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<b>Citation</b>	Hamilton, C. M., Thogmartin, W. E., Radeloff, V. C., Plantinga, A. J., Heglund, P. J., Martinuzzi, S., & Pidgeon, A. M. (2015). Change in agricultural land use constrains adaptation of national wildlife refuges to climate change. <i>Environmental Conservation</i> , 42(1), 12-19. doi:10.1017/S0376892914000174
<b>DOI</b>	10.1017/S0376892914000174
<b>Publisher</b>	Cambridge University Press
<b>Version</b>	Version of Record
<b>Terms of Use</b>	<a href="http://cdss.library.oregonstate.edu/sa-termsfuse">http://cdss.library.oregonstate.edu/sa-termsfuse</a>

# Change in agricultural land use constrains adaptation of national wildlife refuges to climate change

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Date submitted: 25 September 2013; Date accepted: 25 March 2014; First published online 22 May 2014

## SUMMARY

Land-use change around protected areas limits their ability to conserve biodiversity by altering ecological processes such as natural hydrologic and disturbance regimes, facilitating species invasions, and interfering with dispersal of organisms. This paper informs USA National Wildlife Refuge System conservation planning by predicting future land-use change on lands within 25 km distance of 461 refuges in the USA using an econometric model. The model contained two differing policy scenarios, namely a ‘business-as-usual’ scenario and a ‘pro-agriculture’ scenario. Regardless of scenario, by 2051, forest cover and urban land use were predicted to increase around refuges, while the extent of range and pasture was predicted to decrease; cropland use decreased under the business-as-usual scenario, but increased under the pro-agriculture scenario. Increasing agricultural land value under the pro-agriculture scenario slowed an expected increase in forest around refuges, and doubled the rate of range and pasture loss. Intensity of land-use change on lands surrounding refuges differed by regions. Regional differences among scenarios revealed that an understanding of regional and local land-use dynamics and management options was an essential requirement to effectively manage these conserved lands. Such knowledge is particularly important given the predicted need to adapt to a changing global climate.

**Keywords:** climate change, conservation estate, econometric modelling, land-use change

## INTRODUCTION

Agriculture is the predominant land use across the globe and is expected to increase in area and intensity as the size of the

human population increases (Tilman *et al.* 2001; Ellis *et al.* 2010). Effective management of natural resources must thus include humans as a component of the system (Folke *et al.* 2010). Land use and appropriation of natural resources for human use are among the primary drivers of biodiversity loss and environmental degradation (Foley *et al.* 2005). The need to feed a growing human population is expected to simplify the landscape as natural spaces are removed in the process of agricultural intensification (Tilman *et al.* 2001). Agricultural landscapes are expected to face added pressure with increased demand for bioenergy feedstocks and wind-energy generation resulting in loss of grasslands and other habitat (Lu *et al.* 2009; Dale *et al.* 2011; Wright & Wimberly 2013).

Protected-area networks are the primary tool used to stem biodiversity loss caused by human-induced changes. Protected areas cover >12% of the Earth’s land surface (Hansen & DeFries 2007), however they are affected by their surroundings through ecological flows, and land use surrounding them constrains their effectiveness (Hansen & DeFries 2007; Franklin & Lindenmayer 2009). Unfortunately, many protected areas are surrounded by intensive or intensifying land use, diminishing their conservation value (Radeloff *et al.* 2010).

A major concern threatening future functioning of protected-area networks is climate change because of its potential impact on ecological systems and species (Griffith *et al.* 2009; Heller & Zavaleta 2009). Climate change causes the extirpation of range-restricted species, changes demographic rates, facilitates the occurrence and spread of disease and invasive species, and fosters habitat loss (Mawdsley *et al.* 2009). These impacts are similar to those of land-use change (Lepczyk *et al.* 2008), and the synergistic effects of land use and climate change may pose a severe threat to biodiversity conservation in general and protected areas specifically.

