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Biorenewable materials for water remediation

Mishra, Kirti; Siwal, Samarjeet Singh; Sithole, Thandiwe; Singh, Nirankar; Hart, Phil; Thakur, Vijay Kumar

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Review

Biorenewable materials for water remediation: The central role of cellulose in achieving sustainability

Kirti Mishra^a, Samarjeet Singh Siwal^{a,b,c,*}, Thandiwe Sithole^b, Nirankar Singh^a, Phil Hart^d, Vijay Kumar Thakur^{c,e,*}

^a Department of Chemistry, M.M. Engineering College, Maharishi Markandeshwar (Deemed to be University), Haryana 133207, India

^b Department of Chemical Engineering, University of Johannesburg, Doornfontein 2088, South Africa

^c Biorefining and Advanced Materials Research Center, Scotland's Rural College, Edinburgh EH9 3JG, United Kingdom

^d Renewable and Sustainable Energy Research Centre, Technology Innovation Institute, Abu Dhabi 9639, United Arab Emirates

e School of Engineering, University of Petroleum & Energy Studies, Dehradun 248007, India

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ABSTRACT

As the population increases and manufacturing grows, greenhouse gas and other harmful emissions increase. Contaminated with chemicals such as dyes, pesticides, pharmaceuticals, oil, heavy metals or radionuclides, wastewater purification has become an urgent issue. Various technologies exist that can remove these contaminants from wastewater sources, but they often demand high energy and/or high cost, and in some cases produce contaminant laden sludge that requires safe disposal. The need for methods which are less capital intensive, less operationally costly and more environmentally friendly is suggested. Cellulose-based materials have emerged as promising candidates for wastewater treatment due to their renewability, low cost, biodegradability, hydrophilicity, and antimicrobial property. In this review article, we focussed on developing sustainable and biodegradable cellulose-based materials for wastewater treatment. This article deals with cellulose-based materials' scope and their conversion into valuable products like hydrogel, aerogel, cellulose composites, and nanocellulose. The cellulose-based materials have no harmful environmental impact and are plentiful. The modified cellulose-based materials applying as membrane, adsorbent, sorbent, and beads to purify the wastewater were discussed. Finally, the challenges and future prospects of cellulose-based materials for wastewater treatment were considered, emphasizing their potential to be sustainable and eco-friendly alternatives to traditional materials used in wastewater treatment.

Introduction

Water pollution is a major environmental challenge threatening public health and ecosystem sustainability. Wastewater treatment is essential for removing contaminants from water before it is released into the environment (Hernández-Delgado, 2015; Brooks et al., 2016). Traditional materials used in wastewater treatment, such as activated carbon and metal oxides, have several limitations, including high cost, limited availability, and environmental concerns (Sheoran et al., 2022; Siwal et al., 2022c; Kaur et al., 2023a). Cellulose, the most abundant natural polymer, is a promising alternative for wastewater treatment applications due to its renewability, sustainability, biodegradability, and low cost (Sanyang et al., 2016; Tu et al., 2021). In recent years, significant progress has been made in the synthesis, processing, and applications of cellulose-based materials for wastewater treatment (da Silva et al., 2023;

* Corresponding authors.

E-mail addresses: samarjeet6j1@gmail.com (S.S. Siwal), Vijay.Thakur@sruc.ac.uk (V.K. Thakur).

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Motloung et al., 2023; Nawaz et al., 2023). After cellulose, lignin is the second most abundant natural biopolymer and is used in various applications like paper making and the production of ethanol (Thakur et al., 2014).

From ancient times, the human being was highly dependent on hydrocarbon fuel resources. These non-renewable resources by definition deplete over time, and their use leads to serious environmental concerns like global warming and depletion of the ozone layer. Much interest has been taken in developing sustainable biomass to overcome this situation, mainly made up of lignin, cellulose, and hemicellulose (Mussatto and Dragone, 2016; Eqbalpour et al., 2023; Sigmond et al., 2023). Cellulose-based materials are readily available and are biodegradable, renewable, and economical-friendly. Cellulose is converted into water-soluble materials like cellulose acetate (hydroxypropyl) and methylcellulose. Cellulose consists of 1 000–15 000 glucose (β) units arranged linearly in a chain to form a polymer. Cellulose is a semi-rigid polymer having 150 MPa strength (Wang et al., 2012; Bin Bakri and Rahman, 2022; Rana et al., 2023b). The nano-constituents in cellulose can be extracted using mechanical and chemical methods. New forms of cellulose, like cellulose nanocrystals (CNCs) and cellulose nanofibrils (CNFs), can be extracted by the development of nanotechnology. By the breakdown of amorphous domains, rod-shaped CNCs are formed. CNFs comprise side-by-side amorphous and crystalline parts of cellulose and are flexible, long, and thin. Nowadays, nanocellulose is an active candidate/field for study within material science (Habibi, 2014; Shen et al., 2020).

Microorganisms also produce cellulose, which is extracellular and known as bacterial cellulose. In addition to starch, alginate, xylenes, chitosan, and silk fibroin, this bacterial cellulose can be used to form biocomposites. In addition, bacterial cellulose is also utilized in bioelectronics and biomedicine (Li et al., 2018b). These cellulose-based materials can also be used in various applications like treating wastewater, catalysis, energy storage, pharmaceuticals, etc. For water treatment, cellulose-based materials are utilized as flocculants, adsorbents, and in oil-water separating membranes. Further, it was found that cellulose fibres obtained from *Canna indica* waste, could be used to treat domestic wastewater after valorisation (Sharma et al., 2023). Pine needle based cellulose materials were reported in wastewater treatment as well (Rana et al., 2023a). Further, various cellulose-based materials are used for the removal of heavy metal by adsorption, for example sugarcane bagasse for removal of Cu(II), Cu(IV) and Hg ions from wastewater (Karnitz et al., 2007), wood sawdust to remove Cd(II) ion, and sugarcane bagasse and coconut tree sawdust to remove Zn(II), Pb(II), and Cu(II) from polluted water (Pranata Putra et al., 2014). These examples show the sustainability and opportunity of cellulose-based materials in wastewater treatment. This review focuses on converting cellulose into valuable products for wastewater treatment methods often rely on energy-intensive processes and generate large amounts of sludge and chemical waste. There is an increasing interest in developing sustainable and efficient materials for wastewater treatment to address these concerns. This article explores the scope of cellulose-based sustainable materials, highlighting their processing techniques and various applications in wastewater treatment.

Scope of cellulose-based materials

Cellulose is a versatile natural polymer with a wide range of properties and applications. The use of cellulose-based materials has been explored in fields including biomedical, energy, packaging, and environmental applications (Shao et al., 2023; Won et al., 2023).

In wastewater treatment, cellulose-based materials have shown great potential as sustainable and eco-friendly alternatives to traditional materials. The scope of cellulose-based materials for wastewater treatment includes developing novel materials, modifying existing materials, and integrating them with other technologies to enhance their performance. Cellulose-based materials can be used as adsorbents, membranes, and catalysts to remove various pollutants from wastewater, including heavy metals, organic compounds, and dyes (Liu et al., 2020; Sayyed et al., 2021). The scope of cellulose-based materials for wastewater treatment also includes their use in decentralized wastewater treatment systems, where low-cost and efficient materials are required. Cellulose-based materials can be produced from various natural and renewable sources, such as agricultural waste, wood, and bacterial cellulose, which highlights their potential within sustainable and circular economy applications (Ribul et al., 2021; Siwal et al., 2021a; Mujtaba et al., 2023).

Various sources provide energy, like solar, wind, hydrothermal, and fuel cells in daily life (Kasaeian et al., 2023; Lee et al., 2023). However, the main problem associated with these sources is the high cost of equipment and catalysts (Armor, 1999; Etim et al., 2020; Mishra et al., 2023b). Cellulose is inexpensive and has low lifetime carbon emissions, hence, they are widely used in the economic treatment of wastewater (Sheoran et al., 2022, 2023). Cellulose-based materials are naturally available and can remove both inorganic and organic pollutants from wastewater using methods such as chemical precipitation, ultrafiltration, evaporation, and adsorption (Peng et al., 2020; Siwal et al., 2022a).

Research shows that approximately 10 000 t cellulose can be obtained per year from plant stalks, straw, shells, plant fibres, grasses, etc. (Azizi Samir et al., 2005). The primary cellulose sources are bacteria, algae, wood, tunicates, plants, etc. (Moon et al., 2011). Cellulose is a polymer in which glucose is the monomer attached by 1,4 β -linkage. Due to some functional groups like carboxylic acid and hydroxyl, cellulose can combine with some organic species. These organic species provide extra active groups that make cellulose suitable for diverse applications. These functionalized groups help the modification of cellulose into nanoparticles (NPs) with metal-organic frameworks (MOFs), graphene oxide (GO), Ag NPs etc. (Liao et al., 2018; Askari et al., 2022). Cellulose-based materials for water treatment are popular due to their chemical inertness, rich surface chemistry, and large surface area. Due to these properties, nanocellulose is particularly suitable for making filters and membranes that exhibit good ability to remove pollutants from contaminated water (Voisin et al., 2017). For the treatment of drinking water, nanofiltration membranes are very suitable. Cellulose-based membranes are often selected due to their relatively low cost, biodegradability, and energy saving compared to petrochemical-based membranes (Li et al., 2022). For water treatment, coagulation or flocculation is widely used, in which coagulants (chemicals) are added to polluted water to cause neutralization of charged particles present in water. Neutralized particle aggregation

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occurs followed by precipitation of impurities. Coagulation processes driven by chemical additives produce large amounts of alkaline sludge and create secondary pollutants. To overcome this situation, cellulose can be used as an eco-friendly coagulant alternative (Abdelhamid and Mathew, 2021a). To adsorb the metal ions in contaminated water, functional groups in cellulose play a significant role. Further, to enhance the rate of adsorption, cellulose materials are modified by amination (Araki et al., 2001), sulphonation (Septevani et al., 2020), phosphorylation, xanthation (Pillai et al., 2013), and thiols (Choi et al., 2020). These modifications also strengthen the bond between metal ions and adsorbents. Ongoing research and developments in the field of cellulose-based materials for wastewater treatment are expected to yield new and innovative solutions for addressing water pollution challenges.

Processing of cellulose-based materials into valuable products

Processing methods vary depending on the specific cellulose-based material, the desired products, and the planned applications. Investigators and enterprises are actively developing novel processing approaches seeking to improve the value and versatility of cellulose-based materials, widening their utilization in different sectors. Natural polymers like cellulose have many applications within textiles, fuel, catalysis, paper, and pharmaceuticals (Mülhaupt, 2013; Nasrollahzadeh et al., 2021). Research has uncovered opportunities for the conversion of these polymers into novel nanomaterials, such as CNCs and CNFs, which show excellent and unique characteristics superior to other cellulose. For instance, cellulose's surface area increases from 4 to 500 m²/g by converting it into nano papers using a mechanical approach (Banavath et al., 2011). In water treatment removing pollutants like heavy metals, organic materials, and pharmaceutical waste are critical steps to make it suitable for drinking or other agricultural use (Wang, 2019). Nanocellulose has a large surface area and high strength, making it particularly suitable for selectively permeable membranes and water purification filters. Cellulose-based material can also be converted into other valuable products such as hydrogels, aerogels, nanocomposites, and membranes.

