

## ***Reducing Livestock Effects on Public Lands in the Western United States as the Climate Changes: A Reply to Svejcar et al.***

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1                   **Reducing Livestock Effects on Public Lands in the Western United States**  
2                                   **as the Climate Changes: A Reply to Svejcar et al.<sup>1</sup>**

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38 <sup>1</sup> Svejcar et al. 2014 (see Literature cited)

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41 Beschta et al. (2013) synthesized the ecological effects of climate change and ungulate  
42 grazing on western public lands, grounding their recommendations in ecological considerations  
43 and federal agency legal authority and obligations. Svejcar et al. (2014) suggest that Beschta et  
44 al. (2013) neither “present a balanced synthesis of the scientific literature” nor “reflect the  
45 complexities associated with herbivore grazing.” Svejcar et al. (2014) “dispute the notion that  
46 eliminating [livestock] grazing will provide a solution to problems created by climate change,”  
47 although we made no such claim. Instead, Beschta et al. (2013: p. 474) indicate that removal or  
48 reduction of livestock across large areas of public land will reduce a pervasive ecological stress,  
49 diminishing cumulative impacts on these ecosystems under climate change. We respond to three  
50 livestock grazing issues raised by Svejcar et al. (2014): (1) legacy vs. contemporary effects, (2)  
51 fuels reduction and fire effects, and (3) grazing complexity and restoration.

52 **1) Contemporary and legacy livestock use has caused combined effects.**

53 Livestock effects began soon after their introduction to semi-arid ecosystems west of the  
54 Rockies, which had evolved in an absence of large herds of ungulates (Mack and Thompson  
55 1982). Contemporary grazing impacts (as described in Beschta et al. 2013) compound “legacy”  
56 effects, including: altered fire regimes; biological soil crust loss, soil loss, and compaction;  
57 altered composition, structure, and function of upland, riparian, and stream biological  
58 communities; altered streamflow regimes; and reduced food-web support and physical habitat for  
59 terrestrial and aquatic biota (Blackburn 1984; Belsky et al. 1999; Kauffman and Pyke 2001;  
60 Belnap and Lange 2003; Fleischner 2010). Combined legacy and current grazing effects have  
61 left many streams with degraded riparian vegetation, accelerated bank erosion, widened and/or  
62 incised stream channels, and altered water quality (increased temperatures and sediment loads).  
63 These changes have many negative biological effects, including those on imperiled resident and

64 anadromous fish (NRC 1996; NRC 2002). Because the legacy effects of livestock were  
65 significant and extensive, contemporary grazing studies tend to *underestimate* ecological  
66 impacts, as they compare changes within already diminished systems (Fleischner 1994).

67 While some livestock impacts (e.g., soil loss or channel incision) may not be fully reversible  
68 in short timeframes, recovery of native plant communities and soil functions, which underpin  
69 terrestrial ecosystems, often occur when the causes of degradation are removed or reduced.  
70 Despite changes in public land grazing practices over time, evidence indicates that contemporary  
71 livestock use thwarts ecological recovery. Cessation of livestock grazing can result in recovery  
72 of soil properties (Kauffman et al., 2004), riparian vegetation (Hough-Snee et al. 2013 and  
73 **Figure 1**), and channel morphology (Opperman and Merenlender 2004 and **Figure 1**), relative to  
74 areas that continue to be grazed.

75 Riparian and stream ecosystems (Belsky et al. 1999; NRC 2002) and aspen (*Populous*  
76 *tremuloides*) communities (Seager et al. 2013) are biologically diverse and especially susceptible  
77 to the effects of livestock use. For example, recent studies in Wyoming (Hessl and Graumlich  
78 2002), Nevada (Kay 2003), Montana (Kimble 2007), Oregon (Seager 2010), and Utah (Kay  
79 2011) point to high levels of livestock herbivory over many decades, sometimes in combination  
80 with wild ungulate impacts, as a major factor inhibiting aspen growth from seedling/sprouts into  
81 saplings and trees. These long-term effects hamper the ability of this tree species to persist in  
82 many western ecosystems. Livestock grazing also has widespread effects on the frequency and  
83 distribution of native grasses, forbs, and shrubs, and native wildlife species dependent upon those  
84 plants [e.g., sage-grouse (*Centrocercus urophasianus*); Manier et al. 2013].

