

Escaped farmed salmon and trout in Chile: incidence, impacts, and the need for an ecosystem view

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1 **ESCAPED FARMED SALMON AND TROUT IN CHILE:**
2 **INCIDENCE, IMPACTS, AND THE NEED FOR AN ECOSYSTEM VIEW**

3
4 **Shortened title: Escaped farmed salmonids in chile**

5
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23 **ABSTRACT**

24 The exponential growth of the salmonid farming industry during the last three decades
25 has created conditions for massive escapes of these exotic species into natural environments
26 in southern Chile. Here, we review and update information about salmonid escapes from
27 1993 to 2012 and examine their potential environmental, social, and economic
28 consequences. We estimate that more than one million salmonids escape each year from
29 marine farms, mainly due to weather conditions and technical and operational failures of
30 net-pens. While a decrease in the magnitude of escaped Atlantic and coho salmon have
31 occurred during the last several years, escaped rainbow trout have not followed the same
32 pattern. Rainbow trout have become a greater threat to native ecosystems due to their
33 greater potential to establish self-sustaining naturalized populations. The main ecological
34 effects of escapees are related to short-term predatory effects upon native fish, long-term
35 effects linked to the likelihood of farmed salmon establishing self-sustainable populations,
36 and disease and pathogen transfer to native fauna. More research is needed to identify and
37 develop reliable indicators to estimate the impact of escapees at the ecosystem level in both
38 marine and freshwater systems. An understanding of the mechanisms of coexistence
39 between native fishes and introduced non-native salmonids may be useful to design
40 effective management strategies aimed at protecting native fish from salmonid
41 introductions. A precautionary approach that encourages local artisanal and recreational
42 fisheries to counteract colonization and naturalization of salmon species in southern Chile
43 may constitute another management option.

44 **Key words:** fish farming, *Salmo salar*, *Oncorhynchus kisutch*, *Oncorhynchus mykiss*,
45 exotic species, Chile

46 INTRODUCTION

47

48 Salmon and trout (hereafter salmonids) farming has experienced an exponential growth
49 during recent decades, with Chile and Norway accounting for over 80% of the global
50 salmonid aquaculture production (Figure 1; FAO 2011). In 2006 Chilean salmonid
51 aquaculture reached its highest production, with nearly 640 thousands tons valued at US
52 \$3.8 billion (FAO 2011). This production corresponded mostly to Atlantic salmon *Salmo*
53 *salar* (60.3%), coho salmon *Oncorhynchus kisutch* (17.5%), and rainbow trout *O. mykiss*
54 (22.2%). Even though salmonids were initially introduced in the Southern Hemisphere for
55 recreational purposes in the early 1900's (e.g. rainbow trout and brown trout *S. trutta*);
56 additional introductions (mostly Pacific salmon species) occurred during the 1970's when
57 they were farmed in hatcheries for ranching and aquaculture-fishery purposes (Basulto
58 2003). The expansion of the aquaculture industry in Chile started in the 1980's with
59 salmonids grown to commercial size in net-pens in the inner seas and fjords of the Chiloe
60 Archipelago in the Lakes Region (41-43° S; Figure 2). Currently the aquaculture industry,
61 both in size and number of farming facilities, is rapidly expanding further south (Aysen
62 Region; 44-46° S; Figure 2) (Buschmann et al. 2009, Niklitschek et al. 2013).

63 Salmonid farming phases mirror the cycles used by salmonids during their natural lives.
64 Salmonids inhabit both fresh and marine waters, with freshwater systems playing a key role
65 during early developmental stages. In Chile, the main growth of stocks takes place in the
66 sea (Soto et al. 2001, Rojas & Wadsworth 2007) where they are reared in either square or
67 circular floating net-pens until they attain commercial size (at 1-3 years of age). The current
68 density of salmonids in each net-pen is 16-20 kg m⁻³, although higher densities (~ 30 kg m⁻³
69 ³) were recorded in some facilities before 2008 (X. Rojas, *personal communication*). It has

70 been shown that as the magnitude and number of sites where salmonid farming occurs
71 increases, the potential consequences due to net-pen or farm failure increases, resulting in a
72 higher probability of exotic escapees in the environment (Arismendi et al. 2009, Jensen et
73 al. 2010, Niklitschek et al. 2013).

74 Here, we review and update information about salmonid escapes in Chile during the
75 period of 1993 to 2012, and examine their potential environmental and economic
76 consequences. We also provide a summary of the main factors that influence escapees, as
77 well as discuss mitigation and prevention alternatives. We propose actions to diminish
78 escape risks and highlight some management practices to mitigate negative impacts and
79 enhance those which appear to be positive. This information is fundamental to understand
80 the trade-off between the negative effects of biological invasions upon natural ecosystems
81 and the high economic value of salmonids for aquaculture and recreational purposes in
82 Chile and elsewhere.

83

84 **CAUSES OF SALMONID ESCAPES**

85

86 Several factors can explain the escapes of salmonids from facilities in coastal, marine
87 and freshwater environments, including those of external and internal origins. Among
88 these, external factors include attacks by predators (e.g. Sepúlveda & Oliva 2005, Vilata et
89 al. 2010), theft or vandalism (intentional damage of nets to let salmon escape and then steal
90 the fish), and adverse weather conditions, whereas internal factors are directly related to
91 and under the responsibility of the fish farmer and include failure or neglect during routine
92 fish handling procedures and site maintenance and accidental boat collisions (Sepúlveda et
93 al. 2009). Reports by salmonid farm companies during the period 2004-2009 indicate that