Aerogels

Cellulose-based aerogel is a solid porous material. Kistler (1931; 1932) synthesized aerogels for the first time in 1930 by eliminating liquid in a wet gel via supercritical drying. Preparation of cellulose aerogel requires three steps: dissolving cellulose or a derivative of cellulose, formation of cellulose gel via means of the sol-gel procedure, and drying cellulose gel to retain its 3D assembly (Long et al., 2018). The cellulose aerogel preparation process with their further modification is shown in Fig. 1.

Cellulose aerogels have more compressive strength and superior biodegradability than silica and synthetic polymer-based aerogels. Cellulose aerogel has excellent performance as a separator of oil/water and carrier of metal nanoparticles in adsorption, heat insulation, etc. Depending on the source, cellulose aerogels are classified into three classes: natural cellulose aerogels (including nanocellulose aerogels and bacterial cellulose aerogels), renewed cellulose aerogels, and cellulose derivative aerogels. The properties of these different kinds of cellulose aerogel are listed in Table 1.

Hydrogel

Hydrogels (HGs) are made of a natural or synthetic substance arranged in a three-dimensional network with considerable flexibility. HGs can also retain much water and have a soft rubber-like structure. These properties lead to their use in many applications (Siwal et al., 2022b). The substantial part of HGs is a cross-linked network of various polymeric chains, with the appearance of a three-dimensional mesh as shown in Fig. 2. The empty spaces are filled with water or any other fluid (Ullah et al., 2015).

Cellulose contains many –OH groups that provide useful structures and characteristics. By physical crosslinking in the cellulose solution, cellulose-based hydrogels (CBHs) can be synthesized. Some of the derivatives of cellulose that are water soluble are used in various forms like emulsifiers, surfactants, binding agents, food additives etc., as they are biocompatible. Some specific derivatives of cellulose like hydroxypropyl cellulose (HPC), methylcellulose (MC), hydroxypropyl methylcellulose (HPMC), and carboxymethyl cellulose (CMC) are utilized for HGs synthesis via physical and chemical crosslinking. Physical crosslinked HGs contain electrostatic interaction, H-bonding, and associative polymer-polymer interaction, however, it will break in chemical crosslinking formation and covalent bonds. (Chang and Zhang, 2011).

The CBHs are very popular in removing dyes, heavy metals from contaminated water via adsorption. They are soluble in organic solvents, biodegradable, abundant, cheap, and have a high surface area and adsorption capacity, etc. CBHs can be recycled after use. Their hydrophilic ability makes them a conspicuous candidate for wastewater treatment (Akter et al., 2021). CBHs can be shaped into various physical structures like beads, spheres, blocks, cylinders, nano, microparticles and thin films using different synthesis processes. Multiple approaches to CBH synthesis are shown in Fig. 3.

The CBHs remove pollutants in wastewater, such as heavy metals, dye, and oil. Part of the CBH applications are listed in Table 2. Some CBHs lose their pollutant adsorption ability after regeneration. To investigate this issue, some researchers concluded that regeneration power is regained when the renewal is done in NaOH (Satari et al., 2019). CBHs can be modified by chemical and physical reactions to adjust or increase their adsorption ability. The study of the composition and structure of CBHs shows that they have a positive effect on wastewater treatment (Pathak and Navneet, 2017).

Other cellulose composite

Cellulose and chitosan-based composites in membrane-based techniques have achieved much attention (Thakur and Voicu, 2016). In some cases, cellulose is used as an adsorbent when modified with sodium montmorillonite (NaMMT) and acts as a composite

Table 1

Properties of different kinds of cellulose aerogel.

Material	Drying method	Porosity (%)	Density (g/cm ³)	Specific surface area (m ² /g)	Reference
Natural cellulose aerogel					
Nano-fibrillated cellulose, SiO ₂	Freeze-dried	85.15-96.46	0.055-0.295	11.3-700.1	Fu et al., 2016
CNC, SiO ₂	Ambient pressure drying	-	0.137-0.151	620-688	Li et al., 2017
CNC	Supercritical CO ₂ dried	91-95	0.078-0.155	216-605	Heath and Thielemans, 2010
BC	Supercritical CO ₂ dried	-	0.008	200	Liebner et al., 2010
BC, Silica	Freeze-dried	89-99.6	0.007 - 0.229	129-541.1	Sai et al., 2014
BC, GO	Freeze-dried	99.84-99.92	-	-	Wang et al., 2014
Regenerated cellulose aerogel					
Cellulose fibres	Freeze-dried or Supercritical CO ₂ dried	-	0.01-0.06	80-250	Hoepfner et al., 2008
Cellulose powder	Supercritical CO ₂ dried	91-99	0.009-0.137	120-230	Karadagli et al., 2015
Cotton linter	Supercritical CO ₂ dried	95.5 – 98.1	0.03-0.067	190-328	Pircher et al., 2016
Microcrystalline cellulose	Supercritical CO ₂ dried	91–96	0.06-0.3	200-300	Gavillon and Budtova, 2008
Bagasse	Freeze-dried	84.4 - 94.2	0.088 - 0.236	119–185	Chen et al., 2016b
Waste newspaper	Freeze-dried	98.2 - 98.9	0.017-0.029	296-412	Fan et al., 2017
Paper pulp	Supercritical CO ₂ dried	95	0.12-0.17	363-406	Cai et al., 2009
Cellulose derivative aerogel					
Tri acetyl cellulose	Supercritical CO ₂ dried	96.1 – 99.6	0.005-0.05	229-958	Fang et al., 2016
Carboxymethylcellulose/CNF	Freeze-dried	93.19-96.84	0.05-0.109	-	Chen et al., 2016a
2,2,6,6-tetramethylpiperidine-1-oxyl; (TEMPO)-CNF	Freeze-dried	98.8 - 99.5	0.008 - 0.187	12.72-117.8	Jiang and Hsieh, 2014
Hydrophobic CNF	Freeze-dried	98.5	0.023 2	18.4	Mulyadi et al., 2016
CNF-MA	Freeze-dried	-	$0.011\ 2-0.031\ 5$	19.5	Kim et al., 2015

Notes: CNC, cellulose nanocrystal; BC, bacterial cellulose; GO, graphene oxide; CNF, cellulose nanofibril; TEMPO, 2,2,6,6-tetramethylpiperidine-1-oxylradical; MA, maleic acid.

Table 2

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Applications of cellulose-based hydrogel for removal of pollutants.

Cellulose-based hydrogel	Type of pollutant removed	Synthesis process	Working mechanism	Reference
Carboxymethyl cellulose HG beads	Cu(II), Ni(II), and Pb(II)	Inverse suspension crosslinking	Electrostatic and coordination interactions	Yang et al., 2011
Cellulose-graft-acrylic acid HGs	Cd(II), Ni(II), and Pb(II)	Grifting reaction mechanism	Electrostatic connections and ion exchange	Zhou et al., 2012b
Chitin/cellulose composite HGs	Hg(II), Pb(II), and Cu(II)	Freezing/thawing	Electrostatic and coordination interactions	Tang et al., 2011
Carboxymethyl chitosan/poly (acrylonitrile) HGs	Cu(II), Cd(II), and Co(II)	Crosslinking	Electrostatic interaction	Mohamed et al., 2012
CMC/2-acrylamido-2-methyl propane sulfonic acid HGs	Co(II), Cu(II), Mn(II), and Fe(III)	Copolymerization and crosslinking	Electrostatic and chelating interactions	El-Hag Ali, 2012
TEMPO-oxidized cellulose HGs	Zn(II), Fe(III), Cd(II), and Cs(I)	Nitroxy radical catalysed oxidation	Electrostatic interactions and ion exchange	Isobe et al., 2013
Sugar cane bagasse cellulose and gelatin-based composite HGs	Cu(II)	Crosslinking	Electrostatic and coordination interactions	Maity and Ray, 2017
Cellulose-porous bentonite composite	Azo dye	Crosslinking	Electrostatic interaction	Santoso et al., 2019
Cellulose-based bio-adsorbent	Acid blue, methylene blue dye	Graft copolymerization	Electrostatic interactions and H-bonding	Liu et al., 2015
Chitogen/cellulose HGs	Congo red dye	Freeze-dried	Electrostatic and coordination interactions	Tu et al., 2017
CMC-acrylic acid adsorbent	Methyl orange, disperse blue 2BLN, and malachite green chloride	Graft polymerization	Electrostatic interactions	Zhang et al., 2014a
Carboxymethyl cellulose structured nano-adsorbent	Methyl violet	Sol-gel method	Electrostatic and π - π interactions	Sharma et al., 2020
Lignocellulose-g-poly(acrylic acid)/montmorillonite 3D crosslinked polymeric network HGs	Methylene blue dye	Copolymerization	Electrostatic interactions	Shi et al., 2013
Nanocomposite HG Super adsorbent cellulose-graft-acrylic acid	Crystal violet Methylene blue dye	Graft polymerization Free-radical polymerization	Electrostatic interactions and H-bonding Electrostatic interactions	Mahdavinia et al., 2013 Zhou et al., 2012b

Notes: HGs, hydrogels; CMC, carboxymethyl cellulose.

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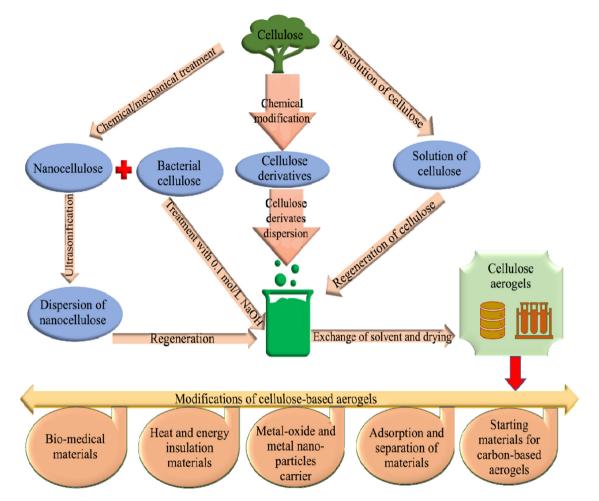


Fig. 1. Fabrication process of cellulose aerogels with their further modifications.

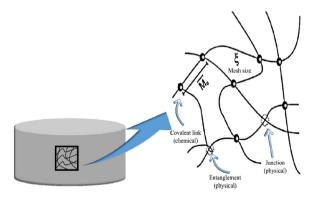


Fig. 2. The structural arrangement of hydrogel (Ullah et al., 2015).

material. NaMMT-modified cellulose composite could be used to eliminate Cr(VI) from aqueous media and be regenerated up to 10 times by using NaOH (Kumar et al., 2012). Chitosan-based cellulose composites can also be fabricated by mixing cellulose with chitosan and used to remove heavy metals from wastewater (Wu et al., 2010; Zhang et al., 2013).

