, comparing both species. Fifteen pairs of neighbouring plants of
22 the two study species of similar size were selected and measured seasonally. The results
23 show that in spring and summer, assimilation rates of *O. drummondii* were significantly
24 higher than those of the native, *A. maritime*, even though the native species had higher
25 photochemical efficiency. Additionally, the non-native species presented better water
26 content regulation than the native one, surely related to better water use efficiency and
27 may be linked to greater root development. The differences in leaf dry matter content
28 values for both species might indicate a different strategy of resource use; with *A.*
29 *maritima* displaying a more conservative strategy and *O. drummondii* presenting a rapid
30 resource acquisition and use strategy as predictors of rapid growth and soil fertility. We
31 conclude that *O. drummondii* utilizes light, water and probably nutrients more efficiently
32 than the native *A. maritima* and suffers lower stress in Mediterranean coastal dunes where
33 water availability is reduced (44 mm from May to October in the study area) and light
34 radiation levels are high.

35

36 **Keywords:** *Achillea maritima*; Invasiveness; Leaf water potential; Mediterranean-type
37 ecosystems; *Oenothera drummondii*; Photosynthesis.

38 **Introduction**

39

40 Coastal dunes are complex and dynamic ecosystems where vegetation is subjected to
41 many stressors including salinity, episodic overwash, water stress, high temperature, high
42 winds, sand burial, substrate instability, low field capacity, and nutrient scarcity as the
43 most common abiotic stressing factors that influence plant community assembly
44 processes in dune systems (Hesp, 1991; Maun, 2009). Additionally, physiological plant
45 activity in Mediterranean-type ecosystems is limited during summer by high
46 temperatures, water scarcity and high radiation levels (Cowling, et al. 2005; Godoy et al.,
47 2011), but it is also limited in winter through the combination of low temperatures with
48 high radiation (Oliveira and Peñuelas, 2004; Flexas et al., 2014). Therefore, in the coastal
49 dunes of the Mediterranean type climate, these stressors reinforce each other intensifying
50 the vegetation stress level. As a consequence of the notable selection pressure imposed
51 by these environmental conditions, the coastal dune vegetation has a high degree of
52 specialization (Novoa et al., 2012), particularly in Mediterranean type climate. In beach
53 and embryo dunes frequent occurrence of intense disturbance events provide invasive
54 species with the opportunity to colonize novel areas (Lortie and Cushman 2007),
55 additionally, in inland dunes, intermediate levels of disturbance and stress can also offer
56 good conditions for the settlement of invasive species (Carboni et al., 2010). That is, the
57 patterns of species invasions on coastal dunes depend on the environmental conditions
58 and the species involved (Gallego-Fernández et al., 2019).

59 Coastal dune environmental conditions impose adaptations to plants that stablish in
60 these areas. For instance, to withstand salt, plants increase leaf thickness and the degree
61 of succulence, while to cope with burial they produce larger seeds and have stolons or
62 rhizomes (Hesp, 1991). Water-limited resources are a primary stressor in coastal dune
63 ecosystems; therefore, many coastal dune plants have adaptations to resist dryness. The
64 adjustments to dryness occur at different plant levels: physiological, reproductive,
65 morphological at the leaf, roots, or the whole plant structure. At the leaf level, for
66 example, plants can adjust leaf rolling, thickness, orientation, hairiness, size or epicular
67 wax. At physiological levels, plants can have osmotic adaptations to increase water use
68 efficiency or different photosynthetic pathways such as C4 and CAM (Hesp, 1991).

69 As a result of the high specialization grade imposed by these harsh environmental
70 conditions, these ecosystems support many threatened and endemic species which are
71 especially sensitive to invasions by non-native plants. One of the greatest threats of

72 natural habitats around the world is the introduction of exotic species (Sax and Gaines,
73 2003) and on coastal dune ecosystem in particular, this has been assessed as high and still
74 increasing (Millennium Ecosystem Assessment, 2005; Novoa et al., 2012), mainly due to
75 the frequent stressor and the existence of open patches free of plant competition (Carboni
76 et al., 2010; Antunes et al., 2018).

77 The success of invasive plants on novel environments depends on several factors,
78 such as the adaptive plasticity of their morphological or physiological traits and the
79 environmental conditions of the new area. Alternatively, numerous studies (Oduor et al.,
80 2016; Liao et al., 2016) support that pre-adaptation or rapid adaptive evolution to new
81 ecological niches could be as important as plasticity to explain the ability of invasive
82 plants to persist and expand in new-introduced habitats. These two processes would not
83 be mutually exclusive. Thus to maximize fitness in response to new environmental
84 conditions a combination of plasticity and rapid genotypic adaption could occur.
85 According to Liao et al. (2016) plasticity plays an important role in invasiveness, but also
86 local genotypic variation in plasticity does.

87 Besides, also the biological interactions (such as plant-plant competition or plant-
88 herbivore interaction) have an ecological impact on original communities and in the
89 ecosystem function which is crucial for the success of plant invasion (Ehrenfeld, 2010;
90 Bottollier-Curtet et al., 2013). Invasive plants commonly compete so effectively in new
91 ecosystems that they displace native species and alter ecosystem function and ecosystem
92 services. Van Kleunen et al. (2010) studied what functional traits were associated with
93 invasiveness (physiology, allocation, growth, size or fitness), considering invasiveness as
94 the capacity of plants to spread through other biota and ecosystems. They found that
95 invasive non-native species had higher values for traits related to performance than non-
96 invasive species. The question of what kinds of traits promote plant invasiveness is of
97 utmost significance for understanding plant success in general, and particularly for
98 understanding the mechanisms of non-native plant invasions. However, studies that
99 suggest that invasive species have advantages over native ones have had irregular results
100 and moreover, the available data of the positive effect of physiological traits on invasive
101 species' fitness are inconsistent.

102 Nonetheless, invasive species expected to have physiological traits (such as higher
103 photosynthetic rates or better stomatal control) than allow them to take advantage of a
104 better physiological status and overcome natives. A comparative study of physiological

105 plant activity in co-occurring invader and native species is a useful tool to improve
106 understanding of the invasiveness of exotic plants. Plant ecophysiological characteristics
107 must be well adapted to their local environmental conditions (Ackerly et al., 2000) so
108 that, the physiological response of plants to local conditions can play a fundamental role
109 in species' ecological distribution (Körner and Diemer, 1987; Brodersen et al., 2008).
110 Physiological adjustments allow species to maintain their fitness under different
111 environmental conditions (Becklin et al., 2016). Among these ecophysiological
112 characteristics, photosynthesis and water status-related traits have been proven to have a
113 considerable influence on the invasive success of non-native species (McDowell, 2002).