The question is: how to reduce anticipated adverse impacts from changing land use and climate (Griffith *et al.* 2009; Sohl *et al.* 2010)? Adaptation can include modifying practices and activities to reduce the vulnerability of ecological systems (Smith *et al.* 2000). Many of the management tools necessary for adaptation are already in use (Mawdsley *et al.* 2009). These tools include increasing the connectivity among protected

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areas, mitigating threats such as pollution and invasive species, interagency coordination, and the promotion of conservation biology concepts, such as increasing reserve size or improving representation in protected area networks (Heller & Zavaleta 2009). The ultimate goal of these activities is to reduce the vulnerability of native species to threats from changing land use and climate (Crossman *et al.* 2012).

A critical component of any adaptation strategy is the comprehensive assessment of conservation threats (Merenlender *et al.* 2009). Assessing the likelihood of future changes and the effectiveness of different policies for mitigating negative effects of those changes has been identified as one of the most pressing concerns facing land managers (Fleishman *et al.* 2011). The USA's Fish and Wildlife Service (FWS) identified the need for rigorous projections of possible future conditions for improving their management of the National Wildlife Refuge System (NWRS; Griffith *et al.* 2009). Adaptation requires examining changes in surrounding lands to improve future management decisions (Lindenmayer *et al.* 2008), and considering alternative futures. The use of scenarios, namely, plausible alternative futures based on a stated set of assumptions, can be an effective tool for evaluating likely changes (Coreau *et al.* 2009). Projections of future land use in the USA indicate that land use will change substantially, with the greatest amount of change in the eastern half of the country (Martinuzzi *et al.* 2013). For the NWRS, conservation-focused scenarios led to increased urban and range lands, and decreased crop and pasture lands by mid-century (Hamilton *et al.* 2013). However, those studies did not explore the effects of economic conditions leading to widespread increases in agricultural land use. Increased agricultural land use is a realistic scenario given projected increases resulting from, for example, increased food supply or bioenergy demands (Dale *et al.* 2011; Tilman *et al.* 2011).

The NWRS is one of two systems of USA federally-managed lands whose primary focus is the conservation of fish and wildlife resources (the National Park Service is the other). Our goal was to evaluate changes in land use surrounding refuges throughout the conterminous USA between 2001 and 2051 under two economic scenarios. We evaluated current and future land-use scenarios within 25 km of all refuges in the conterminous USA to answer four questions:

- (1) What is the current distribution of land uses around the National Wildlife Refuges?
- (2) What is likely change in land use by 2051 under a 'pro-agriculture' scenario?
- (3) How does that change compare with a 'business-as-usual' scenario?
- (4) How does land-use change vary across the USA?

## METHODS

### Study area

We analysed land-use change within 25 km of 461 National Wildlife Refuges (hereafter 'refuges') of the conterminous

USA. We excluded NWRS lands not directly managed by the FWS (namely, cooperatively managed lands) or not specifically designated as refuges in the FWS Cadastral database (<http://www.fws.gov/GIS/data/cadastralDB/index.htm>). Some areas fell within analysed distances for more than one refuge; these areas were counted more than once when calculating values for individual refuges, but were counted only once when calculating values for the refuge system as a whole. The 25-km distance was selected because it is a pertinent scale for extending management to the lands surrounding many of the refuges (Hamilton *et al.* 2013).