Sun et al. (2009) synthesized a composite to adsorb Cu(II), Pb(II), Ni(II), and Cr (VI) by combining cellulose with chitin using ionic liquids that are eco-friendly. Various materials are used to provide robust capability in the adsorbent when removing heavy metals from polluted water. These materials accelerate the adsorption compared to cellulose. Li et al. (2015) fabricated a nanocomposite by mixing TiO₂ with cellulose fibres (CF), i.e., CF/TiO₂ nanocomposite. The CF/TiO₂ nanocomposite exhibited good efficiency for the

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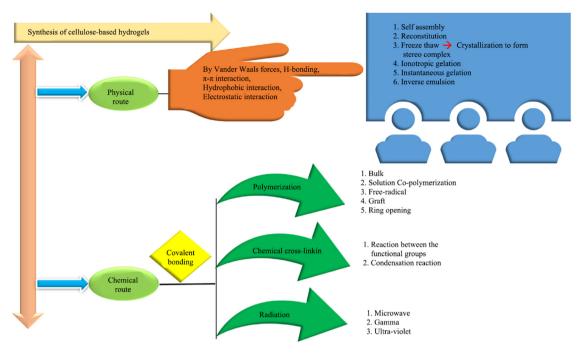


Fig. 3. Synthesis process of cellulose-based hydrogel.

Table 3

Various sources of nanocellulose with their average size.

	Cellulose nanocrystal	Cellulose nanofibril	Bacterial nanocellulose
Origin	Cotton, flax, wood, hemp, straw, etc.	Algae, straw, wood, hemp, tubers, cotton, bacteria, etc.	Alcohols and sugar
Average size (length×width)	(100–1 000) nm×(5–70) nm	Several μ m × (5 – 60) nm	Several μ m × (5 – 70) nm

Source: Wang, 2019.

adsorption of Pb (371 mg/g), which is higher than TiO₂ (20 mg/g). In follow-on work, Li et al. (2016) used CF/TiO₂ nanocomposite with a bed of fibrous membrane to adsorb Pb. For this purpose, he used a single-pass flow approach. The result showed that the adsorption capacity increased 9 times compared to the self-assembled bed and 13 times compared to the bare bed of CF.

Other physical arrangements of cellulose composites were also applied, such as beads, to enhance the adsorption capacity. Chitin/cellulose beads form through the coagulation process in which H_2SO_4 is used as a coagulating agent. Zhou et al. (2004) fabricated chitin/cellulose-based beads to adsorb Cu(II), Pb(II), and Cd(II) from an aqueous solution of thiourea and NaOH. The rate of adsorption of ions is Pd > Cd > Cu. Further, a cellulose composite was fabricated using cellulose acetate (CA) with sulphonated poly(ether imide) and poly(ethylene glycol) 600. This composite was used to adsorb heavy metals like Cd(II), Ni(II), Zn(II), and Cu(II). The order of adsorption is Cd > Zn > Ni > Cu (Nagendran et al., 2008; Barakat, 2011). A cellulose composite was fabricated using chitin with benzo-15-crown 5 in butyl methylimidazolium chloride, which exhibited enhanced adsorption efficiency and high mechanical strength. This is due to the synergistic consequences of combined materials compared to pure form (Mututuvari and Tran, 2014).

Cellulose can also be modified to work in a certain pH range. Qiu et al. (2014) examined ethyl cellulose composite's activity for removing Cr(VI) over a broad pH range of 3 - 11. Hokkanen et al. (2016) fabricated a cellulose composite using hydroxyapatite micro-fibrillated cellulose to adsorbed Cr(VI).

Nanocellulose is a lightweight material obtained from plants consisting of nano-sized cellulose fibrils. Nanocellulose is present in various forms like CNCs, CNFs, etc. Nanocellulose is eco-friendly and water-resistant due to film formation by counter-ion exchange (Rajinipriya et al., 2018). On average, the extraction of nanocellulose is problematic from the plant because of the presence of H-bonding and Vander Waals forces in cellulose. Extracting pure cellulose from plant matter requires pretreatment to remove wax, ash, hemicellulose, lignin, and other waste material. The process of cellulose-based materials extraction requires various steps as shown in Fig. 4. Various nanocellulose materials obtained from multiple sources with diverse particle sizes are shown in Table 3.

Nanocellulose is hydrophilic owing to multiple hydroxyl groups linked to its molecular spine. The existence of these hydroxyl clusters is an attractive characteristic when nanocellulose is used as a supporting representative within hydrophilic polymers. The hydrophilic character of nanocellulose allows dispersal in hydrophobic matrices (Missoum et al., 2013). The incompatibility of nanocel-

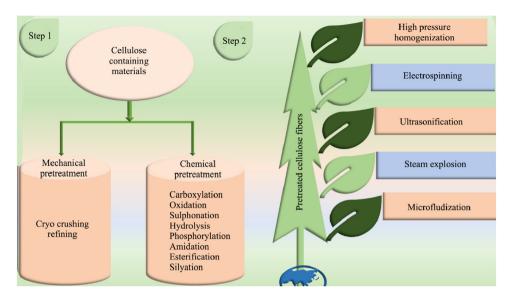


Fig. 4. Process of extraction of cellulose-based materials.

lulose with non-polar matrices disrupts the interfacial relations and eventually results in accumulation and phase partition. Consistent distribution of nanocellulose is important to obtain materials with enhanced mechanical characteristics. Therefore, widening the application of nanocellulose within numerous industrial areas requires further post-production efforts, such as esterification, chemical grafting, silylation, and oxidation, to develop the required surface features. Different methods (Fig. 5) are followed to tune the interfacial chemistry of nanocellulose to create the necessary hydrophilic-hydrophobic steadiness in a besieged environment (Dhali et al., 2021).

New tree-like cellulose-based nanofiber membranes have been synthesized using an electrospinning approach. For this purpose, tetra butyl ammonium chloride in small volumes was added within the CA, which further followed through deacetylation. The fabricated membrane exhibited a small size of pore, resistance toward solvents, and hydrophilicity with good mechanical properties (Zhang et al., 2018). The fabrication process is shown in Fig. 6a. The pore sizes of distinct nanofiber membranes were estimated. Compared to the typical nanofiber membrane, the pore size of a nanofiber membrane with a tree-like network was significantly reduced. As shown in Fig. 6b, the pore size of the common CA nanofiber membrane (C-CA) and deacetylated common CA nanofiber membrane (D-C-CA) membranes was 3.6 and 3.3 µm, while the pore size of CA tree-like nanofiber membrane (T-CA) and deacetylated CA tree-like nanofiber membrane (D-T-CA) membranes was 0.69 and 0.65 μ m. The slight pore size reduction after deacetylation is due to a slight swelling of the nanofibers. The contact angle of C-CA and D-C-CA membranes (Fig. 6c1 and 6c2) were 74° and 11°, respectively. The T-CA membrane (Fig. $6c_3$) had a hydrophilic contact angle of 62° , and the D-T-CA membranes (Fig. $6c_4$) showed excessive hydrophilicity with a measured contact angle of 0° Fig. 6d illustrates the stress-strain arcs of the C-CA nanofiber membrane and T-CA nanofiber membrane before and after deacetylation. All the specimens were hot-pressed at 60 °C and 120 MPa for around 3 min. With respect to mechanical effects, the C-CA membrane retained a tensile strength of 1.9 MPa and elongation-at-break of 0.94 %. The tensile strength and elongation-at-break of deacetylated membrane were 1.83 MPa and 1.69 %, respectively. The different solvent resistance of the D-T-CA and the T-CA membrane is illustrated in Fig. 6e. The T-CA nanofiber membrane appeared to shrink and dissolve, while the D-T-CA nanofiber membrane had no change. This demonstrated that the performance of the D-T-CA nanofiber membrane was stable for separating organic solvents.

The adsorption of pollutants from wastewater depends on many conditions like temperature, pH, and time of contact with the adsorbent. Table 4 illustrates the possible cellulose modifications for wastewater treatment. For wastewater treatment, cellulose-based materials are essential and abundant. These materials show natural binding ability towards different applications. The adsorption of cellulose-based biopolymers can be improved by their pretreatment and then be further modified into valuable products.

Relationship between preparation of cellulose-based materials and water treatment

The relationship between the preparation of cellulose-based materials and water treatment is a fascinating and important intersection of science, technology, and environmental concerns.

Cellulose is the most abundant organic compound on earth and is commonly present in plants and some aquatic animals. It has received scientific and technological attention for its versatile applications in various industries such as the production of food, pharmaceuticals, paper, and textiles, etc., due to their widespread availability, low price, recyclability, biodegradability, and non-toxicity. Cellulose-based materials, often derived from renewable sources such as wood, cotton, or agricultural residues, possess unique properties that make them suitable for water treatment processes. These materials can be prepared and engineered in different forms, such as membranes, filters, adsorbents, and flocculants, to address various and specific water purification challenges.

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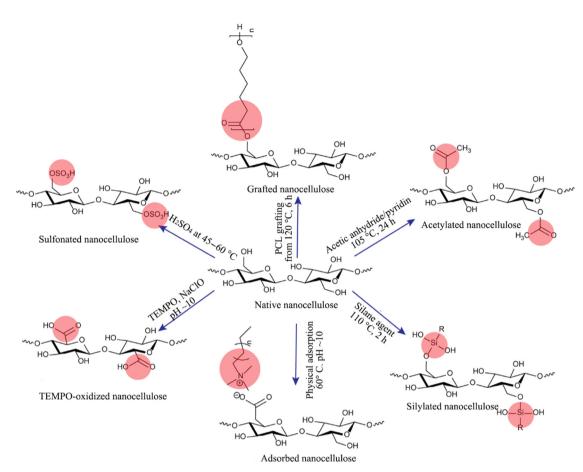


Fig. 5. Common surface modification methods used in nanocellulose production. The presented functional groups are marked in red (Dhali et al., 2021).

Table 4

Various types of modified cellulose composites with their specifications.