94 escape events were primarily caused by severe weather conditions (30%), theft (21%), and
95 structural failure of net-pens and deficient handling incidents (18%; Figure 3). Storms lead
96 to stronger waves and currents, resulting in ripping of the net-pen tethering ropes, breaking
97 of the net-pen mesh, or tipping over of the net-pens (Jensen et al. 2010). Unfortunately, due
98 to the high demand for new farming sites often the selection of a new location does not
99 include proper consideration of potentially adverse environmental conditions (i.e. water
100 currents, winds), increasing the risk of fish escapes during extreme adverse weather
101 conditions. Overall, the causes of salmonid escapes are similar to those reported by
102 producers in other salmonid-producing countries such as Norway, Canada, and Scotland
103 (Valland 2005), but there the responsibility for fish escapes lies mainly with the farmer
104 and/or providers of the equipment and services, including routine site maintenance and fish
105 handling (Melo et al. 2005). Routine net-pen maintenance and fish handling procedures
106 carried out at the farms also pose escape risks, from holes in the nets or transportation of
107 fish among cages. Moreover, carelessness when changing fish or predator nets may result
108 in the escape of fish. The sorting of fish into two or more net-pens through a tube can also
109 lead to the involuntary release of species if the tube is poorly mounted. Moreover,
110 collisions from boats used during operational activities, predator attacks (e.g. sea lions and
111 birds), inadequate manufacturing materials and poor net maintenance increase the
112 probability of escapes (Robles 2002). Lastly, although it is difficult to quantify, intentional
113 damage of nets attributable to fishermen or farmers seeking to benefit from the subsequent
114 captures of escaped salmonids or insurance policies are also factors that may increase
115 escape risks.

116

117 **QUANTIFYING SALMON AND TROUT ESCAPES**

118

119 The real magnitude of salmonid escapes is most likely underestimated, mainly due to the
120 fact that not all escapes are detected or reported. Number of escaped salmonids in Chile
121 have been mostly reported or estimated after large and/or catastrophic events (Soto et al.
122 2001, Thorstad et al. 2008, Niklitschek et al. 2013). In addition to escapes from harsh
123 weather conditions, farmed salmon may escape from marine net pens through persistent
124 low-level leakage (Buschmann et al. 2009, Schröder & García de Leaniz 2011).
125 Unfortunately, information on the number of salmonids that escape from regular leakages
126 in Chile remains poorly documented. Soto et al. (2001) estimated that 1-5% of escapees
127 come from leakages. However, this estimate has not been evaluated directly and remains
128 somewhat speculative, being the threat of leakages insufficiently recognized (Sepúlveda et
129 al. 2009).

130 By consulting insurance companies, Soto et al. (2001) reported an important number of
131 escapes after major storms during 1994 and 1995 (Figure 4, Table S1). Since 2004, the
132 salmon industry in Chile must inform government institutions about every escape event at
133 their facility, but there are no official records of escapes available for the period 1997 to
134 2003. A total of 58 escape events were reported during the period 2004-2012 (Data from
135 Soto et al. 1997 and from National Fisheries Service, Figure 4), accounting for almost 6.5
136 million salmonid escapees, although it is estimated that more than one million salmonids
137 may escape each year in Chilean marine systems (Thorstad et al. 2008).

138 During the 13 yr of escape reports (1993 to 1996 and 2004 to 2012) a total of 3.7 million
139 Atlantic salmon ($289\,600\text{ yr}^{-1}$), 3.1 million coho salmon ($239\,954\text{ yr}^{-1}$) and 4.0 million
140 rainbow trout ($313\,892\text{ yr}^{-1}$) were reported to have escaped from salmon farms located in
141 both the Lakes and Aysen Regions. These amounts of escapes salmonids appear to be

142 similar to those reported in other countries such as Norway and Scotland (440 000 yr⁻¹ and
143 216 000 yr⁻¹, respectively; Thorstad et al. 2008, Jensen et al. 2010) and within a suggested
144 range of escapes in Chile (1-2% of the total production, Niklitschek et al. 2006). However,
145 if we consider the proportion of escaped fish from the total salmonids production, escaped
146 fish in Chile duplicated those levels from Norway, and are similar to Scotland.

147 Overall, the number of salmon and trout reported to escape relative to the total
148 production varies greatly among the three species analysed; it was lowest for Atlantic
149 salmon and highest for coho salmon and rainbow trout: the average escape proportion 1993
150 to 1996 and 2004 to 2012 = (total no. escaped fish reported) / (production by species) were:
151 1.2 t⁻¹ for Atlantic salmon, 2.4 t⁻¹ for rainbow trout, and 2.5 t⁻¹ for coho salmon. Comparing
152 the two time periods (1993-1996 vs. 2004-2012) the escapes of Atlantic salmon and coho
153 salmon decreased in the second period (Atlantic salmon: 374 349 to 251 933 yr⁻¹; coho
154 salmon: 512 413 to 118 860 yr⁻¹) despite a marked increase in production during the second
155 period, especially for Atlantic salmon. On the contrary, the escapes of rainbow trout
156 increased during the second time period (211 669 to 359 324 yr⁻¹). This tendency in
157 rainbow trout escapees is likely due to large-scale escape events in the Aysen Region
158 during 2004, 2007 and 2008 years. The largest episode in 2004 corresponds to a specific
159 event where about 1.8 million rainbow trout and coho salmon escaped due to bad weather
160 conditions. The largest event in year 2007 was mainly associated to the tsunami that took
161 place in April, where more than 1.5 million specimens were reported as escaped from net
162 pens. However, Thorstad et al. (2008) refers to actual figures of about 5 million specimens,
163 which could account for one of the largest escape episodes documented both at national and
164 international levels.