114 *Oenothera drummondii* subsp. *drummondii* is a non-native invasive species of
115 coastal dunes around the world; native to coastal dunes of the Gulf of Mexico and in the
116 last century has colonized coastal dunes all over the world (Dietrich and Wagner, 1988).
117 Nowadays it has spread over coastal areas of different continents, being considered
118 invasive in Spain, Israel, China and Australia and naturalized in many places of the world,
119 shifting to regions with different climatic characteristics (Dietrich 2000; Heyligers 2008;
120 Campos and Herrera 2009; Xu et al. 2012, Dufour-Dror 2013; García de Lomas et al.
121 2015).

122 In coastal dunes of southwestern Spain, *O. drummondii* has produced a strong
123 impact on native communities by modifying the composition and abundance of the
124 species (García de Lomas et al., 2016; Gallego-Fernández et al., accepted). To understand
125 some of the causes of the success of this non-native species we have compared our
126 observations of *O. drummondii* with a species of similar characteristics that has been
127 displaced after the invasion and that used to be highly abundant, *Achillea maritima*. Both
128 are small-sized plants (40-50 cm tall) with reproduction mainly by seeds. The purpose of
129 this study is to establish whether the success of *O. drummondii* can be explained in part
130 by differences from native species in the physiological response to Mediterranean coastal
131 dune environment. To answer this question, we compared the physiological performance
132 of both species measuring the photochemical efficiency, gas exchange, leaf water
133 potential and free proline and pigment content under field conditions over the whole year.
134 All these ecophysiological variables have proven to have a considerable influence in the
135 invasive success of the non-native species (McDowell, 2002) while free proline in leaves
136 has been widely proved to accumulate as a response to numerous stressors (such as
137 drought or cold temperatures). Additionally, weather conditions also induce changes in

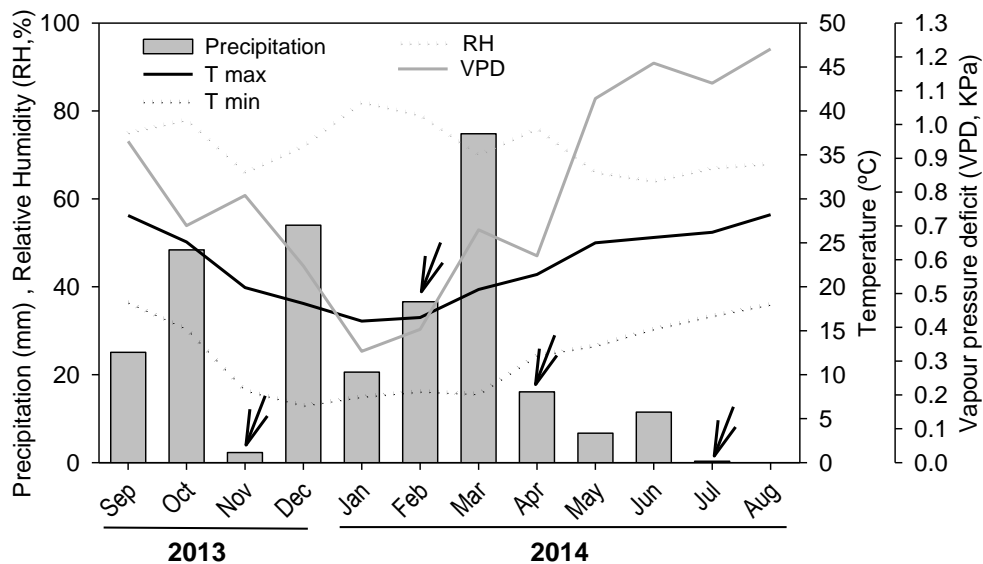
138 the pigment content (Hoffmann and Parsons, 1997; Gratani, 2014, Zunzunegui et al.,
 139 2016). We hypothesized that *O. drummondii* would exhibit a high capacity to adapt to
 140 Mediterranean environmental conditions presenting higher photochemical efficiency,
 141 assimilation rates and water use efficiency than the native species

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Material and methods

145 Study site

146 The study was accomplished on a coastal dune system located in the protected area
 147 “Marismas del Odiel” close to Huelva city, SW Spain (37°09'N, 6°54'W). The climate is
 148 Mediterranean with hot summer; the type is Csa, according to the Köppen-Geiger
 149 classification (Peel et al., 2007). The mean annual temperature is 18.1 °C while the mean
 150 minimum and maximum temperatures are 13.3 and 21.7° C, respectively. Average annual
 151 rainfall is 467 mm, with a pronounced drought period in summer (data from Agencia
 152 Estatal de Meteorología at Huelva Meteorological Station, 30-year record from 1971 to
 153 2000). The year of the study, with precipitation of 296 mm, was 36% dryer than the annual
 154 average (Fig. 1), with remarkably low precipitation during the warmest period of the year
 155 in which the accumulated rainfall from May to October was 44 mm.



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159 **Fig. 1** Monthly precipitation, daily maximum and minimum mean temperatures and air
 160 relative humidity during the study period (2013-14 hydrological cycle). Arrows indicate
 161 sample periods.

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163 Study species

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165 *Oenothera drummondii* subsp. *drummondii* Hook. (*Onagraceae*), is a short-lived leaf
166 perennial species. The main stems are erect to procumbent while the basal side stems are
167 prostrate or ascending to about 50 cm in height, with a strong taproot. Flowers are self-
168 compatible, outcrossing and pollinated by hawkmoths in their native habitat (Wagner et
169 al., 2007). The species reproduces and spreads through small and numerous seeds. In its
170 native range area, the species grows in coastal dunes with humid subtropical and wet
171 tropical climates (Cfa, Aw, Am) according to the Köppen-Geiger classification (Peel et
172 al., 2007). The first records of *O. drummondii* in the southwest of the Iberian Peninsula
173 date from 1957 (Silvestre, 1980) and in the study area it was registered for the first time
174 in 1996 (García Mora com. pers).