Modification in cellulose	Modified material	Removed impurity	Adsorption capacity	Reference
Composite beads	Chitosan	Cu(II)	0.84 mmol/g	Li and Bai, 2005
Nanocomposite	Titanium dioxide	Pb(II)	371.0 mg/g	Li et al., 2015
material	(TiO ₂)			
Composite material	Sodium montmorillonite	Cr(VI)	22.2 mg/g	Kumar et al., 2012
Magnetic nanocomposites of cellulose	Iron oxide	Pb(II)	21.5 mg/g	Xiong et al., 2014
Prepared hydroxyapatite	Calcium	Cr(VI)	2.208 mmol/g	Hokkanen et al., 2016
microfibrillated cellulose composite	hydroxyapatite			
Composite material	Chitin/chitosan, ionic	Cu(II)	0.417 mmol/g	Sun et al., 2009
	liquids	Pb(II)	0.127 mmol/g	
		Zn(II)	0.303 mmol/g	
		Ni(II)	0.225 mmol/g	
		Cr(VI)	0.251 mmol/g	
Cellulose and mercaptobenzothiazole composite	Mercapto benzothiazole	Hg(II)	204.08 mg/g	Krishna Kumar et al., 2013
Ethyl cellulose composites	Polyethylenimine	Cr(VI)	36.8 mg/g	Qiu et al., 2014
Collagen/cellulose beads composite	Bayberry tannin	Pb(II)	1.352 mmol/g	Zhang et al., 2015
Cellulose beads	Chitin, NaOH,	Pb(II)	0.33 mmol/g	Vijayalakshmi et al.,
	H_2SO_4	Cd(II)	0.32 mmol/g	2016
		Cu(II)	0.30 mmol/g	

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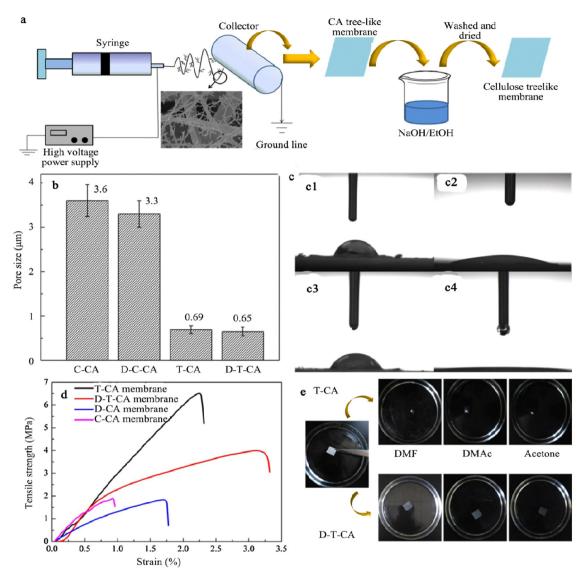


Fig. 6. (a) The new tree-like cellulose-based nanofiber membrane synthesis process; (b) Pore size distributions of various nanofiber membranes; (c) Pictures of the water contact inclination with distinct nanofiber membranes: (c_1) C–CA membrane, (c_2) D-C-CA membrane, (c_3) T-CA membrane, and (c_4) D-T-CA membrane; (d) Tensile-strain curves of various nanofiber membranes; (e) Optical photos of T-CA and D-T-CA membranes absorbed into DMF, DMAc, and acetone (Zhang et al., 2018).

CA: cellulose acetate; C–CA: common cellulose acetate nanofiber membrane; D-C-CA: deacetylated common cellulose acetate nanofiber membrane; T-CA: cellulose acetate tree-like nanofiber membrane; DMF: dimethylformamide; DMAc: Dimethylacetamide.

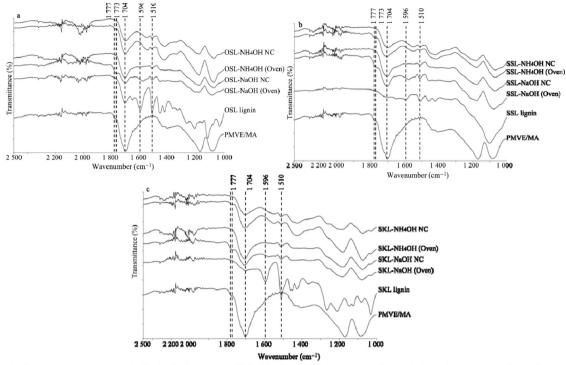
The use of renewable cellulose obtained from agriculture and biomass feedstock in wastewater treatment not only reduces carbon emissions but may also diminish the cost of wastewater treatment. The functionalization of cellulose by the -OH groups can significantly extend its applications in the treatment of wastewater (Peng et al., 2020). In their study of several cellulose-based adsorbents, Peng et al. (2020) confirmed that the various chemical modifications of cellulose by surface functional groups like -COOH, succinate, -NH₂, imidazole, -SH, amidoxime, thioether, -SO₃H, dithiocarbamate, xanthate, phosphate, etc. can significantly enhance their adsorption properties and selectivity towards removal of cations, anions, heavy metals etc. (Rocky et al., 2023).

Interaction mechanism of cellulose-based materials in water treatment

Cellulose-based materials can be characterized using advanced spectroscopy approaches and theoretical calculations to gain insights into their structure, properties, and interaction mechanisms (Mahmoodi et al., 2022). In this section, we summarise how these techniques are utilized.

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OSL: wheat straw organosolv lignin; NC: non-crosslinked hydrogel; PMVE/MA: poly (methyl vinyl ether) co-maleic acid; SSL: wheat straw soda lignin; SKL: softwood kraft lignin.

Fig. 7. Fourier transform infrared spectrum (Domínguez-Robles et al., 2018) OSL: wheat straw organosolv lignin; NC: non-crosslinked hydrogel; PMVE/MA: poly (methyl vinyl ether) co-maleic acid; SSL: wheat straw soda lignin; SKL: softwood kraft lignin.

Infrared (IR) spectroscopy: IR spectroscopy is generally used to study cellulose-based materials to reveal the functional groups present in the material and their relations. By analyzing the absorption bands in the IR spectrum, investigators can determine cellulosespecific peaks, such as those associated with the cellulose backbone (e.g., C–O stretching), hydroxyl groups (-OH), and intermolecular hydrogen bonding (Raucci et al., 2015; Stricher et al., 2021). IR spectroscopy can also explore the influence of processing or changes on cellulose materials.

To remove cationic dye from wastewater, a biocompatible hydrogel was fabricated by crosslinking between lignin and poly (methyl vinyl ether) co-maleic acid (PMVE/MA) via ester linkages. A green synthetic process was used to form hydrogels, and the byproduct of the reaction was H₂O. Fourier transform infrared (FTIR) analysis was applied to establish the chemical property of crosslinking (Fig.7). For the PMVE/MA, the obtained spectra show bands at 1 704 cm⁻¹ for the carbonyl group, 1 777 cm⁻¹ for maleic anhydride. Further, ranges of the hydrogels treated over 80 °C show two peaks of the carbonyl group: a peak at 1 773 cm⁻¹ for the anhydride group and another peak at 1 708 cm⁻¹ for the acid carbonyl group. In addition, a shift in the C=O peak was observed from 1 708 to 1 735 cm⁻¹ showing the esterification process of the crosslinked hydrogels (Domínguez-Robles et al., 2018).

Raman spectroscopy: Raman spectroscopy is used to reveal data about molecular vibrations and crystal structure. Raman spectra can determine cellulose polymorphs (e.g., cellulose I and cellulose II) and show structural changes caused by processing or modifications (Schenzel et al., 2009; Kim et al., 2013). Raman spectroscopy is also sensitive to hydrogen bonding relations and can deliver insights into crystallinity and disorder in cellulose materials.

The X-ray Diffraction (XRD): XRD broadly defines the crystal structure and degree of crystallinity in cellulose-based materials. By studying the diffraction patterns, investigators can specify the structure of cellulose chains and evaluate the crystalline and amorphous areas (Garvey et al., 2005; Agarwal et al., 2017). XRD can also help explain differences in the crystal structure caused by different treatments or changes. Further, cellulose-based composites are fabricated by using a one-step hydrolysis method that consists of Bismuth (Bi) NPs, which are mainly nanoplates of Bi oxychlorides (BiOCl). The BiOCl biochar efficiently removes azo anionic dye and methyl orange dye by photocatalytic degradation and adsorption. Using scanning electron microscope (SEM) and high resolution-SEM, it was shown that the acquired material consisted of homogenous platelet/disk-similar shaped NPs (Fig. 8a) with a size of 15–25 nm and diameter within the scope from 55 to 160 nm, with an average of 95 nm. The pristine alkaline lignin-based biochar had a consistent configuration (Fig. 8b) of a not-densely compacted essence, where irregular cells and ruptures with onsets at the macro-scale appear. SEM images of the materials (Fig. 8c) showed that the network and morphology of either biochar or BiOCl NPs was very consistent, with the former acting as a material towards the NPs. The crystalline structure of the fabricated material was K Mishra S S Siwal T Sithole et al

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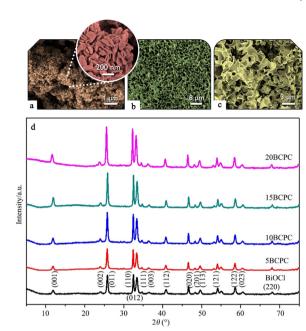


Fig. 8. Scanning electron microscope pictures of (a) BiOCl, (b) lignin-derived biochar, and (c) BiOCl biochar compound 15BCPC. (d) The XRD patterns of pure BiOCl NPs and their composites with various quantities of lignin-derived biochar (Singh et al., 2023).

examined by XRD, as shown in Fig. 8d. The XRD results of the pure BiOCl NPs fit well with the standard tetragonal data of BiOCl (JCPDS No. 85–861), with the individual reflection at (011), (110), and (012) to be the prime ones. As expected, the values of the textural parameters predictable from the N₂ sorption tests were not so high, which is in agreement with analogues described in the literature for the same class of materials. The BiOCl sample offered a specific surface of $19 \text{ m}^2/\text{g}$ and a non-porous structure (total pore volume of 0.037 cm³/g), with the formed "pores" to as a result of the formed inter-particles, inter-clusters, or/and inter-aggregate cages/spaces within the NPs (Singh et al., 2023).

Solid-state nuclear magnetic resonance (NMR) spectroscopy: Solid-state NMR spectroscopy is used to explore cellulose-based materials' local structural circumstances and dynamics. It provides data about molecular mobility, intermolecular interactions, and chemical shifts (El Hariri El Nokab et al., 2022). NMR can examine cellulose carbon environments, for example the cellulose spine and side groups, and observe changes resulting from chemical transformations or processing. In this way, the -OH groups were investigated using the quantitative 31P NMR technique published by two researchers Granata and Argyropoulos (1995). The NMR spectra were attained using a Bruker 500 MHz spectrometer (Billerica, MA, USA). The NMR parameters were 1 024 scans with a pulse delay of 5 s, 90 °C pulse and line boarding of 2 and default baseline correction. Further, a lignin-based bio-adsorbent was fabricated by ammonization and sulphonation of lignin, namely SAPL-1.5, that contained some particular functional groups with spatial cross-linking arrangements. To characterize SAPL-1.5, 31P, 1 H, 13C NMR are used and compared with bare lignin as shown in Fig. 9. The 31P NMR spectroscopy is a simplistic and straight analysis technique for calculating the -OH groups in carboxyl groups, phenolic hydroxyl, aliphatic hydroxyl, etc. for lignin. The intensity of the signals at $1.384-1.371 \times 10^{-4}$ denoting p-hydroxyphenyl components was noticeably improved compared to that of lignin, indicating that the phenol was very successfully grafted onto the lignin. By contrast, the intensity of the signals belonging to Al-OH in phenolated lignin ($1.46-1.49 \times 10^{-4}$) was noticeably decreased, which indicated that phenolate reactions take place in the side chain of lignin in an acidic medium (Wang et al., 2017a).