### Analyses

We quantified future land-use change around the refuges using predictions of the econometric land-use model of Radeloff *et al.* (2012). An advantage of this econometric model was that it permitted evaluating effects of different economic and policy scenarios on land-use change. For instance, an afforestation subsidy will alter the net return for forest land use and result in a change in land use transition probabilities among the different land uses because land owners may decide to convert their land to a use that is more profitable. This model accounted for costs associated with making changes in land use, soil capabilities, county-level net economic returns for particular land uses, and the USA's National Land Cover Dataset (NLCD; <http://www.epa.gov/mrlc/nlcd-2001.html>). With these variables, transition probabilities for urban, forest, range, agriculture, and pasture land uses were calculated between 2001 and 2051 as a multinomial logit model with  $p_{ijk} = F(\beta_{jk}X_i)$ , where  $p_{ijk}$  is the probability of parcel  $i$  changing from use  $j$  to use  $k$  between 1992 and 1997,  $\beta_{jk}$  is a vector of parameters associated with the  $j$ - $k$  transition,  $X_i$  is a matrix of independent variables for parcel  $i$  (current land use, soil capability, county-level net profit for a land use). The variables within  $X_i$  can be adjusted to simulate effects of economic changes or implementation of a policy on land-use transition probabilities (Lubowski *et al.* 2006). The results from the economic estimation postulate probabilistic land-use transition matrices for each county and each land capability class, with each matrix providing transition probabilities for a 5-year time step. We derived 50-year transition probabilities by applying matrix multiplication and combined the inputs in ArcGIS (ESRI 2011) to produce maps of land-use change. The output of the model is at 100-m resolution and gives the probability of a 100-m pixel transitioning from one use to another by 2051.

We considered two different scenarios of future land-use change in this analysis, a 'business-as-usual' scenario, reflecting observed transition probabilities from 1992 to 1997, and a 'pro-agriculture' scenario, providing for a 10% increase in agricultural returns every five years. We chose the pro-agriculture scenario at the request of land managers and regional policy developers, because they considered current economic conditions to be similar to that policy scenario. This subsidy is equivalent to a mean annual increase of +2%

per year over background rates of increasing return (+4.4% per year over 1990–2008; Nickerson *et al.* 2012) and is thus likely an upper limit of what may occur. In addition, the scenario included an assumption that no land entered or left enrolment in the Conservation Reserve Program (CRP). The CRP is the largest conservation programme in the USA, with rental payments paid to landowners to re-establish land cover to improve habitat for wildlife, reduce soil erosion, and improve water quality (<http://www.fsa.usda.gov/FSA/webapp?area=home&subject=copr&topic=crp>). While CRP functions differently from crop land, it is classified as crop in the USA's Department of Agriculture (USDA) National Resources Inventory (NRI, <http://www.nrcs.usda.gov/wps/portal/nrcs/main/national/technical/nra/nri/>) and, therefore, is inseparable within our analyses. We used data from the 1990s NRI because this is the last decade for which the data provided plot-level estimates of land-use change. We assumed no transition from urban landscape since there were no observed occurrences of this type of transition in the observed data (Nusser & Goebel 1997). In addition, we did not include a transition into or from wetland, as wetland areas are regulated by state and federal agencies. Finally, only private lands were deemed subject to change.