175 *Achillea maritima* (L.) Ehrend. & Y.P. Guo is a pioneering herbaceous perennial
176 species, with 20-40 cm of height, from the *Asteraceae* family (Ehrendorfer and Guo,
177 2005). This species grows in sand coast dunes from Europe and Morocco and exerts a
178 stabilizing action on the sandy soils (García Novo and Merino, 1993). In the Iberian
179 Peninsula, the species is in regression, having disappeared, or almost, in some regions
180 (Silván and Campos, 2002; Mayoral, 1999) mainly because of the loss of its natural
181 habitat due to anthropic pressure. Thick white fuzz covers both the stems and the small
182 oval, alternate leaves. The globose yellow flowers are entomophilous and visible from
183 June to September. The selection of this species was based on the fact that it is one of the
184 most abundant native plants in the studied site and has similar functional traits to the
185 invasive species. Moreover, Gallego-Fernandez et al. (2019) found evidence that this
186 species is out-competing *A. maritima* since in dunes invaded by *O. drummondii* the cover
187 of *A. maritima* is 91% lower than in non-invaded dunes.

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189 Experimental design

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191 Field measurements were made in a 50 x 50 m plot located inside the dune zone. In this
192 plot, we randomly selected 15 pairs of neighbouring mature plants of *Oenothera*
193 *drummondii* and *Achillea maritima* of similar size. All pairs were separated more than 2
194 meters from each other. The measurements were carried out with a seasonal periodicity
195 during an annual cycle: November (autumn, with mild temperatures and when the first
196 rains after the summer drought allow plants to recover water status), February (winter,
197 the period of the wet season with the coldest month), April (spring, period with mild but

198 increasing temperatures and at the end of the rainy season) and July (summer, with the
199 highest temperatures and radiation coupled with drought).

200 *Photochemical efficiency of photosystem II:* Chlorophyll fluorescence kinetics was
201 determined in the field on healthy fully expanded leaves using a pulse-amplitude
202 modulation technique through a portable fluorometer (mini-PAM, Walz, Effeltrich,
203 Germany). Fluorescence was excited by a saturating pulse of red light (approximately
204 $3000 \mu\text{mol m}^{-2}\text{s}^{-1}$ for 0.8 s) from a diode (type H-3000 LED, Stanley) attached to a fibre
205 optic. Leaves were exposed to a weak modulated measuring beam from a LED at a
206 frequency of 600 Hz to determine F_0 and F and then to the pulse of saturating red light to
207 assess F_m and F'_m .

208 Maximum photochemical efficiency of PSII (F_v/F_m) was measured on dark-adapted
209 leaves (leaves were artificially maintained 20 minutes in darkness by means of leaf-clips,
210 a time interval considered sufficient to complete the reoxidation of all the reaction centres
211 of photosystem II) from the ratio of variable fluorescence to maximum fluorescence as
212 $F_v/F_m = (F_m - F_0)/F_m$, where F_0 = initial and F_m = maximal fluorescence (Genty et al., 1989).
213 Effective photochemical efficiency of PSII (Φ_{PSII}) was estimated on light-adapted leaves
214 as $\Phi_{\text{PSII}} = (F'_m - F)/F'_m$, where F'_m = maximal and F steady-state fluorescence under actinic
215 irradiance (Genty et al., 1989).

216 Three leaves were measured per plant at 9:30-11:30 h solar time (mean values per
217 plant were used for statistical analysis).

218 *Gas exchange measurements:* Measurement of net photosynthetic or net CO_2
219 assimilation rate (A_n , $\mu\text{mol CO}_2 \text{ m}^{-2}\text{s}^{-1}$), transpiration rate (E , $\text{mmol H}_2\text{O m}^{-2}\text{s}^{-1}$), stomatal
220 conductance rate (g_s , $\text{mol H}_2\text{O m}^{-2}\text{s}^{-1}$) and leaf intercellular CO_2 concentration (C_i , ppm)
221 were taken on mature leaves during clear sunny days, using portable open system,
222 compact infra-red gas exchange analyser (LCi-Portable Photosynthesis, ADC, UK). From
223 these set of data, we calculated instantaneous water use efficiency (WUE_i , $\text{mmolCO}_2 \text{ mol}^{-1}$
224 H_2O) as the ratio of net photosynthesis to transpiration (A_n/E) indicative of the moles of
225 CO_2 assimilated per mol of H_2O lost, and instantaneous carboxylation efficiency (A_n/C_i ,
226 $\text{molCO}_2 \text{ m}^{-2}\text{s}^{-1}$) as the ratio of net assimilation to leaf intercellular CO_2 concentration.

227 Three leaves per plant were measured and the average values per plant calculated.
228 The surface of leaves used for gas exchange measurements was calculated in the
229 laboratory using the image area analyser software Midebmp (Ordiales, Spain, 2000).

230 Measurements were made from 8:30 to 10:30 h, (solar time), to avoid possible midday
231 depression of stomatal conductance and to get maximum net photosynthetic rates.

232 *Leaf water potential:* Plant water status was evaluated through midday leaf water
233 potential (Ψ_{md}) and was determined with a pressure chamber (Manofrigido, Lisbon,
234 Portugal) in the field using terminal shoots which were excised and immediately
235 measured. The readings were made between 12:30 and 14:00 (solar time), when the
236 greatest water deficit of the day occurs and when the minimum values are reached.

237 *Leaf parameters:* Leaves used and scanned for photosynthesis measurements were
238 dried at 70 °C for 48 h and weighted (Leaf mass area (LMA) was calculated as the ratio
239 of dry leaf mass (Md) to fresh leaf area (g m^{-2}).

240 New, fully expanded, healthy leaves were collected from all plants selected for
241 physiological measurements. Leaves for relative water content (RWC) and leaf dry matter
242 content (LDMC) measurements were kept refrigerated and stored in plastic bags until
243 fresh mass (Mf) was recorded within 3 h. After weight, leaves were then hydrated with
244 distilled water to saturation for 24 h at 5°C in the plastic bags. Leaves were dried with
245 tissue paper to remove any surface water, and then immediately weighed to obtain
246 saturated mass (Ms). Finally, leaf samples were then dried at 70 °C for 48 h and weighted
247 (Md). We estimated RWC (%) as follows: $\frac{Mf-Md}{Ms-Md} \times 100$

248 The ratio LDMC (mg g^{-1}), also known as tissue density was calculated as Md/Ms .

249 The rest of the collected leaves were kept in liquid nitrogen in the field and preserved
250 in the laboratory at -24 °C until analysis. Photosynthetic pigments were quantified
251 following Lichtenthaler (1987) on a dry mass basis. Chlorophyll a (Chl a), b (Chl b) and
252 total carotenoid (Car) pigments were extracted with 100 % acetone and determined
253 spectrophotometrically. From these data were calculated the ratio chlorophyll a/b (Chl
254 a/b) and the ratio of total carotenoids to total chlorophylls (Car/Chl),

255 Free proline content was determined colourimetrically by the ninhydrin acid method
256 of Bates et al. (1973).

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258 Statistical analyses

259 Two-way ANOVAs were used to investigate the effects of species and season
260 (independent factors) on ecophysiological variables (dependent factor). Post-hoc Tukey
261 tests were then used for appropriate pair-wise comparisons. Bivariate correlation tests
262 were made between the different variables. A multivariate principal component analysis
263 (PCA) was performed with the physiological and leaf parameters matrix (21 variables x

264 4 seasons x 2 species x 15 plants) to identify which are the variables responsible for the
265 different seasonal response between species. The normality of all measured variables was
266 first checked with the Kolmogorov–Smirnov test. The statistical tests were considered
267 significant at the $P < 0.05$ level. All statistical tests were made with the software package
268 IBM SPSS Statistics for Windows, Version 24 (Armonk, NY, USA).