Theoretical calculations: Theoretical calculations like density functional theory, molecular dynamics, and Monte Carlo simulations, can complete practical spectroscopic approaches. These calculations can be used to understand cellulose-based materials' atomic-level structure, energetics, and mechanical effects (Ramezanzadeh et al., 2018; Qin et al., 2023). By modeling the relations among cellulose chains, solvents, additives, or other molecules, investigators can anticipate and comprehend the behaviour of cellulose-based materials under diverse circumstances.

The hybrid of advanced spectroscopy characterization approaches and theoretical calculations allows a thorough knowledge of cellulose-based materials, including their design, properties, and interaction mechanisms with other materials (Kaur et al., 2023b; Mishra et al., 2023a; Siwal et al., 2023). This understanding is essential for developing enhanced cellulose-based materials and their applications in different areas, such as packaging, textiles, biomedical materials, and renewable energy.

Applications of cellulose and other biobased materials in wastewater treatment

Cellulose-based materials have gained significant attention in recent years as a promising alternative for wastewater treatment due to their biodegradability, abundance, and sustainability. The unique properties of cellulose-based materials, such as their high

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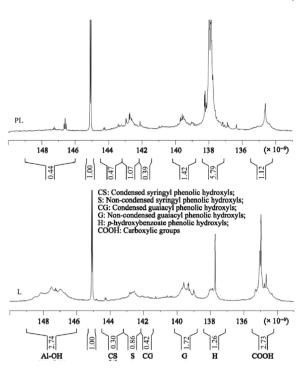


Fig. 9. The 31P nuclear magnetic resonance spectrum of lignin and phenolated lignin (Wang et al., 2017a).

surface area, porous structure, and surface charge, make them suitable for various applications in wastewater treatment (Kabir et al., 2018; Tu et al., 2021). One of the most common applications of cellulose-based materials in wastewater treatment is as adsorbents for removing contaminants such as heavy metals, organic pollutants, and dyes (Sajid et al., 2018). The high surface area of cellulosebased materials allows them to effectively adsorb these contaminants through various mechanisms such as electrostatic attraction, complexation, and ion exchange (Zare et al., 2021).

Cellulose-based materials have been explored as membranes for wastewater treatment. The porous structure of cellulose-based membranes allows for the selective separation of contaminants based on their size, shape, and charge. Furthermore, cellulose-based membranes can be modified to enhance their filtration efficiency and selectivity (Jamshaid et al., 2017; Jalvo et al., 2021). In addition to adsorption and membrane applications, cellulose-based materials have shown promise as catalysts for the degradation of organic pollutants in wastewater. Cellulose-based catalysts can be synthesized by incorporating metal nanoparticles or enzymes onto the surface of cellulose-based materials, which enhances their catalytic activity and stability (Liang et al., 2018). Cellulosebased materials can also be integrated with other technologies to improve their performance in wastewater treatment. For instance, cellulose-based adsorbents can be combined with photocatalysts to enhance their removal efficiency under UV light irradiation.

However, challenges such as the limited mechanical strength, chemical stability, and low reusability of cellulose-based materials must be addressed to achieve practical applications in wastewater treatment (Dong et al., 2021; Wu et al., 2022b). Ongoing research and development in this field aims to overcome these challenges and advance the utilization of cellulose-based materials for wastewater treatment, leading to sustainable and eco-friendly wastewater treatment technologies.

Freshwater can be polluted by various processes including industrial, agricultural, and domestic activity. This water contamination can harm humans and aquatic flora and fauna as it may consist of harmful chemicals like dye and heavy metals. Further, water can be polluted by the oil and petroleum industry, including its various stages like oil production, spilling oil during transportation, refining of oil etc. (Munirasu et al., 2016). These toxic effects imply a requirement for some suitable methods to treat industrial water before releasing it into water bodies. Water contains mainly three types of contaminants: inorganic matter, organic matter, and suspended solid particulate. Various processes like filtration, absorption, adsorption, flocculation etc., are used to treat wastewater. Adsorption mainly removes oils, organic matter, and heavy metals in contaminated water. Cellulose contains many hydroxyl groups on its surface and as such they can entrap chemical species and hence adsorb the pollutants over its surface (Varghese et al., 2019). Solid suspended materials are removed using the flocculation method. Solutions based on natural cellulose are effective, relatively cheap and control environmental pollution by decreasing the emission of carbon (Peng et al., 2020; Sayyed et al., 2021). Various wastewater treatment techniques are shown in Fig. 10 with different cellulose-based nanomaterials.

Removal of dyes

Dyes, used by various industries to colour materials, are the major pollutants that create many problems in humans, and have been linked to cancer and respiratory disorder and are harmful to the ecosystem (Kishor et al., 2021). There are numerous methods

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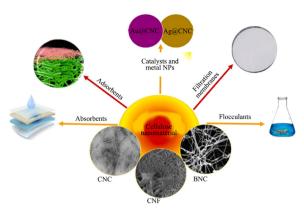


Fig. 10. Schematic representation of various wastewater treatment techniques from cellulose-based nanomaterials (Sayyed et al., 2021).

available to remove dyes from wastewater, but most are costly and some are known to release toxic byproducts. Removing dyes from wastewater using cellulose-based materials has achieved significant attention due to their superior adsorption capability and renewability. The cellulose-based materials are non-toxic, biodegradable, clean (non-polluting), and sustainable method for wastewater treatment (Siwal et al., 2021b; Mishra et al., 2023c).

Different forms of cellulose, like fibres, stalks, bark, seed, shell roots, are used as adsorbents. To eliminate dyes like methylene blue (MB) and basic blue 9, cellulose-based materials like garlic peel (Hameed and Ahmad, 2009), grass waste (Hameed, 2009a), and jackfruit peel (Hameed, 2009b) have been used to make adsorbent with capacities of 142.86, 457.64, 285.73 mg/g, respectively. To remove methylene red (MR) dye from contaminated water, Ahmad et al. (2019) fabricated a cellulose-based material with lemongrass leaf-based activated carbon. The result demonstrates an adsorption capacity of the dye was 76.923 mg/g at the pH over 2 for 5 h. It was found that with the increase in the dye concentration, temperature, and contact time, the MR dye adsorption rate increased. This indicates that the overall process is endothermic and occurs through physical adsorption.

Jawad et al. (2016) fabricated an adsorbent for the adsorption of MB dye from contaminated water. They used H_2SO_4 -treated activated carbon obtained from the coconut leaf. The adsorbent had an irregular and rough surface with many cavities and gave a high adsorption capacity of 149.3 mg/g for MB dye. The adsorption efficiency increased further with the rise of the initial concentration of MB. On the other hand, Ding et al. (2014) fabricated an adsorbent by activated carbon obtained from rice husk to remove rhodamine B dye with an adsorption volume of 478.5 mg/g for 5 h. The pH of the solution did not affect the adsorption of the dye. Chemically activated carbon obtained from jute sticks was used to eliminate brilliant green dye from wastewater with an adsorption volume of 480 mg/g (Asadullah et al., 2010). In another study, guava leaf-based activated carbon was fabricated showing an adsorption capacity of 39.7 mg/g for removing congo red dye (Ojedokun and Bello, 2017). Using the orange peel, Lam et al. (2017) prepared activated carbon with an adsorption volume of 28.5 mg/g for the adsorption of malachite green dye. Further, to adsorb the bromophenol blue and bromothymol blue present in wastewater, raw maize cob are used with an equilibrium time of 125 and 110 min, correspondingly. The adsorption of dye corresponded with the Temkin isotherm model (Dada et al., 2020).

Siddiqui et al. (2019) fabricated a cellulose-based nanocomposite that acted as an adsorptive to eliminate MB dye from polluted water. The nanocomposite comprised antimicrobial Nigella sativa seed-based MnO_2 with black cumin (MnO_2/BC nanocomposite). The adsorption capacity for as-prepared nanocomposite for MB dye was 185.19 mg/g at 318 K and pH of 7. The adsorption procedure proceeds by Langmuir isotherm and endothermic in nature. During the adsorption of MB over the surface of MnO_2/BC nanocomposite, hydrogen bonding and electrostatic attraction occurred between the dye and the surface.

Using exhausted coffee ground powder, rhodamine 6 G and rhodamine B adsorption was demonstrated from wastewater with adsorption capacities of 17.4 and 5.3 mg/g, respectively (Shen and Gondal, 2017). It was found that the adsorption of rhodamine dye over the surface of the adsorbent occurs through hydrophobic, electrostatic, and intermolecular interactions. Platanus oriental leaf powder (Peydayesh and Rahbar-Kelishami, 2015), pineapple leaf powder (Chowdhury et al., 2011), and pine tree leaves (Deniz and Karaman, 2011) were used as adsorbents for the elimination of MB dye, basic green 4, and basic red 46 dye correspondingly. Their maximum removal capacity is 114.94, 48.72, and 71.94 mg/g with a contact time of 70, 150, and 120 min, respectively. Neem leaf powder has also been reported for the adsorption of brilliant green dye, the adsorption capacity is 0.554 mmol/g for 240 min (Bhattacharyya and Sarma, 2003).

Lignin and lignocellulose-based gels are also used to remove dye from wastewater with some modification. Lignin-based lignocellulosic gel with a porous arrangement due to combination with 3-aminopropyltriethoxysilane is used to remove rhodamine B and MB dye. The adsorption capacity for lignin-containing sample is 192 mg/g, and for rhodamine B is 201 mg/g, which is higher than lignin-free materials, the process of adsorption proceeds through the chemisorption process (Zhang et al., 2019a).

Wang et al. (2017c) used a lignin-based hydrogel synthesized by combining it with acrylamide and N-isopropyl acrylamide to eliminate MB dye from wastewater. The process of adsorption was highly reliant on temperature and pH. The hydrogel has regeneration ability and can be used 5 times without performance decrease. For the removal of MB dye, a lignin/graphene-based aerogel was fabricated, in which cross-linking was done by chitosan. The aerogel removed up to \geq 99% MB dye with an adsorption capacity

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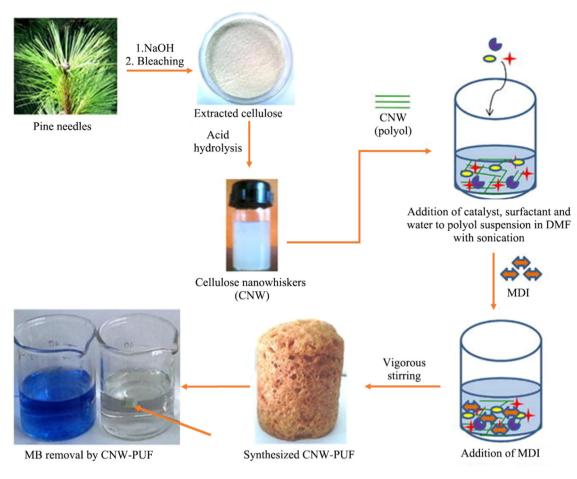


Fig. 11. Fabrication process of CNWs-PUF from pine needles with its approach to removing MB dye (Kumari et al., 2016) MDI: 4,40-diphenylmethane diisocyanate; PUF: polyurethane foam.

of 1 023.9 mg/g (Yan et al., 2019). While through the thermodynamic analysis, it was found that dye adsorption is an endothermic and spontaneous process. During the adsorption, hydrogen bonding, electrostatic interaction and π - π stacking occur between MB dye and lignin-based aerogel.