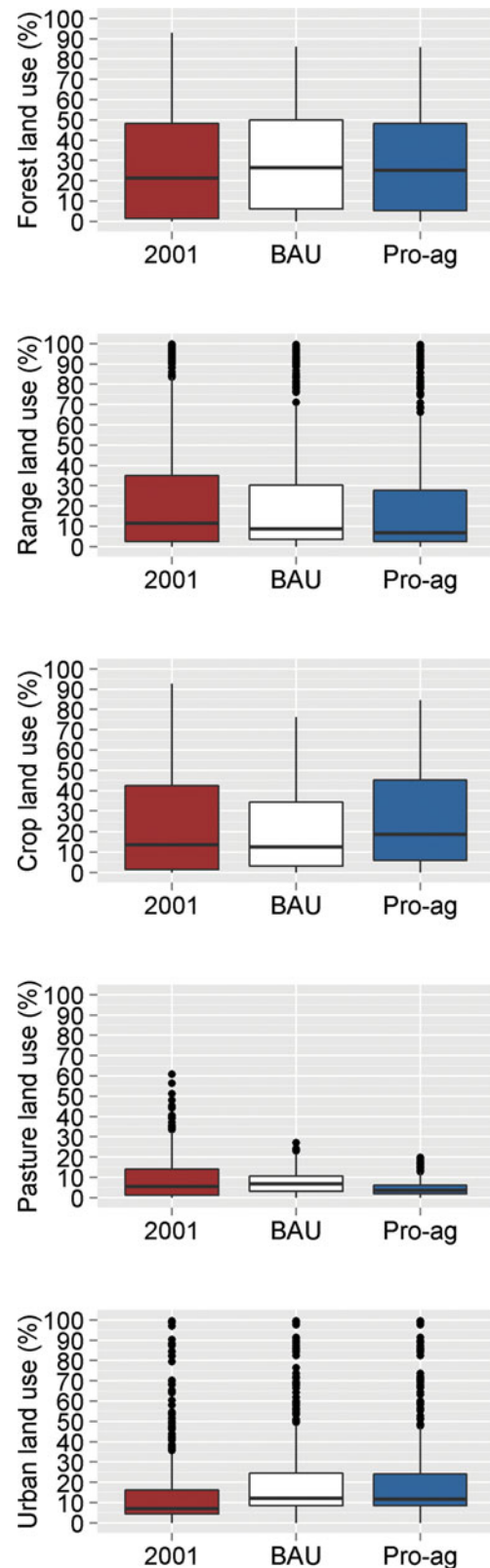
We calculated the number of refuges with > 20% of natural land use (such as forest and/or range) under the different scenarios. We chose this 20% threshold under the premise that corridors between reserves have to be sufficiently large so that the corridors can be found and successfully negotiated by dispersing organisms (Stauffer 1985; Boswell *et al.* 1998); this 20% threshold is likely a minimum amount of habitat needed for most species of concern to the NWRS (With & Crist 1995).

We identified changes in five land-use classes: forest and range (encompassing natural habitat), crop and pasture (denoting intensively used lands), and urban. Finally, we analysed land-use scenarios for the refuge system as a whole, as well as for each of seven FWS administrative regions. Since resource issues can vary regionally, we examined differences among administrative regions. We reported change as the absolute change in area of particular land uses, absolute change in the percentage of particular land uses, and rate of change in extent, relative to the 2001 extent of land use. We based our simulations on the 2001 NLCD data since these data capture land use at the end of the 1990s, namely the decade for which we had land-use transition data from the USDA NRI.