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270 **Results**

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272 Photosynthetic rates ranged close to $20 \mu\text{mol m}^{-2}\text{s}^{-1}$ in the non-native species throughout
273 the year, whereas in the native this variable was seasonal dependent, with pronounced
274 drops in spring and summer (Fig. 2A). *O. drummondii* assimilation rates were
275 significantly higher than those of *A. maritima* in spring and summer ($P < 0.008$), while
276 no differences occurred in autumn and winter. As observed for A_n , and responding to
277 summer stressful conditions, native plants had lower A_n/C_i values in spring ($P < 0.016$)
278 and summer ($P < 0.001$) than non-native plants (Fig. 2B). Native plants showed lower g_s
279 in spring and summer but both species showed similar g_s values after the first rains in
280 autumn and winter (Fig. 2C). Non-native species showed greater WUE_i than native plants
281 during the drought period (Fig. 2D).

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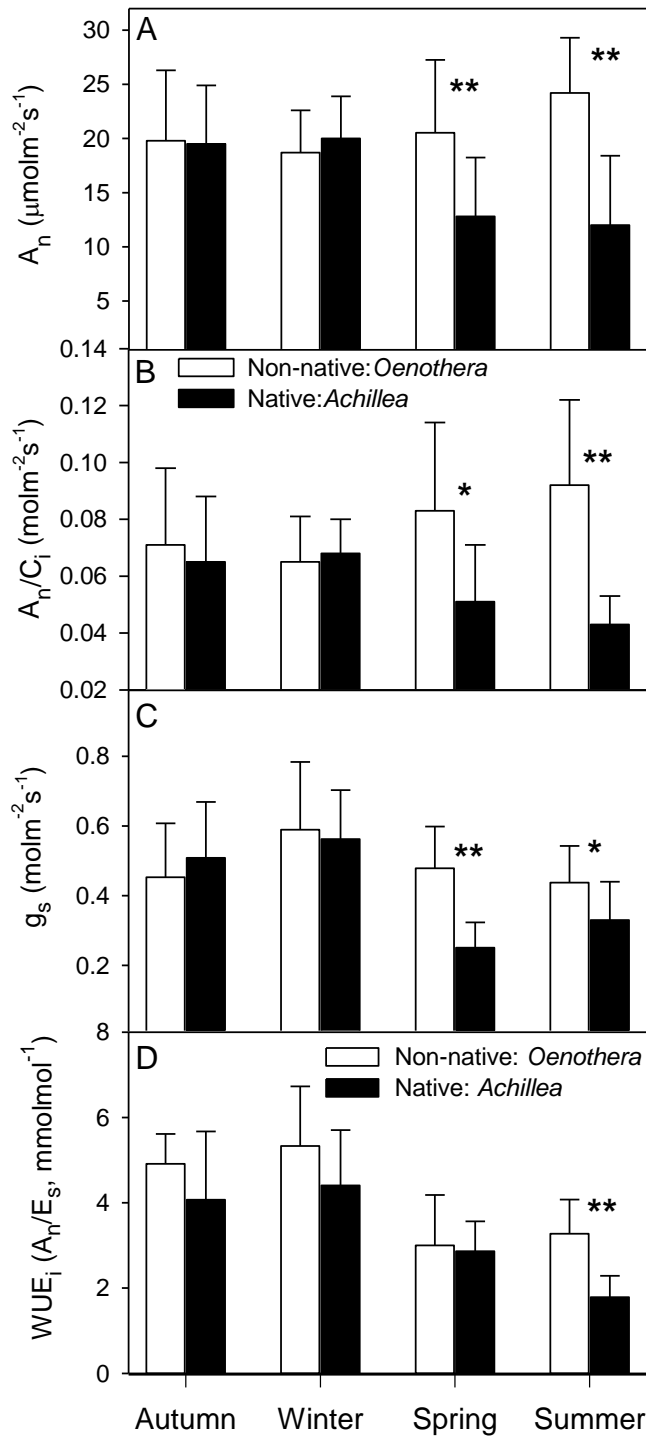
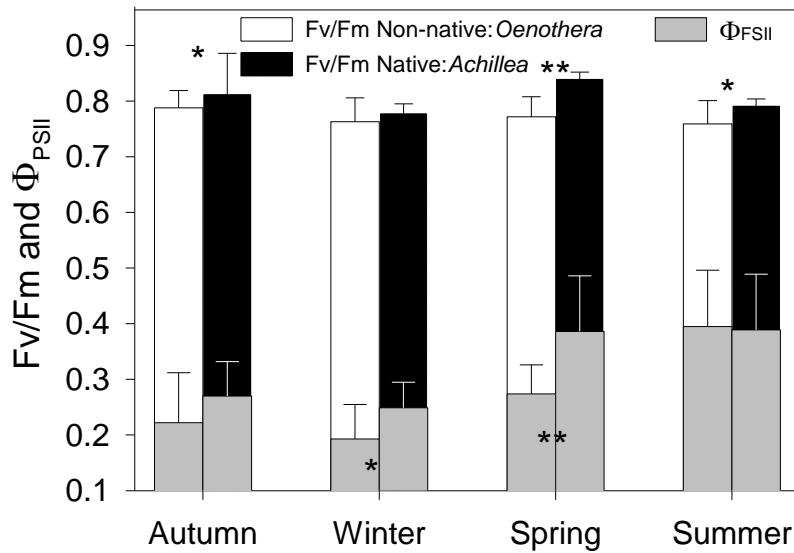


Fig. 2 Mean values (\pm sd) of A_n : photosynthesis rate (A), A_n/C_i : instantaneous carboxylation efficiency (B), g_s : stomatal conductance (C), and WUE_i : instantaneous water use efficiency (D) in the two study species, over the 2013-14 annual cycle. Asterisks denote significance levels for the comparisons between the two species by Student's t-tests (** $P < 0.001$; * $P < 0.05$).