Cellulose nanowhiskers (CNWs) are also used to remove dye from waste and contaminated water. Kumari et al. (2016) produced a novel polyurethane foam (PUF) using CNWs. The acidic hydrolysis of cellulose was done from needles of pine plants and diphenylmethane diisocyanate. The PUF was used to remove MB dye from water effluents. It was observed that MB dye was wholly removed within 20 min at 55 °C. The absorption capacity of PUF was 554.8 mg/g, and it was regenerated and used for many cycles, the complete process of PUF fabrication with its working mechanism is shown in Fig. 11.

Removal of oil

In the oil and petroleum industry, water can be produced and polluted at every step of processing. Within the oil value chain, processing contributes approximately 10 % of polluted discharge with the remaining 90 % coming from transportation, pipelines, offshore platforms, etc. These consequences may be short-term or long-term (Aguilera et al., 2010; DeLaune and Wright, 2011). There is a need to develop more efficient and capable technologies to resolve the pollution problem related to oil spills and as such this area remains an active area of academic and industrial research (Adebajo et al., 2003; Corsellis et al., 2016).

The oil industry has two types of water: produced and processed. Contaminated water obtained during extraction consists of a mix of H_2O , minerals, dispersed oils, chemicals related to production, and dissolved gases (Veil et al., 2004). Compared to processed water, produced water has more mineral ions in dissolved form. Harmful chemicals in the wastewater, like NH_3 , C_6H_5OH , H_2S , benzene, toluene, ethylbenzene, and xylenes are commonly known as BTEX (Munirasu et al., 2016). The oil in the water makes a layer over the surface due to its hydrophobicity. The formation of an oil layer is harmful to aquatic and marine life. Various advanced techniques can remove the oil present in the water. Cellulose-based materials are popular due to their wide availability, low cost, and biodegradable ability.

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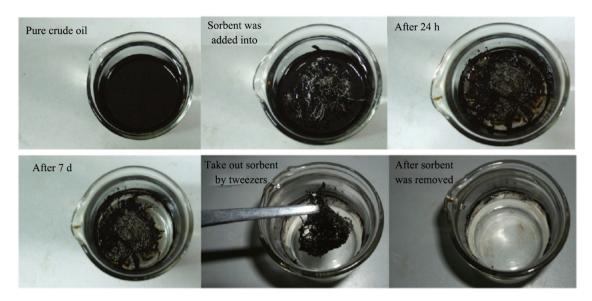


Fig. 12. Oil adsorption process of corn straw modified acetylated cellulose fibres (Li et al., 2013).

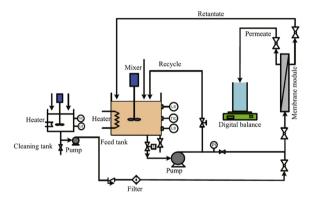


Fig. 13. Diagrammatic representation of cross-flow filtration process as performed in the laboratory (Madaeni et al., 2013).

Li et al. (2013) investigated modified cellulose fibres obtained from corn straw as a sorbent for oil from contaminated water. The cellulose fibres extracted from corn straw were acetylated to enhance their hydrophobicity. This prepared material shows oil sorption capacities of 42.53 g/g for pump oil, 52.65 g/g for diesel oil, and 57.64 g/g for crude oil at 25 °C for 1 h. The oil sorption procedure by corn straw-modified cellulose fibre is shown in Fig. 12. In this method, oil's cellulose fibre absorption is increased using an acetylation process that modifies the lipophilic surface of cellulose fibres.

Nguyen et al. (2013) fabricated cellulose-based aerogels which act as super absorbents for oil from contaminated water. For this purpose, they treated cellulose fibres with NaOH/ NH_2CONH_2 solution. The fibres were placed in a refrigerator for the gelation and then mixed in C_2H_5OH for coagulation. The solvent exchange process was done using deionized water, and the resultant mixture was freeze-dried for two days. After that, aerogel salinization was done using methyl trimethoxy silane (MTMS) over 343.15 K for 2 h. The highly porous, MTMS-doped, low-density cellulose-based aerogel led to superior oil sorbent capacity.

The effect of temperature was also examined. It was found that the higher sorption capability for rice bran crude oil was 24.4 g/g over 313.15 K. The cellulose aerogel regenerated coated with MTMS has a good affinity towards crude oil and acts as a suitable adsorbent for treating oil spills. Besides the adsorption approach, membrane filtration technique is reported as an excellent method to remove oil from wastewater (Zhang et al., 2014b). In the semi-batch process, the infused membrane was continued to be withdrawn, and the oil-improved non-permeate was reprocessed as shown in Fig. 13 (Madaeni et al., 2013).

Using cerium ion as initiator, CA and polyacrylonitrile were combined by a free radical polymerization process, and the ultrafiltration membrane CA-graft polyacrylonitrile (CA-g-PAN) was synthesized (Chen et al., 2009). The prepared CA-g-PAN exhibited superior permeability and excellent antifouling characteristics for the separation of oil/water emulsion.

It was observed that polyacrylonitrile, which is hydrophobic, increased the membrane pore size, resulting in enhanced permeation. The antifouling property of the prepared membrane was examined by flux recovery and flux decay ratio of oil/water emulsion. It was found that CA-g-PAN membranes showed 100 % removal capacity under different experimental conditions. Another group of

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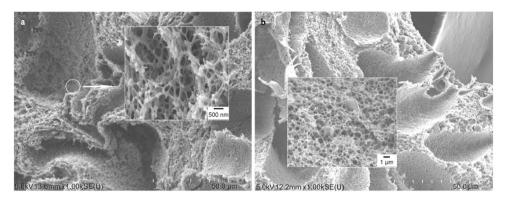


Fig. 14. Scanning electron microscope (SEM) pictures of (a) CA fibre and (b) CA/Z fibre (Ji et al., 2012).

researchers synthesized an ultrathin nanoporous cellulose-based membrane. For this purpose, they performed freeze-extraction of a heavily-diluted cellulose solution. This membrane was used to separate a nano emulsion of oil-in-water. The membrane had a thickness of 80-220 nm and a 10-12 nm cut-off. The membrane showed an ability of water permeation of 1 620 L/(m² h bar) with good size-selective separation characteristics. The efficiency for oil removal from nano emulsion is 96.5 % for the drops with a size of 200.4 nm (Zhou et al., 2014).

Removal of heavy metals

Cellulose-based materials suggest different advantages for heavy metal removal, such as their abundance, renewability, biodegradability, and versatility. Nevertheless, it is essential to assess the specific features of the heavy metals and the wastewater and the optimal processing necessities to maximize heavy metal removal efficiency utilizing cellulose-based materials. Various processes are reported to remove the heavy metals from wastewater such as chemical precipitation (Ku and Jung, 2001), ion exchange (Alyüz and Veli, 2009), membrane filtration (Landaburu-Aguirre et al., 2009), reverse osmosis (Shahalam et al., 2002; Mohsen-Nia et al., 2007), electrodialysis (Nataraj et al., 2007; Sadrzadeh et al., 2009), and adsorption (Park et al., 2007; Zhou et al., 2012a).

The cellulose was modified with maleic anhydride by Zhou et al. (2012a) and employed as an adsorbent to eliminate Hg(II) from wastewater. The outcomes showed that the prepared adsorbent exhibited superior adsorption capacity, i.e., 172.5 mg/g. Further, various experimental conditions like temperature, pH, contact time, were also investigated.

Thirumavalavan et al. (2010) showed the elimination of heavy metal ions from wastewater, like Cu(II), Zn(II), Cd(II), Ni(II), and Pb^{2+} by using three different kinds of cellulose-based materials like orange peel, banana peel, and lemon peel. For this purpose, chemical modification was performed over the lemon peel and cellulose surfaces using activated carbon. The Cu(II) and Ni(II) ions were adsorbed preferentially compared to other ions. While comparing the different adsorbents, it was found that surface-modified lemon peel cellulose exhibited increased adsorption capability. The order of adsorption ability is lemon peel cellulose with activated carbon > lemon peel, using the wet spinning approach.

Ji et al. (2012) fabricated a cellulose composite fibre, i.e., CA/zeolite composite fibres (CA/Z). In this process, zeolite particles were embedded over the surface of CA, which acted as the polymer matrix. The SEM images (Fig. 14) show the morphology of CA and CA/Z fibres. CA fibres have a porous structure with a 300-500 nm pore size, while zeolite particle size is 1 μ m.

As a result, the zeolite particles are retained within the CA/Z fibre network. The result shows that the CA/Z fibres with an average pore size of 24.6 nm allow all the heavy metal ions to pass for direct contact with the zeolite adsorptive site. These CA/Z fibres remove Cu(II) and Ni(II) ions from the wastewater using Langmuir adsorption isotherm. The adsorption capacity of CA/Z fibres is 28.57 mg/g for Cu(II) and 16.95 mg/g for Ni(II) at 25 °C, which can be regenerated over 5 cycles.

To eliminate heavy metals, Wang and Wang (2016) prepared a polyvinyl alcohol (PVA) and CMC hydrogel (PVA/CMC) via a freeze-thaw approach to mix PVA and CMC in different ratios and then frozen and melted in five different cycles. To measure the adsorption capacity of metal ions, i.e., Ag(I), Cu(II), Zn(II) and Ni(II), samples were immersed for 1 day in 0.01 % metal ion solution at 288 K. PVA/CMC showed good adsorption of Ag(I) ion compare with other metal ions because Ag(I) had small size and could quickly diffuses through the polymer network.

Removal of radioactive materials

Nuclear energy is now widely regarded as a green energy source due to the very low emission of greenhouse gases. Nuclear energy has ultra-high energy density and has a key advantage of offering a dispatchable baseload for power generation (Antal and Karhunmaa, 2018; Sovacool et al., 2020; Ahmad, 2021). Due to advancements and potential growth in nuclear energy, concerns arise related to the issue that could harm society due to the release of mining wastewater which contain uranium and other radioactive substances. In the extreme, radioactive materials can also be result within wastewater due to nuclear accidents like as evinced at Chernobyl and Fukushima (Song et al., 2014; Wang et al., 2018; Kong et al., 2020).

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Radioactive materials contained in wastewater can be very harmful to humans and the environment due to radionuclides' ultrahigh biological and chemical toxicity. This toxic effect leads to cancer organ damage with teratogenicity (Zhang et al., 2019b; Wang et al., 2020a). Therefore, there is an urgent requirement for some practical methods that remove these radionuclides from contaminated water. There are several ways to remove the radioactive waste from contaminated water such as ion exchange (Manos and Kanatzidis, 2012), extraction (Ahmad et al., 2020), chemical precipitation (Vigier et al., 2016), adsorption (Zhao et al., 2019), and membrane separation (Luo et al., 2019; Yang et al., 2022). The adsorption of these known methods has advantages of good economics, simplicity, and reusability with high decontamination ability. Various adsorbents are used to decontaminate radioactive wastewater, i.e., polymers, metal-organic framework, graphene, covalent organic frameworks, proteins, MXene, polypeptides etc. However, these synthetic materials have some limitations which may have their own adverse impacts on the environment (Li et al., 2018a; Deshwal et al., 2023).