## RESULTS

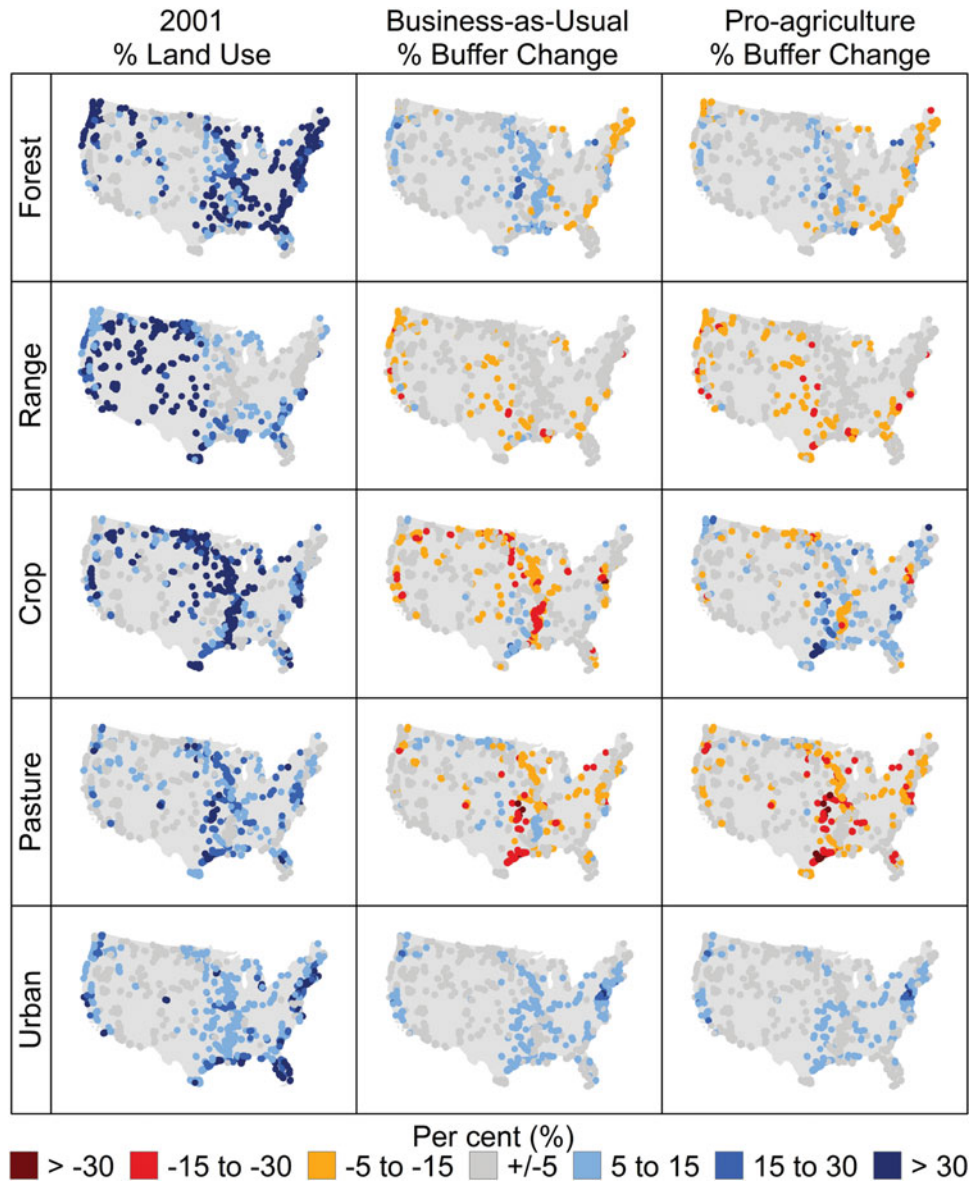
### Use in 2001

Range was the most prevalent land use surrounding refuges in 2001 (31.9 million ha, 31.4% of the area within 25 km of refuges), followed by forest (26.4 million ha, or 26.0% of the area), crops (including CRP; 24.3 million ha, 24.3%), pasture (9.6 million ha, 9.4%) and urban (9 million ha, 8.9%). The prevalent land use varied widely among refuges and exhibited clear regional patterns (Figs 1 and 2). For example, range



**Figure 1** Box plots showing the variability in each land use for starting conditions and under each scenario (BAU = business-as-usual, Pro-ag = pro-agricultural production).

**Figure 2** Initial per cent land use (2001) within 25 km of USA Fish and Wildlife Service National Wildlife Refuges and per cent of buffer area changing by 2051 under each scenario.



was most common near refuges in the Great Plains, while croplands were common near refuges in the Upper Midwest and the Lower Mississippi River.

**Projected landscape changes to the land surrounding NWRS by 2051**

Forest and urban land use increased around refuges, while range and pasture decreased in both future scenarios, but the rates of change for all but urban land use differed considerably by scenario (Table 1). The only land use surrounding refuges to change between scenarios was cropland, decreasing under the business-as-usual scenario and increasing, as expected, under the pro-agriculture scenario (Table 1, Fig. 1). In addition, the pro-agriculture scenario resulted in a nearly 50% smaller forest increase and a doubling of the rate of range and

pasture loss. Urban land-use change rates on lands around refuges slightly decreased under the pro-agriculture scenario (Table 1, Figs 1 and 2). Nearly half of the crop area gained under the pro-agriculture scenario came at the expense of pasture (Table 1, Fig. 1).