331 Maximum photochemical efficiency (F_v/F_m) was lower in the non-native than in the
 332 native species, these differences being significant in spring, summer and autumn (Fig. 3).
 333 Also, Φ_{PSII} was higher in the native species in winter and spring.

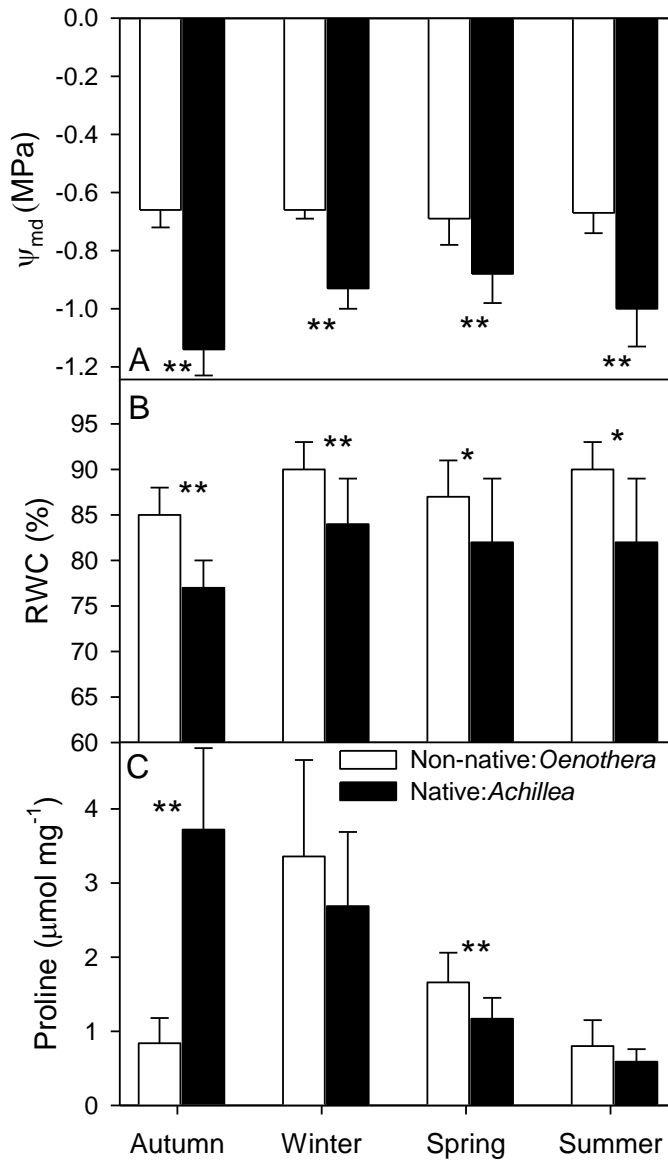
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346 **Fig. 3** Mean values (\pm sd) of maximum (F_v/F_m), and effective (Φ_{PSII}) photochemical
 347 efficiency in the two study species, over the 2013-14 annual cycle. Asterisks denote
 348 significance levels for the comparisons between the two species by Student's t-tests (**P
 349 < 0.001; *P < 0.05).

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Overall, the highest Ψ_{md} of the year of both species were recorded in winter and spring, while summer water deficit decreased, Ψ_{md} reaching the lowest values in autumn (Fig. 4A). The non-native species maintained higher Ψ_{md} than the native one throughout the whole year. This difference was also observed in the RWC values recorded throughout the year, which were always significantly higher in the non-native species (Fig. 4B). Proline content in leaves, as a measure of stress, or an osmotic regulator in plants (Ain-Lhout et al., 2001), was not elevated in any season even in winter when the highest quantities were accumulated. Significant differences between the two species were observed in autumn and spring (Fig. 4C).



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363 **Fig. 4** Mean values (\pm sd) of midday leaf water potential (Ψ_{md}), Relative water content
 364 (RWC) and proline content, of the two study species, over the 2013-14 annual cycle.