165 In freshwater systems the information regarding salmon escapes is even more scarce,
166 with a total of 11 events reported from 2004-2012 accounting for a total of 613 586
167 salmonid escapees, principally from lakes (75% rainbow trout and 25% Atlantic salmon). A
168 positive and strong relationship between the magnitude of salmon production in freshwater
169 facilities and the relative abundance of free-living salmonids including coho, Atlantic,
170 rainbow trout, and Chinook salmon has been described in lakes of southern Chile
171 (Arismendi et al. 2009, Young et al. 2009, 2010, García de Leaniz et al. 2010, Vanhaecke
172 et al. 2012a). In fact, exotic salmonids have been reported as the most abundant fishes in
173 freshwater systems of Chile (Soto et al. 2006). Since there is no documented evidence of
174 successful natural reproduction of Atlantic and coho salmon in Chile, individuals from
175 these species inhabiting freshwater systems appear to be exclusively originated from
176 aquaculture escapes (Soto et al. 2001, Soto et al. 2006, Arismendi et al. 2009, Schröder &
177 García de Leaniz 2011). Also, because no massive escape events have been reported in
178 freshwater systems, the recurrent presence of salmonids in freshwater systems could be
179 explained by frequent operational leakages from salmon farms.

180

181 **ECOLOGICAL CONSEQUENCES OF ESCAPES**

182

183 There is a general concensus among scientists that introduced species directly or
184 indirectly alter the structure and diversity of natural ecosystems (Grosholz 2002, Naylor et
185 al. 2005). Among freshwater introductions, salmonids are considered to be one of the most
186 pervasive exotic species in the world (Pascual et al. 2009). In Chile, the environmental
187 concerns from salmonid escapes have focused on short-term predatory effects upon native
188 fish, long-term effects linked to the probability of farmed salmon in establishing self-

189 sustainable populations, and disease transfer and pathogens (Young et al. 2010, Arismendi
190 et al. 2012, Niklitschek et al. 2013).

191

192 **Displacement of native fishes due to ecological interactions**

193

194 Freshwater systems, including rivers and lakes, and marine systems have been invaded
195 by salmonids and it is possible that their ecosystem-level processes may be affected
196 through trophic cascade effects (Carpenter et al. 1996). Unfortunately, there is scarce
197 information about the state of native fishes before salmonid introductions, which makes an
198 understanding of their impacts more difficult to obtain (García de Leaniz et al. 2010).
199 Based on stomach and stable isotope analyses, several studies conducted in Chile have
200 shown negative effects from salmonid species on native fishes due to predatory and
201 interference competition (Soto et al. 2001, 2006, Penaluna et al. 2009, Arismendi et al.
202 2009, 2012, Young et al. 2009, 2010, García de Leaniz et al. 2010). Collectively, the
203 evidence suggests that salmonid species have detrimental impacts on native fishes in all
204 types of ecosystems including lakes (Soto et al. 2006, Arismendi et al. 2009, García de
205 Leaniz et al. 2010, Habit et al. 2010, Correa & Hendry 2012), rivers (Soto et al. 2006,
206 Penaluna et al. 2009, Arismendi et al. 2012, Vanhaecke et al. 2012a,b), and inner seas (Soto
207 et al. 2001). Lakes in particular, where most of the freshwater phase of salmonid
208 aquaculture occurs, could be particularly sensitive to the impacts of escapes because top
209 predator species may produce a detrimental impact to aquatic biodiversity and species
210 richness (Moyle & Light 1996, García de Leaniz et al. 2010, Vanhaecke et al. 2012a,b).
211 However, a more complete evaluation of the effects of predation and competition on native
212 fauna is prevented by the fact that the basic biology and ecology of native aquatic

213 communities in freshwater, inner seas, and fjords of southern Chile remains poorly
214 understood.

215

216 **Spreading of pathogens and diseases**

217

218 Animal health, especially in response to disease, is another issue to consider when
219 discussing the ecological impacts of salmonid escapes, because exotic salmonids can
220 introduce new pathogens, alter disease patterns, and even act synergistically to increase the
221 impact of other stressors (García de Leaniz et al. 2010, Habit et al. 2010). During the past
222 few years, several aquaculture facilities have been affected by epidemic outbreaks of
223 diseases, favored by the conditions of fish being confined at high densities and the short
224 distance that separates farms (Asche et al. 2010). In addition, escaped salmonids can travel
225 large distances (Melo et al. 2005, Whoriskey et al. 2006, Skilbrei et al. 2009), and hence
226 they become potential vectors for parasites and diseases at a broad scale (Thorstad et al.
227 2008). Epidemiological studies conducted in the Northern Hemisphere (i.e. Ireland,
228 Scotland, Norway, and Canada) suggest that the occurrence of diseases such as Rickettsial
229 septicemia and sea lice (*Caligus* spp.) in both salmonids and native fishes are directly
230 related to higher concentrations of farmed fishes (Krkosek et al. 2005, Naylor et al. 2005).
231 Also, a virus that regularly affects salmon farms in different countries including Chile is the
232 infectious pancreatic necrosis (IPN) virus, which has been detected in all salmon species at
233 all developmental stages (freshwater and ocean phase of aquaculture) as well as in native
234 fishes, molluscs and crustaceans (Rodríguez et al. 2003, Asche et al. 2010). The infectious
235 salmon anemia (ISA) virus has also been documented in salmon farms in Norway, Canada,
236 Scotland, United States, and recently in Chile, causing enormous damage to the industry

237 and the local and national economy (Niklischek et al. 2013). In Chile, the potential
238 transmission of diseases from farmed salmonids to other taxa such as marine birds and
239 mammals is yet unknown. However, preliminary evidence of skin lesions in dolphins has
240 suggested a potential link to the salmonid aquaculture industry (S. Heinrich, *personal*
241 *communication*).