The number of refuges with >20% natural land use (such as forest and/or range) in the surrounding land was stable for the pro-agriculture scenario (increasing from 299 in 2001 to 300 refuges in 2051), but increased under the business-as-usual scenario (from 299 to 319 refuges). The pro-agriculture net change was the result of nearly equal numbers of refuges increasing (12) and decreasing (11) across the 20% threshold of surrounding range and forest land, whereas the increase under the business-as-usual scenario was primarily the result of an increase (21 increases; 1 decrease) (Table 2).

**Table 1** Area (million ha), percentage area, and percentage change under two scenarios within 25 km of the National Wildlife Refuges between 2001 and 2051.

Land use	2001 area (million ha)	% area in 2001	Business-as-usual scenario		Pro-agriculture	
			Area in 2051 (million ha)	% Change	Area in 2051 (million ha)	% Change
Forest	26.4	26.0	29.0	9.7	27.8	5.1
Range	31.9	31.4	30.4	− 4.5	28.7	− 9.9
Crop	24.6	24.3	21.1	− 14.2	27.2	10.2
Pasture	9.6	9.4	7.3	− 24.2	4.3	− 55.1
Urban	9.0	8.9	13.7	52.2	13.6	50.7

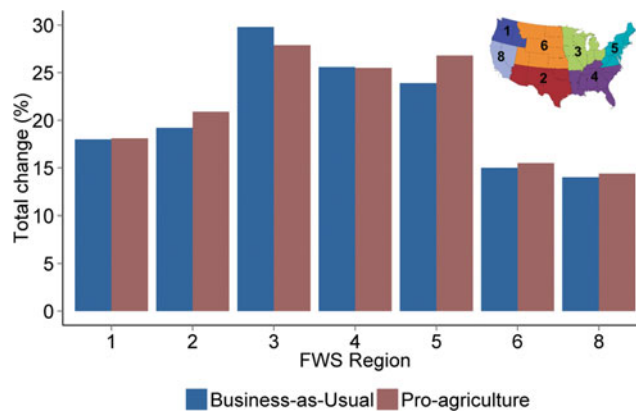
**Table 2** Number of refuges among the total 461 refuges with >20% cover of forest and/or range (natural) land use in 2001, number of refuges that transition above and below the 20% threshold over 50 years, and the total number of refuges above the threshold in 2051.

Scenario	2001	2051		
	Refuges >20% natural cover (n)	Refuges with increased cover (n)	Refuges with decreased cover (n)	Total refuges with > 20% natural cover
Business-as-usual	299	21	1	319
Pro-agriculture	299	12	11	300

### Regional change patterns

High proportions of a specific land use surrounding refuges tended to be clustered regionally, with forest concentrated in the north-western USA and along the Atlantic coast from Georgia to Maine, range in the Great Plains and Rocky Mountains, crops and pasture along the Mississippi River and in the northern Great Plains, and urban along the coasts, with clusters of highest use in the mid-Atlantic region, Florida, the Gulf Coast, and scattered along the west coast of the USA (Fig. 2). Forest land-use change tended to be concentrated regionally, with gains along the Mississippi River and losses in the north-eastern and south-eastern regions. Forest gains tended to align with areas where crops decreased (namely along the Mississippi River) and forest losses where crops increased (in north-eastern and south-eastern USA) (Fig. 2). Pasture exhibited primarily losses, especially along the Mississippi River and in the mid-Atlantic region (Table 1, Fig. 2). Urban use increased east of the Mississippi River and near the west coast, with particularly sharp increases in the mid-Atlantic region. The pro-agriculture scenario resulted in fewer refuges in the upper midwestern USA, with an increase in urban land use within 25 km distance from the refuge (Fig. 2).