365 Asterisks denote significance levels for the comparisons between the two species by

366 Student's t-tests (** $P < 0.001$; * $P < 0.05$).

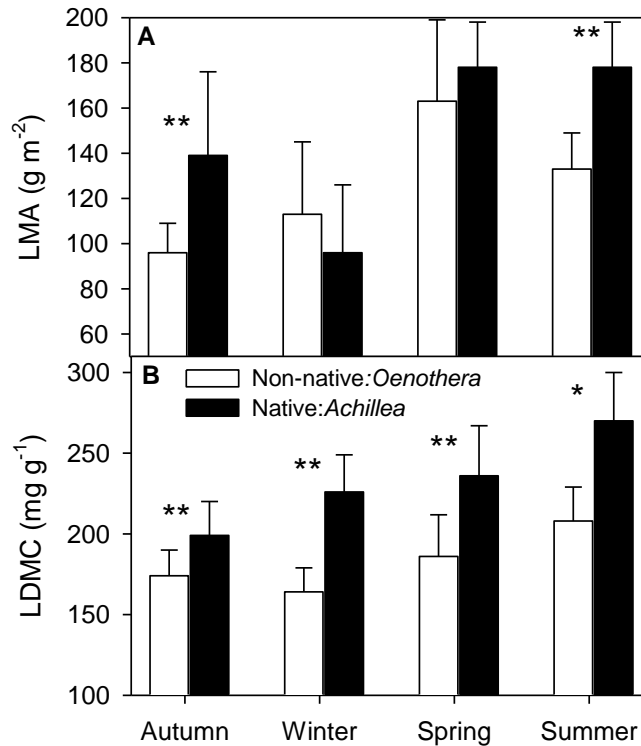
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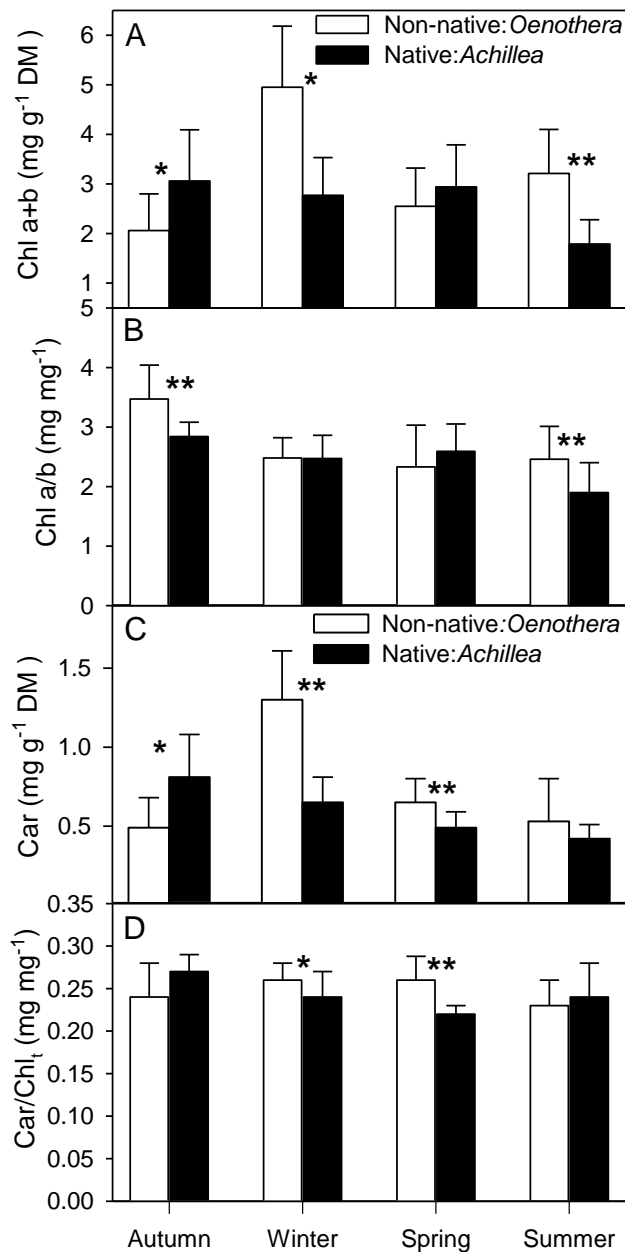
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371 The lowest LDMC values occurred always in *O. drummondii*, while annual
 372 dynamics were similar in both study species, increasing from winter to summer with a
 373 marked drop in autumn (Fig. 5A, B). Compared to the native species, *O. drummondii* had
 374 significantly lower LMA in summer and autumn, while the maximum values in both
 375 species were recorded in spring and summer. LMA and LDMC were correlated, with P
 376 < 0.001 .



377
 378 **Fig. 5** Mean values (\pm sd) of LMA: leaf mass area (A), and LDMC: leaf dry matter content
 379 (B) in the two study species, over the 2013-14 annual cycle. Asterisks denote significance
 380 levels for the comparisons between the two species by Student's t-tests (** $P < 0.001$; * P
 381 < 0.05).

382
 383 The most stressful seasons of Mediterranean climate, winter and summer, were the
 384 periods in which *O. drummondii* accumulated significantly more Chl a+b than the native
 385 species (Fig. 6A). On the contrary, ratio Chl a/b was significantly higher than in the native
 386 in autumn and spring (Fig. 6B). As for the carotenoids, non-native species accumulated
 387 more pigments than the native one, in winter and spring, and also presented higher
 388 Car/Chl ratio (Fig. 6 C, D).



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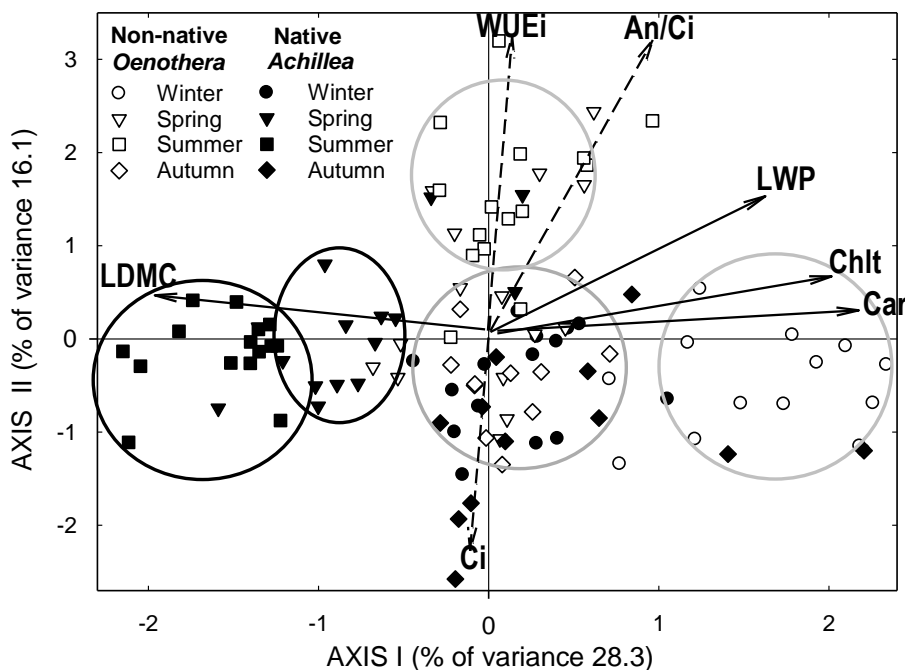
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391 **Fig. 6** Mean values (\pm sd) of the 2 study species in the 4 sampling periods for total
 392 chlorophyll content (A, Chl a+b), chlorophyll a/b ratio (B, Chl a/b), carotenoid content
 393 (C, Car), and total chlorophyll content to total carotenoids ratio (D, Chl/Car), over the
 394 2013-14 annual cycle. Asterisks denote significance levels for the comparisons between
 395 the two species by Student's t-tests (**P < 0.001; *P < 0.05).

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398 The PCA identified two main clusters in the data cloud differentiating the native
 399 from the non-native species, reinforcing the idea of a different physiological annual
 400 strategy to respond to Mediterranean coastal sand habitat (Fig. 7). The PCA provided an
 401 overview of the seasonal physiological pattern of the twenty one-measured variables. The
 402 plane defined by the first two axes of the analysis accounted for 44% of the variation. The
 403 variables more strongly correlated with PCA 1 axis ordered by importance were: LDMC,
 404 Car, Chl a, Chl a+b, Chl b, LWP (leaf water potential), Proline, Φ_{PSII} , LMA; while the
 405 second PCA axis was associated with the variables WUE_i , C_i , A_n/C_i , A_n , E. The PCA
 406 showed that the four seasons and two species segregate according to the two first axes;
 407 this implies that each species presents a different ecophysiological strategy throughout
 408 the year. The first PCA axis runs from *O. drummondii* leaves tending to have a better
 409 water status and larger pigment content than the native species, tending to exhibit higher
 410 LDMC. PCA 2 runs from summer *O. drummondii* leaves ending to have better WUE_i and
 411 carboxylation efficiency.