242

243 **Threats of establishment from escaped salmonid populations**

244

245 Rainbow trout and Chinook salmon escapees may pose the greatest threat to native
246 ecosystems because they have naturalized populations compared to both Atlantic and coho
247 salmon, thus the magnitude of their ecological impacts may increase when they can
248 establish self-sustaining populations (Soto et al. 2006, 2007, Correa & Gross 2008,
249 Arismendi et al. 2009, Arismendi et al. 2011a,b). The successful establishment of self-
250 sustainable populations could be related to a relatively high plasticity of these species (i.e.
251 ability to feed on a broad range of organisms; Becker et al. 2007). Coho, Chinook salmon
252 and rainbow trout have also been part of ranching programs in the past which eventually
253 may also played a role in their establishment, especially in the case of Chinook salmon
254 (Astorga et al. 2008). It is also possible that salmonid escapees might increase the
255 probability of establishing self-sustaining populations when those escapes are greater than
256 from slow leakages or “silent” escapes (Consuegra et al. 2011). According to Consuegra et
257 al. (2011), the invasion success may also depend on propagule pressure. For example,
258 rainbow trout may have achieved high establishment success and expanded more rapidly
259 than other anadromous species (such as brown trout) because its spread is aided by escapes
260 from fish farms (Ciancio et al. 2008).

261 For Atlantic salmon there is no evidence for the establishment of naturalized populations
262 (Soto et al. 2001, 2006, Schröder & García de Leaniz 2011). Indirect evidence suggests that
263 this species fails to establish because escaped individuals do not feed or grow very well in
264 the wild (Soto et al. 2001). This is similar to other systems where efforts to establish
265 Atlantic salmon as a game fish species have failed (Naylor et al. 2005). Considering that
266 Atlantic salmon have traditionally been the highest proportion of farmed salmonid in Chile,
267 the risk of establishment is an ongoing, unresolved question. Similarly to Atlantic salmon
268 there no evidence suggesting that coho salmon may successfully reproduce in the Aysén
269 Region, although some evidence of reproductive individuals migrating upstream has been
270 reported (Becker et al. 2007, Soto et al. 2007). Hence, the possibilities for management and
271 mitigation in these species may be greater than for the other salmonids.

272

273 **SOCIAL AND ECONOMIC EFFECTS OF ESCAPES**

274

275 The conflict between the salmon industry and the artisanal fishing sector is one of the
276 most relevant socio-economical impacts arising from salmon escapes in Chile. Small-scale
277 fishing of escaped salmon could have an important social and economic effect, providing
278 food security and extra income for rural people and low-income families (Arismendi 1997,
279 Soto et al. 2001). For example, during the massive escapes of 1994-95 a large number of
280 local fishers, often women and children, were fishing for salmonids (mostly with gillnets),
281 which were then sold in local markets (Soto et al. 2001). The practice is still quite common
282 in communities near salmon farming locations in the inland seas and lakes. It is important
283 to note however that the application of large quantities of antibiotics in the salmon

284 aquaculture in Chile has environmental implications that potentially impact the health of
285 humans and wildlife (Fortt et al. 1997).

286 Currently salmonids are the property of farm owners even after they have escaped, so the
287 capture and marketing of escaped salmon by artisanal fishermen is considered an illegal
288 practice. During the previously mentioned massive escapes, fishermen created considerable
289 turmoil requesting fishing rights over these escaped salmon, in addition to claiming that
290 native fish resources were affected (Soto et al. 2001). The possibility of an artisanal fishery
291 is somewhat feared by salmon farmers as legalization of a salmon fishery could generate
292 competing products whose standards may not be at the level of those adopted by the
293 farmers' organizations (Niklitchek et al. 2013). It is clear however, that products from the
294 salmon fishery could be oriented to local markets for domestic consumption, whereas
295 production from salmon aquaculture is aimed for export to international markets. However,
296 for salmon farmers opening a fishery of escaped salmon could enhance vandalism and/or
297 theft at the farms. If some of the escaped salmon species are indeed able to establish and
298 develop naturalized populations, we predict a new "battleground" in the marine
299 environment between artisanal fishery and salmon farming.

300 In addition, sport fishing entities are debating the pros and cons of this new species for
301 tourism and business. Hence this creates conflict between those who want to fish and those
302 who would like to eliminate these returning exotic salmon runs.

303

304 **PREVENTION AND MITIGATION OF ESCAPED SALMONIDS**

305

306 **Preventing salmonid escapes**

307

308 Considering that many escapes are due to human mistakes, preventive measures can be
309 effective in a number of cases. A prevention system utilized in aquaculture facilities located
310 in streams is the use of physical barriers to prevent fish from escaping. The barriers are
311 strategically located in critical connection points throughout the facilities, such as pools or
312 tanks containing the fish in water inlets and outflows. Some companies have more efficient
313 systems that minimize the risk of escapes by using recirculation tanks. In these closed and
314 independent systems, salmonids do not come into contact with the outside environment.

315 Ocean grow-out farms pose the largest challenges for the industry. Although anti-
316 predator nets protect against external attacks and may also serve as containment when fish
317 nets tear, there are no actual physical barriers in place. Instead, fish farmers have developed
318 maintenance practices to prevent nets from breaking and releasing fish into the
319 environment. The proper tension of fish and anti-predator nets through anchoring and
320 mooring systems reduces friction between materials and prevents nets from sticking to one
321 another, thereby preventing sea lions from approaching the fish. Other typical practices to
322 reduce escape risks include replacing and maintaining nets and monitoring by divers or the
323 use of video cameras.

324

325 **Mitigating salmonid escapes**

326

327 The aquaculture regulatory framework in Chile includes environmental regulations (the
328 Environmental Impacts Assessment System, SEIA) established in 1997, and the executive
329 decree on environmental norms for aquaculture (RAMA). These regulatory tools and their
330 operational norms affect both licensing and operation of fish farms. The SEIA includes the
331 establishment of contingency plans for escaped salmonids and, according to the RAMA,

332 these plans must follow special guidelines. These guidelines have information about
333 operational procedures and devices to recover escaped salmonids, as well as the obligation
334 to provide immediate detailed information to authorities about escapees. Mandatory
335 reporting of escape incidents was introduced to Chile in 2001, with a national statistics
336 database since 2004. This has enabled a gross assessment of the overall status of the escape
337 problem at an industry-wide scale from year to year, and an evaluation of the causes of
338 escapes.