The percentage of the total area changing land use ranged from 12–30% in the different regions, with the highest rates of change occurring in the eastern USA (Fig. 3). The midwestern USA (Region 3) was the most dynamic region under both scenarios, with the pro-agriculture scenario resulting in less change. The North-east (Region 5) exhibited the largest difference between scenarios with pro-agriculture resulting in more land-use change. The total area of land in forest, range and urban use showed little change regardless



**Figure 3** Regional predictions for the percentage of the area surrounding refuges that will change use under the different scenarios. The USA's Fish and Wildlife Service (FWS) administrative regions (inset) are: (1) Pacific, (2) South-west, (3) Midwest, (4) South-east, (5) North-east, (6) Mountain-prairie, and (8) Pacific-Southwest. Alaska (FWS region 7) was not considered part of the conterminous USA study area.

of scenario. However, crop increases and pasture decreases were substantial and consistent under the pro-agriculture scenario (Appendix 1, Fig. S1, see supplementary material at Journals.cambridge.org/ENC).

The rate of increase of any land use was by far the highest for urban, which grew at a rate of 40–80% in both scenarios (Table 1; Appendix 1, Fig. S1, see supplementary material at Journals.cambridge.org/ENC). Rate of change for the other land uses was more variable. Changes in forest between scenarios were only notable in the (1) Pacific, (2) South-west, (3) Midwest, and (4) South-east FWS regions, with

a lower rate of increase under the pro-agriculture scenario. Even then, growth rates remained positive for regions 1, 2, and 3. The Midwest and North-east regions exhibited increases in range under the business-as-usual scenario, while the other regions experienced losses. Under the pro-agriculture scenario, however, range losses occurred in all regions and the rate of loss was higher. Crop losses occurred in all but the South-west under the business-as-usual scenario but changed to gains in all regions under the pro-agriculture scenario, with the highest rates of increase occurring in the Midwest and North-east regions. Pasture use increased slightly in the Pacific and the Mountain-prairie regions, but decreased dramatically in all other regions under the business-as-usual scenario. Under the pro-agriculture scenario, however, all regions experienced 25–70% decreases in pasture land-use (Appendix 1, Fig. S1, see supplementary material at [Journals.cambridge.org/ENC](http://Journals.cambridge.org/ENC)).

## DISCUSSION

Our analyses predicted substantial land-use change by 2051 in areas surrounding National Wildlife Refuges, particularly for refuges in the eastern USA. Forest and urban use are expected to increase. Regardless of scenario, the largest total area and highest rates of change are projected to be in urban land use, which agrees with previous studies (Pijanowski & Robinson 2011; Radeloff *et al.* 2012). Forest increases are expected to occur at the expense of pasture. The pro-agriculture scenario suggests substantial changes in land-use trajectories from the business-as-usual scenario. In fact, cropland use changed from a likely net decrease to a net increase under the pro-agriculture scenario, range and pasture decreases doubled, and forest increases were only half as large. Interestingly, the number of refuges with >20% natural habitat in the surrounding landscape stayed relatively stable in the pro-agriculture scenario, whereas numbers increased under the business-as-usual scenario.

The high rate of urbanization varied minimally between scenarios. Economic returns from urban uses tend to be much larger than from other uses, explaining why increasing crop prices under the pro-agriculture scenario produced little change in the rate of urbanization. While they did not differ between scenarios, urbanization rates ranged widely among regions, similar to other forecasts of future urbanization (Wear 2011; Radeloff *et al.* 2012). The high rate of urban growth around the National Wildlife Refuges agrees with other studies reporting high rates of housing growth around protected areas (Radeloff *et al.* 2010; Wade & Theobald, 2010). This is troublesome, because once land is converted to urban use, it is unlikely to revert to other uses. Tax- or subsidy-based policies are unlikely to change land-use trajectories for urban use, but strong urban containment policies can, however, significantly reduce urban growth (Woo & Guldmann 2011; Hamilton *et al.* 2013).