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 414 **Fig. 7** Projection on the plane defined by principal component axes (PC) 1 and 2 of
 415 30 individual plants (dots) of *Oenothera drummondii* and *Achillea maritima* from the 21
 416 variables considered. Solid arrows indicate direction and weighing of vectors
 417 representing the more strongly correlated variables with PC1 while short dash

418 indicate the variables correlated with PC2. A_n/C_i : instantaneous carboxylation efficiency,
419 Car: total carotenoid content, C_i : leaf intercellular CO_2 , Chl: total chlorophyll content;
420 LDMC: leaf dry matter content, LWP: leaf water potential, WUE_i : instantaneous water
421 use efficiency.

422

423 **Discussion**

424

425 Our main outcome in this study is that in spite of having higher photosynthetic rates in
426 spring and summer and better water status through the year, the non-native species *O.*
427 *drummondii* had lower photochemical efficiency than the native, *A. maritima*. The better
428 ecophysiological performance of *O. drummondii* appears to be related mostly to capture
429 of resources, such as CO_2 assimilation, and water-use strategy.

430 Respecting maximum photochemical efficiency, and considering the optimum
431 values between 0.75–0.85 in normal plants (Bolhàr-Nordenkampf and Öquist, 1993,
432 Maxwell and Johnson, 2000), the F_v/F_m values in *A. maritima* denote that the native
433 species, with values through the year higher to 0.760, was not under particularly stressing
434 conditions. These data agree with those recorded by Scorce et al. (2019) in the same
435 species in a coastal dune system in Italy. In contrast, Φ_{PSII} was affected by the
436 meteorological conditions, pointing out the species underwent certain dynamic
437 photoinhibition, especially in winter when the lowest values were recorded. When
438 compared with the invasive species, *A. maritima* was photochemically more efficient.
439 Leaf pubescence in *A. maritima* (Ciccarelli et al., 2009), an adaptation of plants to the
440 Mediterranean climate, reflect photosynthetically active radiation and provides protection
441 against UV-BA radiation and inhibits the reduction in photochemical efficiency
442 (Grammatikopoulos et al., 1994; Bisba et al., 1997). Pubescence increases the reflectance
443 of the leaves, improving plant energy balance and therefore diminishing the risk of
444 photoinhibition (Díaz Barradas et al., 1999; Zunzunegui et al., 1999; 2005). So the higher
445 photochemical efficiency values found in *A. maritima* agree with higher F_v/F_m values in
446 pubescence leaves as it has been found in other studies (Savé et al., 2000).

447 The fact that the lowest photochemical efficiency values were measured in winter
448 in the non-native species could be due to the tropical origin that makes it more sensitive
449 to low temperatures as it has been proven in other species with a tropical origin
450 (Zunzunegui et al., 2005). In these species, the risk in photoinhibition might be higher in
451 winter than in summer, as lower temperature might inhibit the photosynthetic electron

452 chain. Nevertheless, despite the lower photochemical efficiency in the invasive species,
453 the photosynthetic rates recorded for this species in spring and summer, were higher than
454 for the native one. This, therefore, indicates the sensitivity of *O. drummondii* to cold but,
455 simultaneously that it must be a species of great photosynthetic capacity because it is able
456 to overcome the native species despite its lower photosynthetic efficiency. That fact
457 would help to explain a 91% lower cover of *A. maritima* in dunes invaded by *O.*
458 *drummondii* (Gallego et al., 2019)

459 The meteorological data show that the harsh summer meteorological conditions,
460 (high temperature and radiation together with low relative humidity and water
461 availability) represent an important element in regulating the seasonal photosynthetic
462 activity. But above all, these results are an important differentiator of the best competitive
463 capacity of the invasive species. We may argue that increasing VPD (0.34 to 1.22 KPa)
464 from winter to summer caused the decreasing trend of g_s and the lower photosynthetic
465 rate of *A. maritima* in summer; since VPD has been considered the main environmental
466 variable controlling stomatal behaviour and causing the decrease of the photosynthetic
467 rate in Mediterranean woody vegetation (Tenhunen et al., 1987; Tognetti et al., 1998). On
468 the contrary, the spring and summer VPD effect on g_s was not patent in the invasive
469 species.

470 Summer conditions affected the variables A_n and A_n/C_i in an opposite way to both
471 species; while in the non-native species the figures increased and the highest annual
472 values were registered, in the native plants they decreased, and the lowest values were
473 recorded. Reductions in instantaneous A_n/C_i suggest the occurrence of a non-stomatal
474 limitation of photosynthesis and some mesophyll limitation on the photosynthesis of
475 studied plants (Silva et al., 2010). Although photosynthesis rate and stomatal conductance
476 have been described to have a linear correlation in response to environmental conditions
477 (Ball et al., 1987) other factors may be acting to increase A_n , as Rubisco kinetic or
478 mesophyll conductance that could explain that the highest annual A_n values recorded in
479 *O. drummondii* were not linked to the highest stomatal conductance values.

480 The high carboxylation efficiency values recorded in *O. drummondii* in summer
481 were supported by higher photosynthetic activity with no significant variations of C_i while
482 the low values recorded in *A. maritima* were supported by high C_i values and low
483 photosynthetic rates.

484 The WUE_i is a major trait of the water economy of plants and of their performance,
485 survival capacity and functioning in arid sites (Damesin et al., 1997; Bacon, 2004).
486 Summer conditions similarly affected WUE_i in both species, as shown by the parallel
487 decline in this variable in spring and summer, especially for *A. maritima*, when the plants'
488 water status decreased to -1.1 MPa. A high evapotranspiration rate in summer was
489 responsible for the reduction in WUE_i especially in the case of *O. drummondii*, where the
490 photosynthetic rate increased during summer. Scorce et al. (2019) also studying *A*
491 *maritima*, recorded summer values of Ψ_{md} as low as -2 MPa, Thus the Ψ_{md} values
492 registered in this study are moderately high. The neighbourhood of the sea might provide
493 moisture to these species during the night and early morning, causing the recorded values
494 in LWP were not as negative as in other Mediterranean species. Nonetheless, *O.*
495 *drummondii* always displayed higher RWC values along with higher Ψ_{md} values than *A.*
496 *maritima*. This fact denoted a different strategy in water use or access and would be
497 indicative of better water status of the invasive species.