339 Different and non-exclusive techniques are used to capture free-living salmonids that
340 have escaped from marine grow-out farms. The most popular technique is try to capture the
341 escaped salmonids with nets or mobile empty cages, often using pellets to attract escaped
342 salmonids to a particular location (Melo et al. 2005). All aquaculture staff is required to
343 work following an escape event, but often the employees often do not have training in how
344 to manage or respond to such events.

345 There is no quantitative information, however, on how effective these prevention
346 techniques and recapture systems are. The background information collected in other
347 producer countries reveals a low success rate for recapture efforts (recaptures amount to < 3
348 percent) (Thorstad et al. 2008). This is likely the case in Chile, because current recapture
349 systems reflect mitigation measures taken by the industry which are both inefficient and
350 insufficient (Melo et al. 2005). The RAMA does not detail specific mitigation and recapture
351 protocols, and producers have to define their own action plan to deal with escaped fish.

352 According to Melo et al. (2005), salmonids not only do not remain in the vicinity of the
353 cages, but they can move up to 3 km in 10 hours, revealing high levels of mobility. In fact,
354 molecular analyses on rainbow trout indicated that the incidence of escapees is widespread
355 (Consuegra et al. 2011). When large-scale escape events often take place under bad weather

356 conditions, time becomes a critical factor and recapture tasks are extremely complex,
357 further reducing the success of such operations. Additionally, action plans are not always
358 enforced due to the limited capacity of local and regional government institutions. An
359 improved knowledge of pattern of movements, behavior and survival rates of escapees
360 would be useful to inform natural resource managers and the fish farming industry.

361 To reinforce the action plans and also the fiscalization by the government institutions it
362 would be useful to determine the specific origin of escaped salmon. Recently some tools
363 have been developed to differentiate potentially escaped from free-living fish, including the
364 detection of manganese concentration from scales (Adey et al. 2009) and stable isotopes of
365 carbon and nitrogen (Schröder & García de Leaniz 2011). Furthermore, some approaches
366 with molecular markers have been used in Chile to distinguish if salmonids in rivers and
367 lakes are descended from specimens introduced for ranching or from individuals escaped
368 from salmon farms (Astorga et al. 2008), or if a genetic admixture occurs between
369 individuals escaping from fish farms and “naturalized” salmonids (Consuegra et al. 2011).
370 Using this approach as a baseline, in the future it might be possible for each aquaculture
371 facility to have a unique and registered genetic marker stored in a database, allowing a
372 posterior cross-comparison with escaped fish and thus allow determining their specific
373 origin.

374 Soto et al. (2001) proposed that a mitigation procedure could be that artisanal fishers try
375 to control escaped salmonids by capturing escapees, especially considering that artisanal
376 fishing commonly occurs around fish farm locations. In addition, there is potential for
377 developing a recreational fishery especially following an escape (Arismendi & Nahuelhual
378 2007). Such fish could be allowed in a take quota (assuming that the fish are safe to eat),
379 complementing the current catch and release approach for trout. Sport fishing could be

380 improved and facilitated around farms to collect and control escaped fish and to provide
381 additional income to local people and fishermen. It is clear that such a fishery must be well-
382 organized to be sure it does not conflict with the industry and/or facilitate more escapes.
383 The promotion of both artisanal and recreational fisheries should be considered only as a
384 mitigation procedure. Although all species of salmonids introduced for aquaculture
385 purposes are already present in both freshwater and marine environments, some of them
386 are not yet reproducing on their own (Soto et al. 2001; Soto et al. 2006; Arismendi et al.
387 2009). The removal of these potential new invasive species through a salmonid-based
388 fishery could certainly decrease the likelihood of new establishments. Artisanal fisheries
389 based on salmonids should be of limited access and highly regulated in order to discourage
390 the promotion of further releases.

391 It is important to have into consideration that artisanal and recreational fisheries should
392 not impact native species. Fortunately, in freshwater systems of southern Chile there are no
393 native fish that could be potentially affected by artisanal or recreational fisheries. Native
394 fish are smaller in size than salmonids, and thus they have a low potential for incidental
395 capture. In general, salmonid-based artisanal fisheries use gillnets, which are highly size-
396 selective. Anglers tend to use a catch-release approach, avoiding negative effects on native
397 fish. In marine environments however, the potential for incidental capture of native species
398 is greater than in freshwater, but these native fish already have preexisting historical
399 fisheries and thus an established commercial value (Soto et al. 2001).

400 Local and regional natural resource managers should be involved to assess free-living
401 and self-sustaining salmonid populations, and should begin discussions with interest groups
402 on the use and management of such populations. To inform these managers, investigators
403 need to evaluate social and economic scenarios involving these potential fisheries. If a

404 monetary value is given to these escaped salmonids then the industry could be held
405 accountable to compensate the local and regional governments.

406

407 **RESEARCH NEEDS**

408

409 As has been identified in this review, there are still several gaps in the knowledge of the
410 impacts and consequences of escaped salmonids in Chile. Thus, considering that salmonid
411 aquaculture is expected to continue to grow, different research needs should be identified
412 including biological, social and economic aspects that could generate useful information for
413 decision makers. One of the most important research needs is to implement a monitoring
414 program to evaluate the abundance and impacts of escaped salmonids. For example, to
415 establish long-term field surveys to estimate the relative importance of fish escaped from
416 the native fishes and and naturalized populations (see for estuarine areas Soto et al. 2001;
417 for lakes Arismendi et al. 2009, 2011; and for streams Soto et al. 2006, 2007). It is also
418 important to evaluate and develop reliable indicators to estimate their impact on native
419 species and ecosystems, as well as their social and economic impact in both marine and
420 freshwater ecosystems (Velásquez et al. 2011). For example, to test broad scale hypothesis
421 about taxonomic homogenization and expansion of introduced species a paired comparison
422 between historical and current presence/absence of both native and introduced species
423 appears to be a useful tool (Marr et al. 2010, 2013).