Refuges, as with other protected areas, are threatened by encroachment of intensive land uses (Griffith *et al.* 2009).

Much of the land currently surrounding the National Wildlife Refuges remains in forest and range, and therefore is potential habitat for wildlife. This is important because the amount of suitable habitat in the matrix surrounding refuges influences population persistence. As the amount of habitat in the vicinity of refuges declines below *c.* 30% of the landscape, fragmentation results in additional threats to population persistence (Fahrig 2001). Thus, even small changes to other land uses may negatively impact the quality of habitat (Gagne & Fahrig 2010a, b; Kuhman *et al.* 2010), but we note that there may be also cases where land use change benefits refuges. Regardless, areas of relatively natural habitat outside protected areas are critical for maintenance of biodiversity (Franklin & Lindenmayer 2009).

Increases in crop under the pro-agriculture scenario were much greater than we had anticipated given that many studies project future decreases in crop area (Pijanowski & Robinson 2011; Wear 2011). It is reasonable to expect intensification of agriculture as the human population continues to grow over the next four decades (Foley *et al.* 2011), and, as agricultural lands are increasingly used as sites for wind power and biofuels in addition to food (Dale *et al.* 2011), it is likely more land will be needed for agricultural production. In fact, other studies have shown increases in agricultural land use for the Upper Midwest and Great Plains regions (Sohl *et al.* 2012).

## CONCLUSIONS

Profitability of agricultural land use could drive substantial changes in land-use patterns near wildlife refuges, with negative impacts on the capacity of the NWRS to maintain wildlife populations. The majority of the increase in crop would likely come at the expense of pasture, but also from forest and range losses. The loss of forest, range and pasture around refuges, which in a semi-natural state provide habitat value and corridors for wildlife, could reduce resilience of the NWRS and complicate climate change adaptation. Habitat loss and fragmentation from increasing crop cover could be exacerbated by continuing growth in urban land use, ultimately resulting in a much more fragmented and complex landscape.

The relative stability of forest and range land area around many refuges may provide opportunities for improving connectivity among protected areas. Given the shifts in suitable climate space that many species are facing, land-use change which reduces landscape connectivity compounds the threat to biodiversity from climate change (Griffith *et al.* 2009; Heller & Zavaleta 2009). Recognizing this compounded threat, one of the tools that the FWS and other land management agencies promote as an adaptive measure to climate change is the establishment of habitat corridors connecting protected areas, removing barriers to dispersal (Griffith *et al.* 2009). Our models may be useful for identifying regions with higher threat of land-use change and, consequently, guide the implementation of conservation actions that promote increased connectivity for the NWRS, thereby increasing the

resilience of the protected area system to climate change. While our data are presented at the national level, we have provided the output for all 461 refuges to FWS managers in spreadsheet form for consideration in their conservation planning efforts. This is a far more relevant scale for taking specific management actions.

Our results indicated regional approaches to policy development and implementation may be at least as important as scenario conditions. Land-use change occurs in a regional context; flexible policy and budgets would allow a network of lands as large as the National Wildlife Refuge System to accommodate regional differences in threat and opportunity. Given that change in the lands surrounding refuges may impact the viability of these refuges, a focus on private land conservation programmes may be useful because these programmes can work proactively with local stakeholders often needing technical assistance in areas such as habitat restoration and conservation planning. Given the regional nature of land-use change, it also may be useful to focus conservation efforts in areas where land-use changes are most threatening to the integrity of the NWRs. Finally, we wish to emphasize that our ‘what if’ scenario, focusing on increasing agricultural land use surrounding wildlife refuges is plausible given recent changes in land use and projections from other research and therefore provides useful information for managers considering adaptation to climate and land use change.

### Supplementary material

To view supplementary material for this article, please visit [Journals.cambridge.org/ENC](http://Journals.cambridge.org/ENC)

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