85 **2) Livestock grazing is not a viable tool for reducing fuels and wildfire effects.**

86 Livestock grazing in western US landscapes altered natural fire regimes by decreasing the  
87 frequency of low-severity fires beginning in the early 1900s (Swetnam and Betancourt 1998),  
88 making large areas prone to invasion by woody species and, in turn, more susceptible to high-  
89 severity fires (Chambers and Pellang 2008). Furthermore, cheatgrass (*Bromus tectorum*), an  
90 annual exotic, spread rapidly throughout the Intermountain West as a result of livestock  
91 movement and overgrazing (Mack, 1986), contributing to more frequent burning. Cheatgrass  
92 dominates nearly 70,000 km<sup>2</sup> in the Great Basin and is a component on an additional 250,000  
93 km<sup>2</sup> (Diamond et al. 2012). Reisner et al. (2013) found that: livestock grazing increases  
94 cheatgrass dominance in sagebrush steppe, reduced grazing may be one of the most effective  
95 means of conserving and restoring imperiled sagebrush ecosystems, and livestock grazing is not  
96 likely a viable tool for reducing cheatgrass dominance because it promotes cheatgrass invasion.

97 **3) Although livestock grazing has complex ecological consequences, large-scale reductions**  
98 **in grazing effects are likely to reduce cumulative ecosystem degradation.**

99 Recognizing the complexity of grazing issues was central to the synthesis and  
100 recommendations included in Beschta et al. (2013). Our analyses provided an integrative view  
101 of that complexity: we discussed three classes of ungulates (domestic, feral, wild), drawing  
102 examples from diverse vegetation types (shrub steppe, desert, conifer forest) and ecological  
103 attributes (such as water quality, hydrology, riparian areas, soils, hydrology, biodiversity).  
104 Nevertheless, compelling reasons exist to single out livestock as a cause of ecological harm to  
105 native plant communities, terrestrial and aquatic habitats, and watershed processes (Belsky et al.  
106 1999; Kauffman and Pyke 2001; Belnap and Lange 2003; NRC 2002). Livestock use is a  
107 principal cause of desertification in arid and semi-arid landscapes (Swetnam and Betancourt  
108 1998; Belnap and Lange 2003; Fleischner 2010). It has the most extensive land-use footprint on

109 western public lands (Beschta et al. 2013), and it continues at major public expense (Vincent  
110 2012). Livestock production also contributes directly and indirectly to greenhouse gases, raising  
111 increasing concern about its climate effects (Ripple et al 2014). The cessation or removal of  
112 factors that cause degradation or prevent recovery is the most effective and robust approach to  
113 ecological restoration (Kauffman et al. 1997). Unlike many stressors, livestock use is subject to  
114 human control.

115 Svejcar et al. (2014) assert that position statements by the American Fisheries Society  
116 (Armour et al. 1991) and the Wildlife Society (2010) “do not advocate removing livestock from  
117 western rangelands.” These position statements, however, as well as those of the Society for  
118 Conservation Biology (Fleischner et al. 1994), conclude that public-land grazing impacts need to  
119 be dramatically reduced to allow recovery of degraded ecosystems—an explicit recommendation  
120 of Beschta et al. (2013). Moreover, these position statements were developed without  
121 consideration of climate change effects.

122 Livestock use of public lands in the West remains a major stressor with effects of increasing  
123 concern under the overarching stressor of climate change. Its removal or reduction is an ecologically  
124 efficient and unambiguous approach for restoring resilience to large areas of these lands (see  
125 synthesis in Beschta et al. 2013). Because livestock grazing has diminished biodiversity and  
126 degraded ecosystems, the burden of proof for maintaining the grazing status quo is on Svejcar et  
127 al (2014). But they offer no evidence that livestock use is compatible with the recovery of  
128 livestock-degraded uplands, riparian areas, or stream systems, or with retention of native species  
129 in arid and semi-arid ecosystems. Absent such evidence, and in the context of a changing  
130 climate, the only rational, effective, and direct alternative for ecologically restoring many  
131 western public lands is to reduce the effects of their most prominent stressor—livestock.

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219 **Figure 1.** A photopoint demonstrating vegetation and stream channel change following removal  
220 of livestock in the Northern Great Basin (Barnhardy Meadows, Hart Mountain National  
221 Antelope Refuge, Oregon). Upper photo was taken October, 1990 after approximately one  
222 century of livestock grazing during which livestock use was managed by the US Fish and  
223 Wildlife Service from 1940-1990. Lower photo was taken August, 2013 following 22 years of  
224 rest from livestock grazing. In this ecosystem, the reestablishment of willows (*Salix* spp.) and  
225 other wetland obligate species, as well as increased aspen recruitment, has occurred. Previously  
226 eroding stream banks have stabilized and stream channels narrowed since the removal of  
227 livestock on the refuge. Photo credits: (upper) Bill Pyle and (lower) Schyler Ries.