Cellulose-based adsorbents are used to overcome the limitations of synthetic materials to treat radioactive wastewater. Cellulosebased materials are biodegradable and hydrophilic with antimicrobial property (Bai et al., 2020; Liu et al., 2021). However, cellulose does not contain sufficient functional groups to remove radioactive elements like uranium. For this purpose, natural cellulose have been modified by many researchers to increase the number of adsorption sites by introducing phosphonate, amine, and dialdehyde groups, which are widely used to capture uranium (Wu et al., 2022a).

In addition, a number of cellulose-based adsorbents derived from natural cellulose fibres like hemp fibres (Bai et al., 2020), bamboo fibres, and sisal fibres (Wang et al., 2020b) have been suggested. These may not be the most effective options due to their low specific surface areas, limited exposed reactive groups, and nonporous morphology. Other cellulose-based materials, such as sugarcane bagasse, lemon peels, pineapple peels, and coconut coir, are used to fabricate adsorbents through milling. These cellulose nanomaterials adsorbents have been used to remove radionuclides from contaminated water.

Rethinasabapathy et al. (2018) fabricated a composite known as FeAC/CMC/POSS by using Fe-amino clay, CMC, and polyhedral oligomeric silsesquioxane. This composite was used to adsorb Cs(I) from contaminated water. FeAC/CMC/POSS shows incredible adsorption ability of 152 mg/g for Cs(I), which may due to their layered morphology (Fig. 15a) and amino groups over the clay surface. In addition, the presence of carboxylate and hydroxyl groups on the surface of CMC might resulte in electrostatic attraction and ion-exchange mechanisms as shown in Fig. 15b.

Banana peels are also used to remove radioactive Th and U from polluted water. In a study, banana peels are converted into nano size (up to 12 nm) via milling, and used to remove radionuclides by binding with amine and carboxylic groups. It was observed that around 0.01-0.3 g of banana peels effectively remove 99.99 % Th(IV) from 100 mg/L, 18.8 mg/L initial concentration, and 70 % U(VI) from an initial concentration of 100 mg/L, 58.8 mg/L from synthetic and real polluted mine water, respectively (Oyewo et al., 2016). Ma et al. (2012) fabricated the ultrafine CNFs from wood pulp as an effective adsorbent for removing radioactive UO_2^{2+} from contaminated water. The wood pulp was chemically oxidated with 2,2,6,6-tetramethylpiperidine-1-oxylradical (TEMPO), sodium bromide, and sodium hypochlorite after mechanical disintegration. When the CNF suspension was added in UO_2^{2+} containing water, an interaction occurs between the -COO- of CNFs and UO_2^{2+} and forms a gel. The highest adsorption power of these ultrafine CNFs for UO_2^{2+} was 167 mg/g.

As shown in Fig. 16, cellulose-based nanomaterials are used in diverse forms to treat contaminated or polluted water, e.g., as a catalyst, adsorbent, membrane, flocculant, etc.

The extracts obtained from plants cells such as bacteria, fungi, actinomycetes, yeast, and algae are used for the synthesis of various bionanomaterials that are less costly and nontoxic (Mohanpuria et al., 2008; Schröfel et al., 2014; Zhang and Liu, 2020). As shown in Fig. 17, the bionanocomposites could give us an appealing substitute for removing radioactive ions of contaminated water from nuclear leakage.

Cellulose-based materials are very useful for developing countries to treat contaminated water because cellulose can be obtained naturally. For the large-scale implementation of these cellulose-based materials for adsorption of dye, heavy metals, radionuclides, pharmaceutical waste, oil removal, etc., more research is required.

Removal of other pollutants

Besides dye, oil, and heavy metal, wastewater also contains various undesirable inorganic pollutants, organic pollutants, and suspended particulate matter. These pollutants can be removed from wastewater using flocculation, filtration, adsorption, etc.

Suopajärvi et al. (2013) explored the flocculation of five different dicarboxyl cellulose (DCC) samples by cumulative levels of carboxylation. The reaction occurs among sodium periodate to prepare the DCC bleached birch chemical wood pulp and sodium chlorite. CNCs were fabricated with different carboxyl amounts, i.e., 0.38, 0.69, 0.75, and 1.75 mmol/g. The CNCs having 1.75 mmol/g carboxyl can remove turbidity (80 %) and chemical oxygen demand (60 %) from wastewater.

Additionally, Suopajärvi et al. (2014) examined a sulphonated CNC as a flocculating agent. Cellulose was attached with $-SO_3H$ group (Liimatainen et al., 2011) and the removal of turbidity was examined using 25 mg/L of $Fe_2(SO_4)_3$ to enhance the flocculation process. It was found that the CNCs had sulphonation and was helpful in removing turbidness from water at 2.5 mg/L.

By using DCC, Zhu et al. (2015) also studied the removal of turbidity from wastewater. Cellulose was prepared by adding NaOH and urea into bamboo pulp for 90 min at -12 °C. The cellulose was mixed with NaIO₄ for 5 h; the extra NaIO₄ was broken down by ethane-1,2-diol, and the remaining solution was cooled at room temperature. By continuing the reaction, DCC is produced, having a carboxyl amount of 2.57 mmol/g. This DCC can remove up to 99.5 % of turbidity from a solution of kaolin (500 mg/L) when 25 mg/L DCC was used with 300 mg/L CaCl₂.

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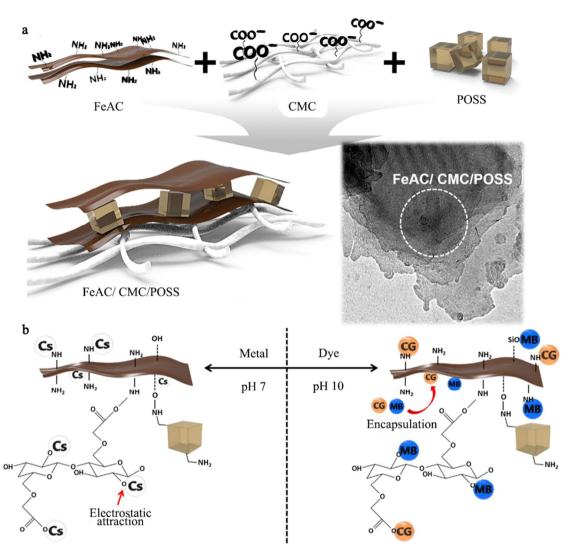


Fig. 15. (a) Schematic formation of FeAC/CMC/POSS and (b) adsorption mechanism of Cs by FeAC/CMC/POSS (Rethinasabapathy et al., 2018) FeAC: Fe-amino clay; CMC: carboxymethyl cellulose; POSS: polyhedral oligomeric silsesquioxane.

Alila and Boufi (2009) used an adsorbent to remove aromatic organic compounds from wastewater. A long hydrocarbon chain was added over a heterogeneous atmosphere to modify cellulose. For this purpose, N, N'-carbodiimidazole and 4,4'-methylene bis (phenyl isocyanate) were used to introduce the hydrocarbon chain in cellulose for modification. The cellulose fibres were modified using linear octyl anhydride and heterogeneous esterification. The modified material exhibited a high adsorption rate of 300 mol/g for organic compounds (aniline, benzene, nitrobenzene, mono or poly-substituted chlorobenzene, etc.) in water (Aloulou et al., 2006).

Pollution of water with F^- , NO_3^- , SO_4^{2-} , PO_4^{3-} , and NO_2^- is harmful to the health of humans. These species are known as anionic species. F^- penetrates the skin of humans and converts hydroxyapatite $[Ca_{10}(PO_4)_6(OH)_2]$ of bones and teeth into fluorapatite $[Ca_5(PO_4)_3F]$ by dissolving in it. To adsorb these anionic pollutants or remove these anionic pollutants, cationic cellulose or cellulose-based nanocomposites are mainly used. Anions like NO_3^- and F^- are effectively adsorbed by the cationic CNFs.

It was reported by Pei et al. (2013) that the adsorption of anionic dyes present in wastewater by CNFs was increased by adding trimethyl ammonium chloride. Sehaqui et al. (2016) fabricated the cationic CNFs using waste pulp residue containing positively charged quaternary ammonium groups. For this purpose, esterification of pulp was done with glycidyl trimethyl ammonium chloride, after that, mechanical disintegration was performed. The obtained material has a 1.2 mmol/g cationic charge amount. The as-prepared quaternary ammonium CNFs showed high adsorption of SO_4^{2-} and PO_4^{3-} as compared to F^- and NO_3^- .

Pollutants, such as antibiotics, pharmaceutical waste, cosmetics, and drugs are soluble in water and known as miscible pollutants. The miscible organic contaminants can be removed from water by adsorption or by breaking down into parts using techniques like UV–visible photocatalysis, redox reactions, and ozonation. For the adsorption of pharmaceutical products like paracetamol, N, N-diethyl-

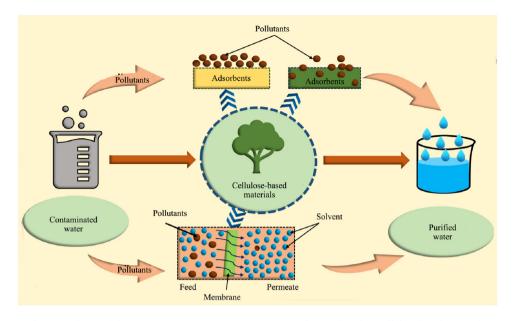


Fig. 16. Scheme of the water/wastewater treatment process mechanism based on cellulose-based materials.

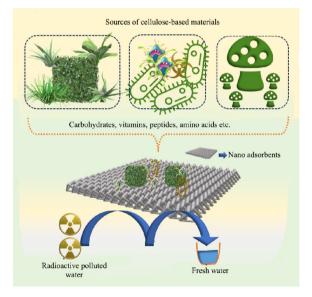


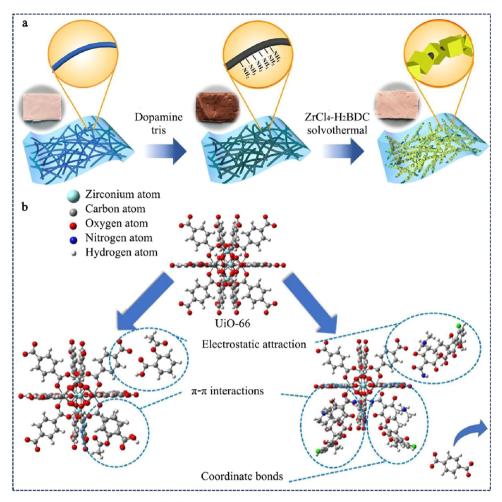
Fig. 17. Cellulose-based materials used for treatment of radioactively contaminated water.

meta-toluamide, sulfamethoxazole, nanocellulose in combination with Jeffamine ED 600 (a block copolymer) was fabricated, the adsorption process of this material based on electrostatic attraction (Herrera-Morales et al., 2017).