424 While additional work is needed to increase our knowledge of the processes underlying
425 the patterns described in this review, more data could improve management of non-native
426 salmonids in areas where they impact native fishes negatively. For example, patterns of
427 apparent coexistence of non-native trout and native fishes in some streams could provide

428 clues for managing invasions in more heavily impacted streams (see conservation status in
429 Campos et al. 1998, Habit et al. 2006). Thus, understanding mechanisms of coexistence
430 between native fishes and introduced non-native salmonids may help in designing effective
431 management strategies that protect both native fishes and important recreational fisheries.
432 Coexistence may also improve our knowledge of the functioning of pelagic and benthic
433 communities in lakes and inner seas of southern Chile, maintaining areas without salmonid
434 farming as reference sites to understand more completely the effects of free-ranging
435 salmonids. Even more important is to monitor lakes without salmon farming, especially
436 lakes without trout which seem to be very scarce (Soto et al. 2006, Correa & Hendry 2012).

437

438 **CONCLUSIONS**

439

440 Aquaculture is continuing to grow and expand worldwide and so is salmonid farming
441 probably everywhere including the major producing countries such as Norway and Chile.
442 Effects of escaped salmonids in regions where they are not native are quite different from
443 where they are native such as Norway, because there major impacts are related to genetic
444 modification of natural populations for Atlantic salmon (Fleming et al. 2000, Thorstad et al.
445 2008). In Chile the effects are mostly related to direct impacts of escaped individuals on
446 native fishes and the local environment, including important social and economic effects
447 related to artisanal and recreational fisheries. The establishment of new salmonid species
448 such as Atlantic and coho salmon, with their potential long-term effects due to naturalized
449 populations, requires urgent action by decision makers.

450 In such a complex situation, with ecological, social, health-related, political and
451 economic implications, all stakeholders must assume their responsibilities. Government

452 agencies must ensure the ecological balance of water systems, minimum escape levels and
453 effective mitigation measures, including regulations to help manage these values. Salmonid
454 farmers must undertake a more proactive prevention role, including: (1) identify critical
455 issues in every stage of salmon farming, so as to establish protocols to prevent salmon
456 escapes; (2) conduct an adequate selection of fish farm sites; (3) design optimal structures
457 for the area's oceanographic conditions; (4) develop and implement special technologies
458 and materials to prevent escapes; and (5) prepare more effective procedures and guidelines
459 for the recapture of escaped fish. In this context, the coupling of aquaculture with fisheries
460 (artisanal and recreational) could help manage the natural resources which both of these
461 activities require, and thus the management of escaped salmonids should be addressed
462 accordingly.

463

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465

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473

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640 **FIGURE CAPTIONS**

641

642 Figure 1. World production (thousands of tons) of salmonid species including rainbow
643 trout, coho and Atlantic salmon between 1990 and 2010 (main producers are shown).

644 Source: FAO aquaculture statistics.

645

646 Figure 2. Salmonid farm locations in the Lakes and Aysen Regions.

647

648 Figure 3. The main technical issues associated with salmonid escapes, according to reports
649 by salmonid farm companies in (a) the Lakes (filled bar) and (b) Aysen (open bars)

650 Regions. Source: National Fisheries Services and Chilean Navy.

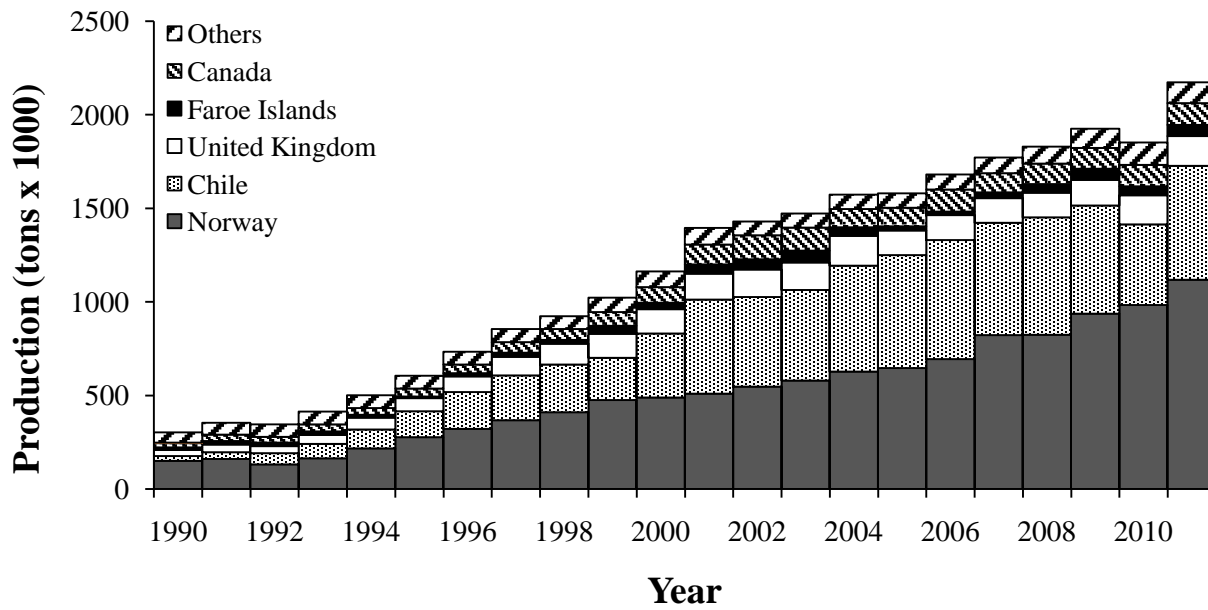
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652 Figure 4. Total salmonid production and reported escapes. Escape data for 1993-1996 from
653 Soto et al. (1997) and for 2004-2012 from National Fisheries Service (unpublished
654 information). Production data for 1993-2012 from Fishing Statistical Yearbooks.