498 Under winter conditions both species presented similar photosynthetic rates but
499 significant differences in hydric status. Photosynthetic activity varies with environmental
500 conditions, and although plants can acclimate to these changes, given the tropical origin
501 of the non-native species, lower photosynthetic rates could be expected in the cold season
502 compared with the native one. Leaves can have different photosynthetic rates under the
503 same environmental conditions due to different stomatal conductance caused by internal
504 or external factors. It has been suggested that changes in the response of cell-wall
505 conductance may be an important factor in temperature acclimation (Makino et al., 1994).
506 The high stomatal conductance in *O. drummondii* leaves along the year could be the
507 underlying factor in its optimal winter response and better water status.

508 Furthermore, the maintenance of high intercellular CO_2 concentration (data not
509 shown) associated with the lowest annual photosynthetic rate in *O. drummondii* suggests
510 the occurrence of non-stomatal limitation of photosynthesis in winter, while in *A.*
511 *maritima*, high intercellular CO_2 concentration values combined with the highest annual
512 photosynthetic rate recorded suggests the occurrence of stomatal limitation in the native
513 species. Also, high A_n/C_i values recorded in *A. maritima* indicate a non-mesophyll
514 limitation on photosynthesis.

515 The highest LDMC and LMA values recorded in the native species leaves in all
516 seasons indicate that this species would be relatively tougher and more resistant to

517 physical stresses than *O. drummondii*, whose leaves have lower LDMC and LMA. Leaf
518 mass area can be interpreted as the cost of light interception at the leaf level (Gutschick
519 and Wiegel, 1988; Poorter et al., 2009). Increases in LMA are often associated with
520 drought exposure and high irradiance (Gratani and Bombelli, 1999). According to these
521 results, higher fitness could be expected in the native *A. maritima*, and although the
522 species presented significant higher photochemical efficiency in the most stressful period
523 of the year, the highest photosynthetic rate was always measured in the invasive species
524 *O. drummondii*. On the other hand, these higher photosynthetic rates in the invasive
525 species are consistent with lower LDMC values recorded, since this functional trait
526 indicates the species' resource use strategy and represents the compensation between
527 assimilation and rapid growth *versus* efficient conservation of resources (Wilson et al.,
528 1999; Hodgson et al., 2011). The results in LDMC values would indicate a different
529 strategy of resource use for both species; with *A. maritima* displaying a more conservative
530 strategy as a predictor of poor environments and *O. drummondii* presenting a rapid
531 resource acquisition and use strategy as a predictor of rapid growth and soil fertility.
532 These outcomes would reinforce the concern that should accompany the invasion by this
533 species due to its rapid resource assimilation and growth which promote its expansion
534 (Garnier et al., 2001; Díaz et al., 2004).

535 Pigment study is important from an ecophysiological perspective as it provides
536 information about productivity, stress or limiting nutrients. Across the most stressful
537 seasons of the year, winter and summer, the invasive species has the advantage over the
538 native of its higher chlorophyll content but also higher Chl a/b ratios in autumn and
539 summer and higher Car/Chl ratios in winter. These variables are both representative of
540 stress. The Chl a/b ratio is indicative of the light absorption capacity of photosystem I to
541 II; declines in this variable are indicative of differential degradation of Chl a concerning
542 Chl b. Under stress conditions, a more rapid destruction of Chl a than Chl b might occur;
543 whereas Car/Chl decreases point out to damage in the photosynthetic apparatus, which is
544 expressed by a faster breakdown of chlorophylls than carotenoids (Lichtenthaler and
545 Buschmann, 2001). This higher pigments concentration can be a response to avoid
546 photochemical damage and an increase in reactive oxygen species due to the combination
547 of low temperatures with high radiation in Mediterranean winter. When the incident light
548 intensity is higher than necessary for photosynthesis reactive intermediates can be
549 produced in excess, leading to oxidative damage (Aro et al., 1993). The presence of

550 photoprotective carotenoids under this combination of winter stressful conditions is
551 believed to be a feature of the photoprotective function of carotenoid pigments (Havaux
552 and Kloppstech, 2001) and consequently, the invasive species could have better stress
553 tolerance or better maintenance of physiological functions under unfavourable
554 environmental conditions.

555 Principal component analysis established the presence of three groups
556 corresponding to seasons (summer of both species and *O. drummondii* winter). According
557 to Wilson et al. (1999), LDMC seems to be the best variable for locating plant species on
558 a resource use on the first axis and also as predictors of the different strategies for the
559 invasive and native species on resources use. The accumulation of Chl and Car, the other
560 variables defining this first axis, can be also related to resources used as optimal growth
561 conditions, as well as sufficient nutrients are positively correlated with pigment contents
562 and levels (del Campo et al., 2000; He et al., 2013). In contrast, axis two seems to separate
563 the two species and seasons according to its water use strategy.

564 In summary, even though *O. drummondii* has a tropical/subtropical origin, our data
565 show that this species is well acclimated to Mediterranean coastal dunes, exhibiting a
566 higher photosynthetic rate and better water performance than the native species *A.*
567 *maritima*. We conclude that *O. drummondii* utilizes light, water and probably nutrients
568 more efficiently than the native *A. maritima* and suffers lower stress in environments with
569 reduced water availability and high light radiation. Our results seem to indicate that the
570 underlying mechanism that confers *O. drummondii* its competitive capacity may be its
571 ability in water uptake, which makes it an effective non-native invasive species; the only
572 limiting factor to control its expansion in the Mediterranean coastal dune could be low
573 temperatures. In the study area, removal projects have been carried out to restore invaded
574 dunes (García-de-Lomas et al., 2016, personal observation) and they have proven the
575 difficulties to maintain the population level below an impact threshold.

576 As the authors, Flores-Moreno and Mole (2013), sustain it is evident that all non-
577 native species are not super plants with high invasive capacity and capable of maintaining
578 a higher fitness than the co-occurring native species in their new ranges. Furthermore,
579 studies comparing non-native and native species performance are contradictory (Daehler,
580 2003; Palacio-López and Gianoli, 2011). However, in the case of *O. drummondii* our
581 findings suggest that it is particularly successful in the Mediterranean range (including
582 other distribution areas with Mediterranean climate, Csa, where this species is expanding

583 such as the coast of Israel and the west coast of Australia). Therefore we would advise
584 managers to take measures for their eradication in the early stages of their appearance in
585 the Mediterranean dune ecosystems since it is extremely successful in these habitats.

586

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588

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