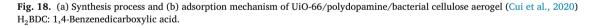
The CNC is also used to remove insecticides like chlorpyrifos from contaminated water. Chemometric examination shows a requirement of $\sim 1.5 \text{ g/L}$ nanocellulose in 20 min to adsorp 5 mg/L chlorpyrifos pesticide, offering an efficiency of 99.3 % (Moradeeya et al., 2017). In addition, to adsorp aspirin and tetracycline hydrochloride, a material consisting of UiO-66, BC, and polydopamine was fabricated. Outstanding adsorption ability was found by material for aspirin (149 mg/g) and tetracycline hydrochloride (184 mg/g). The fabrication process of the material with adsorption mechanism is shown in Fig. 18.

Advantages and disadvantages of cellulose based materials for wastewater treatment

Cellulose-based materials have numerous advantages as they are abundant natural biopolymers with many functional groups. The functional group present in these materials like -COOH, -OH can be used to react with several nanoparticles (NPs) like Ag-NPs, graphene oxide, ZnO NPs, and MOFs, and can be applied to various applications according to requirement. In the case of the water



H2BDC: 1,4-Benzenedicarboxylic acid.



decontamination process, they are helpful in the removal of heavy metals, drugs and cosmetics, radionuclides, oil and organic dyes. The advantages of cellulose-based materials for wastewater treatment are given below.

Cellulose is an abundant naturally occurring biopolymer and can be derived from different origins, such as plant biomass and agricultural scrap (Jha and Kumar, 2019). Its availability drives it as a sustainable and renewable alternative for wastewater processing applications. Cellulose-based materials have outstanding adsorption capabilities, permitting them to effectively withdraw a broad range of pollutants from wastewater (Zhu et al., 2023). They can adsorb contaminants such as heavy metals, dyes, organic compounds, and even pathogens, enhancing water quality. As a biodegradable material, cellulose can efficiently split down over time (George and Sabapathi, 2015). This property is helpful in wastewater treatment as it decreases the environmental effect of the utilized materials and assures proper removal after usage.

Cellulose-based materials are usually cost-effective compared to traditional wastewater treatment processes (Kausar et al., 2023). Since cellulose can be acquired from renewable sources, its raw material costs are low. Further, the manufacturing methods for cellulose-based materials are relatively easy and affordable. Cellulose-based materials can be engineered and altered to improve their performance in wastewater processing (Cai et al., 2017). By combining different functional groups or presenting nanoscale structures, the adsorption capability, selectivity, and general efficiency of cellulose-based materials can be enhanced for specific pollutants. The need for alternative harmful chemicals is also reduced by using cellulose-based materials for water treatment. Some common pollutants present in water and their removal process are shown in Fig. 19 (Abdelhamid and Mathew, 2021b).

The main disadvantage of cellulose-based materials is the generation of large amounts of sludge and its poor moisture resistance. These materials also have poor mechanical properties (Jayarathna et al., 2022). The disadvantages of the cellulose-based materials used in wastewater treatment applications are shown below. Cellulose-based materials can have restricted chemical stability in specific circumstances, particularly when exposed to strong acids, bases, or harsh chemical essentials (Kumar et al., 2017). This can lead

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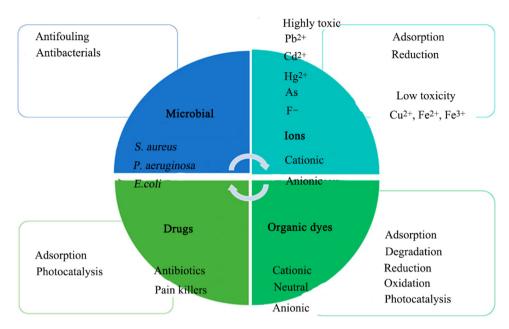


Fig. 19. Some common pollutants in wastewater and their removal process based on cellulose-based materials (Abdelhamid and Mathew, 2021b).

to degradation or decreased performance over time, affecting their efficacy in wastewater treatment. Cellulose-based materials usually contain water which affects their handling and processing (Symington et al., 2009). High water content can raise the materials' weight and bulk, making them challenging to transport and manipulate in large-scale wastewater treatment methods. The adsorption kinetics of cellulose-based materials are slow compared to other adsorbents (Tang et al., 2013). This suggests that extracting pollutants from wastewater may take longer time, needing a longer connection time or more amounts of the material for effective processing. While cellulose-based materials can withdraw a broad range of pollutants, their selectivity for particular pollutants may be determined (Tan et al., 2020). In some circumstances, there may be better choices for targeting specific contaminants. In addition, regeneration of cellulose-based materials after adsorption can be complex. Relying upon the involved specific contaminants and adsorption mechanisms, specialized procedures or additional actions may be required to fully restore the adsorption capability of the material (Ul-Islam et al., 2016).

Consideration of these advantages and disadvantages is essential when considering the appropriateness of cellulose-based materials for wastewater treatment applications. While they suggest considerable advantages, the wastewater's specific requirements and features should be considered to ensure optimal treatment performance.

Economic and environmental considerations

Cellulose has been gaining interest for applications in biomedical (Figueiredo et al., 2018), food (Khan et al., 2018), automobiles (Asadi et al., 2017), electronics (Wang et al., 2017b), personal care (Naderi et al., 2017), and environmental areas(Hamad et al., 2020). Because of their renewability and potential to address environmental challenges, cellulose-based materials have gained much attention for wastewater treatment. This section explores the economic and environmental considerations associated with cellulose-based materials for wastewater treatment. Cost-effectiveness, life cycle analysis, and eco-friendlyness of cellulose-based materials were discussed, highlighting their potential to achieve economic viability and environmental sustainability through their ability and flexibility to be altered or tuned for a specific application (Kim et al., 2018; Muqeet et al., 2020).

Economic considerations

Cost-effectiveness

Being abundantly available and derived from renewable sources, cellulose offers a cost advantage over synthetic materials. The production cost of cellulose-based materials for wastewater treatment can be relatively low, especially when utilizing agricultural waste or recycled cellulose sources (Peng et al., 2020; Williams and Wool, 2000). The potential for cellulose regeneration and reuse further contributes to cost-effectiveness.

Scalability

The scale-up of bio-resource retrieval processes is challenging due to budgetary limitations upon energy and resources. While many chemical routes are available for retrieving useful value products from waste substances, the economics must be sustainable. These limitations are also present in biomass re-utilization (Tiwary et al., 2015; Di Vaio et al., 2023). For example, with the production

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of CNC from waste biomass, different techniques are already available that work well at the bench scale. In contrast, not all can be scaled up to industrial production. It is thus essential to evaluate the feasibility of any future development by analysis and in practical pilot scale experiments to test if the salient characteristics of the bench scale chemical processes are scalable (Bui et al., 2018). Such investigations expose any emergent difficulties that occur during scale-up. Cellulose extraction processes and material fabrication techniques are well-established, enabling low risk, efficient, and reliable production on a commercial scale (Banerjee et al., 2010; Islam et al., 2017).

Integration with existing infrastructure

Cellulose-based materials can be integrated into existing wastewater treatment infrastructure without significant modifications. This adaptability reduces the capital investment required for implementing cellulose-based solutions, making them economically viable options for upgrading or retrofitting existing treatment systems (Dittenber and GangaRao, 2012).

Environmental consideration

Renewable and biodegradable property

Cellulose-based materials are derived from renewable sources such as wood, agricultural waste, and recycled paper, reducing reliance on non-renewable resources (Escursell et al., 2021). Moreover, cellulose is biodegradable, ensuring that the materials used in wastewater treatment do not contribute to long-term environmental pollution (Farghali et al., 2023).

Reduced energy consumption

The production and use of cellulose-based materials for wastewater treatment typically requires less energy than conventional treatment processes (Mohamed et al., 2017). The lightweight cellulose fibres and membranes reduces energy requirements for transportation, installation, and operation. Additionally, using cellulose-based adsorbents can minimize the need for energy-intensive regeneration processes.

Reduced chemical usage

The inherent properties of cellulose, such as its adsorption capacity and ability to facilitate biological processes, enable the removal of pollutants without excessive chemical dosing (Giachini et al., 2020). This reduces the environmental impact associated with chemical usage and the generation of chemical sludge.

Life cycle analysis

Life cycle analysis (LCA) provides a comprehensive assessment of the environmental impacts associated with cellulose-based materials throughout their entire life cycle, from raw material extraction to end-of-life disposal or recycling (Anastas and Lankey, 2000). LCA indicated that cellulose-based materials for wastewater treatment exhibits lower environmental burdens compared to conventional materials, particularly in terms of greenhouse gas emissions, energy consumption, and waste generation (Tsang et al., 2019).

Cellulose-based sustainable materials have been shown to offer economic advantages and environmental benefits for wastewater treatment applications. Their cost-effectiveness, scalability, compatibility with existing infrastructure, and their renewability and biodegradability contribute to their financial viability (Li et al., 2022). The reduced energy consumption, decreased chemical usage, and favourable environmental performance highlighted by LCA emphasizes the eco-friendly attributes of cellulose-based materials. These considerations make cellulose-based sustainable materials a good choice for achieving economic and ecological sustainability in wastewater treatment.

Conclusions

Cellulose-based sustainable materials have emerged as a promising alternative for wastewater treatment due to their unique properties and sustainability. Cellulose-based materials have shown potential in various applications, such as adsorption, membrane separation, catalysis, and integration with other technologies. For wastewater treatment, cellulose-based materials are very efficient and suitable as they are biodegradable, low in cost, and eco-friendly. The biomass-based materials are versatile as they effectively remove many pollutants, like oil, dye, heavy metals, suspended particulate matter, etc., from contaminated or pollutant water. In addition, these cellulose-based materials have superior characteristics like low-cost, biodegradable, and renewable, making them suitable for treating wastewater. In this review article, we discussed the scope of their use and their conversion into a valuable product in wastewater treatment, mainly focussing on removing dye, oil, heavy metal, turbidity. However, challenges remain to be addressed to achieve practical applications in wastewater treatment, including limited mechanical strength, chemical stability, and low reusability. The main challenges associated with water treatment are scaling up from laboratory to industrial level and cutting the cost of raw resources and industrial preparation. Ongoing research in this field aims to overcome these challenges and advance the utilization of cellulose-based materials for wastewater treatment, leading to sustainable and eco-friendly wastewater treatment technologies. Further research is required to explore novel, innovative biomass-based sustainable materials for multifunctional purposes with some alternative biomass raw materials. Overall, the development of cellulose-based sustainable materials for wastewater treatment has the potential to contribute to the goal of achieving a more sustainable and cleaner environment. The continued research and development in this field will be critical in enhancing the performance and practicality of cellulose-based materials for wastewater treatment, paving the way for a more sustainable future.

Declaration of Competing Interest

There are no conflicts to declare.

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