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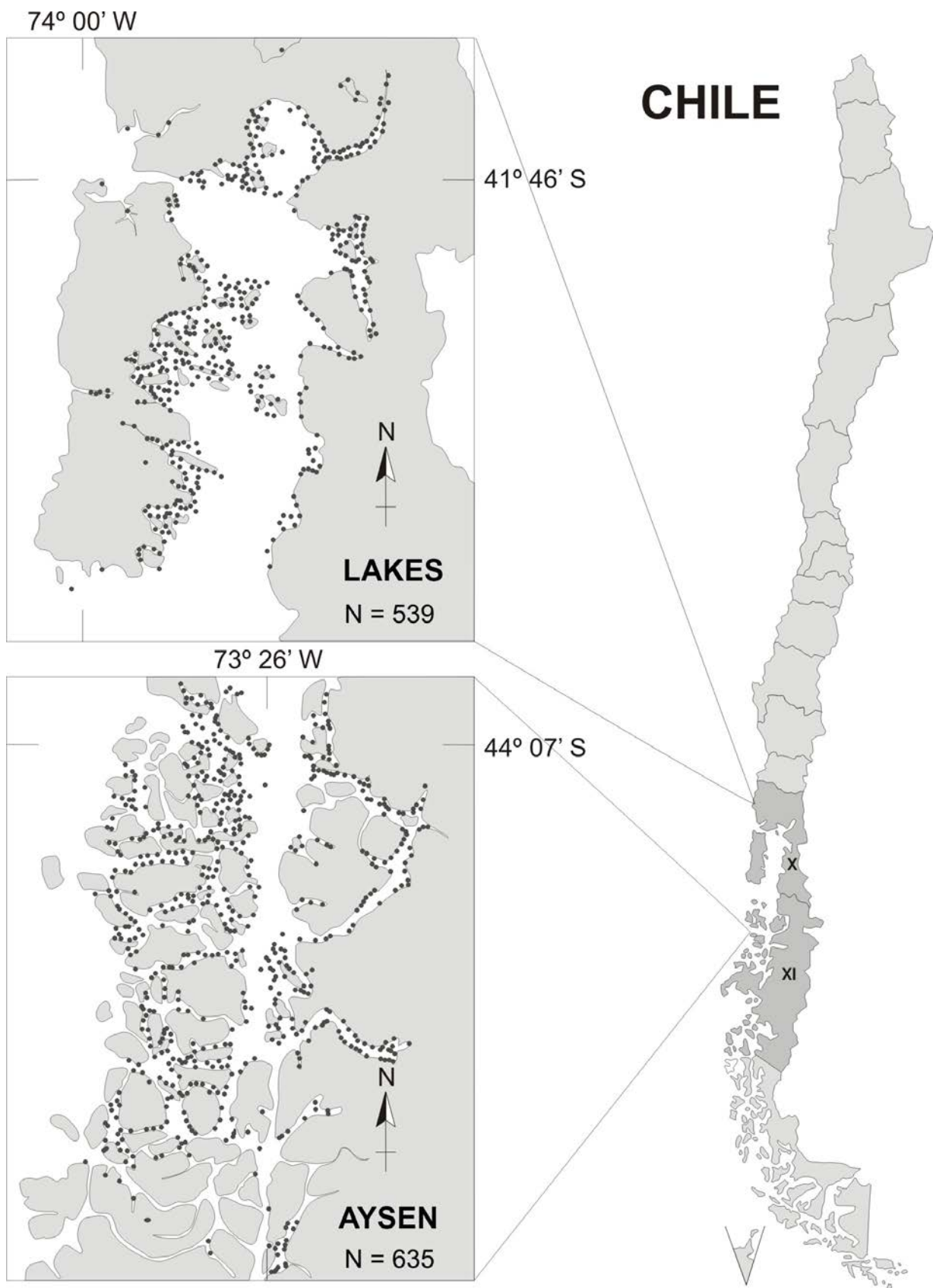
656 Figure 1

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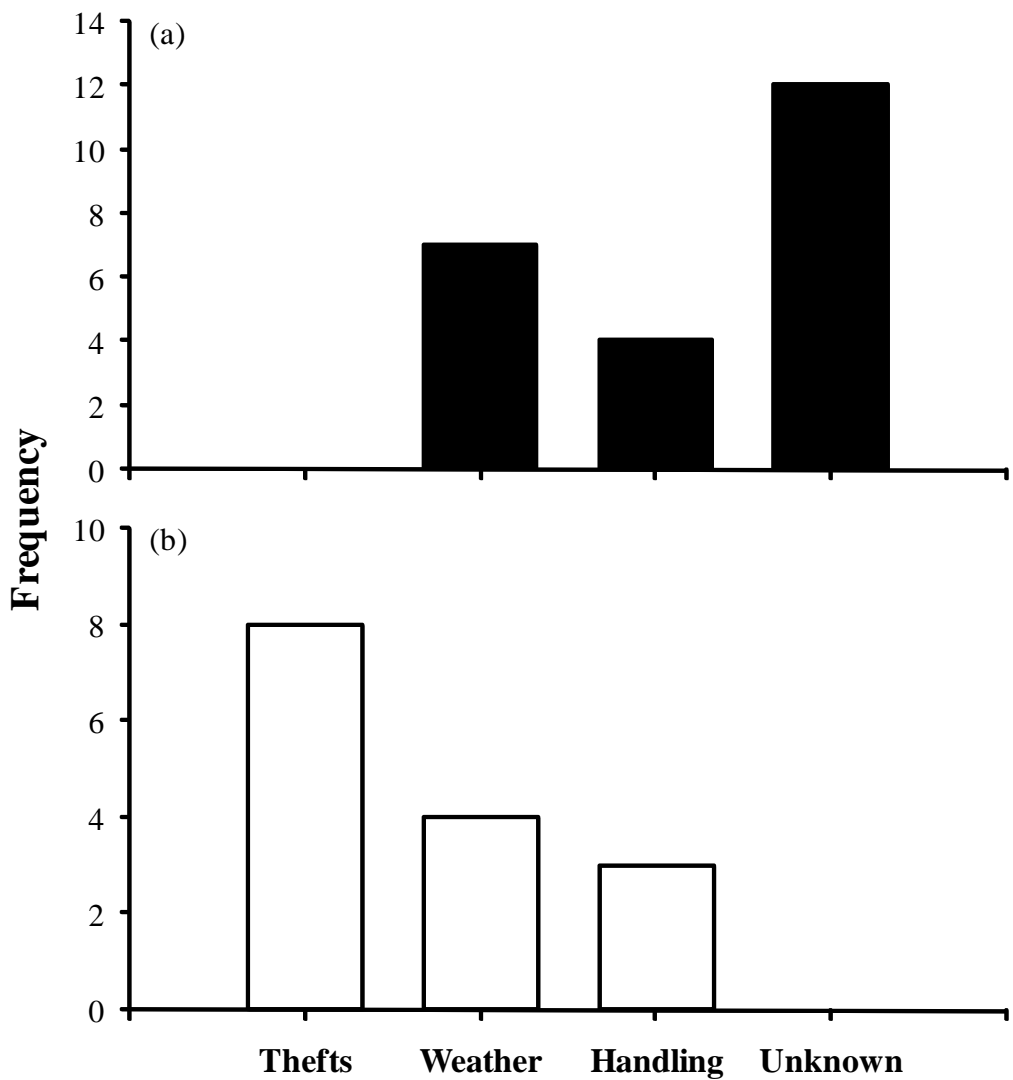
658 Figure 2

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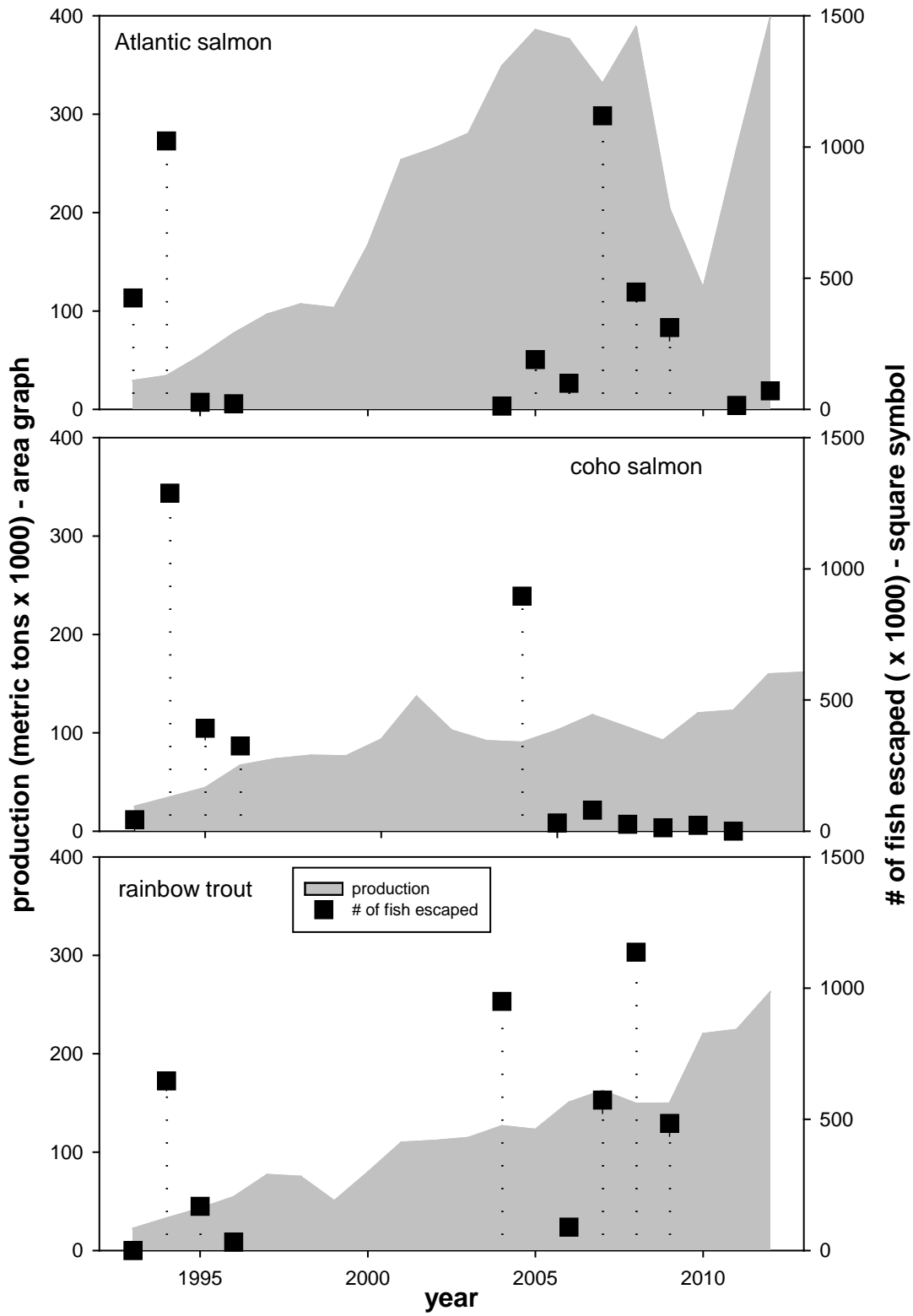


682 Figure 3

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684 Figure 4



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